

1 **Using flow cytometry and light-induced fluorescence technique to characterize the**  
2 **variability and characteristics of bioaerosols in springtime at Metro Atlanta, Georgia**

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22 **Abstract**

23 The abundance and speciation of primary biological aerosol particles (PBAP) is important for  
24 understanding their impacts on human health, cloud formation and ecosystems. Towards this, we have  
25 developed a protocol for quantifying PBAP collected from large volumes of air with a portable wet-walled  
26 cyclone bioaerosol sampler. A flow cytometry (FCM) protocol was then developed to quantify and  
27 characterize the PBAP populations from the sampler, which were confirmed against epifluorescence  
28 microscopy. The sampling system and FCM analysis were used to study PBAP in Atlanta, GA over a two-  
29 month period and showed clearly defined populations of nucleic acid containing particles: Low Nucleic  
30 Acid-content particles above threshold (LNA-AT), and High Nucleic Acid-content particles (HNA) likely  
31 containing wet-ejected fungal spores, and pollen. We find that daily-average springtime PBAP  
32 concentration (1 to 5 $\mu$ m diameter) ranged between  $1.4 \times 10^4$  and  $1.1 \times 10^5$  m<sup>-3</sup>. The LNA-AT population  
33 dominated PBAP during dry days ( $72 \pm 18\%$ ); HNA dominated the PBAP during humid days and following  
34 rain events, where HNA comprised up to 92% of the PBAP number. Concurrent measurements with a  
35 Wideband Integrated Bioaerosol Sensor (WIBS-4A) showed that FBAP and total FCM counts are similar;

36 HNA (from FCM) moderately correlated with ABC type FBAP concentrations throughout the sampling  
37 period (and for the same particle size range, 1-5  $\mu\text{m}$  diameter). However, the FCM LNA-AT population,  
38 possibly containing bacterial cells, did not correlate with any FBAP type. The lack of correlation of any  
39 WIBS FBAP type with the LNA-AT suggest airborne bacterial cells may be more difficult to  
40 unambiguously detect with autofluorescence than currently thought. Identification of bacterial cells even in  
41 the FCM (LNA-AT population) is challenging, given that the fluorescence level of stained cells at times  
42 may be comparable to that seen from abiotic particles. HNA and ABC displayed highest concentration on  
43 a humid and warm day after a rain event (4/14), suggesting that both populations correspond to wet-ejected  
44 fungal spores. Overall, information from both instruments combined reveals a highly dynamic airborne  
45 bioaerosol community over Atlanta, with a considerable presence of fungal spores during humid days, and  
46 LNA-AT population dominating bioaerosol community during dry days.

## 47 **Introduction**

48 Primary biological aerosol particles (PBAP), also called bioaerosols, are comprised of airborne  
49 microbial cells (e.g. bacteria, diatoms), reproductive entities (e.g. pollen, fungal spores), viruses and  
50 biological fragments. Bioaerosols are ubiquitous, with potentially important impacts on human health,  
51 cloud formation, precipitation, and biogeochemical cycles (Pöschl, 2005; Hoose et al., 2010; DeLeon-  
52 Rodriguez et al., 2013; Morris et al., 2014; Longo et al., 2014; Fröhlich-Nowoisky et al., 2016;  
53 Myriokefalitakis et al., 2016). Despite their low number concentration relative to abiotic particles, PBAP  
54 possess unique functional and compositional characteristics that differentiate them from abiotic aerosol.  
55 For example, certain PBAP constitute the most efficient of atmospheric ice nucleators, affecting the  
56 microphysics of mixed phase clouds and precipitation (Hoose and Möhler, 2012; Sullivan et al., 2017). The  
57 mass and nutrient content of PBAP may suffice to comprise an important supply of bioavailable P to  
58 oligotrophic marine ecosystems (Longo et al., 2014; Myriokefalitakis et al., 2016). In addition, the  
59 concurrence of disease outbreaks during dust storms has been attributed to pathogenic microbes attached  
60 to airborne dust that are subsequently inhaled (Griffin et al., 2003; Ortiz-Martinez et al., 2015; Goudie  
61 2014).

62 Quantification of the concentration and size of PBAP is critical for understanding their environmental  
63 impacts. Measuring PBAP however poses a challenge for established microbiology tools, owing to their  
64 low atmospheric concentration ( $10^3 - 10^6$  cells  $\text{m}^{-3}$  air; Fröhlich-Nowoisky et al., 2016) and wide diversity  
65 of airborne particle types and sizes. For instance, only a fraction of microorganisms (an estimated 5%; Chi  
66 and Li et al., 2007) can be cultured, and cultivation cannot be used to quantify dead organisms, viruses or  
67 fragments, while most culture-independent methods are optimized for more abundant microbial  
68 populations. Epifluorescence microscopy (EPM) is the standard for bioaerosol quantification but is not

69 high-throughput and requires considerable time for quantification of concentration per sample. Flow  
70 cytometry (FCM) is an analysis technique based on the concurrent measurement of light scattering and  
71 fluorescence intensity from single particles (Wang et al., 2010). FCM requires a liquid suspension of  
72 bioparticles that flows through an optical cell and interrogated with a series of laser beams. Each sample is  
73 pretreated with stains that targeting specific macromolecules (e.g. DNA/RNA) which subsequently  
74 fluoresce when excited by the FCM lasers. The resulting scattering and fluorescent light emissions are then  
75 detected by an array of sensors to allow the differentiation of biological and abiotic (e.g. dust) particles  
76 according to the characteristic specific to the stain used. FCM has proved to be as reliable as EPM, but with  
77 the advantage of lower uncertainty, higher quantification efficiency and requiring considerably less time  
78 and effort than EPM per sample (Lange et al., 1997). FCM is frequently used in biomedical research to  
79 quantify eukaryotic cell populations, and in microbiology to quantify a wide variety of yeast and bacterial  
80 cells (Nir et al., 1990; Van Dilla et al., 1983). FCM is also used to study environmental samples, e.g., to  
81 differentiate low nucleic acid (LNA) from high nucleic acid (HNA) phytoplankton in aquatic environments  
82 (Wang Y. et al 2010; Müller et al., 2010). Despite its advantages, FCM has seen little use in the bioaerosol  
83 field to date, owing in part to the challenges associated with collecting sufficient PBAP mass for robust  
84 counting statistics to be obtained (Chen and Li, 2005; Liang et al., 2013). Chen and Li (2005) determined  
85 that for counting purposes, the SYTO-13 nucleic acid stain is the most effective (among five different  
86 nucleic acid stains studied) for determining reliable concentration of bioaerosols.

87 Light Induced Fluorescence (LIF) is an increasingly utilized technique for bioaerosol quantification,  
88 and it relies on measuring the autofluorescence intensity of specific high yield fluorophores (e.g.,  
89 Nicotinamide Adenine Dinucleotide – NADH co-enzyme, flavins and amino acids like Tryptophan and  
90 Tyrosine) present in PBAP. The major advantage of the technique is that it is fully automated, does not  
91 require a liquid suspension (i.e., it directly senses particles suspended in air) and it provides high frequency  
92 measurements (~1 Hz) make it ideal for continuous monitoring and operation in highly variable  
93 environments (e.g., aircraft operation). Particles detected by LIF, called Fluorescent Biological Aerosol  
94 Particles (FBAP), although not equal to PBAP, may still constitute a large fraction of the biological particles  
95 (Healy et al., 2014; Gosselin et al., 2016). Using LIF, FBAP diurnal cycles showing maximum  
96 concentrations during evenings and minimum around middays, especially in heavily vegetated  
97 environments have been observed. This behavior has been related to known temperature and relative  
98 humidity release mechanism of certain fungal spore species (Wu et al., 2007; Gabey et al., 2010; Tropak  
99 and Schnaiter, 2013). Huffman et al. (2010) used a UV-Aerodynamic Particle Sizer (UV-APS) to show that  
100 the concentration and frequency of occurrence of 3µm FBAP particles at Mainz, Germany (semi-urban  
101 environment) exhibited a strong diurnal cycle from August through November: with a first peak at ~  
102  $1.6 \times 10^4 \text{ m}^{-3}$  at mid-morning (6-8 am) followed by a constant profile ( $\sim 2\text{-}4 \times 10^4 \text{ m}^{-3}$ ) throughout the rest of

103 the day. Similar studies in urban and densely vegetated environments suggest a notable difference in the  
104 size distributions, diurnal behavior and FBAP loading between the two environments. Gabey et al., 2011  
105 found that the FBAP in Manchester, UK follow a characteristic bimodal distribution with peaks at 1.2 $\mu\text{m}$   
106 and 1.5 – 3.0  $\mu\text{m}$ . As in Mainz, the concentration of larger particles peaks in the mid-morning, ranges from  
107 0 to 300  $\text{L}^{-1}$ , and the 1.2 $\mu\text{m}$  peak is linked to traffic activity. However, at the Borneo tropical rain forest  
108 FBAP concentrations peak during the evening with a robust 2-3 $\mu\text{m}$  population and concentrations ranging  
109 from 100 to 2000  $\text{L}^{-1}$  (Gabey et al., 2010).

110 LIF-based observations (e.g. UV-APS, WIBS), combined with measurements of molecular tracers (e.g.  
111 mannitol and arabitol) and endotoxin measurements provide a more complete picture of PBAP emissions.  
112 Gosselin et al. (2016) applied this approach during the BEACHON-RoMBAS field campaign. A clear  
113 correlation between FBAP and the molecular markers is seen, indicating an increase of fungal spores during  
114 rain events. FBAP concentrations and molecular marker-inferred (arabitol and mannitol; Bauer et al., 2008  
115 approach) fungal spore concentrations were within the same order of magnitude. WIBS-3 cluster  
116 (determined using Crawford et al., 2015) linked to fungal spores gave concentrations 13% lower than those  
117 derived from molecular marker concentrations during rain events. During dry events, FBAP and molecular  
118 markers derived fungal spore concentrations were poorly correlated. It is currently unknown the degree to  
119 which all types of PBAP are consistently detected by LIF over different time of the year and different  
120 environments; it is likely, however, that for certain classes of bioparticles (e.g., pollen and fungi) the  
121 detection efficiency using LIF is relatively high. However, the low intrinsic fluorescence intensity of  
122 bacteria and high variability of thereof in relation to metabolic state may lead to their misclassification as  
123 non-biological particles (Hernandez et al., 2016).

124 For LIF-based quantification of PBAP to be effective, it requires the intrinsic fluorescence of biological  
125 material to exceed that of non-biological matter. Depending on the type, metabolic state and species, PBAP  
126 autofluorescence may vary orders of magnitude and therefore LIF may not always be able to differentiate  
127 between biological and abiotic particles. For example, Tropak and Schnaiter (2013) showed that laboratory-  
128 generated mineral dust, soot and ammonium sulfate may be misclassified as FBAP. To address  
129 misclassification, Excitation Emission Matrices (EEMs) have been developed for biomolecules (e.g.  
130 tryptophan, tyrosine, riboflavin) and non-biological (e.g. Pyrene, Naphthalene, Humic Acid) molecules.  
131 EEMs provide the wavelength-dependent fluorescence emission spectra as a function of the excitation  
132 wavelength and are used to assign spectral modes to known fluorophores. The structure of EEMs is  
133 important for identifying molecules that are unique to PBAP and allow their identification by LIF; it is this  
134 principle upon which detectors in commercial FBAP measurements (e.g. WIBS, UV-APS) are based upon.  
135 Comparison of EEMs from biological and non-biological molecules show that even when biomolecules

136 have higher autofluorescence intensity than non-biologicals in the LIF detection range, interferences from  
137 non-biological compounds (e.g. polycyclic aromatic hydrocarbons and soot) from combustion emissions  
138 can influence LIF detection (Pöhlker et al., 2012). Considerable work remains on determining which  
139 detector(s) or combination thereof provides an unambiguous identification of bioaerosols and related  
140 subgroups (e.g. bacteria, fungal spores, pollen). Towards this, an aerobiology catalog of pure cultures has  
141 been developed for the WIBS-4 showing that instrument-to-instrument variability in fluorescence detection  
142 poses a considerable challenge, as applying common detection thresholds across instruments leads to  
143 considerable differences in PBAP concentration and composition (Hernandez et al., 2016).

144 Another important issue for LIF-based quantification of PBAP is the impact of atmospheric oxidants,  
145 UV and other stressors on the fluorescence intensity of PBAP. Pan et al. (2014) tested the effect of relative  
146 humidity and ozone exposure in the autofluorescence spectra of octapeptide aerosol particles using an UV-  
147 APS connected to a rotating drum. Octapeptides, organic molecules containing eight amino acids and  
148 present in cells, were used as a proxy to study the aging of tryptophan and results suggest bioaerosols  
149 exposure to typical ozone (~150ppb) concentrations decrease tryptophan fluorescence intensity and PBAP  
150 detection. Multiple stressors can be affecting bioaerosols LIF detection so such issues need to be thoroughly  
151 explored to understand PBAP detection efficiency over the wide range of atmospheric conditions and PBAP  
152 population composition (Toprak and Schnaiter, 2013; Hernandez et al., 2016).

153 The aims of the study are to (i) develop an effective and reliable FCM detection and quantification  
154 protocol for bioaerosol; (ii) apply the protocol to understand bioaerosol populations and their variability  
155 during different meteorological conditions, and, (iii) compare FCM and WIBS-4A results to have a better  
156 understanding of PBAP day-to-day variability. To our knowledge, this study is the first to develop a FCM  
157 protocol to identify and quantify well-defined speciated bioaerosols populations from samples collected  
158 from a modified state-of-the-art biosampler. LIF sampling of bioaerosol side-by-side with established and  
159 quantitative biology tools (FCM and EPM) was conducted to assess the LIF detection capabilities toward  
160 different bioaerosol populations and under atmospherically-relevant conditions during this study. Atlanta  
161 is selected as a case study for PBAP sampling, as it provides a highly populated urban environment  
162 surrounded by vast vegetative areas; this and the broad range of temperature and humidity ensures a wide  
163 range of PBAP population composition, state and concentrations. All the samples collected are compared  
164 side-by-side to concurrent WIBS-4A data collected over the same time period.

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## 168 2. Instrumentation and Methodology

### 169 2.1 Bioaerosol Sampler

170 Sampling was performed using the SpinCon II (InnovaPrep LLC, Inc.) portable wet-walled cyclone  
171 aerosol sampler. Aerosol is collected by inertial impaction with a recirculating liquid film in the cyclone;  
172 evaporative losses are compensated so that the sample volume is maintained constant during a sample cycle.  
173 The particle collection efficiency for 1 $\mu$ m, 3 $\mu$ m, 3.5 $\mu$ m and 5.0 $\mu$ m particles is about 47.3 $\pm$ 2.1%,  
174 56.1 $\pm$ 3.9%, 14.6  $\pm$  0.6 and 13.8  $\pm$  2.2%, respectively (Kesavan et al., 2015). However, the experiments  
175 conducted using 1 $\mu$ m PSL and 3 $\mu$ m PSL, 3.5 $\mu$ m oleic acid and 5.0 $\mu$ m oleic acid particles not necessarily  
176 quantify the collection efficiency of biological particles in this size range. Even with a lower collection  
177 efficiency than any impingement sampler, SpinCon effectively collects larger amounts of biological  
178 particles owing to its high volumetric flow rate, which is a considerable advantage (Kesavan et al., 2015).  
179 However, the stress caused by the high flow rate of the SpinCon may affect cell viability. Santl-Temkiv et  
180 al. (2017) recently studied the SpinCon retention efficiency from sea water samples and for *P.agglomerans*  
181 populations from pure cultures ( $\sim 10^5$  cells mL<sup>-1</sup>). After 1 hour of sampling, the SpinCon was found to  
182 retain 20.6 $\pm$ 5.8% of the *P.agglomerans* concentration and 55.3 $\pm$ 2.1% of the sea water microbial  
183 concentration.

184 In our study, the biosampler was run at 478L min<sup>-1</sup> for 4hr sampling cycles. Phosphate-buffered saline  
185 (PBS) 1X pH 7.4 solution was used and the instrument compensated for water evaporation by supplying  
186 Milli-Q water to maintain the PBS concentration constant. Upon termination of each sampling cycle, the  
187 instrument was programmed to dispense the sample in a 15mL centrifuge tube. Then, 10 $\mu$ l of formalin (37  
188 wt.% formaldehyde) per mL of solution was added to every sample for preservation and samples were  
189 stored at 4°C. Given the long sampling times and the low concentration of PBAP, the fluid supply system  
190 of the instrument was modified and a cleaning protocol (CP) has been developed, which is described below.

191 The SpinCon II water and PBS supply bags used in the commercial instrument were replaced by two  
192 2L autoclavable Nalgene bottles (Thermo Scientific Inc.) with antimicrobial tubing, connectors and a small  
193 HEPA filter connected to vent and prevent coarse and submicron particles contamination (Figure 1). Bottles  
194 were autoclaved and filled with Milli-Q water and PBS, beforehand sterilized with 0.2 $\mu$ m pore bottle top  
195 filters (Thermo Fisher Inc.) and transferred inside a biosafety cabinet. An aliquot of each fluid obtained  
196 after preparation was evaluated for sterility by EPM and FCM.

