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Atmospheric Chemistry and Physics Discussions

Fluxes and sources of nutrients and trace metals atmospheric deposition in the northwestern Mediterranean

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Abstract

Total atmospheric deposition was collected on a weekly basis over 3.5-yr (March 2008-October 2011) at a remote coastal site on the west coast of Corsica Island. Deposition time series of macro and micro-nutrient (N, P, Si, Fe), and trace metals (As, Cr, Cu, Mn, Ni, V, Zn) are investigated in terms of variability and source apportionment (from fluxes of proxies for aerosol sources (Al, Ti, Ca, Na, Mg, S, Sr, K, Pb)). The highest fluxes are recorded for Si, P, then Fe for nutrients, and for Zn and Mn for trace metals. For the majority of elements, data show some weeks with high episodic fluxes, except for N, Cr and V which present the lowest variability. Twelve intense mineral dust deposition events are identified during the sampling period. The contribution of these events to the fluxes of Fe and Si represents 52% and 57% of their total fluxes, respectively, confirming the important role of these sporadic dust events on the inputs of these elements. For N and P, the contribution of these intense dust deposition events is lower and reaches 10 and 15%, respectively. Out of these most intense events, positive matrix factorization (PMF) was applied to our total deposition database in order to identify the main sources of nutrients and trace metals deposited. Results show that P deposition is mainly associated to anthropogenic biomass burning inputs. For N deposition, inputs associated to marine sources (maybe associated to the reaction of anthropogenic N on NaCl particles) and anthropogenic sources are quasi-similar. A good correlation is obtained between N and S fluxes, supporting a common origin associated to the inorganic secondary aerosol, i.e. ammonium sulfate and ammonium nitrate. For trace metals, their origin is very variable: with a large contribution of natural dust sources for Ni or Mn and on the contrary of anthropogenic sources for V and Zn.

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1. Introduction

The Mediterranean Sea is a semi-enclosed basin situated at the interface between contrasted continental areas of three continents, namely southern Europe, northern Africa and the Middle East, which coastal areas are heavily populated. Thus, the Mediterranean basin continuously receives anthropogenic aerosols from industrial and domestic activities from all around the basin and other parts of Europe (Sciare et al., 2008; Becagli et al., 2012). In addition to deposition from this anthropogenic background, seasonal inputs from biomass burning occur mainly during dry summers (Chester et al., 1996; Guieu et al., 1997), and strong deposition pulses of mineral dust from the Sahara are superimposed (Guerzoni et al. 1999a), with some 'extreme events' with dust deposition fluxes as high as 22 g m^{-2} as recorded in 2004 (Bonnet and Guieu, 2006) on very short time scales of a few hours to a few days.

A number of key elements for marine biota are associated to those inputs. Thus, several authors showed that the atmospheric deposition of aerosols constitutes the main source of major nutrients, as N, P or Fe to the surface open waters of the Mediterranean Sea in the summer/autumn period when surface water stratification prevents inputs from deep water by vertical mixing (Guerzoni et al., 1999a; Bonnet and Guieu, 2006; Krom et al., 2010; Pulido-Villena et al., 2010). Besides the classical nutrients (N, P and Fe), the aerosols also carry trace metals (hereafter called TMs) such as Cr, Cu, Ni, Mn or Zn that are known to have a biological role, often as cofactors or part of cofactors in enzymes and as structural elements in proteins (Morel and Price, 2003). Recent study of Ridame et al. (2011) suggests so that the trace metals released by Saharan dust could stimulate nitrogen fixation in summer in Mediterranean Sea. This assumption is supported by the works on Tovar-Sanchez (2014) which show that the trace metals concentrations in surface microlayer in Mediterranean Sea is correlated with the atmospheric deposition of mineral dust. However, it has been also suggested that the atmospheric deposition of particulate pollutants are responsible for the contamination of the Mediterranean waters in trace metals (Bethoux et al., 1990; Guerzoni et al., 1999b). Gallisai et al. (2014) show also negative effects of dust deposition on chlorophyll, coinciding with regions under a large influence of aerosols from European origin.

Thus, the partitioning/mixing between anthropogenic vs. natural atmospheric inputs is critical to estimate and predict the role of the atmospheric deposition on marine biosphere and associated services. However in the Mediterranean Sea, the existing database on atmospheric

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fluxes of nutrients and trace metals remain quite limited. Most of studies are focused on total deposition of dust and/or macro-nutrients as P and N (e.g. Markaki et al., 2010). This approach do not include the variety of nutrients and do not enable to distinguish the origin of nutrient-bearing particles. Moreover, the studies on trace metals deposition (Cd, Pb…) often show an influence of local sources (Guieu et al., 2010), limiting the reliability of these data. At the difference of atmospheric deposition, the source apportionment of particles in Mediterranean, from PMF method, has been highly investigated in recent works and showed a large spatial variability of source contributions (Becagli et al., 2012 and 2017; Calzolai et al., 2015; Amato et al., 2016; Diapouli et al., 2017). The signature of continental pollution sources was observed even in remote area, as central Mediterranean islands (Calzolai et al., 2015). Yet, PM concentrations and source are probably different of sources of deposited particles which depends on aerosols size distribution, precipitation pattern… Thus, in a context of anthropogenic changes, it is crucial to distinguish between anthropogenic vs natural atmospheric inputs of nutrients in order to assess how the evolution of chemical atmospheric forcing will modify the marine nutrient cycling.

Here we show a 3.5-yr time continuous series of macro and micro-nutrient (N, P, Si, Fe), trace metals (As, Cr, Cu, Mn, Ni, V, Zn) and source tracers (Al, Ti, Ca, Na, Mg, S, Sr, K, Pb) total deposition fluxes in a remote coastal site in Corsica. Between March 2008 and October 2011, a monitoring station has been operated with a weekly sampling time for total bulk deposition. In order to assess the contribution of sources in the fluxes of nutrients, a work on the source apportionment of various nutrients and TMs was carried out from these data (PMF method). A specific attention was also given on the different types of extreme atmospheric events which are relevant regarding the biogeochemistry in the Mediterranean Sea. They include Saharan events and intense summer storms that trigger the washout of the atmosphere over an altitude of several thousands of meters in a short time.

2. Material and methods

Sampling site and protocol

Total bulk deposition (i.e. dry + wet deposition) was sampled weekly from March 2008 to October 2011 using an open collector for nutrients and TMS analyses, except for N which the analyses started from November 2009. Sampling was conducted at Cap Cuittone (42.44°N,

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8.66°E, 190 m above sea level). The site of sampling is on the Mediterranean coast of the National Corsica Park (Parc Naturel Régional de Corse) at 16 km to the SSE of Calvi, the main city in the part of the island which has no important industry (~5500 inhabitants), and about 3.5 km N of the village of Galeria (~350 inhabitants). Consequently, the data from this site could be considered representative of the open western Mediterranean Sea.

The sampler is a 120-mm diameter PTFE Teflon® funnel (collection aperture 0.0113 m^2) machined on a circular base with a thread adapted to 500 mL polypropylene (PP) Nalgene® bottle neck. The funnel stem is a long tapered stem in such a way that it soaks in the preloaded acid for limiting the evaporation of collected water. All the deposition sampling materials (Teflon®-PTFE capped funnels, 500 mL and 60 mL Nalgene® PP bottles, and 60 mL PP boxes with a screwing cap for bottle's corks) were thoroughly washed with hydrochloric acid at the ultra-clean laboratory of LISA following a protocol adapted for ultra-clean sampling (Heimburger et al., 2012). Before deployment, the sampling bottles are preloaded with 50 mL of hydrochloric acid (2%v/v) and weighed. Each funnel coupled to its bottle preloaded was deployed on the site at 2 m high. The height was controlled by a spirit level to ensure that the funnel aperture was horizontally leveled. Each week, before collection, the internal surface of the funnel was rinsed with 60 mL of 2%v/v ultrapure hydrochloric acid in ultrapure water, taking care to flush all the surface. The rinsing solution was collected in the sample bottle to be removed and replaced by a new 500 mL PP bottle for the next week. The funnel was replaced by a new one every 6 months. Field blanks were performed at the funnel installation and removal, but also from time to time by repeating twice the sampling procedure. One permanent staff of the Corsica National Park was carefully advised and performed sampling during all the monitoring period.

Chemical analyses

In the laboratory, total atmospheric deposition sampled bottles were weighted. The amount of rainwater collected in the funnel was deduced by subtracting added acid solution (i.e. 110 mL) to the sample total mass found in the bottle. Each sample was shaken and then 15 mL were immediately transferred into a PE sampling vial to measure the size distribution of the particulate phase. The rest of the sample was filtered before analysis with acid washed Nuclepore® polycarbonate filters (0.2-µm porosity). The filters were analysed by wavelength dispersive X-ray fluorescence (PW-2404 spectrometer by PANalytical™) for the particulate

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elemental concentrations for elements from Na to Pb, including macro (P) and micro-nutrient (Fe, Si), trace metals (As, Cr, Cu, Mn, Ni, V, Zn), and source tracers (Al, Ti, Ca, Na, Mg, S, Sr, K, Pb). Analyses of the filtered aqueous sample were performed by Inductively Coupled Plasma-Atomic Emission Spectrometry (ICP-AES, Spectro ARCOS Ametek®) coupled with a CETAC ultrasonic nebulizer for dissolved elemental concentrations of as many trace metals as possible (altogether 45 elements were analysed; Desboeufs et al., 2014). Due to the time between collection and analyses, the preservation of N speciation was not guaranteed and in consequence data are expressed as total N. The total dissolved inorganic N concentrations were obtained by adding NH_4^+ concentrations and NO_3^- and NO_2^- concentrations determined by ionic chromatography (Professional IC 850 by Metrohm®). Field blank concentrations represent at the maximum 22% of major nutrients concentrations (max for Fe) and 19% of trace metals concentrations (max for V). Field blank concentrations are subtracted to samples concentrations collected in the same period.

The weekly elemental deposition fluxes were calculated from concentrations of all chemical species measured in dissolved and particulate samples by considering the sampler area and the total liquid volume (preloading + rinsing + rain). The total elemental deposition fluxes were estimated by adding particulate and dissolved fluxes except for N assumed totally acid-soluble. Atmospheric nitrogen exists in particulate phase but also as gaseous species (NO_X, HNO_3, NH_3). In our study, the used bulk collector has a design very close to the one of bulk collectors used during ADIOS project which are not optimized to collect gaseous nitrogen by dry deposition (Markaki et al., 2008). However, wet deposition including both washed-out particulate and gaseous nitrogen, measured N fluxes in this study will be considered mainly representative of bulk deposition of aerosol particles and wet deposition of gaseous N.

