

# Exploring sources of biogenic secondary organic aerosol compounds using chemical analysis and the FLEXPART model

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**Abstract.** Molecular tracers in secondary organic aerosols (SOA) can provide information on origin of SOA, as well as regional scale processes involved in their formation. In this study nine carboxylic acids, eleven organosulfates (OSs) and two nitrooxy organosulfates (NOSs) were determined in daily aerosol particle filter samples from Vavihill measurement station in southern Sweden during June and July 2012. Several of the observed compounds are photo-oxidation products from biogenic volatile organic compounds (BVOCs). Highest average mass concentrations were observed for carboxylic acids derived from fatty acids and monoterpenes ( $12.3 \pm 15.6$  and  $13.8 \pm 11.6$  ng/m<sup>3</sup>, respectively). The FLEXPART model was used to link 9 specific surface types to single measured compounds. It was found that the surface category “sea and ocean” was dominating the air mass exposure (56%) but contributed to low mass concentration of observed chemical compounds. A principal component (PC) analysis identified four components, where the one with highest explanatory power (49%) displayed clear impact of coniferous forest on measured mass concentration of a majority of the compounds. The three remaining PCs were more difficult to interpret, although azelaic, suberic, and pimelic acid were closely related to each other but not to any clear surface category. Hence, future studies should aim to deduce the biogenic sources and surface category of these compounds. This study bridges micro level chemical speciation to air mass surface exposure on the macro level.

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## 1 Introduction

Carbonaceous aerosols are abundant in ambient air around the world and account for 40% of the European PM<sub>2.5</sub> mass (Putaud et al., 2010). The carbonaceous aerosol fraction has severe effects on human health as well as a profound effect on the Earth climate system (Dockery et al., 1993; Pope et al., 1995). During summer, carbonaceous aerosols are mainly of biogenic origin, emitted either through primary emissions or gas-phase oxidation products from biogenic volatile organic compounds (BVOCs) (Genberg et al., 2011; Yttri et al., 2011). BVOCs are primarily emitted from plants as a tool for communication and to handle biotic and abiotic stress (Laothawornkitkul et al., 2009; Monson et al., 2013; Penuelas and Llusia, 2003; Sharkey et al., 2008). The emissions of BVOCs tend to increase with increasing temperature and photosynthetically active radiation (PAR) (Guenther et al., 1995; Guenther et al., 1993; Hakola et al., 2003). Global BVOC emissions are dominated by isoprene (C<sub>5</sub>H<sub>8</sub>) and monoterpenes (C<sub>10</sub>H<sub>16</sub>) (Laothawornkitkul et al., 2009). Isoprene is emitted from a variety of plants, however mainly from deciduous forests and shrubs which may account for more than 70% of the emissions (Guenther et al., 2006). Monoterpenes are largely emitted from coniferous trees like pine and spruce, but also from some deciduous trees, such as birch (Mentel et al., 2009). The most abundant monoterpenes in the boreal forests include  $\alpha$ -pinene,  $\beta$ -pinene,  $\Delta^3$ -carene and limonene (Hakola et al., 2012; Räsänen et al., 2008).

Biogenic secondary organic aerosols (BSOA) are formed by photo-oxidation of BVOCs, a process which tends to lower the saturation vapor pressure of the oxidation products relative to that of the BVOCs, thus forcing the gas-phase products to partition in the aerosol phase. BSOA has been shown to dominate over combustion source aerosols during summer (Genberg et al., 2011; Yttri et al., 2011). Yttri et al. (2011) performed source apportionment at four sites in Scandinavia during August 2009 and found that the biogenic contribution to the carbonaceous aerosol dominated (69-86%) at all four sites. Genberg et al. (2011) performed a one year source apportionment at one site in southern Sweden where they apportioned 80% of the summer-time carbonaceous aerosol to biogenic sources. Gelencser et al. (2007) also reported biogenic source dominance (63-76%) of the carbonaceous aerosol at 6 sites in south-central Europe during summer. Castro et al. (1999) observed a maximum and minimum in SOA in Europe during summer and winter, respectively. The relative SOA contribution was higher in rural forest and ocean measurement sites compared to urban sites (Castro et al., 1999).

BSOA consists of a myriad of organic compounds. Small (carbon number: C<sub>3</sub>-C<sub>6</sub>) and larger (C<sub>7</sub>-C<sub>9</sub>) dicarboxylic acids are highly hydrophilic and hygroscopic which have shown to result in potential strong climate effect due to their cloud condensation properties (Cruz and Pandis, 1998; Kerminen, 2001). Dicarboxylic acid contribution to carbon mass has been estimated to 1-3% in urban and semi-urban areas and up to 10% in remote marine areas (Kawamura and Ikushima, 1993; Kawamura and Sakaguchi, 1999). Primary aerosol sources of dicarboxylic acids in atmospheric aerosols include ocean emissions, engine exhausts and biomass burning (Kawamura and Kaplan, 1987; Kundu et al., 2010; Mochida et al., 2003). However, the main source of dicarboxylic acids are oxidation/photo-oxidation processes of VOCs (Zhang et al., 2010). These VOC precursors may originate from both anthropogenic and biogenic sources (Mochida et al., 2003). However,

BVOCs constitute more than 50% of all atmospheric VOCs, which is approximately equal to 1150 Tg carbon  $y^{-1}$  (Guenther et al., 1995; Hallquist et al., 2009).

Organosulfates (OSs) and nitrooxy organosulfates (NOSs) are low-volatility SOA products that in recent years have gained increased attention due to their potential properties as tracers for atmospheric ageing of aerosols in polluted air masses (Hansen et al., 2015; Hansen et al., 2014; Kristensen, 2014; Kristensen and Glasius, 2011; Nguyen et al., 2014). Many of these compounds are formed from isoprene and monoterpene oxidation products that react with sulfuric acid in the aerosol phase (Iinuma et al., 2007; Surratt et al., 2010; Surratt et al., 2007b). Since atmospheric sulfuric acid is mainly of anthropogenic origin (Zhang et al., 2009), presence of OSs from biogenic organic precursors thus indicates an effect of anthropogenic influence on BSOA (Hansen et al., 2014). Recently, OSs from anthropogenic organic precursors such as alkanes and PAHs have also been discovered (Riva et al., 2016; Riva et al., 2015). Tolocka and Turpin (2012) estimated that OSs could comprise up to 10% of the total organic aerosol mass in the U.S.

Many carboxylic acids and OSs originate from biogenic sources, however, the exact vegetation types emitting the precursor are poorly explored (Mochida et al., 2003; Tolocka and Turpin, 2012). Coniferous forests, deciduous forests, arable land, pastures etc. are all examples of potential BVOCs sources. Information on specific land surface type BVOCs and BSOA emissions is potentially crucial if an increased understanding should be reached on how land-use changes will affect organic aerosol levels and composition. van Pinxteren et al. (2010) demonstrated how air mass exposure to land cover affected the measured size-resolved organic carbon (OC), elemental carbon (EC) and inorganic compounds at a receptor site in Germany by using the HYSPLIT model. Yttri et al. (2011) measured one dicarboxylic acid (pinic acid), four OSs and two NOSs at four locations in Scandinavia and connected this measurement data to the FLEXPART model (Stohl et al., 2005) footprint of specific surface landscape types. They used thirteen types of surface landscapes and found that the two NOSs (MW 295 and MW 297, both formed from monoterpenes) correlated with air mass exposure to mixed forest (Yttri et al., 2011).

