A potential source of atmospheric sulfate from O_2 -induced SO_2 oxidation by ozone

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Abstract. It was formerly demonstrated that O_2SOO^- forms at collisions rate in the gas-phase as a result of SO_2 reaction with O_2^- . Hereby, we present a theoretical investigation of the chemical fate of O_2SOO^- by reaction with O_3 in the gas-phase, based on *ab initio* calculations. Two main mechanisms were found for the title reaction, with fundamentally different products: (i) formation of a van der Waals complex followed by electron transfer and further decomposition to $O_2 + SO_2 + O_3^-$ and (ii) formation of a molecular complex from O_2 switching by O_3 , followed by SO_2 oxidation to SO_3^- within the complex. Both reactions are exergonic, but separated by relatively low energy barriers. The products in the former mechanism would likely initiate other SO_2 oxidations as shown in previous studies, whereas the latter mechanism closes a path wherein SO_2 is oxidized to SO_3^- . The latter reaction is atmospherically relevant since it forms the SO_3^- ion, hereby closing the SO_2 oxidation path initiated by O_2^- . The main atmospheric fate of SO_3^- is nothing but sulfate formation. Exploration of the reactions kinetics

15 indicates that the path of reaction (ii) is highly facilitated by humidity. For this path, we found an overall rate constant of 4.0×10^{-11} cm³ molecule⁻¹ s⁻¹ at 298 K and 50% relative humidity. The title reaction provides a new mechanism for sulfate formation from ion-induced SO₂ oxidation in the gas-phase and highlights the importance of including such mechanism in modeling sulfate-based aerosol formation rates.

1 Introduction

The chemistry of sulfur is highly important in the atmosphere. Through its oxidation products, sulfur participates in the formation of secondary atmospheric aerosols, clouds and acid rain. Sulfur dioxide (SO₂), the most abundant sulfur-containing molecule in the atmosphere, is known to react both in the gas-phase and in multiphase oxidation processes following different

- 5 mechanisms to form sulfate as the final oxidation species. The main SO₂ oxidizers in the gas-phase include the hydroxyl radical (OH) (Seinfeld and Pandis, 2016), stabilized Criegee intermediates (Welz et al., 2012; Mauldin III et al., 2012; Vereecken et al., 2012) and atmospheric ions (Fehsenfeld and Ferguson, 1974; Enghoff et al., 2012; Tsona et al., 2015). The main routes for SO₂ heterogeneous/multiphase oxidation include reactions with mineral dust (Harris et al., 2013), O₃ and H₂O₂ in cloud droplets (Caffrey et al., 2001; Hoyle et al., 2016; Harris et al., 2012; Hegg et al., 1996), NO₂ and O₂ in aerosol water and on
- 10 CaCO₃ particles (Cheng et al., 2016; Wang et al., 2016; Zhang et al., 2018; Yu et al., 2018; Zhao et al., 2018). In the gas-phase, the SO₂ oxidation by OH and Criegee intermediates leads to SO₃ that ultimately forms H₂SO₄ (Larson et al., 2000), whereas reactions with ions are generally more complex. In the aqueous phase, SO_4^{2-} is formed as the terminal oxidation species. Sulfate is known to be the main driving species in atmospheric aerosols formation and its formation is critical in the determination of aerosol formation rates (Nieminen et al., 2009; Sipilä et al., 2010; Kuang et al., 2008; Kulmala et al., 2000). The role of ions
- 15 in this formation has been well established (Yu, 2006; Yu and Turco, 2000, 2001; Enghoff and Svensmark, 2008; Kirkby et al., 2011;Wagner et al., 2017; Yan et al., 2018), although relatively minor compared to the mechanism involving neutral particles, exclusively (Eisele et al., 2006; Manninen et al., 2010; Kirkby et al., 2011; Hirsikko et al., 2011; Wagner et al., 2017).

The immediate products of SO₂ oxidation by ions are mostly sulfur oxides ions intermediates (Fehsenfeld and Ferguson, 1974;

- 20 Möhler et al., 1992; Bork et al., 2012; Tsona et al., 2014) that are susceptible of triggering new reactions or recombining with oppositely charged counterparts to form neutral species. Some of these ions, namely SO₃⁻, SO₄⁻, and SO₅⁻, were detected at relatively high concentrations in the ambient atmosphere (Ehn et al., 2010) and in chamber experiments of SO₂ ionic oxidation studies (Nagato et al., 2005; Hvelplund et al., 2013; Kirkby et al., 2011; Kirkby et al., 2016). The chemical fate of most sulfur oxides anions is relatively known. Bork et al. showed that SO₃⁻ can form SO₃, the precursor for H₂SO₄, through electron
- transfer to ozone (O₃) (Bork et al., 2012). SO₃⁻ can equally react with O₂ to form SO₅⁻ whose atmospheric outcome by reaction with O₃ is H₂SO₄ formation (Bork et al., 2013). It was also speculated from chamber studies that SO₅⁻ could form and stabilize clusters with sulfuric acid in the gas-phase (Kirkby et al., 2011). Reliable predictions of the outcomes of these ions require an exact knowledge of their chemical structures since interactions between molecules or ions depend both on their physical and chemical properties. A previous study demonstrated that two forms of SO₄⁻ separated by a high energy barrier may exist in the
- 30 atmosphere (Tsona et al., 2014): the sulfate radical ion henceforth indicated as SO_4^- , and the peroxy form, O_2SOO^- , in which the O_2S-OO^- bond nature is more dative than covalent. Formerly, the two isomers were often misleadingly attributed exclusively to the sulfate radical ion, the most stable form of SO_4^- . However, their reactive properties greatly differ (Fehsenfeld and Ferguson, 1974).

The formation mechanisms of SO_4^- in the gas-phase have been largely unknown but, recent studies showed that this ion can be formed by SO_5^- reaction with O_3 (Bork et al., 2013) and in a O_2SOO^- isomerization process catalyzed by NO (Tsona et al., 2018). SO_4^- can also be produced during the chemical transformation of organic compounds, triggered by sulfate salts (Noziere

- 5 et al., 2010), whereas O₂SOO⁻ is formed at collision rates upon SO₂ reaction with O₂⁻ (Fahey et al., 1982; Tsona et al., 2014). The sulfate radical ion is believed to react with unsaturated compounds to form organosulfates, a major component of secondary organic aerosol (Surratt et al., 2007; Surratt et al., 2008; Schindelka et al., 2013). Using first-principles calculations, SO₄⁻ was shown to act as a catalyst in SO₂ oxidation to SO₃ by O₃ in the gas-phase and hence, plays a role in atmospheric aerosol formation (Tsona et al., 2015; Tsona et al., 2016). The chemistry of O₂SOO⁻ is largely unknown, although potentially
- 10 important for sulfur chemistry and atmospheric aerosol formation. Fehsenfeld and Ferguson found that O_2SOO^- can be decomposed by NO_2 into NO_3^- and SO_3 (Fehsenfeld and Ferguson, 1974) and a recent study showed that in the presence of nitrogen oxides ($NOx = NO_2 + NO$), O_2SOO^- can be converted into sulfate (Tsona et al., 2018). In mildly polluted environments, the concentration of O_3 can be few orders of magnitude higher than that of NOx and the chemical fate of O_2SOO^- would then also greatly depend on collisions with O_3 . In such environments, O_2SOO^- could experience much more collisions with O_3 than
- 15 with NOx.

Hereby, we investigate the reaction between $O_2SOO^{-}...(H_2O)_{0-1}$ and O_3 using *ab initio* calculations. By determining the reactions thermodynamics and kinetics, we examine the possible pathways for the reaction and propose the most probable outcome of $O_2SOO^{-}...(H_2O)_{0-1}$ based on O_3 reaction. Implications of the most relevant pathways in aerosol formation are discussed.

2 Methods

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2.1 Geometry optimizations, thermochemical and charge analysis

As the substrate in this study is a radical anion, all stationary points in the energy surface were optimized using density functional theory (DFT) based on the UM06-2X density functional (Zhao and Truhlar, 2008) and the aug-cc-pVTZ basis set (Dunning Jr et al., 2001), setting the charge to -1 and the spin multiplicity to 2. The use of UM06-2X implies using unrestricted wavefunctions to describe the quantum state of the system. Spin contamination often arises from unrestricted density functional theory (DFT) calculations and it is not guaranteed that the electronic states from these calculations are eigenstates of the \hat{S}^2 operator. The spin contamination was then evaluated for all electronic states as $\Delta S = \langle \hat{S}^2 \rangle - \langle \hat{S}^2 \rangle_{ideal}$, where $\langle \hat{S}^2 \rangle$ is actual value of the expectation value of the \hat{S}^2 operator from DFT calculations and $\langle \hat{S}^2 \rangle_{ideal}$ is the ideal expectation value. For systems

30 explored in this study, $\langle \hat{S}^2 \rangle_{ideal} = 0.75$.

The UM06-2X functional has successfully proven to be adequate for reactions involving transition state (TS) configurations (Elm et al., 2012, 2013a, b). Harmonic vibrational frequencies analysis on the optimized structures were performed (at 298 K and 1 atm) using the UM06-2X/aug-cc-pVTZ method under the harmonic oscillator-rigid rotor approximation. These calculations ensured that the obtained stationary points were minima or TS and, also, provided the thermal contributions to the Gibbs free energy and the enthalpy.

