



1       **The control of anthropogenic emissions contributed to 80% of the decrease in**  
2                               **PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017**

3       Ziyue Chen<sup>1,2</sup>, Danlu Chen<sup>1</sup>, Meipo Kwan<sup>3</sup>, Bin Chen<sup>4</sup>, Nianliang Cheng<sup>5</sup>, Bingbo Gao<sup>6\*</sup>,  
4       Yan Zhuang<sup>1</sup>, Ruiyuan Li<sup>1</sup>, Bing Xu<sup>7\*</sup>

5       <sup>1</sup>College of Global and Earth System Science, Beijing Normal University, 19 Xijiekou  
6       Street, Haidian, Beijing 100875, China.

7       <sup>2</sup>Joint Center for Global Change Studies, Beijing 100875, China.

8       <sup>3</sup>Department of Geography and Geographic Information Science, University of Illinois at  
9       Urbana-Champaign, Urbana, IL 61801, USA.

10      <sup>4</sup>Department of Land, Air and Water Resources, University of California, Davis, CA 95616, USA

11      <sup>5</sup>Chinese Research Academy of Environmental Sciences, Beijing 100012, China.

12      <sup>6</sup>College of Land Science and Technology, China Agriculture University, Tsinghua East Road,  
13      Haidian District, 100083, China.

14      <sup>7</sup>Ministry of Education Key Laboratory for Earth System Modeling, Department of Earth System  
15      Science, Tsinghua University, Beijing 100084, China

16      \*To whom correspondence should be addressed. Email: gaobb@reis.ac.cn or  
17      bingxu@tsinghua.edu.cn

18      **Abstract**

19      With the completion of the Beijing Five-year Clean Air Action Plan by the end of 2017, the annual  
20      mean PM<sub>2.5</sub> concentrations in Beijing dropped dramatically to 58.0 µg/m<sup>3</sup> in 2017 from 89.5 µg/m<sup>3</sup> in  
21      2013. However, controversies exist to argue that favorable meteorological conditions in 2017 that  
22      helped pollution dispersion were the major factor for such rapid decrease in PM<sub>2.5</sub> concentrations. To  
23      comprehensively evaluate this five-year plan, we employed Kolmogorov-Zurbenko (KZ) filtering and a  
24      WRF-CMAQ model to quantify the relative contribution of meteorological conditions and the control  
25      of anthropogenic emissions to PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017. For these five years, the  
26      relative contribution of emission-reduction measures to the decrease of PM<sub>2.5</sub> concentrations in Beijing  
27      calculated by KZ filtering and the WRF-CMAQ model was 80.6% and 78.6% respectively. The  
28      WRF-CMAQ model further revealed that local and regional emission-reduction measures contributed  
29      to 53.7% and 24.9% of the PM<sub>2.5</sub> reduction respectively. For local emission-reduction measures, the



30 regulation of coal boilers, increasing clean fuels for residential use, industrial restructuring, the  
31 regulation of raise dust and vehicle emissions contributed to 20.1 %, 17.4%, 10.8%, 3.0 % and 2.4% of  
32  $PM_{2.5}$  reduction respectively. Both models suggested that the control of anthropogenic emissions  
33 contributed to around 80% of the total decrease in  $PM_{2.5}$  concentrations in Beijing, indicating that  
34 emission control was crucial for the notable improvement in air quality in Beijing from 2013 to 2017.  
35 Therefore, such long-term air quality clean plan should be continued for the future years to further  
36 reduce  $PM_{2.5}$  concentrations in Beijing. Considering that different emission-reduction measures exert  
37 distinct effects on  $PM_{2.5}$  reduction and existing emission-reduction measures work poorly to reduce  
38 ozone concentrations, future strategies for emission-reduction should be designed and implemented  
39 accordingly.

40 **Keywords:  $PM_{2.5}$  reduction, anthropogenic emissions, meteorological conditions,**  
41 **Kolmogorov-Zurbenko (KZ) filtering, WRF-CMAQ**



## 42 1 Introduction

43 In December 2012, a heavy haze episode occurred in Beijing, during which the highest hourly  $PM_{2.5}$   
44 concentrations once reached  $886 \mu\text{g}/\text{m}^3$ , a historical record. The extremely high  $PM_{2.5}$  concentrations  
45 led to long-lasting black and thick fogs, which not only significantly influenced people's daily life  
46 (low-visibility induced traffic jam), but also exerted strong negative influences on public health  
47 (Brunekreef et al., 2002; Dominici et al., 2014; Nel et al., 2005; Zhang et al., 2012; Qiao et al., 2014).  
48 Since then, severe haze episodes have frequently occurred in Beijing and other regions in China (Chan  
49 et al., 2008; Huang, R., et al., 2014; Guo et al., 2014; Zheng et al., 2015), and  $PM_{2.5}$  pollution has  
50 become one of the most concerned environmental issues in China. Since 2013, a national network of  
51 ground stations for monitoring hourly  $PM_{2.5}$  concentrations has been established gradually, including  
52 35 ground observation stations in Beijing, which provide important support for proper management and  
53 in-depth investigation of  $PM_{2.5}$  concentrations. Meanwhile, for effectively reducing local  $PM_{2.5}$   
54 concentrations, the local government proposed the Beijing Five-year Clean Air Action Plan  
55 (2013-2017). This plan suggested the specific aim that the annual mean  $PM_{2.5}$  concentrations in Beijing  
56 should be reduced from  $89.5 \mu\text{g}/\text{m}^3$  in 2013 to  $60 \mu\text{g}/\text{m}^3$  in 2017 and included a series of  
57 emission-reduction measures, including shutting down heavily polluting factories, restricting traffic  
58 emissions and replacing coal fuels with clean energies. Furthermore, for reducing high  $PM_{2.5}$   
59 concentrations during severe haze episodes, Beijing Municipal Government published the "Heavy Air  
60 Pollution Contingency Plan" in 2012, and further revised this plan in March 2015. According to this  
61 plan, a series of contingent emission reduction measures should be implemented according to the  
62 severeness of  $PM_{2.5}$  pollution episodes. By the end of 2017, these long-term and contingent  
63 emission-reduction measures had worked together to reduce the annual mean  $PM_{2.5}$  in Beijing to  $58.0$   
64  $\mu\text{g}/\text{m}^3$ , indicating a great success of  $PM_{2.5}$  management during the past five years.

65 In addition to anthropogenic emissions, the strong meteorological influences on  $PM_{2.5}$  concentrations in  
66 Beijing have been widely acknowledged (Cheng et al., 2017; Chen, Z. et al., 2016, 2017, 2018; UNEP,  
67 2016; Wang et al., 2014; Zhao et al., 2013). For instance, Chen, Z et al. (2016) found that for 2014,  
68 more than 180 days in Beijing experienced a dramatic AQI (Air Quality Index) change ( $\Delta\text{AQI}>50$ ),  
69 compared with the previous day. Considering the total emission of airborne pollutants for a mega city  
70 hardly change significantly on a daily basis, the rapid variation of meteorological conditions in Beijing  
71 was one important driver for the dramatic change of daily air quality in Beijing. In this case, there



72 arises the controversy that meteorology, instead of emission-reduction measures, made a major  
73 contribution to the remarkable reduction of  $PM_{2.5}$  concentrations in Beijing from 2013 to 2017. With  
74 the completion of the five-year plan, it is highly necessary to quantify the relative contribution of  
75 meteorological conditions and emission-reduction measures to the remarkable decrease of  $PM_{2.5}$   
76 concentrations in Beijing.

