



1 **High-resolution quantification of atmospheric CO₂ mixing ratios in the Greater Toronto Area,**
2 **Canada**

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13



14 Abstract

15 Many stakeholders are seeking methods to reduce carbon dioxide (CO₂) emissions in urban areas,
16 however reliable, high-resolution inventories are required to guide these efforts. We present the
17 development of a high-resolution CO₂ inventory available for the Greater Toronto Area and
18 surrounding region in southern Ontario, Canada (area of $\sim 2.8 \times 10^5$ km², 26 % of the province of
19 Ontario). The new SOCE (Southern Ontario CO₂ Emissions) inventory is available at the 2.5 x 2.5 km
20 spatial and hourly temporal resolution and characterizes emissions from seven sectors: Area,
21 Residential natural gas combustion, Commercial natural gas combustion, Point, Marine, On-road and
22 Off-road. To assess the accuracy of the SOCE inventory, we developed an observation-model
23 framework using the GEM-MACH chemistry-transport model run on a high-resolution grid with 2.5
24 km grid spacing coupled to the Fossil Fuel Data Assimilation System (FFDAS) v2 inventories for
25 anthropogenic CO₂ emissions and the European Center for Medium-Range Weather Forecasts
26 (ECMWF) land carbon model C-TESSSEL for biogenic fluxes. A run using FFDAS v2 for the southern
27 Ontario region was compared to a run in which its emissions were replaced by the SOCE inventory.
28 Simulated CO₂ mixing ratios were compared against in situ measurements made at four sites in
29 southern Ontario, Downsview, Hanlan's Point, Egbert and Turkey Point, in three winter months,
30 January-March, 2016. Model simulations had better agreement with measurements when using the
31 SOCE inventory emissions versus other inventories, quantified using a variety of statistics such as
32 correlation coefficient, root mean square error and mean bias. Furthermore, when run with the SOCE
33 inventory, the model had improved ability to capture the typical diurnal pattern of CO₂ mixing ratios,
34 particularly at the Downsview, Hanlan's Point and Egbert sites. In addition to improved model-
35 measurement agreement, the SOCE inventory offers a sectoral breakdown of emissions, allowing
36 estimation of average time-of-day and day-of-week contributions of different sectors. Our results
37 show that at night, emissions from Residential and Commercial natural gas combustion and other



38 Area sources can contribute > 80 % of the CO₂ enhancement while during the day emissions from the
39 On-road sector dominate, accounting for >70 % of the enhancement.

40 1.0 Introduction

41 Urban areas are sites of dense population and the intensity of human activities (such as
42 transportation, industry and residential and commercial development) makes them hot-spots for
43 anthropogenic carbon dioxide (CO₂) emissions. While occupying only 3 % of the total land area, urban
44 areas are locations of residence for 54 % of the global population and are the source of 53 – 87 % of
45 anthropogenic CO₂ emissions globally (IPCC-WG3, 2014; WHO, 2015). When considering Canada
46 alone, the urban population accounts for an even larger fraction of the total (81 % in 2011) (Statistics
47 Canada, 2011) while urban areas cover only 0.25 % of the land area (Statistics Canada, 2009).
48 Recognizing their influence on the global carbon budget, many urban areas are seeking methods to
49 reduce their anthropogenic CO₂ emissions. The Greater Toronto Area (GTA) in southeastern Canada,
50 for example, has committed to the *Change is in the Air* initiative as well as being a part of the *C40 Cities*
51 *Climate Leadership Group*, both of which call to reduce CO₂ emissions 30 % below 1990 levels by 2020
52 (C40 Cities, 2016; Framework for Public Review and Engagement, 2007). However, in order to
53 effectively guide anthropogenic CO₂ mitigation strategies, reliable inventories are needed,
54 particularly at high spatial and temporal resolution, to gain a better understanding of the carbon
55 cycle (Gurney et al., 2009; Patarasuk et al., 2016). To our knowledge, the only spatially disaggregated
56 CO₂ inventories available for use in the GTA are the EDGAR v.4.2 (Emission Database for Global
57 Atmospheric Research) CO₂ inventory (available at annual, 0.1° x 0.1° resolution) (EDGAR, 2010) and
58 the FFDAS v2 (Fossil Fuel Data Assimilation System) CO₂ inventory (available at hourly, 0.1° x 0.1°
59 resolution) (FFDAS, 2010), both which are limited in their spatial and/or temporal resolution and
60 therefore are not well-suited for the quantification and understanding of CO₂ emissions at the urban



61 scale. The Canadian national CO₂ inventory, on the other hand, is only available at the provincial level
62 (Environment Canada, 2012).

63 Efforts to develop emission inventories at the fine spatial and temporal resolution required
64 for urban-scale understanding of CO₂ emissions has been driven both by policy- and science-related
65 questions (Gurney et al., 2009; Patarasuk et al., 2016). From a policy perspective, improving CO₂
66 emission quantification is essential to independently evaluate whether anthropogenic mitigation
67 regulations are being effectively implemented. From a scientific perspective, gaining information
68 about anthropogenic CO₂ emissions from urban areas has been primarily motivated by atmospheric
69 CO₂ inversions, which are used to better understand the global carbon cycle (Gurney et al., 2009;
70 Patarasuk et al., 2016). Regardless of the motivation, quantification of CO₂ source/sink processes
71 currently uses two techniques: the bottom-up approach and the top-down approach. In the bottom-
72 up approach, local-scale activity level information is combined with appropriate emission factors to
73 infer emission rates. This method has been used widely to develop many inventories (EDGAR, 2010;
74 FFDAS, 2010; Gurney et al., 2009) but is limited by the accuracy of the input parameters. Conversely,
75 in the top-down approach, inverse modelling is used to exploit the variability in atmospheric mixing
76 ratios of CO₂ to identify the source/sink distributions and magnitudes; this method is limited by
77 insufficient mixing ratio data and uncertainties in simulating atmospheric transport (Pillai et al.,
78 2011). Given current policy needs, a strategy using solely bottom-up or top-down approaches is likely
79 insufficient to evaluate CO₂ emissions but rather a synthesis of the two methodologies is required
80 (Miller and Michalak, 2016). Successful examples of high-resolution CO₂ inventory development are
81 available on the urban scale, such as the Airparif inventory in Ile-de-France (publicly available at
82 <http://www.airparif.asso.fr/en/index/index>) and in Indianapolis, Los Angeles, Salt Lake City and
83 Phoenix through the Hestia project (Gurney et al., 2012), on the national scale, such as in China (Zhao
84 et al., 2012), and on the global scale (Wang et al., 2013). However, to our knowledge, there are



85 currently no studies that have quantified Canadian CO₂ emissions at the fine spatial and temporal
86 resolution required for urban analyses in Canada.

87 In an effort to address this gap, this study was focused on quantifying CO₂ emissions at a fine
88 spatial and temporal resolution in the GTA and southern Ontario (we expanded the inventory beyond
89 the urban area of the GTA so we could exploit information on CO₂ mixing ratios collected at rural
90 areas in central and south-western Ontario, proving additional sites for inventory validation). We
91 present the new high-resolution Southern Ontario CO₂ Emissions (SOCE) inventory, which quantifies
92 CO₂ emissions from seven source sectors (On-road, Off-road, Area, Point, Marine, Residential, and
93 Commercial natural gas combustion) at 2.5 km x 2.5 km spatial and hourly temporal resolution for
94 an area covering ~26 % of the province of Ontario (~2.8 x 10⁵ km²). The SOCE inventory was used in
95 combination with the Environment and Climate Change Canada (ECCC) GEM-MACH chemistry-
96 transport model to simulate CO₂ mixing ratios in a domain including south-eastern Canada and the
97 northeastern USA (hereafter referred to as the “PanAm domain”) for comparison with in situ
98 measurements made by the Southern Ontario Greenhouse Gas Network. Until now, estimates of
99 anthropogenic CO₂ emissions in the GTA were available only from the EDGAR v.4.2 (EDGAR, 2010)
100 and the FFDAS v2 (FFDAS, 2010) inventories, which have very different annual totals for this region
101 (1.36 x 10⁸ vs. 1.05x 10⁸ tonnes CO₂, respectively). Therefore, we expect the results of this work will
102 improve our ability to quantify the emissions of CO₂ in the entire domain as well as the relative
103 contributions of different sectors, providing a more detailed characterization of the carbon budget in
104 the GTA.

105 **2.0 Methods**

106 *2.1 Geographic Domain*

107 The geographic focus of this study was the GTA in southern Ontario, Canada. The GTA is the largest
108 urban area in Canada; it comprises five municipalities, Toronto, Halton, Durham, Peel and York,



109 which together have a population exceeding 6 million (Statistics Canada., 2012b). Although the GTA
110 comprises only 0.07 % of Canadian land area, it represents over 17 % of the total population as a
111 result of rapid urbanization over the past few decades (Statistics Canada., 2012b). Therefore, high-
112 resolution characterization of CO₂ emissions can help integrate climate policy with urban planning.
113 This region is home to a network of measurement sites providing long-term, publicly available
114 datasets of atmospheric CO₂ mixing ratio measurements, *Sect. 2.2* (Environment Canada, 2011) which
115 can be used to evaluate model outputs and inventory estimates. In 2016 the government of Ontario
116 released a Climate Change Action Plan, which includes an endowment given to the Toronto
117 Atmospheric Fund of \$17 million to invest in strategies to reduce greenhouse gas pollution in the
118 GTA (Ontario, 2016). Therefore this research can provide timely information on the carbon budget
119 in the GTA and help to implement effective reduction strategies.