197 The cleaning protocol (CP) of the biosampling system consists of two phases. During phase one, all  
198 acrylic windows and the outside of the collector/concentrator were cleaned with ethanol 70 wt. %. Then,  
199 the instrument inlet, outlet, and the inside of the collector/concentrator was cleaned with ethanol 70 wt. %.  
200 In the second phase, the SpinCon II inlet was connected to a HEPA filter to provide a particle-free source

201 of air to the sampling system; the instrument was then washed with ethanol 70 wt.%, 10 wt.% bleach  
202 solution, PBS and Milli-Q H<sub>2</sub>O, respectively. The wash consisted of a rinse, a 2 minutes sample and filling  
203 the instrument collector/concentrator with the fluid in use (i.e., bleach solution, ethanol, PBS and Milli-Q  
204 H<sub>2</sub>O). The collector/concentrator was drained after 1 minute. The above were repeated for the remaining  
205 fluids, taking 5 minutes per fluid. Overall, the CP requires 45 minutes; upon completion, a blank is obtained  
206 to constrain the residual contamination levels after cleaning (described below). Finally, the HEPA filter  
207 was disconnected, instrument inlets and outlets were sealed and the inlet tube was cleaned with ethanol 70  
208 wt.% to be ready for rooftop sampling. SpinCon II was rinsed with ethanol 70wt.% after each sampling  
209 episode and the cleaning protocol was applied before each sample.

210 Several blanks were obtained to quantify the levels of PBAP contamination in the fluids and sampler,  
211 and to ensure that they were sufficiently low to not bias the detection, identification and quantification of  
212 the PBAP. Furthermore, an instrument blank was obtained after a CP to constrain residual particles, by  
213 running the sampler for 2 minutes, while sampling air with a HEPA filter connected to the inlet of the  
214 SpinCon II. Another blank was collected to characterize any contamination of biological particles from the  
215 supply of PBS and water in the SpinCon II. This was done by operating the SpinCon II for a 4hr period  
216 with a HEPA filter connected to the inlet which completely cleans the air entering the wet cyclone from  
217 any bioparticles. All blanks were analyzed directly via FCM (Sect. 2.3) and EPM.

218 The volumetric flow rate within the SpinCon II was routinely calibrated by a VT100 Hotwire Thermo-  
219 anemometer (Cole Palmer Inc.) using a 3-hole round duct transverse approach. A 1 ¼" OD tube with the  
220 same diameter as the SpinCon II inlet was designed with 3 holes. Each hole was 60° apart from the other  
221 and the holes were perpendicular to the axial air flow direction of the tube. (Supplementary Information,  
222 Figure S1). Triplicates of flow rate measurements were taken in each hole at the center of the tube and  
223 averaged to determine SpinCon II volumetric flow rate ( $478.0 \pm 6.4 \text{ L min}^{-1}$ ).

## 224 **2.2 Flow Cytometry**

225 During this study, a BD Accuri C6 flow cytometer (BD Bioscience Inc.) was used for Flow Cytometry.  
226 The instrument quantifies suspended cells in aqueous medium at three flow velocity modes (slow, medium  
227 and fast flow at 14, 35 and 66  $\mu\text{L min}^{-1}$ , respectively). It excites particles with a 488nm laser and possesses  
228 four fluorescence detectors: FL1 (533±30nm), FL2 (585±40nm), FL3 (> 670nm) and FL4 (675±25nm),  
229 which make it possible to analyze the fluorescence from multiple dyes concurrently. In this study, 2.5  $\mu\text{M}$   
230 SYTO-13 nucleic acid probe was added to the fixed samples and incubated for 15min in the dark at room  
231 temperature to stain biological particles. Additionally, 10 $\mu\text{L}$  of 15 $\mu\text{m}$  polystyrene bead suspension was  
232 added to the 1mL total volume samples as an internal standard for PBAP concentration and size  
233 quantification. The BD Accuri C6 was cleansed before each use with 0.2 $\mu\text{m}$  filtered Milli-Q water in fast

234 mode for 10min; background particle counts were typically reduced to  $1\mu\text{L}^{-1}$ . At the beginning of every  
235 experiment, a 1mL blank of the atmospheric sample without SYTO-13 and beads was analyzed, used in  
236 quantification calculations (Sect. 3.1). Each sample was run in slow mode for 5min. After each sample, the  
237 instrument was flushed with  $0.2\mu\text{m}$  filtered Milli-Q water in slow flow for 1 minute (important for robust  
238 quantification of the typically low concentrations of the atmospheric samples). SYTO-13 fluorescence  
239 intensity was quantified by the FL1-A detector and used in combination with other parameters (FSC-A &  
240 SSC-A) to constrain the PBAP populations present. FSC-A measured forward ( $0^\circ \pm 13^\circ$ ) scattering and is  
241 used to characterize the size of particles; SSC-A measured the side ( $90^\circ \pm 13^\circ$ ) scattering and is used to  
242 characterize the internal complexity of particles. The SSC-A scattering intensity is a function of the cellular  
243 granularity or density of the internal structures (e.g. nucleus, mitochondria, ribosomes), the sphericity and  
244 size of the particles. Compared to spherical particles of the same size, elongated particles tend to yield a  
245 broader distribution of side scattering intensities (Mage et al., 2019; Mathaes et al., 2013). Although side  
246 scattering intensity increases with particle size, it has not been commonly used to measure cell size (Tzur  
247 et al., 2011). Overall, SSC-A scattering intensity will be proportional to the amount scattering caused by  
248 the internal structures and the cell membrane, which ultimately depends on the refractive index of each cell  
249 (Muller et al., 2010). Side scattering has been effective to distinguish cells of different complexities (e.g.  
250 monocytes and granulocytes; Shapiro, 2005).

251 A 80,000 unit intensity FSC-H threshold (default FSC-H threshold value suggested by the manufacturer  
252 to minimize the effect of noise) was set in the instrument during data acquisition to minimize the effects of  
253 noise on bioparticle counts. The FSC-H channel (where H denotes height), measures single-particle forward  
254 scattering (FSC) intensity based on the peak (maximum point) of the voltage pulse curve recorded when a  
255 single particle goes through the interrogation point in the flow cytometer, whereas FSC-A, where A denotes  
256 area, measures single-particle FSC intensity based on the area below the curve of the recorder pulse. When  
257 the 80,000 unit FSC-H threshold is defined, only signals with an intensity greater than or equal to threshold  
258 value will be processed, and this could affect the statistics and detection efficiency of the flow cytometer  
259 toward small particles ( $\leq 1\mu\text{m}$ ). Experiments conducted with  $1.0\mu\text{m}$  polystyrene beads suspension  
260 (Supplemental information; Figure S16) have shown that  $1.0\mu\text{m}$  beads have FSC-H intensities above the  
261 80k threshold, no particle losses is observed, and beads estimated concentration agree with the reported by  
262 the manufacturer ( $\sim 6 \times 10^7 \text{ mL}^{-1}$ ; Life Technologies, Inc.) The FCM data from each sample was analyzed  
263 using the Flow Jo software (<https://www.flowjo.com/solutions/flowjo>) to gate and quantify bioparticles  
264 population. The same procedure was used to analyze the PBS, Milli-Q water and blanks.

265

### 266 **2.3 LIF detection of PBAP**



267 The WIBS-4A (referred to henceforth as “WIBS”) is a single biological particle real time sensor, which  
268 measures particle light scattering and autofluorescence in an approximately 0.5 – 15µm particle range  
269 (www.dropletmeasurement.com). Particles are initially sized using the 90-degree side-scattering signal  
270 from a 635 nm continuous-wave diode laser. The scattering intensity is directly related to particle diameter  
271 and was calibrated prior to deployment using polystyrene latex sphere calibration standards (PSL with 0.8,  
272 0.9, 1.0, 1.3, 2.0, 3.0 µm diameter, Thermo Scientific Inc.). The WIBS optical size therefore refers to PSL  
273 material with a real refractive index of 1.59. Healy et al. (2012) determined WIBS-4 counting efficiency by  
274 aerosolizing standardized concentrations of PSL sphere of specific sizes (e.g. 0.3, 0.4, 0.56, 0.7, 0.9 and  
275 1.3µm) and compared WIBS-4 total counts against PSL counts detected by the condensation particle  
276 counter (CPC). Results show WIBS-4 possesses a 50% counting efficiency for 0.5µm particles and detects  
277 100% of the PSL particles above 0.7µm when it is compared to the CPC counts. The 280nm and 370nm  
278 pulsed Xenon flashtube UV lights in the WIBS cause the particles to autofluoresce (i.e., excite the  
279 chromophores preexisting in the PBAP and do not rely on a fluorescent dye as done in FCM). Then,  
280 fluorescent emissions are measured at three wavelength channels, which following the nomenclature of  
281 Perring et al. (2015) are: (i) channel A (“FL1\_280” in previous studies; Robinson et al., 2013), which refers  
282 to the detected emission between 310-400nm after excitation at 280nm, (ii) channel B (“FL2\_280” in  
283 previous studies), which refers to the detected emission between 420-650nm after excitation at 280nm, and,  
284 (iii) channel C (“FL2\_370” in previous studies), which refers to the detected emission between 420-650nm  
285 after excitation at 370nm. The resulting autofluorescence from 280nm excitation is affected by the presence  
286 of tryptophan, tyrosine and phenylalanine amino acids in the PBAP (Pöhlker et al., 2012). Similarly, the  
287 resulting autofluorescence from the 370nm excitation is influenced by the presence of riboflavin and co-  
288 enzyme Nicotinamide Adenine Dinucleotide Phosphate (NAD(P)H) within the cells.

289 Biological and non-biological particles can be discriminated by using a fluorescent intensity threshold;  
290 here the threshold is determined with the Gabey et al. (2010) method and with modifications by Perring et  
291 al. (2015) as follows. Particles with fluorescence intensities below the fluorescence threshold in all channels  
292 are categorized as non-fluorescent (NON-FBAP). Particles that fluoresce above the threshold in only one  
293 channel are named with a single letter (e.g. A, B or C); particles that fluoresce in two channels are named  
294 with the two channel letters (e.g. AB, AC or BC), while particles that fluoresce in all channels are  
295 categorized as type ABC. Furthermore, the total FBAP concentration is defined as the sum of the  
296 concentration in the seven FBAP categories defined above. This approach was applied by Hernandez et al.,  
297 (2016) to pure culture PBAP (bacteria, fungal spores, pollen) to study their correspondence to FBAP types;  
298 bacteria tend to be detected by type A, and fungal spores and pollen by type AB and ABC. However,  
299 bioaerosol classification is instrument-specific and particle size dependent (Hernandez et al., 2016; Savage  
300 et al., 2017). Multiple environments have been studied using the Perring et al. 2015 FBAP types, including

301 rural, urban and highly vegetated locations. In the Southeastern US, the total FBAP concentration range  
302 from  $2 \times 10^4$  to  $8 \times 10^4$   $\text{m}^{-3}$ , constituting 3-24% of the total supermicron particle number between 1 and  $10 \mu\text{m}$   
303 diameter. In the highly vegetated Rocky Mountains, ABC type particles are enhanced during rainy days  
304 (during or post-rain events) to  $\sim 65\%$  of the total FBAP, owing to the release of wet-ejected fungal spores  
305 following precipitation (Gosselin et al., 2016). On the contrary, in the highly populated city of Nanjing,  
306 China all FBAP types, except type C, correlated with black carbon concentrations, suggesting a strong  
307 interference by combustion sources (Yu et al., 2016). A detailed explanation of the above-mentioned studies  
308 using Perring et al. 2015 approach is also included in the section SI.20 of the supplemental information.

## 309 **2.4 Location of sampling site and sampling frequency**

310 Bioaerosol sampling was conducted between April 7 and May 15, 2015 at the rooftop sampling  
311 platform of the Ford Environmental Sciences and Technology (ES&T) building at the Georgia Institute of  
312 Technology campus in Atlanta, GA. The site, which was located at the heart of a major urban environment,  
313 is surrounded by dense forested areas in the southeastern USA: the Oconee National Forest (South East),  
314 the Chattahoochee National Forest (North), and the Talladega National Forest (West). The WIBS was  
315 operating continuously throughout the same period, sampling bioaerosol from a 15 ft. long and  $\frac{1}{4}$  in. ID  
316 conductive tubing inlet fixed 8 ft. above the sampling platform floor. The SpinCon II was placed in the  
317 platform during sampling episodes with its inlet facing South. Three 4-hour samples per week were  
318 collected with the Spincon II sampler over the 5-week period (4 h sampling between 10am and 5pm; Table  
319 1). Meteorological data acquired from the same platform provided wind speed, wind direction, relative  
320 humidity (RH), temperature, total hourly rain and UV radiation index with a 1min resolution.

## 321 **3. Data processing and Analysis**

### 322 **3.1 FCM data processing**

323 All blanks collected showed contamination levels that did not exceed 1% of the PBAP quantified in the  
324 subsequent atmospheric samples. The 2-minute instrument blanks obtained after the CP and the HEPA filter  
325 washes was  $1.06 \times 10^3 \pm 7.37 \times 10^2$   $\text{mL}^{-1}$  and  $9.22 \times 10^2 \pm 1.24 \times 10^2$   $\text{mL}^{-1}$ , respectively, which are negligible  
326 accumulations compared to the  $2.55 \times 10^5 \pm 1.14 \times 10^5$   $\text{mL}^{-1}$  average PBAP concentration quantified in the  
327 atmospheric samples. The concentration of PBAP in the blanks was also confirmed with microscopy (not  
328 shown). Based on this, we are confident that the CP protocol and procedure to replace the working fluids  
329 ensured sterility of the biosampler before each sampling.

330 FCM analysis of the samples was carried out as follows. We obtain the fluorescence intensity (from  
331 each of the 4 fluorescence detectors), forward scattering and side scattering intensity for all the particles  
332 suspended in the samples. A gating procedure was used to determine the fluorescence levels associated  
333 with detecting only particles containing SYTO-13 (hence, a PBAP) and background fluorescence from non-

334 stained particles. The procedure (Supplemental information, SI.2 and SI.3) consists of 3 steps: (a)  
335 fluorescence threshold determination, (b) population gating, and, (c) biological/non-biological particle  
336 discrimination in the population(s) within the threshold (e.g. LNA PBAP, Section 4.1). The fluorescence  
337 threshold was determined using an atmospheric sample without SYTO-13 collected before each FCM  
338 analysis, as a blank. Based on the fluorescence responses obtained, we determine the FL1-A fluorescence  
339 intensity value for which 99.5% or 99.9% of the (unstained) particles of the blank autofluoresce below the  
340 chosen value. This FL1-A intensity, called “fluorescence threshold”, was determined for each sample  
341 (supplementary information, Figure S2a and S2b). The determination of the fluorescence threshold  
342 involved selecting the most conservative value that maximizes inclusion of biological particles and  
343 minimizes the inclusion of non-biological particles, including those that may be subject to background  
344 fluorescence or unspecific binding of SYTO-13 (Diaz et al., 2010; Müller et al., 2010). We found out that  
345 threshold values for the 99.9% approach were substantially higher than 99.5% approach in multiple  
346 sampling events and comparable to the fluorescence intensities observed for stained pure cultures ( $\sim 10^5$   
347 units), which means that the 99.9% threshold values will miscount pure cultures as non-biological.  
348 Consequently, we set the fluorescence threshold to the highest fluorescence intensity value observed by the  
349 99.5% approach (41,839 units; supplementary information, Figure S2b), applied it to all collected samples;  
350 henceforth named the 42k FL1-A threshold. The 42k threshold value aims to minimize any abiotic  
351 interference as it maximizes biological particles quantification. A fixed value has been chosen and applied  
352 to all samples given that having a different threshold value for each sampling event may result in  
353 quantification biases as bioaerosols with strong autofluorescence (e.g. pollen, fungal spores) can increase  
354 the threshold value and affect PBAP quantification in the population(s) within the threshold. The BD Accuri  
355 C6 flow cytometer used for the analysis of the samples maintains constant pre-optimized photomultiplier  
356 voltages and amplifier gain settings. As a result, the fluorescence intensity of particles is consistent from  
357 day-to-day, and the fluorescence intensity of a specific biological particle population having the same  
358 metabolic state and physiological characteristics must not show day-to-day variability  
359 ([www.bdbiosciences.com](http://www.bdbiosciences.com)). Under the 42k threshold approach PBAP concentrations in the population(s)  
360 within the threshold (e.g. LNA, Section 4.1) can be overestimated by up to a 0.5%. Furthermore, FCM  
361 experiments conducted with unprocessed Arizona Test Dust (ATD) show that the FL1\_A intensity  
362 distribution of SYTO-13 stained ATD particles is very similar to unstained ATD particles, and 100% of the  
363 SYTO-13 stained ATD particles stay below the 42k threshold (supplementary information, Figure S14a and  
364 S14b), supporting the 42k threshold effectiveness to filter out abiotic particles.

365 Once the FL1-A threshold was determined, plots of FL1-A vs. SSC-A and FL1-A vs. FSC-A are used  
366 to define clusters of bioparticles with fluorescence that exceed the FL1-A threshold and a characteristic  
367 optical size (obtained from the FSC-A intensity) or particle internal complexity (obtained from the SSC-A

368 intensity). FL1-A vs. SSC-A plots were used to define the populations of bioparticles for PBAP  
369 quantification as clusters using SSC-A parameter were more defined and showed better spatial resolution  
370 than using FSC-A parameter. The limits of each population were also determined with Flow Jo  
371 ([www.flowjo.com](http://www.flowjo.com)), using 2% contour plots (supplemental information; Figure S3) generated by equal  
372 probability contouring (i.e., 50 contour levels so that the same number of cells fall between each pair of  
373 contour lines). Populations above the FL1-A threshold value (41,839 FL1-A units) were considered  
374 biological (Section 4.1; e.g. HNA); the particles in the population within the threshold value (Section 4.1;  
375 e.g. LNA) having a FL1-A intensity greater than 41,839 units were counted as biological to determine the  
376 PBAP counts in the population. The total PBAP counts were considered as all particles counts having FL1-  
377 A fluorescence intensity above the determined threshold value minus the 15 $\mu$ m beads internal standard  
378 having FL1-A fluorescence intensity above the determined threshold value. The 15 $\mu$ m beads of known  
379 concentration and particle size allows for calibrating the optical size (supporting information, SI.7) of the  
380 bioparticles, as well as their concentration and departure from sphericity. The 15 $\mu$ m beads population  
381 showed fluorescence intensities comparable to the determined fluorescence threshold after been stained  
382 with SYTO-13 as it is known that molecular stains can be adsorbed on the surface of polystyrene beads  
383 (Eckenrode et al., 2005; Rödiger et al., 2011). The relatively high fluorescence intensity of the 15 $\mu$ m beads  
384 show populations within the threshold value (e.g. LNA, Section 4.1) cannot be rule out as being affected  
385 by unspecific staining of abiotic particles. However, populations above the threshold value (e.g. HNA,  
386 Section 4.1) should not be affected by such abiotic interferences.

