

1 **Aerosol optical properties and direct radiative forcing based on**
2 **measurements from the China Aerosol Remote Sensing Network**
3 **(CARSNET) in eastern China**

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35 **Abstract**

36 Aerosol pollution in eastern China is an unfortunate consequence of the region's rapid
37 economic and industrial growth. Here, sunphotometer measurements from seven sites in the
38 Yangtze River Delta (YRD) from 2011 to 2015 were used to characterize the climatology of
39 aerosol microphysical and optical properties, calculate direct aerosol radiative forcing (DARF),
40 and classify the aerosols based on size and absorption. Bimodal size distributions were found
41 throughout the year, but larger volumes and effective radii of fine-mode particles occurred in
42 June and September due to hygroscopic growth and/or cloud processing. Increases in the
43 fine-mode in June and September caused $AOD_{440nm} > 1.00$ at most sites, and annual mean
44 AOD_{440nm} values of 0.71-0.76 were found at the urban sites and 0.68 at the rural site. Unlike
45 north China, the $AOD_{440 nm}$ was lower in July and August (~0.40–0.60) than in January and
46 February (0.71–0.89) due to particle dispersion associated with subtropical anticyclones in
47 summer. Low volumes and large bandwidths of both fine- and coarse-mode aerosol size
48 distributions occurred in July and August because of biomass burning. Single scattering
49 albedos at 440 nm ($SSA_{440 nm}$) from 0.91 to 0.94, indicated particles with relatively strong to
50 moderate absorption. Strongly absorbing particles from biomass burning with a significant SSA
51 wavelength dependence were found in July and August at most sites while coarse particles in
52 March to May were mineral dust. Absorbing aerosols were distributed more-or-less
53 homogeneously throughout the region with absorption aerosol optical depths at 440 nm
54 ~0.04-0.06, but inter-site differences in the absorption Angström exponent indicate a degree of
55 spatial heterogeneity in particle composition. The annual mean DARF was -93 ± 44 to $-79 \pm$
56 $39 W/m^2$ at the Earth's surface and $\sim -40 W/m^2$ at the top of the atmosphere (for the solar
57 zenith angle range of 50 to 80 degrees) under cloud free conditions. The fine-mode composed
58 a major contribution of the absorbing particles in the classification scheme based on SSA, fine
59 mode fraction, and extinction Angström exponent. This study contributes to our understanding
60 of aerosols and regional climate/air quality, and the results will be useful for validating satellite
61 retrievals and for improving climate models and remote sensing algorithms.

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64 **1. Introduction**

65 Aerosols can have important effects on the Earth's climate over regional to global scales,
66 but there are still uncertainties in the strengths and significance of these impacts (Hansen et
67 al.2000; Solomon et al., 2007; Schwartz and Andreae, 1996). Aerosols affect the radiative
68 balance of the Earth–atmosphere system by directly scattering and absorbing solar radiation
69 (Charlson et al., 1992; Ackerman and Toon, 1981), and they can affect climate indirectly
70 through aerosol–cloud interactions (Twomey et al., 1984; Albrecht et al., 1989; Li et al., 2016).

71 The physical and optical properties of aerosol particles determine their radiative effects,
72 and information on these properties can be used to predict and assess global and regional
73 changes in the Earth's climate (Eck et al., 2005; Myhre et al., 2009; IPCC, 2013; Panicker et
74 al., 2013). Long-term, ground-based observations have contributed greatly to our
75 understanding of the spatial variations in aerosols and their effects on the Earth's climate
76 (Holben et al., 2001; Kaufman et al., 2002; Sanap and Pandithurai, 2014; Li et al., 2016).
77 Ground-based monitoring networks have been established worldwide—for instance,
78 AERONET (Aerosol Robotic Network) (Holben et al.,1998; Goloub et al., 2007), SKYNET
79 (SKYrad Network) (Takamura et al., 2004), EARLINET (European aerosol Lidar network)
80 (Pappalardo et al., 2014) and the GAW-PFR Network (Global Atmosphere Watch
81 Programmer-Precision Filter Radiometers) (Wehrli, 2002; Estelles et al., 2012). In China,
82 CARSNET (the China Aerosol Remote Sensing NETwork) and CSHNET (the Chinese Sun
83 Hazemeter Network) were established to obtain data on aerosol optical characteristics (Che et
84 al., 2009a, 2015b; Xin et al., 2007). High-frequency, ground-based measurements of aerosol
85 optical properties made at these stations have improved our understanding of the sources,
86 transport, and diurnal variations of air pollutants, and they have provide insights into the
87 aerosols' effects on climate. Ground-based observations are also useful for the validation of
88 satellite retrievals (Holben et al., 2017; Xie et al., 2011).

89 Most of the ground-based studies of the optical properties of aerosols in China have been
90 conducted in urban regions that have been undergoing rapid economic development. Those
91 sites typically have had high aerosol loadings and in many cases serious environmental

92 problems (Cheng et al., 2015; Pan et al., 2010; Xia et al., 2013; Wang et al., 2015; Che et al.,
93 2015a). Detailed information on aerosol optical depth (AOD), the types of aerosols, and
94 especially the size and absorption properties of ambient populations over a wide sampling of
95 regions is needed to understand the effects of aerosols on the Earth's climate and the
96 environment (Giles et al., 2011; Che et al., 2009b; Wang et al., 2010; Zhu et al., 2014). In
97 particular, the aerosol direct radiative forcing is sensitive to the aerosol radiation absorptivity
98 (Haywood and Shine, 1995). Therefore, it is important to understand the connections between
99 the aerosol types and absorption properties because that information can be used for
100 comparisons and validation of chemical transport models and satellites (Lee et al., 2010).

101 The Yangtze River Delta (YRD) region in eastern China has recently undergone rapid
102 economic growth, and the loadings of aerosols in the region can be very high during heavy
103 pollution episodes (Fu et al., 2008; Zhang et al., 2009). Studies of aerosol optical properties in
104 eastern China have contributed to our understanding of local air quality and regional climate
105 impacts (Duan and Mao, 2007; Pan et al., 2010; Ding et al., 2016). In the YRD, investigations
106 of aerosol optical properties have been conducted in Nanjing, Hefei, Shanghai, Shouxian and
107 Taihu (Zhuang et al. 2014; Li et al., 2015; Wang et al., 2015; He et al., 2012; Lee et al., 2010;
108 Cheng et al., 2015; Xia et al., 2007). Those studies mostly involved sampling at single sites
109 ~100 km apart from one other without synchronous observations, and many have been of
110 relatively short duration, and so there remains a need for more extensive ground-based
111 measurements.

112 For the present study, sampling was conducted over a period of several years to better
113 characterize the climatology of the aerosol microphysical and optical properties, including
114 aerosol absorptivity, and to improve estimates of direct aerosol radiative forcing. For these
115 studies, sun photometer measurements were made at 3 min intervals from 2011 to 2015 at
116 seven CARSNET sites (one densely populated urban site, five urban center sites in smaller
117 cities, and one rural site) ~10–40 km apart in the YRD. The dense network of ground-based
118 sun- and sky-scanning spectral radiometers improves the temporal and spatial coverage of the
119 data, and that has enabled us to capture small-scale variations in the aerosols. The results not

120 only contribute to our understanding of regional climate and local air quality impacts, but they
121 also will be useful for validation of satellite data and improving the performance of models and
122 remote sensing algorithms in the future.

123 This paper is organized as follows: Section 2 describes the sites, the methods used for
124 retrieving the aerosol optical properties and their uncertainties, and the calculation of aerosol
125 direct radiative forcing from the retrieved aerosol optical parameters. Section 3 presents the
126 aerosol microphysical properties, optical properties, and calculations of direct radiative forcing.
127 The aerosol type classification is also presented based on the aerosol optical parameters.
128 Then a brief discussion is made about the analysis of this study. Section 4 is the conclusions.

129 **2. Site descriptions, measurements and methods**

130 Figure 1 shows the locations of the seven CARSNET sites of the YRD, and detailed
131 information on the sites is included in Table 1. Hangzhou is a densely populated urban site with
132 a heavy vehicular traffic, and it is affected by various types of anthropogenic emissions. LinAn,
133 Fuyang, Jiande, Xiaoshan and Tonglu are urban center sites in smaller cities, and they are all
134 affected to varying degrees by anthropogenic activities, especially pollutants from industries
135 and agriculture. The rural site of ChunAn has a small population, and there are few industrial
136 sources nearby so the effects from local or regional pollution are relatively small.

137 Sun photometers (CE-318, Cimel Electronique, Paris, France) were installed at each of
138 the seven sites and operated from 2011 to 2015. These instruments were standardized and
139 calibrated using CARSNET reference instruments (Che et al., 2009a), which in turn were
140 periodically calibrated at Izaña, Tenerife, Spain in conjunction with the AERONET program.
141 The cloud-screened AODs (based on the work of Smirnov et al. 2000) at 340, 380, 440, 500,
142 670, 870, 1020 and 1640 nm with uncertainties less than 0.01 (Eck et al., 1999) were obtained
143 using ASTPwin software (Cimel Electronique). The water vapor expressed as precipitable
144 water in the column was derived from the 940 nm channel with uncertainties less than 10%
145 (Eck et al., 1999). Daily averages and statistical analysis were calculated for days on which
146 instantaneous AOD measurements were made more than ten times (Che et al., 2015b). The

147 extinction Angström exponent values (EAE) were calculated from AOD values at 440 and 870
148 nm.

149 Aerosol microphysical properties were retrieved from the almucantar sky irradiance
150 measurements in conjunction with measured spectral AOD, following the methods of Dubovik
151 and King (2000) and Dubovik et al. (2002, 2006). The dataset contained information on (1)
152 volume size distributions in 22 size bins for particle radii 0.05–15 μm ; (2) fine and coarse mode
153 aerosol effective radii; and (3) aerosol optical properties—including the wavelength dependent
154 single-scattering albedo (SSA), the complex refractive index, the absorption AOD (AAOD), and
155 the absorption Angström exponent (AAE). For the retrieval process, the surface albedo (SA)
156 was interpolated/extrapolated to 440, 670, 870, and 1020 nm from the daily Moderate
157 Resolution Imaging Spectroradiometer (MODIS) surface reflectance product of MCD43C3
158 (<https://ladsweb.modaps.eosdis.nasa.gov/>). Following the procedures of Dubovik et al.
159 (2002, 2006), all particles smaller than 0.992 μm were considered fine mode particles while
160 those larger than 0.992 μm were considered coarse mode. And the effective radii for the total
161 ($R_{eff,t}$), fine ($R_{eff, \text{fine}}$) and coarse ($R_{eff, \text{coarse}}$) mode aerosols are calculated as follows:

$$162 \quad R_{eff} = \frac{\int_{r_{min}}^{r_{max}} r^3 \frac{dN(r)}{d \ln r} d \ln r}{\int_{r_{min}}^{r_{max}} r^2 \frac{dN(r)}{d \ln r} d \ln r} \quad (1)$$

163 Where r_{min} is 0.05, 0.05, and 0.992 μm and r_{max} is 15, 0.992, and 15 μm for the total, fine
164 and coarse mode aerosols, respectively.

