



1 Atmospheric pollution concentrations over the Eastern

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2 Mediterranean during summer – A review
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Abstract. The Eastern Mediterranean (EM) is one of the regions in the world where elevated concentrations of primary and secondary gaseous air pollutants have been reported frequently, mainly in summer. This review discusses published studies of the atmospheric dispersion and transport conditions characterizing this region during the summer, followed by a description of some essential studies dealing with the corresponding concentrations of air pollutants such as ozone, carbon monoxide, total reactive nitrogen, methane and sulfate aerosols observed there.

19 The interlaced relationship between the downward motion of the subsiding air aloft induced by 20 global circulation systems affecting the EM and the depth of the Persian Trough, a low-pressure trough that extends from the Asian monsoon at the surface controlling the spatio-temporal 21 22 distribution of the mixed boundary layer during summer is discussed. The strength of the wind flow within the mixed layer and its depth affect much the amount of pollutants transported and 23 24 determine the potential of the atmosphere to disperse contaminants off their origins in the EM. 25 The reduced mixed layer and the accompanying weak westerlies, characterizing the summer in this region, lead to reduced ventilation rates, preventing an effective dilution of the contaminants. 26 27 Several studies pointing at specific local (e.g. ventilation rates) and regional peculiarities (long-28 range transport) enhancing the building up of pollutant concentrations are presented.

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1 Tropospheric ozone concentrations observed in the summer over the EM are among the highest 2 over the Northern Hemisphere. The three essential processes controlling its formation (i.e., longrange transport of polluted air masses, dynamic subsidence at mid-tropospheric levels, and 3 4 stratosphere-to-troposphere exchange) are reviewed. Airborne campaigns and satellite-borne initiatives have indicated that the concentration values of reactive nitrogen identified as precursors 5 in the formation of ozone over the EM were found to be 2 to 10 times higher than in the 6 hemispheric background troposphere. Several factors favor sulfate particulate abundance over the 7 8 EM. Models, aircraft measurements, and satellite derived data, have clearly shown that sulfate has 9 a maximum during spring and summer over the EM. The carbon monoxide (CO) seasonal cycle, as obtained from global background monitoring sites in the EM is mostly controlled by the 10 tropospheric concentration of the hydroxyl radical (OH), and therefore demonstrates high 11 concentrations over winter months and the lowest during summer when photochemistry is active. 12 Modeling studies have shown that the diurnal variations in CO concentration during the summer 13 result from long-range CO transport from European anthropogenic sources, contributing 60 to 80% 14 of the boundary-layer CO over the EM. The values retrieved from satellite data enable us to derive 15 the spatial distribution of methane (CH_4) , identifying August as the month with the highest levels 16 over the EM. The outcomes of a recent extensive examination of the distribution of methane over 17 the tropospheric Mediterranean Basin, as part of the Chemical and Aerosol Mediterranean 18 Experiment (ChArMEx) program, using model simulations and satellite measurements is coherent 19 20 with other previous studies. Moreover, this methane study provides some insights on the role of 21 the Asian monsoon anticyclone in controlling the variability of CH₄ pollutant within mid-to-upper 22 tropospheric levels above the EM in summer.

23

24 1 Introduction

The relationship between atmospheric pollutant concentrations and large-scale atmospheric circulation systems have been examined over the past decades (e.g., Davis and Kalkstein, 1990; Dayan et al., 2008). This strong relationship and its issuing dispersion condition at several scales, and climatically related variables such as air pollutants, is presented in this work as part of the Chemistry-Aerosol Mediterranean Experiment (ChArMEx; http://charmex.lsce.ipsl.fr).





1 However, a first major drawback in attributing pollutant concentrations to variations in large-scale 2 atmospheric circulation arises from the fact that changes in removal processes and upwind 3 emissions are not necessarily concurrent with variations in circulation. Some efforts were 4 undertaken, mainly through coupled climate-chemistry models to treat and analyze at the same time, the changes in general circulation and atmospheric chemistry (Hein et al., 2001; Dastoor and 5 6 Larocque, 2004). Moreover, secondary pollutants such as tropospheric ozone result basically from 7 photochemical reactions among precursors and as such, are controlled by air mass characteristics 8 such as temperature and humidity, and cloud cover/solar radiation. Accordingly, changes in trace 9 gases concentrations are modified by exposure to differing air masses as governed by changes in 10 atmospheric circulation. A second substantial shortcoming in trying to associate changes in pollutant concentration to

A second substantial shortcoming in trying to associate changes in pollutant concentration to variation in circulation patterns is their different life span and distribution. For example, durable greenhouse gases (GHG) such as methane (CH₄) and carbon dioxide (CO₂) as compared to aerosols which are most relevant for short spatial and temporal scales (Andreae, 2001; Voulgarakis et al., 2010). Radiative forcing of aerosols is of much higher spatial variability than GHG forcings due to the relatively short aerosol lifetime (daily-weekly scale) compared to that of GHG (monthlyyearly scale).

Lelieveld et al. (2002) studied air pollutant transport over the EM in summer time. They report that the synoptic flow is controlled by the strong east-west pressure difference between the Azores high and the Asian monsoon low, with additional influence in the upper troposphere from the Tibetan anticyclone. This yields a contrasted situation in the tropospheric column with European influence in the lowermost troposphere, a much longer-range transport from Asia and North America at mid-tropospheric levels, and a major impact from Asia in the upper troposphere and lower stratosphere.

In this review, centered over the Eastern Mediterranean (EM) during the summer season, we first describe the atmospheric dynamic conditions favoring the building up of tropospheric pollutant concentrations. Secondly, we propose a synthesis of the essential studies on air pollutant concentrations including ozone (O₃), sulfate aerosols (SO₄), total reactive nitrogen (NOy), carbon monoxide (CO), and methane (CH₄). The sources of the data reported include in-situ observations,





- 1 balloon-sounding, aircraft and space-borne observations as well as model data which results, in
- 2 terms of dynamics, are mostly updated over 1948-2016 on availability.
- 3

4 2 Summer atmospheric dynamic conditions favoring the building up of tropospheric 5 pollutants concentrations

6 Different spatial and temporal scales of motion affect pollutant transport and dispersion: the 7 microscale, mesoscale, synoptic scale, and macro-, or global scale. At the scale of a few months, 8 the planetary boundary layer is relatively well mixed. However, on shorter timescales and near the 9 Earth's surface (where pollutants are emitted), transport and dispersion are often limited by 10 atmospheric conditions. In this section, we will focus on the global and synoptic scale processes 11 that favor a potential accumulation of pollutants in the EM troposphere.

12 2.1 Global and synoptic scales inducing subsiding conditions over the Eastern Mediterranean

In general, the atmospheric conditions over the EM are persistent during the summer and subject 13 to two essential processes. The first is the cool advection at shallow tropospheric layers caused by 14 the strong, dry north Etesian winds generated by the Persian Trough (PT). This surface low 15 pressure trough extends from the Asian monsoon through the Persian Gulf and further, along 16 southern Turkey to the Aegean Sea (Figs. 1 and 2). The second is the dynamic subsidence 17 generated by several global-scale processes: the African Monsoon as part of the subtropical 18 descending branch of the Hadley cell (Fig. 3 left), the Asian Monsoon as part of the Walker cell 19 20 (Fig. 3 right) and subsidence caused by the negative relative vorticity characterizing this region, 21 during summer, as explained further on.

However, subsidence is neither restricted to mid-tropospheric levels nor solely associated to the descending branch of these both general circulation cells. In summer, at higher atmospheric layers, air masses converge and subside over the EM as contributed by both an anticyclonic curvature caused by the eastward progression of the Subtropical high, and an anticyclonic wind shear as related to the position of the Subtropical Jet. Under these circumstances the southeastern part of the EM is exposed to the southern flank of the jet and therefore prone to negative shear vorticity. Although shear vorticity is an order of magnitude smaller than planetary vorticity, nearby jet streak





- 1 makes this relative vorticity component significant due to the strong change in wind speeds across
- 2 the jet. Contribution of both components enhances negative vorticity resulting in a total long- term
- 3 mean negative vorticity of -1 to -3 10^{-5} s⁻¹ at 200 hPa (~12 km above sea level; a.s.l.) featuring the
- 4 summer season over the EM (Fig. 4).
- 5 The contribution of the above-mentioned dynamic subsidence generated by all processes results
- 6 in positive Omega values, defined as the Lagrangian rate of change in pressure with time,
- 7 indicating a downward air motion over the whole EM with its highest core of maximum subsidence
- 8 over Crete as depicted over mid-tropospheric levels (500 hPa geopotential height) (Fig. 5).

9 Following the subsidence caused by the large-scale downward motion, the warming and drying up

10 is manifested by the delimiting sharp decrease in relative humidity over the EM Basin (Fig. 6).

Based on National Centers for Environmental Prediction/National Center for Atmospheric Research (NCEP/NCAR) reanalysis for 2000–2012, Lensky and Dayan (2015) have recently shown that the coincidence of negative vorticity advection aloft accompanied by cold horizontal advection, at lower tropospheric levels, featuring the EM during PT synoptic conditions drives the wind flow out of the thermal wind balance inducing a vertical downward motion (Figs. 2 and 7).

Rodwell and Hoskins (1996) used an idealized model to explain the strong descent on the 16 17 subtropics attributed to the African Monsoon Hadley cell, and suggested that the monsoon-forced 18 adiabatic descent may result in clear air and, therefore, a local diabatic enhancement which effectively doubles the strength of descent. Ziv et al. (2004) found that the cool advection 19 associated to the PT (Fig. 2) and the subsidence related to both descending branches of the African 20 21 and Asian Monsoons (Fig. 3) are interrelated and tend to balance each other. They suggest that 22 this compensation mechanism explains the reduced day to day temperature variations over the 23 eastern part of the basin in summer (Fig. 8).

However, this monotonic regime is interrupted by the occurrence of hot day events resulting from
an expansion of the Subtropical High from North Africa towards the EM which are prone for
elevated concentration of air pollutants. Harpaz et al. (2014) found that such episodes are confined
to the lower 4 km and controlled by the intensity of the negative temperature advection rather than
by the prevailing subsidence.





1 2.2 Atmospheric mixing layer depth over the Eastern Mediterranean coastal zone

2 The vertical velocity involved in the mixing process within the turbulent layer near the surface and 3 specifically its depth are important parameters in determining air pollutant concentrations at 4 shallow tropospheric levels (Zhang and Rao, 1999). The changes in the mixing layer depth (MLD, i.e. the height of the convective atmospheric boundary layer marked by the base of a thermal 5 inversion) is governed by several factors: surface heating (Holtslag and Van Ulden, 1983), 6 7 horizontal advection determined by the intensity of the sea breeze in coastal areas (McElroy and Smith, 1991; Lensky and Dayan, 2012), local terrain over the continent (Kalthoff et al., 1998), and 8 9 the strength of the subsiding atmospheric air mass capping the mixed layer, defined by the temperature profile within this stable layer and synoptic scale vertical motion (Dayan et al., 1988). 10 Since this turbulent layer is mainly governed by synoptic-scale circulation patterns, both the 11 12 surface synoptic systems and their associated upper tropospheric conditions should be taken into consideration for understanding the behavior of the MLD over the EM basin and its adjacent 13 14 coastal region.

Within the EM, most of the studies on the relationship between synoptic circulation and the 15 structure of the MLD over the continental EM were conducted in Israel, the southeastern part of 16 the basin. In particular, several studies were undertaken to characterize the spatial and temporal 17 behavior of the MLD (Neumann, 1952; Halevy and Steinberger, 1974; Rindsberger, 1974, 1976; 18 Dayan et al., 1988; Glaser et al., 1993; Lieman and Alpert, 1993; Dayan et al., 1996; Dayan and 19 20 Rodnizki, 1999; Dayan et al., 2002; Ziv et al., 2004) using sounding measurements at the Israel Meteorological Service permanent site in Beit-Dagan (31.99°N, 34.82°E, 39 m a.s.l.), 8 km 21 22 southeast of Tel Aviv and at other sporadic sounding sites.

