

Point-by-point reply to the Referee

Re-review of Luo et al. submission:

The authors have addressed a substantial amount of both reviewers concerns. However, there are a few points that I think are critical that have not been adequately addressed. I still think it is very misleading to call any of these background aerosols. The authors are using this as a designation for aerosols that are not influenced by sea fog or dust, but there is a history in the literature of what a background aerosol is, and this is not an appropriate usage, even with a definition in the manuscript.

Reply: Thank you for your suggestion. ‘Background aerosol’ has been widely used to describe the aerosol collected in a clear site relative to highly polluted areas (Allan et al., 2006; Niemi et al., 2006; Arndt et al., 2017). Following previous definitions of ‘background aerosol’, we changed the relevant sentences to ‘Compared with the ECSs, which were strongly influenced by anthropogenic emissions, the NWPO (open ocean) was relatively clear. Hereafter, we define “background aerosol” as aerosol collected in the NWPO without influence from dust and sea fog during the investigation period.’

It appears based on the authors reply that there are no true experimental blanks. They measured the values on the filters alone, i.e., a filter blank, but they never deployed those filters into their setup on the cruise to determine a methodological blank. This is a serious problem as these types of collections on ship can quite easily have handling blanks due to ship exhaust and not being in a clean laboratory.

Reply: Thank you for pointing this out; it is our mistake. We did not state the process of measuring blanks clearly. We rewrote the relevant procedures in Section 2.2.1:

‘A total of eight filters of the same type as those used to collect samples were taken as blanks. Before storage, the blank filter was placed on the filter holder, then the filter holder with the blank filter was subsequently installed in the TSP sampler on the top of the ship under the vacuum motor power-off for 5 min, after which the blank filter was retrieved. All blank filters and aerosol samples were stored at -20°C during the sampling periods and underwent the same extraction procedures.’

The average concentration for the aerosols is not useful given the incredibly wide range in aerosol concentrations reported.

Reply: Referee is right. The range of N_r concentrations in aerosol was large, ranging by a factor of 40, which was large but within the ranges of previous studies (from ~ 20 to > 100) for marine aerosol on the coast, marginal sea and open ocean (Yeatman et al., 2001; Nakamura et al., 2005; Morin et al., 2009; Miyazaki et al., 2011; Zhu et al., 2013; Violaki et al., 2015), due to meteorological conditions (such as sea fog, rain and dust storm), the sources of air masses (such as anthropogenic sources) and the distance from land. Compared with other studies, the upper and lower bounds of our data, averages and standard deviation have been listed in Table 1. The significant differences at the $p < 0.05$ level between different years are shown in Figure 3.

In addition, the filter blank value should be used to correct the reported values.

Reply: The reported data have been corrected by subtracting the filter blank. To clarify this, we added the following sentence to Section 2.2.1: ‘The data presented here have been corrected for blanks.’

Ammonium is normally fine mode aerosol, except in coastal areas with mixed pollution and marine aerosols, where it frequently is present in the coarse mode (e.g., Yeatman, S.G., Spokes, L.J., Dennis, P.F. & Jickells, T.D. 2001. Comparisons of aerosol nitrogen isotopic composition at two polluted coastal sites.). I do not think it is appropriate to use deposition velocities for TSP samples given the potential for fine mode ammonium nitrate due to high concentrations pollutants (i.e., nitrate in the fine mode), and coarse mode ammonium. Everything related to the dry deposition N fluxes should be removed. Knowing the actual size of each aerosol type in a complex highly polluted yet marine area is critical, it is not okay to just use the typical assumptions. All of section 3.3 should be removed as well as Figure 8.

Reply: While we appreciate the reviewer’s comment, we disagree with it. The deposition velocity is not only affected by the aerosol size distribution, but is also regulated by meteorological conditions (Duce et al., 1991; Hoppel et al., 2002) and underlying surface (Piskunov, 2009). To clearly communicate our assumptions and the potential bias, more descriptions about the deposition velocity have been added to Section 2.3 Paragraph 4. In addition, to promote the value of the rare

dataset, the reasonable estimation under the explicit assumption should be acceptable with uncertainties. Thus, we preferred to keep the following paragraph:

‘In addition to particle size, V_i was controlled by the meteorological conditions (wind speed and relative humidity; Duce et al., 1991; Hoppel et al., 2002) and underlying surface (smooth or rough; Piskunov, 2009). The V_i simulated by the model varied from 0.01 to 10 cm s^{-1} under wind speeds from 5 to 30 m s^{-1} and particle size from 0.1 to 100 μm (Hoppel et al., 2002). In our observations, wind speed ranged from 0.1 to 18 m s^{-1} , with relative humidity ranging from 40 to 100% (for 2014 data, see Luo et al. (2016) and for 2015 data see Fig. S2), which was variable, thus prohibiting an accurate deposition estimate. Moreover, we did not obtain information about the size distribution. Previous studies showed that most of the NO_3^- is distributed in supermicron size (ranging from 1 to 10 μm) in marine aerosol, with a small fraction in submicron size (ranging from 0.1 to 1 μm). On the other hand, NH_4^+ is mainly distributed in submicron size and only partly in supermicron size (Nakamura et al., 2005; Baker et al., 2010; Jung et al., 2013), except in coastal areas with mixed pollution and marine aerosols, where NH_4^+ was present in the coarse mode and NO_3^- in the fine mode (Yeatman et al., 2001). In our case, the weather conditions, such as fog and dust, further affected the size distribution of aerosol N_r (Mori et al., 2003; Yao and Zhang, 2012; Hsu et al., 2014). Therefore, in this study, the deposition velocities of water-soluble nitrogen species were set to 2 cm s^{-1} for NO_3^- , 0.1 cm s^{-1} for NH_4^+ and 1.0 cm s^{-1} for WSON, which have also been widely used to estimate the marine aerosol N_r dry deposition (Nakamura et al., 2005; Jung et al., 2013; Luo et al., 2016). However, bearing in mind, that any use of fixed deposition velocities to calculate the depositional flux of aerosol N_r may cause under- or over-estimation.’

The propagated errors for WSON concentrations are huge – this needs to be visible in all figures and its impact on interpretation should be accounted for.

Reply: The uncertainties (~ 20 to > 100%) caused by error propagation during the calculation of WSON have been thoroughly discussed in previous studies (Mace and Duce, 2002; Cornell et al., 2003; Cape et al., 2011; Lesworth et al., 2010; Zamora et al., 2011). However, no error values were provided. Following the referee’s suggestion, the error bar for WSON concentration has been made visible in corresponding figures and the error has been taken into account in associated WSON

discussions. We also rewrote the Section 2.3 Paragraph 1 to clarify the uncertainties of calculated WSON:

‘The calculated WSON was subject to relatively large and variable uncertainties by error propagation, since $[\text{NO}_3^-]$ and $[\text{NH}_4^+]$ were generally much higher than WSON concentrations in most observations. Such error propagation was a common and unavoidable problem (Mace and Duce, 2002; Cornell et al., 2003; Cape et al., 2011; Lesworth et al., 2010; Zamora et al., 2011). In previous studies, data points with high relative uncertainties were excluded ($>100\%$, Lesworth et al., 2010; Zamora et al., 2011) and the negative values were taken as zero (Mace and Duce, 2002; Cornell et al., 2003; Violaki et al., 2015). Following previous studies, we excluded both the negative data and those with high relative uncertainties to reduce the uncertainty of mean WSON.’

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List of the relevant changes made in the manuscript and indicated in red text

1. We redefined background aerosol in Section 2.1 Paragraph 2.
2. We rewrote the blank procedures in Section 2.2.1.
3. The data standard deviations have been added to Table1.
4. “The data presented here have been corrected for blanks.” has been added in Section 2.2.1.
5. More descriptions about the deposition velocity have been added to Section 2.3 Paragraph 4.
6. We rewrote Section 2.3 Paragraph 1 to clarify the uncertainties of calculated WSON, and recalculated the concentrations of WSON in Table 1, and recalculated the fluxes of WSON in Table 3 and Figure 8.
7. Error bars for WSON concentration have been made visible in Figures 2g and 2h.

Sources of reactive nitrogen in marine aerosol over the Northwest Pacific Ocean in spring

Li Luo^{1,3}, Shuh-Ji Kao^{2*}, Hongyan Bao², Huayun Xiao^{1,3}, Hongwei Xiao^{1,3}, Xiaohong Yao⁴, Huiwang Gao⁴, Jiawei Li⁵, Yangyang Lu²

¹Jiangxi Province Key Laboratory of the Causes and Control of Atmospheric Pollution, East China University of Technology, Nanchang 330013, China

²State Key Laboratory of Marine Environmental Science, Xiamen University, Xiamen 361102, China

³School of Water Resources and Environmental Engineering, East China University of Technology, Nanchang 330013, China

⁴Key laboratory of Marine Environmental Science and Ecology, Ministry of Education, Ocean University of China, Qingdao 266100, China

⁵Key Laboratory of Regional Climate-Environment for Temperate East Asia, Institute of Atmospheric Physics, Chinese Academy of Sciences, Beijing 100029, China

Correspondence to: Shuh-Ji Kao (sjkao@xmu.edu.cn)

Abstract Atmospheric deposition of long-range transport of anthropogenic reactive nitrogen (N_r , mainly comprised of NH_3 , NO_y and water-soluble organic nitrogen (WSON)) from continents may have profound impact on marine biogeochemistry. On the other hand, surface ocean dissolve organic nitrogen (DON) may also contribute to aerosol WSON in the overlying atmosphere. Despite of the importance of off-continent dispersion and N_r interactions at the atmosphere-ocean boundary, our knowledge of the sources of various nitrogen species in the atmosphere over the open ocean remains limited due to insufficient observations. We conducted two cruises in the spring of 2014 and 2015 from the coast of China through the East China Seas (ECSs, i.e. the Yellow Sea and East China Sea) to the open ocean (i.e. the Northwest Pacific Ocean, NWPO). Concentrations of water-soluble total nitrogen (WSTN), NO_3^- and NH_4^+ , as well as the $\delta^{15}N$ of WSTN and NO_3^- in marine aerosol were measured during both cruises. In the spring of 2015, we also analysed the concentrations and $\delta^{15}N$ of NO_3^- and the DON of surface sea water (SSW; at a depth of 5 m) along the cruise track. Aerosol NO_3^- , NH_4^+ and WSON decreased logarithmically (1-2 orders of magnitude) with distance from the shore, reflecting strong anthropogenic emission sources of NO_3^- , NH_4^+ and WSON in China. Average aerosol NO_3^- and NH_4^+ concentrations were significantly higher in 2014 (even in the remote NWOP) than in 2015 due to the stronger wind field in 2014, underscoring the role of the Asian winter monsoon in the seaward transport of anthropogenic NO_3^- and NH_4^+ . However, the background aerosol WSON over the NWPO in 2015 (13.3 ± 8.5 nmol m^{-3}) was similar to that in 2014 (12.2 ± 6.3 nmol m^{-3}), suggesting an additional non-anthropogenic WSON source in the open ocean. Obviously, marine DON emissions should be considered in model and field assessments of net atmospheric WSON deposition in the open ocean. This study contributes information on parallel isotopic marine DON composition and aerosol N_r datasets, but, more research is required to explore complex N_r sources and deposition processes in order to advance our understanding of anthropogenic influences on the marine nitrogen cycle and nitrogen exchange at land-ocean and atmosphere-ocean interfaces.

