



1 **Personal exposure to PM_{2.5} emitted from typical anthropogenic sources in**
2 **Southern West Africa (SWA): Chemical characteristics and associated**
3 **health risks**

4

5 Hongmei Xu^{1*,2,3,4}, Jean-François Léon², Cathy Liousse^{2*}, Benjamin Guinot², Véronique
6 Yoboué⁵, Aristide Barthélémy Akpo⁶, Jacques Adon², Kin Fai Ho⁷, Steven Sai Hang Ho³,
7 Lijuan Li³, Eric Gardrat², Zhenxing Shen¹, Junji Cao³

8

9 ¹*Department of Environmental Science and Engineering, Xi'an Jiaotong University, Xi'an,*
10 *China*

11 ²*Laboratoire d'Aérologie, Université de Toulouse, CNRS, Toulouse, France*

12 ³*SKLLQG, Key Lab of Aerosol Chemistry & Physics, Institute of Earth Environment, Chinese*
13 *Academy of Sciences, Xi'an, China*

14 ⁴*Collaborative Innovation Center of Atmospheric Environment and Equipment Technology,*
15 *Jiangsu Key Laboratory of Atmospheric Environment Monitoring and Pollution Control*
16 *(AEMPC), Nanjing University of Information Science & Technology, Nanjing, China*

17 ⁵*Laboratoire de Physique de l'Atmosphère, Université Felix Houphouët-Boigny, Abidjan,*
18 *Côte d'Ivoire*

19 ⁶*Laboratoire de Physique du Rayonnement, Université Abomey-Calavi, Abomey-Calavi,*
20 *Bénin*

21 ⁷*JC School of Public Health and Primary Care, The Chinese University of Hong Kong, Hong*
22 *Kong, China*

23

24 *Corresponding authors:*

25 *Hongmei Xu, E-mail: xuhongmei@xjtu.edu.cn*

26 *Cathy Liousse, E-mail: cathy.leal-liousse@aero.obs-mip.fr*

27

28 **Abstract**

29 Urbanization is a strongly emerging issue in Southern West African (SWA) region.
30 There is a general lack of understanding about the personal exposure to fine particulate matter
31 ($PM_{2.5}$), its chemical components and health risks related to the various anthropogenic
32 sources in this region. In the current study, personal exposure to $PM_{2.5}$ (PE $PM_{2.5}$) sampling
33 was for the first time carried out in dry season (January) and wet season (July) of 2016 to
34 characterize PE $PM_{2.5}$ from Domestic Fires (DF) for women and Waste Burning (WB) for
35 students in Abidjan, Côte d'Ivoire and Motorcycle Traffic (MT) for drivers in Cotonou, Benin.

36 The average PE $PM_{2.5}$ mass concentrations were 331.7 ± 190.7 , 356.9 ± 71.9 and
37 $242.8 \pm 67.6 \mu g m^{-3}$ at DF, WB and MT for women, students and drivers, which were 2.4, 10.3
38 and 6.4 times of the ambient $PM_{2.5}$ concentrations, respectively. Mean concentrations of PE
39 $PM_{2.5}$ at DF ($358.8 \pm 100.5 \mu g m^{-3}$), WB ($494.3 \pm 15.8 \mu g m^{-3}$) and MT ($335.1 \pm 72.1 \mu g m^{-3}$)
40 were much elevated in dry season, 15% higher than that at DF and 55% higher at both WB
41 and MT. The changes in PE $PM_{2.5}$ can be attributed to the source emissions, meteorological
42 factors and personal activities. The results also show that geological material (35.8%, 46.0%
43 and 42.4%) and organic matter (34.1%, 23.3% and 24.9%) were always the major
44 components in PE $PM_{2.5}$ at DF, WB and MT sites. It is worth noting that the contribution to
45 PE $PM_{2.5}$ from heavy metals was higher at WB (1.0%) than at DF (0.7%) and MT (0.4%),
46 which was influenced by the waste burning emission strongly, leading to the highest heavy
47 metal non-cancer risks for students (5.1 and 4.8 times of women and drivers' non-cancer
48 risks).

49 In organic species of PE $PM_{2.5}$, some fingerprints can be used to quantify the exposure
50 concentrations and trace the source contributions from local typical anthropogenic sources to
51 different samples. Women exposure concentration to polycyclic aromatic hydrocarbons
52 (PAHs) in $PM_{2.5}$ at DF ($77.4 \pm 47.9 ng m^{-3}$) was 1.6 times that for students at WB (49.9 ± 30.7
53 $ng m^{-3}$) and 2.1 times for drivers at MT ($37.0 \pm 7.4 ng m^{-3}$), which is related to the solid fuels
54 burning and grilling meat activities, resulting in 5 times higher of cancer risk safety threshold
55 (1×10^{-6}) to women. Phthalate esters (PAEs), commonly used as plasticizers in many products,
56 were observed to be extremely high in student exposure $PM_{2.5}$ samples ($1380.4 \pm 335.2 ng m^{-3}$)
57 at WB site, owing to the waste burning emission obviously. Drivers exposure to fossil fuel
58 emission (especially traffic) markers-hopanes in PE $PM_{2.5}$ at MT ($50.9 \pm 7.9 ng m^{-3}$) was 2.0-
59 2.3 times higher than women at DF ($17.1 \pm 6.4 ng m^{-3}$) and students at WB ($15.6 \pm 6.1 ng m^{-3}$),
60 correlating with the elevated exposure to traffic emissions for drivers.

61 Overall, the study shows that wood combustion, waste burning, fugitive dust and motor



62 vehicle emissions dominated PE $PM_{2.5}$ mass and contributed to its toxicities mainly. Heavy
63 metals and organic chemicals in PE $PM_{2.5}$ in SWA brought about Pb and Mn non-cancer
64 health risks for students at WB site and serious PAHs cancer risks for women at DF site via
65 inhalation pathway. This study provides basic data and initial perspective of $PM_{2.5}$ personal
66 exposure and health risk assessment in underdeveloped area to encourage the government to
67 improve the air quality and living standard of residents in this region.

68

69 **Keywords:** personal exposure to $PM_{2.5}$; domestic fires; waste burning; motorcycle traffic;
70 West Africa

71



72 1. Introduction

73 The southern West Africa (SWA) region has experienced an economic upturn and
74 increasingly significant anthropogenic air pollutant emissions during the last few years and
75 causes serious air pollution (IMF, 2017; Norman et al., 2007). Fine particulate matter (PM_{2.5}
76 with equivalent aerodynamic diameters $\leq 2.5 \mu\text{m}$) is one of the major concerns of
77 international organizations and publics because of the health effects associated with exposure
78 levels, health of individuals and pollutant emission sources (Bruce et al., 2000; Chen et al.,
79 2013; Owili et al., 2017). Owili et al. (2017) found that the four types of ambient PM_{2.5},
80 including mineral dust, anthropogenic pollutant, biomass burning and mixture aerosols are
81 significantly associated with under-five and maternal mortality in Africa. However, studies
82 on PM_{2.5}, especially direct personal exposure to PM_{2.5} (not stationary sampling) and its health
83 effects are still limited in low income countries in this region.

84 Since the 1990s, several international campaigns have been performed in Africa. Some
85 of them were mainly focused on the particles or aerosols, for example DECAFE (Lacaux et
86 al., 1995), EXPRESSO (Delmas et al. 1999; Ruellan et al., 1999), SAFARI-1992 (Lindesay et
87 al., 1996), SAFARI-2000 (Swap et al., 2002), AMMA (Léon et al., 2009; Liousse et al., 2010;
88 Marticorena et al., 2010) and INDAAF (Ouafo-Leumbe et al., 2017). As we known, the
89 Africa is the largest source of mineral dust particles from the Sahara Desert and unpaved road
90 surfaces (Laurent et al., 2008; Marticorena et al., 2010; Reeves et al., 2010), carbonaceous
91 aerosols originated from wild fires (mainly savannah fires) as well (Capes et al., 2008;
92 Gaudichet et al., 1995). Therefore, these campaigns were more biased towards the natural
93 sources of aerosols in Africa. Liousse et al. (2014) have shown the increase of the relative
94 importance of particulate emissions from domestic fires and fossil fuel combustion in Africa.
95 In previous literature, the major contributions to the aerosol chemistry in the dry season in
96 northern Benin were dust (26%-59%), primary organic matters (POC, 30%-59%), elemental
97 carbon (EC, 5%-9%) and water soluble inorganic ions (3%-5%) (Ouafo-Leumbe et al., 2017).
98 This poses serious health questions for people who frequent the city on a daily basis.
99 However, there is still limited literature on the health effects of personal exposure to PM_{2.5}
100 emitted from the typical anthropogenic sources in the emerging cities in Africa.

101 The main anthropogenic emission sources of PM_{2.5} in SWA include domestic wood
102 burning, fossil fuel combustion, unregulated traffic and industries, waste burning and road
103 dust associated to human activities. An ongoing project in Africa-DACCIWA (Dynamics-
104 Aerosol-Chemistry-Cloud Interactions in West Africa) aims at quantifying the influence of
105 anthropogenic and natural emissions on the atmospheric composition over South West Africa



106 and assessing their impact on human, ecosystem health and agricultural productivity, which
107 will be communicated to policy-makers, scientists, operational centres, students and general
108 publics. The current work involved in the framework of the Work Package 2 “Air Pollution
109 and Health” of DACCIWA is trying to link emission sources, air pollution and health impacts
110 over representative differentiated urban sources: domestic fires and waste burning in Abidjan
111 (Ivory Coast) and two-wheel vehicle traffic emission in Cotonou (Benin) for different target
112 populations.

113 Smoking meat (fish and pork) by biomass fuels (mainly wood) is an important diet
114 pattern for residents of coastal countries in SWA area. Many female workers without any
115 personal health protection are engaged in roasting activity. They are directly exposed to
116 extremely PM_{2.5} pollution from wood burning and smoking meat, causing very serious health
117 problems. Urbanization leads to explosive population growth and rural depopulation in SWA,
118 resulting in a large amount of urban domestic waste. The biggest landfill in Abidjan focused
119 in this study received more than 1,000,000 t waste per year (Adjiri et al., 2015). A mass of
120 garbage lacks processing capacity and reasonable treatment method, resulting in a large
121 amount of air pollutants emitted during the combustion and stacking of waste, which
122 damages the living environment and health condition of the populations in Abidjan,
123 especially for children (UNEP, 2015). Moreover, in many low-income countries, motorbike
124 taxis are a major mode of local transportation (Assamoi and Liousse, 2010). In Benin,
125 motorbike taxi drivers (mainly male) represented almost 2.5% of the total population of
126 Benin in 2002 (Lawin et al., 2016). As they spend many hours in the middle of traffic every
127 day, these drivers are highly exposed to traffic-related PM_{2.5} pollution over years.

128 Major chemical components in PM_{2.5}, like OC, ions and EC mentioned above, not only
129 strong impact on PM_{2.5} physicochemical characteristics, but also affect its health risks.
130 Typical trace toxic chemicals, such as heavy metals and polycyclic aromatic hydrocarbons
131 (PAHs) can be attached on PM_{2.5}, which would cause various health problems for humans
132 (Cao et al., 2012; WHO, 1998; Xu et al., 2015). For example, Pb is neuro-developmental
133 metal, affecting children health and mental development seriously (USEPA, 2006; Xu et al.,
134 2017). PAHs can be teratogenic and carcinogenic for humans strongly (Tang et al., 2008). Up
135 to now, only few studies have investigated PM_{2.5} chemical compositions of the personal
136 exposure samples, and little is known regarding the sources and health risks of personal
137 exposure PM_{2.5} in SWA region. This poses a challenge to formulation of strategies aimed at
138 mitigating PM_{2.5} pollution and its health effects in this area.

139 Therefore, our study relies on the portable device sampling PM_{2.5} personal exposure



140 samples in SWA area in 2016 for the purpose of 1) characterizing the personal exposure to
141 $PM_{2.5}$ as variation of different local typical anthropogenic $PM_{2.5}$ sources by the chemical
142 component analysis and $PM_{2.5}$ mass balance; 2) identifying potential pollution sources to
143 different exposed populations by fingerprint organic markers; 3) evaluating the personal
144 exposure to $PM_{2.5}$ health risks by the U.S. EPA health risk assessment model. This
145 information will provide scientific understanding of the personal exposure to $PM_{2.5}$ in SWA
146 and try to arouse the government's attention to protect residents there from various
147 anthropogenic sources.

148

149 **2. Materials and methods**

150 *2.1. Site description and participants selection*

151 Personal exposure to $PM_{2.5}$ (hereafter defined as PE $PM_{2.5}$) filter samples were collected
152 using portative devices in the polluted atmosphere of different source environments,
153 including Domestic Fires (DF) for women, Waste Burning (WB) for students both in Abidjan,
154 Côte d'Ivoire, and Motorcycle Traffic (MT) for drivers in Cotonou, Benin (Figure 1). Abidjan
155 ($5^{\circ}20'$ N, $4^{\circ}1'$ W) is the economic capital of Côte d'Ivoire with 6.5 million inhabitants in
156 2016. It is characterized by a high level of industrialisation and urbanization in SWA area.
157 Cotonou ($6^{\circ}21'$ N, $2^{\circ}26'$ W) is the largest city and economic center of Benin, with about 1.5
158 million inhabitants in 2016. Both cities experience a tropical wet and dry climate, with
159 relatively constant air temperature (24-30 °C) and average relative humidity (RH) above 80%
160 throughout the year.

161 DF site in Abidjan is located in the market of Yopougon-Lubafrique ($5^{\circ}19.7'$ N, $4^{\circ}6.4'$
162 W) in a large courtyard with about 25 fireplaces (Figure 2). The fuels used are essentially
163 hevea wood (one kind of rubber trees) locally. Several adult female workers were employed
164 to grilling meat or roasting peanuts from 06:00 to 15:00 UTC (working time) in the working
165 day. In this study, we selected two healthy, non-smoking female workers (average age of 32.5
166 years old) to investigate personal exposure to $PM_{2.5}$ from domestic fire and related sources,
167 such as grilling (Figure 2). WB site in Abidjan is near the public landfill of Akouédo ($5^{\circ}21.2'$
168 N, $3^{\circ}56.3'$ W), which has received all the waste produced from Abidjan for the last 50 years
169 (Figure 2). We selected two healthy, non-smoking primary school students (average age of 11
170 years old) who live and study next to WB site (within 100 m straight-line distance) to
171 determine the personal exposure to $PM_{2.5}$ from waste burning (spontaneous combustion at
172 high air temperature condition and combustion by the landfill workers sometimes) emission
173 at landfill and other daily sources. MT site in Cotonou is located in the Dantokpa area ($6^{\circ}22.1'$



174 N, 2°25.9' E), one of the biggest markets in western Africa (Figure 2). It is largely dominated
175 by a mass of motorcycle traffic (two-wheel vehicles powered by petrol, also named zemidjan
176 in local language) and a small quantity of other motor vehicles emissions. We chose two
177 healthy, non-smoking male motorcycle drivers (average age of 50 years old) to survey PM_{2.5}
178 personal exposure from motorcycle emission and related sources (such as road dust).

179 Two women (woman A and B) involved in this study at DF are both in charge of
180 cooking at home by charcoal and butane gas, and cleaning house in daily life (Figure S1abc).
181 One of the student participators (student A, boy, 8 years old) at WB doesn't cook at home by
182 himself (energy sources for cooking are charcoal and liquefied petroleum gas (LPG)) (Figure
183 S1ac), but the other student (student B, girl, 14 years old) is usually responsible for cooking
184 at home by burning solid fuel, i.e., wood (Figure S1d). Two motorcycle drivers (driver A and
185 B) focused in this study at MT are both working for a local motorcycle operation company,
186 whose working time is usually from 6:30 to 10:30, 12:00 to 17:00 and 18:30 to 21:00 UTC.
187 They are driving on road almost all the working time and go back home for meals. They
188 don't cook at home by themselves (energy source for cooking is charcoal) (Figure S1a).

189

190 2.2. Personal exposure to PM_{2.5} samples collection and QA/QC

191 12-hour integrated (daytime: 7:30 to 19:30 UTC; nighttime: 19:30 to 7:30 on the next
192 day UTC) PE PM_{2.5} samples were collected during the dry season (from January 6th to 11th)
193 and wet season (from July 5th to 10th), 2016 in two major southwestern African cities
194 mentioned above (Figure 1). PE PM_{2.5} sampling was conducted during three consecutive days
195 with the same type participants synchronously, using the PEM (Personal Environmental
196 Monitor) sampling devices with SKC pump (SKC Inc., USA) at a flow rate of 10 liter per
197 minute (lpm). The PEM PM_{2.5} sampling head worn in the breathing zone of participants in
198 this study. Samples were collected on 37 mm pre-baked quartz filters (800 °C, 3 hours,
199 QM/A®, Whatman Inc., UK). A total of 72 personal exposure samples, including 24 samples
200 (12 daytime + 12 nighttime samples) for women at DF, 24 (12 + 12) for students at WB and
201 24 (12+12) for drivers at MT, were collected in this study. Moreover, 12 PE PM_{2.5} field
202 blanks (one field blank for each participant in one season, collected on the second day of the
203 three consecutive sampling days) were sampled in this study as well.

204 In order to verify the comparability of personal exposure samples and data caused by not
205 identical sampling devices, 10 pairs of PM_{2.5} samples were synchronously collected by two
206 sets of actual PEMs with SKC pumps. The comparison results led to a significant correlation
207 between the PM_{2.5} mass concentrations obtained from two sampling devices



208 $(y=0.986x+0.189, R^2=0.974, P<0.0001)$. Identical membrane (quartz fiber) and analytical
209 treatments were used in this study. After sampling, the filter samples were placed in Petri
210 dishes, sealed with parafilm and stored in a $-20\text{ }^{\circ}\text{C}$ freezer to prevent loss of mass through
211 volatilization prior to analysis. Blank values were used to account for any artifacts caused by
212 gas absorption and subtract the background $\text{PM}_{2.5}$ and chemical compositions concentrations
213 in this area.

214 We report the meteorological observations during the dry (December 2015 to March
215 2016) and wet (April to July 2016) seasons at the sampling places in Table 1. Meteorological
216 data are retrieved from the NOAA Global Surface Summary of the Day I (GSOD) at the
217 airports of each cities, namely Felix Houphouet Boigny Airport (Abidjan) and Cardinal
218 Bernadin Gantin International Airport (Benin). We give the daily average air temperature,
219 wind speed and rainfall accumulation in Table 1.

220

221 2.3. $\text{PM}_{2.5}$ gravimetric and chemical analysis

222 PE $\text{PM}_{2.5}$ filter samples were analyzed gravimetrically for mass concentrations with a
223 high-precision electronic microbalance (Sartorius MC21S, Germany) at Laboratoire
224 d'Aérodologie (Toulouse, France) before and after sampling in the weighing room after
225 equilibration at $20\text{--}23\text{ }^{\circ}\text{C}$ and the RH of 35%–45% for 24-hour. The absolute errors between
226 replicate weights were less than 0.015 mg for blank filters and 0.020 mg for sampled filters.

227 Total carbon (TC) was determined on 0.5 cm^2 punch-out of the filters by a carbon
228 analyzer (Ströhlein Coulomat 702C, Germany) at the Observatoire Midi-Pyrenees (OMP,
229 Toulouse, France). The quartz filter samples were subjected to a thermal pretreatment step
230 (kept at $60\text{ }^{\circ}\text{C}$ for 20 mins) in order to remove the volatile organic compounds (VOCs) and
231 eliminate water vapor. Subsequently the filters were combusted at $1200\text{ }^{\circ}\text{C}$ under O_2 and
232 detected as CO_2 in the carbon analyzer. Elemental carbon (EC) was obtained using a two-step
233 thermal method: step 1 consisted in a pre-combustion at $340\text{ }^{\circ}\text{C}$ under O_2 for 2 h in order to
234 remove organic carbon (OC); step 2 consisted in the oxidation of the remaining EC at
235 $1200\text{ }^{\circ}\text{C}$ under O_2 . The difference (TC-EC) yielded OC concentration (Benchrif et al., 2018;
236 Cachier et al., 2005).