The atmospheric noon-time mixed layer during the summer over the EM region is featured by a persistent elevated inversion base formed by a clear boundary line separating two differing air masses, a cool and humid mass above ground capped by a much warmer and subsiding dry air. The MLD is controlled by the interlaced relationship between the downward motion of the subsiding air aloft and the depth of the PT at the surface (Fig. 9).

Due to the existing correlation between the MLD featuring the PT and air pollution episodes over
the EM evidenced in previous studies (Dayan and Graber 1981; Dayan et al., 1988; Koch and





- 1 Dayan, 1992), this barometric system was classified into three essential types (Fig. 10) defined by
- 2 the surface-pressure difference between Nicosia (35.16°N, 33.36°E, 149 m a.s.l.) in Cyprus and
- 3 Cairo (30.1°N, 31.4°E, 75 m a.s.l.) in Egypt, and the temperature at 850-hPa in Beit-Dagan
- 4 (Israel):
- 5 a) Moderate PT: with a horizontal pressure, gradient Nicosia-Cairo within 1.0–1.9 hPa, or even
- 6 greater than 1.9 hPa in case the 850 hPa temperature exceeds 20° C; this surface pressure pattern
- 7 occurs usually in tandem with a weak zonal flow aloft, on the northern edge of the subtropical high
- 8 pressure system;
- b) Shallow PT: with a horizontal pressure gradient (Nicosia–Cairo) less than 0.9 hPa topped by an
 upper atmospheric ridge;
- 11 c) Deep PT: deepening of the PT described in b), due to a temporary retreat of the subtropical high,
- 12 featured by a cyclonic flow aloft; there, the pressure gradient (Nicosia–Cairo) exceeds 2.0 hPa and
- 13 the 850-hPa temperature is less than 20° C.

Analyses of upper air measurements carried out regularly at Beit-Dagan, in the central coastal 14 15 plain of Israel, point at significant differences of the MLD for the several modes of the PT. The overall summer mean noon time mixing depth values for 1981-84 is 764 ±320 m (Dayan et al., 16 17 1988). A classification along the modes defined above resulted in mean and standard deviation 18 values of 428 ± 144 m and 1010 ± 214 m for the shallow and deep mode respectively (Koch and Dayan, 1992). The spatial distribution of the mixing depth is rather homogeneous under deep PT 19 conditions over the central coastal plain of Israel as compared to the shallow mode where its value 20 21 is kept almost uniform above sea-level while penetrating inland. However, the thickness of the mixed layer changes markedly, for the first kilometers, when moving onshore from the coastline 22 inland. The parabolic profile for inland MLD variation, suggested on theoretical grounds (Stunder 23 and Sethuraman, 1985) and observed empirically by Berman et al. (1999) indicates that interior 24 25 land sites tend to develop the highest average daytime MLD with values falling rapidly as the 26 coastline is approached and reaching minimum values less than 500 m offshore.

Due to the important implication of this behavior on the building up concentration of air pollutants,
the lateral variance of the mixing depth was tested for part of the upper air measurements
performed at 4 sites concurrently during the 1981-1984 campaign (Dayan et al., 1988). These sites





- 1 on a west-east transect were: Nizanim (31.7°N, 34.63°E, 10 m a.s.l.) on the southern coastal shore
- 2 of Israel; Beit-Dagan (31.99°N, 34.82°E, 39 m a.s.l.) on the coastal plain; Ruchama (31.5°N,
- 3 34.7°E, 210 m a.s.l.) ~20 km inland in the northern Negev Desert; and Jerusalem (31.77°N,
- 4 35.21° E, 786 m a.s.l.). The average thickness of the mixed layer when moving from the coast
- 5 inland is reduced by 350 m while reaching Jerusalem (Fig. 11). This result is consistent with
- 6 Halevy and Steinberger (1974). However, airplane measurements indicate that the top of the mixed
- 7 layer over the Dead Sea, located ~80 km inland, is within 1400–1600 m a.s.l.
- This marked onshore decrease in the MLD has an important implication by influencing the pollution transported from the EM inland (Luria et al., 1984; Alper-Siman Tov et al., 1997; Luria et al., 1996). Dayan et al. (1988) has tested the longitudinal variance of the MLD North-South vertical cross section on 21 summer noon-time upper air measurements performed simultaneously at 3 sites ~60 km apart along the Israeli coast: Beit-Dagan, Ruchama and Revivim (31.04° N, 34.72° E, 286 m a.s.l.). Their respective MLD mean was evaluated to 896, 781 and 805 m a.s.l. with a standard deviation of 439, 401 and 329 m a.s.l., respectively.
- A comparison between these 3 MLDs revealed that the mixed layer at Beit-Dagan is significantly thicker than either at Ruchama or Revivim. This finding is explained by the greater distance of the southern sites (i.e., Ruchama and Revivim) from the cyclonic core of the PT (which persists in summer to the northeast of Israel) as well as the decreased distance from the anticyclonic center of the North African subtropical high (which persists during all seasons to the southwest of Israel). The lateral and longitudinal spatial MLD above-mentioned values indicate that the most reduced summer MLDs are expected over the southeastern coast of the EM.
- Characterizing the structure of the MLD spatial variation offshore over the EM basin is of importance for getting a better insight on the processes which control the dispersion of contaminants over the sea. Few investigators (Gamo et al. (1982) and Kuwagata et al. (1990) for Japan; Stunder and Sethuraman (1985) for the United States; Gryning (1985) for Denmark) have analyzed the spatial variations of the atmospheric mixing layer in coastal areas. Similar studies as related to the EM Basin are quite limited and deal also mainly on the conditions not directly located over the open sea but rather at sites distant from the coast line.





1 In a 2006-2011 study based on a remote sensing tool, the European Center for Medium Range 2 Weather Forecasts (ECMWF) (model and radiosonde observations launched at Thessaloniki's airport (Greece, 40.6°N, 22.9°E, 10 m a.s.l.) ~1km from the coastline, Leventidu et al. (2013) 3 found the MLD seasonal cycles peak with a summer maximum of 1400, 1800, and 2100 m a.s.l. 4 in June, July and August, respectively. In the unique study of this type we are aware of, Dayan et 5 6 al. (1996) have evaluated the spatial and seasonal distribution of the MLD over the whole Mediterranean Basin. Based on ~65000 air measurements from 45 radiosonde stations within and 7 8 surrounding the basin from spring 1986 through winter 1988, the MLD was derived from the 9 potential temperature gradient measured within the boundary layer and the capping stable layer above it. As expected, the summer values prove to be generally higher over land and minimum 10 over the most eastern and western limits of the Mediterranean Basin (Fig. 12). They concluded 11 that the distance from the coastline and topography are the main factors influencing the spatial 12 distribution of the MLD. The steep gradient in MLD values observed as moving onshore is 13 14 consistent with the elevated summer values in Thessaloniki (Greece) reported by Leventidu et al. (2013). 15

Moreover, Dayan et al. (1996) found that the most striking temporal effect on MLD distribution 16 17 over the basin is caused by synoptic weather systems and the intensity of the sea-breeze along the coast. The diminishing of the MLD over the Mediterranean Basin as moving from its center 18 eastwards toward the EM coast they have observed is consistent with the unique series of 19 measurements of the temperature profiles performed during the summer of 1987 near Ashdod 20 Harbor (31.82°N, 34.65°E), some 40 km south of Tel-Aviv (Israel) at 2 to 22 km from shore using 21 22 a tethered balloon where prominent inversion bases of 350 to 600 m a.s.l. were observed (Barkan 23 and Feliks, 1993). Moreover, such limited MLD values over the sea were obtained in the airborne Gradient in Longitude of Atmospheric constituents above the Mediterranean basin (GLAM) 24 25 campaign in August 2014 (Zbinden et al., 2016): the MLD over the sea measured in the period 6-26 10 August 2014 was approx. 800 m a.s.l. over Crete diminishing to about 400-500 m a.s.l. over 27 Cyprus.

28 2.3 Diurnal behavior of the mixing layer depth over the Eastern Mediterranean coastal zone

The diurnal behavior of the mixed-layer depth is assessed in the coastal plain of the EM based on routine radiosonde ascents that are, unfortunately, of coarse temporal resolution. The MLD reaches





- 1 a summer minimum at 750 m and a winter maximum at1800 m. The hourly maximum MLD is
- 2 between 23:00 UTC and 05:00 UTC for all seasons and decreases gradually toward its minimal
- 3 value at 18:00 UTC.

4 However, since this cycle is governed mainly by synoptic weather systems, and the strength of the 5 sea-breeze, this behavior would be more significant for the summer. During this season, the variation of the mixed-layer height due to diurnal variations of solar radiation and local terrain 6 7 effects is not obstructed by large-scale variations caused by frequent transitions between different synoptic configurations, as featured by other seasons. Consequently, MLD variation is most 8 9 evident during the summer, mainly controlled by the daily sea-breeze cycle and heat fluxes that are most intensive then. The layer minimal depth, along the coast, is usually observed during late 10 afternoon hours when heat fluxes dissipate rapidly and the wind speed of the cool sea breeze 11 12 reaches its minimal rate. This process results in a decrease of the marine turbulent boundary layer 13 (Dayan and Rodnizki, 1999).

In series of measurements in June and August 1987 conducted over the sea near the Israeli coast, Barkan and Feliks (1993) found a prominent diurnal oscillation in the inversion base. During daytime, the inversion base moved downward reaching its minimum height (250–450 m a.s.l.) in the afternoon and evening. During the night, it moved upward reaching its maximum height in the late night and early morning (450–850 m a.s.l.). These diurnal variations are due to the intensity of the sea and land breezes as shown by Feliks (1992).

20 **2.4 Ventilation rates over the Eastern Mediterranean coastal zone**

21 The rate of potential transport of atmospheric pollutants is strongly affected by the strength of the 22 wind flow within the mixed layer and by its depth. Further along the EM coastal zone, both coastalbreeze circulation and large-scale synoptic situation regulate the mean MLD. Assessing the 23 atmospheric dispersion conditions is commonly derived from the ventilation rates calculation. This 24 term is the MLD multiplied by the mean wind speed in the mixed layer, representing the potential 25 26 of the atmosphere to dilute and transport contaminants away from a source region. Matvev et al. (2002) have calculated over 1948-1999 the mean and standard deviation of the mixing depth, wind 27 28 speed and long-term range of ventilation rates at the Israel Meteorological Service sounding site





- 1 in Beit-Dagan (Israel). Their results in Table 1 clearly show that the reduced ventilation rates
- 2 characterizing the EM during summer inhibit an efficient dispersion of pollutants.
- 3 Table 1: Monthly long-term means (LTM) and standard deviation (S.D.) of the mixing layer depth
- 4 (MLD), wind speed and range of ventilation rates over Beit-Dagan in the central coast of the EM.
- 5 LTM and S.D. values for MLD for the years: 1955-1968 (Rindsberger 1974), 1981-1984 (Dayan
- 6 et al. 1988), 1987-1989 (Dayan and Rodniski 1999). LTM and S.D. values for wind speeds are
- 7 from the NCEP/NCAR Reanalysis Project (NOAA- CIRES Climate Diagnostic Center) for a 51-
- 8 year data record: 1948-1999 from their Web site at http://www.esrl.noaa.gov/psd/ (from Matvev
- 9 et al. (2002), permission requested from Elsevier).

