

1 **The control of anthropogenic emissions contributed to 80% of the**
2 **decrease in PM_{2.5} concentrations in Beijing from 2013 to 2017**

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20 **Abstract**

21 With the completion of the Beijing Five-year Clean Air Action Plan by the end of
22 2017, the annual mean PM_{2.5} concentration in Beijing dropped dramatically to 58.0
23 μg/m³ in 2017 from 89.5 μg/m³ in 2013. However, controversies exist to argue that
24 favorable meteorological conditions in 2017 were the major driver for such rapid
25 decrease in PM_{2.5} concentrations. To comprehensively evaluate this five-year plan, we
26 employed Kolmogorov-Zurbenko (KZ) filter and WRF-CMAQ to quantify the
27 relative contribution of meteorological conditions and the control of anthropogenic
28 emissions to PM_{2.5} reduction in Beijing from 2013 to 2017. For these five years, the
29 relative contribution of emission-reduction to the decrease of PM_{2.5} concentrations
30 calculated by KZ filtering and WRF-CMAQ was 80.6% and 78.6% respectively. KZ
31 filtering suggested that short-term variations of meteorological and emission

32 conditions contributed majorly to rapid changes of PM_{2.5} concentrations in Beijing.
33 WRF-CMAQ revealed that the relative contribution of local and regional
34 emission-reduction to PM_{2.5} decrease in Beijing was 53.7% and 24.9% respectively.
35 For local emission-reduction measures, the regulation of coal boilers, increasing use
36 of clean fuels for residential use and industrial restructuring contributed to 20.1 %,
37 17.4% and 10.8% of PM_{2.5} reduction respectively. Both models suggested that the
38 control of anthropogenic emissions accounted for around 80% of the PM_{2.5} reduction
39 in Beijing, indicating that emission-reduction was crucial for air quality enhancement
40 in Beijing from 2013 to 2017. Consequently, such long-term air quality clean plan
41 should be continued in the following years to further reduce PM_{2.5} concentrations in
42 Beijing.

43 **Keywords: PM_{2.5}, anthropogenic emissions, meteorological conditions,**
44 **Kolmogorov-Zurbenko (KZ) filtering, WRF-CMAQ**

45 **1 Introduction**

46 In January 2013, persistent haze episodes occurred in Beijing, during which the highest
47 hourly PM_{2.5} concentration once reached 886 $\mu\text{g}/\text{m}^3$, a historic high record.
48 High-concentration PM_{2.5} led to long-lasting black and thick fogs, which not only
49 significantly influenced people's daily life (low-visibility induced traffic jam), but also posed
50 a severe threat to public health (Brunekreef et al., 2002; Dominici et al., 2014; Nel et al.,
51 2005; Zhang et al., 2012; Qiao et al., 2014). Since then, severe haze episodes have frequently
52 been observed in Beijing and other regions across China (Chan et al., 2008; Huang, R., et al.,
53 2014; Guo et al., 2014; Zheng et al., 2015), and PM_{2.5} pollution has become one of the most
54 concerned environmental issues in China. Consequently, a national network for monitoring
55 hourly PM_{2.5} concentrations has been established gradually, including 35 ground observation
56 stations in Beijing, which provide important support for better understanding and managing
57 PM_{2.5} concentrations. To effectively mitigate PM_{2.5} pollution, Beijing Municipal
58 Government released "Beijing Five-year Clean Air Action Plan (2013-2017)" with a series of
59 long-term emission-reduction measures, including shutting down heavily polluting factories,
60 restricting traffic emissions and replacing coal fuels with clean energies, and "Heavy Air
61 Pollution Contingency Plan" with a series of contingent emission-reduction measures during
62 heavy pollution episodes. By the end of 2017, these long-term and contingent
63 emission-reduction measures worked jointly to reduce the annually mean PM_{2.5}
64 concentration in Beijing from 89.5 $\mu\text{g}/\text{m}^3$ in 2013 to 58.0 $\mu\text{g}/\text{m}^3$ in 2017, indicating a great
65 success of PM_{2.5} management during the past five years. The notable decrease of PM_{2.5}
66 concentrations attracted nationwide attentions and growing studies have been conducted to
67 understand spatio-temporal characteristics (Shao et al., 2018; Sun et al., 2019; Wang et al.,
68 2019), sources (Chen et al., 2019; Xu et al., 2019; Cheng, J. et al., 2019) and health effects
69 (Liang et al., 2019) of PM_{2.5} variations in Beijing from 2013 to 2017. These studies revealed
70 that air quality in Beijing was improved significantly in 2017 in terms of annual mean PM_{2.5}
71 concentrations, polluted days and pollution durations. Furthermore, despite different outputs,
72 both source apportionment during pollution episodes based on collected samples (Shao et al.,
73 2019; Xu et al., 2019; Chen et al., 2019) and long-term model simulation based on regional
74 and local emission inventories (Cheng, J. et al., 2019) suggested that local and regional
75 anthropogenic emissions (e.g. coal combustion and vehicle emissions) were the major
76 influencing factors for long-term and short-term PM_{2.5} variations in Beijing.

77 In addition to anthropogenic emissions, the strong meteorological influences on PM_{2.5}
78 concentrations in Beijing have been widely acknowledged (Zhao et al., 2013; Wang et al.,
79 2014; UNEP, 2016; Cheng et al., 2017; Chen et al., 2017; Sun et al., 2019). For instance, for
80 2014, more than 180 days in Beijing experienced a dramatic daily AQI (Air Quality Index)
81 change ($\Delta\text{AQI}>50$) (Chen, Z. et al., 2016). Considering that anthropogenic emissions for a
82 mega city unlikely changed significantly on a daily basis, rapid variations of meteorological
83 conditions were one major driver for the dramatic change of daily air quality in Beijing. In
84 winter 2017, strong northwest winds led to favorable meteorological conditions for PM_{2.5}
85 diffusion and low PM_{2.5} concentrations in Beijing. This raised the controversy that
86 meteorological conditions, instead of emission-reduction, accounted for the remarkable
87 PM_{2.5} reduction in Beijing. In this case, with the completion of the five-year plan, it is highly
88 necessary to quantify the relative contribution of meteorological conditions and
89 emission-reduction to the notable decrease in PM_{2.5} concentrations in Beijing from 2013 to
90 2017.

91 In recent years, growing studies have been conducted to investigate meteorological and
92 anthropogenic influences on long-term PM_{2.5} variations. Based on Goddard Earth Observing
93 System (GEOS) chemical transport model (GEOS-Chem), Yang et al (2016) revealed that
94 the relative contribution of meteorological conditions to PM_{2.5} variations in Eastern China
95 from 1985 to 2005 was 12%. Based on a multiple general linear model (GLM), Gui et al.
96 (2019) quantified that meteorological conditions accounted for 48% of PM_{2.5} variations in
97 Eastern China from 1998 to 2016. Based on a stepwise multiple linear regression (MLR)
98 model, Zhai et al. (2019) quantified the relative contribution of meteorology to PM_{2.5}
99 variations from 2013 to 2018 in Beijing-Tianjin-Hebei region, Yangtze River Delta, Pearl
100 River Delta and Sichuan Basin and Fenwei plain was 14%, 3%, 19%, 27% and 23%
101 respectively. Through a two-stage hierarchical clustering method, Zhang et al. (2018)
102 calculated that the relative contribution of meteorological conditions to heavy pollution
103 episodes within the Beijing-Tianjin-Hebei region was larger than 50% from 2013 to 2017.
104 These studies quantified the overall meteorological influences on long-term PM_{2.5} variations
105 using different statistical models and chemical transport models (CTMs). However, due to
106 strong interactions between individual meteorological factors, traditional statistical methods
107 such as correlation analysis and linear regression may be biased significantly when
108 quantifying meteorological influences on PM_{2.5} concentrations (Chen et al., 2017). On the

109 other hand, the accuracy of CTMs can be influenced largely by the uncertainty in emission
110 inventories (Xu et al., 2016) and deficiency of heterogeneous/aqueous processes (Li et al.,
111 2011). Therefore, multiple advanced models should be comprehensively considered to better
112 quantify meteorological influences on PM_{2.5} concentrations (Pearce et al., 2011).

113 To evaluate this five-year clean-air plan, we employ an advanced statistical model,
114 Kolmogorov-Zurbenko (KZ) filtering, which is advantageous of filtering meteorological
115 influences on long-term time series of airborne pollutants, and a CTM model, WRF-CMAQ,
116 which is advantageous of quantifying the relative contribution of different emission sources,
117 to comprehensively investigate the relative contribution of meteorological conditions and
118 emission-reduction to PM_{2.5} reduction in Beijing from 2013 to 2017 respectively. In this light,
119 this research provides important insight for better designing and implementing successive
120 clean air plans in the future to further mitigate PM_{2.5} pollution in Beijing.