237 To extract the water-soluble inorganic ions from the quartz filters, 1/4 of the filter was
238 placed in a separate 15 mL vials containing 10 mL distilled-deionized water ($18.2\text{ M}\Omega$
239 resistivity). The vials were placed in an ultrasonic water bath and shaken with a mechanical
240 shaker for 45 min ($15\text{ min} \times 3$ times) to extract the ions. The extracts were filtered through
241 $0.45\text{ }\mu\text{m}$ pore size microporous membranes. After that, three anions (Cl^- , NO_3^- and SO_4^{2-}) and



242 five cations (Na^+ , NH_4^+ , K^+ , Mg^{2+} and Ca^{2+}) in aqueous extracts of the filters were
243 determined by an ion chromatograph (IC) analyzer (Dionex-600, Dionex, Sunnyvale, CA,
244 USA), which was equipped with an AS11-HC anion column and a CS12 cation column for
245 separation. Details of the IC method are described in Bahino et al. (2018) and Cachier et al.
246 (2005).

247 One element: Fe (representing earth's crust emission) and ten heavy metals: V, Cr, Mn,
248 Co, Ni, Cu, Zn, Sb, Ba and Pb in PE $\text{PM}_{2.5}$ samples were determined by Energy Dispersive
249 X-Ray Fluorescence (ED-XRF) spectrometry (the PANalytical Epsilon 5 ED-XRF analyzer,
250 the Netherlands) on 1/4 of filters in this study as well. The relative errors for all measured
251 elements were $< 6\%$ between NIST Standard Reference Material (SRM) 2783 and our ED-
252 XRF results, which is well within the required range of error, demonstrating the accuracy of
253 ED-XRF. Replicate analysis of one quartz-fiber filter sample (five times) yielded an
254 analytical precision between 5.2%-13.9%. Details of the ED-XRF measurements are
255 described in Brouwer (2003) and Xu et al. (2012).

256 0.1-1.0 cm^2 punch-outs aliquots from the quartz filters were used to quantify the organic
257 compounds, including polycyclic aromatic hydrocarbons (PAHs), phthalate esters (PAEs) and
258 hopanes (see the specific organic species and their abbreviations measured in this study in
259 Table 5) by an in-injection port thermal desorption-gas chromatography/mass spectrometry
260 (TD-GC/MS) method. The approach has the advantages of shorter sample preparation time ($<$
261 1 min), minimizing of contaminations from solvent impurities, and higher sensitivity,
262 compared with the traditional solvent extraction-GC/MS method. The detail analytical
263 procedures have been reported in previous publications (Ho and Yu, 2004; Ho et al., 2008,
264 2011; Xu et al., 2013, 2016a). The results of the blank analyses showed only trace
265 contamination levels ($< 5.0\%$) of PE $\text{PM}_{2.5}$ samples concentrations.

266

267 2.4. Health risk assessment model

268 As we known, heavy metals and toxic organic species are associated with negative
269 personal exposure health effects (Škrbic et al., 2016; Val et al., 2013; Wang et al., 2017a; Xu
270 et al., 2018a). In this study, four heavy metals (Mn, Ni, Zn and Pb) and all PAHs and PAEs
271 species in PE $\text{PM}_{2.5}$ were selected to determine the personal exposure inhalation health risks.
272 The carcinogenic and non-carcinogenic health risks of $\text{PM}_{2.5}$ chemical species were
273 calculated according to the U.S. EPA health risk assessment model (USEPA, 2004, 2011).
274 The average daily exposure dose (D) via inhalation was estimated to assess the risk by the
275 equations (1) as follows:



276
$$D = (C \times R \times EF \times ED \times cf) / (BW \times AT) \quad (1)$$

277 the definitions and recommended values of parameters are shown in Table 2.

278 A hazard quotient (HQ) for non-cancer risk of heavy metals in PE PM_{2.5} samples can be
279 obtained from equation (2):

280
$$HQ = D/RfD \quad (2)$$

281 the threshold value of RfD indicates whether there is an adverse health effect during a certain
282 period. Hazard index (HI) can be obtained by summing up the individual HQ to estimate the
283 total non-cancer risks. If the HI < 1, then non-carcinogenic effect is impossible; HI ≥ 1,
284 adverse health effect might likely appear (Hu et al., 2012).

285 The incremental lifetime cancer risk (ILCR) of PAHs and PAEs in personal exposure
286 PM_{2.5} samples can be calculated by multiplying the cancer slope factor (CSF) of PAHs and
287 PAEs with D as equation (3):

288
$$ILCR = D \times CSF \quad (3)$$

289 for cancer risk, the value of 1×10^{-6} is an internationally accepted as the precautionary or
290 threshold value above which the risk is unacceptable (Jedrychowski et al., 2015).

291 It is worth noting that, among the nineteen PAHs, BaP has been used as an indicator of
292 PAHs carcinogenicity (Wang et al., 2006). The carcinogenic health risk of PAH species can
293 be assessed by [BaP]_{eq} instead (Yassaa et al., 2001) by equation (4):

294
$$\Sigma[BaP]_{eq} = \Sigma (C_i \times TEF_i) \quad (4)$$

295 Besides, the carcinogenic risk for PAEs was assessed by DEHP, which is identified as a
296 possible carcinogen to humans by the International Agency for Research on Cancer (IARC)
297 (IARC, 1982; Li et al., 2016). The definitions and recommended values of the parameters in
298 equations (2-4) are also shown in Table 2 and Table 3.

299

300 2.5. Questionnaire and time-activity diary

301 Questionnaire (Appendix A-C) and time-activity diary (Appendix D) were collected
302 from each participant during the sampling period, respectively, to fully grasp the basic
303 information, potential exposure sources and activities of participants. In the questionnaire,
304 personal information, family status, dermatological, asthma symptoms, medical history,
305 current health status and so on were first asked from each participant. Besides, the questions
306 for women include: (1) living habits and environment (past and current living conditions,
307 general living habits, cooking habits and domestic fuel type/usage); (2) work environment
308 and travel habits (workplace, work nature, working time and daily travel mode/time); and (3)
309 affected by the burning of domestic solid fuels and roasting meat. The questions for students



310 include: (1) living habits and environment (past and current living conditions, general living
311 habits, participation in household duties, the family cooking habits and domestic fuel
312 type/usage; distance from home to WB site); (2) school environment and travel habits (school
313 location and related environment and daily travel mode/time); and (3) affected by the burning
314 of waste and household air pollution source. The questions for drivers include: (1) living
315 habits and environment (past and current living environments, general living habits,
316 participation in household duties, the family cooking habits and domestic fuel type/usage); (2)
317 work environment and travel habits (motorcycle power type, driving conditions, working
318 time and daily travel mode/time); and (3) affected by the motorcycle emission and household
319 air pollution source.

320 The time-activity diaries requested the participants to mark on half an hour basis
321 (sleeping time excluded) to assess each microenvironment time spending and detailed
322 activities.

323

324 **3. Results and discussion**

325 *3.1. Personal exposure to PM_{2.5} and its chemical compositions*

326 *3.1.1. PE PM_{2.5} mass concentration*

327 The average personal exposure to PM_{2.5} (PE PM_{2.5}) mass concentrations were
328 331.7±190.7, 356.9±71.9 and 242.8±67.6 µg m⁻³ for women at Domestic Fires (DF), students
329 at Waste Burning (WB) and drivers at Motorcycle Traffic (MT) respectively in 2016 in
330 Southern West Africa (SWA). Among these three types of subjects, the average
331 concentrations of PE PM_{2.5} for women and students were quite similar, ~40% higher than the
332 drivers. PE PM_{2.5} ranged from 106.2 µg m⁻³ (nighttime in dry season, January 7th) to 1164.7
333 µg m⁻³ (daytime in wet season, July 5th) for women at DF; from 37.8 µg m⁻³ (nighttime in wet
334 season, July 8th) to 1137.0 µg m⁻³ (daytime in dry season, January 11th) for students at WB;
335 and from 65.0 µg m⁻³ (nighttime in wet season, July 11th) to 648.5 µg m⁻³ (daytime in dry
336 season, January 15th) for drivers at MT. The ranges and standard deviations of PE PM_{2.5}
337 concentrations were extremely large, especially for women, because there are direct
338 combustion sources close around the women workers in this study. Moreover, the variations
339 of personal physics activities and air pollution source intensities lead to a drastic fluctuation
340 for PE PM_{2.5}.

341 The average mass concentrations of PE PM_{2.5} were 358.8±100.5, 494.3±15.8 and
342 335.1±72.1 µg m⁻³ in dry season (January), and 304.6±284.5, 219.5±71.3 and 150.6±10.4 µg
343 m⁻³ in wet season (July) for women at DF, students at WB and drivers at MT, respectively



344 (Table 4). Compared to dry season, the reduction rate of PE $PM_{2.5}$ for women at DF in wet
345 season was approximately 15%, while the sharp reductions were observed for students and
346 drivers at a similar level by more than 50%. PE $PM_{2.5}$ concentrations reducing could be
347 attributed to the occurrence of increased levels of rainfall in wet season in SWA (Table 1),
348 which causes the large reduction of road dust exposed to drivers and limits the garbage
349 spontaneous combustion significantly around students. Moreover, large scale transport of
350 mineral dust and combustion aerosols emitted by savannah wild fires contribute significantly
351 to the aerosol load during the dry season (Djossou et al., 2018), which is more important at
352 WB and MT than at DF (crowded community environment).

353 PE $PM_{2.5}$ mass concentrations in the daytime were much higher than those at night, no
354 matter in dry or wet season (Table 4 and Figure 3). The 12-hour averaged PE $PM_{2.5}$
355 concentrations show day/night (D/N) ratios of 3.4 (3.8 in dry season and 3.1 in wet season),
356 2.7 (2.8 and 2.5) and 2.4 (1.5 and 3.3) for women at DF, students at WB and drivers at MT,
357 respectively. Intensive human activities during the day, such as solid fuel combustion, waste
358 combustion or motor vehicle emission around the different group subjects, enhance the levels
359 of $PM_{2.5}$ exposure in the daytime. For example, lower $PM_{2.5}$ personal exposure level for
360 students at night at WB can be explained also by the fact that the participants in this study
361 usually spent most of the time indoors at night with limited physical activity, leading them to
362 be able to stay a distance and/or shelter from obvious emission sources (waste combustion)
363 outdoors. Besides, big fluctuations of D/N ratios for drivers were observed, with lower value
364 in dry season and higher in wet season. Relatively lower D/N ratio probably attributes to
365 nighttime driving (18:30 to 21:00 UTC) after dinner, which enhances their $PM_{2.5}$ exposed
366 levels from vehicle emission and road dust. Much higher D/N ratios in wet season attribute to
367 the increase in precipitation in wet season in Cotonou (Table 1), especially at night (Sealy et
368 al., 2003), leading to the lower exposure for drivers at night after aerosol scavenging and the
369 less driving time in wet season because of the unfavorable weather.

370 The average PE $PM_{2.5}$ levels are compared to the weekly ambient $PM_{2.5}$ concentrations
371 (Djossou et al., 2018) obtained in the same area and similar sampling period. The average PE
372 $PM_{2.5}$ were 3.0 and 2.0 times the ambient values found at DF, and 6.1 and 8.8 times at MT in
373 dry and wet seasons, respectively. The highest PE $PM_{2.5}$ to ambient (A) (PE/A) ratios were
374 found at WB, i.e., 10.3 in dry and 10.5 in wet season. Such large PE/A ratios are due to the
375 impact of waste combustion on the respiratory exposure of residents in this area, especially
376 on children; on the other hand, high PE/A ratios attribute to the fact that WB site is located in
377 the poorest region of Abidjan, where the extremely simple stove and wood used at home



378 (Figure S1d), leading to a very high personal $PM_{2.5}$ exposure level indoors during the cooking
379 time in this area (especially for student B who was in charge of cooking at home sometimes
380 recorded in the activity logging and questionnaire). Meanwhile, the ambient $PM_{2.5}$ sampling
381 equipment at WB was neither fixed very close to (blue marker in Figure 1C) nor in the
382 downwind direction of the landfill (Djossou et al., 2018), which direct suggests the huge
383 differences between the ambient and personal exposure $PM_{2.5}$ concentrations.

384 Moreover, we also compare the daytime PE $PM_{2.5}$ mass concentrations with the daytime
385 ambient $PM_{2.5}$, which collected in the same area and exactly the same dates as the personal
386 exposure sampling period. The average daytime women PE $PM_{2.5}$ were 3.7 and 1.2 times of
387 the ambient $PM_{2.5}$ at DF in dry and wet seasons, respectively, which was similar as the
388 finding from the comparison with the weekly $PM_{2.5}$ mentioned above. But for the students at
389 WB and drivers at MT, PE/A ratios were both much smaller than those compared with the
390 weekly ambient $PM_{2.5}$, 5.1 and 7.0 for the students at WB, 1.9 and 3.3 for the drivers at MT
391 in dry and wet seasons, respectively. PE/A ratios for students were also showed the highest
392 values. PE/A ratios observed all above 1.0 and the great variability of $PM_{2.5}$ mass
393 concentrations between personal exposure and ambient samples imply that fix-point sampling
394 is likely to underestimate the $PM_{2.5}$ personal exposure and consequent human health hazards,
395 and confirm the importance of portative PE $PM_{2.5}$ sampling to $PM_{2.5}$ health risk assessment
396 again.

397 3.1.2. PE $PM_{2.5}$ chemical compositions

398 Table 4 summarizes the average concentrations of PE $PM_{2.5}$ chemical compositions,
399 including carbon fractions (OC and EC), water-soluble inorganic ions and several heavy
400 metals. Total carbon (TC) was the most important chemical species in PE $PM_{2.5}$, accounting
401 for $24.4\% \pm 4.5\%$, $16.6\% \pm 2.0\%$ and $17.8\% \pm 4.9\%$ of PE $PM_{2.5}$ mass of women, students and
402 drivers, respectively. High level of OC proves the strong contribution of combustion sources
403 to PE $PM_{2.5}$ in SWA in this study (Djossou et al., 2018; Ouafu-Leumbe et al., 2017). The OC
404 and EC concentrations varied significantly, ranging from 28.3 to 460.0, 8.0 to 189.9 and 14.7
405 to $65.1 \mu\text{g m}^{-3}$ for OC and 1.5 to 31.1, 0.8 to 35.1 and 1.9 to $18.2 \mu\text{g m}^{-3}$ for EC for women,
406 students and drivers, separately. The OC concentration ($83.2 \mu\text{g m}^{-3}$) and percentage (24.4%)
407 in women PE $PM_{2.5}$ samples were the highest among the three types of exposed participants,
408 due to their close contact with the ignition and the direct burning of the solid fuels (wood in
409 this study) and roasting meat at the workplace, and even cooking at home, etc. However, the
410 EC concentrations and proportions for these three targets were similar ($8.4\text{-}10.5 \mu\text{g m}^{-3}$ and
411 $3.0\%\text{-}3.5\%$), meaning that EC is less affected by human activities related to combustion



412 sources in this study.

413 The OC and EC ratio (OC/EC) has been used to determine emission and transformation
414 characteristics of carbonaceous aerosols (Cao et al., 2008). OC/EC averaged 9.9 ± 5.3 for
415 women at DF, 6.1 ± 0.7 for students at WB, and 5.8 ± 2.7 for drivers at MT. The previous
416 studies (Cao et al., 2008; Li et al., 2009; Tian et al., 2017) suggested that average OC/EC
417 characterizes 1.1 as motor vehicle exhaust, 2.7 as coal combustion and 9.0 as biomass
418 burning. The OC/EC in the present study points out that biomass burning emission was the
419 main contributor to carbonaceous aerosols for women at DF, and the mixed emissions from
420 biomass and coal burning, even or/and motor vehicle exhaust dominated the carbonaceous
421 aerosol sources for students at WB and drivers at MT. The OC/EC was almost always higher
422 during wet than dry season, which may be related to the fact that the higher relative humidity
423 in wet season favors the formation of secondary organic carbon (SOC) (Huang et al., 2014).
424 Drivers' daytime OC/EC shows relatively low (average: 3.7) and stable ratios in wet and dry
425 seasons, indicating that motor vehicle exhaust was the most important source to drivers' OC
426 and EC in the daytime, consistent with the working environment of motorcycle drivers in this
427 study. Personal exposure of women displays the higher (average: 13.9) and more scattered
428 OC/EC than those collected from students and drivers in wet season (Figure 4), resulting
429 from the particularly high and dramatic changes individual exposure to obvious carbonaceous
430 aerosol sources (wood burning and grilling).

431 In a previous study of Djossou et al. (2018) about OC/EC at ambient (A) environment,
432 OC/EC in personal exposure samples were about 1.2 and 2.5 times of the ambient values in
433 dry and wet seasons for women at DF, 1.7 and 2.8 times for students at WB, and 1.1 and 2.0
434 times for drivers at MT. Therefore, the higher OC/EC values in personal exposure samples
435 resulted from some specific individual's activities and potentially contaminated
436 microenvironments (Crist et al., 2008; Meng et al., 2009). From the results we can also see
437 that the influences of precipitation and other meteorological factors on OC/EC of the ambient
438 samples are less than personal exposure samples (Dry season OC/EC was more comparable
439 between the ambient and personal exposure samples in this study).

440 The average concentrations of total measured water soluble inorganic ions were
441 23.6 ± 12.8 , 35.5 ± 18.3 and 22.7 ± 5.0 $\mu\text{g m}^{-3}$ for women at DF, students at WB and drivers at
442 MT, accounting for $8.5\pm 1.0\%$, $12.1\pm 2.7\%$ and $11.9\pm 0.4\%$ of PE $\text{PM}_{2.5}$ mass,
443 respectively. Unlike of the ion compositions in polluted cities of China (SO_4^{2-} , NO_3^- and
444 NH_4^+ were the most abundant ions in ambient $\text{PM}_{2.5}$, accounting for 50%-90% of measured
445 ions and ~30% of $\text{PM}_{2.5}$ mass) (Xu et al., 2016b; Zhang et al., 2013), Ca^{2+} , a marker of



446 fugitive dust, was the most abundant ion, accounting for ~28% (from 25.3% to 29.3%) of
447 ions in this study, following by Cl^- , SO_4^{2-} and K^+ for women at DF, Na^+ SO_4^{2-} and Cl^-
448 for students at WB, SO_4^{2-} , Na^+ and NO_3^- for drivers at MT. It can thus be seen that the particle
449 resuspension by personal activities is the main contributor to PE $\text{PM}_{2.5}$ in SWA (Chen et al.,
450 2017; Xu et al., 2015). The day and night variations of Ca^{2+} contribution to total ions also
451 illustrate this conclusion (day=30.6% and night=22.8%). SO_4^{2-} forms primarily through
452 atmospheric oxidation of SO_2 emitted mainly from coal and diesel combustion (Seinfeld and
453 Pandis, 2006, Xu et al., 2016b). As the second most enriched ion, the average proportion of
454 SO_4^{2-} was 17.7%, which implies that purification raw coal and diesel (Wang et al., 2013) can
455 be applied in this area to lead to lower sulfur emissions and therefore decrease the personal
456 exposure to SO_4^{2-} in $\text{PM}_{2.5}$. Drivers' SO_4^{2-} exposure levels were 33% and 40% higher than
457 women and students respectively, which may indirectly indicate that the emission of SO_4^{2-}
458 precursor SO_2 is higher in Cotonou or targeted drivers affected by vehicle emissions are
459 exposed to high SO_2 or SO_4^{2-} , especially from the diesel vehicles.