10

	MLD		Wind Speed	1	Ventilation Rates
Month	LTM	S.D.	LTM	S.D.	Range of LTM
	(m)	(m)	(m s- ¹)	(m s- ¹)	$(m^2 s^{-1})$
January	1700	950	7.0	4.75	1688 - 31138
February	1830	1135	7.5	5.00	1738 - 37063
March	1790	860	7.0	4.75	2093 - 31138
April	850	660	5.5	3.75	333 - 13968
May	750	525	4.5	3.25	280 - 9880
June	810	470	5.5	2.25	1105 - 9920
July	870	450	5.0	1.65	1365 - 8780
August	820	395	4.5	1.50	1275 - 7290
September	930	510	5.0	2.10	1218 - 10224
October	1650	910	4.5	3.50	740 - 20480





November	1500	940	5.5	4.50	560 - 24400
December	1700	990	6.0	4.50	1065 - 28245

1 2.5 Air mass origins over the Eastern Mediterranean

Studies of the long-range transport (LRT) of pollution by trajectory models help us interpret and 2 better define the movement and removal processes affecting atmospheric concentrations. Although 3 4 changes in wind direction are observed on a diurnal and seasonal basis depending on the synoptic conditions affecting the region, the prevailing wind flows over the EM are from the west towards 5 6 the east. Therefore, air pollutants emitted from upwind sources to the west of the EM will reach 7 the EM and will be added to those emitted locally. Indeed, numerous observational and modeling 8 studies have confirmed that the EM is affected by the long-range transport of air pollutants originating from Europe (e.g., Dayan, 1986; Luria et al., 1996; Wanger et al., 2000; Erel et al., 9 2002, 2007 and 2013, Matvev et al., 2002; Rudich et al., 2008; Drori et al., 2012). 10

To get an insight on the LRT over the EM, the Air Resources Laboratory's trajectory model (GAMBIT- Gridded Atmospheric Multi-Level Backward Isobaric Trajectories; Harris, 1982) was applied over 1978-1982 (Dayan, 1986). The duration of each trajectory was chosen as 5-days backward in time enabling the tracing of air masses originating from Europe, the Mediterranean Basin (MB), North Africa and the Near East close to the EM central coast of Israel (Fig. 13).

The input wind data for GAMBIT, which originate at the National Meteorological Center (NMC) 16 17 in Suitland, Maryland (USA), consist of gridded wind components at standard pressure levels produced by a global atmospheric model. The GAMBIT calculates trajectories to each receptor 18 19 twice daily at standard pressure surfaces. The 850-hPa level (~1500 m a.s.l.) was chosen as the 20 most representative of the transport layer. This level is selected as the intermediary level between the surface wind regime and the regime of upper winds relatively free from local surface effects. 21 22 Trajectory direction was divided into five distinctive geographical classes as shown in Figure 13. 23 Respective occurrences and seasonal distributions can be summarized as follows:

Long fetch of maritime air masses from northwest Europe crossing the Mediterranean Sea,accounting for 36%, was the most frequent on average and fell evenly throughout the whole year;





- 1 2) Northeast continental flow that originated in Eastern Europe, accounting for 30%, was the most
- 2 frequent during the summer season;
- 3 3) Southeast flow from the Arabian Peninsula, accounting for 5%, was infrequent, occurring
- 4 mainly during the autumn;
- 5 4a) Southwest flow along the North African coast, accounting for 11%, was the most frequent
- 6 during late winter and spring; and
- 7 4b) South-southwest flow from inland North Africa was accounting for 7%, with a late winter and
- 8 spring maximum.
- 9 Therefore, 1) and 2) trajectory types are indeed predominant with a summer maximum occurrence
 10 (>66%) over the EM.
- 11

12 **3.** Summer atmospheric pollutant concentrations

The EM is one of the regions in the world where elevated concentration of primary and secondary gaseous air pollutants has been reported frequently. This region is influenced not only by local atmospheric dispersion conditions but also by the ability of the atmosphere to inherit a significant proportion of pollutants from European sources and mineral dust that arrive from two major source regions: the North African Sahara and the Arabian deserts. However, in general, mineral dust does not affect the EM during summer (e.g., Dayan et al., 1991; Moulin et al., 1998; Sciare et al., 2003) and they are not in the scope of this study focused on summer conditions.

After reviewing the atmospheric dispersion and transport conditions characterizing the EM during the summer, a summary of the essential results published over the last decade dealing with trace gases and anthropogenic sulfate aerosol concentrations over this region is presented. These studies demonstrate how the above described global and synoptic scale processes control the extent of trans-boundary transport of air pollutants and chemical composition and concentrations over the EM.

26 **3.1 Processes controlling O₃ formation**





- 1 Most tropospheric O₃ formation occurs when nitrogen oxides (NOx), carbon monoxide (CO) and
- 2 volatile organic compounds (VOCs) react in the atmosphere in the presence of sunlight. Due to
- 3 cloud free conditions, high incoming solar radiation characterizes the EM during summer
- 4 (Lelieveld et al., 2002) which enhances the building-up of O₃ concentrations.
- 5 Numerous researchers have identified the EM as a hot spot" of summertime tropospheric ozone
- 6 (e.g., Stohl et al., 2001; Roelofs et al., 2003; Zbinden et al., 2013; Zanis et al., 2014; Doche et al.,
- 7 2014; Safieddine et al., 2014).

8 Zbinden et al. (2013) derived the climatological profiles and column contents of tropospheric O_3 9 from the Measurements of Ozone by Airbus Aircraft program (MOZAIC) over the mid-northern latitudes (24°N to 50°N) over the 1994-2009 period. Among the 11 most visited sites by the 10 MOZAIC aircrafts, is the EM cluster which comprises 702 profile data from Cairo (31.39°E, 11 30.10°N) Egypt and Tel-Aviv (34.89°E, 32.00°N) Israel airports from which monthly means were 12 derived. The O₃ volume mixing ratio obtained were converted to Dobson units (DU) and validated 13 against coincident ozonesonde profiles. Considering all sites, the EM reaches the largest 14 tropospheric O₃ column concentration of 43.2 DU in July, in agreement with the summer extreme 15 16 found by Ziemke et al. (2011), pointing at the favorable photochemical conditions characterizing 17 this region.

Zanis et al. (2014) identified a summertime pool with high O₃ concentrations in the mid-18 troposphere over the EM over the 1998-2009 period as derived from the ERA-Interim reanalysis 19 O₃ data, Tropospheric Emission Spectrometer (TES) satellite O₃ data, and simulations with the 20 EMAC (ECHAM5-MESSy) atmospheric chemistry-climate model. They indicated that the high 21 O3 pool over the mid-troposphere is controlled by the downward transport from the upper-22 troposphere and lower-stratosphere over this part of the MB, which is characterized by large-scale 23 subsidence. This subsidence is regulated by the Asian Monsoon as described in Section 2.1. 24 25 Furthermore, Zanis et al. (2014) based on previous case studies (e.g., Galani et al., 2003; Akritidis et al., 2010) and climatological studies (e.g., Sprenger and Wernli, 2003; James et al., 2003) and 26 27 their own results deduced that the mechanism leading to high tropospheric O₃ over the EM consists of two essential consecutive phases. Enrichment of stratospheric O₃ into the upper-troposphere via 28 29 a stratosphere-to-troposphere transport process. In the second stage, these ozone-rich air masses 30 are transported downward by the strong summertime subsidence characterizing this region.





1 Doche et al. (2014) analyzed tropospheric O_3 concentrations for the 2007-2012 period as observed 2 over the MB by the Infrared Atmospheric Sounding Interferometer (IASI). They identified an abrupt west-east O₃ gradient in the lower troposphere over the MB with the highest concentrations 3 observed over its eastern part. These concentrations were observed at mid-tropospheric layers 4 (3 km) caused by subsiding ozone-rich air masses from the upper-troposphere, typifying summer. 5 6 A clear and consistent seasonal variability emerges from their measurements, showing a maximum of the 3-km partial column O_3 concentration in July (Fig. 14). This is consistent with the study of 7 8 Tyrlis and Lelieveld (2013) who found that the key global and synoptic driving factors yielding to 9 high O₃ concentrations in July is the strength of the Etesian winds and the subsidence over the EM, which is temporally very well correlated with the Indian monsoon. 10 11 Based on IASI measurements and the Weather Research and Forecasting Model with Chemistry (WRF-Chem), Safieddine et al. (2014) have shown that the air column of the first 2 km above 12 ground is enriched by anthropogenic O_3 . Above 4 km, O_3 is mostly originating from outside the 13 MB by LRT process or generated through stratosphere-to-troposphere exchange characterizing the 14

15 eastern part of the MB during the summer.

16 Air masses from surrounding regions in the EM atmosphere have a great impact on surface O_3 concentrations. In a recent study, Myriokefalitakis et al (2016) have investigated the contribution 17 of LRT on O₃ and CO budget in the EM basin, using a global chemistry transport model (CTM), 18 the TM4-ECPL, driven by ECMWF Interim re-analysis project (ERA-Interim) meteorology 19 20 (Daskalakis et al., 2015). They found that about 8% of surface O₃ concentrations are affected by local anthropogenic emissions, whereas subsiding air masses from the free-troposphere and 21 22 horizontal transport from surrounding regions provide about 38% and 51% of O_3 sources, 23 respectively, into the EM-MLD. Although elevated O_3 concentrations over the EM during the 24 summer are mainly attributed to LRT of polluted air masses originating from Europe and lingering 25 over the MB, its enhancement as a secondary pollutant is also caused by its precursors emitted along the coasts of the EM. Consequently, several studies dealing with O₃ concentrations measured 26 27 over coastal sites surrounding the EM and its inland penetration are presented.

O₃ measurements were performed over the northern coast of Crete, at Finokalia (35°30'N, 25°70'E)
70 km northeast of Heraklion, from September 1997 to September 2000 and from a rural area
(40°32'N, 23°50'E) close to Thessaloniki in the north of Greece from March 2000 to January 2001.





1 Based on these measurements, Kouvarakis et al. (2002) pointed out the existence of a well-defined 2 seasonal cycle in boundary layer O₃, with a summer maximum both above the Aegean Sea and at Finokalia. They indicated that LRT is the main factor accounting for the elevated O₃ levels above 3 the EM. This finding is consistent with the 1997–2004 surface O₃ time series at Finokalia (Crete) 4 of Gerasopoulos et al. (2005) who investigated the mechanisms that control O_3 levels and its 5 variability. They identified transport from the European continent as the main mechanism 6 controlling the O_3 levels in the EM, especially during summer when O_3 reaches a July maximum 7 8 of 58 ±10 ppbv. Moreover, on a larger regional scale, Kourtidis et al. (2002) used ozonesonde 9 ascents, lidar observations, ship cruises, and aircraft flights, to show that south and southwestern synoptic flows associated with Saharan dust events result in lower O₃ above the planetary 10 boundary layer by 20–35 ppbv as compared to these during northerly flows, which transport air 11 from continental Europe. Kleanthous et al. (2014) based on sixteen years of O_3 concentrations 12 measured at the EMEP Agia Marina Xyliatou rural station (Cyprus) and 3 other remote marine 13 14 sites, over the western, central and eastern parts of the island have shown that local precursors contribute to only about 6% (~3 ppbv) of the observed O₃. However, elevated concentrations of 15 this secondary pollutant occurring in summer are attributed to LRT of air masses mainly 16 originating from northerly and westerly directions. The summer average annual maximum of 54.3 17 18 ± 4.7 ppb was observed to be related to the transport of polluted air masses from Near-Asia, East 19 and Central Europe toward Cyprus.

Despite the prevailing synoptic meteorological conditions featuring the EM in summer, the differing pathways of the LRT of polluted air masses can affect differently the buildup of pollutants concentrations. To investigate such changes, Wanger et al. (2000) performed a comprehensive study that included 150 hours of instrumented aircraft monitoring flights comparing two events of air mass transport (September 1993 and June 1994) representing two distinct types of LRT. This airborne study comprised flight paths taken approximately 70 km offshore paralleled to the Israeli coastline and 180 km in length with Tel-Aviv in the center.

These flights were performed during midday under westerly wind flow conditions and at an altitude of 300 m a.s.l. (well within the mixed layer). While both wind flow conditions were nearly similar through the measurement periods and along the 180-km flight path, the air mass sampled in September 1993 was much "cleaner" than the one sampled in June 1994. The O₃ concentration





averaged for the first campaign was 39 ±7 ppb, against 48 ±9 ppb in the second period. Therefore,
Wanger et al. (2000) suggested the two air masses travelled different paths before reaching the EM
region. Model simulation showed that the pollution sources in southern Europe and the Balkans
did not affect the EM coasts in September 1993, contrarily to the synoptic conditions and
simulation results for the June 1994 period where the winds over the EM tended to be
northwesterly and thus forcing the polluted air masses toward the coasts of the EM.