Dry vs. wet deposition

The speciation between wet and dry deposition is a critical parameter to estimate the potential dissolved fluxes of nutrients. Precipitation (mm) was estimated on the site from the amount of water in the sample. The precipitation occurrences are in agreement with the rainfall records on Calvi airport which is distant by about 15 km. Since they are more representative of local rainfall, precipitation estimated from our samples were used for the attribution of deposition fluxes to wet vs. dry deposition. Wet deposition was considered when rainfall was larger than 1 mm during the sampling period. The threshold value of 1 mm

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integrates the uncertainties on the weighing of samples in order to ascertain that the rainfall was real. Samples which present no precipitation or rainfall lower than 1 mm, are considered as dry deposition. In consequence, dry deposition is assimilated to wet deposition when happening the same week as a precipitation event. This method underestimates dry deposition, and provides a lower estimates of deposition dry event number vs total deposition event number.

Positive Matrix Factorization (PMF)

Multivariate statistical methods, such as factor analysis, are widely used to identify source signatures and explore source–receptor relationships using the trace element compositions of atmospheric aerosols (e.g., Polissar et al., 2001, Calzolai et al., 2015) and precipitation (Keeler et al., 2006; Gratz et al., 2013). Since many sources emit characteristic relative amounts of certain trace elements, source–receptor techniques can be used with an understanding of these elemental signatures to identify the major sources influencing a given receptor site.

We applied EPA PMF v5.0 (Norris et al., 2014) to the matrices of tracers, nutrient and TMs total deposition measurements. PMF is a multivariate statistical technique that uses weighted least-squares factor analysis to decouple the matrix of observed values (X) into two matrices representing the factor scores (G) and factor loadings (F), as represented by the equation $X = GF + E$, where E is the residual matrix representing the difference between observed and predicted values (Paatero and Tapper, 1994; Paatero, 1997). Prior to applying PMF, we used the weekly deposition fluxes and we replaced fluxes reported as less than median detection limit (MDL) with the median value. The uncertainties for each samples correspond to the sum of uncertainties in sample collection (i.e. 10%) and analytical measurement (standard deviation of three replicate analysis for each sample). We included all valid samples, excluding the samples that we identified as extreme outliers, i.e. samples corresponding to dust events and high As deposition (12 samples) (see section extreme events). The deposition fluxes for 21 elements is used; i.e macro and micro-nutrient (N, P, Si, Fe) and TMs (As, Cr, Cu, Mn, Ni, V, Zn) and tracer elements (Al, Ti, Ca, Na, Mg, sea-salt S (ssS), Pb, K, and excess S (Sexc). The estimation of ssS fluxes is obtained from Na fluxes on the basis of typical seawater S/Na ratio (Henderson and Henderson, 2009) and excS fluxes in subtracting ssS to total S fluxes. Since S was used as sources tracers, the discrimination between ssS and excS enabled to have a best

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constrain on signature of sources. Elements with a signal-to-noise (S/N) ratio <5 were categorised as "weak" (i.e. As and N) and hence down-weighted so that the user-provided uncertainty was increased by a factor of three (Norris et al., 2014). The variability in the PMF solution was estimated using a block bootstrap technique, which calculates the stability of the model solution by randomly re-sampling blocks of the input dataset and computing the variability between model solutions. We applied 100 bootstrap runs to the PMF base run with the lowest Q value. We determined the final factor profiles based on our ability to identify all the factors, the robustness of Q values, the ability of the model to replicate measured results, and the bootstrap results.

3. Results and discussion

The 3.5-yr time-series of weekly fluxes (195 samples) for nutrients, TMs, major signature elements (Al, Na, S and K) and precipitation are presented in Figure 1. Corresponding time-series of other signature elements (Ti, Mg, Sr, Pb) are available in supplement with the total atmospheric fluxes data. The highest fluxes are recorded for Si, P then Fe for major nutrients and for Zn and Mn for trace metals. 51% of the samples, i.e. 99 samples, sustained at least one event of precipitation during the week of sampling and are here referenced as wet deposition. In our set of 195 samples, 21 presented a rainfall higher than 20 mm and the highest weekly rainfall recorded is 29 mm. However, no systematic link is observed between the biggest rain event and the nutrients of metals fluxes.

The results emphasize large differences in timing of deposition fluxes between the studied elements. But for all the elements, data display some weeks with high episodic fluxes. Due to the sporadic character of specific events such as dust storms or forest fires giving rise to high deposition events, it is known that the fluxes of elements associated to these sources are often important on a short period. For example, for elements such as aluminium associated with dust events, a half or more of the annual deposition flux may occur in one event of a few days or even hours (Guieu et al., 2010), and high deposition events (>1 g m^{-2}) are responsible for the inter-annual variability of the Al deposition flux in the western Mediterranean basin (Löye-Pilot and Martin, 1996). As a consequence, the fluxes linked to these extreme events can dominate and hide the influence of more continuous emission sources. In our dataset, this is the cases of Si, Fe and As, for which 25% of total fluxes on 3.5 years is delivered by 1 to 3 weekly samples, whereas for the majority of nutrient and TMs, 25% of total fluxes are

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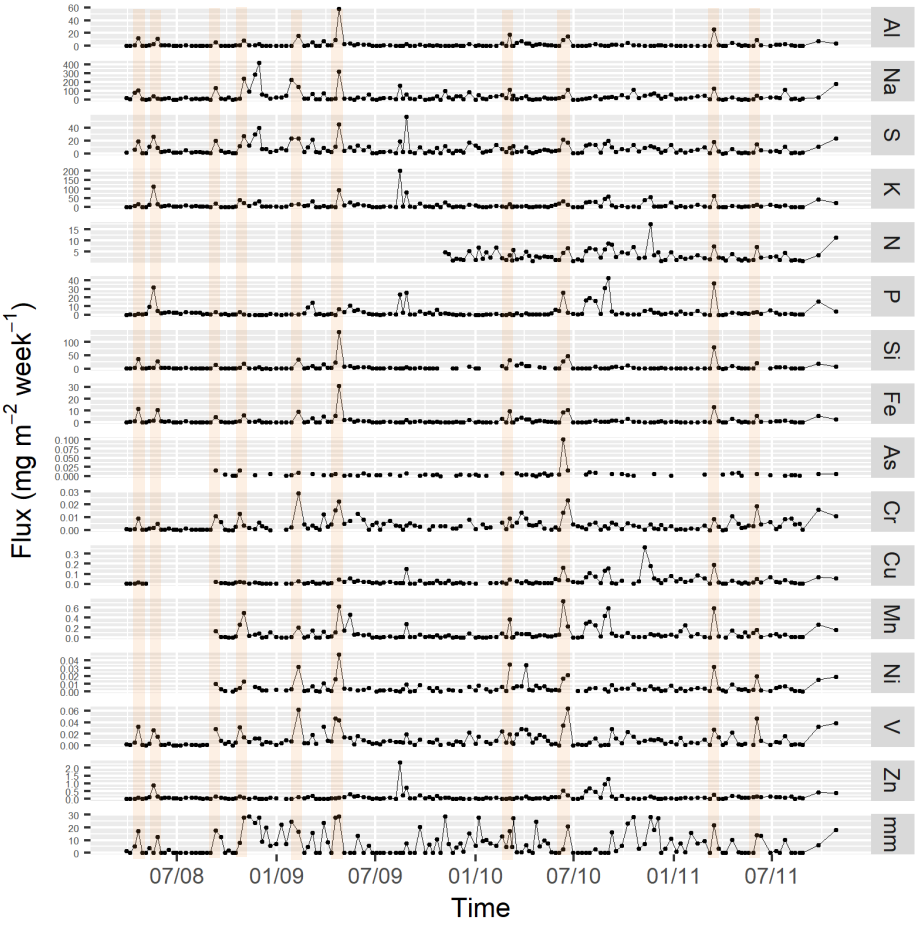
219 constituted from the 5 to 8 highest events. The most obvious case is for As which 23% of the
220 total flux is obtained in only one week during June 2010 (0.1 mg m^{-2} $week^{-1}$). This event
221 corresponds to one event of wet deposition of 7 mm, i.e. no particularly intense rain, and is
222 concomitant with high fluxes for the other studied elements.
223



224
225 **Figure 1: Temporal variability of bulk weekly fluxes from March 2008 to October 2011 for main markers,**
226 **nutrients and trace metals, and rainfall on the same period. The 10 most intense dust event are displayed in**
227 **the boxes in orange.**
228
229 *Seasonal variability*
230 Monthly total and wet fluxes have been estimated to investigate the seasonal variability of
231 the measured elements inputs over the northwestern Mediterranean (Figure 2). A large

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variability in the monthly deposition fluxes of all the elements is observed in agreement with the episodic pattern of weekly inputs. Nutrients deposition presents a clear seasonal pattern: P with the major deposition fluxes in summer, and N in winter, whereas the main fluxes are observed in spring for Fe, Si, Cr, Ni and V. For As, excluding June, which shows its highest monthly mean flux due to the intense event of June 2010, the maximum of fluxes are recorded at the end of summer and beginning of autumn. For Mn, no clear seasonality is observed. A monthly flux predominates in August and November for Zn and Cu, respectively, reaching at least twice the other monthly fluxes. For all the elements, the wet deposition predominates the total fluxes between October and April in agreement with the highest rainfall recording during this period, whereas dry deposition is the main way of input in May, July and August. Our results are in agreement with the seasonal pattern observed in the 1980's for Si and Fe deposition at Capo Cavallo, 8 km more North on the Corsican coast (Bergametti et al., 1989). The maximum of deposition during spring is explained by the concomitance of rainfall and high dust concentrations, whereas Si and Fe atmospheric aerosols concentrations present their maximum in summer during the dry season. On the contrary, for the elements mainly associated to dry deposition i.e. Zn, P and Cr, Bergametti et al. (1989 and 1992) found that the highest deposition is typically associated with the period of their highest aerosols concentrations in summer. This is not the case for Cr in our results, which follows the Si and Fe behavior. At the difference of our Corsica site, no clear seasonal variability is observed for the deposition fluxes recorded at Cap Ferrat, 170 km more NNE on the French continental coast, a site affected by the anthropogenic influences from continental Europe (Pasqueron de Fommervault et al., 2015). That could be the case for Mn atmospheric fluxes on our site.