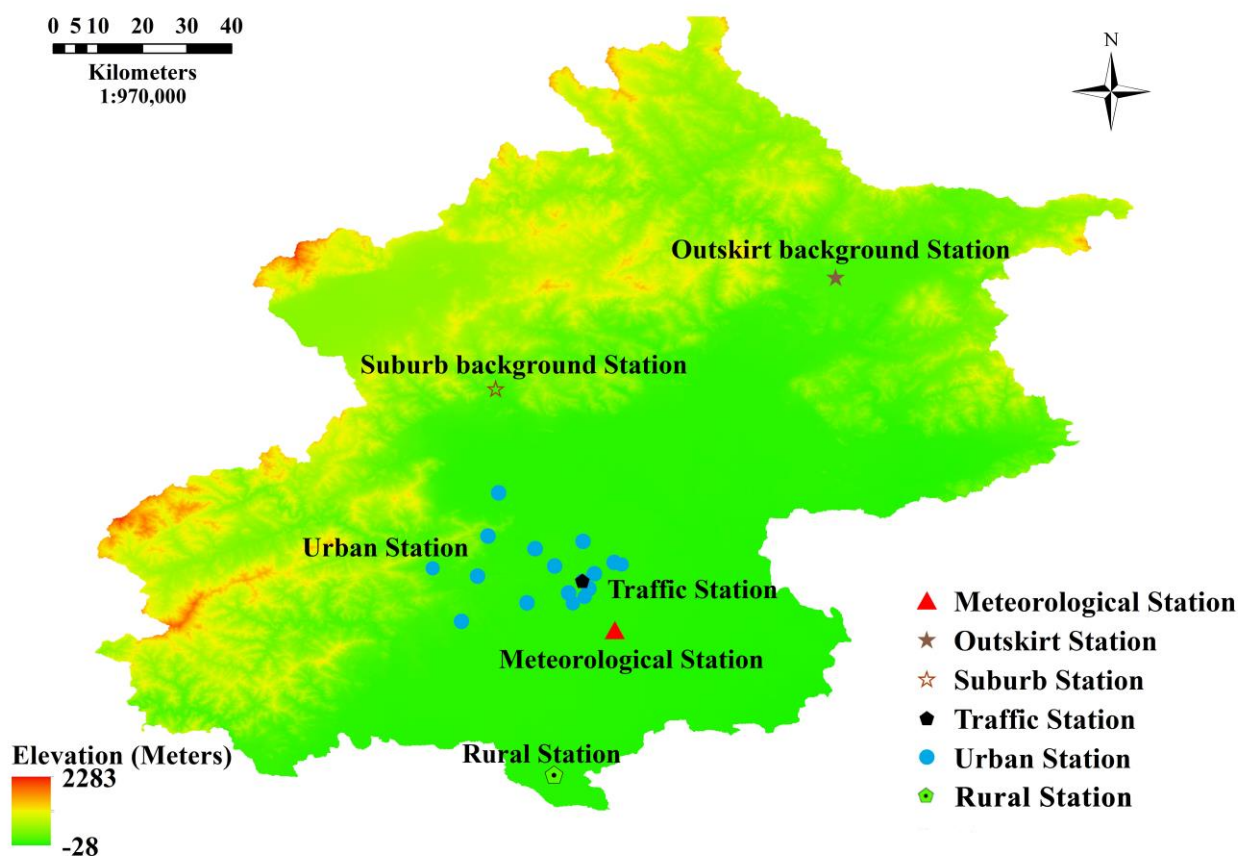
121 This manuscript is structured as follows: Firstly, major data sources, including PM_{2.5} and
122 meteorological data, and emission inventories, employed for this research are briefly
123 introduced. Secondly, the principle and parameter setting of two models, KZ filtering and
124 WRF-CMAQ, and model verification are explained. In the result section, the relative
125 contribution of meteorological conditions and anthropogenic emissions to PM_{2.5} variations in
126 Beijing from 2013 to 2017 calculated using both models is presented. In the discussion and
127 conclusion part, implementations of this research and suggestions for further improving air
128 quality in Beijing are given.

129 **2 Data Sources**

130 **2.1 PM_{2.5} and meteorological data**

131 In this study, hourly PM_{2.5} concentration data were acquired from the website PM25.in
132 (www.PM25.in), which collects official data provided by China National Environmental
133 Monitoring Center (CNEMC). Beijing has established an advanced air quality monitoring
134 network with 35 ground stations across the city. Considering the major contribution of
135 industry and traffic-induced emissions in urban areas, we selected all twelve urban stations
136 to analyze spatio-temporal variations of PM_{2.5} concentrations and quantify their influencing
137 factors. In addition to these urban stations, we selected two background stations, the

138 DingLing Station located in the suburb and the MiYun Reservoir Station located in the outer
139 suburb, one transportation station (the Qianmen station) located close to a main road, and
140 one rural station (the Yufa Station) that is far away from central Beijing for the following
141 analysis. The DingLing and MiYun Reservoir Station were chosen as background stations by
142 the Ministry of Environmental Protection of China. These two stations receive limited
143 influence from anthropogenic emissions due to their location in suburban and outer suburban
144 areas. The Qianmen transportation station received more influences from vehicle emissions.
145 Long-term variations of PM_{2.5} concentrations in different types of stations provide a useful
146 reference for comprehensively understanding the effects of emission-reduction measures on
147 PM_{2.5} decrease in Beijing from 2013 to 2017. Meteorological data for this research were
148 collected from the Guanxiangtai Station (GXT,54511, 116.46° E, 39.80° N), Beijing and
149 downloaded from the Department of Atmospheric Science, College of Engineering,
150 University of Wyoming (<http://weather.uwyo.edu/upperair/sounding.html>). Both PM_{2.5} and
151 meteorological data were collected from January 1st, 2013 to December 31st, 2017. The
152 locations of these selected stations are shown in Fig 1.



153

154

Fig 1. Locations of different ground monitoring stations.

155 **2.2 Emission inventories**

156 For this research, we employed both regional and local emission inventories for running
157 model simulation. Multi-resolution Emission Inventory for China, MEIC,
158 (<http://meicmodel.org/>) provided by Tsinghua University, were employed as the regional
159 emission inventories. MEIC has been widely employed and verified as a reliable emission
160 inventory by a diversity of studies (Hong et al., 2017; Saikawa et al., 2017; Zhou et al., 2017;
161 etc.). For simulating five-year PM_{2.5} concentrations, MEIC from 2013 to 2017 are required.
162 Since official MEIC 2017 has yet been available, we employed a strategy from previous
163 studies (Chen et al., 2019; etc) and updated MEIC 2016 for simulating emission-reduction
164 scenarios and PM_{2.5} concentrations in 2017 by considering official 2017 emission-reduction
165 plans (e.g. the target of coal combustion reduction) required by the local government.

166 Different from regional emission inventories, local emission inventories are usually
167 produced independently by local institutions. The Beijing local-emission inventory
168 employed for this research was produced and updated by Beijing Municipal Research
169 Institute of Environmental protection, fully according to the requirement of MEP on the
170 production of local emission inventories within Beijing-Tianjin-Hebei region. This Beijing
171 local-emission inventory from 2013 to 2017 was produced by synthesizing local
172 environmental statistical data and reported emission data, carrying out field investigations
173 and conducting a series of estimation according to Beijing Five-year Clean Air Action Plan.
174 As shown in table 1, it is highly consistent with other official statistical data, such as the
175 Annual report from National Environmental Statistics Bulletin
176 (http://www.mee.gov.cn/gzfw_13107/hjtj/qghjtjgb/) and “2+26” Center for Air Pollution
177 Prevention and Control, and has been formally employed for the implementation of recent
178 “2017 Air Pollution Prevention and Management Plan for the Beijing-Tianjin-Hebei Region
179 and its Surrounding Areas” (MEP, 2017).

