Assessment of wood burning versus fossil fuel contribution to wintertime black carbon and carbon monoxide concentrations in Athens, Greece

Authors Responses to referees' comments

The authors acknowledge both referees for critically reading the manuscript and for their contribution in improving and clarifying this study. The authors have compiled the responses as follows. Reviews from Referee 1 and Referee 2 are in blue and green font respectively, and have been grouped based on the section of the manuscript they refer to. Author responses are in black font numbered with [A0, A1, A2 ...]. Italics and quotations are used for the information added in the revised manuscript.

General Comments

- Reviewer 1: This manuscript addresses up-to-date scientific questions within the scope of the journal, and may indeed be considered as relevant for the special issue dedicated to ChArMEx. Its overall presentation (including the title, the abstract and the figures) is appropriate, clear and globally well structured. It presents results of wintertime BC and CO source apportionment results obtained for Athens, Greece. To do so, authors claim they are using two different and independent methodologies. However, the "CO/NOx ratio" approach appears to be irrelevant in the present case, so that outputs are not used for the purpose of the study. My only major comment is related to this latter issue, and I would recommend presenting the use of the "CO/NOx ratio" approach in another way (or to simply skip it).
- Referee 2: The manuscript attributes the concentration-contribution of wood burning to air pollution in Athens and with that handles an important health related subject that needs attention. For the source apportionment of black carbon, the well-known technique based on wavelength dependence of the aerosol light absorption coefficient is used after application of necessary compensations. For CO apportionment two models are described. Model 1 relies on known emission ratios of NOx and CO and assumed similar atmospheric lifetimes. This model is not trusted by the authors and according to the authors this linear model always over-estimates the wood burning contribution. The results from model 1 are not researched enough to give recommendations to other scientific studies. It remains unclear why this model is included in this paper. A short discussion in the introduction could be enough.
- A0: Both referees major comment concerns the use and presentation of the CO-NOx linear model used for CO source apportionment. From their comments, we understand that the reason why the model was used seemed unclear since at the end, we give more confidence to the second model.

However, we believe that since the CO/NOx ratio has been used in the past as a diagnostic ratio to characterize different type of emission sources (Fujita et al., 1992; Ravindra et al., 2006; Wahlina et al., 2001), it is interesting to discuss the output of the model 1 with respect to the output of the model 2. Additional, this model could be of use in the case where no absorption measurements are available. This could be for instance the case at monitoring stations part of national networks where only regulated air pollutants are often measured. In order to make a more clear introduction to the reason why we have selected to use this model, we have added in the manuscript, section 2.4.1, the following: "*The CO/NOx ratio has been used in the past as a diagnostic to characterize different type of emission sources* (Fujita et al., 1992; Ravindra et al., 2006; Saurer et al., 2009; Wahlina et al., 2001). *It can serve as a useful tool for apportioning CO concentrations at monitoring stations part of national networks where only regulated air pollutants are often measured.*"

We have also added some discussion about the conditions limiting the use of the model. Indeed, our results suggest that CO-NOx should be used with more caution in environments dominated by aged air masses. As a matter of fact, at the NOA urban background site, the difference between both models is of a factor of about 1.5, which can be considered rather acceptable given the level of uncertainty associated with source apportionment methods. Higher difference between both models are observed for the DEM suburban site. Finally, since this type of model is being used in other studies, we suggest that it is a benefit for future studies to keep the results of this comparative assessment against model 2 in order to have documented the results of the current evaluation.

The following discussion was added in the manuscript (P9L36): "The wood burning contribution to CO estimated using the CO-NO_x linear model is higher compared to the estimations from CO- BC_{ff} - BC_{wb} model, of a factor of about 1.5 for NOA, which can be considered as "acceptable" given the uncertainties associated with both models. A higher overestimation compared to model 2 of a factor of about 2.4 is found at DEM suburban station, which could be explained by the fact

that the site is characterized by more aged air masses compared to the urban background NOA station. Consequently, ambient CO/NO_x ratios might differ more significantly from emission ratios at DEM suburban station. These results suggest that CO-NO_x linear model probably overestimate wood burning contribution to CO, especially in environments characterized by aged air masses were photochemical loss of NOx cannot be considered as negligible.