### 387 **3.2 WIBS data processing**

388 15-minute average total aerosol and FBAP size distributions were obtained from the WIBS. FBAP was  
389 distinguished from the total aerosol using the Gabey et al. (2010) “trigger threshold” approach, which is  
390 applied as follows. First, the average “electronic fluorescence noise” and its standard deviation is  
391 determined for each channel (A, B, C) performing the Force Trigger (FT) calibration which consist to  
392 operate the WIBS without flowing air through the system. The FT calibration, carried out every 24hr, is  
393 critical for determining the lowest particle autofluorescence levels that robustly exceeds instrument  
394 electronic noise. FT calibrations measured the particle-free air background autofluorescence in the three  
395 WIBS channels (e.g. A, B, C), and measurements recorded the fluorescence intensity for 500 excitation  
396 flash events (Ziemba et al., 2016; Tropak and Schnaiter, 2013; Gabey et al., 2010). The threshold for each  
397 detector is then equal to the average fluorescence plus 2.5 times its standard deviation; particles with  
398 fluorescence intensities above this threshold value are classified as FBAP. Then, Perring et al. (2015)  
399 approach (Section 2.3) is applied to determine the combination of thresholds that provide the maximum  
400 concentration of PBAP and minimal interference from abiotic particles, which still remains an area of active  
401 research. It is important to note that the Gabey et al. (2010) threshold approach and the Perring et al. (2015)

402 FBAP types were applied to the WIBS-4A data and should not be directly compared to FBAP  
403 quantifications performed by the WIBS-3 in previous studies, owing to the channel A and B overlap on the  
404 latter. A detailed comparison between WIBS-3 and WIBS-4 models, as well as PBAP detection by both  
405 models, is further discussed in the supplemental information (SI.20).

406 In this study, thresholds for each channel were determined daily, and the total particle concentration,  
407 FBAP types (e.g. A, B, C, AB, BC, AC, ABC) concentrations and the total FBAP concentration (sum of  
408 the seven FBAP types) were used. From the data, 4h-averaged size distributions (using 15-minute average  
409 data) were generated for the total particles and all FBAP types in the 1-10 $\mu$ m range during the time SpinCon  
410 II run. Subsequently, WIBS overall sampling efficiency (aspiration efficiency + transport efficiency) was  
411 calculated using the Particle Losses Calculator (Von der Weiden et al., 2009) and applied to the 1-10 $\mu$ m  
412 size distributions for the sampling characteristics in our setup (15ft. sampling line with ¼ in. ID and 2.3 L  
413 min<sup>-1</sup> flow rate; Figure S4a). The sampling efficiency was calculated to be 67% for 5 $\mu$ m particles, with  
414 larger losses as size increased to 10 $\mu$ m. (supplemental information, FigureS4b). FCM and WIBS total  
415 particles and PBAP comparison was constrained to the 1 to 5 $\mu$ m range being the size overlap of both  
416 techniques. Also, the fractional composition of FBAP (based on number concentrations) was calculated to  
417 characterize its daily variability (Section 4.2), and compared against the daily variability of PBAP from the  
418 FCM analysis (Section 4.4).

## 419 **4. Results and Discussion**

### 420 **4.1 FCM biopopulation identification and quantification**

421 When the FCM results are plotted in terms of FL1-A fluorescence intensity versus SSC-A scattering  
422 intensity, four populations (Figure 2) emerge above the detection thresholds: low nucleic acid (LNA)  
423 particles, high nucleic acid (HNA) particles, pollen and the 15 $\mu$ m internal standard beads. EPM and SEM  
424 pictures (Supplementary Figures S5, S6, and S7) confirm the presence of these heterogeneous populations.  
425 SYTO-13 stains DNA and RNA, and the resulting single-cell FL1-A fluorescence intensity (Figure 2) is  
426 directly proportional to its nucleic acid content (Lebaron et al., 2001; Troussellier et al., 1999; Comas-Riu et  
427 al., 2002). Previously, SYTO-13 has effectively distinguish between HNA and LNA bacterioplankton and  
428 phytoplankton populations in fresh and seawater samples, and results are comparable to SYBR green II and  
429 SYBR green I, more specific DNA probes (Wang et al., 2010; Bouvier et al., 2007; Lebaron et al., 2001).  
430 However, corresponding populations in atmospheric PBAP have not been identified before. The SSC-A  
431 scattering intensity in Figure 2 changes as function of size, composition (e.g. cell refractive index) and  
432 complexity of the cell (e.g. internal structures or surface irregularities), and the strongest SSC-A intensity  
433 corresponding to the largest, most complex particles. Below we focus on each population to further  
434 understand the identified populations of biological particles.

435 The HNA size distributions are dominated by 3-5 $\mu\text{m}$  particles (mean diameter:  $4.15 \pm 0.06 \mu\text{m}$ ;  
436 Supplemental Information, Figure S10) and the total concentration moderately correlated with RH. HNA  
437 were virtually non-existent during several extended dry periods (days with average RH < 70% during  
438 sampling, e.g. 4/9, 4/22 and 5/15) and well defined during periods of high humidity, especially after rain  
439 events (days with average RH > 70% and T > 18 °C during sampling episode; e.g. 4/7, 4/14, 4/15). Both of  
440 these characteristics suggest that HNA particles correspond to wet-ejected fungal spores (e.g., from the  
441 Ascospores and Basidiospores genus; Oliveira et al., 2009; Li and Kendrick, 1995). The LNA size  
442 distributions are dominated by 2-4  $\mu\text{m}$  particles (mean diameter:  $2.99 \pm 0.06\mu\text{m}$ ; Supplemental Information,  
443 Table S1) and dominated Atlanta PBAP composition during dry days. Many individual bacteria are likely  
444 in around 1 $\mu\text{m}$ , but the observed LNA particles are within the median aerodynamic diameter of culturable  
445 bacteria ( $\sim 4\mu\text{m}$ ) in continental sites (Despres et al., 2012). Bacteria in the atmosphere can be co-emitted  
446 together with larger particles (e.g. soil, plant fragments) and occasionally they are observed as clumps of  
447 bacteria cells (Burrows et al., 2009). In addition, several bacterial species observed in the atmosphere  
448 (Delort and Amato, 2018; Monier and Lindow, 2003; Baillie and Read, 2001) are within this sizes range  
449 (e.g., *Sphingomonas spp.*: 1.0 - 2.7 $\mu\text{m}$ ; *Methylobacterium spp.*: 1- 8  $\mu\text{m}$ , *Pseudomona syringae*:  $\sim 2.5\mu\text{m}$ ,  
450 and *Bacillus anthracis*: 3-10 $\mu\text{m}$ ), supporting LNA population may represent single or agglomerated  
451 bacterial cells. However, it is clear that heterogeneous populations will probably contain multiple types of  
452 microorganisms and that may be the case in the LNA population.

453 It is known that pollen may burst into tiny fragments when is suspended in water (e.g., Augustin et al.,  
454 2012; Taylor et al., 2007), potentially increasing the concentration of LNA particles and biasing  
455 concentrations. Although 0.2 $\mu\text{m}$  – 5 $\mu\text{m}$  pollen fragments can be generated upon rupture, pollen (e.g. Birch,  
456 Ryegrass, Oak, Olive) mainly breaks apart into submicron fragments by hydrolysis and favors  
457 fragmentation into small submicron (<1 $\mu\text{m}$ ) particles (Taylor et al., 2007; Bacsı et al., 2006; Grote et al.,  
458 2003), not considered in our FCM analysis. An additional factor to consider in pollen fragmentation is the  
459 number of fragments generated per pollen grain. FCM applied to ragweed pollen suggests a 1:2 pollen-to-  
460 pollen fragments concentration ratio (Supplementary information, Table S2). Also, calculations based upon  
461 FCM-derived ragweed pollen and pollen fragments concentrations during this study (considering the total  
462 pollen mass added to the sample, 15 $\mu\text{m}$  mean diameter previously determined by Lin et al. (2013) and unit  
463 density) suggest approximately 67% of the ragweed pollen grains were intact after hydration and that each  
464 fragmented grain generates  $\sim 5$  pollen fragments; in agreement with Bacsı et al. (2006), 35% of ragweed  
465 pollen fragments upon hydration. Overall, ragweed pollen results suggest FCM experiments do not have a  
466 considerable impact in pollen fragmentation and that pollen fragmentation will have a negligible effect on  
467 LNA concentrations. Ragweed pollen is one of the most abundant wind-driven pollen species in the United  
468 States and its emission peaks during fall, but can be also present during late spring and summer. It is

469 representative of the pollen species we see in the Atlanta area (Darrow et al., 2012) and results suggest  
470 pollen fragmentation would not generate a substantial amount of fragments. The low collection efficiency  
471 of SpinCon toward large particles (<14% for diameters above 5 $\mu$ m) and that pollen concentrations in our  
472 samples are generally two orders of magnitude lower than LNA concentrations (Figure S22; supplemental  
473 information) suggest a negligible effect of pollen fragments in LNA biological particle quantification.  
474 Pollen concentrations are 100-1000 times lower than bacteria concentrations in the atmosphere (Hoose et  
475 al.,2010). At least 100 supermicron (>1 $\mu$ m) pollen fragments will have to be released per pollen grain to  
476 considerably influence the LNA population, which has not been observed. Also, EPM results showed intact  
477 pollen and limited amounts of small debris among the particles identified in the atmospheric samples  
478 collected for this study. Particles with fluorescence intensities above the FL1-A threshold value in the LNA  
479 population were counted as biological, giving us the PBAP counts within the LNA population and will be  
480 referred henceforth as the “LNA-AT” population (Figure 2), where “AT” refers to above threshold.

481

482 The LNA population shows SYTO-13 fluorescence intensities that are about one order of magnitude  
483 lower than the HNA population, and the fluorescence intensity difference is consistent across all sampling  
484 events. Based on Bouvier et al.2007, cell populations with different metabolic activity (e.g. active and non-  
485 active), when detected by FCM, should observe a decrease in fluorescence intensity in consecutive  
486 sampling events if transition from the HNA to the LNA population, or vice-versa if transition from LNA to  
487 HNA population. The fluorescence intensity of the LNA and HNA populations show small variation  
488 throughout the sampling events (LNA-AT:  $7.38 \times 10^4 \pm 1.39 \times 10^4$ ; HNA:  $6.72 \times 10^5 \pm 2.30 \times 10^5$ ; Table  
489 S3) and no anticorrelation is observed in the studied parameters (FSC-A, SSC-A, FL1-A), which supports  
490 we have in fact two distinctive population of bioaerosols (Supplemental Information; Figures S23 and S15).

491 A population of strongly fluorescing and very large particles (10-20 $\mu$ m, avg. average geometric mean  
492 diameter  $12.3 \pm 1.7\mu$ m) was identified (Figure 2). This population also strongly autofluoresces in the FCM  
493 when SYTO-13 was not added to the sample (SI.7, Figure S11). All together this indicates a population of  
494 pollen particles, as they are known to contain cell wall compounds (i.e., phenolic compounds, carotenoid  
495 pigments, Phenylcoumarin) that fluoresce more strongly than the proteins and cytosolic compounds  
496 responsible for bacteria/fungi autofluorescence (Pöhlker et al., 2012; Hill et al., 2009; Pöhlker et al., 2013).  
497 The pollen population was not well-defined during all sampling events; whenever present, pollen was  
498 characterized by concentrations ( $\sim 10^2 \text{ m}^{-3}$ ) consistent with reported values (Despres et al., 2012), which are  
499 also much lower than LNA-AT and HNA concentrations. As a result, pollen population was systematically  
500 gated using a perfect square between  $10^6$  and  $10^8$  intensity units in the FL1-A vs. SSC-A plot for each  
501 atmospheric sample. LNA-AT, HNA and pollen counts, acquired by the 42k threshold approach were used  
502 to calculate liquid-based ( $\text{mL}^{-1}$  of sample solution) and air-based ( $\text{m}^{-3}$  of air) concentrations for each

503 bioaerosol population as detailed in the Supplemental Information. The total PBAP concentration on each  
504 sample consisted of all non-bead particles above the 42k fluorescence threshold given that a non-negligible  
505 biological particle concentration was not constrained in the gated populations. Even though the 2% contour  
506 plots effectively allowed population gating,  $16.5 \pm 7.3\%$  of the total PBAP are not attributed the identified  
507 populations. The biological particles not constrained by FlowJo 2% gating, henceforth named as the  
508 “unclassified” bioparticles, showed the highest concentrations when both HNA and LNA populations are  
509 densely populated (4/16, 4/28 and 5/14; Figure 5). The lowest concentrations were observed when just the  
510 LNA population is identified (4/9, 4/22, 5/15; Figure 5) and when the LNA and HNA populations are  
511 identified after the rain event on 4/14. The observed behavior shows that the unclassified bioparticle  
512 concentrations is linked to the heterogeneity of the biological populations and the concentration of the gated  
513 populations (e.g. HNA, LNA and Pollen). The “unclassified” bioparticles concentration ranges from  $8.1 \times$   
514  $10^2 \text{ m}^{-3}$  to  $1.3 \times 10^4 \text{ m}^{-3}$  (avg.  $4.2 \times 10^3 \pm 3.3 \times 10^3$ ) and they are not constrained to a specific size range.  
515 Most of the unclassified bioparticles are far from the centroids of the gated populations. They can indeed  
516 be formed by fragmentation or accretion, or also be related to plant debris (i.e., irregular bioparticles) that  
517 are characterized by a very broad size, internal complexity and nucleic acid content distributions. In  
518 addition, we must note that additional concentration corrections are required owing to the sampling  
519 efficiency of the SpinCon II, but will be considered in sections 4.3 and 4.4.

520 Before SpinCon II sampling efficiency corrections are applied, FCM total particle concentrations range  
521 from  $2.6 \times 10^4 \text{ m}^{-3}$  to  $2.9 \times 10^5 \text{ m}^{-3}$ , with increasing concentrations toward the end of the sampling period.  
522 In addition, total PBAP concentration averaged  $2.4 \times 10^4 \pm 1.1 \times 10^4 \text{ m}^{-3}$  (coefficient of variation, CV, 13%;  
523 defined as the standard deviation over a triplicate FCM measurements over the average concentration).  
524 LNA-AT ranged between  $6.8 \times 10^2$  and  $2.9 \times 10^4 \text{ m}^{-3}$  (average:  $1.1 \times 10^4 \text{ m}^{-3}$ ; CV: 20%), HNA(fungal spores)  
525 between  $4.7 \times 10^3$  and  $1.9 \times 10^4 \text{ m}^{-3}$  (average:  $1.1 \times 10^4 \text{ m}^{-3}$ ; CV: 15%) when above the detection limit (n=12),  
526 and pollen from  $1.3 \times 10^2$  to  $1.2 \times 10^3 \text{ m}^{-3}$  (average:  $3.6 \times 10^2 \text{ m}^{-3}$ ; CV: 21%). These concentration levels are  
527 consistent with microscopy-based studies in urban environments for bacteria (e.g.,  $1.7 \times 10^4 \pm 1.3 \times 10^4 \text{ m}^{-3}$   
528 in springtime Birmingham, UK; (Harrison et al., 2005); fungal spores ( $1.8 \times 10^4 \pm 1.1 \times 10^4 \text{ m}^{-3}$  in Vienna,  
529 Austria between April-June; Bauer et al., 2008); and pollen (between  $5.69 \times 10^2 \text{ m}^{-3}$  to  $6.144 \times 10^3 \text{ m}^{-3}$  in  
530 Medellin, Colombia; Guarín et al., 2015). Also, additional experiments performed in September 2015,  
531 described in Figure S7 of the supplemental information (supplemental information, SI.6), showed that EPM  
532 and FCM-based quantifications agree within an order of magnitude. This is consistent with Lange et al.  
533 (1997), whom also found that FCM gives higher quantifications than EPM microscopy when studying *P.*  
534 *aeruginosa* pure cultures and airborne bacteria collected from a swine confinement building in Iowa, USA.



535 To better understand SYTO-13 fluorescence intensity differences between the identified (e.g. LNA-  
536 AT, HNA and pollen) populations in the atmospheric samples and their metabolic/stress state, FCM  
537 experiments were conducted with air-isolated bacteria (F8 strain; De Leon Rodriguez, 2015), ragweed  
538 pollen and yeast (*S. cerevisiae*; Y55 strain) mixtures to compare the SYTO-13 fluorescence intensity and  
539 the scattering properties of the pure cultures to those seen in the atmospheric samples. Pure culture  
540 experiments aimed to: (1) serve as positive controls to ensure SYTO-13 effectively stains bacteria, fungi  
541 and pollen, and (2) acquire reference fluorescence and scattering properties on each pure culture population.  
542 Pure cultures and atmospheric samples are summarized in Tables S3, S4 (supplementary information; FCM  
543 pure culture experiments) respectively. The LNA-AT population showed SYTO-13 fluorescence intensity  
544 up two orders of magnitude lower than F8 bacteria. The HNA population showed an order of magnitude  
545 lower SYTO-13 fluorescence intensity than Y55 HNA yeast, and, within the same magnitude for the LNA  
546 Y55 yeast. The HNA and LNA yeast populations in the pure culture experiments (Figure S13a) have one  
547 order of magnitude difference in FL1-A fluorescence intensity and may represent yeast populations with  
548 different metabolic states. Atmospheric and ragweed pollen populations had similar SYTO-13 fluorescence  
549 intensities and Figure S13c shows pollen fluorescence intensity may go up to  $10^8$ . The lower SYTO-13  
550 fluorescence intensity of the atmospheric populations may be related to genetic material degradation from  
551 exposure to atmospheric stressors; depending on the physiological characteristics of each population (Zhen  
552 et al., 2013; Amato et al., 2015). Our results also agree with Guindulian et al. (1997), showing that *E.coli*  
553 overnight cultures have higher SYTO-13 fluorescence intensity than starved *E.coli* population. Overall,  
554 FCM pure culture results suggest microbes starve in the atmosphere, leading to a possible reduction or  
555 leakage of the amount genetic material enclosed within each cell. Sampling can also stress cells, even  
556 disrupt the wall/membrane of the cell and lead to genetic material leakage (Zhen et al., 2013).