460 Generally, Na^+ and Cl^- were the third and fourth ranked ions. The sampling sites in SWA
461 cities in this study are all close to the sea and are affected by sea salt particles strongly. It's
462 also worth noting that biomass burning marker- K^+ (Kang et al., 2004; Zhang et al., 2014b)
463 displays a high absolute average concentration ($3.4 \mu\text{g m}^{-3}$) and percentage (14.5%) in
464 women PE $\text{PM}_{2.5}$ samples, confirming their distinct exposure from biomass burning during
465 the roasting at the workplace. As we know, NO_3^- derives from NO_x emitted mainly from
466 motor vehicle exhaust (especially gasoline vehicle), industry and power plants (Seinfeld and
467 Pandis, 2006; Xu et al., 2016b). Additional consideration here: industry is not well-developed
468 in this area (much less industry in Cotonou than Abidjan) and is not the main source of $\text{PM}_{2.5}$
469 (Ouafo-Leumbe et al., 2017). It suggests that motor vehicle emission contributed to drivers'
470 exposure obviously in this study, comparing with women at DF and students at WB,
471 consistent with the conclusion about SO_4^{2-} above.

472 The concentrations of 10 targeted heavy metals, including V, Cr, Mn, Co, Ni, Cu, Zn, Sb,
473 Ba and Pb, can be found in Table 4. The total concentrations of these 10 elements were
474 1.4 ± 0.3 , 3.9 ± 6.5 and $0.8 \pm 0.2 \mu\text{g m}^{-3}$ for women at DF, students at WB and drivers at MT
475 during the sampling period, accounting for $0.7\% \pm 0.4\%$, $1.0\% \pm 1.2\%$ and $0.4\% \pm 0.1\%$ of PE
476 $\text{PM}_{2.5}$, correspondingly. The heavy metal exposed concentration of students was 1.8 and 3.9
477 times higher than those for women and drivers, resulting mainly from the garbage
478 combustion at landfill which emit extremely high level of heavy metals as we known (Wang
479 et al., 2017b). The D/N ratios ranged from 0.8 to 2.1 for women and drivers, but averaged 4.0



480 in dry season and 7.0 in wet season for students. There are two reasons for this phenomenon:
481 the first reason could be related with the intense physical activities from the students and
482 strong disturbances from landfill workers in the daytime at landfill; the second reason is that
483 spontaneous combustion of waste occurs frequently during the day, because of less
484 precipitation and higher air temperature at daytime. Ba, Zn and Mn were found to be the
485 dominant heavy metals, ~73% of elemental concentration in all samples. It is worth
486 mentioning that Ba took up a decisive advantage over other elements, accounting for more
487 than half of all the elements for students. Because Ba is usually added to rubber and plastic
488 products to improve acid and alkali resistance, at the same time these products are the main
489 components of the garbage at landfill in this area (Feng et al., 2006). Zn and Mn ranked the
490 first and second places for drivers at MT, which are mainly derived from the motor oil
491 additive, tyre wear and brake pads worn (Zhao and Hopke, 2006).

492

493 *3.2. Mass balance of personal exposure to PM_{2.5}*

494 Calculation of mass balance for PM_{2.5} is an effective method to figure out the principal
495 components in PM_{2.5} and for its source discrimination (Gokhale et al., 2008). PE PM_{2.5} mass
496 in this study can be classified into six parts: organic matter (OM), EC, water-soluble
497 inorganic ions, geological material (GM), heavy metals and unknown part (Figure 5). The
498 first five main parts can explain 78.3% to 90.6% of total PE PM_{2.5} mass concentrations in this
499 study. Unknown part may include water and other undetected substances in PE PM_{2.5}. For
500 OM, since the chemical composition of the aerosol organic fraction is largely unknown,
501 conversion factor 1.4 (1.4 corrects the organic carbon mass for other constituent associated
502 with the organic carbon molecule) is generally used (Turpin and Lim, 2001) to shift OC to
503 OM by equation (5):

504
$$\text{OM} = 1.4 \times \text{OC} \quad (5)$$

505 based on the equation (5), OM accounted for 34.1%±6.3%, 23.3%±2.8% and 24.9%±6.9% of
506 PE PM_{2.5} mass for women at DF, students at WB and drivers at MT, respectively, indicating
507 that there are distinct sources to PE PM_{2.5} OC for women at DF. According to the
508 questionnaires, the combustion sources, such as roasting meat/peanuts and burning wood,
509 should be the sources to OC for women personal exposure samples, which consistent with the
510 results mentioned above.

511 In addition, Fe has been widely used to estimate the upper limit of GM in previous
512 studies (Taylor and McLennan, 1985). Fe constitutes about 4.0% of the Earth's crust in dust
513 of the earth's crust (Cao et al., 2005). The amount of GM is calculated by equation (6):



514 $GM = (1/4.0\%) \times Fe$ (6)
515 based on this equation, it is found that GM contributed $35.8\% \pm 2.1\%$, $46.0\% \pm 3.7\%$ and
516 $42.4\% \pm 4.7\%$ of PE $PM_{2.5}$ mass concentrations for women at DF, students at WB and drivers
517 at MT, respectively. Fugitive dust, including road dust resuspension from disturbance of
518 motor vehicles and humans, construction dust from uncovered construction sites and the dust
519 related to burning activities, could be the domination sources to PE $PM_{2.5}$ in this study. OM
520 and GM show the almost identical proportions (34.1% and 35.8%) of PE $PM_{2.5}$ mass for
521 women at DF. GM percentages for students and drivers were approximately 10% and 7%
522 higher than that for women. Therefore, the fugitive dust related contributions are the most
523 important sources for PE $PM_{2.5}$ in this less developed area, meaning that there are nearly half
524 PE $PM_{2.5}$ contribution sources of students and drivers attributable to human physical
525 activities. As mentioned above, it is surprising to note that the secondary formed ions (SO_4^{2-} ,
526 NO_3^- and NH_4^+), even total measured water-soluble inorganic ions show the exceedingly low
527 proportions in PE $PM_{2.5}$ for all subjects. This reconfirms the limited contribution to PE $PM_{2.5}$
528 from secondary formation ionic sources.

529 From Figure 5, evident diurnal distinguishes are observed in two major chemical
530 compositions (OM and GM) in this study. We can see that GM exhibits the lower proportion
531 at night (35.3%) than daytime (47.5%), indicating its close relationship with human activities.
532 For different seasons, we find the higher GM for each type of participant in dry season,
533 because of the harmattan haze bringing mineral dust and the lake of precipitation increasing
534 road dust resuspension. Moreover, OM shows the equal or lower proportion in PE $PM_{2.5}$ at
535 daytime (25.0%) that nighttime (30.0%), which is mainly related with the meteorological
536 parameters (they affect the formation of secondary organic carbonaceous aerosol) and
537 combustion source variations between day and night. There is an exception in the last case,
538 i.e., OM proportion at daytime women PE $PM_{2.5}$ was much higher (50.8%) than nighttime
539 (38.2%) in wet season, due to the influence from the damp wood burning at the working time.
540 As we know, the damp wood burning emits more smoke (PM) than dry wood (Shen et al.,
541 2013) or change in emission factors (Keita et al., 2018).

542

543 4. Fingerprint organic species in personal exposure to $PM_{2.5}$

544 In this section, we use organic fingerprint markers that indicate specific emission
545 sources to further investigate the sources and detailed characteristics of PE $PM_{2.5}$ for different
546 populations. Unlike PE $PM_{2.5}$ mass concentration variations (students > women > drivers),
547 organic fingerprint measured in this study, such as PAHs, PAEs and hopanes (Table 5), show



548 different concentration orders in this study. The average PM_{2.5}-bound PAHs, PAEs and
549 hopanes mass concentrations were 54.8±20.3, 986.8±82.2 and 27.9±1.0 ng m⁻³ in this study,
550 respectively, showing a very serious PM_{2.5} organic pollution in SWA region. The descending
551 following orders were women > students > drivers for PAHs, students > women > drivers for
552 PAEs, and drivers > women > students for hopanes (Table 5 and Figure 6).

553 4.1. PAHs

554 The total quantified PAHs (ΣPAHs) accounted for 0.12%-0.21% of PE PM_{2.5} mass
555 concentration. BbF was the most abundant PAH for women at DF, followed by BaP and IcdP.
556 The average BbF concentration (the marker of low temperature combustion, such as wood
557 burning) was 11.6±19.2 ng m⁻³, accounting for up to approximately 15.0% of the ΣPAHs for
558 women PE samples (Table 5) (Wang et al., 2006). While the most abundant PAHs for
559 students at WB and drivers at MT were IcdP (6.4±4.5 ng m⁻³) and BghiP (6.4±0.5 ng m⁻³),
560 respectively, which indicate the contributions from the waste incineration or high temperature
561 fuel combustion (gasoline vehicle emission) (Baek et al., 1991; Wang et al., 2006). The
562 ΣPAHs average concentrations in wet season increased 326% and 52% for women at DF
563 (125.4±54.8 ng m⁻³) and drivers at MT (44.6±10.8 ng m⁻³) than those in dry season (29.4±5.6
564 and 29.4±4.4 ng m⁻³ respectively), while the ΣPAHs decreased 42% in wet season (36.8±15.7
565 ng m⁻³) compared with dry season (62.9±45.0 ng m⁻³) for students at WB. The dramatic
566 increase in women's exposure to PAHs is mainly due to the increase in humidity (moisture
567 content) of the wood used for grilling meat in wet season, resulting in PAH emissions sharp
568 raising from wood combustion (Shen et al., 2013). The restraint of waste combustion in wet
569 season is the main factor in the decrease of students' exposure to PM_{2.5}-bound PAHs at
570 landfill, in accordance with PE PM_{2.5} mass seasonal change pattern. PE PAHs concentrations
571 were measured in Cotonou in the previous study (Fanou et al., 2006), the result showed that
572 the level of total PAHs associated with particles ranged from 76.21 to 103.23 ng m⁻³ for 35
573 taxi-moto drivers in March 2001. The PAH levels determined in this study for drivers at MT
574 site was 50%-64% lower than the values in Fanou et al. (2006) study, suggesting that the
575 motorbike driver exposure to PAHs in this region has improved.

576 As shown in Figure 6A, differing from the almost unchanged diurnal variation (daytime >
577 nighttime) of PE PM_{2.5} and its major chemical components discussed above, PE PAHs show
578 unstable diurnal variations for these three types of target populations: 1) Women at DF: the
579 daytime concentrations during the wet and dry seasons were both higher than those at
580 nighttime, suggesting that women's intensive combustion activities at daytime (roasting meat
581 and burning wood) strongly impacted the PE PAHs. The average D/N ratios were 1.7 in dry



582 season with the 12-hour average Σ PAHs of $37.4 \pm 25.1 \text{ ng m}^{-3}$ for daytime and $21.4 \pm 17.2 \text{ ng}$
583 m^{-3} for nighttime and 3.5 in wet season with $195.6 \pm 121.9 \text{ ng m}^{-3}$ for daytime and 55.3 ± 44.3
584 ng m^{-3} for nighttime; 2) Students at WB: nighttime PE PAHs were higher in dry season and
585 lower in wet season compared with daytime levels, with the average D/N ratios of 0.7 and 1.8,
586 respectively. The higher concentrations of combustion markers-BbF and BeP were observed
587 during the day, while the higher concentrations of gasoline vehicle emission markers-DahA
588 and BghiP were found at night (Baek et al., 1991; Wang et al., 2006), which was related to
589 the garbage truck for waste transportation from city to the landfill during night. Moreover, we
590 should also note that the impact of garbage truck emission was offset by $\text{PM}_{2.5}$ wet deposition
591 during the wet season; 3) Drivers at MT: we are surprised to see that the average dry season
592 D/N ratio was 0.8 with the Σ PAHs of $26.3 \pm 7.6 \text{ ng m}^{-3}$ for daytime and $32.5 \pm 13.8 \text{ ng m}^{-3}$ for
593 nighttime, and the average wet season D/N ratio was 0.3 with $21.9 \pm 8.4 \text{ ng m}^{-3}$ for daytime
594 and $67.3 \pm 23.7 \text{ ng m}^{-3}$ for nighttime, respectively. The high nighttime Σ PAHs concentrations
595 and low D/N ratios for drivers in this study may be explained by the possibility that there are
596 potential combustion sources (PAH sources) around participant drivers (especially around
597 drivers' home at night) in Cotonou, Benin at night rather than the motor vehicle exhaust,
598 especially in wet season (combustion emission marker BaP was the highest PAH species at
599 night in wet season), although the drivers exposed to the traffic emissions during the night
600 working time (18:30 to 21:00 UTC). Further studies are required to confirm the findings and
601 figure out the reasons. Even so, the highest PAH individual species for drivers PE samples
602 was BghiP (gasoline vehicle emission marker) in both wet and dry seasons, proving the
603 obvious influence from the motor vehicle emissions to motorcycle drivers (Baek et al., 1991;
604 Wang et al., 2006).

605 Diagnostic ratios of PAHs with similar molecular weights have been widely used in
606 source identification (Tobiszewski and Namiesnik, 2012; Yunker et al., 2002). In our study,
607 the average values of $\text{BeP}/(\text{BeP}+\text{BaP})$ and $\text{IcdP}/(\text{IcdP}+\text{BghiP})$ were 0.47 and 0.52 for women
608 at DF, 0.51 and 0.52 for students at WB, and 0.64 and 0.34 for drivers at MT, respectively
609 (Figure 7), showing that the impacts of different atmospheric pollution sources on the
610 different type participants are very significant, and that the diagnostic ratios of PAHs can be
611 applied to identify the source of PAHs in PE $\text{PM}_{2.5}$ effectively. The average $\text{BeP}/(\text{BeP}+\text{BaP})$
612 ratio ranged from 0.47 to 0.64, comparable with those reported in Guangzhou (0.41-0.72) and
613 Xi'an (0.59-0.73) of China (Li et al., 2005; Xu et al., 2018b) and lower than that reported in
614 Shanghai (all samples > 0.70), China (Feng et al., 2006), implying the low oxidability of the
615 PAHs in SWA cities (less developed than Chinese cities). PAHs in drivers' PE samples are



616 more prone to aging (the average ratio was 1.3-1.4 times of those for women and students)
617 because of the re-suspension of road dusts where PAHs are attached to (longer residence
618 lifetime) and longer outdoor activity time (more sunlight); and more fine and ultra-fine
619 particles-bound PAHs from high-temperature combustion in motor vehicular engine, which
620 are more easily photochemical oxidation in the air (Baek et al., 1991; Lima et al., 2005). The
621 difference of BeP/(BeP+BaP) ratios in dry and wet seasons is not obvious and no fixed rule.
622 However, this ratio exhibits a significant day-night change, with the values of 0.59 at daytime
623 and 0.49 at nighttime. It means that more beneficial meteorological conditions at daytime
624 (such as more sunlight) and stronger individual physical activity (increasing the time of
625 particulate re-suspension) are more conducive to the aging of PM_{2.5} and its bounded PAHs.
626 Moreover, IcdP/(BghiP+IcdP) of < 0.2, 0.2-0.5 and > 0.5 represent petrogenic, petroleum
627 combustion and a mix of grass, wood, and coal combustions, respectively (Yunker et al.,
628 2002). The quite low ratios for drivers at MT (0.34) indicates that the PAHs in those samples
629 were mainly produced from motor vehicle emissions (petroleum combustion), while grass,
630 wood and coal combustions were more dominant for women at DF (0.52) and students at WB
631 (0.52) (Figure 7). IcdP/(IcdP+BghiP) ratio in all samples from our study shows not
632 significant seasonal variation.

633

634 4.2. Phthalate esters (PAEs)

635 Phthalate esters are widely used plasticizers in plastic materials and can be released into
636 the air from the matrix evaporation and plastics combustion (Gu et al., 2010; Wang et al.,
637 2017a). The personal exposure levels of PAEs could be mainly attributed to the usage of the
638 household products, painting material at home, plastic waste incineration and municipal
639 sewage release (Zhang et al., 2014a). The total concentrations of six phthalate esters (DMP,
640 DEP, DBP, BBP, DEHP and DNOP) and one plasticizer (DEHA) (named ΣPAEs for all these
641 seven species) were 882.0±193.3, 1380.4±335.2 and 698.1±192.4 ng m⁻³, respectively, for
642 women at DF, students at WB and drivers at MT (Table 5). DEHP was the most dominant
643 PAE species, followed by DBP in this study for all the three kinds of participants. DEHP is
644 mainly used as a plasticizer for polyvinyl chloride (PVC). And together with DBP, they are
645 the most widely used phthalate esters globally (Meng et al., 2014). The average DEHP and
646 DBP concentrations were 543.6 and 304.6 ng m⁻³, accounting for up to approximately 55.1%
647 and 30.9% of the ΣPAEs, respectively (Figure 6B). The elevated ΣPAEs for students at WB
648 in this study are mostly result from combustion of the plastic products at landfill. The results
649 in this study are similar as the previous studies carried out in Xi'an, Tianjin of China (Kong



650 et al., 2013; Wang et al., 2017a). The Σ PAEs ranged from 376.6 to 1074 ng m⁻³ in outdoors,
651 and from 469.2 to 1537 ng m⁻³ in student classrooms in Xi'an (Wang et al., 2017a), in which
652 DEHP and DBP were also the dominant species, totally accounting for 68% and 73% of the
653 Σ PAEs in outdoor and indoor environments, respectively.

654 The average concentrations of the Σ PAEs for women at DF, students at WB and drivers
655 at MT were comparable in dry season. But the average concentrations were 927.2±154.9,
656 1929.8±340.4 and 594.6±16.6 ng m⁻³ in wet season in this study, 1.1, 2.3 and 0.7 times of the
657 Σ PAEs values in dry season (Figure 6B). A significant increase in students PE Σ PAEs at WB
658 in wet season can be attributed to the enhanced PAEs emission in the day (3173.6±1028.3 ng
659 m⁻³), consistent with the findings on PE PM_{2.5} above. Dry and wet seasons led to almost
660 similar PAEs profiles with different day and night variations (Figure 6B). The average D/N
661 ratios of the Σ PAEs in dry season show limited changes, with the values of 1.0, 1.0 and 1.3,
662 respectively, for women, students and drivers; while 1.1, 4.6 and 0.7 in wet season.
663 Noticeably different D/N ratios between two seasons observed in this study for students at
664 WB is interrelated with the human activities (specially related to the plastic material
665 emissions) and the subdued waste spontaneous combustion resulting from diurnal variations
666 of meteorological parameters (more precipitation at night in wet season) mentioned in Sect.
667 3.1.1.

668

669 4.3. Hopanes

670 Hopanes have been used as markers for fossil fuel combustion, especially for petroleum
671 combustion (Simoneit, 1999; Wang et al., 2009). The average concentration of drivers who
672 exposed to the eight hopanes (Σ hopanes) in this study was 50.9±7.9 ng m⁻³, 2.0 and 2.3 times
673 higher than for women at DF (17.1±6.4 ng m⁻³) and students at WB (15.6±6.1 ng m⁻³) (Table
674 5), respectively, which proves the extremely high driver personal respiratory exposure
675 contribution from the motor vehicle emissions (gasoline combustion) in this study. Then, it is
676 important to note that numbers of automobiles are rapidly increasing in SWA cities, which
677 further exacerbates the air pollution and related health problems there. The Σ hopanes show
678 the unobvious seasonal variations for three kinds of exposure participants, i.e., 0.9, 1.8 and
679 0.7 times Σ hopane concentrations were observed in dry season of those in wet season.
680 Although the Σ hopane concentrations were changeable in this study among three sites, the
681 distribution of individual species of hopanes were similar for each participant. $\alpha\beta$ -NH and
682 $\alpha\beta$ -HH were two dominant hopanes in all PE PM_{2.5} samples, with the average concentrations



683 of 6.0 and 6.5 ng m⁻³ and the percentages of 21.4% and 23.3% of the Σhopanes, respectively
684 (Table 5 and Figure 6C).