- Model 2 for CO apportionment is based on multiple linear regression between CO and BC. This technique is well-known but in the conclusion section it is written that the method is new; or is the application to CO in combination to wood burning new? If new this fact should receive more attention earlier in the paper. The manuscript is not pushing scientific boundaries, but contains important numbers, e.g. 30% of BC is wood burning related.
 - We have taken into account the referee comments and added the following in the revised manuscript, in order to address what is the novelty about the method used (P6L36). "While multiple linear regressions are known techniques for source apportionment, they have not yet been applied to investigate wood burning contribution to CO using the aethalometer's model results. The considerable increase in measurements carried out using aethalometers makes this technique an interesting methodology for apportioning CO concentrations."

Specific comments

Abstract

- line 21: occasional significant impacts of long-range transport are not really discussed/demonstrated in the manuscript.
 - A1: The sentence "and are only occasionally affected by long-range transport" has been removed from the manuscript.
- 30% to the observed eBC and the CO concentrations (: : :.)' this doesn't read well and is confusing for CO.
 - A2: Corrected in the revised manuscript (P1L23).

1 Introduction

- Page 2, lines 16-23: it is not clear within which periods the discussed increases/decreases were observed (e.g., lines 16-18: a constant increase of 30% every year since 2012 ? or an increase of 30% for the period 2012-201x, compared to which period ? : : :).
 - A3: Precisions on the time periods discussed have added in the revised manuscript.

2 Material and Methods

2.1 Sampling Sites

- Page 3, line 9: how much "relatively far" from major roads?
 - A4: The station is at about 500 m from major roads (this information was added in the revised manuscript). Within this radius of 500 m around the site, there are mainly pedestrian streets parks, hills. This is the reason why the site can be considered as an <u>urban background</u> and not an urban traffic site.

2.2 Measurements of aerosol light absorption and carbon monoxide

2.2.1 Aerosol light absorption and equivalent black carbon

- Page 3, lines 18-26: it may be worthy to indicate more clearly this data correction procedure was applied to AE42 (and AE31) datasets only.
 - A5: Added in the revised manuscript
- Also, what could be the uncertainties related to the use of f_values that weren't estimated for this individual instrument /site?
- For compensation parameter f values given in Drinovec et al. (2015) were used. But Drinovec et al. describe that filter loading effects change with location and time. The values of Drinovec differ a lot from Sandradewi et al (2008) (reference in manuscript but not listed in References, make sure to find the correct paper) and Zotter et al (2017). The latter paper, that is known to the authors confirms the importance of proper compensation. I would like to see a more worked out compensation correction.
 - A6: We agree with both reviewers that *f* parameter depend on location and time and that uncertainties related to the use of *f* values should be estimated. *f* compensation parameter is expressed as: $f = a^{*}(1 w0) + 1$, with w0, the aerosol single scattering albedo and *a*, a constant parameter varying in the range 0.82–0.88 for the different wavelengths (950–370 nm). As a result, *f* depends mainly on the single scattering albedo. Because no simultaneous measurements of SSA were available, we chose

default values based on Drinovec et al.(2015) because they were estimated for an urban environment where SSA is expected to be lower than background or remote environments characterized by aged aerosols. The f-values given in Drinovec et al., (2015) correspond to SSA values of about 0.75. Later measurements during the Athens smog ACTRIS JRA1 campaign indicated that wintertime SSA value at NOA exhibits an average value of 0.8 ± 0.05 . In order to estimate uncertainties related to the chosen f values, absorption coefficients calculated with f values taken from Drinovec et al., (2015) were compared with those using f values calculated for w0=0.8 (see Figure 1). Differences are found to be lower than 1%. Therefore, we estimate that on average the shadowing effect was correctly accounted for and therefore did not change the values in the manuscript. The following discussion has been added on the revised manuscript: "The compensation parameter f_{λ} is a parameter that mainly depends on the single scattering albedo of aerosol (SSA). Because no simultaneous scattering coefficient measurements were available, f_A values given in Drinovec et al. (2015) for an urban site and characteristic of a single scattering albedo of about 0.75 were used for this purpose."