557 Pollen, HNA and LNA-AT atmospheric populations showed different SYTO-13 fluorescence  
558 intensities. Pollen showed the highest fluorescence intensity, followed by the HNA and LNA-AT (fraction  
559 of LNA above threshold; Figure 2) populations, respectively (Figure 2; Table S4). Guindulian et al. (1997)  
560 FCM results with starved bacterioplankton from seawater samples treated with DNase/RNase showed  
561 SYTO-13 fluorescence intensity can be related to the DNA content of starved bacterioplankton due to the  
562 low amount of RNA enclosed in starved cells. Taking in consideration our results and previous studies, we  
563 can suggest that Pollen, LNA-AT and HNA populations in the atmospheric samples are differentiated by their  
564 DNA content, which can in part explain SYTO-13 fluorescence intensity difference between them. We also  
565 acknowledge DNA sequestration by bacteria, fungal spores and pollen may differ and their cell membrane  
566 characteristics will ultimately determine how much stress the cells will sustain before they completely  
567 rupture. SYTO-13 is a highly permeable stain and effectively detects nucleic acids (DNA and RNA) of  
568 bacteria endospores and vegetative cells (Comas Riu et al., 2002). Fungal spores have also been effectively

569 stained by DNA/RNA probes (Bochdansky et al., 2017; Chen and Li et al., 2005), but some fungal spores  
570 might not be equally stained due to their harder cell wall, and chromatin-binding of DNA (Standaert-Vitse  
571 et al., 2015). Future work is needs to further study this.

## 572 **4.2 WIBS total concentration and FBAP daily variability**

573 WIBS-4A collected data continuously throughout the period; for comparison against the SpinCon  
574 II 4h liquid batch samples, WIBS data was averaged to the SpinCon II sampling times (Table 1). WIBS  
575 total particle concentration (1-5 $\mu$ m diameter) ranged from  $2.0 \times 10^5$  to  $1.0 \times 10^6$  m<sup>-3</sup> in agreement with  
576 observed particle concentrations in previously studied urban environments during Spring/Summer months  
577 like Helsinki, Finland (UV-APS avg.  $1.6 \times 10^5$  m<sup>-3</sup>; Saari et al., 2014) and Karlsruhe, Germany (WIBS-4  
578 avg.  $6.9 \times 10^5$  m<sup>-3</sup>; Tropak and Schnaiter et al., 2013). 4h average total particles concentrations in Figure 3a  
579 show particle concentrations declined during rain episodes (during or post-rain: e.g. 4/15, 4/16, 4/28, 4/29,  
580 4/30) as wet removal of PBAP is most efficient. However, during dry (no rain) episodes total particle  
581 concentrations built up in the atmosphere. To better understand the day-to-day variability of different FBAP  
582 types, the seven Perring et al. (2015) FBAP categories (e.g. Type A, B, C, AB, AC, BC and ABC) were  
583 studied plus the NON-FBAP type constituting particles that do not fluoresce in any channel (e.g. channel  
584 A, B, C). NON-FBAP concentrations are one order of magnitude higher than FBAP concentrations, and  
585 NON-FBAP, hence traced WIBS total particles throughout all sampling events (Figure 3a). Total FBAP  
586 concentrations also show similar behavior to the total particle concentration (Figure 3a) and it suggests non-  
587 biological particles can be biasing the total FBAP concentration. The variability of the total FBAP  
588 concentration is mainly linked to type A and type B concentrations as overall they constitute the two largest  
589 fractions to the total FBAP concentration (Figure 3b), and both FBAP types have previously misidentified  
590 non-biological particles as FBAP (Tropak and Schnaiter et al., 2013; Yu et al., 2016). As a result, our study  
591 considers the total FBAP concentration as the upper limit, and ABC type concentration as the lower limit  
592 of FBAP concentration in Metro, Atlanta. Type B dominates the FBAP fractional composition (Figure 3b),  
593 which has been linked to possible non-biological interferences from black carbon (Yu et al., 2016) and  
594 polycyclic aromatic hydrocarbons (PAHs) emitted from combustion sources. Total FBAP fraction ranges  
595 from 16% and 43%, and ABC fraction ranges from 1.3% and 9.2% of the total particles in the 1 to 5 $\mu$ m  
596 size range. ABC type fractions and ABC type concentrations are within the values observed by Tropak and  
597 Schnaiter (2013) using WIBS-4 in Karlsruhe, Germany; averaging  $2.9 \times 10^4$  m<sup>-3</sup> (when considering the sum  
598 of AC and ABC types) and constituting 7% of total coarse mode particles (0.8 $\mu$ m-16 $\mu$ m).

599 ABC type concentrations show an interesting variability throughout the 15 sampling events, as  
600 ABC reaches its maximum concentration on 4/14, on a warm and humid day after a rain event, concurrently

601 when the FCM HNA population also reaches its highest concentration – strongly suggesting ABC particles  
602 are fungal spores. (Figure 3a, Table 1). Furthermore, WIBS high resolution data in Figure S24 shows the  
603 enhancement of AB and ABC type right after the beginning of the rain event on 4/13 (6pm; night before  
604 sampling on 4/14) and is not correlated to NON-FBAP concentrations; FBAP concentration enhancement  
605 previously linked to wet-ejected fungal spores (Huffman et al., 2013; Gosselin et al., 2016). Gosselin et al.  
606 (2016) used WIBS-3 in the Rocky Mountains, Colorado showing ABC type fractional composition  
607 enhances after rain events to dominate the total FBAP composition and the enhancement is correlated to  
608 mannitol and arabitol concentrations (fungal spore tracers), which have been previously linked to  
609 Ascomycota and Basidiomycota spores emitted by the wet-ejection mechanism (Elbert et al., 2007). In  
610 addition, ABC type constitute a considerable fraction (~20%) of total FBAP during dry days in the Rocky  
611 Mountains possible because such highly vegetative environments maintain a high background of fungal  
612 spores (Huffman et al., 2013). However, urban environments like Metro Atlanta are not necessary  
613 dominated by fungal spores and its FBAP composition will be affected by the biological sources close to  
614 city (e.g. forests), local emissions and meteorology. The overall FBAP composition in metro Atlanta  
615 (Figure 3b) is dominated by type B (avg. fraction:  $33 \pm 9\%$ ), type A (avg. fraction:  $22 \pm 5\%$ ) and type AB  
616 (avg. fraction:  $22 \pm 5\%$ ) particles. Type ABC constitute  $12 \pm 6\%$  of the total FBAP and it reaches 30% on  
617 4/14, comparable to values observed by Gosselin et al., 2016 in the Rocky Mountains. The dominance of  
618 type B particles has been observed in the polluted atmosphere of Nanjing, China using WIBS-4A were type  
619 B constituted ~45% of the total PBAP and type B ( $\sim 2 \times 10^6 \text{ m}^{-3}$ ) concentrations were up to two orders of  
620 magnitude higher than type A concentrations ( $\sim 5 \times 10^4 \text{ m}^{-3}$ ) suggesting a high likelihood of interference  
621 from abiotic particle sources. However, Metro-Atlanta shows much lower total particle concentrations than  
622 Nanjing, China ( $\sim 10^7 \text{ m}^{-3}$ ) and type A and type B concentrations are within the same order of magnitude.  
623 Furthermore, Perring et al. (2015) have shown type B particles constitute a considerable fraction of the total  
624 supermicron particles across the United States, being ~15% and ~25% over (altitude >100m) the  
625 Southeastern US and Southwestern US, respectively. Total particle and NON-FBAP size distributions in  
626 Figure 3c peaked at  $\sim 1 \mu\text{m}$ . Similarly, types A, B, AB size distributions (Figure 3d) peaked close to  $1 \mu\text{m}$   
627 showing that interferences by non-biological particles cannot be rule out. However, ABC type size  
628 distribution (light blue line, Figure 3d) is dominated by 3-5 $\mu\text{m}$  particles and ABC type particles may have  
629 come from a different source to other FBAP types as they get enhanced after rain events (e.g. 4/14; Table  
630 1). Yu et al. (2016) also observed 4-6 $\mu\text{m}$  ABC type particles in the highly polluted Nanjing, China, but  
631 ABC type bimodal size distributions showed a peak between 1-2 $\mu\text{m}$  and a second peak between 4-6 $\mu\text{m}$ . In  
632 addition, ABC type number fractions in Nanjing, China correlated to black carbon mass fractions  
633 suggesting a considerable influence by combustion related particles and no rain events occurred during the  
634 sampling period. The difference between Metro Atlanta and Nanjing, China ABC type size distributions

635 suggest ABC type is not influenced by combustion related particles in Metro Atlanta. Overall, results show  
636 FBAP concentration (1-5 $\mu\text{m}$ ) ranges from  $10^4$  - $10^5$   $\text{m}^{-3}$  in metro Atlanta and wet-ejected fungal spores  
637 concentration, detected by ABC type, can constitute up to 30% of the FBAP (1-5  $\mu\text{m}$ ) after rain events.

### 638 **4.3 Correlation of HNA population with ABC type**

639 A quantitative comparison between WIBS-4A total particle and FCM total particle concentrations  
640 was subsequently performed and we focused the analysis to the 1 to 5 $\mu\text{m}$  size range as SpinCon sampling  
641 efficiency is reduced significantly above 5 $\mu\text{m}$  ( $\leq 14\%$ ; Kesavan et al., 2015). WIBS-4A and FCM total  
642 particle concentrations differed by about one order of magnitude (for optical diameter,  $d_o$ , greater than  
643 1.5 $\mu\text{m}$ ) and particle concentration difference increased for particles with  $d_o < 1.5 \mu\text{m}$  as shown in the size  
644 distribution (geometrically averaged across the 15 SpinCon II sampling events) in Figure 4a. The largest  
645 difference between WIBS-4A and uncorrected FCM size distributions seems to be related to SpinCon II  
646 having a cutoff size close to 1 $\mu\text{m}$ , reducing significantly its sampling efficiency. Even with the observed  
647 difference in the magnitude of the concentrations between the two techniques, ABC type and HNA  
648 concentrations traced throughout all the sampling events and are moderately correlated ( $R^2 = 0.40$ , P-value  
649 = 0.016; Figure 4b) and showed similar size distributions in the 1 to 5 $\mu\text{m}$  range as shown in Figure S12a.  
650 HNA and ABC type were both dominated by 3-5 $\mu\text{m}$  particles and it seems both are detecting the same  
651 type of biological particles. In addition, AB type showed a weak correlation with HNA concentrations ( $R^2$   
652 = 0.17), but their size distributions differed as type AB peaks close to  $\sim 1\mu\text{m}$  (Figure 3d). ABC is the only  
653 FBAP type showing a considerable correlation to the HNA population, and LNA-AT population is not  
654 correlated with any FBAP type. Overall, ABC type and HNA correlation is an important step forward to  
655 better understand the effectiveness of WIBS-4A FBAP categories to provide speciated PBAP  
656 concentrations in urban areas. ABC type particles have shown substantial concentrations ( $10^4$  - $10^5$   $\text{m}^{-3}$ ;  
657 Perring et al., 2015; Ziemba et al., 2016) across the US. The highest ABC fraction of the total FBAP was  
658 observed in Panhandle, Florida during an airborne study among multiple environments studied using  
659 WIBS-4A to sample from the California coast to central Florida, suggesting ABC type particles are  
660 ubiquitous in the US (Perring et al., 2015). Previous studies (Healy et al., 2014, Huffman et al., 2013) have  
661 shown correlations between LIF technology (e.g. WIBS-4 and UV-APS) fluorescence channels and fungal  
662 spores number concentrations, especially during fungal spores invigoration after rain events. Healy et al.  
663 (2014) used WIBS-4 in Killarney National Park, Ireland (e.g. high vegetative rural area) finding correlations  
664 between channel B (FL2;  $R^2 = 0.29$ ) and channel C (FL3;  $R^2 = 0.38$ ) concentrations and fungal spores  
665 concentrations (collected by Sporewatch impactor and quantified by microscopy). Gosselin et al. (2016)  
666 observed stronger correlations between fungal spores (inferred from mannitol and arabitol concentrations)  
667 and WIBS-4 concentrations in the Rocky Mountains, but our study in Atlanta, GA was carried out in  
668 completely different environment (e.g. highly-populated urban environment). Now for the first time FCM

669 HNA population have shown a correlation with WIBS-4A ABC type and suggests ABC type category  
670 detects wet actively ejected fungal spores in Metro Atlanta (e.g. urban area). In addition, recent WIBS-4A  
671 experiments using pure cultures have shown ABC type detects well several fungal spores (e.g. *Aspergillus*  
672 *Versicolor* & *Botrytis spp.*) and small pollen grains, but detection may vary across instruments (Hernandez  
673 et al., 2016).

674 FCM concentrations were corrected based on correction factors (CF) calculated upon the  
675 comparison of ABC and HNA size distributions (1 to 5 $\mu$ m) for each sampling event given (1) ABC type  
676 and HNA population similar size distributions and number concentrations (1 to 5 $\mu$ m) correlation, and, (2)  
677 WIBS-4A provides us representative concentrations of airborne particle concentrations in Metro Atlanta  
678 after sampling losses being corrected (Section 3.2). Concentration correction factors were determined for  
679 each sampling episode by taking the quotient of ABC type to HNA concentrations over the 1-5 $\mu$ m size  
680 range. The resulting size-dependent correction factor (Figure S12b) was then applied to the FCM size  
681 distributions, giving the “corrected FCM” bioaerosol data (between 1 and 5  $\mu$ m). Figure 4a shows that the  
682 corrected FCM total particle average size distribution traces WIBS-4A size distribution, allowing us to  
683 correct for SpinCon II low collection efficiency and to better constrain the magnitude of FCM  
684 concentrations. Our approach to calculate the estimated collection efficiency (ECE) considers all the  
685 processes that affect the concentration of PBAP, from collection to final quantification in the FCM. Figure  
686 S12b compares Kesavan et al. (2015) collection efficiencies determined for SpinCon I and the estimated  
687 collection efficiency calculated upon the CF calculation ( $ECE = 1/CF$ ) and shows the ECE of the SpinCon  
688 II is lower than Kesavan et al. (2015) below 3 $\mu$ m and performs better for particles above 3 $\mu$ m, but above  
689 3 $\mu$ m Kesavan et al (2015) collection efficiency is within the uncertainty of our calculations. Our lower ECE  
690 values (Figure S12b) for particles below 3 $\mu$ m can be related to SpinCon sampling time as Kesavan et al.  
691 (2015) experiment were conducted in a short period of time (e.g. 10-15 min) and ours took place for 4 h  
692 The main mechanisms leading to below 3 $\mu$ m particle losses could be their re-aerosolization over time being  
693 lost through the blower exhaust of the SpinCon II (Figure 1). Also, coagulation of small particles over time  
694 can not be rule out, but future work is needs to study it. Although SpinCon/FCM results correction based  
695 on the HNA and ABC type size distributions comparison effectively constrain the efficiency of the  
696 SpinCon/FCM analysis in this study, corrections are limited to the 1 to 5 $\mu$ m size range and must  
697 acknowledge that the specific sampling may stress cells and affect their detection.

698

#### 699 **4.4 PBAP populations after collection/detection corrections**

700 After correction through the application of the ABC correction factors, FCM total particle  
701 concentrations (1 to 5 $\mu$ m avg.:  $5.5 \times 10^5 \pm 5.1 \times 10^5$  m<sup>-3</sup>; Figure 5a) are within the same order of magnitude

702 as WIBS-4A concentrations (1 to 5 $\mu$ m avg.:  $5.4 \times 10^5 \pm 2.9 \times 10^5$  m<sup>-3</sup>; Figure 3a), and continue to exhibit  
703 substantial variability. The HNA (e.g. fungal spores) population showed a substantial invigoration during  
704 three sampling events (4/7, 4/14, 4/15; Figure 5a and 5b). To better understand the role of meteorology on  
705 PBAP composition, 24 hr-averaged temperature and relative humidity were used to express the prevailing  
706 temperature and relative humidity (RH) during each sampling event, considering the residence time of  
707 microorganisms (e.g. bacteria and fungal spores) before sampling. Sampling events were classified into  
708 four regimes based on the average diurnal (24hr avg.) relative humidity and ambient temperature, with T=  
709 18 °C (65 °F) to differentiate between warm and cold days, and, RH = 70% to differentiate between humid  
710 and dry days. During the 15 sampling days, temperature ranged from 10.4°C to 31.2°C, and RH varied from  
711 19.0% to 97.0% in Atlanta, GA (Look Table S4; supplemental information) The temperature and RH  
712 threshold values were chosen based on the observations and understanding that a combination of  
713 temperature and RH within these threshold values can significantly impact bioaerosol composition. For  
714 instance, humid and warm conditions may lead to the invigoration of fungal spores by wet ejection from  
715 plants (Ingold, 1971), on contrary, PBAP will get stressed when exposed to warm and dry conditions. The  
716 sampling times, RH, ambient temperature and meteorological categories of each SpinCon II sample is  
717 presented in Table 1.