120 *2.2 The Southern Ontario Greenhouse Gas Network*

121 Measurements of ambient CO₂ dry air mixing ratios began in 2005 in southern Ontario at the Egbert
122 station followed by the Downsview station (2007), Turkey Point station (2012) and Hanlan's Point
123 station (2014), Figure 1. Egbert is located ~75 km north-northwest of Toronto in a rural area,
124 Downsview is located ~20 km north of downtown core of the city of Toronto in a populated suburban
125 area, Turkey Point is located to the south-west of the GTA in a rural area on the north shore of Lake
126 Erie, and Hanlan's Point is located on Toronto Island, just south of the city of Toronto on the shore of
127 Lake Ontario. Site details and instrument types used can be found in Table 1. CO₂ measurements are
128 collected as a part of ECCC's Greenhouse Gas Observational Program. The measurement procedure
129 follows a set of established principles and protocols outlined by a number of international agencies
130 through recommendations of the *Meeting on Carbon Dioxide, Other Greenhouse Gases, and Related*
131 *Measurement Techniques*, coordinated by the World Meteorological Organization (WMO) every 2
132 years.



133 The atmospheric CO₂ observational program Egbert is based on non-dispersive infrared
134 (NDIR) methodology and fine-tuned for high precision measurements (Worthy et al., 2005). A
135 detailed description of the NDIR observational system can be found in Worthy et al, (2005). The
136 atmospheric CO₂ observational programs at Turkey Point, Downsview, and Hanlan's Point are based
137 on Cavity Ring-Down Spectrometer (CRDS). Each Picarro CRDS system is calibrated in the ECCC
138 central calibration facility in Toronto before deployment to the field. The response function of the
139 analyzer is determined against 3 calibrated standards tanks (Low, Mid, High). The working (W) and
140 target (T) tanks assigned to the system are also included in the injection sequence and calibrated. At
141 each site, ambient measurements are made using two sample lines placed at the same level. Each
142 sample line has separate dedicated sample pumps and driers (~ -30°C). Pressurized aluminum 30 L
143 gas cylinders are used for the working and target tanks. The sample flow rate of the ambient and
144 standard tank gases is set at ~300 cc/min. The injection sequence consists of a target and working
145 tanks sequentially passed through the analyzer for 10 minutes each every 2 days. The ambient data
146 from line1 is passed through the analyzer for 18 hours followed by Line2 for 6 hours. The
147 Line1/Line2 sequence repeats one time before the target and working tanks are again passed
148 through the system. The working and target tanks are calibrated on site at least once per year against
149 a single transfer standard transported between the sites and the central laboratory facility in
150 Toronto. The CO₂ measurements from both the NDIR and CRDS analytical systems have a precision
151 of around 0.1 ppm based on one-minute averages and are accurate to within 0.2 ppm.

152 *2.3 GEM-MACH chemistry- transport model*

153 In this project, we used the GEM-MACH (Global Environmental Multi-scale–Modelling Air quality and
154 CHemistry) chemistry–transport model (CTM) (Gong et al., 2015; Moran et al., 2013; Pavlovic et al.,
155 2016; Talbot et al., 2008) to link surface emission estimates and atmospheric mixing ratios. GEM-
156 MACH is an on–line CTM embedded within the Canadian weather forecast model GEM (Côté et al.,



157 1998a; Côté et al., 1998b). The configuration of GEM-MACH used in our study has 62 vertical levels
158 from the surface to ~ 1.45 hPa on a terrain-following staggered vertical grid for a log-hydrostatic
159 pressure coordinate. The thickness of the lowest layer was 40 m. The PanAm domain used in our
160 simulations, which includes central and southern Ontario, as well as western Quebec and the
161 northeastern USA, is shown in Figure 1. The PanAm domain has 524×424 grid cells in the horizontal
162 on a rotated latitude-longitude grid with 2.5-km grid spacing and covers an area of approximately
163 $1310 \text{ km} \times 1060 \text{ km}$ (total domain area is $1.39 \times 10^6 \text{ km}^2$). A 24-hour forecasting period was used
164 with a 60-second time step for each integration cycle. Meteorological fields (wind, temperature,
165 humidity, etc.) were re-initialized every 24 hours (i.e., after each model integration cycle); chemical
166 fields were carried forward from the previous integration cycle (i.e., perpetual forecast). Hourly
167 meteorological and chemical boundary conditions were provided by the ECCC operational 10-km
168 GEM-MACH air quality forecast model (Moran et al., 2015).

169 In our study, we simulated two scenarios of CO_2 surface fluxes, indicated by the sum of the following:

170 Scenario 1:

- 171 • Anthropogenic fossil fuel CO_2 emissions within the province of Ontario estimated by the SOCE
172 inventory, available at $2.5 \text{ km} \times 2.5 \text{ km}$ spatial and hourly temporal resolution, as described
173 in *Sect. 2.4*
- 174 • Anthropogenic fossil fuel CO_2 emissions estimated by the FFDAS v2 inventory (FFDAS, 2010)
175 outside of the province of Ontario (province of Quebec and USA), available at $0.1^\circ \times 0.1^\circ$
176 spatial and hourly temporal resolution
- 177 • Biogenic CO_2 fluxes from the C-TESSSEL land surface model, as described in *Sect. 2.5*

178

179



180 Scenario 2:

- 181 • Anthropogenic fossil fuel CO₂ emissions estimated by the FFDAS v2 inventory (FFDAS, 2010)
- 182 for the entire domain, available at 0.1° x 0.1° spatial and hourly temporal resolution
- 183 • Biogenic CO₂ fluxes from the C-TESSEL land surface model, as described in *Sect. 2.5*

184 CO₂ is not a usual chemical species considered by GEM-MACH but a set of special inert tracer fields
185 were added to GEM-MACH for this project to account for CO₂ concentration fields associated with
186 difference source sectors and the lateral boundaries. The CO₂ boundary conditions set at the lateral
187 and top edges of the domain were obtained from the Monitoring Atmospheric Composition and
188 Climate (MACC) global inversion, v.10.2 (<http://www.copernicus-atmosphere.eu/>). Model simulated
189 specific humidity (q , kg/kg) was used to convert estimated CO₂ mixing ratios to dry air mixing ratios.
190 CO₂ dry air mixing ratios are hereafter referred to CO₂ mixing ratios.

191 *2.4 High-Resolution SOCE inventory development*

192 The high-resolution SOCE inventory was constructed primarily from a pre-existing carbon monoxide
193 (CO) inventory developed by the Pollutant Inventories and Reporting Division (PIRD) of ECCC as part
194 of the 2010 Canadian Air Pollutant Emissions Inventory (APEI). The CO inventory is a comprehensive
195 national anthropogenic inventory that includes emissions from area sources, point sources, on-road
196 mobile sources and off-road mobile sources, including aircraft, locomotive and marine emissions for
197 base year 2010 (Moran et al., 2015). This annual inventory at the provincial level compiled by PIRD
198 was transformed into model-ready emissions files using the Sparse Matrix Operator Kernel
199 Emissions (SMOKE, <https://www.cmascenter.org/smoke/>) emissions processing system for spatial
200 allocation (distribution of non-point source emissions to 2.5 km x 2.5 km (roughly 0.02° x 0.02°
201 resolution) using spatial surrogate fields) and temporal allocation (conversion of inventory annual
202 emission rates into hourly values) (Moran et al., 2015). More detailed information about the CO



203 inventory compilation and subsequent processing has been provided elsewhere (Environment
204 Canada, 2013; Moran et al., 2015; PIRD, 2016).

205 The objective of our work was to calculate CO₂ emissions based on this processed, model-
206 ready CO inventory for Ontario grid cells using sector-specific emission ratios estimated by the
207 Canadian National Inventory Report (NIR) (Environment Canada, 2012). Emission sources within
208 each sector of the CO inventory are classified by SCC (Source Classification Code) and were converted
209 to NFR (Nomenclature for Reporting) for accurate cross-reference with the NIR CO₂ and CO
210 estimates. Provincial totals for CO₂ and CO are estimated based on the NFR sources that are included
211 in the sector, producing the following sector-averaged CO₂:CO ratio:

$$212 \quad CO_{2(\text{sector},kt)} = CO_{(\text{sector},kt)} * \frac{CO_{2(\text{Ontario total},kt)}}{CO_{(\text{Ontario total},kt)}} \quad \text{Eq. (1)}$$

213 This sector-averaged CO₂:CO ratio is used to convert the APEI-based CO model-ready gridded
214 emissions fields into CO₂ emissions fields at the same spatial and temporal resolution. A detailed
215 outline of this conversion is presented for each sector in the following subsections. Unless otherwise
216 noted, temporal allocation of emissions in each sector is based on estimates made available by
217 SMOKE.