Figure 1 Scatter plot between absorption calculated using f_values of 1.17 (ssa=0.8) and 1.203 (Drinovec et al.2015) in the shadowing effect correction algorithm.

- Finally, what could be the impact of the PM10 cut-off, compared to the PM2.5 used at the other site?
- A7: At NOA, TSP were collected and not PM10 as indicated in the original version of the manuscript. This information has been corrected in the revised manuscript. As the inlet includes curved tubing, a significant aerosol loss of the coarse fraction is expected. However, as indicated in several studies, BC mainly contributes to PM1 (Laborde et al., 2013; Wang et al., 2015), therefore differences in the eBC concentrations due to the different aerodynamic diameters of sampled aerosols are expected to be negligible. (added in the revised manuscript to support the fact that BC is mainly related to fine particles)
- About the instruments At DEM (AE33) at NOA(AE42). I understand that AE33 and AE31 are compared. At page 3 line29: 'and data from AE31 aethalometer with : : :: : to AE-42, which operates continuously in parallel with the AE33 at DEM station. So what instruments are compared and where is the AE-31 located?
 - A8: Since we could not compare directly AE33 and AE42, we decided to compare AE33 with an AE31-Aethalometer (with identical measurement settings to the AE42). Both instruments were running simultaneously from 1st August to 30th September 2014 at NCSR Demokritos station. We modified the following sentence for more clarity: "The results indicated a very good agreement between the absorption measurements (Mm⁻¹) from the AE33 and AE31 instruments after compensation, with R²=0.79, a small intercept of -0.15 and a slope of 0.97."
- The R-squared of 0.79 is not very convincing for aethalometers, I would like to see the plot. The intercept (what is the unit? Inverse Mm or ng/m3?) of 0.15 is interesting.
 - A9: The plot is shown below. The unit is Mm⁻¹ and this information has been added in the revised manuscript. The intercept is -0.15.



Figure 2 Linear regression between absorption coefficients at 880 nm measured simultaneously by AE33 and AE31 and corrected from loading and scattering effects using the dual Spot Technolgy and the Weingartner procedure respectively, from 1st August to 30th September 2014 at NCSR Demokritos station.

- Page 4, line 6-7: please indicate whether this value was also obtained using the 1.64 "ACTRIS correction factor" (as used by Zanatta et al., 2016)?
 - A10: As suggested by the reviewer, in the revised version of this manuscript the default C-factor for both aethalometers has been corrected with an additional correction factor. The need to use a compensation factor on top of the default value has also been confirmed by parallel measurements performed between AE31 and AE33 and a multiangle absorption photometer (MAAP) (Model 5012, Thermo Electron Group,Waltham, MA, USA) during 2011 at the DEM station (Diapouli et al., 2017). Absorption measurements have therefore been corrected with a factor of 1.64 as used in the ACTRIS community and proposed by the reviewer. This additional correction of our absorption measurements. The new corrected MAC at 880 nm value is therefore 4.6. However, correction on absorption coefficient and MAC compensate one another, and therefore final eBC values remain unchanged. The following changes have been made on the revised manuscript:

"The value of 3.5 was used for C_0 as recommended in Zanatta et al., (2016)......"As a result, the eBC mass concentration was derived in this study by multiplying the b_{abs} coefficient at 880 nm with a constant value of mass absorption cross-section (MAC) of 4.6 m² g⁻¹ (determined from the comparison with simultaneous measurements at DEM of elemental carbon)"

"Assuming an absorption Ångstrom exponent of 1.0, the MAC used here is 6.13 when adjusted to 637 nm. Our MAC value is at the lower limit of the values reported by Zanatta et al., (2016), for nine rural background stations across Europe (7.5-13.3 m2 g-1, calculated for 637 nm), and within the range of values reported by Hitzenberger et al., (2006) for an urban background site in Vienna (5.9 -7.5 at 637 nm)."

