A European aerosol phenomenology-6: Scattering properties of atmospheric aerosol particles from 28 ACTRIS sites

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1 Abstract

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3 This paper presents the light scattering properties of atmospheric aerosol particles measured over 4 the past decade at 28 ACTRIS observatories which are located mainly in Europe. The data include 5 particle light scattering (σ_{sp}) and hemispheric backscattering (σ_{bsp}) coefficients, scattering Ångström 6 exponent (SAE), backscatter fraction (BF) and asymmetry parameter (g). An increasing gradient of 7 σ_{sp} is observed when moving from remote environments (Arctic/mountain) to regional and to urban 8 environments. At regional level in Europe, σ_{sp} also increases when moving from Nordic and Baltic 9 countries and Western Europe to Central/Eastern Europe whereas no clear spatial gradient is 10 observed for other station environments. The SAE does not show a clear gradient as a function of 11 the placement of the station. However, a West to East increasing gradient is observed for both 12 regional and mountain placements suggesting a lower fraction of fine-mode particle in Western/Southwestern Europe compared to Central and Eastern Europe where the fine-mode 13 14 particles dominate the scattering. The q does not show any clear gradient by station placement or 15 geographical location reflecting the complex relationship of this parameter with the aerosol 16 particles physical properties. Both the station placement and the geographical location are important factors affecting the intra-annual variability. At mountain sites, higher σ_{sp} and SAE values 17 are measured in the summer due to the enhanced boundary layer influence and/or new particles 18 19 formation episodes. Conversely, the lower horizontal and vertical dispersion during winter leads to 20 higher σ_{sp} values at all low altitude sites in Central and Eastern Europe compared to summer. 21 These sites also show SAE maxima in the summer (with corresponding *q* minima). At all sites, both 22 SAE and g show a strong variation with aerosol particle loading. The lowest values of g are always 23 observed together with low σ_{sp} values, indicating a larger contribution from particles in the smaller 24 accumulation mode. During periods of high σ_{sp} values, the variation of g is less pronounced 25 whereas the SAE increases or decreases, suggesting changes mostly in the coarse aerosol 26 particle mode rather than in the fine mode. Statistically significant decreasing trends of σ_{sp} are 27 observed at 5 out of the 13 stations included in the trend analyses. The total reductions of σ_{sp} are 28 consistent with those reported for PM_{2.5} and PM₁₀ mass concentrations over similar periods across 29 Europe.

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31 **1. Introduction**

Atmospheric aerosol particles are recognized as an important atmospheric constituent which have demonstrated effects on climate and health. The radiative forcing of aerosol particles, estimated as -0.9 [-1.9 to -0.1] W/m² (IPCC, 2014), has two competing components: a cooling effect from most particle types and a partially offsetting warming contribution from black carbon (BC) particle light absorption of solar radiation. The aerosol cooling is the dominant effect; thus aerosol particles counteract a substantial portion of the warming effect from well-mixed greenhouse gases (GHGs). This process is driven by the scattering properties of most aerosol particle types (e.g. secondary sulphate and nitrate particles, mineral and organic matter), which reduce the amount of solar
 radiation reaching the Earth's surface, instead reflecting it back into space thus modifying the
 Earth's radiative balance.

4 However, the high temporal and spatial variability of atmospheric aerosol particles, due to the 5 wide variety of aerosol sources and sinks, together with their short and variable lifetimes (hours to weeks in the planetary boundary layer) and spatial non-uniformity, constitute the largest 6 7 uncertainties in the estimation of the total radiative forcing. Reducing these uncertainties is 8 mandatory in view of the global warming the planet has experienced over the past 50 years. In 9 fact, there is evidence suggesting that the observed (and projected) decrease in emissions of 10 anthropogenic aerosol particles in response to air quality policies will eventually exert a positive 11 aerosol effective radiative forcing at the top of the atmosphere (Rotstayn et al., 2013). Thus, 12 current emission controls could both enhance climate warming while improving air quality (e.g. 13 Stohl et al., 2015).

14 The measurements of aerosol particle optical properties, such as light scattering and 15 absorption, together with measurements of their physical and chemical properties, are fundamental 16 in order to better understand the current trade-off between the impacts of aerosols on 17 environmental health and the Earth's climate. In recent decades, several international projects 18 have provided important information on atmospheric particle properties worldwide. Near-surface in 19 situ observations of aerosol particle properties are being made worldwide under the GAW/WMO 20 (Global Atmosphere Watch; http://www.wmo.int/pages/prog/arep/gaw/gaw home en.html) 21 program and are complemented with policy-oriented programs such as IMPROVE (Interagency 22 Monitoring of Protected Visual Environments; http://vista.cira.colostate.edu/Improve/) in the United 23 States and EMEP (European Monitoring and Evaluation Programme; http://www.emep.int/) in 24 Europe. Additional information specifically targeting advanced aerosol particle properties have 25 been obtained in Europe using information from the European research infrastructure ACTRIS (Aerosols, Clouds, and Trace gases Research InfraStructure; http://www.actris.eu) and from short-26 27 term RTD (Research and Technological Development) projects such as EUCAARI (European 28 Integrated Project Aerosol Cloud Climate and Air Quality Interactions: on 29 http://www.cas.manchester.ac.uk/resprojects/eucaari/).

The implementation of the GAW program in Europe is performed under ACTRIS in regard to the advanced observation of aerosol particle properties. ACTRIS provides harmonized measurements of different (physical, chemical and optical) aerosol properties in a systematic way at major observation sites across Europe. More than 60 measuring sites worldwide are currently providing ground-based in situ aerosol particle light scattering measurements (EBAS database; www. http://ebas.nilu.no/) and this number has increased substantially in the last decade. However, EBAS also includes data from the IMPROVE network nephelometers, which latter are

operated at ambient conditions with no size cut, as a result of which these IMPROVE data are not
 directly comparable to the ACTRIS dataset discussed in this investigation.

3 The objective of this work is to integrate the total aerosol light scattering coefficient (σ_{sp}) and hemispheric backscattering coefficient (σ_{bsp}) measurements performed over several years at the 4 5 ground based in situ ACTRIS stations. A total of 28 stations (26 European + 2 non-European) are included in order to document the variability in near-surface aerosol particle light scattering across 6 7 the ACTRIS network. Moreover, at some of the ACTRIS stations more than 10 years of σ_{sp} data are available, allowing us to perform trend analyses. The study of the trend of σ_{sp} is important 8 9 given that a decreasing or increasing trend of σ_{sp} over time would be indicative of the effectiveness 10 of the air quality control measures. In fact, many studies have shown that the concentrations of 11 particulate matter (PM), and other air pollutants such as sulphur dioxide (SO₂) and carbon 12 monoxide (CO), have clearly decreased during the last 20 years in many European countries 13 (Barmpadimos et al., 2012; Cusack et al., 2012; EEA, 2013; Querol et al., 2014; Guerreiro et al., 14 2014; Pandolfi et al., 2016, Tørseth et al., 2012, among others).

15 Previous studies presenting multi-site ground-based in situ aerosol particle optical 16 measurements were, for example, performed by Delene and Ogren (2002), Sherman et al. (2015), 17 Collaud Coen et al. (2013) and Andrews et al. (2011). Delene and Ogren (2002) and Sherman et al. (2015) reported on the variability of aerosol particle optical properties at four North American 18 19 surface monitoring sites. Collaud Coen et al. (2013) presented long term (>8-9 years) aerosol 20 particle light scattering and absorption measurements performed at 24 regional/remote 21 observatories located mostly in the United States (although 5 are located in Europe). Andrews et 22 al. (2011) reported aerosol particle optical measurements performed at 12 mountain top 23 observatories (4 of which are located in Europe, 5 in the United States and Canada and 3 in Asia).

24 Our work is focused mainly on European observatories and is aimed at presenting a 25 representative phenomenology of aerosol particle light scattering coefficients measurements at 26 ACTRIS stations. Thanks to the establishment of European monitoring networks and/or research 27 projects, five papers relating to aerosol phenomenology have been published in Europe: Van 28 Dingenen et al. (2004) and Putaud et al. (2004), respectively, on the physical and chemical 29 characteristics of particulate matter (PM) at the kerbside, urban, rural and background sites in 30 Europe; Putaud et al. (2010) on the physical and chemical characteristics of PM measured at 60 31 sites across Europe; Cavalli et al. (2016) on the harmonized concentrations of carbonaceous 32 aerosols at ten regional background sites in Europe; and Zanatta et al. (2016) presenting a 33 climatology of BC optical properties at nine European regional background sites. The importance 34 of these studies and of the present work rests on the premise that a reliable assessment of the 35 physical, chemical and optical properties of aerosol particles at a European scale is of crucial 36 importance for an accurate estimation of the radiative forcing of atmospheric aerosols. This work is the first European phenomenology study dedicated to the light scattering properties of aerosol 37

particles measured in situ at near-surface ground-based observatories. Moreover, the trend analyses presented can be used to evaluate how the European mitigation strategies adopted to improve air quality have impacted aerosol particle optical properties.

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5 2. Experimental

6 **2.1 Atmospheric Observatories**

7 Figure 1 shows the location of the observatories which are grouped according to their geographical 8 locations, a grouping employed in other European phenomenology studies (e.g. Putaud et al., 9 2010). Observatory information (country, code, coordinates, altitude, geographical location, among 10 others) and measurement periods are summarized in Table 1. The observatories are also divided 11 into five different categories depending on their placement within each geographical sector. Arctic: includes stations located in the Arctic/sub-Arctic region; mountain: includes those observatories 12 13 located at more than 985 m above sea level (the lowest altitude among the mountain observatories 14 included here); coastal: includes observatories located close to the coast (<1-4 km); regional/rural: 15 includes those observatories that are representative of large regional areas; urban/sub-urban: 16 includes observatories located in a background of an urban or suburban area. Two non-European 17 stations are also included; one Antarctic site and one mountain site in Bolivia. Given that this work 18 mainly focuses on European ACTRIS observatories, the results from these two non-European 19 stations are reported in the Supporting Information.

The altitude of the mountain stations considered here range between 985 m at HPB and 5240 m at CHC (cf. Table 1). Some of the mountain stations included in this investigation have already been included in the work of Andrews et al. (2011), namely IZO, JFJ, CMN, and BEO. Moreover, the FKL, HPB, JFJ, MHD and PAL stations have been included in the study by Collaud Coen et al. (2013). Both studies presented in situ aerosol particle optical measurements taken at these stations. The main results of these previous investigations are summarized in the results section.

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28 2.2 Scattering measurements

29 2.2.1 Instruments

The measurements of σ_{sp} and σ_{bsp} included in this study were obtained from TSI and Ecotech integrating nephelometers (Table 1). These optical instruments measure the amount of light scattered by particles in the visible spectrum and provide σ_{sp} and σ_{bsp} coefficients of the sampled aerosols. The most common nephelometers in the ACTRIS program are the TSI3563 and the Ecotech AURORA3000 nephelometers, both of which provide both σ_{sp} and σ_{bsp} . Model TSI3563 measures σ_{sp} and σ_{bsp} at 450, 550 and 700 nm whereas the Ecotech AURORA3000 measures at

1 450, 525 and 635 nm. Other models used are the M9003 from Ecotech (SIR and CMN) and the RR 2 (Radiance Research) nephelometer model M903 (FKL) measuring σ_{sp} at 520 nm and 532 nm, respectively. Due to the non-homogeneity of the angular distribution of light intensity of model 3 M9003 (cf. Müller et al., 2009), the light source was changed at SIR in 2013 with the 4 5 AURORA3000 light source and at CMN in 2009 with an opal glass light source. After the change of the light sources, both nephelometers were examined at the World Calibration Center for Aerosol 6 7 Physical properties in Leipzig and both performed very well (personal communication from CMN 8 and SIR data providers). The detailed description of the main characteristics and the working 9 principle of the integrating nephelometers can be found e.g. in Müller et al. (2011) for the Ecotech 10 AURORA3000 and in Anderson and Ogren (1998) for the model TSI 3563.

11 Recommended quality assurance procedures during on-site operation, as described in GAW 12 (WMO-GAW Report, 2016), help to ensure the quality and comparability of the data. The 13 nephelometers included in this investigation are regularly calibrated using span gas and are zero 14 adjusted using particle-free air. Additionally, most of the integrating nephelometers employed in 15 ACTRIS have undergone a schedule of performance checks at the World Calibration Center for 16 Aerosol Physics of ACTRIS/GAW.

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18 **2.2.2 Data treatment**

19 Data used in this investigation include hourly averaged Level 2 aerosol particle scattering data 20 downloaded from the ACTRIS/EBAS Data Centre web portals (www.actris.nilu.no; 21 www.ebas.nilu.no; last downloads August 2017). The σ_{sp} and σ_{bsp} data reported to EBAS and used 22 in this work are referenced to standard T (273.15 °C) and P (1013 hPa) conditions. Data 23 consistency is critical when comparing many years' worth of data from different stations. In this 24 work, the Level 2 scattering data were further reviewed in order to ensure a high quality of the data 25 presented. There are however station-to-station differences (e.g. sizecut, RH control, wavelength, 26 data processing, etc.) which are addressed in the sections below.

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28 **2.2.2.1 Truncation correction**

29 Data from the integrating nephelometers used here are corrected for non-ideal illumination of the 30 light source (deviation from a Lambertian distribution of light) and for truncation of the sensing 31 volumes in the near-forward (around 0-10°) and near-backward (around 170-180°) directions 32 (Müller et al., 2009 and Anderson and Ogren, 1998). Correction schemes have been provided by 33 Müller et al. (2009; 2011) for the RR M903 and Ecotech models M9003 and AURORA3000, and by 34 Anderson and Ogren (1998) for the TSI3563. These schemes consist of a simple linear correction 35 based on the scattering Angström exponent (SAE) determined from the raw nephelometer data to 36 take account of the size-distribution-dependent truncation error. It has been demonstrated that these simple correction schemes are accurate for a wide range of atmospheric aerosols and that
the uncertainties in the corrections are not expected to be larger than 2% for an aerosol particle
population with a single scattering albedos (SSA) greater than 0.8 (Bond et al., 2009).

4 The majority of the σ_{sp} data in the EBAS database are corrected for non-ideal illumination 5 and for truncation by the data providers. Exceptions are the scattering data submitted for KOS, MHD, PLA, CMN, FKL and SIR. Scattering data from KOS, MHD and PLA were corrected in this 6 7 work using the correction scheme provided by Anderson and Ogren (1998) (cf. Table S1 of the 8 Supporting Material). The σ_{sp} data collected at CMN, FKL and SIR are not corrected because the 9 nephelometers deployed at these three stations provide scattering only at one wavelength, thus 10 preventing the estimation of the SAE. Given that the nephelometer correction factors vary as a 11 function of SAE, the assumption of a constant correction factor to correct the 1- λ scattering data 12 could introduce undesired noise. Moreover, at SIR and CMN, the σ_{sp} is measured with the single 13 wavelength Ecotech nephelometer model M9003 (until 2013 at CMN). The correction curve from 14 Müller et al. (2009; Figure 4) provides a correction factor of around 0.97 to 1.0 for the M9003 for a 15 SAE of around 1.5 to 2. Using the TSI3563 scattering measurements performed at CMN during 16 2014-2015, we estimated a mean SAE of around 2 for CMN (cf. Table S5). Thus, given the rather 17 small effect of the correction factor estimated for the Ecotech M9003, scattering data from CMN 18 and SIR were not corrected in this work. At FKL the nephelometer models RR M903 (until 2011) 19 and Ecotech 1000 (from 2012) were used (cf. Table 1). To the best of our knowledge, no correction 20 scheme has been provided for the Ecotech 1000. Moreover, at FKL, the inlet was changed many 21 times (cf. Table 1) and the correction factors provided in the literature are a strong function of the 22 size cut-off used. For these reasons, scattering data collected at FKL are not corrected in this 23 investigation.