218 *2.4.1 Area emissions*

219 Area emissions are mostly small stationary sources that represent diffuse emissions that are not
220 inventoried at the facility level. In the APEI CO inventory, the major emission sources in the Area
221 sector include emissions from public electricity and heat production (1A1a), residential and
222 commercial plants (1A4a and 1A4b), stationary agriculture/forestry/fishing (1A4c), iron and steel
223 production (2C1), and pulp and paper (2D1). The NIR estimates an Ontario total from these (and
224 other minor sources) of 23,455 kt CO₂ and 218.8 kt CO, producing a CO₂:CO ratio of 107.2 kt CO₂/kt



225 CO. This ratio was applied to every Area sector grid cell belonging to Ontario in the domain to convert
226 sector CO emissions to CO₂ emissions.

227 *2.4.2 Point emissions*

228 Point emissions are stationary sources in which emissions exit through a stack or identified exhaust.
229 In the APEI CO inventory, the major emission sources in the Point sector include public electricity
230 and heat production (1A1a), stationary combustion in manufacturing industries and construction
231 (1A2f), chemical industry (2B5a), pulp and paper (2D1), iron and steel production (2C1) and other
232 metal production (2C5). Unlike the Area sector, we found that applying a single CO₂:CO ratio to every
233 facility did not produce realistic CO₂ emissions due to the significant variability in combustion
234 efficiency (and thus CO₂:CO ratio). Therefore, we used ECCC Facility Reported Data (Environment
235 Canada, 2015) to identify the geocoded location and annual average CO₂:CO for 48 individual facilities
236 in Ontario (Table S1) and applied the specific CO₂:CO ratios to the grid cells where the facilities were
237 located. In addition, stack height of individual facilities were included in the emission model to
238 optimize plume rise. All other point sources (minor facilities) were scaled by a sector average CO₂:CO
239 ratio of 313.1 kt CO₂/kt CO, calculated from Ontario total CO₂ and CO point-source emissions from
240 the NIR. Temporal allocation of emissions in the Point sector are based on facility level operating
241 schedule data collected by ECCC.

242 *2.4.3 On-road mobile emissions*

243 On-road emissions include the emissions from any on-road vehicles (quantified by the Statistics
244 Canada Canadian Vehicle Survey) (Environment Canada, 2013). In the APEI CO inventory, the major
245 emission sources in the On-road sector includes gasoline and diesel-powered light- and heavy-duty
246 vehicles (1A3b). The NIR estimates an Ontario total from these (and other minor on-road sources) of
247 44,458 kt CO₂ and 1508.3 kt CO, producing a CO₂:CO ratio of 29.5 kt CO₂/kt CO. This ratio was applied
248 to every On-road grid cell belonging to Ontario in the domain to convert sector CO emissions to CO₂.



249 Temporal allocation of emissions in the On-road sector is estimated using data collected in the FEVER
250 (Fast Evolution of Vehicle Emissions from Roadways) campaign in 2010 (Gordon et al., 2012a;
251 Gordon et al., 2012b; Zhang et al., 2012).

252 *2.4.4 Off-road mobile emissions*

253 Off-road emissions include the emissions from any off-road vehicles that do not travel on designated
254 roadways, including aircraft, all off-road engines, and locomotives. In the APEI CO inventory, the
255 major emission sources in the Off-road sector include civil aviation (1A3a), railways (1A3c), and
256 agriculture/forestry/fishing: off-road vehicles and other machinery (1A4c). Similar to the Point
257 sector, we found that applying a single CO₂:CO ratio to every grid cell did not produce realistic CO₂
258 emissions for the two airports in the GTA, Pearson International Airport (PIA) and Billy Bishop
259 Toronto City Airport (BBTCA). Therefore, we used air quality assessment reports compiled for each
260 airport (RWDI AIR Inc., 2009; RWDI AIR Inc., 2013) to identify the geocoded location and facility-
261 specific annual average CO₂:CO ratio. Sources of emissions from each airport include aircraft (landing
262 and take-off cycles), auxiliary power units, ground support equipment, roadways, airside vehicles,
263 parking lots, stationary sources and training fires; note that emissions from aircrafts in-transit
264 between airports, which are injected in the free troposphere, have not been included in this
265 inventory (Moran et al., 2015; RWDI AIR Inc., 2009). Based on these two reports, we applied a ratio
266 of 175 kt CO₂/kt CO to the grid cell containing PIA and a ratio of 20 kt CO₂/kt CO to the grid cell
267 containing BBTCA. All other off-road sources belonging to Ontario grid cells were scaled by a sector
268 average CO₂:CO ratio of 7.2 kt CO₂/kt CO, calculated from NIR-reported Ontario total CO₂ and CO
269 emissions.

270 *2.4.5 Marine emissions*

271 Commercial marine emissions include the emissions from any marine vessels travelling on the Great
272 Lakes (quantified by the Statistics Canada *Shipping in Canada*) (Environment Canada, 2013). In the



273 APEI CO inventory, the major emission source in the Marine sector is national navigation (1A3d). The
274 NIR estimates an Ontario total from this source of 729.2 CO₂ and 0.86 kt CO, producing a CO₂:CO ratio
275 of 844.2 kt CO₂/kt CO. This ratio was applied to every marine grid cell in the domain to convert sector
276 CO emissions to CO₂.

277 *2.4.6 Residential and commercial emissions*

278 Residential and commercial CO₂ emissions reflect on-site combustion of natural gas for electricity
279 and heating, a source that we found was not included in the APEI CO inventory because of the high
280 efficiency of the furnaces and resulting low CO emissions. To include the CO₂ emissions from these
281 on-site furnaces, we used the Statistics Canada 2012 Report on Energy Supply and Demand to
282 quantify the amount of natural gas consumed by residential and commercial buildings in Ontario,
283 7969.6 gigalitres (GJ) and 4895.7 GJ respectively (Statistics Canada, 2012a). We used an emission
284 factor of 1879 g CO₂/m³ natural gas combustion (Environment Canada, 2012) to estimate CO₂
285 emissions from residential and commercial on-site furnaces in Ontario to be 1.5 x 10⁷ tonnes and 9.2
286 x 10⁶ tonnes, respectively. These two emission totals were spatially allocated using a “capped-total
287 dwelling” spatial surrogate developed by ECCC and temporally allocated using the SMOKE emissions
288 processing system (Moran et al., 2015).

289 *2.5 Biogenic flux*

290 The net ecosystem exchange (NEE) fluxes used in our simulations were provided by the land surface
291 component of the ECMWF forecasting system, C-TESSSEL (Bousetta et al., 2013). Fluxes are extracted
292 at the highest available resolutions, 15 x 15 km and 3 hour for January and February 2016 and 9 x 9
293 km and 3 hour for March. These data are interpolated in space and time to be consistent with our
294 model resolution. With our main priority being understanding anthropogenic emissions in the GTA,
295 we chose to analyze a period where the biogenic CO₂ flux is minimized and therefore this paper
296 focuses on three winter months in 2016, January to March inclusive.



297 3.0 Results and Discussion

298 3.1 The SOCE inventory

299 Figure 1 displays the PanAm domain total anthropogenic CO₂ emissions estimated by the SOCE
300 inventory for the province of Ontario portion ($\sim 0.02^\circ \times 0.02^\circ$) and by the FFDAS v2 inventory ($0.1^\circ \times$
301 0.1°) (FFDAS, 2010) for the remainder of the domain. Regions of high emissions typically correspond
302 to population centers, for example the GTA in Ontario, Montreal and Quebec City in Quebec, and
303 Chicago, Boston and New York City (amongst others) in the USA. Emissions from highways and major
304 roadways are only clear in the province of Ontario (at higher spatial resolution) but industrial and
305 large scale area sources are evident across the entire domain.

306 The total CO₂ emissions can be separated into contributions from the seven sectors in the
307 province of Ontario described in *Sect. 2.4*. Figure 2 shows the anthropogenic CO₂ contributions from
308 the Area sector, Residential and Commercial sector, Point sector, Marine sector, On-road sector and
309 Off-road sector, focused on southern Ontario and the GTA. If we consider emissions from a domain
310 including the area solely around the GTA (indicated by the black-box in Figure 2a), the total CO₂
311 emissions estimated by the SOCE inventory is 94.8 Mt CO₂ per year, Table 2. Figures 2a and b display
312 the CO₂ emissions from the Area sector and from Residential and Commercial natural gas combustion
313 in southern Ontario. These two sectors combined represent the largest source of CO₂ in the black-
314 box area (41.6 Mt CO₂/year, contributing 43.9 % of the total). The majority of these emissions are
315 concentrated in the GTA and surrounding urban areas as a result of a significant portion of the
316 population (64 %) being reliant on natural gas for heat production (Statistics Canada, 2007; Statistics
317 Canada, 2012a). Figure 2c represents emissions from the Point sector, contributing 24.4 Mt CO₂/year,
318 25.7 % of the total. The largest point source emitters are located on the western shore of Lake Ontario
319 (Hamilton and surrounding areas) as this area is one of the most industrialized regions of the country
320 with intensive metal production activities. Figures 2d, e and f display CO₂ emissions from various



321 transportation sectors, Marine, On-road, and Off-road respectively, which together contribute more
322 than 30 % of total CO₂ emissions in the area within the black box. While emissions from marine
323 activity are minimal, those from On-road and Off-road sources are significant (25.0 % and 5.3 %,
324 respectively), concentrating on the major highways connecting the various population centres of the
325 GTA to the downtown core, as well as at PIA located within the city.

