

## Point-by-point response to review comments on manuscript acp-2017-582 “Quantifying black carbon light absorption enhancement by a novel statistical approach”

By Cheng Wu et al.

We thank the two anonymous reviewers for their constructive comments to improve the manuscript. Our point-by-point responses to the review comments are listed below. Changes to the manuscript are marked in blue in the revised manuscript. The marked manuscript is submitted together with this response document.

### Anonymous Referee #1

**R1-Q1.** This paper proposed a new approach to characterize the light absorption enhancement of BC and reported its application to the observed results based on an aethalometer and a thermo-optical carbon analyzer in Pearl River Delta area. This manuscript includes sufficient originality, and the topic seems to fit the journal. However, there are several concerns on the accuracy of the light absorption measurements, definition of  $E_{abs}$ , and discussions on the observed optical properties. I believe that the points below should be addressed before considering the publication in ACP.

**Author’s Response:** Thanks for the very insightful and detailed comments. Please see below for point-by-point response to reviewers’ comments.

### General comments

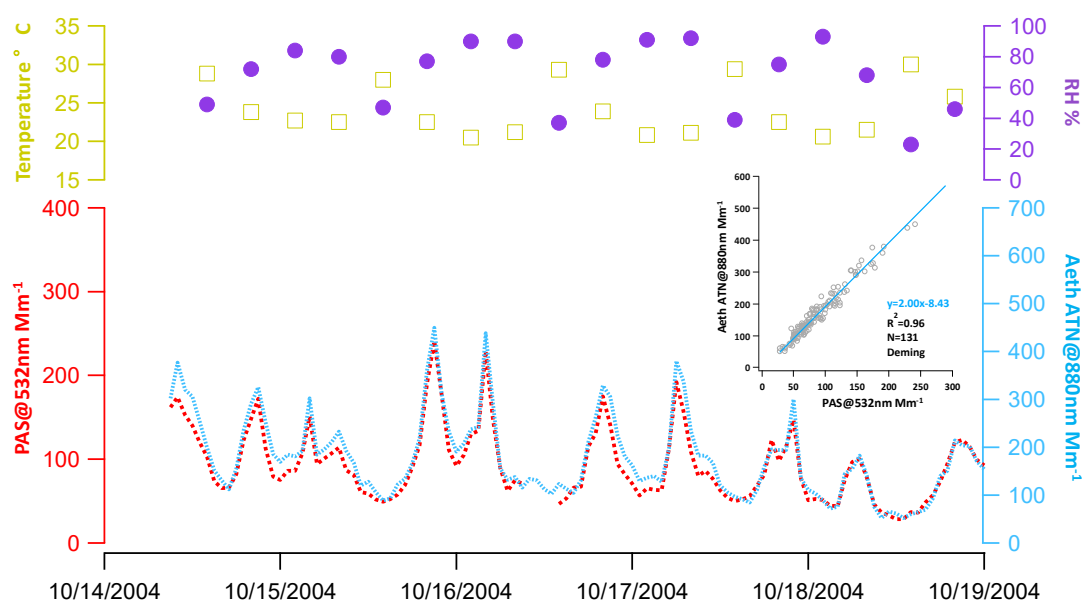
**R1-Q2.** 1) Accuracy of the aethalometer measurements. Although the authors used the correction scheme reported by Weingartner et al. (2003) to correct the artifacts due to aerosol loading, filter matrix and scattering effect, it is not clear that the light absorption for “coated-BC” can be accurately measured (or corrected), especially under “high RH” conditions. For example, Arnott et al (2003) reported that a filter-based photometer has large potential artifacts above 80% RH. The authors need to give more detailed information on accuracy of aethalometer measurements.

**Author’s Response:** We add the following content in section 2.1 to elaborate the factors that can affect the accuracy of Aethalometer in details.

Besides these artifacts, RH is also a source of  $\sigma_{abs}$  measurement uncertainty. Elevated RH is not only a driving force of increased  $\sigma_{abs}$  due to the hygroscopic growth of particles, but also a factor affecting ambient  $\sigma_{abs}$  measurements. Previous studies found  $\sigma_{abs}$  by PAS exhibit a systematic decrease when  $RH > 70\%$  (Arnott et al., 2003; Kozlov et al., 2011). Water evaporation was found as the major cause for the biased PAS  $\sigma_{abs}$  measurements under high RH (Raspert et al., 2003; Lewis et al., 2009; Langridge et al., 2013). Filter-based measurements are also affected under high RH conditions. For example, Arnott et al. (2003) observed erratic responses by particle soot absorption photometer (PSAP) as RH changed. The main reason is traced to the hydrophilic cellulose membrane, which serves to reinforce the quartz filter used in PSAP. The fibers can swell and shrink as RH changes, causing unwanted light attenuation signal. The PTFE-coated glass-fiber tape has become available since 2012 for the recent model of Aethalometer to minimize the RH interference (Magee-Scientific, 2017). A study by Schmid et al. (2006) reported dependency of PSAP  $\sigma_{abs}$  on RH, but found negligible effect of RH on Aethalometer performance. It is also worth noting that RH in the Aethalometer optical chamber may be lower than the ambient RH due to the

slightly elevated temperature inside the instrument. The magnitude of RH difference was found similar between different instruments: 20% for the Aethalometer (Schmid et al., 2006) and 15% for the nephelometer (Guyon et al., 2004). The RH in the Aethalometer optical chamber was not measured in this study. We expected its level to be slightly lower than the ambient RH. Cappa et al. (2008) found  $\sigma_{abs}$  measurements by PSAP and PAS maintained a high linearity even under high RH conditions (65-91%). Inter-comparison studies demonstrated that with proper corrections, Aethalometer  $\sigma_{abs}$  measurements agree well with those by PAS (Ajtai et al., 2011). During the inter-comparison study of an Aethalometer (AE-16) and a PAS in Guangzhou (Wu et al., 2009), good correlation was found ( $R^2=0.96$ ) as shown in Figure S1. These comparison results imply that the Aethalometer results are linearly correlated with PAS measurements and RH has a limited interference on Aethalometer measurements.

The Aethalometer measurements correlated well with the PAS results during a field campaign in Guangzhou as evidenced by the high  $R^2$  (0.96) shown in Figure R-1. The ambient RH during the comparison period ranged from 20% to 95%. Since the PAS inlet was equipped with a dryer while Aethalometer was sampling without a dryer, the high  $R^2$  proved that RH did not affect the correlation between Aethalometer and PAS. The results proved that Aethalometer can achieve comparable precision of PAS for hourly resolution data. We add tests in section 4.1 to evaluate the robustness of MRS against systematic biases. The results show that MRS is insensitive to systematic biases and details can be found in the response to R2-Q5.



**Figure R-1.** Comparison of collocated Aethalometer and PAS at Guangzhou (Oct 2004). Both PAS and Aethalometer (AE-16) were equipped with  $PM_{2.5}$  inlets. RH of the sampled air was controlled to be <45% for PAS. Aethalometer sampling was conducted without RH control.

**R1-Q3.** 2) Definitions of the  $MAE_p$  and  $E_{abs}$  The  $MAE_p$  obtained by the method in this study is MAE value for primary emission source. I think that the  $MAE_p$ , which was obtained in this study, includes the effects of “lensing effect” due to co-emitted OC. Therefore, the definition of  $E_{abs}$  values should be

different with the  $E_{\text{abs}}$  in the previous studies, at least, using TD technique. Detailed explanation on this point should be added.

**Author's Response:** Thanks for raising this point, which we did not make it clear in our manuscript. Yes, we agree with the reviewer that the nature of  $\text{MAE}_{\text{p,TD}}$  is different from  $\text{MAE}_{\text{p}}$  at emission source. The following contents are added to clarify this point.

In introduction:

It's also worth noting that  $\text{MAE}_{\text{p}}$  by the TD approach is different from the  $\text{MAE}_{\text{p}}$  at the emission source. First, the morphology of thermally denuded BC particles (compact aggregates) is different from that of freshly emitted BC particles (chain-like aggregates). Second, most of the coatings is removed with the TD denuded soot particles, but freshly emitted soot particles usually come with a thin coating of OC formed from condensation of OC vapors as the temperature drops from engine to the ambient air. As a result, the  $\text{MAE}_{\text{p}}$  by TD approach is expected to be lower than the  $\text{MAE}_{\text{p}}$  of emission source. In this sense, the TD approach may not be a perfect "time machine" to reverse the aging process for  $E_{\text{abs}}$  determination.

In section 4.1:

As mentioned in section 1, the definition of  $\text{MAE}_{\text{p}}$  by the TD approach is different from the  $\text{MAE}_{\text{p}}$  of emission source. The TD  $\text{MAE}_{\text{p}}$  is expected to be slightly lower than the  $\text{MAE}_{\text{p}}$  of emission source. Therefore, the corresponding  $E_{\text{abs}}$  are slightly different and it should be cautioned when comparing MRS-derived  $E_{\text{abs}}$  with  $E_{\text{abs}}$  by the TD approach and Mie simulations.

**R1-Q4.** 3) The effect of biomass burning

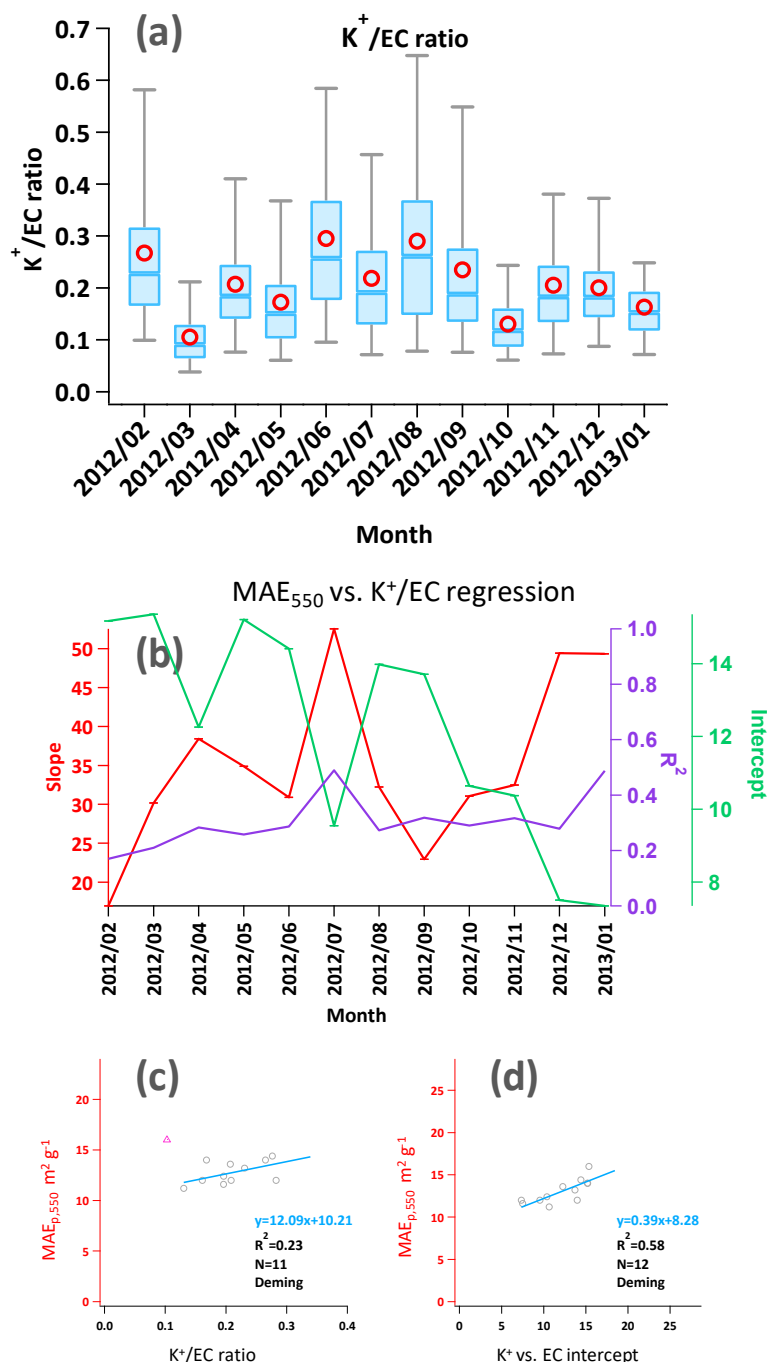
3-1) Lines 371-374: "During the rainy season when oceanic prevailing wind dominates, BC from BB emission in Southeast Asia can reach PRD through long range transport (LRT), resulting in an elevated  $\text{K}^+/\text{EC}$  ratio and  $\text{MAE}_{550}$ , which might be a combination of a thicker coating when freshly emitted from BB sources and enhanced coating during LRT."

=> If so, I think the  $\text{MAE}_{\text{p}}$  values should be higher in summer (rainy season). Please consider about this point.

**Author's Response:** BB is one aerosol source affecting  $\text{MAE}_{\text{p},550}$  but not the dominating source. As shown in Figure 8 in the revised manuscript, the data points are scattered and the  $R^2$  (0.33) is relatively low. The intercept in Figure 8 represents the part of MAE that are not explained by  $\text{K}^+/\text{EC}$  ratio. To further clarify this point, we calculate  $\text{K}^+/\text{EC}$  ratio and  $\text{MAE}_{550}$  -  $\text{K}^+/\text{EC}$  regression intercept of each month, then we compare the correlation of  $\text{MAE}_{\text{p},550}$  vs.  $\text{K}^+/\text{EC}$  ratio and  $\text{MAE}_{\text{p},550}$  vs.  $\text{K}^+/\text{EC}$  intercept. As shown in Figure R-2,  $\text{MAE}_{\text{p},550}$  has higher correlation with  $\text{K}^+/\text{EC}$  intercept ( $R^2=0.58$ ) than with  $\text{K}^+/\text{EC}$  ratio ( $R^2=0.23$ ). These results imply that BB is one of the contributors to the  $\text{MAE}_{\text{p},550}$  variations, but unlikely to be the dominating one. As a result, in some summer months, the  $\text{MAE}_{\text{p},550}$  is higher than the winter months but some months are not. Other BC sources and BB combustion conditions may affect the overall  $\text{MAE}_{\text{p},550}$ . The relevant content is revised as follow:

During the rainy season when oceanic wind prevails, BC from BB emission in Southeast Asia can reach PRD through long range transport (LRT), resulting in an elevated  $\text{K}^+/\text{EC}$  ratio and  $\text{MAE}_{550}$ . The Deming regression intercept (11.89) in Figure 8 represents the MAE without the BB effect. This non-BB  $\text{MAE}_{550}$  ( $11.89 \text{ m}^2 \text{ g}^{-1}$ ) is only slightly lower than  $\text{MAE}_{\text{p},550}$  ( $13 \text{ m}^2 \text{ g}^{-1}$ ) obtained in section 4.3, implying that a large fraction of

MAE<sub>p,550</sub> could not be explained by the BB source. Additional evidence was obtained through examining regression relationships of MAE<sub>p,550</sub> with K<sup>+</sup>/EC month-by-month (Figure S17b). Correlation of monthly MAE<sub>p,550</sub> vs. K<sup>+</sup>/EC ratio yield a R<sup>2</sup> of 0.23 (Figure S17c). In contrast, a much higher correlation (R<sup>2</sup>=0.58) was observed (Figure S17d) between MAE<sub>p,550</sub> and non-BB MAE<sub>550</sub> (i.e., K<sup>+</sup>/EC intercepts from Figure S17b). These results imply that BB is one of the contributors to the MAE<sub>p,550</sub> variations, but unlikely to be the dominating one.



**Figure R-2.** (a) Monthly variations of K<sup>+</sup>/EC ratio from 2012 Feb to 2013 Jan at NC site. (b) Monthly regressions between MAE<sub>550</sub> and K<sup>+</sup>/EC with slope in red, intercept in green and R<sup>2</sup> in purple. (c) regressions between monthly MAE<sub>p,550</sub> and K<sup>+</sup>/EC. (d) regression between monthly MAE<sub>p,550</sub> and intercepts from (b).

**R1-Q5.** 3-2) Lines 384-388: “Since the monthly average AAE in wintertime didn’t exceed 1.2 (Table S3), the variations of AAE in the PRD are likely more associated with thicker coatings rather than the contribution of BrC. The results also imply that attempts on BrC absorption attribution for the PRD dataset presented in this study could be risky, considering that elevation of AAE is actually dominated by coating (Lack and Langridge, 2013).”

=> Because higher  $E_{abs}$  values were observe in summer, the coating would be thicker in summer. The suggestion that “the higher AAE in winter are likely more associated with thicker coatings” is not reasonable. Please consider about this point.

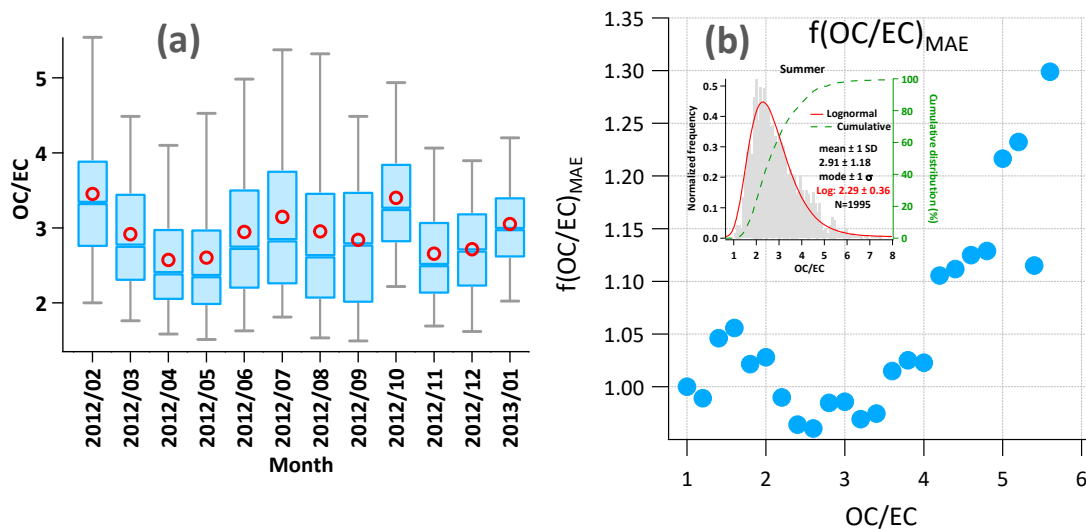
**Author’s Response:** The point we intend to make here is that since the seasonal variations of  $AAE_{470-660}$  is small, such variations are likely induced by the coating effect as shown in the Mie simulation rather than the presence of BrC. The corresponding text is rephrased as below to improve the clarity.

Since the monthly average  $AAE_{470-660}$  in wintertime did not exceed 1.2 (Table S3), the variations of  $AAE_{470-660}$  in the PRD are more likely associated with coatings rather than the contribution of BrC.

**R1-Q6.** 4) The effect of relative humidity (RH) on optical properties As mentioned above, the authors need to show some evidences suggesting that the obtained positive correlation between  $f(RH)$  (i.e., MAE) and RH is not due to the artifact of RH on aethalometer measurements. If the measurements are assumed to be accurate, the observed higher  $E_{abs}$  (and MAE) and lower AAE values in summer may be explained only by RH, considering that the RH may be higher in summer and RH for air masses from Southeast Asia. Which one of higher OC/EC ratio and higher RH do you suggest as a main contributor for observed higher  $E_{abs}$  in summer?

**Author’s Response:** The study by Schmid et al. (2006) found PSAP  $\sigma_{abs}$  shows dependency on RH, but the effect of RH on Aethalometer performance is neglectable. See our response to R1-Q2 for more detailed discussion on RH effect.

The contribution of elevated OC/EC ratio to the increase of  $E_{abs}$  might not be as important as RH in the summer time. First, the OC/EC ratios in summer is only slightly higher than other seasons (Figure R-3a). Second, as shown by the  $f(OC/EC)_{MAE}$  plot (Figure R-3b), OC/EC induced MAE enhancement only occurred when  $OC/EC > 4$ , which corresponded to the data with the highest 20% OC/EC ratio. As a result, the OC/EC induced MAE enhancement was only important for episodes of high OC/EC hours, which have a limited temporal coverage.



**Figure R-3.** (a) Monthly variations of OC/EC (b) MAE<sub>550</sub> enhancement as a function of OC/EC.

## Specific comments

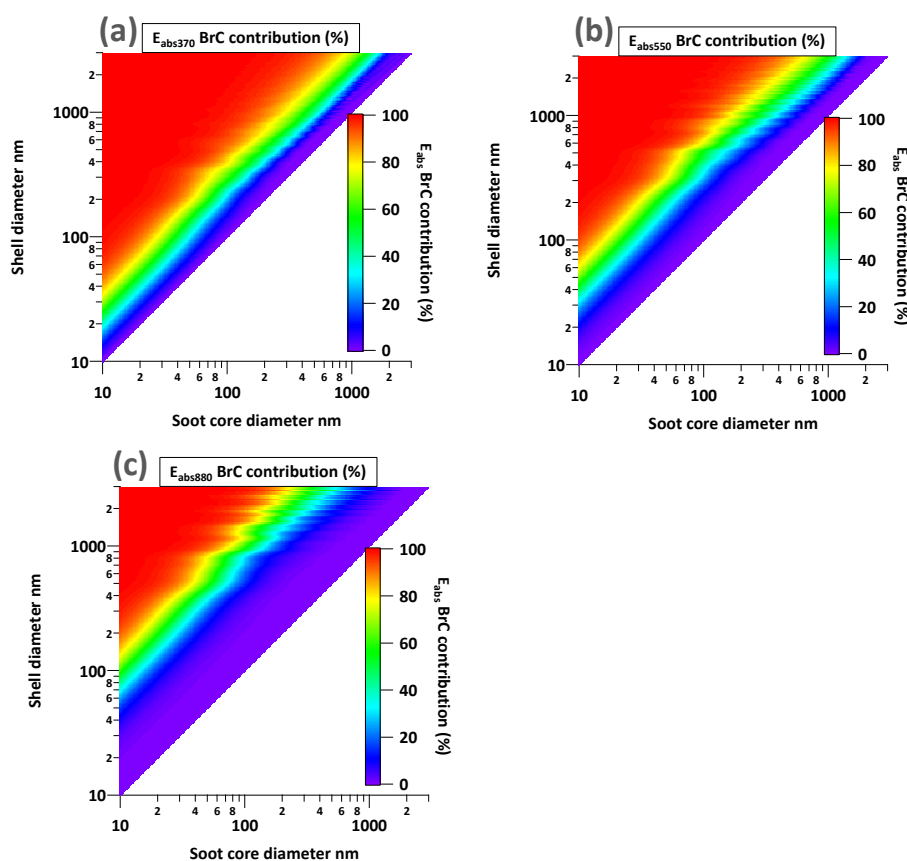
**R1-Q7.** 1) Line 235-236 “Thus, a BrC coating (brown shell) scenario is also considered in Mie simulation following the wavelength dependent RI suggested by Lack and Cappa (2010).” => Because the results of Mie simulation should be varied depending on the RI values given here, I recommend adding the range of RI values.

**Author’s Response:** Suggestion taken. The corresponding content is revised as follows:

Thus, a BrC coating (brown shell) scenario is also considered in Mie simulation following the wavelength-dependent RI suggested by Lack and Cappa (2010), which ranges from  $1.55-0.059i$  (370 nm) to  $1.55-0.0005i$  (950 nm).

**R1-Q8.** 2) Section 3.2.4, 1st paragraph. For the brown shell scenario, I think that both of lensing effect and light absorption by BrC contribute to the  $E_{abs}$  values. I recommend adding the fraction of each contribution.

**Author’s Response:** We agree with the reviewer on this point. Fractional contribution from BrC is discussed as shown below. The following figure is added in SI (Figure S7).



**Figure R-4.** Mie simulated BrC absorption contribution to total  $E_{abs}$  (lensing effect + BrC absorption) in the brown shell scenario. (a) 370 nm (b) 550 nm (c) 880 nm.

The following text is added in the revised manuscript (section 3.2.4):

The main reason behind is that in the brown shell scenario, both lensing effect and BrC absorption contribute to  $E_{abs}$ . As shown in Figure S7, the BrC absorption contribution to total  $E_{abs}$  strongly depends on coating thickness and is insensitive to soot core diameters. When the coating is relatively thin ( $<5$  nm for  $\lambda@370$  nm,  $<15$  nm for  $\lambda@550$  nm and

<40 nm for  $\lambda@880$  nm), BrC absorption contribution to the total  $E_{\text{abs}}$  is less than 20%. As the coating increases to a certain level (~15 nm for  $\lambda@370$  nm, ~35 nm for  $\lambda@550$  nm and ~90 nm for  $\lambda@880$  nm), BrC absorption contribution is comparable to the lensing effect contribution, each contributing ~50% to the total  $E_{\text{abs}}$ . When the BrC coating is sufficiently thick (>30 nm for  $\lambda@370$  nm, >90 nm for  $\lambda@550$  nm and >110 nm for  $\lambda@880$  nm), BrC absorption dominates the  $E_{\text{abs}}$  contribution. As a result, if BrC coating is indeed present in ambient samples, a strong wavelength dependent  $E_{\text{abs}}$  could be observed, since a BrC coating of 30 nm would be enough to induce a large amount of detectable  $E_{\text{abs}}$  in the UV range.

**R1-Q9.** 3) Section 4.1, 2nd paragraph The annual average  $MAE_p$  value ( $13 \text{ m}^2/\text{g}$ ) is larger than those estimated for bare BC. It may indicate that the obtained  $MAE_p$  values include the effects of “lensing effect” due to co-emitted OC, as mentioned above. If so, caution on the definition of  $MAE_p$  in each study should be taken during comparison with other studies. Also, it is better to add information on the wavelength used in each study.

**Author’s Response:** The authors fully agree with the reviewer that the difference in  $MAE_p$  definition should be emphasized. Following contents are added in section 4.1 to clarify this point.

$MAE_p$  by MRS represents the  $MAE_p$  at the emission source, which is different from the  $MAE_p$  by the TD approach for two reasons. First, the morphology of thermally denuded BC particles (compact aggregates) is different from that of freshly emitted BC particles (chain-like aggregates). Second, most of the coatings are removed for TD denuded soot particles, but freshly emitted soot particles usually come with a thin coating of OC formed from condensation of OC vapors as the temperature drops from the flame to the ambient air. As a result, the MRS-derived  $MAE_p$  is expected to be higher than the  $MAE_p$  by the TD approach.

The MAE mentioned from literature was scaled to 550 nm for comparison. The following text is added to improve the clarity.

For comparison purpose, MAE measured at original wavelength and MAE scaled to 550 nm following the  $\lambda^{-1}$  assumption are both shown in Table S1. The MAE comparisons discussed below are MAE at 550 nm.

**R1-Q10.** 4) Section 4.5 How did the absorption coefficients for dry conditions determine to obtained  $f(\text{RH})$ ?

**Author’s Response:** The sampled air into the Aethalometer was not dried in this study. As the sampling covered a whole year, the RH also spanned a range sufficiently large to study the RH effect. The MAE data obtained at  $\text{RH}=30\%$  (as shown in Figure 9a of the revised manuscript, the starting point of the  $f(\text{RH})$  curve) are used as the dry condition to calculate  $f(\text{RH})$ .

**R1-Q11.** 5) Figure 9 What do to want to suggest from 30:1 line in Fig. 9?

**Author’s Response:** The 30:1 line is to serve as a reference line with an integer slope that is close to the regressed slope through the origin.

## References

Arnott, W. P., Moosmuller, H., Sheridan, P. J., Ogren, J. A., Rasp, R., Slaton, W. V. 2003. Photoacoustic and Filter-Based Ambient Aerosol Light Absorption Measurements: Instrument Comparisons and the Role of Relative Humidity. *J. Geophys. Res.-Atmos.*, 108: 4034–4045.