180

181 **Table 1. The comparison of local environmental statistical data used for this research**
 182 **and other official statistical data in 2017 (unit: 10k tons)**

	SO ₂	NO _x	CO	VOC	NH ₃	PM ₁₀	PM _{2.5}	BC	OC
Statistical data for this research	1.38	10.15	49.54	13.47	3.20	14.74	3.92	0.17	0.44
National Environmental Statistics Bulletin	1.38	12.16	52.03	24.24	3.26	14.68	3.91	0.22	0.41
“2+26” center for air pollution prevention and control	0.89	9.24	48.98	13.93	3.16	13.82	3.72	0.19	0.46

183 **3 Methods**

184 A key step for quantifying the relative contribution of anthropogenic emissions to PM_{2.5}
 185 variations is to properly filter meteorological influences on PM_{2.5} concentrations, which is
 186 highly challenging and rarely investigated by previous studies. Therefore, we employed both
 187 a statistical method and a CTM to comprehensively evaluate the role of anthropogenic
 188 emissions and meteorological conditions in the decrease of PM_{2.5} concentrations in Beijing
 189 from 2013 to 2017.

190 **3.1 Kolmogorov-Zurbenko (KZ) filtering**

191 Since meteorological conditions exert a strong influence on PM_{2.5} concentrations in Beijing,
 192 the removal of seasonal signals from time series of meteorological factors produces data sets
 193 suitable for understanding the trend of PM_{2.5} concentrations mainly influenced by
 194 anthropogenic factors (Eskridge et al., 1997). To better analyze the trend of time series data
 195 without the disturbances from other major influencing variables, a statistical method
 196 Kolmogorov-Zurbenko (KZ) filtering was proposed by Rao et al. (1994). The KZ filter is
 197 advantageous of removing high-frequency variations in data sets through iterative moving
 198 average. Eskridge et al. (1997) compared four major approaches for trend detection,
 199 including PEST, anomalies, wavelet transform, and the KZ filter, and suggested that KZ
 200 achieved higher confidence in detecting long-term trend than other models. Due to its
 201 reliable performance in trend detection in complicated ecosystems, the KZ filter has been
 202 increasingly employed to remove seasonal signals of meteorological conditions and extract
 203 long-term trend of airborne pollutants (Zurbenko, et al., 1996; Eskridge, et al., 1997; Kang,

204 et al., 2013; Ma et al., 2016; Cheng, N et al., 2019). One potential limitation of the KZ filter
205 is that iterative moving average (m) may impose an influence on detecting abrupt variations.
206 Therefore, Zurbenko et al. (1996) proposed an enhanced KZ filter that employed a dynamic
207 variable m that decreased with the increase in changing rate. For this research, we employed
208 this dynamic m to produce an adjusted time-series of PM_{2.5} concentrations in Beijing by
209 removing large inter-annual and seasonal variations in meteorological conditions. The
210 principle of the KZ filter is briefly introduced as follows.

211 The raw time-series of airborne pollutants can be decomposed as:

$$212 \quad X(t) = E(t) + S(t) + W(t) \quad (1)$$

$$213 \quad X_b(t) = E(t) + S(t) \quad (2)$$

$$214 \quad E(t) = KZ_{365,3}(X) \quad (3)$$

$$215 \quad S(t) = KZ_{15,5}(X) - KZ_{365,3}(X) \quad (4)$$

$$216 \quad W(t) = X(t) - KZ_{15,5}(X) \quad (5)$$

217 **Where $X(t)$ is the original time series of airborne pollutants, $E(t)$ is the long-term trend component,**
218 **$S(t)$ is the seasonal component, $W(t)$ is the short-term (synoptic-scale) component or residue. $KZ_{i,j}(X)$**
219 **indicates KZ filtering on the original dataset X with a moving window size of i and j iterations.**

220 $X_b(t)$ stands for the base component, the sum of the long-term and seasonal component,
221 presenting steady trend variation. $E(t)$ is mainly affected by long-term anthropogenic
222 emission and climate change. $S(t)$ is mainly influenced by the seasonal variation of emission
223 and meteorological conditions. $W(t)$ is caused by short-term and small-scale shifts of
224 emissions and meteorological conditions.

225 The long-term trend component $E(t)$ processed by KZ filtering still contains the influence of
226 meteorological conditions, which can be removed by multiple regression models. Multiple
227 linear relationships are established for the residue and baseline component respectively using
228 meteorological factors strongly correlated with airborne pollutants.

229 We examined correlations between seasonal PM_{2.5} concentrations in Beijing and a set of
230 meteorological factors, including temperature, wind speed, wind direction, precipitation,
231 relative humidity, solar radiation, evaporation and air pressure. Due to limited space, detailed
232 correlations between PM_{2.5} concentrations and individual meteorological factors in Beijing

233 are not presented here and readers can refer to previous studies for more information (Chen
 234 et al., 2017; 2018). The correlation analysis revealed that wind speed, relative humidity,
 235 temperature and solar radiation were strongly and significantly correlated with PM_{2.5}
 236 concentrations in Beijing (as shown in Table 2), which was consistent with findings from
 237 other studies (Sun et al., 2013; Wang et al., 2018).

238 **Table 2. Major meteorological factors strongly correlated with seasonal PM_{2.5}**
 239 **concentrations in Beijing (Chen et al., 2017)**

Spring	Summer	Autumn	Winter
	RHU**(0.648)	RHU**(0.587)	RHU**(0.738)
RHU**(0.532)	SSD**(-0.447)	SSD**(-0.509)	SSD**(-0.715)
	TEM**(0.554)	WIN**(-0.468)	WIN**(-0.558)

240 ****Correlation is significant at the 0.01 level (2 tailed);**

241 **RHU: Relative humidity; SSD: Sunshine Duration; TEM: Temperature; WIN: Wind speed**

242 Therefore, we further established multiple linear regression equations between PM_{2.5}
 243 concentrations and wind speed, relative humidity, temperature and solar radiation as follows.

$$244 \quad W(t) = \alpha_0 + \sum \alpha_i w_i(t) + \varepsilon_w(t) \quad (6)$$

$$245 \quad X_b(t) = b_0 + \sum b_i x_i(t) + \varepsilon_b(t) \quad (7)$$

$$246 \quad \varepsilon(t) = \varepsilon_w(t) + \varepsilon_b(t) \quad (8)$$

247 **Where $w_i(t)$ and $x_i(t)$ stand for the different short-term and baseline component of the i^{th}**
 248 **meteorological factor. ε_w and ε_b is the regression residue of the short-term and baseline**
 249 **component. $\varepsilon(t)$ indicates the total residue, including the short-term influence of local emission**
 250 **and meteorological factors neglected during the regression process and other noises.**

251 Next, KZ filtering was conducted on the $\varepsilon(t)$ for its long-term component $\varepsilon_E(t)$. After the
 252 influence of meteorological variations was filtered, the reconstructed time series of airborne
 253 pollutants $X_{LT}(t)$ was calculated as the sum of $\varepsilon_E(t)$ and the average value of $E(t)$, $\overline{E(t)}$.

$$254 \quad X_{LT}(t) = \overline{E(t)} + \varepsilon_E(t) \quad (9)$$

255 After KZ filtering, the relative contribution of meteorological conditions to PM_{2.5} variations
 256 can be calculated as follows:

257
$$P_{contrib} = \frac{K_{org} - K}{K_{org}} \times 100\% \quad (10)$$

258 **Where $P_{contrib}$ is the relative contribution of meteorological conditions to $PM_{2.5}$ variations in Beijing,**
259 **K_{org} is the variation slope of the original $PM_{2.5}$ time series; K is the variation slope of adjusted $PM_{2.5}$**
260 **time series with filtered influences from meteorological variations.**

261 **3.2 WRF-CMAQ model**

262 We employed WRF-CMAQ for simulating the effects of emission-reduction on the decrease
263 of $PM_{2.5}$ concentrations. WRF-CMAQ includes three models: The middle-scale meteorology
264 model (WRF), the source emission model (SMOKE) (<http://www.cmascenter.org/smoke/>)
265 and the community multiscale air quality modeling system (CMAQ)
266 (<http://www.cmascenter.org/CMAQ>). The center of the CMAQ was set at coordinate 35°N,
267 110°E and a bi-directional nested technology was employed, producing two layers of grids
268 with a horizontal resolution of 36 km and 12 km respectively. The first layer of grids with
269 36km resolution and 200×160 cells covered most areas in East Asia (including China, Japan,
270 North Korea, South Korea, and other countries). The second layer of grids with 12km
271 resolution and 120×102 cells covered the North China Plain (including the
272 Beijing-Tianjin-Hebei region, Shandong and Henan Province). The vertical layer was
273 divided into 20 unequal layers, eight of which were of a less-than-1km distance to the
274 ground for better featuring the structure of atmospheric boundary. The height of the ground
275 layer was 35m.