- I'm not sure, if it is interesting that 'BC is historically defined from aethalometer measurements at 880nm'. The important message should be that the whole spectral de- pendence approach depends on fixation somewhere. This is done at 880nm because it is believed that at that wavelength the MAC for wood burning and fossil fuel combustion is very close. Otherwise the DEC MAC cannot be applied at NOA. The whole fractitioning is based on the wavelength dependence that is somewhere fixed (Equation 10). The reader should be convinced of the choice that is supported by literature.
 - A11: The following sentence was added in the revised manuscript: "At 880 nm, no significant difference in MAC at 880 nm between eBC originating from traffic or woodburning emission is expected (Zotter et al., 2017)"

- It is written '(MAC): : :. (determined from the comparison with concurrent measurements at DEM of elemental carbon)'. A bit later a reference to Diapouli et al. (2014) is included. Does this paper include the 7.5 m2 g-1? .
 - A12: The paper of Diapouli et al. (2014) does not include absorption measurements, therefore no MAC value is presented. As indicated in the manuscript, the paper of Diapouli et al. (2014) includes an extensive description of EC/OC measurements at DEM.
- The angstrom exponent for absorption is measured why do the authors assume an exponent of 1.0 in line 6 (p4)?
 - A13: As the reviewer mentions the absorption exponent is measured for both sites and exhibits spatial and temporal variability, with an average value superior to 1. However, we decided to use an exponent of 1 to calculate the MAC value to 637 nm for a better intercomparison with Zannata et al. 2016 paper, where MAC values from different sites were adjusted to 637 nm assuming an absorption Ångstrom exponent of 1.0.

2.2.2 Carbon monoxide and nitrogen oxides

- Page 4-5, BC and CO source apportionment: please discuss here possible interference from coal combustion emissions
 - A14: Added in the manuscript. "It should also be noted that coal-burning organic aerosol is known to significantly absorb light at near UV wavelengths (Yang et al., 2009) and may thus interfere with b_{abs}(λ_{UV})_{wb}. Lignite coal is the single most important local energy source in Greece (Kavouridis, 2008). However, interferences from coal use are expected to be very low, as the lignite-fired power plants are located far away from Athens (>200 km distance)."

2.3 Source apportionment of black carbon from fossil fuel and wood burning combustion.

- P4 line21 Reference to Sandradewi (please include the correct one in references).
 - A15: Added
- Page 5, BC source apportionment: please justify/discuss a bit more the choice of _wb = 2 by comparison with value recently proposed by Zotter et al., 2017.
- Sandradewi discussed different Angstrom exponents depending on the chosen wavelengths. This wavelength dependence should be discussed in light of the choices given in line 32, or refer to other studies that use same wavelengths. The 470 nm channel was broken in that Sandradewi study, why does this study start at 470 nm (line 32) P5 top para. Exponents 0.9 (traffic) and 2.0 (wood) 'were used, based on the range of values.. reported'. The value of 2.0 is disqualified by Zotter et al., 2017, because is leads to differences with radiocarbon results. The exponents are crucial to the method, 'based on' should be worked out.
 - A16: In a previous study using the aethalometer model at DEM (Diapouli et al., 2017), 0 calculations were made for different values of a_{wb}, in the range 1.1-3.0, by a step of 0.1. In order to identify an acceptable range of values for awb, the calculated babs(950)ff were correlated with NOx data, which are mainly related to fossil fuel combustion emissions. Values of a_{wb} below 1.7 produced either no correlation or weak correlations and were therefore not considered acceptable values (Pearson coefficients below 0.7). On top of that, during fire events, values of angstrom exponent up to 2 have been observed at DEM (Figure 3). We expect therefore that angstrom exponent from biomass burning to be at least as high as 2. In view of these results, awb=2 has been selected for the study. Discussion about the choice of awb has been added in the revised manuscript: "The application of the model requires the selection of suitable Ångström exponents for fossil fuel (α_{ff}) and wood burning (α_{wb}), since one of the greatest uncertainties of the model is associated with the a priori assumed α values for both types of emissions. Reported Ångström exponents range between 0.8-1.1 for pure traffic. For wood burning a wider range of values has been observed (0.9-3.5), even though awb equal to 2.0 has long been considered a typical value for wood burning aerosol (Favez et al., 2009; Fuller et al., 2014; Herich et al., 2011; Petit et al., 2014; Sciare et al., 2011). Recently, Zotter et al., (2017) recommended to use α_{ff} =0.9 and α_{wb} =1.68, obtained by fitting the model outputs (calculated with the absorption coefficients at 470 and 950 nm) against the fossil fraction of EC derived from 14C measurements. At DEM site, a previous study showed that values of awb below 1.7 were not appropriate for the specific site (Diapouli et al., 2017). On top of that, during fire events, Ångström exponent values up to 2 have been observed at DEM. Taking into consideration these results, absorption Ångström exponents of 0.9 and 2.0 for pure traffic (α_{ff}) and wood burning (α_{wb}), respectively, were used in this study"
 - "Even though different pairs of near-UV and near-IR wavelengths can be used, it is recommended to use the pair 470 nm versus 950 nm. The choice of 470 against 370 is