# Quantifying black carbon light absorption enhancement by a novel statistical approach

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## Abstract

Black carbon (BC) particles in the atmosphere can absorb more light when coated by non-absorbing or weakly absorbing materials during atmospheric aging, due to the lensing effect. In this study, the light absorption enhancement factor,  $E_{\text{abs}}$ , was quantified using one year's measurement of mass absorption efficiency (MAE) in the Pearl River Delta region (PRD). A new approach for calculating primary MAE ( $\text{MAE}_p$ ), the key for  $E_{\text{abs}}$  estimation, is demonstrated using the Minimum Residual Squared (MRS) method, exploring the inherent source independency between BC and its coating materials. A unique feature of  $E_{\text{abs}}$  estimation by the MRS approach is its insensitivity to systematic biases in EC and  $\sigma_{\text{abs}}$  measurements. The annual average  $E_{\text{abs}550}$  is found to be  $1.50 \pm 0.48$  ( $\pm 1$  S.D.), exhibiting a clear seasonal pattern with higher values in summer and lower in the winter. Elevated  $E_{\text{abs}}$  in the rainy summer season is likely associated with aged air masses dominating from marine origin, along with long-range transport of biomass burning influenced air masses from Southeast Asia.  $E_{\text{abs}}$

31 induced by hygroscopic growth at elevated RH could be as high as 1.3. Core-shell Mie simulations  
32 along with measured  $E_{abs}$  and Angstrom absorption exponent (AAE) constraints suggest that in the  
33 PRD, the coating materials are unlikely to be dominated by brown carbon and the coating thickness is  
34 higher in the rainy season than the dry season. A negative correlation is found between  $AAE_{470-660}$  and  
35 RH, suggesting a dominant particle size of  $D_{core} = 130$  nm and  $D_{shell}/D_{core}$  range of 2 to 4.

36

## 37 **1 Introduction**

38 Originating from incomplete combustion, black carbon (BC) is a crucial constituent of  
39 atmospheric aerosols, and is an air pollutant itself, having an adverse health impacts on humans (Suglia  
40 et al., 2008). BC has also been recognized as the third most important climate forcer due to its broad  
41 light absorbing capability across the UV-Vis-IR spectrum (IPCC, 2013). BC can alter the climate in a  
42 variety of ways, including by direct forcing (Bond et al., 2011), affecting cloud cover (Koch and Del  
43 Genio, 2010) and precipitation (Tao et al., 2012), reducing the albedo of snow and ice (Hansen and  
44 Nazarenko, 2004) and causing surface dimming (Wild, 2011). The climate effects of BC can be global  
45 or regional (Ramanathan and Carmichael, 2008). A recent study found BC can modify planetary  
46 boundary layer meteorology, and thus enhance local pollution indirectly (Ding et al., 2016). However,  
47 due to its variable optical characteristics induced during atmospheric aging, large uncertainties still  
48 exist in estimating the radiative forcing from BC. Optical properties of BC can be predicted by  
49 knowing the mass concentration, mixing state and size distribution, which collectively serve as the  
50 cornerstone for modeling the climate effect of BC. In 3D modeling studies, to conserve computational  
51 resources, the mass absorption efficiency (MAE) or mass absorption cross-section (MAC) is widely  
52 used to convert black carbon mass concentration to light absorption coefficient ( $\sigma_{abs}$ ). MAE is a  
53 quantity to describe the light absorption ability per unit EC mass:

$$54 \quad MAE (m^2 g^{-1}) = \frac{\text{absorption coefficient } \sigma_{abs} (Mm^{-1})}{EC \text{ mass concentration } (\mu g m^{-3})} \quad (1)$$

55 As a fundamental input parameter, MAE has a critical impact on BC's radiative forcing  
56 estimation in climate modeling studies. Mixing state is one of the governing factors affecting MAE.

57 Light absorption of soot particles is enhanced when coated with non-absorbing materials (Fuller et al.,  
 58 1999) or weakly absorbing materials (Lack and Cappa, 2010) during atmospheric aging. The coating  
 59 materials can focus more light onto the soot core through the lensing effect, resulting in elevated MAE  
 60 (Wang et al., 2017). Strong correlations between MAE and the number/volume fraction of coated  
 61 particles have been reported in urban areas like Tokyo (Naoe et al., 2009), Shenzhen (Lan et al., 2013)  
 62 and Xi'an (Wang et al., 2014), implying that the elevated MAE observed at these locations was mainly  
 63 due to the elevated fraction of coated of soot particles. Total absorption ( $\sigma_{abs,t}$ ) of coated particles can  
 64 be separated into two parts: primary absorption ( $\sigma_{abs,p}$ ) due to the uncoated soot core alone, and extra  
 65 absorption ( $\sigma_{abs,c}$ ) due to lensing effect of the coating (Bond et al., 2006; Jacobson, 2006; Liu et al.,  
 66 2016a) and the presence of secondarily formed brown carbon (BrC) (Lack and Cappa, 2010; Liu et al.,  
 67 2016b).

$$68 \quad \sigma_{abs,t} = \sigma_{abs,p} + \sigma_{abs,c} \quad (2)$$

69 The absorption enhancement factor ( $E_{abs}$ ) then can be defined as ratio of the total absorption and  
 70 primary absorption coefficients or the corresponding MAE values:

$$71 \quad E_{abs} = \frac{\sigma_{abs,t}}{\sigma_{abs,p}} = \frac{MAE_t}{MAE_p} \quad (3)$$

72 Where  $MAE_p$  represents the ratio of  $\sigma_{abs,p}/EC$  for uncoated soot particles, similar to the concept of  
 73 the primary OC/EC ratio in the EC tracer method:

$$74 \quad MAE_p = \frac{\sigma_{abs,p}}{EC} \quad (4)$$

75 And the MAE of coated BC can be defined as:

$$76 \quad MAE_t = \frac{\sigma_{abs,t}}{EC} \quad (5)$$

77 Thus, elevated MAE induced by coating during atmospheric aging results in an  $E_{abs}$  larger than 1.

78 Previous model studies suggest that absorption by aged soot particles can be 1.5 times greater than  
 79 fresh soot (Fuller et al., 1999; Bond et al., 2006). Laboratory studies have demonstrated that soot  
 80 particles coated with SOA (Saathoff et al., 2003; Schnaiter et al., 2005) and sulfuric acid (Zhang et al.,  
 81 2008; Khalizov et al., 2009) can increase  $E_{abs}$ . An artificial coating experiment by Shiraiwa et al. (2010)

82 found an  $E_{\text{abs}}$  of 2 for graphite particles growing in diameter from 185 to 370 nm. A laboratory study  
83 by McMeeking et al. (2014) found that in the presence of BrC, light absorption enhancement is more  
84 pronounced at the shorter wavelength. A recent chamber study coupling actual ambient air with seed  
85 BC particles implies that the timescale for  $E_{\text{abs}}$  reaching 2.4 is only 5 hours in Beijing but 18 hours in  
86 Houston (Peng et al., 2016). Field studies conducted in recent years have also substantiated enhanced  
87 light absorption in Canada (Knox et al., 2009; Chan et al., 2011), US (Lack et al., 2012b), UK (Liu et  
88 al., 2015) and Japan (Nakayama et al., 2014; Ueda et al., 2016). In contrast, field studies in California,  
89 US (Cappa et al., 2012) found a weaker light absorption enhancement (6% on average). A recent study  
90 suggests the mass ratio of non-BC content to BC particles determines the occurrence of the absorption  
91 enhancement of black-carbon particles (Liu et al., 2017).

92 Two approaches are widely used to determine  $E_{\text{abs}}$  from ambient measurements. The first approach  
93 removes the coating materials on particles physically using a thermal denuder (TD) (Lack et al., 2012a)  
94 or by aerosol filter filtration-dissolution (AFD) (Cui et al., 2016b). The TD approach is briefly  
95 discussed here. Coating materials can be removed by TD at a working temperature around 200 to  
96 300 °C (depending on the charring characteristics of aerosols at the sampling site) to measure  $\sigma_{\text{abs},p}$ ,  
97 which are cycled with measurements of  $\sigma_{\text{abs},t}$  (without passing through TD), allowing  $E_{\text{abs}}$  to be  
98 obtained from the ratio of  $\sigma_{\text{abs},t}/\sigma_{\text{abs},p}$  following Eq.3. The major advantage of the TD approach is  
99 its ability to provide highly time resolved measurements (minutes). A photo-acoustic spectrometer  
100 (PAS) is commonly used with TD for detection to satisfy its high time resolution demands. As an in-  
101 situ technique, PAS eliminates the artifacts associated with filter-based methods (Weingartner et al.,  
102 2003; Coen et al., 2010) and is often considered as the reference instrument for light absorption  
103 coefficient determination (Arnott et al., 2003; Arnott et al., 2005). One limitation of the TD approach is  
104 that a universal optimal operation temperature does not exist. If the temperature is too low, the coating  
105 cannot be fully removed, and charring can occur if the TD temperature is too high, leading to biased  
106 results. Another issue is particle loss due to TD, which can be ~ 20% and needs to be taken into account  
107 (Ueda et al., 2016). It's also worth noting that  $\text{MAE}_p$  by the TD approach is different from the  $\text{MAE}_p$   
108 at the emission source. First, the morphology of thermally denuded BC particles (compact aggregates)  
109 is different from that of freshly emitted BC particles (chain-like aggregates). Second, most of the

110 coatings is removed with the TD denuded soot particles, but freshly emitted soot particles usually  
111 come with a thin coating of OC formed from condensation of OC vapors as the temperature drops  
112 from engine to the ambient air. As a result, the MAE<sub>p</sub> by TD approach is expected to be lower than  
113 the MAE<sub>p</sub> of emission source. In this sense, the TD approach may not be a perfect “time machine” to  
114 reverse the aging process for E<sub>abs</sub> determination.

115 The second approach is the MAE ratio method, which is also stated in Eq. 3. The key to this method  
116 is determining an appropriate MAE<sub>p</sub> that can represent the MAE from primary emissions. One  
117 approach is to adopt the reference MAE<sub>p</sub> from the literature but it may fail to represent the actual  
118 MAE<sub>p</sub> at a specific sampling site, since MAE<sub>p</sub> varies temporally and spatially. For example, MAE<sub>p</sub> of  
119 diesel soot was found to be 7.1 m<sup>2</sup>g<sup>-1</sup> at 532 nm (Adler et al., 2010). A much higher MAE<sub>p</sub> (16 m<sup>2</sup>g<sup>-1</sup>  
120 at 530 nm) was observed from natural gas flaring (Weyant et al., 2016). MAE<sub>p</sub> of biomass burning  
121 (BB) samples is highly varied due to a wide range of fuel types and combustion conditions (Reid et  
122 al., 2005; Roden et al., 2006). A range from 6.1 to 80.8 m<sup>2</sup>g<sup>-1</sup> was reported for BB MAE<sub>p</sub> at 550 nm  
123 (Pandey et al., 2016). Without the knowledge of source contributions, it is not feasible to derive a  
124 representative MAE<sub>p</sub> for E<sub>abs</sub> estimation. The other commonly used approach is to determine MAE<sub>p</sub>  
125 from the dependency of MAE on the number fraction of coated soot particles measured by SP2 (Lan  
126 et al., 2013). Since MAE (y axis) is positively correlated with the number fraction of coated soot  
127 particles (x axis), MAE<sub>p</sub> can be determined by extending the regression line to x=0. It is worth noting  
128 that this approach provides only a rough approximation of E<sub>abs</sub> since the parameter used here (coated  
129 soot particles number fraction) ignores other main drivers of light absorption enhancement (e.g.  
130 coating thickness). As a result, this approach is only valid for a period of measurements, for which  
131 coating thickness is relatively constant and the MAE variations are dominated by coated soot particles  
132 number fraction.

133 However, the high cost of the TD-PAS system and SP2 limit the field measurement of E<sub>abs</sub> around  
134 the world. In addition, long-term E<sub>abs</sub> measurements by a TD-PAS system and SP2 are not easily  
135 achieved and rarely reported. On the other hand, an Aethalometer and RT-ECOC analyzer can be  
136 effectively deployed for long term measurements and E<sub>abs</sub> estimation, at a relatively lower cost. In this  
137 study, based on one year of hourly MAE measurements (with the field carbon analyzer and

138 Aethalometer) at a suburban site in the Pearl River Delta (PRD) region of China, quantification of  
139 MAE<sub>p</sub> is demonstrated by a novel statistical approach, the Minimum R squared method (MRS) (Wu  
140 and Yu, 2016). The aim of this study is to demonstrate the capability of E<sub>abs</sub> estimation using a year-  
141 long dataset from cost-effective instrumentation. The seasonal variability of MAE, AAE and E<sub>abs</sub> in  
142 the PRD region are characterized and their dependency on air mass origin, biomass burning and RH  
143 are discussed. Abbreviations used in this study are summarized in Table 1 for a quick lookup.

## 144 **2 Ambient measurements**

145 Sampling was conducted from Feb 2012 to Jan 2013 at the suburban Nancun (NC) site (23° 0'11.82"N,  
146 113°21'18.04"E). NC, situated on the top of the highest peak (141 m ASL) in Guangzhou's Panyu  
147 district, is located at the geographic center of the PRD region, making it a representative location for  
148 average atmospheric mixing characteristics of city clusters in the PRD region. Light absorption  
149 measurements were performed by a 7-λ Aethalometer (AE-31, Magee Scientific Company, Berkeley,  
150 CA, USA). The Aethalometer was equipped with a 2.5 μm cyclone with a sampling flow rate of 4 L  
151 min<sup>-1</sup>. Weingartner's algorithm (Weingartner et al., 2003) was adopted to correct the sampling artifacts  
152 (aerosol loading, filter matrix and scattering effect) rooted in filter based method. A customized  
153 Aethalometer data processing program (Wu, 2017a) with graphical user interface was developed to  
154 perform data correction and detailed descriptions can be found in the SI (The program is available  
155 from <https://sites.google.com/site/wuchengust>). Details of the Aethalometer setup and data correction  
156 can be found in our previous paper (Wu et al., 2013).

157 EC mass concentrations were determined by a real time ECOC analyzer (Model RT-4, Sunset  
158 Laboratory Inc., Tigard, Oregon, USA). The sunset carbon analyzer was sampling on hourly cycles at  
159 a flow rate of 8 Lmin<sup>-1</sup> with a PM<sub>2.5</sub> sharp-cut cyclone inlet. For each measurement hour, the first  
160 45min were for sample collection and the remaining 15 min for thermal-optical analysis. OC is  
161 volatilized first by step-wise temperature ramping in an oxygen-free atmosphere while in the second  
162 stage EC is combusted in the presence of oxygen. Laser transmittance is applied to correct the charring  
163 artifact during the OC stage.

164 Considering a measurement **precision** of 5% for the Aethalometer (Hansen, 2005) and 24% for the  
165 RT-ECOC analyzer (Bauer et al., 2009), the propagated relative **precision** of  $E_{abs}$  ( $E_{abs,Unc}$ ) is 35%  
166 following Eq. S1&S2 in the SI. It should be noted that  $E_{abs,Unc}$  is mainly attributed to the  
167 measurement **precision** of EC by the RT-ECOC analyzer. Since the measurement **precision** of the RT-  
168 ECOC analyzer estimated by Bauer et al. (2009) is obtained from field measurement at an environment  
169 (EC below  $1 \mu\text{g m}^{-3}$ ) where EC is much lower than the present study (annual average EC  $2.63 \mu\text{g m}^{-3}$ ),  
170 the  $E_{abs,Unc}$  of 35% should be considered as an upper limit for the present study.

171 Light scattering was measured by an integrating nephelometer (Aurora-1000, Ecotech, Melbourne,  
172 Australia). Water soluble ions were measured by MARGA (The instrument for Measuring AeRosols  
173 and GAses)(ten Brink et al., 2007). **Both instruments are equipped with a PM<sub>2.5</sub> inlet to remove the**  
174 **coarse particles.**

## 175 **2.1 Uncertainties of MAE determination**

176 Two major uncertainties associated with the  $\sigma_{abs}$  and EC determination techniques should be taken  
177 into account when comparing MAE across different studies. For the  $\sigma_{abs}$  determination technique,  
178 photo-acoustic spectroscopy (PAS) is an in-situ technique free from filter based artifacts, but its  
179 application is limited by its high cost. The filter based optical transmittance method (e.g., Aethalometer  
180 and Multi Angle Absorption Photometer, MAAP) is the most widely used technique around the world,  
181 but data correction is needed to minimize the bias from artifacts due to the loading effect, matrix effect  
182 and scattering effect (Weingartner et al., 2003; Arnott et al., 2005; Schmid et al., 2006; Virkkula et al.,  
183 2007; Coen et al., 2010; Drinovec et al., 2017; Saturno et al., 2017). **Besides these artifacts, RH is also**  
184 **a source of  $\sigma_{abs}$  measurement uncertainty. Elevated RH is not only a driving force of increased  $\sigma_{abs}$**   
185 **due to the hygroscopic growth of particles, but also a factor affecting ambient  $\sigma_{abs}$  measurements.**  
186 **Previous studies found  $\sigma_{abs}$  by PAS exhibit a systematic decrease when  $\text{RH} > 70\%$  (Arnott et al., 2003;**  
187 **Kozlov et al., 2011). Water evaporation was found as the major cause for the biased PAS  $\sigma_{abs}$**   
188 **measurements under high RH (Raspert et al., 2003; Lewis et al., 2009b; Langridge et al., 2013). Filter-**  
189 **based measurements are also affected under high RH conditions. For example, Arnott et al. (2003)**  
190 **observed erratic responses by particle soot absorption photometer (PSAP) as RH changed. The main**  
191 **reason is traced to the hydrophilic cellulose membrane, which serves to reinforce the quartz filter used**  
192 **in PSAP. The fibers can swell and shrink as RH changes, causing unwanted light attenuation signal.**  
193 **The PTFE-coated glass-fiber tape has become available since 2012 for the recent model of**

194 Aethalometer to minimize the RH interference (Magee-Scientific, 2017). A study by Schmid et al.  
195 (2006) reported dependency of PSAP  $\sigma_{abs}$  on RH, but found negligible effect of RH on Aethalometer  
196 performance. It is also worth noting that RH in the Aethalometer optical chamber may be lower than  
197 the ambient RH due to the slightly elevated temperature inside the instrument. The magnitude of RH  
198 difference was found similar between different instruments: 20% for the Aethalometer (Schmid et al.,  
199 2006) and 15% for the nephelometer (Guyon et al., 2004). The RH in the Aethalometer optical chamber  
200 was not measured in this study. We expected its level to be slightly lower than the ambient RH. Cappa  
201 et al. (2008) found  $\sigma_{abs}$  measurements by PSAP and PAS maintained a high linearity even under high  
202 RH conditions (65-91%). Inter-comparison studies demonstrated that with proper corrections,  
203 Aethalometer  $\sigma_{abs}$  measurements agree well with those by PAS (Ajtai et al., 2011). During the inter-  
204 comparison study of an Aethalometer (AE-16) and a PAS in Guangzhou (Wu et al., 2009), good  
205 correlation was found ( $R^2=0.96$ ) as shown in Figure S1. These comparison results imply that the  
206 Aethalometer results are linearly correlated with PAS measurements and RH has a limited interference  
207 on Aethalometer measurements. In our study, careful corrective measures (Wu et al., 2013) are  
208 conducted for the Aethalometer  $\sigma_{abs}$  data treatment to minimize these artifacts. But such artifacts  
209 still cannot be fully eliminated.

210 For the EC determination, different thermal optical analysis (TOA) protocols can impact the  
211 measurement variability and thus MAE. As shown in Table S1, MAE for the same samples at Fresno  
212 varied from 6.1 to 9.3  $\text{m}^2 \text{g}^{-1}$ , depending on which EC analysis protocol was applied (Chow et al.,  
213 2009). Studies in the PRD found that discrepancies in measured EC by different analysis protocols  
214 could be as large as a factor of 5 (Wu et al., 2012; Wu et al., 2016a), which adds to the uncertainty for  
215 the MAE estimation. In addition, EC by TOA is also different from refractory BC (rBC) reported by  
216 the laser induced incandescence (LII) technique (e.g. single particle soot photometer, SP2). For  
217 example, two studies in Toronto (Knox et al., 2009; Chan et al., 2011) both used the PAS for  $\sigma_{abs}$   
218 measurement but different techniques for EC mass determination, resulting in very different MAE  
219 results. LII instruments are usually calibrated with a commercially available surrogate (e.g. fullerene)  
220 since direct calibration with ambient soot is not easy to achieve. Laborde et al. (2012) indicates that  
221 the incandescence response of SP2 exhibits a dependency on soot type (15% between fullerene and  
222 denuded diesel soot particles; 14% between biomass burning and denuded diesel soot particles). Due  
223 to the absence of widely accepted reference materials for EC, the uncertainties in EC determination



224 will exist in the foreseeable future. All these uncertainties, including the uncertainty of rBC mass  
225 determination by SP2, uncertainty of EC in TOA, the discrepancy between SP2 rBC and TOA EC and  
226 the discrepancy of  $\sigma_{abs}$  between filter transmission and photo-acoustic methods, can contribute to the  
227 differences in MAE listed in Table S1.

228 Systematic bias in MAE (e.g. overestimation of  $\sigma_{abs}$  and variability of EC mass by different  
229 TOA protocols) discussed above have little effect on  $E_{abs}$  estimation by MRS. As shown in Eq. 3,  $E_{abs}$   
230 is the ratio of  $MAE_t$  to  $MAE_p$  or  $\sigma_{abs,t}$  to  $\sigma_{abs,p}$ , thus most of the bias in EC mass or  $\sigma_{abs}$  is  
231 cancelled out during the  $E_{abs}$  calculation. More details are discussed in section 4.1.

## 232 **3 Methodology**

### 233 **3.1 MAE<sub>p</sub> estimation by MRS from the ambient data**

234 In this section, a new approach for MAE<sub>p</sub> estimation is introduced for  $E_{abs}$  determination, which  
235 requires the knowledge of differentiating  $\sigma_{abs,p}$  and  $\sigma_{abs,c}$  portions in  $\sigma_{abs,t}$ . The idea of  
236 decoupling  $\sigma_{abs,t}$  into  $\sigma_{abs,p}$  and  $\sigma_{abs,c}$  is conceptually similar to decoupling OC into primary OC  
237 (POC) and secondary OC (SOC) in the EC tracer method as shown in Table 2. In the EC tracer method,  
238 if  $(OC/EC)_p$  is known, POC can be determined from OC (Turpin and Huntzicker, 1991). The role of  
239 MAE<sub>p</sub> here is similar to the role of  $(OC/EC)_p$ , the primary OC/EC ratio in the EC tracer method (a  
240 comparison is given in Table 2). If MAE<sub>p</sub> (average MAE from primary emission sources) is known,  
241  $E_{abs}$  can be obtained from the ratio of  $MAE_t/MAE_p$  (Eq. 3). Therefore, the key for  $E_{abs}$  estimation is to  
242 derive an appropriate MAE<sub>p</sub>. It is worth noting that MAE<sub>p</sub> here does not represent MAE from a single  
243 or specific primary emission source, instead it reflects an average and effective MAE that has taken  
244 consideration of various primary emission sources. Thus, the MAE<sub>p</sub> is conceptually analogous to  
245  $(OC/EC)_p$  in the EC tracer method, in which the primary ratio reflects an overall ratio from primary  
246 emission sources rather than from a single primary source.

247 The Minimum R squared method (MRS) explores the inherent independency between  
248 pollutants from primary emissions (e.g., EC) and products associated with secondary formation  
249 processes (e.g., SOC,  $\sigma_{abs,c}$ ) to derive the primary ratios (e.g.,  $(OC/EC)_p$ , MAE<sub>p</sub>) in the EC tracer

250 method (Wu and Yu, 2016). When applying MRS for light absorption enhancement estimation, MRS  
251 is used to explore the inherent independency between EC and  $\sigma_{abs,c}$ , which is gained during  
252 atmospheric aging after emission. An example of MAE<sub>p</sub> estimation by MRS is shown in Figure 1.  
253 Firstly, the assumed MAE<sub>p</sub> value is varied continuously in a reasonable range (0.01 to 50 m<sup>2</sup> g<sup>-1</sup> as  
254 shown in Figure 1). Then at each hypothetical MAE<sub>p</sub>,  $\sigma_{abs,c}$  can be calculated by Eq. 6 (a combination  
255 of Eq. 2&4) using EC and  $\sigma_{abs,t}$  from ambient measurements.

$$256 \quad \sigma_{abs,c} = \sigma_{abs,t} - MAE_p \times EC \quad (6)$$

257 Accordingly, for each hypothetical MAE<sub>p</sub>, a correlation coefficient value (R<sup>2</sup>) of  $\sigma_{abs,c}$  vs.  
258 EC (i.e., R<sup>2</sup>( $\sigma_{abs,c}$ , EC)) can be obtained. The series of R<sup>2</sup>( $\sigma_{abs,c}$ , EC) values (y axis) are then plotted  
259 against the assumed MAE<sub>p</sub> values (x axis) as shown by the red curve in Figure 1. The physical meaning  
260 of this plot can be interpreted as follows. The  $\sigma_{abs,p}$  is the fraction of light absorption owing to  
261 primary emitted soot particles. As a result,  $\sigma_{abs,p}$  is well correlated with EC mass. In contrast, the  
262  $\sigma_{abs,c}$  is the fraction of light absorption gained by the lensing effect of the coating on particles after  
263 emission. The variability of  $\sigma_{abs,c}$  mainly depends on the coating thickness of the soot particles.  
264 Consequently,  $\sigma_{abs,c}$  is independent of EC mass. Since variations of EC and  $\sigma_{abs,c}$  are independent,  
265 the assumed MAE<sub>p</sub> corresponding to the minimum R<sup>2</sup>(EC,  $\sigma_{abs,c}$ ) would then represent the most  
266 statistically probable MAE<sub>p</sub> of the tested dataset.

267 A computer program (Wu, 2017b) in Igor Pro (WaveMetrics, Inc. Lake Oswego, OR, USA)  
268 was developed to facilitate MRS calculation with a user friendly graphical user interface. Another two  
269 Igor Pro based computer programs Histbox (Wu, 2017c) and Scatter Plot (Wu, 2017d) are used for  
270 generating histograms, box plots and scatter plots (with Deming regressions) presented in this study.  
271 Detailed descriptions of these computer programs can be found in the SI and the computer programs  
272 are available from <https://sites.google.com/site/wuchengust>.

### 273 **3.2 Mie simulation**

274 It can be informative to model a single soot particle using Mie theory (Bohren and Huffman,  
275 1983) and understand the theoretical range and variability of the soot particle's optical properties.

276 Three types of mixing state are widely employed for parameterization: internal mixing, external  
277 mixing and core-shell. To better represent the real situation (coating due to the aging process), a core-  
278 shell model is considered in the Mie calculation (Figure S2), which is more realistic than a volume  
279 mixture model (Bond et al., 2006). An aerosol optical closure study in the North China Plain (NCP)  
280 found that the core-shell model can provide better performance than assuming purely internal mixing  
281 and external mixing (Ma et al., 2012). A morphology study using Scanning Transmission X-ray  
282 Microscopy found that core-shell is the dominating mixing state in ambient samples (Moffet et al.,  
283 2016). It should be noted that the core-shell model assumption still has its own limitations. A single  
284 particle soot photometer (SP2) study by Sedlacek et al. (2012) reported a negative lag time between  
285 the scattering and incandescence signals in samples influenced by biomass burning, implying a near  
286 surface location of soot relative to non-absorbing materials. Near surface type mixing of soot has also  
287 been observed in Tokyo, but accounted for only 10% of total mixed soot containing particles (Moteki  
288 et al., 2014). Considering the domination of core-shell type particles in the ambient environment, the  
289 core-shell assumption in our optical model is sufficient to approximate the real situation.