276 We employed ARW-WRF3.2 to simulate the meteorological field. The setting of the center
277 and the bidirectional nest for WRF and CMAQ was similar. There were 35 vertical layers for
278 WRF and the outer layer provided boundary conditions of the inner layer. The
279 meteorological background field and boundary information with a FNL resolution of 1°×1°
280 and temporal resolution of 6h were acquired from NCAR (National Center for Atmospheric
281 Research, <https://ncar.ucar.edu/>) and NCEP (National Centers for Environmental Prediction)
282 respectively. The terrain and underlying surface information was obtained from the USGS
283 30s global DEM (<https://earthquake.usgs.gov/>). The outputs from WRF were interpolated to
284 the region and grid of CMAQ using the Meteorology-Chemistry Interface Processor (MCIP,
285 <https://www.cmascenter.org/mcip>). The meteorological factors used for this model included
286 temperature, air pressure, humidity, geopotential height, zonal wind, meridional wind,
287 precipitation, boundary layer heights and so forth. An estimation model for terrestrial

288 ecosystem MEGAN (<http://ab.inf.uni-tuebingen.de/software/megan/>) was employed to
289 process the natural emissions. Multi-resolution Emission Inventory for China, MEIC
290 0.5°×0.5° emission inventory (<http://www.meicmodel.org/>) and Beijing emission inventory
291 (<http://www.cee.cn/>) provided anthropogenic emission data. We input the processed natural
292 and anthropogenic emission data into the SMOKE model and acquired comprehensive
293 emission source files.

294 Scenario simulation is employed to estimate the contribution of emission-reduction to the
295 variation of PM_{2.5} concentrations.

$$296 \quad P_{contrib} = \frac{C - C_{base}}{C} \times 100\% \quad (11)$$

297 **Where $P_{contrib}$, C and C_{base} are the contribution rate of emission-reduction to PM_{2.5}**
298 **concentrations, simulated PM_{2.5} concentrations under the emission-reduction scenario, and**
299 **simulated PM_{2.5} concentrations under the baseline scenario respectively.**

300 To evaluate the relative contribution of meteorological conditions and different
301 emission-reduction measures to the decrease of PM_{2.5} concentrations, we designed two
302 baseline experiments and four sensitivity experiments. For the first baseline experiment, we
303 employed the actual meteorological data in 2013. For the second baseline experiment, we
304 employed the actual meteorological data in 2017 and emission inventory in 2017. Since no
305 emission-reduction measures were conducted in 2013, the first baseline experiment was used
306 to estimate the relative contribution of meteorological conditions to the variation of PM_{2.5}
307 concentrations. By comparing the first and second baseline experiment, the relative
308 contribution of all emission-reduction measures to the variation of PM_{2.5} concentrations can
309 be quantified. For the first sensitivity experiment, we employed the actual meteorological
310 conditions in 2013 and emission inventory in 2017 and compared the simulation result with
311 the baseline experiment, which demonstrated the relative contribution of meteorological
312 concentrations to PM_{2.5} reduction in Beijing from 2013 to 2017. Since the WRF-CMAQ
313 simulation simply considers PM_{2.5} concentrations and meteorological conditions in 2013 and
314 2017 without considering their variation process from 2013 to 2017, KZ filtering may
315 perform better in quantifying the relative contribution of meteorological variations to PM_{2.5}
316 reduction in Beijing. However, the output from this sensitivity experiment served as a useful
317 reference for cross-verifying the output from the KZ filtering. For the remaining three
318 sensitivity-simulation experiments, we added the reduced emission amount induced by one

319 specific emission-reduction measure to the actual emission amount in 2017 and kept other
320 parameters unchanged, and thus quantified the relative contribution of one specific
321 emission-reduction measure to PM_{2.5} reduction in Beijing from 2013 to 2017. Consequently,
322 we quantified the relative contribution of three major emission-reduction measures to PM_{2.5}
323 reduction in Beijing (Table 3).

Table 3. The design and materials for two baseline and four sensitivity experiments using WRF-CMAQ

ID	Meteorological Data	Emission-reduction measures	Simulation Year	Major purposes
Baseline Experiment1	2013	No emission-reduction Measures	2013	2013 baseline scenario
Baseline Experiment2	2017	All emission-reduction Measures	2017	2017 baseline scenario
Sensitivity Experiment 1	2013	All emission-reduction Measures	2017	The relative contribution of meteorological variations to the decrease of PM _{2.5} concentrations in Beijing from 2013 to 2017
Sensitivity Experiment 2	2017	All emission-reduction measures except for industrial restructuring	2017	The relative contribution of industrial restructuring to the decrease of PM _{2.5} concentrations in Beijing from 2013 to 2017
Sensitivity Experiment 3	2017	All emission-reduction measures except for the regulation of coal boilers	2017	The relative contribution of the regulation of coal boilers to the decrease of PM _{2.5} concentrations in Beijing from 2013 to 2017
Sensitivity Experiment 4	2017	All emission-reduction measures except for increasing clean fuels for civil use	2017	The relative contribution of increasing clean fuels for civil use to the decrease of PM _{2.5} concentrations in Beijing from 2013 to 2017

For emission data, all experiments employed Beijing local emissions inventory in 2017 for Beijing and regional emission inventory in 2017 for other regions.

MEIC 2017 was acquired based on our update of MEIC 2016 according to official 2017 emission-reduction targets required by the local government.

327 **3.3 Model verification**

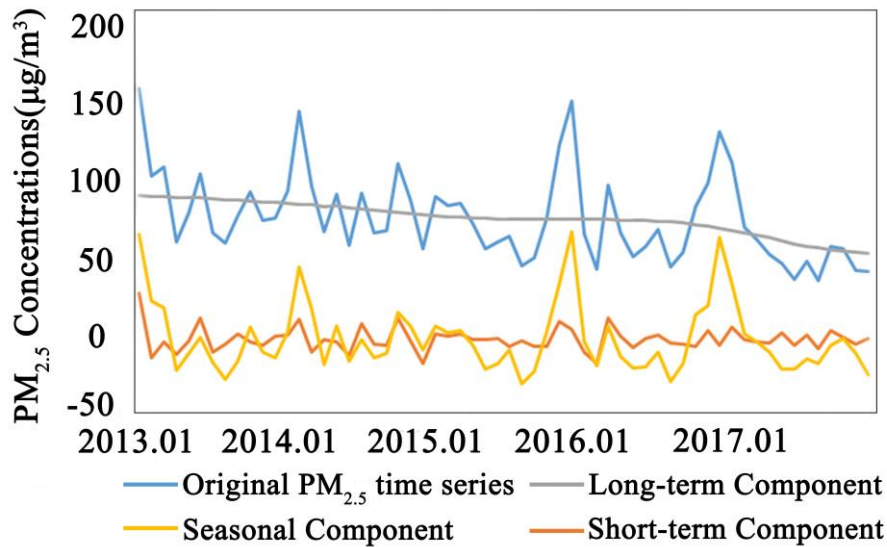
328 **3.3.1 Verification of KZ filtering**

329 For each station, the original time series of PM_{2.5} data was processed by the KZ filter
330 and the relative contribution of the long-term, seasonal and short-term component to
331 the total variance is shown as Table 4. The sum of the long-term, seasonal and
332 short-term component contributed 93.6~95.3% to the total variance in different
333 stations respectively. The larger the total variance, the three components are more
334 independent to each other. The total variance close to 100% suggests that a majority
335 of meteorological influences has been considered and effectively removed. As shown
336 in Table 4, the large value of the total variation in all stations indicated a satisfactory
337 output from the KZ filtering.

338 Specifically, the relative contribution of the seasonal component (ranging from
339 9%-23.8%) and short-term component (ranging from 66.8%-83.8%) was much larger
340 than that of the long-term component (ranging from 1.2%-3.5%), suggesting that
341 seasonal and short-term variations of meteorological and emission factors exerted a
342 major influence on the rapid change of PM_{2.5} concentrations in Beijing. The
343 decomposed long-term, seasonal and short-term component from the original time
344 series of mean urban PM_{2.5} concentrations in Beijing from 2013 to 2017 are
345 demonstrated as Fig 2. According to Fig 2, the notable peaks of decomposed seasonal
346 and short-term component were highly consistent with the peaks of PM_{2.5}
347 concentrations in the original time-series, which further proved the dominant
348 influence of seasonal and short-term variations of meteorological and anthropogenic
349 factors on the temporal changes of PM_{2.5} concentrations in Beijing.