165 The inversion algorithm used for calculating the aerosol volume distribution ($dV/\ln r$)
166 assumed a homogeneous distribution of non-spherical aerosol particles as in the work of
167 Dubovik (2006); this approach has been widely applied in studies of many different areas of
168 the world. The SSA was retrieved using only $AOD_{440\text{nm}} > 0.40$ measurements to avoid the
169 large uncertainties inherent in low AOD retrievals (Dubovik et al. 2002, 2006). Real and
170 imaginary parts of refractive index at 440, 675, 870, and 1020 nm were constrained to the
171 ranges of 1.33–1.60 and 0.0005–0.50, respectively (Dubovik and King, 2000; Che et al.,
172 2015b). The complex refractive index is assumed independent of particle size. This

173 assumption is valid for fine or coarse-dominated cases, however, could cause some errors in
174 SSA and particle size retrievals for mixed aerosol scenarios (Xu and Wang, 2015; Xu et al
175 2015). The AOD, AAOD, and AAE are related to one another as shown in equations (2) and
176 (3):

$$177 \quad \text{AAOD}(\lambda) = [1 - \text{SSA}(\lambda)] \times \text{AOD}(\lambda) \quad (2)$$

$$178 \quad \text{AAE} = -\text{dln}[\text{AAOD}(\lambda)]/\text{dln}(\lambda) \quad (3)$$

179 The inversion algorithms mentioned above have been used for AERONET and CARSNET,
180 and the accuracies of the volume size distribution were 15-25% for $0.1 \mu\text{m} \leq r \leq 7.0 \mu\text{m}$ and
181 25-100% for $r < 0.1 \mu\text{m}$ and $r > 7 \mu\text{m}$. The accuracies for both AOD and AAOD are ~ 0.01 . The
182 errors for the total, fine and coarse mode SSA are about 0.030, 0.037 and 0.085, respectively.
183 The imaginary and real parts of the complex refractive index for the AOD at 440 nm > 0.40 and
184 a solar zenith angle $> 50^\circ$ have errors of $\sim 0.0025\text{--}0.0042$ and 0.04, respectively (Dubovik et
185 al., 2000; Li et al., 2015a).

186 The direct aerosol radiative forcing (DARF) values in units of W/m^2 were calculated using
187 the radiative transfer module in the AERONET inversion (García et al., 2008; 2012) under the
188 assumption of cloud-free conditions. The DARF is defined as the difference in the shortwave
189 radiative fluxes between the two energy levels including and excluding aerosol effects at the
190 Earth's surface (bottom of the atmosphere, BOA) and the top of the atmosphere (TOA) in
191 equations (4) and (5) as follows:

$$192 \quad \text{DARF}_{\text{TOA}} = F_{\text{TOA}}^{\uparrow 0} - F_{\text{TOA}}^{\uparrow} \quad (4)$$

$$193 \quad \text{DARF}_{\text{BOA}} = F_{\text{BOA}}^{\downarrow} - F_{\text{BOA}}^{\downarrow 0} \quad (5)$$

194 where F and F^0 represent the broadband fluxes with and without aerosols at BOA and TOA,
195 respectively. The arrows in these equations indicate the direction of the fluxes for the
196 downward and upward cases. Defined this way, a negative value for DARF indicates aerosol
197 cooling effects while positive values imply warming, both at the BOA and the TOA.

198 In the radiative transfer module used here, the flux calculations accounted for absorption
199 and multiple scattering effects using the Discrete Ordinates (DISORT) approach (Stamnes et
200 al., 1988; Nakajima and Tanaka, 1988). The solar broadband fluxes from 0.2 to 4.0 μm were
201 simulated by using information on aerosol properties (size distribution, spectral AOD, SSA,
202 and phase function) obtained from the ground-based measurements. The spectral refractive
203 indices (both real and imaginary parts) were interpolated/extrapolated from the values
204 retrieved at four distinct wavelengths (440, 670, 870, 1020 nm) from the ground-based sun
205 photometers. Likewise, the spectral dependence of surface reflectance was
206 interpolated/extrapolated from surface albedo values used in the aerosol property retrieval
207 process for the same wavelengths.

208 The integrated effects of atmospheric aerosol scattering and absorption, gaseous
209 absorption, and molecular scattering and underlying surface reflection effects were evaluated
210 using the Global Atmospheric Model (GAME) code (Dubuisson et al., 1996; Roger et al., 2006).
211 In the GAME code, gaseous absorption (mainly H_2O , CO_2 , and O_3), is calculated from the
212 correlated k-distribution (Lacis and Oinas, 1991). The instantaneous column water vapor
213 content was retrieved by the absorption differential method from the 0.94 mm channel
214 (Smirnov et al., 2004). The total ozone content was taken from monthly climatology values
215 based on the Total Ozone Mapping Spectrometer (TOMS) measurements. The GAME model
216 accounts for spectral gaseous absorption; that is, ozone in the ultraviolet-visible spectral range
217 (0.20–0.35 μm and 0.5–0.7 μm) and water vapor in the shortwave infrared spectrum (0.8–3.0
218 μm).

219 The flux calculations were performed for a multi-layered atmosphere with the US standard
220 1976 atmosphere model for gaseous distributions and single fixed aerosol vertical distribution
221 (exponential with an aerosol height of 1 km) (Gacia et al., 2008). As these authors have
222 pointed out, solar fluxes calculated using the module described above show excellent
223 agreement with ground-based measurements of solar radiation (slope of 0.98 ± 0.00 and bias
224 of $-5.32 \pm 1.00 \text{ W/m}^2$) with a correlation of 99%. There is a small overestimation of $+9 \pm 12$
225 Wm^{-2} of the observed solar radiation at the surface in global terms, and this corresponds to a

226 relative error of $+2.1 \pm 3.0\%$. The differences range from $+14 \pm 10 \text{ Wm}^{-2}$ to $+6 \pm 13 \text{ Wm}^{-2}$ for
227 urban-industrial and biomass burning aerosols, respectively. The errors are expected to be of
228 the same magnitude at the TOA, since the same methodology and inputs are used at both
229 levels (gaseous and aerosol distribution, radiative model, etc).

230 **3. Results and discussion**

231 **3.1 Aerosol microphysical properties: particle radius and volume size distributions**

232 Figure 2 shows the monthly aerosol size distribution ($dV/d\ln r$) for all seven sites in the
233 YRD. The annual mean values for the effective radii of the total particles (R_{eff}) were $\sim 0.30 \mu\text{m}$,
234 and the average annual volume was $\sim 0.18 \mu\text{m}^3/\mu\text{m}^2$ (Table 1). The fine-mode effective radii
235 averaged $\sim 0.16 \mu\text{m}$ in the YRD with a fractional volume of $0.10\text{--}0.11 \mu\text{m}^3/\mu\text{m}^2$ while the
236 coarse-mode average effective radii were $\sim 2.2 \mu\text{m}$ with a fractional volume $\sim 0.08 \mu\text{m}^3/\mu\text{m}^2$.
237 These results show that there was a larger contribution of fine-mode particles to the aerosol
238 volume compared with the coarse mode at all sites. The total and fine-mode aerosol volumes
239 and effective radii showed small differences from the densely populated urban site (Hangzhou),
240 to the urban center sites in smaller cities (Xiaoshan, Fuyang, LinAn, Tonglu, Jiande) or the
241 rural site (ChunAn), and this reflects a generally homogeneous distribution of the aerosol in the
242 YRD. The coarse-mode aerosol volumes also showed small differences among site, but the
243 range of effective radii varied $2.16\text{--}2.30 \mu\text{m}$.

244 Higher volumes and larger effective radii of the fine-mode were observed in June and
245 September at most sites, and again there were relatively small differences among sites; the
246 exception was Xiaoshan where the month-to-month differences in these variables were less
247 pronounced (Fig. 2). The increases in submicron particles during the summer may have been
248 caused in part by the hygroscopicity of the aerosols. In this regard, Saha et al. (2004) reported
249 the volume of accumulation mode particles increased faster than the coarse-mode under high
250 relative humidity conditions. At our sites, the precipitation in June and September was greater
251 than in the other months because of the “Meiyu flood period” and the “typhoon and autumn
252 rain period”, respectively. Indeed, the major aerosol components, which include sulfate, nitrate,

253 ammonium, and organic compounds, can cause severe haze-fog events during high relative
254 humidity conditions. Fine particles containing sulfate, nitrate, and ammonium are hygroscopic,
255 and their sizes are strongly affected by the relative humidity (Fu et al., 2008; Shen et al., 2015;
256 Li et al., 2015b; Huang et al., 2016). Additionally, broad fine-mode distributions may result from
257 the occurrence of fog or low-altitude cloud dissipation events (Eck et al. 2012; Li et al., 2010,
258 2014). Eck et al. (2012) also pointed out that a large range of fine-mode aerosol sizes may
259 result from cloud processing, and that also could contribute to a shoulder of larger size
260 particles in the accumulation mode, especially in regions where sulfate and other
261 water-soluble aerosols exist. Another interesting observation is that the fine-mode aerosol
262 volume in July was relatively low at all sites, and is coincident with lower relative humidity in
263 July (~60%) compared with that in June (~80%) in the YRD. Therefore, the hygroscopic
264 effects on fine particles in July evidently are not as obvious as in June or September.

265 High volumes for coarse-mode aerosol occurred in March to May at all sites, and this
266 suggests that more large particles occurred in spring than in the other seasons. The most
267 likely explanation for this is the presence of mineral dust in the YRD region at that time of year.
268 High PM₁₀ mass concentrations of Hangzhou during 2012-2015 showed 78.5 ± 15.4 , $84.7 \pm$
269 11.3 , and $83.6 \pm 15.5 \mu\text{g}/\text{m}^3$ in March, April, and May, respectively (see the supplement), which
270 is consistent to the results of relative large coarse-mode aerosol volumes in this study. As Cao
271 et al. (2009) pointed out, fugitive dust can account for about one-third of PM₁₀ mass
272 concentration in Hangzhou, and that mainly came from re-suspended road dust and
273 construction soil. The long-range transportation of dust in spring from northern/northwestern
274 China could also contribute to the high coarse-mode aerosol volume. For instance, Fu et al.
275 (2014) and Sun et al. (2017) found that dust particles could be transported long distances, and
276 the impacts were apparent in the YRD. Low volumes and large bandwidths of coarse-mode
277 aerosol were found in July and August at all sites, and that may have been due to the wet
278 removal of coarse particles by the heavy precipitation in June. In July and August, strong
279 convection associated with subtropical anticyclone can disperse the aerosol. These results
280 also are consistent with previous reports showing minimum PM₁₀ concentrations in summer of
281 the YRD region (Cao et al., 2009). It was found the PM₁₀ mass concentrations of Hangzhou

282 during 2012-2015 were relative low with values 53.2 ± 6.8 , and $56.7 \pm 4.4 \mu\text{g}/\text{m}^3$ in July and
283 August, respectively (see the supplement) Sun et al. (2013) found that aerosol size
284 distributions can broaden under unstable weather conditions, and in July and August, the
285 Weather in the YRD can become unstable due to the high temperatures, and this could be
286 another factor that contributed to the large bandwidth of both fine- and coarse-mode particles
287 in our study. However, these possible connections should be re-visited in the future.

288 **3.2 Aerosol optical properties: AOD and EAE**

289 The arithmetic mean annual values for $\text{AOD}_{440\text{nm}}$ at the six urban sites (Hangzhou,
290 Xiaoshan, Fuyang, LinAn, Tonglu, Jiande) were 0.71-0.76 and 0.68 at the rural site of ChunAn
291 (Table 1). The difference in AOD between urban and rural sites was <10%, and this indicates
292 there were widespread anthropogenic impacts on the aerosol populations in the YRD and that
293 the high particle concentrations extend beyond the local to the regional scale. Nonetheless,
294 the $\text{AOD}_{440\text{nm}}$ generally decreased from the east coast to inland areas towards the west (0.76
295 at Hangzhou, 0.73 at LinAn, 0.71 at Jiande, and 0.68 at ChunAn), and this can be explained by
296 stronger anthropogenic impacts in the more urbanized east. The high $\text{AOD}_{440\text{nm}}$ at Hangzhou
297 was likely the result of the greater industrial activity and higher population density in the
298 eastern part of that metropolitan region; both of those factors could lead to larger aerosol
299 emissions compared with the less populated urban and rural sites. The coarse-mode AOD
300 values were just ~ 0.06 to 0.08 , and ratios of the fine-mode $\text{AOD}_{440\text{ nm}}$ to the total $\text{AOD}_{440\text{ nm}}$
301 varies from 0.89-0.91 at the sites, and therefore, fine-mode particles clearly were the main
302 contributors to light extinction in the region. The less coarse mode fraction of total aerosol
303 extinction ($\sim 10\%$) indicated that the contribution of coarse particles to aerosol loading in the
304 YRD region is not as obvious as in other north/northeast China region (Zhang et al., 2012).