290 As shown in Figure S2, fresh emitted soot particles are chain-like aggregates of small spheres  
291 (30~50 nm). After the aging process, soot particles are coated with organic and inorganic materials.  
292 Sufficient evidence has shown that the coating not only results in particle size growth, but also makes  
293 the soot core become more compact due to its collapse (Alexander et al., 2008; Zhang et al., 2008;  
294 Lewis et al., 2009a), especially under high RH conditions (Leung et al., 2017). Since the spherical like  
295 core and shell favor Mie simulation, both core and shell are considered as spheres in the Mie  
296 calculation.

297 To investigate the spectrum properties of soot particles, 11 wavelengths (370, 405, 470, 520,  
298 532, 550, 590, 660, 781, 880 and 950 nm) are considered in calculations to cover wavelengths in the  
299 most frequently used absorption measurement instruments. A refractive index (RI) of  $1.85 - 0.71i$  is  
300 adopted for soot core (Bond and Bergstrom, 2006) and 1.55 for non-absorbing coating (clear shell) in  
301 the Mie calculation for all wavelengths. Studies suggest a group of organic matter (OM), known as  
302 Brown Carbon (BrC), can absorb solar radiation at UV wavelengths (Kirchstetter et al., 2004). Thus,  
303 a BrC coating (brown shell) scenario is also considered in Mie simulation following the wavelength-

304 dependent RI suggested by Lack and Cappa (2010), which ranges from 1.55-0.059i (370 nm) to 1.55-  
305 0.0005i (950 nm). A modeling study by Bond et al. (2006) indicates that absorption amplification is  
306 not sensitive to the RI, thus the result below is not expected to be sensitive to the RI variability. Due  
307 to the spherical assumption of the BC core, a constant particle density is adopted for simplicity instead  
308 of size dependent particle density. But it is worth noting that in reality, the effective density of soot  
309 varies with particle size due to the morphology change during particle aging (Tavakoli and Olfert,  
310 2014; Dastanpour et al., 2017). Both core diameters ( $D_{\text{core}}$ ) and shell diameters ( $D_{\text{shell}}$ ) are constrained  
311 in the range of 10 ~ 3000 nm in the model simulations. The Mie calculations are implemented with a  
312 customized program (Wu, 2017e) written in Igro Pro (WaveMetrics, Inc. Lake Oswego, OR, USA)  
313 and it is available from <https://sites.google.com/site/wuchengust>. It should be noted that the core-shell  
314 type mixing state of particles is still rare in 3D atmospheric models like WRF-Chem (Matsui et al.,  
315 2013; Nordmann et al., 2014) due to computational cost limitation.

### 316 **3.2.1 Mie modeled absorption angstrom exponent (AAE)**

317 Absorption Angstrom Exponent (AAE) is a widely used parameter that describes the  
318 wavelength dependence of aerosol light absorption (Moosmuller et al., 2011), which can be written  
319 explicitly as

$$320 \quad AAE(\lambda_1, \lambda_2) = -\frac{\ln(\sigma_{abs,\lambda_1}) - \ln(\sigma_{abs,\lambda_2})}{\ln(\lambda_1) - \ln(\lambda_2)} \quad (7)$$

321 It is well known that ambient soot particles exhibit an AAE close to unity (Bond, 2001).  
322 Modeled variability in  $AAE_{470-660}$  of bare soot particles is shown in Figure S3. For soot particles with  
323  $D_{\text{core}} < 200$  nm,  $AAE_{470-660}$  is very close to 1 and decreases significantly for particles with  $D_{\text{core}} > 200$   
324 nm. Considering a typical  $D_{\text{core}}$  of fresh emitted soot particles smaller than 200 nm (Rose et al., 2006;  
325 China et al., 2013), the model results confirm the frequently observed AAE close to 1 from ambient  
326 measurements (Kirchstetter et al., 2004). Modeled variability in  $AAE_{470-660}$  of soot particles coated by  
327 non-absorbing substances (clear shell) and weakly absorbing materials (brown shell) is shown in  
328 Figure 2. Elevated AAE to ~2 is observed in the clear shell scenario (Figure 2a and 3b) for the most  
329 probable soot core particle sizes (<200 nm), which agrees well with a previous model study (Lack and

330 Cappa, 2010), implying that elevated AAE cannot be exclusively attributed to mixing with BrC. AAE  
331 elevation is more pronounced in the brown shell scenario. For soot particles with  $D_{\text{core}} < 200$  nm, brown  
332 shell AAE<sub>470-660</sub> can easily reach 3 for a coating of  $D_{\text{shell}}/D_{\text{core}}=3$  (Figure 2c and 2d). These high AAE  
333 results are consistent with the previous model study (Lack and Cappa, 2010) and could partially  
334 explain the high AAE observed in measurement studies (Kirchstetter et al., 2004; Hoffer et al., 2006),  
335 since the presence of externally mixed BrC particles also contribute to the wavelength dependent light  
336 absorption.

### 337 **3.2.2 Mie modeled single scattering albedo (SSA)**

338 Variability in modeled SSA of soot particles coated by non-absorbing substances and weakly  
339 absorbing materials (e.g. BrC) is shown in Figure S4. For particles with  $D_{\text{core}} < 200$  nm and  $D_{\text{shell}}/D_{\text{core}}$   
340  $< 3$ , the SSA increases gradually (up to  $\sim 0.9$ ) with a thicker coating and behaves similarly between  
341 clear shell and brown shell scenarios.

### 342 **3.2.3 Mie modeled mass absorption efficiency (MAE)**

343 MAE is a useful indicator for soot mixing state. Variability in MAE of bare soot particles as a  
344 function of particle size at a wavelength of 550 nm is illustrated in Figure S5. The magnitude of MAE  
345 is sensitive to the soot density assumption, especially for particles  $< 200$  nm (Figure S5), but the overall  
346 trend of particle size dependency is similar between different density scenarios. MAE peaks at a  
347 particle size of 200 nm and decreases dramatically for larger particles. In our MAE calculation, a soot  
348 density of  $1.9 \text{ g cm}^{-3}$  is adopted, as suggested by Bond and Bergstrom (2006). The purpose of adopting  
349 constant density is to simplify the MAE calculation. It should be noted that the effective density of  
350 soot core is highly variable in ambient environments. For example, a study in Beijing (Zhang et al.,  
351 2016b) found a value of  $1.2 \text{ g cm}^{-3}$ . A recent chamber study found the effective density of soot can  
352 evolve from  $0.43$  to  $1.45 \text{ g cm}^{-3}$  during aging as coated by m-Xylene oxidation products (Guo et al.,  
353 2016). A study by a single-particle aerosol mass spectrometer in Guangzhou found the effective  
354 density of soot increased with particle size in the range of 400 to 1600 nm (Zhang et al., 2016a). The  
355 MAE of coated particles from different core/shell diameter combinations are shown in Figure S6. For

356 thickly coated particles, the MAE in the clear shell scenario varied as  $D_{\text{shell}}/D_{\text{core}}$  increased, but the  
357 MAE of brown shell scenario increased quasi-monotonously with  $D_{\text{shell}}/D_{\text{core}}$ .

### 358 **3.2.4 Mie modeled light absorption enhancement factor ( $E_{\text{abs}}$ )**

359  $E_{\text{abs}}$  is a better indicator for soot mixing state than MAE since it does not rely on the soot density  
360 assumption and is more suitable for comparing Mie simulations with ambient measurements. Modeled  
361 variability in  $E_{\text{abs}}$  of soot particles coated by non-absorbing substances and weakly absorbing materials  
362 (e.g. BrC) is shown in Figure 3a and 3c respectively.  $E_{\text{abs}}$  is not only sensitive to the core/shell diameter  
363 combination, but also behaves very differently on the clear and brown shell assumptions. For the clear  
364 shell scenario, when  $D_{\text{coat}}/D_{\text{core}} < 2$ ,  $E_{\text{abs}}$  does not exceed 2 for particles with different soot core sizes,  
365 but for the same  $D_{\text{coat}}/D_{\text{core}}$ , a larger soot core size yields a higher  $E_{\text{abs}}$  (Figure 3b, cross-sections of  
366 Figure 3a). If  $D_{\text{coat}}/D_{\text{core}} > 2$ ,  $E_{\text{abs}}$  could be 3 to 5 for particles with a soot core smaller than 200 nm, but  
367 for particles with a soot core larger than 200 nm, the  $E_{\text{abs}}$  is limited to  $\sim 2$  as shown in Figure 3b. For  
368 the brown shell scenario,  $E_{\text{abs}}$  increased quasi-monotonically with  $D_{\text{coat}}/D_{\text{core}}$ , and this trend is similar  
369 for different soot core sizes (Figure 3d). **The main reason behind is that in the brown shell scenario,**  
370 **both lensing effect and BrC absorption contribute to  $E_{\text{abs}}$ . As shown in Figure S7, the BrC absorption**  
371 **contribution to total  $E_{\text{abs}}$  strongly depends on coating thickness and is insensitive to soot core diameters.**  
372 **When the coating is relatively thin ( $< 5$  nm for  $\lambda@370$  nm,  $< 15$  nm for  $\lambda@550$  nm and  $< 40$  nm for**  
373  **$\lambda@880$  nm), BrC absorption contribution to the total  $E_{\text{abs}}$  is less than 20%. As the coating increases to**  
374 **a certain level ( $\sim 15$  nm for  $\lambda@370$  nm,  $\sim 35$  nm for  $\lambda@550$  nm and  $\sim 90$  nm for  $\lambda@880$  nm), BrC**  
375 **absorption contribution is comparable to the lensing effect contribution, each contributing  $\sim 50\%$  to**  
376 **the total  $E_{\text{abs}}$ . When the BrC coating is sufficiently thick ( $> 30$  nm for  $\lambda@370$  nm,  $> 90$  nm for  $\lambda@550$**   
377 **nm and  $> 110$  nm for  $\lambda@880$  nm), BrC absorption dominates the  $E_{\text{abs}}$  contribution. As a result, if BrC**  
378 **coating is indeed present in ambient samples, a strong wavelength dependent  $E_{\text{abs}}$  could be observed,**  
379 **since a BrC coating of 30 nm would be enough to induce a large amount of detectable  $E_{\text{abs}}$  in the UV**  
380 **range. Another major difference between the clear and brown shell scenarios is that, for thickly coated**  
381 **particles (e.g.  $D_{\text{coat}}/D_{\text{core}} > 2$ ), the brown shell can yield a much higher  $E_{\text{abs}}$  than the clear shell.**

382 Both primary soot size distribution and coating thickness can affect the absorption  
383 enhancement of ambient BC particles. Ambient measurements by LII found soot particle number and  
384 mass modes peaking at 110 nm and 220 nm, respectively, in the PRD (Huang et al., 2011b). A study  
385 in Shanghai found similar results (70 nm for number concentrations and 200 nm for mass  
386 concentrations)(Gong et al., 2016). Considering that the LII technique is specific for BC mass  
387 determination which is independent of BC mixing state, the size distribution reported by LII can  
388 represent the size distribution of the BC core. A study using a Micro Orifice Uniform Deposit Impactor  
389 (MOUDI) found a EC mass size distribution in the PRD exhibiting three modes peaking at ~300, ~900  
390 and ~5000 nm (Yu et al., 2010), implying a substantial coating of BC particles, and a diameter  
391 amplification of 3. BC sizing by LII is based on volume equivalent diameter (VED), while MOUDI is  
392 based on aerodynamic diameter. As a result, these two techniques do not necessarily yield similar sizes,  
393 even for the bare soot particles. The conversion between these two types of diameters involves the  
394 knowledge of particle density and morphology (drag force). A recent closure study on BC mixing state  
395 in the PRD region suggests  $\sigma_{abs}$  is dominated by coated soot particles in the range of 300~400 nm  
396 (Tan et al., 2016). Considering the dominant BC core distribution measured by SP2 (110 nm), the  
397 upper limit of  $E_{abs}$  in the PRD is roughly estimated as ~2 for the clear shell scenario (Figure 3b).

## 398 4 Results and discussions

### 399 4.1 Annual measurement statistics

400 The frequency distribution (log-normal) of  $\sigma_{abs550}$  is shown in Figure 4a, with an annual average ( $\pm 1$   
401 S.D.) of  $42.65 \pm 30.78 \text{ Mm}^{-1}$ . A log-normal distribution is also found in the EC mass concentration  
402 (Figure 4b), with an annual average of  $2.66 \pm 2.27 \mu\text{g m}^{-3}$ . Figure 4c demonstrates the yearlong  
403 frequency distribution of  $MAE_{550}$  at the NC site. The annual average  $MAE_{550}$  is  $18.75 \pm 6.16 \text{ m}^2 \text{ g}^{-1}$  and  
404 the peak ( $\pm 1$  S.D.) of the lognormal fit is  $15.70 \pm 0.22 \text{ m}^2 \text{ g}^{-1}$ . A good correlation is observed between  
405  $\sigma_{abs}$  and EC mass ( $R^2=0.92$ ) as shown in Figure 4d, and the color coding indicates a MAE dependency  
406 on RH (the RH effect will be discussed in section 4.5). Annual average  $AAE_{470-660}$  is  $1.09 \pm 0.13$  (Figure  
407 S8a), indicating that soot is the dominant absorbing substance in the PRD and the brown shell scenario

408 shown in the Mie simulation is unlikely to be important. Annual mean SSA<sub>525</sub> is 0.86±0.05 (Figure  
 409 S8c), similar to previous studies in the PRD (Jung et al., 2009; Wu et al., 2009). For comparison  
 410 purpose, MAE measured at original wavelength and MAE scaled to 550 nm following the  $\lambda^{-1}$   
 411 assumption are both shown in Table S1. The MAE comparisons discussed below are MAE at 550 nm.  
 412 MAE<sub>550</sub> by previous studies at various locations was found to cover a wide range, from 5.9 to 61.6 m<sup>2</sup>  
 413 g<sup>-1</sup>. Annual average observed MAE<sub>550</sub> at NC (18.75 m<sup>2</sup> g<sup>-1</sup>) is higher than many studies shown in Figure  
 414 5, e.g., Shenzhen (Lan et al., 2013), Beijing (Yang et al., 2009), Mexico city (Doran et al., 2007) and  
 415 Fresno (Chow et al., 2009).

416 As shown in Figure 1, the annual average MAE<sub>p,550</sub> estimated by MRS is 13 m<sup>2</sup> g<sup>-1</sup>. MAE<sub>p</sub> by  
 417 MRS represents the MAE<sub>p</sub> at the emission source, which is different from the MAE<sub>p</sub> by the TD  
 418 approach for two reasons. First, the morphology of thermally denuded BC particles (compact  
 419 aggregates) is different from that of freshly emitted BC particles (chain-like aggregates). Second, most  
 420 of the coatings are removed for TD denuded soot particles, but freshly emitted soot particles usually  
 421 come with a thin coating of OC formed from condensation of OC vapors as the temperature drops from  
 422 the flame to the ambient air. As a result, the MRS-derived MAE<sub>p</sub> is expected to be higher than the  
 423 MAE<sub>p</sub> by the TD approach. The estimated MAE<sub>p,550</sub> is higher than a previous study in Guangzhou  
 424 (7.44 m<sup>2</sup> g<sup>-1</sup>) (Andreae et al., 2008), but comparable to Xi'an (11.34 m<sup>2</sup> g<sup>-1</sup>) (Wang et al., 2014) and  
 425 Toronto (9.53~12.57 m<sup>2</sup> g<sup>-1</sup>) (Knox et al., 2009). The annual average E<sub>abs550</sub> by MRS following Eq. 3  
 426 is estimated to be 1.50±0.48 (mean ± 1 S.D.).

427 It should be noted that the E<sub>abs</sub> estimation approach demonstrated here is insensitive to the MAE  
 428 bias (e.g. overestimation of  $\sigma_{abs}$  and variability of EC mass by different TOA protocols) discussed in  
 429 section 2.1, because bias in EC mass or  $\sigma_{abs}$  is cancelled out in the E<sub>abs</sub> calculation (Eq. 3), since E<sub>abs</sub>  
 430 is the ratio of  $\sigma_{abs,t}$  to  $\sigma_{abs,p}$ . To investigate the performance of the MRS approach in response to  
 431 systematic bias in EC and  $\sigma_{abs}$ , two simple tests are conducted as shown in Figures S9 and S10 by  
 432 adding systematic biases to  $\sigma_{abs550}$  and EC. Test A represents a situation when  $\sigma_{abs}$  is overestimated  
 433 and EC is underestimated. The biased data are marked as  $\sigma'_{abs550}$  and EC' respectively, as shown  
 434 below:

$$435 \quad \sigma'_{abs550} = \sigma_{abs550} \times 2 \quad (8)$$

$$436 \quad EC' = EC \times 0.7 \quad (9)$$



437 As a result, the average  $MAE_{550}$  changed from 18.75 to 53.58  $m^2 g^{-1}$  and  $MAE_p$  changed from 13 to 37  
438  $m^2 g^{-1}$  (Figure S9). However,  $E_{abs}$  by ratio of averages remain the same (1.44).

439 In Test B, EC by different TOA protocols are compared to investigate the effect of different EC  
440 determination approaches while  $\sigma_{abs550}$  remains unchanged. EC by IMPROVE TOR protocol is  
441 calculated from NIOSH TOT EC following an empirical formula for suburban sites derived from a 3-  
442 year OCEC dataset in PRD (Wu et al., 2016a):

$$443 \quad EC_{IMP\_TOR} = 2.63 \times EC_{NSH\_TOT} + 0.05 \quad (10)$$

444 As shown in Figure S10,  $MAE_{550}$  changed from 18.75 to 7.02  $m^2 g^{-1}$  and  $MAE_p$  changed from 13 to 5  
445  $m^2 g^{-1}$ , but  $E_{abs}$  remain almost the same (1.40). Result of Test B implies that although EC is  
446 operationally defined, the discrepancy of EC between TOA protocols did not weaken the role of EC  
447 serving as a tracer for primary emissions in MRS application. These examples demonstrate that  
448 systematic biases in  $\sigma_{abs550}$  and EC have no effects on  $E_{abs}$  estimation by the MRS approach.

449 As mentioned in section 1, the definition of  $MAE_p$  by the TD approach is different from the  
450  $MAE_p$  of emission source. The TD  $MAE_p$  is expected to be slightly lower than the  $MAE_p$  of emission  
451 source. Therefore, the corresponding  $E_{abs}$  are slightly different and it should be cautioned when  
452 comparing MRS-derived  $E_{abs}$  with  $E_{abs}$  by the TD approach and Mie simulations. The  $E_{abs}$  could vary  
453 by location, depending on the coating thickness and size distribution of the primary aerosols. After  
454 undergoing atmospheric aging, the  $E_{abs}$  can be increased during transport from emission source to rural  
455 areas. The magnitude of the  $E_{abs}$  found at the NC site is comparable to other locations such as Boulder  
456 (Lack et al., 2012a) (1.38), London (Liu et al., 2015) (1.4), Shenzhen (Lan et al., 2013) (1.3), Yuncheng  
457 (Cui et al., 2016b) (2.25), Jinan (Chen et al., 2017) (2.07) and Nanjing (Cui et al., 2016a) (1.6) and is  
458 higher than studies in California (Cappa et al., 2012) (1.06), as listed in Table 3. Spectrum  $E_{abs}$  are  
459 calculated from 370 to 950 nm as shown in Figure S11.  $E_{abs}$  in the PRD exhibits a weak wavelength  
460 dependence, with slightly higher  $E_{abs}$  at the shorter wavelength (e.g.  $E_{abs370} = 1.55$ ) and is relatively  
461 lower in the IR range (e.g.  $E_{abs950} = 1.49$ ).

## 462 **4.2 Monthly characteristics of MAE, AAE and SSA**

463 Monthly variations of  $MAE_{550}$  at the NC site are shown in Figure 6a and Table S2, revealing distinct  
464 patterns of higher  $MAE_{550}$  in summer and lower in winter. On the other hand,  $AAE_{470-660}$  is lower in  
465 summer and higher in winter (Figure 6b and Table S3). Monthly  $SSA_{525}$  varied from 0.83 to 0.90  
466 without a clear seasonal pattern, as shown in Figure S12 and Table S4.  $MAE_{p,550}$  estimation for  
467 individual months is shown in Figure 6a (the purple line) and monthly  $E_{abs550}$  is calculated accordingly  
468 following Eq. 3 (Figure 6c).  $E_{abs550}$  shows clear seasonal variations, with higher values from April to  
469 August (1.52~1.97 as shown in Table S5) and relatively lower values from September to March  
470 (1.24~1.49). The highest enhancement is found in August (1.97). Factors affecting variation of  $E_{abs550}$   
471 are discussed in the following sections, including air mass origin, biomass burning and RH.

## 472 **4.3 The effect of air mass origin**

473 It's of interest to understand the seasonal variations of optical properties in the PRD. Hourly backward  
474 trajectories for the past 72 hours were calculated using NOAA's HYSPLIT (Hybrid Single Particle  
475 Lagrangian Integrated Trajectory, version 4) model (Draxier and Hess, 1998) from Feb 2012 to Jan  
476 2013 as shown in Figure S13. Cluster analysis was conducted using MeteoInfo (Wang, 2014). By  
477 examining the total spatial variance (TSV), the number of clusters was determined to be four as shown  
478 in Figure S14. Cluster 1 (C1) represents continental air masses from the north, accounting for 44.4%  
479 of total trajectories. C2 (22.8%) represents marine air masses coming from the South China Sea. C3  
480 represents air masses from the east (Taiwan island). C4 (15.8%) represents transitional air masses  
481 coming from the east coastline of China. As shown in Figure 7,  $E_{abs550}$  from C2 (1.78) is higher than  
482 other clusters (1.30 – 1.42). Further Wilcoxon-Mann-Whitney tests show that  $E_{abs550}$  from C2 is  
483 significantly higher than  $E_{abs550}$  from C1, C3 and C4 (Figure S15), implying that particles from the  
484 South China Sea cluster is likely more aged than other clusters. Air mass origin in the PRD is  
485 dominated by C2 from Apr to Aug (Figure S16a) as a result of the South China Sea monsoon in the  
486 rainy season. In contrast, the dry season is ruled by continental air masses from the north (C1) due to  
487 the influence of the northeast monsoon.  $E_{abs550}$  from C2 varied from 1.67 to 2.19, but was always

488 higher than  $E_{\text{abs}550}$  from C1 and C3 during the rainy season (Figure S16b). As a result, the domination  
489 of aged air mass from the vast ocean is one of the reasons for the much higher  $E_{\text{abs}550}$  found in the rainy  
490 season.

#### 491 **4.4 The effect of biomass burning**

492 Biomass burning (BB) and vehicular emission are the two major sources of soot particles. BC  
493 from biomass burning emission, depending on the fuel type and burning condition, may have a higher  
494 OC/EC ratio and a thicker coating, resulting in a higher MAE than vehicular emission (Shen et al.,  
495 2013; Cheng et al., 2016). In this study, the influence of BB on optical properties is investigated using  
496 the  $K^+$ /EC ratio as a BB indicator. As shown in Figure 8,  $MAE_{550}$  is positively correlated with the  
497  $K^+$ /EC ratio, which exhibits a clear seasonal pattern that is higher in the rainy season and lower in the  
498 dry season (Figure S17a). Southeast Asia has the highest fire emission density globally due to the high  
499 biofuel consumption along with frequent fire activity in this region (Aouizerats et al., 2015), making  
500 Southeast Asia a large contributor to BC emissions (Jason Blake, 2014). During the rainy season when  
501 oceanic wind prevails, BC from BB emission in Southeast Asia can reach PRD through long range  
502 transport (LRT), resulting in an elevated  $K^+$ /EC ratio and  $MAE_{550}$ . The Deming regression intercept  
503 (11.89) in Figure 8 represents the MAE without the BB effect. This non-BB  $MAE_{550}$  ( $11.89 \text{ m}^2 \text{ g}^{-1}$ ) is  
504 only slightly lower than  $MAE_{p,550}$  ( $13 \text{ m}^2 \text{ g}^{-1}$ ) obtained in section 4.3, implying that a large fraction of  
505  $MAE_{p,550}$  could not be explained by the BB source. Additional evidence was obtained through  
506 examining regression relationships of  $MAE_{p,550}$  with  $K^+$ /EC month-by-month (Figure S17b).  
507 Correlation of monthly  $MAE_{p,550}$  vs.  $K^+$ /EC ratio yield a  $R^2$  of 0.23 (Figure S17c). In contrast, a much  
508 higher correlation ( $R^2=0.58$ ) was observed (Figure S17d) between  $MAE_{p,550}$  and non-BB  $MAE_{550}$  (i.e.,  
509  $K^+$ /EC intercepts from Figure S17b). These results imply that BB is one of the contributors to the  
510  $MAE_{p,550}$  variations, but unlikely the dominating one.

511 Many studies have found that BB influenced samples exhibit elevated AAE due to the presence  
512 of wavelength dependent light absorbing substances like BrC and HUmic-Like Substances (HULIS)  
513 (Kirchstetter et al., 2004; Hoffer et al., 2006; Sandradewi et al., 2008; Herich et al., 2011; Pokhrel et  
514 al., 2017). It is of interest to investigate whether elevated AAE observed in the PRD during the dry

515 season is associated with BB influence. As shown in Figure S18,  $AAE_{370-470}$  and  $AAE_{470-660}$  did not  
516 correlate with the BB indicator,  $K^+/EC$  ratio. These results suggest that the elevated AAE observed in  
517 the PRD wintertime is unlikely to be dominated by the BB effect. Beside the independency between  
518  $AAE_{470-660}$  and  $K^+/EC$  ratio, the measured  $AAE_{470-660}$  range also implies that BB is not the major  
519 driving force of  $AAE_{470-660}$  variations. The limited light absorption contribution from BrC in RPD  
520 region is observed in a recent study (Yuan et al., 2016) , which suggest an upper limit of BrC  
521 contribution of 10% at 405 nm in the winter time using the AAE approach. As discussed in our Mie  
522 simulation (section 3.1) and a previous study (Lack and Cappa, 2010), coating of non-absorbing  
523 materials onto soot particles can increase AAE up to 2. Since the monthly average  $AAE_{470-660}$  in  
524 wintertime did not exceed 1.2 (Table S3), the variations of  $AAE_{470-660}$  in the PRD are more likely  
525 associated with coatings rather than the contribution of BrC. The results also imply that attempts on  
526 BrC absorption attribution for the PRD dataset presented in this study could be risky, considering that  
527 elevation of AAE is actually dominated by coating (Lack and Langridge, 2013).

#### 528 **4.5 The effect of relative humidity (RH) on optical properties**

529 Soot particles are relatively hydrophobic when freshly emitted, but tend to gain hygroscopicity  
530 during atmospheric aging. Hygroscopic growth of coated laboratory generated model BC was reported  
531 by McMeeking et al. (2011). Growth of ambient BC particle size by a factor of 1.4-1.6 under high RH  
532 has been observed in a UK study (Liu et al., 2013). Located in the subtropical zone, RH plays an  
533 important role on aerosol optical properties in the PRD region. The yearlong measurements at the NC  
534 site provide a unique opportunity to investigate the effect of RH on aerosol optical properties, since  
535 most existing ad hoc studies in the PRD only last for months. Liquid water content (LWC) was  
536 calculated using the E-AIM (model 2) thermodynamic model (Clegg et al., 1998). As shown in Figure  
537 S19, LWC on average accounted for a significant fraction (44%) of non-EC  $PM_{2.5}$  mass, making it an  
538 important component of  $PM_{2.5}$  mass and due to high RH in the PRD. Previously, hygroscopic growth  
539 was only considered for particle scattering in the IMPROVE formula for chemically resolved light  
540 extinction budget studies. In this study  $f(RH)$  of MAE was obtained from yearlong measurements as  
541 shown in Figure 9a for  $RH = 30 \sim 100\%$  and color coded for LWC. It clearly shows that  $MAE_{550}$

542 measured in NC is positively correlated with RH and the enhancement can be fitted by a polynomial  
543 equation. When RH is close to 100%, the LWC can account for 70% of PM<sub>2.5</sub> mass. The maximum  
544  $f(\text{RH})$  can reach 1.3, which is higher than the value found in Beijing (1.2) (Wu et al., 2016b), **but lower**  
545 **than a numerical study (1.35) (Nessler et al., 2005)**. These results reveal that a large contribution of  
546  $E_{\text{abs}}$  is coming from high LWC under high RH in the PRD region. Because RH has a clear diurnal  
547 pattern, it can affect the diurnal pattern of  $E_{\text{abs}}$  in the PRD. Since the RH effect on  $E_{\text{abs}}$  is rarely  
548 considered in existing climate models, the inclusion of RH effect can reduce the uncertainty for  
549 assessing BC's climate effect.