In this study, a comprehensive measurement campaign was conducted in order to investigate sources and levels of BSOA. 38 sequential 24h filter samples were analyzed for 9 species of carboxylic acids, 11 species of OSs and 2 species of NOSs at a rural background station in southern Sweden. FLEXPART model simulations at the time and location of the observations were then used to estimate the potential origin of the aerosols sampled.

## 2 Methods

### 2.1 Location and sampling

The Vavihill measurement station is a rural background station in southern Sweden (56°01' N, 13°09' E, 172 m.a.s.l.) within ACTRIS (Aerosols, Clouds and Trace gases Research Infrastructure) and EMEP (European Monitoring and Evaluation Programme). The surrounding landscape consists of pastures, mixed forest and arable land. The largest nearby cities are Helsingborg (140 000 inhabitants), Malmö (270 000 inhabitants) and Copenhagen (1 990 000 inhabitants) at a distance of 25, 45 and 50 km, respectively. These cities are in the west and southwest direction from the measurement station. Previous

observations have shown that air masses from continental Europe are usually more polluted than air masses from the north and westerly direction, i.e. Norwegian Sea and Atlantic Ocean (Kristensson et al., 2008).

38 filter samples of aerosols were collected at the Vavihill field station in southern Sweden from 10<sup>th</sup> of June to 18<sup>th</sup> of July 2012. Aerosols were collected on 150 mm quartz fibre filters (Advantec) using a high-volume sampler (Digitel, DHA-80) with a PM<sub>1</sub> inlet. The filters were heated to 900 °C for four hours prior to sampling, with the purpose of removing adsorbed organic compounds from the filters. The sampling air flow was 530 litres per minute and total sampling time per filter was 24 hours. Sampled filters were wrapped in aluminium foil and stored at -18 °C until extraction.

## 2.2 BSOA analysis

The method for extraction and analysis is based on previous studies (Hansen et al., 2014; Kristensen and Glasius, 2011; Nguyen et al., 2014) and thus only described briefly here. For extraction each filter was placed in a beaker and spiked with 15 µL of a 100 µg/mL recovery standard (camphoric acid). The filter was covered with 90% acetonitrile with 10% MilliQ water and extracted in a cooled ultrasound bath for 30 min. The extract was filtered through a Teflon filter (0.45 µm pore size, Chromafil) and evaporated until dryness using a rotary evaporator. The sample was then re-dissolved twice in 0.5 mL 3% acetonitrile, 0.1% acetic acid, and stored in a refrigerator (3-5 °C) until analysis. The samples were analysed with an Ultra High Performance Liquid Chromatograph (UHPLC, Dionex) coupled to a quadrupole Time-Of-Flight Mass Spectrometer (q-TOF-MS, Bruker Daltonics) through an electro-spray ionisation (ESI) inlet. The UHPLC stationary phase was an Acquity T3 1.8µm (2.1 × 100 mm) column from Waters, and the mobile phase consisted of eluent A: 0.1% acetic acid in MilliQ water and eluent B: Acetonitrile with 0.1% acetic acid. The operational eluent flow was 0.3 mL/min and an 18 min multistep gradient was applied: From 1 min to 10 min eluent B increased from 3% to 30%, then eluent B increased to 90% during 1 min, where it was held for 1 min, before eluent B was increased further to 95% (during 0.5 min) kept here for 3.5 min before reduction to 3% (during 0.5 min) for the remaining 0.5 min of the analysis. The ESI-q-TOF-MS instrument was operated in negative ionisation mode with a nebulizer pressure of 3.0 bar and a dry gas flow of 8 L/min. All data were acquired and processed using Bruker Compass software. The analyzed dicarboxylic acids, OSs and NOSs are summarized in Table 1 and 2, respectively. Authentic standards were used for identification and quantification of all carboxylic acids, while OSs and NOSs were identified based on their MS/MS loss of HSO<sub>4</sub><sup>-</sup> ( $m/z=97$ ) and an additional neutral loss of HNO<sub>3</sub> ( $u=63$ ) in the case of NOSs. This work focused on identification of OSs from biogenic organic precursors, since OSs from alkanes and PAHs had not been discovered at the time of the analysis. OSs and NOSs were quantified using surrogate standards of OS 250 derived from β-pinene (synthesized in-house), octyl sulfate sodium salt (≥95% Sigma-Aldrich) or D-mannose-6-sulfate sodium salt (≥90% Sigma-Aldrich) based on their retention times in the UHPLC-q-TOF-MS system (Table 2). A linear or quadratic relationship between peak area and concentration was demonstrated for all standards and surrogates, and the correlation coefficients, R<sup>2</sup>, of all calibration curves were better than 0.98 (n = 7 data points).

The analytical uncertainty was estimated to be <20% for carboxylic acids and <25% for OSs and NOSs. The uncertainty of the absolute concentrations of OSs and NOSs are higher than carboxylic acids due to lack of authentic standards.

### 2.3 Auxiliary measurements and analysis

PM<sub>2.5</sub> was measured with one hour time resolution using a tapered element oscillating microbalance (TEOM, Thermo, 8500 FDMS), and estimated uncertainty was less than 25%. Geographical air mass origin was analyzed with the Hybrid Single Particle Lagrangian Integrated Trajectory Model (HYSPLIT) model (Draxler and Hess, 1998; Stein et al., 2015). Gridded meteorological data from the Centre of Environmental Predictions (NCEP) Global Data Assimilation System (GDAS) were used as input by the trajectory model. Back-trajectories were calculated at an hourly frequency 120-hour backward in time and the trajectories started 100 m above ground at the Vavihill measurement site. For each filter sample, 24 trajectories were used since the sampling time was 24 hours.

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### 2.4 Source apportionment

The concentration and chemical composition of an aerosol sample depends on the trajectory of the sampled air mass in the days preceding the observation (whether or not it gets in contact with a source of aerosols or of aerosol precursors), but also on other meteorological factors such as the temperature and the amount of solar radiation (which control the chemical reactions that lead to production, destruction and transformation of aerosols), and the occurrence of precipitation, which can lead to a rapid scavenging of aerosol particles.

A formal source apportionment would typically involve using a complex chemistry-transport model, able to account for the most important of these factors, and comparing this model result's with the observations to validate or refute hypotheses on the origin of the aerosols. The size of our observation dataset is unfortunately too limited for such an exercise to provide meaningful results. Instead, we opted for a much simpler approach: We first used the FLEXPART model to compute back-trajectories corresponding to the air masses sampled. We then used these back-trajectories to estimate the exposure of each sample to various land-surface types. Finally, we analyzed the relations between the surface type exposures and the aerosols chemical composition of the samples to deduce information about the origin of the sampled aerosols.