The case of N deposition is specific, since the N deposition flux corresponds mainly to total aerosol and wet gaseous deposition inputs in our samples. The general pattern for N with highest fluxes in winter could be linked to the thermal instability of the ammonium nitrate, which is the dominant form of N in aerosol particles associated to a decrease of rain events during the hot season, and to extremely typical intense nitrate episodes recorded from November to March in the western Mediterranean basin associated to maximum wet deposition (Querol et al., 2009).

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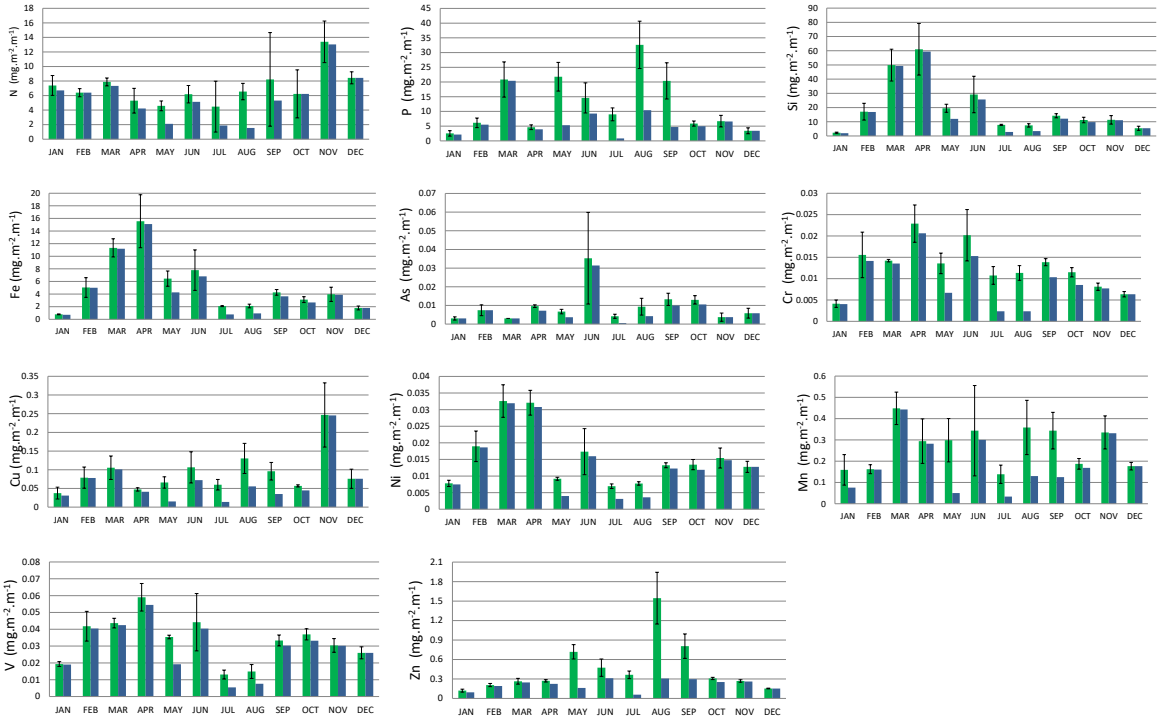

Figure 2: Temporal variation of monthly total (green bars) and wet (blue bars) deposition during the sampling period March 2008–October 2011. Bars indicate standard deviations over the weekly values available over the period.

Inter-annual variability

The average annual total deposition fluxes for the major nutrients and trace metals during the 3.5 years of sampling are presented in table 1. Among major nutrients, the most abundant nutrients in bulk deposition is Si followed by P and N which have fluxes in the same order of magnitude. The highest annual fluxes recorded for N in comparison to Fe is due to the sporadic pattern of Fe fluxes in comparison to N that shows more regular weekly fluxes. For trace metals, the highest annual fluxes are observed for Zn, Mn and Cu whereas the other trace metals have fluxes smaller by one order of magnitude. Except for Ni, the standard deviation on the mean fluxes are larger than 15%, and reach more than 50% for P, As and Cu, meaning a large inter-annual variability of their deposition, in agreement with the high recorded sporadic weekly fluxes for these elements. Our results are compared with other fluxes in Corsica (Table 1) as reported in the literature. Data show that for trace metals, the recorded values are in the same order of magnitude of previous measurements in Corsica. On the

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contrary, for the major elements as Fe, Si and N except for P, our deposition flux values are much lower than the previous ones obtained in Corsica (Table 1) and globally in the western Mediterranean (Bonnet et al., 2006; Pasqueron de Fommervault et al., 2015). A net decrease in N deposition is also observed between the 1990's and now in Europe (Waldner et al., 2014). The only element with highest deposition fluxes in comparison to the literature is P, suggesting an increase in atmospheric fluxes for this element. Keeping in mind that dry deposition events can be underestimated by our method, the wet fluxes predominate the total deposition fluxes (≥ 64%) for the majority of elements except for P and Zn, for which less than half of the total flux is associated to precipitation. This is in agreement with the seasonality of deposition of these elements which is high in summer when the contribution of dry deposition is the highest.

Table 1: Left part, annual total, wet and dry deposition fluxes (mg m^{-2} y^{-1}) of major nutrients and trace elements, and relative contribution (%) of wet periods on the total fluxes measured at Capo Cuittone, Corsica, between March 2008 and October 2011. Right part, annual deposition fluxes at various sites in Corsica available in the literature.

Ele-ment	Total Flux		*Wet Flux*	*Dry Flux*		Capo Cavallo[a]	Pirio[b]	Ostrioni[c]	Bavella[d]
	Average	Range	*Average*	*Average*	*% wet*	Fev. 1985 to Nov. 1987	Jan. 1995 to Mar. 1997	Jun. 2001 to May 2002	1984-1986
N	143 ±61	81-167	*107.6*	*19.9*	*84%*			355-377*	644-766
P	149 ±79	114-253	*73.3*	*75.4*	*49%*	24.2-40.1		43.2	
Fe	67 ±10	65-77	*56.2*	*10.5*	*84%*	395-406	118-156	1188	
Si	246 ±61	197-280	*206.7*	*39.3*	*84%*				
As	0.14 ±0.07	0.10-0.20	*0.10*	*0.06*	*64%*				
Cr	0.16 ±0.04	0.12-0.19	*0.11*	*0.05*	*69%*				
Cu	1.06 ±1.07	0.44-2.16	*0.72*	*0.33*	*68%*	2.3-3.7	0.7-1.4		
Mn	3.7 ±1.4	2.5-5.2	*2.15*	*1.20*	*64%*	12.5-15.2	6.3-6.5		
Ni	0.21 ±0.01	0.14-0.24	*0.16*	*0.03*	*84%*		0.4		
V	0.4 ±0.06	0.34-0.46	*0.33*	*0.07*	*83%*				
Zn	5.7 ±1.9	3.74-7.42	*2.51*	*3.20*	*44%*		4.2-6.1	6	

[a]: total bulk deposition from Bergametti et al. (1987 and 1992) and Remoudaki et al. (1991)
[b]: total bulk deposition from Ridame et al. (1999)
[c]: total bulk deposition from Guieu et al. (2010) and Markaki et al. (2010)
[d]: total wet deposition from Löye-Pilot et al. (1990)
** data obtained between jun. 2001 and May 2003*

Mineral dust deposition fluxes

The annual deposition fluxes of soil dust have been estimated from Al fluxes, considering an amount of Al of 7% (Guieu et al., 2010). The results show that the mean annual dust flux ranges from 1.39 to 1.94 g m^{-2} y^{-1}. Typically, 70% of the annual flux is related to 3 or 4 dust deposition

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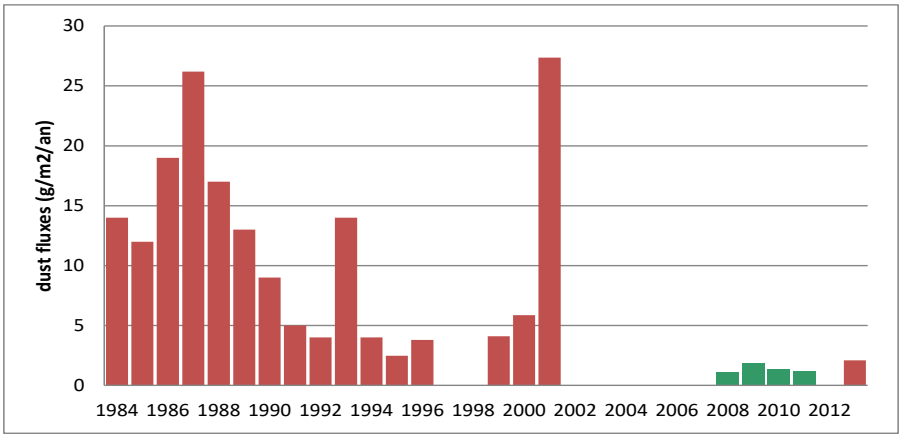
304 events during the year, in agreement with the sporadic pattern of dust input over the year
305 (Löye-Pilot et al., 1996). For example, a weekly maximum of 0.82 g m^{-2} w^{-1} is recorded during
306 the last week of April 2009, representing 44% of the total flux for this year. It is well known
307 that the intense dust events lead to a very high intra and inter-annual variability. Al deposition
308 data recording in Corsica show a range of dust fluctuation for the period 1985-2002 in Corsica
309 being to 4 to 28 g m^{-2} y^{-1} (Guieu et al., 2010), higher value being always associated with very
310 intense events (>10 g m^{-2}). Our values between 2008 and 2011 are lower than the range
311 previously measured between the 1980's and early 2000's, probably because no intense dust
312 event (>1 g m^{-2}) has been recorded during the sampling period (Figure 3). This trend is
313 consistent with the low annual deposition mass fluxes observed by Vincent et al. (2016)
314 recorded in Corsica and more generally in the western Mediterranean between 2011 and 2013
315 (max 2.1 g m^{-2} yr^{-1}). This result is also consistent with the decreasing trend in PM_{10}
316 concentrations over the Mediterranean region due to the decrease of dust contribution (Pey
317 et al., 2013). As mentioned by these authors, this is probably due to the variation in large scale
318 atmospheric circulation affecting dust atmospheric contents (lower values of the NAO indices
319 during the last two decades) as pointed by Moulin et al. (1997). However, the kind of
320 deposition collectors and the sampling sites being different, we cannot exclude effects of
321 sampling conditions on the obtained results.
322