550 The  $\text{AAE}_{470-660}$  dependency on RH is shown in Figure 9b. When RH is low (e.g. 30%), the  
551  $\text{AAE}_{470-660}$  is around 1.25 and decreases to 1.10 as RH increases to 50%.  $\text{AAE}_{470-660}$  remains around  
552 1.12 when RH is 50-70%. Then  $\text{AAE}_{470-660}$  decreases again when RH is higher than 70% and can reach  
553 1 when RH is close to 100%. Since a higher RH results in hygroscopic growth and larger particle  
554 diameters, the negative correlation between  $\text{AAE}_{470-660}$  and RH provides a clue on soot particles'  
555 primary diameter and mixing state. As shown in the Mie simulation in Figure 2b, for a particle with  
556  $D_{\text{core}}$  of 130 nm and  $D_{\text{shell}}/D_{\text{core}}$  of 2 to 4,  $\text{AAE}_{470-660}$  decreases as the coating increases, and the decrease  
557 tapers off when  $D_{\text{shell}}/D_{\text{core}} = 3$ . The  $D_{\text{core}}$  obtained here (130nm) is comparable with  $D_{\text{core}}$  obtained  
558 from SP2 measurements (110nm) in the PRD (Huang et al., 2011a).

## 559 **4.6 Implications for mixing state**

560 Quantitative direct measurements of BC mixing state and coating thickness are still challenging.  
561 SP2 can estimate the coating thickness using a lag-time approach or a Mie calculation approach can  
562 be employed, but both methods have a limited range in coating thickness and uncertainties arise from  
563 the assumptions made during the retrieval. For example, recent studies found that the mass equivalent  
564 diameter of soot core measured by SP2 could be underestimated due to density assumptions (Zhang et  
565 al., 2016b). Although size distribution measurement is not available in this study, clues of mixing state  
566 still can be derived from bulk measurements of optical properties. As discussed in section 4.4.1,  
567 elevated  $E_{\text{abs}550}$  observed in the rainy season is associated with aged air masses from a marine origin.  
568 To probe the possible mixing state difference between dry and rainy season,  $E_{\text{abs}550}$ ,  $\text{SSA}_{525}$  and

569 AAE<sub>470-660</sub> are used to narrow down the possible core-shell size range as shown in Figure S20. Monthly  
 570 averages with one standard deviation of AAE<sub>470-660</sub>, SSA<sub>525</sub> and E<sub>abs550</sub> are used as constraints to extract  
 571 the intersecting core-shell size range from Figure 2a, Figure S4 and Figure 3a. **January and August**  
 572 **data are used to represent two different scenarios: elevated AAE<sub>470-660</sub> (1.19±0.11) with lower E<sub>abs550</sub>**  
 573 **(1.31±0.32) in dry season and low AAE<sub>470-660</sub> (1.04±0.09) with elevated E<sub>abs550</sub> (1.97±0.71) in rainy**  
 574 **season.** The results show that January and August have a very different core-shell size range: in  
 575 January, the core and shell range are 100 ~ 160 nm and 120 ~ 250 nm, respectively; in August, the  
 576 core and shell range are 120 ~ 165 nm and 170 ~ 430 nm, respectively. This confirms again that the  
 577 soot particles in the rainy season are likely to have a thicker coating than in the dry season.

## 578 **5 Caveats of the MRS method in its applications to ambient data**

579 The data in this study is dominated by BC absorption that did not show much influence from  
 580 BrC. However, extra care should be taken if the samples exhibit substantial BrC signature (e.g.  
 581 AAE>2). Such situations are equivalent to the two-source scenarios discussed in our previous paper  
 582 on the MRS method (Wu and Yu, 2016) and the major findings are described below. Two types of  
 583 two-source scenarios are considered: two correlated primary sources (scenario A) and two independent  
 584 primary sources (scenario B). In scenario A in which both BC and primary BrC are dominated by BB,  
 585 using BC as a solo tracer to calculate the primary ratio (MAE<sub>p</sub>) still works. In scenario B in which BC  
 586 and primary BrC are independent, using BC alone to determine a single primary MAE<sub>p</sub> could lead to  
 587 a considerable bias in E<sub>abs</sub> estimation. Alternatively, if a reliable primary BrC tracer is available, the  
 588 corresponding MAE<sub>p,BrC</sub> can be determined by MRS. With the knowledge of MAE<sub>p,BrC</sub> and MAE<sub>p,BC</sub>, light  
 589 absorption by BC and BrC can be calculated separately and the E<sub>abs</sub> can be determined using Eq. (11) :

$$590 \quad E_{abs} = \frac{\sigma_{abs,t}}{\sigma_{abs,p,BC} + \sigma_{abs,p,BrC}} = \frac{\sigma_{abs,t}}{MAE_{p,BC} \times EC + MAE_{p,BrC} \times BrC} \quad (11)$$

591 However, the implementation of Eq.11 is challenging due to the complexity in the chemical  
 592 composition of BrC. For example, a recent study found that the 20 most absorbing BrC chromophores  
 593 account for ~50% BrC light absorption and there is not a single compound contributing more than 10%  
 594 (Lin et al., 2016), making it difficult to choose a single compound as the BrC tracer. In addition, time

595 resolved measurement of BrC chromophores has yet to emerge. As a result, for scenario B (sample  
596  $AAE > 2$  & primary BrC variations independent of BC), estimation of  $E_{abs}$  by MRS is not practical at  
597 this stage due to the lack of required input data. Using BC alone to determine a single primary  $MAE_p$   
598 could lead to a considerable bias and should be avoided.

## 599 **6 Conclusions**

600 In this study, a novel statistical approach is proposed and its application on ambient data is  
601 demonstrated using one-year hourly OC and EC data coupled with Aethalometer measurements.  
602 Unlike conventional  $E_{abs}$  determination approaches that require expensive instrumentation (e.g. TD-  
603 PAS, VTDMA, SP2), this new approach employs widely deployed instruments (field carbon analyzer  
604 and Aethalometer). The key of this new approach involves calculating  $MAE_p$  by the Minimum R  
605 Squared (MRS) method (Wu and Yu, 2016). It is found that  $E_{abs}$  estimation by MRS is insensitive to  
606 systematic biases in EC and  $\sigma_{abs}$  measurements. The annual average  $MAE_{p,550}$  estimated by MRS is  
607  $13 \text{ m}^2 \text{ g}^{-1}$  and annual average  $MAE_{550}$  is  $18.75 \pm 6.16 \text{ m}^2 \text{ g}^{-1}$ , suggesting an annual average enhancement  
608 factor ( $E_{abs550}$ ) of  $1.50 \pm 0.48$ . This value is within the upper limit of  $E_{abs}$  ( $\sim 2$ ) by core-shell Mie  
609 simulations considering the typical soot size distribution and coating thickness in the PRD.

610 Both  $MAE_{p,550}$  and  $E_{abs}$  show distinct seasonal variations, implying the complexity of soot  
611 particle mixing state variations in this region. The elevated summertime  $E_{abs550}$  in the PRD is found to  
612 be associated with the domination of aged air masses from the South China Sea, along with the long-  
613 range transport of biomass burning influenced air masses from Southeast Asia. Hygroscopic growth  
614 with elevated RH contributes to  $E_{abs}$  as well, which could be as high as 1.3. A negative correlation is  
615 found between  $AAE_{470-660}$  and RH, suggesting a dominant particle size with a  $D_{core}$  of 130 nm and  
616  $D_{shell}/D_{core}$  range of 2 to 4. Core-shell size ranges narrowed down by  $E_{abs550}$  and  $AAE_{470-660}$  constraints  
617 suggest that soot particles in the rainy season are likely to have thicker coatings than in the dry season.

## 618 **Data availability**

619 OC, EC, inorganic ions and  $\sigma_{abs}$  data used in this study are available from corresponding authors  
620 upon request.

621

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1013 Table 1. Abbreviations.

1014

Abbreviation	Definition
AAE <sub>470-660</sub>	Absorption Angstrom Exponent between 470 and 660 nm
BB	Biomass burning
BrC	Brown Carbon
D <sub>core</sub> , D <sub>shell</sub>	Particle diameter of core/shell
E <sub>abs550</sub>	Light absorption enhancement factor at 550 nm
$\sigma_{abs550}$	Light absorption coefficient at 550 nm
$\sigma_{abs,t}$	Total light absorption coefficient of a coated particle
$\sigma_{abs,p}$	Primary light absorption coefficient attributed to the soot core alone of a coated particle
$\sigma_{abs,c}$	Extra light absorption coefficient due to the lensing effect of coating on the soot core
LII	Laser induced incandescence technique for soot measurement
LWC	Liquid water content
MAE <sub>550</sub>	Mass absorption efficiency at 550 nm, also known as mass absorption cross-section (MAC)
MAE <sub>p,550</sub>	Primary MAE of freshly emitted soot particles at 550 nm
MAAP	Multi Angle Absorption Photometer
MOUDI	Micro Orifice Uniform Deposit Impactor
MRS	Minimum R squared method
PAS	Photo acoustic spectrometer
PRD	Pearl River Delta region, China
SP2	Single particle soot photometer
SSA	Single scattering albedo
TD	Thermal denuder
TOA	Thermal optical analysis
TSV	Total spatial variance in backward trajectories cluster analysis

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1016 Table 2. Comparison of MRS application on  $(OC/EC)_p$  (for SOC estimation) and  $MAE_p$  (for  $E_{abs}$  estimation).  
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	MRS in EC tracer method for SOC estimation (Wu and Yu, 2016)	MRS in EC tracer method for $E_{abs}$ estimation (this study)
Key parameter of fresh EC particles to be determined	$\left(\frac{OC}{EC}\right)_p = \frac{POC}{EC}$	$MAE_p = \frac{O\sigma_{abs,p}}{EC}$
Input quantities for MRS from measurements	OC, EC (tracer)	$\sigma_{abs,t}$ , EC (tracer)
Variable to be decoupled by the tracer	$OC = POC + SOC$ $= \left(\frac{OC}{EC}\right)_p \times EC + SOC$	$\sigma_{abs,t} = \sigma_{abs,p} + \sigma_{abs,c}$ $= \left(\frac{O\sigma_{abs,t}C}{EC}\right)_p \times EC + \sigma_{abs,c}$
Ambient measurement at its closest to fresh emissions	Minimum $R^2$ (SOC, EC) $SOC = OC - \left(\frac{OC}{EC}\right)_p \times EC$	Minimum $R^2$ ( $\sigma_{abs,c}$ , EC) $\sigma_{abs,c} = \sigma_{abs,t} - MAE_p \times EC$
Graph	<p>Minimum <math>R^2</math> <math>(OC/EC)_p = 2.26</math></p>	<p>Minimum <math>R^2</math> <math>(OC/EC)_p = 13</math></p>

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1019 Table 3. Comparison of  $E_{\text{abs}}$  between various studies.

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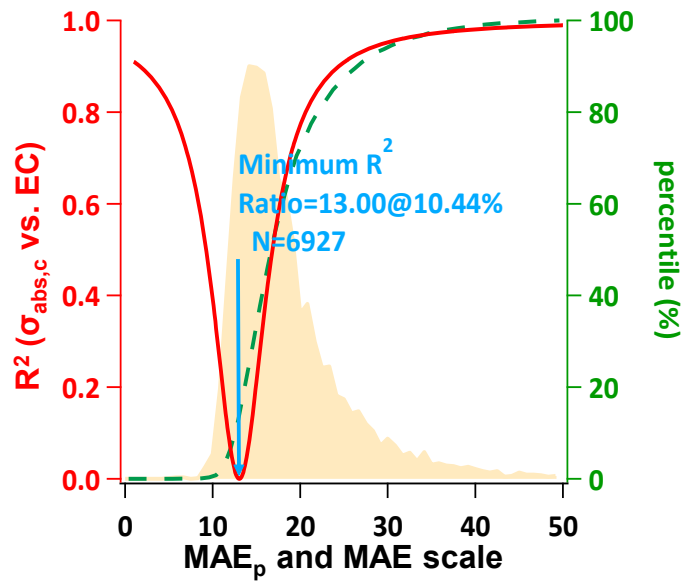
Location	Type	Sampling Duration	$\lambda$ (nm)	Instrument	$E_{\text{abs}}$	Method	Reference
Guangzhou, China	Suburban	2012.2-2013.1	550	AE+OCEC	1.50±0.48	MAE	This study
Xi'an, China	Urban	2012.12-2013.1	870	PAS	1.8	MAE	(Wang et al., 2014)
Shenzhen, China	Urban	2011.8-9	532	PAS	1.3	MAE	(Lan et al., 2013)
Jinan, China	Urban	2014.2	678	OCEC	2.07 ± 0.72	AFD	(Chen et al., 2017)
Nanjing, China	Suburban	2012.11	532	PAS	1.6	MAE	(Cui et al., 2016a)
Boulder, USA	Forest fire	2010.9	532	PAS	1.38	TD 200°C	(Lack et al., 2012a)
London, UK	Rural	2012.2	781	PAS	1.4	TD 250°C	(Liu et al., 2015)
California, USA	Rural	2010.6	532	PAS	1.06	TD 250°C	(Cappa et al., 2012)
Noto Peninsula, Japan	Rural	2013.4-5	781	PAS	1.22	TD 300°C	(Ueda et al., 2016)
Yuncheng, China	Rural	2014.6-7	678	OCEC	2.25 ± 0.55	AFD	(Cui et al., 2016b)
San Jose, Costa Rica	Rural	2006 winter	1064	SP2	1.3	Mie+SP2	(Schwarz et al., 2008)

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AE: Aethalometer ; OCEC: OCEC analyzer; PAS: photo acoustic spectrometer; SP2: Single particle soot photometer; TD: Thermal denuder AFD: filter filtration-dissolution

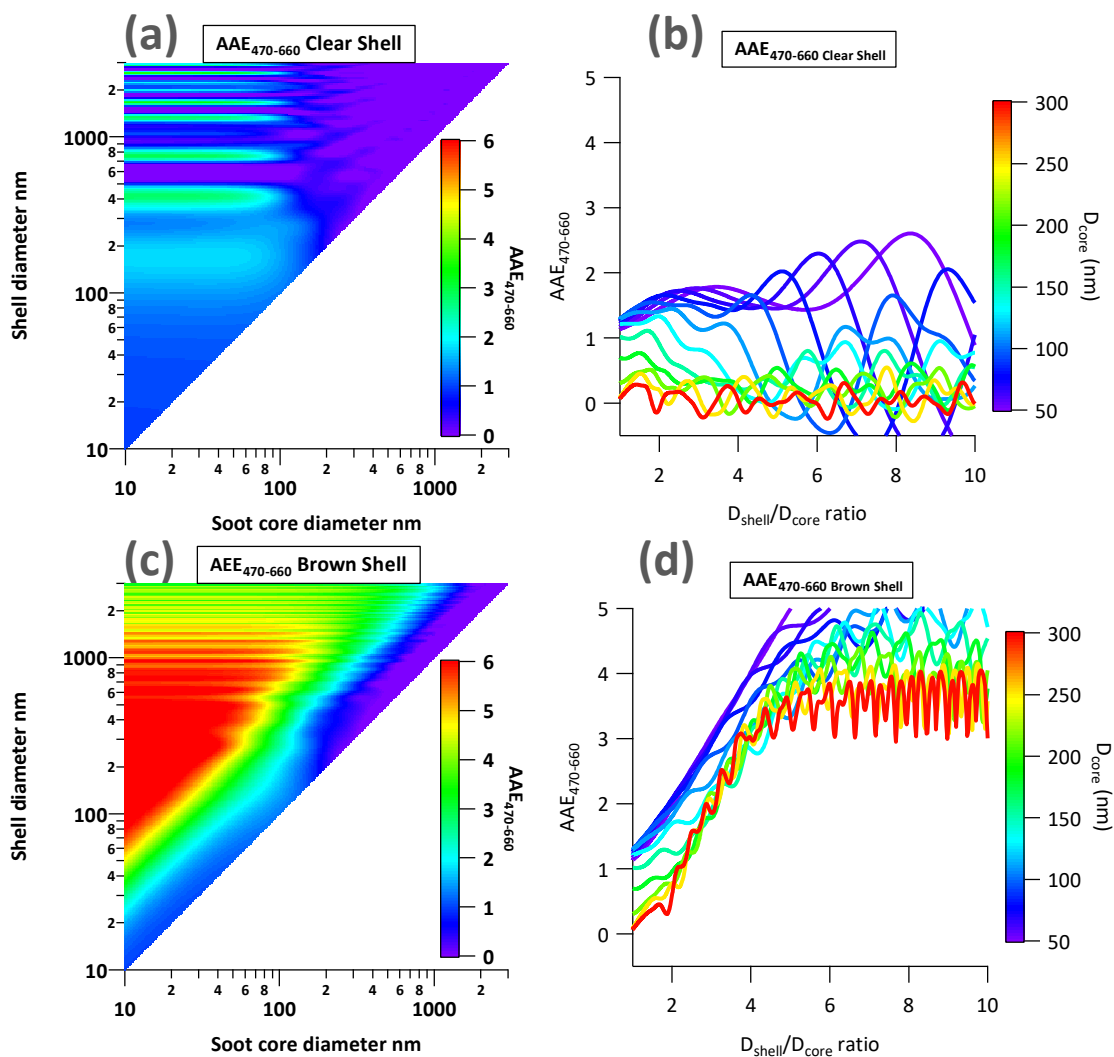
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1025 Figure 1. Minimum R squared (MRS) plot for calculating  $MAE_p$  at 550 nm. The red curve is the correlation result between  
 1026  $\sigma_{abs,c}$  ( $\sigma_{abs,t} - EC * MAE_p$ ) and EC mass. The shaded area in light tan represents the frequency distribution of observed  
 1027 MAE. The dashed green line is the cumulative distribution of observed MAE.



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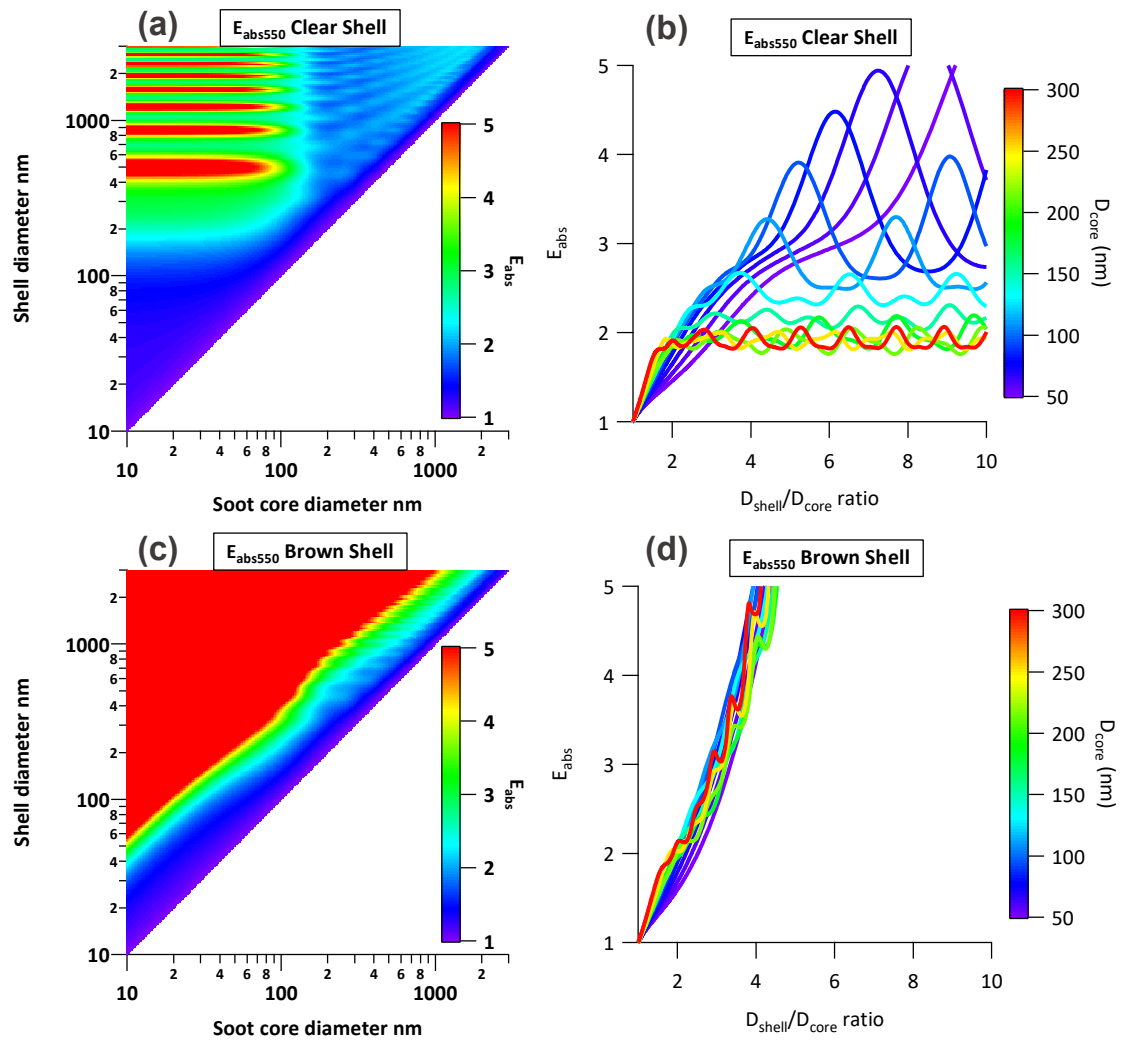
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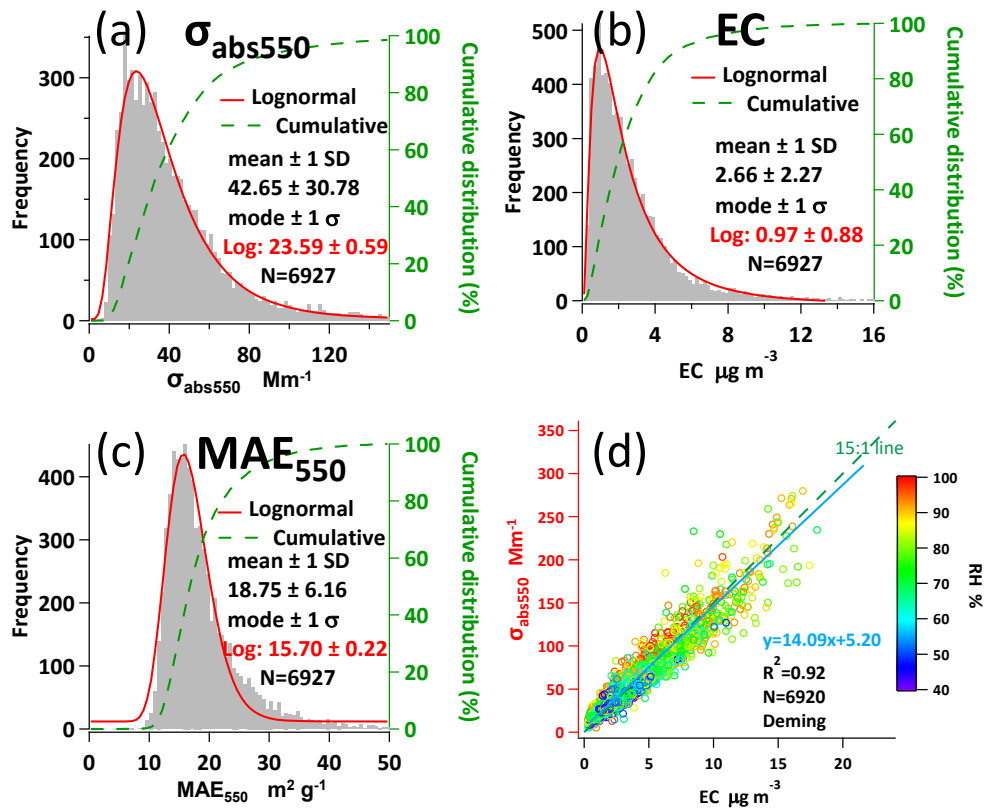
Figure 2. Mie simulated size dependency of soot particles AAE<sub>470-660</sub>. (a) Combination of different clear shell (y axis) and core diameters (x axis). The color coding represents the AAE<sub>470-660</sub> of a particle with specific core and clear shell size; (b) Cross-sections views of (a). The color coding represents different D<sub>core</sub> in the range of 50 ~ 300 nm. (c)&(d) Similar to (a)&(b) but from the brown shell scenario.



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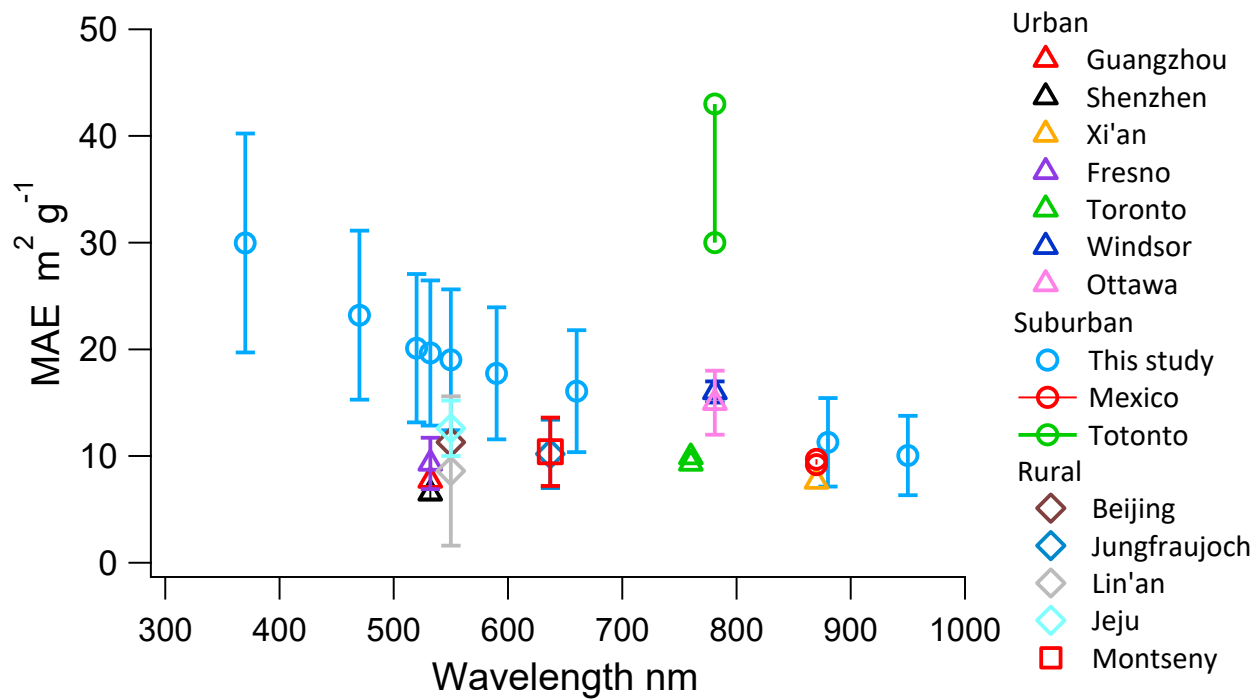
Figure 3. Mie simulated size dependency of soot particles  $E_{abs}$  at wavelength 550 nm. (a) Combination of different clear shell (y axis) and core diameters (x axis). The color coding represents the  $E_{abs}$  of a particle with specific core and clear shell size; (b) Cross-sections views of (a). The color coding represents different  $D_{core}$  in the range of 50 – 300 nm. (c)&(d) Similar to (a)&(b) but from the brown shell scenario.