5 Gibbs free energy and the enthalpy.

Transition states structures were initially located by scanning the reactants configurations. The best TS guesses out of the scans were then refined using the synchronous transit quasi Newton method (Peng et al., 1996), and the final TS structures underwent internal reaction coordinate calculations (Fukui, 1981) to ensure they connected the reactants to desired products.

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The electronic energies of the UM06-2X/aug-cc-pVTZ optimized geometries were corrected with the UCCSD(T) method (Purvis and Bartlett, 1982) in conjunction with the aug-cc-pVTZ basis set. The Gibbs free energies, G, of all relevant species were then calculated as

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$$G = E_{\text{UCCSD}(T)} + G_{\text{therm}}$$
(1)

where $E_{\text{UCCSD}(T)}$ is the electronic energy calculated with the UCCSD(T)/aug-cc-pVTZ method and G_{therm} is the thermal contribution to the Gibbs free energy, calculated at the UM06-2X/aug-cc-pVTZ level of theory. All geometry optimizations, harmonic vibrational frequencies analysis and electronic energies correction calculations were carried out in the Gaussian 09 package (Frisch et al., 2016).

To analyse the distribution of the excess electronic charge over different species and fragments in the optimized systems, we used the Atoms-in-Molecules charge partitioning method presented by Bader (Bader, 1998). This is an intuitive way of dividing the molecules of a system into atoms, which are purely defined in terms of electronic charge density. The Bader

25 charge partitioning assumes that the charge density between atoms of a molecular system reaches a minimum, which is an ideal place to separate atoms from each other. As input, this method requires electronic density and nuclear coordinates from electronic structure calculations. We used the approach implemented in the algorithm developed by Henkelman and co-workers, which has been shown to work well both for charged and water-containing systems (Tang et al., 2009; Bork et al., 2011; Henkelman et al., 2006).

30 2.2 Reactions kinetics

Regardless of the presence of water, the reaction between O_2SOO^- and O_3 begins by forming different van der Waals prereactive intermediates, depending on the orientation of the reactants at impact. The pre-reactive intermediate could either decompose to different species or react further through a transition state configuration to form new products: The traditional approach to determine the rate constant of reaction (R1) relies on the steady-state approximation and leads to the following equation:

$$k = k_{\text{coll}} \frac{k_{\text{reac}}}{k_{\text{reac}} + k_{\text{evap}}} \tag{2}$$

where k_{coll} is the collision frequency for O₂SOO⁻-O₃ collisions, k_{evap} is the rate constant for the evaporation of the pre-reactive intermediate back to initial reactants, and k_{reac} is the unimolecular rate constant for the reaction of the pre-reactive intermediate to the products. Moreover, assuming that $k_{evap} >> k_{reac}$, the rate constant of reaction (R1) becomes $k = K_{eq}k_{reac}$ over a range of temperatures, with K_{eq} being the equilibrium constant of formation of the pre-reactive intermediate from the reactants. This consideration is, however, not valid for reactions with submerged barrier, since the pre-reactive intermediate seldom thermally equilibrates. For such reactions, a two-transition state theory has been introduced, treating two distinct transition state bottlenecks that define the net reactive flux (Klippenstein et al., 1988; Georgievskii and Klippenstein, 2005; Greenwald et al., 2005). The first bottleneck, the "outer" transition state, occurs in the association of the initial reactants to form the pre-reactive intermediate, whereas the second bottleneck, the "inner" transition state, occurs in the transformation of the pre-reactive intermediate to the products. Based on this theory, the overall rate constant (*k*) for a reaction channel is expressed in terms of the outer (k_{out}) and inner (k_{in}) rate constants as:

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$$\frac{1}{k} = \frac{1}{k_{\text{out}}} + \frac{1}{k_{\text{in}}} \tag{3}$$

The outer transition state is treated by the long-range transition state theory approach (Georgievskii and Klippenstein, 2005), while the inner transition state is resolved by the transition state theory. It was shown that for interactions between ions and neutral molecules, due to their long-range attraction, the collision cross section is larger than would be measured from the physical dimensions of the colliding species (Kupiainen-Määttä et al., 2013). To account for this phenomenon, the outer rate constant was determined from the ion-dipole parametrization of Su and Chesnavich who performed trajectory simulations of collisions between a point charge and a rigidly rotating molecule (Su and Chesnavich, 1982). This parametrization is equivalent

to a Langevin capture rate constant $(k_{\rm L})$ scaled by a temperature-dependent term and was found to provide good agreement

30 with experiments (Kupiainen-Määttä et al., 2013). It is given as

$$k_{\rm out} = k_{\rm L} \times \left(\frac{(x+0.5090)^2}{10.526} + 0.9754\right) \tag{4}$$

where $k_{\rm L} = q\mu^{-1/2}(\pi\alpha/\varepsilon_0)^{1/2}$, $x = \mu_{\rm D}/(8\pi\varepsilon_0\alpha k_{\rm B}T)^{1/2}$, q is the charge of the ion, μ is the reduced mass of the colliding species, ε_0 is the vacuum permittivity, α and $\mu_{\rm D}$ are the polarizability and dipole moment of the neutral molecule (ozone), $k_{\rm B}$ is Boltzmann's constant, and T is the absolute temperature. The inner rate constant can be written as:

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$$k_{\rm in} = \frac{k_{\rm B}T}{hc^0} \times \exp\left(-\frac{\Delta G^{\#}}{RT}\right) \tag{5}$$

where $\Delta G^{\#}$ is the Gibbs free energy barrier separating the pre-reactive intermediate and the products, *h* is the Planck's constant, *R* is the molar gas constant, and c^0 is the standard gas-phase concentration. The constants and variables in Eq. (4) and Eq. (5)

10 are given in centimetre-gram-second (CGS) system of units and International System (SI) units, respectively. Details on these units are given in the Supplement.

3 Results and discussion

Starting with optimized structures of O₂SOO⁻···(H₂O)₀₋₁ and O₃ shown in Fig. S1, a series of geometry optimizations were performed on the O₂SOO⁻···(H₂O)₀₋₁ + O₃ system, taking into account different spatial orientations of the reactants at impact. These optimizations led to two main chemical processes, depending on the initial orientation of the reactants, with potentially different outcomes. The first process is the formation of a van der Waals complex followed by its direct decomposition to other species. The second process is the low-lying formation of a molecular complex in which the SO₂ entity of O₂SOO⁻···(H₂O)₀₋₁ is oxidized to SO₃⁻.

20 3.1 Cluster formation and decomposition of O₂SOO^{-...}(H₂O)₀₋₁

As O_3 approaches $O_2SOO^{-...}(H_2O)_{0-1}$ towards the oxygen atoms of the peroxy fragment or the oxygen atom of the SO₂ moiety, the immediate outcome of $O_2SOO^{-...}(H_2O)_{0-1}$ and O_3 collisions is the formation of the van der Waals $O_3...O_2SOO^{-...}(H_2O)_{0-1}$ complex in which O_3 interacts with O_2SOO^{-} . Among the different stable configurations found upon optimizations, we solely report the most stable one with respect to the Gibbs free energy, which is henceforth denoted RC1 and RCW1 for the

- 25 unhydrated and monohydrated, respectively, shown in Fig. 1. Exploration of the RC1 and RCW1 structures reveals that O₂SOO⁻...(H₂O)₀₋₁ basically keeps its configuration upon clustering with O₃. The spin contaminations for RC1 and RCW1 are negligible, being 0.0086 and 0.0081, respectively. The electronic energies of formation of RC1 and RCW1 are -5.1 and -4.6 kcal mol⁻¹, respectively. Despite these complexes may form at 0 K, the Gibbs free energies of their formation under atmospheric pressure and 298 K (4.5 and 4.7 kcal mol⁻¹, respectively) indicate that their formation is endergonic under atmospherically
- 30 relevant conditions. These Gibbs free energy values indicate that, if formed, these complexes would not live long and will rather decompose either to initial reactants or to different species. Hence, the Gibbs free energies of formation of RC1 and

RCW1 define the lowest states at which O_2SOO^- can interact with O_3 to allow electron transfer and O_2S-OO decomposition, and thus represent the energy barrier towards $O_2 + SO_2 + O_3^-$ formation. Inspecting the vibrational modes of RC1 and RCW1, two vibrations are found that would clearly lead to the dissociation of O_2SOO^- within the complex. The analysis of the charge distribution over $O_3 \cdots O_2SOO^- \cdots (H_2O)_{0-1}$ shows that the extra electron initially located on O_2SOO in the reactants has migrated

5 to the O_3 molecule in the van der Waals product complex, as can be observed in Fig. S2. This is as expected, given the high electronegativity of O_3 relative to those of O_2 and SO_2 (Rothe et al., 1975). The charge distribution over the different atoms of the optimized complex is weakly affected by the presence of water, as previously demonstrated by Bork and co-workers (Bork et al., 2011).

The most likely fates of RC1 and RCW1 are, therefore, decomposition into O₂, SO₂ and O₃⁻ as follows:

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$$O_2SOO^{-...}(H_2O)_{0-1} + O_3 \leftrightarrow O_3^{...}O_2SOO^{-...}(H_2O)_{0-1} \rightarrow O_2 + SO_2 + O_3^{--} + (H_2O)_{0-1}$$
(R2)

The numerical values of the formation energies of all intermediate species in reaction (R2) are given in Table 1 and the energy surfaces are plotted in Fig. 2. RC1 and RCW1 decompositions are highly exergonic at 298 K, occurring with -18.1 and -16.7 kcal mol⁻¹ Gibbs free energy changes, respectively. These processes are, therefore, likely to occur in the atmosphere upon

formation of $O_3 \cdots O_2 SOO^{-} \cdots (H_2 O)_{0-1}$.