7 The summer synoptic and dynamic conditions prevailing over the EM supply the essential ingredients for the building up of O_3 concentrations. Based on the similar climatic conditions 8 9 between the Los Angeles Basin and the EM, Dayan and Koch (1996) proposed a theoretical description of the cyclic mechanism in summer, leading to fumigation (i.e., a downward dispersion 10 of an enriched O₃ cloud toward the ground) further inland from the EM coast. Under the deep 11 12 mode of the PT, stronger westerly winds, acting as a weak cold front (Fig. 15, panel A1) penetrate far inland, undercutting the mixed layer polluted by O_3 from the previous day (Fig. 15, panel A2). 13 In this way, part of the mixed layer containing O_3 is pushed upward and isolated from the ground. 14 15 If the pressure gradient weakens on the following day, the western flow weakens (Fig. 15, panel B1). The cooling effect of the cool and moist marine air is consequently reduced and the convective 16 17 boundary layer inflates rapidly. When the top of the mixed layer reaches the elevated O_3 cloud, the latter is penetrated by convective currents (Fig. 15, panel B2) and parts of the cloud are 18 19 entrained toward the ground in this fumigating process.

20 Elevated O_3 concentrations (>117 ppb) were measured at inland rural sites during the 1988-1991 early summer months (Peleg et al., 1994). Based on air mass back-trajectory analyses, these 21 22 elevated O₃ mixing ratios at rural sites over central Israel were found only in case of air masses 23 overpassing Tel Aviv metropolitan area. Furthermore, the very low ratio of SO_2/NO_x clearly 24 indicates that O₃ precursors such as nitrogen oxides (NOx, the sum of NO and NO₂), carbon 25 monoxide (CO), and volatile organic compounds (VOC) originate mainly from fossil-fuel combustion from mobile sources (Nirel and Dayan, 2001). These pollutants are subjected to 26 27 chemical and photochemical transformations in the presence of solar radiation and atmospheric 28 free radicals to form O₃.

Over central Israel, the main source for these precursors emitted along the Israeli coastline is transportation (Peleg et al., 1994). Since O₃ formation and other secondary pollutants takes several





1 hours, significant transport and mixing occur simultaneously with the chemical reactions (Seinfeld, 2 1989; Kley, 1997). Thus, increasing urban and commercial activity along the highly populated Israeli coastal region, together with expanding transportation activity in the Gaza region, was 3 4 found to strongly deteriorate inland air quality and, specifically, to cause increasingly elevated inland O_3 levels. Model results showed that traffic emissions during the morning rush hour from 5 6 the Tel Aviv metropolitan area contribute about 60% to the observed O₃ concentrations (Ranmar 7 et al., 2002). Moreover, their study showed the summer season features a shallow mixed layer and 8 weak zonal flow, leading to poor ventilation rates which restrict O₃ dispersion efficiency. These 9 poor ventilation rates result in the slow transport of O₃ precursors, enabling their photochemical transformation under intense solar radiation during their travel from the EM coast inland. 10 However, elevated O₃ concentrations are not limited to the summer over the EM. Dayan and Levy 11

(2002) found 103 "high-ozone days" where O₃ is >80 ppbv for at least 2 hours based on 24 Israeli
sites over 1997-1999. From their O₃ temporal analyses they concluded that the highest values are
more frequent during the transitional (spring and autumn) seasons (65% of 103 days) than during
the summer season (35%).

16 Based on the recent remote sensing tools in conjunction with meteorological observations and 17 models, we conclude on the three essential processes that control the O_3 concentration during summer at various tropospheric levels over the EM: 1/ in the shallow troposphere, the horizontal 18 transport of O₃-enriched air masses from eastern continental Europe to the region controlled by 19 20 the position and strength of the Azores High and the PT; 2/ the dynamic subsidence at midtropospheric levels; and 3/ the stratosphere-to-troposphere exchange in the upper troposphere. At 21 22 the surface of the EM coast, during transitional seasons, high O_3 episodes are associated with hot 23 and dry air masses originating east of Israel, where O_3 precursor emissions are negligible, 24 demonstrating that high O₃ levels are more dependent on air mass characteristics than on upwind 25 precursor emissions.

26 **3.2 Particulate sulfate (SO4) abundance**

Globally, the two main particulate sulfate (SO₄) precursors are sulfur dioxide (SO₂) from anthropogenic sources and volcanoes, and dimethyl sulfide (DMS) from biogenic sources,





- 1 especially marine plankton. In the EM atmosphere, particulate sulfate contributes more than 50%
- 2 to the submicron aerosol mass (Bardouki et al., 2003a, b; Sciare et al., 2005).
- 3 Several factors favor particulate sulfate abundance over the EM. The homogeneous conversion of 4 gaseous SO₂ to particulate sulfate is rather slow, i.e., about 1-3% per hour (Meagher et al., 1981). 5 Wet deposition chiefly governs the atmospheric lifetime of sulfate, estimated to be up to 6 days on a global average (Chin et al., 2000). Due to rainless conditions and associated wet deposition in 6 summer, and the slow dry deposition velocity of sulfate aerosol (~ 0.01- 0.4 cm s⁻¹), sulfate 7 aerosols account for 50-90% of the total sulfur in transported air masses toward the EM (Mate et 8 9 al., 2002). Two additional factors favor late spring and summer particulate sulfate regional abundance. First is the intense radiant energy emitted by the sun under clear sky conditions that 10 leads to an efficient oxidation of SO₂ to SO₄ via hydroxyl radical (OH) as the predominant oxidant 11 12 during daytime (Mihalopoulos et al., 2007). Second is the prevailing summertime westerly winds that transport sulfate-rich air masses from sources over central Europe before significant removal 13 occurs. This is consistent with Sciare et al. (2003) who measured particulate non-sea-salt SO₄ 14 15 during a one-month experiment in summer 2000 at a background site on Crete, which resulted to a high average concentration of $6 \mu g m^{-3}$ (~62 nmole m⁻³) for air masses originating from Turkey 16 17 and Central Europe.

Another source of sulfate aerosols is ship emissions, which contribute substantially to atmospheric pollution over the summertime Mediterranean region. Marmer and Langmann (2005) based on a regional atmospheric-chemistry model and a radiation model found that the summer mean sulfate aerosol column burden over the Mediterranean is 7.8 mg m $^{-2}$, 54% originating from ship emissions.

23 Concentrations of sulfate-rich air masses have been measured intermittently at various downwind ground sites in Israel, the easternmost Mediterranean region, from an instrumented aircraft for a 24 25 10-year period between 1984 and 1993 by Luria et al. (1996). They found that the concentration of particulate sulfate observed during the summer was relatively high compared with other world 26 locations. The highest values occasionally exceed 500 nmole m⁻³. Wintertime levels were in the 27 range of 50-100 nmole m⁻³. In their airborne study, Wanger et al. (2000) measured an averaged 28 SO_4 concentration of 38 ±7 nmole m⁻³ in their first series of measurement between 5 and 9 29 September 1993, and up to 108 ± 63 nmole m⁻³ between 15 and 21 June, 1994. The annual average, 30





calculated in Luria et al. (1996), is 100 ± 15 nmole m⁻³, which is twice as high as predicted for the 1 region by a global model and as high as reported for some of the most polluted regions in the USA. 2 They pointed to several indicators suggesting that the origin of the particulate sulfate over the EM 3 region is not from local sources but the result of LRT. The indicators include the lack of correlation 4 between SO_4 and primary pollutants, the high SO_4 to total sulfur values, the origin of the air mass 5 6 back trajectories and the fact that similar levels were observed during concurrent periods at 7 different sites. Throughout their study, a higher concentration of SO₄ was found during the 8 afternoon hours, especially during the summer and at the inland locations.

9 In another attempt to quantify the sulfur flux arriving at Israel's western coast from Europe and the Israeli pollution contribution to the air masses leaving its eastern borders towards Jordan, 10 Matvev et al. (2002) conducted 14 research flights at an altitude of approximately 300 m above 11 12 ground level, measuring sulfur dioxide and particulate sulfates during the summer and autumn seasons. Two different legs were performed for each research flight: the first over the 13 Mediterranean Sea, west of the Israeli coast and the second along the Jordan Valley. Their results 14 15 have shown that the influx of sulfur (S) reaching the Israeli coast from Europe varied in the range of 1–30 mg S h^{-1} , depending on the measuring season. The SO₄ level in the incoming LRT air 16 17 masses was at least 50% of the total sulfur content. The contribution of the local pollutant sources to the outgoing easterly fluxes also varied strongly per season. The Israeli sources contributed an 18 average of 25 mg S h⁻¹ to the total pollution flux during the early and late summer as compared to 19 only approximately $9 \text{ mg S} h^{-1}$ during the autumn period. Synoptic analysis indicates that 20 conditions during the summer in Israel favor the accumulation of pollution species above the MB 21 22 from upwind European sources. This season is characterized by weak zonal flow within a shallow 23 mixed layer that lead to poor ventilation rates, limiting an efficient dispersion of these pollutants 24 during their transport eastward. Under these summer conditions, in-flux local contribution and the 25 total out-flux of these pollutants are elevated as opposed to other seasons. To illustrate, during the 26 autumn, the EM is usually subjected to weak easterly winds, interrupted at times by strong westerly 27 wind flows inducing higher ventilation rates. Such autumnal meteorological conditions and the lack of major emitting sources eastwards of Israel result in lower sulfur budgets to and from Israel. 28

An estimate of the yearly flux showed that approximately 0.06 Tg S arrived at the Israeli coast from the west (Matvev et al., 2002). This is approximately 15% of the pollution leaving Europe





- 1 towards the EM. The outgoing flux towards Jordan contributed by local sources was calculated to
- 2 be 0.13 Tg S per year, i.e. almost all the sulfur air pollution emitted in Israel. The results of the
- 3 flux rates for the sulfur compounds over Israel are summarized in Table 2 for the different research
- 4 flights and field campaigns. These latter results show that the uppermost fluxes from the west were
- 5 for the early summer time, averaging 0.19 Tg y^{-1} . During this season, the levels doubled the
- 6 averages for late summer $(0.085 \text{ Tg y}^{-1})$ and were over five times the average levels measured for
- 7 the autumn (0.035 Tg y⁻¹). The wide range in fluxes derived is explained by the varying distance
- 8 from the polluted coastline

9

10





- 1 Table 2. Compilation from Rudich et al. (2008) of sulfate particulate concentrations and yearly
- 2 fluxes in [a] Luria et al. (1996), [b] Wanger et al. (2000) and [c] Matvev et al. (2002).

			Yearly		
	Measurement	Conc. Avg	Flux		
Regions	Periods	(nmole m ³)	(Tg y ⁻¹) (*)	Authors	
	July-Aug. 1984,1986	86	0.08		
	May-June 1989	70	0.06		
Judean mountains	July-Aug. 1987,1988	103	0.09	[a]	
	July-Aug. 1990	128	0.12		
	May, July 1990, 1991	85	0.08	[a]	
Sea of Galilee	AugSept. 1993	87	0.03		
	Dec. 1993	71	0.07		
North coastal plain	June 1993	106	0.12		
	Sept. 1993	38	0.08		
	June 1994 (**)	108	0.22	[b]	
Eastern Mediterranean coast	June 1998	105	0.16		
	Sept. 1996	26	0.04	[c]	
	Nov. 1995	21	0.03		
	L				

3 (*) conversion from nmole m^{-3} to yearly fluxes was along Matvev et al. (2002) taking into

4 account the vector component of onshore wind speed, length of flight leg, and the MLD.

5 (**) the June 1994 flight has been performed during a highly-polluted month over Israel.