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1045 Figure 4. Measured annual statistics of  $\sigma_{abs550}$ , EC and MAE<sub>550</sub>. (a) Annual frequency distribution of light absorption at  
 1046 550 nm. The red curve represents the fitting line for a log-normal distribution. (b) Annual frequency distribution of EC  
 1047 mass concentration (c) Frequency distribution of Mass absorption efficiency (MAE) at 550 nm. (d) Scatter plot of light  
 1048 absorption (550 nm) and EC mass. The slope represents MAE<sub>550</sub>. The blue regression line is by Deming regression. The  
 1049 color coding represents RH.  
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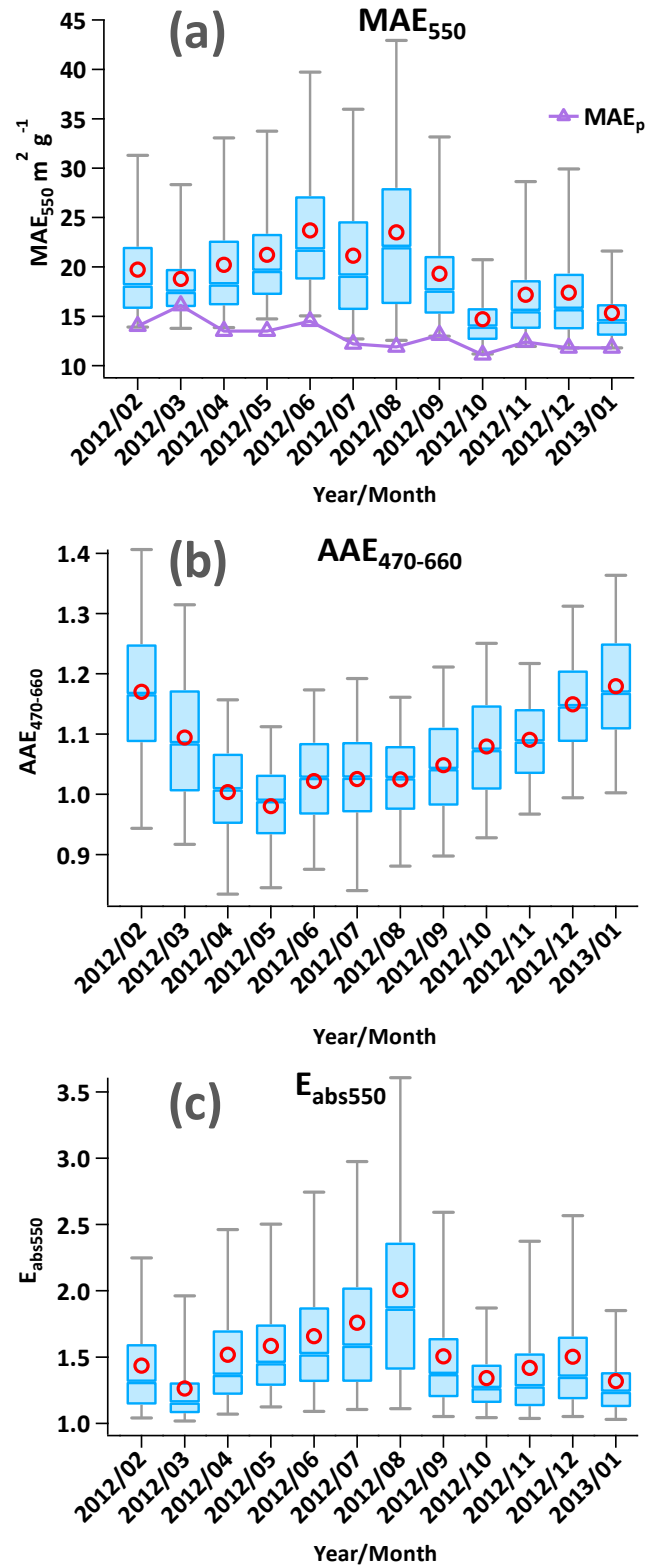
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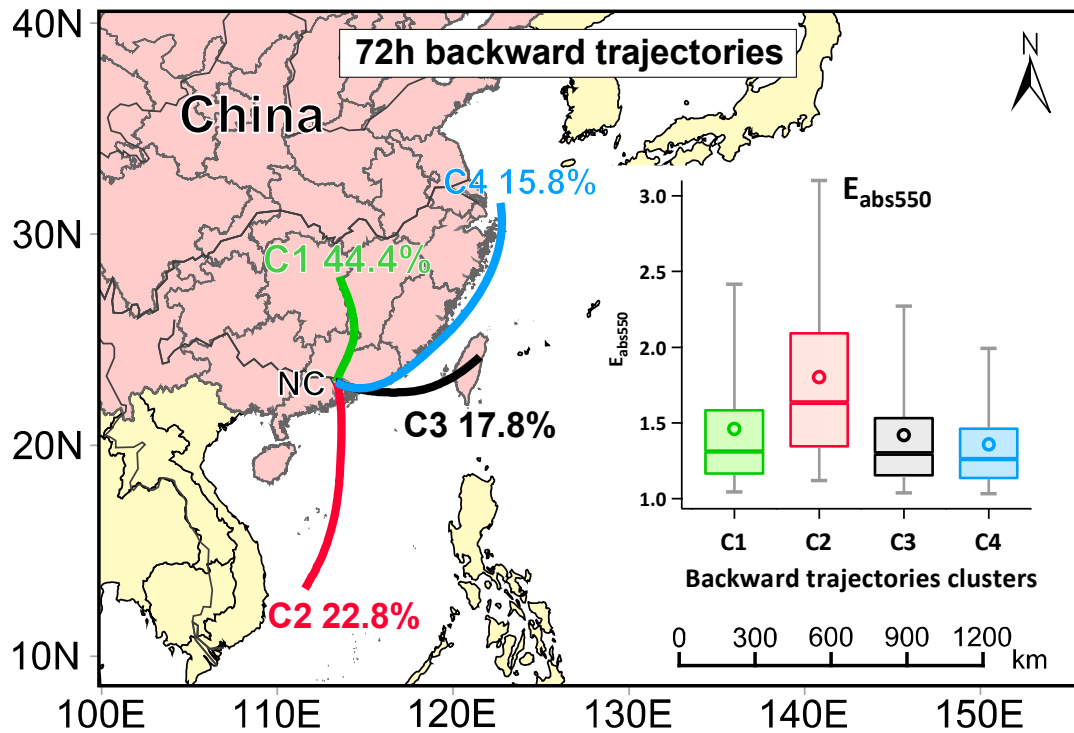
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Figure 5. Comparison of spectral MAE measurements from this study with previous studies. Triangle, circle and rhombus represent urban, suburban and rural respectively. Details and reference can be found in Table S1. The whiskers represent one standard deviation.



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1056 Figure 6. Measured monthly variations of (a) MAE<sub>550</sub>, the purple line represents MAE<sub>p</sub> estimated by MRS (b) AAE<sub>470-660</sub>  
 1057 and (c) E<sub>abs550</sub>. Red circles represent the monthly average. The line inside the box indicates the monthly median. Upper and  
 1058 lower boundaries of the box represent the 75<sup>th</sup> and the 25<sup>th</sup> percentiles; the whiskers above and below each box represent  
 1059 the 95<sup>th</sup> and 5<sup>th</sup> percentiles.

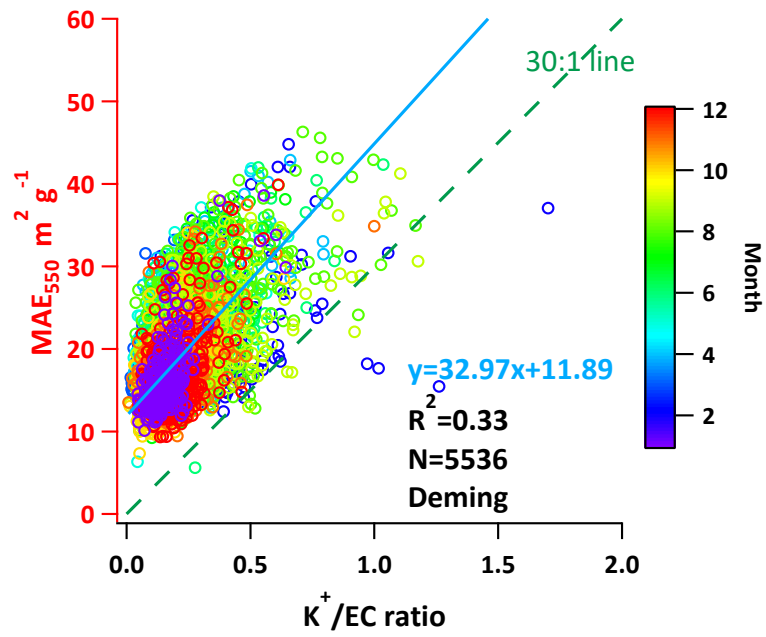


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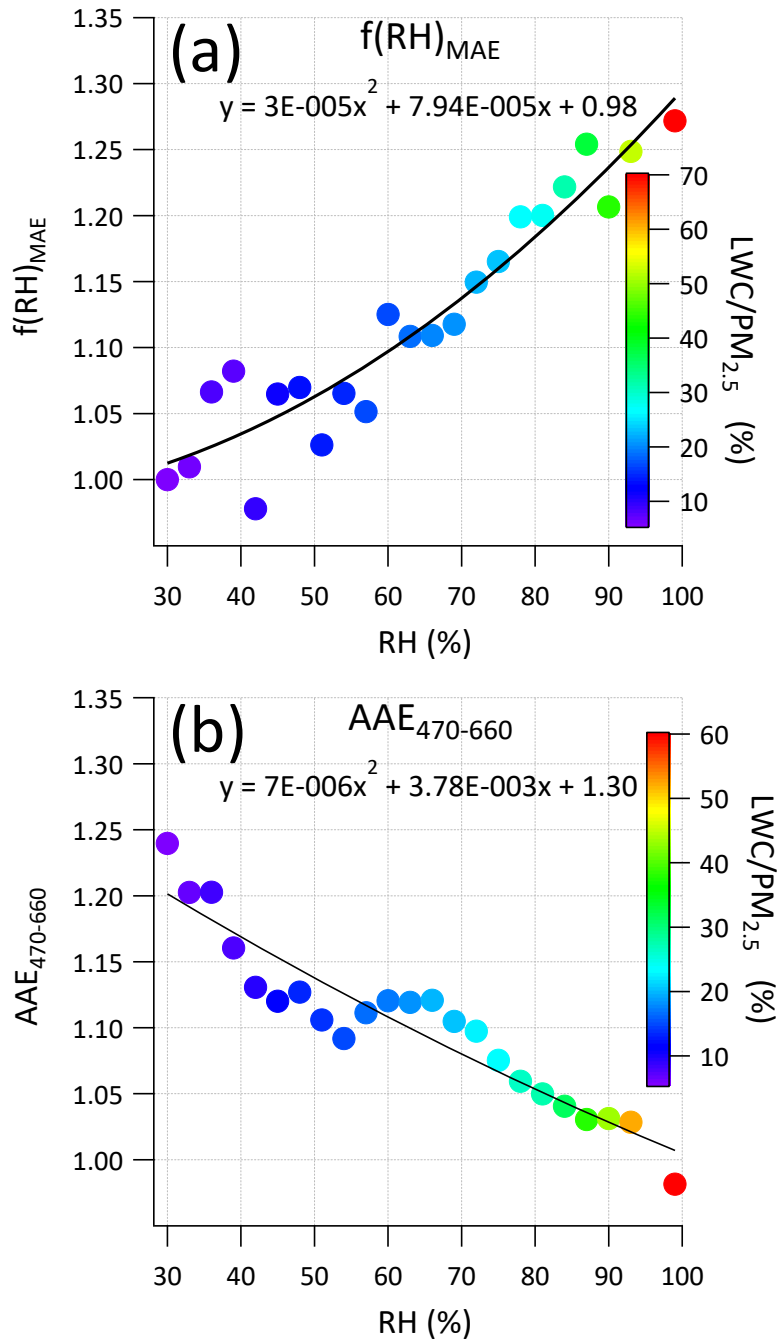
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Figure 7. Average backward trajectories arriving at 100 m at NC site for four clusters (2012 Feb - 2013 Jan).  $E_{abs550}$  by different clusters are shown in the box plot.



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1064 Figure 8. MAE<sub>550</sub> dependency on biomass burning indicator  $K^+/EC$  ratio. The color coding represents months. The intercept  
 1065 represents MAE without biomass burning effect. The 30:1 line serves as a reference line with an integer slope that  
 1066 is close to the regressed slope through the origin.



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Figure 9. Optical properties dependency on RH determined from one year's sampling data at NC site. (a) Hygroscopic growth factor ( $f(RH)$ ) of EC MAE<sub>550</sub> (b)  $AAE_{470-660}$  as a function of RH.

1 *Supplement of*

2 **Quantifying black carbon light absorption enhancement by**  
3 **a novel statistical approach**

4

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23 This SI contains five tables and twenty-five figures.

24

## 25 **Uncertainty of $E_{abs}$ estimation**

26 The uncertainty of  $E_{abs}$  estimation depends on uncertainty propagation from MAE uncertainty,  
27 which can be calculated from (Harris, 2010):

$$28 \quad MAE_{Unc} = MAE \times \sqrt{\left(\frac{\sigma_{abs,Unc}}{\sigma_{abs}}\right)^2 + \left(\frac{EC_{Unc}}{EC}\right)^2} \quad S1$$

$$29 \quad E_{abs,Unc} = E_{abs} \times \sqrt{\left(\frac{MAE_{Unc}}{MAE}\right)^2 + \left(\frac{MAE_{p,Unc}}{MAE_p}\right)^2} \quad S2$$

30

## 31 **Descriptions of customized programs used in this study for data analysis and** 32 **visualization**

33 Several computer programs were developed to meet specific research purpose in this study. All  
34 the programs are based on Igor Pro ([www.wavemetrics.com](http://www.wavemetrics.com)) that provides a friendly GUI. Brief  
35 descriptions are given below.

36

### 37 **MRS program (Igor Pro based)**

38 The program (Figure S21) is written in Igor Pro (WaveMetrics, Inc. Lake Oswego, OR, USA)  
39 to feasible MRS calculation via a user-friendly GUI. The MRS application is not limited in  
40 SOC estimation, but can also be extended to other applications (e.g.  $E_{abs}$  estimation) as long as  
41 a reliable tracer is available.

42 MRS calculation can be done by different temporal cycles (batch calculation): by year, by  
43 year&season, by season, by year&month, by month, by year&month&hour. Data filter is also  
44 available to calculate MRS on a specific subset of data.

45 The program is available from <https://sites.google.com/site/wuchengust>.

46

### 47 **Mie program and source code written in Igor Pro**

48 A computer program (Figure S22) written in Igor Pro (WaveMetrics, Inc. Lake Oswego, OR,  
49 USA) for Mie scattering calculation. Both BHMIE and BHCOAT (coated particles)  
50 algorithms(Bohren and Huffman, 1983) are included. The program is also capable of batch  
51 calculation for both algorithms. Available from <https://sites.google.com/site/wuchengust>.



52

### 53 **Aethalometer data processing program (Igor Pro based)**

54 This handy tool (Figure S23) can perform different corrections (e.g. Weingartner, Virkkula) on  
55 Aethalometer data. Raw Aethalometer data suffers from several artifacts including filter matrix  
56 effect (multiple scattering), loading effect (shadowing) and scattering effect. Careful  
57 corrections are needed for reporting light absorption coefficient from attenuation measurement.  
58 This Igor based program can directly import Aethalometer raw data and perform corrections  
59 (algorithm can be selected by user). Results can be exported to .csv files. Extra information  
60 including statistics of sensor voltage from each channel, sampling flow rate, etc are plotted for  
61 a quick QA/QC check. Available from <https://sites.google.com/site/wuchengust>.

62

### 63 **Histbox program (Igor Pro based)**

64 A handy tool (Figure S24) to generate histogram and box plots with many powerful features.  
65 Data can be sorted by different time scale and batch plotting is available. Available from  
66 <https://sites.google.com/site/wuchengust>.

67

### 68 **Scatter plot program**

69 Scatter plot (Figure S25) is a handy tool to maximize the efficiency of data visualization in  
70 atmospheric science. The program includes Deming, WODR and York algorithm for linear  
71 regression, which consider uncertainties in both X and Y, that is more realistic for atmospheric  
72 applications. It is Igor based, and packed with lots of useful features for data analysis and graph  
73 plotting, including batch plotting, data masking via GUI, color coding in Z axis, data filtering  
74 and grouping. Available from <https://sites.google.com/site/wuchengust>.

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**Table S1.** Comparison of Mass absorption efficiency (MAE) at various locations. For literature MAE values at different wavelengths rather than 550 nm, an estimated MAE<sub>550</sub> is given in the brackets following equations given by Moosmuller et al. (2011) assuming AAE of 1.

Location	Type	Sampling Duration	Inlet	$\lambda$ (nm)	$\sigma_{\text{abs}}$ Instrument	EC determination protocol	$\sigma_{\text{abs}} \pm 1$ S.D. (Mm <sup>-1</sup> )	EC mass ( $\mu\text{g m}^{-3}$ )	estimated MAE <sub>p</sub> * (m <sup>2</sup> g <sup>-1</sup> )	observed MAE (m <sup>2</sup> g <sup>-1</sup> )		Reference
										arithmetic mean $\pm 1$ S.D.	Gaussian fit	
Guangzhou, China	Suburban	2012.2-2013.1	PM <sub>2.5</sub>	550	AE	NIOSH_TOT	42.65±29.41	2.66±2.27	13*	18.75±6.16	16.16	This study
Shenzhen, China	Urban	2011.8-9	PM <sub>2.5</sub>	532	PAS	LII	25.4±19.0	4.0±3.1	/	6.5±0.5[6.29±0.48]	/	(Lan et al., 2013)
Xi'an, China	Urban	2012.12-2013.1	PM <sub>2.5</sub>	870	PAS	LII	/	8.8±7.3	7.17[11.34]	/	7.62[12.05]	(Wang et al., 2014)
Xi'an, China	Urban	2013.2	PM <sub>2.5</sub>	532	PAS	LII	/	/	/	14.6±5.6	12.7	(Wang et al., 2017)
Guangzhou, China	Urban	2004.10	PM <sub>2.5</sub>	532	PAS	NIOSH_TOT	91±60	7.1	7.7[7.44]	/	/	(Andreae et al., 2008)
Fresno, USA	Urban	2005.8-9	PM <sub>2.5</sub>	532	PAS	IMPROVE_A_TOR	5.06	1.01	/	6.1±2.5[5.9±2.42]	/	(Chow et al., 2009)
						NIOSH_TOT	/	0.58	/	9.3±2.4[8.99±2.32]	/	
T1, Mexico city, Mexico	Suburban	2006.3	PM <sub>2.5</sub>	870	PAS	NIOSH_TOT	/	/	/	9.2~9.7***[14.55~15.34]	/	(Doran et al., 2007)
Tokyo, Japan	Suburban	2005.8	PM <sub>2.5</sub>	565	PSAP	IMPROVE_A_TOR	30.43±20.41	2.9±2.13	11±1	/	/	(Naoue et al., 2009)
Pasadena, USA	Urban	2010.5-6	PM <sub>2.5</sub>	532	AM	NIOSH_TOT	3.8±3.4	0.6~0.7	5.7[5.51]	/	/	(Thompson et al., 2012)
Toronto, Canada	Urban	2006.12-2007.1	PM <sub>2.5</sub>	760	PAS	NIOSH_TOT	/	/	6.9~9.1** [9.53~12.57]	9.3~9.9[12.85~13.68]	/	(Knox et al., 2009)
Toronto, Canada	Suburban						3~6	0.10~0.14	/	30~43[42.6~61.06]	/	
Windsor, Canada	Urban	2007.8	PM <sub>2.5</sub>	781	PAS	LII	4.4±2.9	0.27±0.23	/	16±1[22.72±1.42]	/	(Chan et al., 2011)
Ottawa, Canada	Urban						26±17	1.7±0.9	/	15±3[21.3±4.26]	/	
Beijing, China	Rural	2005.3	/	550	AE	NIOSH_TOT	/	/	9.5	11.3	/	(Yang et al., 2009)
Montseny, Spin (Mediterranean)	Rural	2009.11-2010.10	PM <sub>10</sub>	637	MAAP	NIOSH_TOT	2.8±2.2	0.271±0.215	/	10.4[12.04]	/	(Pandolfi et al., 2011)
Jungfrauoch, Switzerland	Rural (high alpine)	2007.2-3	/	637	MAAP	LII	/	/	/	10.2±3.2[11.81±3.71]	/	(Liu et al., 2010)
Lin'an, China	Rural	1999.11	PM <sub>2.5</sub>	550	PSAP	NIOSH_TOT	23±14	3.4±1.7	/	8.6±7.0	/	(Xu et al., 2002)
Jeju Island, Korea	Coastal Rural, (East China Sea)	2001.4	PM <sub>10</sub>	550	PSAP	NIOSH_TOT	/	/	/	12.6±2.6	/	(Chuang et al., 2003)
Maldives	Oceanic rural	1999.2-3	PM <sub>3</sub>	550	PSAP	EGA	62±34	2.5±1.4	6.6	8.1	/	(Mayol-Bracero et al., 2002)

\*Determined by Minimum R Squared method; \*\* Median values;

AE: Aethalometer ; PAS photo acoustic spectrometer; MAAP: Multi Angle Absorption Photometer; PSAP: particle soot absorption photometer; AM: albedo meter; LII: Laser induced incandescence

**Table S2.** Statistics of monthly MAE<sub>550</sub>.

<b>Month</b>	<b>95th</b>	<b>75th</b>	<b>50th</b>	<b>25th</b>	<b>5th</b>	<b>Mean</b>	<b>Max</b>	<b>Min</b>	<b>S.D.</b>	<b>N</b>
Feb-2012	31.24	22.00	18.12	15.74	13.92	19.66	47.73	11.74	5.66	529
Mar-2012	26.51	19.63	17.45	15.91	13.71	18.46	45.56	10.98	4.30	651
Apr-2012	33.06	22.66	18.24	16.11	13.85	20.21	48.29	6.01	6.23	595
May-2012	33.24	23.25	19.59	17.16	14.82	21.07	46.66	6.33	5.62	528
Jun-2012	35.52	25.86	21.28	18.57	14.99	22.95	49.07	5.62	6.66	315
Jul-2012	33.93	23.77	18.81	15.58	12.71	20.51	49.22	9.23	6.79	587
Aug-2012	40.75	27.72	21.85	16.14	12.51	23.09	49.95	9.75	8.56	545
Sep-2012	30.75	20.86	17.52	15.24	12.97	18.99	46.44	10.39	5.63	674
Oct-2012	20.72	15.84	13.95	12.60	11.18	14.70	34.09	7.34	3.21	715
Nov-2012	26.45	18.10	15.43	13.70	11.89	16.75	39.32	8.34	4.72	495
Dec-2012	28.57	19.04	15.73	13.66	11.78	17.18	47.39	9.33	5.47	585
Jan-2013	21.53	16.24	14.47	13.03	11.80	15.29	43.19	7.16	3.77	708

**Table S3.** Statistics of monthly AAE<sub>470-660</sub>.

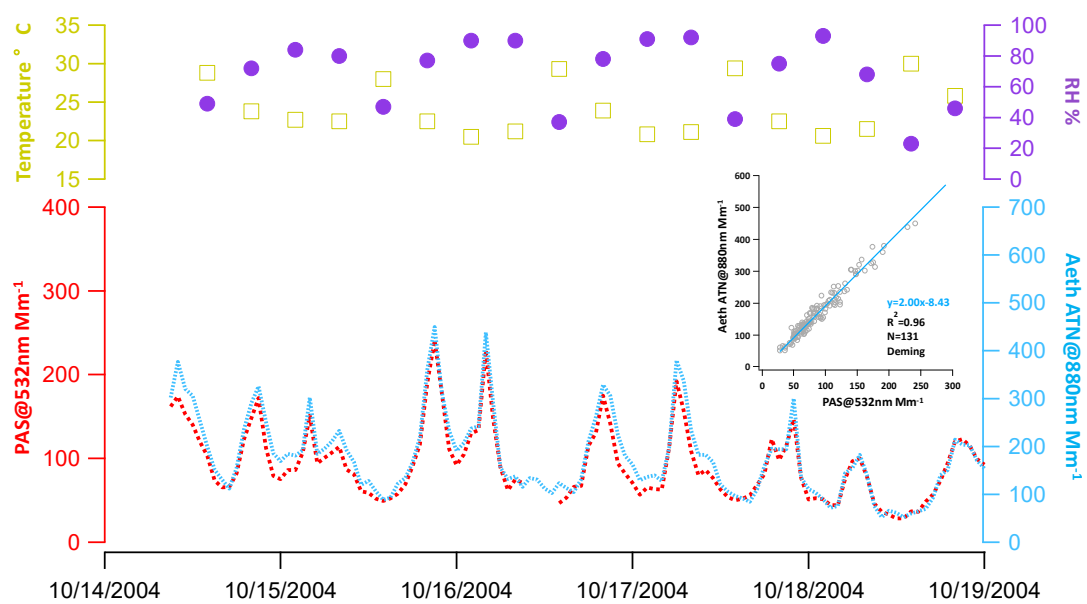
<b>Month</b>	<b>95th</b>	<b>75th</b>	<b>50th</b>	<b>25th</b>	<b>5th</b>	<b>Mean</b>	<b>Max</b>	<b>Min</b>	<b>S.D.</b>	<b>N</b>
Feb-2012	1.42	1.26	1.18	1.11	0.96	1.19	1.72	0.86	0.14	529
Mar-2012	1.33	1.18	1.10	1.02	0.93	1.10	1.50	0.65	0.12	651
Apr-2012	1.19	1.08	1.02	0.95	0.78	1.01	1.76	0.15	0.14	595
May-2012	1.13	1.05	1.00	0.94	0.84	0.99	1.24	0.39	0.10	528
Jun-2012	1.18	1.10	1.04	0.97	0.90	1.04	1.29	0.78	0.09	315
Jul-2012	1.22	1.11	1.04	0.98	0.83	1.04	1.43	0.20	0.13	587
Aug-2012	1.18	1.10	1.04	0.99	0.90	1.04	1.31	0.69	0.09	545
Sep-2012	1.23	1.13	1.06	1.00	0.91	1.07	1.40	0.64	0.11	674
Oct-2012	1.27	1.16	1.09	1.02	0.93	1.09	1.40	0.85	0.10	715
Nov-2012	1.24	1.16	1.11	1.05	0.97	1.11	1.52	0.79	0.08	495
Dec-2012	1.33	1.22	1.16	1.10	1.01	1.16	1.42	0.77	0.09	585
Jan-2013	1.38	1.27	1.18	1.12	1.01	1.19	1.66	0.93	0.11	708