323
324 **Figure 3: Time-series of dust fluxes (g m^{-2} an^{-1}) at various locations in Corsica issued from Löye-Pilot and Martin (1996) for**
325 **years between 1984 and 1994, from Ridame et al. (1999) for 1995 and 1996, from Guieu et al. (2010) for 2001-2002, from**
326 **Vincent et al. (2016) for 2013 (in red) and from this work for 2008-2011 (in green).**
327

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The case of high deposition events

Over our sampling period, the average weekly dust deposition is 0.028 ± 0.07g m^{-2} w^{-1}. In order to identify the outlier dust events in the data set, we selected the samples with weekly fluxes higher than the last 95th percentile of data, i.e. the 5 % of highest values. Thus, weekly dust fluxes > 0.104 g m^{-2} w^{-1} are considered as the extreme dust events. From this threshold, 12 samples are isolated and correspond to 10 dust events (2 events concern two successive sampling periods in April 2009 and June 2010). All these events correspond to wet deposition periods (Figure 1). Observations of remote sensing data from MODIS or SEVIRI confirm that these events are due to intense Saharan dust plumes associated with clouds (not shown). Eight of these events happened in spring (between March and June), 1 in fall and 1 in winter, in agreement with the seasonal pattern of dust plumes in western Mediterranean and in particular in Corsica (Bergametti et al., 1989; Salvador et al., 2014). The first week of one of the two dust events which fall in two sampling periods corresponds also with the high episode of As deposition (0.1 mg m^{-2} wk^{-1}) recorded in June 2010. This weekly As flux represents the annual flux measured in a remote environment like Scandinavia forest (0.1 mg m^{-2} yr^{-1} on average between 2002-2005; Hovmand et al., 2008). It is much larger than the weekly fluxes recorded in an urban environment as Venice where the reported fluxes range from 0.7 to 367 µg m^{-2} wk^{-1} between 2005-2010 (Morabito et al., 2014). The As/Al ratio (0.011) for this event show a large enrichment in As (x16) in comparison to the average of other intense dust events (7.10^{-4}), which are in agreement with the crustal ratio (Mason and Moore, 1982). The identified sources of atmospheric particulate As are coal-fired industries, waste-incineration, oil refining, mining and fossil fuel combustion (Wai et al., 2016). In the given sample, an enrichment in comparison to the other intense dust events is also observed for P (x12), Sr (x9), Cu and Zn (x6) whereas no significant enrichment is observed during the second sampling week of the dust event. Besides dust and marine aerosol, the biomass burning and fossil fuel combustion are the main sources of anthropogenic Cu, P, Sr and Zn (Mahowald et al., 2008; Nava et al., 2015). Moreover, the particulate filter corresponding with this event is brown-grey, showing a probable mixing between dust and black carbon (not shown). During the As-rich deposition week, the back-trajectories show that the air masses come from mainly from South in concomitance with a high dust intrusion in the western Mediterranean basin (Figure 4), no intense biomass burning event is recorded during this period on the pathways of back-

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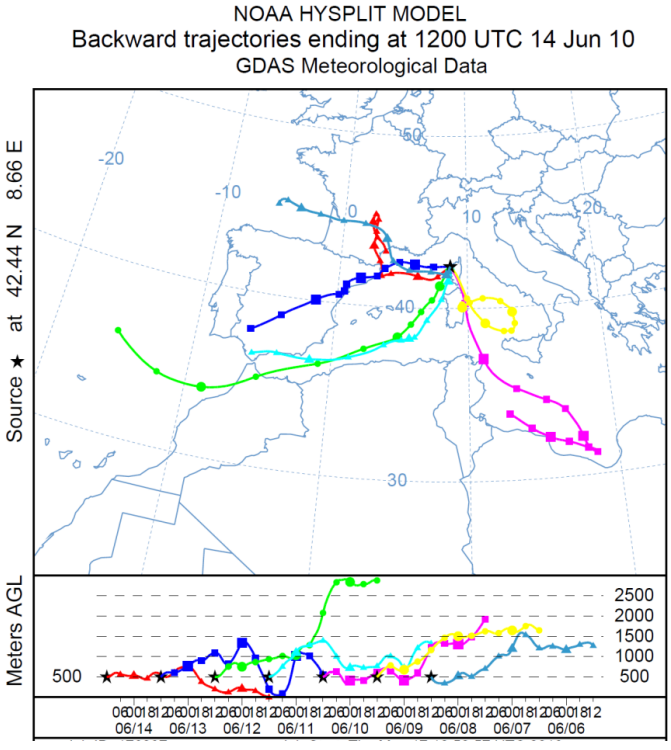
359 trajectories . These observations suggest that deposition from a likely local combustion source
360 occurred before or at the beginning of the dust deposition event.



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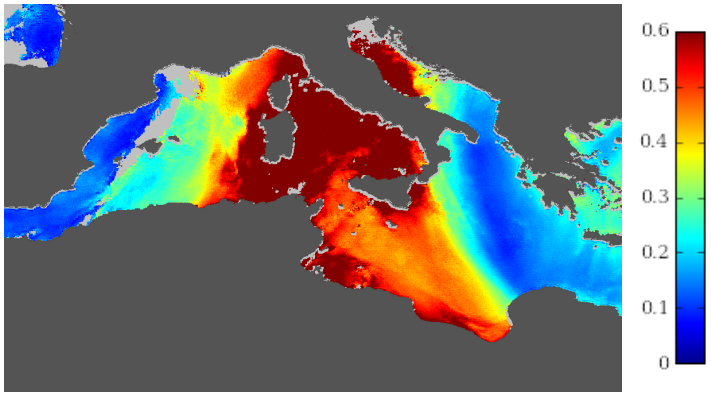

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363 **Figures 4: (a) Pathways of 3-day back-trajectories at 500 m from NOAA HYSPLITT model for the week of As-rich event and**
364 **(b) MSG/SEVIRI daily (daytime) mean aerosol optical depth oven Ocean for the 10th of June 2010 in agreement with the**
365 **intrusion of an intense Saharan dust event during this week.**

366

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The dust flux associated to these most intense dust deposition events represents 56% of the total dust flux on the 3.5 years of recording. The contribution of dust events on the fluxes Fe and Si represents 52% and 57% of their total fluxes respectively. Our results confirm the important role of these sporadic dust events on the inputs of these elements. In agreement with previous observations, Si and Fe fluxes present also a good correlation with Al fluxes (R^2= 0.97 and 0.96, respectively) and mean mass ratios (Si/Al = 2.5 and Fe/Al = 0.57) are typical of Saharan dust (Formenti et al., 2008 and 2011), supporting more generally the important role of dust deposition on Si and Fe inputs.

For N and P, the contribution of the outlier dust events is lower and reaches 10 and 15% respectively, and even 11% for P if the As-dust mixed event is excluded. That means that other sources than soil dust dominate the fallouts of these species. For trace metals, the high dust deposition events represent around 1/3 of total fluxes for Cr, Mn, Ni and V, whereas the contribution is low for As (10% without the intense event), Cu (16% and even 12% excluding As-dust mixed event) and Zn (9% and even 6% excluding As-dust mixed event). Keeping in mind that no high dust deposition event >1 g m^{-2} has been recorded during our 2008-2011 period of sampling, our data confirm that African dust wet deposition constitutes the major atmospheric source for Fe and Si to the northwestern Mediterranean and an important source for Cr, Mn, Ni and V (1/3 of their total fluxes).

Source apportionment and background deposition

In order to perform a source apportionment by the PMF method, we excluded the 12 samples corresponding to the high African dust deposition events in order to address background atmospheric deposition. We evaluated PMF solutions with two to six factors. Finally, a solution with 4 factors has been chosen since it is the optimum solution coupling a good agreement with our understanding of sources identification and the indicator of PMF optimization. The 4-factor solution was the most stable with a sharper decrease in the Q/Qexp trend and a constant global minimum Q value among 100 random runs. In terms of the stability of the PMF analysis, all factors of the 4-factor solution were reproduced in 100% of bootstrap runs, demonstrating that this solution was stable. No correlation between 4 factors has been observed, indicating that they represented distinct sources. The 4-factor solution enables to apportion the results between 4 sources: a dust factor (related to Si, Al, Ca, Fe and Ti), a marine aerosol factor (related to Na, Mg, Sr), an anthropogenic source factor (related to Cr, Pb, V, N,

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399 Sexc) and a biomass burning aerosol factor (related to K, P, Zn, Cu and Mn) (Figure 5). This
400 source identification is based on the presence of signature elements as well as on the ratio of
401 the elements in the source profile. The source identification was also supported by the
402 seasonal contributions of each source to the atmospheric fluxes shown in Fig. 5. The
403 identification of factors was supported by the previous works on source apportionment of
404 aerosol particles in Mediterranean (Calzolai et al., 2015, Becagli et al., 2012 and 2017).
405 However, even if the source oil-combustion rich in Ni-V has been typically observed in central
406 Mediterranean, no PMF solution (2- to 6-factor solutions) did enable us extracting a Ni-V
407 factor, corresponding to a ship plume signature. Even if this source could be important for
408 aerosol concentrations (Becagli et al., 2017), it does not seem to be important for deposition
409 in Corsica. The marine factor is marked by Na, ssS, Mg and Sr, at least 60% of their fluxes
410 corresponding to this source. The elemental ratio obtained for this source (Mg/Na =0.13;
411 K/Na =0.064; Ca/Na =0.070, Sr/Na =$8.5x10^{-4}$) are in agreement with the typical elemental ratio
412 in seawater (Mg/Na =0.12; K/Na =0.037; Ca/Na =0.038; Sr/Na =$7x10^{-4}$; Bowen, 1979). It is
413 known that sea salt aerosol concentrations are a function of surface wind speeds (O'Downd
414 et al., 1993). The seasonal contribution of this source is consistent with a wind production in
415 fall and winter, with a maximum of deposition in agreement with the maximum of rain in fall.
416 Na represents on the total mass at least 15% of marine, dust and biomass burning source due
417 to the high influence of marine environment on the Corsica Island. The factor identified as
418 dust source, marked by Al, Fe, Si, Ti, is in agreement with the typical seasonal variation of dust
419 deposition in Corsica with a maximum in spring and in fall (Bergametti et al., 1989). Moreover,
420 the elemental ratios (Si/Al =2.7; Fe/Al = 0.72; Ti/Fe = 0.12) correspond with a Saharan dust
421 signature (Si/Al between 2 and 4 and Ti/Fe between 0.1 and 0.15; Formenti et al., 2014). The
422 biomass burning/waste source is mainly characterized by Cu, K, P and Zn. K is commonly
423 associated to waste/biomass burning or wood combustion (Dall'osto et al., 2013, Nava et al.,
424 2015). The maximum of this source deposition in summer, in spite of a minimum of rain,
425 corresponds with the intense forest fires observed in Mediterranean in this period and which
426 the extend impacts all the basin (Bossioli et al., 2016). Finally, Cr, Pb, V, Sexc and N are the
427 characterizing elements found in the anthropogenic source. Even with >4-factor solutions, no
428 profile distinguishing N, Sexc and metals is emphasized by PMF, suggesting a common source.
429 Thus, the major contribution of N and Sexc in mixing with metals suggest that this source
430 correspond with the secondary aerosols formed in air masses issued from combustion sources

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(traffic, industrial). The two sources of combustion identified by PMF, i.e. biomass burning and anthropogenic sources, have previously been observed in background aerosols in Corsica (Arndt et al., 2017; Claeys et al. 2017).