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25 2.2.2.2 Relative humidity

26 The integrating nephelometer measurements within ACTRIS and WMO-GAW should be performed 27 at a low relative humidity (RH<40%) in order to avoid enhanced scattering due to water uptake of 28 aerosol particles and in order to make the measurements comparable. For the Ecotech integrating 29 nephelometers, the RH threshold can be set by using a processor-controlled automatic heater 30 inside the instrument. At some mountain sites, where whole air is sampled (cf. Table 1), the natural 31 temperature difference between the outside and inside air dries cloud droplets to the aerosol phase 32 when a cloud is present at the station. RH is also controlled by de-humidifying in the inlet pipe, as 33 reported in GAW report 226, to ensure a sampling RH of less than 40%. This recommendation is 34 intended to ensure that the data are comparable across the network, as measurements would 35 otherwise would be a strong function of the highly variable sample RH. Currently, at the majority of 36 ACTRIS observatories, the aerosol particle light scattering measurements are performed at a RH

below 40%. However, given that at some stations the 40% RH threshold is sometimes exceeded,
 we selected in this work a RH threshold of 50% in order to improve the data coverage.

3 Estimating the aerosol particle light scattering enhancement due to an increase of RH from 4 40% to 50% is difficult using the data available here because the σ_{sp} measurements at a RH>40% 5 are not evenly distributed over the measurement periods, with the majority of the stations registering a RH higher than 40% during the summer. Moreover, the chemical composition of 6 7 atmospheric aerosol particles is an important factor determining the magnitude of the scattering 8 enhancement due to water uptake, which can then change from one site to another (e.g. 9 Fierz-Schmidhauser et al., 2010a,b; Zieger et al., 2014, 2017). However, the scattering 10 enhancement due to a change in RH between 40% and 50% should be small and will not exceed 11 few percent even for more hygroscopic particles (e.g. Fierz-Schmidhauser et al., 2010a,b). Table 12 S2 in the Supporting Material reports the percentage of hourly σ_{sp} values collected in the range 40%<RH<50% whereas the frequency distributions of the measured RH are shown in Figure S1. 13

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15 2.2.2.3 Available wavelengths

16 In this work we present and discuss the σ_{sp} , backscatter fraction (BF) and asymmetry parameter 17 (q) measurements obtained using the green wavelength of the integrating nephelometers. The 18 available wavelengths ranged from 520 nm (2 stations; CMN and VHL) to 550 nm (18 stations). 19 Other wavelengths used are 525 nm (6 stations) and 532 nm (used at FKL until 2010; cf. Table 2). An exception is SIR, where only σ_{sp} values at 450 nm are available. The measurements of σ_{sp} 20 21 reported here are not adjusted to 550 nm, which is generally the most common wavelength (e.g. 22 And rews et al., 2011) because of the different data availability of σ_{sp} and SAE at the measuring 23 stations. As discussed in the following sections, the SAE is calculated for σ_{sp} data higher than 0.8 24 Mm^{-1} , thus leading to different data coverage for σ_{sp} and SAE and preventing the adjustment of all 25 measured σ_{so} to 550 nm. Moreover, the SAE is not available at FKL and SIR (or at CMN until 26 2014) thus preventing any wavelength adjustment at these stations. Using the mean SAE 27 calculated at those stations, where σ_{sp} is measured at wavelengths in addition to 550 nm (cf. Tables S4 and S5 in Supporting material), we estimate differences in the σ_{sp} values of less than 28 29 6% after adjusting to 550 nm. At FKL and SIR, where the SAE is not available, and assuming a 30 reasonable SAE range between 1.5 and 1.0, the difference due to the adjustment to 550 nm is 4.9-31 3.0% at FKL and 26-18% at SIR. The higher difference at SIR is due to the fact that measurements 32 at this station are performed at 450 nm. Finally, at CMN, the effect of the adjustment of σ_{sp} to 550 33 nm (from 520 nm) using a mean SAE of 2 (calculated using the $3-\lambda$ nephelometer data from 2014; cf. Table S5) is below 10%. 34

1 **2.2.2.4 Inlet size cut changes**

It should be noted that any comparison of the σ_{sp} and SAE values among the different stations and the presented trend analyses could be slightly biased by the different particle size cuts upstream of the integrating nephelometers used in this work (cf. Table 1). Currently, all ACTRIS integrating nephelometers measure whole air or PM₁₀, with the exception of SIR, where the PM₁ inlet is used. Whole air is currently measured at mountain observatories (BEO, CMN, JFJ, PUY, CHC), one coastal observatory (MHD) and one urban observatory (UGR) (cf. Table 1).

8 At some stations, the inlet was changed from whole air to PM₁₀ at some point, namely at OPE, FKL 9 and TRL. Given the lower scattering efficiency of aerosol particles larger than 10 µm, no important 10 differences in the aerosol particle optical parameters should be expected between aerosol particles 11 sampled with a whole air and a PM₁₀ cut-off. At the other stations the inlet was changed during the 12 measurement period from a cut-off lower than 10 µm (1 µm at KPS; 2.5 µm or 5 µm at PAL, MSA and MAD) to PM₁₀. For PAL (where a median SAE of around 1.8 was measured; cf. Paragraph 3.2 13 14 and Table S5), Lihavainen et al. (2015a) assumed that the inlet changes (from PM_5 to $PM_{2.5}$ in 15 2005 and from PM_{2.5} to PM₁₀, cf. Table 1) had only minor effects on scattering because the number 16 concentration of coarse particles is very low at PAL. Similarly, the KPS observatory registers 17 among the highest SAE values observed in the network (median value of around 2; cf. Paragraph 18 3.2 and Table S5) suggesting an aerosol particle size distribution dominated by fine particles. 19 Moreover, at KPS, the inlet was changed in April 2008, less than 1.5 years after the measurements 20 commenced, and thus likely has also a minor effect in the trend analyses and climatology 21 performed at this site over the period 2006 to 2014. Two stations (MSA and MAD) changed the 22 inlet from a PM_{2.5} diameter cut-off to PM₁₀. For these two Southern European stations the inlet 23 change may have had an effect on the SAE, especially during Saharan dust outbreaks, which are 24 however sporadic events. Finally, the FKL observatory was removed from the trend analysis because the inlet was changed from whole air to PM_{10} in 2009, from PM_{10} to PM_1 in 2011 and 25 again from PM₁ to PM₁₀ in 2013 (cf. Table 1). These events likely had a major effect on the 26 27 measured particle optical properties.

A sensitivity study (not shown) was performed to assess the effect of the inlet changes on the SAE values measured at the aforementioned stations. We looked at the climatology of SAE for different inlet sizes and for different time periods (with and without inlet size changes) and we did not observe any obvious change in the climatology as a function of size cut due to interannual variability. Thus, despite the differences in the particle diameter cut-off, the comparison between the different stations seems feasible.

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1 **2.2.3 Calculation of aerosol particle intensive optical properties**

2 Starting from the spectral σ_{sp} measurements performed at the ACTRIS observatories, three 3 intensive aerosol particle optical parameters can be estimated, namely; the scattering Angström 4 exponent (SAE), the backscattering fraction (BF) and the asymmetry parameter (g). These 5 intensive properties do not depend on the PM mass concentration and are directly related to aerosol particle properties such as size, shape, size distribution and chemical composition. The 6 7 SAE can be considered as a proxy for the aerosol particle size range with a higher (lower) SAE 8 associated with predominance of fine (coarse) aerosol particles (e.g. Seinfeld and Pandis, 1998; 9 Esteve et al., 2012; Valenzuela et al., 2015 among others). The BF and g parameters are 10 calculated quantities that influence the variability of the radiative forcing efficiency and that 11 represent the angular light scattering of aerosol particles. For computational efficiency, the angular 12 light scattering is often represented by a single value (BF, σ_{sp}/σ_{bsp} or g) (Andrews et al., 2006).

13 The SAE characterizes the wavelength dependency of σ_{sp} and it can be calculated as follows:

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$$SAE = -\frac{\log\left(\frac{\sigma_{sp}^{\lambda_1}}{\sigma_{sp}^{\lambda_2}}\right)}{\log\left(\frac{\lambda_1}{\lambda_2}\right)} \qquad (Eq. 1)$$

Here, the SAE is derived from a multispectral log linear fit based on the three nephelometer wavelengths. The SAE depends on the particle size distribution and takes values greater than 2 when the light scattering is dominated by fine particles (radii $\leq 0.5 \ \mu m$ as e.g. in Schuster et al. (2006)), while it is lower than one when the light scattering is dominated by coarse particles (Seinfeld and Pandis, 1998; Schuster et al., 2006).

20 The asymmetry parameter (g) (Andrews et al., 2006; Delene and Ogren, 2002) describes the 21 probability that the radiation is scattered in a given direction and it is defined as the cosine-22 weighted average of the phase function. Thus, g yields information regarding the amount of 23 radiation that a particle scatters in the forward direction compared to the backward direction. 24 Theoretically, the values of g can range from -1 for only back scattering to +1 for complete forward 25 scattering, with a value of 0.7 commonly used in radiative transfer models. The g parameter can be estimated from the backscatter fraction (BF), which is the ratio of σ_{bsp} and σ_{sp} (Andrews et al., 26 27 2006):

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$$g = -7.14(BF)^3 + 7.46(BF)^2 - 3.96(BF) + 0.9893$$
. (Eq. 2)

29

30 2.2.4 Data coverage

Table S3 in the Supporting Material reports the percentage [%] of data coverage at the 28 ACTRIS stations included in this study. Removed data include data flagged as non-valid by the data 1 providers (instrument failure, calibration periods, unspecified contamination or local influence, etc.) 2 or obtained at a RH of greater than 50%. The data coverage for the extensive measured aerosol 3 particle optical properties (σ_{sp} and σ_{bsp}) is generally high, ranging from around 60% to 95%. 4 Exceptions are the σ_{so} measurements at CMN in the blue (450 nm) and red (700 nm) wavelengths which have much less data coverage compared to the green wavelength because the three 5 6 wavelength nephelometer was implemented at CMN in 2014. Consequently, also the SAE and g 7 have low data coverage at CMN. Moreover, lower data coverage (< 40%) was registered at PLA 8 and VHL.

9 The data coverage for the intensive aerosol particle optical properties (SAE and g) is generally lower compared to the data coverage of σ_{sp} and $\sigma_{\text{bsp}}.$ This is because the intensive 10 optical properties are calculated from hourly σ_{sp} and σ_{bsp} data higher than 0.8 Mm⁻¹ to avoid noise 11 12 in the calculations. As a consequence, the data coverage of the intensive properties is lower at 13 those stations measuring low σ_{sp} and σ_{bsp} values (e.g. mountain and remote sites). For example, at 14 JFJ, the SAE and g data coverages are around 54% and 22%, respectively. At TRL, these values 15 are even lower, at 21% and 1%, respectively. However, as reported in Table S3, at the majority of 16 the stations the data coverage of SAE and g is higher than 60%.

17

19

18 **3. Results/Discussion**

20 **3.1 Variability of** σ_{sp}

Figure 2 shows the box-and-whiskers plots of σ_{sp} measured at the stations included in this investigation. In Figure 2, the observatories are grouped based on their placement and ordered according to their geographical location. Table S4 and Figure S2 in the Supplementary Material report, respectively, the statistics of σ_{sp} (mean, standard deviation, minimum and maximum values and 5th, 25th, 50th, 75th, and 95th percentiles) and frequency and cumulative frequency distributions.

In each geographical sector, an increasing gradient of σ_{sp} is generally observed when moving from mountain to regional and to urban sites. Thus, the σ_{sp} values measured at mountain sites are lower than the measurements made at other locations (coastal to urban) even if exceptions are observed in some sectors.

A large range of σ_{sp} coefficients is observed across the network, ranging from median values lower than 10 Mm⁻¹ to values higher than 40 Mm⁻¹. Overall, the lowest σ_{sp} is on average measured at remote stations because of either: a) their altitude, for example JFJ is located in Central Europe at more than 3500 m a.s.l. and CHC in Bolivia is at around 5300 m a.s.l. (cf. Figure S3), or b) because of their large distance from pollution sources, for example the Arctic ZEP and PAL stations, TRL station (cf. Figure S3) and some regional sites in the Nordic and Baltic sector such

as BIR and SMR. Higher σ_{sp} values (medians > 40 Mm⁻¹) are on average registered at more polluted sites, such as some urban sites in Southern Europe (UGR and DEM), some regional sites in Eastern and Central Europe (KPS and IPR, respectively) and one coastal site in the Nordic and Baltic sector (PLA).

5 The observed variation is consistent with the differences in particulate matter (PM) mass 6 concentrations, PM chemical composition, particle number concentration and absorption 7 coefficients observed across Europe, as described for example by Putaud et al. (2010), Asmi et al. 8 (2011) and Zanatta et al. (2016).

9 Figures 3a and 3b show the relationship between the mean particle number concentration 10 measured at different stations during 2008 to 2009 (and reported in Asmi et al. (2011)) and the 11 mean σ_{sp} measured over the same period (where available). As reported in Figure 3, good 12 correlations are observed between N50 (Figure 3a: mean/median particle number between 50 nm 13 and 500 nm) and N100 (Figure 3b: mean/median particle number between 100 nm and 500 nm) 14 and mean σ_{sp} . Figure 3c shows the relationship (for some stations) between absorption coefficients 15 reported in Zanatta et al. (2016) and the total scattering. The good correlations reported in Figure 16 3c (especially high for the winter and autumn periods) suggest an increase of both scattering and 17 absorption coefficients with increasing aerosol loading. Figure 3c also reports the mean single scattering albedo (SSA). On average lower SSA is observed at IPR, whereas higher SSA is 18 19 observed at the Nordic and Baltic VHL and BIR observatories.

20 Finally, at all stations included in this work, the skewness of the σ_{sp} distributions (cf. Table 21 S4) is higher than one and ranges between 1.4 at PLA and 10.6 at TRL (skewness calculated from 22 hourly averaged data). The skewness can be used to evaluate the asymmetry of a distribution. 23 Positive skewness is usually observed for parameters which are defined to be positive and it 24 indicates that the tail on the right side of the distribution is longer or fatter than that on the left side. 25 Thus, for a right-skewed distribution, the mass of the distribution is concentrated on the left, and 26 there is a higher probability of measuring a high value compared to a left-skewed distribution. For 27 example Querol et al. (2009) used the skewness to assess the importance of Saharan dust 28 outbreaks on PM₁₀ levels measured at different sites across the Mediterranean basin. They found 29 a positive correlation between the calculated skewness and the net dust contribution to the 30 measured PM₁₀ concentration (i.e. the strength of dust pollution episodes; cf. Fig. 6 in Querol et al., 31 2009). Figure S2 in the Supporting Material shows the frequency and cumulative frequency 32 distributions for σ_{sp} for each station, evidencing the presence of these right-skewed tails.