**Table S4.** Statistics of monthly SSA<sub>525</sub>.

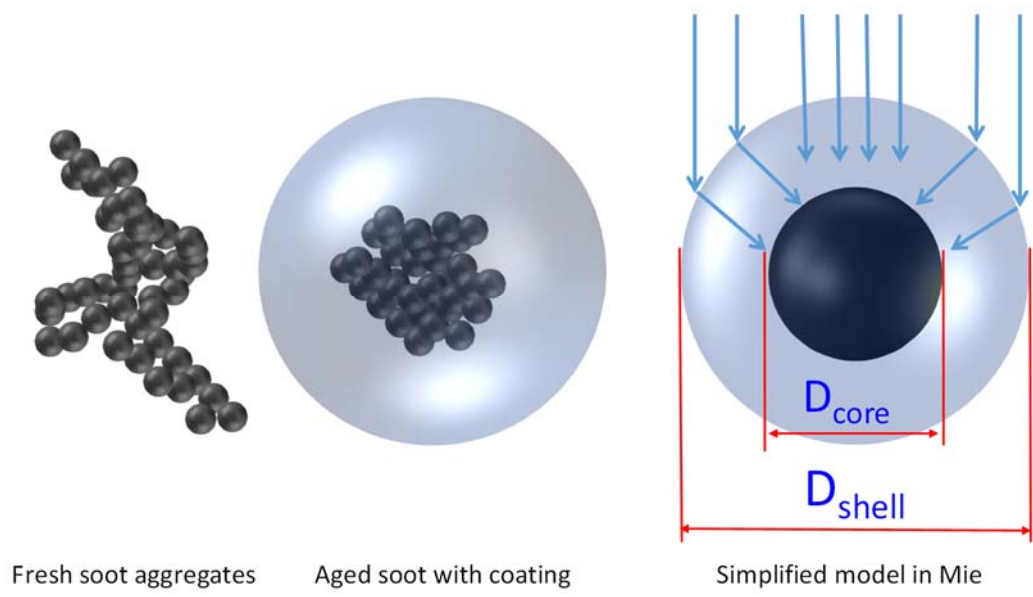
<b>Month</b>	<b>95th</b>	<b>75th</b>	<b>50th</b>	<b>25th</b>	<b>5th</b>	<b>Mean</b>	<b>Max</b>	<b>Min</b>	<b>S.D.</b>	<b>N</b>
Feb-2012	0.91	0.89	0.87	0.84	0.79	0.86	0.94	0.65	0.04	526
Mar-2012	0.91	0.89	0.86	0.83	0.77	0.85	0.95	0.42	0.05	648
Apr-2012	0.92	0.89	0.86	0.83	0.76	0.85	0.94	0.45	0.06	552
May-2012	0.92	0.90	0.87	0.83	0.74	0.85	0.94	0.45	0.06	527
Jun-2012	0.92	0.89	0.86	0.81	0.74	0.84	0.95	0.64	0.06	310
Jul-2012	0.91	0.87	0.83	0.79	0.71	0.83	0.95	0.57	0.06	580
Aug-2012	0.94	0.92	0.89	0.85	0.79	0.88	0.96	0.67	0.05	536
Sep-2012	0.93	0.91	0.88	0.84	0.75	0.87	0.96	0.55	0.06	672
Oct-2012	0.94	0.93	0.91	0.89	0.84	0.90	0.96	0.66	0.03	715
Nov-2012	0.91	0.90	0.87	0.83	0.75	0.86	0.94	0.18	0.06	495
Dec-2012	0.91	0.89	0.86	0.82	0.74	0.85	0.94	0.66	0.05	585
Jan-2013	0.91	0.89	0.87	0.85	0.80	0.86	0.93	0.64	0.04	708

**Table S5.** Statistics of monthly E<sub>abs550</sub>.

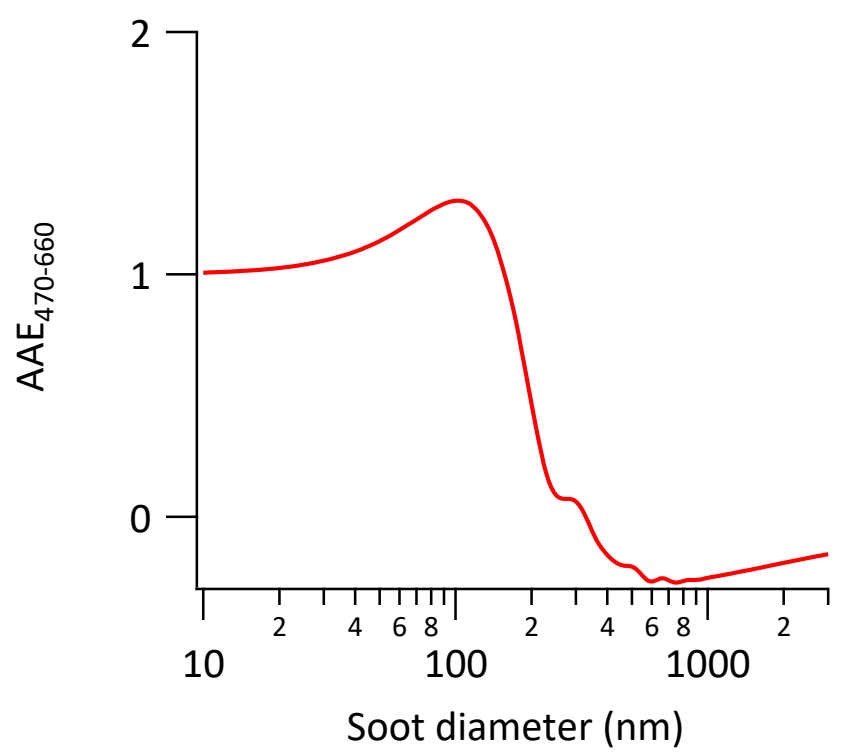
<b>Month</b>	<b>95th</b>	<b>75th</b>	<b>50th</b>	<b>25th</b>	<b>5th</b>	<b>Mean</b>	<b>Max</b>	<b>Min</b>	<b>S.D.</b>	<b>N</b>
Feb-2012	2.23	1.59	1.31	1.14	1.04	1.43	3.41	1.00	0.40	501
Mar-2012	1.76	1.30	1.15	1.07	1.02	1.24	2.83	1.00	0.26	466
Apr-2012	2.46	1.70	1.37	1.21	1.07	1.52	3.58	1.00	0.45	576
May-2012	2.48	1.73	1.45	1.28	1.12	1.57	3.46	1.02	0.41	520
Jun-2012	2.47	1.80	1.49	1.30	1.09	1.61	3.38	1.01	0.45	305
Jul-2012	2.83	1.97	1.57	1.30	1.10	1.71	4.03	1.01	0.55	568
Aug-2012	3.45	2.34	1.86	1.40	1.11	1.97	4.20	1.00	0.71	528
Sep-2012	2.36	1.62	1.36	1.20	1.05	1.48	3.54	1.00	0.42	636
Oct-2012	1.87	1.44	1.27	1.15	1.04	1.34	3.07	1.00	0.28	683
Nov-2012	2.14	1.51	1.27	1.13	1.04	1.38	3.17	1.00	0.38	461
Dec-2012	2.46	1.64	1.35	1.18	1.05	1.49	4.02	1.00	0.46	555
Jan-2013	1.85	1.39	1.24	1.12	1.03	1.31	3.66	1.00	0.32	672



**Figure S1.** Comparison of collocated Aethalometer and PAS at Guangzhou (Oct 2004). Both PAS and Aethalometer (AE-16) were equipped with PM<sub>2.5</sub> inlets. RH of the sampled air was controlled to be <45% for PAS. Aethalometer sampling was conducted without RH control.

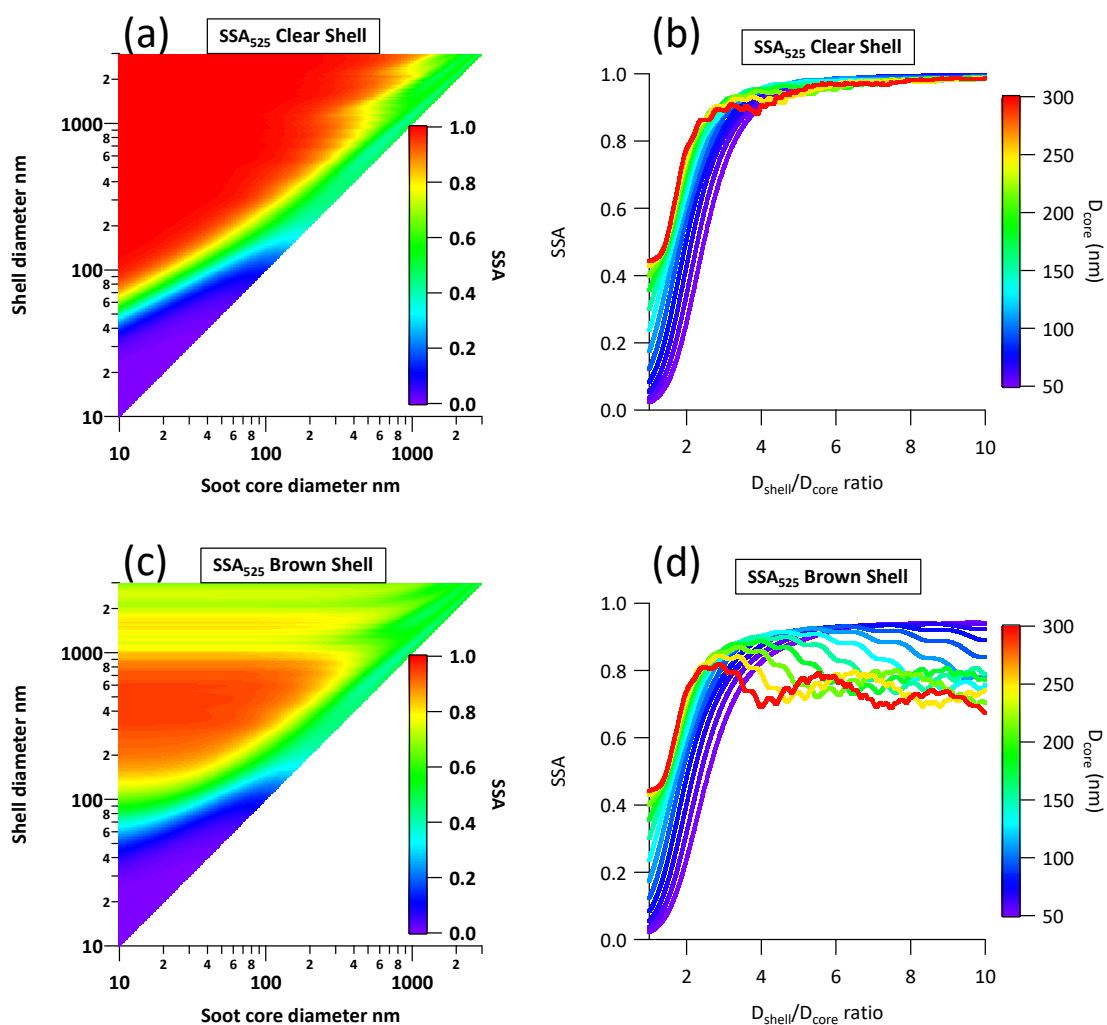


**Figure S2.** Schematic of the aging effect on light absorption. More light is absorbed by the soot particle core due to the lensing effect of the coating materials.

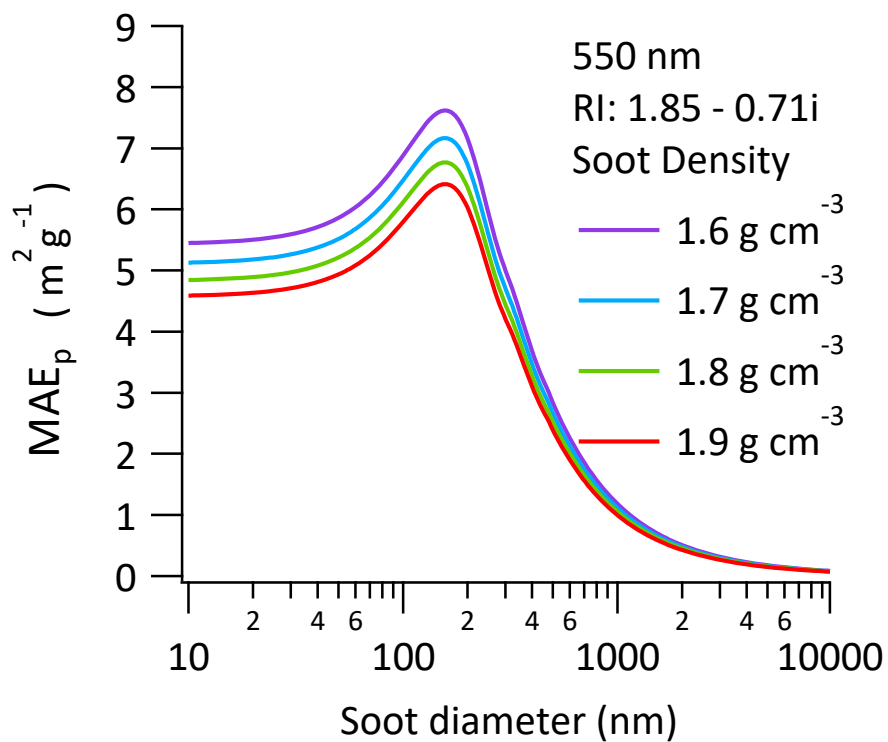


**Figure S3.** Mie simulated AAE<sub>470-660</sub> of a bare soot particle as a function of diameter with a Refractive index of 1.85 – 0.71i.

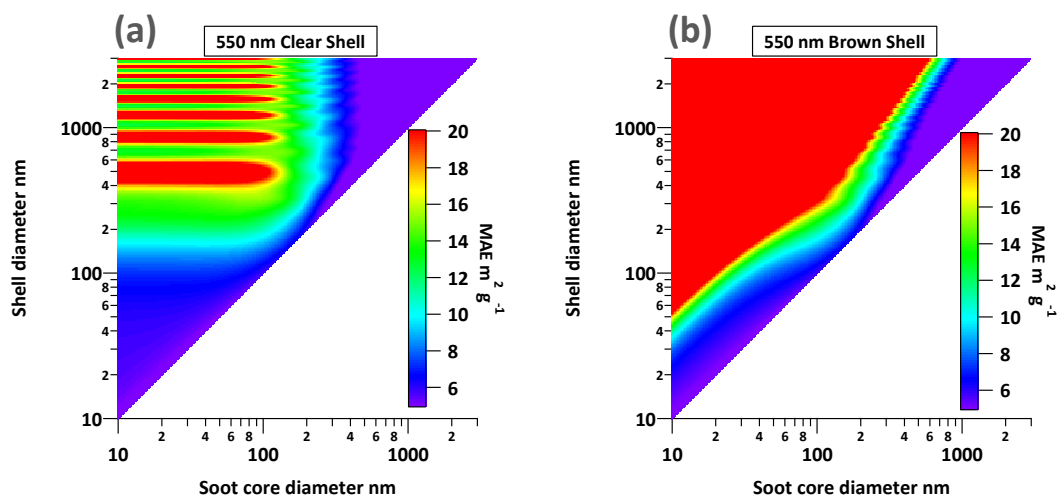




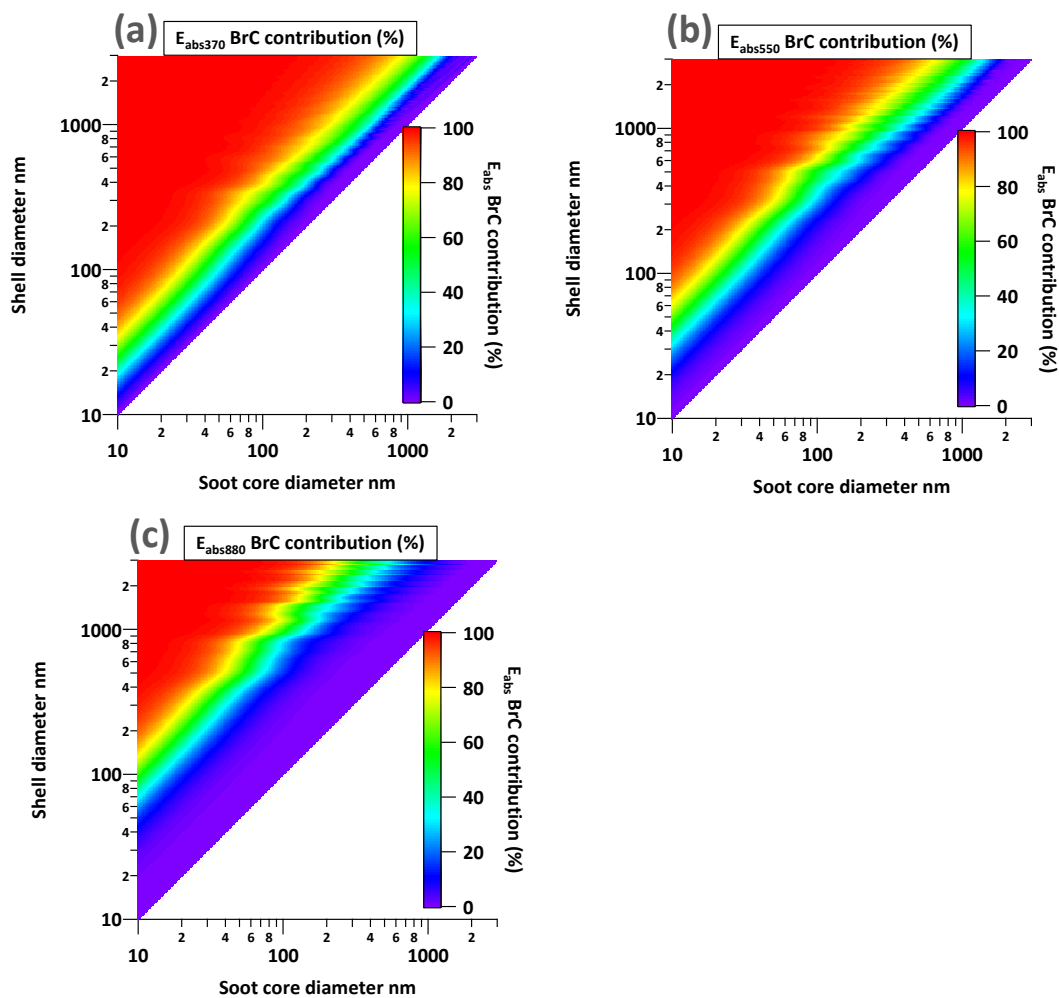
**Figure S4.** Mie simulated size dependency of soot particles SSA at wavelength 525 nm. (a) Combination of different clear shell (y axis) and core diameters (x axis). The color coding represents the SSA of a particle with specific core and clear shell size; (b) Cross-sections views of (a). The color coding represents different  $D_{\text{core}}$  in the range of 50 – 300 nm. (c)&(d) Similar to (a)&(b) but from the brown shell scenario.



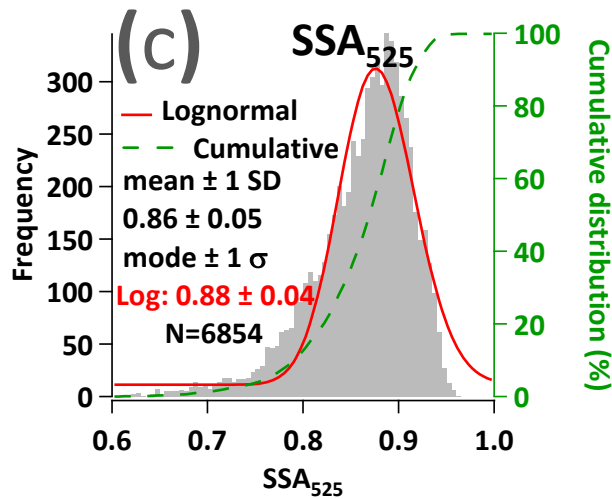
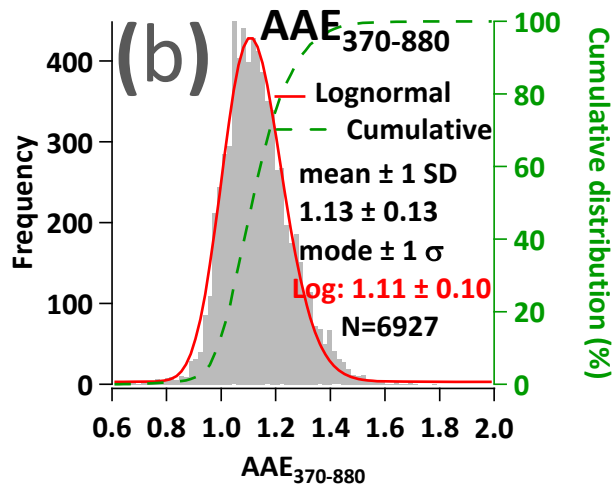
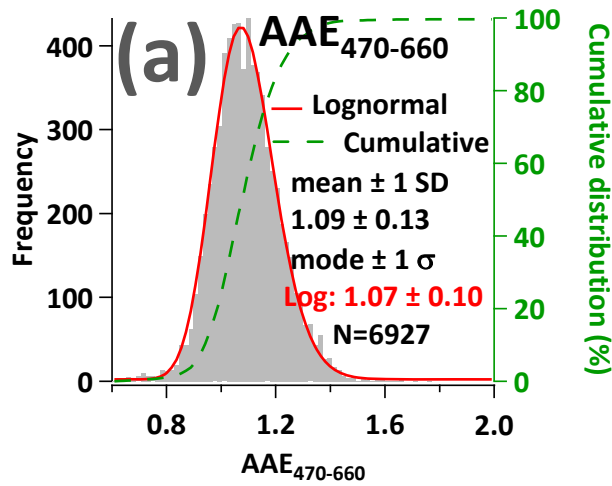
**Figure S5.** Mie simulated mass absorption efficiency (MAE<sub>p</sub>) of a bare soot particle as a function of diameter at a wavelength of 550nm. Refractive index is 1.85 – 0.71i and density varied from 1.6 to 1.9 g cm<sup>-3</sup>.



**Figure S6.** Mie simulated mass absorption efficiency (MAE) of a bare soot particle as a function of diameter at a wavelength of 550nm. Refractive index is  $1.85 - 0.71i$  and density is  $1.9 \text{ g cm}^{-3}$  for the soot core. Refractive index for clear coating is 1.55. Refractive index for brown coating is wavelength dependent adopted from Lack and Cappa (2010).



**Figure S7.** Mie simulated BrC absorption contribution to total  $E_{abs}$  (lensing effect + BrC absorption) in the brown shell scenario. (a) 370 nm (b) 550 nm (c) 880 nm.

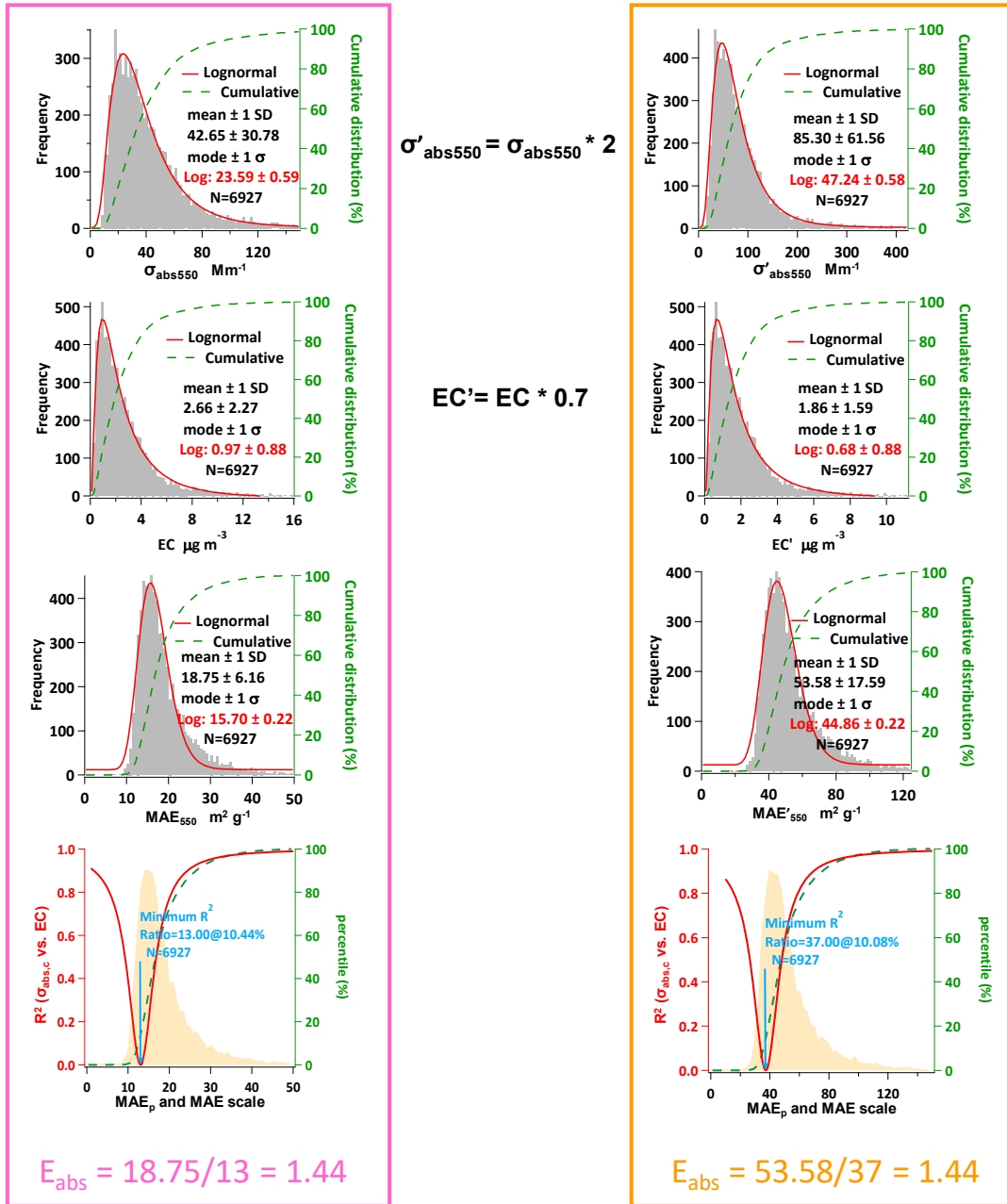


**Figure S8.** Measured annual statistics of AAE and SSA. (a) Annual frequency distribution of AAE<sub>470-660</sub>. (b) Annual frequency distribution of AAE<sub>370-880</sub>. (c) Annual frequency distribution of SSA<sub>525</sub>. The red line represents lognormal fitting curve.

# Test A

Original data

Systematically biased data

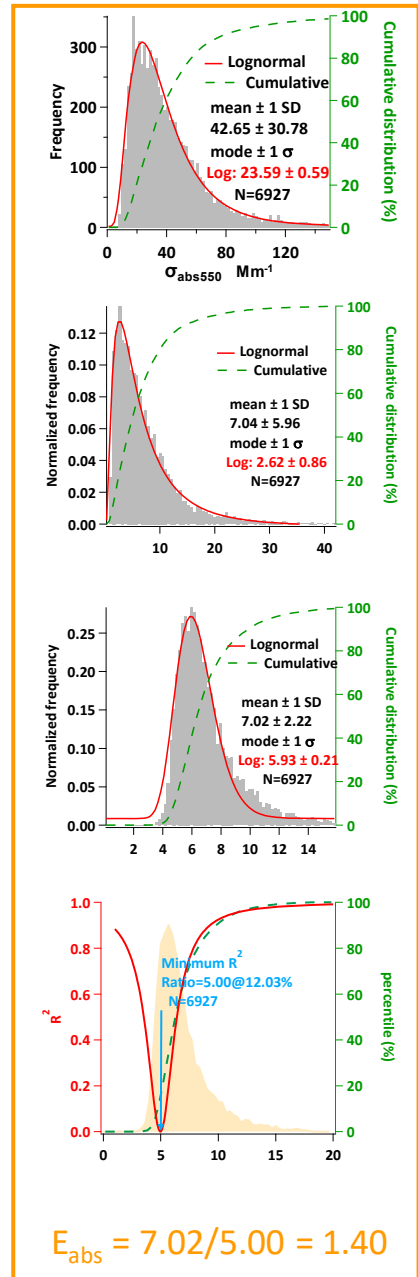
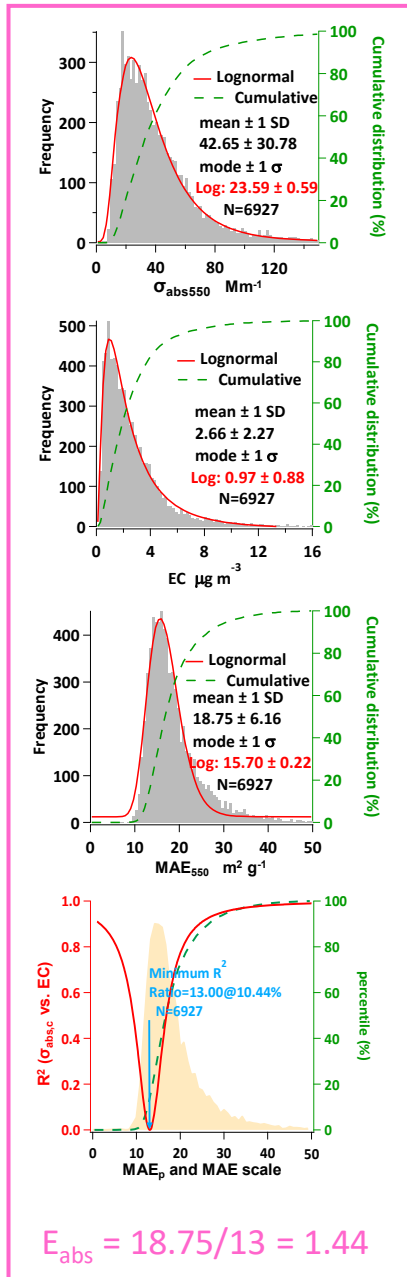


**Figure S9.** Comparison of  $E_{abs}$  from original data and systematically biased data (Test A). It should be noted that the  $E_{abs}$  shown here is ratio of averages, which is different from the annual average  $E_{abs}$  calculated from average of ratios.