350 **Table 4. The relative contribution of different components to the total variance of**
 351 **original time series of PM_{2.5} concentrations from 2013-2017 at different stations**

Stations	Long-term component (%)	Seasonal component (%)	Short-term component (%)	Total variance(%)
Yufa	2.1	23.8	66.8	94.0
Miyun Reservoir	1.4	9.0	83.8	95.2
Dingling	1.6	11.0	81.3	94.9
Qianmen	2.7	12.7	78.5	95.1
Olympic center	2.1	11.9	80.0	95.3
Xiangshan	1.2	10.3	83.4	94.9
Huayuan	2.2	15.9	75.6	93.7
Yungang	2.1	15.1	76.5	93.6
WanShouxigong	1.6	14.2	78.2	94.0
Dongsi	1.6	12.3	80.0	94.0
TianTan	2.1	13.2	78.6	93.8
NongZhanguan	1.8	13.7	78.6	94.1
Gucheng	1.8	13.5	78.5	93.7
Guanyuan	1.6	12.6	79.8	94.0
BeiBuxinqu	1.7	13.8	78.4	93.9
WanLiu	3.5	11.9	78.2	93.6



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Fig 2. The long-term, seasonal and short-term component extracted from the original time series of mean urban PM_{2.5} concentrations in Beijing from 2013 to 2017

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3.3.2 Verification of WRF-CMAQ

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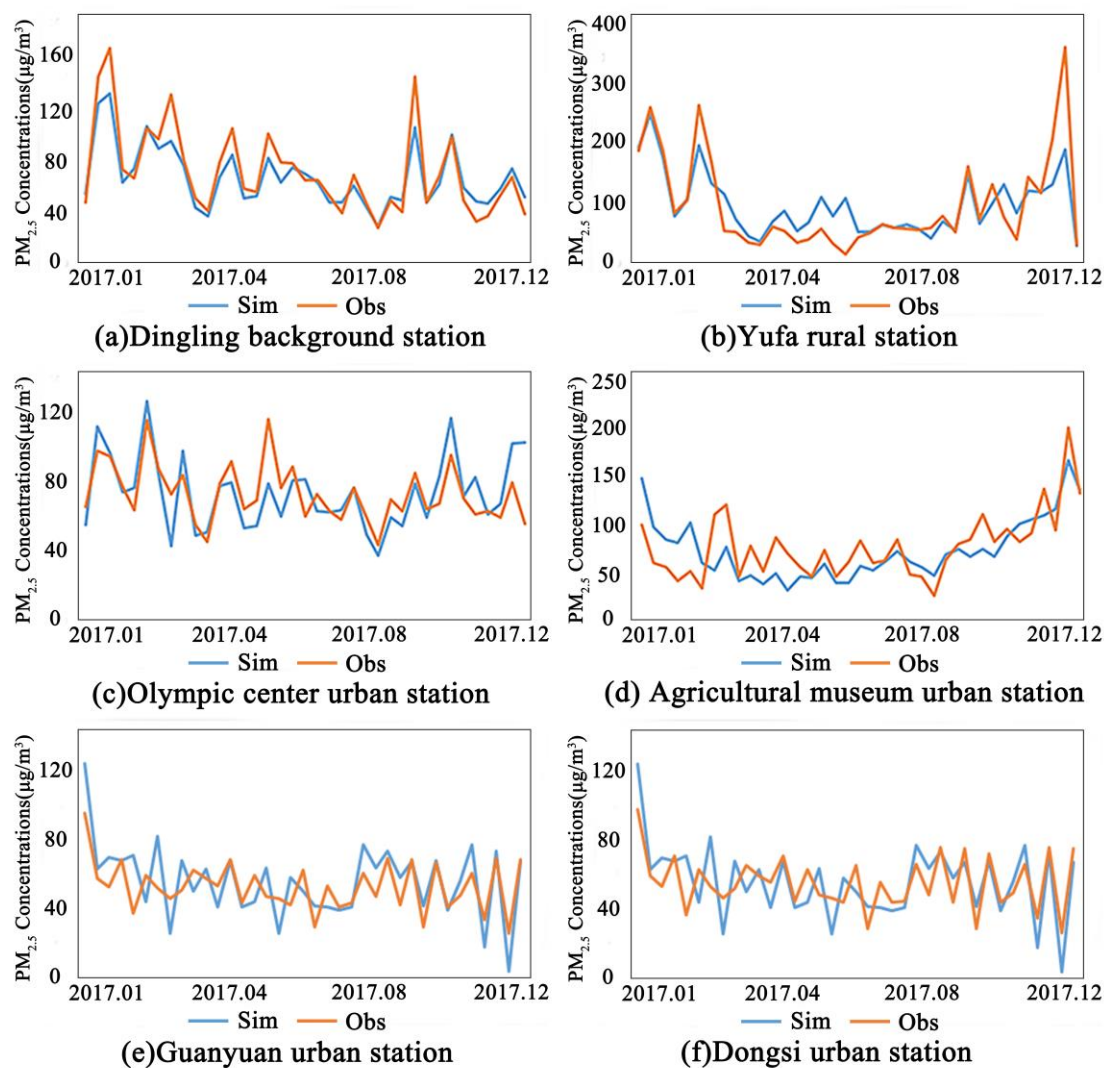
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We employed the emission inventory and meteorological data for 2017 to verify the accuracy of WRF-CMAQ simulation. For six stations of different types (DingLing background station, Yufa rural station, Olympic Center urban station, Guanyuan urban station, Dongsu urban station and Agricultural museum urban station), we compared the observed and estimated PM_{2.5} concentrations and presented the comparison result as Fig 3. According to Fig 3, the general trend of the simulated PM_{2.5} concentrations was consistent with that of the observed PM_{2.5} concentrations. For six stations, the correlation coefficient R, normalized mean bias (NMB), normalized mean error (NME), mean fractional bias (MFB) and mean fractional error (MFE) between observed and simulated data was 0.63~0.91, -6%~6%, 26%~40%, -5%~7%, and 27%~46% respectively, indicating a satisfactory simulation output (EPA, 2005; Boylan et al., 2006). However, as shown in Figure 3, WRF-CMAQ may notably underestimate PM_{2.5} concentrations during heavy pollution episodes due to unified parameter setting for long-term simulation, the uncertainty in emission inventories, and especially insufficient chemical reaction mechanisms, which is a common challenge for CTM-based PM_{2.5} simulation (Li et al., 2011). For instance, without considering heterogeneous/aqueous reactions between multiple precursors, CTMs failed to approach the maximum PM_{2.5} concentrations during severe haze episodes

374 and the simulation accuracy was dramatically improved by including proper
375 descriptions of heterogeneous/aqueous reactions into CTMs (Chen, D. et al. 2016).
376 With more finer-scale emission inventories and better descriptions of reaction
377 mechanisms between precursors, the accuracy of PM_{2.5} simulation can be improved
378 significantly.



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380 **Fig 3. The comparison between observed and WRF-CMAQ simulated PM_{2.5}**
381 **concentrations in 2017 in six stations across Beijing**

382 **4 Results**

383 **4.1 The relative contribution of emission-reduction and meteorological variations** 384 **to the decrease of PM_{2.5} concentrations in Beijing from 2013 to 2017**

385 **4.1.1 Estimation based on KZ filtering**

386 Through KZ filtering, the adjusted time-series of PM_{2.5} concentrations with filtered
387 meteorological variations was acquired. Next, for each station, the actual PM_{2.5}
388 variations and adjusted PM_{2.5} variations without the disturbance of meteorological
389 variations from 2013 to 2017 were calculated respectively (as shown in Table 5).
390 Based on this, the relative contribution of emission-reduction and meteorological
391 conditions to PM_{2.5} reduction in Beijing from 2013 to 2017 can be quantified.

392 The original and KZ-processed time series of PM_{2.5} concentrations were illustrated
393 using one background station, one rural station and four urban stations (Fig 4). As
394 shown in Fig 4, most abrupt variations in the original time series of PM_{2.5}
395 concentrations have been smoothed through KZ filtering and the generally decreasing
396 trend of PM_{2.5} variations from 2013 to 2017 caused by anthropogenic emissions can
397 be clearly presented.