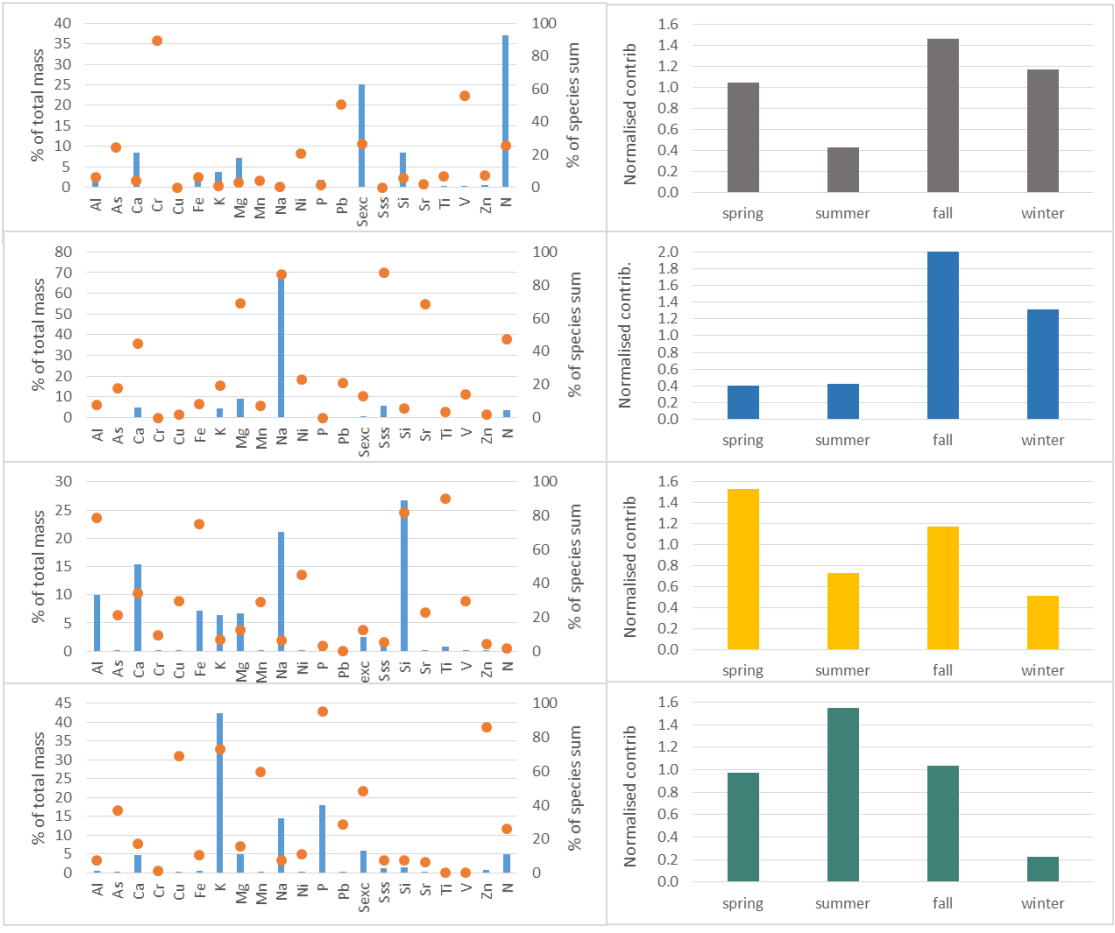

Fig. 5: PMF-derived profiles of the four sources identified. Left column, from top to bottom: (a) anthropogenic, (b) marine, (c) dust and (d) biomass burning including fraction of total mass (blue columns) and fraction of elemental sum (orange circles). Right column: seasonal contributions of these four respective sources.

In figure 6, we show the relative contribution from the identified sources to background deposition flux of nutrients and trace metals. The results show that the combustion sources (biomass burning or anthropogenic) predominates in the background inputs of major nutrients and TMs, except Fe and Si. Even for background deposition, the source apportionment of Fe and Si is quasi-similar to Al (correlation coefficient close to 1 for the

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elemental fluxes even out of intense events and ratio Si/Al and Fe/Al characteristics of mineral dust). These results suggest that even if the PMF apportionment source distributes the fluxes from the 4 sources, in all the cases, the fluxes of Fe and Si are associated to mineral dust sources, most probably Saharan dust.

Concerning major nutrients, P deposition is highly associated to biomass burning inputs out of the most intense dust deposition events. Considering that the dust deposition accounts for 15% of the total P deposition flux (incl. intense dust deposition events + background deposition), almost 85% of P inputs are associated to the deposition of biomass burning/waste/wood-related aerosol. This confirms the importance to consider the biomass combustion source to estimate the role of this element on marine environment in Mediterranean. For N deposition, the inputs associated to marine sources are quasi-similar to the inputs from combustion sources. Thus, almost 50% of N fluxes is explained by the marine source. Several works observed that the depletion of chloride (Cl) and the simultaneous occurrence of NO_3 in sea salt aerosol particles is due to the reaction between NaCl and HNO_3 when maritime and anthropogenic air masses are mixed, in Mediterranean environments (Sellegri et al., 2001; Bardouki et al., 2003; Pey et al., 2009) and in particular in Corsica (Claeys et al., 2017). The contribution of marine source to N deposition is probably due to the deposition of these processed seasalt particles. In the case of the anthropogenic source, the good correlation obtained between N and Sexc supports for these elements a common origin which is associated to the inorganic secondary aerosol, i.e. ammonium sulfate and ammonium nitrate. It is known that the deposition efficiency of particles in the coarse mode, as sea salts, is higher than the one of fine particles, as inorganic secondary aerosols. Our results suggests that the addition of nitrate on sea salt particles could be a key process to estimate the N atmospheric deposition fluxes to Mediterranean surface waters. Recent works suggest that a large part of nitrogen associated to anthropogenic secondary aerosol could be soluble organic nitrogen (Violaki et al., 2015). Thus, the observed difference in sources of deposited N could also mean a difference in N speciation in the fallout (inorganic nitrate vs organic nitrate).

For trace metals, the marine source present the lowest contribution. The biomass burning/waste source is clearly predominant for Cu, Mn and Zn, whereas atmospheric fluxes of Cr and Ni are largely linked to the anthropogenic source. Fu et al. (2017) show that the Cr deposition in Cape Corsica, even during intense dust event is originated from an anthropogenic source, suggesting a contamination by a local source. Even if the Cape Corsica

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and our sites of deposition measurements are distant by about 100 km, both suggest that Cr deposition is controlled by an anthropogenic source. For Zn, Guieu et al. (2010) also showed a large contribution of non-dust source. Our works enable to support their conclusions and to identify rather a biomass combustion source than a fossil fuel or industrial origin. It appears that the deposition of Cu, Mn, Ni and V is influenced at least for 20% by dust deposition out of intense events. That means that for these trace metals, the natural dust inputs can represented up to 50% of annual fluxes.

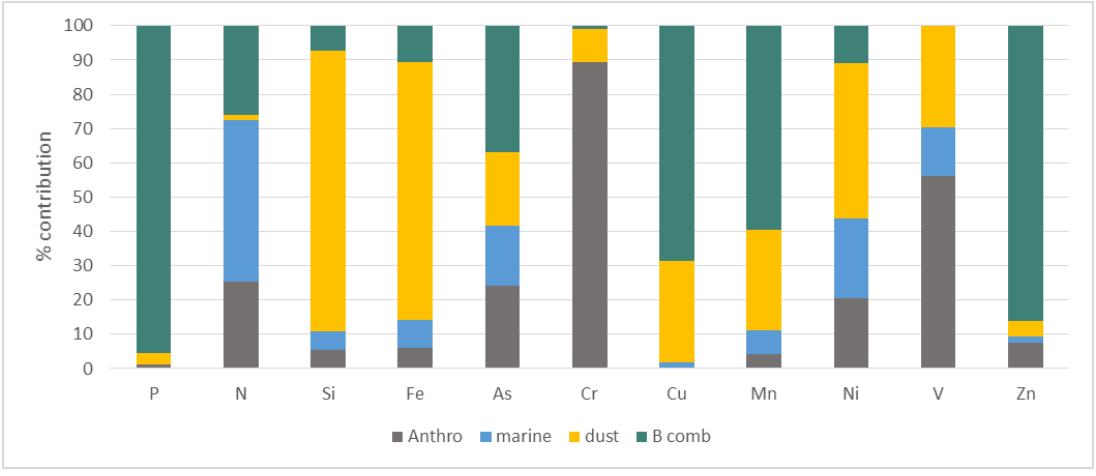

Fig. 6: Relative contribution of each of the 4 identified sources (Antro= anthropogenic, marine, dust and B comb= biomass combustion) to the "background" mass fluxes of nutrients and TMs (i.e. excluding the 12 most intense African dust deposition samples).