77 To this end, we employ different approaches in this paper to comprehensively estimate adjusted  $PM_{2.5}$   
78 concentrations in Beijing while eliminating the influence from the variation in meteorological  
79 conditions and thus quantify the relative contribution of emission-reduction measures to the decrease of  
80  $PM_{2.5}$  concentrations. In this light, this research provides important insight for better designing and  
81 implementing successive clean air plans in the future to further mitigate  $PM_{2.5}$  pollution in Beijing.

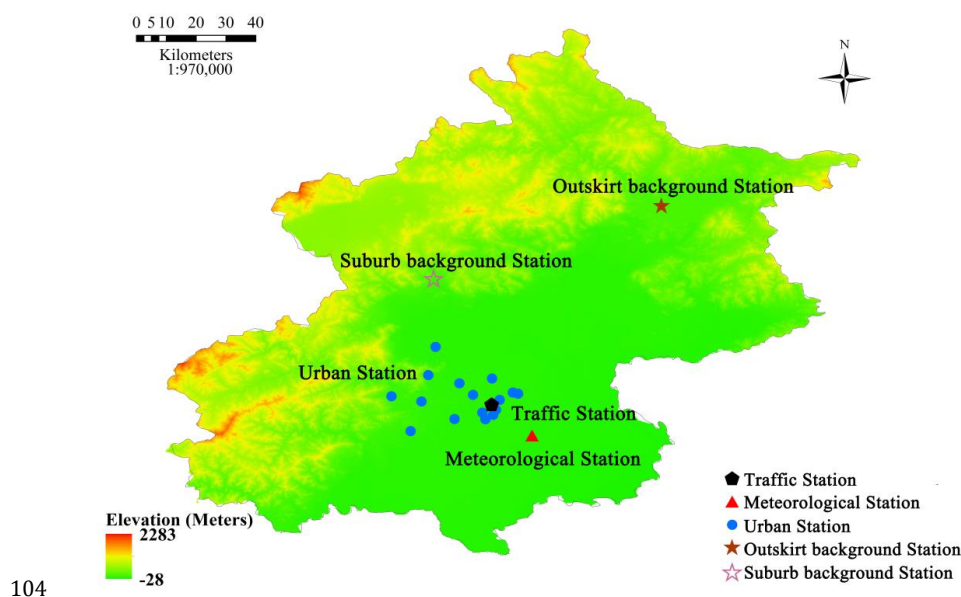
## 82 **2 Data Sources**

### 83 **2.1 $PM_{2.5}$ and meteorological data**

84 In this study, hourly  $PM_{2.5}$  concentration data were acquired from the website PM25.in, which collects  
85 official data provided by China National Environmental Monitoring Center (CNEMC). Beijing has  
86 established an advanced air quality monitoring network with 35 ground stations across the city.  
87 Considering the major contribution of industry and traffic-induced emissions in urban areas, we  
88 selected all twelve urban stations to analyze the variation of  $PM_{2.5}$  concentrations and quantify their  
89 influencing factors. In addition to these urban stations, we also selected two background stations, the  
90 DingLing Station located in the suburb and the MiYun Reservoir Station located in the outer suburb,  
91 one transportation station (the Qianmen station) located close to a main road, and one rural station (the  
92 Yufa Station) which is far away from central Beijing for the following analysis. The DingLing and  
93 MiYun Reservoir Stations were chosen as background stations by the Ministry of Environmental  
94 Protection of China. These two stations receive limited influence from anthropogenic emissions due to  
95 their location in suburban and outer suburban areas. Comparing the variation in  $PM_{2.5}$  concentrations  
96 and its anthropogenic and meteorological driving factors in different type of stations provides a useful  
97 reference for comprehensively understanding the effects of emission-reduction measures on the  
98 reduction of  $PM_{2.5}$  concentrations in Beijing in the past five years. The locations of these selected  
99 stations are shown in Fig 1. Meteorological data for this research were collected from the Guanxiangtai  
100 Station (GXT,54511, 116.46° E, 39.80° N), Beijing and were downloaded from the Department of



101 Atmospheric Science, College of Engineering, University of Wyoming  
102 (<http://weather.uwyo.edu/upperair/sounding.html>). Both the PM<sub>2.5</sub> and meteorological data were  
103 collected from January 1<sup>st</sup>, 2013 to December 31<sup>st</sup>, 2017.



104

105

Fig 1. Locations of different ground monitoring stations.

## 106 2.2 Emission inventories

107 For this research, we employed both regional and local emission inventories for running model  
108 simulation. Multi-resolution Emission Inventory for China, MEIC, (<http://meicmodel.org/>) provided by  
109 Tsinghua University, were employed as the regional emission inventories. MEIC has been widely  
110 employed and verified as a reliable emission inventory by a diversity of studies (Hong et al., 2017;  
111 Saikawa et al., 2017; Zhou et al., 2017; etc.). Different from regional emission inventories, local  
112 emission inventories are usually produced independently by local institutes. The Beijing local-emission  
113 inventories employed for this research is produced and updated by Beijing Municipal Research  
114 Institute of Environmental protection fully according to the requirement of MEP on the production of  
115 local emission inventories within the Beijing-Tianjin-Hebei region. This local-emission inventory is  
116 produced by synthesizing local environmental statistical data and reported emission data, carrying out  
117 field investigations and conducting a series estimation according to Beijing Five-year Clean Air Action



118 Plan. This Beijing local-emission inventory has been formally employed for the implementation of  
119 recent “2017 Air Pollution Prevention and Management Plan for the Beijing-Tianjin-Hebei Region and  
120 its Surrounding Areas” (MEP, 2017).

### 121 **3 Methods**

122 A key step for quantifying the relative contribution of anthropogenic emissions to the decrease of PM<sub>2.5</sub>  
123 concentrations is to properly filter meteorological influences on PM<sub>2.5</sub> concentrations, which is highly  
124 challenging and rarely investigated by previous studies. Therefore, we employed both a statistical  
125 method and a chemical transport model in this study to comprehensively evaluate the role of  
126 anthropogenic emissions and meteorological conditions in the decrease of PM<sub>2.5</sub> concentrations in  
127 Beijing during the past five years.

#### 128 **3.1 Kolmogorov-Zurbenko (KZ) filtering**

129 Since meteorological conditions exert a strong influence on PM<sub>2.5</sub> concentrations in Beijing, the  
130 removal of seasonal signals from time series of meteorological factors results in data sets suitable for  
131 understanding the trend of PM<sub>2.5</sub> concentrations mainly influenced by anthropic factors (Eskridge et al.,  
132 1997). To better analyze the trend of time series data without the disturbances from large variations of  
133 influencing variables, a statistical method called Kolmogorov-Zurbenko (KZ) filtering was proposed  
134 by Rao et al. (1994). The KZ filter is advantageous in removing high-frequency variations in the data  
135 set based on the iterative moving average. Eskridge et al. (1997) compared four major approaches for  
136 trend detection, including PEST, anomalies, wavelet transform, and the KZ filter, and suggested that  
137 the confidence in detecting long-term trend of the KZ filter was much higher than that of the other  
138 methods. Due to its reliable performance in trend detection in complicated ecosystems, the KZ filter  
139 has frequently been employed to remove seasonal signals of meteorological conditions and extract  
140 long-term trend of airborne pollutants (Zurbenko, et al., 1996; Eskridge, et al., 1997; Kang, et al.,  
141 2013). One potential limitation of the KZ filter is that iterative moving average ( $m$ ) may impose an  
142 influence on detecting abrupt changes of variations. Therefore, Zurbenko et al. (1996) proposed an  
143 enhanced KZ filter that employed a dynamic variable  $m$  that decreases with the increase in changing  
144 rate, which is employed in this study to estimate the modified PM<sub>2.5</sub> concentrations in Beijing by  
145 removing large seasonal variations in meteorological conditions. The principle of the KZ filter is  
146 briefly introduced as follows.