Table 5. Estimated relative contribution of emission-reduction and meteorological variations to PM_{2.5} reduction in Beijing from 2013 to 2017 using KZ filter

Stations	PM _{2.5} concentrations in 2013($\mu\text{g}\cdot\text{m}^{-3}$)	PM _{2.5} concentrations in 2017 ($\mu\text{g}\cdot\text{m}^{-3}$)	Adjusted PM _{2.5} concentrations in 2017($\mu\text{g}\cdot\text{m}^{-3}$)	PM _{2.5} Decrease rate ($\mu\text{g}\cdot\text{m}^{-3}\cdot\text{m}^{-1}$) ¹	Adjusted PM _{2.5} Decrease rate ($\mu\text{g}\cdot\text{m}^{-3}\cdot\text{m}^{-1}$) ²	Contribution of emission reduction (%) ³	Contribution of meteorological variations (%) ⁴
Yufa	111.1	69.7	74.6	-0.78	-0.63	80.4	19.7
Miyun Reservoir	58.8	44.8	47.0	-0.40	-0.33	82.8	17.2
Dingling	69.6	47.1	50.6	-0.54	-0.44	80.8	19.2
Qianmen	103.9	64.0	68.9	-0.81	-0.69	85.0	15.0
Olympic center	90.4	57.2	61.7	-0.68	-0.55	80.8	19.2
Xiangshan	77.0	59.3	60.3	-0.46	-0.39	83.9	16.1
Huayuan	101.5	64.4	69.2	-0.77	-0.63	81.9	18.1
Yungang	91.8	60.2	64.0	-0.69	-0.55	79.6	20.4
WanShouxigong	93.7	62.0	66.8	-0.64	-0.50	78.2	21.8
Dongsi	94.9	62.4	67.5	-0.62	-0.49	78.9	21.1
TianTan	92.3	58.4	64.6	-0.68	-0.55	80.2	19.9
NongZhanguan	92.2	59.9	65.9	-0.66	-0.53	80.3	19.8
Gucheng	92.7	61.4	65.9	-0.65	-0.50	77.6	22.4
Guanyuan	89.6	59.5	64.6	-0.60	-0.48	79.6	20.4
BeiBuxinqu	86.6	59.5	63.3	-0.60	-0.45	75.2	24.8
WanLiu	98.1	56.2	60.4	-0.87	-0.73	84.2	15.8

399 ¹ PM_{2.5} decrease rate: the fitted variation slope of original monthly average PM_{2.5} time series;

400 ² Adjusted PM_{2.5} decrease rate: the fitted variation slope of adjusted monthly average PM_{2.5} time series;

401 ³ Contribution of emission reduction = 1 - Contribution of meteorological variations;

402 ⁴ Contribution of meteorological variations = (PM_{2.5} decrease rate - Adjusted PM_{2.5} decrease rate) / PM_{2.5} decrease rate.

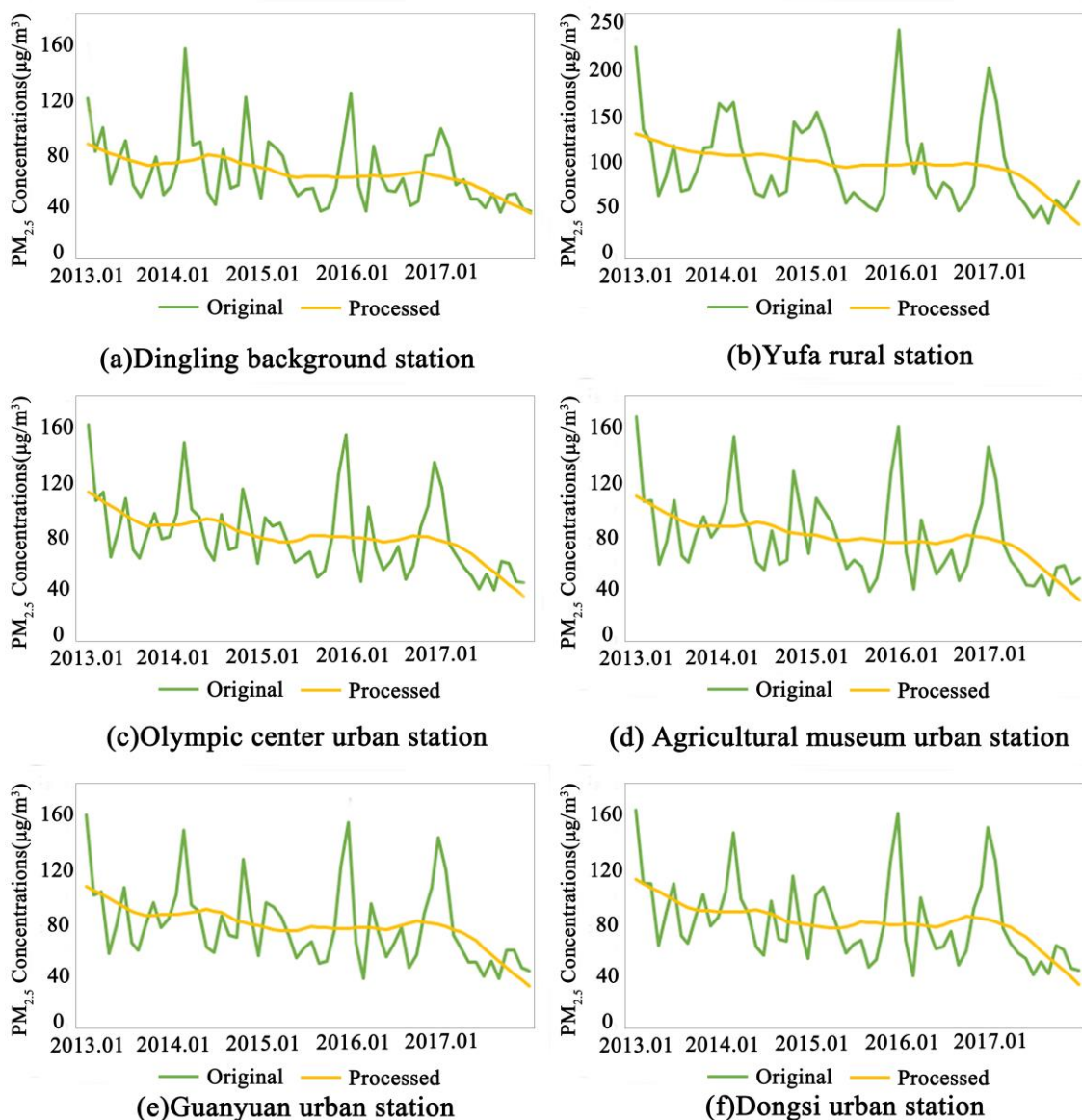


Fig 4. The comparison of original and KZ processed time series of PM_{2.5} concentrations in six stations from 2013 to 2017

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406 According to Table 5, the annual mean PM_{2.5} concentration in Beijing in 2017 was

407 35.6% lower than that in 2013. By filtering the influence of meteorological variations,

408 the adjusted annual mean PM_{2.5} concentration in Beijing in 2017 decreased by 31.7%

409 when compared to that in 2013, indicating that the variation in meteorological

410 conditions exerted a moderate influence on PM_{2.5} reduction from 2013 to 2017.

411 Meteorological conditions in Beijing were generally favorable for PM_{2.5} dispersion

412 during the five-year period, especially the latter half of 2017, when there was a high

413 frequency of strong northerly winds and much lower wintertime PM_{2.5} concentrations

414 than previous years.

415 For the winter of 2017, frequent windy weather and successive clean sky had a strong
416 influence on the reduction of PM_{2.5} concentrations in Beijing. This led to a hot debate
417 concerning whether the notable decrease in PM_{2.5} concentrations was mainly
418 attributed to the favorable meteorological conditions or emission-reduction. Table 5
419 suggests that the control of anthropogenic emissions contributed to 75.2%~85.0% of
420 PM_{2.5} decrease in the five-year period, indicating that emission-reduction worked
421 effectively in all rural, urban and background stations. On average, the relative
422 contribution of emission-reduction and meteorological variations to PM_{2.5} reduction
423 in Beijing from 2013 to 2017 was 80.6% and 19.4% respectively. Therefore, in spite
424 of more favorable meteorological conditions, properly designed and implemented
425 emission-reduction measures were the dominant driver for the remarkable decrease of
426 PM_{2.5} concentrations in Beijing from 2013 to 2017.

427 **4.1.2 Estimation based on WRF-CMAQ**

428 In addition to the KZ filter, we also employed WRF-CMAQ to estimate the relative
 429 contribution of emission-reduction and meteorological conditions to the decrease of
 430 PM_{2.5} concentrations in Beijing. The result is shown in Table 6.