#### 2.4.1 Footprint computations

For each observation, seven days footprints (i.e. sensitivity of the observations to surface processes) are computed, using the FLEXPART Lagrangian particle dispersion model in its version 10.0 (Seibert and Frank, 2004; Stohl et al., 2005). The response functions are computed hourly, seven days backward, on a 0.2°x0.2° grid ranging from 30°N to 65°N and from 2°W to 32°E. Only one (surface) layer is used, ranging from the surface to 400 m altitude. This choice of a relatively thick surface layer is a compromise between the necessity to account for a maximum of the aerosol production, which doesn't occur only at the earth (or canopy) surface, and the fact that the higher the altitude, the more mixed the air. This setting also means that we do not compute the sensitivity of the observations to aerosol production/destruction above 400 m. Even though aerosol formation occurs throughout the whole troposphere (de Reus et al., 2000), it would be impossible, with our simple model approach, to distinguish in-situ aerosol production from long-range transport.

Each footprint was computed based on the dispersion seven days backward in time of 100 000 particles. An average particle size of 250 nm was used, with a size distribution parameter (“dsigma”) of 12.5, meaning that 68% of the total particles mass is in a 250/12.5 nm to 250\*12.5 nm range. Previous particle-size measurements at Vavihill measurement station have shown a distribution around a mean of ~100 nm (Kristensson et al., 2008). The particles density was set to 1500 kg m<sup>-3</sup>. We briefly discuss the impact of these selected parameters in Section 3.4. FLEXPART configuration files are provided in Supplementary Information.

#### 2.4.2 Land surface type exposures

To compute the exposure of each sample to different land surface types, we coupled the information from the footprints to the CORINE 2012 land cover map (Copernicus, 2012). CORINE 2012 is a high resolution (250x250 m) map of the land surface types in the European Union (44 land surface categories, to which we added a “sea and ocean” category). The exposure  $E_i$  of one observation to the land type  $i$  is given by  $E_i = \sum_j f_j^i R_j$ , where  $j$  is one pixel of the domain,  $f_j^i$  is the fraction of the land surface type  $i$  in that pixel, and  $R_j$  is the sensitivity of the observation to that pixel (i.e. the value of the footprint at that location), divided by the height of the surface layer (400 m) and by the size of the grid-cell. It is important to remember that since aerosol formation/destruction along the particles trajectories is not accounted for in the FLEXPART simulations (except for deposition processes), these land surface exposures are not a proper source apportionment, but are only a tool to interpret the observations.

#### 2.4.3 Principal Component Analysis (PCA)

In order to deduce potential sources of measured BSOA compounds a principal component analysis (PCA) was performed on measured chemical compounds together with air mass exposure to the landscape surface types derived from the FLEXPART model. The principle of PCA is that if measured parameters from the same source are strongly correlated they are treated as one principal component (PC), i.e. PCA identifies variables that have a prominent role by analysis of correlation and variance. PCA has been an extensively used tool in order to reduce the complexity of atmospheric data and has been applied in several studies on aerosol chemical composition (Almeida et al., 2006; Chan and Mozurkewich, 2007; Ito et al., 2004; Nyanganyura et al., 2007; van Pinxteren et al., 2010; van Pinxteren et al., 2014; Viana et al., 2006; Wehner and Wiedensohler, 2003). PCA with VARIMAX rotation was performed by using the software SPSS (version 23, IBM). VARIMAX rotation was chosen due to its property of producing uncorrelated principal components (PCs) which aids interpretation of the data. In PCA, it is of good practice to transform all variables into a standardized format (i.e. Z-score), however the PCA solution from the standardized variables did not differ from the unstandardized one, hence unstandardized variables were used in the analysis. Extracted factors were varied from 2-6 in order to achieve the best logical and physical interpretation of the derived factors. The most interpretable result was found using 4 extracted factors.

### 3 Results and discussion

### 3.1 Variations and features in BSOA compounds

A total of 9 organic acids, 11 OSs and 2 NOSs of anthropogenic and biogenic origin were determined in the samples (Tables 1 and 2). All organic acids were quantified with authentic standards whereas the other compounds were quantified with surrogates (see experimental section). On average, the total mass of the organic chemical species from filters contributed to 0.3% ( $\pm 0.2\%$ , standard deviation) to  $PM_{2.5}$ . However, it is worth noting that the particles were sampled through a  $PM_1$  inlet, which may have excluded a considerable portion of the mass collected on filters compared to the  $PM_{2.5}$  mass measured by the TEOM. On the other hand, it has been shown that  $PM_1$  can comprise up to 90% of  $PM_{2.5}$  in rural locations during summertime (Gomiscek et al., 2004). Since no gravimetric analysis of filters was performed, no information on the total mass loading of  $PM_1$  is available.

In Table 3 and Fig. 1A concentrations of observed compounds during the sampling period are given. The compounds have been merged into groups based on their likely precursors in Fig. 1A (see Tables 1 and 2). It should be noted that pimelic acid, in Table 1 listed as having cycloheptene as a suggested precursor (i.e. to be of anthropogenic origin), also can be synthesized from salicylic acid (Müller, 1931), which is a compound naturally found in plants. Hence, whether the main formation route of pimelic acid is anthropogenic or natural is unclear. On the other hand, adipic acid is rarely found naturally and is originally synthesized from benzene (Tuttle Musser, 2000). Table 3 summarizes concentration ranges, means and standard deviations (SDs) for individual dicarboxylic acids, OSs and NOSs. In general the organic acids from monoterpenes and fatty acids dominate the total concentration over the entire period, where the concentration of acids from monoterpenes range from 1.7 to 49.0  $ng\ m^{-3}$  and the concentration of organic acids from fatty acids range from 0.03 to 64.1  $ng\ m^{-3}$ . The concentration of isoprene-derived OSs ranges from 0.34 to 21.6  $ng\ m^{-3}$  over the sampling period and dominates over the monoterpene-derived OSs. This pattern has also been observed in other studies in the Nordic countries (Yttri et al., 2011), and is in line with high emissions of isoprene during summer. The NOSs are low in average concentration (NOS 295=0.12 $\pm$ 0.11  $ng/m^3$ , NOS 297=0.05 $\pm$ 0.03  $ng/m^3$ ), and are lower than the observed mean concentration by Yttri et al. (2011) from the summer of 2011 (NOS 295=0.74  $ng/m^3$ , NOS 297=1.2  $ng/m^3$ ). This could be due to differences in aerosol sources and surrogate standards for quantification between the two studies.

The fatty acid-derived azelaic acid was found to be the most abundant dicarboxylic acid with a concentration range from 0.03 to 55.3  $ng/m^3$  (mean=10.5 $\pm$ 13.8  $ng/m^3$ ). Hyder et al. (2012) who measured 9 dicarboxylic acids in aerosol samples obtained at the Vavihill measurement station 2008-2009 also found azelaic acid to be the most prominent with peak concentration during summer (16.2  $ng/m^3$ ). The concentration of the anthropogenic acids is low (mean $\approx$ 2  $ng/m^3$ ) except during the 27<sup>th</sup> of June and the 6<sup>th</sup> of July when the concentration reaches 19.6 and 16.0  $ng/m^3$ , respectively. The spike in concentration of anthropogenic acids during these two days is caused by an increase in the concentration of adipic acid.