1 The Aerosol Optical Depth (AOD), the vertical integral over an atmospheric column of the incident 2 light scattered and absorbed by aerosols is often used to estimate aerosol loading in the atmosphere. Nabat et al. (2013) compared AOD from several model data to satellite derived data for the period 3 4 2003-2010 over the Mediterranean region. They found that the seasonal cycle obtained from the Monitoring Atmospheric Composition and Climate (MACC) reanalysis model which includes 5 MODIS AOD assimilation at 550 nm resembles much the satellite-derived AOD variability and 6 7 have the best spatio-temporal correlation compared to AERONET stations. Based on these models 8 and satellite derived data, Nabat et al. (2013) have clearly shown that particulate sulfate, as one of 9 the aerosol types, has a maximum during spring and summer over the eastern MB (Fig. 16). 10 Matvev et al. (2002) performed airborne measurements along a 150-km line west of the Israeli coast. They derived an annual flux of the order of 0.06 Tg yr⁻¹ of (dry) sulfur across the 11 corresponding surface. Given the observed ratio of sulfate to total sulfur of 40-90% in the region 12 13 (Matvev et al., 2002; Sciare et al., 2003), the annual flux of SO₄ based on field measurements is 0.024-0.054 Tg y⁻¹. Rudich et al. (2008) used satellite data to estimate the pollution transport 14 toward the EM. MODIS Terra- and Aqua-derived estimates of the annual sulfate flux along the 15 same transect are 0.038 and 0.040 Tg y⁻¹, respectively, in the middle of the range from field 16 observations. 17

Rudich et al. (2008) also found that MODIS-based estimates of the sulfate flux agree reasonably 18 well with the Goddard Chemistry Aerosol Radiation and Transport (GOCART) model simulations 19 20 of anthropogenic sulfate, as shown in Figure 17 for seasonal averages. Compared to Terra, the GOCART estimates are about 85% and 30% higher in winter and spring, respectively, but 10-21 22 25% lower in summer and autumn than the MODIS-based estimates. The annual sulfate flux from the GOCART model is 0.181 Tg y⁻¹, about 18% larger than the MODIS estimate of 0.153 Tg y⁻¹. 23 GOCART simulations are in excellent agreement with MODIS/Aqua estimates in summer and 24 fall, but about 50% higher in winter and spring. The GOCART simulated annual flux of 25 0.201 Tg y⁻¹ is about 25% higher than the MODIS/Aqua estimate of 0.159 Tg y⁻¹. Based on the 26 comparison of the two instruments, the model results, and the consistency with the aircraft 27 measurements, they suggested that both MODIS instruments can be used for estimating the flux 28 of pollution based on their daily AOD retrievals. 29

30 3.3 Local formation and long-range transport of total reactive nitrogen (NOy)





1 Research studies measuring inorganic reactive nitrogen compounds over marine areas in general, 2 and more specifically over the EM basin are scarce (Lawrence and Crutzen, 1999; Corbett et al., 3 1999; Veceras et al., 2008). Measurements of nitrogen dioxide (NO₂), nitric (HNO₃) and nitrous (HNO₂) acids undertaken with instrumentation onboard a research vessel in the Aegean Sea 4 between 25 to 29 July 2000 revealed typical NO₂ concentrations of 4–6 ppbv with a broad 5 maximum of 20-30 ppbv. The level of NO₂ was relatively high during the night and low during 6 the day due to enhanced photochemical activity, vertical mixing and the daily wind characteristics. 7 8 Extreme NO₂ concentration were caused by up slope wind bringing air from marine traffic 9 emissions trapped within the marine atmospheric boundary layer. The concentration of both nitric and nitrous acids in ambient air of the Aegean Sea was low, below 50 pptv. Večeřa et al. (2008) 10 explained these results by the lack of precursors for these acids (Cohen et al., 2000), the high solar 11 irradiation leading to HNO₃ dissociation, and the reaction of nitric acid with sodium chloride 12 aerosol. 13

NOy, the total reactive nitrogen (NO + NO₂ + HNO₃) identified as precursors in the O₃ formation was measured by Wanger et al. (2000) during two summer airborne campaigns over the EM at an altitude of about 300 m (well within the MLD) using a high-sensitivity NO-NOy analyzer (TEII 42 S, chemiluminescence method, ± 0.1 ppbv sensitivity). In the first campaign of September 1993, characterized by cleaner air mass conditions, an average NOy concentration of 1.0 ± 0.6 ppbv was measured as compared to 3.9 ± 1.8 ppbv sampled during the June 1994 campaign.

The Mediterranean Intensive Oxidant Study (MINOS) campaign, performed in the summer of 2001, allowed Lelieveld et al. (2002) to examine the air pollution conditions at shallow and mid-21 tropospheric layers over the EM Basin. During this experiment, elevated concentrations, typically 23 0.1 to 0.2 ppbv, of nitric oxide (NO) in the upper troposphere and only about 20 pptv within the 24 MLD were observed at the Finokalia station.

Ranmar et al. (2002) addressed the dynamics of transboundary air pollution, where transportation emissions (such as NOx and VOC) originating from Israeli major coastal sources impact the onshore mixing layer. Their aim was to highlight the daily influence of rush-hour traffic emissions from these coasts onto the inland. They focused on the transportation sources in the Tel Aviv and Gaza Strip areas, the main coastal urban areas. This study included numerical simulations, 1994-1997 O₃ airborne measurements, and June-September ground-based measurements within 1999-





2000. A comparison of the coastal NOy spectra (Tel Aviv) to the inland NOy profiles (Modiin,
31.5°N, 35.0°E, 300 m a.s.l.) and Jerusalem discerned different dynamics for these locations.
While the initial levels in Tel Aviv were higher during the morning rush hour emissions, they were
subjected to a noticeable bleaching by the late morning sea breeze in comparison to inland
locations, which leveled off at relatively higher midday concentrations. This may indicate, in the
absence of any alternative NOy source, that the early morning NOx produced by transportation
sources in Tel Aviv is transported inland, providing additional NOy to the regions along its path.

Dayan et al. (2011) measured the atmospheric concentrations of NOx in three urban centers (Tel-8 Aviv, Jerusalem and Beer-Sheva) in Israel before, during, and after the Day of Atonement (varying 9 in the period September-October) for nine consecutive years (2000-2008). This permitted to 10 inquire into the meaning and implications of local versus long-range sources of pollution and to 11 12 probe the effect of meteorological conditions (synoptic systems, ventilation rates, height of surface boundary layer, and wind speed) on the level of pollution. During the Day of Atonement, all traffic 13 and most of the industrial activities cease in the Jewish populated parts of the country. They found 14 15 out that NOx concentrations during this day are significantly lower than the days before and after, reflecting NOx short residence time in the atmosphere and the strong influence of local emission 16 17 sources. Moreover, the difference in concentrations of NOx in different Days of Atonement cannot be accounted for either by variations in ventilation rates, different synoptic conditions, or by height 18 of surface boundary layer, and wind speed. They are determined by the combination of local and 19 20 regional sources of emission and the meteorological conditions which control their transport from these nearby sources to the sampling station. This is consistent with Bruggemann et al. (2009) who 21 22 showed that during week-days the concentrations of elemental and total carbon, certain metals, 23 and the ultra fine $(0.05-0.14 \,\mu\text{m})$ and coarse $(1.2-10 \,\mu\text{m})$ fractions of suspended particulate matter in a German urban center are higher than during weekends when traffic density is 30% lower. 24

Beside cruises of research vessels, airborne campaigns, and ground truth measurements, satelliteborne initiatives have been undertaken to get a better insight on the reactive nitrogen concentrations over the EM. Marmer et al. (2009) used the OMI instrument (Boersma et al., 2007) as an observation tool to measure atmospheric NO₂ column concentrations in order to validate ship emission inventories over the MB. Figure 18 shows the average OMI NO₂ tropospheric columns (gridded to $0.125^{\circ} \times 0.125^{\circ}$) over the Mediterranean Sea for June-August 2006. The most prominent





- 1 feature here is the elevated NO₂ monthly mean. Under cloud free conditions, typical values ranged
- 2 from 1.2 to $2.0 \ 10^{15}$ molecules cm⁻² over the north-eastern African coast, the EM coast, the
- southern coast of Turkey and the whole Aegean Sea, as compared to over 6 10^{15} molecules cm⁻²
- 4 for European inland congested regions. Based on OMI NO2 tropospheric columns and the GEOS-
- 5 Chem chemistry transport model, Vinken et al. (2014) attributed the elevated NO₂ columns regions
- 6 over the Mediterranean to NO₂ emissions along ship tracks.

7 **3.4 Carbon Monoxide sources and pathways**

8 Carbon monoxide (CO) has a global-average lifetime of about two months in the troposphere and 9 its molecular weight is close to that of air. This molecule is considered as an excellent tracer for 10 pollution sources and pollution pathways through the troposphere. In addition to production by 11 chemical oxidation in the atmosphere, CO is emitted from biomass burning, man-made sources, 12 vegetation, and ocean. The CO seasonal cycle is mainly governed by the concentration of the 13 hydroxyl radical (OH) in the troposphere (Novelli et al., 1992) and is expected to be the lowest in 14 the summer when photochemistry is active and the highest during late winter or spring.

CO was measured and used as a tracer in a chemistry transport model (Lelieveld and Dentener, 15 16 2000) during the summer 2001 MINOS campaign (Lelieveld et al., 2002). The model diagnosed 17 CO from anthropogenic sources in different parts of Europe, North America, and Asia. Trajectory calculations in the lower troposphere identified western and eastern Europe as the main source 18 emissions. Consequently, model simulations were performed for August 2001 over Sardinia 19 (40°N, 8°E) in the western Mediterranean (WM) and over Crete (35°N, 25°E). Considering the 20 negligible impact of local pollution sources, the high CO levels observed over Crete, in excess of 21 150 ppby, were surprising. The model results indicated that regions surrounding the Mediterranean 22 such as southern Italy, Greece, Serbia, Macedonia, the Middle East, and North Africa contribute 23 relatively little to the CO pollution, typically about 20%. Furthermore, Lelieveld et al. (2002) 24 25 found that the EM is affected by CO polluted air emitted from Eastern Europe, Poland, the Ukraine, and Russia. This pollution flow, east of the Carpathian Mountains, is channeled over the Black 26 27 Sea and the Aegean Sea, and contributes 60 to 80% of the boundary-layer CO over the EM. Their 28 model results are consistent with aircraft measurements, showing that the entire Mediterranean 29 lower troposphere is polluted.





1 In the free troposphere, where westerly winds predominate, they revealed a quite different situation 2 as compared to concentrations measured within the MLD. The mid-tropospheric CO measurements were ~75-80 ppbv. From their model tracer analysis, the largest contribution over 3 the Mediterranean is found originating from Asia (40 to 50%). The CO typical lifetime 4 $(\sim 2 \text{ months})$ enables air mass to circumnavigate the globe which results in a low variability of its 5 concentrations. Lelieveld et al. (2002) found that contributions by pollution from western and 6 eastern Europe to mid-tropospheric CO were only about 10%. 7 Drori et al. (2012) conducted a study to locate the various CO sources converging from Europe, 8

North Africa and the Middle East and quantify their respective contributions to the EM. 9 Background CO concentrations are monitored regularly over the southern part of Israel in Sde-10 Boker (WIS Station Negev Desert: 31.13°N, 34.88°E, 400 m) as part of the NOAA Earth System 11 Research Laboratory Global Monitoring Division (ESRL/GMD) which aims at representing the 12 EM. While comparing the seasonal cycle of Sde-Boker to other European ESRL/GMD background 13 sites (see Table 3), one essential feature is eminent from their results represented in Figure 19: CO 14 15 concentrations are high over winter months, decreasing abruptly during April and increasing again from November. A second maximum is observed during August compared to July and September 16 17 (Drori et al., 2012).

18

19 Table 3. Locations and elevations of ESRL/GMD sites

Code	Name	Latitude	Longitude	Elevation (m)	Country
WIS	WIS Station Negev Desert	31.13	34.88	400.0	Israel
HUN	Hegyhatsal	46.95	16.65	248.0	Hungary
LMP	Lampedusa	35.52	12.62	45.0	Italy
BSC	Black Sea Constanta	44.17	28.68	3.0	Romania
OXK	Ochsenkopf	50.03	11.80	1022.0	Germany
BAL	Baltic Sea	55.35	17.22	3.0	Poland
MHD	Mace Head County Galway	53.33	-9.90	5.0	Ireland

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1 To get an insight on the spatial distribution of CO concentrations over the EM, the Version 4 2 MOPITT level-2 CO retrievals (Deeter et al., 2010) were employed by Drori et al. (2012) using a priori information for MOPITT V4 CO retrievals based on MOZART simulation climatology. 3 While in-situ measurements represent a small volume at a specific layer, MOPITT retrievals can 4 be regarded as averaged value over several layers (e.g. the 900 hPa retrieval is averaged from the 5 6 surface up to the 700 hPa level). The averaging kernel profile obtained for two retrievals near Sde-Boker ESRL/GMD station shows that, during the day, the 900 hPa retrieval sharply peaks at the 7 8 same level, indicating that there is a good sensitivity to lower tropospheric concentration. During 9 the night, the 900 hPa retrieval moderately peaks at its level with the same contribution from the 10 800 hPa level pointing at a slightly less sensitivity to the lower tropospheric concentration (not 11 shown).