The limiting step in reaction (R2) is the formation of RC1 and RCW1, whose formation energies indicated above can then be considered as the only barrier to the formation of $O_2 + SO_2 + O_3^-$. This leads to overall rate constants (according to Eq. 5) of 1.4×10^{-10} and 9.9×10^{-11} cm³ molecule⁻¹ s⁻¹ at 298 K for the unhydrated and monohydrated reaction, respectively. Both reactions

20 1.4×10^{-10} and 9.9×10^{-11} cm³ molecule⁻¹ s⁻¹ at 298 K for the unhydrated and monohydrated reaction, respectively. Both reactions are, in principle, collision-limited and the effect of hydration on the kinetics is found to be negligible. The atmospheric relevance of reaction (R2) has been determined earlier (Bork et al., 2012; Bork et al., 2013; Enghoff et al., 2012).

3.2 O₂SOO^{-...}(H₂O)₀₋₁ reaction with O₃

- 25 When O₃ approaches O₂SOO^{-...}(H₂O)₀₋₁ from the sulfur atom side, the formation of a more stable cluster than found above prevails. The incoming O₃ molecule strongly interacts with O₂SOO^{-...}(H₂O)₀₋₁ by forming a coordination bond with the sulfur atom and hereby, inducing the ejection of the O₂ molecule that remains in interaction with the remainder of the system. This process leads to the formation of the O₂···O₂S–O₃^{-...}(H₂O)₀₋₁ complex which further transforms, through an intramolecular SO₂ oxidation, into SO₃^{-...}(H₂O)₀₋₁ + 2O₂ according to the following equation:
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$$O_2SOO^{-...}(H_2O)_{0-1} + O_3 \leftrightarrow O_2^{...}O_2S - O_3^{-...}(H_2O)_{0-1} \rightarrow SO_3^{-...}(H_2O)_{0-1} + 2O_2$$
(R3)

The configurations of the most stable intermediate structures in reaction (R3) are given in Fig. 1. The charge analysis on this system indicates that the electronic charge on the pre-reactive complex is essentially on two oxygen atoms of the O_3 moiety that is coordinated to SO_2 as can be seen in Fig. 3(a). The net charge on these two oxygen atoms is 1.04e, whereas the net charge on the free O_2 molecule is 0.01e. The latter value shows that the O_2 molecule formed in the pre-reactive complex has

- 5 no unpaired electrons, and hence is a singlet. Although the charge is still on the O_3 moiety in the transition state configuration, it is mostly located on the oxygen atom bound to sulfur (**Fig. 3(b**)). The net charge on the two outer oxygen atoms of the O_3 moiety in the transition state has substantially decreased to 0.30e while the charge on the free O_2 molecule has slightly increased to 0.04e. The strong decrease in the charge of the two outer oxygen atoms of O_3 from the pre-reactive complex to the transition state suggests that the O_2 molecule to form in the product will likely be a singlet. In the products (**Fig. 3(c**)), the old free O_2
- 10 molecule has a net charge of 1.99e, whereas the charge on the newly formed O_2 is 0.06e. The 1.99e charge on the old free O_2 molecule indicates the presence of unpaired electrons in its configuration, meaning that the singlet O_2 has been transformed into triplet. This clearly shows that a spin flip has occurred in O_2 during further optimization of the products. The newly formed O_2 with 0.06e charge is obviously a singlet. This analysis shows for the unhydrated system that the singlet O_2 initially formed in the pre-reactive complex transforms into triplet in the products state, while a new singlet O_2 is also formed. In the
- 15 monohydrated system, the singlet O_2 initially formed in the pre-reactive complex remains as singlet in the products state and a triplet O_2 is also released. Overall, the two forms of O_2 (singlet and triplet) are formed in the studied reaction, despite following different mechanisms. Although the water molecule in the monohydrated system does not retain any electric charge, it most likely stabilizes the initially formed singlet O_2 and prevents the spin flip.
- 20 Though the necessity to determine the electronic structure of the O_2 molecule in the singlet sate $({}^{1}\Delta_{g})$ has been demonstrated to be useful (Buttar and Hirst, 1994), obtaining a reliable electronic energy for $O_2({}^{1}\Delta_{g})$ is difficult (Drougas and Kosmas, 2005). An alternative approach to determine this energy is to add the experimental energy spacing (22.5 kcal mol⁻¹) between triplet and singlet states of O_2 to the computed electronic energy of the triplet O_2 (Schweitzer and Schmidt, 2003; Drougas and Kosmas, 2005). We used this approach to determine the energies of the products of reaction (R3). The numerical values of the
- formation energies of all intermediate species in reaction (R3) are thus given in Table 1 and the energy surfaces are plotted in Fig. 2. The most stable optimized structures of $O_2 \cdots O_2 S - O_3^{-} \cdots (H_2 O)_{0-1}$ according to our calculations are denoted as RC2 and RCW2 for the unhydrated and monohydrated systems, respectively and are shown in Fig. 1. Regardless of the presence of water, the $O_2 \cdots O_2 S O_3^{-}$ configuration results from O_2 being switched by O_3 in the $O_2 S OO^{-}$ molecular ion. In the optimized $O_2 \cdots O_2 S - O_3^{-}$ structure, the O3 atom of O3 points towards the S atom of O2 SOO⁻, forming S-O3 bonds at distances of 1.90 and
- 30 1.87 Å in the absence and presence of water, respectively. These bonds are coordination bonds in nature since the S-O covalent bond in e.g., 1.43 Å in SO₃ and 1.46 Å in H₂SO₄. The S-O3 coordination bond distances in RC2 and RCW2 are shorter by 0.04 and 0.03 Å than O₂S-OO⁻ bond distances in unhydrated and monohydrated O₂SOO⁻ forms. This indicates stronger interaction between O₃ and SO₂, and hence higher stability of O₂···O₂S-O₃⁻ relative to O₂SOO⁻.

The formations of RC2 and RCW2 are highly exergonic, with Gibbs free energy changes at 298 K of -14.7 and -12.4 kcal mol⁻¹, respectively. These values, with corresponding electronic energies and enthalpies are shown in Table 1. These Gibbs free energy changes for the formation of RC2 and RCW2 are about 19 kcal mol⁻¹ lower than those of RC1 and RCW1 at similar conditions, indicating the higher stability of RC2 and RCW2, and the highly favourable switching reaction at ambient

- 5 conditions. SO₂ oxidation can readily occur within the O₂···O₂S–O₃⁻···(H₂O)₀₋₁ cluster and lead to SO₃⁻ formation. In principle, to form the products of reaction (R3), the O3 atom of the O₃ moiety in O₂···O₂S–O₃⁻···(H₂O)₀₋₁ transfers, through transition state configurations, to SO₂ and form SO₃⁻ followed by the ejection of the O₂ molecule. The transition states are denoted TS2 and TSW2 for the unhydrated and monohydrated systems, respectively, and their structures are presented in Fig. 1. While RC2 and RCW2 are formed with similar Gibbs free energies within 2 kcal mol⁻¹ difference, the formation Gibbs free energies of
- 10 their transition states at similar conditions greatly differ. TS2 is located at 10 kcal mol⁻¹ Gibbs free energy above RC2, while TSW2 is located at -4 kcal mol⁻¹ below RCW2. It is speculated that the low energy barrier in the monohydrated reaction is the result of a strong stabilisation of the transition state due to hydration, with the S-O3 bonds in RCW2 and TSW2 shorter by ~0.03 Å than in RC2 and TS2. Another reason for this substantial drop in energy barrier is the difference in the electronic configurations of the two outer oxygen atoms of the O₃ moiety in the two transition states that form O₂ with different
- 15 multiplicities in the products.

Based on the TS2 energy, the unimolecular decomposition of $O_2 \cdots O_2 S - O_3^-$ at 298 K in the absence of water was found to occur at a rate constant of 3.1×10^5 s⁻¹, corresponding to an atmospheric lifetime of $3.3 \ \mu$ s. Both this short lifetime and the negative energy barrier of the monohydrated reaction indicate that $O_2 \cdots O_2 S - O_3^-$ would not live long enough to experience collisions with other atmospheric oxidants. It should be noted that few to no collisions with nitrogen can, however, be achieved. It follows

with other atmospheric oxidants. It should be noted that few to no collisions with nitrogen can, however, be achieved. It follows that the most likely outcome of $O_2 \cdots O_2 S - O_3^-$ is decomposition to the products of reaction (R3), which are formed with about -46 kcal mol⁻¹ overall Gibbs free energy at 298 K. The net reaction is an O_2^- -initiated SO₂ oxidation to SO₃⁻ by O₃.

The spin contamination for electronic states in reaction (R3) is quite significant, being 1.0122, 1.4666, and 2.0374 for the pre-25 reactive complex, transition state and product, respectively, and is almost insensitive to the presence of water. The actual values of the expectation values of the \hat{S}^2 operator for all electronic states obtained from our calculations are given in the Supplement, along with their cartesian coordinates. The high values of spin contamination likely reflect the formation of O₂ with different multiplicities within the system. As the charge analysis indicates, starting with singlet O₂ in the pre-reactive complexes of reaction (R3), both singlet and triplet O₂ are formed in the final products.