Correlations between the different compounds was investigated by Pearson correlation. All Pearson r-coefficients are given in Table 4. In general, the biogenic compounds (derived from isoprene and monoterpenes) correlated well ( $r\geq 0.8$ ) with each other. The only exception was OS 250, which showed low to medium correlation with the other compounds.

Three dicarboxylic acids (azelaic, pimelic and suberic acid) correlated well with each other ( $r > 0.87$ ). It is likely that the fatty acid-derived dicarboxylic acids has a different origin than isoprene- and monoterpene-generated acids, a conclusion that also was reached in a previous study (Hyder et al., 2012). It was expected that adipic acid would show good agreement with pimelic acid since they are both suggested to be of anthropogenic origin. However, this correlation was poor ( $r = 0.16$ ) and is believed to be explained by two strong concentration peaks in adipic acid (27<sup>th</sup> of June and 6<sup>th</sup> of July, Fig. 1A) with no corresponding peak in pimelic acid. Removing these two concentration peaks led to a better agreement between the two acids ( $r = 0.67$ ).

### 3.2 Air mass surface exposure

Figure 1B displays the exposures of the samples to the nine largest surface categories as percentage contribution and Tables 5 and 6 present the mean exposures and a correlation matrix for the investigated surface types. These surface categories are explained in more detail in the supplemental information. The “sea and ocean” category is dominating the exposure with an average of 56% ( $\pm 16\%$ ). This is hardly surprising since a majority of the incoming air mass is from the westerly region where the North Atlantic Ocean, North Sea and Norwegian Sea are situated. The second most common surface exposure is from “non-irrigated arable land” (mean = 19%  $\pm 8\%$ ). This is a common land type in continental Europe which is anti-correlated ( $r = -0.84$ ) to the “sea and ocean” surface category. The fact that several land-based surface categories anti-correlated to the “sea and ocean” category may be an indicator of the model working properly. The category “other” has a significant contribution to the total exposure (mean = 8%  $\pm 3\%$ ), but it groups 34 surface categories and is therefore difficult to interpret beyond the common fact that all these categories are land masses. It is important to remember that these exposures should not be read as a representation of the contribution of the land surface types to the production of the aerosols measured. For that, an estimation of the aerosol production (or transformation) associated to each surface category would be required. However, correlating the land surface exposures to the measured aerosol time series can provide an indication on the origin of the aerosols.

During a period of increased concentrations of molecular BSOA compounds (6<sup>th</sup> to 8<sup>th</sup> of July) the air mass was more exposed to land surface categories such as “non-irrigated arable land”, “coniferous forest”, “broad-leaved forest” and “pastures” on the expense of “sea and ocean” (Fig. 1A-B). Further, the category “other” is also increased during this particular period. Within the “other” category, “mixed forest”, “complex cultivation patterns”, “land principally occupied by agriculture, with significant areas of natural vegetation” and “transitional woodland/shrub” are dominating (more information about the surface categories can be found on the CORINE database homepage) (EEA, 2016). This particular concentration increase is caused by the fatty acid-derived organic acids, monoterpene-derived organic acids and isoprene-derived OSs (Fig. 1A). The concentration of  $PM_{2.5}$  does not provide any explanation of the cause of the high concentrations, since  $PM_{2.5}$  is in general high during the entire campaign period. Both the HYSPLIT and FLEXPART model revealed that arriving air masses during this period mainly had an origin from continental Europe (Fig. 2). As stated earlier, it has been



observed that air masses arriving from this direction usually carry more PM and OSs than from other directions (Nguyen et al., 2014; Kristensson et al., 2008).

The period of increased concentrations of molecular BSOA compounds (6<sup>th</sup> to 8<sup>th</sup> of July) are in large contrast to the “clean periods” observed during 12<sup>th</sup>-16<sup>th</sup> of June and 16<sup>th</sup>-18<sup>th</sup> of July (Fig. 1A-B). In particular, the latter period shows very low values of molecular BSOA compounds and a corresponding “sea and ocean” exposure of 79-86%. Hence, “sea and ocean” exposure does not seem to contribute to the measured mass of molecular BSOA compounds. Similarly, the “non-irrigated arable land” contributes to a significant fraction during 16<sup>th</sup>-18<sup>th</sup> of July (8-12%) and most probably does not contribute to the mass of measured BSOA species either.

### 10 3.3 Connection between surface type and measured species

To further investigate the impact of surface types on measured BSOA species a principal component analysis (PCA) was conducted as described in the Methods section. A 4 principal component (PC) VARIMAX-rotated solution was chosen. This solution explained 80.3% of the total variance. Table 7 shows the individual parameter contribution to the respective PC. PC1 accounts for 49.1% of the total variance and has strong positive contributions from several of the monoterpene-derived dicarboxylic acids and both monoterpene- and isoprene-derived OSs and NOSs. The strongest positive surface category in PC1 is “coniferous forest”, suggesting that the species with a bold number in PC1 within Table 7 are originating, or that their mass concentration have a positive response, from coniferous forest. Coniferous forests are mainly known as large-scale emitters of monoterpenes. Despite this, the PCA illustrates that isoprene oxidation products are positively correlated to this surface category. Steinbrecher et al. (1999) observed negligible emissions of isoprene from common conifers as Scots pine (*Pinus sylvestris*) and common juniper (*Juniperus communis*). However, they found significant emissions from Norway spruce (*Picea abies*) which may explain some of the isoprene derived compounds in this study. Although the less strong positive contribution of 0.53, isoprene emitting “broad-leaved forest” may also have contributed to the above described pattern in PC1.

PC2 accounts for 14.9% of the total variation and can roughly be classified as surface categories with low contribution to measured BSOA compounds. Six of the ten investigated surface categories show strong positive contribution to PC2 while many of the measured compounds show low and in some cases negative contribution to PC2. The observed pattern of high “sea and ocean” and “non-irrigated arable land” exposure when the mass concentration of BSOA compounds was low, further strengthens the explanation of PC2.

PC3 accounts for 9.3% of the total variance. The main contributors are suberic acid, azelaic acid and pimelic acid. They are all similar in chemical structure, although suberic and azelaic acid probably originate from fatty acids while pimelic acid likely is of anthropogenic origin (Table 1). Further, azelaic acid has been found to be involved in the triggering of the plant immune system (Jung et al., 2009). Hyder et al. (2012), who also found these three acids to be highly correlated in ambient aerosol, inferred that pimelic acid was either produced from the same source as suberic and azelaic acid or that pimelic acid is produced by continued oxidation of suberic and azelaic acid down to lower carbon numbered acids. None of the land

surface categories displayed high contribution to PC3: “broad-leaved forest” had the highest contribution of 0.21 while the other forest category, “conifer forest”, had a one order of magnitude lower contribution of -0.04.

PC4 accounted for 6.9% of the total variance and is harder to interpret than the previous three PCs. The anthropogenic derived adipic acid has a positive PC contribution (0.59) as well as the surface categories “sparsely vegetated areas” (0.86) and “moors and heath” (0.85). The used land cover maps reveals that both “sparsely vegetated areas” and “moors and heath” are mainly found in Norway and northern Sweden, i.e. in the north and north-westerly direction of Vavihill measurement station. The overall interpretation of PC4 is difficult since adipic acid is thought to be of anthropogenic origin but, in this case, seem to correlate with landscape surface types that are sparsely populated and are associated with low human activity (i.e. “sparsely vegetated areas” and “moors and heath”).