147 The raw time-series data of airborne pollutants can be decomposed as:

$$148 \quad X(t) = E(t) + S(t) + W(t) \quad (1)$$

$$149 \quad X_b(t) = E(t) + S(t) \quad (2)$$

$$150 \quad E(t) = KZ_{365,3}(X) \quad (3)$$

$$151 \quad S(t) = KZ_{15,5}(X) - KZ_{365,3}(X) \quad (4)$$

$$152 \quad W(t) = X(t) - KZ_{15,5}(X) \quad (5)$$

153 Where  $X(t)$  is the original time series of airborne pollutants,  $E(t)$  is the long-term trend component,  $S(t)$   
154 is the seasonal variation,  $W(t)$  is the residue or synoptic-scale (short-term) variations.  $KZ_{i,j}(X)$   
155 indicates a KZ filtering on the original dataset  $X$  with a moving window size of  $i$  and  $j$  iterations.

156  $X_b(t)$  stands for the base component, the sum of the long-term trend component and seasonal variation,  
157 presenting steady trend variation.  $E(t)$  is mainly effected by long-term anthropogenic emission and  
158 climate change.  $S(t)$  is mainly influenced by the seasonal variation of emission factors and  
159 meteorological conditions. The residue  $W(t)$  is caused by short-term and small-scale shifts of emissions  
160 and meteorological conditions.

161 The long-term trend component  $E(t)$  processed by KZ filtering still contains the influence of  
162 meteorological conditions, which can be removed by multiple regression models. Multiple linear  
163 relationships are established for the residue and baseline component respectively using strongly  
164 correlated meteorological factors.

165 We conducted correlation analysis between  $PM_{2.5}$  concentrations and a series of meteorological  
166 factors, including temperature, wind speed, wind direction, precipitation, relative humidity, solar  
167 radiation, evaporation and air pressure. The correlation analysis revealed that wind speed, relative  
168 humidity, temperature, solar radiation and air pressure were strongly and significantly correlated with  
169  $PM_{2.5}$  concentrations in Beijing, which was consistent with the findings from previous studies (Sun et  
170 al., 2013; Chen, Z., et al., 2017, 2018; Wang et al., 2018). Therefore, we further established multiple  
171 linear regression equations between  $PM_{2.5}$  concentrations and wind speed, relative humidity,  
172 temperature and solar radiation as follows.

$$173 \quad W(t) = a_0 + \sum a_i w_i(t) + \varepsilon_w(t) \quad (6)$$

$$174 \quad X_b(t) = b_0 + \sum b_i x_i(t) + \varepsilon_b(t) \quad (7)$$

$$175 \quad \varepsilon(t) = \varepsilon_w(t) + \varepsilon_b(t) \quad (8)$$



176 Where  $w_i(t)$  and  $x_i(t)$  stand for the different synoptic-scale variations and baseline component of the  
177  $i^{\text{th}}$  meteorological factor.  $\varepsilon_w$  and  $\varepsilon_b$  is the regression residue of the synoptic-scale variations and  
178 baseline component.  $\varepsilon(t)$  indicates the total residue, including the short-term influence of local  
179 emission sources, meteorological influences neglected during the regression and noise.

180 Next, KZ filtering is conducted on the  $\varepsilon(t)$  for its long-term component  $\varepsilon_E(t)$ . After the  
181 variation of meteorological influences was filtered, the reconstructed time series of airborne pollutants  
182  $X_{LT}(t)$  was calculated as the sum of  $\varepsilon_E(t)$  and the average value of  $E(t)$ ,  $\overline{E(t)}$ .

$$183 \quad X_{LT}(t) = \overline{E(t)} + \varepsilon_E(t) \quad (9)$$

184 After KZ filtering, the relative contribution of meteorological conditions to the variation in  $PM_{2.5}$   
185 concentrations can be calculated as follows:

$$186 \quad P_{contrib} = \frac{K_{org} - K}{K_{org}} \times 100\% \quad (10)$$

187 Where  $P_{contrib}$  is the relative contribution of meteorological conditions to the variation of  $PM_{2.5}$   
188 concentrations in Beijing,  $K_{org}$  is the variation slope of the original  $PM_{2.5}$  time series;  $K$  is the  
189 variation slope of adjusted  $PM_{2.5}$  time series after meteorological variations are removed.

### 190 3.2 WRF-CMAQ model

191 We employed the WRF-CMAQ model for simulating the effects of emission-reduction measures on the  
192 reduction of  $PM_{2.5}$  concentrations. The WRF-CMAQ model includes three models: The middle-scale  
193 meteorology model (WRF), the source emission model (SMOKE) (<http://www.cmascenter.org/smoke/>)  
194 and the community multiscale air quality modeling system (CMAQ)  
195 (<http://www.cmascenter.org/CMAQ>). The center of the CMAQ was set at coordinate 35°N, 110°E and  
196 a bi-directional nested technology was employed, producing two layers of grids with a horizontal  
197 resolution of 36 km and 12 km respectively. The first layer of grids with 36 km resolution and 200×160  
198 cells covered most areas in East Asia (including China, Japan, North Korea, South Korea, and other  
199 countries). The second layer of grids with 12 km resolution and 120×102 cells covered the North China  
200 Plain (including the Beijing-Tianjin-Hebei region, and Shandong and Henan Provinces). The vertical  
201 layer was divided into 20 unequal layers, eight of which were of a distance of less than 1 km to the  
202 ground for better featuring the structure of atmospheric boundary. The height of the ground layer was  
203 35 m.





204 We employed ARW-WRF3.2 to simulate the meteorological field. The setting of the center and the  
205 bi-directional nest for the WRF was similar to that of the CMAQ as mentioned above. There were 35  
206 vertical layers for the WRF and the outer layer provided boundary conditions of the inner layer. The  
207 meteorological background field and boundary information with a FNL resolution of  $1^\circ \times 1^\circ$  and  
208 temporal resolution of 6h were acquired from NCAR (National Center for Atmospheric Research,  
209 <https://ncar.ucar.edu/>) and NCEP (National Centers for Environmental Prediction) respectively. The  
210 terrain and underlying surface information was obtained from the USGS 30s global DEM  
211 (<https://earthquake.usgs.gov/>). The output from the WRF model was interpolated to the region and grid  
212 of the CMAQ model using the Meteorology-Chemistry Interface Processor (MCIP,  
213 <https://www.cmascenter.org/mcip>). The meteorological factors used for this model includes  
214 temperature, air pressure, humidity, geopotential height, zonal wind, meridional wind, precipitation,  
215 boundary layer heights and so forth. An estimation model for terrestrial ecosystem MEGAN  
216 (<http://ab.inf.uni-tuebingen.de/software/megan/>) was employed to process the natural emissions.  
217 Anthropogenic emission data were from the Multi-resolution Emission Inventory for China, MEIC  
218  $0.5^\circ \times 0.5^\circ$  emission inventory (<http://www.meicmodel.org/>) and Beijing emission inventory  
219 (<http://www.cee.cn/>). We input the processed natural and anthropogenic emission data into the SMOKE  
220 model and acquired comprehensive emission source files.