Table1.Geographical location and annual arithmetic mean optical parameters for aerosols from seven sites in the Yangtze River Delta.

	Hangzhou	Xiaoshan	Fuyang	LinAn	Tonglu	Jiande	ChunAn
Site type	Urban	Suburban	Suburban	Suburban	Suburban	Suburban	Rural
Longitude (°E)	120.19	120.25	119.95	119.72	119.64	119.27	119.05
Latitude (°N)	30.26	30.16	30.07	30.23	29.80	29.49	29.61
Altitude (m)	41.9	14.0	17.0	139	46.1	88.9	171.4
^a N _{day}	485	180	217	562	498	480	439
^b N _{inst.}	2052	752	906	2410	2255	1952	1731
^c AOD _{440nm}	0.76±0.42	0.76±0.43	0.76±0.45	0.73±0.44	0.71±0.41	0.73±0.40	0.68±0.38
^c AOD _{fine(440nm)}	0.68±0.42	0.69±0.41	0.69±0.44	0.66±0.43	0.64±0.41	0.66±0.40	0.61±0.38
^c AOD _{coarse(440nm)}	0.08±0.06	0.07±0.06	0.07±0.06	0.07±0.07	0.07±0.06	0.07±0.07	0.06±0.05
^d EAE _{440-870 nm}	1.29±0.26	1.37±0.24	1.32±0.24	1.29±0.27	1.30±0.26	1.32±0.28	1.22±0.25
^c SSA _{440nm}	0.91±0.06	0.93±0.04	0.94±0.04	0.93±0.05	0.92±0.04	0.92±0.05	0.94±0.03
^e SSA _{670nm}	0.92±0.06	0.91±0.06	0.93±0.06	0.92±0.05	0.93±0.05	0.92±0.07	0.94±0.03
^f SSA _{870nm}	0.90±0.07	0.90±0.07	0.91±0.08	0.91±0.06	0.91±0.06	0.90±0.08	0.93±0.04
^g SSA _{1020nm}	0.89±0.08	0.89±0.08	0.89±0.09	0.90±0.07	0.90±0.07	0.90±0.09	0.92±0.05
^c AAOD _{440 nm}	0.06±0.05	0.05±0.04	0.04±0.04	0.05±0.04	0.05±0.04	0.06±0.04	0.04±0.03
^d AAE _{440-870 nm}	1.13±0.46	0.88±0.42	0.85±0.43	0.98±0.35	1.11±0.49	1.16±0.44	0.93±0.31
^c Reff _i (μm)	0.30±0.10	0.29±0.09	0.30±0.09	0.29±0.10	0.29±0.10	0.29±0.09	0.30±0.10
^c Reff _{fine} (μm)	0.16±0.04	0.16±0.03	0.17±0.04	0.16±0.04	0.16±0.04	0.17±0.04	0.17±0.04
^c Reff _{coarse} (μm)	2.21±0.40	2.26±0.35	2.30±0.39	2.24±0.44	2.19±0.41	2.16±0.39	2.27±0.42
^c Volume(μm ³ /μm ²)	0.19±0.09	0.19±0.09	0.19±0.09	0.18±0.09	0.17±0.09	0.18±0.09	0.17±0.07
^c Volume _{fine} (μm ³ /μm ²)	0.10±0.06	0.11±0.06	0.11±0.07	0.10±0.06	0.10±0.06	0.10±0.06	0.10±0.06
^c Volume _{coarse} (μm ³ /μm ²)	0.09±0.06	0.08±0.05	0.08±0.06	0.08±0.05	0.08±0.06	0.08±0.07	0.07±0.05
^c DARF-BOT(W/m ²)	-93±44	-84±41	-80±40	-81±39	-79±39	-82±40	-74±34
^c DARF-TOA(W/m ²)	-35±20	-36±21	-37±21	-36±21	-35±20	-35±21	-40±19

306 Table 1 (Continued)

307 ^a Number of available observation days.

308 ^b Number of instantaneous observations.

309 ^c Optical parameters at a wavelength of 440 nm.

310 ^d Angström exponents between 440 and 870 nm.

311 ^e Optical parameters at a wavelength of 670 nm.

312 ^f Optical parameters at a wavelength of 870 nm.

313 ^g Optical parameters at a wavelength of 1020 nm.

314

315 Unlike one peak $AOD_{440\text{ nm}}$ distribution found in June–August over northern and
316 northeastern China (Che et al., 2015), the monthly averaged AODs at 440 nm at all seven
317 sites showed two peaks, one in June and the other in September (Fig.3) with mean values of
318 $\sim 1.26 \pm 0.50$ (FMF ~ 0.93) and $\sim 1.03 \pm 0.57$ (FMF ~ 0.95), respectively. The FMF is defined
319 as the fine mode particle $AOD_{440\text{ nm}}$ fraction ($AOD_{\text{fine}(440\text{ nm})}/AOD_{440\text{ nm}}$). Low $AOD_{440\text{ nm}}$ values
320 of $\sim 0.40 - 0.60$ were found in July and August throughout the region, and these low values are
321 consistent with the discussion above concerning aerosol microphysical characteristics. That is,
322 hygroscopic effects and/or cloud processing of fine-mode aerosol particles caused greater
323 extinction in June and September compared with July and August. Zhang et al. (2015)
324 reported the hygroscopic particles under high relative humidity conditions could cause strong
325 aerosol light scattering in the Yangtze River Delta. Other meteorological factors also may have
326 played a role in the variations in $AOD_{440\text{ nm}}$ during summer because in July and August,
327 subtropical high-pressure systems prevail, and the planetary boundary layer (PBLH) at
328 Hangzhou is deep, $\sim 1.5\text{--}2.0$ km (Sun et al., 2017). The large PBLH associated with
329 subtropical anticyclone favor aerosol dispersion, and this can help explain the relatively low
330 aerosol extinction observed in July and August. In January and February, high $AOD_{440\text{ nm}}$
331 values (0.71-0.89) were observed at all sites, and this can be attributed to emissions from
332 residential heating and the stability of the atmosphere, which can cause the near surface
333 accumulation of aerosol particles. The high $AOD_{440\text{ nm}}$ values in winter also are consistent with
334 studies by Cao et al. (2009) who reported that the PM_{10} mass concentrations at Hangzhou
335 were highest in winter. According to the PM_{10} mass concentration measurements during
336 2012-2015 at Hangzhou, it was found that the PM_{10} mass concentrations was about 82.2 ± 9.5 ,
337 59.1 ± 6.0 , 82.8 ± 13.9 , and 92.3 ± 14.3 $\mu\text{g}/\text{m}^3$ in spring, summer, autumn, and winter,
338 respectively (see the supplement).

339 The extinction Angström exponent ($EAE = -d\ln[EAOD(\lambda)]/d\ln(\lambda)$) can be regarded as an
340 indicator of aerosol size; that is, $EAE_{440\text{ nm}-870\text{ nm}} > 1.00$ typically indicates that the aerosol
341 particles are small. The mean extinction Angström exponent at all seven CARSNET sites was
342 higher than 1.20 throughout the year (Table 1), which means that small particles were
343 predominant. This finding is consistent with the reported dominance of small particles from

344 anthropogenic emissions and agricultural activity in the region (Tan et al., 2009).

345 **3.3 Aerosol optical properties: single-scattering albedo**

346 The SSAs at 440 nm at our seven sites in YRD region varied from 0.91 to 0.94 (Table 1),
347 and boxplots of the monthly SSAs at wavelengths of 440, 670, 870 and 1020 nm are shown in
348 Figure 4. Eck et al. (2005) reported that the SSAs at 440 nm from AERONET retrievals were
349 confined to a relatively narrow range of values globally from ~ 0.82 to 0.98. Therefore, the SSA
350 values in this study may be explained by moderately to strongly absorbing aerosols from
351 industrial emissions and other anthropogenic sources. The $SSA_{440\text{ nm}}$ at Hangzhou site was
352 0.91 ± 0.06 , which is lower than that at the rural ChunAn site (0.94 ± 0.03). As Dubovik et al.
353 (2000, 2002, 2006) reported, the SSA depends on two factors—particle size and composition.
354 From Table 1, one can see that the differences in effective radii for total, fine and coarse mode
355 particles between Hangzhou and ChunAn are quite small, and therefore, the differences in
356 SSAs between the two sites can best be explained by differences in composition. Furthermore,
357 the differences in SSAs between these sites indicate that there was a higher percentage of
358 absorbing aerosols at urban site than the rural one.

359 The SSAs for seven sites showed significant month-to-month variations. The increased
360 scattering (light-absorbing) effects seen in June can be attributed to hygroscopic growth, which
361 can modify aerosol properties greatly (Xia et al., 2007). The presence of light-absorbing dust
362 aerosols in spring and absorbing aerosols from biomass burning in August were probably
363 responsible for the differences in SSA values observed between those months because of the
364 distinct differences in the intensive optical properties of dust and biomass burning products
365 (Yang et al., 2009). At Hangzhou, the monthly average SSA values at 440 nm were relatively
366 high in February ($\sim 0.94 \pm 0.05$) and June ($\sim 0.92 \pm 0.06$) and more moderate in March ($\sim 0.90 \pm$
367 0.06) and August ($\sim 0.89 \pm 0.09$). In comparison, the differences in monthly SSA values at the
368 rural ChunAn site were smaller, only varying from 0.92 to 0.95. We conclude from the temporal
369 patterns of the SSAs that the types of aerosols at the urban/suburban sites were more variable
370 than at the rural sites.

371 The SSA wavelength dependence is a function of the specific absorption/scattering
372 properties of different aerosol types (Sokolik and Toon, 1999; Eck et al., 2010). The SSA for
373 mineral dust particles typically shows a strong wavelength dependence from 440 to 1020 nm
374 with a low value at 440 nm due to iron oxide absorption (Cheng et al., 2006; Dubovik et al.,
375 2002). In spring, especially in March, the SSA was obviously lower at shorter wavelengths than
376 at the longer ones, and this implies absorption by dust particles. This conclusion is consistent
377 with the discussion above concerning the impact of dust on aerosol size distributions. In
378 addition, there was a significant decrease in SSA at shorter wavelengths in July and August at
379 most sites, and this supports the presence of aerosol particles with strong absorption,
380 especially at infrared wavelengths. The decreases in those months can be explained by
381 strongly absorbing aerosols from biomass burning or possibly industrial emissions. As Ding et
382 al. (2013a, b) and Wang and Zhang (2008) have reported, plumes from agricultural burning
383 typically contain light-absorbing carbonaceous aerosols, and these pollutants can seriously
384 impair air quality. Indeed, aerosols from biomass burning were more than likely responsible for
385 the low in SSAs found in our study during July and August.

386 **3.4 Aerosol optical properties: AAOD and AAE**

387 The AAODs at 440 nm at the seven sites were similar ~ 0.04-0.06 (Fig. 5a and Table 1),
388 and therefore, absorbing aerosols apparently were widely distributed throughout the YRD.
389 There were large uncertainties in the AAOD_{440 nm}, however; in fact, the standard deviations
390 (0.03 to 0.05) were comparable to the means, and this reflects the large temporal variability in
391 absorbing aerosol particle loadings. The average AAOD_{440 nm} at Hangzhou was about 0.02
392 higher than that at ChunAn, and this shows that the relative proportion of absorbing particles at
393 the urban area was larger than at the rural site, presumably due to greater anthropogenic
394 emissions. From Fig. 5, one can see that the monthly AAOD_{440 nm} at the urban sites from
395 March to November usually exceeded 0.05, which implies there were more absorbing species
396 in spring and autumn compared with winter (December to February) when the AAOD_{440 nm}
397 tended to be lower, < 0.05. This result suggests that the relative abundances of light-absorbing
398 particles were lower in winter compared with other seasons, and this is different from many

399 regions in northern China where the AAOs are highest in winter due to the emission of
400 absorbing particles from residential heating and other sources (Zhao et al., 2015). At the
401 ChunAn, the variations in AAOs were smoother than the more heavily impacted urban sites,
402 and the monthly mean AAOs at ChunAn were < 0.05 throughout the year.