685 Compared with D/N ratios of the ΣPAHs and ΣPAEs, hopanes exhibit a more stable
686 diurnal trend in this study, namely, daytime concentrations were always greater than
687 nighttime, owing to the obvious traffic emissions during the day. For women at DF, D/N ratio
688 was both 2.0 in dry and wet seasons, with the Σhopanes of 24.0±11.1 and 12.2±5.0 ng m⁻³ for
689 daytime and nighttime in dry season, and 21.4±17.5 and 10.9±3.6 ng m⁻³ in wet season.
690 Emphasize that D/N ratio of the Σhopane for drivers at MT presents the highest value (11.5)
691 for all the detected chemical species in this study, with the concentrations of 78.0±19.1 and
692 44.9±16.4 ng m⁻³ for daytime and nighttime in dry season, and 74.2±16.3 and 6.5±1.7 ng m⁻³
693 in wet season. We notice that the daytime concentrations were comparable for drivers
694 between two seasons, while the nighttime hopanes in wet season were washed away by
695 rainfall mostly, resulting in a very large drop in concentration levels.

696 Therefore, although PAHs, PAEs and hopanes are not abundant components in PE PM_{2.5},
697 these fingerprint organics can more accurately trace the contribution of air pollution sources
698 to PM_{2.5}. The PAHs, PAEs and hopanes representing emissions from combustion sources,
699 plastics emissions and fossil fuel combustion emissions (gasoline vehicles) respectively are
700 very well matched to the potential air pollution sources around these three type participants in
701 this study. The results not only indicate that the PM_{2.5} respiratory exposure was strongly
702 contributed from the environmental pollution sources and individual activities, but also prove
703 the successful application of organic tracers in human exposure study.

704

705 *5. Health risk assessment of personal exposure to PM_{2.5}*

706 Non-cancer risks of four heavy metals (Mn, Ni, Zn and Pb) and cancer risks of PAHs
707 and PAEs via inhalation exposure way for women at DF, students at WB and drivers at MT
708 are shown in Table 6. In general, the non-carcinogenic risks of Mn and Pb were relatively
709 higher than those of Ni and Zn, but still well-behind the international threshold value (1.0).
710 Among those four metals, Hazard Quotient (HQ) of Pb in wet season for students at WB was
711 the highest (2.95E-02), which suggests that Pb non-carcinogenic risk to children is obvious in
712 that area compared with other participants and metals. Except that Ni shows the stable wet
713 season greater non-carcinogenic risk than dry season for all three kinds targets, there was no
714 stable change in dry/wet season risks of other components. Summing up these four metals,
715 Hazard Index (HI) values for women at DF, students at WB and drivers at MT in dry and wet
716 seasons were also represented in Table 6. Dry/wet season ratios of HI were 0.9, 0.5 and 2.3



717 for women, students and drivers, suggesting that the non-cancer risk of personal exposure to
718 metals in PM_{2.5} in dry season was much higher than that in wet season for drivers, owing to a
719 mass of fugitive dust on the road in dry season. Moreover, the yearly average HI levels were
720 8.06E-03, 4.13E-02 and 8.68E-03 for women at DF, students at WB and drivers at MT,
721 respectively, showing the highest non-cancer health risks from the heavy metals in PM_{2.5} for
722 students, 5.1 and 4.8 times of women and drivers. Overall, Mn, Zn, Ni, Pb and HI were all
723 within the safety limit for all populations involved in this study, pointing out the negligible
724 non-cancer health risks of heavy metals in PM_{2.5} in SWA region.

725 In Table 6, the ILCRs of PAHs were all beyond 1×10^{-6} (international acceptable level),
726 suggesting non-negligible cancer risks of PAHs for women at DF, students at WB and drivers
727 at MT whenever dry or wet season. Meanwhile, the ILCRs of PAEs were all below 1×10^{-6} ,
728 well within the safety limit of cancer risk. For all types of targets, PE PM_{2.5}-bound PAHs and
729 PAEs in wet season were more likely to cause cancer risks than dry season; thus, the seasonal
730 changes, mainly due to increased humidity, result in an increased personal exposure cancer
731 risks to toxic organic species in PM_{2.5}. In dry season, the average ILCR values of PAHs were
732 comparable for women and drivers, both ~50% lower than those for students, implying the
733 high toxicity originated from the waste burning sources and high sensitivity to juveniles. In
734 wet season, PAHs exhibit the highest ILCR for women at DF, 2.5 and 2.7 times of those for
735 students and drivers, respectively. It can be seen that the domestic wood burning and grilling
736 meat can trigger nearly ten times safe limit of cancer risks to target women in this study. The
737 cancer risks of PAEs show the similar pattern in dry and wet seasons, with the descending
738 order of students at WB > women at DF > drivers at MT (Yang et al., 2011). The
739 carcinogenic risks of PAEs for drivers in traffic environment was the lowest, much lower (45%
740 and 76% lower in dry and wet seasons) than PAEs for students who are close to the source of
741 waste incineration. In a word, the ILCRs of PAHs exceeded the threshold value of 1×10^{-6} for
742 all the participants, indicating that the carcinogenic PAHs are a threat to the individual's
743 health and subsequently alerting a need of effective control in SWA. However, PAEs show
744 limited carcinogenic risks in this study, but the effect of waste burning source to students is
745 needed to pay more attention and reasonable controls for both PM_{2.5}-bound heavy metals and
746 organic compounds.

747 In addition, it should be noted that the non-cancer and cancer risks could be potentially
748 underestimated since many toxic chemical components could not be detected in this study. It
749 is concluded based on the data that different targets present different levels of risks from
750 different chemical species in PE PM_{2.5} from various air pollution sources. We must pay



751 attention to heavy metal non-cancer health risks via inhalation way, especially Pb and Mn for
752 students at WB site as well as PAHs cancer risks for women at DF site in wet season in SWA
753 region.

754

755 **6. Conclusions**

756 We explore the chemical characteristics and health risks of personal exposure to $PM_{2.5}$
757 (PE $PM_{2.5}$) from different typical anthropogenic air pollution sources in Southern West Africa.
758 Our study finds that organic matter and geological material are the almost identical
759 proportions (34.1% and 35.8%) for women at domestic fire site. Nearly half contribution to
760 PE $PM_{2.5}$ for students at waste burning site and drivers at motorcycle traffic site comes from
761 fugitive dust. Therefore, the primary source (mainly dust) is the most important source for PE
762 $PM_{2.5}$ in these undeveloped regions. The contribution to PE $PM_{2.5}$ from heavy metals was
763 higher for students (1.0%), owing to the waste burning emissions strongly, leading to the
764 highest non-cancer risk among these three kinds of participants, as well as the extremely high
765 PAEs concentrations (indicator of plastic emissions). PE $PM_{2.5}$ -bound PAHs concentration for
766 women at domestic fire site was 1.6 times for students and 2.1 times for drivers, which is
767 mainly attributed to the wood burning and grilling meat activities, resulting in approximately
768 five times higher of international cancer risk safe limit (nearly ten time of threshold value in
769 wet season). Drivers' exposure to hopanes in PE $PM_{2.5}$ was 2.0-2.3 times higher than women
770 and students, correlating with the elevated traffic emissions on road environment well.

771 This work can be regarded as the first attempt in underdeveloped country of Africa at the
772 current condition, although there are some drawbacks, such as relatively short sampling
773 period and a limited number of participants. More investigations on personal exposure and
774 related potential health effects by cohort study method will be considered in the further. The
775 policy implication of our findings is that developing and implementing appropriate
776 preventive and control measures on different $PM_{2.5}$ anthropogenic sources in different regions
777 are appropriate, such as using dry wood for barbecues for the female workers and improving
778 waste treatment equipment at landfill as soon as possible to reduce waste inorganized stack
779 and open combustion.

780

781 **Acknowledgments**

782 This study was supported by the Natural Science Foundation of China (NSFC)
783 (41503096) and has received funding from the European Union 7th Framework Programme
784 (FP7/2007-2013) under Grant Agreement no. 603502 (EU project DACCIWA: Dynamics-



785 Aerosol-Chemistry-Cloud Interactions in West Africa). Supports from open fund by Jiangsu
786 Key Laboratory of Atmospheric Environment Monitoring and Pollution Control (KHK1712),
787 and a Project Funded by the Priority Academic Program Development of Jiangsu Higher
788 Education Institutions (PAPD), and open fund by State Key Laboratory of Loess and
789 Quaternary Geology, Institute of Earth Environment, CAS (SKLLQG1722) are also thanked.
790 In addition, the authors thank all the personal exposure sampling participants who helped a
791 lot in this study.

792

793 Author Contributions

794 H.X. and C.L. conceived and designed the study. H.X., J.-F.L., C.L. and B.G.
795 contributed to the literature search, data analysis/interpretation and manuscript writing. J.-F.L.,
796 C.L., B.G., V.Y., A.A., K.H., S.H., Z.S. and J.C. contributed to manuscript revision. H.X., J.-
797 F.L., E.G., J.A and L.L. carried out the particulate samples collection and chemical
798 experiments, analyzed the experimental data.

799

800 Additional Information

801 Fig. S1 accompany this manuscript can be found in Supplementary Information.

802

803 Competing financial interests

804 The authors declare no competing financial interests.

805

806 References

- 807 Adjiri, O. A., Mafou, C. K., and Konan, P. K.: Impact of Akouedo landfill (Abidjan-Côte d'Ivoire) on the
808 populations: socio-economic and environmental study, available at: [http://www.ijias.issr-](http://www.ijias.issr-journals.org/abstract.php?article=IJIAS-15-235-01)
809 [journals.org/abstract.php?article=IJIAS-15-235-01](http://www.ijias.issr-journals.org/abstract.php?article=IJIAS-15-235-01) (last access: 24 May 2017), 2015.
- 810 Assamoi, E. and Liousse, C.: Focus on the impact of two wheel vehicles on African combustion aerosols
811 emissions, *Atmos. Environ.*, 44, 3985-3996, 2010.
- 812 Baek, S. O., Fleld, R. A., Goldstone, M.E., Kirk, P.W., Lester, J.N., and Perry, R.: A review of atmospheric
813 polycyclic aromatic hydrocarbons: sources, fate and behavior, *Water Air Soil Poll.*, 60, 279-300, 1991.
- 814 Bahino, J., Yoboué, V., Galy-Lacaux, C., Adon, M., Akpo, A., Keita, S., Liousse, C., Gardrat, E., Chiron,
815 C., Ossouhou, M., Gnamien, S., and Djossou, J.: A pilot study of gaseous pollutants' measurement
816 (NO_2 , SO_2 , NH_3 , HNO_3 and O_3) in Abidjan, Côte d'Ivoire: contribution to an overview of gaseous
817 pollution in African cities, *Atmos. Chem. Phys.*, 18, 5173-5198, 2018.
- 818 Benchrif, A., Guinot, B., Bounakhla, M., Cachier, H., Damnati, B., and Baghdad, B.: Aerosols in Northern
819 Morocco: Input pathways and their chemical fingerprint, *Atmos. Environ.*, 174, 140-147, 2018.
- 820 Brouwer, P.: Theory of XRF: Getting acquainted with the principles. PANalytical BV, the Netherlands,
821 2003.
- 822 Bruce, N., Perez-Padilla, R., and Albalak, R.: Indoor air pollution in developing countries: a major
823 environmental and public health challenge. *Special Theme-Environment and Health*, B. World Health
824 *Organ.*, 78, 1078-1092, 2000.
- 825 Cachier, H., Aulagnier, F., Sarda, R., Gautier, F., Masclat, P., Besombes, J.-L., Marchand, N., Despiou, S.,
826 Croci, D., Mallet, M., Laj, P., Marinoni, A., Deveau, P.-A., Roger, J.-C., Putaud, J.-P., Dingenen, R. V.,



- 827 Dell'Acqua, A., Viidanoja, J., Santos, S. M.-D., Liousse, C., Cousin, F., Rosset, R., Gardrat, E., and
828 Galy-Lacaux, C.: Aerosol studies during the ESCOMPTE experiment: an overview, *Atmos. Res.*, 74,
829 547-563, 2005.
- 830 Cao, G., Zhang, X., Gong, S., and Zheng, F.: Investigation on emission factors of particulate matter and
831 gaseous pollutants from crop residue burning, *J. Environ. Sci.*, 20, 50-55, 2008.
- 832 Cao, J. J., Xu, H. M., Xu, Q., Chen, B. H., Kan, H. D.: Fine particulate matter constituents and
833 cardiopulmonary mortality in a heavily polluted Chinese city, *Environ. Health Persp.*, 120, 373-378,
834 2012.
- 835 Cao, J. J., Rong, B., Lee, S. C., Chow, J. C., Ho, K. F., Liu, S. X., and Zhu, C. S.: Composition of indoor
836 aerosols at emperor Qin's terra-cotta museum, Xi'an, China, during summer, 2004. *China Part.*, 3,
837 170-175, 2005.
- 838 Capes, G., Johnson, B., McFiggans, G., Williams, P. I., Haywood, J., and Coe, H.: Aging of biomass
839 burning aerosols over West Africa: Aircraft measurements of chemical composition, microphysical
840 properties, and emission ratios, *J. Geophys. Res.-Atmos.*, 113, D23,
841 <https://doi.org/10.1029/2008JD009845>, 2008.
- 842 Chen, R. J., Zhao, Z. H., and Kan, H. D.: Heavy smog and hospital visits in Beijing, China, *Am. J. Resp.*
843 *Crit. Care.*, 188, 1170-1171, 2013.
- 844 Chen, X. C., Jahn, H. J., Engling, G., Ward, T. J., Kraemer, A., Ho, K. F., Yim, S. H. L., and Chan, C. Y.:
845 Chemical characterization and sources of personal exposure to fine particulate matter (PM_{2.5}) in the
846 megacity of Guangzhou, China, *Environ. Pollut.*, 231, 871-881, 2017.
- 847 Crist, K. C., Liu, B., Kim, M., Deshpande, S. R., and John, K.: Characterization of fine particulate matter
848 in Ohio: Indoor, outdoor, and personal exposures, *Environ. Res.*, 106, 62-71, 2008.
- 849 Delmas, R. A., Druilhet, A., Cros, B., Durand, P., Delon, C., Lacaux, J. P., Brustet, J. M., Serca, D., Affre,
850 C., and Guenther, A.: Experiment for regional sources and sinks of oxidants (EXPRESSO): an
851 overview, *J. Geophys. Res.-Atmos.*, 104, 30609-30624, 1999.
- 852 Djossou, J., Léon, J.-F., Akpo, A. B., Liousse, C., Yoboué, V., Bedou, M., Bodjrenou, M., Chiron, C., Galy-
853 Lacaux, C., Gardrat, E., Abbey, M., Keita, S., Bahino, J., Touré N'Datchoh, E., Ossohou, M., and
854 Awanou, C. N.: Mass concentration, optical depth and carbon composition of particulate matter in the
855 major southern West African cities of Cotonou (Benin) and Abidjan (Côte d'Ivoire), *Atmos. Chem.*
856 *Phys.*, 18, 6275-6291, 2018.
- 857 Fanou, L. A., Mobio, T. A., Creppy, E. E., Fayomi, B., Fustoni, S., Møller, P., Kyrtopoulos, S., Georgiades,
858 P., Loft, S., Sanni, A., Skov, H., Øvrebø, S., and Autrup, H.: Survey of air pollution in Cotonou,
859 Benin-air monitoring and biomarkers, *Sci. Total Environ.*, 358, 85-96, 2006.
- 860 Feng, S. L., Wang, X. M., Wei, G. J., and Peng, P. A.: The leaching behavior and genotoxic assessment of
861 heavy metals in the municipal solid waste incineration bottom ashes from Macao, *Asian J. Ecotoxicol.*,
862 4, 330-335, 2006.
- 863 Gaudichet, A., Echalar, F., Chatenet, B., Quiseft, J. P., Malingre, G., Cachier, H., Buat-Menard, P., Artaxo,
864 P., and Maenhaut, W.: Trace elements in tropical African Savanna biomass burning ae'rosols, *J.*
865 *Atmos. Chem.*, 22, 19-39, 1995.
- 866 Gokhale, S., Kohajda, T., and Schlink, U.: Source apportionment of human personal exposure to volatile
867 organic compounds in homes, offices and outdoors by chemical mass balance and genetic algorithm
868 receptor models, *Sci. Total Environ.*, 407, 122-138, 2008.
- 869 Gu, Z. P., Feng, J. L., Han, W. L., Wu, M. H., Fu, J. M., and Sheng, G. Y.: Characteristics of organic matter
870 in PM_{2.5} from an e-waste dismantling area in Taizhou, China, *Chemosphere*, 80, 800-806, 2010.
- 871 Ho, S. S. H. and Yu, J. Z.: In-injection port thermal desorption and subsequent gas chromatography-mass
872 spectrometric analysis of polycyclic aromatic hydrocarbons and n-alkanes in atmospheric aerosol
873 samples, *J. Chromatogr. A*, 1059, 121-129, 2004.
- 874 Ho, S. S. H., Chow, J. C., Watson, J. G., Ng, L. P. T., Kwok, Y., and Ho, K. F.: Precautions for in-injection
875 port thermal desorption-gas chromatography/mass spectrometry (TD-GC/MS) as applied to aerosol
876 filter samples, *Atmos. Environ.*, 45, 1491-1496, 2011.
- 877 Ho, S. S. H., Yu, J. Z., Chow, J. C., Zielinska, B., Watson, J. G., Sit, E. H., and Schauer, J. J.: Evaluation of
878 an in-injection port thermal desorption-gas chromatography/mass spectrometry method for analysis of
879 non-polar organic compounds in ambient aerosol samples, *J. Chromatogr. A*, 1200, 217-227, 2008.
- 880 Hu, X., Zhang, Y., Ding, Z. H., Wang, T. J., Lian, H. Z., Sun, Y. Y., and Wu, J. C.: Bioaccessibility and
881 health risk of arsenic and heavy metals (Cd, Co, Cr, Cu, Ni, Pb, Zn and Mn) in TSP and PM_{2.5} in
882 Nanjing, China, *Atmos. Environ.*, 57, 146-152, 2012.
- 883 Huang, R.-J., Zhang, Y., Bozzetti, C., Ho, K., Cao, J., Han, Y., Daellenbach, K., Slowik, J., Platt, S.,