221 Scenario simulation is employed to estimate the contribution of emission-reduction to the variation in  
222  $PM_{2.5}$  concentrations.

$$223 \quad P_{contrib} = \frac{C - C_{base}}{C} \times 100\% \quad (11)$$

224 Where  $P_{contrib}$ ,  $C$  and  $C_{base}$  are the contribution rate of emission reduction to  $PM_{2.5}$  concentrations,  
225 the simulated  $PM_{2.5}$  concentrations under the emission reduction scenario and simulated  $PM_{2.5}$   
226 concentrations in the baseline scenario respectively.

227 To evaluate the relative contribution of meteorological conditions and different emission-reduction  
228 measures to the decrease of  $PM_{2.5}$  concentrations, we designed two baseline experiments and six  
229 sensitivity experiments. For the first baseline experiment, we employed the actual meteorological data  
230 in 2013. For the second baseline experiment, we employed the actual meteorological data in 2017 and  
231 emission inventory in 2017. Since no emission-reduction measures were conducted in 2013, the first  
232 baseline experiment was used for model verification and estimating the relative contribution of  
233 meteorological variations to the variation of  $PM_{2.5}$  concentrations. By comparing the first and second



234 baseline experiment, the relative contribution of all emission-reduction measures to the variation of  
235  $PM_{2.5}$  concentrations can be quantified. For the first sensitivity experiment, we employed the actual  
236 meteorological conditions in 2013 and emission inventory in 2017 and compared the simulation result  
237 with the baseline experiment, which demonstrated the relative contribution of meteorological variations  
238 to a  $PM_{2.5}$  reduction in Beijing during the past five years. Since the WRF-CMAQ simulation simply  
239 considered the  $PM_{2.5}$  concentrations and meteorological conditions in 2013 and 2017 without  
240 considering their variation process from 2013 to 2017, KZ filtering may perform better in quantifying  
241 the relative contribution of meteorological variations to a  $PM_{2.5}$  reduction in Beijing. However, the  
242 output from this sensitivity experiment serves as a useful reference for understanding the reliability of  
243 the output from the KZ filtering. For the remaining five sensitivity-simulation experiments, we added  
244 the reduced emission amount induced by one specific emission-reduction measure to the actual  
245 emission amount in 2017 and kept other parameters unchanged, which quantified the relative  
246 contribution of one type of emission sources to the  $PM_{2.5}$  reduction in Beijing during the past five years.  
247 Therefore, we acquired the influence of the relative contribution of each emission source on  $PM_{2.5}$   
248 reduction in Beijing (Table 1).



Table 1. The design and material for the baseline and five sensitivity experiments using WRF-CMAQ model

ID	Meteorological Data	Emission-reduction measures	Simulation Year	Major purposes
Baseline Experiment1	2013	No emission-reduction Measures	2013	2013 baseline scenario
Baseline Experiment2	2017	All emission-reduction Measures	2017	2017 baseline scenario
Sensitivity Experiment 1	2013	All emission-reduction Measures	2017	The relative contribution of meteorological variations to the decrease of PM <sub>2.5</sub> concentrations from 2013 to 2017
Sensitivity Experiment 2	2017	All emission-reduction measures except for industrial restructuring	2017	The relative contribution of industrial restructuring to the decrease of PM <sub>2.5</sub> concentrations from 2013 to 2017
Sensitivity Experiment 3	2017	All emission-reduction measures except for the regulation of coal boilers	2017	The relative contribution of the regulation of coal boilers to the decrease of PM <sub>2.5</sub> concentrations from 2013 to 2017
Sensitivity Experiment 4	2017	All emission-reduction measures except for increasing clean fuels for civil use	2017	The relative contribution of increasing clean fuels for civil use to the decrease of PM <sub>2.5</sub> concentrations from 2013 to 2017
Sensitivity Experiment 5	2017	All emission-reduction measures except for the regulation of vehicle emissions	2017	The relative contribution of the regulation of vehicle emissions to the decrease of PM <sub>2.5</sub> concentrations from 2013 to 2017

For emission data, all experiments employed Beijing local emissions inventory in 2017 for Beijing and regional emission inventory in 2017 for other regions.



### 251 **3.3 Model verification**

#### 252 **3.3.1 Verification of the KZ filtering**

253 For each station, the original time series of  $PM_{2.5}$  data was processed by the KZ filter and the relative  
254 contribution of the long-term trend, seasonal variation and short-term variation to the total variance  
255 was shown as Table 2. The sum of the long-term trend, seasonal variation and short-term variation  
256 contributed to more than 93.6~95.3% of the total variance for different stations respectively. The larger  
257 the total variance, the three components are more independent to each other. According to Table 2, the  
258 large value of the total variation for each station indicated a satisfactory result from the KZ filtering.  
259 The relative contribution of short-term variation was much larger than that of the seasonal and  
260 long-term variation, suggesting that short-term variations of meteorological conditions and emission  
261 conditions exerted a strong influence on the rapid variation in  $PM_{2.5}$  concentrations in Beijing. This  
262 result is consistent with findings from previous studies (Chen et al., 2016; Ma et al., 2016).

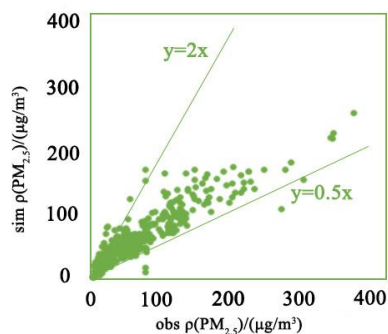


263 **Table 2. The relative contribution of different components to the total variance of original time**  
264 **series of PM<sub>2.5</sub> concentrations from 2013-2017 at different stations**

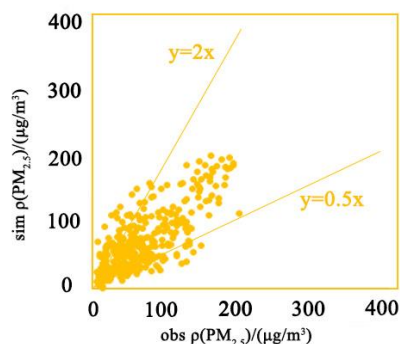
Stations	Long-term Trend(%)	Seasonal Variation(%)	Short-term Variation(%)	Total variance(%)
Yufa	2.1	23.8	66.8	94.0
Miyun Reservoir	1.4	9.0	83.8	95.2
Dingling	1.6	11.0	81.3	94.9
Qianmen	2.7	12.7	78.5	95.1
Olympic center	2.1	11.9	80.0	95.3
Xiangshan	1.2	10.3	83.4	94.9
Huayuan	2.2	15.9	75.6	93.7
Yungang	2.1	15.1	76.5	93.6
WanShouxigong	1.6	14.2	78.2	94.0
Dongsi	1.6	12.3	80.0	94.0
TianTan	2.1	13.2	78.6	93.8
NongZhanguan	1.8	13.7	78.6	94.1
Gucheng	1.8	13.5	78.5	93.7
Guanyuan	1.6	12.6	79.8	94.0
BeiBuxinqu	1.7	13.8	78.4	93.9
WanLiu	3.5	11.9	78.2	93.6