1 Introduction

Atmospheric transport and deposition of anthropogenic reactive nitrogen (Nr) to global oceans have increased considerably since the industrial revolution (Duce et al., 2008). Due to accumulated atmospheric Nr deposition, the stoichiometric relationship between nitrogen and phosphorus in the upper North Pacific Ocean (where nitrogen is the limiting nutrient in surface ocean) has been significantly altered (Kim et al., 2011). Such alterations may in turn impact pristine oceanic ecosystems and biogeochemical cycles. The Nr species deposited in the ocean include inorganic reduced nitrogen species (NH_3 and NH_4^+), oxidized nitrogen species (HNO_3 and NO_3^-), and organic nitrogen compounds (Erisman et al., 2002). The depositional fluxes (both dry and wet) of atmospheric Nr to global oceans have been studied previously through models (Duce et al., 2008; Doney et al., 2010). Recent model (Kanakidou et al., 2012) and observational (Altieri et al., 2014, 2016) studies have also reported that the ocean may be a source of atmospheric WSON and NH_3 . Nevertheless, field observations in the open ocean remain scarce; thus, more observations and new approaches, such as stable nitrogen isotopic composition studies, are urgently needed to trace the sources of Nr and investigate Nr exchange at the atmosphere-ocean interface.

Using organic nitrogen compounds, Cape et al. (2011) revealed several possible sources of WSON in the atmosphere, including livestock and animal husbandry, fertilisers, vehicle exhaust, biomass burning, secondary pollutants, and marine biological sources. Cape et al. (2011) also explicitly noted that complex atmospheric chemical process may obscure source identification for individual organic nitrogen compounds in atmospheric WSON. Stochastic analysis coupled with molecular characterization using Fourier-transform ion cyclotron resonance mass spectrometry revealed that biological organic nitrogen in surface sea water can be a source of atmospheric WSON over the open ocean (Wozniak et al., 2014; Altieri et al., 2016). Similar conclusions have been drawn from the positive correlation between marine aerosol WSON concentration and wind speed during a cruise in the Northwest Pacific Ocean (Luo et al., 2016).

The stable nitrogen isotopic composition ($\delta^{15}\text{N}$, $\delta^{15}\text{N}(\text{‰})_{\text{sample}} = ((^{15}\text{N}/^{14}\text{N})_{\text{sample}} / (^{15}\text{N}/^{14}\text{N})_{\text{standard}} - 1) \times 1000$) may be used to discriminate the sources of atmospheric NO_x and NH_x . This approach (i.e. the use of $\delta^{15}\text{N}\text{-NO}_x$) has successfully distinguished fossil fuel-burning NO_x from soil biogenic-activity NO_x (Felix and Elliott, 2014), as well as coal combustion emissions (Felix et al., 2012) from vehicle exhaust (Walters et al., 2015). Similarly, atmospheric NH_x can be measured and traced using $\delta^{15}\text{N}\text{-NH}_x$ (Freyer, 1978; Heaton, 1987; Jickells, 2003; Altieri et al., 2014). However, direct measurements of atmospheric $\delta^{15}\text{N}\text{-WSON}$ are currently highly impractical due to difficulties in completely separating organic and inorganic nitrogen. Via isotope mass conservation, a few previous studies have reported $\delta^{15}\text{N}\text{-WSON}$ values in precipitation collected from urban, rural and remote regions ranging from -7.3 to $+7.3\text{‰}$ (Cornell et al., 1995), which is consistent with values from precipitation sampled in a metropolis surrounded by agricultural areas in southern Korea (-7.9 to $+3.8\text{‰}$, with annual means of $+0.3\text{‰}$ and $+0.2\text{‰}$ in 2007 and 2008, respectively; Lee et al., 2012), but lower than the $\delta^{15}\text{N}\text{-WSON}$ values (-0.5 to $+14.7\text{‰}$, with a median of $+5\text{‰}$ and non-significant seasonal variation) reported in precipitation over the United states East Coast area

(Russell et al., 1998). Compared with those for precipitation, aerosol $\delta^{15}\text{N}$ -WSON values reported in various rural regions in the United Kingdom cover a wider range (mainly caused by low values; the range covers -14.6 to $+12.5\%$, with medians of -2 and -5% for the fine and coarse mode, respectively; Kelly et al., 2005) (Fig. S1). Using $\delta^{15}\text{N}$, it is more difficult to identify the sources of atmospheric WSON than it is to identify NO_x and NH_x sources. However, the relatively uniform $\delta^{15}\text{N}$ values (+2.2 to +5.4%) of dissolved organic nitrogen (DON) in surface sea water worldwide (Knapp et al., 2005; Knapp et al., 2011) enable the use of the isotope end member mixing approach for primary WSON aerosol; unfortunately, no cruises to date have undertaken parallel marine aerosol sampling and SSW $\delta^{15}\text{N}$ identification.

In terms of the hemispheric wind field, the East Asian monsoon transition from winter (October to April) to summer (May to September) influences the entire East Asian region. During the East Asian winter monsoon period, strong cold air masses mobilize rapidly through north-eastern China to the NWPO; in contrast, summer monsoon air masses arise primarily from the tropical Pacific Ocean (Wang et al., 2003). Air masses originating from China in winter have been reported to contain higher concentrations of NO_x and NH_x than air masses arising from remote Pacific Ocean regions in summer (Kunwar et al., 2014). Monitoring over the NWPO at Hedo Island and Ogawasara Island also shows that the dry deposition of aerosol NO_3^- and NH_4^+ varies inter-annually by a factor of 2-5 due to variable monsoon intensity (<http://www.eanet.asia/>). Moreover, dust storms occur frequently during monsoonal transition periods, and long-range transport in the upper- and mid-troposphere through northern China to the remote Pacific Ocean (Yang et al., 2013); these dust plumes contain abundant crustal elements in addition to NO_x and NH_x (Duce et al., 1980; Kang et al., 2009). In order to evaluate the seaward gradient of atmospheric N_r concentrations and explore the sources and fates of atmospheric NO_3^- , NH_4^+ and WSON from China (which features the largest emissions of such species worldwide), we conducted cruises from China to the NWPO during spring, when the East Asian monsoon transition period occurs; cruises were complete during two different years to allow comparison.

In this study, we measured WSTN, NO_3^- and NH_4^+ concentrations, as well as $\delta^{15}\text{N}$ -WSTN and $\delta^{15}\text{N}$ - NO_3^- in marine aerosols collected over the ECSs and the NWPO during the spring of 2014 and 2015. The concentrations and $\delta^{15}\text{N}$ of DON and NO_3^- in SSW (surface sea water, collected at a depth of 5 m) were analysed in parallel along the cruise track in 2015. The purposes of this study were (1) to investigate the spatial distributions of concentrations of various N_r species in marine aerosol from the ECSs to the NWPO, (2) to explore possible sources of atmospheric WSON in marine environments and (3) to advance our understanding of atmospheric N_r transport at the land-ocean boundary and potential N_r exchanges between the atmosphere and the ocean.

2 Material and Methods

2.1 Sampling and background weather during cruises

Total suspended particulate (TSP) samples were collected using a high-volume sampler (TE-5170D; Tisch Environmental,

Inc.) with Whatman[®]41 cellulose filters during two research cruises (Fig. 1) aboard the R/V *Dongfanghong II*. The first cruise (Fig. 1a) spanned from 17 March to 22 April 2014 (44 samples were collected in total; detailed sampling information can be found in Luo et al., 2016), and the second cruise (Fig. 1b) lasted from 30 March to 3 May 2015 (38 samples were collected in total; detailed sampling information, including the date, time period and location for each sample are listed in Table S1). To avoid self-contamination from the research vessel, the TSP sampler was installed on the top of the tower at the ship head, and aerosols were sampled only during travel. More information about self-contamination from the ship exhaust can be found in Luo et al. (2016). Both cruises were undertaken during the East Asian monsoon transition period. The 2-month average (March and April) wind streamlines at 1000 hPa over the NWPO show that the wind speed ranged from 2 to 6 m s⁻¹ in 2014 (Fig. 1a) and from 1 to 3 m s⁻¹ in 2015 (Fig. 1b). In general, the wind was stronger in 2014 than in 2015 over the open ocean during the sampling periods.

Meteorological data, including wind speed, direction, relative humidity (RH) and ambient temperature, are shown in Figure S2 for the 2015 cruise (data for the 2014 cruise are reported in Luo et al., 2016). In the ECSs, both cruises encountered sea fog, which inevitably influenced aerosol sampling and aerosol chemistry. Because of the analogous weather conditions experienced during the two cruises, we used the techniques of Luo et al. (2016) to classify the 2015 marine aerosol samples into three types (namely, sea-fog-modified aerosol (orange triangles) collected in the ECSs, dust aerosol (pink circles), and background aerosol (black squares) sampled in the NWPO; Fig. 1) based on the meteorological conditions (Fig. S2), concentrations of aluminium (data not shown), and the Lidar browse images from NASA (Fig. S3). Compared with the ECSs, which were strongly influenced by anthropogenic emissions, the NWPO (open ocean) was relatively clear. Hereafter, we define “background aerosol” as aerosol collected in the NWPO without influence from dust and sea fog during the investigation period.

To examine the relationships between the isotopic compositions of WSON in marine aerosol and DON in SSW, we collected SSW at a depth of 5 m (sampling locations are shown in Fig. 1b as open blue circles) using Niskin bottles during the 2015 cruise. The SSW samples were filtered using a 0.22 µm MILLEX•GP filter and kept frozen at -20 °C in 50 mL 450 °C pre-combusted brown glass tubes until analysis.

2.2 Chemical analyses

2.2.1 NO₃⁻ and NH₄⁺ in marine aerosol

The marine aerosol samples were extracted in Milli-Q water (with specific resistivity of 18.2 MΩ cm⁻¹) following Luo et al. (2016). The aerosol extracts were analysed using an ion chromatograph (model ICS-1100 for anions and model ICS-900 for cations) equipped with a conductivity detector (ASRS-ULTRA) and suppressor (ASRS-300 for the ICS-1100 and CSRS-300 for the ICS-900). The precision was better than 5% for all ionic species. Details of the analytical processes can be found in Hsu et al. (2014). Only five of the aerosol samples contained detectable NO₂⁻, and these accounted for < 1% of the WSTN.

