

1       **The control of anthropogenic emissions contributed to 80% of the**  
2       **decrease in PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017**

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19      **Abstract**

20      With the completion of the Beijing Five-year Clean Air Action Plan by the end of  
21      2017, the annual mean PM<sub>2.5</sub> concentration in Beijing dropped dramatically to 58.0  
22      μg/m<sup>3</sup> in 2017 from 89.5 μg/m<sup>3</sup> in 2013. However, controversies exist to argue that  
23      favorable meteorological conditions in 2017 were the major driver for such rapid  
24      decrease in PM<sub>2.5</sub> concentrations. To comprehensively evaluate this five-year plan, we  
25      employed Kolmogorov-Zurbenko (KZ) filter and WRF-CMAQ to quantify the  
26      relative contribution of meteorological conditions and the control of anthropogenic  
27      emissions to PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017. For these five years, the  
28      relative contribution of emission-reduction to the decrease of PM<sub>2.5</sub> concentrations  
29      calculated by KZ filtering and WRF-CMAQ was 80.6% and 78.6% respectively. KZ  
30      filtering suggested that short-term variations of meteorological and emission  
31      conditions contributed majorly to rapid changes of PM<sub>2.5</sub> concentrations in Beijing.

32 WRF-CMAQ revealed that the relative contribution of local and regional  
33 emission-reduction to PM<sub>2.5</sub> decrease in Beijing was 53.7% and 24.9% respectively.  
34 For local emission-reduction measures, the regulation of coal boilers, increasing use  
35 of clean fuels for residential use and industrial restructuring contributed to 20.1 %,  
36 17.4% and 10.8% of PM<sub>2.5</sub> reduction respectively. Both models suggested that the  
37 control of anthropogenic emissions accounted for around 80% of the PM<sub>2.5</sub> reduction  
38 in Beijing, indicating that emission-reduction was crucial for air quality enhancement  
39 in Beijing from 2013 to 2017. Consequently, such long-term air quality clean plan  
40 should be continued in the following years to further reduce PM<sub>2.5</sub> concentrations in  
41 Beijing.

42 **Keywords:** PM<sub>2.5</sub>, anthropogenic emissions, meteorological conditions,  
43 **Kolmogorov-Zurbenko (KZ) filtering, WRF-CMAQ**

## 44 **1 Introduction**

45 In January 2013, persistent haze episodes occurred in Beijing, during which the highest  
46 hourly PM<sub>2.5</sub> concentration once reached 886  $\mu\text{g}/\text{m}^3$ , a historic high record.  
47 High-concentration PM<sub>2.5</sub> led to long-lasting black and thick fogs, which not only  
48 significantly influenced people's daily life (low-visibility induced traffic jam), but also posed  
49 a severe threat to public health (Brunekreef et al., 2002; Dominici et al., 2014; Nel et al.,  
50 2005; Zhang et al., 2012; Qiao et al., 2014). Since then, severe haze episodes have frequently  
51 been observed in Beijing and other regions across China (Chan et al., 2008; Huang, R., et al.,  
52 2014; Guo et al., 2014; Zheng et al., 2015), and PM<sub>2.5</sub> pollution has become one of the most  
53 concerned environmental issues in China. Consequently, a national network for monitoring  
54 hourly PM<sub>2.5</sub> concentrations has been established gradually, including 35 ground observation  
55 stations in Beijing, which provide important support for better understanding and managing  
56 PM<sub>2.5</sub> concentrations. To effectively mitigate PM<sub>2.5</sub> pollution, Beijing Municipal  
57 Government released "Beijing Five-year Clean Air Action Plan (2013-2017)" with a series of  
58 long-term emission-reduction measures, including shutting down heavily polluting factories,  
59 restricting traffic emissions and replacing coal fuels with clean energies, and "Heavy Air  
60 Pollution Contingency Plan" with a series of contingent emission-reduction measures during  
61 heavy pollution episodes. By the end of 2017, these long-term and contingent  
62 emission-reduction measures worked jointly to reduce the annually mean PM<sub>2.5</sub>  
63 concentration in Beijing from 89.5  $\mu\text{g}/\text{m}^3$  in 2013 to 58.0  $\mu\text{g}/\text{m}^3$  in 2017, indicating a great  
64 success of PM<sub>2.5</sub> management during the past five years. The notable decrease of PM<sub>2.5</sub>  
65 concentrations attracted nationwide attentions and growing studies have been conducted to  
66 understand spatio-temporal characteristics (Shao et al., 2018; Sun et al., 2019; Wang et al.,  
67 2019), sources (Chen et al., 2019; Xu et al., 2019; Cheng, J. et al., 2019) and health effects  
68 (Liang et al., 2019) of PM<sub>2.5</sub> variations in Beijing from 2013 to 2017. These studies revealed  
69 that air quality in Beijing was improved significantly in 2017 in terms of annual mean PM<sub>2.5</sub>  
70 concentrations, polluted days and pollution durations. Furthermore, despite different outputs,  
71 both source apportionment during pollution episodes based on collected samples (Shao et al.,  
72 2019; Xu et al., 2019; Chen et al., 2019) and long-term model simulation based on regional  
73 and local emission inventories (Cheng, J. et al., 2019) suggested that local and regional  
74 anthropogenic emissions (e.g. coal combustion and vehicle emissions) were the major  
75 influencing factors for long-term and short-term PM<sub>2.5</sub> variations in Beijing.

76 In addition to anthropogenic emissions, the strong meteorological influences on PM<sub>2.5</sub>  
77 concentrations in Beijing have been widely acknowledged (Zhao et al., 2013; Wang et al.,  
78 2014; UNEP, 2016; Cheng et al., 2017; Chen et al., 2017; Sun et al., 2019). For instance, for  
79 2014, more than 180 days in Beijing experienced a dramatic daily AQI (Air Quality Index)  
80 change ( $\Delta\text{AQI}>50$ ) (Chen, Z. et al., 2016). Considering that anthropogenic emissions for a  
81 mega city unlikely changed significantly on a daily basis, rapid variations of meteorological  
82 conditions were one major driver for the dramatic change of daily air quality in Beijing. In  
83 winter 2017, strong northwest winds led to favorable meteorological conditions for PM<sub>2.5</sub>  
84 diffusion and low PM<sub>2.5</sub> concentrations in Beijing. This raised the controversy that  
85 meteorological conditions, instead of emission-reduction, accounted for the remarkable  
86 PM<sub>2.5</sub> reduction in Beijing. In this case, with the completion of the five-year plan, it is highly  
87 necessary to quantify the relative contribution of meteorological conditions and  
88 emission-reduction to the notable decrease in PM<sub>2.5</sub> concentrations in Beijing from 2013 to  
89 2017.

90 In recent years, growing studies have been conducted to investigate meteorological and  
91 anthropogenic influences on long-term PM<sub>2.5</sub> variations. Based on Goddard Earth Observing  
92 System (GEOS) chemical transport model (GEOS-Chem), Yang et al (2016) revealed that  
93 the relative contribution of meteorological conditions to PM<sub>2.5</sub> variations in Eastern China  
94 from 1985 to 2005 was 12%. Based on a multiple general linear model (GLM), Gui et al.  
95 (2019) quantified that meteorological conditions accounted for 48% of PM<sub>2.5</sub> variations in  
96 Eastern China from 1998 to 2016. Based on a stepwise multiple linear regression (MLR)  
97 model, Zhai et al. (2019) quantified the relative contribution of meteorology to PM<sub>2.5</sub>  
98 variations from 2013 to 2018 in Beijing-Tianjin-Hebei region, Yangtze River Delta, Pearl  
99 River Delta and Sichuan Basin and Fenwei plain was 14%, 3%, 19%, 27% and 23%  
100 respectively. Through a two-stage hierarchical clustering method, Zhang et al. (2018)  
101 calculated that the relative contribution of meteorological conditions to heavy pollution  
102 episodes within the Beijing-Tianjin-Hebei region was larger than 50% from 2013 to 2017.  
103 These studies quantified the overall meteorological influences on long-term PM<sub>2.5</sub> variations  
104 using different statistical models and chemical transport models (CTMs). However, due to  
105 strong interactions between individual meteorological factors, traditional statistical methods  
106 such as correlation analysis and linear regression may be biased significantly when  
107 quantifying meteorological influences on PM<sub>2.5</sub> concentrations (Chen et al., 2017). On the

108 other hand, the accuracy of CTMs can be influenced largely by the uncertainty in emission  
109 inventories (Xu et al., 2016) and deficiency of heterogeneous/aqueous processes (Li et al.,  
110 2011). Therefore, multiple advanced models should be comprehensively considered to better  
111 quantify meteorological influences on PM<sub>2.5</sub> concentrations (Pearce et al., 2011).

112 To evaluate this five-year clean-air plan, we employ an advanced statistical model,  
113 Kolmogorov-Zurbenko (KZ) filtering, which is advantageous of filtering meteorological  
114 influences on long-term time series of airborne pollutants, and a CTM model, WRF-CMAQ,  
115 which is advantageous of quantifying the relative contribution of different emission sources,  
116 to comprehensively investigate the relative contribution of meteorological conditions and  
117 emission-reduction to PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017 respectively. In this light,  
118 this research provides important insight for better designing and implementing successive  
119 clean air plans in the future to further mitigate PM<sub>2.5</sub> pollution in Beijing.