326 *3.2 Comparison of the SOCE inventory with other inventories*

327 The EDGAR v4.2 inventory estimates CO₂ emissions on an annual basis and by sector based on
328 Selected Nomenclature for Air Pollution (SNAP) sub-sectors while FFDAS v2 provides hourly mean
329 grid cell totals. Table 2 shows a comparison between the sectoral CO₂ estimates of the SOCE and
330 EDGAR v4.2 inventories (SNAP sectors were grouped to correspond to SOCE sectors, Table S2) as
331 well as the domain total estimated by the FFDAS v2 inventory for the area surrounding the GTA (the
332 black-box area outlined in Figure 2a). There is a significant discrepancy between the CO₂ emissions
333 estimated by the SOCE and EDGAR v4.2, inventories both in the relative sectoral contributions as well
334 as domain total (percent difference >35 %). The largest sectoral discrepancies are in the Point and
335 the On-road sectors, where the EDGAR v4.2 inventory estimates a contribution 1.9 and 1.7 times
336 larger than that of the SOCE inventory, respectively. Figure 3 shows a comparison of the spatial
337 distribution of the CO₂ inventory predicted by a) FFDAS v2, b) EDGAR v4.2, and c) SOCE (aggregated
338 to 0.1° x 0.1° to match the resolution of EDGAR v4.2 and FFDAS v2) for the GTA area. Figure 3 reveals
339 that the largest differences between the SOCE inventory and the EDGAR v4.2 inventory is the CO₂
340 emissions in the GTA; EDGAR v4.2 predicts much higher emissions in the GTA (in some grid cells,
341 differences are an order of magnitude), particularly in the downtown core relative to the SOCE
342 inventory.

343 Although there is no sectoral breakdown in the FFDAS v2 inventory, the domain total around
344 the GTA can be compared to that of the SOCE inventory, Table 2. Unlike the comparison with the



345 EDGAR v4.2 inventory, there is a closer agreement between the FFDAS v2 inventory and the SOCE
346 inventory (difference of $\sim 10\%$). The comparison plots in Figure 3 show a good agreement of the
347 spatial variability of emissions in the GTA between the FFDAS v2 and SOCE inventories; however, the
348 gradient between urban and rural areas is not as sharp in the SOCE inventory as it is in the FFDAS v2
349 inventory. Furthermore, emissions along the western shore of Lake Ontario (Hamilton and the
350 surrounding areas) are predicted to be larger in the SOCE inventory relative to FFDAS v2. The FFDAS
351 v2 inventory was interpolated to $0.02^\circ \times 0.02^\circ$ using a mass conservative interpolation scheme to
352 allow the production of a difference plot of the two inventories, SOCE minus FFDAS v2, shown in
353 Figure S1. The difference plot reveals the largest divergence between the inventories occurs in the
354 GTA and Ottawa, with the FFDAS v2 inventory estimating $>1000 \text{ g CO}_2/\text{second}$ ($\sim 30 \text{ kt CO}_2/\text{year}$)
355 more than the SOCE inventory in some grid cells. In addition to similar spatial variability, the FFDAS
356 v2 and SOCE inventories also have similar temporal variability. Figure S2 shows the diurnal profile
357 of estimated emissions from January-March for both the FFDAS v2 and SOCE inventories for the
358 black-box area in the PanAm domain. Both inventories allocate the highest emissions between 08:00
359 and 18:00 and the lowest emission between 00:00 and 5:00, however the amplitude of the diel cycle
360 is higher in SOCE, and emissions in the morning are as high as in the afternoon. FFDAS allocates a
361 relatively larger proportion of the emissions to the 15:00 – 19:00 period.

362 *3.3 Preliminary analyses using the SOCE, FFDAS v2 and EDGAR v4.2 inventories with FLEXPART*

363 To investigate the impact of the differing inventories on ambient mixing ratios, preliminary analyses
364 were run with footprints generated by the FLEXPART driven by GEM meteorology and products were
365 compared against the measured CO_2 gradient between the Downsview and TAO (43.7°N , 79.4°W , a
366 temporary site decommissioned in January 2016) stations in the year 2014. Observed gradients
367 ranged from $+20$ to -10 ppm. Figure S3 displays the measured and modelled CO_2 gradients. These
368 results show that when the EDGAR v4.2 inventory was used, simulated CO_2 gradients were



369 consistently overestimated by ~10-60 ppm relative to observations. Conversely, when the SOCE
370 inventory was used, a higher level of agreement was obtained between simulated mixing ratios and
371 measurements; however, none of the model simulations were able to capture times when the
372 gradient was negative ($CO_{2,TAO} > CO_{2,Downsview}$), an effect we believe to be due to the TAO inlet being
373 ~60 m above ground level and surrounded by many high-rise buildings creating canyon flows and
374 turbulence which are not properly accounted for in GEM at this resolution. These factors contributed
375 to the decommissioning of TAO in January 2016. The poor performance of our model system when
376 using the EDGAR v4.2 inventory to simulate CO_2 mixing ratios was also found by a study quantifying
377 on-road CO_2 emissions in Massachusetts, USA (Gately et al., 2013). In this study, EDGAR emission
378 estimates were found to be significantly larger than any other inventory by as much as 9.3 million
379 tons, or >33 %. The difference in estimates between the EDGAR v4.2 and the SOCE inventories is
380 likely explained by their underlying differences in methodology. Being a global product and not
381 specifically designed for mesoscale applications, the EDGAR v4.2 inventory estimates CO_2 emissions
382 based on country-specific activity data and emission factors, however the spatial proxies used to
383 disaggregate the data are not always optimal. A study performed by McDonald et al. (2014) showed
384 that the use of road density as a spatial proxy for vehicle emissions in EDGAR v4.2 causes an
385 overestimation of emissions in population centers (McDonald et al., 2014). Given the much larger
386 emission estimates for On-road CO_2 from EDGAR v4.2 (Table 2), this also seems to be an issue in the
387 GTA. Based on this large discrepancy, the EDGAR v4.2 inventory was not further used in this study
388 and we focussed on the inventories developed for regional scale studies.

389 When similar preliminary analyses were run with FLEXPART footprints using the FFDAS v2
390 inventory, Figure S3, good agreement was observed with CO_2 gradients measured between the
391 Downsview and TAO stations. We are confident that the enhanced measurement agreement between
392 the FFDAS v2 and SOCE relative to EDGAR v4.2 is due to improved methodology; spatial allocation of
393 emissions in FFDAS v2 is achieved through the use of satellite observations of nightlights from human



394 settlements from the U.S. Defense Meteorological Satellite Program Operational Linescan System
395 (DMSP-OLS).

396 Beyond the differences in methodology for estimating and allocating emissions, it is
397 important to note that the emissions reported in Table 2 by the FFDAS v2, SOCE and EDGAR v4.2
398 inventories also fundamentally differ in time period quantified. The emissions reported for both
399 FFDAS v2 and the SOCE are based on emissions from three winter months (January-March 2010)
400 extrapolated for the entire year. However, emissions from EDGAR v4.2 are annual averages of all
401 twelve months of 2010. Since CO₂ emissions in the GTA are higher in the winter months relative to
402 the summer months because of increased building and home heating, it is likely that the average
403 annual estimates of SOCE and FFDAS v2 are slightly overestimated. This does not affect the relative
404 agreement between SOCE and FFDAS v2 however it does further increase the divergence between
405 the EDGAR v4.2 and SOCE and FFDAS v2 inventories. Following this and the improved agreement
406 with observations, the FFDAS v2 inventory was used with the SOCE inventory for all subsequent
407 modelling analyses.

408 *3.4 Simulation of CO₂ mixing ratios in the Greater Toronto Area*

409 We used the GEM-MACH CTM and the SOCE and FFDAS v2 inventories to simulate hourly CO₂ mixing
410 ratios in the PanAm domain. The model framework was evaluated for a continuous three-month
411 period, January-March 2016 using four sampling locations in the GTA, Figure 1 (note that
412 measurements for the Hanlan's Point site were not available until January 14, 2016). Figure 4
413 displays afternoon (12:00-16:00 EST) measured and simulated CO₂ mixing ratios produced with the
414 SOCE and FFDAS v2 inventories for the two emissions scenarios described in *Sect. 2.3* for the month
415 of February (Figures S4 and S5 show the same figure for other months). We chose to present only
416 afternoon data as this is the time of day when the mixed layer is likely to be the most well-developed;
417 nighttime and morning data showed largest variations in observations as a result of the shallow



418 boundary layer causing surface emissions to accumulate within the lowest atmospheric layers
419 (Breon et al., 2015; Chan et al., 2008; Gerbig et al., 2008). During the night, atmospheric mixing ratios
420 are most sensitive to vertical mixing, an atmospheric process that is difficult to model for stable
421 boundary layers.