- 884 Canonaco, F., Zotter, P., Wolf, R., Pieber, S., Bruns, E., Crippa, M., Ciarelli, G., Piazzalunga, A.,
885 Schwikowski, M., Abbazade, G., Schnelle-Kreis, J., Zimmermann, R., An, Z., Szidat, S.,
886 Baltensperger, U., El Haddad, I., and Prévôt, A.: High secondary aerosol contribution to particulate
887 pollution during haze events in China, *Nature*, 514, 218-222, 2014.
- 888 IARC: Di (2-ethylhexyl) phthalate. International Agency for Research on Cancer (IARC) Monographs on
889 the Evaluation of Carcinogenic Risks to Humans 29, 269-294, 1982.
- 890 IMF: World Economic Outlook Update, July 2017: A Firming Recovery, available at:
891 [https://www.imf.org/en/Publications/WEO/Issues/2017/07/07/world-economic-outlook-update-july-](https://www.imf.org/en/Publications/WEO/Issues/2017/07/07/world-economic-outlook-update-july-2017)
892 2017.
- 893 Jedrychowski, W. A., Perera, F. P., Majewska, R., Mrozek-Budzyn, D., Mroz, E., Roen, E. L., Sowa, A.,
894 and Jacek, R.: Depressed height gain of children associated with intrauterine exposure to polycyclic
895 aromatic hydrocarbons (PAH) and heavy metals: The cohort prospective study, *Environ. Res.*, 136,
896 141-147, 2015.
- 897 Kang, C. M., Lee, H. S., Kang, B. W., Lee, S. K., and Sunwoo, Y.: Chemical characteristics of acidic gas
898 pollutants and PM_{2.5} species during hazy episodes in Seoul, South Korea, *Atmos. Environ.*, 38, 4749-
899 4760, 2004.
- 900 Keita, S., Liousse, C., Yoboué, V., Dominutti, P., Guinot, B., Assamoi, E.-M., Borbon, A., Haslett, S. L.,
901 Bouvier, L., Colomb, A., Coe, H., Akpo, A., Adon, J., Bahino, J., Doumbia, M., Djossou, J., Galy-
902 Lacaux, C., Gardrat, E., Gnamien, S., Léon, J. F., Ossouhou, M., N’Datchoh, E. T. and
903 Roblou, L.: Particle and VOC emission factor measurements for anthropogenic sources in West Africa,
904 *Atmos. Chem. Phys.*, 18(10), 7691-7708, doi:10.5194/acp-18-7691-2018, 2018.
- 905 Kong, S., Ji, Y., Liu, L., Chen, L., Zhao, X., Wang, J., Bai, Z., and Sun, Z.: Spatial and temporal variation
906 of phthalic acid esters (PAEs) in atmospheric PM₁₀ and PM_{2.5} and the influence of ambient
907 temperature in Tianjin, China, *Atmos. Environ.*, 74, 199-208, 2013.
- 908 Lacaux, J. P., Brustet, J. M., Delmas, R., Menaut, J. C., Abbadie, L., Bonsang, B., Cachier, H., Baudet, J.,
909 Andreae, M. O., and Helas, G.: Biomass burning in the tropical savannas of IvoryCoast: an overview
910 of the field experiment Fire of Savannas (FOS/DECAFE 91), *J. Atmos. Chem.*, 22, 195-216, 1995.
- 911 Laurent, B., Marticorena, B., Bergametti, G., Le'on, J. F., and Mahowald, N. M.: Modeling mineral dust
912 emissions from the Sahara Desert using new surface properties and soil database, *J. Geophys. Res.-*
913 *Atmos.*, <https://doi.org/10.1029/2007JD009484>, 2008.
- 914 Lawin, H., Agodokpessi, G., Ayelo, P., Kagima, J., Sonoukon, R., Ngahane, B. H. M., Awopeju, O.,
915 Vollmer, W. M., Nemery, B., Burney, P., and Fayomi, B.: A cross-sectional study with an improved
916 methodology to assess occupational air pollution exposure and respiratory health in motorcycle taxi
917 driving, *Sci. Total Environ.*, 550, 1-5, 2016.
- 918 Léon, J.-F., Derimian, Y., Chiapello, I., Tanré, D., Podvin, T., Chatenet, B., Diallo, A. and Deroo, C.:
919 Aerosol vertical distribution and optical properties over M'Bour (16.96° W; 14.39° N), Senegal from
920 2006 to 2008, *Atmos. Chem. Phys.*, 9(23), 9249-9261, doi:10.5194/acp-9-9249-2009, 2009.
- 921 Li, C. L., Fu, J. M., Sheng G. Y., Bi, X. H., Hao, Y. M., Wang, X. M., and Mai, B. X.: Vertical distribution
922 of PAHs in the indoor and outdoor PM_{2.5} in Guangzhou, China, *Build. Environ.*, 40, 329-341, 2005.
- 923 Li, H. L., Song, W. W., Zhang, Z. F., Ma, W. L., Gao, C. J., Li, J., Huo, C. Y., Mohammed, M. O. A., Liu, L.
924 Y., Kannan, K., and Li, Y. F.: Phthalates in dormitory and house dust of northern Chinese cities:
925 Occurrence, human exposure, and risk assessment, *Sci. Total Environ.*, 565, 496-502, 2016.
- 926 Li, X., Wang, S., Duan, L., Hao, J., and Nie, Y.: Carbonaceous aerosol emissions from household biofuel
927 combustion in China, *Environ. Sci. Technol.*, 43, 6076-6081, 2009.
- 928 Lima, A. L. C., Farrington, J. W., and Reddy, C. M.: Combustion-derived polycyclic aromatic
929 hydrocarbons in the environment - A review, *Environ. Forensics*, 6, 109-131, 2005.
- 930 Liu, X. T., Zhai, Y. B., Zhu, Y., Liu, Y. N., Chen, H. M., Li, P., Peng, C., Xu, B. B., Li, C. T., and Zeng, G.
931 M.: Mass concentration and health risk assessment of heavy metals in size-segregated airborne
932 particulate matter in Changsha, *Sci. Total Environ.*, 517, 215-221, 2015.
- 933 Lindesay, J. A., Andreae, M. O., Goldammer, J. G., Harris, G. W., Annegarn, H. J., Garstang, M., Scholes,
934 R. J., and van Wilgen, B. W.: The IGBP/IGAC SAFARI-92 field experiment: background and
935 overview, *J. Geophys. Res.*, 101, 521-530, 1996.
- 936 Liousse, C., Assamoi, E., Criqui, P., Granier, C., and Rosset, R.: African combustion emission explosive
937 growth from 2005 to 2030, *Environ. Res. Lett.*, 9, 156-166, 2014.
- 938 Liousse, C., Guillaume, B., Grégoire, J. M., Mallet, M., Galy, C., Pont, V., Akpo, A., Bedou, M., Castéra, P.,
939 Dunggall, L., Gardrat, E., Granier, C., Konaré, A., Malavelle, F., Mariscal, A., Mieville, A., Rosset, R.,
940 Serça, D., Solmon, F., Tummon, F., Assamoi, E., Yoboué, V., and Van Velthoven, P.: Updated African



- 941 biomass burning emission inventories in the framework of the AMMA-IDAF program, with an
942 evaluation of combustion aerosols, *Atmos. Chem. Phys.*, 10, 9631-9646, 2010.
- 943 Marticorena, B., Chatenet, B., Rajot, J. L., Traore, S., Coulibaly, M., Diallo, A., Kone, I., Maman, A.,
944 Ndiaye, T., and Zakou, A.: Temporal variability of mineral dust concentrations over West Africa:
945 analyses of a pluriannual monitoring from the AMMA Sahelian Dust Transect, *Atmos. Chem. Phys.*,
946 10, 8899-8915, 2010.
- 947 Meng, Q. Y., Spector, D., Colome, S., and Turpin, B.: Determinants of indoor and personal exposure to
948 PM_{2.5} of indoor and outdoor origin during the RIOPA study, *Atmos. Environ.*, 43, 5750-5758, 2009.
- 949 Meng, X. Z., Wang, Y., Xiang, N., Chen, L., Liu, Z. G., Wu, B., Dai, X. H., Zhang, Y. H., Xie, Z. Y., and
950 Ebinghaus, R.: Flow of sewage sludge-borne phthalate esters (PAEs) from human release to human
951 intake: implication for risk assessment of sludge applied to soil, *Sci. Total Environ.*, 476, 242-249,
952 2014.
- 953 Nisbet, I. C. T. and Lagoy, P. K.: Toxic equivalency factors (TEFs) for polycyclic aromatic hydrocarbons
954 (PAHs), *Regul. Toxicol. Pharm.*, 16, 290-300, 1992.
- 955 Norman, R., Barnes, B., Mathee, A., and Bradshaw, D.: Estimating the burden of disease attributable to
956 indoor air pollution from household use of solid fuels in South Africa in 2000, *S. Afr. Med. J.*, 97,
957 764-771, 2007.
- 958 Ouafou-Loumbé, M.-R., Galy-Lacaux, C., Lioussé, C., Pont, V., Akpo, A., Doumbia, T., Gardrat, E., Zouiten,
959 C., SighaNkamdjou, L., and Ekodeck, G. E.: Chemical composition and sources of atmospheric
960 aerosols at Djougou (Benin), *Meteorol. Atmos. Phys.*, 1-19, [https://doi.org/10.1007/s00703-017-0538-](https://doi.org/10.1007/s00703-017-0538-5)
961 5, 2017.
- 962 Owili, P. O., Lien, W.-H., Muga, M. A., and Lin, T. H.: The associations between types of ambient PM_{2.5}
963 and under-five and maternal mortality in Africa, *Inter. J. Env. Res. Pub. Heal.*, 14, 359, 1-20, 2017.
- 964 Reeves, C. E., Formenti, P., Afif, C., Ancellet, G., Attié, J.-L., Bechara, J., Borbon, A., Cairo, F., Coe, H.,
965 Crumeyrolle, S., Fierli, F., Flamant, C., Gomes, L., Hamburger, T., Jambert, C., Law, K. S., Mari, C.,
966 Jones, R. L., Matsuki, A., Mead, M. I., Methven, J., Mills, G. P., Minikin, A., Murphy, J. G., Nielsen,
967 J. K., Oram, D. E., Parker, D. J., Richter, A., Schlager, H., Schwarzenboeck, A., and Thouret, V.:
968 Chemical and aerosol characterisation of the troposphere over West Africa during the monsoon period
969 as part of AMMA, *Atmos. Chem. Phys.*, 10, 7575-7601, 2010.
- 970 Ruellan, S., Cachier, H., Gaudichet, A., Masclat, P., and Lacaux, J. P.: Airborne aerosols over central Africa
971 during the experiment for regional sources and sinks of oxidants (EXPRESSO), *J. Geophys. Res.-*
972 *Atmos.*, 104, 673-690, 1999.
- 973 Sealy, A., Jenkins, G. S., and Walford, S. C.: Seasonal/regional comparisons of rain rates and rain
974 characteristics in West Africa using TRMM observations, *J. Geophys. Res.-Atmos.*, 108, 4306,
975 <https://doi.org/10.1029/2002JD002667>, 2003.
- 976 Seinfeld, J. H. and Pandis, S. N.: Atmospheric chemistry and physics: from air pollution to climate change,
977 John Wiley & Sons (last access: 9 August 2017), 2006.
- 978 Shen, G., Xue, M., Wei, S., Chen, Y., Wang, B., Wang, R., Lv, Y., Shen, H., Li, W., Zhang, Y., Huang, Y.,
979 Chen, H., Wei, W., Zhao, Q., Li, B., Wu, H. and TAO, S.: The influence of fuel moisture, charge size,
980 burning rate and air ventilation conditions on emissions of PM, OC, EC, Parent PAHs, and their
981 derivatives from residential wood combustion, *J. Environ. Sci. (China)*, 25(9), 1808-1816, 2013.
- 982 Simoneit, B. R.: A Review of biomarker compounds as source indicators and tracers for air pollution,
983 *Environ. Sci. Pollut. R.*, 6, 159-169, 1999.
- 984 Škrbic, B. D., Ji, Y. Q., Durisic-Mladenovic, N., and Zhao, J.: Occurrence of the phthalate esters in soil and
985 street dust samples from the Novi Sad city area, Serbia, and the influence on the children's and adults'
986 exposure, *J. Hazard. Mater.*, 312, 272-279, 2016.
- 987 Swap, R. J., Annegarn, H. J., and Otter, L.: Southern African regional science initiative (SAFARI 2000)
988 summary of science plan, *S. Afr. J. Sci.*, 98, 119-124, 2002.
- 989 Tang, D. L., Li, T.-Y., Liu, J. J., Zhou, Z. J., Yuan, T., Chen, Y.-H., Rauh, V. A., Xie, J., and Perera, F.:
990 Effects of prenatal exposure to coal-burning pollutants on children's development in China, *Environ.*
991 *Health Persp.*, 116, 674-679, 2008.
- 992 Taylor, S. R. and McLennan, S. M.: The continental crust: Its composition and evolution, *Phys. Earth*
993 *Planet. In.*, 42, 196-197, 1985.
- 994 Tian, J., Ni, H. Y., Cao, J. J., Han, Y. M., Wang, Q. Y., Wang, X. L., Chen, L. W. A., Chow, J. C., Watson, J.
995 G., Wei, C., Sun, J., Zhang, T., and Huang, R. J.: Characteristics of carbonaceous particles from
996 residential coal combustion and agricultural biomass burning in China. *Atmos. Pollut. Res.* 8, 521-527,
997 2017.



- 998 Tobiszewski, M. and Namiesnik, J.: Review: PAH diagnostic ratios for the identification of pollution
999 emission sources, *Environ. Pollut.*, 162, 110-119, 2012.
- 1000 Turpin, B. J. and Lim, H. J.: Species contributions to PM_{2.5} mass concentrations: Revisiting common
1001 assumptions for estimating organic mass, *Aerosol Sci. Tech.*, 35, 602-610, 2001.
- 1002 UNEP: available at: <https://environmentlive.unep.org/publication/country/C%3%B4te%20d'Ivoire/unep>,
1003 2015.
- 1004 USEPA, Integrated risk information system. Di(2-ethylhexyl)phthalate (DEHP) (CASRN 117-81-7).
1005 <http://www.epa.gov/iris/subst/0014.htm>, 1997.
- 1006 USEPA, Risk Assessment Guidance for Superfund Volume I: Human Health Evaluation Manual (Part E,
1007 Supplemental Guidance for Dermal Risk Assessment). Office of Superfund Remediation and
1008 Technology Innovation U.S. Environmental Protection Agency Washington, DC, 2004.
- 1009 USEPA, Special report on lead pollution, 2006.
- 1010 USEPA, Exposure factors handbook. National Center for Environmental Assessment, Washington, DC;
1011 EPA/600/R-09/052F. Available from the National Technical Information Service, Springfield, VA,
1012 2011.
- 1013 Val, S., Lioussé, C., Doumbia, E. H. T., Galy-Lacaux, C., Cachier, H., Marchand, N., Badel, A., Gardrat, E.,
1014 Sylvestre, A., and Baeza-Squiban, A.: Physico-chemical characterization of African urban aerosols
1015 (Bamako in Mali and Dakar in Senegal) and their toxic effects in human bronchial epithelial cells:
1016 description of a worrying situation, *Part. Fibre Toxicol.*, 10, 10, [https://doi.org/10.1186/1743-8977-10-](https://doi.org/10.1186/1743-8977-10-10)
1017 10, 2013.
- 1018 Wang, G., Kawamura, K., Xie, M., Hu, S., Cao, J., An, Z., Weston, J. G., and Chow, J. C.: Organic
1019 Molecular Compositions and Size Distributions of Chinese Summer and Autumn Aerosols from
1020 Nanjing: Characteristic Haze Event Caused by Wheat Straw Burning, *Environ. Sci. Technol.*, 43,
1021 6493-6499, 2009.
- 1022 Wang, G. H., Kawamura, K., Lee, S. C., Ho, K. F., and Cao, J. J.: Molecular, seasonal, and spatial
1023 distributions of organic aerosols from fourteen Chinese cities, *Environ. Sci. Technol.*, 40, 4619-4625,
1024 2006.
- 1025 Wang, J. Z., Guinot, B., Dong, Z. B., Li, X. P., Xu, H. M., Xiao, S., Ho, S. S. H., Liu, S. X., and Cao, J. J.:
1026 PM_{2.5}-bound polycyclic aromatic hydrocarbons (PAHs), oxygenated-PAHs and phthalate esters (PAEs)
1027 inside and outside middle school classrooms in Xi'an, China: Concentration, characteristics and
1028 health risk assessment, *Aerosol Air Qual. Res.*, 17, 1811-1824, 2017a.
- 1029 Wang, Y., Cheng, K., Wu, W. D., Tian, H. Z., Yi, P., Zhi, G. R., Fan, J., and Liu, S. H.: Atmospheric
1030 emissions of typical toxic heavy metals from open burning of municipal solid waste in China, *Atmos.*
1031 *Environ.*, 152, 6-15, 2017b.
- 1032 Wang, Y., Zhang, Q. Q., He, K., Zhang, Q., and Chai, L.: Sulfate-nitrate-ammonium aerosols over China:
1033 Response to 2000-2015 emission changes of sulfur dioxide, nitrogen oxides, and ammonia, *Atmos.*
1034 *Chem. Phys.*, 13, 2635-2652, 2013.
- 1035 WHO. Environmental Health Criteria 202: Selected non-heterocyclic polycyclic aromatic hydrocarbons.
1036 World Health Organization, Geneva, 1998.
- 1037 Xu, H. M., Guinot, B., Cao, J. J., Li, Y. Q., Niu, X. Y., Ho, K. F., Shen, Z. X., Liu, S. X., Zhang, T., Lei, Y.
1038 L., Zhang, Q., Sun, J., and Gao, J. J.: Source, health risk and composition impact of outdoor very fine
1039 particles (VFPs) to school indoor environment in Xi'an, Northwestern China, *Sci. Total Environ.*, 612,
1040 238-246, 2018a.
- 1041 Xu, H. M., Guinot, B., Ho, S. S. H., Li, Y. Q., Cao, J. J., Shen, Z. Z., Niu, X. Y., Zhao, Z. H., Liu, S. X.,
1042 Lei, Y. L., Zhang, Q., and Sun, J.: Evaluation on exposures to particulate matter in a junior secondary
1043 school: First comprehensive study on health risks and effective inflammatory responses in
1044 Northwestern China, *Environ. Geochem. Hlth.*, 40, 849-863, 2018b.
- 1045 Xu, H. M., Sonke, J. E., Guinot, B., Fu, X. W., Sun, R. Y., Lanzanova, A., Candaudap, F., Shen, Z. X., and
1046 Cao, J. J.: Seasonal and annual variations in atmospheric Hg and Pb isotopes in Xi'an, China, *Environ.*
1047 *Sci. Technol.*, 51, 3759-3766, 2017.
- 1048 Xu, H. M., Ho, S. S. H., Gao, M. L., Cao, J. J., Guinot, B., Ho, K. F., Long, X., Wang, J. Z., Shen, Z. X.,
1049 Liu, S. X., Zheng, C. L., and Zhang Q.: Microscale spatial distribution and health assessment of
1050 PM_{2.5}-bound polycyclic aromatic hydrocarbons (PAHs) at nine communities in Xi'an, China, *Environ.*
1051 *Pollut.*, 218, 1065-1073, 2016a.
- 1052 Xu, H. M., Cao, J. J., Chow, J. C., Huang, R.-J., Shen, Z. X., Chen, L. W. A., Ho, K. F., and Watson, J. G.:
1053 Inter-annual variability of wintertime PM_{2.5} chemical composition in Xi'an, China: evidences of
1054 changing source emissions, *Sci. Total Environ.*, 545-555, 2016b.