Biogeochemical applications

The typical N:P molar ratio in seawater required by marine phytoplankton corresponds to the Redfield ratio of 16. This ratio is generally higher in Mediterranean surface seawater, with values ranging from 20 to 100 on the basin (Ribera d'Alcala et al., 2003). The atmospheric input to the Mediterranean Sea displays a high N:P ratio for dissolved or soluble inorganic forms (Herut and Krom, 1996), which could be one possible reason of the high N:P ratio in Mediterranean deep sea waters (Markaki et al., 2010; Krom et al., 2010).

In our data set, the yearly deposition mass fluxes measured for N and P are quasi equivalent (0.14-0.15 g m^{-2} yr^{-1}; Table 1). However, weekly measurements show a very large variability in P fluxes, contrary to N. Hence, a large variability in the N:P molar ratio is observed in the

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atmospheric inputs at short time scales. A mean value of 35 is observed for the molar ratio but the weekly ratio range from 0.4 to 1220. The observed N enrichment in comparison to the Redfield ratio is in agreement with previous observations showing a preponderance of N relative to P in the atmospheric deposition over the Mediterranean Sea (Markaki et al., 2010). However, a detailed analysis shows that the atmospheric ratio is equal or higher than Redfield ratio only for 25% of samples, and higher than 160 only for 3 samples (4%). This value reaches 36% for the wet periods. The higher ratio observed in wet deposition could be linked to a wash-out effect of the gaseous N species (as NOx, NH_3..) by rain (Ochoa-Hueso et al., 2011). The intense dust events present an average ratio of 3.5, which is lower than previously reported for dust deposition (between 30 and 70; Morales-Baquero et al., 2013). However, this value is consistent with the typical N:P ratio in Saharan dust aerosols which is around to 2.5. The highest N:P ratio are generally due to the reactivity of dust aerosol with gaseous nitric acid to form nitrate on dust particles (Desboeufs et al., 2014). Our data suggest that the effect of mixing between dust and nitric acid do not commonly affect atmospheric dust deposited in Corsica. On the contrary, the mixing between sea salt and nitric acid identified with the source apportionment could be a source of N during the fall and winter period, when the biomass burning source is negligible. It appears also that the lowest N:P ratio are mainly observed from May to September (Figure 7). During this period, the atmospheric deposition becomes the main sources of nutrients since the Mediterranean is highly stratified and the surface is depleted in nutrients. Thus, in these conditions, the atmospheric inputs will be deficient in N relative to phytoplankton requirements. Studies show that phytoplankton growth in western Mediterranean waters is usually limited by a lack of phosphate, rather than nitrate in summer (Lazzari et al., 2016; Richon et al., 2017), even if phosphorus addition experiments have indicated also N limitation in this period (Tanaka et al., 2011). Our results suggest that the role of atmospheric aerosol input will be rather favorable in case of P-starvation of surface seawater. However, even if the N:P ratio from this study were obtained with comparable deposition collectors than previous literature (e.g. Markaki et al., 2010), it has to be kept in mind the deposition collectors were not optimized for gaseous N fluxes measurements, and the N:P ratio could be underestimated. The temporal evolution of marine N and P concentrations since 1985 has shown a high sensitivity to anthropogenic atmospheric deposition and they are expected to decline in the coming decades due to mitigation/control of pollutant emissions (Moon et al. 2016). Due to the high contribution of anthropogenic

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532 deposition sources on P fluxes emphasized by our study, it is important to include precise
533 anthropogenic P emissions to estimate the impact of atmospheric deposition on carbon fluxes
534 and phytoplankton biomass in the future (Richon et al., 2017).
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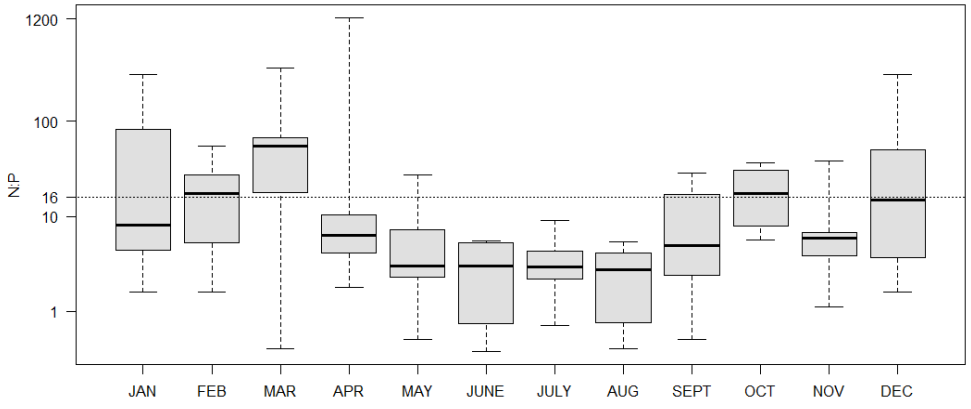

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537 **Figure 7: Box plots of monthly molar N:P ratio in deposition samples, showing the third quartile (Q3) and first quartile (Q1)**
538 **range of the data and minimum and maximum of data. For the sake of comparison, The N:P scale is logarithmic and the**
539 **Redfield ratio is displayed.**

540 ## 4. Conclusion

541 In a context of anthropogenic changes, in order to assess how the evolution of chemical
542 atmospheric forcing will modify the marine nutrient cycling, it is crucial to distinguish between
543 anthropogenic vs natural atmospheric inputs of nutrients to the oligotrophic Mediterranean
544 surface waters. We monitored elemental atmospheric deposition on a weekly basis over 3.5-
545 years (March 2008-October 2011) at a coastal site on the western coast of Corsica. The
546 contribution of 4 different source types to the fallout of nutrients and trace metals was
547 determined by statistical PMF method, namely desert dust, seasalt, anthropogenic activities,
548 and biomass combustion sources. The data show that Si and Fe fluxes are typically related to
549 African dust deposition, with fluxes dominated by high dust deposition events. A typical Si:Al
550 ratio of 2.5 is obtained whatever the samples. That shows that Al is a good proxy to estimate
551 the Si fluxes in Mediterranean region since Si is often not measured when X-Ray fluorescence
552 spectrometry is not available due its lost by HF digestion during the protocol of chemical
553 analysis. Our results on the mineral dust fallout is of the same order of magnitude that of 2013
554 at another site in Corsica (Vincent et al., 2016) and confirm the fact that dust deposition has

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strongly decreased in Corsica compared to the 1980's and 1990's, with no observed occurrence of a high dust deposition event larger than 1 g m^{-2} in 2008-2011 and 2013.

Atmospheric fluxes of Cu, Mn, Ni and V are also associated at least at 50% to mineral dust deposition, whereas half of atmospheric fluxes is issued either from biomass burning particles deposition (Cu and Mn), either from fossil fuel combustion (V), either both (Ni). The anthropogenic/combustion sources govern the atmospheric fluxes of major nutrient N and P, with a predominance of biomass combustion source for P and secondary aerosols for N. Dust deposition is contributing around 15% of deposited P at the yearly time scale. Our result show that these combustion sources need to be considered in P deposition modelling. Finally, Zn or Cr deposition is very largely associated to continuous combustion sources.

This work is a first tentative assessment of the origin of nutrients and trace metals deposited in the western Mediterranean. Of course, our study is not sufficient to apprehend the spatial variability of the influence of the identified source types over the basin. It needs to be supported by other studies of source apportionment on deposition samples in the region.

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References

Amato, F., Alastuey, A., Karanasiou, A., Lucarelli, F., Nava, S., Calzolai, G., Severi, M., Becagli, S., Gianelle, V. L., Colombi, C., Alves, C., Custadio, D., Nunes, T., Cerqueira, M., Pio, C., Eleftheriadis, K., Diapouli, E., Reche, C., Minguillon, M. C., Manousakas, M. I., Maggos, T., Vratolis, S., Harrison, R. M., and Querol, X.: AIRUSE-LIFE+: a harmonized PM speciation and source apportionment in five southern European cities, Atmos. Chem. Phys., 16, 3289-3309, doi:10.5194/acp-16-3289-2016, 2016.

Arndt, J., Sciare, J., Mallet, M., Roberts, G. C., Marchand, N., Sartelet, K., Sellegri, K., Dulac, F., Healy, R. M., and Wenger, J. C.: Sources and mixing state of summertime background aerosol in the north-western Mediterranean basin, Atmos. Chem. Phys., 17, 6975-7001, doi:10.5194/acp-17-6975-2017, 2017.

Bardouki, H., Liakakou, H., Economou, C., Sciare, J., Smolik, J., Zdimal, V., Eleftheriadis, K., Lazaridis, M., Dye, C., and Mihalopoulos, N.: Chemical composition of size-resolved atmospheric aerosols in the

© Author(s) 2018. CC BY 4.0 License.

592 eastern Mediterranean during summer and winter, Atmos. Environ., 37, 195-208,
593 doi:10.1016/S1352-2310(02)00859-2, 2003.

594 Becagli, S., Sferlazzo, D. M., Pace, G., di Sarra, A., Bommarito, C., Calzolai, G., Ghedini, C., Lucarelli, F.,
595 Meloni, D., Monteleone, F., Severi, M., Traversi, R., and Udisti, R.: Evidence for heavy fuel oil
596 combustion aerosols from chemical analyses at the island of Lampedusa: a possible large role of ships
597 emissions in the Mediterranean, Atmos. Chem. Phys., 12, 3479-3492, doi:10.5194/acp-12-3479-
598 2012, 2012.

599 Becagli, S., Anello, F., Bommarito, C., Cassola, F., Calzolai, G., Di Iorio, T., di Sarra, A., Gómez-Amo, J.-
600 L., Lucarelli, F., Marconi, M., Meloni, D., Monteleone, F., Nava, S., Pace, G., Severi, M., Sferlazzo, D.
601 M., Traversi, R., and Udisti, R.: Constraining the ship contribution to the aerosol of the central
602 Mediterranean, Atmos. Chem. Phys., 17, 2067-2084, doi:10.5194/acp-17-2067-2017, 2017.

603 Bergametti, G., Apports de matière par voie atmosphérique à la Méditerranée Occidentale: aspects
604 géochimiques et météorologiques, PhD, 296 pp., Univ. Paris VII, Paris, 1987.

605 Bergametti, G., Dutot, A. L., Buat-Ménard, P., Losno, R., and Remoudaki, E.: Seasonal variability of the
606 elemental composition of atmospheric aerosol particles over the Northwestern Mediterranean,
607 Tellus, 41B, 353-361, doi:10.1111/j.1600-0889.1989.tb00314.x, 1989.