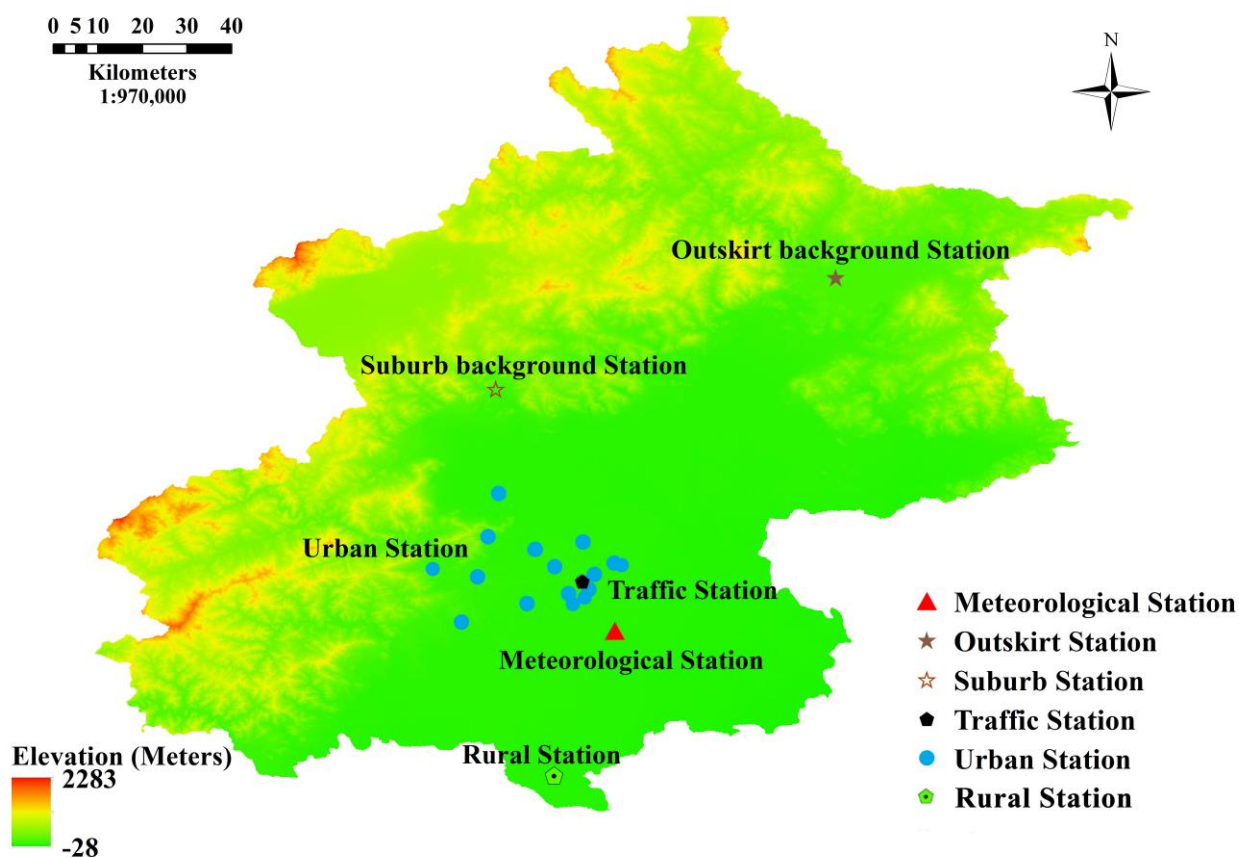
120 This manuscript is structured as follows: Firstly, major data sources, including PM<sub>2.5</sub> and  
121 meteorological data, and emission inventories, employed for this research are briefly  
122 introduced. Secondly, the principle and parameter setting of two models, KZ filtering and  
123 WRF-CMAQ, and model verification are explained. In the result section, the relative  
124 contribution of meteorological conditions and anthropogenic emissions to PM<sub>2.5</sub> variations in  
125 Beijing from 2013 to 2017 calculated using both models is presented. In the discussion and  
126 conclusion part, implementations of this research and suggestions for further improving air  
127 quality in Beijing are given.

## 128 **2 Data Sources**

### 129 **2.1 PM<sub>2.5</sub> and meteorological data**

130 In this study, hourly PM<sub>2.5</sub> concentration data were acquired from the website PM25.in  
131 ([www.PM25.in](http://www.PM25.in)), which collects official data provided by China National Environmental  
132 Monitoring Center (CNEMC). Beijing has established an advanced air quality monitoring  
133 network with 35 ground stations across the city. Considering the major contribution of  
134 industry and traffic-induced emissions in urban areas, we selected all twelve urban stations  
135 to analyze spatio-temporal variations of PM<sub>2.5</sub> concentrations and quantify their influencing  
136 factors. In addition to these urban stations, we selected two background stations, the

137 DingLing Station located in the suburb and the MiYun Reservoir Station located in the outer  
138 suburb, one transportation station (the Qianmen station) located close to a main road, and  
139 one rural station (the Yufa Station) that is far away from central Beijing for the following  
140 analysis. The DingLing and MiYun Reservoir Station were chosen as background stations by  
141 the Ministry of Environmental Protection of China. These two stations receive limited  
142 influence from anthropogenic emissions due to their location in suburban and outer suburban  
143 areas. The Qianmen transportation station received more influences from vehicle emissions.  
144 Long-term variations of PM<sub>2.5</sub> concentrations in different type of stations provide a useful  
145 reference for comprehensively understanding the effects of emission-reduction measures on  
146 PM<sub>2.5</sub> decrease in Beijing from 2013 to 2017. Meteorological data for this research were  
147 collected from the Guanxiangtai Station (GXT,54511, 116.46° E, 39.80° N), Beijing and  
148 downloaded from the Department of Atmospheric Science, College of Engineering,  
149 University of Wyoming (<http://weather.uwyo.edu/upperair/sounding.html>). Both PM<sub>2.5</sub> and  
150 meteorological data were collected from January 1<sup>st</sup>, 2013 to December 31<sup>st</sup>, 2017. The  
151 locations of these selected stations are shown in Fig 1.



152

153

**Fig 1. Locations of different ground monitoring stations.**

## 154 2.2 Emission inventories

155 For this research, we employed both regional and local emission inventories for running  
156 model simulation. Multi-resolution Emission Inventory for China, MEIC,  
157 (<http://meicmodel.org/>) provided by Tsinghua University, were employed as the regional  
158 emission inventories. MEIC has been widely employed and verified as a reliable emission  
159 inventory by a diversity of studies (Hong et al., 2017; Saikawa et al., 2017; Zhou et al., 2017;  
160 etc.). For simulating five-year PM<sub>2.5</sub> concentrations, MEIC from 2013 to 2017 are required.  
161 Since official MEIC 2017 has not been available yet, we employed a strategy from previous  
162 studies (Chen et al., 2019; etc) and updated MEIC 2016 for simulating emission-reduction  
163 scenarios and PM<sub>2.5</sub> concentrations in 2017 by considering official 2017 emission-reduction  
164 plans (e.g. the target of coal combustion reduction) required by the local government.

165 Different from regional emission inventories, local emission inventories are usually  
166 produced independently by local institutions. The Beijing local-emission inventory  
167 employed for this research was produced and updated by Beijing Municipal Research  
168 Institute of Environmental protection, fully according to the requirement of MEP on the  
169 production of local emission inventories within Beijing-Tianjin-Hebei region. This Beijing  
170 local-emission inventory from 2013 to 2017 was produced by synthesizing local  
171 environmental statistical data and reported emission data, carrying out field investigations  
172 and conducting a series of estimation according to Beijing Five-year Clean Air Action Plan.  
173 As shown in table 1, it is highly consistent with other official statistical data, such as the  
174 Annual report from National Environmental Statistics Bulletin  
175 ([http://www.mee.gov.cn/gzfw\\_13107/hjtj/qghjtjgb/](http://www.mee.gov.cn/gzfw_13107/hjtj/qghjtjgb/)) and “2+26” Center for Air Pollution Prevention  
176 and Control, and has been formally employed for the implementation of recent “2017 Air  
177 Pollution Prevention and Management Plan for the Beijing-Tianjin-Hebei Region and its  
178 Surrounding Areas” (MEP, 2017).