### 265 3.3.2 Verification of the WRF-CMAQ

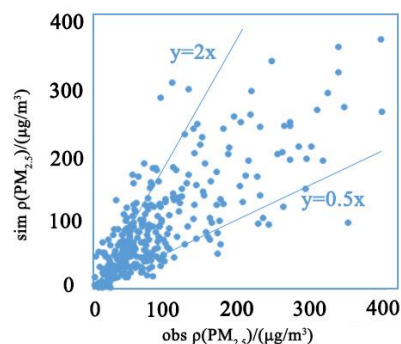
266 We employed the emission inventory and meteorological data for 2013 to verify the accuracy of the  
267 WRF-CMAQ model. For three different stations (the DingLing background station, the Yufa rural  
268 station and the Olympic Center urban station), we compared the observed and estimated PM<sub>2.5</sub>  
269 concentrations (Fig 2). According to Fig 2, the general trend of the simulated PM<sub>2.5</sub> concentrations was  
270 similar to that of the observed value. A general agreement was found between the simulated and  
271 observed data with more than 85% of data points falling into the siege area of 1:2 and 2:1 lines.  
272 WRF-CMAQ slightly underestimated PM<sub>2.5</sub> concentrations due to the uncertainty in the emission  
273 inventory, meteorological field simulation errors and insufficient chemical reaction mechanisms. For  
274 three stations, the correlation coefficient R, normalized mean bias (NMB), normalized mean error  
275 (NME), mean fractional bias (MFB) and mean fractional error (MFE) between observed and simulated  
276 data was 0.69~0.74, 11%~17%, 20%~27%, -21%~-17%, and 15%~27% respectively, indicating a  
277 satisfactory simulation output (EPA, 2005; Boylan et al., 2006)



(a) Dingling background station



(b) Olympic center urban station



(c) Yufa rural station

278 Fig 2. The comparison between observed and WRF-CMAQ simulated PM<sub>2.5</sub> concentrations

## 279 4 Results

### 280 4.1 The relative contribution of emission-reduction measures and meteorological 281 variations to the decrease of PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017

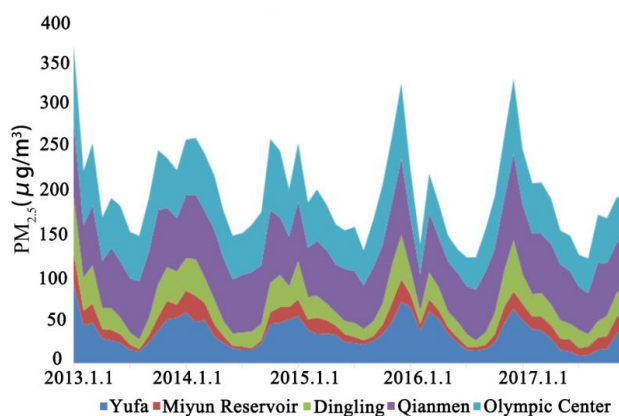
#### 282 4.1.1 Estimation based on KZ filtering

283 Through KZ filtering, the original time-series of PM<sub>2.5</sub> concentrations and adjusted time-series of PM<sub>2.5</sub>  
284 concentrations with filtered meteorological variations were acquired. Based on these, for each station,  
285 the actual variation of PM<sub>2.5</sub> concentrations and the adjusted variation in PM<sub>2.5</sub> concentrations without  
286 the influence of meteorological variations from 2013 to 2017 were calculated (as shown in Table 3),  
287 which indicate the relative contribution of anthropogenic emissions and meteorological conditions to

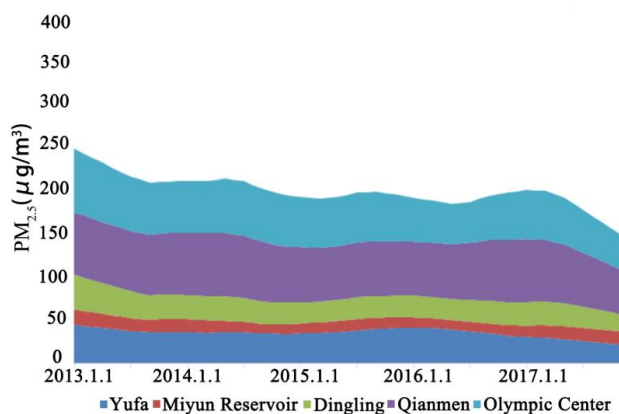


288 the decrease in  $PM_{2.5}$  concentrations in Beijing during the five-year period.

289 The original time series of  $PM_{2.5}$  concentrations and adjusted time series of  $PM_{2.5}$  concentrations  
290 processed using KZ filtering were illustrated using one urban station, one rural station, one  
291 transportation station, and two background stations (Fig 3). As shown in Fig 3, the most abrupt  
292 variations in  $PM_{2.5}$  concentrations have been smoothed through KZ filtering.



a. Original time series of  $PM_{2.5}$  concentrations from 2013 to 2017



b. Processed time series of  $PM_{2.5}$  concentrations from 2013 to 2017 using KZ filter

293 **Fig 3. The comparison of original and KZ processed time series of  $PM_{2.5}$  concentrations in**  
294 **Beijing from 2013 to 2017**

295 According to Table 3, the annual mean  $PM_{2.5}$  concentrations in Beijing in 2017 was 35.6% lower than  
296 that in 2013. By filtering the influence of meteorological variations, the adjusted annual mean  $PM_{2.5}$   
297 concentrations in Beijing in 2017 decreased by 31.7% when compared to that in 2013, indicating that



298 the variation in meteorological conditions exerted a moderate influence on the reduction of  $PM_{2.5}$   
299 concentrations during the past five years. Meteorological conditions in Beijing were generally  
300 favorable for  $PM_{2.5}$  dispersion during the five years, especially the latter half of 2017, when there was a  
301 high frequency of strong Northerly winds and much lower wintertime  $PM_{2.5}$  concentrations than  
302 previous years.

303 For the winter of 2017, frequent windy weather and successive clean sky had a strong influence on the  
304 reduction of  $PM_{2.5}$  concentrations in Beijing. This led to a hot debate concerning whether the notable  
305 decrease in  $PM_{2.5}$  concentrations was largely due to the favorable meteorological conditions or  
306 emission-reduction measures. Table 3 suggests that emission-reduction measures contributed to  
307 75.2%~85.0%  $PM_{2.5}$  decrease in the five-year period, indicating that emission-reduction measures  
308 worked effectively in all rural, urban and background stations. On average, the relative contribution of  
309 anthropogenic emissions and meteorological variations to  $PM_{2.5}$  reduction in Beijing from 2013 to  
310 2017 was 80.6% and 19.4% respectively. Therefore, in spite of more favorable meteorological  
311 conditions, properly designed and implemented emission-reduction measures were the dominant driver  
312 for the remarkable decrease of  $PM_{2.5}$  concentrations in Beijing during the past five years.