even more critical as explained in Zotter et al (2017) since interference of VOCs or other absorbing non-BC particles can interfere with measurements with the 370 nm channel of aethalometer."



Figure 3 eBC and angstrom exponent measured at DEM station during summer forest fires.

2.4 Source apportionment of carbon monoxide from fossil fuel and wood burning combustion 2.4.2 Model 2: CO BCwb BCff multiple linear regression model 6

- P6 line 20 'the hypothesis of negligible photochemical chemistry is validated.' Where is it validated please include reference(s). Negligible what does that mean, negligible for the scale considered in this study? Or is the assumption that BC and CO have similar lifetimes? This para needs to be worked out to convince the reader that model2 is superior to model 1.
 - A17 "The sentence was corrected to "the hypothesis of negligible photochemical chemistry is met for the time scale considered in this study"

3. Results and Discussion

- 3.1 Levels and diurnal variations of black carbon and carbon monoxide
 - Page 6, line 36 Page 7, line 7: the expected diurnal cycle of the intensity of emissions could be discussed more deeply here.
 - A18: Added in the manuscript: Based on traffic volume data (Grivas et al., 2012), a first peak in the emissions from transportation is expected around 08:00 when people commute to work, followed by a plateau from 08:00-18:00, and a secondary peak until 21:00, after when traffic is decreasing. Wood burning emissions from residential heating are expected to increase during the evening, when temperatures drop and people are back-home.

7

3.2 Source apportionment of BC and diurnal variability

- Page 7 line 9 'eBC in PM2.5' but at NOA a PM10 sampling head is installed, right?
 - A19: This information has been corrected on the revised manuscript.
- Page7 line 11. Apparently the 1.25 percentile of a dataset can be used for background. Really would like to read that paper. Please include Kondo et al., 2006 in the references
 - A20: We have added the reference to the paper of (Kondo et al., 2006), that investigated temporal variations of elemental carbon in Tokyo. The same approach was used in (Verma et al., 2010) for determining the BC background concentration of black carbon in Guangzhou, China.

- P7L14 'relatively short lifetime of BC' please compare to P6L20
 - A21: We agree that there was a contradiction between statement in P7L14 and P6L20. In P6L20 we wanted to point out that both CO and BC are not chemically reactive, whereas in P7L14 we were referring to deposition losses. In order to avoid any misunderstanding we removed the sentence "*relative short lifetime*" in p7 I14.
- P7L34 how is the relative standard deviation defined in this case.
 - A22: It is defined as the interval of confidence in the coefficient values (slope and intercept) of the linear regression as calculated by Igor Pro software. The values are automatically calculated for each fitting.