179

180 **Table 1. The comparison of local environmental statistical data used for this research**  
 181 **and other official statistical data in 2017 (unit: 10k tons)**

	SO <sub>2</sub>	NO <sub>x</sub>	CO	VOC	NH <sub>3</sub>	PM <sub>10</sub>	PM <sub>2.5</sub>	BC	OC
<b>Statistical data for this research</b>	1.38	10.15	49.54	13.47	3.20	14.74	3.92	0.17	0.44
<b>National Environmental Statistics Bulletin</b>	1.38	12.16	52.03	24.24	3.26	14.68	3.91	0.22	0.41
<b>“2+26” center for air pollution prevention and control</b>	0.89	9.24	48.98	13.93	3.16	13.82	3.72	0.19	0.46

## 182 **3 Methods**

183 A key step for quantifying the relative contribution of anthropogenic emissions to PM<sub>2.5</sub>  
 184 variations is to properly filter meteorological influences on PM<sub>2.5</sub> concentrations, which is  
 185 highly challenging and rarely investigated by previous studies. Therefore, we employed both  
 186 a statistical method and a CTM to comprehensively evaluate the role of anthropogenic  
 187 emissions and meteorological conditions in the decrease of PM<sub>2.5</sub> concentrations in Beijing  
 188 from 2013 to 2017.

### 189 **3.1 Kolmogorov-Zurbenko (KZ) filtering**

190 Since meteorological conditions exert a strong influence on PM<sub>2.5</sub> concentrations in Beijing,  
 191 the removal of seasonal signals from time series of meteorological factors produces data sets  
 192 suitable for understanding the trend of PM<sub>2.5</sub> concentrations mainly influenced by  
 193 anthropogenic factors (Eskridge et al., 1997). To better analyze the trend of time series data  
 194 without the disturbances from other major influencing variables, a statistical method  
 195 Kolmogorov-Zurbenko (KZ) filtering was proposed by Rao et al. (1994). The KZ filter is  
 196 advantageous of removing high-frequency variations in data sets through iterative moving  
 197 average. Eskridge et al. (1997) compared four major approaches for trend detection,  
 198 including PEST, anomalies, wavelet transform, and the KZ filter, and suggested that KZ  
 199 achieved higher confidence in detecting long-term trend than other models. Due to its  
 200 reliable performance in trend detection in complicated ecosystems, the KZ filter has been  
 201 increasingly employed to remove seasonal signals of meteorological conditions and extract  
 202 long-term trend of airborne pollutants (Zurbenko, et al., 1996; Eskridge, et al., 1997; Kang,



203 et al., 2013; Ma et al., 2016; Cheng, N et al., 2019). One potential limitation of the KZ filter  
204 is that iterative moving average ( $m$ ) may impose an influence on detecting abrupt variations.  
205 Therefore, Zurbenko et al. (1996) proposed an enhanced KZ filter that employed a dynamic  
206 variable  $m$  that decreased with the increase in changing rate. For this research, we employed  
207 this dynamic  $m$  to produce an adjusted time-series of PM<sub>2.5</sub> concentrations in Beijing by  
208 removing large inter-annual and seasonal variations in meteorological conditions. The  
209 principle of the KZ filter is briefly introduced as follows.

210 The raw time-series of airborne pollutants can be decomposed as:

$$211 \quad X(t) = E(t) + S(t) + W(t) \quad (1)$$

$$212 \quad X_b(t) = E(t) + S(t) \quad (2)$$

$$213 \quad E(t) = KZ_{365,3}(X) \quad (3)$$

$$214 \quad S(t) = KZ_{15,5}(X) - KZ_{365,3}(X) \quad (4)$$

$$215 \quad W(t) = X(t) - KZ_{15,5}(X) \quad (5)$$

216 **Where  $X(t)$  is the original time series of airborne pollutants,  $E(t)$  is the long-term trend component,**  
217  **$S(t)$  is the seasonal component,  $W(t)$  is the short-term (synoptic-scale) component or residue.  $KZ_{i,j}(X)$**   
218 **indicates KZ filtering on the original dataset  $X$  with a moving window size of  $i$  and  $j$  iterations.**

219  $X_b(t)$  stands for the base component, the sum of the long-term and seasonal component,  
220 presenting steady trend variation.  $E(t)$  is mainly affected by long-term anthropogenic  
221 emission and climate change.  $S(t)$  is mainly influenced by the seasonal variation of emission  
222 and meteorological conditions.  $W(t)$  is caused by short-term and small-scale shifts of  
223 emissions and meteorological conditions.

224 The long-term trend component  $E(t)$  processed by KZ filtering still contains the influence of  
225 meteorological conditions, which can be removed by multiple regression models. Multiple  
226 linear relationships are established for the residue and baseline component respectively using  
227 meteorological factors strongly correlated with airborne pollutants.

228 We examined correlations between seasonal PM<sub>2.5</sub> concentrations in Beijing and a series of  
229 meteorological factors, including temperature, wind speed, wind direction, precipitation,  
230 relative humidity, solar radiation, evaporation and air pressure. Due to limited space, detailed  
231 correlations between PM<sub>2.5</sub> concentrations and individual meteorological factors in Beijing

232 are not presented here and readers can refer to previous studies for more information (Chen  
 233 et al., 2017; 2018). The correlation analysis revealed that wind speed, relative humidity,  
 234 temperature and solar radiation were strongly and significantly correlated with PM<sub>2.5</sub>  
 235 concentrations in Beijing (as shown in Table 2), which was consistent with findings from  
 236 other studies (Sun et al., 2013; Wang et al., 2018).

237 **Table 2. Major meteorological factors strongly correlated with seasonal PM<sub>2.5</sub>**  
 238 **concentrations in Beijing (Chen et al., 2017)**

Spring	Summer	Autumn	Winter
	RHU**(0.648)	RHU**(0.587)	RHU**(0.738)
RHU**(0.532)	SSD**(-0.447)	SSD**(-0.509)	SSD**(-0.715)
	TEM**(0.554)	WIN**(-0.468)	WIN**(-0.558)

239 **\*\*Correlation is significant at the 0.01 level (2 tailed);**

240 **RHU: Relative humidity; SSD: Sunshine Duration; TEM: Temperature; WIN: Wind speed**

241 Therefore, we further established multiple linear regression equations between PM<sub>2.5</sub>  
 242 concentrations and wind speed, relative humidity, temperature and solar radiation as follows.

$$243 \quad W(t) = \alpha_0 + \sum \alpha_i w_i(t) + \varepsilon_w(t) \quad (6)$$

$$244 \quad X_b(t) = b_0 + \sum b_i x_i(t) + \varepsilon_b(t) \quad (7)$$

$$245 \quad \varepsilon(t) = \varepsilon_w(t) + \varepsilon_b(t) \quad (8)$$

246 **Where  $w_i(t)$  and  $x_i(t)$  stand for the different short-term and baseline component of the  $i^{\text{th}}$**   
 247 **meteorological factor.  $\varepsilon_w$  and  $\varepsilon_b$  is the regression residue of the short-term and baseline**  
 248 **component.  $\varepsilon(t)$  indicates the total residue, including the short-term influence of local emission**  
 249 **and meteorological factors neglected during the regression process and other noises.**

250 Next, KZ filtering was conducted on the  $\varepsilon(t)$  for its long-term component  $\varepsilon_E(t)$ . After the  
 251 variation of meteorological influences was filtered, the reconstructed time series of airborne  
 252 pollutants  $X_{LT}(t)$  was calculated as the sum of  $\varepsilon_E(t)$  and the average value of  $E(t)$ ,  $\overline{E(t)}$ .

$$253 \quad X_{LT}(t) = \overline{E(t)} + \varepsilon_E(t) \quad (9)$$

254 After KZ filtering, the relative contribution of meteorological conditions to PM<sub>2.5</sub> variations  
 255 can be calculated as follows:

256 
$$P_{contrib} = \frac{K_{org} - K}{K_{org}} \times 100\% \quad (10)$$

257 **Where  $P_{contrib}$  is the relative contribution of meteorological conditions to  $PM_{2.5}$  variations in Beijing,**  
258  **$K_{org}$  is the variation slope of the original  $PM_{2.5}$  time series;  $K$  is the variation slope of adjusted  $PM_{2.5}$**   
259 **time series with filtered influences from meteorological variations.**

### 260 **3.2 WRF-CMAQ model**

261 We employed WRF-CMAQ for simulating the effects of emission-reduction on the decrease  
262 of  $PM_{2.5}$  concentrations. WRF-CMAQ includes three models: The middle-scale meteorology  
263 model (WRF), the source emission model (SMOKE) (<http://www.cmascenter.org/smoke/>)  
264 and the community multiscale air quality modeling system (CMAQ)  
265 (<http://www.cmascenter.org/CMAQ>). The center of the CMAQ was set at coordinate 35°N,  
266 110°E and a bi-directional nested technology was employed, producing two layers of grids  
267 with a horizontal resolution of 36 km and 12 km respectively. The first layer of grids with  
268 36km resolution and 200×160 cells covered most areas in East Asia (including China, Japan,  
269 North Korea, South Korea, and other countries). The second layer of grids with 12km  
270 resolution and 120×102 cells covered the North China Plain (including the  
271 Beijing-Tianjin-Hebei region, Shandong and Henan Province). The vertical layer was  
272 divided into 20 unequal layers, eight of which were of a less-than-1km distance to the  
273 ground for better featuring the structure of atmospheric boundary. The height of the ground  
274 layer was 35m.