The anomalous high concentration observed at the WIS ESRL/GMD Sde-Boker station, and calculated by the MOZART-4 chemistry-transport model during August (Fig. 19), might be limited to lower levels, and therefore averaging over several layers might hide this signal. MOPITT retrievals are highly variable both temporally and spatially, especially during winter and spring. This variability stems from the highly variable boundary layer height (300–2000 m) featuring the spring and cold fronts sweeping the EM during winter.

Furthermore, Drori et al. (2012) compared the in-situ measurements at Sde-Boker and CO 18 retrieved from MOPITT to a chemical transport model. For this sake, the MOZART-4 global 19 20 chemical transport model of the troposphere (Emmons et al., 2010) was used. CO sources included direct emissions and secondary production from hydrocarbons oxidation, while CO sinks included 21 22 a reaction with OH and dry deposition. The seasonal cycle of surface CO at Sde-Boker simulated 23 by MOZART averaged for five consecutive years shows a similar pattern exhibiting high CO 24 concentration during winter months, reaching a maximum in spring, and decreasing sharply around 25 May. The CO concentration in mid-summer months (i.e., July and August) surpasses those of the early summer (i.e., May-June) (Fig. 19). 26

To attribute the CO sources affecting the EM, Drori et al. (2012) partitioned these sources using a tagging method into five types: anthropogenic, biogenic, fire, chemical production and ocean. The total and specific sources 2006–2007 times-series of MOZART CO concentration at 30° N and 33.75° E are shown in Figure 20, except for ocean, that are negligible and thus not shown. Both





1 biogenic (green line) and biomass burning sources (red line) have a minor contribution. Biogenic 2 sources are characterized by a distinct seasonal cycle with high contribution over winter and low 3 daily variability. Biomass burning has no defined seasonal signature and contributes on an episodic event basis. CO from chemical production (orange) contributes substantially (50-80 ppb) with a 4 defined seasonal cycle: low during winter and autumn and high during summer featured by a low 5 6 daily variability. Anthropogenic sources were found to be the main contributor to the total CO (purple, 50–180 ppb). As expected, their seasonal cycle is featured by winter elevated 7 concentrations decreasing during spring, slightly increasing during summer; and decreasing again 8 9 during autumn. The daily variability is high and similar to the total CO daily variability. Comparing the daily variability of the various sources, Drori et al. (2012) concluded that 10 11 anthropogenic sources mainly govern total CO daily variability over the EM.

12 To further attribute the CO surface daily variation, Drori et al. (2012) tagged the anthropogenic sources for the three northern continents, i.e., North America, Europe, and Asia. Figure 21 shows 13 the results of these anthropogenic sources attribution to the CO surface. European anthropogenic 14 sources contribute substantially (10-80 ppb) to local CO concentrations with the greatest daily 15 variability all year round. Asian and North American sources are in the same order of magnitude 16 17 (10–25 ppb) with low daily variability during most of the year, and very small variability during summer. Obviously, daily summer CO variations in the EM are mainly caused by European 18 19 anthropogenic sources. The seasonal cycle of the European contribution is very similar to the 20 seasonal cycle of total CO, featured by a high concentration in winter, spring, and autumn and a lower summer concentration. The contribution of European emissions to CO surface 21 22 concentrations is comparable to that from EM local emissions.

Drori et al. (2012) found, however, that local and European emission contributions to local CO concentrations are generally negatively correlated, meaning that either local or European sources are dominant, except during summer, when both sources affect simultaneously the local CO concentration. A possible explanation for the positive summer correlation might be explained by the short range of air mass transport caused by the dominant summer synoptic system, i.e., the Persian trough in its weak mode recirculating local and European emissions, and by the fact that summer chemical production is a major CO source over the EM.





1 Another recent modeling study focused on CO concentrations was conducted by Myriokefalitakis 2 et al., (2016). They compared and validated model results against in-situ observations at the surface, in the MLD and in the free troposphere (between 850 hPa and the tropopause) in the 3 countryside and remote atmosphere over Europe for 2008. This study analyzes the total CO budget 4 and the partial contribution of regional anthropogenic, biogenic and biomass burning CO 5 emissions in the EM. The budget calculated for 2008 in the EM MLD, using a basic simulation 6 relying on anthropogenic emissions and meteorology, points at a load of 0.6 Tg of CO, a chemical 7 production of 10 Tg yr⁻¹, primary emissions in the region of 8 Tg yr⁻¹ and a dry deposition flux of 8 3 Tg yr⁻¹. Moreover, Myriokefalitakis et al (2016) found that subsidence from higher atmospheric 9 layers typifying the EM summer is an important CO source (12 Tg yr⁻¹) in the EM free 10 troposphere. At the surface, anthropogenic local emissions in the EM was found to contribute by 11 18% to surface CO levels on an annual average. Over Cairo, out of the total surface CO 12 concentration, roughly 32% are contributed by anthropogenic sources. These EM CO 13 concentration results are consistent with previous modelling studies (e.g., Kanakidou et al., 2011; 14 Drori et al., 2012; Im and Kanakidou, 2012). 15

16 **3.5 Methane concentrations**

CH₄ concentration in the atmosphere is of natural and anthropogenic origins. It is the most 17 affecting greenhouse gas after water vapor and CO₂ due to its high global warming potential as 18 contributed by its infrared absorption and long atmospheric lifetime. CH₄ emissions are primarily 19 20 caused by microbiological decay of organic matter under depleted of dissolved oxygen conditions in wetlands, followed by decomposition of solid waste and enteric fermentation from domestic 21 22 livestock. As for the geologic sources, a total geological CH₄ flux of 53 ± 11 Tg yr⁻¹ was suggested, 23 which accounts for 7-10% of the total global CH₄ budget (Etiope et al., 2008). The geological formations contributing to CH₄ over the Greater Area of the EM (25° N- 50° N, 5° E- 55 °E) are 24 mud volcanoes with essential hot spots located over eastern Romania, the Black Sea, Central and 25 Eastern Azerbaijan and the Caspian Sea. 26

Georgoulias et al. (2011) used data from the SCIAMACHY (Scanning Imaging Absorption Spectrometer for Atmospheric Cartography) instrument on board the European environmental satellite ENVISAT (Bovensmann et al., 1999). SCIAMACHY's spectral near-infrared nadir measurements are sensitive to CH₄ and CO₂ concentration changes at all atmospheric altitudes,





1 including the MLD where the signal emitted from the surface source is the largest. Annual, 2 seasonal and monthly spatial distribution of CH₄ were displayed for 2003 and 2004 based on the analysis of Weighting Function Modified Differential Optical Absorption Spectroscopy (WFM-3 4 DOAS) version 1.0 (Schneising et al., 2009) dry air column-averaged mole fractions, denoted as XCH_4 (ppb). The reflectivity of water surfaces is very low, therefore Georgoulias et al. (2011) 5 6 presented their results over land only by depicting the concentration of CH₄ along the coasts of the EM Basin. To reduce the noise inserted by the single pixel retrieval error and the temporal and 7 8 spatial sparsity of the data, the data were averaged on $1^{\circ} \times 1^{\circ}$ monthly mean grids. Annual, summer 9 and August spatial distributions for 2003 are displayed on Fig. 22 top, mid and bottom panel, 10 respectively.

11 Those maps illustrate an eminent seasonal variation with a summer maximum in XCH₄ levels 12 observed in both consecutive years (2004 not shown). The northeastern African coast exhibits the highest XCH₄ values, with a hot spot over the Nile's delta in Egypt in summer and August. The 13 lowest XCH₄ levels along the Arabian Peninsula, the Zagros Mountain and eastern Anatolia 14 15 mountain barrier coincide spatially with high altitude areas. To examine to what extent the warm period affects the annual, seasonal, and latitudinal patterns, Georgoulias et al. (2011) further 16 17 proceeded to a monthly analysis. They observed an increase in XCH₄ levels during the summer season, August being the month with the highest levels of $1775-1780 \pm 24$ ppb for both 2003 and 18 19 2004. The highest values are concentrated in the northeastern part of the area primarily in July-August. From July to September, there is a shift of high XCH₄ levels from higher to lower latitudes. 20 Despite the abundance of mud volcanoes over the Greater Area of the Eastern Mediterranean 21 22 region, Georgoulias et al. (2011) ruled out the possibility that the CH₄ total columns from 23 SCIAMACHY (2003-2004) measured over these EM regions were attributed to volcano eruptions.

Ricaud et al. (2014) presented a thorough analysis of atmospheric CH₄ distributions over the MB
in the troposphere, as part of the Chemical and Aerosol Mediterranean Experiment (ChArMEx)
program, using both satellite measurements and model simulations. For this sake, they analyzed
space-borne measurements from (i) the Thermal And Near infrared Sensor for carbon
Observations-Fourier Transform Spectrometer (TANSO-FTS) instrument on the Greenhouse
gases Observing SATellite (GOSAT) satellite, (ii) the Atmospheric InfraRed Spectrometer (AIRS)
on the AURA platform and (iii) the Infrared Atmospheric Sounder Interferometer (IASI)





- 1 instrument aboard the MetOp-A platform. These space-borne tools were used in conjunction with
- 2 the results obtained from three global models: the Chemical Transport Model (CTM) MOCAGE
- 3 (Teyssedre et al., 2007), the Chemical Climate Models (CCMs) CNRM-AOCCM, (Michou et al.,
- 4 2011) and LMDz-OR-INCA (Hourdin et al.,2006).
- 5 The sensitivity of the space-borne sensors is mainly located in the upper tropospheric layers
- 6 peaking around 300 hPa with an envelope as defined by the half-width at half-maximum of the
- 7 averaging kernels (see Figure 23) from 400 to 200 hPa. Consequently, the comparisons between
- 8 measurements and model outputs of CH₄ is mainly concentrated on the layer around 300 hPa for
- 9 AIRS and GOSAT, or considering the total column for IASI.
- In summer, the horizontal distribution of CH₄ in the upper troposphere shows a clear longitudinal gradient between the East and the West of the Mediterranean Basin, both in the space-borne measurements and in the model calculations (Figure 24). There is a maximum of CH₄ in the Eastern MB compared to the Western MB, both considering the upper tropospheric layer and the total column information. The difference between the East and the West of the MB has been calculated within all the datasets and the seasonal variations has been investigated (Figure 25). This clearly shows that the East-West difference peaks in summer, mainly in August.

The long-range transport conditions in the upper troposphere differ over both parts of the Mediterranean Basin. In the western part, whatever the season considered, air masses are basically coming from the west. However, in the EM, apart from the westerlies influence, air masses are also originating from Northern Africa and the Arabic Peninsula (Ziv et al., 2004; Liu et al., 2009), and even farther away, from Asia.

22 To further examine the origin of air masses reaching the Eastern MB, a six-day backtrajectory from the point at 33°N, 35°E located in the EM (red filled circle in Fig. 26) was 23 calculated, considering vertical movement, using the British Atmospheric Data Centre (BADC) 24 trajectory service (http://badc.nerc.ac.uk/community/trajectory/) for July-August over 2001-2010 25 every 12 hours. The position of the gravity center of all trajectories (i.e. the maximum in the 26 probability density function) is displayed every 24 h in Figure 26 at 850 (red stars), 700 (orange), 27 28 500 (green), 300 (blue) and 200 hPa (yellow). For this purpose, data from ECMWF archive (2.5 29 degree/pressure levels) were used in the calculation.