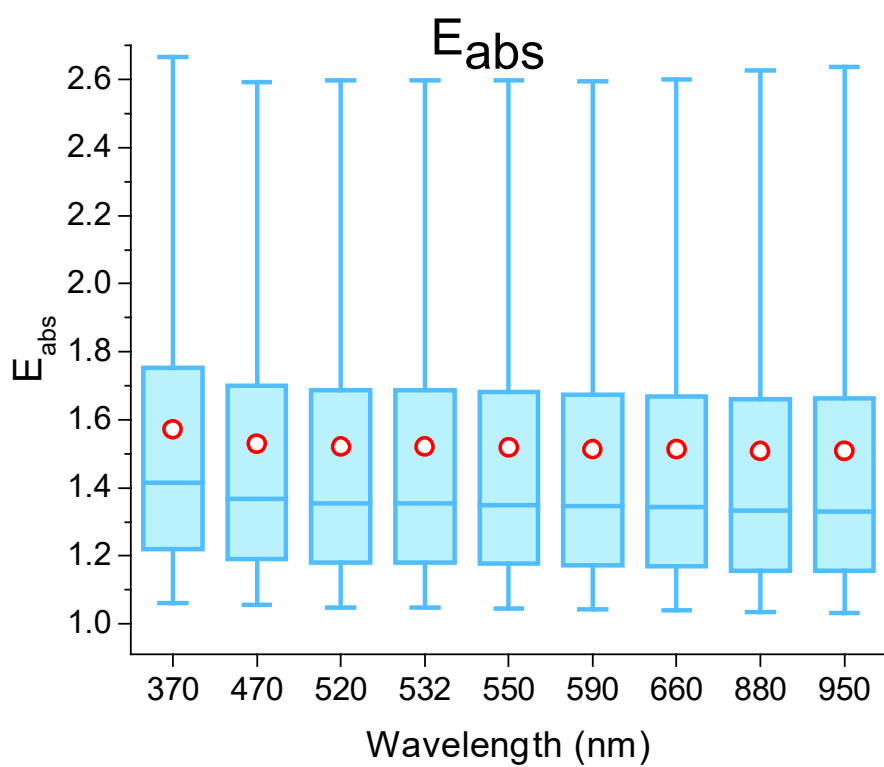
# Test B

Original data

Systematically biased data



**Figure S10.** Comparison of  $E_{\text{abs}}$  from data using NIOSH EC and data using IMPROVE EC (Test B). It should be noted that the  $E_{\text{abs}}$  shown here is ratio of averages, which is different from the annual average  $E_{\text{abs}}$  calculated from average of ratios.



Wavelength (nm)	370	470	520	532	550	590	660	880	950
$E_{\text{abs}}$ mean	1.55	1.51	1.50	1.50	1.50	1.49	1.49	1.48	1.49
$E_{\text{abs}}$ S.D.	0.48	0.47	0.47	0.47	0.48	0.47	0.48	0.49	0.49

**Figure S11.** Spectrum annual average  $E_{\text{abs}}$  from 370 to 950 nm.



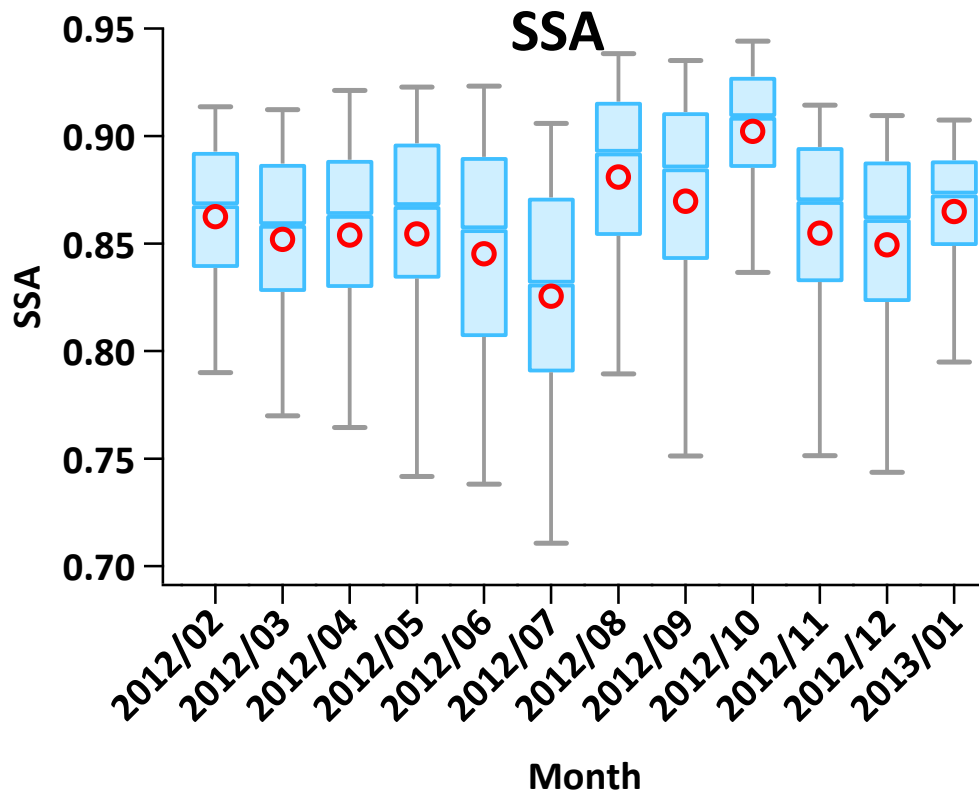
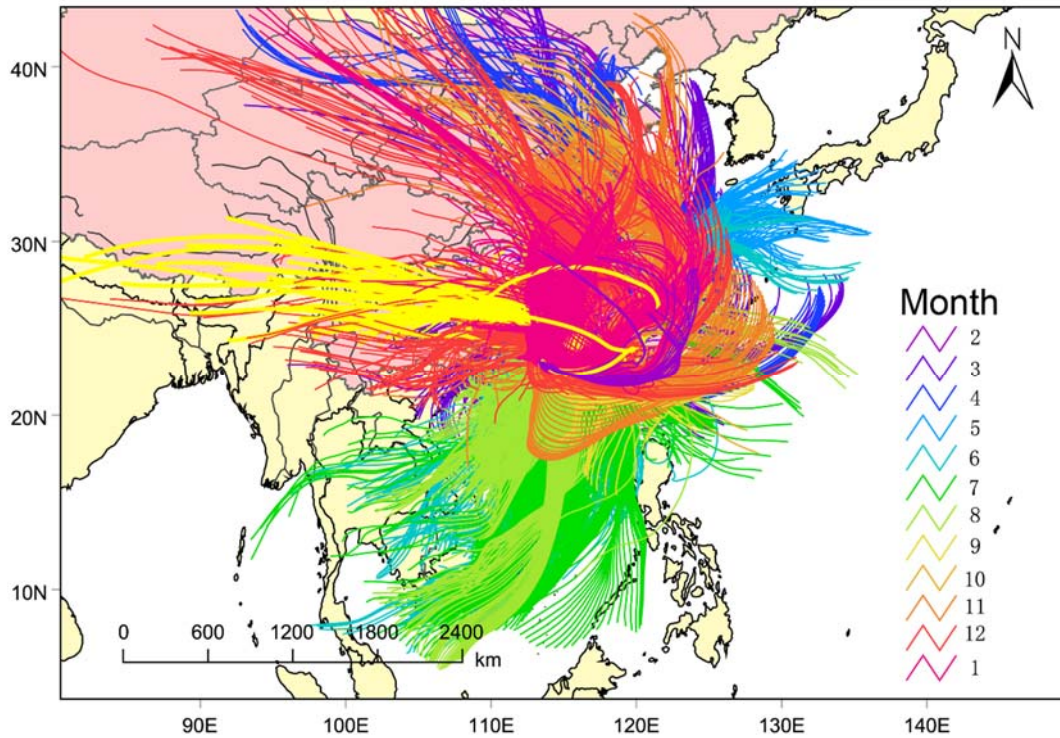
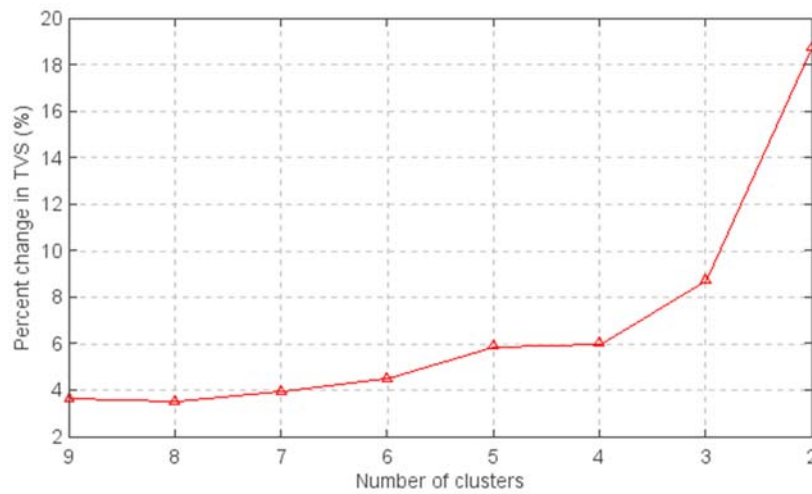


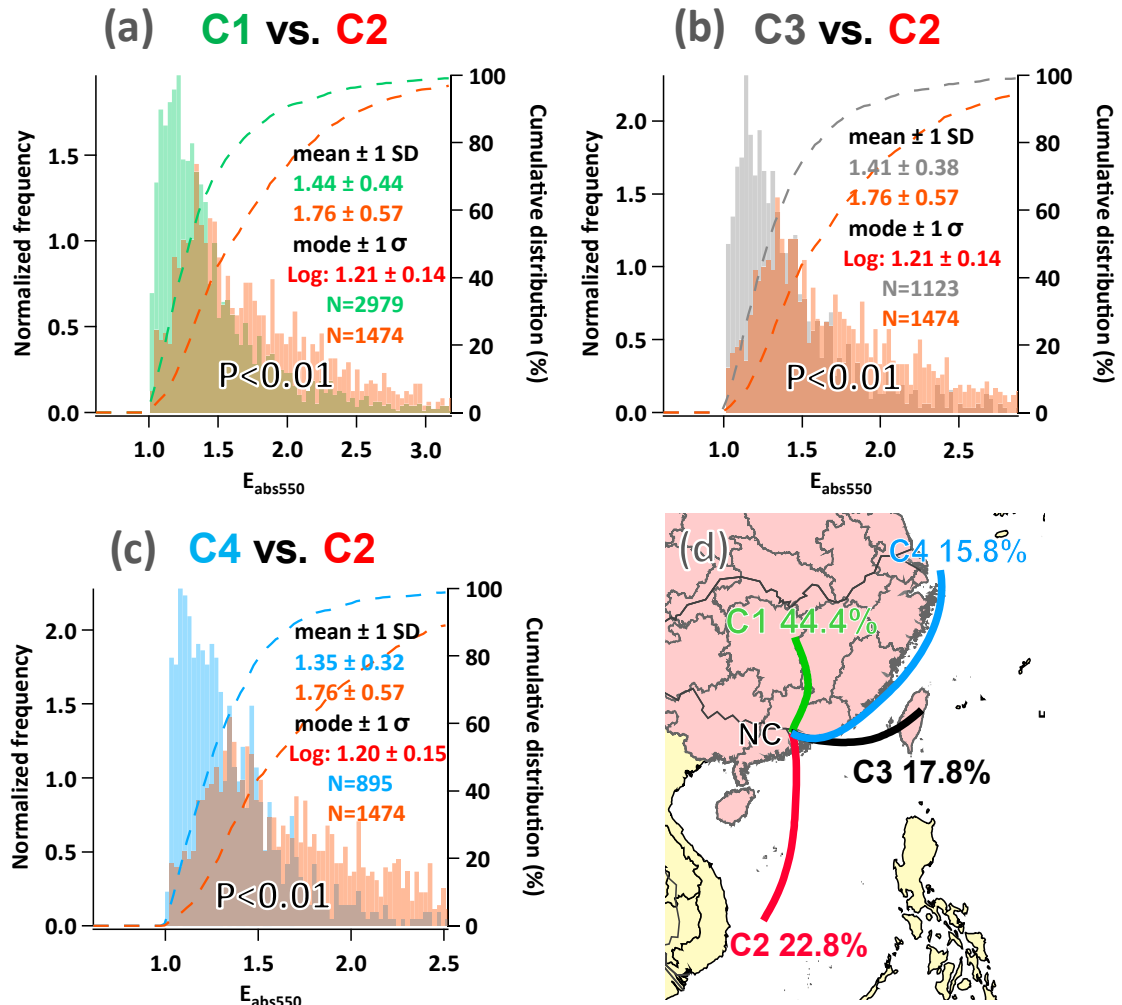
Figure S12. Measured monthly variations of SSA<sub>525</sub>.



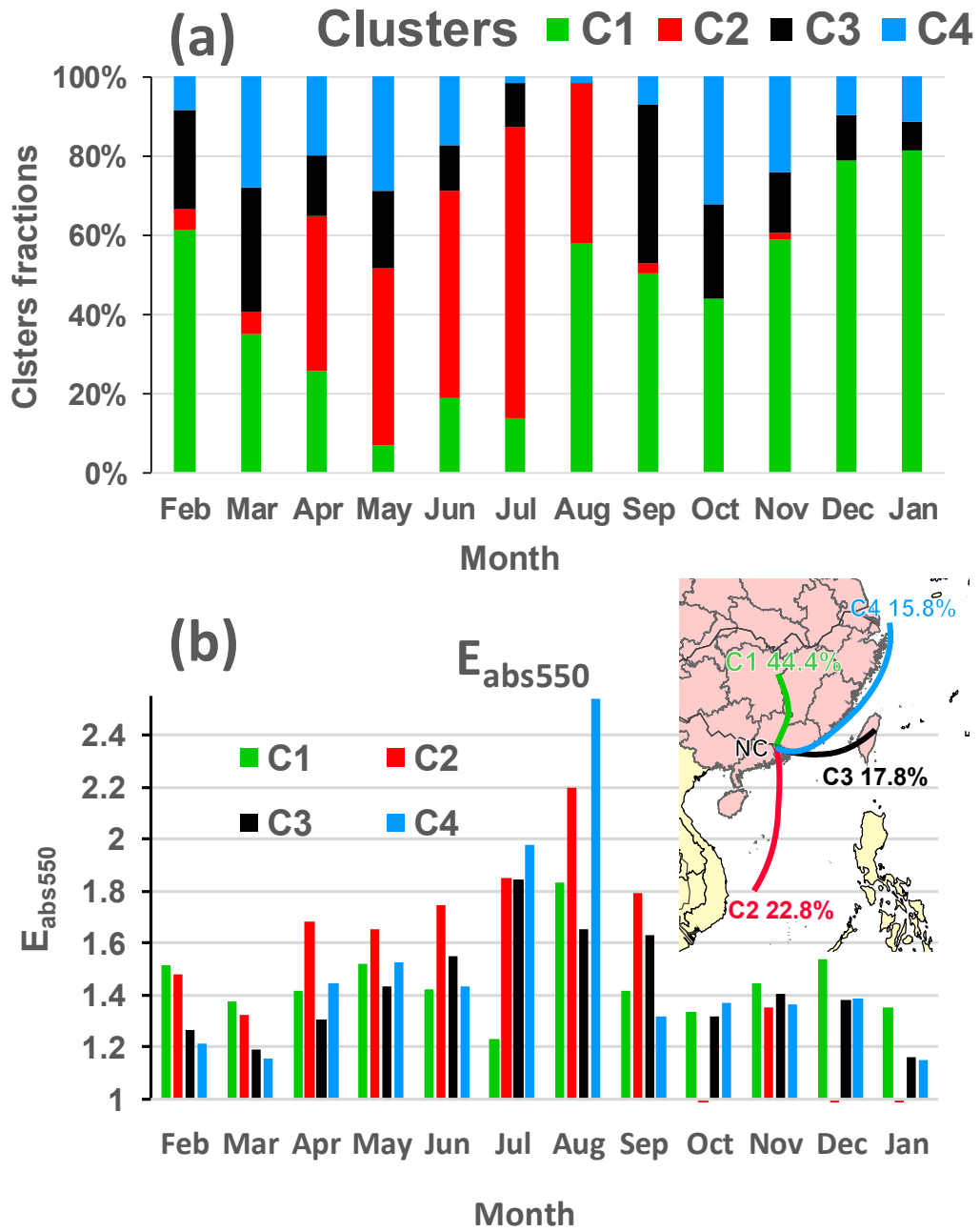
**Figure S13.** Hourly back trajectories for the past 72 hours calculated using NOAA's HYSPLIT model from Feb 2012 to Jan 2013. The color coding represents different months.

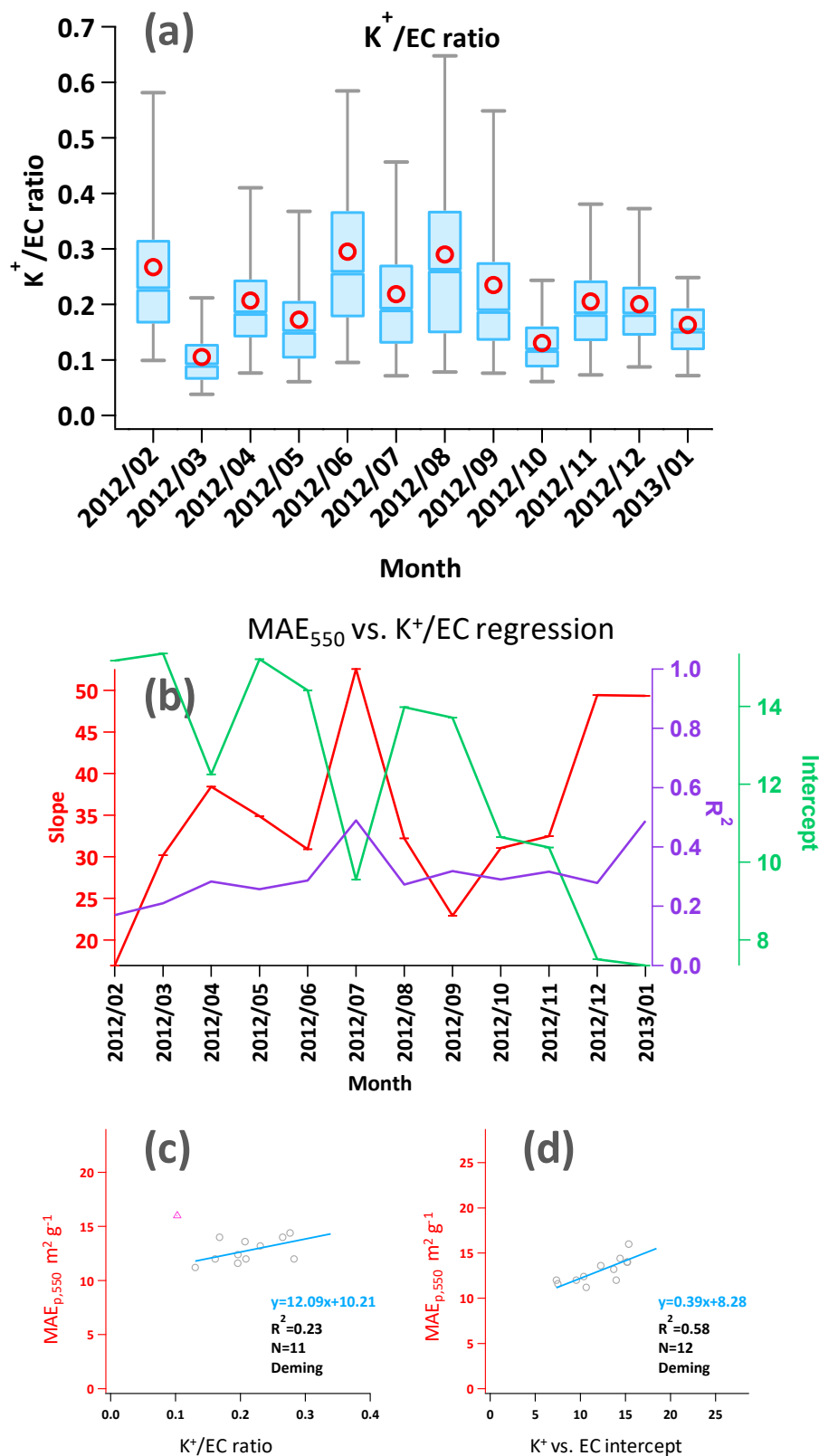


**Figure S14.** Total spatial variance (TSV) as a function of number of clusters in back trajectories clustering analysis.

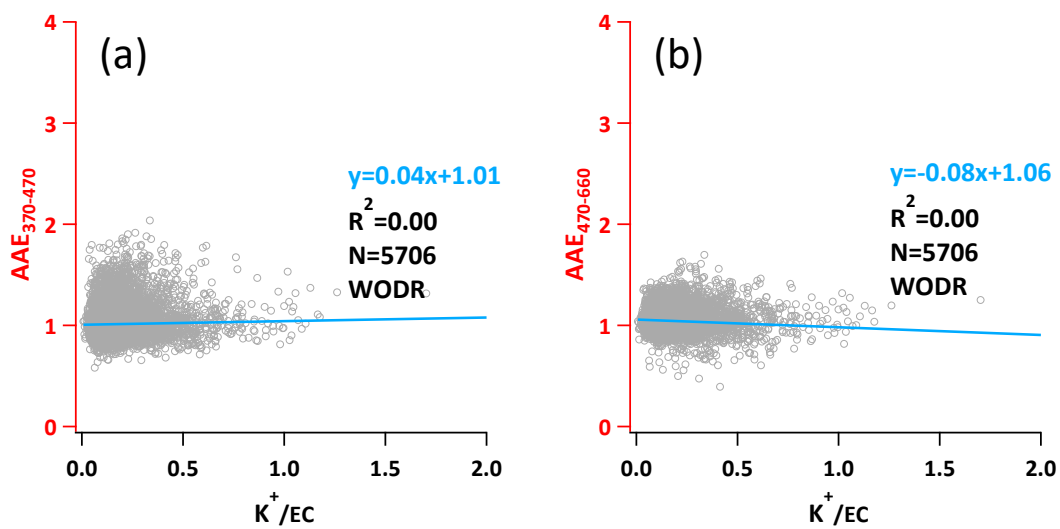


**Figure S15.** Frequency distributions of  $E_{abs550}$  by different air mass clusters. P is calculated by Wilcoxon-Mann-Whitney tests.





**Figure S17.** (a) Monthly variations of K<sup>+</sup>/EC ratio from 2012 Feb to 2013 Jan at NC site. (b) Monthly regressions between MAE<sub>550</sub> and K<sup>+</sup>/EC with slope in red, intercept in green and R<sup>2</sup> in purple. (c) regressions between monthly MAE<sub>p,550</sub> and K<sup>+</sup>/EC. (d) regression between monthly MAE<sub>p,550</sub> and intercepts from (b).



**Figure S18.** Correlations of AAE with K<sup>+</sup>/EC ratio (biomass burning indicator). (a) AAE from 370 – 470 nm. (b) AAE from 470 – 660 nm.

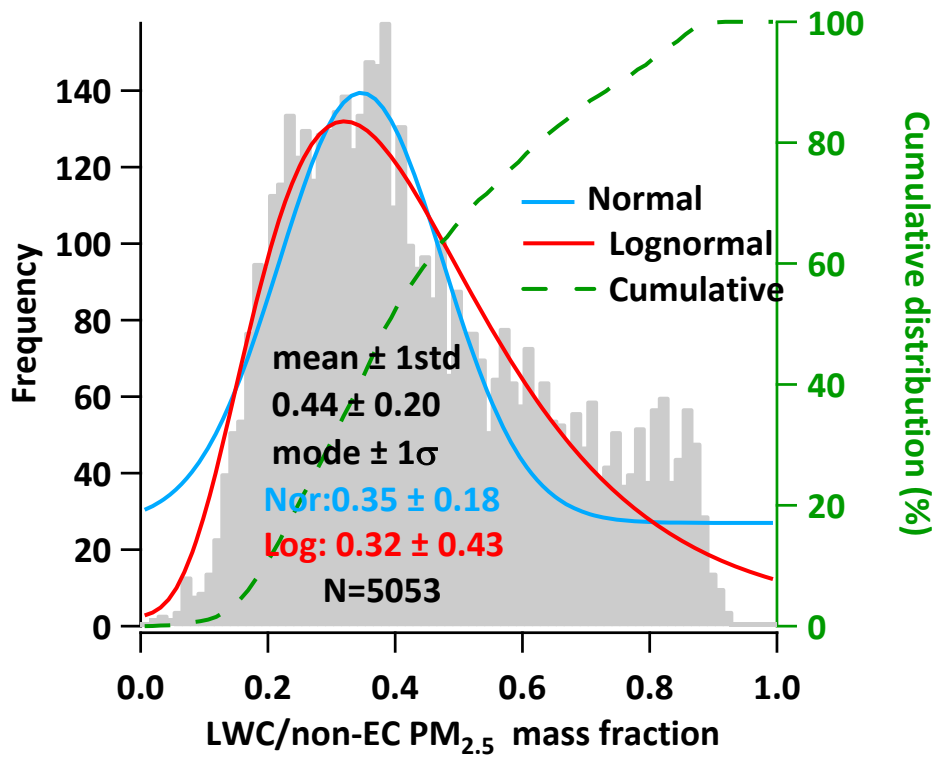
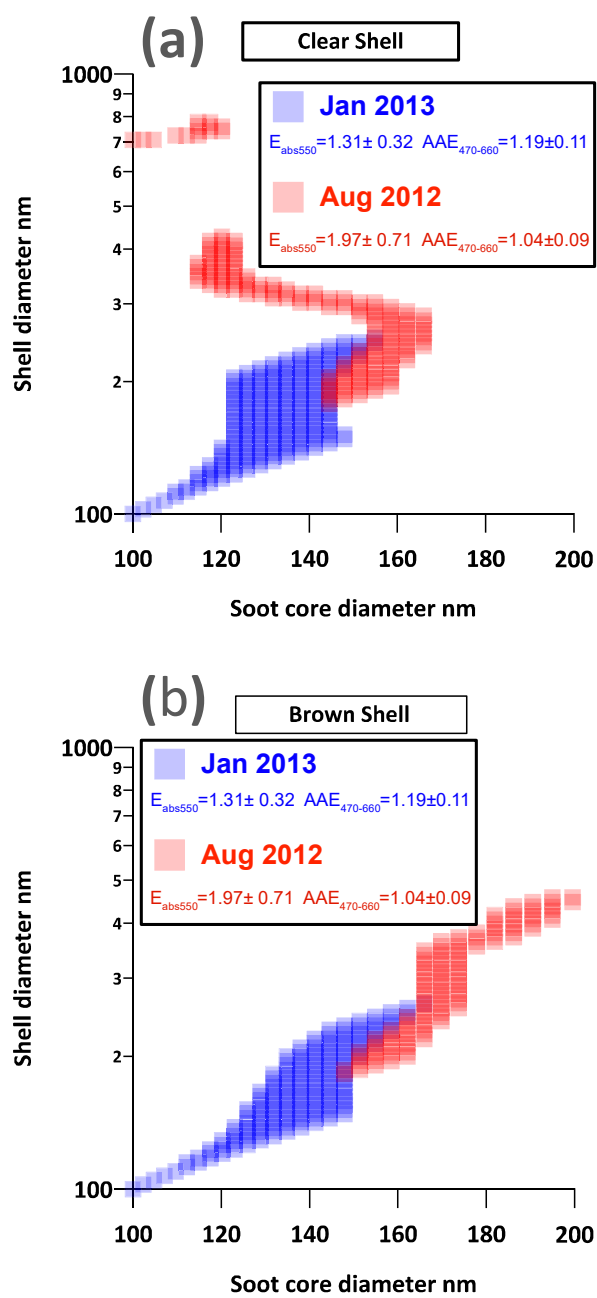


Figure S19. Annual frequency distribution of LWC/non-EC PM<sub>2.5</sub> mass fraction.