403 The AAE can be viewed as an indicator of the type of dominant absorbing aerosol
404 particles, which include black carbon, organic matter, and mineral dust (Giles et al., 2012).
405 Generally, an AAE < 1 indicates mixing, coating, and coagulation of black carbon with organic
406 and inorganic materials; an AAE close to 1 indicates absorbing black carbon aerosols from the
407 fossil fuel burning; and an AAE > 1.10 indicates absorbing aerosols mainly from the biomass
408 burning or mineral dust (Russell et al., 2010; Bergstrom et al., 2007; Lack and Cappa, 2010).
409 The annual mean AAEs at Hangzhou, Xiaoshan, Fuyang, LinAn, Tonglu, Jiande and ChunAn
410 were 1.13 ± 0.46 , 0.88 ± 0.42 , 0.85 ± 0.43 , 0.98 ± 0.35 , 1.11 ± 0.49 , 1.16 ± 0.44 and $0.93 \pm$
411 0.31 , respectively (Table 1), and for discussion purposes, the seven sites were grouped into
412 three categories based on their average AAEs. The mean AAE values at Xiaoshan and
413 Fuyang were < 1.00 , which suggests that coated black carbon particles dominated at these
414 two sites. However, this also could be due to measurement uncertainties due to smaller
415 numbers of samples from those sites or to slightly larger values of the imaginary part of the
416 refractive index at longer wavelengths for certain particles (Bergstrom et al., 2007). There
417 would need to more observations to confirm the low AAEs at these two sites.

418 The AAE values ~ 1.00 at LinAn and ChunAn indicate that the absorbing aerosol
419 population was mainly composed of black carbon from fossil fuel burning. LinAn has a more
420 developed economy compared with Tonglu and Jiande, and in comparison, LinAn has many
421 more motor vehicles and is more heavily impacted by industrial emissions and fossil fuel
422 combustion. In contrast, the ChunAn site is located in the Qiandao Lake National Water
423 Resources Protection Zone where biomass burning and industrial activities are banned. Thus,
424 emissions from motor vehicles are probably the main source of absorbing carbon aerosols at
425 ChunAn. Finally, the AAE values at Hangzhou, Tonglu and Jiande were > 1.10 , indicating a
426 predominance of light-absorbing aerosols from either biomass burning or mineral dust.

427 Hangzhou has a population of ~10 million and more than one million vehicles, and the city can
428 be impacted by fugitive dust and biomass burning emissions (Cao et al., 2009). Tonglu and
429 Jiande have small economies compared with Hangzhou, but there is more agricultural
430 production near those two sites, and they can be impacted by biomass burning. These
431 inter-site differences in the AAEs reflect a degree of spatial heterogeneity in the distributions of
432 absorbing aerosols even though the AODs were relatively similar, and these differences likely
433 result from the many types of emission sources that can impact the sites.

434 **3.5 Direct Aerosol radiative forcing at the Earth's surface and TOA**

435 In this study, only clear sky direct aerosol radiative forcings could be investigated because
436 the aerosol microphysical and optical parameters were derived from ground-based retrievals
437 under free cloud conditions. For the direct aerosol radiative forcing calculations, the solar
438 fluxes are only evaluated for solar zenith angles (SZA) between 50° and 80°, which is where
439 the solar geometry conditions are the most appropriate for retrieving the aerosol properties
440 (Dubovik et al., 2000, 2002).

441 The annual direct DARF-BOA values under clear conditions for Hangzhou, Xiaoshan,
442 Fuyang, LinAn, Tonglu, Jiande and ChunAn were -93 ± 44 , -84 ± 40 , -80 ± 40 , -81 ± 39 , -79
443 ± 39 , -82 ± 40 and $-74 \pm 34 \text{ W/m}^2$, respectively (Figure 6a). The DARF-BOA at Hangzhou was
444 $\sim -20 \text{ W/m}^2$ lower than that at the rural ChunAn site. The DARFs at the Earth's surface and
445 TOA are governed by the aerosol microphysical and optical properties, primarily particle size
446 distributions, AODs and SSAs. The large negative DARF-BOA values at Hangzhou can be
447 attributed to the high aerosol extinction ($\text{AOD}_{440 \text{ nm}} \sim 0.76$) and small absorption (SSA at 440
448 $\text{nm} \sim 0.91$). These calculations indicate that effects of those particles on radiative fluxes can
449 cause significant surface cooling at that urban site. In comparison, at the rural ChunAn site,
450 less negative DARF-BOA values result from lower aerosol extinction ($\text{AOD}_{440 \text{ nm}} \sim 0.68$) and
451 higher scattering ($\text{SSA}_{440 \text{ nm}} \sim 0.94$).

452 The monthly DARF-BOA at the seven sites was most strongly negative in June, followed
453 by March and September. The strong cooling effect at the surface in June was due to the high
454 aerosol extinction (in section 3.2) and in particular the high volumes of fine-mode particles as

455 shown in Fig.2. Weakly absorbing particles with $SSA_{440\text{ nm}} \sim 0.90-0.95$ (Fig.6) also would
456 reduce the flux of solar radiation to the surface. A large surface cooling effect also was found in
457 March and April. Although the AOD in spring was lower than in winter, there were more
458 strongly absorbing aerosols with relatively smaller SSAs at shorter wavelengths then, and this
459 can be attributed to the presence of mineral aerosol particles. These coarse mode particles
460 with high volumes and large radii are common in spring (see section 3.1), and they can absorb
461 and scatter solar radiation and in so doing cool the surface. Surface cooling in September can
462 be explained by high aerosol extinction that resulted from the high volumes of weakly
463 absorbing fine-mode particles ($SSA_{440\text{ nm}} \sim 0.90-0.95$). These results indicate that the
464 attenuation of the solar radiation by the aerosols leads to significant surface cooling over the
465 YRD.

466 The DARF-TOA annual mean values under clear conditions were $\sim -40\text{ W/m}^2$ at all sites
467 (Figure 7a), and these negative DARF-TOA values indicate that the aerosols caused cooling
468 whole earth-atmosphere system in the YRD. This is different from the case in
469 north/northeastern China where the instantaneous DARF-TOA value can be positive in winter
470 due to the high surface reflectance of short wavelength radiation combined with atmospheric
471 heating caused by absorbing aerosols (Zhao et al., 2015, Che et al., 2014).

472 The monthly DARF-TOA means under clear conditions varied smoothly during two
473 periods: (1) from January to May, (2) from October to December (Fig. 7). The DARF-TOA
474 values were found approximate -40 W/m^2 at all sites in above two periods. However, the
475 monthly DARF-TOA means in June and September exceeded -40 W/m^2 at most sites, which
476 indicated more cooling effects on the whole earth-atmosphere system due to large aerosol
477 extinctions in YRD. The DARF-TOA means were about -20 W/m^2 at the seven sites in
478 July/August, which suggests weak cooling at that time. In contrast, the DARF-TOA values
479 under clear conditions at Shenyang (urban area of Northeastern China), Beijing (urban area of
480 Northern China), and Xianghe (rural area of Northern China) showed the negative peak during
481 June to August due to the large aerosol extinctions in summer season (Zhao et al., 2015; Xia
482 et al., 2016). As noted in section 3.3, the SSA was low in July and August, and that was

483 attributed to absorbing particles from biomass burning. Ding et al. (2016) found that large
484 quantities of black carbon can be emitted from biomass burning in the YRD during the summer,
485 and because these particles strongly absorb at infrared wavelengths, they can cause heating
486 up the atmosphere and resulting in lower negative ARF-TOA. Indeed, the positive DARF-TOA
487 we found under clear conditions from April to October were mainly due to the effects of
488 strongly absorbing particles; that is, the SSAs at 440 nm were < 0.80 , and they showed a
489 strong decrease with wavelength (not shown here) when DARF-TOA values were positive.
490 Moreover, strongly absorbing aerosol particles can heat the atmosphere column and the TOA
491 at the same time.

492 **3.6 Aerosol type classification based on the aerosol optical properties**

493 In previous studies, aerosol types have sometimes been simply classified as dust (high
494 AOD, low AE) and anthropogenic aerosols (high AOD, high AE) (Che et al., 2009a, b; Wang et
495 al., 2010). However, aerosol direct radiative forcing is affected by the absorptivity of aerosol
496 and the underlying surface conditions (Haywood and Shine, 1995), and there are advantages
497 to better characterization of the aerosol populations. For example, the classification of
498 ground-based aerosol types can be used to compare and validate aerosol types in chemical
499 transport models and satellites retrievals (Lee et al., 2010). Thus, it is advantageous to
500 categorize aerosols as absorbing or non-absorbing based on ground-based optical
501 parameters, including SSA, fine-mode fraction of AOD, and EAE etc. In this study, we used the
502 SSA, FMF, and EAE values to classify the fine and coarse mode particles from each site into
503 eight groups of particles following the method of Zheng et al. (2017). The eight types of
504 particles were (I) highly-absorbing fine-mode particles ($AE > 1.2$, $SSA \leq 0.85$); (II)
505 moderately-absorbing fine-mode particles ($AE > 1.2$, $0.85 \leq SSA < 0.9$); (III) slightly-absorbing
506 fine-mode particles ($AE > 1.2$, $0.9 \leq SSA < 0.95$); (IV) weakly-absorbing fine-mode particles
507 ($AE > 1.2$, $SSA > 0.95$); (V) mixed-absorbing particles ($0.6 \leq AE < 1.2$, $SSA \leq 0.95$); (VI) mixed
508 slightly absorbing particles ($0.6 \leq AE < 1.2$, $SSA > 0.95$); (VII) strongly absorbing coarse mode
509 particles—mainly dust ($AE \leq 0.6$, $SSA \leq 0.95$); (VIII) weakly-absorbing coarse-mode particles
510 ($AE \leq 0.6$, $SSA > 0.95$).

511

512 From Figure 8 and 9, one can see that the absorbing fine-mode particles (Type I, II and III)
513 accounted for ~30 to 50% of the aerosol in the YRD region with the FMF ~0.86-0.94%. The
514 percentages of “highly-absorbing fine-particles” (Type I) was obviously larger at Hangzhou
515 (~8%) (FMF~0.90) than the smaller city sites (~2-3%), and at the ChunAn rural site, the
516 percentages of Type I particles was only ~0.16% (FMF~0.91). This indicates that there were
517 greater emissions of strong absorbing aerosols from sources such as biomass burning and/or
518 urban/industrial activities at the urban site compared with the rural one. The proportion of
519 weakly-absorbing fine-mode particles (Type IV) varied from ~20 to 30% for all sites, and the
520 FMF varies from 0.89~0.95 at same time, which suggests the second largest aerosol type in
521 the area is weakly-absorbing fine-mode particles. The percentage of mixed absorbing particle
522 (Type V) was about ~25-26% both at Hangzhou and the rural ChunAn site, which is slightly
523 higher than that at small city sites (~10-20%) where the FMF of these particles was ~0.81-0.89.
524 The higher FMF at Hangzhou was probably due to the complex aerosol emission sources that
525 impact that megacity while at ChunAn, which is surrounded by mountains, the basin
526 topography promotes particle mixing. The mixed slightly absorbing particles (Type VI) showed
527 the highest percentages of the total aerosol number at Fuyang (18.05%) and ChunAn
528 (15.32%), and the FMF of this group varied from 0.84-0.91 at all sites. The contribution of Type
529 VI mixed slightly absorbing particles at Hangzhou was ~11.49%, and the FMF (0.86) there was
530 not as high as at ChunAn (0.91). The proportion of strongly absorbing coarse mode
531 particles—mainly dust (Group VII) was only ~1.04% of the total particle count at Hangzhou
532 while at the other sites the percent abundances were < 1%, and the FMF for these particles
533 were ~0.50-0.63 at all sites. These patterns show that the YRD region is different from regions
534 in north China, including Beijing, where dust particles contribute significantly to the coarse
535 mode absorption (Zheng et al., 2017). The percentage of weakly-absorbing coarse-mode
536 particles (Group VIII) at all sites was < 0.54%, which shows that this aerosol type was rare. In
537 addition, the FMF of Group VIII particles was 0.5–0.8 with large uncertainties at all sites.
538 Overall, this analysis of aerosol types shows that the aerosol absorption is relatively weak in
539 the YRD region, and the fine-mode makes up an especially large contribution of the absorbing

540 particles at Hangzhou.