275 We employed ARW-WRF3.2 to simulate the meteorological field. The setting of the center  
276 and the bidirectional nest for WRF and CMAQ was similar. There were 35 vertical layers for  
277 WRF and the outer layer provided boundary conditions of the inner layer. The  
278 meteorological background field and boundary information with a FNL resolution of 1°×1°  
279 and temporal resolution of 6h were acquired from NCAR (National Center for Atmospheric  
280 Research, <https://ncar.ucar.edu/>) and NCEP (National Centers for Environmental Prediction)  
281 respectively. The terrain and underlying surface information was obtained from the USGS  
282 30s global DEM (<https://earthquake.usgs.gov/>). The outputs from WRF were interpolated to  
283 the region and grid of CMAQ using the Meteorology-Chemistry Interface Processor (MCIP,  
284 <https://www.cmascenter.org/mcip>). The meteorological factors used for this model included  
285 temperature, air pressure, humidity, geopotential height, zonal wind, meridional wind,  
286 precipitation, boundary layer heights and so forth. An estimation model for terrestrial

287 ecosystem MEGAN (<http://ab.inf.uni-tuebingen.de/software/megan/>) was employed to  
288 process the natural emissions. Multi-resolution Emission Inventory for China, MEIC  
289 0.5°×0.5° emission inventory (<http://www.meicmodel.org/>) and Beijing emission inventory  
290 (<http://www.cee.cn/>) provided anthropogenic emission data. We input the processed natural  
291 and anthropogenic emission data into the SMOKE model and acquired comprehensive  
292 emission source files.

293 Scenario simulation is employed to estimate the contribution of emission-reduction to the  
294 variation of PM<sub>2.5</sub> concentrations.

$$295 \quad P_{contrib} = \frac{C - C_{base}}{C} \times 100\% \quad (11)$$

296 **Where  $P_{contrib}$ ,  $C$  and  $C_{base}$  are the contribution rate of emission-reduction to PM<sub>2.5</sub>**  
297 **concentrations, simulated PM<sub>2.5</sub> concentrations under the emission-reduction scenario, and**  
298 **simulated PM<sub>2.5</sub> concentrations in the baseline scenario respectively.**

299 To evaluate the relative contribution of meteorological conditions and different  
300 emission-reduction measures to the decrease of PM<sub>2.5</sub> concentrations, we designed two  
301 baseline experiments and four sensitivity experiments. For the first baseline experiment, we  
302 employed the actual meteorological data in 2013. For the second baseline experiment, we  
303 employed the actual meteorological data in 2017 and emission inventory in 2017. Since no  
304 emission-reduction measures were conducted in 2013, the first baseline experiment was used  
305 to estimate the relative contribution of meteorological conditions to the variation of PM<sub>2.5</sub>  
306 concentrations. By comparing the first and second baseline experiment, the relative  
307 contribution of all emission-reduction measures to the variation of PM<sub>2.5</sub> concentrations can  
308 be quantified. For the first sensitivity experiment, we employed the actual meteorological  
309 conditions in 2013 and emission inventory in 2017 and compared the simulation result with  
310 the baseline experiment, which demonstrated the relative contribution of meteorological  
311 concentrations to PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017. Since the WRF-CMAQ  
312 simulation simply considers PM<sub>2.5</sub> concentrations and meteorological conditions in 2013 and  
313 2017 without considering their variation process from 2013 to 2017, KZ filtering may  
314 perform better in quantifying the relative contribution of meteorological variations to PM<sub>2.5</sub>  
315 reduction in Beijing. However, the output from this sensitivity experiment serves as a useful  
316 reference for cross-verifying the output from the KZ filtering. For the remaining three  
317 sensitivity-simulation experiments, we added the reduced emission amount induced by one

318 specific emission-reduction measure to the actual emission amount in 2017 and kept other  
319 parameters unchanged, and thus quantified the relative contribution of one specific  
320 emission-reduction measure to PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017. Consequently,  
321 we quantified the relative contribution of three major emission-reduction measures to PM<sub>2.5</sub>  
322 reduction in Beijing (Table 3).

323

**Table 3. The design and materials for two baseline and four sensitivity experiments using WRF-CMAQ**

<b>ID</b>	<b>Meteorological Data</b>	<b>Emission-reduction measures</b>	<b>Simulation Year</b>	<b>Major purposes</b>
<b>Baseline Experiment1</b>	2013	No emission-reduction Measures	2013	<b>2013 baseline scenario</b>
<b>Baseline Experiment2</b>	2017	All emission-reduction Measures	2017	<b>2017 baseline scenario</b>
<b>Sensitivity Experiment 1</b>	2013	All emission-reduction Measures	2017	The relative contribution of meteorological variations to the decrease of PM <sub>2.5</sub> concentrations in Beijing from 2013 to 2017
<b>Sensitivity Experiment 2</b>	2017	All emission-reduction measures except for industrial restructuring	2017	The relative contribution of industrial restructuring to the decrease of PM <sub>2.5</sub> concentrations in Beijing from 2013 to 2017
<b>Sensitivity Experiment 3</b>	2017	All emission-reduction measures except for the regulation of coal boilers	2017	The relative contribution of the regulation of coal boilers to the decrease of PM <sub>2.5</sub> concentrations in Beijing from 2013 to 2017
<b>Sensitivity Experiment 4</b>	2017	All emission-reduction measures except for increasing clean fuels for civil use	2017	The relative contribution of increasing clean fuels for civil use to the decrease of PM <sub>2.5</sub> concentrations in Beijing from 2013 to 2017

324

For emission data, all experiments employed Beijing local emissions inventory in 2017 for Beijing and regional emission inventory in 2017 for other regions.

325

MEIC 2017 was acquired based on our update of MEIC 2016 according to official 2017 emission-reduction targets required by the local government.

### 326 **3.3 Model verification**

#### 327 **3.3.1 Verification of KZ filtering**

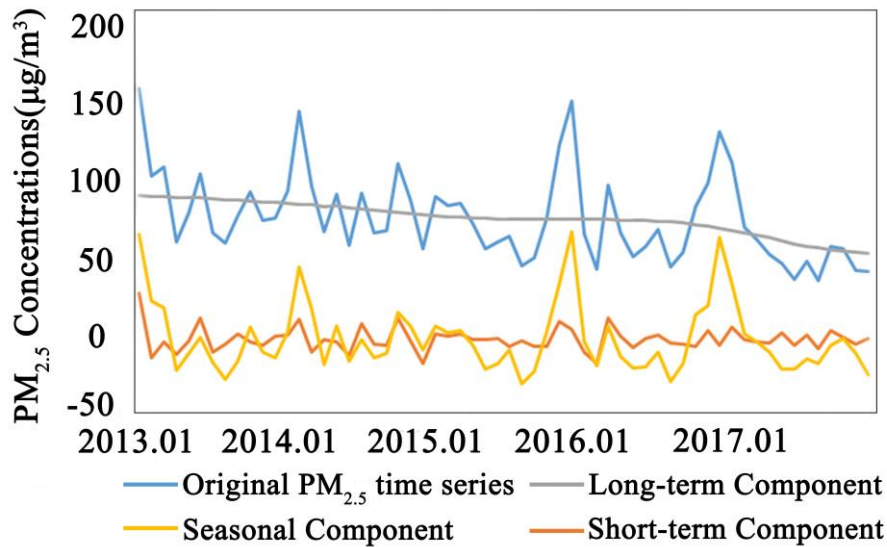
328 For each station, the original time series of PM<sub>2.5</sub> data was processed by the KZ filter  
329 and the relative contribution of the long-term, seasonal and short-term component to  
330 the total variance is shown as Table 4. The sum of the long-term, seasonal and  
331 short-term component contributed to more than 93.6~95.3% of the total variance in  
332 different stations respectively. The larger the total variance, the three components are  
333 more independent to each other. The total variance close to 100% suggests that a  
334 majority of meteorological influences has been considered and effectively removed.  
335 As shown in Table 4, the large value of the total variation in all stations indicated a  
336 satisfactory output from the KZ filtering.

337 Specifically, the relative contribution of the seasonal component (ranging from  
338 9%-23.8%) and short-term component (ranging from 66.8%-83.8%) was much larger  
339 than that of the long-term component (ranging from 1.2%-3.5%), suggesting that  
340 seasonal and short-term variations of meteorological and emission factors exerted a  
341 major influence on the rapid change of PM<sub>2.5</sub> concentrations in Beijing. The  
342 decomposed long-term, seasonal and short-term component from the original time  
343 series of mean urban PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017 are  
344 demonstrated as Fig 2. According to Fig 2, the notable peaks of decomposed seasonal  
345 and short-term component were highly consistent with the peaks of PM<sub>2.5</sub>  
346 concentrations in the original time-series, which further proved the dominant  
347 influence of seasonal and short-term variations of meteorological and anthropogenic  
348 factors on the temporal changes of PM<sub>2.5</sub> concentrations in Beijing.