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The overall rate constants of reaction (R3), determined at 298 K using Eq. (3), are 1.3×10^{-14} and 8.0×10^{-10} cm³ molecule⁻¹ s⁻¹ for the unhydrated and monohydrated reactions, respectively. The values of the different components (k_{out} and k_{in}) are listed in Table S1 of the Supplement. It is observed from Table S1 that the inner transition sate provides the dominant bottleneck to the

rate constant of the unhydrated reaction, whereas the outer transition state provides the dominant bottleneck to the rate constant of the monohydrated reaction.

The effective effect of water on the rate constant can be evaluated by taking into account the stability of O₂SOO⁻···H₂O (which is formed at the entrance channel of the reaction in the presence of water before colliding with O₃) and the equilibrium vapor pressure of water. Starting from the definition of the reaction rate for the hydrated reaction,

$$J_{(R3w)} = k_{(R3w)} \times [O_2 SOO^{-\dots} H_2 O] \times [O_3]$$
(6)

$$=k_{(R_{3W})}^{\text{eff}} \times [O_2 \text{SOO}^-] \times [O_3]$$
(7)

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where $k_{(R3w)}$ is the overall rate constant for the hydrated reaction, determined using Eq. (3), $k_{(R3w)}^{eff}$ is the effective reaction rate constant calculated as $k_{(R3w)}^{eff} = k_{(R3w)} \times K_{eq} \times p_{H2O}$. Keq is the equilibrium constant for the O₂SOO⁻ + H₂O \leftrightarrow O₂SOO⁻ ... H₂O reaction and p_{H2O} the actual water vapor pressure. Details on the determination of K_{eq} and p_{H2O} are given in the Supplement.

At 298 K and 50 % relative humidity, the effective rate constant of the monohydrated reaction is 1.7×10⁻¹⁰ cm³ molecule⁻¹ s⁻¹, four orders of magnitude higher than the rate constant of the unhydrated reaction. Therefore, water plays a catalytic role on the kinetics of reaction (R3). The net rate constant of reaction (R3) can be obtained by weighing the rate constants of the unhydrated and monohydrated reactions to corresponding equilibrium concentrations of O₂···O₂S–O₃⁻ hydrates. Using the law of mass action, we find that O₂···O₂S–O₃⁻ mostly exists as a dry species, constituting 77% of the total population, whereas the monohydrated species forms 23% of the total population. The net rate constant of reaction (R3) is then determined to be 4.0×10⁻¹¹ cm³ molecule⁻¹ s⁻¹ at 298 K.

Considering only the unhydrated process of reaction (R3), the rate constant is 4-5 orders of magnitude lower than the rate constant obtained for the $SO_2 + O_3^- \rightarrow SO_3^- + O_2$ reaction (Fehsenfeld and Ferguson, 1974; Bork et al., 2012). Despite this

- 25 difference, the oxidation process follows a similar mechanism to the one presented by Bork et al. for the $SO_2 + O_3^- \rightarrow SO_3^- + O_2$ reaction, consisting of the oxygen transfer from O_3 to SO_2 (Bork et al., 2012). The discrepancy between the two results is associated with the effect of the presence of the O–O fragment initially coordinated to SO_2 in the current study, which tends to stabilize the $O_2 \cdots O_2 S O_3^-$ pre-reactive complex. The presence of the O-O fragment seemingly deactivates SO_2 for the upcoming O transfer from O_3 to form SO_3^- . However, this situation is rapidly reversed with the presence of water as the reaction
- 30 becomes much faster, proceeding nearly at collision rate.

3.3 Further chemistry

In real atmospheric and ionized conditions, despite O_2 has lower electron affinity than O_3 , it would likely ionize faster than O_3 owing to its much higher concentration. Considering for example chamber experiments, upon interaction of ionizing particles with the gas, electrons can transfer from one species to another and, e.g., O_2^- can form and rapidly hydrate within one

5 nanosecond (Svensmark et al., 2007; Fahey et al., 1982). Furthermore, Fahey et al. showed that $O_2^{-}\cdots(H_2O)_{0-1}$ association reaction with SO₂ is faster than the electron transfer from $O_2^{-}\cdots(H_2O)_{0-1}$ to O_3 (Fahey et al., 1982). This means that in an ionized environment containing O_2 , O_3 , and SO₂, the formation of O_2S –OO⁻ resulting from SO₂ and O_2^{-} association will happen faster than O_3^{-} formation. O_2S –OO⁻ would react thereafter with O_3 and the following stepwise process could take place

$$O_2^{-} \cdots (H_2 O)_{0-1} + SO_2 \to O_2 S - OO^{-} \cdots (H_2 O)_{0-1}$$
 (R4a)

$$O_2S-OO^{-}(H_2O)_{0-1} + O_3 \rightarrow O_2 O_2S-O_3^{-}(H_2O)_{0-1}$$
 (R4b)

$$O_2 \cdots O_2 S - O_3 \cdots (H_2 O)_{0-1} \to SO_3 \cdots (H_2 O)_{0-1} + 2O_2$$
 (R4c)

Net:
$$O_2^{-\cdots}(H_2O)_{0.1} + SO_2 + O_3 \rightarrow SO_3^{-\cdots}(H_2O)_{0.1} + 2O_2$$
 (R4)

The Gibbs free energy change of this net reaction at 298 K is about -40 kcal mol⁻¹ more negative than that of the $SO_2 + O_3^- \rightarrow SO_3^- + O_2$ reaction at similar conditions. Given that the intermediate steps of reaction (R4) are significantly fast, this reaction is believed to be an important process in most environments of SO_2 ion-induced oxidation to SO_3^- or more oxidized species. The limiting step in the process of reaction (R4) is reaction (R4c) for which an energy barrier has to be overcome before the products are released.

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 SO_3^- is an identified stable ion detected in the atmosphere and in experiments (Ehn et al., 2010; Kirkby et al., 2011; Kirkby et al., 2016). The chemical fate of SO_3^- is fundamentally different from that of SO_3 that forms H_2SO_4 by hydration. Likely outcomes of SO_3^- are hydrolysis, electron transfer by collision with O_3 , reaction with O_2 and H_2O and, possibly, radicals, according to the following equations

25

$$SO_3^- + H_2O \rightarrow SO_3^- - H_2O \rightarrow H_2SO_4 + e^-$$
 (R5)

$$SO_3^- + O_3 \rightarrow SO_3 + O_3^- \tag{R6}$$

$$SO_3^- + O_2 + H_2O \rightarrow HSO_4^- + HO_2$$
 (R7)

$$SO_3^- + OH \rightarrow HSO_4^-$$
 (R8)

30

Fehsenfeld and Ferguson showed that H_2SO_4 formation could occur in the SO_3 -···H₂O cluster, releasing a free electron (Reaction (R5)) (Fehsenfeld and Ferguson, 1974). Owing to the high electron affinity of O₃ relative to SO₃ (Rothe et al., 1975), the electron can transfer from SO₃ to O₃ and lead to the formation of SO₃, the precursor for sulfuric acid in the atmosphere.

Moreover, the free electron released and the O_3^- formed in reactions (R5) and (R6), respectively, are potential triggers of new SO₂ oxidations with implication in aerosol formation (Svensmark et al., 2007; Enghoff and Svensmark, 2008; Bork et al., 2013). Reactions (R7) and (R8) are potential outcomes for SO₃⁻ as well, forming the highly stable HSO₄⁻ species that would terminate the oxidation process of SO₂ initiated by a free electron. Reactions (R5)–(R8) are likely competitive processes upon

5 SO₃⁻ formation in the gas-phase, and their different rates would determine the number of SO₂ oxidations induced by a free electron. However, they have no other fate than HSO₄⁻ or H₂SO₄, the most oxidized sulfur species in the atmosphere, which both share many properties and play a central role in atmospheric particle formation. Experimental studies have shown that in atmospheres heavily enriched in SO₂ and O₃, a free electron could initiate SO₂

Experimental studies have shown that in atmospheres heavily enriched in SO₂ and O₃, a free electron could initiate SO₂ oxidation and induce the formation of $\sim 10^7$ cm⁻³ sulfates in the absence of UV light, clearly indicating the importance of other

- 10 ionic SO₂ oxidation mechanisms than UV-induced (Enghoff and Svensmark, 2008). To evaluate the importance of the mechanism presented in this study in the formation of sulfate, it is necessary to identify the scavengers that terminate the SO₂ oxidation initiated by O₂⁻. Possible scavengers include radicals, NOx, acids, cations and other particles. The main ones are likely NOx, OH, HO₂ and organic acids, which lead to the formation of the stable NO₃⁻, HSO₄⁻, and CO₃⁻ species. If the ion concentration was known, the contribution of reaction (R4) to H₂SO₄ formation could be determined by comparing its
- 15 formation rates from ionic and electrically neutral mechanisms. Alternatively, it can be assumed that reaction (R4) is terminated when the ion cluster hits a scavenger. The free electron which acts as catalyst is then scavenged. The average catalytic turnover number (TON) is defined as (Kozuch and Martin, 2012):

$$TON = \frac{\text{concentration of limiting reacted molecules}}{\text{concentration of the catalyst}}$$
(8)