33

34 **3.1.1** σ_{sp} at Arctic/Antarctic observatories

The Arctic (ZEP and PAL; cf. Fig 2) and Antarctic (TRL; cf. Figure S3) monitoring stations are located in undisturbed environments with minimal influence from the local settlement because they 1 are located above the inversion layer. The mean σ_{sp} values measured at ZEP and TRL are by far 2 the lowest across the network, whereas higher σ_{sp} values are measured at PAL. PAL is located in 3 a remote continental area characterized by the absence of large local and regional pollution sources (e.g. Aaltonen et al., 2006). However, Lihavainen et al. (2015a) reported that high values 4 5 of the absorption coefficient and low values of the single scattering albedo at PAL are related to 6 continental air masses from lower latitudes. Despite this, the mean σ_{sp} at PAL is among the lowest 7 in the ACTRIS network and is comparable to the mean σ_{sp} observed at the JFJ and CHC 8 mountaintop observatories (cf. Table S4).

9

10 **3.1.2** σ_{sp} at mountain observatories

11 Differences can be observed among stations with similar environments but different geographical 12 locations. For mountain observatories, a clear gradient is not observed when moving from West to 13 Southeast Europe, because the altitude of the station is an important parameter contributing to the σ_{sp} measured at these observatories. Among the mountain stations a higher mean σ_{sp} is on 14 15 average measured at HPB and IZO (cf. Table S4). The HPB station is likely to be more influenced 16 by the PBL than other mountain stations due to its lower altitude (Nyeki et al., 2012; Collaud Coen 17 et al., 2017), whereas IZO is largely influenced by Saharan dust outbreaks transporting dust 18 toward the station (e.g. Rodriguez et al., 2011) thus increasing σ_{sp} . In fact, at IZO, the median value of σ_{sp} is among the lowest measured at these mountain sites (around 7 Mm⁻¹; cf. Table S4) 19 20 indicating that the sporadic but extremely intense pollution episodes due to Saharan mineral dust 21 outbreaks strongly affect the mean σ_{sp} at this station.

22 Despite their placement at higher altitudes, both CMN and BEO (more than 2 km a.s.l.) register 23 similar σ_{sp} values compared to PUY and MSA (around 1.5. km a.s.l.) likely because the effect of 24 important regional pollution sources (i.e. the Po Valley for CMN) affecting, under favourable 25 meteorological conditions, these Central and Eastern European observatories (i.e. Marenco et al., 26 2004). Conversely, the region around the MSA observatory is sparsely populated and the station is 27 isolated from large urban and industrial agglomerations (i.e. Pandolfi et al., 2014; Ripoll et al., 28 2014; Ealo et al., 2016). PUY observatory is surrounded by a protected area with fields and forests 29 and previous works have shown that the influence of the Clermont-Ferrand city on the PUY 30 measurements remains too small to be detected (i.e. Asmi et al. 2011).

The lowest median σ_{sp} values at mountain sites are on average measured at JFJ, probably due to the higher altitude of this station compared to other mountain stations included in this work and/or its distance from important pollution sources. Moreover, Collaud Coen et al. (2017) reported a low PBL influence at this site due to the location of the station in a dominant position within the whole mountainous massif. CHC (cf. Figure S3) registers higher median σ_{sp} values compared to JFJ despite its location at around 5300 m a.s.l. likely due to the influence of the emissions from the city

of La Paz (3600 m a.s.l.), located around 30 km from the CHC site, and the local topography,
 which facilitates the uplift of air masses toward the CHC observatory (Collaud Coen et al., 2017).

3

4 3.1.3 σ_{sp} at coastal observatories

5 The PLA coastal station registered σ_{sp} values which are higher compared to both other Nordic and 6 Baltic stations and other coastal sites (e.g. MHD and FKL) and which are amongst the highest in 7 Europe. Kecorius et al. (2016) have shown that ship emissions in the Baltic Sea contribute strongly 8 to pollution levels at PLA and that up to 50% of particles arriving at PLA are generated by 9 processes and emissions, including shipping, taking place in areas upwind of the station. 10 Moreover, Asmi et al. (2011) presented a number of similarities in particle number concentrations 11 measured at PLA to those measured at some Central European sites, such as IPR, which are due 12 to the influence of multiple source areas (cf. Figure 3). It should be noted however, that the period 13 with available σ_{sp} measurements is very short at PLA (cf. Table 1 and Figure 7) and the data 14 coverage is also low (cf. Table S3). Consequently, more measurements at this site are needed in 15 order to confirm the σ_{sp} values reported there. The other two coastal stations (MHD and FKL) 16 register median σ_{sp} values in the upper range of σ_{sp} measured across the network, mostly due to 17 the contribution of marine aerosols in winter and mineral dust in summer at MHD and FKL, 18 respectively (cf. Section 3.5).

19

20 3.1.4 σ_{sp} at regional/rural observatories

21 Regional sites exhibit a large variability in σ_{sp} coefficients across Europe, with the lowest values 22 measured at BIR and SMR (Nordic and Baltic) and the highest at IPR (central Europe) and KPS 23 (Eastern Europe). Thus, a gradient is observed in σ_{sp} when moving from West to East Europe. At 24 both IPR and KPS, the frequent winter time episodes, linked to stable air due to strong thermal 25 inversions, affect the level of pollution at these sites (e.g. Putaud et al., 2014; Molnár et al., 2016). 26 It is known that at the IPR station, even though it lies several tens of kilometres away from large 27 pollution sources, is located in an area (the Po Valley) which is one of the most polluted regions in 28 Europe (e.g. van Donkelaar et al., 2010). The VHL observatory registers an on average higher σ_{sp} 29 compared to PAL and compared to the BIR and SMR regional sites likely because VHL is located 30 closer to the European continent and it is consequently more affected by polluted continental air 31 masses. Moreover, the emissions from the densely populated areas of Helsingborg and Malmö 32 and the city of Copenhagen, located 25 km to the west, 50 km to the south and 45 km to the south-33 east, respectively, could also explain the relatively high σ_{sp} measured at VHL (Kecorius et al., 34 2016). The σ_{sp} values at a regional level in Western Europe (OPE and CBW) are on average 35 higher compared to those measured in the Nordic and Baltic regions and lower compared to those measured at a regional level in Southern Europe (MSY). The relatively higher σ_{sp} values measured at MSY are due to both the contaminated air transported by the sea breeze from the close metropolitan area of Barcelona to the mountains and the frequent Saharan dust outbreaks (i.e. Pandolfi et al., 2011; 2014a).

5

6 3.1.5 σ_{sp} at urban/sub-urban observatories

7 Among the urban background sites, lower σ_{sp} values are measured at MAD and SIR compared to 8 DEM and UGR. The low σ_{sp} at MAD during the period presented here (only 2014 data are available 9 for MAD) could be related to the reduced formation of secondary nitrate aerosols due to the limited 10 availability of ammonia in this urban environment (Revuelta et al., 2014). However, it should be 11 noted that winter episodes with high secondary nitrate concentrations are not uncommon in Madrid 12 and we are presenting here only one year of measurements for this station. On the other hand, 13 secondary inorganic aerosol concentrations recorded at the SIR sub-urban observatory can be 14 considered as representative of a large geographical zone, given the rather flat orography of the 15 Parisian basin. At UGR, the accumulation, mainly in winter, of fine particles from traffic, domestic 16 heating and the burning of biomass explains the relatively higher σ_{sp} (e.g. Lyamani et al., 2012; 17 Titos et al., 2017). Traffic emissions, the high level of formation of secondary sulphate and organic 18 aerosols in the summer and the transport of dust from Africa are the main contributory factors to 19 the high σ_{sp} at DEM where high PM_{2.5} and PM₁₀ values are usually measured as compared to other 20 important Mediterranean cities (e.g.: Diapouli et al., 2017; Eleftheriadis et al., 2014; Karanasiou et 21 al. 2014; Querol et al., 2009).

22

23 3.2 Variability of SAE

Figure 4 shows the box-and-whiskers plots of the SAE calculated at the different stations. Table S5 and Figure S4 in the Supplementary Material report the statistics of the SAE and frequency and cumulative frequency distributions, respectively.

27

28 **3.2.1 Variability of SAE by geographical sector**

29 The SAE shows a large variability across the geographical sectors considered in this study (Figure 30 4). On average, independently from of the station setting, the highest SAE is observed at the 31 Central and Eastern European observatories (cf. Table S5) with station-averaged values of 1.88 ± 32 0.49 and 1.88 ± 0.53, respectively. The high SAE values in Central and Eastern Europe indicate 33 clearly the predominance of fine particles. In fact, high PM_{2.5}/PM₁₀ ratios, indicative of the presence 34 of small particles, are typical for rural lowland sites in Central Europe (e.g. Spindler et al., 2010; 35 EMEP, 2008). Figure S4 shows that the frequency plots of SAE data have very similar unimodal 36 delta-like distributions and the variability of the SAE within the 5th to the 95th percentile range is much lower than that of the other European regions, suggesting a greater homogeneity in some microphysical properties of atmospheric particles such as size. Exceptions are the CMN, JFJ and BEO mountain sites, where left-tailed SAE distributions are observed, likely due to the reduced effect of fine particles from the PBL in winter and an increase in the relative importance of coarse mineral dust, sea salt particles as well as aged aerosols compared to lower altitude stations in the same geographical sector.

7 On average, the SAE is lower for all other geographical sectors. Station-averaged mean SAE 8 of around 1.60 ± 0.61 , 1.25 ± 0.86 and 1.36 ± 0.67 are observed in the Nordic and Baltic, Western 9 and Southwestern sectors, respectively. Exceptions are however observed. For example, at CBW 10 (Western Europe) the median SAE reaches values of around 2.1. Indeed, both polluted air masses 11 from the industrialized zones of the Benelux countries and clean air masses from the sea 12 contribute to the presence of aerosol particles at this site (Crumeyrolle, et al., 2010). Moreover, 13 CBW is surrounded by several large cities at distances of approximately 20 to 40 km from the 14 station, which may have contributed to the high SAE values measured in this geographical 15 location. Asmi et al. (2011) have also shown that background particle number concentrations at 16 CBW are much higher than, for example, at BIR.

Median SAE values close to one or lower, indicative of the fact that the σ_{sp} is dominated by 17 18 large particles, are observed at remote sites in Western Europe (MHD), Southwestern Europe 19 (IZO) and the Nordic and Baltic (ZEP) and Antarctic (TRL) regions. A low SAE at MHD has already 20 been reported by Vaishya et al. (2011, 2012) and justified by the frequent presence, mainly in 21 winter, of coarse-mode sea-salt particles, since mineral dust particles can be ruled out. In fact, air 22 masses originating from dust sources are infrequent at these sites. Similarly, the low SAE 23 observed at ZEP and TRL can be associated with the presence of coarse sea-salt particles (e.g. 24 Zieger et al., 2010 for ZEP). Conversely, the SAE obtained at IZO is mainly due to the frequent 25 presence of mineral dust particles from African deserts (e.g. Rodríguez et al., 2011). Very similar 26 bi-modal frequency distributions are observed at MHD and IZO, showing a pronounced left peak 27 indicative of the high probability of encountering coarse particles at these sites. BIR and PLA also 28 show an enhanced left peak in the SAE frequency distributions likely due to the presence of coarse 29 marine aerosols at these sites.

30

31 **3.2.2 Variability of SAE by station type**

32 Unlike σ_{sp} , the SAE does not show any clear gradient when moving from mountainous to 33 regional/urban sites in each geographical sector. For example, at mountain sites the median SAE 34 ranges between around 0.7 at IZO (Southwest Europe) and values higher than two at JFJ and 35 CMN (Central Europe). As reported by Zieger et al. (2012) a SAE value of around 2 prevails for 36 most of the time at JFJ and can be regarded as the typical background under non-dusty conditions. 37 Thus, the SAE values at JFJ and CMN can be considered as representative of Central Europe's 38 free troposphere, especially in winter when the PBL emissions at these sites are reduced. This high variability of SAE at mountain sites was also reported by Andrews et al. (2011) with values 39

from 11 mountaintop stations worldwide ranging from less than one to more than two. Moreover, Bourcier et al. (2012) have shown that coarse particles are transported more efficiently at high altitude by the higher wind speed, thus probably also contributing to the observed variability of SAE at mountain sites.

5 Also at coastal sites (PLA and MHD), the SAE shows large variability, with higher SAE 6 measured at PLA compared to MHD, confirming a higher effect of anthropogenic emissions at PLA 7 compared to MHD.

An increasing gradient of SAE is observed when moving from regional/rural observatories in the Northwest of Europe to regional/rural observatories in the east of Europe. Among these stations, the lowest SAE is observed at VHL (Nordic and Baltic) and MSY (Southwestern Europe), whereas, as already observed, central and eastern regional areas are characterized by high SAE values. This gradient is also driven by the importance of sea salt or dust particle contributions affecting more the Northwestern and Southwestern European countries compared to countries in Central and Eastern Europe.

Among the urban sites, MAD registers the lowest median SAE (1.47) compared to UGR (1.69) and DEM (1.60). The lower SAE at MAD could be explained, as already noted, by the reduced formation of secondary inorganic aerosols during the available measurement period. Moreover, re-suspended dust from vehicles could also explain the lower SAE observed at the MAD observatory.

20

21 3.3 Variability of g

22 The asymmetry parameter is widely used in radiative transfer models because it provides 23 information regarding how much radiation is scattered back compared to the amount of radiation 24 scattered in the forward direction. Figure 5 shows the box-and-whiskers plots of g calculated at the 25 different stations. Table S6 and Figure S5 in the Supporting Material report the statistics of g and 26 the frequency and cumulative frequency distributions, respectively. Given that q is calculated from 27 BF using Equation 2 (Section 2.2.3), we report in Figure S6 of the Supporting material the box-and-28 whiskers plots of BF, whereas Table S7 reports the statistics of BF. Figure 5 and Figure S6 are 29 symmetrical, with a lower BF corresponding to a higher g.

30

31 **3.3.1 Variability of** *g* **by geographical sector**

Unlike the SAE, the *g* parameter does not show any clear gradient when moving from the west to the east of Europe. Slightly higher *g* values are observed in Western Europe (station-averaged mean *g* of 0.61 ± 0.08) compared to Central and Eastern Europe (mean $g = 0.59 \pm 0.07$ and 0.57 ± 0.06, respectively). These differences in the g values, even if small, are consistent with the opposite gradient observed for SAE, this latter being smaller in Western Europe. However, the station-averaged *g* in Central and Eastern Europe is similar to the mean *g* observed in the Nordic and Baltic regions (mean $g = 0.58 \pm 0.08$) and in Southwestern Europe (mean $g = 0.57 \pm 0.06$).

1 Thus, contrary to the SAE, a clear relationship between aerosol size and *g* is not observed. The 2 possible reasons for this are reported below.

3

4 **3.3.2 Variability of** *g* by station type

At some mountain sites higher median g values are observed relative to the g values obtained at 5 regional or urban locations. This is the case for example for IZO compared to MSY, UGR and MAD 6 7 in the Southwestern European sector and for HPB and JFJ compared to IPR, MPZ and KOS in 8 Central Europe. However, exceptions are observed. For example at CMN, where the median g 9 value (only 2 years of data are available) is the lowest in the Central European sector and among 10 the lowest observed in this study. On average, g values range between 0.49 and 0.64 at mountain 11 sites, with a mean value of 0.58 ± 0.05 . This value is consistent with the mean value of 0.61 ± 0.05 . 12 reported by Andrews et al. (2011) at the mountain sites included in their work.