608 Bergametti, G., Remoudaki, E., Losno, R., Steiner, E., Chatenet B., and Buat-Ménard, P.: Sources,
609 transport and deposition of atmospheric phosphorus over the northwestern Mediterranean, J.
610 Atmos. Chem., 14, 501-513, doi:10.1007/BF00115254, 1992.

611 Bethoux J.-P., Courau, P., Nicolas, E., Ruiz-Pino, D.: Trace-metal pollution in the Mediterranean-sea.
612 Oceanol. Acta, 13, 481-488, 1990.

613 Bonnet, S., and Guieu, C.: Atmospheric forcing on the annual iron cycle in the western Mediterranean
614 Sea: A 1-year survey, J. Geophys. Res.-Oceans, 111, 2006.

615 Bossioli, E., Tombrou, M., Kalogiros, J., Allan, J., Bacak, A., Bezantakos, S., Biskos, G., Coe, H., Jones, B.
616 T., Kouvarakis, G., Mihalopoulos, N., and Percival, C. J.: Atmospheric composition in the Eastern
617 Mediterranean: Influence of biomass burning during summertime using the WRF-Chem model,
618 Atmos. Environ., 132, 317-331, doi:10.1016/j.atmosenv.2016.03.011, 2016.

619 Bowen, H. J. M.: Environmental chemistry of the elements, Academic Press, 1979.

620 Calzolai, G., Nava, S., Lucarelli, F., Chiari, M., Giannoni, M., Becagli, S., Traversi, R., Marconi, M., Frosini,
621 D., Severi, M., Udisti, R., di Sarra, A., Pace, G., Meloni, D., Bommarito, C., Monteleone, F., Anello, F.,
622 and Sferlazzo, D. M.: Characterization of PM10 sources in the central Mediterranean, Atmos. Chem.
623 Phys., 15, 13939-13955, doi:10.5194/acp-15-13939-2015, 2015.

624 Chester, R., Nimmo, M., and Keyse, S.: The influence of saharan and middle eastern desert-derived
625 dust on the trace metal composition on Mediterranean aerosols and rainwaters: an overview, in: The
626 impact of Desert dust across the Mediterranean, edited by: Guerzoni, S., and Chester, R., Kluwer
627 Academic Publishers, Dordrecht, 253-273, 1996.

628 Claeys, M., Roberts, G., Mallet, M., Arndt, J., Sellegri, K., Sciare, J., Wenger, J., and Sauvage, B.: Optical,
629 physical and chemical properties of aerosols transported to a coastal site in the western
630 Mediterranean: a focus on primary marine aerosols, Atmos. Chem. Phys., 17, 7891-7915,
631 doi:10.5194/acp-17-7891-2017, 2017.

632 Dall'Osto, M., Querol, X., Amato, F., Karanasiou, A., Lucarelli, F., Nava, S., Calzolai, G., and Chiari, M.:
633 Hourly elemental concentrations in PM2.5 aerosols sampled simultaneously at urban background
634 and road site during SAPUSS – diurnal variations and PMF receptor modelling, Atmos. Chem. Phys.,
635 13, 4375-4392, doi:10.5194/acp-13-4375-2013, 2013.

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636 Desboeufs, K., Leblond, N., Wagener, T., Bon Nguyen, E., and Guieu, C.: Chemical fate and settling of
637 mineral dust in surface seawater after atmospheric deposition observed from dust seeding
638 experiments in large mesocosms, Biogeosci., 11, 5581-5594, doi:10.5194/bg-11-5581-2014, 2014.

639 Diapouli, E., Manousakas, M. I., Vratolis, S., Vasilatou, V., Pateraki, S., Bairachtari, K. A., Querol, X.,
640 Amato, F., Alastuey, A., Karanasiou, A. A., Lucarelli, F., Nava, S., Calzolai, G., Gianelle, V. L., Colombi,
641 C., Alves, C., Custodio, D., Pio, C., Spyrou, C., Kallos, G. B., and Eleftheriadis, K.: AIRUSE-LIFE +:
642 estimation of natural source contributions to urban ambient air PM10 and PM2.5 concentrations in
643 southern Europe - implications to compliance with limit values, Atmos. Chem. Phys., 17, 3673-3685,
644 doi:10.5194/acp-17-3673-2017, 2017.

645 Formenti, P., Rajot, J. L., Desboeufs, K., Caquineau, S., Chevaillier, S., Nava, S., Gaudichet, A., Journet,
646 E., Triquet, S., Alfaro, S., Chiari, M., Haywood, J., Coe, H., and Highwood, E.: Regional variability of the
647 composition of mineral dust from western Africa: Results from the AMMA SOP0/DABEX and DODO
648 field campaigns, J. Geophys. Res.-Atmos., 113, *D00C13,* doi:10.1029/2008JD009903, 2008.

649 Formenti, P., Schütz, L., Balkanski, Y., Desboeufs, K., Ebert, M., Kandler, K., Petzold, A., Scheuvens, D.,
650 Weinbruch, S., and Zhang, D.: Recent progress in understanding physical and chemical properties of
651 African and Asian mineral dust, Atmos. Chem. Phys., 11, 8231-8256, 10.5194/acp-11-8231-2011,
652 2011.

653 Formenti, P., Caquineau, S., Desboeufs, K., Klaver, A., Chevaillier, S., Journet, E., and Rajot, J. L.:
654 Mapping the physico-chemical properties of mineral dust in western Africa: mineralogical
655 composition, Atmos. Chem. Phys., 14, 10663-10686, doi:10.5194/acp-14-10663-2014, 2014.

656 Fu, Y., Desboeufs, K., Vincent, J., Bon Nguyen, E., Laurent, B., Losno, R., and Dulac, F.: Estimating
657 chemical composition of atmospheric deposition fluxes from mineral insoluble particles deposition
658 collected in the Western Mediterranean region, Atmos. Meas. Tech., 10, 4389–4401,
659 doi:10.5194/amt-10-4389-2017, 2017.

660 Gallisai, R., Peters, F., Volpe, G., Basart, S., and Baldasano, J. M.: Saharan Dust Deposition May Affect
661 Phytoplankton Growth in the Mediterranean Sea at Ecological Time Scales, PLoS ONE, 9, e110762,
662 doi:10.1371/journal.pone.0110762, 2014.

663 Gratz, L. E., Keeler, G. J., Morishita, M., Barres, J. A., and Dvonch, J. T.: Assessing the emission sources
664 of atmospheric mercury in wet deposition across Illinois, Sci. Total Environ., 448, 120-131,
665 doi:10.1016/j.scitotenv.2012.11.011, 2013.

666 Guerzoni, S., Chester, R., Dulac, F., Herut, B., Loÿe-Pilot, M.-D., Measures, C., Migon, C., Molinaroli, E.,
667 Moulin, C., and Rossini, P.: The role of atmospheric deposition in the biogeochemistry of the
668 Mediterranean Sea, Progr. Oceanog., 44, 147-190, doi:10.1016/S0079-6611(99)00024-5, 1999a.

669 Guerzoni, S., Molinaroli, E., Rossini, P., Rampazzo, G., Quarantotto, G., De Falco, G., and Cristini, S.:
670 Role of desert aerosol in metal fluxes in the Mediterranean area, Chemosphere, 39, 229-246, 1999b.

671 Guieu, C., Chester, R., Nimmo, M., Martin, J.-M., Guerzoni, S., Nicolas, E., Mateu, J., and Keyse, S.:
672 Atmospheric input of dissolved and particulate metals to the northwestern Mediterranean, Deep Sea
673 Res. II, 44, 655-674, doi:10.1016/S0967-0645(97)88508-6, 1997.

674 Guieu, C., Loÿe-Pilot, M. D., Benyahya, L., and Dufour, A.: Spatial variability of atmospheric fluxes of
675 metals (Al, Fe, Cd, Zn and Pb) and phosphorus over the whole Mediterranean from a one-year
676 monitoring experiment: Biogeochemical implications, Mar. Chem., 120, 164-178,
677 doi:10.1016/j.marchem.2009.02.004, 2010.

678 Heimburger, A., Losno, R., Triquet, S., Dulac, F., and Mahowald, N. C. G.: Direct measurements of
679 atmospheric iron, cobalt, and aluminum-derived dust deposition at Kerguelen Islands, Global
680 Biogeochem. Cycles, 26, doi:10.1029/2012gb004301, 2012.

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681 Herut B., and Krom, M.: Atmospheric Input of Nutrients and Dust to the SE Mediterranean. In: Guerzoni
682 S., Chester R. (eds) The Impact of Desert Dust Across the Mediterranean. Environmental Science and
683 Technology Library, vol 11. Springer, Dordrecht, 1996.

684 Hovmand, M. F., Kemp, K., Kystol, J., Johnsen, I., Riis-Nielsen, T., and Pacyna, J. M.: Atmospheric heavy
685 metal deposition accumulated in rural forest soils of southern Scandinavia, Environ. Poll., 155, 537-
686 541, doi:10.1016/j.envpol.2008.01.047, 2008.

687 Keeler, G.J., Landis, M.S., Norris, G.A., Christianson, E.M., and Dvonch, J.T.: Sources of mercury wet
688 deposition in Eastern Ohio, USA, Environ. Sci. Technol., 40, 5874-5881, doi:10.1021/es060377q,
689 2006.

690 Krom, M. D., Emeis, K. C., and Van Cappellen, P.: Why is the Eastern Mediterranean phosphorus
691 limited?, Prog. Oceanog., 85, 236-244, doi:10.1016/j.pocean.2010.03.003, 2010.

692 Lazzari, P., Solidoro, C., Salon, S., and Bolzon, G.: Spatial variability of phosphate and nitrate in the
693 Mediterranean Sea: A modeling approach, Deep Sea Res. I, 108, 39-52,
694 doi:10.1016/j.dsr.2015.12.006, 2016.

695 Loÿe-Pilot, M. D., and Martin, J. M.: Saharan Dust Input to the Western Mediterranean: An Eleven Years
696 Record in Corsica, in: The Impact of Desert Dust Across the Mediterranean, edited by: Guerzoni, S.,
697 and Chester, R., Kluwer, 191-199, 1996.