541 **3.7 Discussion**

542 Compared with previous studies on the climatology of aerosol microphysical and optical
543 properties in China, large volumes and effective radii of fine-mode aerosol, as well as high
544 AODs at 440 nm were found at most sites in June and September but not in July or August.
545 This is remarkably different from studies conducted in Northeastern/Northern China where
546 these properties showed the maximum values in July or August (Eck et al., 2005; Zhao et al.,
547 2013; Che et al., 2015b). This reflects the distinctive climatology of aerosol microphysical and
548 optical properties in the YRD region of Eastern China. This information on the distributions and
549 variations of aerosol microphysical and optical properties obtained in our study should be
550 taken into account in the validation of satellite retrievals and aerosol modeling studies in the
551 future.

552 The $AOD_{440\text{ nm}}$ values in our study (0.71- 0.76) are similar to other urban areas of China,
553 such as Shenyang (0.75), Beijing (0.76), Tianjin (0.74), Shanghai (0.70), and Hefei (0.69) (Pan
554 et al., 2010; He et al., 2012; Zhao et al., 2013; Che et al., 2015b, Liu et al., 2017). This
555 indicates that high aerosol loadings caused by anthropogenic activities occur over many urban,
556 suburban, and even rural areas of eastern China. The $AOD_{440\text{ nm}}$ at the rural ChunAn site
557 (~ 0.68) is $\sim 2\text{--}6$ times higher than in other rural sites in China, such as Longfengshan (0.35;
558 northeastern China), Xinglong (0.28, northern China), Akedala (0.20, northwestern China) and
559 Shangri-La (0.11, southwestern China) (Wang et al., 2010; Che et al., 2011; Zhu et al., 2014;
560 Che et al., 2015b). Therefore, strong aerosol effects on light extinction occur not only in urban
561 areas but also in much or possibly all of the YRD. The $SSA_{440\text{ nm}}$ in our study ranged from
562 0.91–0.94, which is similar to other regions of China, such as Wuhan (0.92), Beijing (0.89) and
563 Xinglong (0.92) (Wang et al., 2015; Xin et al., 2014; Zhu et al., 2014); this suggests aerosol
564 particles in the YRD are slight to relatively strong absorbers of violet/indigo wavelengths.

565 Although the AAE has been used as an indicator of the dominant absorbing aerosol type,
566 the SSA, FMF, and EAE also used to classify the absorption characteristics of fine and coarse

567 mode particles. It should also be mentioned that uncertainties in the AAE calculations due to
568 uncertainties in the SSAs may have contributed to the observed differences between sites
569 (Bergstrom et al., 2007; Giles et al. 2011). Ideally, the aerosol physical and chemical
570 characteristic measurements will be combined at some point to more definitively classify the
571 aerosol types. In addition, even though Zhuang et al. (2017) pointed that the DARFs are not
572 very sensitive to vertical profiles under clear sky condition, future research should take into
573 account the vertical distributions of aerosols to more accurately assess the direct aerosol
574 radiative forcing effects.

575 The useful observations at most sites in this study were made on 439~562 days, but
576 smaller numbers of observation were made in Xiaoshan and Fuyang because of instrument
577 failures, 180 and 217 days, respectively. There were fewer than 15 days of useful data for
578 October and November at Fuyang, and for June, July, September, and November at
579 Xiaoshan. While the available data do provide insights into the aerosol characteristics at
580 the sites, more extended observations should be conducted at the two sites in future.

581 **4. Conclusions**

582 A detailed study of aerosol microphysical and optical properties retrieved from
583 synchronous ground-based sun photometer observations was conducted at seven sites in the
584 Yangtze River Delta region of Eastern China from 2011 to 2015. The aerosol were classified
585 into eight types, and calculations were made for the direct aerosol radiative forcing (DARF) at
586 the top and bottom of the atmosphere. The conclusions of the study can be summarized as
587 follows:

588 A relatively homogeneous distribution of aerosol microphysical properties was found for a
589 megacity, five small cities, and a rural site in the YRD region. High particle volumes of
590 coarse-mode aerosol occurred in March to May, which reflects the existence of large mineral
591 particles from springtime dust storms. High volumes and large effective radii of fine-mode
592 aerosol in June and September were found at most sites, and this was attributed to aerosol
593 hygroscopicity and cloud processing. The low volumes and large bandwidth of both fine- and

594 coarse-mode aerosol found in July and August at all sites was explained by the wet removal of
595 coarse particles during the heavy precipitation in June and the influences of subtropical
596 anticyclones in summer.

597 The $AOD_{440\text{ nm}}$ generally decreased from the east coast (0.76–Hangzhou) to the inland
598 areas towards the west (0.68–ChunAn), and this can be explained by anthropogenic impacts
599 of the more urbanized regions in the east YRD region. The $AOD_{440\text{ nm}}$ values at urban and rural
600 sites of YRD were 0.68–0.76, and the fine-mode fractions were ~ 0.90 , which indicates that
601 fine-mode particles were more important than the coarse mode for the light extinction. The
602 difference in AODs between the urban and rural sites was less than 10%, and this can be
603 explained by somewhat stronger effects of anthropogenic activities in the urban area. The
604 monthly averaged AODs at 440 nm showed peaks in June and September that resulted from
605 increases fine-mode aerosol particles. However, AODs at 440 nm in July and August were the
606 lowest over the year, and that was related to conditions favorable for aerosol dispersion. The
607 mean extinction Angström exponent was > 1.20 all year indicating that small particles were
608 predominant in the region. The SSAs at 440 nm varied from 0.91 to 0.94 at the urban and rural
609 sites, indicating that the aerosol particles were moderately absorbing, and this is almost surely
610 a result of impacts from the high industrial emissions and other anthropogenic activities in the
611 region. There was an obvious wavelength dependence for the SSA in July and August, and
612 aerosols absorbed most strongly at infrared wavelengths. The AAODs at 440 nm at the seven
613 sites were ~ 0.04 – 0.06 , which suggests that absorbing aerosols are distributed more-or-less
614 homogeneously throughout the YRD region. The averaged $AAOD_{440\text{ nm}}$ at Hangzhou is about
615 0.02 higher than that at ChunAn, which indicates the relative proportion of absorbing particles
616 in the urban area is larger compared with the rural area. The $AAOD_{440\text{ nm}}$ in winter was < 0.05 ,
617 which suggests that light absorption by the particles was low compared with the other seasons.
618 The geographical variability in the distributions of the AAEs suggests that the absorbing
619 aerosols possibly have different optical characteristics related to the local emission sources in
620 the YRD.

621 The DARF-BOA at Hangzhou was $\sim -20\text{ W/m}^2$ lower than that at the rural ChunAn site,

622 which shows stronger aerosol cooling at the megacity. The monthly DARF-BOA was strongly
623 negative in June due to the high aerosol extinction and especially the high fine-mode volume.
624 The DARF-TOAs under clear conditions were $< -40 \text{ W/m}^2$, but about -20 W/m^2 in July/August
625 which suggests weaker cooling in mid summer. The DARF-TOAs were positive from April to
626 October when the $\text{SSA}_{440 \text{ nm}} < 0.80$, and the greater effects at shorter wavelengths was likely
627 due to emissions of carbonaceous particles from the burning of crop residues.

628 The SSA, FMF, and EAE values were used to classify the particles as absorbing or
629 non-absorbing. Relatively large emissions of strongly absorbing aerosols in the Hangzhou
630 urban area was the result of biomass burning and/or urban/industrial activities. The aerosol
631 type classification showed overall that the aerosol absorption is weak to moderate in the YRD,
632 and the fine-mode has a large contribution to the higher percentage of absorbing particles at
633 the Hangzhou site.

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960 Figure captions

961 Figure 1. Locations and elevations of the seven CARSNET sites in the Yangtze River Delta.

962 Figure 2. Temporal variations in the aerosol volume-size distributions at (a) Hangzhou, (b)
963 Xiaoshan, (c) Fuyang, (d) LinAn, (e) Tonglu, (f) Jiande and (g) ChunAn.