349 **Table 4. The relative contribution of different components to the total variance of**  
 350 **original time series of PM<sub>2.5</sub> concentrations from 2013-2017 at different stations**

<b>Stations</b>	<b>Long-term component (%)</b>	<b>Seasonal component (%)</b>	<b>Short-term component (%)</b>	<b>Total variance(%)</b>
<b>Yufa</b>	2.1	23.8	66.8	94.0
<b>Miyun Reservoir</b>	1.4	9.0	83.8	95.2
<b>Dingling</b>	1.6	11.0	81.3	94.9
<b>Qianmen</b>	2.7	12.7	78.5	95.1
<b>Olympic center</b>	2.1	11.9	80.0	95.3
<b>Xiangshan</b>	1.2	10.3	83.4	94.9
<b>Huayuan</b>	2.2	15.9	75.6	93.7
<b>Yungang</b>	2.1	15.1	76.5	93.6
<b>WanShouxigong</b>	1.6	14.2	78.2	94.0
<b>Dongsi</b>	1.6	12.3	80.0	94.0
<b>TianTan</b>	2.1	13.2	78.6	93.8
<b>NongZhanguan</b>	1.8	13.7	78.6	94.1
<b>Gucheng</b>	1.8	13.5	78.5	93.7
<b>Guanyuan</b>	1.6	12.6	79.8	94.0
<b>BeiBuxinqu</b>	1.7	13.8	78.4	93.9
<b>WanLiu</b>	3.5	11.9	78.2	93.6





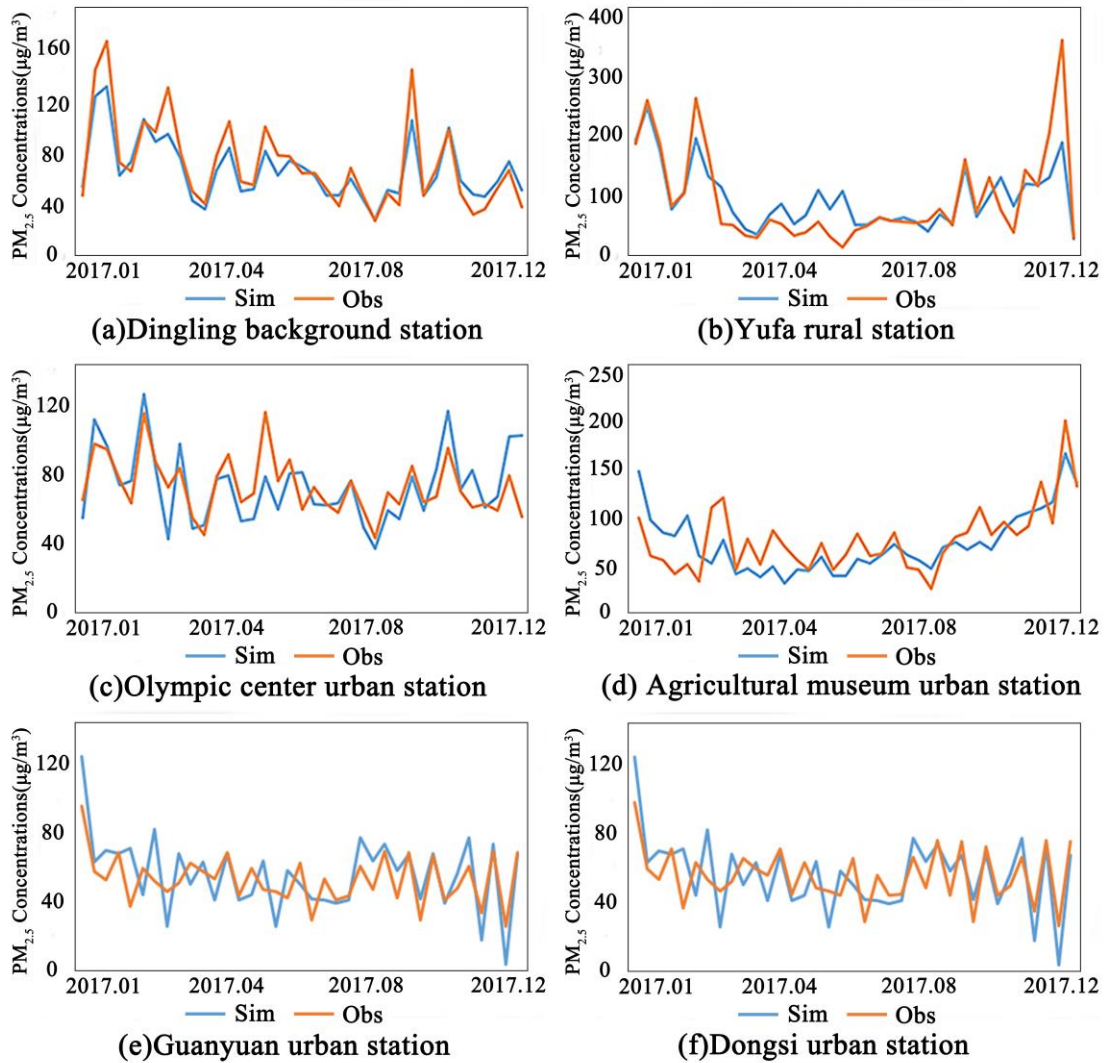
351

352 **Fig 2. The long-term, seasonal and short-term component extracted from the original**  
 353 **time series of mean urban PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017**

354 **3.3.2 Verification of WRF-CMAQ**

355 We employed the emission inventory and meteorological data for 2017 to verify the  
 356 accuracy of WRF-CMAQ simulation. For six stations of different types (DingLing  
 357 background station, Yufa rural station, Olympic Center urban station, Guanyuan urban  
 358 station, Dongsu urban station and Agricultural museum urban station), we compared  
 359 the observed and estimated PM<sub>2.5</sub> concentrations and presented the comparison result  
 360 as Fig 3. According to Fig 3, the general trend of the simulated PM<sub>2.5</sub> concentrations  
 361 was consistent with that of the observed PM<sub>2.5</sub> concentrations. For six stations, the  
 362 correlation coefficient R, normalized mean bias (NMB), normalized mean error  
 363 (NME), mean fractional bias (MFB) and mean fractional error (MFE) between  
 364 observed and simulated data was 0.63~0.91, -6%~6%, 26%~40%, -5%~7%, and  
 365 27%~46% respectively, indicating a satisfactory simulation output (EPA, 2005;  
 366 Boylan et al., 2006). However, as shown in Figure 3, WRF-CMAQ may notably  
 367 underestimate PM<sub>2.5</sub> concentrations during heavy pollution episodes due to unified  
 368 parameter setting for long-term simulation, the uncertainty in emission inventories,  
 369 and especially insufficient chemical reaction mechanisms, which is a common  
 370 challenge for CTM-based PM<sub>2.5</sub> simulation (Li et al., 2011). For instance, without  
 371 considering heterogeneous/aqueous reactions between multiple precursors, CTMs  
 372 failed to approach the maximum PM<sub>2.5</sub> concentrations during severe haze episodes

373 and the simulation accuracy was dramatically improved by including proper  
 374 descriptions of heterogeneous/aqueous reactions into CTMs (Chen, D. et al. 2016).  
 375 With more finer-scale emission inventories and better descriptions of reaction  
 376 mechanisms between precursors, the accuracy of PM<sub>2.5</sub> simulation can be improved  
 377 significantly.



378

379 **Fig 3. The comparison between observed and WRF-CMAQ simulated PM<sub>2.5</sub>**  
 380 **concentrations in 2017 in six stations across Beijing**

381 **4 Results**

382 **4.1 The relative contribution of emission-reduction and meteorological variations**  
383 **to the decrease of PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017**