The complexity in PC4 may be caused by the concentration peaks in adipic acid that occurred the 27<sup>th</sup> of June and 6<sup>th</sup> of July (Fig. 1A). During the 27<sup>th</sup> of June, the air mass mainly arrived from the Atlantic Ocean and southern Norway, while the air mass during 6<sup>th</sup> of July mainly originated from the Baltic countries and central Europe (partially illustrated in Fig. 2). Removing the two concentration peaks in adipic acid gave a different PCA solution. Adipic acid now falls into the same PCA as pimelic, suberic and azelaic acid with PC contributions of 0.52, 0.66, 0.70 and 0.73, respectively. Further, the new PC solution show that the aforementioned acids are associated to “pastures” (PC contribution=0.82), “discontinuous urban fabric” (0.84), “non-irrigated arable land” (0.82), “broad-leaved forest” (0.81), “sea and ocean” (0.69) and the “other” category (0.66). Hence, the nature of adipic acid remains unclear since it shows good agreement with the other acids when concentration peaks are removed, implying that adipic is derived from fatty acids or salicylic acid. On the other hand, including the concentration peaks, neither this study nor the study by Hyder et al. (2012) found any strong correlation between adipic and pimelic acid. It can be speculated whether the observed concentration peaks in adipic acid have their explanation in local emission sources of benzene or cyclohexene, followed by a fast oxidation into adipic acid. Future studies should repeat the presented methodology to focus on heavy anthropogenic influenced surface categories (i.e. cities, industries etc.) and their impact on anthropogenic acids and newly discovered anthropogenic OSs (Riva et al., 2016; Riva et al., 2015).

### 3.4 Uncertainties and limits

In this study, our analysis approach relies on two steps: first the calculation of the exposures, using FLEXPART, and then the estimation of land type contributions using a PCA analysis. Both steps suffer from uncertainties which limit the robustness of our results:

The longer the back-trajectories used in FLEXPART, the larger the error is likely to be. On the other hand, shorter back-trajectories lead to neglecting a larger proportion of “older” aerosols. We tested the impact of the footprint length choice on the exposure time series by repeating the analysis with footprints of 3 and 5 days (instead of 7 days in our default setup). Overall, the exposures are not significantly affected, except for the exposure to the “sea and ocean” surface type during the 8-10 July peak, which show an uncertainty of 6% (Fig. S1).

Besides the length of the simulations, a number of FLEXPART settings can impact the results. The size of the aerosols particles has a strong impact on the lifetime of the aerosols in the atmosphere, and therefore on the footprints. We have repeated the experiment with mean aerosol sizes of 50 nm and 1  $\mu\text{m}$ , and the results of the PCAs remained reasonably similar (Table S1 and S2). This is mainly because the PCA is sensitive to correlations, and not to absolute values.

5 The calculation of the observation exposures is based on the assumption that the measured aerosol compositions scale linearly with the aerosol production within the back-plume of the observation. This is not the case in reality: processes such as coagulation, nucleation, chemical reactions between aerosols and surrounding reactive gas species, photo-dissociation and wet and dry deposition (removal of aerosols from the atmosphere by the rain and by gravitational settling) alter the aerosol composition and concentration all along the air mass trajectory. Our approach also ignores the influence of aerosol particles  
10 (or precursors) older than seven days on the observations. Accounting adequately for all these processes would require a comprehensive (much heavier) aerosol model, which is out of the scope of this study. This mainly means that our approach cannot be used to quantify the aerosol production associated to, for example, a specific forest type.

The main limit to the PCA analysis is the shortness of the time series. In particular, there is only one strong event during the campaign (6-8 July), which is not enough for drawing strong conclusions. Our study can however be regarded as a proof-of-  
15 concept: computing FLEXPART footprints is relatively easy and lightweight, and could be performed routinely. The conclusions of a PCA analysis are likely to be a lot more robust with longer time series with more observations included, and/or multi-sites observation campaigns (provided that the footprints of the different sites overlap sufficiently).

#### 4 Conclusions

20 Nine carboxylic acids along with eleven organosulfates (OSs) and two nitrooxy organosulfates (NOSs) were analyzed from 38 daily aerosol samples sampled at Vavihill measurement station in southern Sweden during June and July 2012. Most of the measured compounds can be considered as photo-oxidation products from biogenic volatile organic compounds (BVOCs), hence derived from terrestrial plants. The FLEXPART model was used to identify exposure of the aerosol samples to several different surface categories. For easier interpretation, the study was focused on four potential source-  
25 specific components using 22 chemical species and the nine largest surface categories. The “sea and ocean” category was found to dominate the exposure, and other important categories were “non-irrigated arable land” and “pastures”. A principal component analysis (PCA) of four principal components (PC) was used to explore the impact and connection of surface categories on mass concentration of measured biogenic secondary organic aerosol compounds. It was found that coniferous forest had a positive effect on several of the measured monoterpene-derived compounds. The remaining three PCs were  
30 harder to interpret, however future studies should aim to investigate the sources of azelaic, suberic and pimelic acids which dominate in mass concentration but showed no clear correlation to surface categories.

This study demonstrates the interest of using an atmospheric transport model in aerosol source apportionment on specific chemical compounds. With the presented methodology it is possible to connect single chemical tracer compounds to potential local and long range aerosol sources, i.e. surface categories. More advanced applications may include particle age

estimation and its relation to surface categories; this could be achieved by measuring first- and second-generation BVOC oxidation products and relate these to its measureable gas-phase precursor.

## 5 Data availability

- 5 All data are accessible through the supporting information.

## Author contribution

Johan Martinsson designed the study, compiled all data, performed the PCA and wrote most of the paper. Guillaume Monteil ran the FLEXPART simulations. Moa K. Sporre ran the HYSPLIT simulations. Anne Maria Kaldal Hansen and Marianne  
10 Glasius ran the chemical analysis. Adam Kristensson, Erik Swietlicki and Kristina Eriksson Stenström assisted in the writing process.

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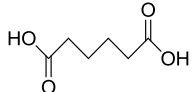
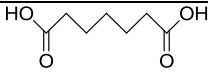
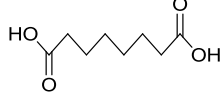
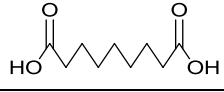
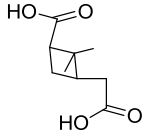
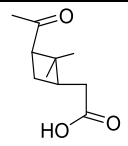
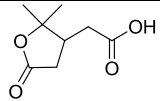
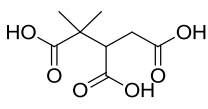
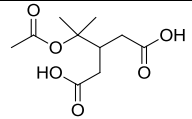
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**Table 1: Analyzed organic acids in the Vavihill aerosol samples. Measured m/z, molecular formula, possible molecular structure, suggested precursor and assigned precursor class. a) Hatakeyama et al. (1987), b) Stephanou and Stratigakis (1993), c) Kawamura and Gagosian (1987), d) Szmigielski et al. (2007), e) Ma et al. (2007), f) Claeys et al. (2009).**