431 **Table 6. Estimated relative contribution of emission-reduction and meteorological variations to**
 432 **PM_{2.5} reduction in Beijing from 2013 to 2017 using WRF-CMAQ**

Stations	Contribution of meteorological variations (%)	Contribution of emission-reduction(%)
Yufa	21.9	78.2
Miyun Reservoir	20.8	79.2
Dingling	21.7	78.3
Qianmen	21.2	78.8
Olympic center	21.2	78.8
Xiangshan	20.3	79.7
Huayuan	21.2	78.8
Yungang	21.2	78.8
WanShouxigong	21.2	78.8
Dongsi	21.2	78.8
TianTan	21.2	78.8
NongZhanguan	21.2	78.8
Gucheng	22.2	77.8
Guanyuan	21.2	78.8
BeiBuxinqu	22.2	77.8
WanLiu	22.2	77.8

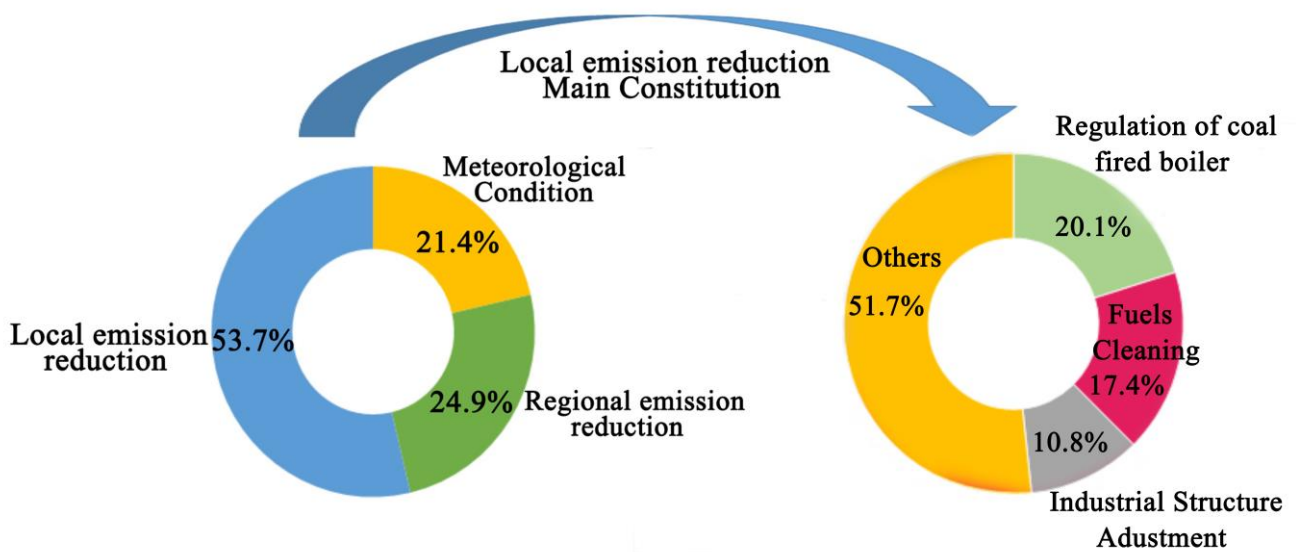
433 Based on WRF-CMAQ, the relative contribution of meteorological variations to the
 434 decrease in PM_{2.5} concentrations in Beijing ranged from 20.3% to 22.2% in different
 435 stations, whilst emission-reduction accounted for about four-fifths of PM_{2.5} reduction
 436 from 2013 to 2017. It is worth mentioning that WRF-CMAQ is a grid-based model
 437 and thus the calculated contribution of meteorological variations for some stations
 438 located in the same grid was the same. Instead, station-based KZ filtering led to more
 439 reliable analysis for each station and can better distinguish the differences between
 440 multiple stations. Furthermore, WRF-CMAQ simply considered the differences

441 between meteorological conditions in 2013 and 2017 without considering their
442 variations during the five-year period while the KZ filtering analyzed the entire time
443 series of PM_{2.5} and meteorological data from 2013 to 2017. The averaged relative
444 contribution of meteorological variations to PM_{2.5} reduction in Beijing calculated
445 using WRF-CMAQ was 21.4%, very similar to the 19.4% calculated using KZ
446 filtering. The slightly larger meteorological contribution calculated using
447 WRF-CMAQ might be attributed to that WRF-CMAQ simply considered the
448 favorable meteorological conditions in 2017 whilst KZ fully considered the long-term
449 meteorological variations from 2013 to 2017.

450 Since KZ filtering is fully based on observed data, and simply considers the influence
451 of time-series meteorology data on PM_{2.5} variations, less uncertainty is involved. The
452 accuracy of KZ filtering is influenced mainly by the variations of PM_{2.5}-meteorology
453 interactions in different areas and seasons. On the other hand, CTMs (e.g.
454 WRF-CMAQ or WRF-CAMx) consider both meteorological conditions (mainly
455 large-scale meteorological data for model simulation, not as accurate as local
456 observed meteorological data) and anthropogenic emissions for estimating PM_{2.5}
457 concentrations under different emission scenarios. The accuracy of these models are
458 not only decided by proper understanding of PM_{2.5}-meteorology interactions, but also
459 the reliability of emission inventories and proper descriptions of reaction mechanisms
460 for PM_{2.5} production, especially during heavy pollution episodes, which is a major
461 challenge for current model simulation. Consequently, KZ filtering provides a more
462 reliable method for researchers and decision makers to understand the relative
463 importance of emission-reduction and meteorological conditions in recent PM_{2.5}
464 reduction in Beijing. Meanwhile, similar outputs from WRF-CMAQ simulation
465 provide complementary evidence for the fact that anthropogenic emissions exerted a
466 much stronger influence on PM_{2.5} concentrations than meteorological conditions. In
467 addition to the combined effects of all emission-reduction measures, we further
468 employed WRF-CMAQ to quantify the relative contribution of different
469 emission-reduction measures to the decrease in PM_{2.5} concentrations in Beijing from
470 2013 to 2017.

471 **4.2 The relative contribution of different emission-reduction measures to the**
472 **decrease in PM_{2.5} concentrations in Beijing**

473 The observed annual average PM_{2.5} concentration in Beijing in 2017 was 58 mg/m³,
474 compared with 89.5 µg/m³ in 2013. Based on WRF-CMAQ simulation,
475 meteorological conditions contributed 6.7 µg/m³ whilst the control of anthropogenic
476 emissions contribute contributed 24.7 µg/m³ to the total PM_{2.5} reduction of 31.5 µg/m³
477 in Beijing from 2013 to 2017. Specifically, local and regional emission-reduction
478 accounted for 16.9 µg/m³ and 7.8 µg/m³ of PM_{2.5} reduction. Local emissions and
479 regional transport took up 68.4% and 31.6% of total anthropogenic emissions in
480 Beijing. This result is consistent with our recent study (Chen et al., 2019). Chen et al.
481 (2019) investigated four pollution episodes in Beijing in 2013, 2016, 2017 and 2018
482 respectively and found that local emissions accounted for 69.3%, 76.8%, 49.5% and
483 88.4% of total emissions in Beijing respectively. Except for the moderate pollution
484 episode in 2017, local emissions caused more than two thirds of anthropogenic
485 emissions in Beijing. Therefore, local emissions played a dominant role for PM_{2.5}
486 variations in Beijing in both long-term run and heavy pollution episodes. According to
487 three emission-reduction scenarios designed, the regulation of coal boilers had the
488 most significant effect on PM_{2.5} reduction in Beijing and resulted in a decrease of 6.3
489 µg/m³. Meanwhile, increasing clean fuels for residential use and industrial
490 restructuring also exerted strong influences on PM_{2.5} reduction and contributed to a
491 decrease of 5.5 µg/m³ and 3.4 µg/m³ respectively. The three major strategies
492 accounted for around half of the total effects of emission-reduction on PM_{2.5}
493 variations in Beijing.