13 Figure S7 in the Supporting Material reports the mean SAE (ordered from low to high values 14 for each station setting) and g at each station used in this work together with the SAE-g scatter 15 plot. Figure S7 shows that no clear relationship between q and SAE is observed. For example, the 16 TRL and MHD observatories register among the highest q values observed in the network which is 17 consistent with the very low SAE measured at these stations because of the frequent presence of coarse-mode sea-salt particles (cf. Figure 4). However, high g values, similar to TRL and MHD, are 18 19 also observed at stations such as PLA, BIR, JFJ and DEM, which are dominated on average by 20 fine aerosol particles (with SAE values similar to or higher than 1.5). Similarly, similar g values are 21 observed at IZA and PUY or HPB despite the differences in SAE values at these observatories.

22 Differences in the shape of the particle number size distribution, particle shape and chemical 23 composition (e.g. refractive index, RI) are likely factors contributing to the poor relationship 24 observed between g and SAE. The Mie theory of polydisperse spherical particles predicts that the 25 BF is lower and g correspondingly higher for coarse-mode aerosol particles (for which the SAE will be low) compared to fine-mode particles. However, some studies deploying integrating 26 27 nephelometers have found that the BF can be higher for coarse-mode aerosol particles (such as 28 mineral dust) than for fine-mode aerosol particles (Carrico et al., 2003; Doherty et al., 2005). 29 Doherty et al. (2005) suggested that an under-correction for the σ_{sp} truncation of the forward-30 scattered radiation (which is relatively larger for coarse particles) could bias the calculated BF 31 toward high values. Moreover, the shape of the particle number size distribution is another factor 32 affecting the BF and SAE. Thus, differences in the relative fractions of the fine and coarse modes 33 could also drive the BF-SAE relationship. In fact, the SAE is most sensitive to the presence of 34 coarse-mode aerosol particles compared to the BF, which is most sensitive to small accumulation-35 mode particles (Delene and Ogren, 2002; Collaud Coen et al., 2007). Thus, depending on the 36 shape of the particle number size distribution, the BF and SAE values might or might not correlate.

The refractive index (RI), which is strongly related to the chemical composition of the particles, is another important variable that can affect g (e.g. Marshall et al., 1995). In the work of Hansen and Travis (1974; Figure 12) the authors showed that, for a given particle diameter, the g

parameter non-linearly decreased with increasing real RI. Thus, coarse-mode particles with a given RI could have an asymmetry parameter similar to or lower than that of fine particles with lower RI. Recently, Obiso et al. (2017) confirmed the findings of Hansen and Travis (1974), showing also that for fine particles a perturbation in the RI of 20% has a larger effect on *g* than a similar relative perturbation of particle shape. Obiso et al. (2017) also showed that a variation of the RI for coarse particles can have a small effect on the mass scattering efficiency of a particle and its spectral dependence, and consequently also on SAE.

8

9 3.4 Seasonal variability

Figures 6, 7 and 8 present the annual cycles of σ_{sp} , SAE and *g*, respectively, at each site. The annual cycles for the non-European CHC and TRL stations are reported in Figure S8 in the Supporting Material. Overall, strong seasonal cycles of σ_{sp} and intensive aerosol particle optical parameters are observed at the majority of the stations, although exceptions are observed. The analysis of the annual cycles is presented below separately for different station settings.

15

16 **3.4.1 Seasonal variability at Arctic observatories**

17 ZEP and PAL observatories present quite different annual cycles of σ_{sp} . At ZEP, the highest σ_{sp} is 18 observed in late winter and in spring whereas the lowest σ_{sp} is observed in the summer. The σ_{sp} 19 values increase in late winter and spring due to the Arctic Haze phenomenon, i.e. layers with 20 enhanced concentrations of aerosols and precursor gases in the Arctic troposphere caused by 21 anthropogenic sources and long-range transport (i.e. Engvall et al., 2008; Ström et al., 2003). 22 Ström et al. (2003) have shown that, during winter and spring, the aerosol particle accumulation-23 mode dominates. Conversely, in summer, this mode is significantly smaller and Aitken-mode-sized 24 aerosols dominate the size distribution. The change in the aerosol size distribution between 25 winter/spring and summer is likely the cause of the observed variations of σ_{sp} and g at ZEP, the 26 latter being slightly larger in late winter and spring compared to the summer. At PAL observatory, 27 an on average higher σ_{sp} is observed in spring/summer compared to autumn/winter. As reported by 28 Lihavainen et al. (2015a), low values of σ_{sp} in autumn and early winter can be related to frequent 29 precipitation events, whereas the high values of σ_{sp} in summer are probably related to biogenic 30 organic aerosols from natural sources. At PAL, the monthly variation of SAE and g is rather 31 pronounced: SAE (g) increases (decreases) in summer compared to winter, indicating the 32 predominance of relatively smaller particles during the warmest months. Lihavainen et al. (2015a) 33 observed that the seasonal variations in intensive aerosol optical properties at PAL are related to 34 both the transport of different air masses at this remote site depending on the season, and the 35 enhanced formation of BSOA (biogenic secondary organic aerosols) in the summer. Lihavainen et 36 al. (2015a) also reported a lower single scattering albedo in winter compared to summer at PAL 37 due to a significant contribution from light absorbing carbon, mostly from residential wood

1 2 combustion. Thus, they have shown that aerosol particles observed in the summer at PAL have the potential to cool the atmosphere more efficiently than those observed during winter.

3

4 **3.4.2** Seasonal variability at mountain observatories

5 At the mountain stations (PUY, HPB, JFJ, CMN, BEO, MSA and IZO), the σ_{sp} peaks in 6 spring/summer whereas lower σ_{sp} values are measured in autumn/winter. Similar findings were, for 7 example, already reported by Nyeki et al. (1998) for JFJ and summarized by Andrews et al. (2011) 8 for many mountain top stations worldwide and by Pandolfi et al. (2014) for MSA station. Different 9 factors contribute to the σ_{sp} increase in spring/summer at the mountaintop observatories, such as 10 the increase of the boundary layer height and the stronger upslope winds during the warmest 11 months. Moreover, specific events, such as Saharan mineral dust outbreaks, may contribute to the 12 increased σ_{sp} observed at mountain stations in spring/summer, especially in Southern Europe (e.g. Pey et al., 2013; Pandolfi et al., 2014; Rodríguez et al., 2011). At IZO, σ_{sp} peaks strongly in July-13 14 August because of the very high influence of African mineral dust at this station during these 15 months (e.g. Alastuey et al., 2005; Diaz et al., 2006; Rodríguez et al., 2015). At the mountaintop 16 CHC observatory (cf. Figure S8), σ_{sp} progressively increases during the dry season, from May to 17 October, reaching lower values during the rainy season (from December to April). Moreover, during the dry season, the new particle formation events, taking place at CHC with one of the highest 18 19 frequencies reported in the literature so far (Rose et al. 2015), can introduce very small particles 20 that grow to nucleation and the Aitken mode.

21 At the mountain stations, both SAE and σ_{sp} are on average higher in summer compared to 22 the winter period, thus suggesting a higher anthropogenic influence at these sites during the 23 warmest months. The summer SAE increase is more evident at some mountain stations, e.g. HPB, 24 CMN and BEO, compared to other mountain stations such as JFJ and MSA. The less pronounced 25 SAE seasonal variation at JFJ was related by Bukowiecki et al. (2016) to the rather constant composition of the JFJ aerosol. At MSA in Southwestern Europe, the observed less pronounced 26 27 seasonal cycle of SAE could be due to the contribution of Saharan dust in spring/summer, which 28 contrasts with the PBL transport of fine particles observed at other mountain sites during the warm 29 season. At IZO, the SAE reaches its lowest values during July-August in conjunction with the peak 30 frequency of dust events (Rodríguez et al., 2015).

Overall, the *g* parameter shows an opposite seasonal cycle compared to the SAE at almost all mountain stations, with the exception of JFJ and BEO, where *g* slightly increases with SAE in the summer. At almost all mountain stations, the seasonal variations of SAE and *g* are less pronounced compared to the seasonal variation of σ_{sp} , indicating a larger seasonal variation in the extensive aerosol optical properties than in the intensive properties. At CHC, the SAE decreases as the σ_{sp} increases when moving from the wet to the dry season, indicating an increasing effect of coarse particles on the σ_{sp} during the dry season. At PUY, σ_{sp} peaks from March to September and this increase is accompanied by a small increase in SAE. Venzac et al. (2009) and Boulon et al.
(2011) have shown that PUY is more often influenced by the free troposphere or residual layers in

3 4

5 **3.4.3 Seasonal variability at coastal observatories**

winter and spring compared to the summer season.

6 A very different seasonal variation of σ_{sp} is observed at the two coastal observatories, MHD and 7 FKL (at PLA, the lack of spring/summer measurements prevents the analysis of the annual cycles). 8 The σ_{sp} at MHD (Western Europe) peaks in winter, whereas a higher σ_{sp} is observed in summer at 9 FKL (Southeastern Europe). At FKL, where no intensive optical aerosol properties are available, 10 the higher σ_{sp} in summer can be associated with mineral dust storm events, such as reported by 11 Vrekoussis et al. (2005). However, mineral dust storms in the Mediterranean are not the only 12 reason for the observed increased σ_{sp} in the summer at FKL. In fact, as for example reported by 13 Kalivitis et al. (2011), ammonium sulphate and particulate organic matter, whose concentrations 14 increase in summer in the Mediterranean Basin, can also be assumed to be important contributors to σ_{sp} during the warm season. At MHD, the higher σ_{sp} in winter is related to the higher contribution 15 16 of wind-speed-generated sea-salt particles in the marine boundary layer during winter time 17 (Vaishya et al., 2011). At MHD, the SAE (g) is higher (lower) in summer compared to winter. O'Connor et al. (2008) and Vaishya et al. (2011, 2012) showed that the background marine aerosol 18 19 level measured at MHD contains a strong and significant seasonal cycle with sea salt dominating 20 in winter and biogenic organic aerosols dominating at the submicron scale in summer. This is 21 consistent with the observed seasonal cycles of SAE and g reported here for MHD.

22

23 **3.4.4 Seasonal variability at regional/rural observatories**

24 Regional observatories in Central and Eastern Europe show marked seasonal cycles of both 25 extensive and intensive aerosol particle optical properties. In these regions, less horizontal and 26 vertical pollutant dispersion in winter, due to a higher frequency of stagnant conditions and 27 temperature inversions, play an important role in the accumulation of aerosols. As a consequence, the σ_{sp} is much higher in winter compared to summer. SAE and g also show marked seasonal 28 cycles in these regions, with the SAE (g) being higher (lower) in summer compared to winter. Ma et 29 30 al. (2014) have shown that, at MPZ, an increased SAE in summer is mainly explained by the 31 variation of the particle number size distribution. Thus, high concentrations in spring and summer 32 of small particles during new particle formation and subsequent growth periods cause the observed 33 increase of SAE (and correspondingly a decrease of g) during the warmest months.

At regional sites in the Nordic and Baltic region, the monthly variation of σ_{sp} is on average less pronounced compared to the Central or Eastern European stations, especially at BIR and SMR. This is likely due to the placement of these stations in remote areas with a different meteorology (e.g. less pronounced PBL variations) where on average much lower σ_{sp} values are measured compared to other European sites. Moreover, this could also indicate the importance of anthropogenic sources such as domestic heating in Central and Eastern Europe in winter.
 However, both SAE and *g* show marked seasonal cycles at these Nordic and Baltic observatories,
 similar to those reported for Central and Eastern European observatories with higher (lower) SAE
 (*g*) in summer compared to winter.

5 Differences are observed in the annual cycle of σ_{sp} at a regional level in Southwestern 6 Europe (represented by the MSY observatory) where higher σ_{sp} values are registered in summer. 7 At the MSY regional site (located at around 720 m a.s.l.), the higher efficiency of the sea breeze in 8 transporting pollutants from the urbanized/industrialized coastline toward regional elevated inland 9 areas during the warmer season mainly explains the summer increase in aerosol particle mass 10 concentration and scattering coefficient observed at this site (e.g. Pandolfi et al., 2011). Moreover, 11 the enhanced formation of secondary sulphate and organic matter in the summer, together with 12 frequent Saharan mineral dust outbreaks, strongly contribute to the observed seasonal cycle for 13 σ_{sp} and the intensive properties at the MSY site. The σ_{sp} peak observed at MSY in March is due to 14 the winter pollution episodes typical of the western Mediterranean Basin (WMB) (e.g. Pandolfi et 15 al., 2014a and references therein). During these episodes, the accumulation of pollutants close to the emission sources is favoured by anticyclonic conditions coupled with strong atmospheric 16 17 inversions. During such conditions, pollutants accumulate in the PBL and can subsequently reach 18 the MSY station when the PBL height increases.

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20 **3.4.5 Seasonal variability at urban/sub-urban observatories**

21 Among the urban sites, marked variations of σ_{sp} and the intensive properties are observed at UGR 22 and DEM. At the urban UGR site, the mean aerosol type is very different in winter compared to 23 summer. As evidenced by the seasonal cycles of SAE and g, aerosol particles are generally finer 24 during the winter at UGR compared to the summer season, as already observed for example by 25 Lyamani et al. (2010; 2012) and Titos et al. (2012). This is likely due to the accumulation of fine 26 particles, mainly from traffic, domestic heating and biomass burning, favoured by stagnant 27 conditions and atmospheric inversions during winter. In summer, the higher frequency of Saharan 28 mineral dust outbreaks at this site increases the mean size of the particles during the warmest 29 months. At the DEM urban observatories, the high σ_{sp} values measured in spring are linked to 30 Saharan dust outbreaks, as also supported by the seasonal cycles of SAE and g which show the 31 lowest and highest, respectively, values in spring.

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33 **3.5 SAE and** g vs. σ_{sp} relationships

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Figure 9 shows the relationships between σ_{sp} and SAE and between σ_{sp} and g at each station. Mean SAE and g are calculated for each σ_{sp} bin and the bin size at each station is calculated following the Freedman–Diaconis rule:

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$$Bin size = 2 \frac{IQR(x)}{\sqrt[3]{n}}$$

(Eq. 3)

where IQR(x) is the interquartile range of the data and *n* is the number of observations in the sample *x*. These graphs help in understanding which aerosol type on average dominates the particle light scattering, depending on the degree of scattering measured. It should be noted that, in Figure 9, the number of samples available at each station is not evenly distributed among the considered bins. Figure S9 in the Supplementary Material shows, for some stations, the SAE- σ_{sp} pairs coloured by the number of samples in each bin to highlight how the samples are distributed amongst the bins.

9

1

10 **3.5.1** *g*- σ_{sp} relationships

11 The asymmetry parameter *g* shows the lowest values for very low σ_{sp} , suggesting the 12 predominance of small fine-mode particles. Andrews et al. (2011) reported similar *g*- σ_{sp} 13 relationships at different mountain sites and suggested that the removal of large particles by cloud 14 scavenging or by deposition during transport could explain the observed low *g* values in a clean 15 atmosphere. They also suggested that the formation of new particles followed by 16 condensation/coagulation could generate small but optically active particles. Here, we show that 17 this behaviour of BF or *g* as a function of σ_{sp} was observed at all sites, not only at mountain sites.