Precursor class	Name	Measured m/z	Molecular formula	Possible structure	Suggested precursor
Anthropogenic	Adipic acid	145.050	C <sub>6</sub> H <sub>10</sub> O <sub>4</sub>		Cyclohexene <sup>a</sup>
	Pimelic acid	159.065	C <sub>7</sub> H <sub>12</sub> O <sub>4</sub>		Cycloheptene <sup>a</sup>
Fatty acid-derived	Suberic acid	173.081	C <sub>8</sub> H <sub>14</sub> O <sub>4</sub>		Unsaturated fatty acid <sup>b,c</sup>
	Azelaic acid	187.097	C <sub>9</sub> H <sub>16</sub> O <sub>4</sub>		Unsaturated fatty acid <sup>b,c</sup>
1. generation Monoterpene	Pinic acid	185.081	C <sub>9</sub> H <sub>14</sub> O <sub>4</sub>		α-/β-pinene <sup>d,e</sup>
	Pinonic acid	183.102	C <sub>10</sub> H <sub>16</sub> O <sub>3</sub>		α-/β-pinene <sup>d,e</sup>
	Terpenylic acid	171.065	C <sub>8</sub> H <sub>12</sub> O <sub>4</sub>		α-pinene <sup>f</sup>
2. generation Monoterpene	3-methyl-1,2,3-butane-tricarboxylic acid (MBTCA)	203.055	C <sub>8</sub> H <sub>12</sub> O <sub>6</sub>		α-pinene <sup>d</sup>
	Diaterpenylic acid acetate (DTAA)	231.086	C <sub>10</sub> H <sub>16</sub> O <sub>6</sub>		α-pinene <sup>f</sup>

5 **Table 2: Analyzed organosulfates (OSs) and nitrooxy organosulfates (NOSs) in the Vavihill aerosol samples. Measured m/z, molecular formula, possible molecular structure, suggested precursor and assigned precursor class. a) Surratt et al. (2007a), b) Schindelka et al. (2013), c) Olson et al. (2011), d) Shalamzari et al. (2013), e) Gomez-Gonzalez et al. (2008), f) Surratt et al. (2008), g) Hettiyadura et al. (2015), h) (Surratt et al. (2010)). The OSs and NOSs were quantified with D-mannose 6-sulfate (1),  $\beta$ -pinene OS 250 (2) or octyl sulfate (3).**

Precursor class	Name	Measured m/z	Molecular formula	Possible structure	Suggested precursor
Isoprene/ Anthropogenic	OS 140 <sup>1</sup>	138.970	C <sub>2</sub> H <sub>4</sub> O <sub>5</sub> S		Glycolaldehyde <sup>a</sup>
	OS 154 <sup>1</sup>	152.985	C <sub>3</sub> H <sub>6</sub> O <sub>5</sub> S		Hydroxyacetone <sup>a</sup> /Methacrolein <sup>b</sup> / Methyl vinyl ketone <sup>b</sup>
	OS 156 <sup>1</sup>	154.961	C <sub>2</sub> H <sub>4</sub> O <sub>6</sub> S		Glycolic acid <sup>c,d</sup> / Methyl vinyl ketone <sup>b</sup>
	OS 170 <sup>1</sup>	168.979	C <sub>3</sub> H <sub>6</sub> O <sub>6</sub> S		Methylglycolic acid <sup>c,d</sup>
	OS 200 <sup>1</sup>	198.991	C <sub>4</sub> H <sub>8</sub> O <sub>7</sub> S		2-methylglyceric acid <sup>a,e</sup>
Isoprene	OS 212 <sup>1</sup>	210.991	C <sub>5</sub> H <sub>8</sub> O <sub>7</sub> S		Isoprene <sup>f,g</sup>
	OS 214 <sup>1</sup>	213.007	C <sub>5</sub> H <sub>10</sub> O <sub>7</sub> S		Isoprene <sup>f</sup>
	OS 216 <sup>1</sup>	215.021	C <sub>5</sub> H <sub>12</sub> O <sub>7</sub> S		C <sub>5</sub> -epoxydiols from isoprene (IEPOX) <sup>h</sup>
Monoterpene	OS 250 <sup>2</sup>	249.080	C <sub>10</sub> H <sub>18</sub> O <sub>5</sub> S		$\alpha$ - $\beta$ -pinene and limonene <sup>f</sup>
	OS 268 <sup>2</sup>	267.053	C <sub>9</sub> H <sub>16</sub> O <sub>7</sub> S		Limonene <sup>f</sup>
	OS 280 <sup>2</sup>	279.054	C <sub>10</sub> H <sub>16</sub> O <sub>7</sub> S		$\alpha$ - $\beta$ -pinene <sup>f</sup>
Monoterpene NOS	NOS 295 <sup>3</sup>	294.062	C <sub>10</sub> H <sub>17</sub> O <sub>7</sub> NS		$\alpha$ - $\beta$ -pinene, Limonene <sup>a,f</sup>
	NOS 297 <sup>2</sup>	296.044	C <sub>9</sub> H <sub>15</sub> O <sub>8</sub> NS		Limonene <sup>f</sup>

**Table 3: Ranges of concentrations, means and standard deviation (SD) of the analyzed compounds in aerosol samples collected at the Vavihill measurement station 10<sup>th</sup> June to 18<sup>th</sup> of July 2012.**

<b>Compound</b>	<b>N</b>	<b>Minimum (ng m<sup>-3</sup>)</b>	<b>Maximum (ng m<sup>-3</sup>)</b>	<b>Mean (ng m<sup>-3</sup>)</b>	<b>±SD (ng m<sup>-3</sup>)</b>
Adipic acid	36	0.03	19.27	1.76	3.87
Pimelic acid	36	0.02	1.21	0.38	0.28
Suberic acid	31	0.05	9.03	2.45	2.42
Azelaic acid	35	0.03	55.27	10.52	13.83
Pinic acid	38	0.28	4.71	1.31	1.04
Pinonic acid	38	0.82	10.66	2.89	2.00
Terpenylic acid	38	0.72	8.86	2.57	1.87
DTAA	38	0.04	5.67	0.84	1.23
MBTCA	38	0.38	29.42	6.18	7.00
OS 140	38	0.02	0.28	0.11	0.07
OS 154	38	0.15	2.95	0.76	0.64
OS 156	32	0.02	2.35	0.65	0.61
OS 170	38	0.08	0.78	0.33	0.17
OS 200	38	0.06	2.02	0.41	0.40
OS 212	38	0.16	4.63	0.91	0.95
OS 214	38	0.06	3.08	0.50	0.58
OS 216	38	0.06	5.83	0.63	1.07
OS 250	38	0.02	3.48	0.51	0.64
OS 268	38	0.01	0.48	0.13	0.12
OS 280	32	0.01	0.70	0.09	0.17
NOS 295	38	0.02	0.53	0.12	0.11
NOS 297	37	0.01	0.18	0.05	0.03

**Table 4: Correlation matrix displaying the Pearson product-moment coefficient (r) for measured chemical species. Colours represent degree of correlation: yellow: [0.7-0.8]; green: [0.8-0.9]; red: [0.9-1.0].**