20

The concentration of the catalyst can be approximated to the concentration of the scavengers and, considering that at most atmospheric conditions $[O_3]>[SO_2]$, SO₂ is the limiting species in reaction (R4). Equation (8) can then be re-written as

$$TON \approx \frac{[SO_2]}{[OH] + [HO_2] + [NO_x] + [organic acids]}$$
(9)

25

The catalytic efficiency of SO2 ion-induced oxidation is then given as

$$J_{\rm ion} = k_{\rm ion} \times \text{TON} \tag{10}$$

30 Where k_{ion} is the ion production rate. Depending on the tropospheric temperature and altitude, measurements at the Cosmics Leaving Outdoor Droplets chamber at CERN found $k_{ion} = 2-100$ cm⁻³ s⁻¹, covering the typical ionization range in the troposphere (Franchin et al., 2015). We assume nearly pristine conditions with [SO₂] = 5 ppb = 1.2×10^{11} molecule cm⁻³, [NOx] = 200 ppt = 4.9×10^9 molecule cm⁻³, [OH] = 5.0×10^5 molecule cm⁻³ (day and night average), and [HO₂] = 10^8 molecule cm⁻³ (Dusanter et al., 2009; Holland et al., 2003). Noting that formic acid and acetic acid are the most abundant organic acids in the atmosphere, their concentrations are considered in Eq. (9) as representative examples for organic acids, [organic acids] = 110 ppt = 2.7×10^9 molecule cm⁻³ (Le Breton et al., 2012; Baasandorj et al., 2015). We then determine J_{ion} in the range 3.2×10^1 – 1.6×10^3 cm⁻³ s⁻¹. The rate of the UV-induced SO₂ oxidation by OH is

5

$$J_{\rm UV} = k_{\rm UV} \times [\rm{SO}_2] \times [\rm{OH}] \tag{11}$$

With $k_{\rm UV} = 1.3 \times 10^{-12} \text{ cm}^3$ molecule⁻¹ s⁻¹ (Atkinson et al., 2004), $J_{\rm UV} = 7.9 \times 10^4$ molecule cm⁻³ s⁻¹, and the proportion of H₂SO₄ formed from ion-induced oxidation can be estimated from the following equation

10

$$\frac{[H_2SO_4]_{\text{ion}}}{[H_2SO_4]_{\text{total}}} = \frac{J_{\text{ion}}}{J_{\text{UV}} + J_{\text{ion}}}$$
(12)

We find that the contribution of ion-induced SO₂ oxidation to H₂SO₄ formation can range from 0.1 to 2.0% of the total formation rate. This estimate could be improved by considering also the SO₂ oxidation by Criegee Intermediates, another 15 important channel for H₂SO₄ formation.

3 Conclusions

This study highlights the role of the superoxide ions (O_2^-) in SO₂ oxidation. Our previous study demonstrated that SO₂ interacts with O_2^- and forms O_2SOO^- whose atmospheric fate remains unelucidated (Tsona et al., 2014). In this study, we used ab initio

- 20 calculations to assess the chemical fate of O_2SOO^- by collisions with O_3 . Regardless of the presence of water, two main mechanisms are observed, leading to fundamentally different products. The first mechanism is characterized by electron transfer followed by O_2SOO^- decomposition, leading to O_3^- formation and releasing SO_2 . The chemistry of $SO_2 + O_3^-$ has been explored elsewhere. The second mechanism is characterized by SO_2 oxidation and proceeds through formation of a pre-reactive complex that subsequently reacts to form the products by overcoming a relatively low energy barrier. The overall reaction, O_2^-
- 25 + SO₂ + O₃ \rightarrow SO₃⁻ + 2O₂, is faster and more energetically favourable than the SO₂ + O₃⁻ \rightarrow SO₃⁻ + O₂ reaction, thereby highlighting the positive role of O₂⁻ in SO₂ ionic oxidation. Hence, the two reactions may compete in chamber experiments and in the atmosphere.

While for the electron transfer and O_2SOO^- decomposition process the reaction is hindered by the presence of water, the oxidation reaction is catalysed instead as the rate constant is increased by 6 orders of magnitude with the presence of water.

30 Weighing the rate constants of unhydrated and monohydrated reactions to the equilibrium concentrations of hydrates of corresponding pre-reactive complexes leads to the net rate constant of 4.0×10^{-11} cm³ molecule⁻¹ s⁻¹ at 298 K for the oxidation reaction. Hence, this reaction proceeds nearly at collision rate. The main species (SO₃⁻) in the end products of the studied

reaction has been proved to form both in the atmosphere and in experiments, where it definitely plays a role in atmospheric sulfur chemistry and particle formation. The contribution of this mechanism to the total atmospheric sulfuric acid formation is estimated. The studied reaction further deepens the understanding of ion-induced SO_2 oxidation, with implications in aerosol formation.

5 Author contributions

NTT and LD designed the work. NTT performed all calculations and analysed the data. NTT wrote the whole manuscript and LD edited it.

Competing interests

The authors declare that they have no conflict of interest.

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References

Atkinson, R., Baulch, D. L., Cox, R. A., Crowley, J. N., Hampson, R. F., Hynes, R. G., Jenkin, M. E., Rossi, M. J., and Troe, J.: Evaluated kinetic and photochemical data for atmospheric chemistry: Volume I - gas phase reactions of Ox, HOx, NOx and SOx species, Atmos. Chem. Phys., 4, 1461-1738, doi:10.5194/acp-4-1461-2004, 2004.

20 Baasandorj, M., Millet, D. B., Hu, L., Mitroo, D., and Williams, B. J.: Measuring acetic and formic acid by proton-transferreaction mass spectrometry: sensitivity, humidity dependence, and quantifying interferences, Atmos. Meas. Tech., 8, 1303-1321, doi:10.5194/amt-8-1303-2015, 2015.

Bader, R. F. W.: A Bond Path: A Universal Indicator of Bonded Interactions, J. Phys. Chem. A, 102, 7314-7323, doi:10.1021/jp981794v, 1998.

Bork, N., Kurtén, T., Enghoff, M., Pedersen, J. O. P., Mikkelsen, K. V., and Svensmark, H.: Ab initio studies of $O_2^{-}(H_2O)_n$ and $O_3^{-}(H_2O)_n$ anionic molecular clusters, n≤ 12, Atmos. Chem. Phys., 11, 7133-7142, doi:10.5194/acp-11-7133-2011, 2011.

¹⁵

Bork, N., Kurtén, T., Enghoff, M., Pedersen, J. O. P., Mikkelsen, K. V., and Svensmark, H.: Structures and reaction rates of the gaseous oxidation of SO₂ by an $O_3^{-}(H_2O)_{0.5}$ cluster – a density functional theory investigation, Atmos. Chem. Phys., 12, 3639-3652, doi:10.5194/acp-12-3639-2012, 2012.

Bork, N., Kurtén, T., and Vehkamäki, H.: Exploring the atmospheric chemistry of O₂SO₃⁻ and assessing the maximum turnover number of ion-catalysed H₂SO₄ formation, Atmos. Chem. Phys., 13, 3695-3703, doi:10.5194/acp-13-3695-2013, 2013.

Buttar, D., and Hirst, D. M.: Ab initio quantum chemistry study of the gas-phase reaction of ClO with HO₂, J. Chem. Soc. Faraday Trans., 90, 1811-1817, doi:10.1039/FT9949001811, 1994.

Caffrey, P., Hoppel, W., Frick, G., Pasternack, L., Fitzgerald, J., Hegg, D., Gao, S., Leaitch, R., Shantz, N., Albrechcinski, T., and Ambrusko, J.: In-cloud oxidation of SO₂ by O₃ and H₂O₂: Cloud chamber measurements and modeling of particle growth, J. Geophys. Res.-Atmos., 106, 27587-27601, doi:10.1029/2000jd900844, 2001.

10

Cheng, Y., Zheng, G., Wei, C., Mu, Q., Zheng, B., Wang, Z., Gao, M., Zhang, Q., He, K., Carmichael, G., Pöschl, U., and Su, H.: Reactive nitrogen chemistry in aerosol water as a source of sulfate during haze events in China, Sci. Adv., 2, e1601530, doi:10.1126/sciadv.1601530, 2016.

Drougas, E., and Kosmas, A. M.: Computational studies of (HIO₃) isomers and the HO₂+IO reaction pathways, J. Phys. Chem. 15 A, 109, 3887-3892, doi:10.1021/jp044197j, 2005.

Dunning Jr, T. H., Peterson, K. A., and Wilson, A. K.: Gaussian basis sets for use in correlated molecular calculations. X. The atoms aluminum through argon revisited, J. Chem. Phys., 114, 9244-9253, doi:10.1063/1.1367373, 2001.

Dusanter, S., Vimal, D., Stevens, P. S., Volkamer, R., Molina, L. T., Baker, A., Meinardi, S., Blake, D., Sheehy, P., Merten, A., Zhang, R., Zheng, J., Fortner, E. C., Junkermann, W., Dubey, M., Rahn, T., Eichinger, B., Lewandowski, P., Prueger, J., and Holder, H.: Measurements of OH and HOs concentrations during the MCMA-2006 field campaign - Part 2: Model

20 and Holder, H.: Measurements of OH and HO₂ concentrations during the MCMA-2006 field campaign - Part 2: Model comparison and radical budget, Atmos. Chem. Phys., 9, 6655-6675, doi:10.5194/acp-9-6655-2009, 2009.