The parameter *g* then increases with increasing σ_{sp} , indicating a shift of the particle number size distribution toward the larger end of the accumulation mode. Delene and Ogren (2002), Andrews et al. (2011), Pandolfi et al. (2014) and Sherman et al. (2015) showed that the BF tends to decrease with increasing aerosol loading, consistent with the observed increase of *g*. For comparison with previous works, Figure S10 in the Supplementary Material shows the BF- σ_{sp} relationships for all observatories, evidencing the aforementioned BF decrease with increasing σ_{sp} .

The shift of the particle number size distribution toward the large end of the fine mode with increasing σ_{sp} is probably the main reason causing the observed increase of g (and the decrease of BF, cf. Figure S10). A possible explanation for this shift is a progressive aging of atmospheric aerosol particles. Then, at the majority of stations, the variation of g is less pronounced during periods of high particle mass concentration, suggesting changes mostly in the coarse aerosol particle mode rather than in the fine mode.

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31 **3.5.2 SAE-**σ_{sp} relationships

As reported in Figure 9, at some stations the SAE progressively increases with σ_{sp} in the σ_{sp} range where the *g* parameter also increases. The increase of both *g* and SAE with σ_{sp} , observed for example at the Nordic and Baltic regions, and Central and Eastern European observatories, could be related to the different effects that different particle sizes have on the SAE and *g*. A progressive increase of SAE with σ_{sp} would suggest an increasing relative importance of fine aerosol particles. The origin of these fine particles is probably different depending on the location of the measuring site. For the remote PAL site, for example, Lihavainen et al. (2015b) observed an increase of both

1 σ_{sp} and SAE with increasing temperature due to the increasing rate of formation of BSOA with 2 increasing ambient temperature, thus likely driving the σ_{sp} -SAE relationships reported in Figure 9 3 for PAL. The BSOA from gas-to-particle formation over regions substantially lacking in 4 anthropogenic aerosol sources, such as the European boreal region (Tunved et al., 2006), 5 probably contribute strongly to the σ_{sp} -SAE relationships observed at other Nordic and Baltic sites, such as SMR. At polluted sites, such as those located in Central and Eastern Europe, the 6 7 anthropogenic aerosol emissions and active secondary aerosol production in the region (e.g. Ma et 8 al., 2014) are probably driving the σ_{sp} -SAE relationships reported in Figure 9.

9

10 For higher σ_{sp} , the σ_{sp} -SAE relationships change and a progressive shift toward relatively 11 larger particles is on average observed with increasing σ_{sp} . However, at the majority of 12 Northwestern, Central and Eastern European stations, the SAE maintains values around, or higher 13 than, 1.5 at high particle loads, indicating that the high σ_{sp} is dominated by fine particles. An 14 exception is MHD, where the SAE increases with increasing σ_{sp} , maintaining values on average 15 lower than 1.4 at high particle loads (cf. Figure 9). As already observed, the low SAE at MHD is 16 mainly due to the predominance of sea-salt coarse particles at this site (Vaishya et al., 2011). Conversely, at some sites in Southern Europe (e.g. MSA, MSY, IZO, DEM) the SAE reaches 17 18 values of around one or lower for high particle loads, indicating that, at these stations, the high σ_{sp} 19 is dominated by mineral dust coarse particles mainly from African deserts. Exceptions are two 20 urban sites in Southwestern Europe (UGR and MAD) where fine particles, probably generated for 21 the most part by traffic (and also from biomass burning at UGR) on average dominate the highest 22 measured σ_{sp} values.

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24 Similar σ_{sp} -SAE relationships to those reported in Figure 9 were observed by Andrews et al. 25 (2011) at mountain sites and by Delene and Ogren (2002) at marine sites. Among the lowest SAE 26 are observed at IZO, the station closest to the African continent. Interestingly, at IZO, the SAE shows the highest gradient for σ_{sp} coefficients in the range of 0 to 50 Mm⁻¹ whereas the gradient is 27 much lower for σ_{sp} values higher than 50 Mm⁻¹, with the SAE being almost constant for σ_{sp} higher 28 29 than 100 Mm⁻¹. The IZO station is often in the free troposphere and high loadings at this station are 30 only registered during Saharan dust events, thus it is virtually only the mineral dust that is 31 measured at IZO. Normally, the long-range transport mineral dust particles do not represent a 32 significant fraction of the particle population above 10 µm because of their short lifetimes, which 33 likely explains the constant SAE observed at the IZO site under high aerosol loading.

35 **3.6 Trends**

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Trends of σ_{sp} , SAE and BF are studied for those stations having more than 8 years' worth of data (13 observatories). Among the ACTRIS stations, PAL, SMR, MHD, HPB, IPR, JFJ and UGR have more than 10 years of data, whereas at PUY, MPZ, CMN, BEO, KPS and IZO, 8 or 9 years of data

1 are available. These stations are included in order to improve the spatial coverage, as is the case 2 in Collaud Coen et al. (2013). The Theil Sen statistical estimator (Theil, 1950; Sen, 1968) is used 3 here to determine the regression parameters of the data trends, including slope, uncertainty in the 4 slope and p-value. The Theil Sen method provides similar results to the Mann-Kendall test and it is 5 implemented for example in the Openair Package available for R software (Carslaw, 2012; Carslaw and Ropkins, 2012). The applied method yields accurate confidence intervals, even with 6 7 non-normal data, and it is less sensitive to outliers and missing values (Hollander and Wolfe, 8 1999). Monthly means are used for trend analysis and the data are corrected for seasonal effects.

9 The data coverage for σ_{sp} is higher than 70% at all stations included in the trend analyses 10 with the exception of IZO, where the σ_{sp} data coverage is 55%. For SAE, the data coverage is 11 higher than 65% at all sites with the exception of PAL (54%), PUY (59%) and IZO (52%). For BF, 12 the data coverage is higher than 65% with the exception of PAL (26%), PUY (43%), BEO (47%) 13 and IZO (27%). At the remote (PAL) or mountain stations (PUY, BEO and IZO), the percentage for 14 the intensive aerosol particle optical properties is lower because there is a higher probability of 15 measuring σ_{sp} lower than the threshold (0.8 Mm⁻¹) selected for the calculation of SAE and BF.

Table 2 reports the trends observed for σ_{sp} , SAE and BF at the thirteen observatories included in this analysis. Magnitude and statistical significance of the trends for these parameters are reported in Table S8 in the Supporting Material. In Table 2, comparisons with the previous trend analysis results presented by Collaud Coen et al. (2013) for aerosol particle optical properties and by Asmi et al. (2013) for particle number concentrations are also reported.

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22 **3.6.1 Trends of** σ_{sp}

23 Overall, a statistically significant decreasing trend for σ_{sp} is observed at around 50% of the stations 24 considered here (Table 2). Significantly, decreasing trends for σ_{sp} are observed at the two Nordic 25 and Baltic observatories (PAL for the period 2000 to 2010 and SMR); at two observatories (HPB 26 and IPR) out of the five observatories in Central Europe; and at the two observatories in 27 Southwestern Europe (IZO and UGR). The trends are not statistically significant in Western (MHD 28 and PUY) and Eastern (BEO and KPS) Europe. The highest magnitude for the σ_{sp} trend [Mm⁻¹/yr] 29 (cf. Table S8 in the Supplementary Material) is observed at the polluted IPR observatory. 30 Conversely, the lowest magnitude is observed at the remote PAL observatory.

For the periods considered in this work, the total reductions (TR) for σ_{sp} range between approximately 30% (SMR) and 60% (IZO). The high TR observed at IZO might be affected by the intensity and frequency of Saharan dust outbreaks at this site. However, estimating the effects of these events at IZO is beyond the scope of this study. Overall, the observed decreasing trends of

1 σ_{sp} are consistent with a uniform decrease in the aerosol optical depth observed in Europe 2 (AERONET data in Li et al., 2014).

3 The observed statistically significant and decreasing trends of σ_{sp} are consistent with the 4 demonstrated reduction of PM concentration in the atmosphere in Europe in recent decades 5 thanks to the implementation of European/national/regional/local mitigation strategies. These 6 decreasing trends are also consistent with the trends in the aerosol chemistry derived from 7 observations in urban environments in Europe (e.g. EEA, 2013; Barmpadimos et al., 2011; Titos et 8 al., 2014; Pandolfi et al., 2016), regional and remote environments in the western Mediterranean 9 (Cusack et al., 2012; Pandolfi et al., 2016) and in general with trends derived for the aerosol 10 chemistry across Europe (Tørseth et al., 2012). Recently, Collaud Coen et al. (2013) showed that 11 trends in σ_{sp} are observed at most of the US continental sites and that these trends are generally consistent with the strong SO₂ and PM reductions observed in the United States (Asmi et al., 2013; 12 13 EPA, 2011). Conversely, in Europe, the strong decreasing trend observed for SO₂ (e.g. Tørseth et 14 al., 2012; Henschel et al., 2013) and, with a lower spatial homogeneity and statistical significance, 15 for PM_{2.5} (e.g. EEA, 2016) is not observed for aerosol optical properties. As reported in Collaud 16 Coen et al. (2013) the reasons that at some of the European sites no significant trends are 17 observed, might be related to the spatial inhomogeneities and under-representation of continental 18 Europe PBL sites (e.g. Laj et al., 2009) and/or the timing of the SO₂ and PM trends for the United 19 States and Europe. In Europe, the emission reductions were greater for the period 1980 to 2000 20 compared to the period 2000 to 2010 (e.g. Colette et al., 2016; Tørseth et al., 2012; Manktelow et 21 al., 2007), thus the measurements of optical particle properties in Europe may not go back far 22 enough to reflect the time period with the largest emission reductions. Tørseth et al. (2012) 23 reported average reductions for ambient sulphate and nitrate mass concentrations in Europe of 24 -12% and -1%, respectively, during 2000 to 2009 compared to -24% and -7%, respectively, 25 during 1990 to 2000. These authors also reported statistically significant decreases of the PM₁₀ 26 and PM_{25} mass concentrations at around 50% of European sites, with total reductions of -18% 27 and -27%, for PM₁₀ (24 sites) and PM_{2.5} (13 sites), respectively, during 2000 to 2009. A direct 28 comparison between the stations included in this work and those included in the study of Tørseth et al. (2012) is not possible because of the different timings of the reported σ_{sp} and PM mass 29 30 concentration measurements. At those stations, where a significant decreasing trend for σ_{sp} is 31 observed and considering a period of 10 years (even if not coincident for all stations), the total 32 reduction for σ_{sp} in Europe is around -35% (cf. Table S8), consistent with the trend reported by 33 Tørseth et al. (2012) for PM in Europe.

Quite good agreement, although again likely biased by the different timings, is also observed when comparing the PM mass concentration and σ_{sp} trends by geographical sector. A significant total reduction of around -40 to -30% was reported for PM₁₀ and PM_{2.5} in the Nordic and Baltic sector by Tørseth et al. (2012; cf. Figure 7 in Tørseth et al. (2012)), in close agreement with the

1 statistically significant total decrease of σ_{sp} of around -34% reported for PAL during 2000 to 2010 2 (cf. Table S8). In the western sector (MHD), the decreasing trend for PM_{2.5} during 2000 to 2009 was insignificant (-10 to 0%) as reported here for σ_{sp} during the period 2001 to 2010. In the central 3 4 sector, statistically significant decreases for the PM_{2.5} and PM₁₀ mass concentrations ranging 5 between -20% and -40% were observed during a 10 year period (2000 to 2009) and the total 6 reduction for σ_{sp} ranged between -38% (HPB) and around -48% (IPR). In the Southwest 7 European sector the total reduction for σ_{sp} is around -32% (at UGR) and -60% (at IZO), whereas 8 Tørseth et al. (2012) reported decreases of around -20 to -40% for the PM₁₀ mass concentration 9 in the same geographical sector.

10 To further confirm the observed close agreement between the PM trends reported in the 11 literature and the trends of σ_{sp} detailed in this work, Table S9 in the Supporting Material reports the 12 comparison between σ_{sp} and PM₁₀ and/or PM_{2.5} mass concentration trends calculated at those 13 stations where simultaneous σ_{sp} and PM mass concentration measurements are available. As 14 reported in Table S9, both the observed total reductions and the statistical significance levels of the 15 trends are very similar for σ_{sp} and PM₁₀.

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17 3.6.2 Trends of SAE and BF

The trends for SAE are estimated for three different quantities, namely the SAE is calculated using 18 19 the three wavelengths (b-g-r), using the blue and the green wavelengths (b-g) and using the green 20 and red wavelengths (q-r). For the periods considered in this work (in bold in Table 2), the SAE 21 calculated using the three wavelengths (b-g-r) shows statistically significant trends at five sites. At 22 PAL (Nordic and Baltic), PUY (Western Europe) and BEO (Eastern Europe) decreasing trends are 23 observed, whereas increasing trends are observed at HPB (Central Europe) and UGR 24 (Southwestern Europe). Uniform negative trends of the columnar Ångström exponent from 25 AERONET data were reported by Li et al. (2014) across Europe and these trends were ascribed to 26 reduced fine-mode anthropogenic emissions. The positive SAE trend observed at HPB and UGR 27 would suggest a shift of the accumulation-mode particles toward smaller sizes and/or a change in 28 the coarse aerosol mode. For example, the SAE increase at UGR might be explained by a 29 progressive relative importance of fine particle emissions driven by a progressive reduction of 30 coarse particles, for example from construction/demolition works due to the economic crisis which 31 affected Spain from 2008 (e.g. Lyamani et al., 2011; Querol et al., 2014; Pandolfi et al., 2016). In 32 fact, Titos et al. (2014) reported a statistically significant decreasing trend for the PM₁₀ fraction 33 during the period 2006 to 2010 whereas no trend was observed for the PM₁ fraction. Moreover, at 34 UGR, a statistically significant increasing trend is also observed for the SAE calculated using the 35 green and red wavelengths (g-r), which are likely more sensitive to the coarser particle mode, 36 whereas the trend was non-statistically significant for the SAE at b-g wavelengths.

1 The possible change in the coarse aerosol mode at UGR is likely also the cause of the 2 observed statistically significant increasing trend of BF (cf. Table 2), given that a positive trend of 3 BF would be consistent with a shift of the accumulation-mode particles toward smaller sizes. 4 Similarly, statistically significant increasing trends for both SAE and BF are also observed at SMR 5 (SAE b-g) and HPB. Statistically significant increasing trends of BF are also observed at the other Nordic and Baltic stations (PAL) and at PUY (Western Europe), where the SAE shows statistically 6 7 significant decreasing trends, and at IPR (Central Europe) where the trend of SAE is insignificant. 8 Thus, overall, the trends of BF are positive at all stations where BF measurements are available. 9 The opposite sign of the trends for SAE and BF at PAL and PUY could be due to the different 10 effects that the different particle sizes have on SAE and g or a progressive change in the mean 11 diameter of the fine-mode aerosols. Further research involving, for example, size distribution data 12 and a Mie calculation could help in understanding the differences observed in some cases 13 between SAE and BF (or g).

14 Recently, Korras-Carraca et al. (2015) have shown that the column integrated g from Modis-Terra showed widely statistically significant positive trends (2002-2010) with stronger increases 15 observed in the eastern and southern Black Sea, as well as over the Baltic and Barents seas. 16 17 Moreover, both Modis-Terra and Modis-Aqua produce positive trends of q in the eastern 18 Mediterranean Sea and the eastern coast of the Iberian Peninsula. Positive trends for g 19 correspond to negative trends for BF. The difference observed in our work could be due to the 20 different variability often observed between near-surface measurements and column integrated 21 measurements which can confound the relationship between surface and column optical properties 22 (e.g. Bergin et al., 2000; Lyamani et al., 2010), although it has been shown that a mid-altitude 23 station might be globally representative of the whole atmospheric column (Chauvigne et al., 2016).