7

- P7L39 '0.184' please include units if appropriate
 - A23: Added

3.3.1 Using BCwb and BCff as tracers of fossil fuel and wood burning sources

- P8L4 the value 0.00137 is 0.7% of the best estimate 0.184. This is very small compared to values in Table 2. Please include discussion.
 - A24: 0.00137 is the standard deviation of the rff coefficient given from the multiple regression fitting for DEM station. This is very small compared to the range of values found in the literature (and given in Table 2) that have been calculated using different methodologies and for different type of transport fleet (size distribution and age of vehicle fleet, fuel consumption, environmental performance etc.). We assume that DEM and NOA experience similar vehicle fleet mix, and this is the reason why, the regression model was run for NOA using a fixed r'ff according to values found in DEM station (and not based on the literature). As a result, the error estimation of rff was made based on the standard deviation of the model for DEM, and not on values from the literature.
 - Why is 0.00137 an useable value and why is the resulting uncertainty of 25% for the emission ratio 'rather reasonable' (not scientific terminology)
 - A25: 0.00137 was the uncertainty calculated for DEM station. When fixing r'ff at NOA, we did a sensitivity analysis using r'ff values ranging from the lowest (0.184-0.00137) to the highest (0.184+0.00137) around the determined *r'ff*, and the resulting r'wb variated as much as 25%. We corrected the uncertainty in eq.21 and removed the statement "can be considered rather reasonable" from the revised manuscript.
- P8L13 background or intercept values are 109 and 147 how do these differences related to 'cannot differ significantly' line 1 of this page?
 - A26: Considering a uniform/similar vehicle fleet in terms of type of vehicle and driving patterns, the <u>emission ratio</u> between BC/CO from traffic should not differ. However, absolute <u>concentrations</u> of CO vary between the urban and suburban sites.
- P8L14 background concentrations of CO with a reference to Goldstein and Schade (2000), this work contains some informations on background but not on CO. Howshould the reader interpret the reference, please modify.
 - A27: The reference was about background estimation using regression's intercept values. In order to avoid any confusion, we removed the reference
- P8L14etc The resulting background concentration are in very good agreement with : : :1.25 percentile. Really want to learn more. For me it sounds like abracadabra.
 - A28: The good-agreement between this value based on the literature and our results from the multiple regression models further supports the use of the 1.25 percentile for the calculation of the background concentrations.
- Page 8, line 25 (% COwb): please discuss these percentage regarding previous studies/results. A29: While there have been numerous studies in the last years investigating the contribution of different sources to black carbon surface concentrations, similar studies are very limited for carbon monoxide. Saurer et al., 2009 used the stable isotope composition of CO (δ¹³C and δ¹⁸O) for the characterization of different CO sources at 3 sites in Switzerland during winter (along with other indicators for traffic and wood combustion such as NO_x-concentration and aerosol light absorption at different wavelengths) and estimated the wood burning contribution to night-time CO concentrations at 70%, 49% and 29% for a village site dominated by domestic heating, a site close to a motorway and a rural site respectively. These differences reflect the spatial variability in the wood burning use within the same region depending on the type of site, as well as between

countries depending on the country and the heating practices. This discussion is added in the revised manuscript.

3.3.2 Comparison the CO BCwb BCff linear model vs the CO NOx linear model

- P8L41 'using a best fit line' If this is a fit how was the data selected? This was not clear from the references literature.
 - A30: In the reference literature there are no explanation on the methodology used to draw these two slopes. In this study, in order to o draw the minimum and maximum slopes, the 10th percentile and 90th percentile of (CO-CObgd)/NOx ratio have been calculated. To draw the minimum slope, fitting was applied for data where CO/NOx ratio was below the calculated 10th percentile. To draw the maximum slope, fitting was applied to data where CO/NOX ratio was above the calculated 90th percentile. However, these fitted lines are just indicative of the expected range of values of CO/NOX ratios for each emission source. This information has been added in the revised manuscript.