A total of eight filters of the same type as those used to collect samples were taken as blanks. Before storage, the blank filter was placed on the filter holder, then the filter holder with the blank filter was subsequently installed in the TSP sampler on the top of the ship under the vacuum motor power-off for 5 min, after which the blank filter was retrieved. All blank filters and aerosol samples were stored at -20°C during the sampling periods and underwent the same extraction procedures. The data presented here have been corrected for blanks.

2.2.2 NO_3^- in SSW

The SSW NO_3^- concentration was measured using a chemiluminescence method (Braman and Hendrix, 1989). Briefly, the solution containing NO_3^- was injected into a heated solution of acidic Vanadium (III), in which the NO_3^- was reduced to nitric oxide (NO) to be measured by a NO_x analyser (MODEL T200U, Teledyne Technologies Incorporated, USA). Working standards were injected after every 10 samples. The relative standard deviation for the standard replicate was $< 5\%$. The concentration of NO_2^- in the SSW was below the $0.1\text{ }\mu\text{mol L}^{-1}$ detection limit throughout the cruise, as reported previously (Adornato et al., 2005).

2.2.3 WSTN in marine aerosol and total dissolve nitrogen in SSW

Aerosol WSTN and SSW total dissolved nitrogen (TDN, i.e. $\text{NO}_3^- + \text{NH}_4^+ + \text{DON}$) were measured using the alkaline potassium persulfate oxidation method to convert WSTN and TDN to NO_3^- (Luo et al., 2016; Knapp et al., 2005). The NO_3^- content of the digested solution was then measured via chemiluminescent detection (Braman and Hendrix, 1989). To verify the WSTN and TDN oxidation efficiency, N-containing organic and inorganic compound standards (specifically, glycine, urea, ethylene diamine tetraacetic acid, and ammonium sulphate) were prepared in solution at a concentration of $800\text{ }\mu\text{M-N}$ for oxidation analysis. The recoveries of the N-containing compound standards under oxidation by alkaline potassium persulfate were within $95 \sim 105\%$ ($n = 6$).

2.2.4 Stable nitrogen isotope

The $\delta^{15}\text{N-NO}_3^-$ was analysed using the denitrifier method described by Sigman et al. (2001) and Casciotti et al. (2002), which has been widely used to analyse the $\delta^{15}\text{N}$ in NO_3^- in aerosol, rainwater and sea water (Altieri et al., 2013; Buffam and McGlathery, 2003; Gobel et al., 2013; Hastings et al., 2003; Sigman et al., 2005), as well as that in NO_3^- in solutions digested with alkaline potassium persulfate (Knapp et al., 2005; Knapp et al., 2010; Knapp et al., 2011; Knapp et al., 2012). Detailed stable nitrogen isotope analysis procedures can be found in Archana et al. (2016) and Yang et al. (2014). Briefly, NO_3^- was reduced to N_2O by the denitrifying bacteria *Pseudomonas aureofaciens* (ATCC 13985); then, the stable nitrogen isotope of N_2O was analysed using a GasBench II connected to a continuous flow isotope ratio mass spectrometer (IRMS, Thermo Delta V Advantage). Two international standards, namely USGS34 and IAEA-N3 (Böhlke et al., 2003), and two NO_3^- laboratory

working standards were used to verify instrument stability. After the WSTN and TDN were oxidized to NO_3^- , the $\delta^{15}\text{N}$ -WSTN and $\delta^{15}\text{N}$ -TDN were analysed using the same procedures employed for NO_3^- . The pooled standard deviations for replicates were $\pm 0.2\text{‰}$, $\pm 0.5\text{‰}$ and $\pm 0.5\text{‰}$ for $\delta^{15}\text{N}$ - NO_3^- , $\delta^{15}\text{N}$ -WSTN and $\delta^{15}\text{N}$ -TDN, respectively.

2.3 Data analysis

The concentrations of WSON in marine aerosol, which could not be measured directly (as mentioned previously), were calculated using the following equation:

$$[\text{WSON}] = [\text{WSTN}] - [\text{NO}_3^-] - [\text{NH}_4^+], \quad (1)$$

where $[\text{WSON}]$, $[\text{WSTN}]$, $[\text{NO}_3^-]$ and $[\text{NH}_4^+]$ are the molar concentrations (nmol N m^{-3}) of the given water-soluble nitrogen species in marine aerosol. The calculated WSON was subject to relatively large and variable uncertainties by error propagation since $[\text{NO}_3^-]$ and $[\text{NH}_4^+]$ were generally much higher than WSON concentrations in most observations. Such error propagation was a common and unavoidable problem (Mace and Duce, 2002; Cornell et al., 2003; Cape et al., 2011; Lesworth et al., 2010; Zamora et al., 2011). In previous studies, data points with high relative uncertainties were excluded ($>100\%$, Lesworth et al., 2010; Zamora et al., 2011) and the negative values were taken as zero (Mace and Duce, 2002; Cornell et al., 2003; Violaki et al., 2015). Following previous studies, we excluded both the negative data and those with high relative uncertainties to reduce the uncertainty of mean WSON.

Reduced nitrogen (RN, i.e. $\text{NH}_4^+ + \text{WSON}$) and the $\delta^{15}\text{N}$ -RN in the aerosol were calculated via mass balance:

$$[\text{RN}] = [\text{WSTN}] - [\text{NO}_3^-], \quad (2)$$

$$\delta^{15}\text{N-RN} = (\delta^{15}\text{N-WSTN} \times [\text{WSTN}] - \delta^{15}\text{N-NO}_3^- \times [\text{NO}_3^-]) / [\text{RN}], \quad (3)$$

where $[\text{WSTN}]$ and $[\text{NO}_3^-]$ are the molar concentrations (nmol N m^{-3}) of the given water-soluble nitrogen species in the marine aerosol. The average propagated standard error for RN was 9% for both 2014 and 2015, the propagated error for the calculation of $\delta^{15}\text{N-RN}$ was $\pm 0.6\text{‰}$.

Similar to aerosol WSON, the SSW DON concentration and $\delta^{15}\text{N}$ -DON were calculated using the following equations:

$$[\text{DON}] = [\text{TDN}] - [\text{NO}_3^-], \quad (4)$$

$$\delta^{15}\text{N-DON} = (\delta^{15}\text{N-TDN} \times [\text{TDN}] - \delta^{15}\text{N-NO}_3^- \times [\text{NO}_3^-]) / [\text{DON}], \quad (5)$$

where $[\text{TDN}]$ and $[\text{NO}_3^-]$ are the molar concentrations ($\mu\text{mol N L}^{-1}$) of the given species in SSW. The standard error propagated through the DON calculation was 5.3%. Since the average $[\text{NH}_4^+]$ in SSW at the selected sites during the 2015 cruise (12 sites and 23 samples) was $0.05 \mu\text{M}$, which is much less than DON in μM level, thus, $[\text{NH}_4^+]$ was neglected in Equations 4 and 5. Unfortunately, most of the NO_3^- concentrations in the SSW samples were $< 0.5 \mu\text{mol L}^{-1}$, which is too low for the measurement of $\delta^{15}\text{N-NO}_3^-$. We attempted to evaluate the interference from nitrate in the $\delta^{15}\text{N-DON}$ calculations. For all the SSW samples on average, NO_3^- comprised 5.7% of the total NO_3^- plus DON; the $\delta^{15}\text{N-NO}_3^-$ in SSW ranged from +8.2 to +16.4‰ in our measurements, and the bias of the calculated $\delta^{15}\text{N-DON}$ varied from +0.5 to +0.9‰.

The dry deposition nitrogen fluxes were calculated as follows:

$$F = C_i \times V_i, \quad (7)$$

where C_i is the concentration of a given water-soluble nitrogen species in the aerosol, and V_i is the given dry deposition velocity of the given nitrogen species. In addition to particle size, V_i was controlled by the meteorological conditions (wind speed and relative humidity; Duce et al., 1991; Hoppel et al., 2002) and underlying surface (smooth or rough; Piskunov, 2009). The V_i simulated by the model varied from 0.01 to 10 cm s^{-1} under wind speeds from 5 to 30 m s^{-1} and particle size from 0.1 to 100 μm (Hoppel et al., 2002). In our observations, wind speed ranged from 0.1 to 18 m s^{-1} , with relative humidity ranging from 40 to 100% (for 2014 data, see Luo et al. (2016) and for 2015 data see Fig. S2), which was variable, thus prohibiting an accurate deposition estimate. Moreover, we did not obtain information about the size distribution. Previous studies showed that most of the NO_3^- is distributed in supermicron size (ranging from 1 to 10 μm) in marine aerosol, with a small fraction in submicron size (ranging from 0.1 to 1 μm). On the other hand, NH_4^+ is mainly distributed in submicron size and only partly in supermicron size (Nakamura et al., 2005; Baker et al., 2010; Jung et al., 2013), except in coastal areas with mixed pollution and marine aerosols, where NH_4^+ was present in the coarse mode and NO_3^- in the fine mode (Yeatman et al., 2001). In our case, the weather conditions, such as fog and dust, further affected the size distribution of aerosol Nr (Mori et al., 2003; Yao and Zhang, 2012; Hsu et al., 2014). Therefore, in this study, the deposition velocities of water-soluble nitrogen species were set to 2 cm s^{-1} for NO_3^- , 0.1 cm s^{-1} for NH_4^+ and 1.0 cm s^{-1} for WSON, which have also been widely used to estimate the marine aerosol Nr dry deposition (Nakamura et al., 2005; Jung et al., 2013; Luo et al., 2016). However, bearing in mind, that any use of fixed deposition velocities to calculate the depositional flux of aerosol Nr may cause under- or over-estimation.

3 Results and Discussion

3.1 Spatial and temporal variations of water-soluble nitrogen species in the aerosol

Overall, significant logarithmic decreases occurred from the shore seaward for all water-soluble nitrogen species and WSTN in both 2014 and 2015 (Fig. 2). The seaward gradient was caused primarily by continental emissions influenced by sea fog (Luo et al., 2016); thus, concentrations were high in the ECSs (orange triangles in Fig. 2) and low offshore in the NWPO background aerosol (black squares in Fig. 2). Dust aerosol (pink circles in Fig. 2) appeared sporadically in the NWPO and generally featured higher NH_4^+ and NO_3^- (but not WSON) values (Table 1 and Fig. 2).

The measured WSTN concentrations in TSP varied from 21 to 2411 nmol m^{-3} (Table 1 and Fig. 2a and 2b), lower than those in PM_{10} sampled during spring in Xi'an, China (which ranged from 786 to 3000 nmol m^{-3} ; Wang et al., 2013), but higher than those in TSP sampled in Sapporo, Japan (which ranged from 20.9 to 108.6 nmol m^{-3} ; Pavuluri et al., 2015), Okinawa Island (which ranged from 5 to 216 nmol m^{-3} ; Kunwar and Kawamura, 2014), and the North Pacific (which ranged from 1.4

to 64.3 nmol m^{-3} in May-July; Hoque et al., 2015). This wide range of aerosol WSTN content illustrates the influence of the distance between sampling locations and emission sources (Matsumoto et al., 2014), seasonality (Kunwar and Kawamura, 2014), and meteorological conditions, such as sea fog (Luo et al., 2016).