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25 **3.6.3 Comparison with previous trend analyses**

26 Table 2 shows the comparison, over the same time periods, between the trend analyses performed 27 in this work and the analyses presented by Collaud Coen et al. (2013) for aerosol particle optical properties and by Asmi et al. (2013) for particle number concentrations (N_{LDL-500}, N₂₀₋₅₀₀ and N₁₀₀-28 29 ₅₀₀). An agreement with the results from Collaud Coen et al. (2013) is observed for JFJ, where 30 consistent insignificant trends are detected for the three periods reported in Collaud Coen et al. 31 (2013). For MHD, we observed a non-significant increasing trend for σ_{sp} during 2001 to 2010, 32 whereas Collaud Coen et al. (2013) reported a statistically significant increasing trend for the same 33 period. At PAL, a non-statistically significant trend for σ_{sp} is observed both in the current work and 34 in Collaud-Coen et al. (2013) for the period 2001 to 2010, whereas we observe a statistically 35 significant decreasing trend for the period 2000 to 2010. Moreover, at PAL, we observe a statistically significant decreasing trend for SAE during the two common periods which were 36 37 insignificant in Collaud Coen et al. (2013). It should be noted that Collaud Coen et al. (2013)

1 reported an insignificant SAE trend at PAL using the Mann-Kendall test whereas they reported 2 statistically significant decreasing trends using the GLS/ARB (generalized least square trends with 3 either autoregressive or block bootstrap confidence intervals) and LMS (least-mean square) 4 methods, consistent with our work. These differences are thus likely due to the relative short period 5 used in these trend analyses and the different sensitivity of the methods used to missing values or the presence of outliers, especially at PAL, where σ_{sp} is very low (cf. Figure 2). For example, in this 6 7 work, the SAE calculated for PAL during the year 2007 was removed from the trend analysis due 8 to the presence of too many extremely high SAE values, likely explaining the difference observed 9 in SAE compared to the work of Collaud Coen et al. (2013). Moreover, here we use de-10 seasonalized monthly means for trend analyses whereas Collaud-Coen et al. (2013) used de-11 seasonalized medians with a different time granularity (3 days), likely affecting the comparison, 12 especially over relatively short periods.

13 A comparison of trends analysis results between σ_{sp} and the particle number concentration is not 14 straightforward as the σ_{sp} measurements are more sensitive to the particle number concentration in 15 the upper end of the fine mode than to smaller particles. For example, Asmi et al. (2013) reported 16 that, globally, no strong similarities were observed between σ_{sp} and particle number concentration 17 (N) trends and that the N trends are controlled by particles in the larger range of the Aitken mode and the smaller range of the accumulation mode, i.e. ca. 50-150 nm diameter. In this work, as 18 19 reported in Table 2, the statistically significant decreasing trend reported for N during the period 20 2001 to 2010 is not observed for σ_{sp} . However, differences are also observed at PAL between N20 21 and N100, mainly due to the fact that the DMPS measurements at PAL had long gaps during 22 periods with unusually low concentrations, thus effectively removing low concentrations from the 23 trend analysis (Asmi et al., 2013).

24

25 **3.6.4 Daytime and night time trend analyses at mountain sites**

26 Finally, the analysis of the trends during daytime (08:00 to 16:00 GMT) and night time (21:00 to 27 05:00 GMT) by season at the mountain stations are also analysed (Table 3). This analysis could 28 provide information about changes in σ_{sp} during periods when the mountain stations are likely 29 affected by the PBL (e.g. daytime and/or summer) or by the residual layer (e.g. night time in 30 summer) or when the mountain stations are representative of the free troposphere (e.g. night time 31 in winter). Consistent with what is reported in Table 2 for σ_{sp} , the trends are insignificant at JFJ, 32 PUY, CMN and BEO irrespective of the time of the day or season. The decreasing trends observed 33 at HPB, also reported in Table 2, are statistically significant only during autumn, irrespective of the 34 time of day. Conversely, the trend observed for σ_{sp} at IZO reported in Table 2, is not observed on 35 splitting the analysis by time of day and/or season.

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1 Conclusions

This investigation presented the near-surface in situ σ_{sp} (aerosol particle light scattering), SAE 3 4 (scattering Angström exponent), BF (backscatter fraction) and g (asymmetry parameter) 5 measurements obtained over the past decade at 28 atmospheric observatories which are part of the ACTRIS Research Infrastructure, with most belonging to the GAW network. Results show a 6 7 large variability of both extensive and intensive aerosol particle optical properties across the 8 network, which is consistent with the previously reported variability observed for other aerosol 9 particle properties such as particle mass concentration, particle number concentration and 10 chemical composition. Main findings can be summarized as follows:

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12 An increasing gradient of σ_{sp} is observed when moving from remote environments 13 (Arctic/mountain) to regional and to urban environments. At regional level in Europe, σ_{sp} also increases when moving from Nordic and Baltic countries and Western Europe to 14 15 Central/Eastern Europe whereas no clear spatial gradient is observed for other station 16 environments. For example, the lack of a clear spatial gradient of σ_{sp} measured at mountain 17 observatories is likely due to the different altitudes of the observatories in the different 18 geographical sectors considered in this study. Among the European mountain observatories a 19 relationship was observed between station altitude and the median σ_{sp} , this latter being the 20 highest at the station located at the lower altitude and vice versa.

- 21 Overall, the highest σ_{sp} values are measured at low altitude observatories in Central and 22 Eastern Europe and at some urban observatory in Southern Europe whereas the lowest σ_{sp} 23 values are observed at mountain stations and at Arctic and Antarctic observatories. Low σ_{sp} 24 levels, comparable to those measured at mountain sites, are also observed at the majority of 25 the regional Nordic and Baltic observatories. The σ_{sp} values in Western Europe are on average 26 higher compared to those measured in the Nordic and Baltic regions and lower compared to 27 those measured at a regional level in Southern Europe. Some exceptions to these general 28 features are however observed.
- The SAE does not show any clear gradient as a function of the placement of the station.
 However, a West to East increasing gradient is observed for both regional and mountain
 placements suggesting a lower fraction of fine-mode particle in Western/Southwestern Europe
 compared to Central and Eastern Europe where the fine-mode particles dominate the
 scattering.
- In fact, in Central and Eastern Europe, independently of the station placement, the SAE is
 among the highest observed across the network, indicating a large predominance of fine
 particles. In these regions, the SAE is even higher in summer compared to winter, suggesting
 the shift toward the small end of the aerosol particle size distribution likely linked to new particle
 formation events during the warmest months. On average SAE is lower in the Nordic and Baltic
 and western geographical sectors (likely due to the contribution from coarse-mode sea salt

- particles), and southern sectors (likely because of the presence of mineral dust particles from
 African deserts), compared to Central and Eastern Europe.
- The *g* does not show any clear gradient by station placement or geographical location
 reflecting the complex relationship of this parameter with the aerosol particles properties such
 as size distribution, particle shape or refractive index.
- Slightly higher g values are observed in Western Europe compared to Central and Eastern
 Europe. These differences in the g values, even if small, are consistent with the opposite
 gradient observed for SAE, this latter being smaller in Western Europe. However, the stationaveraged g in Central and Eastern Europe is similar to the mean g observed in the Nordic and
 Baltic regions and in Southwestern Europe. Thus, contrary to the SAE, a clear relationship
 between aerosol size and g is not observed.
- 12 Seasonal cycles for σ_{sp} , SAE and g are observed in all geographical sectors and explained by 13 different factors. The seasonal cycles are especially marked at a regional level in Central and 14 Eastern Europe where winter time episodes linked with stable air and thermal inversions favour 15 the accumulation of pollutants. In these European regions the SAE (g) is higher (lower) in 16 summer compared to winter due to variations in particle number size distribution due to the 17 enhanced formation of small and optically active particles during new particles formation and subsequent growth. Clear annual cycles are also observed at mountain sites where σ_{sp} is 18 19 higher in summer because of the enhanced influence of the boundary layer. In some cases, the 20 SAE (g) is also high (low) in summer at mountain sites indicating a higher PBL anthropogenic 21 influence during the warmer months and/or new particles formation episodes. In the Nordic and 22 Baltic regions, the seasonal variation of σ_{sp} is less pronounced compared to Central and 23 Eastern Europe, likely due to the different meteorology and less pronounced PBL variations. 24 Despite the relatively small σ_{sp} seasonal cycles in the Nordic and Baltic regions, SAE (g) 25 increases (decreases) in these regions in summer compared to the winter period likely due to a 26 season-dependent transport of air masses at these remote sites and the enhanced formation of 27 secondary organic aerosols previously observed at these sites during the warmest months. At 28 coastal sites in Northwestern Europe, the presence of sea-salt particles in winter also 29 contributes to the observed pronounced seasonal cycles of SAE and g. in Southern Europe the 30 seasonal cycles are strongly driven by the enhanced formation of secondary sulphate and 31 organic matter in the summer, together with frequent Saharan mineral dust outbreaks.
- ³² The analyses of the systematic variabilities of SAE and g as a function aerosol loading (σ_{sp}) ³³ reveal some common patterns. At all stations, *g* shows the lowest values at very low σ_{sp} likely ³⁴ because of the formation of new particles in a clean atmosphere followed by ³⁵ condensation/coagulation with, as a consequence, the generation of small but optically active ³⁶ particles. The *g* value then sharply increases with increasing σ_{sp} , indicating the shift of the ³⁷ particle number size distribution toward the larger end of the accumulation mode. Then, during ³⁸ periods of high σ_{sp} values, the variation of *g* is less pronounced at the majority of the stations,

1 contrary to the SAE, which increases or decreases, suggesting changes mostly in the coarse 2 aerosol particle mode rather than in the fine mode. At the majority of Northwestern, Central and 3 Eastern European stations, the SAE maintains high values at high σ_{sp} values, indicating that 4 the high σ_{sp} is dominated by fine particles. Conversely, at some sites in Southern Europe the SAE reaches values of around one or lower for high particle loads, indicating that, at these 5 6 stations, the high σ_{sp} is dominated by mineral dust coarse particles mainly from African 7 deserts. Exceptions are two urban sites in Southwestern Europe where fine particles, probably 8 generated for the most part by traffic (and also from biomass burning) on average dominate the 9 highest measured σ_{sp} values.

- 10 The analyses of the trends reported in this investigation provide evidence that both extensive 11 and intensive aerosol optical properties have significantly changed at some of the locations 12 included here over the last 10 and 15 years. The σ_{sp} decreasing trends reported here are 13 statistically significant at 5 out of 13 stations included in the analyses. These 5 stations are 14 located in the Nordic and Baltic regions, and the central and southwestern sectors. Conversely, σ_{sp} trends which are decreasing are not statistically significant in Western and Eastern Europe. 15 16 Statistically significant decreasing trends of SAE are observed at 3 out of 10 observatories 17 included in the analysis: one site in the Nordic and Baltic sector and two mountain sites in the 18 western and eastern sectors. These negative trends could be ascribed to reduced fine-mode 19 anthropogenic emissions, as already observed in the literature for columnar SAE in Europe. 20 Conversely, at two stations (one mountain site in Central Europe and one urban site in 21 Southwestern Europe), the SAE shows a statistically significant increasing trend, suggesting a 22 shift in the accumulation-mode particles toward smaller sizes and/or a change in the coarse 23 aerosol mode. At the remaining 5 observatories, the reported SAE trends are not statistically 24 significant. The backscatter fraction shows a statistically significant increasing trend at 5 out of 25 the 9 sites where BF measurements are available. At three stations (the mountain site in 26 Central Europe, the urban site in Southwestern Europe and one of the two sites in the Nordic 27 and Baltic sector), both BF and SAE increase, suggesting consistent evidence of a shift in the 28 accumulation-mode particles toward a smaller size. Conversely, at the other site in the Nordic 29 and Baltic sector and at one mountain site in the western sector the BF increases whereas the 30 SAE decreases.
- A general agreement is observed between the trend analyses preformed in this work and the
 analyses presented in a previous work confirming the general decreasing trends observed for
 σ_{sp} in Europe. However, some differences are also observed and likely due to the relative short
 periods used in these trend analyses and the different sensitivity of the methods used to
 missing values or presence of outliers. (Mann-Kendall or Theil-Sen vs. GLS/ARB or MLS;
 means vs. medians; different time granularity)
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1 In conclusion, this investigation provides a clear and useful picture of the spatial and temporal 2 variability of the surface in situ aerosol particle optical properties in Europe. The results presented here give a comprehensive view of the particle optical properties and provide a reliable analysis of 3 4 aerosol optical parameters for model constraints. In addition, the analyses presented here suggest 5 findings that may need additional investigation. For example, the fact that at some of the stations 6 the trend of σ_{sp} changes in terms of both statistical significance and sign depending on the period used, suggests that trend analyses are necessary in the future when longer-duration records will 7 8 be available. Moreover, the fact that at some sites BF and SAE show different signs in their trends 9 suggests that further analysis is needed to better understand how other aerosol parameters, such 10 as particle size distribution and mean diameter, affect the relationships between BF and SAE.

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12 Acknowledgments

13 This project has received funding from the European Union's Horizon 2020 research and 14 innovation programme under grant agreement No 654109, ACTRIS (project No. 262254), ACTRIS-15 PPP (project No 739530). We thank the International Foundation High Altitude Research Stations 16 JFJ (Jungfraujoch) aand Gornergrat (HFSJG), which made it possible to carry out the experiments 17 at the High Altitude Research JFJ Station and the support of MeteoSwiss within the Swiss program 18 of the Global Atmosphere Watch (GAW) of the WMO. MAD (Madrid) station is co-financed by the 19 PROACLIM (CGL2014-52877-R) project. SMR (Hyytiälä) station acknowledges BACCHUS (project 20 No. 603445), CRAICC (project No. 26060) and Academy of Finland (project No. 3073314). UGR 21 (Granada) station is co-financed by the Spanish Ministry of Economy and Competitiveness through 22 project CGL2016-81092-R. Measurements at MSY (Montseny) and MSA (Montsec) stations were 23 supported by the MINECO (Spanish Ministry of Economy, Industry and Competitiveness) and 24 FEDER funds under the PRISMA project (CGL2012-39623-C02/00) and under the HOUSE project 25 (CGL2016-78594-R), by the MAGRAMA (Spanish Ministry of Agriculture, Food and Environment) and by the Generalitat de Catalunya (AGAUR 2014 SGR33, AGAUR 2017 SGR41 and the 26 27 DGQA). Measurements at IZO (Izaña) were supported by the AEROATLAN project (CGL2015-17 28 66229-P), co-funded by the Ministry of Economy and Competitiveness of Spain and the European 29 Regional Development Fund. Station KOS (Košetice) is supported by the Ministry of Education, 30 Youth and Sports of the Czech Republic within the project for support of national research infrastructure ACTRIS - participation of the Czech Republic (ACTRIS-CZ - LM2015037). 31 32 Measurements at PUY (Puy de Dôme) were partly supported by CNRS-INSU, University Clermont-Auvergne, OPGC and the french CLAP program. PAL (Pallas) station acknowledges KONE 33 34 Foundation, Academy of Finland (project No. 269095 and No. 296302). CHC (Mt. Chacaltaya) 35 station received support from Institut de Recherche pour le Développement (IRD) under both 36 Jeune Equipe program attributed to LFA and support to ACTRIS-FR program. CHC received 37 grants from Labex OSUG@2020 (Investissements d'avenir - ANR10 LABX56). Marco Pandolfi is 38 funded by a Ramón y Cajal Fellowship (RYC-2013-14036) awarded by the Spanish Ministry of 39 Economy and Competitiveness. The authors would like to express their gratitude to D. C. Carslaw

1	and K. Ropkins for providing the OpenAir software used in this paper (Carslaw and Ropkins, 2012;
2	Carslaw, 2012). We also thank the co-editor Andreas Petzold and two anonymous reviewers for
3	their constructive comments.
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Table captions:

3 Table 1: List of ACTRIS observatories providing aerosol particle scattering measurements

Table 2: Trends of aerosol particle scattering coefficient (σ_{sp}), scattering Ångström exponent (SAE) and backscatter fraction (BF). Three trends for SAE are reported: SAE calculated as a linear fit using; three wavelengths (b-g-r), using the blue and green wavelengths (b-g) and using the green and red wavelengths (g-r). Trend results are reported for the whole period available at each station until 2015 (bold) and for the periods reported in Collaud Coen et al. (2013) and Asmi et al. (2013). Trends are considered as statistically significant for a p-value of <0.05. Statistically significant increasing or decreasing trends are highlighted with up (1) and down (1) red and green arrows, respectively. Non-statistically significant increasing or decreasing trends are highlighted with up (\uparrow) and down (\downarrow) grey arrows, respectively. Grey coloured table cells highlight stations included in this work but not included in the works of Collaud Coen et al. (2013) or Asmi et al. (2013). \$: parameters removed in this work and in the work of Collaud Coen et al. (2013) because of measurement gaps, low data coverage or break points for one or more wavelengths. #: Only available for 2014-2015; ± not available.