1 Based on these studies focused on the EM, Ricaud et al. (2014) proposed a scheme displaying the 2 transport mechanism (Fig. 27) representing the several stages process: (1) capturing of lower 3 tropospheric pollutants, including CH₄, in the Asian monsoon; (2) pollutants ascent to the upper 4 troposphere by the Asian monsoon; (3) accumulation of pollutants within the Asian monsoon in the upper troposphere; (4) long- range transport and large-scale repartition of pollutants in the 5 upper troposphere from the Asian Monsoon Anticyclone (AMA) to the Middle East and North 6 7 Africa; (5) Subsiding air masses yielding to the build-up of pollutants at mid-tropospheric layers 8 above the EM.

9 4. Conclusions and Perspectives

This review demonstrates the significant progress made in understanding the atmospheric pollution over the MB. Measurements from space-borne and aircraft instruments and outputs from chemistry-climate models and chemistry transport models clearly revealed that the general atmospheric dynamic summer conditions characterizing the eastern MB differ much from the western ones. The impact of the different meteorological regimes together with the seasonal variabilities of the emissions of various atmospheric pollutants result in a longitudinal gradient between the eastern and the western MBs.

Several new campaigns have been recently organized to give more insights in the understanding 17 of the processes occurring in the Eastern and Western parts of this basin. The TRAnsport and Air 18 Quality (TRAQA) campaign (Attié et al., 2014) held in 2012-2014 was dedicated to the 19 export/import of pollutants from the French continent to the Mediterranean Sea by means of 20 21 balloon and airborne measurements. The Gradient in Longitude of Atmospheric constituents above the Mediterranean basin (GLAM) campaign (Ricaud et al., manuscript submitted) held in August 22 23 2014 was dedicated to the study of the gradient of chemical constituents (pollutants and 24 greenhouse gases) from Toulouse (France) to Larnaca (Cyprus) and the impact of the AMA on the 25 EM pollutant levels.

Surface background stations in the Eastern MB (e.g., Crete, Greece and Larnaca, Cyprus) and in
the Western MB (e.g., Menorca, Spain and Lampedusa, Italy) deployed even more instruments to
obtain a wide variety of atmospheric parameters (meteorology, chemistry, dynamics, radiation,
etc.). These campaigns were organized in close relationship with modelling studies (forecasts, and





- 1 re-analyses) and space-borne observations. New airborne campaigns are under analysis, e.g. OMO
- 2 in summer 2015, or in project (RIMES in summer 2018) in order to quantify the export of the
- 3 Asian pollutants to the EM basin chemical constituents loading.
- 4 Concurrently to these intensive experiments, new sites have been instrumented. In early 2015, the
- 5 Agia Marina Xyliatou EMEP rural background air quality station sited at 532 m in altitude in the center of Cyprus (35°02'17"N, 33°03'28"E), and operated since October 1996 (Kleanthous et al., 6 2014), has been augmented with a package of atmospheric chemistry and physics monitoring 7 instruments thanks to the Cyprus Institute and French laboratories, in order to initiate an enhanced 8 9 atmospheric chemistry observation period of several years in the easternmost MB. Unmanned aircraft vehicles are also deployed on a regular basis to document the lower troposphere above the 10 station and the German TROPOS institute has deployed a full set of aerosol-cloud-water vaper 11 12 remote sensing instrument for almost a year in October 2016. This unprecedented experimental effort is expected to bring information on the variability of new compounds and processes with a 13 focus on VOCs and secondary and carbonaceous aerosols and their origins, and on interactions 14 15 between aerosols and the water vapor cycle.

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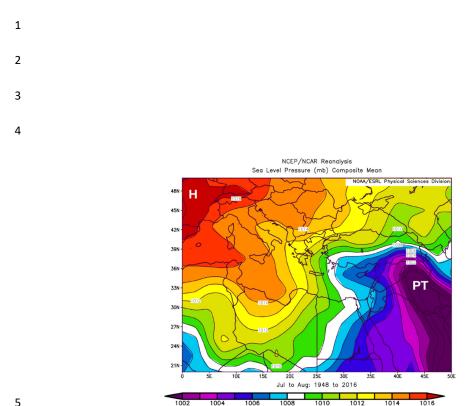




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6 Figure 1. Composite long term mean Sea Level Pressure (hPa) for July-August over 1948-2016.

"PT" indicates the Persian Trough position. "H" indicates the Anticyclone position. Source: NCEP 7

Reanalysis data provided by the NOAA/OAR/ESRL PSD, Boulder, Colorado, USA, 8

9 http://www.esrl.noaa.gov/psd/





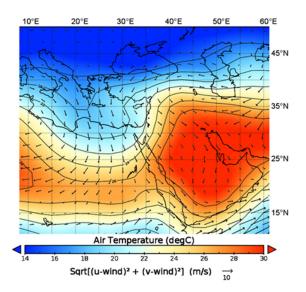


Figure 2. NCEP/NCAR reanalysis composite long-term mean temperature at 850 hPa (~1500 m,
above sea level or a.s.l.) with wind vectors, averaged over 1948-2016 for July-August. Note the
southward penetration of the European cold air over the Mediterranean Basin. This cold air mass
is transported at shallow tropospheric layers towards the EM by the Etesian north-westerlies
characterizing the Persian trough. Source: NCEP Reanalysis data provided by the
NOAA/OAR/ESRL PSD, Boulder, Colorado, USA, http://www.esrl.noaa.gov/psd/.



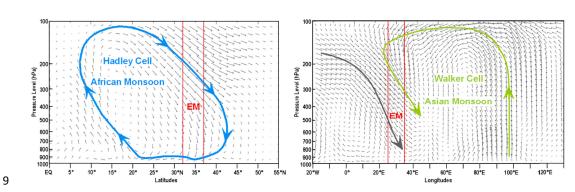


Figure 3. (Left) Closed Hadley cell circulation of the African Monsoon depicted by the vertical
cross section of wind vectors for July-August averaged over the 30-40°E longitudinal band. (Right)
Closed Walker cell circulation of the Asian Monsoon depicted by the vertical cross section of wind





- 1 vectors for July-August averaged over the 20-35°N latitudinal band. The two figures are based on
- 2 the NCEP/NCAR long-term averages (1948-2016) with the position of the Eastern Mediterranean
- 3 (EM) in red. Source: NCEP Reanalysis data provided by the NOAA/OAR/ESRL PSD, Boulder,
- 4 Colorado, USA, http://www.esrl.noaa.gov/psd/.

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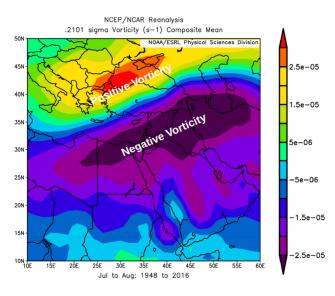
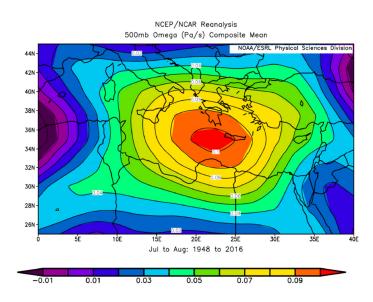


Figure 4. NCEP/NCAR Reanalysis long-term averages (1948-2016) of the relative vorticity at 200
hPa (~12 km a.s.l.) for July-August. The relative vorticity vector is generally perpendicular to the
ground, positive when the vector points upward, negative when it points downward. Note the
negative relative vorticity region located over the southeastern Mediterranean as a result from both
shear and curvature negative relative vorticity. Relative vorticity units are 10⁻⁵ s⁻¹. Source: NCEP
Reanalysis data provided by the NOAA/OAR/ESRL PSD, Boulder, Colorado, USA,
http://www.esrl.noaa.gov/psd/.



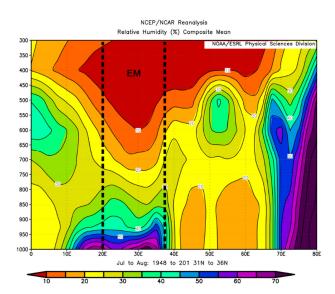




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Figure 5. NCEP/NCAR reanalysis long-term averages of Omega (Pa s⁻¹) at 500 hPa (~5.5 km 2 3 a.s.l.) designating vertical motion for July to August 1948-2016. The maximum subsidence of 0.1 Pa s-1 is equivalent to a downward air motion of ~1.5 cm s⁻¹. Source: NCEP Reanalysis data 4 NOAA/OAR/ESRL 5 provided by the PSD, Boulder, Colorado, USA, http://www.esrl.noaa.gov/psd/. 6

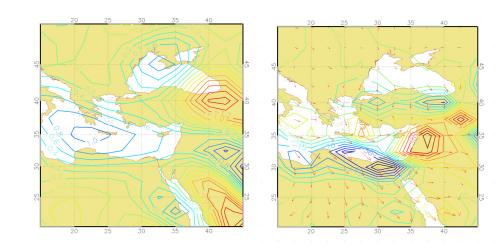
7







- 1 Figure 6. Long- term mean vertical cross section of relative humidity, averaged over the 31-36° N
- 2 latitudinal band for July-August 1948-2016 with the Eastern Mediterranean position (EM), in
- 3 dashed black lines. Source: NCEP Reanalysis data provided by the NOAA/OAR/ESRL PSD,
- 4 Boulder, Colorado, USA, http://www.esrl.noaa.gov/psd/.



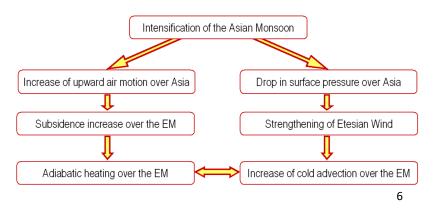
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7 Figure 7. (Left) Blue contours display positive Omega values (cm s-1) representing the vertical descending air motion at a mid-tropospheric level (700 hPa) (~3 km a.s.l.) pointing at a core of 8 1 cm s-1 located over Crete. Red contours are negative Omega values. (Right) Blue contours 9 display cold advection calculated as multiplication of the horizontal thermal gradient by the wind 10 vector. Red contours indicate warm advection, both at 995 hPa level, equivalent to about 140 m 11 a.s.l at 12:00 UTC during PT summer synoptic conditions. Source: NCEP Reanalysis data for 12 2000-2012, provided by the NOAA/OAR/ESRL PSD, Boulder, Colorado, USA, 13 http://www.esrl.noaa.gov/psd/. 14







- 7 Figure 8. Schematic of the proposed mechanism during intensification of the Asian Monsoon
- 8 (adapted from Ziv et al., (2004), permission requested from John Wiley and Sons).
- 9

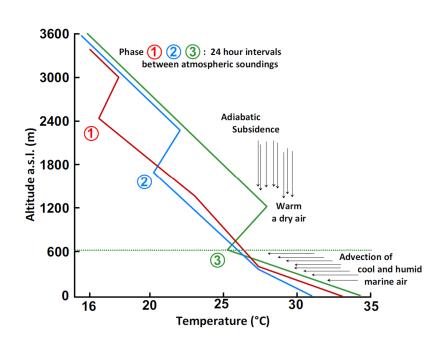


Figure 9. Successive schematic sounding thermal profiles indicating the downward motion of adiabatic subsidence accompanied by a weakening of the Persian Trough which restricts the MLD to shallow layer of the atmosphere (phases 1-3 are 24 h intervals between each sounding at Beit-Dagan, Israel); (from Dayan et al. (1988), ©American Meteorological Society. Used with permission).





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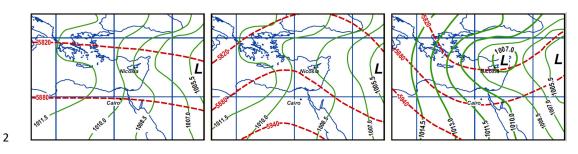


Figure 10. Typical synoptic charts showing the three modes: (a) moderate, (b) shallow, and (c)
deep mode of the Persian Trough as defined by the surface-pressure differences between Nicosia
(Cyprus) and Cairo (Egypt), and their associated upper level conditions. Solid lines are isobars of
sea level pressure with 1.5-hPa intervals. Dashed lines are contours at 500-hPa level with 60-m
intervals (from Dayan et al. (2002), ©American Meteorological Society. Used with permission).