718 Humid and warm days (4/7, 4/14 and 4/15; light green shaded areas in Figure 5a) were characterized  
719 by well-defined HNA and LNA-AT populations. These sampling episodes had the highest average HNA  
720 (fungal spore) concentration ( $4.0 \times 10^4 \pm 1.3 \times 10^4$  m<sup>-3</sup>) among the four meteorological regimes and during  
721 these sampling events HNA constituted  $\geq 77$  % of the total PBAP. Among the humid and warm days (Figure  
722 5a and 5b), average LNA-AT, HNA and “unclassified” bioaerosol compositions were 6.1%, 84.0% and  
723 9.9%, respectively of the total PBAP number. Also, the humid and warm days occurred after rain events,  
724 which can be linked directly to the strong fungal spore invigoration (Huffman et al., 2013). Before sampling,  
725 early morning precipitation occurred during 4/14 and 4/15, as well as during the night of 4/6. Precipitation  
726 did not occur during sampling in any of the humid and warm days. The FCM results (Figure S15a-c) that  
727 display the PBAP population between 4/7 and 4/9 show a disappearance of the (HNA) fungal spore  
728 population during the transition from a “humid and warm” day (4/7) to a “dry and warm” day (4/9). Figure  
729 5b shows how the HNA contribution to the total PBAP goes down on 4/8 when RH decreases and is  
730 undetected on 4/9. Furthermore, Figure 6a-c shows FL1 vs. SSC-A plots for 4/14 to 4/16 consecutive  
731 sampling periods, where a marked increase in the LNA-AT concentration from 4/15 to 4/16 goes together  
732 with a striking decrease in the HNA concentration. HNA fraction went down from 92.0% to 34.1% of the  
733 total PBAP and LNA-AT concentration went up from  $3.8 \times 10^3$  m<sup>-3</sup> to  $2.9 \times 10^4$  m<sup>-3</sup>. Humid and Warm days  
734 had the lowest averaged PBAP concentration ( $4.6 \times 10^4 \pm 9.8 \times 10^3$  m<sup>-3</sup> in the 1 to 5 $\mu$ m range) among the

735 four meteorological regimes, a possible effect of the bioaerosols being lost by wet scavenging, resulting in  
736 the enhancement of fungal spore contribution to the total PBAP number concentrations. The unclassified  
737 biological particles concentration also showed its lowest contribution ( $2.9 \times 10^3 \text{ m}^{-3}$ ; 9.9%) to the total  
738 PBAP number concentration during these events, when the HNA and LNA populations are best identified  
739 by the 2% contour plots.

740 Cold and humid days (4/16 and 4/29; light yellow shaded areas in Figure 5a) also showed well-defined  
741 HNA population, and HNA contributed on average to  $29.5 \pm 6.5 \%$  of the total PBAP concentration (1 to  
742  $5\mu\text{m}$ ). On 4/16 drizzling took place by the end of the sampling period, but no accumulated rainfall was  
743 measured by the meteorological station. However, on 4/29, accumulated rainfall averaged 0.04in. from  
744 11:55 AM to 2:20 PM (Figure S21). The similar HNA concentration between “Humid and Warm” and  
745 “Humid and Cold” days seen in Figure 5a and the lower contribution of HNA to the total PBAP during the  
746 “Humid and Cold” days may be linked to previously suggested bacteria emissions by droplet soil impaction  
747 during rain events (Joung et al., 2017). Bacteria emission by soil impaction can increase airborne LNA-  
748 AT concentration and HNA (fungal spores) will have a lower contribution to the total PBAP even when the  
749 fungal spore concentration is high during rain events. Both cold and humid days showed a considerable  
750 difference in LNA-AT contributions to the total PBAP concentration. On 4/16 and 4/29 LNA-AT  
751 constituted 45.2% and 65.3% of the total PBAP concentration, respectively (Figure 5b). The difference in  
752 the LNA-AT contribution to the total PBAP can be linked to the intensity of precipitation, as it shapes the  
753 composition (e.g. size and types) of microbes suspended in the atmosphere during the different stages of a  
754 rainfall (e.g. before, on set, during and after a rainfall; Yue et al., 2016).

755 Six of the fifteen sampling days were classified as warm and dry (4/8, 4/9, 4/22, 5/13, 5/14, 5/15; light  
756 orange shaded areas in Figure 5a) and it did not rain before or during any of these days (Table 1). During  
757 warm and dry days, HNA had the lowest averaged concentration ( $8.7 \times 10^3 \pm 1.2 \times 10^4 \text{ m}^{-3}$ ) among the four  
758 meteorological categories. In addition, during three dry and warm days (4/9, 4/22 and 5/15) the HNA  
759 population was undetected. This behavior can be related to the fact that high RH drives fungal spore  
760 emissions by wet ejection, but soil wetness could also affect emissions because the HNA population was  
761 detected in other warm and dry days with comparable RH (Huffman et al., 2013; Gosselin et al., 2016). The  
762 air mass trajectories reaching Atlanta during each sampling event could also affect the biological particles  
763 composition. For example, on 4/22, when the HNA was undetected, the 500m and 100m 72 h backward air  
764 mass trajectories reaching Atlanta came from the NW (US/Canada border) at high altitudes and do not  
765 spend more than 24h near surface. This air mass could affect bioaerosol composition with minimal  
766 influence from local bioaerosol emissions. However, the enhancement or the depletion of the HNA  
767 population have not been linked to specific air masses trajectories. Overall, warm and dry days prevail

768 during springtime in Atlanta and LNA-AT contribution (avg.:  $3.4 \times 10^4 \pm 2.5 \times 10^4 \text{ m}^{-3}$ ) may represent the  
769 bioaerosol background of Atlanta.

770 Four of the fifteen sampling days (4/21, 4/23, 4/28 and 4/30; light blue shaded areas in Figure 5a) were  
771 characterized by cold and dry conditions (Table 1). PBAP were dominated by LNA-AT during these events,  
772 as can see in Figure 7a-c, where LNA population are the dominant contributors to PBAP number. HNA  
773 population was diminished in Figure 7a (4/21) & Figure 7c (4/23) during cold and dry days and disappeared  
774 in Figure 7b during a warm and dry day. Overall, HNA was detected during cold and dry days, but showed  
775 lower contributions to the total PBAP number concentration than humid days. Among cold and dry days,  
776 the PBAP population (1 to 5  $\mu\text{m}$ ) was composed on average of  $72.6 \pm 10.1\%$  LNA-AT and  $16.5 \pm 8.2\%$   
777 HNA. Cold and dry days had on average the highest LNA-AT ( $5.3 \times 10^4 \pm 1.8 \times 10^4 \text{ m}^{-3}$ ) and total PBAP  
778 ( $7.3 \times 10^4 \pm 2.0 \times 10^4 \text{ m}^{-3}$ ) number concentrations (1 to 5  $\mu\text{m}$ ) among the four meteorological categories,  
779 reaching the PBAP maximum concentration on 4/23 (Figure 5a).

780

#### 781 **4.5 PBAP day-to-day variability in Metro Atlanta: FCM vs. WIBS**

782 Although WIBS and FCM possess different methodologies, they show similar trends providing a  
783 good understanding of the daily variability of PBAP in Metro Atlanta. FCM PBAP fraction (1 to 5  $\mu\text{m}$ )  
784 ranges from 3.8% to 69.2% of the total particles and the highest PBAP fraction (69.2%) and HNA  
785 concentration is observed on 4/14 ( $5.25 \times 10^4 \pm 5.89 \times 10^3 \text{ m}^{-3}$ ). The total FBAP fraction (1 to 5  $\mu\text{m}$ ) ranges  
786 from 16% to 43%, but it reaches its maximum on 4/15. However, ABC fraction of the total WIBS particle  
787 concentration ranges from 1.3% to 9.2% and it reaches its maximum on 4/14. Even when the magnitudes  
788 of the PBAP and FBAP fractions differ on average by a factor of  $\sim 2$  throughout the sampling period, both  
789 techniques agree an enhancement in the total biological particles takes place between 4/14 to 4/16. Given  
790 the uncertainty of the two methodologies, it is remarkable that there is such agreement between WIBS and  
791 FCM results.

792 Among the four meteorological categories, humid and warm days characterize for showing the  
793 highest HNA, A type, AB type and ABC type concentrations suggesting that A and AB types may also be  
794 related to wet-ejected fungal spores in Metro Atlanta; this possibly explains why the ABC fraction of the  
795 total FBAP in 4/7 is not as high as on 4/14 and 4/15 (Figure 3b), and differs with the behavior observed by  
796 the HNA population on 4/7. The LNA-AT population does not show a correlation to any specific FBAP  
797 type and shows its highest concentrations during dry and cold days. In addition, LNA-AT concentrations are  
798 anticorrelated with type B concentrations (Figure S19, correlation coefficient,  $r = -0.59$ ;  $R^2 = 0.30$ ) during  
799 dry (both cold and warm) days, when LNA-AT dominates the total PBAP concentration. Given that type B  
800 particles have been previously correlated to abiotic particles (e.g. black carbon) in urban environments (Yue



801 et al., 2017), LNA-AT and type B anticorrelation suggests that LNA-AT particles may in fact represent a  
802 heterogeneous bioaerosol population. That LNA-AT is not correlated with any FBAP type gives rise to two  
803 possibilities: (1) if LNA-AT population is mainly composed of bacteria or agglomerated bacteria, then it is  
804 possible that they are detected by multiple FBAP types and is not attributed specifically to one of them; (2)  
805 the intrinsic fluorescence of LNA-AT particles is too low and a high fraction of them is abiotic. It is  
806 challenging to determine what PBAP types each WIBS FBAP type is mainly detecting. Based on WIBS-  
807 4A results in Metro Atlanta, ABC type detects wet-ejected fungal spores, but still unclear what PBAP types  
808 are detect by the other FBAP types or if they just capture a high fraction of non-biological particles. FBAP  
809 types and WIBS total particles correlations in Figure S17 show all FBAP types are correlated to WIBS total  
810 particles, but ABC and AB types show the lowest correlations (type AB:  $R^2 = 0.101$ ; type ABC:  $R^2 =$   
811  $0.1266$ ).

812 Figure 8 shows FCM total PBAP (black line), ABC type (light green), FL1(Channel A; dark green  
813 line) and total FBAP (blue line) concentrations, where the FL1 concentration ( [FL1] ) constitutes the sum  
814 of the number concentrations of types A, AB, AC, and ABC ([FL1] = [A] + [AB] + [AC] + [ABC]; Gabey  
815 et al., 2011; Healy et al., 2014). Throughout the April-May 2015 sampling events, total PBAP  
816 concentrations (1 to 5 $\mu$ m) were mainly constrained between the FL1 and ABC type concentrations  
817 suggesting FL1 and ABC type represent the upper and lower bound PBAP concentrations in Metro Atlanta,  
818 respectively. It also important to highlight that FCM PBAP concentrations are closer to the ABC type  
819 concentrations before April 16 when the HNA population dominates, but then after April 16 FCM PBAP  
820 concentrations are closer to FL1 concentrations when LNA-AT starts to dominate the total PBAP  
821 concentration. In addition, Figure 8 shows that total FBAP (sum of type A, B, C, AB, AC, ABC) exceeds  
822 the (corrected) PBAP concentrations in Metro Atlanta.

## 823 **5. Conclusions**

824 In this study we presented the development and testing of an effective FCM protocol to identify and  
825 quantify bioaerosol populations. The FCM protocol, designed to constrain any particle accumulation due  
826 to cleaning or by fluid supplies, successfully quantified the day-to-day variability of bioaerosols in the  
827 Atlanta Metro area. It is the first FCM study to detect well-defined LNA (low nucleic acid) and HNA (high  
828 nucleic acid) atmospheric biological populations under different meteorological scenarios. FCM results  
829 show dynamic bioaerosol populations in Atlanta leading to a 84.0% of HNA (wet-ejected fungal spores)  
830 and 6.1% LNA-AT contribution to the PBAP number (1 to 5 $\mu$ m range), respectively, during humid and  
831 warm days after rain events. However, LNA-AT dominates warm and cold dry days, constituting 72% of  
832 the PBAP number concentration.

833 WIBS-4A and SpinCon II collocated sampling showed that the HNA and ABC type concentrations are  
834 well correlated ( $R^2=0.40$ ) and display similar size distributions. We therefore conclude that both  
835 instruments detect the same particles, and used empirical collection/detection efficiency factors to correct  
836 the FCM size distributions and concentrations in the 1 to 5 $\mu\text{m}$  diameter range. WIBS-4A and FCM results  
837 suggest Metro Atlanta PBAP concentrations range between  $10^4 - 10^5 \text{ m}^{-3}$  (1 to 5 $\mu\text{m}$ ) and they can constitute  
838 a substantial fraction of coarse mode particle concentration (WIBS-4A: 43%; FCM: 69%), comparable to  
839 the PBAP coarse mode fraction in highly vegetated environments. The FCM LNA-AT population, possibly  
840 containing bacterial cells, did not correlate to any FBAP type. The fact that the LNA-AT population is not  
841 correlated with a specific FBAP type suggests it may be particularly challenging to use LIF techniques to  
842 distinguish bioaerosols with low intrinsic autofluorescence from non-biological particles, especially given  
843 the heterogeneities introduced by the large biodiversity of airborne microbes. The possible influence of  
844 abiotic particles in the LNA-AT population can also explain the lack of correlation between LNA-AT and  
845 FBAP types given that the FCM threshold approach does not ensure total exclusion of abiotic particles. In  
846 addition, the unspecific binding of SYTO-13 to abiotic particles cannot be ruled out in the LNA-AT  
847 population. FCM comparison between atmospheric and pure culture samples showed lower SYTO-13  
848 fluorescence intensities in the atmospheric samples and suggests a degradation in the genetic material of  
849 PBAP, possibly caused by the limited nutrients and strong stress prevailing in the atmosphere, which further  
850 challenge the ability of LIF to distinguish LNA-AT.

851 In summary, this study has shown for the first time that FCM can effectively identify, quantify and  
852 study the daily variability of heterogeneous PBAP populations (e.g. HNA, LNA-AT and pollen) with  
853 different genetic material content in atmospheric samples. We also show that a number of FCM and WIBS-  
854 4A populations are largely correlated and therefore can be used to identify the nature of the FBAP detected  
855 in the latter. Our results finally show that the detection and quantification of bacterial cells in atmospheric  
856 samples remains a challenging task and is best achieved through the combination of techniques.

857

## 858 **Acknowledgments**

859 We acknowledge support from a Georgia Power Faculty Scholar chair, a Cullen-Peck Faculty Fellowship,  
860 a Dreyfus Foundation Postdoctoral Fellowship in Environmental Chemistry, NASA, a NASA Earth System  
861 Science Fellowship (grant no. 80NSSC17K0434). We also acknowledge support from project PyroTRACH  
862 (ERC-2016-COG) funded from H2020-EU.1.1. - Excellent Science - European Research Council (ERC),  
863 project ID 726165. We also thank Prof. Rodney Weber for helpful suggestions on the SpinCon II flow  
864 calibration.

865

866 **Competing interests**

867 The authors declare no competing interests.

868

869 **Author contributions**

870 AN, AN, KK and MB conceived the study. AN, NDLR, SW developed the modified biosampler. AN and  
871 NDLR developed the FCM analysis and sampling protocol. AN, NDLR carried out measurements, and SW  
872 helped support with analysis of the biological samples. LZ, BA provided the WIBS and helped with its  
873 setup and initial data analysis procedure. AN, AN worked on the analysis, write codes to interpret the data  
874 and developed the analysis protocol to combine the FCM and WIBS analysis outlined here. AN and AN  
875 wrote the paper, and all authors contributed with comments and modified text.