8

- P9Line4 informs us that the ratio is larger than : : :please explain
 - A31: Wood burning lines from figure 11, exhibit slopes of 20 and 25. However we do not expect to have 100% contribution of wood burning at any time of the day. We can therefore estimate that r_{wb} is superior to both these values. Modifications in the manuscript: "Nevertheless, based on "wood burning" lines from Fig. 11, and assuming that emission ratios from wood burning are similar between NOA and DEM, we estimate a r_{wb} ratio for the area of Athens, larger than 25 ppbv ppbv 1."
- P9L9 'values found in the literature' please include references
 - A32: We removed the sentence 'values found in the literature'. The references are given in the following sentence.
- P9L10 2-3% where should I look to see the supporting material?
 - A33: An additional column in Table 4 with the COwb% resulting from the sensitivity analysis test has been added in the revised manuscript.

4. Conclusion

- Page 9, line 25-26: here, it sounds like wind speed is controlling the diurnal patterns. Please consider rephrasing this sentence.
 - A34: Modified in the manuscript: "Both BC and CO displayed a clear bimodal diurnal pattern, in which morning peaks were observed due to morning inversion and rush-hour traffic, while evening peaks were attributed to combustion sources (evening traffic rushhour, residential heating) combined with the effects of a shallow nocturnal boundary layer. Highest concentrations were observed during low wind speeds, suggesting that both combustion products were not related to regional transport but instead originated from sources within Athens."

Acknowledgements References

- Figure axes: please homogenize the use of "BC" / "eBC".
 A35: Corrected
- Figure 7, right panel: legend of the y-axis seems inaccurate
 - A36: Corrected
- Table 4 Regression Slope between model 1 and model 2: what model outcomes are regressed? Are we looking at COwb/totalCO?
 - o A37: Yes, this information is added in the Table
- Typos-suggestions P3 line26 'this purpose' ! for loading compensation (corrected), P4 L9 ratios were (corrected), P4 line 25 lambda is bold (corrected) in equation P4 eq 5 lambda1 should be lambda2 in Denominator (corrected). P7L40 last ff should be sub (corrected), P8L28 diurnal variabilities : : :are (corrected), Comparison of A and B Figure 7 caption or axis titles are wrong for right bottom figure (corrected)

Diapouli, E., Kalogridis, A.-C., Markantonaki, C., Vratolis, S., Fetfatzis, P., Colombi, C., and Eleftheriadis, K. (2017). Annual Variability of Black Carbon Concentrations Originating from Biomass and Fossil Fuel Combustion for the Suburban Aerosol in Athens, Greece. Atmosphere *8*, 234.

Drinovec, L., Močnik, G., Zotter, P., Prévôt, A.S.H., Ruckstuhl, C., Coz, E., Rupakheti, M., Sciare, J., Müller, T., Wiedensohler, A., et al. (2015). The "dual-spot" Aethalometer: an improved measurement of aerosol black carbon with real-time loading compensation. Atmos Meas Tech *8*, 1965–1979.

Favez, O., Cachier, H., Sciare, J., Sarda-Estève, R., and Martinon, L. (2009). Evidence for a significant contribution of wood burning aerosols to PM2.5 during the winter season in Paris, France. Atmos. Environ. *43*, 3640–3644.

Fujita, E.M., Croes, B.E., Bennett, C.L., Lawson, D.R., Lurmann, F.W., and Main, H.H. (1992). Comparison of Emission Inventory and Ambient Concentration Ratios of CO, NMOG, and NOx in California's South Coast Air Basin. J. Air Waste Manag. Assoc. *42*, 264–276.

Fuller, G.W., Tremper, A.H., Baker, T.D., Yttri, K.E., and Butterfield, D. (2014). Contribution of wood burning to PM10 in London. Atmos. Environ. *87*, 87–94.

Grivas, G., Cheristanidis, S., and Chaloulakou, A. (2012). Elemental and organic carbon in the urban environment of Athens. Seasonal and diurnal variations and estimates of secondary organic carbon. Sci. Total Environ. *414*, 535–545.

Herich, H., Hueglin, C., and Buchmann, B. (2011). A 2.5 year's source apportionment study of black carbon from wood burning and fossil fuel combustion at urban and rural sites in Switzerland. Atmospheric Meas. Tech. *4*, 1409–1420.

Kavouridis, K. (2008). Lignite industry in Greece within a world context: Mining, energy supply and environment. Energy Policy *36*, 1257–1272.