384 **4.1.1 Estimation based on KZ filtering**

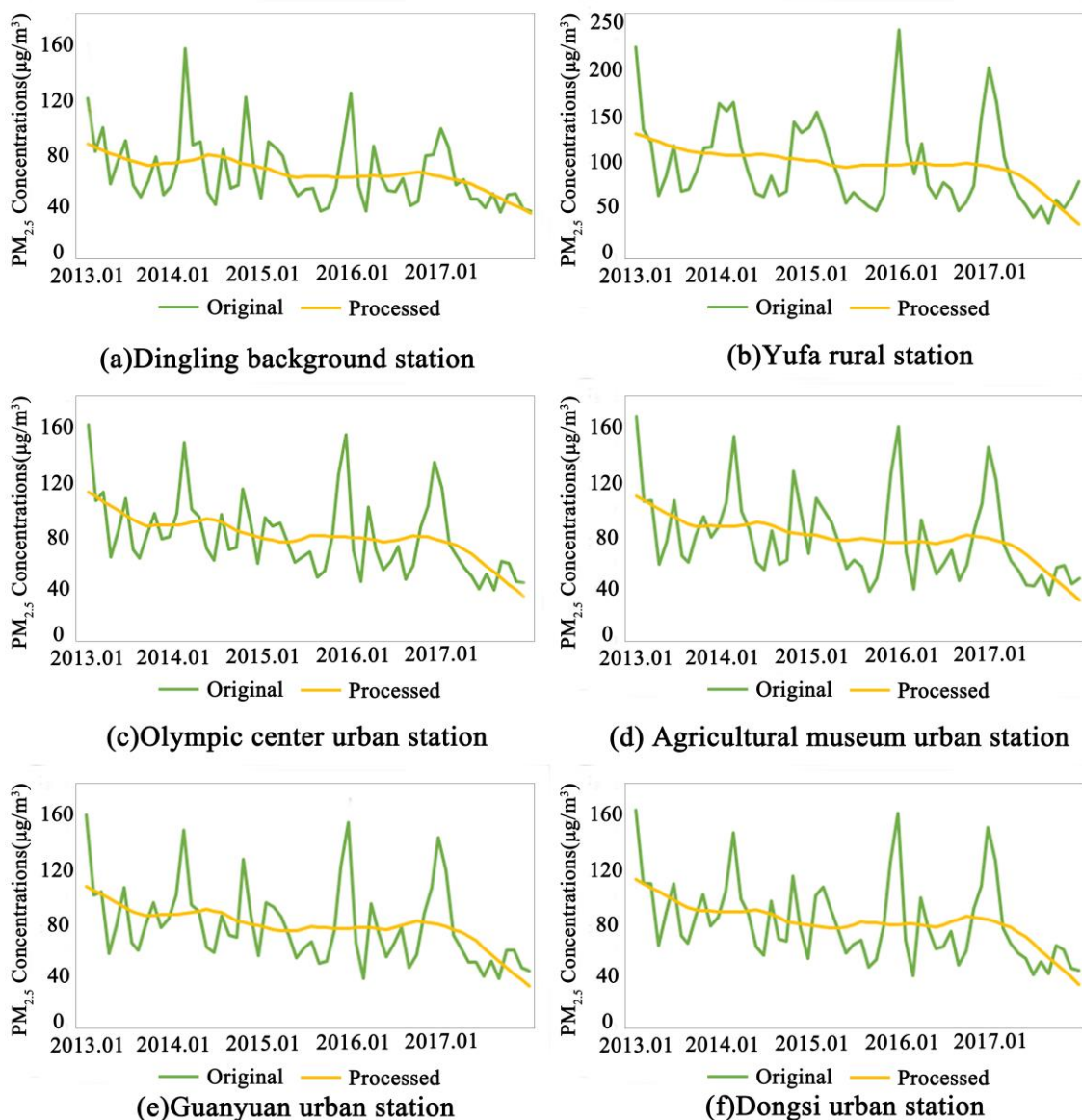
385 Through KZ filtering, the adjusted time-series of PM<sub>2.5</sub> concentrations with filtered  
386 meteorological variations was acquired. Next, for each station, the actual PM<sub>2.5</sub>  
387 variations and adjusted PM<sub>2.5</sub> variations without the disturbance of meteorological  
388 variations from 2013 to 2017 were calculated respectively (as shown in Table 5).  
389 Based on this, the relative contribution of emission-reduction and meteorological  
390 conditions to PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017 can be quantified.

391 The original and KZ-processed time series of PM<sub>2.5</sub> concentrations were illustrated  
392 using one background station, one rural station and four urban stations (Fig 4). As  
393 shown in Fig 4, most abrupt variations in the original time series of PM<sub>2.5</sub>  
394 concentrations have been smoothed through KZ filtering and the generally decreasing  
395 trend of PM<sub>2.5</sub> variations from 2013 to 2017 caused by anthropogenic emissions can  
396 be clearly presented.

Table 5. Estimated relative contribution of emission-reduction and meteorological variations to PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017 using KZ filter

Stations	PM <sub>2.5</sub> concentrations in 2013( $\mu\text{g}\cdot\text{m}^{-3}$ )	PM <sub>2.5</sub> concentrations in 2017 ( $\mu\text{g}\cdot\text{m}^{-3}$ )	Adjusted PM <sub>2.5</sub> concentrations in 2017( $\mu\text{g}\cdot\text{m}^{-3}$ )	PM <sub>2.5</sub> Decrease rate ( $\mu\text{g}\cdot\text{m}^{-3}\cdot\text{m}^{-1}$ ) <sup>1</sup>	Adjusted PM <sub>2.5</sub> Decrease rate ( $\mu\text{g}\cdot\text{m}^{-3}\cdot\text{m}^{-1}$ ) <sup>2</sup>	Contribution of emission reduction (%) <sup>3</sup>	Contribution of meteorological variations (%) <sup>4</sup>
Yufa	111.1	69.7	74.6	-0.78	-0.63	80.4	19.7
Miyun Reservoir	58.8	44.8	47.0	-0.40	-0.33	82.8	17.2
Dingling	69.6	47.1	50.6	-0.54	-0.44	80.8	19.2
Qianmen	103.9	64.0	68.9	-0.81	-0.69	85.0	15.0
Olympic center	90.4	57.2	61.7	-0.68	-0.55	80.8	19.2
Xiangshan	77.0	59.3	60.3	-0.46	-0.39	83.9	16.1
Huayuan	101.5	64.4	69.2	-0.77	-0.63	81.9	18.1
Yungang	91.8	60.2	64.0	-0.69	-0.55	79.6	20.4
WanShouxigong	93.7	62.0	66.8	-0.64	-0.50	78.2	21.8
Dongsi	94.9	62.4	67.5	-0.62	-0.49	78.9	21.1
TianTan	92.3	58.4	64.6	-0.68	-0.55	80.2	19.9
NongZhanguan	92.2	59.9	65.9	-0.66	-0.53	80.3	19.8
Gucheng	92.7	61.4	65.9	-0.65	-0.50	77.6	22.4
Guanyuan	89.6	59.5	64.6	-0.60	-0.48	79.6	20.4
BeiBuxinqu	86.6	59.5	63.3	-0.60	-0.45	75.2	24.8
WanLiu	98.1	56.2	60.4	-0.87	-0.73	84.2	15.8

398 <sup>1</sup> PM<sub>2.5</sub> decrease rate: the fitted variation slope of original monthly average PM<sub>2.5</sub> time series;399 <sup>2</sup> Adjusted PM<sub>2.5</sub> decrease rate: the fitted variation slope of adjusted monthly average PM<sub>2.5</sub> time series;400 <sup>3</sup> Contribution of emission reduction = 1 - Contribution of meteorological variations;401 <sup>4</sup> Contribution of meteorological variations = (PM<sub>2.5</sub> decrease rate - Adjusted PM<sub>2.5</sub> decrease rate) / PM<sub>2.5</sub> decrease rate.



**Fig 4. The comparison of original and KZ processed time series of  $PM_{2.5}$  concentrations in six stations from 2013 to 2017**

402  
 403  
 404  
 405 According to Table 5, the annual mean  $PM_{2.5}$  concentration in Beijing in 2017 was  
 406 35.6% lower than that in 2013. By filtering the influence of meteorological variations,  
 407 the adjusted annual mean  $PM_{2.5}$  concentration in Beijing in 2017 decreased by 31.7%  
 408 when compared to that in 2013, indicating that the variation in meteorological  
 409 conditions exerted a moderate influence on  $PM_{2.5}$  reduction from 2013 to 2017.  
 410 Meteorological conditions in Beijing were generally favorable for  $PM_{2.5}$  dispersion  
 411 during the five-year period, especially the latter half of 2017, when there was a high  
 412 frequency of strong northerly winds and much lower wintertime  $PM_{2.5}$  concentrations  
 413 than previous years.

414 For the winter of 2017, frequent windy weather and successive clean sky had a strong  
415 influence on the reduction of PM<sub>2.5</sub> concentrations in Beijing. This led to a hot debate  
416 concerning whether the notable decrease in PM<sub>2.5</sub> concentrations was mainly  
417 attributed to the favorable meteorological conditions or emission-reduction. Table 5  
418 suggests that the control of anthropogenic emissions contributed to 75.2%~85.0% of  
419 PM<sub>2.5</sub> decrease in the five-year period, indicating that emission-reduction worked  
420 effectively in all rural, urban and background stations. On average, the relative  
421 contribution of emission-reduction and meteorological variations to PM<sub>2.5</sub> reduction  
422 in Beijing from 2013 to 2017 was 80.6% and 19.4% respectively. Therefore, in spite  
423 of more favorable meteorological conditions, properly designed and implemented  
424 emission-reduction measures were the dominant driver for the remarkable decrease of  
425 PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017.

426 **4.1.2 Estimation based on WRF-CMAQ**

427 In addition to the KZ filter, we also employed WRF-CMAQ to estimate the relative  
 428 contribution of emission-reduction and meteorological conditions to the decrease of  
 429 PM<sub>2.5</sub> concentrations in Beijing. The result is shown in Table 6.