422 The time series comparisons at all four sites demonstrate the model's general ability to
423 capture variability in observations of CO₂, albeit with better skill for the Downsview and Egbert sites
424 (this is particularly clear when we look at model-measurement difference plots, Figure S6). The
425 model is able to capture many extreme events of mixing ratio increases and decreases, such as
426 February 11-14, 2016 at the Downsview site; however, some short time periods are poorly simulated,
427 such as January 21-23, 2016 at Hanlan's Point, when the model significantly overestimated measured
428 CO₂. Generally, mixing ratios simulated by the FFDAS v2 inventory are similar or larger than those
429 produced when the SOCE inventory is used, with differences most noticeable at the Downsview and
430 Hanlan's Point sites. This was expected as the difference plot shown in Figure S1 reveals that the
431 SOCE and FFDAS v2 inventories diverge the most in the GTA (where the Downsview and Hanlan's
432 Point sites are located) and are more similar in rural areas (where the Turkey Point and Egbert sites
433 are located).

434 Measured CO₂ mixing ratios have a typical diurnal pattern, in which mixing ratios are higher
435 at night and lower during the day, despite higher emissions during the day. This results from the daily
436 cycle of the mixed layer, which is shallow at night due to thermal stratification and deepens during
437 the day due to solar heating of the surface. Figure 5 displays the measured and modelled mean
438 diurnal profile of CO₂ at the four sites in our network using data from January-March, 2016 (note
439 difference in y-axis scale for urban vs. rural sites). At all four sites, the shapes of the modelled and
440 measured mixing ratios throughout the day agree very well, suggesting that the GEM meteorology in
441 our framework is capturing the diurnal variation in emissions and the boundary layer evolution. At



442 the Downsview site, there is a very strong agreement between the modelled and measured diurnal
443 profiles when using the SOCE inventory, whereas the FFDAS v2 simulated profile largely
444 overestimates mixing ratios, particularly at nighttime. This is consistent with the FFDAS inventory
445 having larger emissions than the SOCE inventory during the night (Fig S2). At the Hanlan's Point site,
446 a difference of ~ 5 ppm CO₂ is observed when using the SOCE inventory relative to measurements;
447 however, similar to the Downsview site, the FFDAS v2 simulated profile has a larger difference of
448 ~ 10 ppm CO₂. At both the Egbert and Turkey Point sites, the use of both inventories similarly
449 overestimates the diurnal pattern of CO₂ mixing ratios by ~ 3 -5 ppm, again likely a result of the
450 similarities of these two inventories at these sites, Figure S1. At all four sites, it is possible that some
451 of the biases that are observed in simulated CO₂ mixing ratios may arise from inaccuracies in the
452 boundary CO₂ provided by MACC; this aspect was not, however, further explored in this study.

453 *3.5 Quantifying model-measurement agreement*

454 Figure 6 shows scatter plots of afternoon (12:00-16:00 EST) modelled versus measured CO₂ mixing
455 ratios from January- March, 2016 at the four sites used in this study. The top row shows the
456 correlation between measured and modelled mixing ratios using the SOCE inventory and the bottom
457 row shows the correlation using the FFDAS v2 inventory. It is immediately clear that there is a
458 stronger model-measurement correlation at the Downsview and Egbert sites ($R > 0.75$) relative to
459 that of Hanlan's Point or Turkey Point ($R < 0.53$). The difficulty with accurately simulating CO₂ mixing
460 ratios at Hanlan's Point and Turkey Point may arise from their proximity to shorelines, Hanlan's Point
461 to Lake Ontario and Turkey Point to Lake Erie (see Figure 1). Differential heating of land versus water
462 near these lakes creates pressure gradients driving unique circulation patterns (Burrows, 1991; Sills
463 et al., 2011). These circulation patterns are difficult for models to capture and therefore may
464 contribute to the relatively poor correlation observed at Hanlan's Point and Turkey Point.



465 It is also clear from Figure 6 that simulating CO₂ mixing ratios at the Egbert and Turkey Point
466 sites using either the FFDAS v2 or the SOCE inventory results in similar performance, likely because
467 the emissions estimated by the two inventories are similar in the vicinity of these two rural sites (see
468 also Figure 5). However at both the Downsview and Hanlan's Point sites, using the SOCE inventory
469 provided a slightly higher correlation and reduced RMSE and MB relative to using the FFDAS v2
470 inventory. The improvement by using the SOCE inventory is likely a result of both the improved
471 spatial resolution (2.5 km versus 10 km), and therefore more accurate allocation of emissions to grid
472 cells, and also a better estimation of emission magnitudes, as large differences are shown in Figures
473 3 and S1.

474 *3.6 Sectoral contributions to simulated CO₂ mixing ratios*

475 One of the major advantages of the SOCE inventory over the FFDAS v2 inventory is the availability of
476 sectoral emission estimates. Figure 7 displays the sectoral percent contributions to diurnal CO₂
477 mixing ratio enhancements (calculated as local CO₂ mixing ratios above the MACC estimated
478 background) for the Downsview station in February 2016 averaged by the day of week (Figures S7
479 and S8 displays the same for other months). This figure clearly demonstrates the importance of Area
480 emissions (defined here as the sum of the Area + Residential natural gas combustion + Commercial
481 natural gas combustion) to simulated CO₂ mixing ratios, reaching ~80 % contribution in the early
482 morning and late evening, consistent with times when emissions from home heating are the
483 dominant source of CO₂. Contributions from Area emissions decrease to ~35 % midday, which
484 coincides with when emissions from other sources, such as On-road, gain importance. In the midday,
485 emissions from the On-road sector can contribute ~50 %, which is consistent with transportation
486 patterns of the times when the population is travelling to and from work and other activities. The
487 relative contributions to CO₂ mixing ratios from point source emissions is quite variable during the
488 course of a day and week, but generally seems to increase in the early morning and evening and can



489 contribute a significant portion of total CO₂ emissions (up to ~20 %). Figure 7 indicates that biogenic
490 sources of CO₂ play a negligible role during January-March in the GTA. Recent studies, however, have
491 shown the importance of the biospheric contribution (up to ~132-308 g CO₂ km⁻² s⁻¹) to measured
492 CO₂ in urban environments during the growing season (Decina et al., 2016). Therefore, this finding
493 supports the importance of modelling CO₂ in the wintertime in cities like the GTA to avoid
494 complications associated with biospheric contributions. The new ability to understand the sectoral
495 contributions to CO₂ mixing ratios in the GTA and southern Ontario has implications from a policy
496 perspective; with recent initiatives to curb CO₂ emissions, understanding from which sector the CO₂
497 is being emitted could be useful to assess how effective applied mitigation efforts have been or where
498 to target future efforts. These efforts could be complemented by running simulations with additional
499 tracers, such as carbon monoxide (CO), nitrogen oxides (NO_x), or stable carbon isotopes (¹²C and ¹³C)
500 to gain further insight.

501 *4.0 Conclusions*

502 We presented the SOCE inventory for southern Ontario and the GTA, the first, to our knowledge, high-
503 resolution CO₂ inventory for southern Ontario and for a Canadian metropolitan region. The SOCE
504 inventory provides CO₂ emissions estimates at 2.5 km x 2.5 km spatial and hourly temporal
505 resolution for seven sectors: Area, Residential natural gas combustion, Commercial natural gas
506 combustion, Point, Marine, On-road and Off-road. When compared against two existing CO₂
507 inventories available for southern Ontario, the EDGAR v4.2 and the FFDAS v2 inventories, using
508 FLEXPART footprints, the SOCE inventory had improved model-measurement agreement; FFDAS v2
509 agreed well with in situ measurements, but the EDGAR v4.2 inventory systematically overestimated
510 mixing ratios. We developed a model framework using the GEM-MACH chemistry-transport model
511 on a high-resolution 2.5 km x 2.5 km grid coupled to the SOCE and FFDAS v2 inventories for
512 anthropogenic CO₂ emissions and C-TESSSEL for biogenic CO₂ fluxes. We compared output simulations



513 to observations made at four stations across southern Ontario and for three winter months, January
514 – March, 2016. Model-measurement agreement was strong in the afternoon using both
515 anthropogenic inventories, particularly at the Downsview and Egbert sites. Difficulty in capturing
516 mixing ratios at the Hanlan’s Point and Turkey Point sites was hypothesized to be a result of their
517 close proximity to shorelines (Lake Ontario and Lake Erie, respectively) and the model’s inability to
518 capture the unique circulation patterns that occur in those environments. Generally, across all
519 stations and months, simulations using the SOCE inventory resulted in higher model-measurement
520 agreement than those using the FFDAS v2 inventory, quantified using R, RMSE and mean bias. In
521 addition to improved agreement, the primary advantage of the SOCE inventory over the FFDAS v2
522 inventory is the sectoral breakdown of emissions; using average day of week diurnal mixing ratio
523 enhancements, we were able to demonstrate that emissions from area sources can contribute >80 %
524 of CO₂ mixing ratio enhancements in the early morning and evening with on-road sources
525 contributing >50 % midday. The applications of the SOCE inventory will likely show future utility in
526 understanding the impacts of CO₂ reduction efforts in southern Ontario and identify target areas
527 requiring further improvement.