Table 3: Daytime (08:00 to 16:00 GMT) and night time (21:00 to 05:00 GMT) σ_{sp} trends by season calculated for the periods considered in this work. Sp: Spring; Su: Summer; Au: Autumn; Wi: Winter. Trends are considered as statistically significant at a p-value of <0.05. Statistically significant increasing or decreasing trends are highlighted with up (\uparrow) and down (\downarrow) red and green arrows, respectively. Non-statistically significant increasing or decreasing trends are highlighted with up (\uparrow) and down (\downarrow) grey arrows, respectively.

Observatory name/setting (1)	Country	Observatory code	Lat, Long	Altitude [m a.s.l.]	Geographical location	Inlet	Nephelometer model	Period (a)
Arctic observ	atories							
Zeppelin (ZEP)	Svalbard (Norway)	NO0042G	78.9067 N, 11.8883 E	474	Nordic and Baltic	PM ₁₀	TSI3563	07/2010 -12/2014
Pallas (PAL)	Finland	FI0096G	67.97 N, 24.12 E	565	Nordic and Baltic	PM ₅ ; PM _{2.5} ; PM ₁₀ (<i>b</i>)	TSI3563	02/2000 –12/2015
Antarctic obs	ervatories							
Troll (TRL)	Antarctica	NO0058G	-72.0167 N, 2.5333 E	1309	Antarctica	whole air; PM ₁₀ (c)	TSI3563	02/2007 –12/2015
Mountain obs	ervatories							
Puy de Dome (PUY)	France	FR0030R	45.7667 N, 2.95 E	1465	West	whole air	TSI3563	01/2007 –12/2014
lzaña (IZO)	Spain	ES0018G	28.309 N, -16.4994 E	2373	Southwest	PM ₁₀	TSI3563	03/2008 - 12/2015
Montsec (MSA)	Spain	ES0022R	42.0513 N, 0.44 E	1570	Southwest	PM _{2.5} ; PM ₁₀ (<i>d</i>)	ECOTECH Aurora3000	01/2013 – 12/2015
Jungfraujoch (JFJ)	Switzerland	CH0001G	46.5475 N, 7.985 E	3578	Central	whole air	TSI3563	07/1995 –12/2015
Mt. Cimone (CMN)	Italy	IT0009R	44.1833 N, 10.7 E	2165	Central	whole air	ECOTECH Aurora M9003;	05/2007 –12/2015
							TSI 3563 (<i>e</i>)	
Hohenpeisse nberg (HPB)	Germany	DE0043G	47.8 N, 11.0167 E	985	Central	PM ₁₀	TSI3563	01/2006 -12/2015
Beo Moussala (BEO)	Bulgaria	BG0001R	42.1667 N, 23.5833 E	2971	East	whole air	TSI3563	03/2007 –12/2015
Mt. Chacaltaya (CHC)	Bolivia	BO0001R	-16.2000 N, -68.09999 E	5240	South America	whole air	ECOTECH Aurora3000	01/2012 – 12/2015 (f)
Coastal obser	watorios							
Preila (PLA)	Lithuania	LT0015R	55.35 N, 21.0667 E	5	Nordic and Baltic	PM ₁₀	TSI3563	12/2012 -04/2014
Mace Head	Ireland	IE0031R	53.3258 N, -9.8994 E	5	West	whole air	TSI3563	07/2001 –12/2013
Finokalia (FKL) ⁽²⁾	Greece	GR0002R	35.3167 N, 25.6667 E	250	Southeast	whole air; PM ₁ ; PM ₁₀ (g)	RR M903; Ecotech Aurora1000 (h)	04/2004 –12/2015
Regional/rura	l observatories							
Birkenes II (BIR)	Norway	NO0002R	58.3885 N, 8.252 E	219	Nordic and Baltic	PM ₁₀	TSI3563	07/2009 –12/2015
Hyytiälä (SMR)	Finland	FI0050R	61.85N, 24.2833 E	181	Nordic and Baltic	PM ₁₀	TSI3563	05/2006 -12/2015
Vavihill (VHL) ⁽³⁾	Sweden	SE0011R	56.0167 N, 13.15 E	175	Nordic and Baltic	PM ₁₀	ECOTECH Aurora3000	03/2008 –04/2014
Observatory Perenne (OPE)	France	FR0022R	48.5622 N, 5.505555 E	392	West	whole air; PM ₁₀ (<i>i</i>)	ECOTECH Aurora3000	09/2012 –12/2015
Cabauw (CBW) ⁽⁴⁾	The Netherlands	NL0011R	51.9703 N, 4.9264 E	1	West	PM ₁₀ ()	TSI3563	01/2008 -12/2012
Montseny (MSY)	Spain	ES1778R	41.7667 N, 2.35 E	700	Southwest	PM ₁₀	ECOTECH Aurora3000	01/2010 -12/2015
Košetice (KOS)	Czech Republic	CZ0007R	49.58333N, 15.0833 E	534	Central	PM ₁₀	TSI3563	03/2013 – 12/2015
Melpitz (MPZ) ⁽⁵⁾	Germany	DE0044R	51.53 N, 12.93 E	86	Central	PM ₁₀	TSI3563	01/2007 –12/2015
lspra (IPR)	Italy	IT0004R	45.8 N, 8.6333 E	209	Central	PM ₁₀	TSI3563	01/2004 -12/2014
K-Puszta (KPS)	Hungary	HU0002R	46.9667 N, 19.5833 E	125	East	PM ₁ ; PM ₁₀ (j)	TSI3563	05/2006 -12/2014
	ban observator		40 7000 11				E0075011	
SIRTA	France	FR0020R	48.7086 N,	162	West	PM ₁	ECOTECH	07/2012 -12/2013

(SIR)			2.1589 E				M9003			
Madrid	Spain	ES1778R	40.4627 N,	669	Southwest	PM _{2.5} ;	ECOTECH	01/2014 - 12/2014		
(MAD)	-		-3.717 E			PM ₁₀ (<i>k</i>)	Aurora3000			
Granada	Spain	ES0020U	37.164 N,	680	Southwest	whole	TSI3563	01/2006 - 12/2015		
(UGR)	-		-3.605 E			air				
Athens	Greece	GR0100B	37.9905 N,	270	Southeast	PM ₁₀	ECOTECH	01/2012 - 12/2015		
(DEM)			23.8095 E				Aurora3000			
1 (1) Observatory codes from EBAS: (2) GAW code: FIK: (3) GAW code: VAV: (4) GAW code: CES: (5) GAW code: MEL: (a) start-end of										

(1) Observatory codes from EBAS; (2) GAW code: FIK; (3) GAW code: VAV; (4) GAW code: CES; (5) GAW code: MEL; (a) start-end of measurements; total aerosol particle scattering was used as reference for measurement period; (b) PM₅ (2000-08/2005), PM_{2.5} (08/2005-2007) and PM₁₀ (2008-2015); (c) whole air (2007-2009) and PM₁₀ (2010-2015); (d) PM_{2.5} (2013-03/2014) and PM₁₀ (04/2014-2015); (e) ECOTECH Aurora M9003 during 2007-2013 and TSI 3563 (2014-2015); (f) only measurements performed during the year 2012 were used in this investigation; (g) whole air (2004-2008), PM₁₀ (2009-2011), PM₁ (2011-2012), PM₁₀ (2013-2015); (h) RR M903 during 2004-2011, Ecotech AURORA1000 during 2012-2015; (i) whole air (2012-08/2013) and PM₁₀ (09/2014-2015); (j) PM₁ (2006-04/2008) and PM₁₀ (05/2008-2014); (k) PM₁₀ from 03/2014.

				Trene (This wo			MK Trend (Collaud Coen et al., 2013)						MK Trend (Asmi et al., 2013)
Station period			SAE			BF			SAE				Particle number	
		σ_{sp}	b-g-r	b-g	g-r	БГ	σ_{sp}	b-r	b-g	g-r	BF	Ν	N20 (20-500 nm)	N100 (100-500 nm)
		_					<u>Nordi</u>	c and	d Baltio	2				-
	2000 - 2015	•	+	$\mathbf{+}$	+	•								
	2000 - 2010	♦	+	\$	\$	•	♦	1	\$	\$	1			
PAL	2001 - 2010	₩	+	\$	\$		↓	+	\$	\$	1	↓(10-500 nm)	~	+
	1996 - 2010											(10-500 nm) ↓		
	2006 - 2015	↓		1	1	1								
SMR	1996 - 2011												¥	+
	2001 - 2010												+	+
							<u>k</u>	Veste	ern					•
_	2001 - 2013	+	\$	\$	\$	\$								
MHD	2000 - 2010											🔶 (3-500 nm)		
	2001 - 2010	•	\$	\$	\$	\$	1	\$	\$	\$	\$	🛉 (3-500 nm)		
PUY	2007 - 2014	. ↓	¥	+	+	1			_					
							<u>.</u>	Centr	<u>al</u>					
	2006 - 2015	+	1	+	•	1								
НРВ	2001 - 2010						•	\$	\$	\$	\$			
	2002 - 2010						♦	\$	\$	\$	\$			
	1995 - 2011											🕈 (15-500 nm)		
IPR ⁽¹⁾	2004 - 2014	•	1	•	^	1								
MPZ	2007 - 2015 1997 – 1998	▼	♦	•	+	1								
	and 2004 - 2010													1
	1995 - 2015	↓	\$	\$	\$	\$								
	1995 - 2010	•	\$	\$	\$	\$	1	\$	\$	\$	\$			
JFJ	1996 - 2010		\$	\$	\$	\$	1	\$	\$	\$	\$			
	2001 - 2010	₩	\$	\$	\$	\$	♦	\$	\$	\$	\$	🔶 (10-500 nm)		
	1997 - 2010	•	\$	\$	\$	\$						(10-500 nm)		
CMN	2007 - 2015	₩	#	#	#	#								
	1						<u></u>	Easte	<u>rn</u>			· · · · · · · · · · · · · · · · · · ·		
BEO	2007 - 2015	•	•	•	•	★								
KPS	2006 - 2014	1	₩	+	1	1	Sou	thwe	stern					
IZO	2008 - 2015				•	\$								
UGR	2006 - 2015				 ▲	Ψ ↓								
(1)		llv siar	l	decro	l ocina tro		L ot IF			roportor	l I by Dut	aud et al. (2014	1) for the peri	ad 2002 20

1 Table 3

		SCATTERING								
Station	mania d	day	time	nigh	ttime	24h				
	period	Sp	Su	Sp	Su	Sp	Su			
		Au	Wi	Au	Wi	Au	Wi			
JFJ	1995 - 2015	₩	➡	¥	¥	₩	➡			
51.5	1333 - 2013	1	→	1	₩		+			
HPB	2006 - 2015	₩	₩	\downarrow	+	+	\downarrow			
	2000 - 2013	+	\downarrow	¥	\downarrow	+	\downarrow			
PUY	2006 - 2014	$\mathbf{+}$	₩	\downarrow	\downarrow	+	\downarrow			
101	2000 - 2014	₩	₩	\downarrow	+	+	\downarrow			
CMN	2007 - 2015	₩	1	+	+	+	$-\downarrow$			
Civin	2007 - 2015	↓	₩	↓	₩	♦	¥			
REO	2007 2015	₩	₩	₩	1	₩	₩			
BEO	2007 - 2015	₩	→	₩	1	₩	1			
170	2008 2015	₩	₩	₩	₩	₩	↓			
IZO	2008 - 2015	1	₩	1	¥	+	₩			

Figure captions:

Figure 1: Locations of the 28 ACTRIS stations included in this work.

Figure 2: Total aerosol scattering coefficients in the green divided by station setting. Different colours highlight different geographical locations. At SIR, aerosol scattering was available only at 450 nm. Medians (horizontal lines in the boxes), percentiles 25 and 75 (lower and upper limits of the boxes, respectively) and percentiles 5 and 95 (lower and upper limits of the vertical dashed lines) are reported. Hourly data were used for the statistics.

Figure 3: Relationship between: (a) N50 (mean particle number concentration between 50 nm and 500 nm), (b) N100 (mean particle number concentration between 100 nm and 500 nm), (c) absorption coefficient and mean aerosol particle total scattering coefficient. (a) and (b): data averaged over the period 2008 to 2009. For ZEP, BIR, KOS and PLA aerosol particle scattering measurements were not available during 2008 to 2009 and different periods were used. R² values, highlighted in red, were obtained using the median values. (c) Data averaged as in Zanatta et al. (2016). Figure 3c also reports the geometric mean of SSA.

Figure 4: Scattering Ångström exponent divided by station setting. Different colours highlight different geographical locations. Medians (horizontal lines in the boxes), percentiles 25 and 75 (lower and upper limits of the boxes, respectively) and percentiles 5 and 95 (lower and upper limits of the vertical dashed lines) are reported. Hourly data were used for the statistics.

Figure 5: Asymmetry parameter in the green divided by station setting. Different colours highlight different geographical locations. Medians (horizontal lines in the boxes), percentiles 25 and 75 (lower and upper limits of the boxes, respectively) and percentiles 5 and 95 (lower and upper limits of the vertical dashed lines) are reported. Hourly data were used for the statistics.

- **Figure 6**: Seasonal cycles of σ_{sp} [Mm⁻¹] measured in the green nephelometer wavelength.
- Figure 6: (Continued) Seasonal cycles of σ_{sp} [Mm⁻¹] measured in the green nephelometer wavelength.