876

877 **References**

- 878 Amato, P., Joly, M., Schaupp, C., Attard, E., Möhler, O., Morris, C. E., Brunet, Y., and Delort, A. M.:  
879 Survival and ice nucleation activity of bacteria as aerosols in a cloud simulation chamber, *Atmospheric*  
880 *Chemistry and Physics Discussions*, 15, 4055-4082, 10.5194/acpd-15-4055-2015, 2015.
- 881 Augustin, S., Hartmann, S., Pummer, B., Grothe, H., Niedermeier, D., Clauss, T., Voigtländer, J.,  
882 Tomsche, L., Wex, H., and Stratmann, F.: Immersion freezing of birch pollen washing water,  
883 *Atmospheric Chemistry and Physics Discussions*, 12, 32911-32943, 10.5194/acpd-12-32911-2012, 2012.
- 884 Bacsi, A., Choudhury, B. K., Dharajiya, N., Sur, S., and Boldogh, I.: Subpollen particles: carriers of  
885 allergenic proteins and oxidases, *J Allergy Clin Immunol*, 118, 844-850, 10.1016/j.jaci.2006.07.006,  
886 2006.
- 887 Baillie, L., and Read, T. D.: *Bacillus anthracis*, a bug with attitude!, *Current Opinion in Microbiology*, 4,  
888 78-81, [https://doi.org/10.1016/S1369-5274\(00\)00168-5](https://doi.org/10.1016/S1369-5274(00)00168-5), 2001.
- 889 Bauer, H., Claeys, M., Vermeylen, R., Schueller, E., Weinke, G., Berger, A., and Puxbaum, H.: Arabitol  
890 and mannitol as tracers for the quantification of airborne fungal spores, *Atmospheric Environment*, 42,  
891 588-593, 10.1016/j.atmosenv.2007.10.013, 2008.
- 892 Bauer, H., Schueller, E., Weinke, G., Berger, A., Hitznerberger, R., Marr, I. L., and Puxbaum, H.:  
893 Significant contributions of fungal spores to the organic carbon and to the aerosol mass balance of the  
894 urban atmospheric aerosol, *Atmospheric Environment*, 42, 5542-5549, 10.1016/j.atmosenv.2008.03.019,  
895 2008.
- 896 Bochdansky, A. B., Clouse, M. A., and Herndl, G. J.: Eukaryotic microbes, principally fungi and  
897 labyrinthulomycetes, dominate biomass on bathypelagic marine snow, *The Isme Journal*, 11, 362,  
898 10.1038/ismej.2016.113, 2016.
- 899 Bouvier, T., Del Giorgio, P. A., and Gasol, J. M.: A comparative study of the cytometric characteristics of  
900 high and low nucleic-acid bacterioplankton cells from different aquatic ecosystems, *Environ Microbiol*, 9,  
901 2050-2066, 10.1111/j.1462-2920.2007.01321.x, 2007.
- 902 Burrows, S. M., Elbert, W., Lawrence, M. G., and Pöschl, U.: Bacteria in the global atmosphere – Part 1:  
903 Review and synthesis of literature data for different ecosystems, *Atmos. Chem. Phys.*, 9, 9263-9280,  
904 <https://doi.org/10.5194/acp-9-9263-2009>, 2009.
- 905 Chen, P. S., and Li, C. S.: Bioaerosol characterization by flow cytometry with fluorochrome, *J Environ*  
906 *Monit*, 7, 950-959, 10.1039/b505224f, 2005.
- 907 Chi, M.-C., and Li, C.-S.: Fluorochrome in Monitoring Atmospheric Bioaerosols and Correlations with  
908 Meteorological Factors and Air Pollutants, *Aerosol Science and Technology*, 41, 672-678,  
909 10.1080/02786820701383181, 2007.
- 910 Comas-Riu, J. and Vives-Rego, J. (2002), Cytometric monitoring of growth, sporogenesis and spore cell  
911 sorting in *Paenibacillus polymyxa* (formerly *Bacillus polymyxa*). *Journal of Applied Microbiology*, 92:  
912 475-481. doi:10.1046/j.1365-2672.2002.01549.x
- 913 Crawford, I., Robinson, N. H., Flynn, M. J., Foot, V. E., Gallagher, M. W., Huffman, J. A., Stanley, W.  
914 R., and Kaye, P. H.: Characterization of bioaerosol emissions from a Colorado pine forest: results from

- 915 the BEACHON-RoMBAS experiment, *Atmospheric Chemistry and Physics Discussions*, 14, 2499-2552,  
916 10.5194/acpd-14-2499-2014, 2014.
- 917 Darrow, L. A., Hess, J., Rogers, C. A., Tolbert, P. E., Klein, M., and Sarnat, S. E.: Ambient pollen  
918 concentrations and emergency department visits for asthma and wheeze, *Journal of Allergy and Clinical*  
919 *Immunology*, 130, 630-638.e634, <https://doi.org/10.1016/j.jaci.2012.06.020>, 2012.
- 920 DeLeon-Rodriguez, N., Terry L. Lathem, Luis M. Rodriguez-R, James M. Barazesh, Bruce E. Anderson,  
921 Andreas J. Beyersdorf, Luke D. Ziemba, Michael Bergin, Athanasios Nenes, and Konstantinos T.  
922 Konstantinidis.: Microbiome of the upper troposphere: Species composition and prevalence, effects of  
923 tropical storms, and atmospheric implications, *Proceedings of the National Academy of Sciences* 110,  
924 2575-2580, 2013.
- 925 DeLeon-Rodriguez, N.: Microbes in the atmosphere: prevalence, species composition, and relevance to  
926 cloud formation School of Biology Georgia Institute of Technology, 129 pp.,  
927 (<http://hdl.handle.net/1853/55517>), 2015.
- 928 Delort, A.-M., and Amato, P.: *Microbiology of aerosols*, pp.167-168, 2018.
- 929 Després, V. R., Alex Huffman, J., Burrows, S. M., Hoose, C., Safatov, A. S., Buryak, G., Fröhlich-  
930 Nowoisky, J., Elbert, W., Andreae, M. O., Pöschl, U., and Jaenicke, R.: Primary biological aerosol  
931 particles in the atmosphere: a review, *Tellus B*, 64, 10.3402/tellusb.v64i0.15598, 2012.
- 932 Díaz, M., Herrero, M., García, L. A., and Quirós, C.: Application of flow cytometry to industrial  
933 microbial bioprocesses, *Biochemical Engineering Journal*, 48, 385-407,  
934 <https://doi.org/10.1016/j.bej.2009.07.013>, 2010.
- 935 Eckenrode, H. M., Jen, S.-H., Han, J., Yeh, A.-G., and Dai, H.-L.: Adsorption of a Cationic Dye Molecule  
936 on Polystyrene Microspheres in Colloids: Effect of Surface Charge and Composition Probed by Second  
937 Harmonic Generation, *The Journal of Physical Chemistry B*, 109, 4646-4653, 10.1021/jp045610q, 2005.
- 938 Elbert, W., Taylor, P. E., Andreae, M. O., & Pöschl, U.: Contribution of fungi to primary biogenic  
939 aerosols in the atmosphere: wet and dry discharged spores, carbohydrates, and inorganic ions.,  
940 *Atmospheric Chemistry and Physics*, 7, 4569-4588, 2007.
- 941 Fröhlich-Nowoisky, J., Kampf, C. J., Weber, B., Huffman, J. A., Pöhlker, C., Andreae, M. O., Lang-  
942 Yona, N., Burrows, S. M., Gunthe, S. S., Elbert, W., Su, H., Hoor, P., Thines, E., Hoffmann, T., Després,  
943 V. R., and Pöschl, U.: Bioaerosols in the Earth system: Climate, health, and ecosystem interactions,  
944 *Atmospheric Research*, 182, 346-376, <http://dx.doi.org/10.1016/j.atmosres.2016.07.018>, 2016.
- 945 Gabey, A. M., Gallagher, M. W., Whitehead, J., Dorsey, J. R., Kaye, P. H., and Stanley, W. R.:  
946 Measurements and comparison of primary biological aerosol above and below a tropical forest canopy  
947 using a dual channel fluorescence spectrometer, *Atmospheric Chemistry and Physics*, 10, 4453-4466,  
948 10.5194/acp-10-4453-2010, 2010.
- 949 Gabey, A. M., Stanley, W. R., Gallagher, M. W., and Kaye, P. H.: The fluorescence properties of aerosol  
950 larger than 0.8  $\mu\text{m}$  in urban and tropical rainforest locations, *Atmospheric Chemistry and Physics*, 11,  
951 5491-5504, 10.5194/acp-11-5491-2011, 2011.
- 952 Gosselin, M. I., Rathnayake, C. M., Crawford, I., Pöhlker, C., Fröhlich-Nowoisky, J., Schmer, B.,  
953 Després, V. R., Engling, G., Gallagher, M., Stone, E., Pöschl, U., and Huffman, J. A.: Fluorescent

- 954 bioaerosol particle, molecular tracer, and fungal spore concentrations during dry and rainy periods in a  
955 semi-arid forest, *Atmos. Chem. Phys.*, 16, 15165-15184, 10.5194/acp-16-15165-2016, 2016.
- 956 Goudie, A. S.: Desert dust and human health disorders, *Environment International*, 63, 101-113,  
957 <http://dx.doi.org/10.1016/j.envint.2013.10.011>, 2014.
- 958 Griffin, D. W., Kellogg, C. A., Garrison, V. H., Lisle, J. T., Borden, T. C., and Shinn, E. A.: Atmospheric  
959 microbiology in the northern Caribbean during African dust events, *Aerobiologia*, 19, 143-157,  
960 10.1023/B:AERO.0000006530.32845.8d, 2003.
- 961 Grote, M., Valenta, R., and Reichelt, R.: Abortive pollen germination: A mechanism of allergen release in  
962 birch, alder, and hazel revealed by immunogold electron microscopy, *J. Allergy Clin. Immun.*, 111, 1017-  
963 1023, doi:10.1067/mai.2003.1452, 2003.
- 964 Guarín, F. A., Abril, M. A. Q., Alvarez, A., and Fonnegra, R.: Atmospheric pollen and spore content in  
965 the urban area of the city of Medellin, Colombia, *Hoehnea*, 42, 9-19, 2015.
- 966 Guindulain, T., J. Comas, and J. Vives-Rego: Use of nucleic acid dyes SYTO-13, TOTO-1, and YOYO-1  
967 in the study of *Escherichia coli* and marine prokaryotic populations by flow cytometry, *Applied and*  
968 *environmental microbiology*, 63, 4608 - 4611, 1997.
- 969 Harrison, R. M., Jones, A. M., Biggins, P. D. E., Pomeroy, N., Cox, C. S., Kidd, S. P., Hobman, J. L.,  
970 Brown, N. L., and Beswick, A.: Climate factors influencing bacterial count in background air samples,  
971 *International Journal of Biometeorology*, 49, 167-178, 10.1007/s00484-004-0225-3, 2005.
- 972 Healy, D. A., Huffman, J. A., O'Connor, D. J., Pöhlker, C., Pöschl, U., and Sodeau, J. R.: Ambient  
973 measurements of biological aerosol particles near Killarney, Ireland: a comparison between real-time  
974 fluorescence and microscopy techniques, *Atmos. Chem. Phys.*, 14, 8055-8069, 10.5194/acp-14-8055-  
975 2014, 2014.
- 976 Healy, D. A., O'Connor, D. J., and Sodeau, J. R.: Measurement of the particle counting efficiency of the  
977 "Waveband Integrated Bioaerosol Sensor" model number 4 (WIBS-4), *Journal of Aerosol Science*, 47,  
978 94-99, <http://dx.doi.org/10.1016/j.jaerosci.2012.01.003>, 2012.
- 979 Hernandez, M., Perring, A. E., McCabe, K., Kok, G., Granger, G., and Baumgardner, D.: Chamber  
980 catalogues of optical and fluorescent signatures distinguish bioaerosol classes, *Atmos. Meas. Tech.*, 9,  
981 3283-3292, <https://doi.org/10.5194/amt-9-3283-2016>, 2016.
- 982 Hill, S. C., Mayo, M. W., and Chang, R. K.: Fluorescence of bacteria, pollens, and naturally occurring  
983 airborne particles: excitation/emission spectra, Army Research Lab Adelphi Md Computational And  
984 Information Sciences Directorate, 2009.
- 985 Hoose, C., Kristjánsson, J. E., and Burrows, S. M.: How important is biological ice nucleation in clouds  
986 on a global scale?, *Environmental Research Letters*, 5, 024009, 2010.
- 987 Hoose, C., and Möhler, O.: Heterogeneous ice nucleation on atmospheric aerosols: a review of results  
988 from laboratory experiments, *Atmos. Chem. Phys.*, 12, 9817-9854, 10.5194/acp-12-9817-2012, 2012.
- 989 Huffman, J. A., B. Treutlein, and U. Pöschl: Fluorescent biological aerosol particle concentrations and  
990 size distributions measured with an Ultraviolet Aerodynamic Particle Sizer (UV-APS) in Central Europe,  
991 *Atmospheric Chemistry and Physics*, 10, 3215-3233, 2010.

- 992 Huffman, J. A., Prenni, A. J., DeMott, P. J., Pöhlker, C., Mason, R. H., Robinson, N. H., Fröhlich-  
 993 Nowoisky, J., Tobo, Y., Després, V. R., Garcia, E., Gochis, D. J., Harris, E., Müller-Germann, I., Ruzene,  
 994 C., Schmer, B., Sinha, B., Day, D. A., Andreae, M. O., Jimenez, J. L., Gallagher, M., Kreidenweis, S. M.,  
 995 Bertram, A. K., and Pöschl, U.: High concentrations of biological aerosol particles and ice nuclei during  
 996 and after rain, *Atmospheric Chemistry and Physics*, 13, 6151-6164, 10.5194/acp-13-6151-2013, 2013.
- 997 Ingold, C. T.: Fungal spores. Their liberation and dispersal, Oxford, Clarendon Press., 302 pp. pp., 1971.
- 998 Joly, M., Amato, P., Sancelme, M., Vinatier, V., Abrantes, M., Deguillaume, L., and Delort, A.-M.:  
 999 Survival of microbial isolates from clouds toward simulated atmospheric stress factors, *Atmospheric*  
 1000 *Environment*, 117, 92-98, <https://doi.org/10.1016/j.atmosenv.2015.07.009>, 2015.
- 1001 Joung, Y. S., and Buie, C. R.: Aerosol generation by raindrop impact on soil, 6, 6083,  
 1002 10.1038/ncomms7083, <https://www.nature.com/articles/ncomms7083#supplementary-information>, 2015.
- 1003 Kesavan, J., and Sagripanti, J. L.: Evaluation criteria for bioaerosol samplers, *Environ Sci Process*  
 1004 *Impacts*, 17, 638-645, 10.1039/c4em00510d, 2015.
- 1005 Lange, J. L. T., P S and Lynch, N: Application of flow cytometry and fluorescent in situ hybridization for  
 1006 assessment of exposures to airborne bacteria *Appl. Environ. Microbiol.* , 63, 1557-1563, 1997.
- 1007 Lebaron, P., Servais, P., Agogue, H., Courties, C., and Joux, F.: Does the high nucleic acid content of  
 1008 individual bacterial cells allow us to discriminate between active cells and inactive cells in aquatic  
 1009 systems?, *Appl Environ Microbiol*, 67, 1775-1782, 10.1128/AEM.67.4.1775-1782.2001, 2001.
- 1010 Li, D.-W., and Bryce Kendrick: A year-round study on functional relationships of airborne fungi with  
 1011 meteorological factors, *International Journal of Biometeorology* 39, 74-80, 1995.
- 1012 Liang, L., Engling, G., Cheng, Y., Duan, F., Du, Z., and He, K.: Rapid detection and quantification of  
 1013 fungal spores in the urban atmosphere by flow cytometry, *Journal of Aerosol Science*, 66, 179-186,  
 1014 10.1016/j.jaerosci.2013.08.013, 2013.
- 1015 Lin, H., Gomez, I., and Meredith, J. C.: Pollenkit Wetting Mechanism Enables Species-Specific Tunable  
 1016 Pollen Adhesion, *Langmuir*, 29, 3012-3023, 10.1021/la305144z, 2013.
- 1017 Longo, A. F., Ingall, E. D., Diaz, J. M., Oakes, M., King, L. E., Nenes, A., Mihalopoulos, N., Violaki, K.,  
 1018 Avila, A., Benitez-Nelson, C. R., Brandes, J., McNulty, I., and Vine, D. J.: P-NEXFS analysis of aerosol  
 1019 phosphorus delivered to the Mediterranean Sea, *Geophysical Research Letters*, 41, 4043-4049,  
 1020 10.1002/2014GL060555, 2014.
- 1021 Mage, P. L., Csordas, A. T., Brown, T., Klinger, D., Eisenstein, M., Mitragotri, S., Hawker, C., and Soh,  
 1022 H. T.: Shape-based separation of synthetic microparticles, *Nature materials*, 18, 82, 2019.
- 1023 Mathaes, R., Winter, G., Engert, J., and Besheer, A.: Application of different analytical methods for the  
 1024 characterization of non-spherical micro- and nanoparticles, *International journal of pharmaceutics*, 453,  
 1025 620-629, 2013.
- 1026 Monier, J. M., and Lindow, S. E.: *Pseudomonas syringae* Responds to the Environment on Leaves by Cell  
 1027 Size Reduction, *Phytopathology*, 93, 1209-1216, 10.1094/PHYTO.2003.93.10.1209, 2003.

- 1028 Morris, C. E., Conen, F., Alex Huffman, J., Phillips, V., Poschl, U., and Sands, D. C.: Bioprecipitation: a  
1029 feedback cycle linking earth history, ecosystem dynamics and land use through biological ice nucleators  
1030 in the atmosphere, *Glob Chang Biol*, 20, 341-351, 10.1111/gcb.12447, 2014.
- 1031 Muller, S., and Nebe-von-Caron, G.: Functional single-cell analyses: flow cytometry and cell sorting of  
1032 microbial populations and communities, *FEMS Microbiol Rev*, 34, 554-587, 10.1111/j.1574-  
1033 6976.2010.00214.x, 2010.
- 1034 Myriokefalitakis, S., Nenes, A., Baker, A. R., Mihalopoulos, N., and Kanakidou, M.: Bioavailable  
1035 atmospheric phosphorous supply to the global ocean: a 3-D global modeling study, *Biogeosciences*, 13,  
1036 6519-6543, 10.5194/bg-13-6519-2016, 2016.
- 1037 Nir, R., Yisraeli, Y., Lamed, R., & Sahar, E.: Flow cytometry sorting of viable bacteria and yeasts  
1038 according to beta-galactosidase activity, *Applied and environmental microbiology*, 56, 3861-3866, 1990.
- 1039 Oliveira, M., Ribeiro, H., Delgado, J. L., and Abreu, I.: The effects of meteorological factors on airborne  
1040 fungal spore concentration in two areas differing in urbanisation level, *Int J Biometeorol*, 53, 61-73,  
1041 10.1007/s00484-008-0191-2, 2009.
- 1042
- 1043 Ortiz-Martínez, Mario G., Rosa I. Rodríguez-Cotto, Mónica A. Ortiz-Rivera, Cedric W. Pluguez-Turull,  
1044 and Braulio D. Jiménez-Vélez. Linking endotoxins, African dust PM10 and asthma in an urban and rural  
1045 environment of Puerto Rico. *Mediators of inflammation* 2015.
- 1046 Pan, Y.-L., Santarpia, J. L., Ratnesar-Shumate, S., Corson, E., Eshbaugh, J., Hill, S. C., Williamson, C. C.,  
1047 Coleman, M., Bare, C., and Kinahan, S.: Effects of ozone and relative humidity on fluorescence spectra of  
1048 octapeptide bioaerosol particles, *Journal of Quantitative Spectroscopy and Radiative Transfer*, 133, 538-  
1049 550, <https://doi.org/10.1016/j.jqsrt.2013.09.017>, 2014.
- 1050 Perring, A. E., et al. (2015), Airborne observations of regional variation in fluorescent aerosol across the  
1051 United States, *J. Geophys. Res. Atmos.*, 120, 1153–1170, doi:10.1002/2014JD022495.
- 1052 Pöhlker, C., Huffman, J. A., Förster, J. D., and Pöschl, U.: Autofluorescence of atmospheric bioaerosols:  
1053 spectral fingerprints and taxonomic trends of pollen, *Atmospheric Measurement Techniques*, 6, 3369-  
1054 3392, 10.5194/amt-6-3369-2013, 2013.
- 1055 Pöhlker, C., Huffman, J. A., and Pöschl, U.: Autofluorescence of atmospheric bioaerosols – fluorescent  
1056 biomolecules and potential interferences, *Atmospheric Measurement Techniques*, 5, 37-71, 10.5194/amt-  
1057 5-37-2012, 2012.
- 1058 Pöschl, U.: Atmospheric Aerosols: Composition, Transformation, Climate and Health Effects,  
1059 *Angewandte Chemie International Edition*, 44, 7520-7540, 10.1002/anie.200501122, 2005.
- 1060 Robinson, N. H., Allan, J. D., Huffman, J. A., Kaye, P. H., Foot, V. E., and Gallagher, M.: Cluster  
1061 analysis of WIBS single-particle bioaerosol data, *Atmos. Meas. Tech.*, 6, 337-347,  
1062 <https://doi.org/10.5194/amt-6-337-2013>, 2013.
- 1063 Rödiger, S., Ruhland, M., Schmidt, C., Schröder, C., Grossmann, K., Böhm, A., Nitschke, J., Berger, I.,  
1064 Schimke, I., and Schierack, P.: Fluorescence Dye Adsorption Assay to Quantify Carboxyl Groups on the  
1065 Surface of Poly(methyl methacrylate) Microbeads, *Analytical Chemistry*, 83, 3379-3385,  
1066 10.1021/ac103277s, 2011.