The concentrations of marine aerosol WSTN in the ECSs ranged from 444 to 2411 nmol m^{-3} in 2014 (with a volume-weighted mean of 1136 nmol m^{-3}) and from 92.9 nmol m^{-3} to 1195 nmol m^{-3} in 2015 (with a volume-weighted mean of 287 nmol m^{-3}), which were clearly higher than those in dust aerosol (with volume-weighted means of 242 nmol m^{-3} in 2014 and 154 nmol m^{-3} in 2015) and background aerosol (with volume-weighted means of 85.6 nmol m^{-3} in 2014 and 42.3 nmol m^{-3} in 2015) collected in the NWPO (Table 1). The air mass backward trajectories (see Fig. S5 for 2015 and Luo et al., 2016 for 2014) reveal that the high aerosol water-soluble nitrogen species in the ECSs arose from anthropogenic Nr emissions from eastern China (Gu et al., 2012). In addition, frequent formation of sea fog in the ECSs in spring (Zhang et al., 2009) may also enrich the amount of water-soluble nitrogen in sea-fog-modified aerosol via chemical processing (Luo et al., 2016). The much higher water-soluble nitrogen species in the ECSs marine aerosol (compared to that in the NWPO aerosol) indicates that continental/anthropogenic Nr strongly affected the marine aerosol. However, the amounts of sea-salt ions (such as Na^+) in the ECSs aerosols sampled in both 2014 ($123 \pm 98 \text{ nmol m}^{-3}$; Luo et al., 2016) and 2015 ($151 \pm 164 \text{ nmol m}^{-3}$; Luo et al., unpublished data) were higher than those in land aerosol sampled during spring ($23 \pm 7.8 \text{ nmol m}^{-3}$ in Beijing; Zhang et al., 2013), which implies that those aerosols sampled in the ECSs were also significantly influenced by sea salt. Thus, we define the aerosol collected by ship over the ECSs as marine aerosol.

In the NWPO, higher WSTN values were observed in dust aerosol than in background aerosol in both 2014 and 2015; the dust aerosol WSTN consisted predominantly of NH_4^+ and NO_3^- rather than WSON (Table 1, pink circles in Fig. 2), which implies that dust can carry more NH_4^+ and NO_3^- during long-range transport from East Asia to the NWPO during the Asian winter monsoon in spring. The air mass back trajectories of those dust aerosols arose mainly from high- Nr regions, as evidenced by dust plumes captured by Lidar browse images from NASA (Fig. S3 for 2015 and Luo et al., 2016 for 2014).

In our observations, the concentrations of NH_4^+ and NO_3^- were higher in all aerosols in 2014 than they were in 2015 (Fig. 3a and 3b; Table 1). The difference between the two years was caused by a stronger Asian winter monsoon in 2014. Additionally, the cruise in 2014 (17 March to 22 April) occurred during a period of intensive fossil fuel combustion for heat supply in northern China; in contrast, the 2015 cruise started on 30 March and finished on 3 May, during a decrease in heating demand. The influence of heating on aerosol emissions can be seen in the atmospheric aerosol optical depth over heat-generating areas in China; for example, Xiao et al. (2015) reported that the aerosol optical depth was five times higher during heat generation periods than during non-generation period. The consistent variations between heat supply in northern China and higher NO_3^- in marine aerosol sampled in 2014 than 2015 underscore the influence of anthropogenic NO_x emissions on marine aerosol NO_3^- . Our spring observations in 2014 and 2015 showed average concentrations of NH_4^+ and NO_3^- in background aerosol (black boxes in Fig. 3a and 3b; Table 1) higher than the average concentrations ($3.5 \pm 3.3 \text{ nmol m}^{-3}$ for

NH_4^+ and $2.1 \pm 1.5 \text{ nmol m}^{-3}$ for NO_3^-) reported in the western Pacific Ocean in summer (blue boxes in Fig. 3a and 3b, data from Miyazaki et al., 2011 and Jung et al., 2013), which suggests far-reaching influence of anthropogenic emissions during the monsoon transition. Moreover, concentrations of NH_4^+ and NO_3^- were higher in 2014 due to the stronger Asian winter monsoon, which further supports the idea that the monsoon exerts an important role in annual and seasonal variations in marine aerosol Nr via atmospheric long-range transport.

Unlike NH_4^+ and NO_3^- , WSON concentrations in the background aerosol sampled in the NWPO in 2014 (average = $12.2 \pm 6.3 \text{ nmol m}^{-3}$) were similar to those in 2015 (average = $13.3 \pm 8.5 \text{ nmol m}^{-3}$; black boxes in Fig. 3c). In the open ocean, apart from terrestrial and anthropogenic WSON long-range transport (Mace et al., 2003; Lesworth et al., 2010), the ocean itself is the most likely source of marine aerosol WSON. For instance, in situ observations in the subtropical North Atlantic found that aerosol WSON had strong positive relationships with surface ocean primary productivity and wind speed (Altieri et al., 2016). Another study in the South Atlantic Ocean showed that the WSON in marine aerosol associated with high SSW chlorophyll-*a* was 9 times higher than that associated with low SSW chlorophyll-*a* (Violaki et al., 2015). In our observations, the WSON in the background aerosol (black bar in Fig. 3c) was significantly higher in spring than in summer (blue bar in Fig. 3c), which is consistent with the higher SSW chlorophyll-*a* concentration over the NWPO in spring relative to that in summer (Fig. S6). However, the sources of marine aerosol WSON are a complex mixture composed of primary marine organic N and secondary N-containing organic aerosol. Biogenic organic material in SSW can be injected into the atmosphere to form an ice cloud via bubble bursting at the atmosphere-ocean interface (Wilson et al., 2015); this is probably the primary WSON aerosol source. Volatile organic compounds emitted from the surface ocean can react with NO_x and NH_x in the atmosphere to form secondary N-containing organic aerosol (Fischer et al., 2014; Liu et al., 2015).

3.2 Isotopic composition of nitrogen species

Aerosol $\delta^{15}\text{N-NO}_3^-$ values over the ECSs and NWPO in 2014 and 2015 ranged from -9.2 to $+10.2\text{‰}$ (Table 2 and Fig. 4a and 4b). All the observed $\delta^{15}\text{N-NO}_3^-$ values fell within the ranges previously reported for atmospheric $\delta^{15}\text{N-NO}_3^-$ over land (Elliott et al., 2009; Fang et al., 2011; Felix and Elliott, 2014) and in the marine boundary layer (Altieri et al., 2013; Gobel et al., 2013; Hastings et al., 2003; Morin et al., 2009; Savarino et al., 2013). The mass-weighted mean aerosol $\delta^{15}\text{N-NO}_3^-$ values in 2014 ($+1.6\text{‰}$ in the ECSs, -1.1‰ for dust aerosol and -2.6‰ for background aerosol sampled in the NWPO) were similar to those in 2015 ($+1.9\text{‰}$ in the ECSs, -3.8‰ for dust aerosol and -1.1‰ for background aerosol sampled in the NWPO; Table 2) for all the aerosols, suggesting that the aerosol NO_3^- arose from similar origins and atmospheric chemical pathways in both 2014 and 2015.

The $\delta^{15}\text{N-WSTN}$ values for all the aerosols ranged from -10.7 to $+5.6\text{‰}$ (Table 2, Fig 4c and 4d), which is consistent with the $\delta^{15}\text{N-WSTN}$ ranges reported in precipitation, namely -4.2 to $+12.3\text{‰}$ in the Baltic Sea (Rolff et al., 2008); -8 to $+8\text{‰}$ in Bermuda (Knapp et al., 2010); -4.9 to $+3.2\text{‰}$ in a forest in southern China (Koba et al., 2012); and -12.1 to $+2.9\text{‰}$ in Cheju,

Korea (Lee et al., 2012). In contrast, our results were lower than the $\delta^{15}\text{N}$ -WSTN in TSP sampled in Sapporo, Japan (+12.2 to +39.1‰; Pavuluri et al., 2015) and the Sapporo Forest (+9.0 to +26.0‰; Miyazaki et al., 2014). These authors attributed higher isotopic values to biogenic sources, nitrogenous aerosol aging, and fossil fuel combustion. However, WSTN is, in fact, composed of various nitrogen species, and the relative proportions of NH_4^+ , NO_3^- and WSON to WSTN, coupled with their isotopic compositions (i.e. $\delta^{15}\text{N}$ - NH_4^+ , $\delta^{15}\text{N}$ - NO_3^- and $\delta^{15}\text{N}$ -WSON), jointly mediate variations in aerosol $\delta^{15}\text{N}$ -WSTN (where $\delta^{15}\text{N}$ -WSTN \cdot [WSTN] = $\delta^{15}\text{N}$ - NO_3^- \cdot [NO_3^-] + $\delta^{15}\text{N}$ - NH_4^+ \cdot [NH_4^+] + $\delta^{15}\text{N}$ -WSON \cdot [WSON]). Taking our study as an example, the inconsistent trends in both the positive relationships between $\delta^{15}\text{N}$ -WSTN and $\delta^{15}\text{N}$ - NO_3^- in the ECSs aerosols and NWPO background aerosols (Fig. S7a and S7c) and the negative relationship between $\delta^{15}\text{N}$ -WSTN and the NO_3^- concentration (Fig. S7d and S7f) imply that the $\delta^{15}\text{N}$ of other species (NH_4^+ and WSON) in WSTN affected the $\delta^{15}\text{N}$ -WSTN.

The $\delta^{15}\text{N}$ -WSTN values in 2014 (−10.7 to +1.0‰) were lower than those in 2015 (−5.6 to +5.6‰; Table 2), whereas the NH_4^+ /WSTN ratios were higher in 2014 than in 2015 (Table 1) for all aerosol types. The negative linear relationships between NH_4^+ /WSTN and $\delta^{15}\text{N}$ -WSTN for all aerosol types (Fig. 5a, 5b, and 5c) may be attributed to higher proportions of NH_4^+ in WSTN and negative $\delta^{15}\text{N}$ - NH_4^+ values, which sourced from the anthropogenic and marine emissions (Altieri et al., 2014; Jickells, 2003; Koba et al., 2012; Liu et al., 2014; Xiao et al., 2012; Yeatman et al., 2001). In fact, low $\delta^{15}\text{N}$ - NH_4^+ values have been reported in precipitation in many places, such as Beijing (−33.0 to +14.0‰ with an arithmetic mean of −10.8‰; Liu et al., 2014); Guiyang City in southwestern China (−38.0 to +5.0‰ with an average of −15.9‰; Xiao et al., 2012); Gwangju, Korea (−15.9 to +2.9‰ with volume-weighted means of −6.0‰ in 2007 and −6.8‰ in 2008; Lee et al., 2012); and a forest in southern China (−18.0 to +0.0‰ with a concentration-weighted mean of −7.7‰; Koba et al., 2012). Low atmospheric $\delta^{15}\text{N}$ - NH_4^+ has also been associated with marine air masses (e.g. −8 to −5‰, Jickells et al. 2003; −9 ± 8‰, Yeatman et al., 2001; and −4.1 ± 2.6‰, Altieri et al., 2014). Together, this low atmospheric average $\delta^{15}\text{N}$ - NH_4^+ (−15.9 to −4.1‰) supports our findings of higher NH_4^+ /WSTN and lower $\delta^{15}\text{N}$ -WSTN in aerosol.