495 **Fig 5. The relative contribution of different influencing factors to the decrease of**
 496 **PM_{2.5} concentrations in Beijing from 2013 to 2017**

497 **5 Discussion**

498 By the end of 2017, the Beijing Five-year Clean Air Action Plan (2013-2017) was
 499 completed and achieved its primary goal of reducing the annual average PM_{2.5}
 500 concentration to less than 60 µg/m³. Meanwhile, in November 2017, strong northerly
 501 winds in Beijing resulted in the cleanest winter in the past five years, raising
 502 arguments whether the favorable meteorological conditions were primarily
 503 responsible for PM_{2.5} reduction or whether the significant improvement in air quality
 504 in Beijing was mainly attributed to the control of anthropogenic emissions. In this
 505 case, a quantitative comparison between the influence of meteorological conditions
 506 and emission-reduction on PM_{2.5} reduction is necessary for comprehensively
 507 evaluating the Five-year Clean Air Action Plan. Based on two different approaches,
 508 this research revealed that the control of anthropogenic emissions contributed to
 509 around 80% of PM_{2.5} reductions in Beijing from 2013 to 2017, indicating that the
 510 Five-Year Clean Air Plan exerted a dominant influence on air quality enhancement in
 511 Beijing. The large contribution of some specific emission-reduction measures may be
 512 obscured in the presence of favorable meteorological conditions. For instance, many
 513 residents may attribute the clean winter of 2017 to the notable strong winds without
 514 noticing some of major emission-reduction strategies implemented during this period.

515 A large-scale replacement of coal boilers with gas boilers was conducted in Beijing
516 and its neighboring areas since 2013. As quantified by WRF-CMAQ, the regulation of
517 coal boilers and increasing use of clean fuels for residential use jointly contributed to
518 an $11.8\mu\text{g}/\text{m}^3$ decrease in $\text{PM}_{2.5}$ concentrations, much (almost twice) larger than the
519 $6.7\mu\text{g}/\text{m}^3$ decrease caused by favorable meteorological conditions. In general,
520 although favorable meteorological conditions (e.g., strong winds) may lead to an
521 instant improvement of air quality, regular emission-reduction measures exert a
522 reliable and consistent influence on the long-term reduction of $\text{PM}_{2.5}$ concentrations in
523 Beijing. Given the satisfactory performance of the Five-year Clean Air Action Plan in
524 $\text{PM}_{2.5}$ reduction, such long-term clean air plan should be further designed and
525 implemented in Beijing and other mega cities with heavy $\text{PM}_{2.5}$ pollution.

526 Recently, with growing attention to the completion of the Five-year Clean Air Action
527 Plan, some other studies have also been conducted to evaluate this five-year plan.
528 Cheng, J. et al. (2019) employed a finer-scale and more detailed local
529 emission-inventory and quantified the relative contribution of multiple
530 emission-reduction strategies, including the control of coal-fired boilers, increasing
531 use of clean fuels, optimization of industrial structure, fugitive dust control, vehicle
532 emission control, improved end-of-pipe control, and integrated treatment of VOCs.
533 The relative contribution of these emission-reduction measures to $\text{PM}_{2.5}$ reduction in
534 Beijing from 2013 to 2017 was 18.7%, 16.8%, 10.2%, 7.3%, 6.0%, 5.7% and 0.6%
535 respectively. By contrast, our research revealed that three major emission-reduction
536 measures (the regulation of coal-fired boiler, increasing use of clean fuels and
537 industrial restructuring) contributed 20.1%, 17.4% and 10.8% of total $\text{PM}_{2.5}$ reduction
538 in Beijing from 2013 to 2017, which was very close to Cheng, J et al. (2019)'s
539 findings. Based on finer-scale local emission-inventories with more field-collected
540 emission data, Cheng, J et al. (2019) provided a comprehensive and reliable
541 understanding of the effects of multiple emission-reduction measures on $\text{PM}_{2.5}$
542 reduction in Beijing. The similar outputs from the two studies further proved the
543 reliability of WRF-CMAQ simulation. Meanwhile, Cheng, J et al. (2019) and UNEP
544 (2019) jointly quantified that the total amount of reduction in SO_2 , NO_x , VOCs and
545 direct $\text{PM}_{2.5}$ induced by the control of anthropogenic emissions was 79420t, 93522t,
546 115752t and 44307t respectively, which was the major driver for the notable $\text{PM}_{2.5}$

547 reduction in Beijing from 2013 to 2017.

548 Although the “2+26” regional strategy for air quality improvement in Beijing has
549 become a hotly debated issue and growing emphasis has been placed on the proper
550 design and implementation of regional emission-reduction strategies in Beijing and its
551 surrounding cities, previous studies (Chen et al., 2019; Cheng, J. et al., 2019) and this
552 research proved that local emissions played a dominant role in affecting PM_{2.5}
553 concentrations in Beijing. Specifically, Chen et al. (2019) pointed out that with
554 intensive reduction of coal-fired boilers in Beijing-Tianjin-Hebei region, the relative
555 contribution of vehicle emissions to PM_{2.5} concentrations in Beijing, especially during
556 heavy pollution episodes, could be up to 50%. To further improve air quality in
557 Beijing, stricter regulations on local vehicle emissions, including contingent strategies
558 during pollution episodes (e.g. odd-even license plate policy) and long-term policies
559 (e.g. increasing availability of public transit systems and electric cars) should be a
560 major priority for the next stage clean-air actions.

561 Based on KZ filtering, Cheng, N et al. (2019) and Ma et al. (2016) suggested the
562 seasonal component contributed dominantly to O₃ variations in Beijing. By
563 comparison, this research revealed that the short-term component contributed
564 dominantly to PM_{2.5} variations in Beijing. These findings well explained the
565 phenomenon that ground ozone pollution in Beijing, controlled by seasonal variations
566 of emission and meteorological conditions (especially high-temperature and
567 low-humidity), simply occurred in summer, whilst PM_{2.5} pollution in Beijing,
568 controlled by short-term variations of meteorological and emission factors, might
569 occur in all seasons. Consequently, contingent emission-reduction measures during
570 heavy pollution episodes are an effective approach to offset the short-term
571 deterioration of meteorological conditions and improve local air quality.

572 Despite the major contribution of emission-reduction measures to PM_{2.5} reduction in
573 Beijing, meteorological influences, which contributed to 20% of PM_{2.5} reduction,
574 should also be considered balancedly. In addition to the control of anthropogenic
575 emissions, PM_{2.5} reduction may be realized through meteorological means. For the
576 winter of 2017, strong northwesterly winds led to instant improvement in air quality,
577 suggesting wind was a dominant meteorological factor for the accumulation or

578 dispersion of PM_{2.5} in Beijing. Meanwhile, previous studies (Chen et al., 2017)
579 suggested that increasing wind speeds led to increased evaporation, increased
580 sunshine duration (SSD) and reduced humidity, which further reduced local PM_{2.5}
581 concentrations. In other words, strong winds help reduce PM_{2.5} concentrations
582 through direct and indirect measures. In this light, the forthcoming Beijing
583 Wind-corridor Project, which includes five 500m-width corridors and more than ten
584 80m-width corridors to bring in stronger wintertime northwesterly winds, can be a
585 promising approach for promoting long-term favorable meteorological influences on
586 PM_{2.5} reduction in Beijing.

587 **6 Conclusions**

588 To comprehensively evaluate the effect of the Beijing Five-year Clean Air Action Plan
589 (2013-2017), we quantified the relative contribution of meteorological conditions and
590 the control of anthropogenic emissions to the notable decrease in PM_{2.5} concentrations
591 in Beijing from 2013 to 2017. Based on KZ filtering, we found that meteorological
592 conditions and emission-reduction accounted for 19.4% and 80.6% of the PM_{2.5}
593 reduction in Beijing, respectively. The large short-term component suggested that
594 short-term variations of meteorological and emission factors exerted a dominant
595 influence on the rapid variation of PM_{2.5} concentrations in Beijing. Meanwhile,
596 WRF-CAMQ revealed that meteorological conditions and emission-reduction
597 contributed to 21.4% and 78.6% of PM_{2.5} variations. Specifically, local and regional
598 emission-reduction measures contributed to 53.7% and 24.9% of PM_{2.5} reduction. For
599 three major emission-reduction measures, the regulation of coal boilers, increasing
600 use of clean fuels for residential use and industrial restructuring contributed to 20.1 %,
601 17.4% and 10.8% of PM_{2.5} reduction, respectively. Similar outputs from two models
602 suggested that the control of anthropogenic emissions contributed to around 80% of
603 the total decrease in PM_{2.5} concentrations in Beijing from 2013 to 2017, indicating
604 that the Five-year Clean Air Plan worked effectively and such long-term clean air plan
605 should be continued in the following years to further reduce PM_{2.5} concentrations in
606 Beijing.

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615 **Author contribution**

616 Chen, Z., Gao, B. and Xu, B designed this research. Chen, Z wrote this manuscript.
617 Chen, D., Zhuang, Y, Gao, B and Li, R. conducted data analysis. Chen, D and
618 Zhuang, Y. produced the figures. Kwan, M., and Chen, B helped revise this
619 manuscript.

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