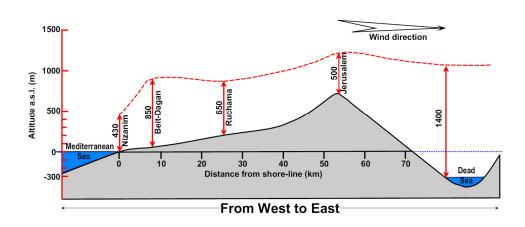
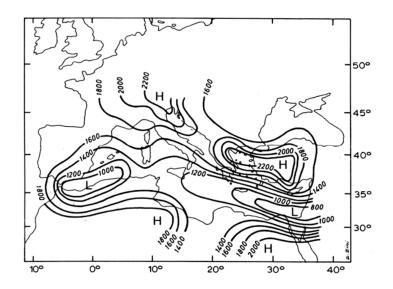


Figure 11. Schematic description of the lateral variation of the mixing layer depth from the
Mediterranean Sea to the Dead Sea (from Dayan et al. (1988), ©American Meteorological Society.
Used with permission).

13





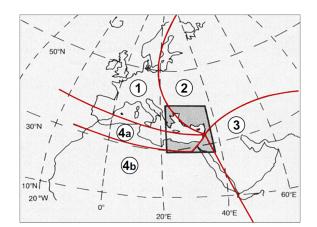


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2 Figure 12. Seasonal map of the mixing layer depth (MLD, in m) for summer (JJA) 1987 over the

- 3 Mediterranean region at 12:00 UTC (from Dayan et al. (1996), permission requested from Kluwer
- 4 Academic Publishers).

5

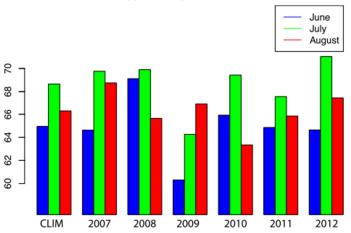


7 Figure 13. Trajectory typing method used to categorize 5-day back-trajectories from the EM region

- 8 at 850 hPa using the Air Resources Laboratory's trajectory model GAMBIT over the 1978-1982
- 9 period (from Dayan (1986), ©American Meteorological Society. Used with permission).







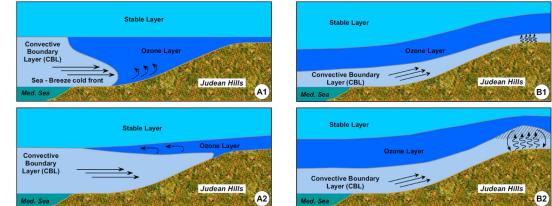
3 km Ozone Concentration (ppb) Monthly Time Series over the Mediterranean

1

2 Figure 14. Monthly means of O₃ concentrations (ppbv) at 3 km partial column measured by IASI

3 in summer (JJA) within the 2007-2012 period over the Mediterranean (IASI morning overpasses).

- 4 Only the observations over the sea are considered in the averages. The monthly means referred to
- 5 as "CLIM" represent the averages over the whole period (reproduced from Doche et al., 2014).

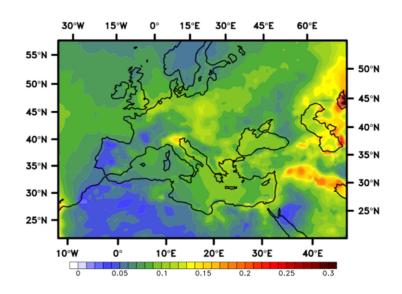


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Figure 15. Scheme of the mechanism causing fumigation of a rich O₃ cloud toward the ground as
moving inland over the EM coast during the weakening of a deep mode of the PT (from Dayan
and Koch (1996), ©American Meteorological Society. Used with permission).

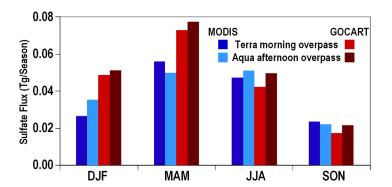






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Fig. 16: Average aerosol optical depth contributed by particulate sulfate validated against
AERONET AOD observations over the period 2003–2009. As mentioned by
http://www.esrl.noaa.gov/gmd/grad/surfrad/aod/, a value of 0.01 corresponds to an extremely
clean atmosphere, and a value of 0.4 to a very hazy condition (The 2003-2010 average AOD over
the Mediterranean Basin is ~0.20 (reproduced from Nabat et al., 2013)

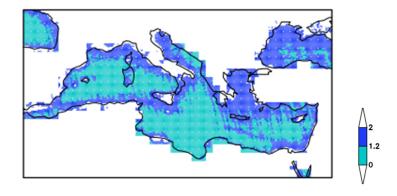


8

9 Figure 17. Comparison of MODIS space-borne and GOCART model derived estimates of the
10 seasonal flux (Tg season⁻¹) of dry sulfate to the EM along the 150-km Israeli coastline. (a)
11 MODIS/Terra overpass and (b) MODIS/Aqua overpass (adapted from Rudich et al., 2008).



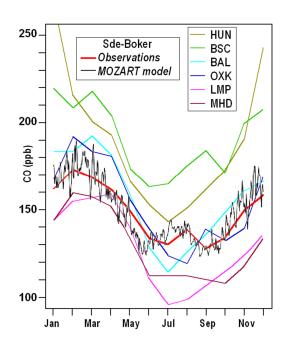




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- 2 Figure 18. Seasonal average over June-August 2006 of NO₂ columns over the Mediterranean Sea
- 3 (1015 molecules cm^{-2}), retrieved from the OMI satellite and considering only maritime pixels
- 4 (reproduced from Marmer et al., 2009).





6

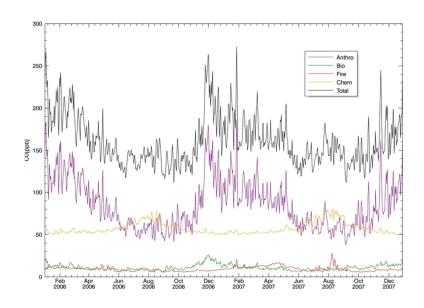
7 Figure 19. CO monthly mean concentrations over 1996-2009 at Sde-Boker (red) and at seven

8 European ESRL/GMD background stations (defined in Table 3, multiple colors), compared to the



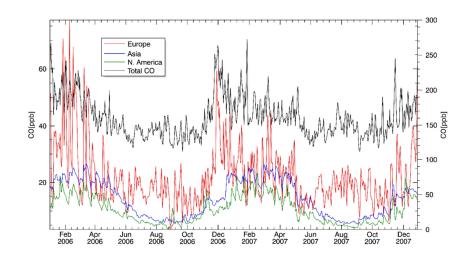


- 1 five-year averaged CO surface concentrations at Sde-Boker (black) over 2003-2007 from the
- 2 MOZART-4 chemistry-transport model (adapted from Drori et al., 2012).
- 3



4

Figure 20. Total CO (black) over Sde-Boker as simulated by MOZART for 2006–2007 and
specific sources contributing to the CO surface: anthropogenic (purple), chemical production
(orange), biogenic (green) and fires (red) (adapted from Drori et al., 2012).







- 1 Figure 21. The European (red), Asian (blue) and North American (green) anthropogenic
- 2 contribution to the total surface CO (black) at Sde-Boker as simulated by MOZART over 2006-
- 3 2007. Distinct continents are scaled on the left vertical axis and total CO on the right vertical axis
- 4 (adapted from Drori et al., 2012).
- 5

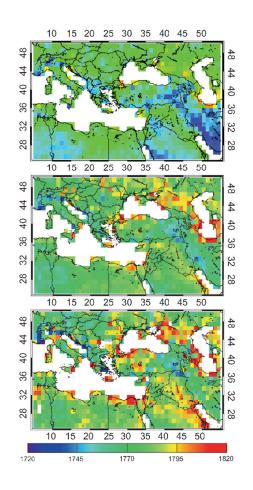


Figure 22. Maps by 1° × 1° resolution of SCIAMACHY WFM-DOAS XCH₄ levels in 2003
including a yearly average (top panel), a summer average (mid-panel) and an August average
(bottom panel), in ppbv (from Georgoulias et al. (2011), permission requested from Taylor and
Francis).





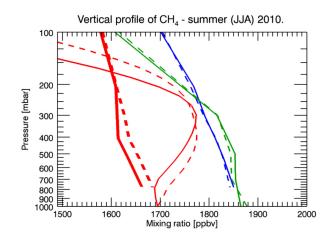
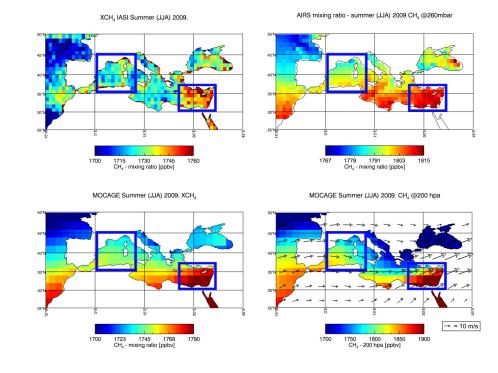




Fig. 23: Summer averaged vertical profiles of CH₄ as measured by AIRS (blue lines) and GOSAT (green lines), and as calculated by MOCAGE (thin red lines) over the Eastern (dashed lines) and Western (solid lines) Mediterranean Basins in summer 2010. Also shown are the seasonally-averaged MOCAGE profiles convolved with the AIRS averaging kernels (thick red lines) for the summer over the Eastern (dashed lines) and Western (solid lines) Mediterranean Basins (adapted from Ricaud et al., 2014).

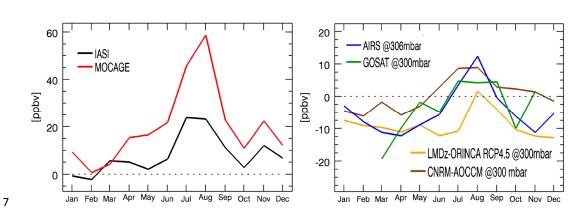






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Figure 24: Fields of CH₄ as calculated by MOCAGE (bottom) and as measured by IASI (top left)
in total column and AIRS (top right) at 260 hPa averaged for summer JJA 2009. Horizontal winds
are from ARPEGE averaged over the same period. The two blue squares represent the West and
East Mediterranean Basins (adapted from Ricaud et al., 2014).







- 1 Figure 25: Seasonal evolution of the difference in CH₄ fields between the Eastern and Western
- 2 MB: (right) around 300 hPa as measured by AIRS (blue) and GOSAT (green) and as calculated
- 3 by LMDz-OR-INCA (yellow) and CNRM-AOCCM (brown), and (left) in total column as
- 4 measured by IASI and calculated by MOCAGE (adapted from Ricaud et al. 2014).
- 5

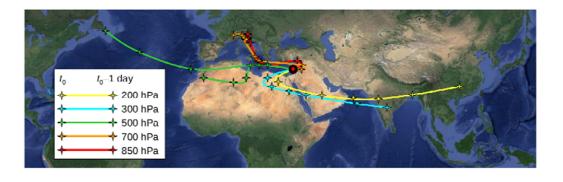
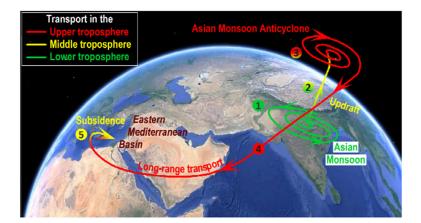




Figure 26: Six-day back-trajectories climatology from the point at 33°N and 35°E located in the
Eastern Mediterranean Basin (red filled circle) derived for July-August over 2001-2010 every 12
hours. The position of the gravity center of each distribution (i.e. the maximum in the probability
density function) at each level is represented every 24 h by a star. (adapted from Ricaud et al.,
2014).







- 1 Figure 27: Schematic representation of the processes impacting the mid-to-upper tropospheric
- 2 pollutants, including CH₄ above the EM in summer (July-August) (adapted from Ricaud et al.,
- з 2014).
- 4
- 5
- 0
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