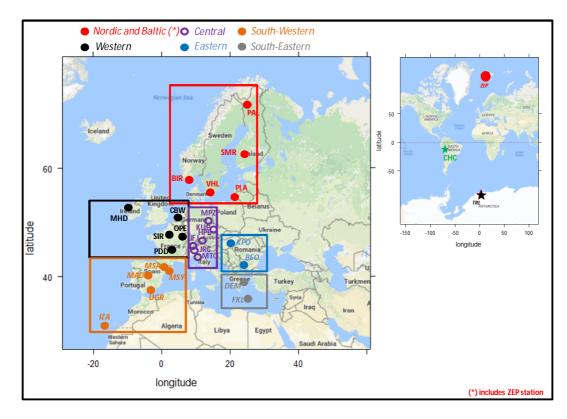
Figure 7: Seasonal cycles of SAE (calculated using the three nephelometer wavelengths).

Figure 7: (Continued) Seasonal cycles of SAE (calculated using the three nephelometer wavelengths).

Figure 8: Seasonal cycles of g (calculated for the green wavelength).

Figure 8: (Continued) Seasonal cycles of g (calculated for the green wavelength).

Figure 9: Scatterplots between σ_{sp} (x-axes) and SAE (right y-axes; red lines) and *g* (left y-axes; black lines). Dashed lines represent median σ_{sp} values at each station. Different colours highlight different geographical locations as in Figures 2, 4 and 5.



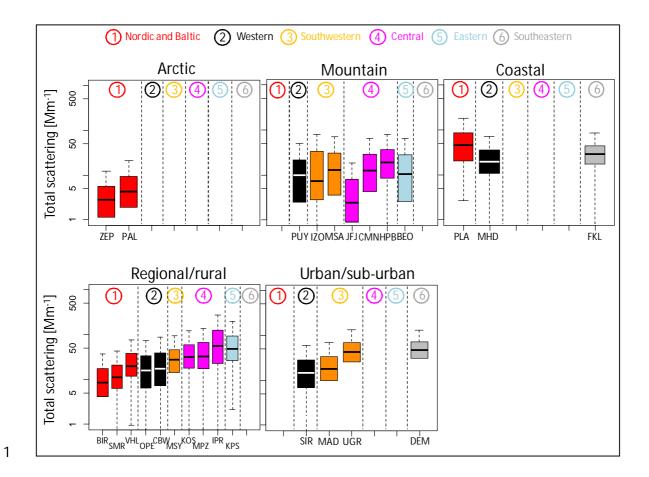
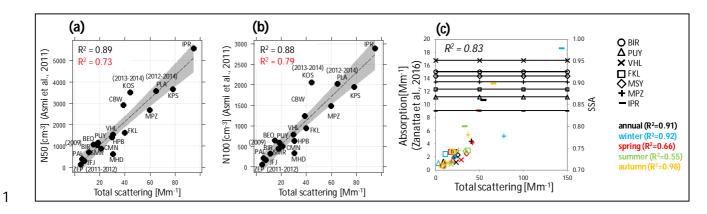
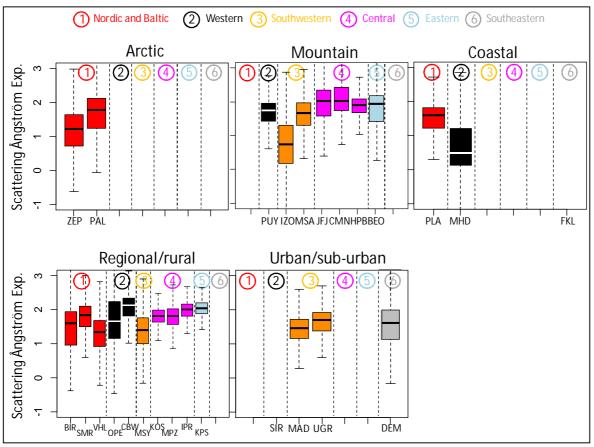


Figure 2









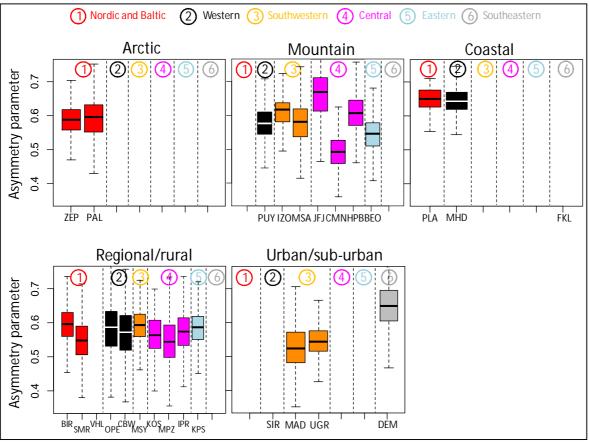
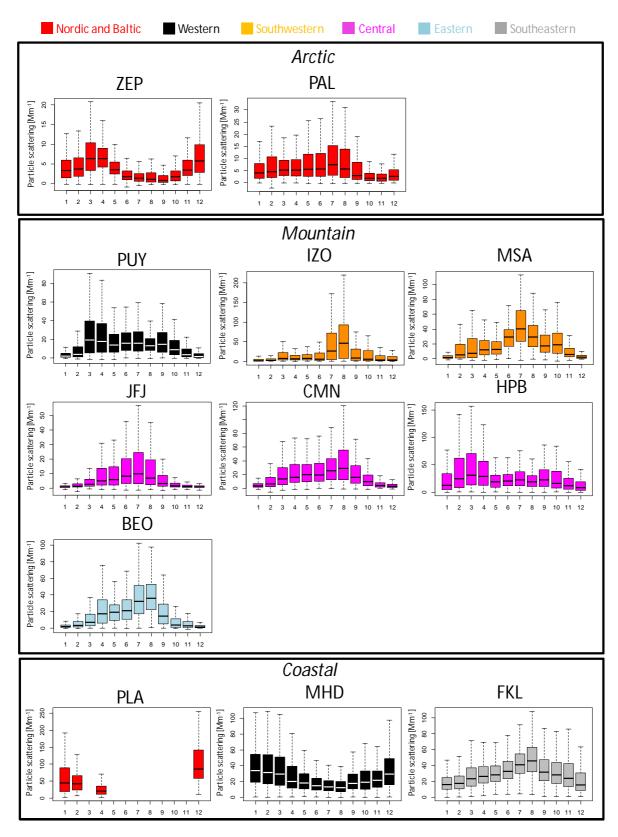
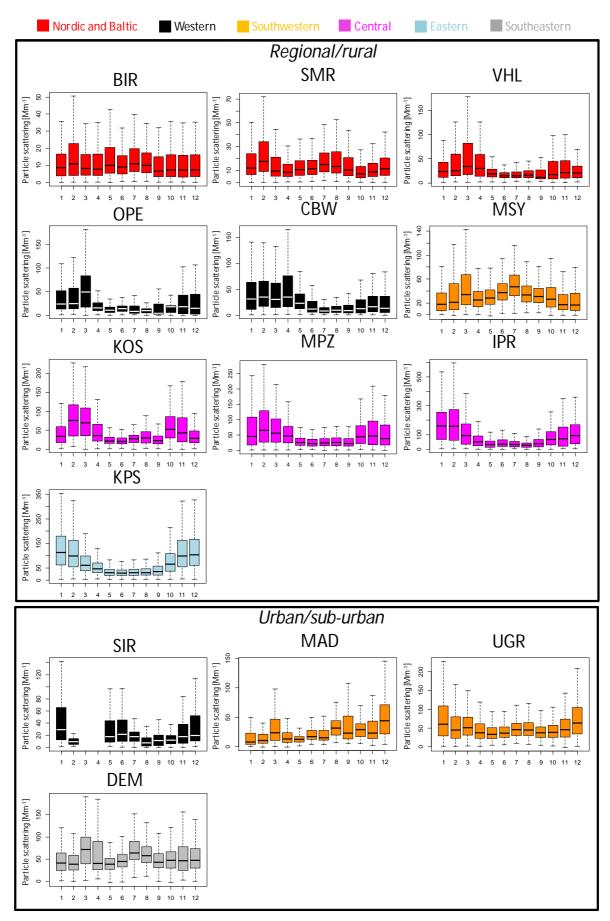
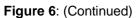
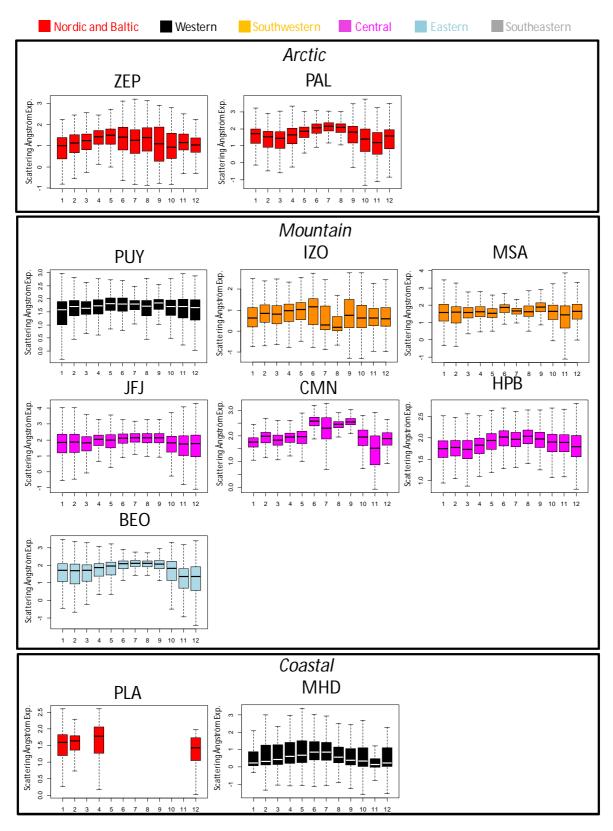


Figure 5











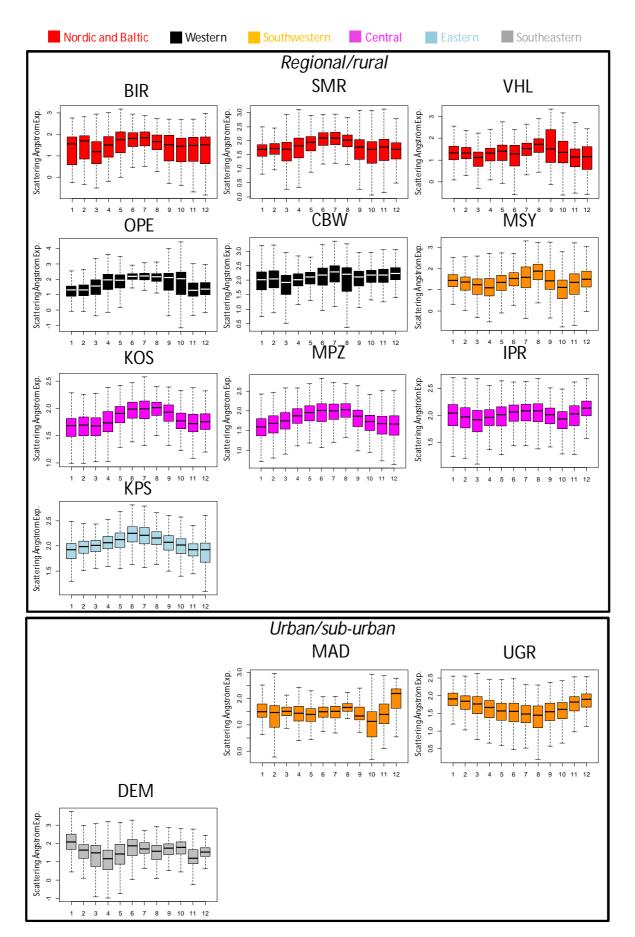
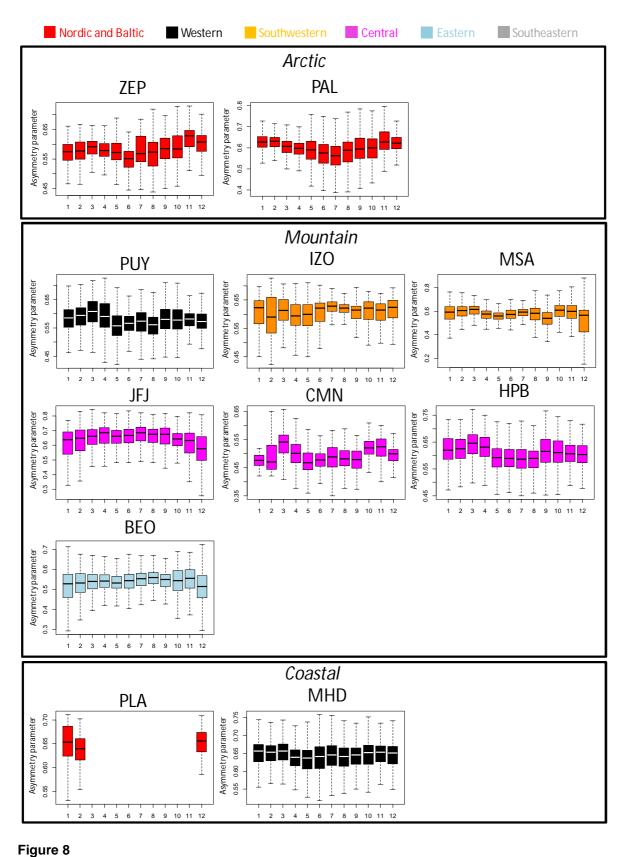


Figure 7: (Continued)



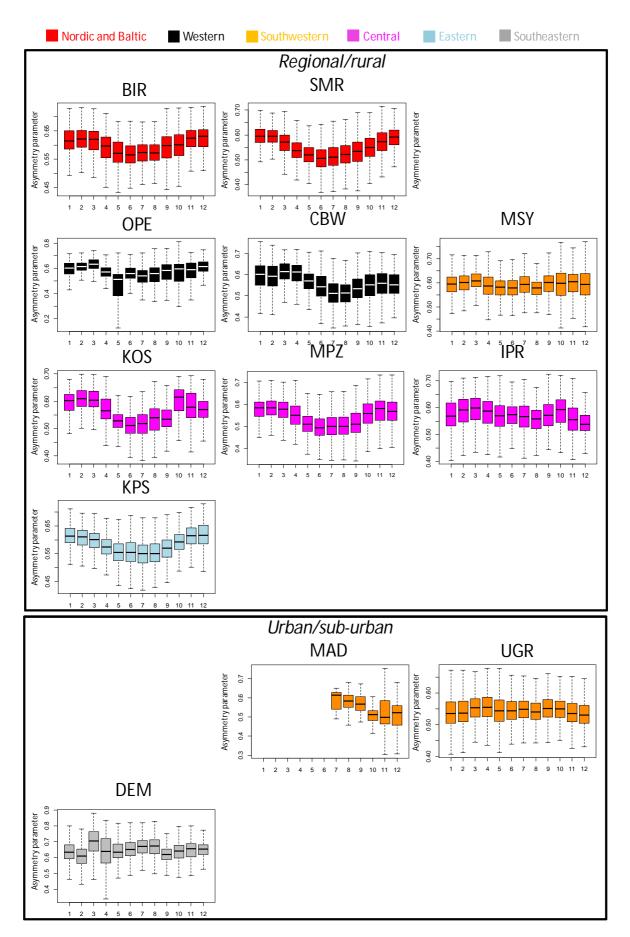


Figure 8: (Continued)