430 **Table 6. Estimated relative contribution of emission-reduction and meteorological variations to**  
 431 **PM<sub>2.5</sub> reduction in Beijing from 2013 to 2017 using WRF-CMAQ**

<b>Stations</b>	<b>Contribution of meteorological variations (%)</b>	<b>Contribution of emission-reduction(%)</b>
<b>Yufa</b>	21.9	78.2
<b>Miyun Reservoir</b>	20.8	79.2
<b>Dingling</b>	21.7	78.3
<b>Qianmen</b>	21.2	78.8
<b>Olympic center</b>	21.2	78.8
<b>Xiangshan</b>	20.3	79.7
<b>Huayuan</b>	21.2	78.8
<b>Yungang</b>	21.2	78.8
<b>WanShouxigong</b>	21.2	78.8
<b>Dongsi</b>	21.2	78.8
<b>TianTan</b>	21.2	78.8
<b>NongZhanguan</b>	21.2	78.8
<b>Gucheng</b>	22.2	77.8
<b>Guanyuan</b>	21.2	78.8
<b>BeiBuxinqu</b>	22.2	77.8
<b>WanLiu</b>	22.2	77.8

432 Based on WRF-CMAQ, the relative contribution of meteorological variations to the  
 433 decrease in PM<sub>2.5</sub> concentrations in Beijing ranged from 20.3% to 22.2% in different  
 434 stations, whilst emission-reduction accounted for about four-fifths of PM<sub>2.5</sub> reduction  
 435 from 2013 to 2017. It is worth mentioning that WRF-CMAQ is a grid-based model  
 436 and thus the calculated contribution of meteorological variations for some stations  
 437 located in the same grid was the same. Instead, station-based KZ filtering led to more  
 438 reliable analysis for each station and can better distinguish the differences between  
 439 multiple stations. Furthermore, WRF-CMAQ simply considered the differences

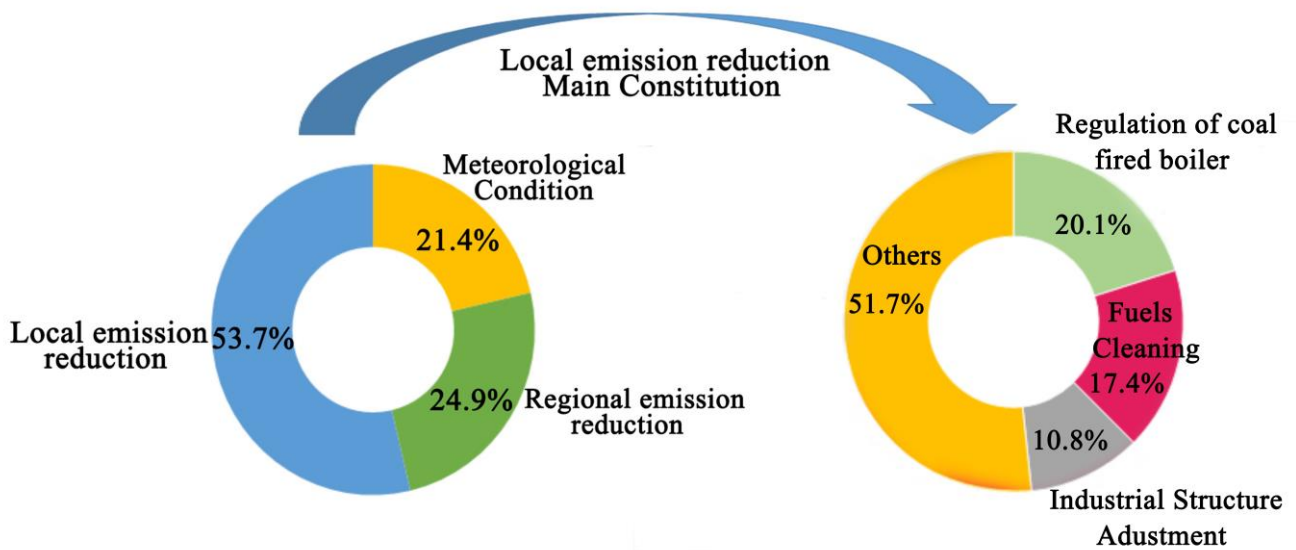
440 between meteorological conditions in 2013 and 2017 without considering their  
441 variations during the five-year period while the KZ filtering analyzed the entire time  
442 series of PM<sub>2.5</sub> and meteorological data from 2013 to 2017. The averaged relative  
443 contribution of meteorological variations to PM<sub>2.5</sub> reduction in Beijing calculated  
444 using WRF-CMAQ was 21.4%, very similar to the 19.4% calculated using KZ  
445 filtering. The slightly larger meteorological contribution calculated using  
446 WRF-CMAQ might be attributed to that WRF-CMAQ simply considered the  
447 favorable meteorological conditions in 2017 whilst KZ fully considered the long-term  
448 meteorological variations from 2013 to 2017.

449 Since KZ filtering is fully based on observed data, and simply considers the influence  
450 of time-series meteorology data on PM<sub>2.5</sub> time series, less uncertainty is involved. The  
451 accuracy of KZ filtering is influenced mainly by the variations of PM<sub>2.5</sub>-meteorology  
452 interactions in different areas and seasons. On the other hand, CTMs (e.g.  
453 WRF-CMAQ or WRF-CAMx) consider both meteorological conditions (mainly  
454 large-scale meteorological data for model simulation, not as accurate as local  
455 observed meteorological data) and anthropogenic emissions for estimating PM<sub>2.5</sub>  
456 concentrations under different emission scenarios. The accuracy of these models are  
457 not only decided by proper understanding of PM<sub>2.5</sub>-meteorology interactions, but also  
458 the reliability of emission inventories and proper descriptions of reaction mechanisms  
459 for PM<sub>2.5</sub> production, especially during heavy pollution episodes, which is a major  
460 challenge for current model simulation. Consequently, KZ filtering provides a more  
461 reliable method for researchers and decision makers to understand the relative  
462 importance of emission-reduction and meteorological conditions in recent PM<sub>2.5</sub>  
463 reduction in Beijing. Meanwhile, similar outputs from WRF-CMAQ simulation  
464 provide complementary evidence for the fact that anthropogenic emissions exerted a  
465 much stronger influence on PM<sub>2.5</sub> concentrations than meteorological conditions. In  
466 addition to the combined effects of all emission-reduction measures, we further  
467 employed WRF-CMAQ to quantify the relative contribution of different  
468 emission-reduction measures to the decrease in PM<sub>2.5</sub> concentrations in Beijing from  
469 2013 to 2017.



470 **4.2 The relative contribution of different emission-reduction measures to the**  
471 **decrease in PM<sub>2.5</sub> concentrations in Beijing**

472 The observed annual average PM<sub>2.5</sub> concentration in Beijing in 2017 was 58 mg/m<sup>3</sup>,  
473 compared with 89.5 µg/m<sup>3</sup> in 2013. Based on WRF-CMAQ simulation,  
474 meteorological conditions contributed 6.7 µg/m<sup>3</sup> whilst the control of anthropogenic  
475 emissions contribute contributed 24.7 µg/m<sup>3</sup> to the total PM<sub>2.5</sub> reduction of 31.5 µg/m<sup>3</sup>  
476 in Beijing from 2013 to 2017. Specifically, local and regional emission-reduction  
477 accounted for 16.9 µg/m<sup>3</sup> and 7.8 µg/m<sup>3</sup> of PM<sub>2.5</sub> reduction. Local emissions and  
478 regional transport took up 68.4% and 31.6% of total anthropogenic emissions in  
479 Beijing. This result is consistent with our recent study (Chen et al., 2019). Chen et al.  
480 (2019) investigated four pollution episodes in Beijing in 2013, 2016, 2017 and 2018  
481 respectively and found that local emissions accounted for 69.3%, 76.8%, 49.5% and  
482 88.4% of total emissions in Beijing respectively. Except for the moderate pollution  
483 episode in 2017, local emissions caused more than two thirds of anthropogenic  
484 emissions in Beijing. Therefore, local emissions played a dominant role for PM<sub>2.5</sub>  
485 variations in Beijing in both long-term run and heavy pollution episodes. According to  
486 three emission-reduction scenarios designed, the regulation of coal boilers had the  
487 most significant effect on PM<sub>2.5</sub> reduction in Beijing and resulted in a decrease of 6.3  
488 µg/m<sup>3</sup>. Meanwhile, increasing clean fuels for residential use and industrial  
489 restructuring also exerted strong influences on PM<sub>2.5</sub> reduction and contributed to a  
490 decrease of 5.5 µg/m<sup>3</sup> and 3.4 µg/m<sup>3</sup> respectively. The three major strategies  
491 accounted for around half of the total effects of emission-reduction on PM<sub>2.5</sub>  
492 variations in Beijing.