313 **Table 3. Estimated relative contribution of emission-reduction and meteorological variations to PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017 using KZ filter**

Stations	PM <sub>2.5</sub> concentrations in 2013 (µg·m <sup>-3</sup> )	PM <sub>2.5</sub> concentrations in 2017 (µg·m <sup>-3</sup> )	Adjusted PM <sub>2.5</sub> concentrations in 2017 (µg·m <sup>-3</sup> )	PM <sub>2.5</sub> Decrease rate (µg·m <sup>-3</sup> ·m <sup>-1</sup> ) <sup>1</sup>	Adjusted PM <sub>2.5</sub> Decrease rate (µg·m <sup>-3</sup> ·m <sup>-1</sup> ) <sup>2</sup>	Contribution of emission reduction (%) <sup>3</sup>	Contribution of meteorological variations (%) <sup>4</sup>
Yufa	111.1	69.7	74.6	-0.78	-0.63	80.4	19.7
Miyun Reservoir	58.8	44.8	47.0	-0.40	-0.33	82.8	17.2
Dingling	69.6	47.1	50.6	-0.54	-0.44	80.8	19.2
Qianmen	103.9	64.0	68.9	-0.81	-0.69	85.0	15.0
Olympic center	90.4	57.2	61.7	-0.68	-0.55	80.8	19.2
Xiangshan	77.0	59.3	60.3	-0.46	-0.39	83.9	16.1
Huayuan	101.5	64.4	69.2	-0.77	-0.63	81.9	18.1
Yungang	91.8	60.2	64.0	-0.69	-0.55	79.6	20.4
WanShouxigong	93.7	62.0	66.8	-0.64	-0.50	78.2	21.8
Dongsi	94.9	62.4	67.5	-0.62	-0.49	78.9	21.1
TianTan	92.3	58.4	64.6	-0.68	-0.55	80.2	19.9
NongZhanguan	92.2	59.9	65.9	-0.66	-0.53	80.3	19.8
Gucheng	92.7	61.4	65.9	-0.65	-0.50	77.6	22.4
Guanyuan	89.6	59.5	64.6	-0.60	-0.48	79.6	20.4
BeiBuxinqu	86.6	59.5	63.3	-0.60	-0.45	75.2	24.8
WanLiu	98.1	56.2	60.4	-0.87	-0.73	84.2	15.8

314 <sup>1</sup> PM<sub>2.5</sub> decrease rate: the fitted variation slope of original monthly average PM<sub>2.5</sub> time series;

315 <sup>2</sup> Adjusted PM<sub>2.5</sub> decrease rate: the fitted variation slope of adjusted monthly average PM<sub>2.5</sub> time series;

316 <sup>3</sup> Contribution of emission reduction = 1 - Contribution of meteorological variations;

317 <sup>4</sup> Contribution of meteorological variations = (PM<sub>2.5</sub> decrease rate - Adjusted PM<sub>2.5</sub> decrease rate) / PM<sub>2.5</sub> decrease rate.



#### 318 4.1.2 Estimation based on WRF-CMAQ model

319 In addition to the KZ filter, we also employed the WRF-CMAQ model to estimate the relative  
320 contribution of emission-reduction measures and meteorological conditions to the decrease of PM<sub>2.5</sub>  
321 concentrations in Beijing. The result is shown in Table 4.

322 **Table 4. Estimated relative contribution of emission-reduction and meteorological variations to**  
323 **PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017 using WRF-CMAQ model**

Stations	Contribution of meteorological variations (%)	Contribution of emission-reduction(%)
Yufa	21.9	78.2
Miyun Reservoir	20.8	79.2
Dingling	21.7	78.3
Qianmen	21.2	78.8
Olympic center	21.2	78.8
Xiangshan	20.3	79.7
Huayuan	21.2	78.8
Yungang	21.2	78.8
WanShouxigong	21.2	78.8
Dongsi	21.2	78.8
TianTan	21.2	78.8
NongZhanguan	21.2	78.8
Gucheng	22.2	77.8
Guanyuan	21.2	78.8
BeiBuxinqu	22.2	77.8
WanLiu	22.2	77.8

324 As Table 4 shows, and based on the WRF-CMAQ model, the relative contribution of meteorological  
325 variations to the decrease in PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017 ranged from 20.3% to  
326 22.2% in different stations, while emission-reduction measures contributed to about four-fifths of the  
327 decrease in PM<sub>2.5</sub> concentrations from 2013 to 2017. It is worth mentioning that the WRF-CMAQ  
328 model was a grid-based model and thus the calculated contribution of meteorological variations for  
329 some stations located in the same grid was the same. Instead, station-based KZ filtering led to more  
330 reliable analysis for each station and can better distinguish the differences between different stations.



331 Furthermore, the WRF-CMAQ model simply considered the differences between the meteorological  
332 conditions in 2013 and 2017 without considering their variations during the past five years while the  
333 KZ filtering analyzed the entire time series of  $PM_{2.5}$  and meteorological data from 2013 to 2017. The  
334 averaged relative contribution of meteorological variations to  $PM_{2.5}$  reduction in Beijing calculated  
335 using the WRF-CMAQ model was 21.4%, very similar to the 19.4% obtained by using KZ filtering.  
336 The slightly larger meteorological contribution calculated using the WRF-CMAQ model might be  
337 attributed to the favorable meteorological conditions in the winter of 2017.

338 Due to its fine spatial resolution and capability in providing a better understanding of the influence of  
339 meteorological conditions on  $PM_{2.5}$  concentrations, KZ filtering provides a more reliable method for  
340 researchers and decision makers to understand the relative importance of emission-reduction measures  
341 and meteorological conditions in recent  $PM_{2.5}$  reduction in Beijing. However, similar results from the  
342 WRF-CMAQ simulation provide complementary evidence for the fact that anthropogenic emissions  
343 exerted a much stronger influence on  $PM_{2.5}$  concentrations than meteorological conditions. In the next  
344 subsection, and based on a detailed local emission inventory, we use the WRF-CMAQ model to further  
345 quantify the relative contribution of different emission-reduction measures to the decrease in  $PM_{2.5}$   
346 concentrations in Beijing.

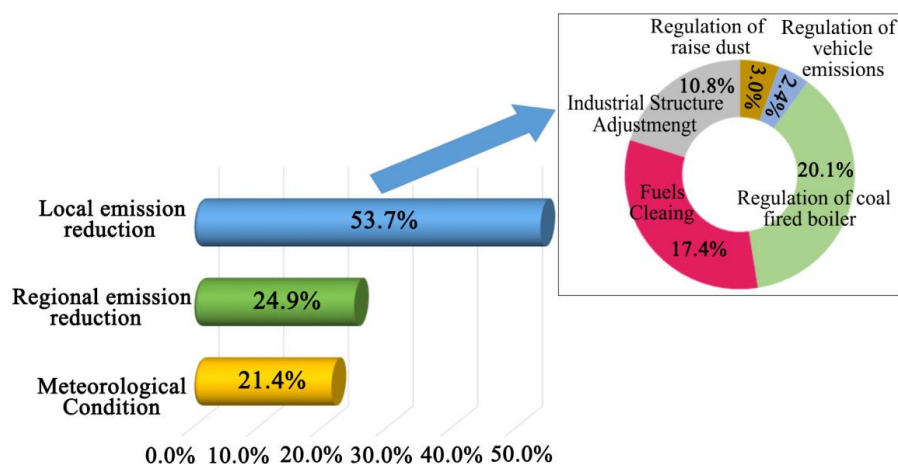
#### 347 **4.2 The relative contribution of different emission-reduction measures to the** 348 **decrease in $PM_{2.5}$ concentrations in Beijing**

349 Based on the WRF-CMAQ model, we simulated the scenario that no emission-reduction measures  
350 were implemented in Beijing from 2013 to 2017 and estimated that with emission-reduction measures,  
351 the total amount of reduction in  $SO_2$ ,  $NO_x$ , VOCs, direct  $PM_{2.5}$  and direct  $PM_{10}$  caused by these  
352 measures was 79000t, 93000t, 116000t, 44000t and 139000t respectively. The amount of reduced  
353 pollutants accounted for 83.2%、42.9%、42.4%、54.7% and 52.4% of the total emission of  $SO_2$ ,  $NO_x$ ,  
354 VOCs, direct  $PM_{2.5}$  and direct  $PM_{10}$  respectively, indicating the remarkable effect of  
355 emission-reduction measures on  $PM_{2.5}$  reduction during the past five years (UNEP,2018).