964 Figure 3. Variations in the total, fine- and coarse-mode $AOD_{440\text{ nm}}$ at (a) Hangzhou, (b)
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968 Figure 4. Variations in the single scattering albedo at 440, 670, 870, and 1020 nm at (a)
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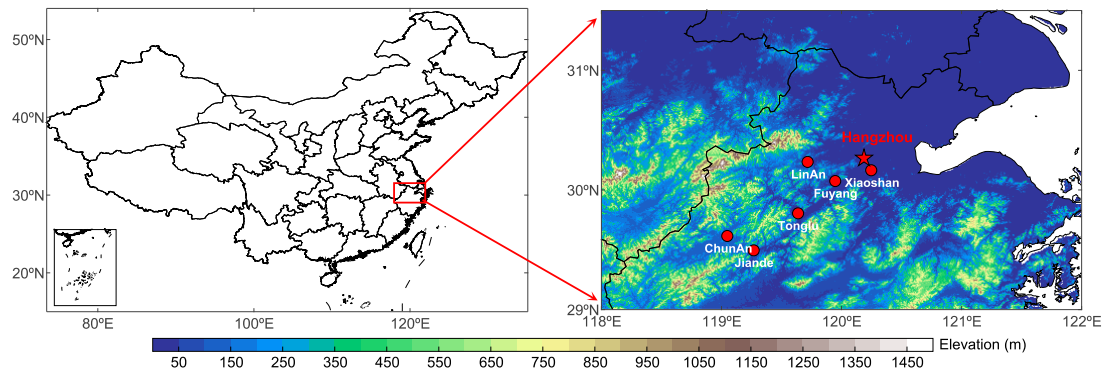
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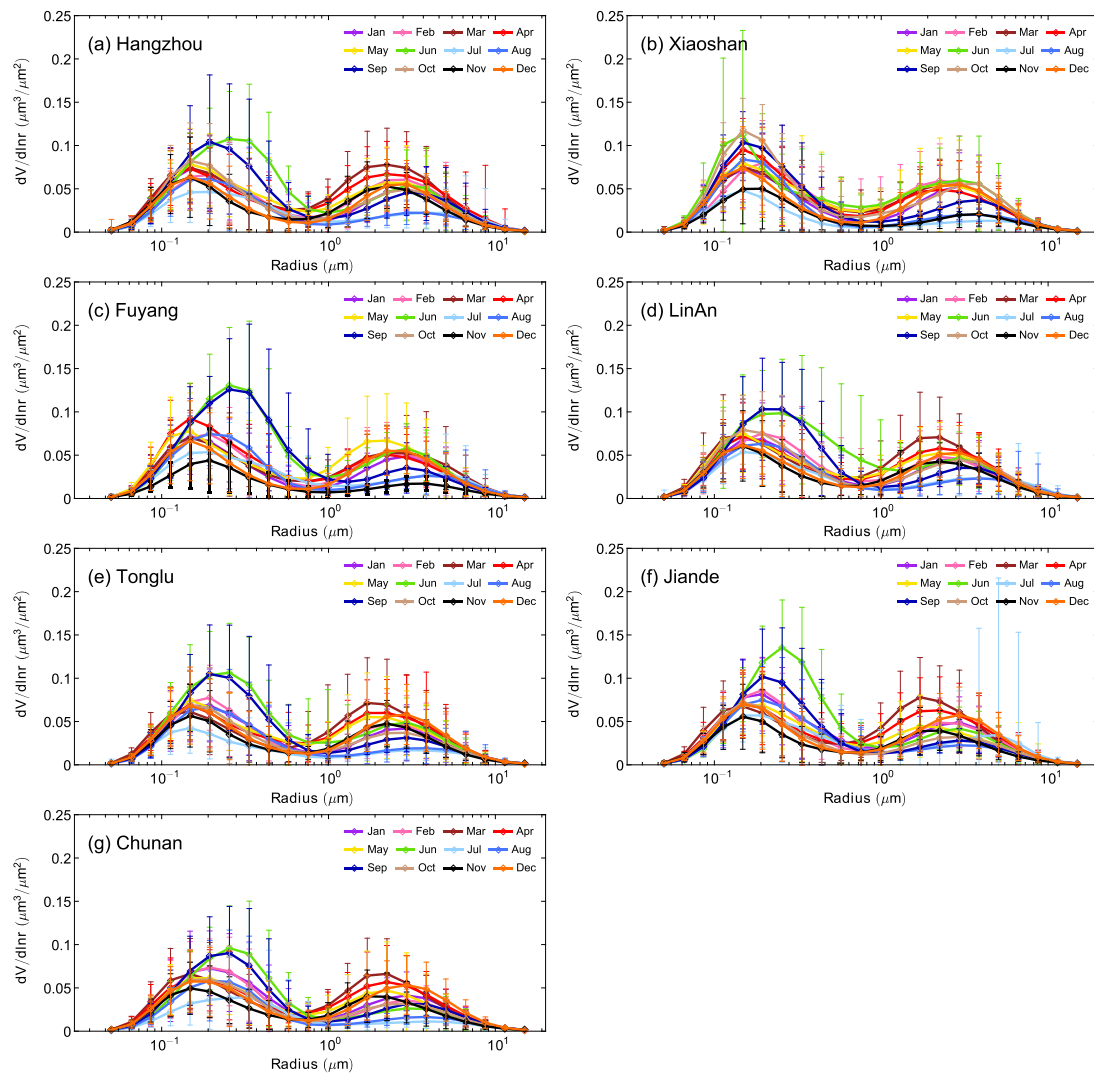
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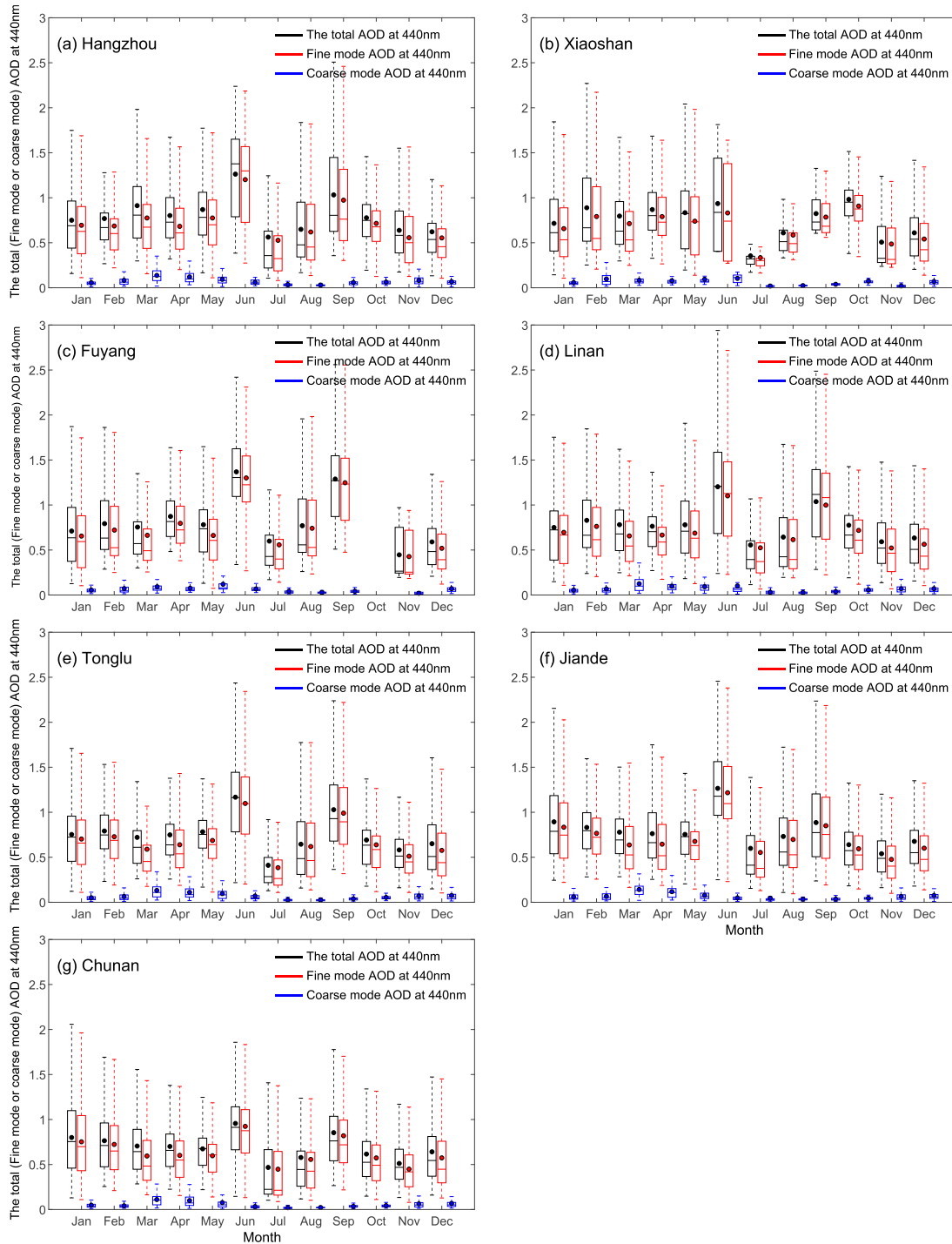