- 1055 Xu, H. M., Guinot, B., Niu, X. Y., Cao, J. J., Ho, K. F., Zhao, Z. H., Ho, S. S. H., and Liu, S. X.:
1056 Concentrations, particle-size distributions, and indoor/outdoor differences of polycyclic aromatic
1057 hydrocarbons (PAHs) in a middle school classroom in Xi'an, China, *Environ. Geochem. Hlth.*, 37,
1058 861-873, 2015.
- 1059 Xu, H. M., Tao, J., Ho, S. S. H., Ho, K. F., Cao, J. J., Li, N., Chow, J. C., Wang, G. H., Han, Y. M., Zhang,
1060 R. J., Watson, J. G., and Zhang, J. Q.: Characteristics of fine particulate non-polar organic compounds
1061 in Guangzhou during the 16th Asian Games: effectiveness of air pollution controls, *Atmos. Environ.*,
1062 76, 94-101, 2013.
- 1063 Xu, H. M., Cao, J. J., Ho, K. F., Ding, H., Han, Y. M., Wang, G. H., Chow, J. C., Watson, J. G., Khol, S. D.,
1064 Qiang, J., and Li, W. T.: Lead concentrations in fine particulate matter after the phasing out of leaded
1065 gasoline in Xi'an, China, *Atmos. Environ.*, 46, 217-224, 2012.
- 1066 Yang, F. X., Jin, S. W., Xu, Y., and Lu, Y. N.: Comparisons of IL-8, ROS and p53 responses in human lung
1067 epithelial cells exposed to two extracts of PM_{2.5} collected from an e-waste recycling area, China,
1068 *Environ. Res. Lett.*, 6, 1-6, 2011.
- 1069 Yassaa, N., Meklati, B. Y., Cecinato, A., and Marino, F.: Particulate n-alkanes, nalkanoic acids and
1070 polycyclic aromatic hydrocarbons in the atmosphere of Algiers City area, *Atmos. Environ.*, 35, 1843-
1071 1851, 2001.
- 1072 Yunker, M. B., Macdonald, R. W., Vingarzan, R., Mitchell, R. H., Goyette, D., and Sylvestre, S.: PAHs in
1073 the Fraser River basin: a critical appraisal of PAH ratios as indicators of PAH source and composition,
1074 *Org. Geochem.*, 33, 489-515, 2002.
- 1075 Zhang, L. B., Wang, F. M., Ji, Y. Q., Jiao, J., Zou, D. K., Liu, L. L., Shan, C. Y., Bai, Z. P., and Sun, Z. R.:
1076 Phthalate esters (PAEs) in indoor PM₁₀/PM_{2.5} and human exposure to PAEs via inhalation of indoor
1077 air in Tianjin, China, *Atmos. Environ.*, 85, 139-146, 2014a.
- 1078 Zhang, R., Jing, J., Tao, J., Hsu, S.-C., Wang, G., Cao, J., Lee, C. S. L., Zhu, L., Chen, Z., Zhao, Y., and
1079 Shen, Z.: Chemical characterization and source apportionment of PM_{2.5} in Beijing: seasonal
1080 perspective, *Atmos. Chem. Phys.*, 13, 7053-7074, 2013.
- 1081 Zhang, T., Cao, J. J., Chow, J. C., Shen, Z. X., Ho, K. F., Ho, S. S. H., Liu, S. X., Han, Y. M., Watson, J. G.,
1082 Wang, G. H., and Huang, R. J.: Characterization and seasonal variations of levoglucosan in fine
1083 particulate matter in Xi'an, China, *J. Air Waste Manag.*, 64, 1317-1327, 2014b.
- 1084 Zhao, W. X. and Hopke, P. K.: Source investigation for ambient PM_{2.5} in Indianapolis, IN, *Aerosol Sci.*
1085 *Tech.*, 40, 898-909, 2006.
- 1086 Zhou, P., Guo, J., Zhou, X. Y., Zhang, W., Liu, L. L., Liu, Y. C., and Lin, K. F.: PM_{2.5}, PM₁₀ and health risk
1087 assessment of heavy metals in a typical printed circuit boards manufacturing workshop, *J. Environ.*
1088 *Sci.*, 26, 2018-2026, 2014.
- 1089



1090 **Figure Caption:**

1091 **Figure 1.** Locations of the sampling sites within the cities. A: Domestic Fires (DF) site at the
1092 Yopougon-Lubafrique market in Abidjan; B: Waste Burning (WB) site at the landfill of
1093 Akeoudo in Abidjan, the location of the long-term sampling site is given by the blue marker;
1094 and C: Motorcycle Traffic (MT) site at Dantokpa area in Cotonou.

1095 **Figure 2.** Pictures showing the sampling sites and corresponding participants: (a) women at
1096 DF; (b) students at WB; (c) drivers at MT.

1097 **Figure 3.** Personal exposure to $PM_{2.5}$ mass concentrations of woman at DF, student at WB
1098 and driver at MT in dry season (January) and wet season (July) of 2016 in SWA area.

1099 **Figure 4** Variations of OC/EC ratios in personal exposure to $PM_{2.5}$ samples for women at DF,
1100 students at WB and drivers at MT (The box plots indicate the average concentration and the
1101 min, 1st, 25th, 50th, 75th, 99th and max percentiles).

1102 **Figure 5.** Personal exposure to $PM_{2.5}$ mass concentration closures for women at DF, students
1103 at WB and drivers at MT in different sampling seasons.

1104 **Figure 6.** Distributions of A: PAHs; B: PAEs; and C: hopanes in $PM_{2.5}$ personal exposure
1105 samples for women at DF, students at WB and drivers at MT in dry and wet seasons of 2016.

1106 **Figure 7.** Correlations between PAHs diagnostic ratios (average ratio points of each type
1107 participant indicate day and night value respectively).

1108



1109 **Figure 1.**

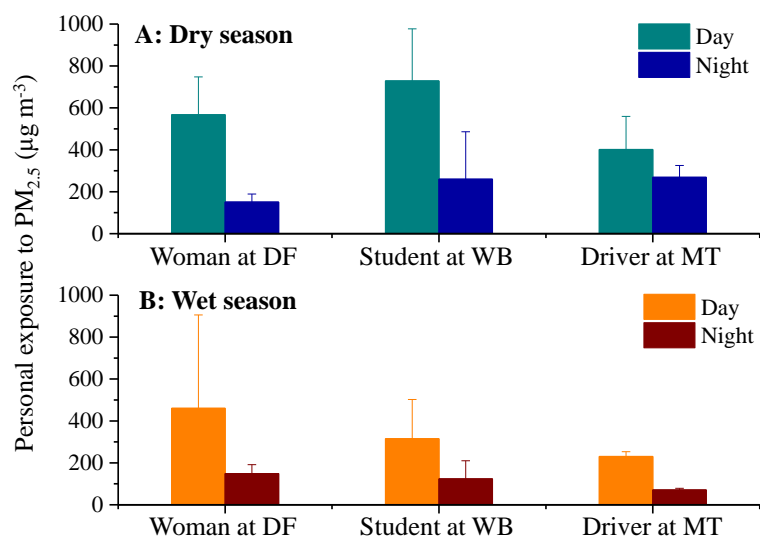
1110



1111

1112 **Figure 2.**

1113

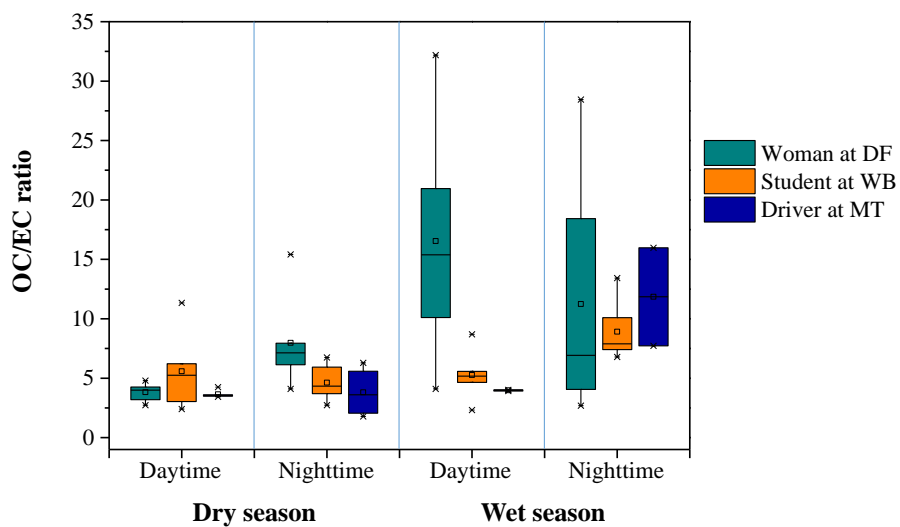


1114

1115

Figure 3.

1116

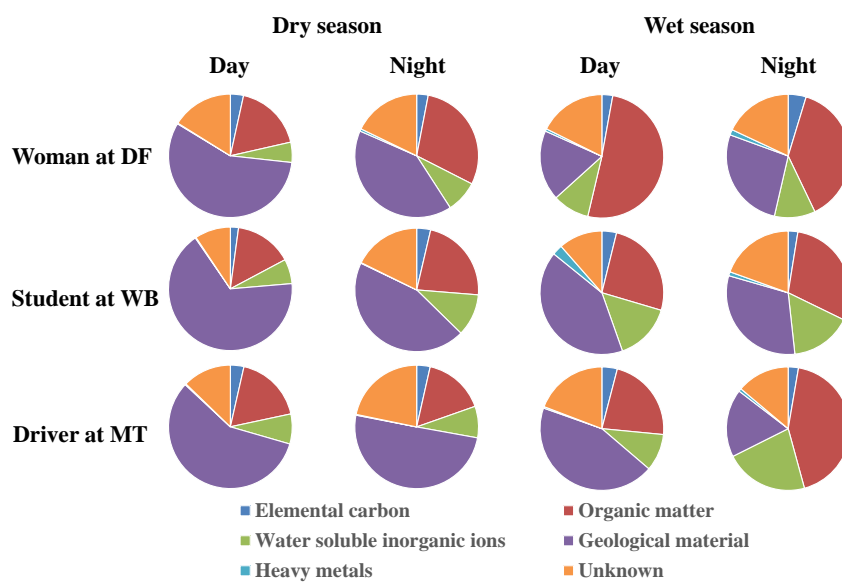


1117

1118

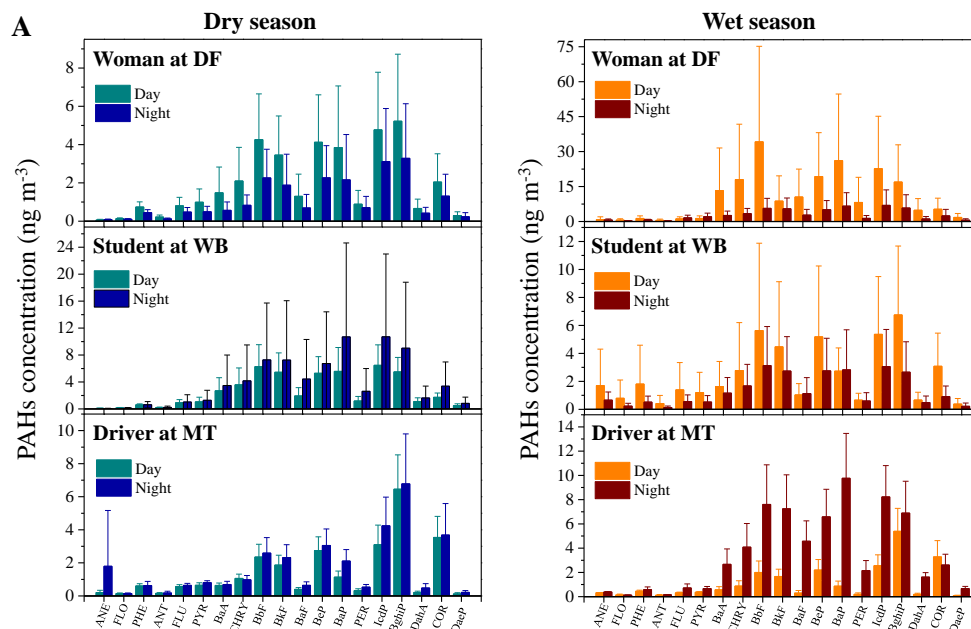
1119

Figure 4.

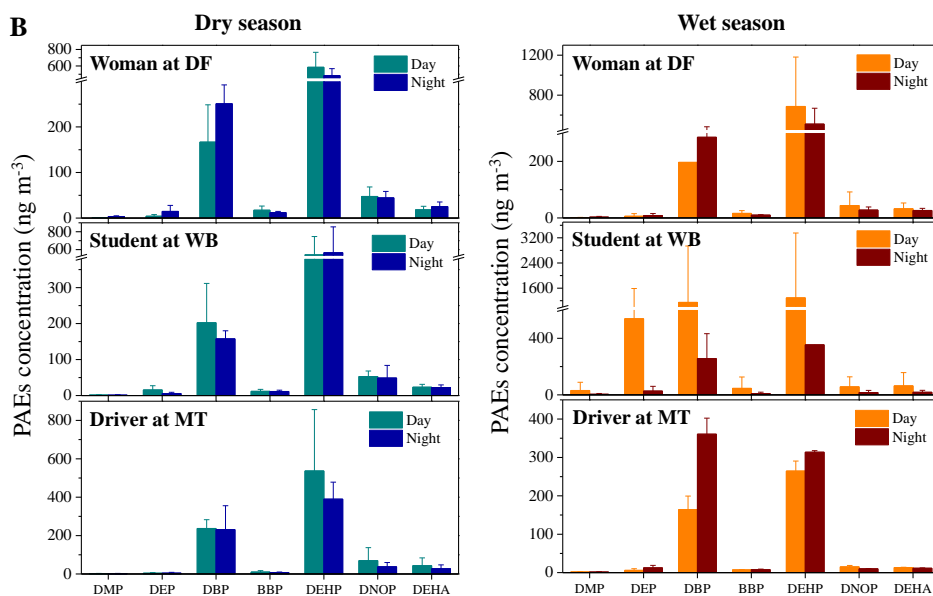


1120
1121
1122

Figure 5.

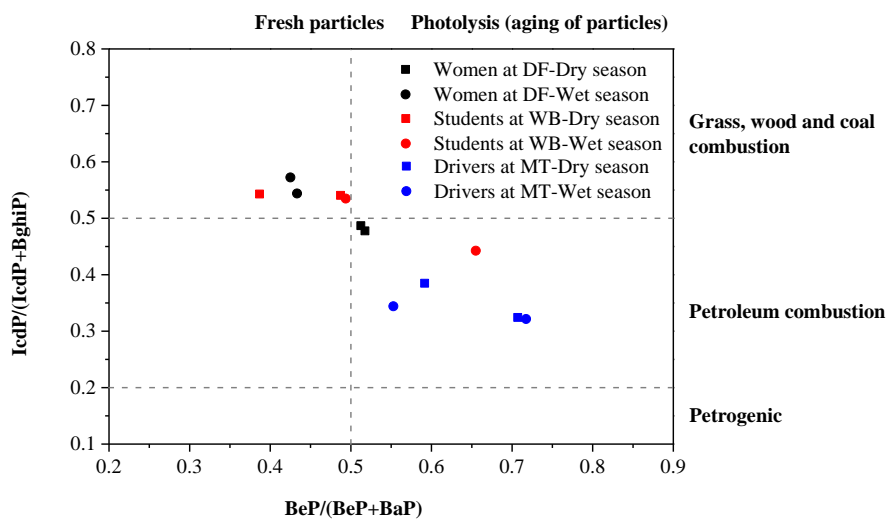


1123



1124

1125



1129

1130

Figure 7.

1131



1132 **Table 1** Meteorological parameters of the studied two cities during the dry (December 2015
1133 to March 2016) and wet (April to July 2016) seasons.

	Season	Abidjan	Cotonou
Mean daily air temperature (°C)	Dry	28.0	28.3
	Wet	27.5	27.7
Total rainfall (mm)	Dry	268	92
	Wet	626	558
Mean wind speed (m s ⁻¹)	Dry	3.0	3.0
	Wet	3.4	4.3

1134



1135 **Table 2** Definitions and recommended values of the parameters in equations (1-4) in this
1136 study.

Parameter	Definition (unit)	Value used in this study (reference)
D	average daily exposure dose ($\text{mg kg}^{-1} \text{day}^{-1}$)	/
C	heavy metals concentrations in equations (ng m^{-3})	/
R	inhalation rate, air volume a child inhaled each day ($\text{m}^3 \text{day}^{-1}$)	16.0 for women and drivers; 15.2 for students (USEPA, 2011)
EF	exposure frequency (day year^{-1})	130 for women and drivers (half working days); 182 for students (half year)
ED	exposure duration (year)	30 for women and drivers (working years); 15 for students (before going to high school)
BW	body weight (kg)	62.5 for women ^a ; 37.5 for students ^a ; 85.0 for drivers ^a
AT	averaging time (day)	30 or 15×365 (non-cancer); 70×365 (cancer)
<i>cf</i>	conversion factor (kg mg^{-1})	10^{-6}
HQ	hazard quotient	/
RfD	reference dose, estimated as the maximum permissible risk on human by daily exposure ($\text{mg kg}^{-1} \text{day}^{-1}$)	Table 3
HI	hazard index	/
ILCR	incremental lifetime cancer risk (ILCR)	/
CSF	cancer slope factor ($\text{mg kg}^{-1} \text{day}^{-1}$) ⁻¹	Table 3
[BaP] _{eq}	equivalent BaP toxicity concentration (ng m^{-3})	/
C _i	individual PAH species concentration (ng m^{-3}) (i means target PAH species)	/
TEF _i	toxicity equivalency factor of each target PAH compound (i means target PAH species)	(Nisbet and Lagoy, 1992)

1137 a: Measured in this study.

1138



1139 **Table 3** Reference dose (RfD) ($\text{mg kg}^{-1} \text{day}^{-1}$) and cancer slope factor (CSF) ($(\text{mg kg}^{-1} \text{day}^{-1})^{-1}$)
1140 via inhalation exposure way used in this study.

	RfD	CSF	Reference
Mn	1.8×10^{-3}	/	Liu et al., 2015
Ni	5.4×10^{-3}	/	Zhou et al., 2014; Liu et al., 2015
Zn	3.0×10^{-1}	/	Zhou et al., 2014
Pb	3.5×10^{-3}	/	Zhou et al., 2014; Hu et al., 2012
BaP	/	3.140	USEPA, 2011
DEHP	/	0.014	USEPA, 1997; Wang et al., 2017a



1141 **Table 4** Statistical analysis (arithmetic mean±standard deviation) of personal exposure to PM_{2.5} mass concentrations and the chemical
 1142 compositions (units: µg m⁻³) during the sampling period in SWA region.