- 1067 S.-L.Von der Weiden, F. D., and S. Borrmann: Particle Loss Calculator - a new software tool for the  
1068 assessment of the performance of aerosol inlet systems, *Atmospheric Measurement Techniques*, 2, 479-  
1069 494, 2009.
- 1070 Saari, S., Reponen, T., and Keskinen, J.: Performance of Two Fluorescence-Based Real-Time Bioaerosol  
1071 Detectors: BioScout vs. UVAPS, *Aerosol Science and Technology*, 48, 371-378,  
1072 10.1080/02786826.2013.877579, 2014.
- 1073 Šantl-Temkiv, T., Amato, P., Gosewinkel, U., Thyrhaug, R., Charton, A., Chicot, B., Finster, K., Bratbak,  
1074 G., and Löndahl, J.: High-Flow-Rate Impinger for the Study of Concentration, Viability, Metabolic  
1075 Activity, and Ice-Nucleation Activity of Airborne Bacteria, *Environmental Science & Technology*, 51,  
1076 11224-11234, 10.1021/acs.est.7b01480, 2017.
- 1077 Savage, N. J., Krentz, C. E., Könemann, T., Han, T. T., Mainelis, G., Pöhlker, C., and Huffman, J. A.:  
1078 Systematic characterization and fluorescence threshold strategies for the wideband integrated bioaerosol  
1079 sensor (WIBS) using size-resolved biological and interfering particles, *Atmospheric Measurement  
1080 Techniques*, 10, 4279-4302, 10.5194/amt-10-4279-2017, 2017.
- 1081 Shapiro, Howard M. *Practical flow cytometry*. John Wiley & Sons, 2005.
- 1082 Standaert-Vitse, A., Aliouat-Denis, C.-M., Martinez, A., Khalife, S., Pottier, M., Gantois, N., Dei-Cas, E.,  
1083 and Aliouat, E. M.: SYTO-13, a Viability Marker as a New Tool to Monitor In Vitro Pharmacodynamic  
1084 Parameters of Anti-Pneumocystis Drugs, *PLoS One*, 10, e0130358-e0130358,  
1085 10.1371/journal.pone.0130358, 2015.
- 1086 Sullivan, S. C., Hoose, C., Kiselev, A., Leisner, T., and Nenes, A.: Initiation of secondary ice production  
1087 in clouds, *Atmos. Chem. Phys. Discuss.*, 2017, 1-22, 10.5194/acp-2017-387, 2017.
- 1088 Taylor, P. E., Jacobson, K. W., House, J. M., and Glovsky, M. M.: Links between pollen, atopy and the  
1089 asthma epidemic, *Int Arch Allergy Immunol*, 144, 162-170, 10.1159/000103230, 2007.
- 1090 Toprak, E., and Schnaiter, M.: Fluorescent biological aerosol particles measured with the Waveband  
1091 Integrated Bioaerosol Sensor WIBS-4: laboratory tests combined with a one year field study,  
1092 *Atmospheric Chemistry and Physics*, 13, 225-243, 10.5194/acp-13-225-2013, 2013.
- 1093 Troussellier, M., Courties, C., Lebaron, P., and Servais, P.: Flow cytometric discrimination of bacterial  
1094 populations in seawater based on SYTO 13 staining of nucleic acids, *FEMS Microbiology Ecology*, 29,  
1095 319-330, [https://doi.org/10.1016/S0168-6496\(99\)00026-4](https://doi.org/10.1016/S0168-6496(99)00026-4), 1999.
- 1096 Tzur A, Moore JK, Jorgensen P, Shapiro HM, Kirschner MW (2011) Optimizing Optical Flow Cytometry  
1097 for Cell Volume-Based Sorting and Analysis. *PLoS ONE* 6(1): e16053.  
1098 <https://doi.org/10.1371/journal.pone.0016053>
- 1099 Van Dilla, M. A., Langlois, R. G., Pinkel, D., Yajko, D., & Hadley, W. K.: Bacterial characterization by  
1100 flow cytometry., *Science*, 220, 620-622, 1983.
- 1101 Wang, Y., Hammes, F., De Roy, K., Verstraete, W., and Boon, N.: Past, present and future applications of  
1102 flow cytometry in aquatic microbiology, *Trends Biotechnol*, 28, 416-424, 10.1016/j.tibtech.2010.04.006,  
1103 2010.

- 1104 Wu, Y.-H., Chan, C.-C., Rao, C. Y., Lee, C.-T., Hsu, H.-H., Chiu, Y.-H., and Chao, H. J.: Characteristics,  
1105 determinants, and spatial variations of ambient fungal levels in the subtropical Taipei metropolis,  
1106 *Atmospheric Environment*, 41, 2500-2509, 10.1016/j.atmosenv.2006.11.035, 2007.
- 1107 Yu, X., Wang, Z., Zhang, M., Kuhn, U., Xie, Z., Cheng, Y., Pöschl, U., and Su, H.: Ambient  
1108 measurement of fluorescent aerosol particles with a WIBS in the
- 1109 Yangtze River Delta of China: potential impacts of combustion-related aerosol particles, *Atmospheric*  
1110 *Chemistry and Physics*, 16, 11337-11348, 10.5194/acp-16-11337-2016, 2016.
- 1111 Yue, S., Ren, H., Fan, S., Sun, Y., Wang, Z., and Fu, P.: Springtime precipitation effects on the  
1112 abundance of fluorescent biological aerosol particles and HULIS in Beijing, *Sci Rep*, 6, 29618,  
1113 10.1038/srep29618, 2016.
- 1114 Yue, S., Ren, H., Fan, S., Wei, L., Zhao, J., Bao, M., Hou, S., Zhan, J., Zhao, W., Ren, L., Kang, M., Li,  
1115 L., Zhang, Y., Sun, Y., Wang, Z., and Fu, P.: High Abundance of Fluorescent Biological Aerosol  
1116 Particles in Winter in Beijing, China, *ACS Earth and Space Chemistry*, 1, 493-502,  
1117 10.1021/acsearthspacechem.7b00062, 2017.
- 1118 Zhen, H., Han, T., Fennell, D. E., and Mainelis, G.: Release of free DNA by membrane-impaired bacterial  
1119 aerosols due to aerosolization and air sampling, *Appl Environ Microbiol*, 79, 7780-7789,  
1120 10.1128/AEM.02859-13, 2013.
- 1121 Ziemba, L. D., Beyersdorf, A. J., Chen, G., Corr, C. A., Crumeyrolle, S. N., Diskin, G., Hudgins, C.,  
1122 Martin, R., Mikoviny, T., Moore, R., Shook, M., Thornhill, K. L., Winstead, E. L., Wisthaler, A., and  
1123 Anderson, B. E.: Airborne observations of bioaerosol over the Southeast United States using a Wideband  
1124 Integrated Bioaerosol Sensor, *Journal of Geophysical Research: Atmospheres*, 121, 8506-8524,  
1125 10.1002/2015JD024669, 2016.
- 1126

1 **Table 1:** Summary of the SpinCon II sampling events, the 24 h. averaged RH, ambient temperature, the  
 2 assigned meteorological category (using Section 4.4 definitions) and the corrected FCM-derived PBAP  
 3 number concentration (1 to 5  $\mu\text{m}$ ) for each sample collected during this study.

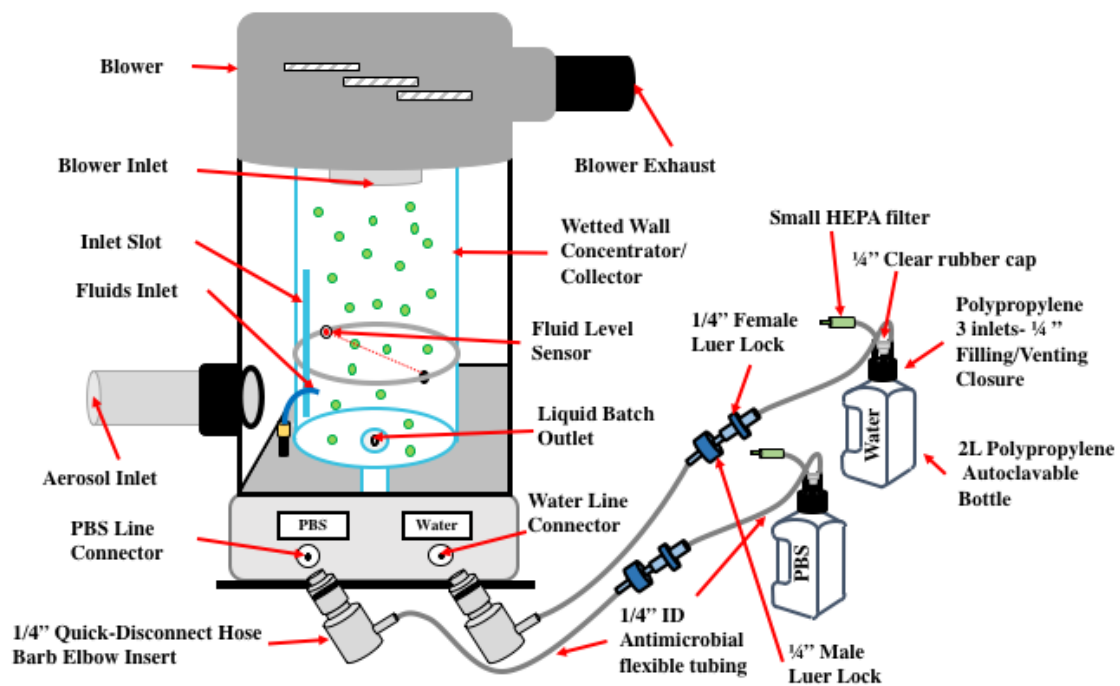
4

<b>Date (starting – ending time)</b>	<b>RH (%)</b>	<b>Temperature (°C)</b>	<b>Meteorological Category</b>	<b>PBAP Concentration (<math>\text{m}^{-3}</math>) 1 to 5<math>\mu\text{m}</math> diameter range</b>
4/7/15 (11:17 - 15:17) *	70.9	21.4	Humid, Warm	$9.282 \times 10^4$
4/8/15 (11:10 - 15:10)	53.6	24.9	Dry, Warm	$5.203 \times 10^5$
4/9/15 (11:15 - 15:15)	53.8	25.3	Dry, Warm	$1.254 \times 10^5$
4/14/15 (11:30 - 15:30) *	76.8	22.5	Humid, Warm	$8.253 \times 10^4$
4/15/15 (11:40 - 15:40) *	83.6	18.9	Humid, Warm	$1.234 \times 10^5$
4/16/15 (10:55 - 14:55)	86.3	12.5	Humid, Cold	$3.399 \times 10^5$
4/21/15 (13:15 - 17:15)	43.2	16.6	Dry, Cold	$4.741 \times 10^5$
4/22/15 (11:25 - 15:25)	38.1	18.8	Dry, Warm	$3.351 \times 10^5$
4/23/15 (11:35 - 15:35)	48.1	16.8	Dry, Cold	$1.708 \times 10^6$
4/28/15 (12:25 - 16:25)	45.3	17.0	Dry, Cold	$4.899 \times 10^5$
4/29/15 (11:55 - 15:55) #	79.4	14.2	Humid, Cold	$4.591 \times 10^5$
4/30/15 (12:10 - 16:10)	57.3	17.4	Dry, Cold	$9.603 \times 10^5$
5/13/15 (10:50 - 14:50)	40.1	23.5	Dry, Warm	$3.680 \times 10^5$
5/14/15 (11:50 - 15:50)	52.3	23.0	Dry, Warm	$4.851 \times 10^5$
5/15/15 (10:19 - 14:19)	64.4	23.1	Dry, Warm	$1.656 \times 10^6$

5 \* Sampling occurred post-rain event.

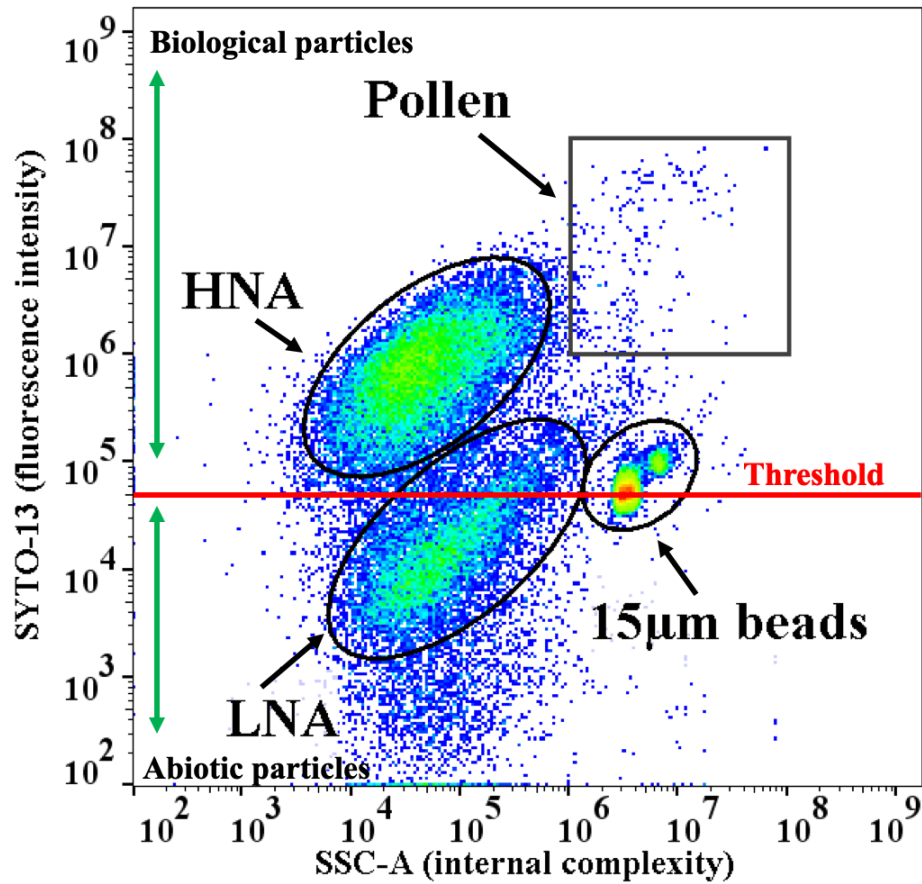
6 # Sampling occurred during a rain event.