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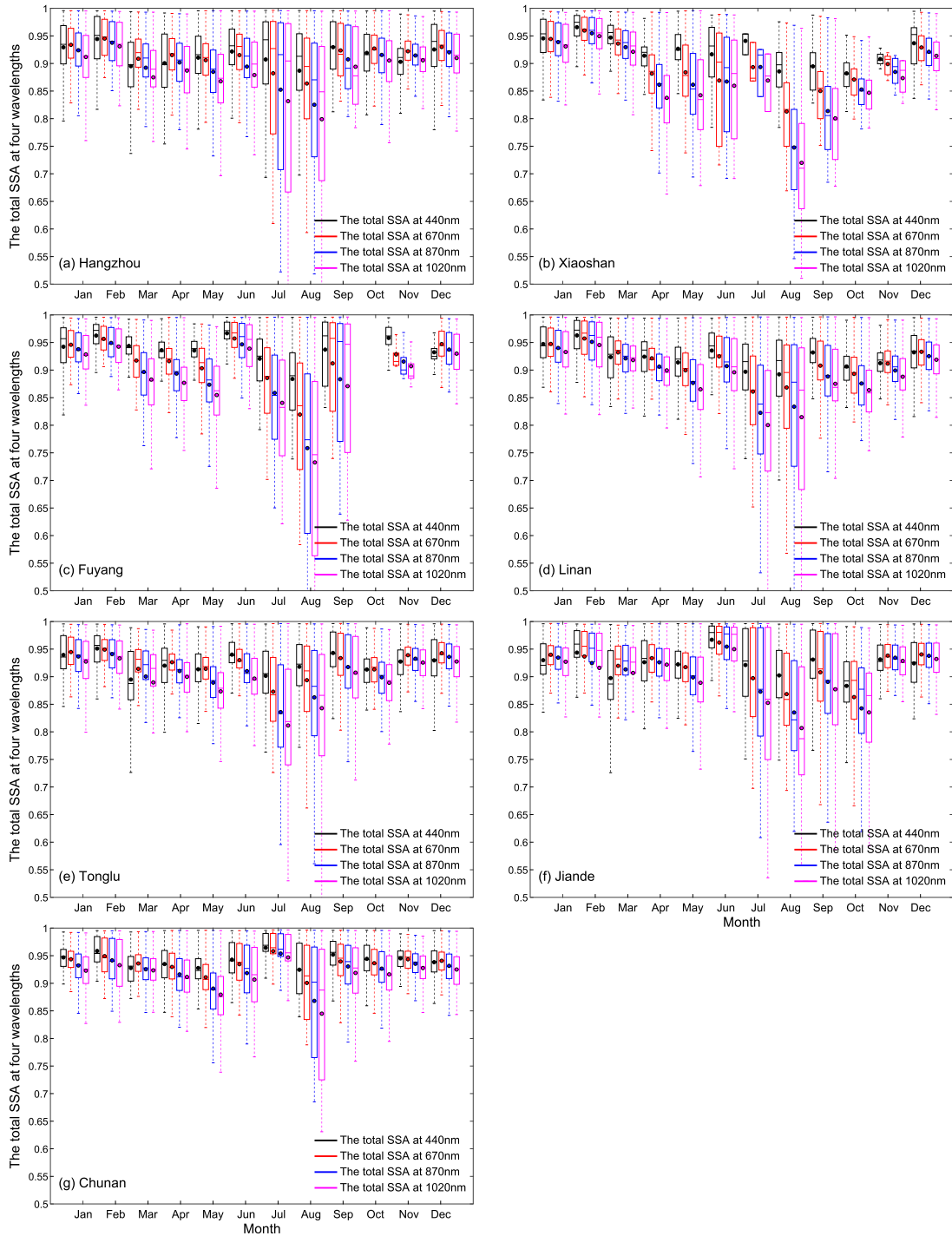
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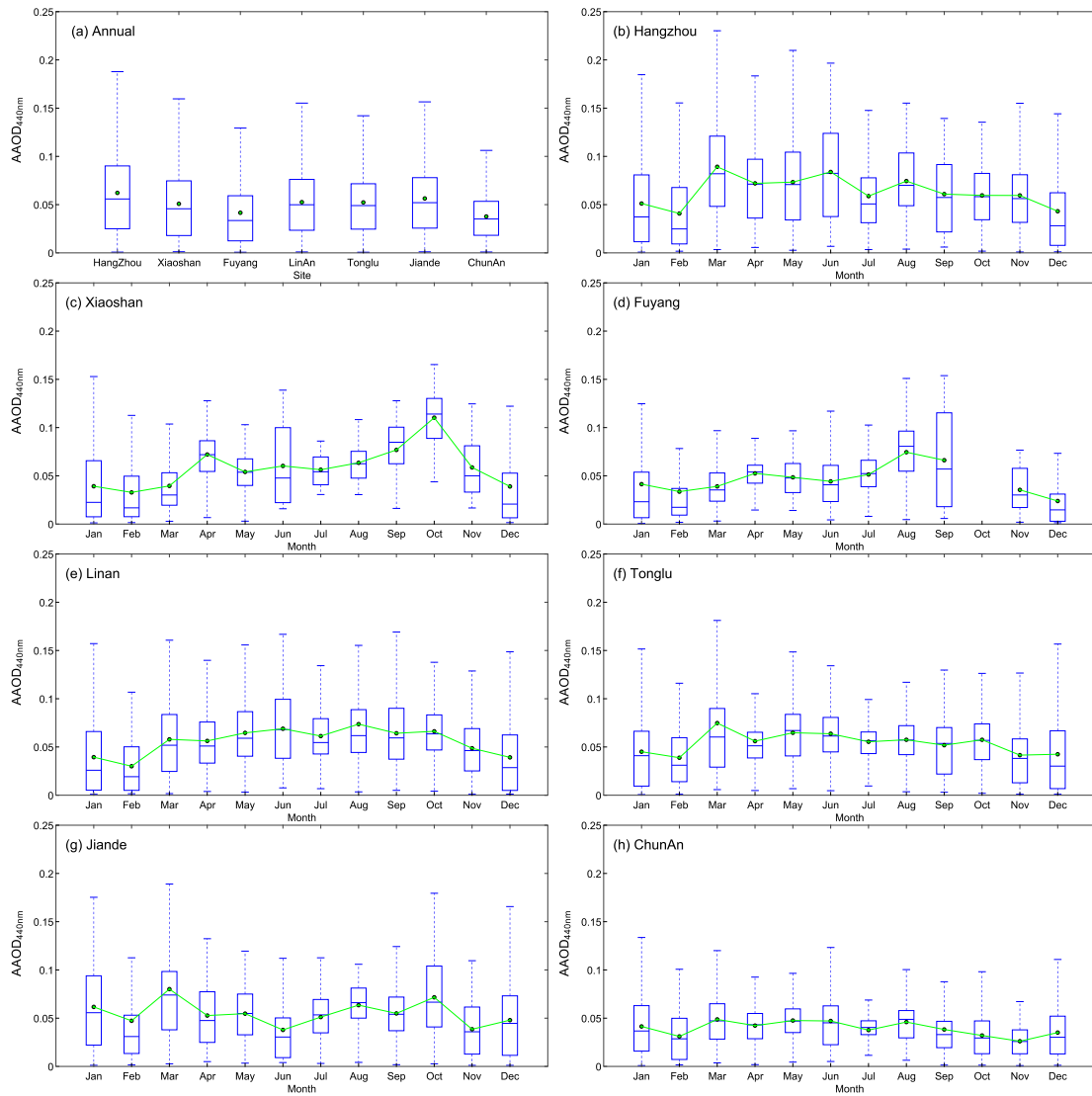
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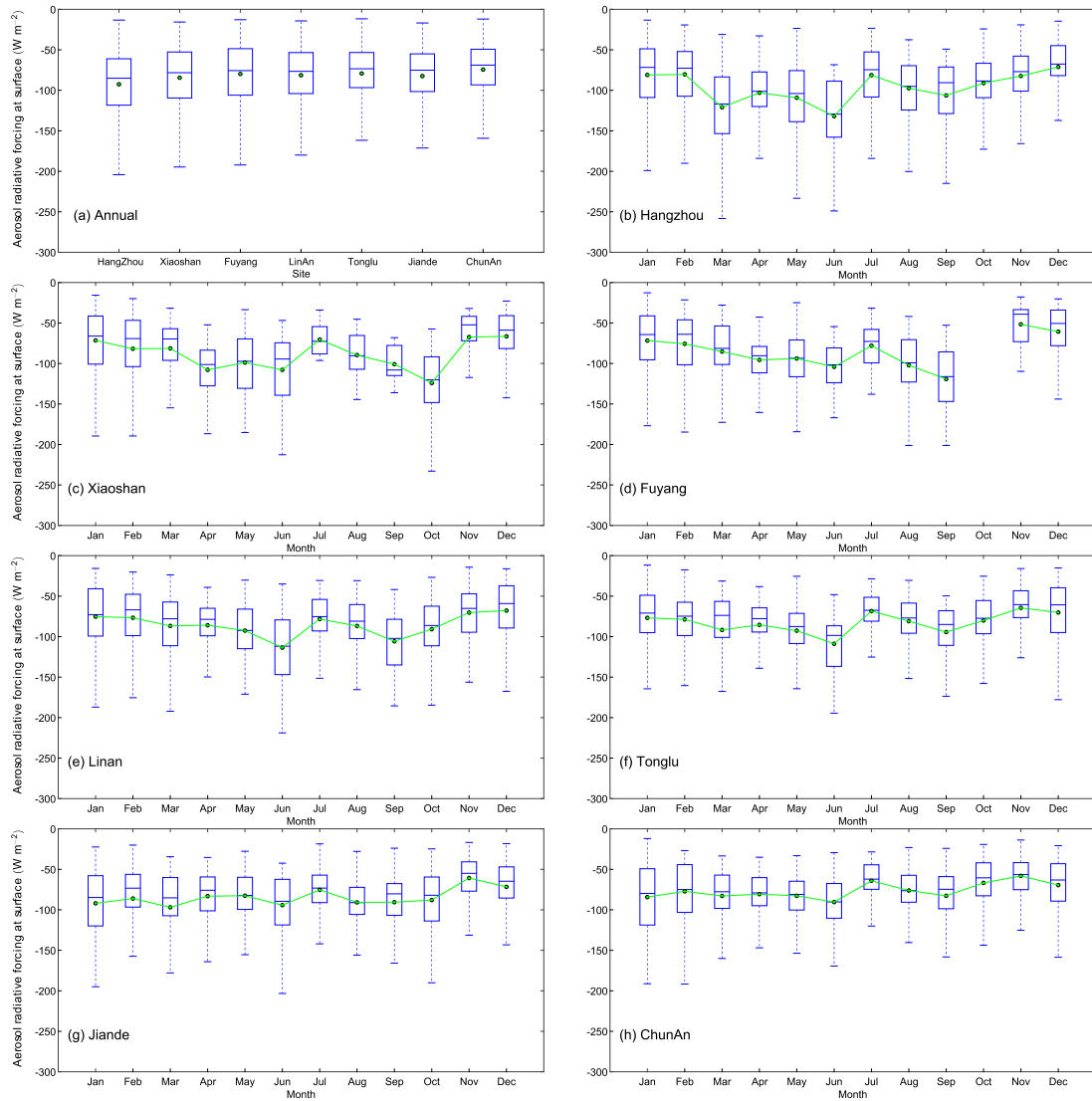
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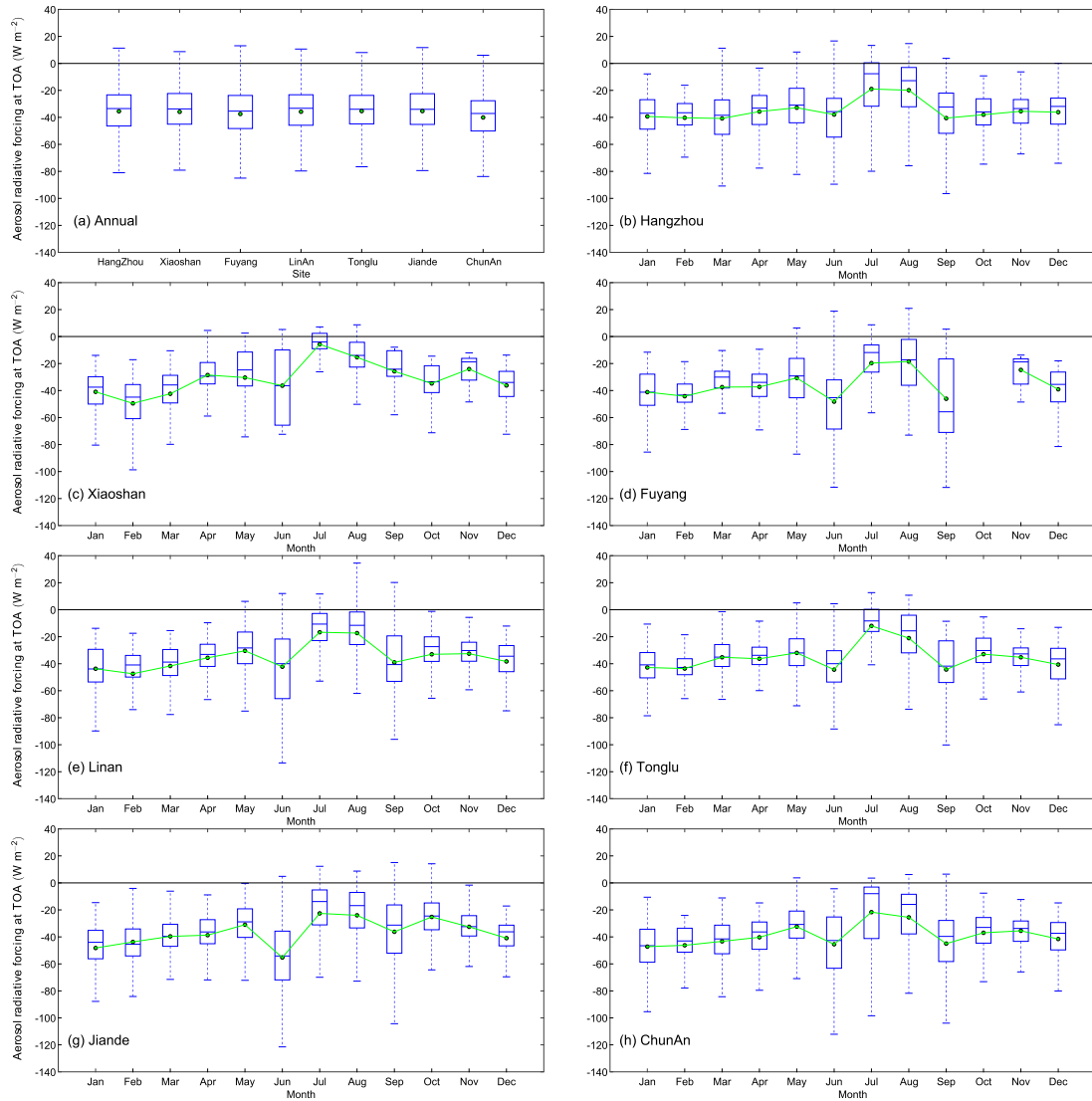
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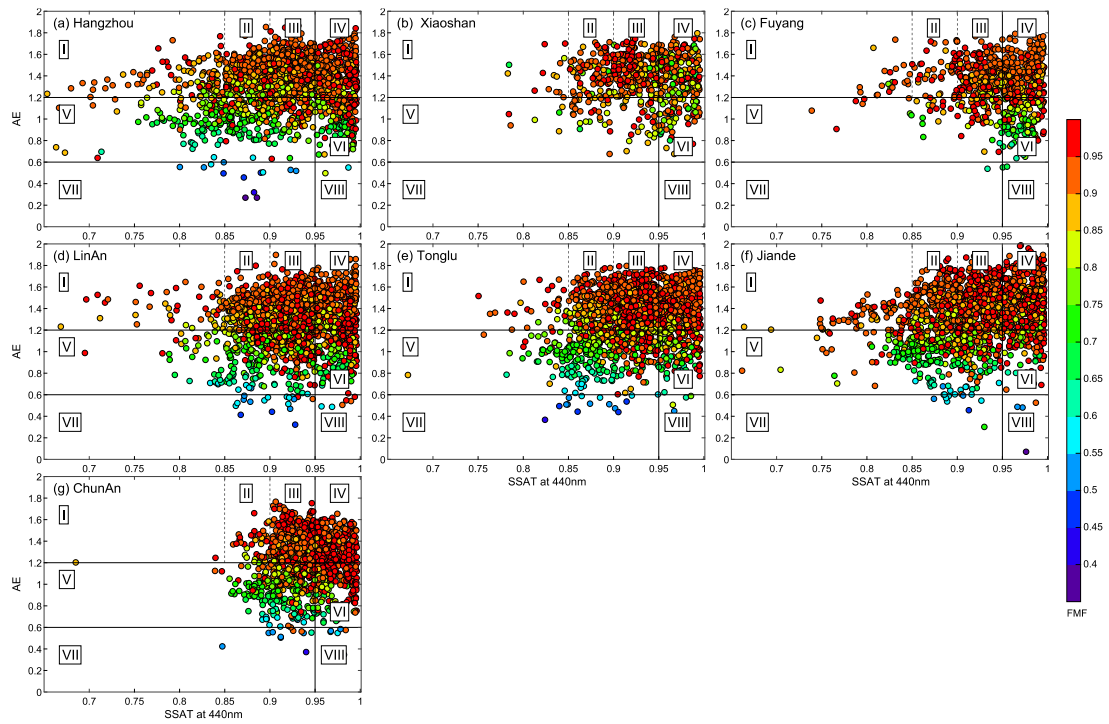
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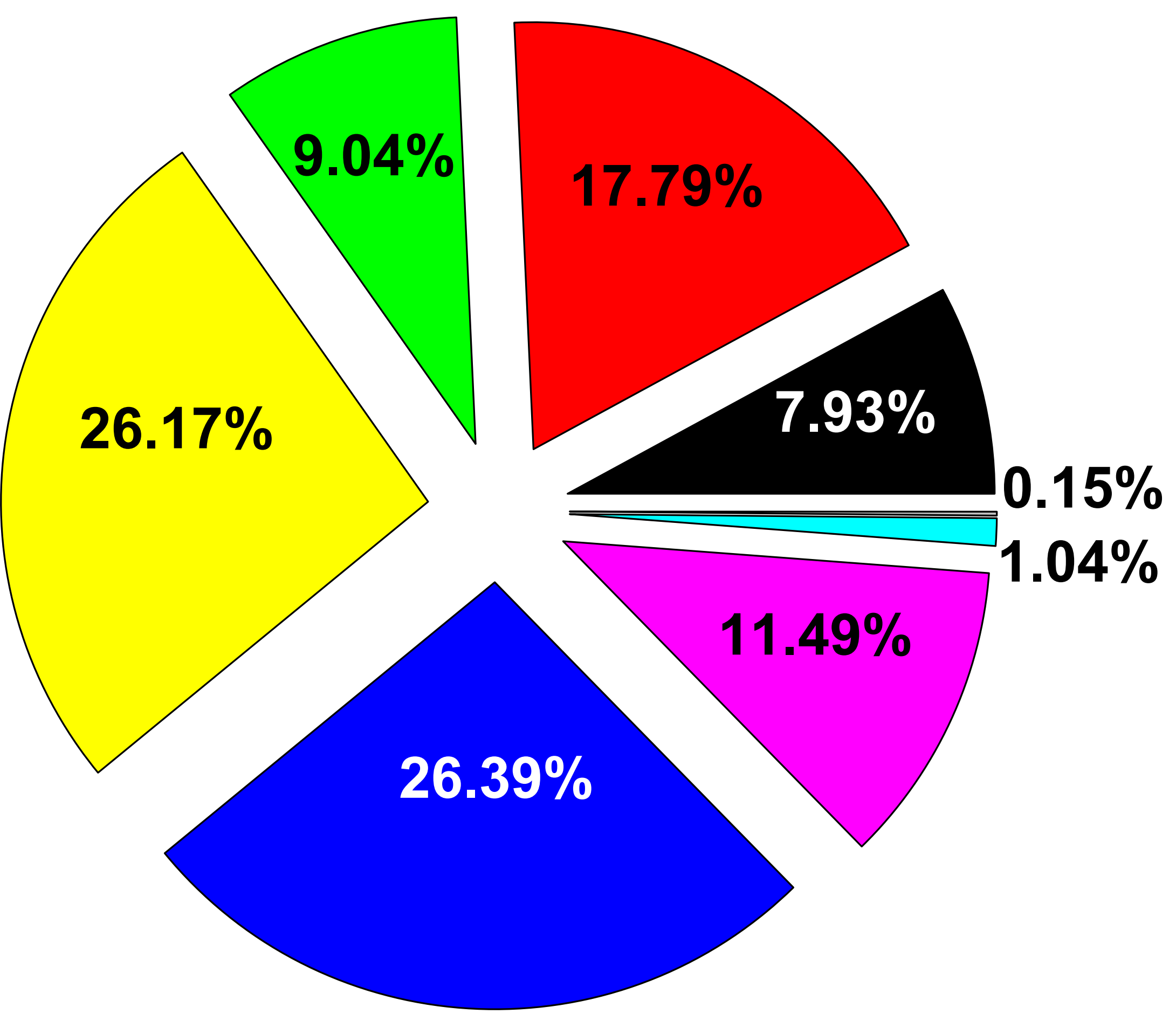