528 *Author Contributions*

529 The SOCE inventory was prepared by Stephanie C. Pugliese, with critical input from Felix Vogel and
530 Jennifer Murphy. The CO inventory which the SOCE inventory is based upon was provided by Mike
531 Moran, Junhua Zhang and Qiong Zheng. The GEM-MACH modelling analyses were performed by
532 Shuzhan Ren and Craig Stroud. The ambient CO₂ data were collected by Douglas Worthy and his team
533 at Environment and Climate Change Canada. The MACC and C-TESSSEL products used in our model
534 simulations were provided by Gregoire Broquet. The data was analyzed and interpreted for
535 publication by Stephanie C. Pugliese. This manuscript was written by Stephanie C. Pugliese, with
536 critical input from Jennifer Murphy, Felix Vogel and Mike Moran.



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681 **Table 1: Summary of atmospheric measurement programs in Southern Canada operated by**
682 **Environment and Climate Change Canada**

Start Date	Site Name	Coordinates	Elevation (asl)	Intake Height	In-situ Instrumentation
March, 2005	Egbert	44.231037N, 79.783834W	251m	3m, 25m*	NDIR
November, 2010	Downsview	43.780491N, 79.468010W	198m	20m	NDIR
November, 2012	Turkey Point	42.635368N, 80.557659W	231m	35m	CRDS
June, 2014	Hanlan's Point	43.612201N 79.388705W	87m	10m	CRDS

683 * At Egbert, a 25 m tower was installed in March 9, 2009

684 NDIR = Non-dispersive infrared

685 CRDS = cavity ring-down spectroscopy

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687



688 **Table 2: Anthropogenic CO₂ emissions for the year 2010 in the black-box area (shown in**
 689 **Figure 2a) by sector. Values in parentheses indicate the percentage contribution of the**
 690 **sector to the total CO₂ emissions in the black-box area.**

Sector	FFDAS v2 [‡] CO ₂ Inventory (Mt CO ₂ /year)	EDGAR v4.2 [#] CO ₂ Inventory (Mt CO ₂ /year)	SOCE CO ₂ Inventory (Mt CO ₂ /year)
Area*	-	46.2 (33.9 %)	41.6 (43.9 %)
Point	-	45.9 (33.7 %)	24.4 (25.7 %)
Marine	-	0.12 (0.10 %)	0.10 (0.10 %)
On-road	-	41.2 (30.2 %)	23.7 (25.0 %)
Off-road	-	2.95 (2.2 %)	5.01 (5.3 %)
Total	104.8	136.4	94.8

691 *Area sector represents the summation of Area + Residential + Commercial natural gas combustion.

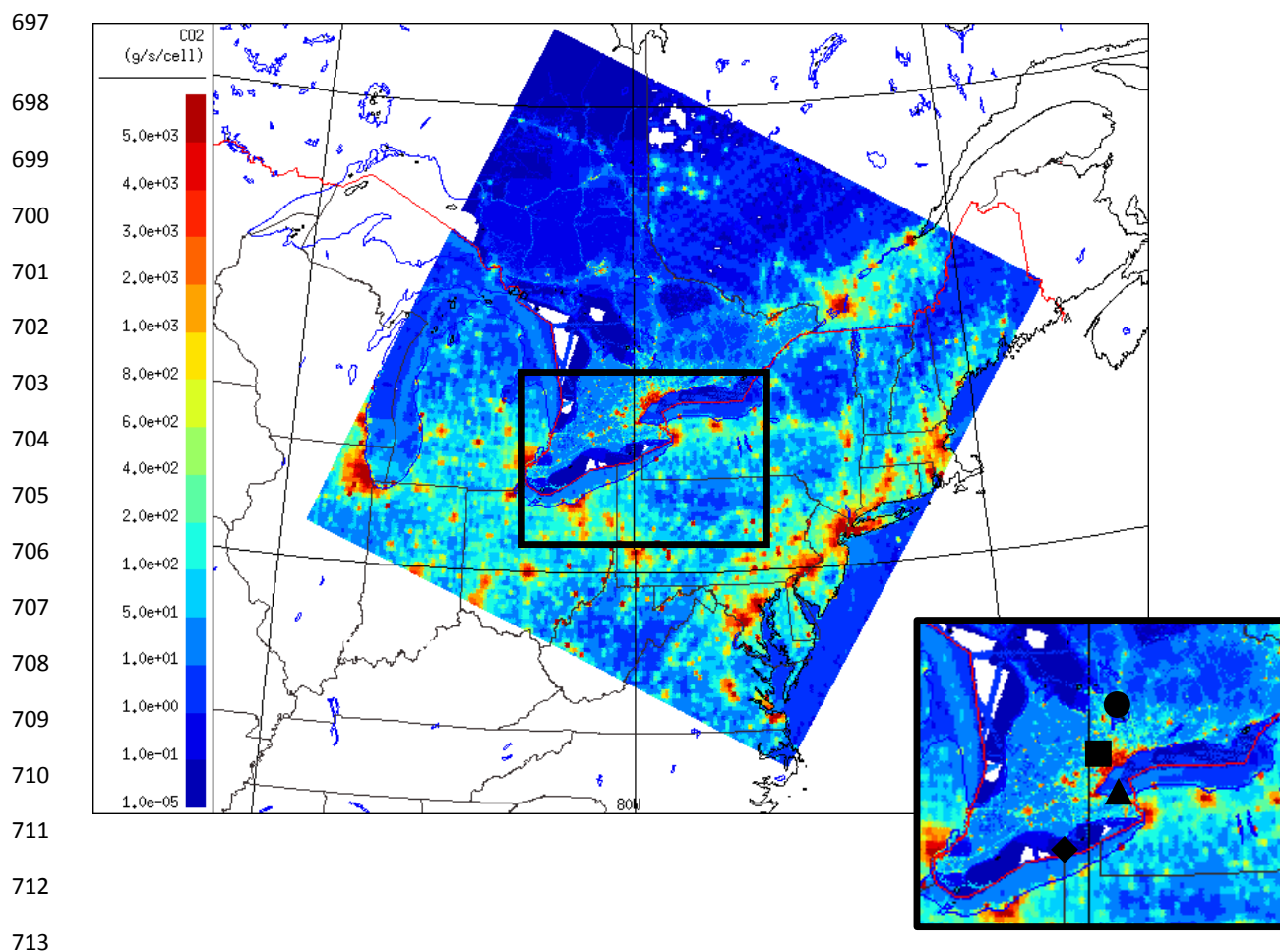
692 [#]The EDGAR inventory v4.2 can be found at <http://edgar.jrc.ec.europa.eu>.

693 [‡]The FFDAS v2 inventory can be found at <http://hpcg.purdue.edu/FFDAS/>.