There were positive linear relationships between the WSON/WSTN ratio and $\delta^{15}\text{N}$ -WSTN for all the aerosols types (Fig. 5d, 5e, and 5f). This implies that the aerosol $\delta^{15}\text{N}$ -WSON may be positive. The WSON in the marine aerosol either originated from terrestrial long-range transport or the DON from SSW and N-containing secondary organic marine aerosol as discussed in Section 3.1. Terrestrial aerosol $\delta^{15}\text{N}$ -WSON was reported in a wide range (−15.0 to +14.7‰) with mean values from −3.7 to +5.0‰, and $\delta^{15}\text{N}$ -WSON in marine aerosol with more positive $\delta^{15}\text{N}$ (Fig. S1). In addition, our own observations for $\delta^{15}\text{N}$ -DON in SSW showed positive $\delta^{15}\text{N}$ in the ECSs (varying from +5.1 to +12.9‰, with an average of +7.9 ± 2.3‰) and NWPO (ranging from +1.9 to +11.6‰, with an average of +5.7 ± 2.0‰; Fig. 6a). At the same time, DON concentration in our observations ranged from 4.4 to 11.8 $\mu\text{mol L}^{-1}$ (Fig. 6b), which is within the DON concentration range reported in global SSW (Knapp et al., 2011; Letscher et al., 2013; Li et al., 2009; Lønborg et al., 2015; Van Engeland et al., 2010). This high DON concentration in SSW may be ejected into the atmosphere during bubble-bursting (Wilson et al., 2015).

To better clarify the sources of aerosol WSON, we removed the aerosol NO_3^- and its $\delta^{15}\text{N}$ effect from the WSTN and its

$\delta^{15}\text{N}$ -WSTN, respectively, by mass balance. The remaining NH_4^+ and WSON was defined as reduced nitrogen ($\text{RN} = \text{NH}_4^+ + \text{WSON}$). The $\delta^{15}\text{N}$ -RN ranged from -11.8 to $+7.4\text{‰}$ for all aerosol types in both 2014 and 2015 (Table 2), which is consistent with the reported $\delta^{15}\text{N}$ -RN in precipitation (-12.6 to $+7.8\text{‰}$) collected in Bermuda, in the Atlantic Ocean (Knapp et al., 2010). The negative relationships between $\delta^{15}\text{N}$ -RN and NH_4^+ concentration (Fig. S8a, S8b, and S8c), and between $\delta^{15}\text{N}$ -RN and the ratios of NH_4^+/RN (Fig. S8d, S8e, and S8f) for all aerosol types further support the low values of $\delta^{15}\text{N}$ - NH_4^+ in marine aerosol. Three important end members, compiled atmospheric $\delta^{15}\text{N}$ - NH_4^+ (-15.9 to -4.1‰ , green bars in Fig. 7) and continental $\delta^{15}\text{N}$ -WSON (-3.7 to $+5.0\text{‰}$, grey bars) versus the $\delta^{15}\text{N}$ -DON observed in SSW in our cruise ($+7.9 \pm 2.3\text{‰}$ in the ECSs and $+5.7 \pm 2.0\text{‰}$ in the NWPO, red dots with error bars), were added to Figure 7 to facilitate discussion. Note that nearly all the aerosol $\delta^{15}\text{N}$ -RN values in 2014 were lower than those in 2015, which may be attributed to higher NH_4^+ concentrations in 2014 than in 2015 (Table 1). Moreover, higher values of $\delta^{15}\text{N}$ -RN can be seen with higher WSON/RN ratios for all aerosol types in Fig. 7. The high values of $\delta^{15}\text{N}$ in continental WSON and marine DON in SSW may cause these positive relationships. Thus, although higher WSON/RN values accompany with higher $\delta^{15}\text{N}$ -RN, we may still not conclude a significant DON in SSW contribution to aerosol WSON. In the open ocean, to some extent, background aerosol WSON was more likely influenced by DON in surface sea water judging by nitrogen isotopic information.

Note that some data points collected in 2015 for the open ocean case in Fig. 7b and 7c fell outside the mixing field, deviating toward higher $\delta^{15}\text{N}$ -RN values; these high $\delta^{15}\text{N}$ -RN values may be attributable to $\delta^{15}\text{N}$ fractionation and ^{15}N enrichment in the WSON during processes such as secondary N-containing organic aerosol formation by the reaction of NH_x or NO_x with organic aerosol (Fischer et al., 2014; Liu et al., 2015), complex atmospheric chemical reactions (i.e. the photolysis of organic nitrogen into ammonium; Paulot et al., 2015), aerosol WSON aging process, and in-cloud scavenging (Altieri et al., 2016). More studies are needed to explore nitrogen transformation processes, especially those focusing on secondary N-containing organic aerosol in the atmosphere from an isotopic perspective.

3.3 N_r dry deposition and its biogeochemical role

The dry deposition of aerosol NH_4^+ , NO_3^- and WSON is summarized in Table 3. The calculated depositional fluxes of water-soluble nitrogen species in the ECSs were significantly higher than those in the NWPO (Fig. 8). The averaged dry depositional fluxes of NH_4^+ and NO_3^- in 2014 were 2 to 5 times higher than those in 2015 for all aerosol types (Table 3). The dry depositional fluxes of NH_4^+ and NO_3^- in dust aerosol were clearly higher than those in background aerosol in the NWPO (Table 3, Fig. 8a, 8b, 8c and 8d). Comparisons of these dry fluxes with other similar studies and estimations of the contribution of atmospheric N_r deposition to primary production in the NWPO are discussed in Luo et al. (2016), specifically for 2014; here, we focus on the influence of atmospheric N_r deposition on the nitrogen cycle in the ocean.

The influences of atmospheric N_r deposition on the marine nitrogen cycle are obvious over long time scales. For example, by analysing concentrations of NO_3^- and phosphorus in sea water over the NWPO from 1980 to 2010, Kim et al. (2010)

reported that the higher N/P ratio in the upper ocean (in contrast to the deep ocean) in the NWPO was caused primarily by the accumulation of atmospheric anthropogenic N_r deposition. Another recent study found higher atmospheric anthropogenic N_r deposition to be associated with lower $\delta^{15}N$ in surface sediment over the NWPO (Kim et al., 2017), the authors also posited that atmospheric anthropogenic N_r deposition can reach as far down as the deep ocean through biological action and lower the $\delta^{15}N$ in surface sediment. The atmospheric $\delta^{15}N$ values for water-soluble nitrogen species in our observations (Table 2) are lower than the $\delta^{15}N-NO_3^-$ in deep ocean water (+5.6‰; unpublished data from Kao); thus, it is possible that atmospheric $\delta^{15}N-N_r$ to lower $\delta^{15}N-NO_3^-$ in the thermocline as mentioned in previous studies (Knapp et al., 2010; Yang et al., 2014). However, it is hard to quantify the contribution of atmospheric $\delta^{15}N-N_r$ to $\delta^{15}N-NO_3^-$ in the thermocline from the perspective of ^{15}N for the following reasons: First, there are large spatial and temporal uncertainties in the dry and wet depositional fluxes of atmospheric N_r . For example, the dry depositional fluxes of NH_4^+ and NO_3^- ($23.1 - 43.3 \mu\text{mol N m}^{-2} \text{ d}^{-1}$; Table 3) in our observations are significantly higher than those in summer ($4.9 \mu\text{mol N m}^{-2} \text{ d}^{-1}$; Jung et al. 2013). Moreover, according to a previous study, wet N_r deposition is 2 – 3 times higher than dry deposition (Jung et al., 2013), but N_r wet deposition in spring over the NWPO is unknown. Second, the $\delta^{15}N$ of N-fixation (–2 to 0‰; summarized by Knapp et al., 2010) is similar to the atmospheric $\delta^{15}N-N_r$ (Table 2), but the fluxes of N-fixation in the global ocean also vary considerably both spatially and temporally (Mulholland and Bernhardt, 2005; Needoba et al., 2007; Karl et al., 1997). Third, ^{15}N fractionation occurs in the complicated marine nitrogen cycle (Knapp et al., 2005, 2010, 2011, 2012), which hampers the use of ^{15}N in estimating the influence of atmospheric deposition N_r on marine $\delta^{15}N$ under our limited current understanding.

4 Conclusions

Concentrations of water-soluble total nitrogen (WSTN), nitrate (NO_3^-) and ammonium (NH_4^+), as well as the stable nitrogen isotopes of $\delta^{15}N$ -WSTN and $\delta^{15}N-NO_3^-$, were measured in marine aerosols sampled between the ECSs and the NWPO in spring 2014 and 2015. Dissolve organic nitrogen (DON) and $\delta^{15}N$ -DON were also analysed in SSW collected at a depth of 5 m along the cruise route in the spring of 2015. The highest concentrations of water-soluble nitrogen species were found in aerosol sampled in the ECSs, which suggests significant influence from anthropogenic emissions on aerosol N_r . The higher NO_3^- and NH_4^+ in all aerosol types in 2014 (relative to 2015) may be attributed to the stronger Asian winter monsoon in 2014, as well as the intensity of residential heating in spring in northern China.

Negative linear relationships were found between the $NH_4^+/WSTN$ ratios and $\delta^{15}N$ -WSTN for all aerosol types. In contrast, positive linear relationships were observed between the WSON/WSTN ratios and $\delta^{15}N$ -WSTN. The distinctive nitrogen species compositions and isotopic compositions suggest that aerosol $\delta^{15}N$ -WSTN values were mediated synergistically by NO_3^- , NH_4^+ , and WSON in our observations. Meanwhile, our isotope mixing model indicates that DON in SSW is likely to be a source of primary WSON in aerosol, especially over the open ocean. Many uncertainties remain

concerning Nr in the marine boundary layer and SSW, let alone Nr exchange at the atmosphere-ocean interface; further study of Nr exchange between the lower atmosphere and upper ocean is needed in the future.

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Table 1. Concentration ranges and means for WSTN, NH_4^+ , NO_3^- , WSON and RN in aerosols.