356 The observed annual average  $PM_{2.5}$  concentrations in Beijing in 2017 was  $58 \mu\text{g}/\text{m}^3$ , compared with  
357  $89.5 \mu\text{g}/\text{m}^3$  in 2013. Based on the WRF-CMAQ simulation, meteorological conditions contributed a  
358 decrease of  $6.7 \mu\text{g}/\text{m}^3$  to the total decrease of  $31.5 \mu\text{g}/\text{m}^3$ . Meanwhile, local and regional  
359 emission-reduction measures contributed  $16.9 \mu\text{g}/\text{m}^3$  and  $7.8 \mu\text{g}/\text{m}^3$  respectively. Amongst the



360 emission-reduction measures implemented in 2017, the regulation of coal boilers had the most  
361 significant effect on  $PM_{2.5}$  reduction in Beijing and resulted in a decrease of  $6.3 \mu\text{g}/\text{m}^3$ . Meanwhile,  
362 increasing clean fuels for residential use and industrial restructuring also exerted strong influences on  
363  $PM_{2.5}$  reduction and contributed to a decrease of  $5.5 \mu\text{g}/\text{m}^3$  and  $3.4 \mu\text{g}/\text{m}^3$  respectively. The relative  
364 contribution of the regulations on raise dust and vehicle emissions was relatively small, leading to a  
365 decrease of  $1.7 \mu\text{g}/\text{m}^3$  in total.



366  
367 **Fig 4. The relative contribution of different influencing factors to the decrease of  $PM_{2.5}$**   
368 **concentrations in Beijing from 2013 to 2017**

## 369 5 Discussion

370 By the end of 2017, the Beijing Five-year Clean Air Action Plan (2013-2017) was completed and  
371 achieved its primary goal of reducing the annual average  $PM_{2.5}$  concentrations to less than  $60 \mu\text{g}/\text{m}^3$ .  
372 Meanwhile, since November 2017, strong northerly winds in Beijing resulted in the cleanest winter for  
373 the past five years, raising arguments about whether the favorable meteorological conditions was  
374 primarily responsible for the  $PM_{2.5}$  reduction or whether the significant improvement in air quality in  
375 Beijing was mainly due to the control of anthropogenic emissions. In this case, a quantitative  
376 comparison between the influence of meteorological conditions and emission-reduction measures on  
377  $PM_{2.5}$  reduction is necessary for comprehensively evaluating the effects of the Five-year Clean Air  
378 Action Plan. Based on two different approaches, results of this study revealed that the control of  
379 anthropogenic emissions contributed to around 80% of the decrease in  $PM_{2.5}$  concentrations in Beijing  
380 from 2013 to 2017, indicating that the Five-Year Clean Air Plan exerted a much stronger influence on



381 the improvement of air quality than meteorological conditions. The large contribution of some specific  
382 emission-reduction measures may be obscured in the presence of favorable meteorological conditions.  
383 For instance, many residents may attribute the clean winter of 2017 to the notable strong winds without  
384 noticing some of the major emission-reduction measures implemented during this period. A large-scale  
385 replacement of coal boilers with gas boilers was conducted in Beijing and its neighboring areas since  
386 2013. As quantified by the WRF-CMAQ model, the regulation of coal boilers and increasing clean  
387 fuels for residential use in total contributed to an  $11.8\mu\text{g}/\text{m}^3$  decrease in  $\text{PM}_{2.5}$  concentrations, much  
388 (almost twice) larger than the  $6.7\mu\text{g}/\text{m}^3$  decrease brought about by favorable meteorological conditions.  
389 In general, although favorable meteorological conditions (e.g., strong winds) may lead to an instant  
390 improvement of air quality, regular emission-reduction measures exert a reliable and consistent  
391 influence on the long-term reduction of  $\text{PM}_{2.5}$  concentrations in Beijing. Given the satisfactory  
392 performance of the Five-year Clean Air Action Plan in  $\text{PM}_{2.5}$  reduction, such kind of long-term clean  
393 air plans should be further designed and implemented in the future.

394 Despite the major contribution of emission-reduction measures to  $\text{PM}_{2.5}$  reduction in Beijing,  
395 meteorological influences, which contributed to 20% of  $\text{PM}_{2.5}$  reduction, should also be considered as  
396 well. In addition to the control of anthropogenic emissions, the  $\text{PM}_{2.5}$  reduction may be realized  
397 through meteorological means. For the winter of 2017, strong northwesterly winds led to instant  
398 improvement in air quality, suggesting wind was a dominant meteorological factor for the  
399 concentration or dispersion of  $\text{PM}_{2.5}$  in Beijing. Meanwhile, previous studies (Chen et al., 2017)  
400 suggested that increasing wind speeds lead to increased evaporation, increased sunshine duration (SSD)  
401 and reduced humidity, which further reduced local  $\text{PM}_{2.5}$  concentrations. In other words, strong winds  
402 help reduce  $\text{PM}_{2.5}$  concentrations through direct and indirect measures. In this light, the forthcoming  
403 Beijing Wind-corridor Project (<http://news.10jqka.com.cn/20170331/c597397500.shtml>), which  
404 includes five 500m-width corridors and more than ten 80m-width corridors to bring in stronger  
405 wintertime northwesterly winds, can be a promising approach for promoting favorable long-term  
406 meteorological influences on  $\text{PM}_{2.5}$  reduction in Beijing.

407 Despite the remarkable decrease in  $\text{PM}_{2.5}$  concentrations, recent ground ozone pollution in Beijing has  
408 aroused growing concerns. In the past decade, ozone concentrations in Beijing demonstrated a notable  
409 increase and ozone even became the dominant pollutant in June 2017 (Cheng et al., 2018). Current  
410 emission-reduction measures, even the wind-corridor project, have been designed and implemented to  
411 simply reduce  $\text{PM}_{2.5}$  concentrations. Meanwhile, ozone concentrations even increased during specific



412 periods with strict emission-reduction measures, indicating that ordinary emission-reduction measures  
413 for PM<sub>2.5</sub> reduction were not suitable for reducing ozone concentrations. Due to complicated and  
414 unpredictable reactions between a diversity of ozone precursors, emission-reduction measures for  
415 reducing one specific precursor may conversely increase ozone concentrations (Cheng et al., 2018).  
416 Given the severe threat ground ozone exerts on public health, future emission-reduction measures  
417 should be comprehensively designed to reduce both ozone and PM<sub>2.5</sub> concentrations.