	Wet season											
	Dry season						Wet season					
	Women at DF		Students at WB		Drivers at MT		Women at DF		Students at WB		Drivers at MT	
	Daytime	Nighttime	Daytime	Nighttime	Daytime	Nighttime	Daytime	Nighttime	Daytime	Nighttime	Daytime	Nighttime
PE	567.0±180.6	150.6±38.5	728.5±248.5	260±226.1	401.3±158.0	269.0±56.1	460.5±445.2	148.6±42.9	315.2±186.9	123.7±86.1	230.4±22.8	70.7±8.1
PM _{2.5}	72.4±24.6	31±5.0	85.0±57.4	40.9±34.4	49.5±12.5	31.8±14.2	189.3±197.8	40.1±9.3	65.2±65.2	28.5±26.8	37.0±3.5	22.2±10.6
OC	19.5±7.3	4.7±2.2	15.0±4.7	8.6±5.7	13.6±3.6	9.0±2.3	11.5±10.8	6.3±3.7	12.3±11.4	3.6±3.6	9.3±0.8	1.9±0.0
EC	91.9±31.1	35.7±6.8	100.0±60.1	49.5±39.5	63.1±16.0	40.8±13.6	200.8±207.1	46.3±7.2	77.4±76.2	32.1±30.3	46.3±4.2	24.1±10.6
Total carbon	4.4±1.3	1.6±0.6	6.5±3.6	6.4±9.4	2.4±0.8	2.2±0.6	8.6±8.4	1.9±1.0	4.6±5.4	1.9±0.7	3.1±0.2	2.3±0.2
Cl ⁻	2.7±0.7	2.2±1.4	5.5±1.3	3.0±0.7	3.7±1.3	2.7±0.5	2.2±0.8	1.6±0.7	5.0±6.0	1.8±1.3	1.6±0.2	1.2±0.1
NO ₃ ⁻	4.0±1.1	1.8±0.6	7.5±2.5	3.6±0.9	7.5±2.5	5.3±0.6	6.8±5.2	2.3±0.8	6.4±5.9	2.3±0.4	5.2±0.3	3.2±0.5
SO ₄ ²⁻	2.9±0.4	1.6±0.3	4.1±1.1	1.9±0.8	3.3±1.1	2.4±0.3	4.2±2.2	4.4±1.7	16.2±17.3	3.3±3.1	3.6±0.2	2.6±0.1
Na ⁺	0.6±0.2	0.4±0.5	1.4±0.4	3.0±4.1	1.1±0.2	0.9±0.2	0.6±0.5	0.1±0.0	0.6±0.2	0.4±0.3	0.7±0.0	0.1±0.0
NH ₄ ⁺	3.2±0.6	1.7±0.6	5.8±4.0	2.2±0.8	1.9±0.4	2.1±0.9	7.6±8.0	1.3±0.8	3.3±4.4	1.3±0.6	1.1±0.0	3.6±1.5
K ⁺	0.6±0.2	0.2±0.1	0.8±0.3	0.3±0.2	0.4±0.2	0.3±0.1	1.1±1.2	0.3±0.1	1.0±0.9	0.3±0.2	0.3±0.0	0.2±0.0
Mg ²⁺	11.0±3.2	3.1±0.9	14.9±4.5	4.9±3.2	10.6±5.5	6.0±1.2	6.6±4.3	3.2±0.8	17.3±13.9	4.5±3.8	6.8±0.3	2.3±0.1
Ca ²⁺	29.3±6.6	12.5±3.7	46.6±15.4	25.2±18.8	30.9±11.9	21.9±3.2	37.6±29.5	15.1±2.2	54.4±50.0	15.8±8.8	22.3±1.0	15.5±1.9
Total ions	14.61±5.25	2.64±0.36	21.17±4.64	4.85±3.30	10.99±6.50	5.90±0.37	3.37±3.34	1.87±0.96	5.07±1.74	1.76±1.24	4.56±0.64	0.57±0.05
Fe	0.04±0.02	0.00±0.00	0.07±0.02	0.02±0.01	0.03±0.02	0.01±0.01	0.01±0.01	0.00±0.00	0.03±0.03	0.01±0.01	0.01±0.00	0.01±0.00
V	0.04±0.02	0.01±0.00	0.06±0.02	0.01±0.01	0.03±0.03	0.01±0.01	0.05±0.02	0.06±0.03	0.31±0.35	0.04±0.05	0.03±0.00	0.03±0.00
Cr	0.18±0.06	0.04±0.03	0.29±0.08	0.07±0.04	0.35±0.12	0.21±0.11	0.14±0.16	0.04±0.00	0.37±0.36	0.06±0.06	0.17±0.02	0.04±0.00
Mn	0.05±0.02	0.01±0.01	0.09±0.02	0.01±0.01	0.05±0.03	0.02±0.02	0.02±0.02	0.02±0.02	0.04±0.05	0.02±0.02	0.02±0.01	0.01±0.00
Co	0.02±0.01	0.00±0.00	0.02±0.01	0.01±0.01	0.02±0.01	0.01±0.01	0.02±0.02	0.03±0.02	0.12±0.14	0.02±0.03	0.02±0.00	0.01±0.00
Ni	0.04±0.01	0.02±0.01	0.14±0.03	0.02±0.01	0.05±0.03	0.03±0.01	0.13±0.07	0.13±0.07	0.67±0.81	0.10±0.09	0.07±0.02	0.06±0.01
Cu	0.40±0.22	0.55±0.73	0.49±0.19	0.15±0.12	0.33±0.16	0.19±0.07	0.51±0.32	0.32±0.17	1.41±1.55	0.26±0.27	0.29±0.04	0.12±0.00
Zn												



Sb	0.02±0.01	0.05±0.02	0.02±0.02	0.00±0.00	0.02±0.04	0.01±0.01	0.12±0.08	0.21±0.18	1.16±1.38	0.22±0.29	0.07±0.04	0.08±0.09
Ba	0.19±0.09	0.16±0.12	0.25±0.11	0.07±0.09	0.22±0.18	0.05±0.07	0.47±0.39	1.02±0.60	6.80±8.30	0.84±1.41	0.18±0.18	0.14±0.01
Pb	0.07±0.03	0.07±0.07	0.17±0.07	0.04±0.03	0.07±0.05	0.02±0.03	0.14±0.02	0.09±0.03	0.92±1.01	0.13±0.18	0.05±0.02	0.03±0.01
Heavy metals	1.05±0.28	0.91±0.80	1.59±0.51	0.40±0.31	1.16±0.66	0.56±0.28	1.62±0.65	1.93±1.10	11.80±13.91	1.69±2.38	0.90±0.26	0.53±0.09



Table 5 Mass concentrations of PE PM_{2.5}-bound PAHs, PAEs and hopanes species for women at DF, students at WB and drivers at MT (ng m⁻³).

Specific species (abbreviation)	Women at DF		Students at WB		Drivers at MT	
	Average	Stdev*	Average	Stdev*	Average	Stdev*
acenaphthene (ACE)	0.4	0.5	0.6	1.2	0.7	1.7
fluorene (FLO)	0.3	0.3	0.3	0.6	0.1	0.0
phenanthrene (PHE)	0.8	0.4	0.9	1.2	0.6	0.1
anthracene (ANT)	0.3	0.2	0.2	0.2	0.2	0.0
fluoranthene (FLU)	1.0	0.4	1.0	0.7	0.6	0.1
pyrene (PYR)	1.2	0.5	1.0	0.5	0.6	0.1
benzo[a]anthracene (BaA)	4.5	8.5	2.2	1.5	1.1	0.5
chrysene (CHR)	6.1	11.2	3.0	1.6	1.8	0.8
benzo[b]fluoranthene (BbF)	11.6	19.2	5.6	2.7	3.6	1.2
benzo[k]fluoranthene (BkF)	4.9	4.2	5.0	2.9	3.3	1.1
benzo[a]fluoranthene (BaF)	3.8	5.3	2.1	2.4	1.5	0.8
benzo[e]pyrene (BeP)	7.7	8.1	5.0	2.5	3.6	0.7
benzo[a]pyrene (BaP)	9.7	12.5	5.5	5.7	3.5	1.6
perylene (PER)	2.8	5.0	1.3	1.4	0.8	0.4
indeno[1,2,3-cd]pyrene (IcdP)	9.4	9.3	6.4	4.5	4.5	0.7
benzo[ghi]perylene (BghiP)	7.8	6.1	6.0	3.6	6.4	0.5
dibenzo[a,h]anthracene (DahA)	1.8	2.2	1.0	0.6	0.6	0.1
coronene (COR)	2.8	1.6	2.3	1.4	3.3	0.4
dibenzo[a,e]pyrene (DaeP)	0.7	0.7	0.5	0.3	0.3	0.1
ΣPAHs	77.4	47.9	49.9	30.7	37.0	7.4
dimethyl phthalate (DMP)	2.2	1.0	9.6	27.9	1.9	0.5
diethyl phthalate (DEP)	8.3	4.1	146.5	517.0	6.8	1.4
di-n-butyl phthalate (DBP)	224.8	90.6	440.7	848.4	248.2	42.1
benzyl butyl phthalate (BBP)	13.8	4.3	19.7	37.3	8.1	2.9
bis(2-ethylhexyl)phthalate (DEHP)	566.4	181.4	688.0	899.1	376.3	144.5
di-n-octyl phthalate (DNOP)	40.9	16.9	43.8	26.2	33.0	31.0
bis(2-ethylhexyl)adipate (DEHA)	25.6	6.0	32.0	41.8	23.8	19.0
ΣPAEs	882.0	193.3	1380.4	335.2	698.1	192.4
17α(H)-22,29,30-trisnorhopane (Tm)	1.3	0.5	1.3	1.9	2.5	0.5
17α(H)-21β(H),30-norhopane (αβ-NH)	4.0	1.2	3.3	4.1	10.6	1.9
17β(H)-21α(H),30-norhopane (βα-NH)	1.5	1.8	1.1	1.5	1.9	0.3
17α(H)-21β(H)-hopane (αβ-HH)	4.3	1.9	3.6	5.4	11.5	2.2
17α(H)-21α(H)-hopane (αα-HH)	0.8	0.2	1.0	2.0	3.6	2.1
17β(H)-21α(H)-hopane (βα-HH)	0.7	0.2	0.8	1.2	2.9	1.2
17α(H)-21β(H),(22S)-homohopane (αβ-S-HH)	2.3	0.7	2.2	2.4	8.9	1.3
17α(H)-21β(H),(22R)-homohopane (αβ-R-HH)	2.2	0.8	2.1	2.1	8.9	1.3
Σhopanes	17.1	6.4	15.6	6.1	50.9	7.9

*: standard deviation



Table 6 Non-cancer risks of heavy metals and cancer risks of PAHs and PAEs via inhalation exposure way in PE PM_{2.5} of women at DF, students at WB and drivers at MT in dry and wet seasons.

	Dry season			Wet season		
	Women	Students	Drivers	Women	Students	Drivers
Non-cancer risk						
Mn	5.71E-03	2.02E-02	1.09E-02	4.83E-03	2.31E-02	4.26E-03
Ni	1.44E-04	5.60E-04	1.77E-04	4.49E-04	2.59E-03	2.00E-04
Zn	1.45E-04	2.15E-04	6.16E-05	1.24E-04	5.45E-04	5.05E-05
Pb	1.75E-03	5.98E-03	9.33E-04	2.97E-03	2.95E-02	7.75E-04
HI	7.74E-03	2.70E-02	1.21E-02	8.37E-03	5.57E-02	5.29E-03
Cancer risk (ILCR)						
PAHs ([BaP] _{eq})	3.13E-06	6.43E-06	3.22E-06	9.33E-06	3.68E-06	3.42E-06
PAEs (DEHP)	2.92E-07	3.36E-07	1.86E-07	3.15E-07	4.86E-07	1.16E-07



Appendix A.

2016

**Assessing Air Pollution Exposures in southern West Africa
- Questionnaire for Women**

1. Participant name: _____
2. Interviewer name: _____
3. Sampling site: _____
4. Address of the interview place: _____

5. Address of the participant home: _____

6. Interview date: ____/____/____ (yyyy/mm/dd)
7. Interview start time: _____ End time: _____

This questionnaire is for research purposes only. Please think carefully and answer all the questions below. Your answers will be kept completely confidential and your personal information will not be disclosed or displayed in any way and any case.

Thank you for your cooperation!



A. Basic Information

【Please choose by or fill in the answer】

A1. Gender: (0) Male (1) Female

A2. Age: _____ years old

A3. Height: _____ cm; Weight: _____ Kg

A4. Marital status:

(0) Single (1) Married

(2) Divorced (3) Widowed

A5. Highest level of education:

(0) Primary school

(1) Junior high school

(2) High school

(3) Undergraduate

(4) Above undergraduate

A6. The total number of family members
(including you): _____

A7. Number of adults (18 years or older;
including you): _____

A8. Family total annual income: _____ West
African Franc / Month

A9. Now professional:

(0) Unemployed

(1) Students

(2) Retired staff

(3) Workers

(4) Farmers

(5) Corporate staff

(6) Civil servants

(7) Housewife

(8) Driver

(9) Others: _____

A10. Work address (if any)?

A11. Your housing type:

(0) Apartment

(1) One-storey house

(2) Other: _____

A12. _____ Floor

A13. Residential area: _____ m²

(0) One room and one hall

(1) Two rooms and one hall

(2) Three rooms and two halls

(3) Others

A14. How long did you move in this house after
it was decorated?

(0) < 3 months (1) 3-6 months

(2) 6-12 months (3) > 12 months

A15. When was your house built? _____(year)

A16. How many years have you lived in this
house? _____ (year)

A17. What material is your house built?

(0) Brick (1) Armored concrete

(2) Timber (3) Other materials

A18. What is the material of the floor in your
house?

(0) Cement (1) Marble

(2) Solid wood (3) Composite wood

(4) Tile (5) Plastic (6) Rock

(7) Brick (8) Bare soil

A19. What is the material of the furniture in
your house?

(0) Solid wood (1) Plastic

(2) Leather (3) Metal

(4) Stone (5) Glass

(6) Cloth (7) Artificial board

A20. Has your house been decorated in the last
year?

(0) Yes (1) No

↳A21. What kind of decoration?

(0) Paint

(1) Change the floor



(2) Add new furniture

(3) Other: _____

A22. Does your house have ventilation equipment?

(0) Yes (1) No

↳ **A23. What kind of equipment?**

(Please select all suitable answers)

(0) Hanging air conditioner

(1) Cabinet air conditioner

(2) Ventilator

A24. How far is your house from the main road?

(0) <20m (1) 20-100m

(2) >100m

A25. Do you smoke?

(0) Yes (1) No

A26. Do you have a smoking history?

(0) Yes (1) No

↳ **A27. How long is your smoking**

history? _____ years

A28. Does your family member smoke (not including you)?

(0) Yes (1) Used to smoke

(2) No

A29. In general, are you influenced by second hand smoke in the following environments often?

(0) Your own home

(1) Working environment

(2) Other' house

(3) Restaurants, bars, supermarkets,

streets and so on

(4) Other: _____

(5) Rarely affected by second hand

smoke

A30. Do you often drink alcohol?

(0) Yes (1) No (2) Not often

A31. Drinking type:

(0) Alcohol (1) Beer

(2) Wine (3) Other

A32. Drinking frequency per week:

(0) < once (1) 1-3 times

(2) > 3 times (3) I don't know

A33. Please describe your health status in general.

(0) Very good (1) Good

(2) Not bad (3) Not good

A34. Do you have a family history of allergies?

(0) Yes (1) No (2) I don't know

A35. Have you been allergic to flowers or animals, food, etc.?

(0) Yes (1) No (2) I don't know

A36. Have you ever had itchy skin and red patches (rashes) lasting more than 6 months?

(0) Yes (1) No (2) I don't know

A37. Do your parents have asthma?

(0) Yes (1) No (2) I don't know

A38. Do you have asthma?

(0) Yes (1) No (2) I don't know

A39. Have you heard any noise or wheeze in your chest (whistle sound) during breathing?

(0) Yes (1) No (2) I don't know

A40. Have you had symptoms of sneezing, runny nose or stuffy nose in the absence of a cold?

(0) Yes (1) No (2) I don't know

A41. Are you diagnosed with high blood pressure by your doctor?

(0) Yes (1) No (2) I don't know

↳ **A42. Are you taking antihypertensive drugs every day?**

(0) Yes (1) No

A43. Are you diagnosed with diabetes by your doctor?

(0) Yes (1) No (2) I don't know



A44. Are you diagnosed with a myocardial infarction by your doctor?

(0) Yes (1) No (2) I don't know



B. Environment

【Please choose by or fill in the answer】

B1. Do you cook at home?

(0) Yes (1) No

↳B2. Cooking frequency per day:

(0) Once (1) Twice

(2) Three times

(3) > Three times

B3. What kind of fuel is used at home for cooking? (Please select all suitable answers)

(0) Natural gas

(1) Coal

(2) Liquefied petroleum gas (LPG)

(3) Electricity

(4) Other: _____

(5) Don't cook at home

B4. Does your kitchen have ventilation equipment?

(0) Yes (1) No

↳B5. What kind of equipment? (Please select all suitable answers)

(0) Kitchen smoke exhaust ventilator

(1) Kitchen ventilator

(2) Chimney

B6. Your kitchen area: _____ m²

B7. Do you usually make smoked fish?

(0) Yes (1) No

C. Travel habits

【Please choose by or fill in the answer】

B8. Where do you usually make smoked fish?

(0) Kitchen at home (1) Yard at home

(2) Outdoor-working place

B9. How many times do you make smoked fish per week? _____ times

B10. How long do you average make the smoked fish each time? _____ min

B11. When do you usually make smoked fish?

(0) Morning

(1) Noon

(2) Afternoon

(3) Evening

B12. Do you raise pet at home?

(0) Yes (1) No

B13. Do you grow flowers or plants at home?

(0) Yes (1) No

B14. Do you use insecticide at home?

(0) Yes (1) No (2) I don't know

B15. What's the open conditions of your windows at home every day?

(0) Wide open < 1h

(1) Wide open > 3h

(2) Half open < 1h

(3) Half open > 1h

(4) Never open

B16. What tool do you use to clean house?

(0) Broom and mop

(1) Electric dust collector



C1. How much time do you spend indoors per day, except sleeping?

(0) > 50% (1) = 50% (2) < 50%

C2. How much sleep do you have daily, including daytime and nighttime? _____ h

C3. What time of the day do you stay in your house in general? (Please select all suitable answers)

Morning		Afternoon			Evening				
8-10 am	10-12 am	12-14 pm	14-16 pm	16-18 pm	18-20 pm	20-22 pm	22-24 pm	Next day 0-4 am	Nexr day 4-8 am
<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>

C4. What time of the day do you stay at your working place in general? (Please select all suitable answers)

Morning		Afternoon			Evening				
8-10 am	10-12 am	12-14 pm	14-16 pm	16-18 pm	18-20 pm	20-22 pm	22-24 pm	Next day 0-4 am	Nexr day 4-8 am
<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>

C5. What is your main travel style when you travel < 3 km from your house?

(0) Walk (1) Bicycle (2) Motorcycle
 (3) Public bus (4) Car (5) Seldom travel

C6. How many hours do you spend on traveling each day?

(0) 0 h (1) 0-0.5 h (2) 0.5-1 h (3) 1-1.5 h
 (4) 1.5-2 h (5) 2-3 h (6) > 3 h

C7. How often do you perform outdoor exercises for longer than 30 minutes per week? _____ times

C8. What kind of transportation do you use when you go to work (if any)? How long does it take?

(0) Walk average time _____ min
 (1) Bicycle, tricycle average time _____ min
 (2) Electric bicycles, motorcycles average time _____ min
 (3) Bus, private car, taxi average time _____ min

The interview is over. Thank you for your cooperation again!



Appendix B.

2016

Assessing Air Pollution Exposures in southern West Africa - Questionnaire for Students

1. Participant name: _____
2. Interviewer name: _____
3. Sampling site: _____
4. Address of the interview place: _____

5. Address of the participant home: _____

6. Interview date: ____/____/____ (yyyy/mm/dd)
7. Interview start time: _____ End time: _____

This questionnaire is for research purposes only. Please think carefully and answer all the questions below. Your answers will be kept completely confidential and your personal information will not be disclosed or displayed in any way and any case.

Thank you for your cooperation!



A. Basic Information

【Please choose by or fill in the answer】

A1. Gender: (0) Male (1) Female

A2. Age: _____ years old

A3. Height: _____ cm

A4. Weight: _____ Kg

A5. Grade: _____ grade

A6. The total number of family members
(including you): _____

A7. Number of adults (18 years or older): _____

A8. Your housing type:

(3) Apartment

(4) One-storey house

(5) Other: _____

↳ A9. _____ Floor

A10. Residential area: _____ m²

(4) One room and one hall

(5) Two rooms and one hall

(6) Three rooms and two halls

(7) Others

A11. How many years have you lived in this
house? _____ (year)

A12. What material is your house built?

(0) Brick (1) Armored concrete

(2) Timber (3) Other materials

A13. What is the material of the floor in your
house?

(0) Cement (1) Marble

(2) Solid wood (3) Composite wood

(4) Tile (5) Plastic (6) Rock

(7) Brick (8) Bare soil

A14. What is the material of the furniture in
your house?

(0) Solid wood (1) Plastic

(2) Leather (3) Metal

(4) Stone (5) Glass

(6) Cloth (7) Artificial board

A15. Has your house been decorated in the last
year?

(1) Yes (1) No

↳ A16. What kind of decoration?

(0) Paint

(1) Change the floor

(2) Add new furniture

(3) Other: _____

A17. Does your house have ventilation
equipment?