		WSTN (nmol m ⁻³)			NH_4^+ (nmol m ⁻³)			NH_4^+ / WSTN	NO_3^- (nmol m ⁻³)			NO_3^- / WSTN	WSON (nmol m ⁻³)			WSON / WSTN	RN (nmol m ⁻³)		
		Range	Mean ^a	Mean ^b	Range	Mean ^a	Mean ^b		Range	Mean ^a	Mean ^b		Range	Mean ^a	Mean ^b		Range	Mean ^a	Mean ^b
ECSs (Sea fog)	2014	444 ~ 2411	1126 \pm 644	1136	228 ~ 777	442 \pm 194	437	0.42 \pm 0.09	160 ~ 1118	536 \pm 300	550	0.48 \pm 0.07	39 ~ 517	170 \pm 76	169	0.12 \pm 0.06	267 ~ 1294	591 \pm 355	585
	2015	92.9 ~ 1195	321 \pm 339	287	25.7 ~ 564	126 \pm 168	113	0.35 \pm 0.12	30.1 ~ 239	93.2 \pm 62.8	87.6	0.35 \pm 0.15	6.9 ~ 392.3	114 \pm 12	83.9	0.32 \pm 0.17	32.6 ~ 956	228 \pm 281	199
NWPO (Dust)	2014	205 ~ 297	245 \pm 36.1	242	94.4 ~ 163	138 \pm 24.2	137	0.56 \pm 0.07	78.6 ~ 145	100 \pm 22.9	98.3	0.41 \pm 0.05	16.9	16.9	16.9	0.08	111 ~ 173	145 \pm 22.8	144
	2015	81.4 ~ 340	147 \pm 77.8	154	20.7 ~ 143	58.3 \pm 39.6	61.4	0.39 \pm 0.13	34.7 ~ 126	57.2 \pm 27.5	58.6	0.41 \pm 0.1	11.0 ~ 80.7	34.9 \pm 26.9	33.7	0.23 \pm 0.13	45.0 ~ 214	89.9 \pm 54.2	95.8
NWPO	2014	31.4 ~ 411	88.8 \pm 74.0	85.6	16.1 ~ 244	54.0 \pm 45.1	51.8	0.60 \pm 0.11	6.4 ~ 166	25.8 \pm 32.0	25.1	0.27 \pm 0.09	3.1 ~ 27.1	12.2 \pm 6.3	11.5	0.17 \pm 0.06	19.2 ~ 245	63.1 \pm 44.5	60.5
	2015	21.0 ~ 68.7	41.7 \pm 12.6	42.3	6.9 ~ 29.5	16.3 \pm 6.0	15.5	0.41 \pm 0.15	2.8 ~ 35.1	12.7 \pm 8.4	13.4	0.29 \pm 0.13	5.1 ~ 40.3	13.3 \pm 8.5	13.4	0.31 \pm 0.15	15.2 ~ 48.2	29.0 \pm 8.2	28.9

a, arithmetic mean

b, volume-weighted mean

Table 2. Ranges and means for stable nitrogen isotopes of WSTN, NO_3^- and RN in aerosols.

		$\delta^{15}\text{N}$ -WSTN			$\delta^{15}\text{N}$ - NO_3^-			$\delta^{15}\text{N}$ -RN		
		Range	Mean ^a	Mean ^b	Range	Mean ^a	Mean ^b	Range	Mean ^a	Mean ^b
ECSs (Sea fog)	2014	-5.3 ~ -1.7	-3.4	-3.4	-1.7 ~ +4.3	+1.0	+1.6	-11.8 ~ -2.5	-7.3	-7.9
	2015	-4.3 ~ +0.9	-1.1	-2.1	-1.3 ~ +10.2	+2.8	+1.9	-5.2 ~ -0.6	-3.3	-3.7
NWPO (Dust)	2014	-6.9 ~ -3.1	-4.7	-4.4	-3.0 ~ +1.3	-1.0	-1.1	-9.4 ~ -5.8	-7.1	-7.3
	2015	-3.0 ~ +0.8	-1.5	-1.3	-7.3 ~ -2.1	-3.9	-3.8	-2.7 ~ +5.5	+0.6	+0.3
NWPO (Bgd.)	2014	-10.7 ~ +1.0	-5.6	-5.5	-7.6 ~ +4.3	-2.3	-2.6	-11.7 ~ +1.5	-6.9	-6.7
	2015	-5.6 ~ +5.6	+0.8	+0.9	-9.2 ~ +1.2	-1.6	-1.1	-5.0 ~ +7.4	+1.9	+1.7

a, arithmetic mean

b, mass-weighted mean

Table 3. Dry deposition fluxes of water-soluble nitrogen species.

		NH ₄ ⁺ (μmol N m ⁻² d ⁻¹)			NO ₃ ⁻ (μmol N m ⁻² d ⁻¹)			WSN (μmol N m ⁻² d ⁻¹)		
		Range	Mean ^a	Mean ^b	Range	Mean ^a	Mean ^b	Range	Mean ^a	Mean ^b
ECSs (Sea fog)	2014	19.7 ~ 67.2	38.1	37.8	277 ~ 1931	926	951	34 ~ 446	147	146
	2015	2.2 ~ 48.7	10.9	9.8	52.1 ~ 412	161	151	5.9 ~ 338.9	98.5	72.5
NWPO (Dust)	2014	8.2 ~ 14.1	11.9	11.9	136 ~ 250	172	170	14.6	14.6	14.6
	2015	1.8 ~ 12.4	5.0	5.3	60.0 ~ 218	98.8	101	9.5 ~ 69.7	30.2	29.1
NWPO (Bgd.)	2014	1.4 ~ 21.1	4.7	4.5	11.0 ~ 287	44.6	43.3	2.7 ~ 23.4	10.5	9.9
	2015	0.6 ~ 2.6	1.4	1.3	4.9 ~ 60.6	21.9	23.1	4.4 ~ 34.8	11.5	11.6

a, arithmetic mean

b, volume-weighted mean

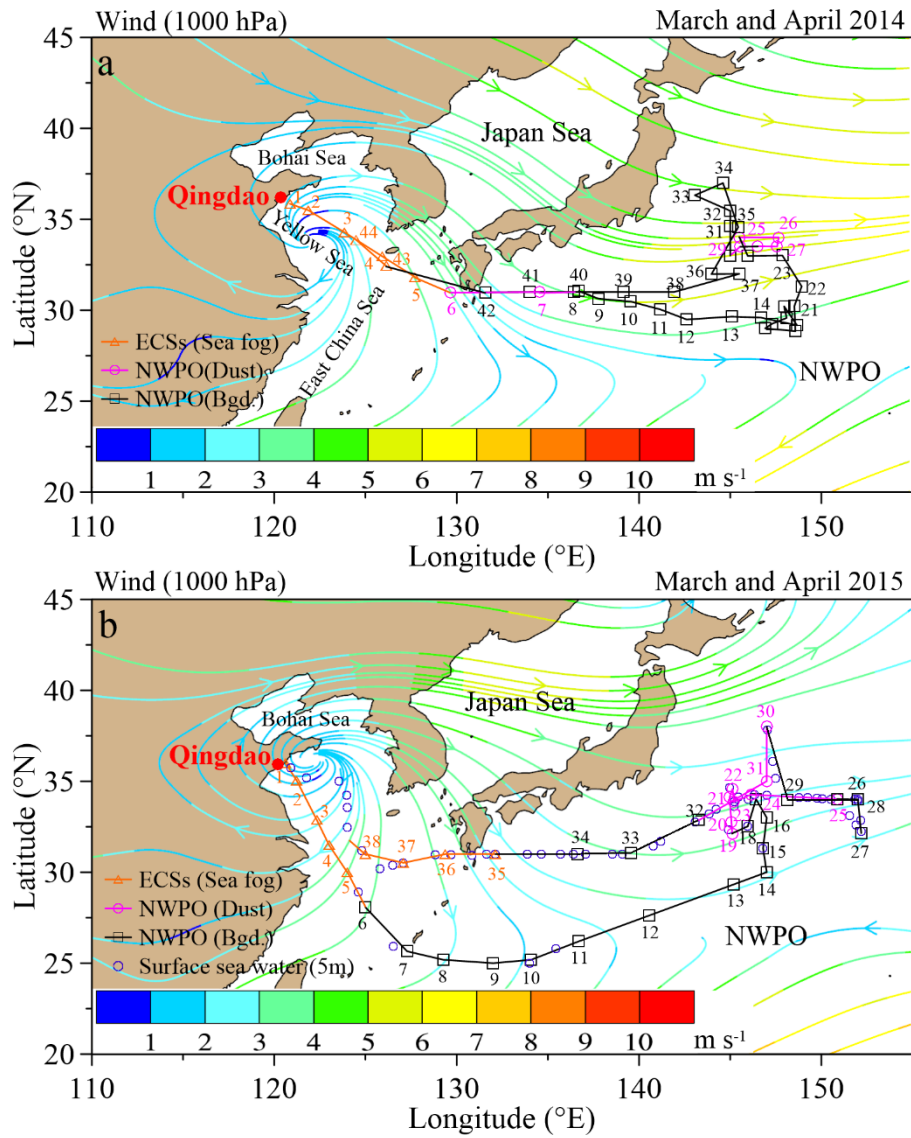


Figure 1. Regional wind streamlines (in m s^{-1}) at 1000 hPa during the Asian winter monsoon period (a, March and April in 2014 and b, March and April in 2015) based on an NCEP dataset. Cruise tracks are also shown (orange, pink, and black indicate sea fog, dust, and background aerosol, respectively). The aerosol number and collection range are shown in orange, pink, and black for sea-fog modified, dust, and background aerosol, respectively. The blue open circles in (b) indicate the locations of surface sea water sample (at a depth of 5 m) during the 2015 cruise.