Kondo, Y., Komazaki, Y., Miyazaki, Y., Moteki, N., Takegawa, N., Kodama, D., Deguchi, S., Nogami, M., Fukuda, M., Miyakawa, T., et al. (2006). Temporal variations of elemental carbon in Tokyo. J. Geophys. Res. Atmospheres *111*, D12205.

Laborde, M., Crippa, M., Tritscher, T., Jurányi, Z., Decarlo, P.F., Temime-Roussel, B., Marchand, N., Eckhardt, S., Stohl, A., Baltensperger, U., et al. (2013). Black carbon physical properties and mixing state in the European megacity Paris. Atmos Chem Phys *13*, 5831–5856.

Petit, J.-E., Favez, O., Sciare, J., Canonaco, F., Croteau, P., Močnik, G., Jayne, J., Worsnop, D., and Leoz-Garziandia, E. (2014). Submicron aerosol source apportionment of wintertime pollution in Paris, France by double positive matrix factorization (PMF2) using an aerosol chemical speciation monitor (ACSM) and a multi-wavelength Aethalometer. Atmos Chem Phys *14*, 13773–13787.

Ravindra, K., Wauters, E., Tyagi, S.K., Mor, S., and Grieken, R.V. (2006). Assessment of Air Quality After the Implementation of Compressed Natural Gas (CNG) as Fuel in Public Transport in Delhi, India. Environ. Monit. Assess. *115*, 405–417.

Saurer, M., Prévôt, A.S.H., Dommen, J., Sandradewi, J., Baltensperger, U., and Siegwolf, R.T.W. (2009). The influence of traffic and wood combustion on the stable isotopic composition of carbon monoxide. Atmospheric Chem. Phys. *9*, 3147–3161.

Sciare, J., d'Argouges, O., Sarda-Estève, R., Gaimoz, C., Dolgorouky, C., Bonnaire, N., Favez, O., Bonsang, B., and Gros, V. (2011). Large contribution of water-insoluble secondary organic aerosols in the region of Paris (France) during wintertime. J. Geophys. Res. Atmospheres *116*, D22203.

Verma, R.L., Sahu, L.K., Kondo, Y., Takegawa, N., Han, S., Jung, J.S., Kim, Y.J., Fan, S., Sugimoto, N., Shammaa, M.H., et al. (2010). Temporal variations of black carbon in Guangzhou, China, in summer 2006. Atmos Chem Phys *10*, 6471–6485.

Wahlina, P., Palmgren, F., and Van Dingenen, R. (2001). Experimental studies of ultrafine particles in streets and the relationship to traffic. Atmos. Environ. *35*, S63–S69.

Wang, Q., Liu, S., Zhou, Y., Cao, J., Han, Y., Ni, H., Zhang, N., and Huang, R. (2015). Characteristics of Black Carbon Aerosol during the Chinese Lunar Year and Weekdays in Xi'an, China. Atmosphere *6*, 195–208.

Yang, M., Howell, S.G., Zhuang, J., and Huebert, B.J. (2009). Attribution of aerosol light absorption to black carbon, brown carbon, and dust in China – interpretations of atmospheric measurements during EAST-AIRE. Atmos Chem Phys *9*, 2035–2050.

Zanatta, M., Gysel, M., Bukowiecki, N., Müller, T., Weingartner, E., Areskoug, H., Fiebig, M., Yttri, K.E., Mihalopoulos, N., Kouvarakis, G., et al. (2016). A European aerosol phenomenology-5: Climatology of black carbon optical properties at 9 regional background sites across Europe. Atmos. Environ. *145*, 346–364.

Zotter, P., Herich, H., Gysel, M., El-Haddad, I., Zhang, Y., Močnik, G., Hüglin, C., Baltensperger, U., Szidat, S., and Prévôt, A.S.H. (2017). Evaluation of the absorption Ångström exponents for traffic and wood burning in the Aethalometer-based source apportionment using radiocarbon measurements of ambient aerosol. Atmos Chem Phys *17*, 4229–4249.