Ehn, M., Junninen, H., Petäjä, T., Kurtén, T., Kerminen, V.-M., Schobesberger, S., Manninen, H., Ortega, I., Vehkamäki, H., and Kulmala, M.: Composition and temporal behavior of ambient ions in the boreal forest, Atmos. Chem. Phys., 10, 8513-8530, doi:10.5194/acp-10-8513-2010, 2010.

25 Eisele, F. L., Lovejoy, E. R., Kosciuch, E., Moore, K. F., Mauldin III, R. L., Smith, J. N., McMurry, P. H., and Iida, K.: Negative atmospheric ions and their potential role in ion-induced nucleation, J. Geophys. Res.-Atmos., 111, D04305, doi:10.1029/2005jd006568, 2006.

Elm, J., Bilde, M., and Mikkelsen, K. V.: Assessment of Density Functional Theory in Predicting Structures and Free Energies of Reaction of Atmospheric Prenucleation Clusters, J. Chem. Theory Comput., 8, 2071-2077, doi:10.1021/ct300192p, 2012.

30 Elm, J., Bilde, M., and Mikkelsen, K. V.: Assessment of binding energies of atmospherically relevant clusters, Phys. Chem. Phys., 15, 16442-16445, doi:10.1039/c3cp52616j, 2013a.

Elm, J., Bilde, M., and Mikkelsen, K. V.: Influence of Nucleation Precursors on the Reaction Kinetics of Methanol with the OH Radical, J. Phys. Chem. A, 117, 6695-6701, doi:10.1021/jp4051269, 2013b.

Enghoff, M., and Svensmark, H.: The role of atmospheric ions in aerosol nucleation-a review, Atmos. Chem. Phys., 8, 4911-4923, doi:10.5194/acp-8-4911-2008, 2008.

Enghoff, M. B., Bork, N., Hattori, S., Meusinger, C., Nakagawa, M., Pedersen, J. O. P., Danielache, S., Ueno, Y., Johnson, M. S., Yoshida, N., and Svensmark, H.: An isotopic analysis of ionising radiation as a source of sulphuric acid, Atmos. Chem. Phys., 12, 5319-5327, doi:10.5194/acp-12-5319-2012, 2012.

Fahey, D., Böhringer, H., Fehsenfeld, F., and Ferguson, E.: Reaction rate constants for $O_2^{-}(H_2O)_n$ ions n=0 to 4, with O_3 , NO, SO₂, and CO₂, J. Chem. Phys., 76, 1799-1805, doi:10.1063/1.443220, 1982.

Fehsenfeld, F., and Ferguson, E.: Laboratory studies of negative ion reactions with atmospheric trace constituents, J. Chem. Phys., 61, 3181-3193, doi:10.1063/1.1682474, 1974.

- 10 Franchin, A., Ehrhart, S., Leppä, J., Nieminen, T., Gagné, S., Schobesberger, S., Wimmer, D., Duplissy, J., Riccobono, F., Dunne, E. M., Rondo, L., Downard, A., Bianchi, F., Kupc, A., Tsagkogeorgas, G., Lehtipalo, K., Manninen, H. E., Almeida, J., Amorim, A., Wagner, P. E., Hansel, A., Kirkby, J., Kürten, A., Donahue, N. M., Makhmutov, V., Mathot, S., Metzger, A., Petäjä, T., Schnitzhofer, R., Sipilä, M., Stozhkov, Y., Tome, A., Kerminen, V. M., Carslaw, K., Curtius, J., Baltensperger, U., and Kulmala, M.: Experimental investigation of ion-ion recombination under atmospheric conditions, Atmos. Chem. Phys.,
- 15, 7203-7216, doi:10.5194/acp-15-7203-2015, 2015. 15

Fukui, K.: The path of chemical reactions-the IRC approach, Acc. Chem. Res., 14, 363-368, doi:10.1021/ar00072a001, 1981.

Georgievskii, Y., and Klippenstein, S. J.: Long-range transition state theory, J. Chem. Phys., 122, doi:10.1063/1.1899603, 2005.

Greenwald, E. E., North, S. W., Georgievskii, Y., and Klippenstein, S. J.: A two transition state model for radical-molecule 20 reactions: A case study of the addition of OH to C_2H_4 , J. Phys. Chem. A, 109, 6031-6044, doi:10.1021/jp058041a, 2005.

Harris, E., Sinha, B., Hoppe, P., Crowley, J. N., Ono, S., and Foley, S.: Sulfur isotope fractionation during oxidation of sulfur dioxide: gas-phase oxidation by OH radicals and aqueous oxidation by H_2O_2 , O_3 and iron catalysis, Atmos. Chem. Phys., 12, 407-423, doi:10.5194/acp-12-407-2012, 2012.

Harris, E., Sinha, B., van Pinxteren, D., Tilgner, A., Fomba, K. W., Schneider, J., Roth, A., Gnauk, T., Fahlbusch, B., Mertes, 25 S., Lee, T., Collett, J., Foley, S., Borrmann, S., Hoppe, P., and Herrmann, H.: Enhanced Role of Transition Metal Ion Catalysis During In-Cloud Oxidation of SO₂, Science, 340, 727-730, doi:10.1126/science.1230911, 2013.

Hegg, D. A., Majeed, R., Yuen, P. F., Baker, M. B., and Larson, T. V.: The impacts of SO₂ oxidation in cloud drops and in haze particles on aerosol light scattering and CCN activity, Geophys. Res. Let., 23, 2613-2616, doi:10.1029/96gl02419, 1996.

Henkelman, G., Arnaldsson, A., and Jónsson, H.: A fast and robust algorithm for Bader decomposition of charge density, 30 Comput. Mater. Sci., 36, 354-360, doi:10.1016/j.commatsci.2005.04.010, 2006.

Hirsikko, A., Nieminen, T., Gagné, S., Lehtipalo, K., Manninen, H. E., Ehn, M., Hörrak, U., Kerminen, V. M., Laakso, L., McMurry, P. H., Mirme, A., Mirme, S., Petäjä, T., Tammet, H., Vakkari, V., Vana, M., and Kulmala, M.: Atmospheric ions and nucleation: a review of observations, Atmos. Chem. Phys., 11, 767-798, doi:10.5194/acp-11-767-2011, 2011.

⁵

Holland, F., Hofzumahaus, A., Schafer, R., Kraus, A., and Patz, H. W.: Measurements of OH and HO₂ radical concentrations and photolysis frequencies during BERLIOZ, J. Geophys. Res.-Atmos., 108, D4, 8246, doi:10.1029/2001jd001393, 2003.

Hoyle, C. R., Fuchs, C., Järvinen, E., Saathoff, H., Dias, A., El Haddad, I., Gysel, M., Coburn, S. C., Tröstl, J., Bernhammer, A. K., Bianchi, F., Breitenlechner, M., Corbin, J. C., Craven, J., Donahue, N. M., Duplissy, J., Ehrhart, S., Frege, C., Gordon,

- 5 H., Höppel, N., Heinritzi, M., Kristensen, T. B., Molteni, U., Nichman, L., Pinterich, T., Prévôt, A. S. H., Simon, M., Slowik, J. G., Steiner, G., Tomé, A., Vogel, A. L., Volkamer, R., Wagner, A. C., Wagner, R., Wexler, A. S., Williamson, C., Winkler, P. M., Yan, C., Amorim, A., Dommen, J., Curtius, J., Gallagher, M. W., Flagan, R. C., Hansel, A., Kirkby, J., Kulmala, M., Möhler, O., Stratmann, F., Worsnop, D. R., and Baltensperger, U.: Aqueous phase oxidation of sulphur dioxide by ozone in cloud droplets, Atmos. Chem. Phys., 16, 1693-1712, doi:10.5194/acp-16-1693-2016, 2016.
- 10 Hvelplund, P., Pedersen, J. O. P., Stochkel, K., Enghoff, M. B., and Kurten, T.: Experimental studies of the formation of cluster ions formed by corona discharge in an atmosphere containing SO₂, NH₃, and H₂O, Int. J. Mass Spectrom., 341, 1-6, doi:10.1016/j.ijms.2013.03.001, 2013.

Kirkby, J., Curtius, J., Almeida, J., Dunne, E., Duplissy, J., Ehrhart, S., Franchin, A., Gagné, S., Ickes, L., Kürten, A., Kupc, A., Metzger, A., Riccobono, F., Rondo, L., Schobesberger, S., Tsagkogeorgas, G., Wimmer, D., Amorim, A., Bianchi, F.,

- 15 Breitenlechner, M., David, A., Dommen, J., Downard, A., Ehn, M., Flagan, R. C., Haider, S., Hansel, A., Hauser, D., Jud, W., Junninen, H., Kreissl, F., Kvashin, A., Laaksonen, A., Lehtipalo, K., Lima, J., Lovejoy, E. R., Makhmutov, V., Mathot, S., Mikkila, J., Minginette, P., Mogo, S., Nieminen, T., Onnela, A., Pereira, P., Petäjä, T., Schnitzhofer, R., Seinfeld, J. H., Sipilä, M., Stozhkov, Y., Stratmann, F., Tomé, A., Vanhanen, J., Viisanen, Y., Vrtala, A., Wagner, P. E., Walther, H., Weingartner, E., Wex, H., Winkler, P. M., Carslaw, K. S., Worsnop, D. R., Baltensperger, U., and Kulmala, M.: Role of sulphuric acid,
- ammonia and galactic cosmic rays in atmospheric aerosol nucleation, Nature, 476, 429-433, doi:10.1038/nature10343, 2011.

Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon, M., Yan, C., Almeida, J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim, A., Bernhammer, A.-K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A., Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kürten, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Peräkylä, O., Piel, F., Petäjä, T., Praplan, A. P., Pringle, K., Rap, A., Richards, N. A. D., Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipilä, M., Steiner, G., Stozhkov, Y., Stratmann, F., Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A. C., Wagner, P. E., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue, N. M., Worsnop, D. R., Baltensperger, U., Kulmala, M., Carslaw, K. S., and Curtius, J.: Ion-induced nucleation of pure biogenic particles, Nature, 533, 521, doi:10.1038/nature17052.2016

30 doi:10.1038/nature17953, 2016.

Klippenstein, S. J., Khundkar, L. R., Zewail, A. H., and Marcus, R. A.: Application of unimolecular reaction-rate theory for highly flexible transition-states to the dissociation of NCNO into NC and NO, J. Chem. Phys., 89, 4761-4770, doi:10.1063/1.455670, 1988.

Kozuch, S., and Martin, J. M. L.: "Turning Over" Definitions in Catalytic Cycles, Acs Catalysis, 2, 2787-2794, doi:10.1021/cs3005264, 2012.

Kuang, C., McMurry, P. H., McCormick, A. V., and Eisele, F. L.: Dependence of nucleation rates on sulfuric acid vapor concentration in diverse atmospheric locations, J. Geophys. Res.-Atmos., 113, D10209, doi:10.1029/2007jd009253, 2008.

Kulmala, M., Pirjola, L., and Mäkelä, J. M.: Stable sulphate clusters as a source of new atmospheric particles, Nature, 404, 66-69, doi:10.1038/35003550, 2000.

Kupiainen-Määttä, O., Olenius, T., Kurtén, T., and Vehkamäki, H.: CIMS sulfuric acid detection efficiency enhanced by amines due to higher dipole moments: a computational study, J. Phys. Chem. A, 117, 14109-14119, doi:10.1021/jp4049764, 2013.

Larson, L. J., Kuno, M., and Tao, F.-M.: Hydrolysis of sulfur trioxide to form sulfuric acid in small water clusters, J. Chem. Phys., 112, 8830-8838, doi:10.1063/1.481532, 2000.

Le Breton, M., McGillen, M. R., Muller, J. B. A., Bacak, A., Shallcross, D. E., Xiao, P., Huey, L. G., Tanner, D., Coe, H., and Percival, C. J.: Airborne observations of formic acid using a chemical ionization mass spectrometer, Atmos. Meas. Tech., 5, 3029-3039, doi:10.5194/amt-5-3029-2012, 2012.

10

5

20

25

Manninen, H. E., Nieminen, T., Asmi, E., Gagné, S., Häkkinen, S., Lehtipalo, K., Aalto, P., Vana, M., Mirme, A., Mirme, S., Hõrrak, U., Plass-Dülmer, C., Stange, G., Kiss, G., Hoffer, A., Törő, N., Moerman, M., Henzing, B., de Leeuw, G., Brinkenberg, M., Kouvarakis, G. N., Bougiatioti, A., Mihalopoulos, N., O'Dowd, C., Ceburnis, D., Arneth, A., Svenningsson, B., Swietlicki, E., Tarozzi, L., Decesari, S., Facchini, M. C., Birmili, W., Sonntag, A., Wiedensohler, A., Boulon, J., Sellegri,

K., Laj, P., Gysel, M., Bukowiecki, N., Weingartner, E., Wehrle, G., Laaksonen, A., Hamed, A., Joutsensaari, J., Petäjä, T., 15 Kerminen, V. M., and Kulmala, M.: EUCAARI ion spectrometer measurements at 12 European sites – analysis of new particle formation events, Atmos. Chem. Phys., 10, 7907-7927, doi:10.5194/acp-10-7907-2010, 2010.

Mauldin III, R. L., Berndt, T., Sipilä, M., Paasonen, P., Petäjä, T., Kim, S., Kurtén, T., Stratmann, F., Kerminen, V.-M., and Kulmala, M.: A new atmospherically relevant oxidant of sulphur dioxide, Nature, 488, 193-196, doi:10.1038/nature11278, 2012.

Möhler, O., Reiner, T., and Arnold, F.: The formation of SO₅⁻ by gas phase ion-molecule reactions, J. Chem. Phys., 97, 8233-8239, doi:10.1063/1.463394, 1992.

Nagato, K., Kim, C. S., Adachi, M., and Okuyama, K.: An experimental study of ion-induced nucleation using a drift tube ion, mobility spectrometer/mass spectrometer and a cluster-differential mobility analyzer/Faraday cup electrometer, J. Aerosol Sci., 36, 1036-1049, doi:10.1016/j.jaerosci.2004.12.006, 2005.

Nieminen, T., Manninen, H., Sihto, S.-L., Yli-Juuti, T., Mauldin III, R. L, Petäjä, T., Riipinen, I., Kerminen, V.-M., and Kulmala, M.: Connection of sulfuric acid to atmospheric nucleation in boreal forest, Environ. Sci. Technol., 43, 4715-4721, doi: 10.1021/es803152j, 2009.

Noziere, B., Ekstrom, S., Alsberg, T., and Holmstrom, S.: Radical-initiated formation of organosulfates and surfactants in atmospheric aerosols, Geophys. Res. Lett., 37, doi:10.1029/2009gl041683, 2010. 30

Peng, C., Ayala, P. Y., Schlegel, H. B., and Frisch, M. J.: Using redundant internal coordinates to optimize equilibrium geometries and transition states, J. Comput. Chem., 17, 49-56, 1996.

Purvis, G. D., and Bartlett, R. J.: A full coupled-cluster singles and doubles model - the inclusion of disconnected triples, J. Chem. Phys., 76, 1910-1918, doi:10.1063/1.443164, 1982.

Rothe, E. W., Tang, S. Y., and Reck, G. P.: Measurement of electron affinities of O₃, SO₂, and SO₃ by collisional ionization, J. Chem. Phys., 62, 3829-3831, doi:10.1063/1.430941, 1975.

Schindelka, J., Iinuma, Y., Hoffmann, D., and Herrmann, H.: Sulfate radical-initiated formation of isoprene-derived organosulfates in atmospheric aerosols, Faraday Discuss., 165, 237-259, doi:10.1039/c3fd00042g, 2013.

5 Schweitzer, C., and Schmidt, R.: Physical mechanisms of generation and deactivation of singlet oxygen, Chem. Rev., 103, 1685-1757, doi:10.1021/cr010371d, 2003.

Seinfeld, J. H., and Pandis, S. N.: Atmospheric Chemistry and Physics: From Air Pollution to Climate Change, 3rd Edition, John Wiley & Sons, Inc., 2016.

Sipilä, M., Berndt, T., Petäjä, T., Brus, D., Vanhanen, J., Stratmann, F., Patokoski, J., Mauldin III, R. L., Hyvarinen, A.-P.,
Lihavainen, H., and Kulmala, M.: The Role of Sulfuric Acid in Atmospheric Nucleation, Science, 327, 1243-1246, doi:10.1126/science.1180315, 2010.

Su, T., and Chesnavich, W. J.: Parametrization of the ion-polar molecule collision rate-constant by trajectory calculations, J. Chem. Phys., 76, 5183-5185, doi:10.1063/1.442828, 1982.

Surratt, J. D., Lewandowski, M., Offenberg, J. H., Jaoui, M., Kleindienst, T. E., Edney, E. O., and Seinfeld, J. H.: Effect of
acidity on secondary organic aerosol formation from isoprene, Environ. Sci. Technol., 41, 5363-5369, doi:10.1021/es0704176, 2007.

Surratt, J. D., Gomez-Gonzalez, Y., Chan, A. W. H., Vermeylen, R., Shahgholi, M., Kleindienst, T. E., Edney, E. O., Offenberg, J. H., Lewandowski, M., Jaoui, M., Maenhaut, W., Claeys, M., Flagan, R. C., and Seinfeld, J. H.: Organosulfate formation in biogenic secondary organic aerosol, J. Phys. Chem. A, 112, 8345-8378, doi:10.1021/jp802310p, 2008.

20 Svensmark, H., Pedersen, J. O. P., Marsh, N. D., Enghoff, M. B., and Uggerhoj, U. I.: Experimental evidence for the role of ions in particle nucleation under atmospheric conditions, Proc. R. Soc. A, 463, 385-396, doi:10.1098/rspa.2006.1773, 2007.

Tang, W., Sanville, E., and Henkelman, G.: A grid-based Bader analysis algorithm without lattice bias, J. Phys. Condens. Matter, 21, doi:10.1088/0953-8984/21/8/084204, 2009.

Tsona, N., Bork, N., and Vehkamäki, H.: Exploring the chemical fate of the sulfate radical anion by reaction with sulfur dioxide in the gas phase, Atmos. Chem. Phys., 15, 495-503, doi:10.5194/acp-15-495-2015, 2015.

Tsona, N. T., Bork, N., and Vehkamäki, H.: On the gas-phase reaction between SO_2 and $O_2^-(H_2O)_{0-3}$ clusters – an ab initio study, Phys. Chem. Chem. Phys., 16, 5987-5992, doi:10.1039/c3cp54715a, 2014.