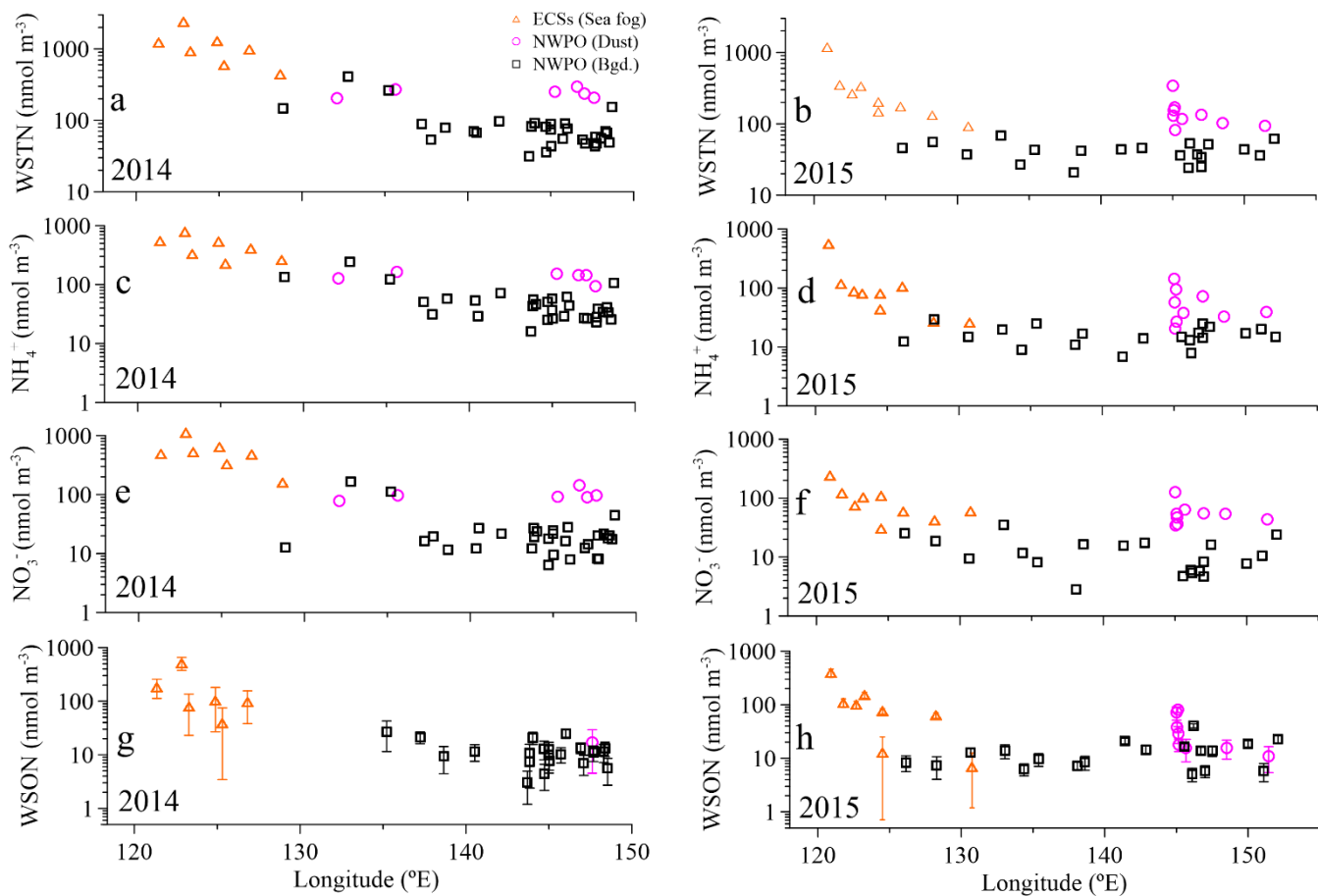


Figure 2. Concentrations of aerosol WSTN (a and b), NH_4^+ (c and d), NO_3^- (e and f), and WSON (g and h) with longitude for the 2014 and 2015 cruises. The orange open triangles denote sea-fog-modified aerosol in the ECSs, the pink circles denote dust aerosol, and the black open squares denote background aerosol in the NWPO. The error bars in (g) and (h) were the uncertainties of WSON caused by error propagation during the calculation.

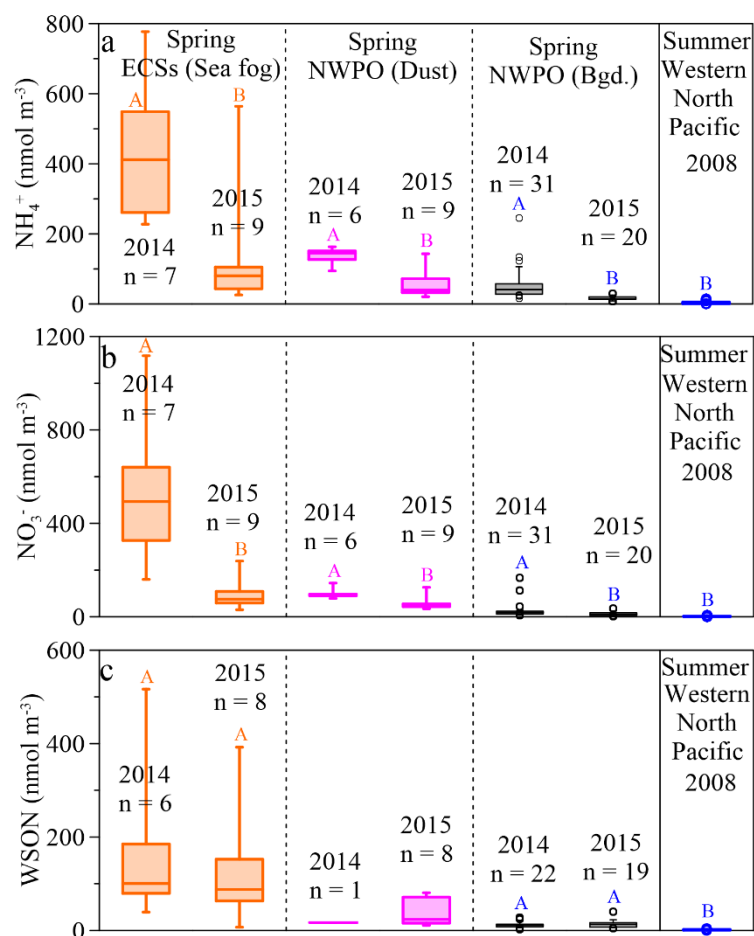


Figure 3. Box plots for spring concentrations of aerosol NH_4^+ (a), NO_3^- (b) and WSON (c) in the ECSs (the orange boxes denote sea-fog-modified aerosol) and NWPO (the pink boxes denote dust aerosol and the black boxes denote background aerosol), and summer aerosol (the blue box denotes NH_4^+ and NO_3^- from Miyazaki et al. (2011) and Jung et al. (2013) and the WSON from Miyazaki et al. (2011)). The large boxes represented the interquartile range from the 25th to 75th percentile, the line inside the box indicates the median value, and the whiskers extend upward to the 90th and downward to the 10th percentiles. Significant differences at the $p < 0.05$ level between different years are marked with coloured uppercase letters.

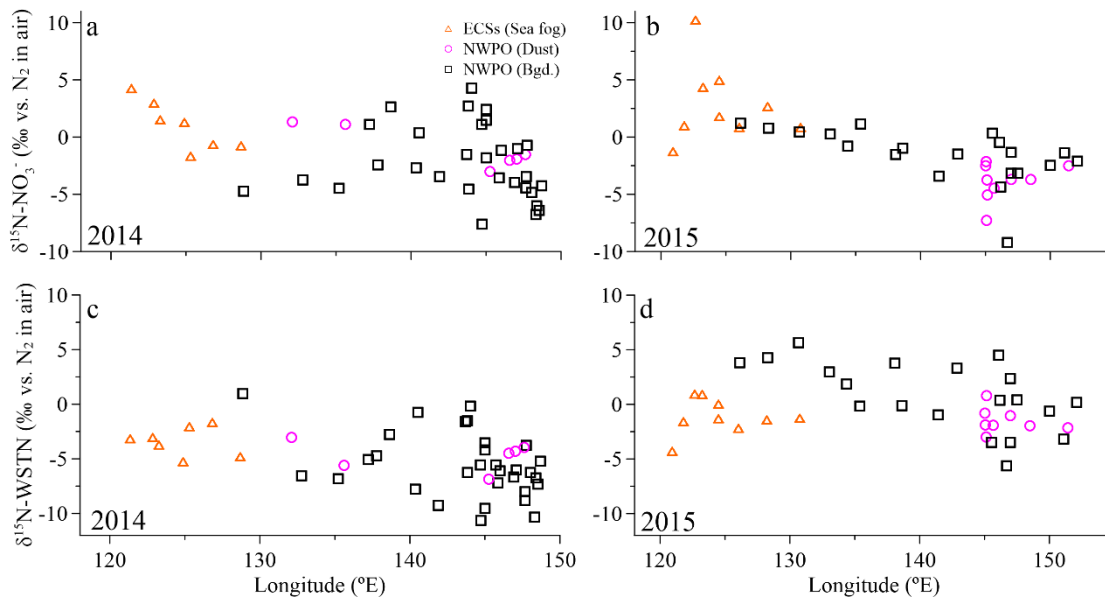


Figure 4. Aerosol $\delta^{15}\text{N-NO}_3^-$ (a and b) and $\delta^{15}\text{N-WSTN}$ (c and d) in the ECSs (the orange open triangles denote for sea-fog-modified aerosol), and in the NWPO (the pink open circles denote dust aerosol and the black open squares denote background aerosol) with longitude during the 2014 and 2015 cruise, respectively.