	Adipic acid	Pimelic acid	Suberic acid	Azelaic acid	Pinic acid	Pinonic acid	Terpenylic acid	DTAA	MBTCA	OS 140	OS 154	OS 156	OS 170	OS 200	OS 212	OS 214	OS 216	OS 250	OS 268	OS 280	NOS 295	NOS 297	
Adipic acid	0.16																						
Pimelic acid	0.02	0.95																					
Suberic acid	0.01	0.87	0.95																				
Azelaic acid	0.25	0.20	0.01	0.20																			
Pinic acid	0.05	0.02	0.32	0.00	0.81																		
Pinonic acid	0.33	0.35	0.18	0.40	0.80	0.39																	
Terpenylic acid	0.35	0.29	0.18	0.37	0.66	0.20	0.89																
DTAA	0.32	0.22	0.06	0.26	0.71	0.29	0.94	0.92															
MBTCA	0.13	0.41	0.27	0.50	0.47	0.06	0.90	0.83	0.70														
OS 140	0.33	0.36	0.22	0.43	0.67	0.19	0.92	0.94	0.93	0.82													
OS 154	0.22	0.36	0.26	0.34	0.62	0.21	0.83	0.87	0.84	0.76	0.92												
OS 156	0.24	0.24	0.00	0.31	0.58	0.21	0.77	0.73	0.80	0.84	0.86	0.83											
OS 170	0.27	0.32	0.19	0.41	0.58	0.10	0.81	0.93	0.86	0.80	0.96	0.93	0.84										
OS 200	0.34	0.35	0.22	0.43	0.65	0.17	0.86	0.97	0.88	0.76	0.97	0.92	0.81	0.98									
OS 212	0.33	0.30	0.18	0.38	0.61	0.15	0.80	0.96	0.82	0.70	0.92	0.89	0.74	0.97	0.98								
OS 214	0.33	0.26	0.21	0.33	0.50	0.06	0.65	0.89	0.68	0.55	0.80	0.79	0.57	0.89	0.91	0.96							
OS 216	0.20	0.00	0.00	0.00	0.48	0.26	0.56	0.45	0.51	0.51	0.54	0.59	0.63	0.53	0.50	0.45	0.31						
OS 250	0.19	0.12	0.00	0.08	0.63	0.36	0.80	0.72	0.87	0.63	0.78	0.67	0.72	0.71	0.69	0.64	0.45	0.55					
OS 268	0.38	0.24	0.15	0.14	0.56	0.08	0.84	0.83	0.93	0.65	0.88	0.77	0.76	0.78	0.82	0.75	0.66	0.39	0.75				
OS 280	0.00	0.00	0.00	0.00	0.56	0.56	0.53	0.38	0.62	0.35	0.44	0.44	0.50	0.32	0.33	0.27	0.09	0.33	0.70	0.55			
NOS 295	0.01	0.14	0.00	0.14	0.53	0.35	0.67	0.57	0.77	0.57	0.68	0.61	0.68	0.59	0.57	0.50	0.31	0.42	0.85	0.70	0.88		
NOS 297																							

**Table 5: Ranges, means and standard deviations (SD) of the FLEXPART surface type exposure of incoming air masses during 10<sup>th</sup> June to 18<sup>th</sup> of July 2012.**

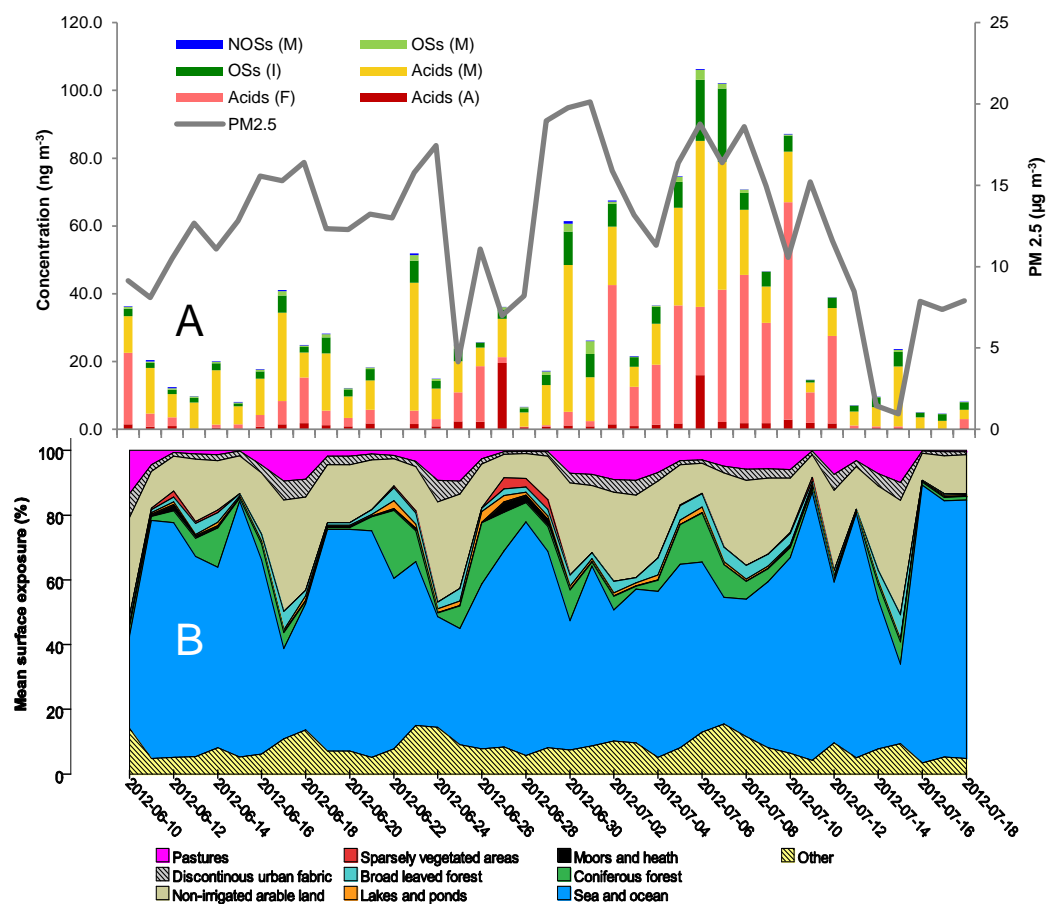
Surface type	N	Minimum (%)	Maximum (%)	Mean (%)	±SD (%)
Pasture	38	0	13	4.4	3.6
Discontinuous urban fabric	38	1	7	2.6	1.7
Non-irrigated arable land	38	7	35	18.8	8.3
Sparsely vegetated areas	38	0	3	0.4	0.9
Broad leaved forest	38	0	8	2.6	1.7
Lakes and ponds	38	0	3	0.9	0.6
Moors and heath	38	0	3	0.5	0.7
Coniferous forest	38	0	22	5.5	5.2
Sea and ocean	38	24.6	86	56.0	16.3
Other	38	3	15	8.3	3.2

**Table 6: Correlation matrix displaying the Pearson product-moment coefficient ( $r$ ) for surface types. Colours represent degree of correlation: yellow:  $[0.7-0.8]$ ; green:  $[0.8-0.9]$ ; red:  $[0.9-1.0]$ .**