(1) Yes (1) No

↳ A18. What kind of equipment? (Please
select all suitable answers)

(0) Hanging air conditioner

(1) Cabinet air conditioner

(2) Ventilator

A19. How far is your house from the main road?

(0) <20m (1) 20-100m

(2) >100m

A20. Does your classroom have ventilation
equipment?

(0) Yes (1) No

↳ A21. What kind of equipment? (Please



select all suitable answers)

(0) Hanging air conditioner

(1) Cabinet air conditioner

(2) Ventilator

A22. How far is your classroom from the main road?

(0) <20m (1) 20-100m

(2) >100m

A23. Do you smoke?

(0) Yes (1) No

A24. Does your family member smoke (not including you)?

(0) Yes (1) Used to smoke

(2) No

A25. In general, are you influenced by second hand smoke in the following environments often?

(0) Your own home

(1) Working environment

(2) Other' house

(3) Restaurants, bars, supermarkets,

streets and so on

(4) Other: _____

(5) Rarely affected by second hand

smoke

A26. Do you often drink alcohol?

(0) Yes (1) No (2) Not often

A27. Mainly drinking type:

(0) Alcohol (1) Beer

(2) Wine (3) Other

A28. Please describe your health status in general.

(0) Very good (1) Good

(2) Not bad (3) Not good

A29. Do you have a family history of allergies?

(0) Yes (1) No (2) I don't know

A30. Have you been allergic to flowers or animals, food, etc.?

(0) Yes (1) No (2) I don't know

A31. Have you ever had itchy skin and red patches (rashes) lasting more than 6 months?

(0) Yes (1) No (2) I don't know

A32. Do your parents have asthma?

(0) Yes (1) No (2) I don't know

A33. Do you have asthma?

(0) Yes (1) No (2) I don't know

A34. Have you heard any noise or wheeze in your chest (whistle sound) during breathing?

(0) Yes (1) No (2) I don't know

A35. Have you had symptoms of sneezing, runny nose or stuffy nose in the absence of a cold?

(0) Yes (1) No (2) I don't know



B. Environment

【Please choose by or fill in the answer】

B1. Do you cook at home?

(0) Yes (1) No

↳ **B2. Cooking frequency per day:**

(0) Once (1) Twice

(2) Three times

(3) > Three times

B3. What kind of fuel is used at home for cooking? (Please select all suitable answers)

(0) Natural gas

(1) Coal

(2) Liquefied petroleum gas (LPG)

(3) Electricity

(4) Other: _____

(5) Don't cook at home

B4. Does your kitchen have ventilation equipment?

(0) Yes (1) No

↳ **B5. What kind of equipment?**

(Please select all suitable answers)

(0) Kitchen smoke exhaust ventilator

(1) Kitchen ventilator

(2) Chimney

B6. Your kitchen area: _____ m²

B7. Does your family usually make smoked fish?

(0) Yes (1) No

B8. How many times does your family eat smoked fish per week? _____ times.

B9. Do you raise pet at home?

(0) Yes (1) No

B10. Do you grow flowers or plants at home?

(0) Yes (1) No

B11. Do you use insecticide at home?

(0) Yes (1) No (2) I don't know

B12. What's the open conditions of your windows at home every day?

(0) Wide open < 1h

(1) Wide open > 3h

(2) Half open < 1h

(3) Half open > 3h

(4) Never open

B13. What's the open conditions of your windows at classroom every day?

(0) Wide open < 1h

(1) Wide open > 3h

(2) Half open < 1h

(3) Half open > 3h

(4) Never open

B14. What tool does your family use to clean house?

(2) Broom and mop

(3) Electric dust collector

B15. What tool does your class use to clean classroom?

(0) Broom and mop

(1) Electric dust collector

**B16. How far does dumps away from your house?
_____ m**

Walking time:

(0) < 5min



- (1) 5-10min
- (2) 10-15min
- (3) 15-20min
- (4) 20-30 min
- (5) > 30 min

B17. How far does dumps away from your school? _____ m

Walking time:

- (0) < 5min
- (1) 5-10min
- (2) 10-15min
- (3) 15-20min
- (4) 20-30 min
- (5) > 30 min

B18. Can you see waste burning at home?

- (0) Yes
- (1) No

↳ **B19. How many times per week?**

_____ times

B20. Can you see waste burning at school?

- (0) Yes
- (1) No

↳ **B21. How many times per week?**

_____ times

B22. Can you smell smoke from waste burning at home?

- (0) Yes
- (1) No

B23. Can you smell smoke from waste burning at school?

C. Travel habits

【Please choose by or fill in the answer】

C1. How much time do you spend indoors per day, except sleeping?

- (0) Yes
- (1) No

B24. How to deal with your home waste?

- (0) Throwing to dumps
- (1) Burning by yourselves
- (2) I don't know

B25. Does waste burning in dumps impact on your live?

- (0) Yes
- (1) No

↳ **B26. What specific performance?**

(Please select all suitable answers)

- (0) Road congestion, inconvenient travel
- (1) Air odor, black smoke filled
- (2) Air pollution, low visibility
- (3) Water pollution, fish and shrimp death

B27. Does waste burning in dumps impact on your health?

- (0) Yes
- (1) No

↳ **B28. What specific performance?**

(Please select all suitable answers)

- (0) Congestion, runny nose
- (1) Dry eyes, tears
- (2) Skin allergies
- (3) Throat dry, inflamed
- (4) Difficulty breathing
- (5) Other: _____



(0) > 50% (1) = 50% (2) < 50%

C2. How much sleep do you have daily, including daytime and nighttime? _____ h

C3. What time of the day do you stay in your house in general? (Please select all suitable answers)

Morning		Afternoon			Evening				
8-10 am	10-12 am	12-14 pm	14-16 pm	16-18 pm	18-20 pm	20-22 pm	22-24 pm	Next day 0-4 am	Nexr day 4-8 am
<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>

C4. What time of the day do you stay at school in general? (Please select all suitable answers)

Morning		Afternoon			Evening				
8-10 am	10-12 am	12-14 pm	14-16 pm	16-18 pm	18-20 pm	20-22 pm	22-24 pm	Next day 0-4 am	Nexr day 4-8 am
<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>

C5. What is your main travel style when you travel < 3 km from your house?

- (0) Walk (1) Bicycle (2) Motorcycle
 (3) Public bus (4) Car (5) Seldom travel

C6. How many hours do you spend on traveling each day?

- (0) 0 h (1) 0-0.5 h (2) 0.5-1 h (3) 1-1.5 h
 (4) 1.5-2 h (5) 2-3 h (6) > 3 h

C7. How often do you perform outdoor exercises for longer than 30 minutes per week? _____ times

C8. What kind of transportation do you use when you go to school? How long does it take?

- (0) Walk average time _____ min
 (1) Bicycle, tricycle average time _____ min
 (2) Electric bicycles, motorcycles average time _____ min
 (3) Bus, private car, taxi average time _____ min

The interview is over. Thank you for your cooperation again!



Appendix C.

2016

Assessing Air Pollution Exposures in southern West Africa - Questionnaire for Drivers

1. Participant name: _____
2. Interviewer name: _____
3. Sampling site: _____
4. Address of the interview place: _____

5. Address of the participant home: _____

6. Interview date: ____/____/____ (yyyy/mm/dd)
7. Interview start time: _____ End time: _____

This questionnaire is for research purposes only. Please think carefully and answer all the questions below. Your answers will be kept completely confidential and your personal information will not be disclosed or displayed in any way and any case.

Thank you for your cooperation!



A. Basic Information

【Please choose by or fill in the answer】

A1. Gender: (0) Male (1) Female

A2. Age: _____ years old

A3. Height: _____ cm; Weight: _____ Kg

A4. Marital status:

(0) Single (1) Married

(2) Divorced (3) Widowed

A5. Highest level of education:

(5) Primary school

(6) Junior high school

(7) High school

(8) Undergraduate

(9) Above undergraduate

A6. The total number of family members
(including you): _____

A7. Number of adults (18 years or older;
including you): _____

A8. Family total annual income: _____ West
African Franc / Month

A9. Which kind of car do you drive when you
work?

(0) Motorcycle (1) Car

(2) Tricycle (3) Others _____

A10. As a driver, how long do you work per
day?

(0) < 1h

(1) 1-3h

(2) 4-6h

(3) 7-9h

(4) 10-12h

(5) >12h

A11. Your housing type:

(6) Apartment

(7) One-storey house

(8) Other: _____

↳A12. _____ Floor

A13. Residential area: _____ m²

(8) One room and one hall

(9) Two rooms and one hall

(10) Three rooms and two halls

(11) Others

A14. How long did you move in this house after
it was decorated?

(0) < 3 months (1) 3-6 months

(2) 6-12 months (3) > 12 months

A15. When was your house built? _____(year)

A16. How many years have you lived in this
house? _____ (year)

A17. What material is your house built?

(0) Brick (1) Armored concrete

(2) Timber (3) Other materials

A18. What is the material of the floor in your
house?

(0) Cement (1) Marble

(2) Solid wood (3) Composite wood

(4) Tile (5) Plastic (6) Rock

(7) Brick (8) Bare soil

A19. What is the material of the furniture in
your house?

(0) Solid wood (1) Plastic

(2) Leather (3) Metal

(4) Stone (5) Glass

(6) Cloth (7) Artificial board

A20. Has your house been decorated in the last
year?

(2) Yes (1) No

↳A21. What kind of decoration?

(0) Paint

(1) Change the floor

(2) Add new furniture



(3) Other: _____

A22. Does your house have ventilation equipment?

(2) Yes (1) No

↳ A23. What kind of equipment?

(Please select all suitable answers)

(0) Hanging air conditioner

(1) Cabinet air conditioner

(2) Ventilator

A24. How far is your house from the main road?

(0) < 20m (1) 20-100m

(2) > 100m

A25. Do you smoke?

(0) Yes (1) No

A26. Do you have a smoking history?

(0) Yes (1) No

↳ A27. How long is your smoking history? _____ (year)

A28. Does your family member smoke (not including you)?

(0) Yes (1) Used to smoke

(2) No

A29. In general, are you influenced by second hand smoke in the following environments often?

(0) Your own home

(1) Working environment

(2) Other' house

(3) Restaurants, bars, supermarkets,

streets and so on

(4) Other: _____

(5) Rarely affected by second hand

smoke

A30. Do you often drink alcohol?

(0) Yes (1) No

(2) Not often (3) Already abstaining

A31. Mainly drinking type:

(0) Alcohol (1) Beer

(2) Wine (3) Other

A32. Drinking frequency per week (If any): _____ times.

A33. Please describe your health status in general.

(0) Very good (1) Good

(2) Not bad (3) Not good

A34. Do you have a family history of allergies?

(0) Yes (1) No (2) I don't know

A35. Have you been allergic to flowers or animals, food, etc.?

(0) Yes (1) No (2) I don't know

A36. Have you ever had itchy skin and red patches (rashes) lasting more than 6 months?

(0) Yes (1) No (2) I don't know

A37. Do your parents have asthma?

(0) Yes (1) No (2) I don't know

A38. Do you have asthma?

(0) Yes (1) No (2) I don't know

A39. Have you heard any noise or wheeze in your chest (whistle sound) during breathing?

(0) Yes (1) No (2) I don't know

A40. Have you had symptoms of sneezing, runny nose or stuffy nose in the absence of a cold?

(0) Yes (1) No (2) I don't know

A41. Are you diagnosed with high blood pressure by your doctor?

(0) Yes (1) No (2) I don't know

↳ A42. Are you taking antihypertensive drugs every day?

(0) Yes (1) No

A43. Are you diagnosed with diabetes by your doctor?

(0) Yes (1) No (2) I don't know

A44. Are you diagnosed with a myocardial



infarction by your doctor?

(0) Yes (1) No (2) I don't know



B. Environment

【Please choose by or fill in the answer】

B1. Do you cook at home?

(0) Yes (1) No

↳ **B2. Cooking frequency per day:**

(0) Once (1) Twice

(2) Three times

(3) > Three times

B3. What kind of fuel is used at home for cooking? (Please select all suitable answers)

(0) Natural gas

(1) Coal

(2) Liquefied petroleum gas (LPG)

(3) Electricity

(4) Other: _____

(5) Don't cook at home

B4. Does your kitchen has ventilation equipment?

(0) Yes (1) No

↳ **B5. What kind of equipment?**

(Please select all suitable answers)

(0) Kitchen smoke exhaust ventilator

(1) Kitchen ventilator

(2) Chimney

B6. Your kitchen area: _____ m²

B7. Do you cook at home?

(0) Yes (1) No

B8. Does your family usually make smoked fish?

(0) Yes (1) No

B9. How many times do you eat smoked fish at home per week? _____ times.

B10. Do you raise pet at home?

(0) Yes (1) No

B11. Do you grow flowers or plants at home?

(0) Yes (1) No

B12. Do you use insecticide at home?

(0) Yes (1) No (2) I don't know

B13. What's the open conditions of your windows at home every day?

(0) Wide open < 1h

(1) Wide open > 3h

(2) Half open < 1h

(3) Half open > 1h

(4) Never open

B14. What tool do you use to clean house?

(4) Broom and mop

(5) Electric dust collector

B15. What kind of road do you usually drive on?

(Please select all suitable answers)

(0) Unsurfaced road (1) Stone road

(2) Asphalt road (3) Cement road

(4) Others: _____

B16. What kind of environment do you usually drive on? (Please select all suitable answers)

(0) Business district (1) Industrial area

(2) Residential area (3) Suburbs

(4) Others: _____

B17. What type of power does your motorcycle use?

(0) Diesel (1) Gasoline

(2) Manpower (3) Electricity

B18. When do you usually work with motorcycle?



(0) Daytime (1) Nighttime

(2) Morning rush hour

(3) Night rush hour

B19. What is the main purpose of your driving motorcycle?

(0) Freight (1) Passenger

(2) Both of above (3) Others _____

B20. How far do you drive motorcycle per day?

_____ Km

B21. How do you feel about the surrounding air

C. Travel habits

【Please choose by or fill in the answer】

C1. How much time do you spend indoors per day, except sleeping?

(0) > 50% (1) = 50% (2) < 50%

C2. How much sleep do you have daily, including daytime and nighttime? _____ h

C3. What time of the day do you stay in your house in general? (Please select all suitable answers)

Morning		Afternoon			Evening				
8-10 am	10-12 am	12-14 pm	14-16 pm	16-18 pm	18-20 pm	20-22 pm	22-24 pm	Next day 0-4 am	Nexr day 4-8 am

C4. What time of the day do you drive cars for work in general? (Please select all suitable answers)

Morning		Afternoon			Evening				
8-10 am	10-12 am	12-14 pm	14-16 pm	16-18 pm	18-20 pm	20-22 pm	22-24 pm	Next day 0-4 am	Nexr day 4-8 am

C5. What is your main travel style when you travel < 3 km from your house?

(0) Walk (1) Bicycle (2) Motorcycle

(3) Public bus (4) Car (5) Seldom travel

C6. How often do you perform outdoor exercises for longer than 30 minutes per week?

_____ times

The interview is over. Thank you for your cooperation again!



Appendix D.

2016

Assessing Air Pollution Exposures in
southern West Africa

DETAILED TIME-ACTIVITY DIARY

Participant name: _____

Start date: __/__/__ Time: __:__

MM/DD/YY

End date: __/__/__ Time: __:__

MM/DD/YY

Interviewer name: _____



DETAILED ACTIVITY DIARY FOR WOMAN

Time	Location (mark all that apply)	Activities (mark all that apply)
x:00 am-x:30 am* Pollution sources <input type="checkbox"/> Environmental tobacco smoke <input type="checkbox"/> Cooking <input type="checkbox"/> smoking fish <input type="checkbox"/> no smoking fish <input type="checkbox"/> Use of cleaning products <input type="checkbox"/> House decoration <input type="checkbox"/> Transportation emissions <input type="checkbox"/> Other (specify): _____	Indoors <input type="checkbox"/> Home <input type="checkbox"/> kitchen <input type="checkbox"/> living room <input type="checkbox"/> bedroom <input type="checkbox"/> courtyard <input type="checkbox"/> Work <input type="checkbox"/> Other (specify): _____ Outdoors <input type="checkbox"/> In transit (specify): _____ <input type="checkbox"/> Other (specify): _____	<input type="checkbox"/> Cooking <input type="checkbox"/> Smoking fish <input type="checkbox"/> Cleaning room <input type="checkbox"/> Washing the clothes <input type="checkbox"/> Taking food <input type="checkbox"/> Watching television <input type="checkbox"/> Taking a rest <input type="checkbox"/> Working at office <input type="checkbox"/> Going out <input type="checkbox"/> Taking exercise <input type="checkbox"/> Shopping <input type="checkbox"/> Visiting friends <input type="checkbox"/> Other (specify): _____

*X refers to the hour.



DETAILED ACTIVITY DIARY FOR STUDENT

Time	Location (mark all that apply)	Activities (mark all that apply)
x:00 am-x:30 am* Pollution sources <input type="checkbox"/> Environmental tobacco smoke <input type="checkbox"/> Cooking <input type="checkbox"/> smoking fish <input type="checkbox"/> no smoking fish <input type="checkbox"/> Use of cleaning products <input type="checkbox"/> House decoration <input type="checkbox"/> Transportation emissions <input type="checkbox"/> Other (specify): _____	Indoors <input type="checkbox"/> Home <input type="checkbox"/> kitchen <input type="checkbox"/> living room <input type="checkbox"/> bedroom <input type="checkbox"/> courtyard <input type="checkbox"/> School classroom <input type="checkbox"/> Other (specify): _____ Outdoors <input type="checkbox"/> In transit (specify): _____ <input type="checkbox"/> Other (specify): _____	<input type="checkbox"/> Cooking <input type="checkbox"/> Smoking fish <input type="checkbox"/> Cleaning room <input type="checkbox"/> Washing the clothes <input type="checkbox"/> Taking food <input type="checkbox"/> Watching television <input type="checkbox"/> Taking a rest <input type="checkbox"/> Studying at school <input type="checkbox"/> Studying at home <input type="checkbox"/> Going out <input type="checkbox"/> Taking exercise <input type="checkbox"/> Shopping <input type="checkbox"/> Playing <input type="checkbox"/> Other (specify): _____

*X refers to the hour.



DETAILED ACTIVITY DIARY FOR DRIVER

Time	Location (mark all that apply)	Activities (mark all that apply)
x:00 am-x:30 am* Pollution sources <input type="checkbox"/> Environmental tobacco smoke <input type="checkbox"/> Cooking <input type="checkbox"/> smoking fish <input type="checkbox"/> no smoking fish <input type="checkbox"/> Use of cleaning products <input type="checkbox"/> House decoration <input type="checkbox"/> Transportation emissions <input type="checkbox"/> Other (specify): _____	Indoors <input type="checkbox"/> Home <input type="checkbox"/> kitchen <input type="checkbox"/> living room <input type="checkbox"/> bedroom <input type="checkbox"/> courtyard <input type="checkbox"/> Other (specify): _____ Outdoors <input type="checkbox"/> In transit (specify): _____ <input type="checkbox"/> In the street <input type="checkbox"/> Other (specify): _____	<input type="checkbox"/> Cooking <input type="checkbox"/> Smoking fish <input type="checkbox"/> Cleaning room <input type="checkbox"/> Washing the clothes <input type="checkbox"/> Taking food <input type="checkbox"/> Watching television <input type="checkbox"/> Taking a rest <input type="checkbox"/> Driving the MOTO <input type="checkbox"/> Driving the car <input type="checkbox"/> Going out <input type="checkbox"/> Taking exercise <input type="checkbox"/> Shopping <input type="checkbox"/> Visiting friends <input type="checkbox"/> Other (specify): _____

*X refers to the hour.