Tsona, N. T., Bork, N., Loukonen, V., and Vehkamäki, H.: A Closure Study of the Reaction between Sulfur Dioxide and the Sulfate Radical Ion from First-Principles Molecular Dynamics Simulations, J. Phys. Chem. A, 120, 1046-1050, doi: 10.1021/acs.jpca.5b12395, 2016.

30

Tsona, N. T., Li, J., and Du, L.: From O_2^- -Initiated SO₂ oxidation to sulfate formation in the gas-phase, J. Phys. Chem. A, 122, 5781-5788, doi:10.1021/acs.jpca.8b03381, 2018.

Vereecken, L., Harder, H., and Novelli, A.: The reaction of Criegee intermediates with NO, RO₂, and SO₂, and their fate in the atmosphere, Phys. Chem. Chem. Phys., 14, 14682-14695, doi:10.1039/c2cp42300f, 2012.

Wagner, R., Yan, C., Lehtipalo, K., Duplissy, J., Nieminen, T., Kangasluoma, J., Ahonen, L. R., Dada, L., Kontkanen, J., Manninen, H. E., Dias, A., Amorim, A., Bauer, P. S., Bergen, A., Bernhammer, A. K., Bianchi, F., Brilke, S., Mazon, S. B.,

- 5 Chen, X., Draper, D. C., Fischer, L., Frege, C., Fuchs, C., Garmash, O., Gordon, H., Hakala, J., Heikkinen, L., Heinritzi, M., Hofbauer, V., Hoyle, C. R., Kirkby, J., Kürten, A., Kvashnin, A. N., Laurila, T., Lawler, M. J., Mai, H., Makhmutov, V., Mauldin Iii, R. L., Molteni, U., Nichman, L., Nie, W., Ojdanic, A., Onnela, A., Piel, F., Quéléver, L. L. J., Rissanen, M. P., Sarnela, N., Schallhart, S., Sengupta, K., Simon, M., Stolzenburg, D., Stozhkov, Y., Tröstl, J., Viisanen, Y., Vogel, A. L., Wagner, A. C., Xiao, M., Ye, P., Baltensperger, U., Curtius, J., Donahue, N. M., Flagan, R. C., Gallagher, M., Hansel, A.,
- 10 Smith, J. N., Tomé, A., Winkler, P. M., Worsnop, D., Ehn, M., Sipilä, M., Kerminen, V. M., Petäjä, T., and Kulmala, M.: The role of ions in new particle formation in the CLOUD chamber, Atmos. Chem. Phys., 17, 15181-15197, doi:10.5194/acp-17-15181-2017, 2017.

Wang, G., Zhang, R., Gomez, M. E., Yang, L., Levy Zamora, M., Hu, M., Lin, Y., Peng, J., Guo, S., Meng, J., Li, J., Cheng, C., Hu, T., Ren, Y., Wang, Y., Gao, J., Cao, J., An, Z., Zhou, W., Li, G., Wang, J., Tian, P., Marrero-Ortiz, W., Secrest, J., Du,

15 Z., Zheng, J., Shang, D., Zeng, L., Shao, M., Wang, W., Huang, Y., Wang, Y., Zhu, Y., Li, Y., Hu, J., Pan, B., Cai, L., Cheng, Y., Ji, Y., Zhang, F., Rosenfeld, D., Liss, P. S., Duce, R. A., Kolb, C. E., and Molina, M. J.: Persistent sulfate formation from London Fog to Chinese haze, Proc. Natl. Acad. Sci., 113, 13630, doi:10.1073/pnas.1616540113, 2016.

Welz, O., Savee, J. D., Osborn, D. L., Vasu, S. S., Percival, C. J., Shallcross, D. E., and Taatjes, C. A.: Direct kinetic measurements of Criegee intermediate (CH₂OO) formed by reaction of CH₂I with O₂, Science, 335, 204-207, doi:10.1126/science.1213229, 2012.

20

35

Yan, C., Dada, L., Rose, C., Jokinen, T., Nie, W., Schobesberger, S., Junninen, H., Lehtipalo, K., Sarnela, N., Makkonen, U., Garmash, O., Wang, Y., Zha, Q., Paasonen, P., Bianchi, F., Sipilä, M., Ehn, M., Petäjä, T., Kerminen, V. M., Worsnop, D. R., and Kulmala, M.: The role of H₂SO₄-NH₃ anion clusters in ion-induced aerosol nucleation mechanisms in the boreal forest, Atmos. Chem. Phys., 18, 13231-13243, doi:10.5194/acp-18-13231-2018, 2018.

25 Yu, F.: From molecular clusters to nanoparticles: second-generation ion-mediated nucleation model, Atmos. Chem. Phys., 6, 5193-5211, doi:10.5194/acp-6-5193-2006, 2006.

Yu, F. Q., and Turco, R. P.: Ultrafine aerosol formation via ion-mediated nucleation, Geophys. Res. Lett., 27, 883-886, doi:10.1029/1999gl011151, 2000.

Yu, F. Q., and Turco, R. P.: From molecular clusters to nanoparticles: Role of ambient ionization in tropospheric aerosol formation, J. Geophys. Res.-Atmos., 106, 4797-4814, doi:10.1029/2000jd900539, 2001.

Yu, T., Zhao, D., Song, X., and Zhu, T.: NO₂-initiated multiphase oxidation of SO₂ by O₂ on CaCO₃ particles, Atmos. Chem. Phys., 18, 6679-6689, doi:10.5194/acp-18-6679-2018, 2018.

Zhang, H., Chen, S., Zhong, J., Zhang, S., Zhang, Y., Zhang, X., Li, Z., and Zeng, X. C.: Formation of aqueous-phase sulfate during the haze period in China: Kinetics and atmospheric implications, Atmos. Env., 177, 93-99, doi:10.1016/j.atmosenv.2018.01.017, 2018.

Zhao, D., Song, X., Zhu, T., Zhang, Z., Liu, Y., and Shang, J.: Multiphase oxidation of SO₂ by NO₂ on CaCO₃ particles, Atmos. Chem. Phys., 18, 2481-2493, doi:10.5194/acp-18-2481-2018, 2018.

Zhao, Y., and Truhlar, D. G.: The M06 suite of density functionals for main group thermochemistry, thermochemical kinetics, noncovalent interactions, excited states, and transition elements: two new functionals and systematic testing of four M06-class functionals and 12 other functionals, Theor. Chem. Account., 120, 215-241, doi:10.1007/s00214-007-0310-x, 2008.

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Figure 1: Optimized structures of the most stable intermediates in the $O_2SOO^2 + O_3$ reaction (a) in the absence and (b) in the presence of a single water molecule. Optimizations were performed at the UM06-2X/aug-cc-pVTZ level of theory. Lengths (in Å) of some descriptive bonds are indicated. The color coding is yellow for sulfur, red for oxygen and white for hydrogen.



Figure 2: Formation Gibbs free energies of the most stable intermediate species in the O₂SOO⁻ + O₃ reaction in the absence and in the presence of water. "W" is the shorthand notation for water. RC1, RC2, TS2, RCW1, RCW2, and TSW2 structures are shown in

5 Fig. 1. P1 = O₂ + SO₂ + O₃⁻, P2 = SO₃⁻ + 2O₂, PW1 = O₂ + SO₂ + O₃⁻ + H₂O and PW2 = SO₃⁻ ····H₂O + 2O₂. Calculations were performed at the UCCSD(T)/aug-cc-pVTZ//UM06-2X/aug-cc-pVTZ level of theory.



Figure 3: Representation of the spin density (in blue color) on intermediate structures in the O₂SOO⁻ + O₃ reaction. The spin density clearly indicates that the extra electron is progressively distributed over all the atoms from (a) the pre-reactive complex through (b) the transition state to (c) the product complex.

Table 1: Electronic energies (ΔE), enthalpies (ΔH_{298K}) and Gibbs free energies (ΔG_{298K}) of the different states in the O₂SOO⁺ + O₃ reaction both in the absence and in the presence of water, calculated relative to the energy of initial reactants at the UCCSD(T)/aug-cc-pVTZ//UM06-2X/aug-cc-pVTZ level of theory.

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Species	ΔE	$\Delta H_{298\mathrm{K}}$	$\varDelta G_{298\mathrm{K}}$
Unhydrated reaction			
$O_2SOO^2 + O_3$	0	0	0
$RC1 = O_3 \cdots O_2 SOO^-$	-5.1	-3.9	4.5
$RC2 = O_2 \cdots O_2 S - O_3^-$	-21.9	-21.0	-14.7
TS2	-11.6	-11.9	-4.6
$SO_3^- + 2O_2$	-35.8	-36.4	-45.3
$O_2 + SO_2 + O_3^-$	-1.7	-3.2	-13.6
Monohydrated reaction			
O_2SOO ····· $H_2O + O_3$	0	0	0
$RCW1 = O_3 \cdots O_2 SOO^{-} \cdots H_2 O$	-4.6	-3.4	4.7
$RCW2 = O_2 \cdots O_2 S - O_3 \cdots H_2 O$	-21.9	-20.9	-12.4
TSW2	-25.3	-25.4	-16.4
SO_3 -····H ₂ O + 2O ₂	-36.7	-37.4	-46.6
$O_2 + SO_2 + O_3 + H_2O$	10.3	7.1	-12.1