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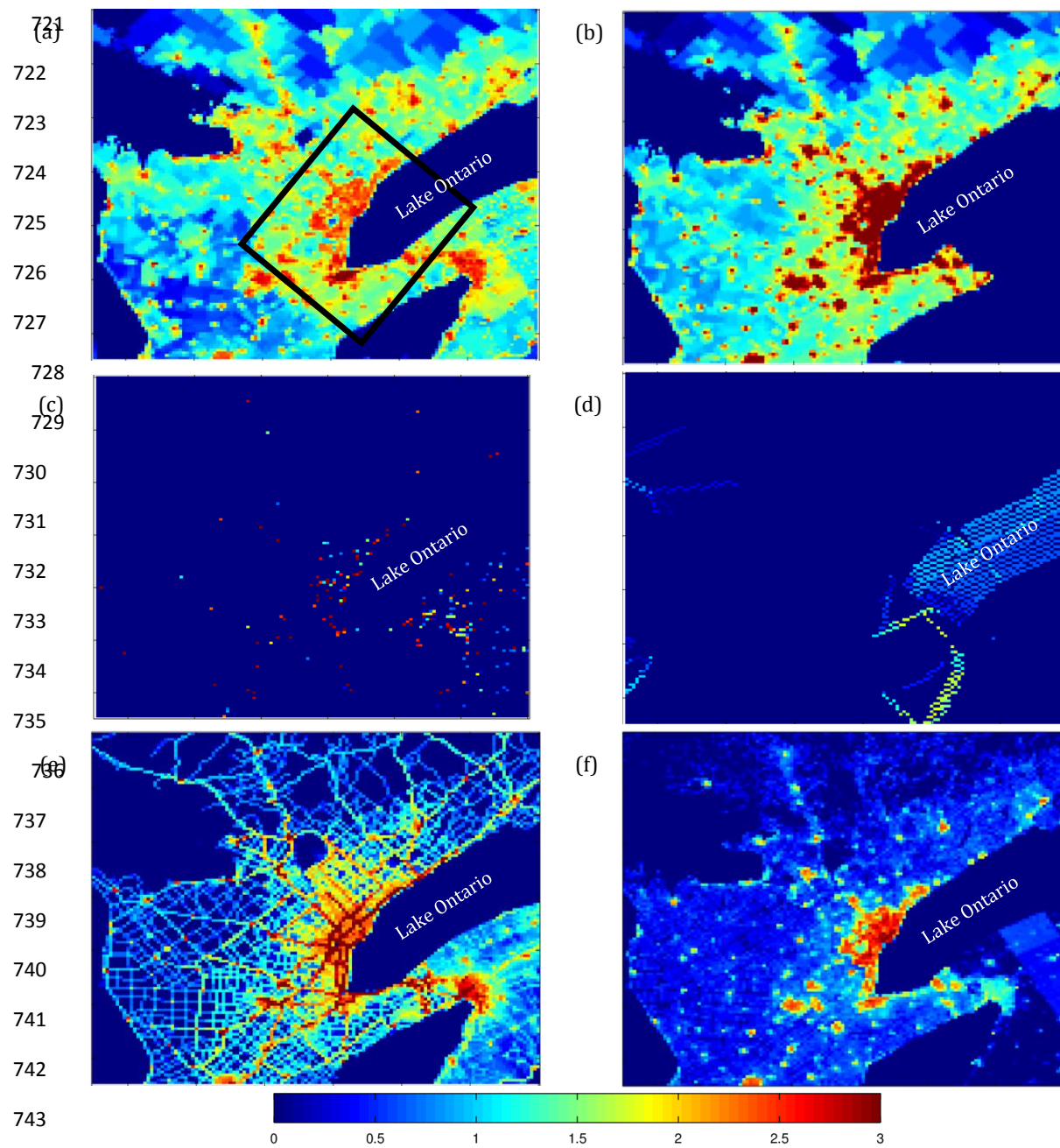
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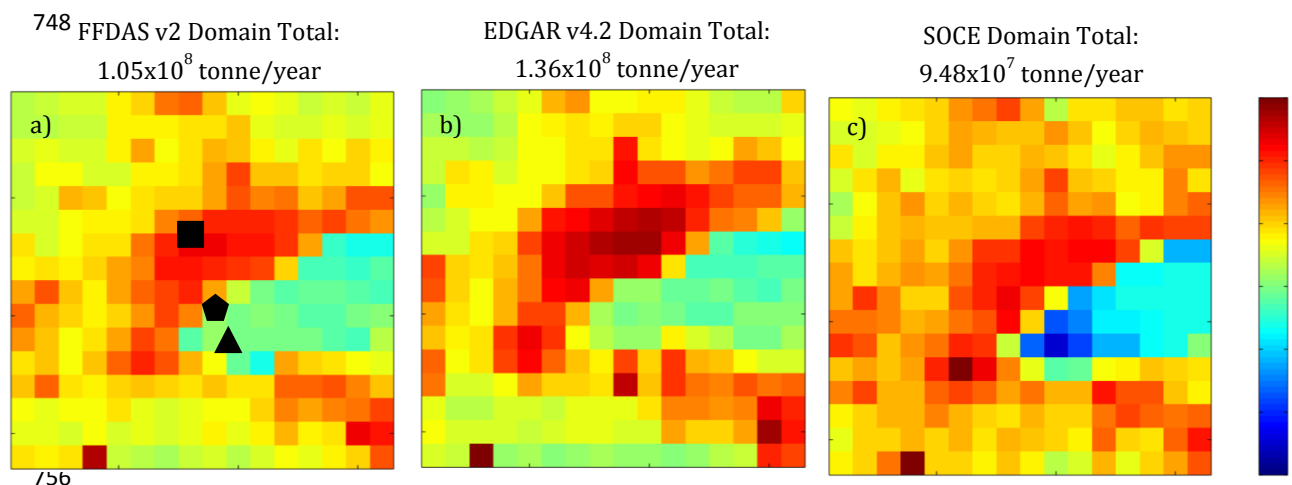


714 **Figure 1: Total anthropogenic CO₂ emissions for a weekday in February 2010 estimated by the**
715 **SOCE inventory for the province of Ontario and by the FFDAS v2 inventory for the remainder**
716 **of the GEM-MACH PanAm domain. Locations of in-situ measurements of CO₂ in the Southern**
717 **Ontario GHG Network are shown in the inset (Downsview = square, Egbert = circle, Hanlan's**
718 **Point = triangle, Turkey Point = diamond). The Downsview and Hanlan's Point sites are both**
719 **located in the GTA. Units: g CO₂/second/grid cell.**

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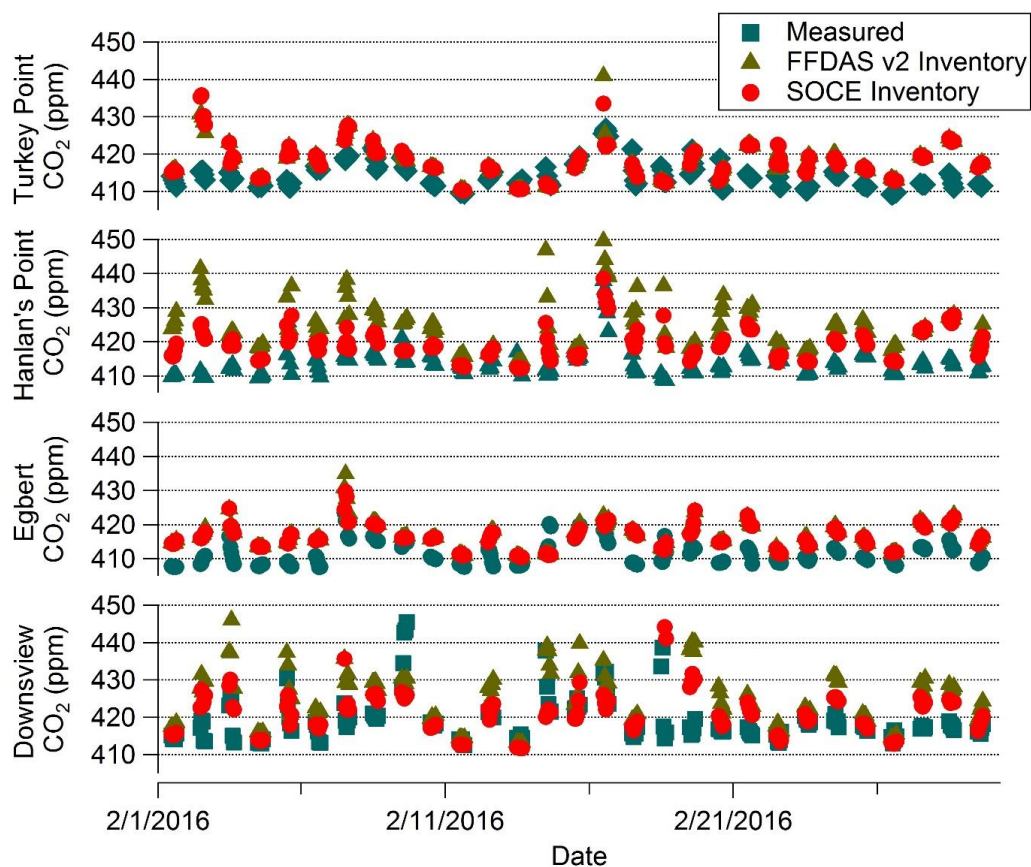


744 **Figure 2: Anthropogenic CO₂ emissions for a weekday in February 2010 in southern Ontario.**
745 **Emissions are estimated by the SOCE inventory for the (a) Area sector; (b) sum of the**
746 **Residential and Commercial sectors; (c) Point sector; (d) Marine sector; (e) On-road sector;**
747 **(f) Off-road sector. Units: log₁₀(g CO₂/second/grid cell).**



757 **Figure 3: Comparison of spatial distribution of annual CO₂ emissions inventories for the black-**
 758 **box area (shown in Figure 2a) at 0.1° x 0.1° resolution. Panel a) shows the FFDAS v2 inventory**
 759 **estimate, Panel b) shows the EDGAR v4.2 inventory estimate and Panel c) shows the SOCE**
 760 **inventory estimate. Units: log₁₀(tonne CO₂/year/grid cell). Domain totals are shown on top of**
 761 **each panel and locations of in-situ measurements of CO₂ for three stations in the Southern**
 762 **Ontario GHG Network are shown on Panel a (Downsview = square, Hanlan's Point = triangle,**
 763 **TAO = pentagon). The other two stations, Egbert and Turkey Point, are located outside this**
 764 **area.**