698 Loÿe-Pilot, M.D., Guieu, C., Ridame, C., Atmospheric bulk fluxes of natural and pollutant metals to the
699 north western Mediterranean; their trend over the past 15 years (1985–2000). Atmospheric
700 Transport and Deposition of Pollutants into the Mediterranean Sea —Final Reports on Research
701 Project. UNEP/MAP, Athens, pp. 35–54. MAP Technical Reports Series, 133, 2001.

702 Loÿe-Pilot, M.D., Martin, J.M. & Morelli, J.: Atmospheric input of inorganic nitrogen to the Western
703 Mediterranean, Biogeochem., 9, 117-134, doi:10.1007/BF00692168, 1990.

704 Mahowald, N., Jickells, T. D., Baker, A. R., Artaxo, P., Benitez-Nelson, C. R., Bergametti, G., Bond, T. C.,
705 Chen, Y., Cohen, D. D., Herut, B., Kubilay, N., Losno, R., Luo, C., Maenhaut, W., McGee, K. A., Okin, G.
706 S., Siefert, R. L., and Tsukuda, S.: Global distribution of atmospheric phosphorus sources,
707 concentrations and deposition rates, and anthropogenic impacts, Global Biogeochem. Cycles, 22,
708 doi:10.1029/2008gb003240, 2008.

709 Markaki, Z., Loÿe-Pilot, M. D., Violaki, K., Benyahya, L., and Mihalopoulos, N.: Variability of atmospheric
710 deposition of dissolved nitrogen and phosphorus in the Mediterranean and possible link to the
711 anomalous seawater N/P ratio, Mar. Chem., 120, 187-194, doi:10.1016/j.marchem.2008.10.005,
712 2010.

713 Mason, B., and Moore, C.B.: Principles of Geochemistry, 1982.

714 Morabito, E., Contini, D., Belosi, F., Stortini, A. M., Manodori, L., and Gambaro, A.: Atmospheric
715 Deposition of Inorganic Elements and Organic Compounds at the Inlets of the Venice Lagoon, Adv.
716 Meteor., 2014, 10, doi:10.1155/2014/158902, 2014.

717 Morales-Baquero, R., Pulido-Villena, E., and Reche, I.: Chemical signature of Saharan dust on dry and
718 wet atmospheric deposition in the south-western Mediterranean region, Tellus B, 65,
719 10.3402/tellusb.v65i0.18720, 2013.

720 Morel, F. M. M., and Price, N. M.: The Biogeochemical Cycles of Trace Metals in the Oceans, Science,
721 300, 944-947, doi:10.1126/science.1083545, 2003.

722 Moulin, C., Lambert, C. E., Dulac, F., and Dayan, U.: Control of atmospheric export of dust from North
723 Africa by the North Atlantic Oscillation, Nature, 387, 691-694, doi:10.1038/42679, 1997.

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Nava, S., Lucarelli, F., Amato, F., Becagli, S., Calzolai, G., Chiari,M., Giannoni, M., Traversi, R., and Udisti, R.: Biomass burning contributions estimated by synergistic coupling of daily and hourly aerosol composition records, Sci. Tot. Environ., 511, 11–20, doi:10.1016/j.scitotenv.2014.11.034, 2015.

Norris, G., Duvall, R., and Brown, S.: EPA Positive Matrix Factorization (PMF) 5.0 Fundamentals & User Guide, EPA, U.S., 2014.

O'Dowd, C. D. and Smith, M. H.: Physicochemical properties of aerosols over the northeast Atlantic: Evidence for wind-speed related submicron sea-salt aerosol production, J. Geophys. Res., 98, 1137–1149, doi:10.1029/92JD02302, 1993.

Ochoa-Hueso, R., Allen, E. B., Branquinho, C., Cruz, C., Dias, T., Fenn, M. E., Manrique, E., Perez-Corona, M. E., Sheppard, L. J., and Stock, W. D.: Nitrogen deposition effects on Mediterranean-type ecosystems: An ecological assessment, Environ. Poll., 159, 2265-2279, doi:10.1016/j.envpol.2010.12.019, 2011.

Paatero, P., and Tapper, U.: Positive matrix factorization — a non-negative factor model with optimal utilization of error-estimates of data values, Environmetrics, 5, 111–126, doi:10.1002/env.3170050203, 1994.

Paatero, P. : Least squares formulation of robust non-negative factor analysis, Chemom. Intell. Lab. Syst., 37, 23-35, 1997.

Pey, J., Querol, X., and Alastuey, A.: Variations of levels and composition of PM10 and PM2.5 at an insular site in the Western Mediterranean, Atmos. Res., 94, 285–299, doi:10.1016/j.atmosres.2009.06.006, 2009.

Polissar, A.V., Hopke, P.K., Poirot, R.L.: Atmospheric aerosol over Vermont: chemical composition and sources, Environ Sci Technol, 35, 4604-4621, doi:10.1021/es0105865, 2001.

Pulido-Villena, E., Rérolle, V., and Guieu, C.: Transient fertilizing effect of dust in P-deficient LNLC surface ocean, Geophys. Res. Lett., 37, L01603, doi:10.1029/2009gl041415, 2010.

Remoudaki, E., Bergametti, G., and Losno, R.: On the dynamic of the atmospheric input of copper and manganese into the Western Mediterranean Sea, Atmos. Environ., 25A, 733-744, doi:10.1016/0960-1686(91)90072-F, 1991.

Ribera d'Alcalà, M., Civitarese, G., Conversano, F., and Lavezza, R. C.: Nutrient ratios and fluxes hint at overlooked processes in the Mediterranean Sea, J. Geophys. Res. - Oceans, 108, 10.1029/2002jc001650, 2003.

Richon, C., Dutay, J.-C., Dulac, F., Wang, R., Balkanski, Y., Nabat, P., Aumont, O., Desboeufs, K., Laurent, B., Guieu, C., Raimbault, P., and Beuvier, J.: Modeling the impacts of atmospheric deposition of nitrogen and desert dust-derived phosphorus on nutrients and biological budgets of the Mediterranean Sea, Prog. Oceanog.doi:10.1016/j.pocean.2017.04.009, 2017.

Ridame, C., Guieu, C., and Löye-Pilot, M. D.: Trend in total atmospheric deposition fluxes of aluminium, iron, and trace metals in the northwestern Mediterranean over the past decade (1985-1997), J. Geophys. Res., 104, 30127-30138, doi:10.1029/1999JD900747, 1999.

Ridame, C., Le Moal, M., Guieu, C., Ternon, E., Biegala, I. C., L'Helguen, S., and Pujo-Pay, M.: Nutrient control of N2 fixation in the oligotrophic Mediterranean Sea and the impact of Saharan dust events, Biogeosci., 8, 2773-2783, doi:10.5194/bg-8-2773-2011, 2011.

Salvador, P., Alonso-Pérez, S., Pey, J., Artinano, B., de Bustos, J. J., Alastuey, A., and Querol, X.: African dust outbreaks over the western Mediterranean Basin: 11-year characterization of atmospheric circulation patterns and dust source areas, Atmos. Chem. Phys., 14, 6759-6775, doi:10.5194/acp-14-6759-2014, 2014.

© Author(s) 2018. CC BY 4.0 License.

768 Sciare, J., Oikonomou, K., Favez, O., Liakakou, E., Markaki, Z., Cachier, H., and Mihalopoulos, N.: Long-
769 term measurements of carbonaceous aerosols in the Eastern Mediterranean: evidence of long-range
770 transport of biomass burning, Atmos. Chem. Phys., 8, 5551-5563, doi:10.5194/acp-8-5551-2008,
771 2008.

772 Sellegri, K., Gourdeau, J., Putaud, J.-P., and Despiau, S.: Chemical composition of marine aerosol in a
773 Mediterranean coastal zone during the FETCH experiment, J. Geophys. Res., 106, 12023,
774 doi:10.1029/2000JD900629, 2001.

775 Tanaka, T., Thingstad, T. F., Christaki, U., Colombet, J., Cornet-Barthaux, V., Courties, C., Grattepanche,
776 J. D., Lagaria, A., Nedoma, J., Oriol, L., Psarra, S., Pujo-Pay, M., and Van Wambeke, F.: Lack of P-
777 limitation of phytoplankton and heterotrophic prokaryotes in surface waters of three anticyclonic
778 eddies in the stratified Mediterranean Sea, Biogeosci., 8, 525-538, doi:10.5194/bg-8-525-2011, 2011.

779 Tovar-Sánchez, A., Duarte, C. M., Arrieta, J. M., and Sañudo-Wilhelmy, S.: Spatial gradients in trace
780 metal concentrations in the surface microlayer of the Mediterranean Sea, Frontiers Mar. Sci., 1,
781 doi:10.3389/fmars.2014.00079, 2014.

782 Vincent, J., Laurent, B., Losno, R., Bon Nguyen, E., Roullet, P., Sauvage, S., Chevaillier, S., Coddeville, P.,
783 Ouboulmane, N., di Sarra, A. G., Tovar-Sanchez, A., Sferlazzo, D., Massanet, A., Triquet, S., Morales
784 Baquero, R., Fornier, M., Coursier, C., Desboeufs, K., Dulac, F., and Bergametti, G.: Variability of
785 mineral dust deposition in the western Mediterranean basin and south-east of France, Atmos. Chem.
786 Phys., 16, 8749-8766, doi:10.5194/acp-16-8749-2016, 2016.

787 Violaki, K., Sciare, J., Williams, J., Baker, A. R., Martino, M., and Mihalopoulos, N.: Atmospheric water-
788 soluble organic nitrogen (WSON) over marine environments: a global perspective, Biogeosci., 12,
789 3131-3140, doi:10.5194/bg-12-3131-2015, 2015.

790 Wai, K.-M., Wu, S., Li, X., Jaffe, D. A., and Perry, K. D.: Global Atmospheric Transport and Source-
791 Receptor Relationships for Arsenic, Environ. Sci. Technol., 50, 3714-3720,
792 doi:10.1021/acs.est.5b05549, 2016.

793 Waldner, P., Marchetto, A., Thimonier, A., Schmitt, M., Rogora, M., Granke, O., Mues, V., Hansen, K.,
794 Pihl Karlsson, G., Zlindra, D. et al..: Detection of temporal trends in atmospheric deposition of
795 inorganic nitrogen and sulphate to forests in Europe, Atmos. Environ., 95, 363-374,
796 https://doi.org/10.1016/j.atmosenv.2014.06.054, 2014.