**Figure S20.** Size range of soot particles constrained by  $E_{\text{abs}}$ ,  $SSA_{525}$  and  $AAE_{470-660}$  from measurements. (a) Clear shell scenario; (b) Brown shell scenario



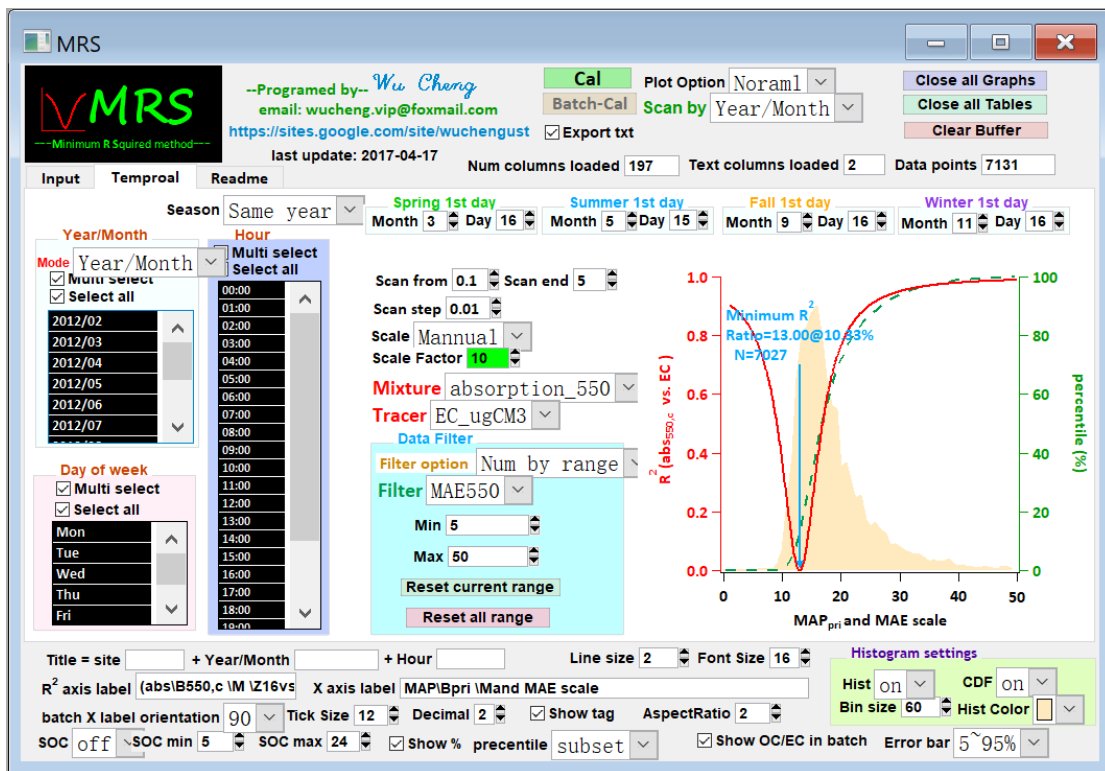
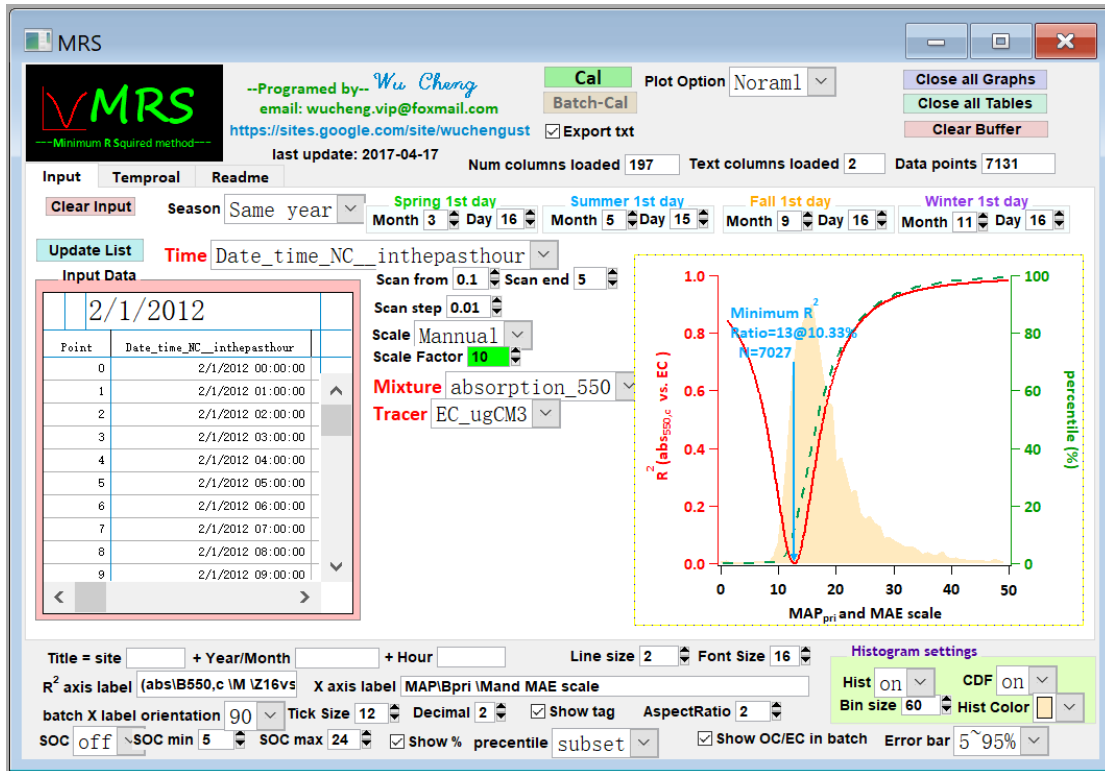
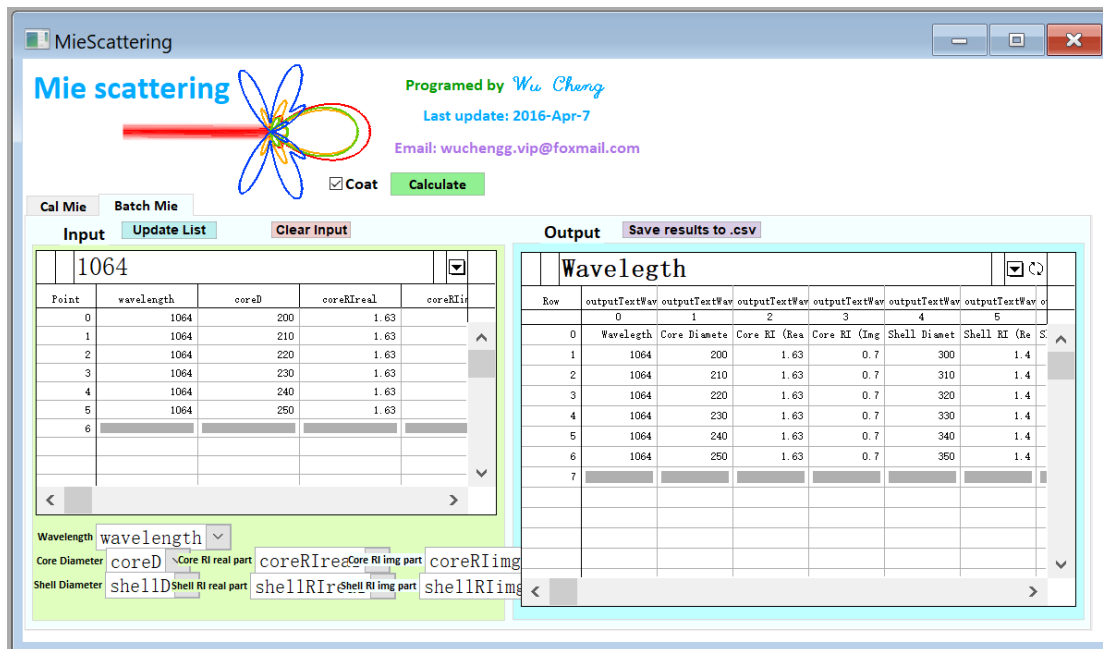
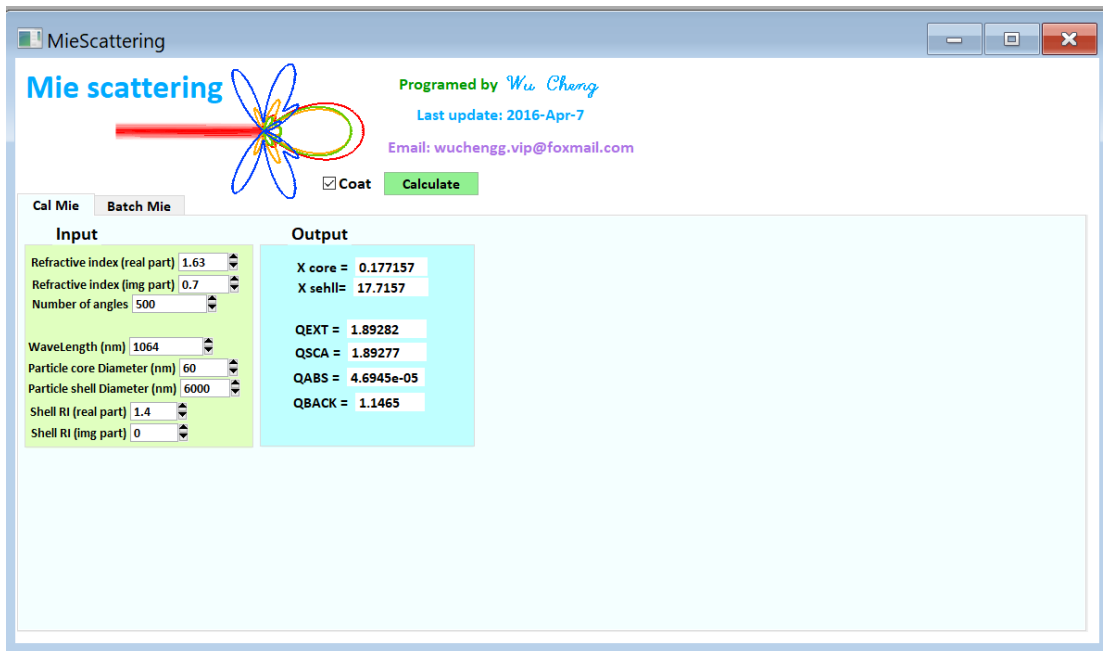
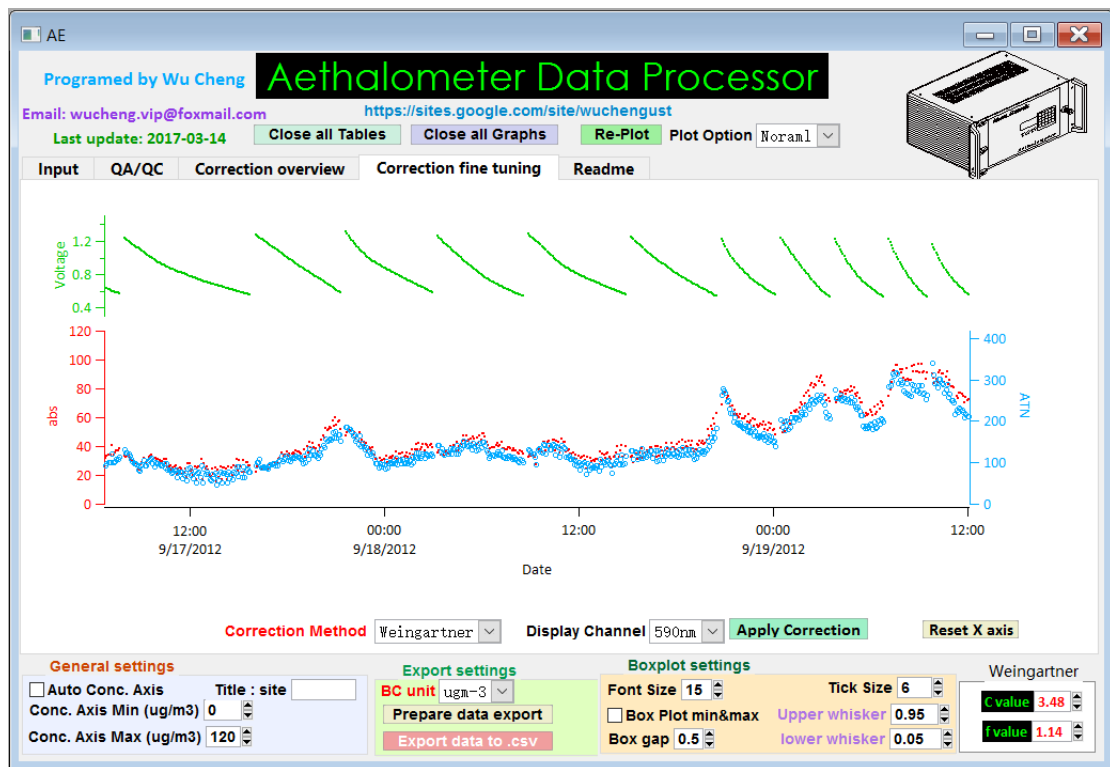
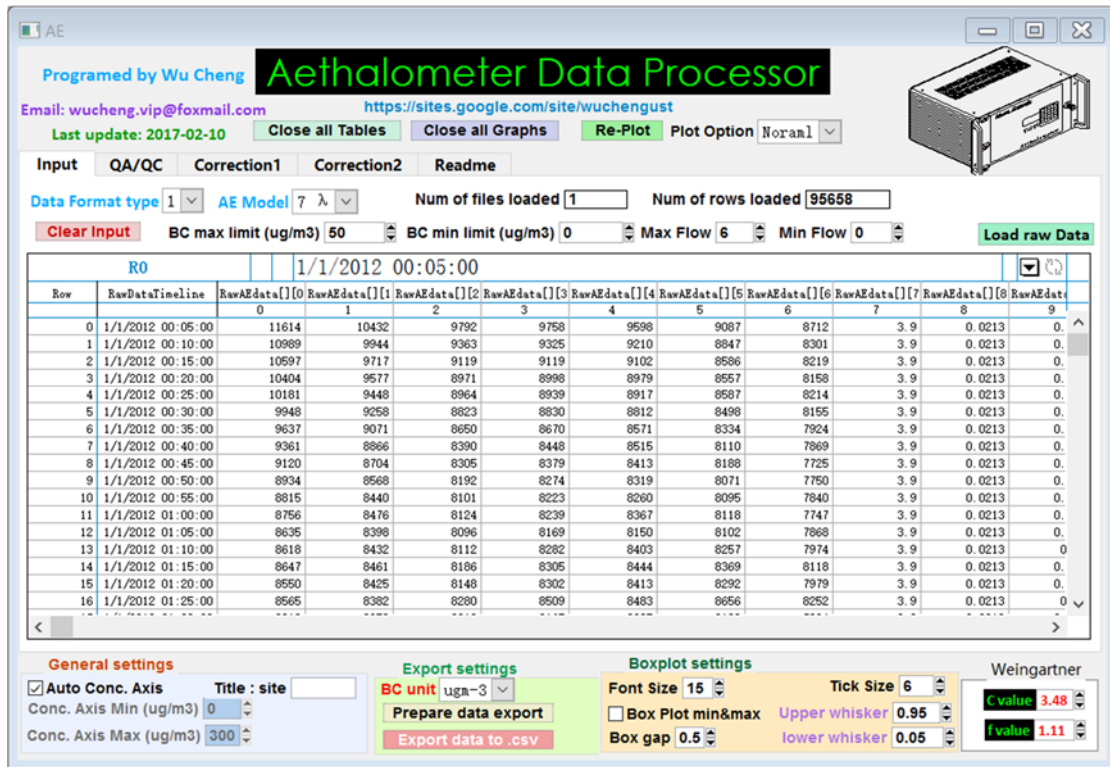


Figure S21. MRS program written in Igr Pro (WaveMetrics, Inc. Lake Oswego, OR, USA). Available from <https://sites.google.com/site/wuchengust>.



**Figure S22.** Mie program written in Igro Pro (WaveMetrics, Inc. Lake Oswego, OR, USA). Available from <https://sites.google.com/site/wuchengust>.



**Figure S23.** Aethalometer data processing program written in Igro Pro (WaveMetrics, Inc. Lake Oswego, OR, USA). Available from <https://sites.google.com/site/wuchengust>.

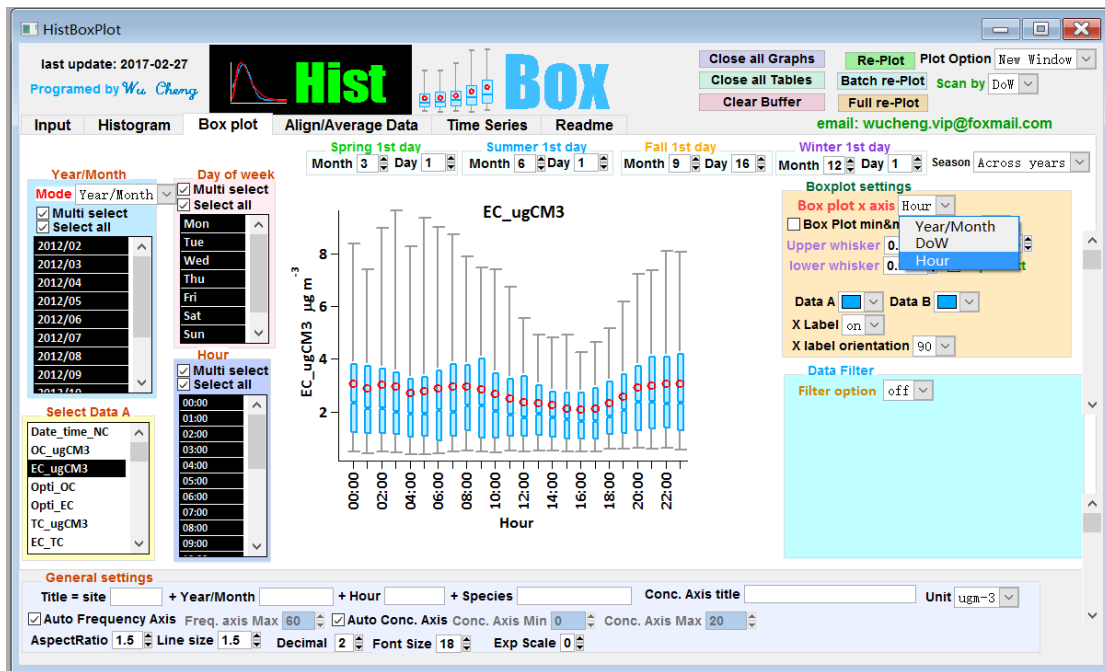
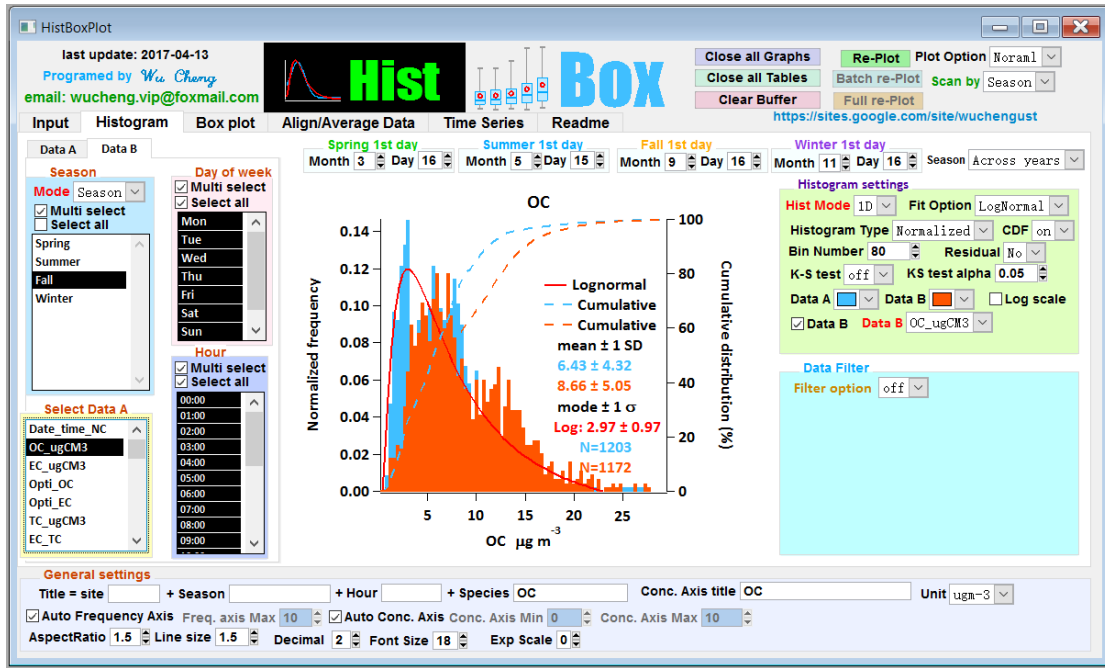


Figure S24. Histbox program written in Igro Pro (WaveMetrics, Inc. Lake Oswego, OR, USA). Available from <https://sites.google.com/site/wuchengust>.

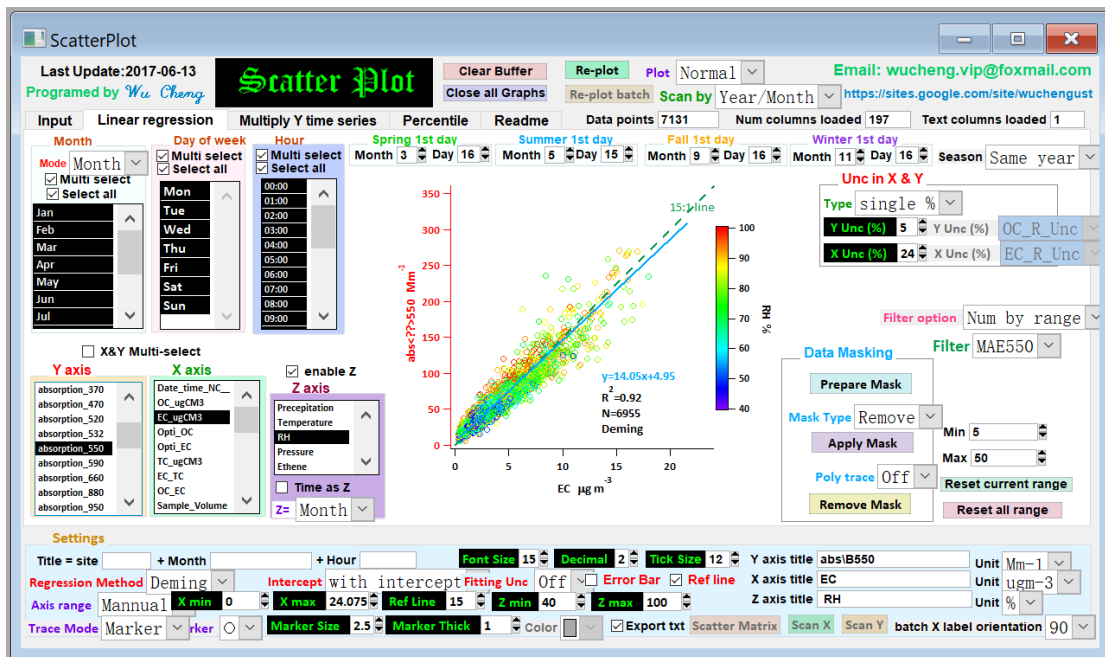
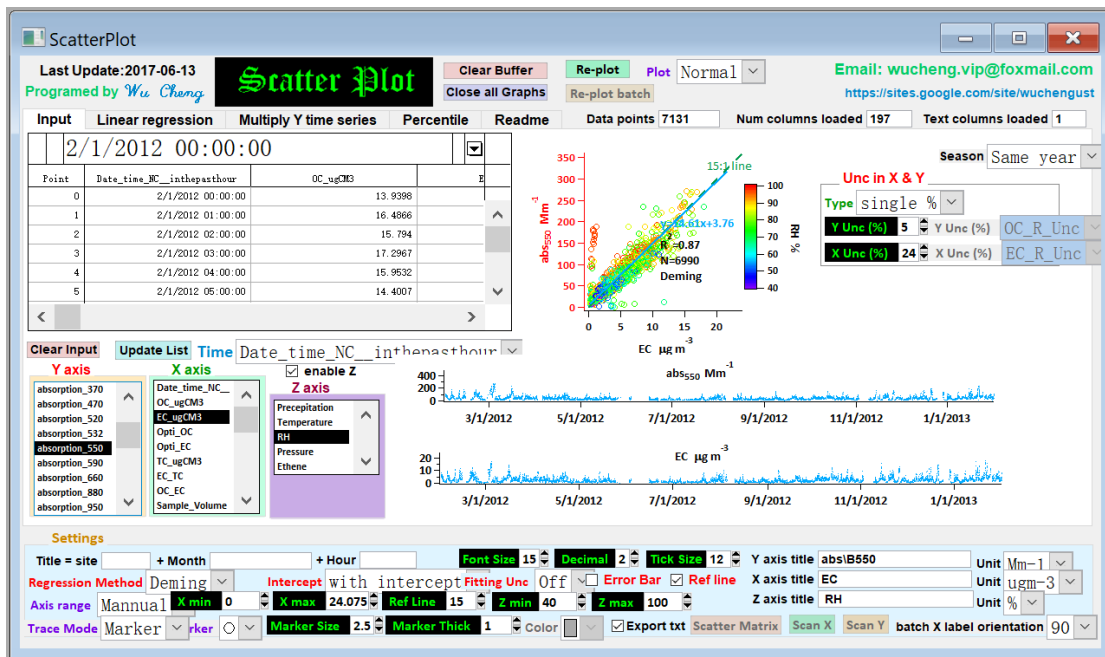


Figure S25. Scatter plot program written in Igro Pro (WaveMetrics, Inc. Lake Oswego, OR, USA). Available from <https://sites.google.com/site/wuchengust>.