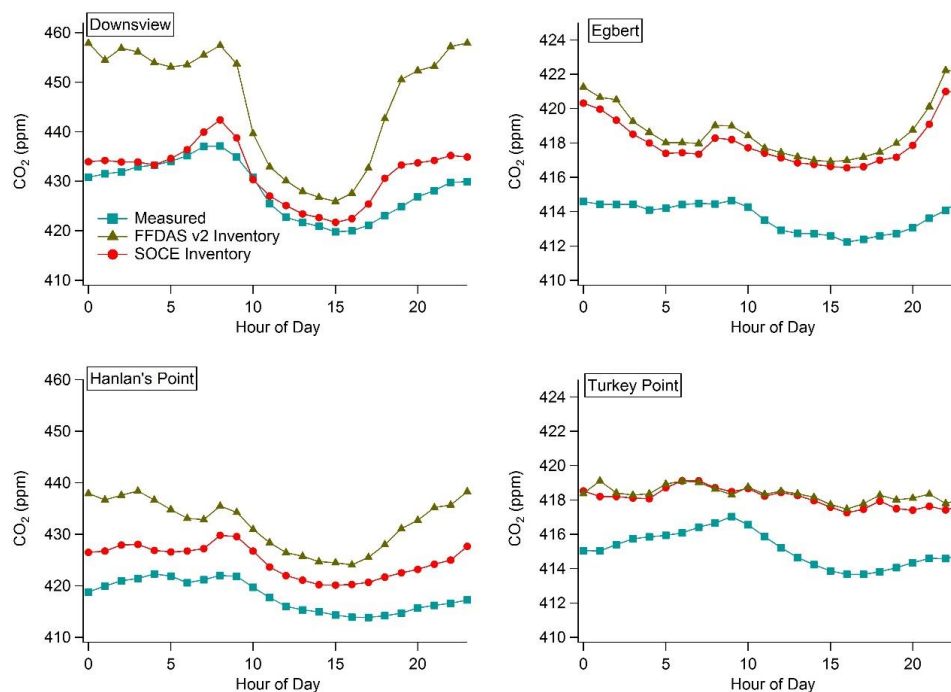
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767 **Figure 4: Time series of measured (blue) and modelled February afternoon (12:00-16:00 EST)**
 768 **CO₂ mixing ratios for the four sites used in this study. The red and gold markers are the**
 769 **modelled mixing ratios when using the SOCE CO₂ inventory and the FFDAS v2 inventory,**
 770 **respectively.**

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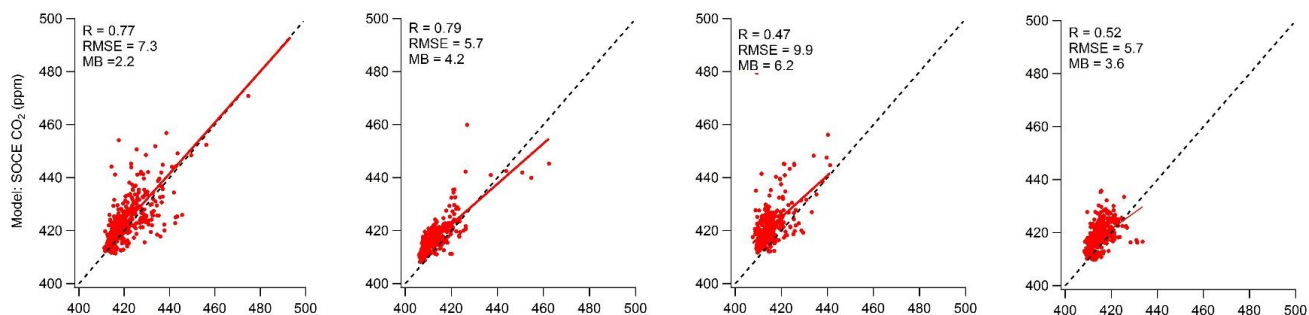
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775 **Figure 5: Time series of mean measured (blue) and modelled diurnal CO₂ mixing ratios at the**
 776 **four sites considered in this study for January – March 2016. The red and gold markers are**
 777 **the modelled diurnal mixing ratios when using the SOCE CO₂ inventory and the FFDAS v2**
 778 **inventory, respectively. Note the difference in scale for urban and rural sites.**

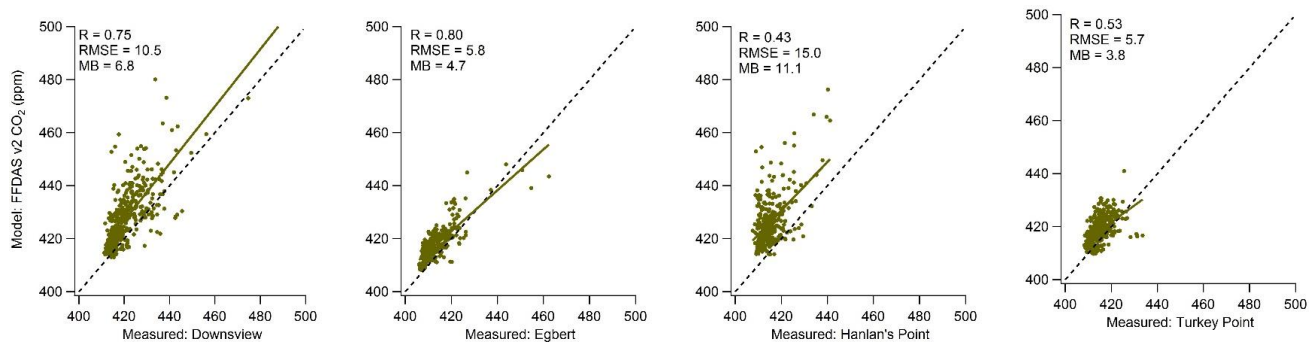
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782 **Figure 6: Scatter plot of the modelled and measured afternoon (12:00-16:00 EST) CO₂ mixing ratios from**
783 **January-March, 2016 at the four monitoring stations used in this study. The top and bottom panels show**
784 **measurement-model correlation when the SOCE inventory and the FFDAS v2 inventory were used,**
785 **respectively. The model vs. measurement Correlation Coefficient (R), root mean square error (RMSE) and**
786 **mean bias (MB) (units: ppm) are provided within each panel. Solid lines are the standard major axis**
787 **regression lines and dashed lines are 1:1 lines shown for reference.**

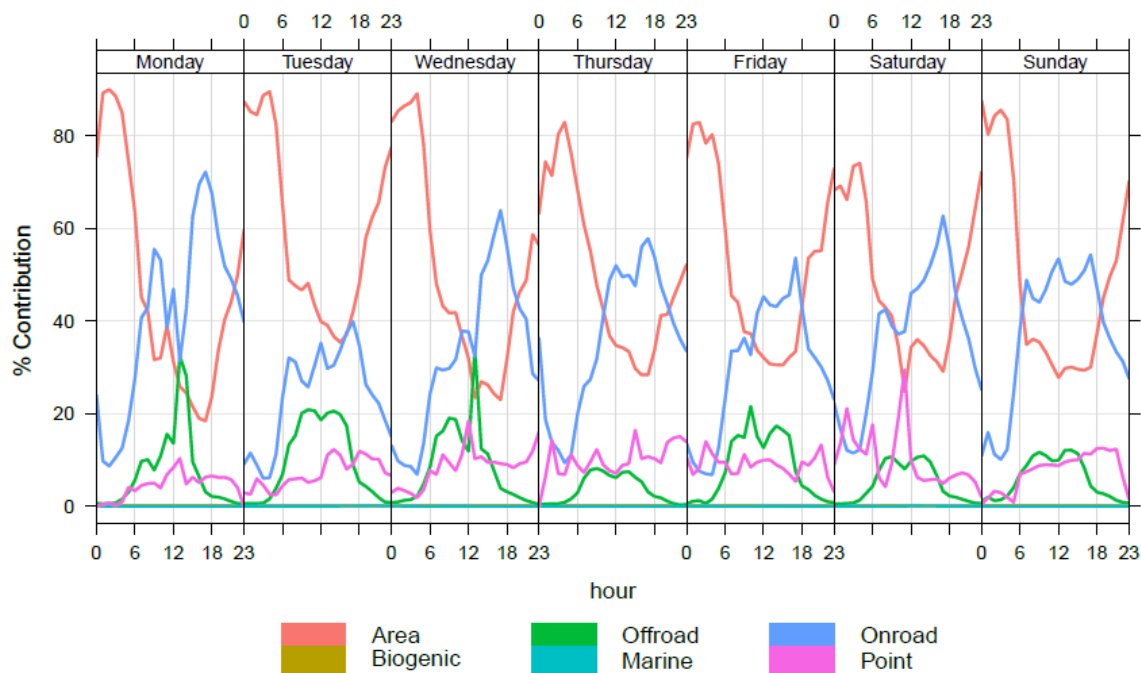


Figure 7: Modelled sectoral percent contributions to diurnal local CO₂ enhancement for February 2016 at Downsview averaged by day of week. Note: Area = Area + Residential natural gas combustion + Commercial natural gas combustion. (Time zone is EST).