494 **Fig 5. The relative contribution of different influencing factors to the decrease of**  
 495 **PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017**

496 **5 Discussion**

497 By the end of 2017, the Beijing Five-year Clean Air Action Plan (2013-2017) was  
 498 completed and achieved its primary goal of reducing the annual average PM<sub>2.5</sub>  
 499 concentration to less than 60 µg/m<sup>3</sup>. Meanwhile, in November 2017, strong northerly  
 500 winds in Beijing resulted in the cleanest winter in the past five years, raising  
 501 arguments whether the favorable meteorological conditions were primarily  
 502 responsible for PM<sub>2.5</sub> reduction or whether the significant improvement in air quality  
 503 in Beijing was mainly attributed to the control of anthropogenic emissions. In this  
 504 case, a quantitative comparison between the influence of meteorological conditions  
 505 and emission-reduction on PM<sub>2.5</sub> reduction is necessary for comprehensively  
 506 evaluating the Five-year Clean Air Action Plan. Based on two different approaches,  
 507 this research revealed that the control of anthropogenic emissions contributed to  
 508 around 80% of PM<sub>2.5</sub> reductions in Beijing from 2013 to 2017, indicating that the  
 509 Five-Year Clean Air Plan exerted a dominant influence on air quality enhancement in  
 510 Beijing. The large contribution of some specific emission-reduction measures may be  
 511 obscured in the presence of favorable meteorological conditions. For instance, many  
 512 residents may attribute the clean winter of 2017 to the notable strong winds without  
 513 noticing some of major emission-reduction strategies implemented during this period.

514 A large-scale replacement of coal boilers with gas boilers was conducted in Beijing  
515 and its neighboring areas since 2013. As quantified by WRF-CMAQ, the regulation of  
516 coal boilers and increasing use of clean fuels for residential use jointly contributed to  
517 an  $11.8\mu\text{g}/\text{m}^3$  decrease in  $\text{PM}_{2.5}$  concentrations, much (almost twice) larger than the  
518  $6.7\mu\text{g}/\text{m}^3$  decrease caused by favorable meteorological conditions. In general,  
519 although favorable meteorological conditions (e.g., strong winds) may lead to an  
520 instant improvement of air quality, regular emission-reduction measures exert a  
521 reliable and consistent influence on the long-term reduction of  $\text{PM}_{2.5}$  concentrations in  
522 Beijing. Given the satisfactory performance of the Five-year Clean Air Action Plan in  
523  $\text{PM}_{2.5}$  reduction, such long-term clean air plan should be further designed and  
524 implemented in Beijing and other mega cities with heavy  $\text{PM}_{2.5}$  pollution.

525 Recently, with growing attention to the completion of the Five-year Clean Air Action  
526 Plan, some other studies have also been conducted to evaluate this five-year plan.  
527 Cheng, J. et al. (2019) employed a finer-scale and more detailed local  
528 emission-inventory and quantified the relative contribution of multiple  
529 emission-reduction strategies, including the control of coal-fired boilers, increasing  
530 use of clean fuels, optimization of industrial structure, fugitive dust control, vehicle  
531 emission control, improved end-of-pipe control, and integrated treatment of VOCs.  
532 The relative contribution of these emission-reduction measures to  $\text{PM}_{2.5}$  reduction in  
533 Beijing from 2013 to 2017 was 18.7%, 16.8%, 10.2%, 7.3%, 6.0%, 5.7% and 0.6%  
534 respectively. By contrast, our research revealed that three major emission-reduction  
535 measures (the regulation of coal-fired boiler, increasing use of clean fuels and  
536 industrial restructuring) contributed 20.1%, 17.4% and 10.8% of total  $\text{PM}_{2.5}$  reduction  
537 in Beijing from 2013 to 2017, which was very close to Cheng, J et al. (2019)'s  
538 findings. Based on finer-scale local emission-inventories with more field-collected  
539 emission data, Cheng, J et al. (2019) provided a comprehensive and reliable  
540 understanding of the effects of multiple emission-reduction measures on  $\text{PM}_{2.5}$   
541 reduction in Beijing. The similar outputs from the two studies further proved the  
542 reliability of WRF-CMAQ simulation. Meanwhile, Cheng, J et al. (2019) and UNEP  
543 (2019) jointly quantified that the total amount of reduction in  $\text{SO}_2$ ,  $\text{NO}_x$ , VOCs and  
544 direct  $\text{PM}_{2.5}$  induced by the control of anthropogenic emissions was 79420t, 93522t,  
545 115752t and 44307t respectively, which was the major driver for the notable  $\text{PM}_{2.5}$

546 reduction in Beijing from 2013 to 2017.

547 Although the “2+26” regional strategy for air quality improvement in Beijing has  
548 become a hotly debated issue and growing emphasis has been placed on the proper  
549 design and implementation of regional emission-reduction strategies in Beijing and its  
550 surrounding cities, previous studies (Chen et al., 2019; Cheng, J. et al., 2019) and this  
551 research proved that local emissions played a dominant role in affecting PM<sub>2.5</sub>  
552 concentrations in Beijing. Specifically, Chen et al. (2019) pointed out that with  
553 intensive reduction of coal-fired boilers in Beijing-Tianjin-Hebei region, the relative  
554 contribution of vehicle emissions to PM<sub>2.5</sub> concentrations in Beijing, especially during  
555 heavy pollution episodes, could be up to 50%. To further improve air quality in  
556 Beijing, stricter regulations on local vehicle emissions, including contingent strategies  
557 during pollution episodes (e.g. odd-even license plate policy) and long-term policies  
558 (e.g. increasing availability of public transit systems and electric cars) should be a  
559 major priority for the next stage clean-air actions.

560 Based on KZ filtering, Cheng, N et al. (2019) and Ma et al. (2016) suggested the  
561 seasonal component contributed dominantly to O<sub>3</sub> variations in Beijing. By  
562 comparison, this research revealed that the short-term component contributed  
563 dominantly to PM<sub>2.5</sub> variations in Beijing. These findings well explained the  
564 phenomenon that ground ozone pollution in Beijing, controlled by seasonal variations  
565 of emission and meteorological conditions (especially high-temperature and  
566 low-humidity), simply occurred in summer, whilst PM<sub>2.5</sub> pollution in Beijing,  
567 controlled by short-term variations of meteorological and emission factors, might  
568 occur in all seasons. Consequently, contingent emission-reduction measures during  
569 heavy pollution episodes are an effective approach to offset the short-term  
570 deterioration of meteorological conditions and improve local air quality.

571 Despite the major contribution of emission-reduction measures to PM<sub>2.5</sub> reduction in  
572 Beijing, meteorological influences, which contributed to 20% of PM<sub>2.5</sub> reduction,  
573 should also be considered balancedly. In addition to the control of anthropogenic  
574 emissions, PM<sub>2.5</sub> reduction may be realized through meteorological means. For the  
575 winter of 2017, strong northwesterly winds led to instant improvement in air quality,  
576 suggesting wind was a dominant meteorological factor for the accumulation or

577 dispersion of PM<sub>2.5</sub> in Beijing. Meanwhile, previous studies (Chen et al., 2017)  
578 suggested that increasing wind speeds led to increased evaporation, increased  
579 sunshine duration (SSD) and reduced humidity, which further reduced local PM<sub>2.5</sub>  
580 concentrations. In other words, strong winds help reduce PM<sub>2.5</sub> concentrations  
581 through direct and indirect measures. In this light, the forthcoming Beijing  
582 Wind-corridor Project, which includes five 500m-width corridors and more than ten  
583 80m-width corridors to bring in stronger wintertime northwesterly winds, can be a  
584 promising approach for promoting long-term favorable meteorological influences on  
585 PM<sub>2.5</sub> reduction in Beijing.

## 586 **6 Conclusions**

587 To comprehensively evaluate the effect of the Beijing Five-year Clean Air Action Plan  
588 (2013-2017), we quantified the relative contribution of meteorological conditions and  
589 the control of anthropogenic emissions to the notable decrease in PM<sub>2.5</sub> concentrations  
590 in Beijing from 2013 to 2017. Based on KZ filtering, we found that meteorological  
591 conditions and emission-reduction accounted for 19.4% and 80.6% of the PM<sub>2.5</sub>  
592 reduction in Beijing, respectively. The large short-term component suggested that  
593 short-term variations of meteorological and emission factors exerted a dominant  
594 influence on the rapid variation of PM<sub>2.5</sub> concentrations in Beijing. Meanwhile,  
595 WRF-CAMQ revealed that meteorological conditions and emission-reduction  
596 contributed to 21.4% and 78.6% of PM<sub>2.5</sub> variations. Specifically, local and regional  
597 emission-reduction measures contributed to 53.7% and 24.9% of PM<sub>2.5</sub> reduction. For  
598 three major emission-reduction measures, the regulation of coal boilers, increasing  
599 use of clean fuels for residential use and industrial restructuring contributed to 20.1 %,  
600 17.4% and 10.8% of PM<sub>2.5</sub> reduction, respectively. Similar outputs from two models  
601 suggested that the control of anthropogenic emissions contributed to around 80% of  
602 the total decrease in PM<sub>2.5</sub> concentrations in Beijing from 2013 to 2017, indicating  
603 that the Five-year Clean Air Plan worked effectively and such long-term clean air plan  
604 should be continued in the following years to further reduce PM<sub>2.5</sub> concentrations in  
605 Beijing.

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614 **Author contribution**

615 Chen, Z., Gao, B. and Xu, B designed this research. Chen, Z wrote this manuscript.  
616 Chen, D., Zhuang, Y, Gao, B and Li, R. conducted data analysis. Chen, D and  
617 Zhuang, Y. produced the figures. Kwan, M., and Chen, B helped revise this  
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