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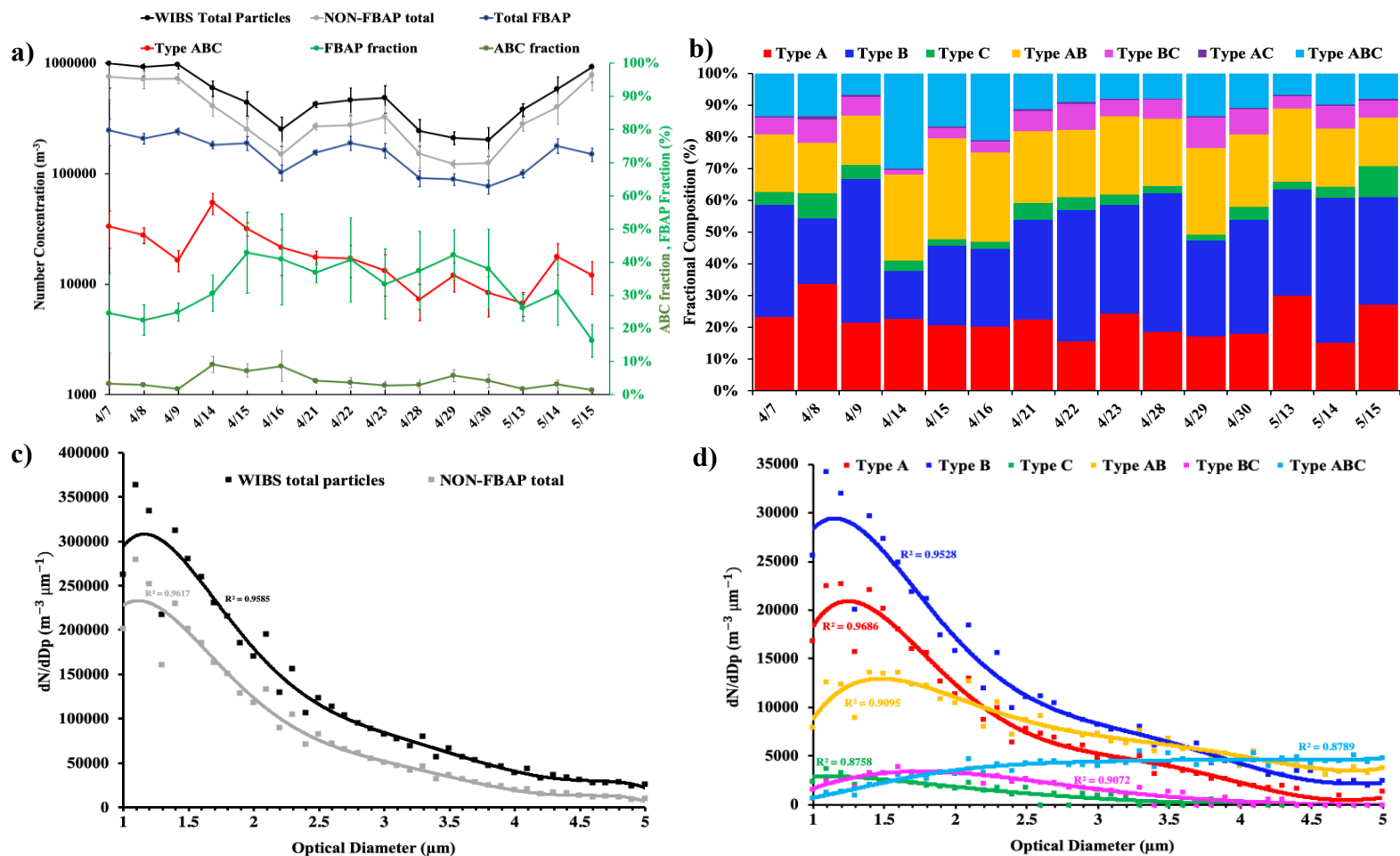
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**Figure 1:** SpinCon II sampling setup including modified fluid supply system with anti-microbial tubing and 2L Autoclavable bottles.



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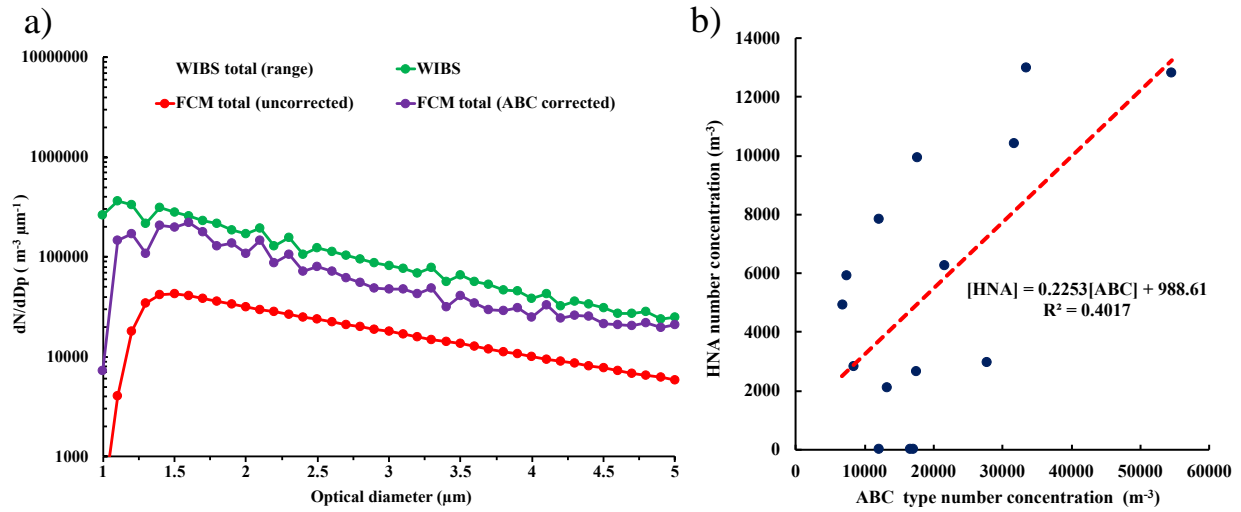
**Figure 2:** FL1-A vs. SSC-A plot used to identify populations in the April 14, 2015 sample including: the 42k threshold line in red and, abiotic particles (below threshold) and biological particles (above threshold) designated regions. In the density plot green and red zones denote the most populated regions. FL1-A in the y-axis shows the fluorescence intensity of each particle in the plot stained with SYTO-13 and SSC-A in the x-axis measures 90° light scattering, related to the internal complexity (e.g. granularity or amount of internal structures) and size of the particles. The fraction of the LNA population above the threshold line is referred as the “LNA-AT” population.



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23 **Figure 3:** WBS-4A 4hr (SpinCon II sampling time) averaged results of WBS total particle, NON-FBAP, total FBAP and type ABC  
 24 concentrations in the left Y-axis and ABC and FBAP fraction in the right Y-axis for each SpinCon II sampling event in (a); 4hr averaged FBAP  
 25 types number concentration fractional composition for each SpinCon II sampling event in (b); 1to5 $\mu\text{m}$  WBS total particles and NON-FBAP size  
 26 distributions in (c) and 1to5 $\mu\text{m}$  size distributions for all FBAP types, except AC type in (d). AC type showed low statistics and constituted less  
 27 than 1% of the total FBAP (not shown). Size distributions in (c) and (d) have been averaged over the 15 SpinCon II sampling events and constitute  
 28 the overall size distributions during rooftop sampling events. Solid lines in (c) and (d) shown 6-degree polynomial regressions performed to FBAP  
 29 the size distributions, including their respective correlation coefficients ( $R^2$ ).

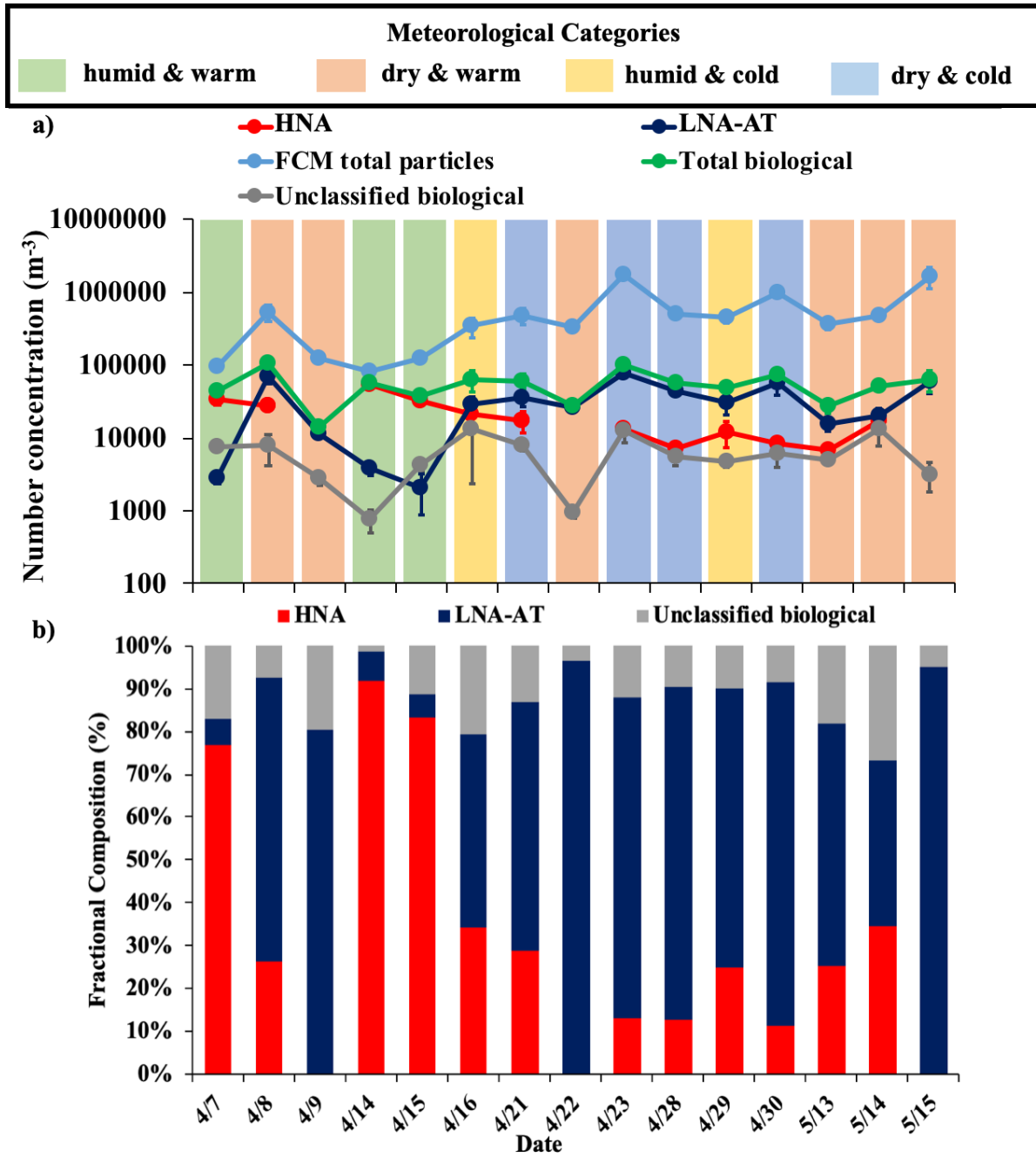
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**Figure 4:** WIBS-4A, FCM uncorrected and FCM (ABC corrected) total particle concentration (1 to 5 $\mu m$ ) average size distributions (geometrically averaged over the 15 SpinCon II sampling events) including WIBS range ( $\pm$  geometric standard deviation factor) in (a); and HNA and ABC type concentration correlation in the 1 to 5 $\mu m$  range in (b) including it linear correlation in red.

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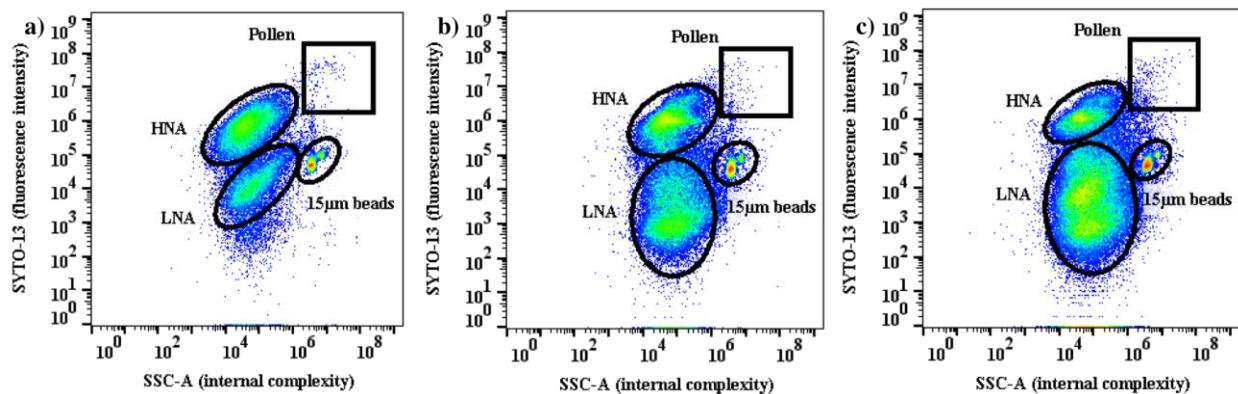


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**Figure 5:** FCM total particle, HNA, LNA-AT and total PBAP number concentrations in the 1 to 5 $\mu$ m range highlighting the prevailing meteorological category during each sampling event in (a); HNA and LNA-AT number concentration fractional compositions for each sampling event in (b).



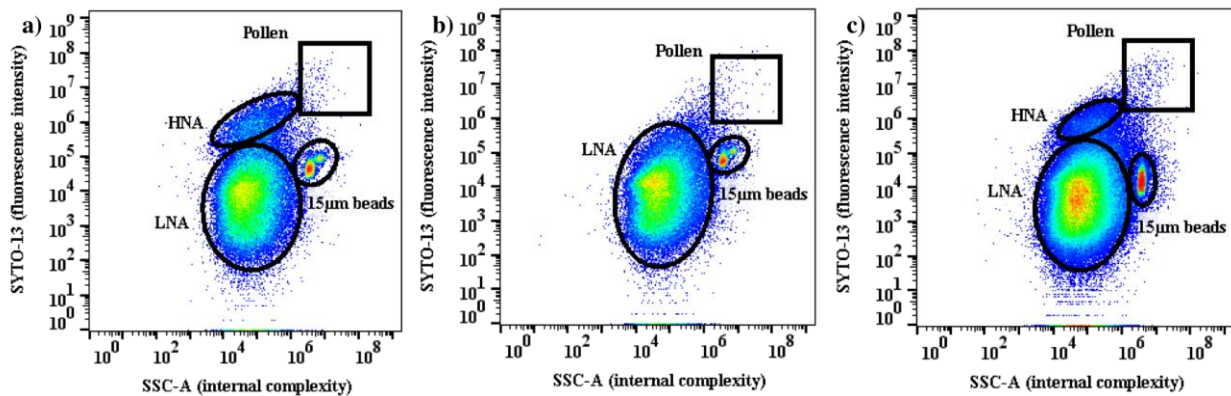
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57 **Figure 6:** FL1-A vs. SSC-A FSC plots for (a) April 14, (b) April 15, and, (c) April 16. This period was  
58 characterized by a transition from humid & warm to humid & cold conditions (diurnal average RH=77%,  
59 T=22.5 °C on 4/14; RH=84%, T=18.9 °C on 4/15, and RH= 86%, T= 12.5 °C on 4/16). The FCM plots  
60 during this transition period show a decrease of fungal population and an increase of the LNA population.  
61 In each population, warmer colors represent higher particle concentrations.

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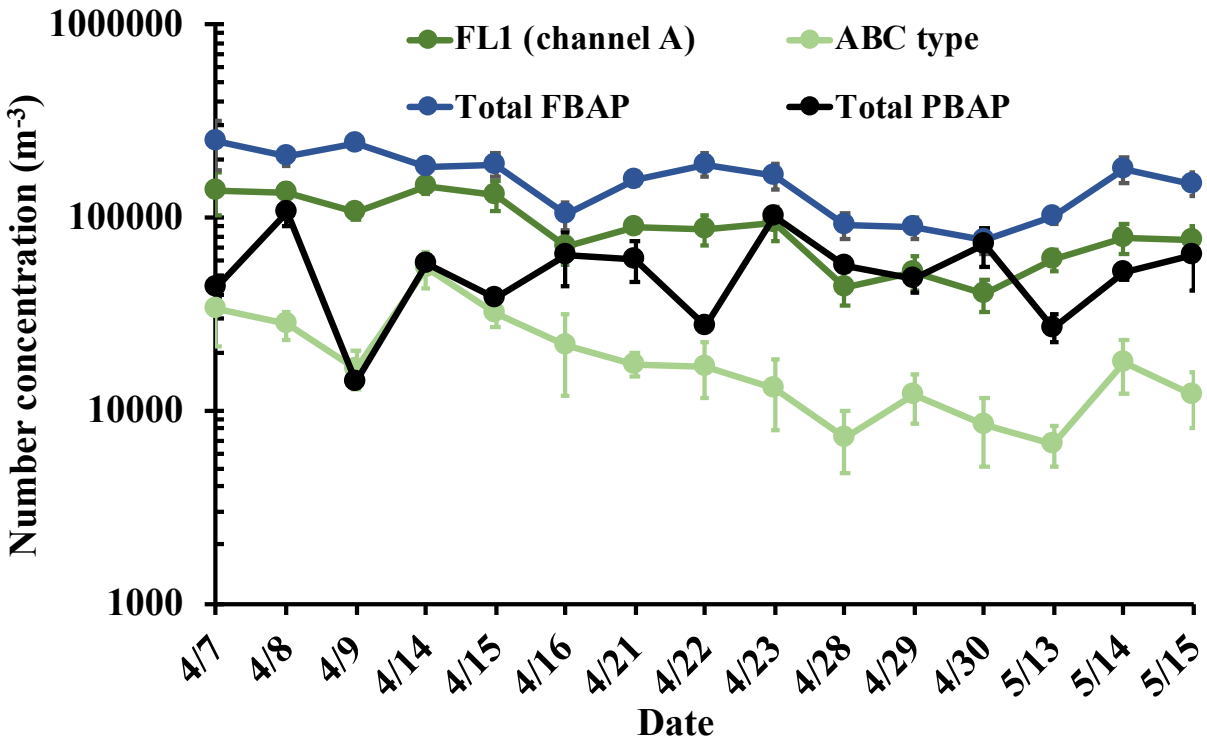


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68 **Figure 7:** Similar to Figure 6, but for (a) April 21, (b) April 22, (c) April 23, which was characterized by  
69 dry and variability in temperature (diurnal average RH=43%, T=16.6 °C on 4/21; RH=41%, T=19.0 °C on  
70 4/22, and, RH= 48%, T= 16.8 °C on 4/23). Note the disappearance of the fungal spore population on the  
71 warmest day (4/22).

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**Figure 8:** WIBS-4A total FBAP, FL1 and ABC type, and FCM total particle number concentrations in the 1 to 5 $\mu$ m range for each sampling event from April 7 to May 15, 2015.