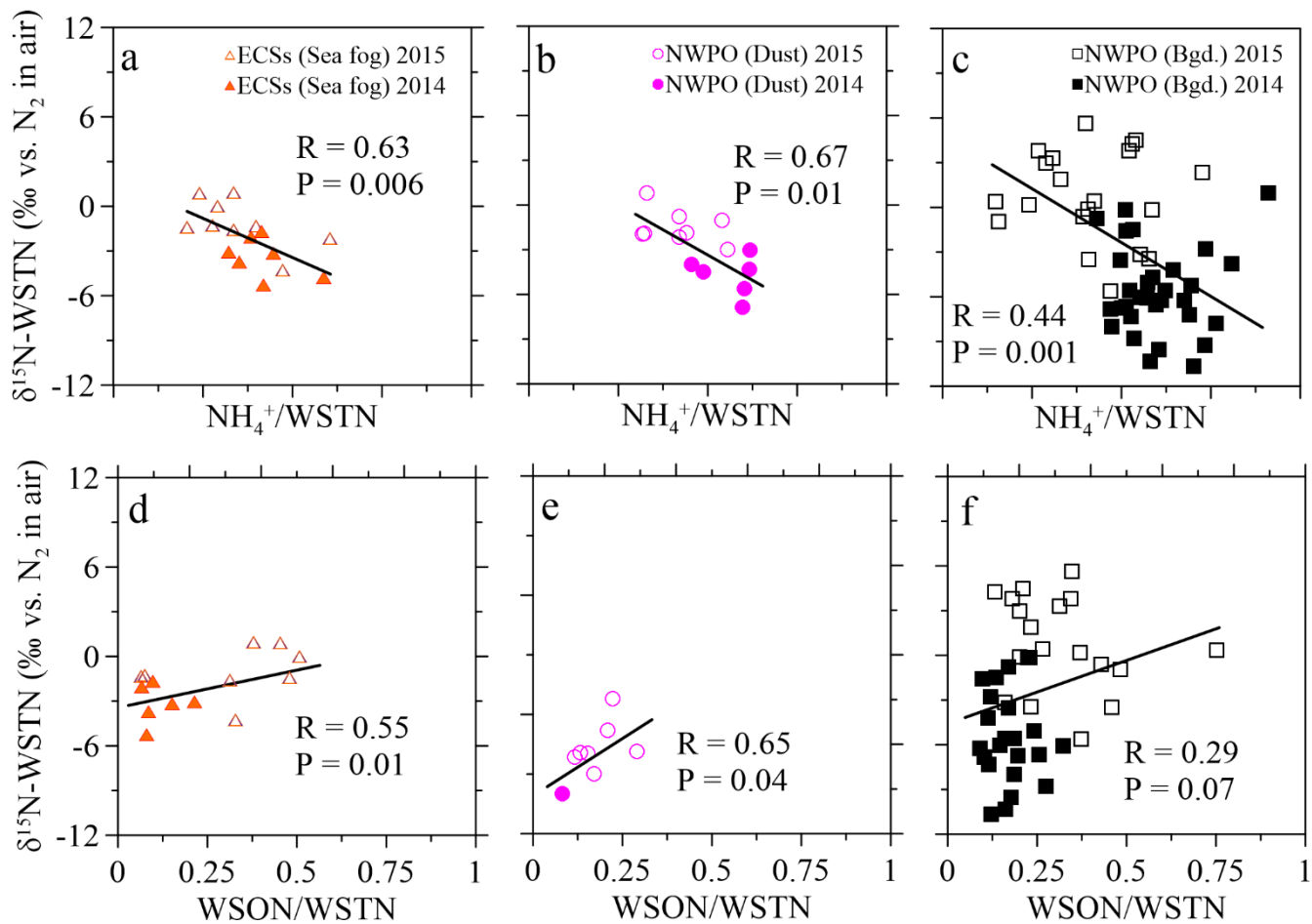


Figure 5. Scatter plots of $\delta^{15}\text{N-WSTN}$ against the $\text{NH}_4^+/\text{WSTN}$ ratio in (a) aerosol sampled in the ECSs, (b) dust aerosol and (c) background aerosol collected in the NWPO. Scatter plots of aerosol $\delta^{15}\text{N-WSTN}$ against the WSON/WSTN ratio in the (d) ECSs, (e) dust aerosol and (f) background aerosol in the NWPO. The solid and open symbols indicate aerosol sampled in 2014 and 2015, respectively.

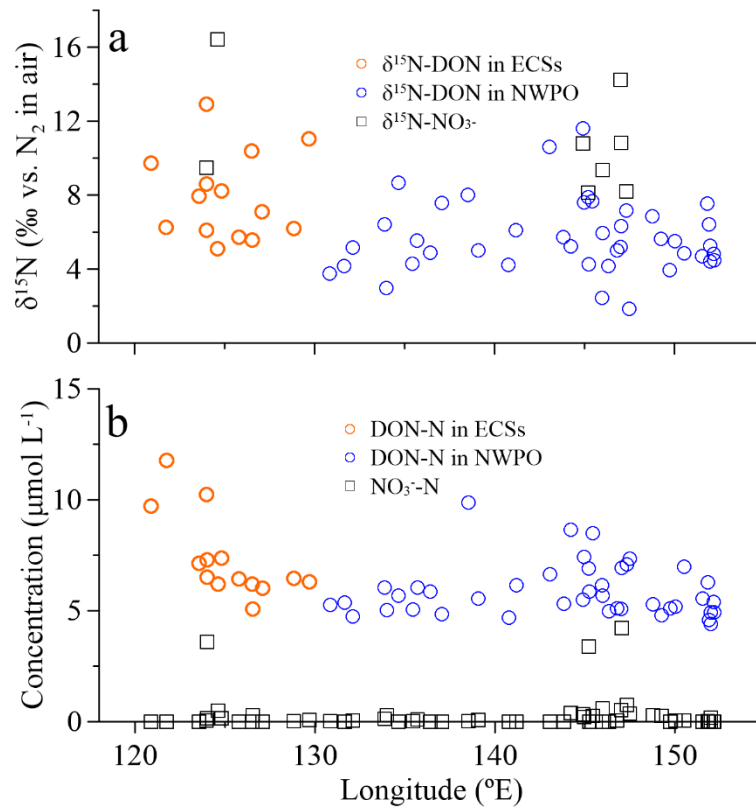


Figure 6. (a) $\delta^{15}\text{N}$ -DON (open circles) and $\delta^{15}\text{N}$ - NO_3^- (black squares), and (b) concentrations of DON (open circles) and NO_3^- (black squares) in SSW with longitude during the 2015 cruise.

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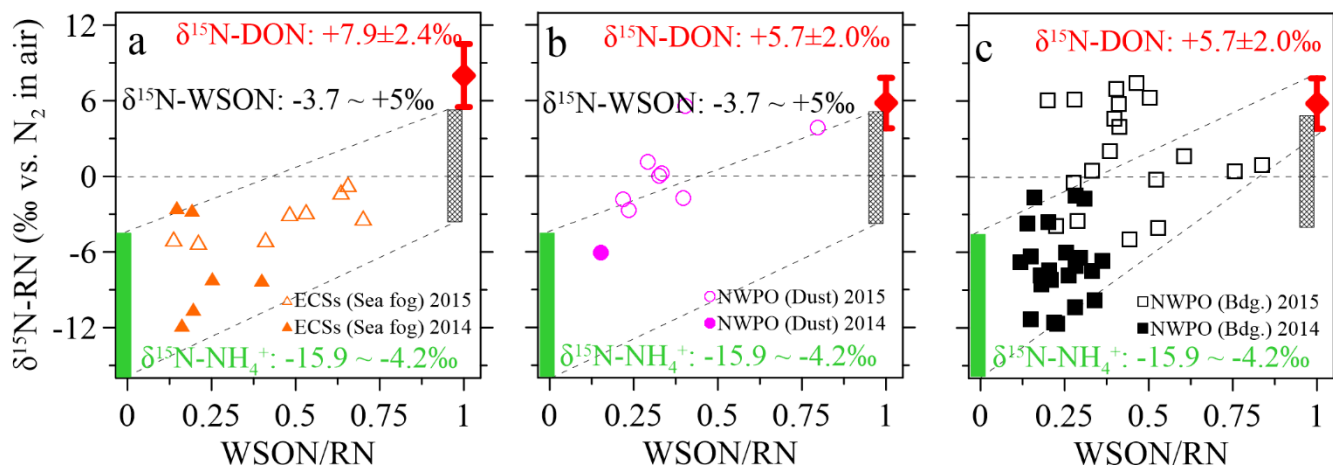


Figure 7. Scatter plots of aerosol $\delta^{15}\text{N-RN}$ against the WSON/RN ratio in the (a) ECSs, (b) dust aerosol in the NWPO, and (c) background aerosol in the NWPO. The green bar indicates the sources of anthropogenic, terrestrial, and oceanic $\delta^{15}\text{N-NH}_4^+$, the gray bar indicates the sources of terrestrial and anthropogenic $\delta^{15}\text{N-WSON}$, and the red bar indicates the $\delta^{15}\text{N-DON}$ in SSW.

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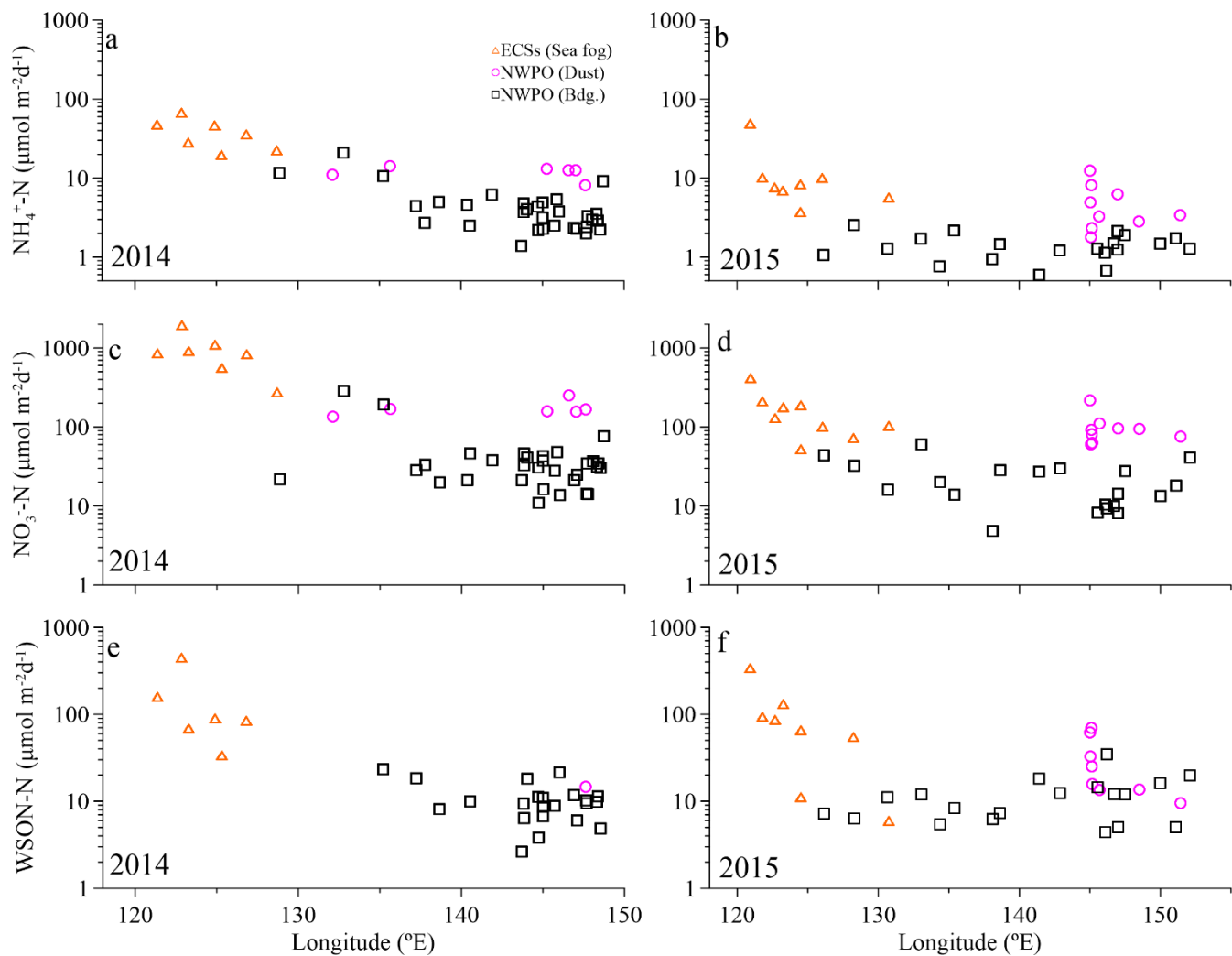


Figure 8. Dry deposition of aerosol NO_3^- -N (a and b), NH_4^+ -N (c and d) and WSON-N (e and f) in the ECSs (orange open triangles) and in the NWPO (pink open circles for dust aerosol and black open squares for background aerosol) along longitude during the 2014 and 2015 cruise, respectively.