	Pasture	Discontinuous urban fabric	Non-irrigated arable land	Sparsely vegetated areas	Broad leaved forest	Lakes and ponds	Moors and heath	Coniferous forest	Sea and ocean	Other
Pasture										
Discontinuous urban fabric	0.92									
Non-irrigated arable land	0.89	0.9								
Sparsely vegetated areas	-0.47	-0.42	-0.49							
Broad leaved forest	0.48	0.32	0.53	-0.13						
Lakes and ponds	0	-0.12	-0.13	0.18	0.2					
Moors and heath	-0.46	-0.4	-0.47	0.98	-0.17	0.14				
Coniferous forest	-0.17	-0.31	-0.22	0.23	0.43	0.8	0.17			
Sea and ocean	-0.84	-0.78	-0.84	0.27	-0.73	-0.31	0.28	-0.28		
Other	0.59	0.57	0.53	-0.16	0.42	0.29	-0.18	0.23	-0.77	

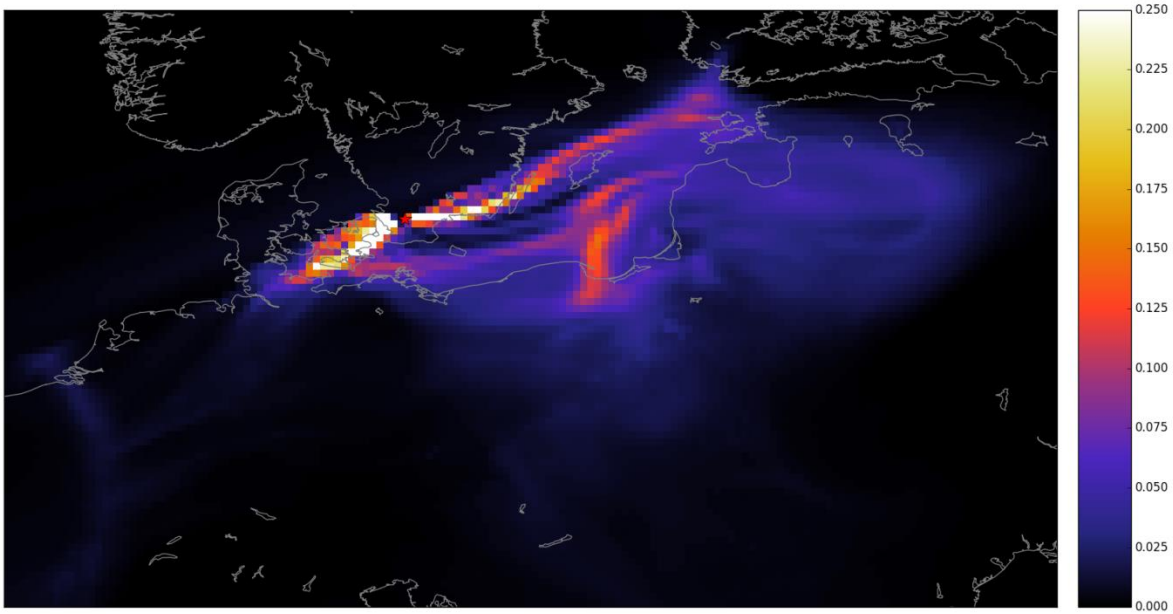
Table 7: Principal component (PC) loadings. The loadings display the variation (between -1 and 1) explained by the PC. Numbers in bold indicates absolute number >0.6. PC1 explained 49.1%, PC2 14.9%, PC3 9.3% and PC4 6.9%.

	Principal Component			
	1	2	3	4
Adipic acid	0.37	-0.25	0.08	0.59
Pimelic acid	0.24	0.20	<b>0.75</b>	-0.21
Suberic acid	0.20	0.26	<b>0.82</b>	-0.19
Azelaic acid	0.21	0.39	<b>0.74</b>	-0.17
Pinic acid	<b>0.70</b>	-0.04	-0.25	0.14
Pinonic acid	0.19	-0.15	-0.37	0.16
Terpenylic acid	<b>0.88</b>	0.29	-0.11	0.04
DTAA	<b>0.93</b>	0.24	0.04	0.09
MBTCA	<b>0.89</b>	0.28	-0.26	-0.02
OS 140	<b>0.76</b>	0.30	0.12	-0.41
OS 154	<b>0.96</b>	0.22	0.04	-0.10
OS 156	<b>0.93</b>	0.06	0.04	-0.14
OS 170	<b>0.79</b>	0.20	-0.17	-0.28
OS 200	<b>0.92</b>	0.18	0.10	-0.12
OS 212	<b>0.95</b>	0.18	0.13	-0.01
OS 214	<b>0.92</b>	0.13	0.15	0.04
OS 216	<b>0.87</b>	0.03	0.26	0.11
OS 250	0.48	-0.06	-0.38	-0.06
OS 268	<b>0.67</b>	0.24	-0.51	-0.18
OS 280	<b>0.87</b>	0.13	-0.20	-0.05
NOS 295	0.43	0.16	<b>-0.69</b>	-0.25
NOS 297	0.59	0.28	-0.48	-0.35
Pastures	0.22	<b>0.85</b>	0.15	-0.37
Discontinuous urban fabric	-0.02	<b>0.92</b>	0.12	-0.29
Non-irrigated arable land	0.20	<b>0.94</b>	0.10	-0.14
Broad leaved forest	0.53	<b>0.77</b>	0.21	0.11
Sparsely vegetated areas	-0.11	-0.10	-0.18	<b>0.86</b>
Lakes and ponds	<b>0.76</b>	0.34	0.02	0.42
Moors and heath	-0.16	-0.04	-0.23	<b>0.85</b>
Coniferous forest	<b>0.79</b>	0.35	-0.04	0.39
Sea and ocean	0.37	<b>0.62</b>	0.27	0.34
Other	<b>0.60</b>	<b>0.65</b>	0.19	0.19



5 **Figure 1:** A) Total concentration of all measured carboxylic acids, organosulfates (OSs) and nitrooxy organosulfates (NOSs) in  $PM_1$  collected at the Vavihill measurement station. The thick grey line displays the  $PM_{2.5}$  concentration. Capital letters in parenthesis in the legend is the precursor class given in Table 1 and 2. A=Anthropogenic, F=Fatty acid, I=Isoprene and M=Monoterpenes. B) FLEXPART generated mean exposure from the nine mean largest surface categories. The exposure is a mean of 3, 5 and 7 days back trajectories. The category “Other” represents the remaining 34 surface categories. More detailed information on the surface categories can be found in the supplemental information.





5 **Figure 2: 120 hour back-trajectory air mass covering the concentration peak dates; 6-8<sup>th</sup> July. FLEXPART are shown in shaded colors. The colorbar displays the FLEXPART footprint, normalized to 1 (the color range has been limited to 0-0.3 to highlight grid points with low but a non-zero contribution). Together, the grid points with a value larger than 0.1 contribute 17% of the total sensitivity while grid boxes with a value larger than 0.01 contribute 81% of the total sensitivity. 120 h back-trajectory was chosen for easier interpretation of the illustration.**