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#### 426 **Author contribution**

427 Chen, Z., Gao, B. and Xu, B designed this research. Chen, Z wrote this manuscript. Chen, D., Zhuang,  
428 Y, Cheng, N. and Li, R. conducted data analysis. Chen, D and Zhuang, Y. produced the figures. Kwan,  
429 M., and Chen, B helped revise this manuscript.



## References

1. Boylan J W, Russell A G.: PM and light extinction model performance metrics, goals and criteria for three-dimensional air quality models. *Atmospheric Environment*, 40 (26): 4946-4959,2006.
2. Brunekreef B, Holgate S.: Air pollution and health *Lancet*, 360:1233-1242,2002.
3. Chan, C.K., Yao, X.: Air pollution in mega cities in China. *Atmos. Environ.*, 42: 1-42,2008.
4. Chen, Z. Y., Cai, J., Gao, B. B., Xu, B., Dai, S., He, B., Xie, X. M.: Detecting the causality influence of individual meteorological factors on local PM<sub>2.5</sub> concentrations in the Jing-Jin-Ji region, *Scientific Reports*, 7, 40735,2017.
5. Chen, Z.Y., Xie, X., Cai, J., Chen, D., Gao, B., He, B., Cheng, N., Xu, B.: Understanding meteorological influences on PM<sub>2.5</sub> concentrations across China: a temporal and spatial perspective, *Atmos. Chem. Phys.*, 18, 5343-5358,2018.
6. Chen, Z.Y., Xu, B., Cai, J., Gao, B.B.: Understanding temporal patterns and characteristics of air quality in Beijing: A local and regional perspective. *Atmospheric Environment*. 127, 303-315,2016.
7. Cheng, N., Chen, Z., Sun, F., Sun, R., Dong, X., & Xie, X.: Ground ozone concentrations over beijing from 2004 to 2015: variation patterns, indicative precursors and effects of emission-reduction. *Environmental Pollution*, 237, 262-274,2018.
8. Cheng,N.L.,Zhang,D.W.,Li,Y.T.: Spatio-temporal variations of PM<sub>2.5</sub> concentrations and the evaluation of emission reduction measures during two red air pollution alerts in Beijing. *Scientific Reports*7, 8220 (2017),doi:10.1038/s41598-017-08895-x,2017.
9. Dominici F,Greenstone M, Sunstein C.: Particulate matter matters *Science*, 344:257-259,2014.
10. Eskridge R E, Ku J Y, Rao S T.: Separating Different Scales of Motion in Time Series of Meteorological Variables, *B. Am. Meteorol. Soc.*, 78, 1473–1483,1997.
11. Guo S, Hu M, Zamora M L.: Elucidating severe urban haze formation in China. *Proceedings of the National Academy of Sciences*, 111(49): 17373-17378,2014.
12. Hong C, Zhang Q, He K.: Variations of China's emission estimates: response to uncertainties in energy statistics. *Atmospheric Chemistry and Physics*, 17(2): 1227-1239,2017.
13. Huang,R.,Zhang,Y.,Bozzetti,C.,Ho,K.F.,Cao,J.J.,Han,Y.: High secondary aerosol contribution to particulate pollution during haze events in China. *Nature*,514: 218–222,2014.
14. Kang, D., Hogrefe, C., Foley, K. L., Napelenok, S. L., Mathur, R., Rao, S. T.: Application of the kolmogorov–zurbenko filter and the decoupled direct 3d method for the dynamic evaluation of a regional air quality model. *Atmospheric Environment*, 80(23), 58-69,2013.



15. Ma, Z. Q., Xu, J., Quan, W. J.: Significant increase of surface ozone at a rural site, north of eastern China. *Atmospheric Chemistry and Physics*, 16(6):3969-3977, 2016.
16. MEP.: 2017 air pollution prevention and management plan for the Beijing-Tianjin-Hebei region and its surrounding areas, 2017.
17. Nel A.: Air pollution-related illness effects of particles *Science*, 308:804-806, 2015.
18. Qiao, L.P., Cai, J., Wang, H.L., Wang, W.L., Zhou, M., Lou, S.R., Chen, R.J., Dai, H.X., Chen, C.H., Kan, H.D.: PM<sub>2.5</sub> Constituents and Hospital Emergency-Room Visits in Shanghai, China. *Environmental Science and Technology*. 48 (17), 10406–10414, 2014.
19. Rao S T and Zurbenko I G.: Detecting and Tracking Changes in Ozone Air Quality. *Air Waste*, 44, 1089–1092, doi:10.1080/10473289.1994.10467303, 1994.
20. Saikawa E, Kim H, Zhong M.: Comparison of emissions inventories of anthropogenic air pollutants and greenhouse gases in China. *Atmospheric Chemistry and Physics*, 17(10): 6393-6421, 2017.
21. Sun Y, Song T, Tang G.: The vertical distribution of PM<sub>2.5</sub>, and boundary-layer structure during summer haze in Beijing. *Atmospheric Environment*, 74(2):413-421, 2013.
22. UNEP.: A Review of Air Pollution Control in Beijing: 1998-2013. United Nations Environment Programme (UNEP), Nairobi, Kenya, 2016.
23. UNEP.: A Review of Air Pollution Control in Beijing: 1998-2017. United Nations Environment Programme (UNEP), Nairobi, Kenya, 2018.
24. US EPA.: Guidance on the use of models and other analyses in attainment demonstrations for the 8-hour ozone NAAQS, North Carolina, US Environmental Protection Agency, 2005.
25. Wang X, Wei W, Cheng S.: Characteristics and classification of PM<sub>2.5</sub> pollution episodes in Beijing from 2013 to 2015. *Science of the Total Environment*, 612:170-179, 2018.
26. Wang, S.X., Xing, J., Zhao, B., Jang, C., Hao, J.M.: Effectiveness of national air pollution control policies on the air quality in metropolitan areas of China. *J. Environ. Sci.* 26 (1):13–22. [http://dx.doi.org/10.1016/s1001-0742\(13\)60381-2](http://dx.doi.org/10.1016/s1001-0742(13)60381-2), 2014.
27. Zhang, Q., He, K., Huo, H.: Cleaning China's air. *Nature*, 484:161-162, 2012.
28. Zhao, B., Wang, S.X., Wang, J.D., Fu, J., Liu, T.H., Xu, J.Y., Fu, X., Hao, J.M.: Impact of national NO<sub>x</sub> and SO<sub>2</sub> control policies on particulate matter pollution in China. *Atmos. Environ.* 77, 453–463, 2013.
29. Zheng, G. J., Duan, F. K., Su, H., Ma, Y. L., Cheng, Y., Zheng, B.: Exploring the severe winter haze in Beijing: the impact of synoptic weather, regional transport and heterogeneous reactions,





Atmos. Chem. Phys., 15, 2969–2983, doi:10.5194/acp-15-2969-2015, 2015.

30. Zhou Y, Zhao Y, Mao P.: Development of a high-resolution emission inventory and its evaluation and application through air quality modeling for Jiangsu Province, China[J]. Atmospheric Chemistry and Physics, 17(1): 211–233, 2017.
31. Zurbenko I, Chen J, Rao S T.: Detecting discontinuities in time series of upper air data: Demonstration of an adaptive filter technique, J. Climate, 9, 3548–3560, 1996.