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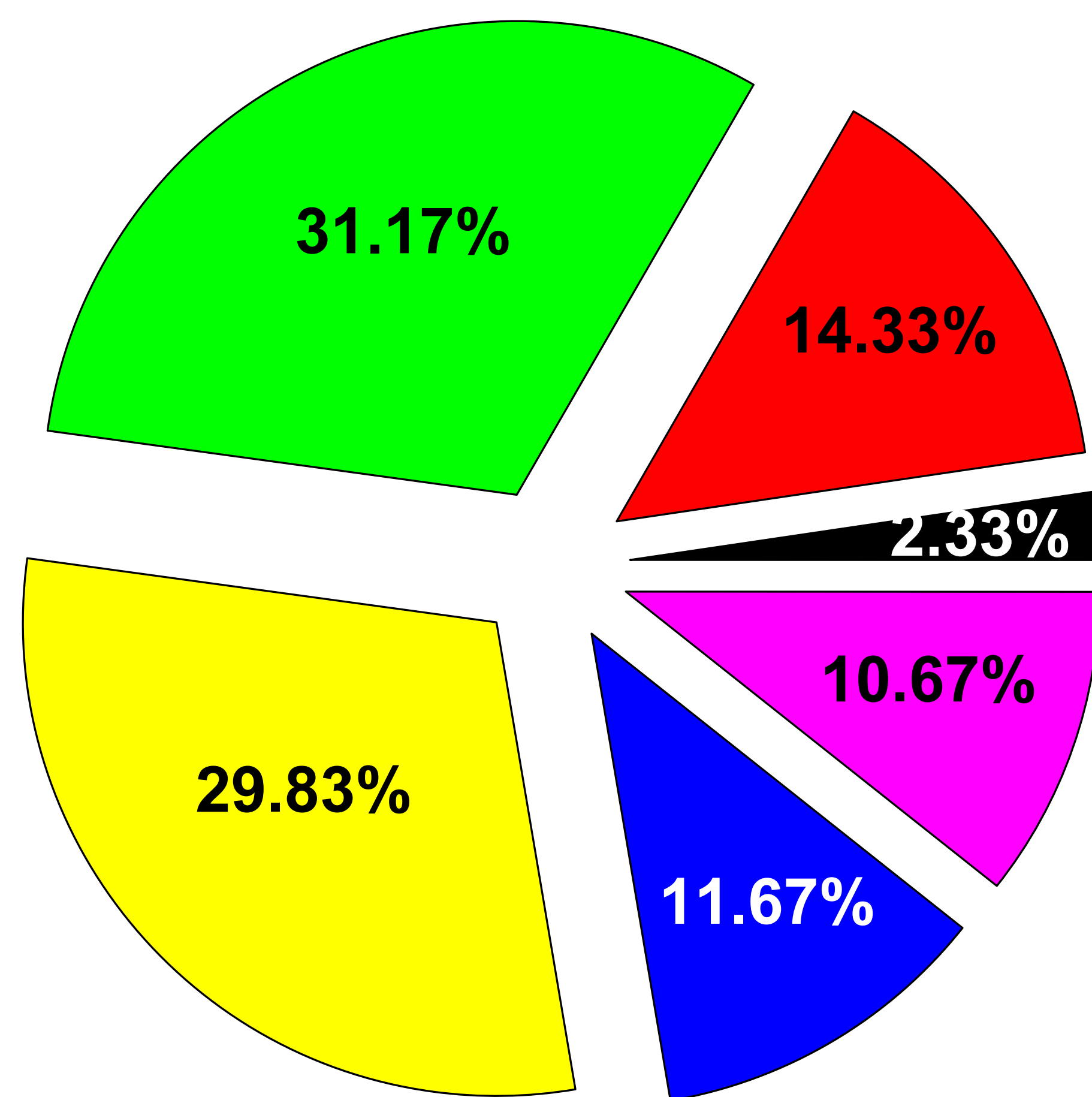
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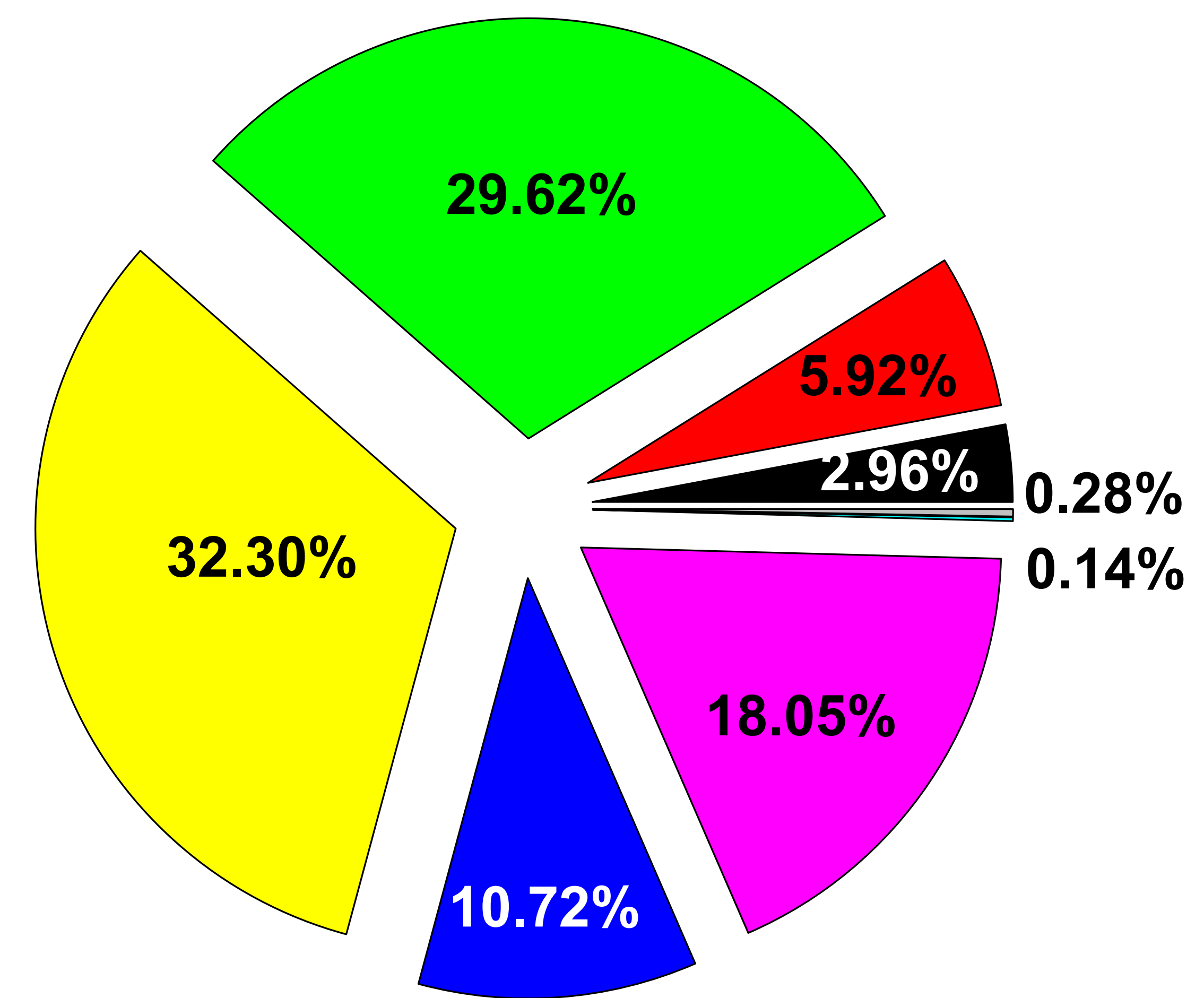
(a) Hangzhou



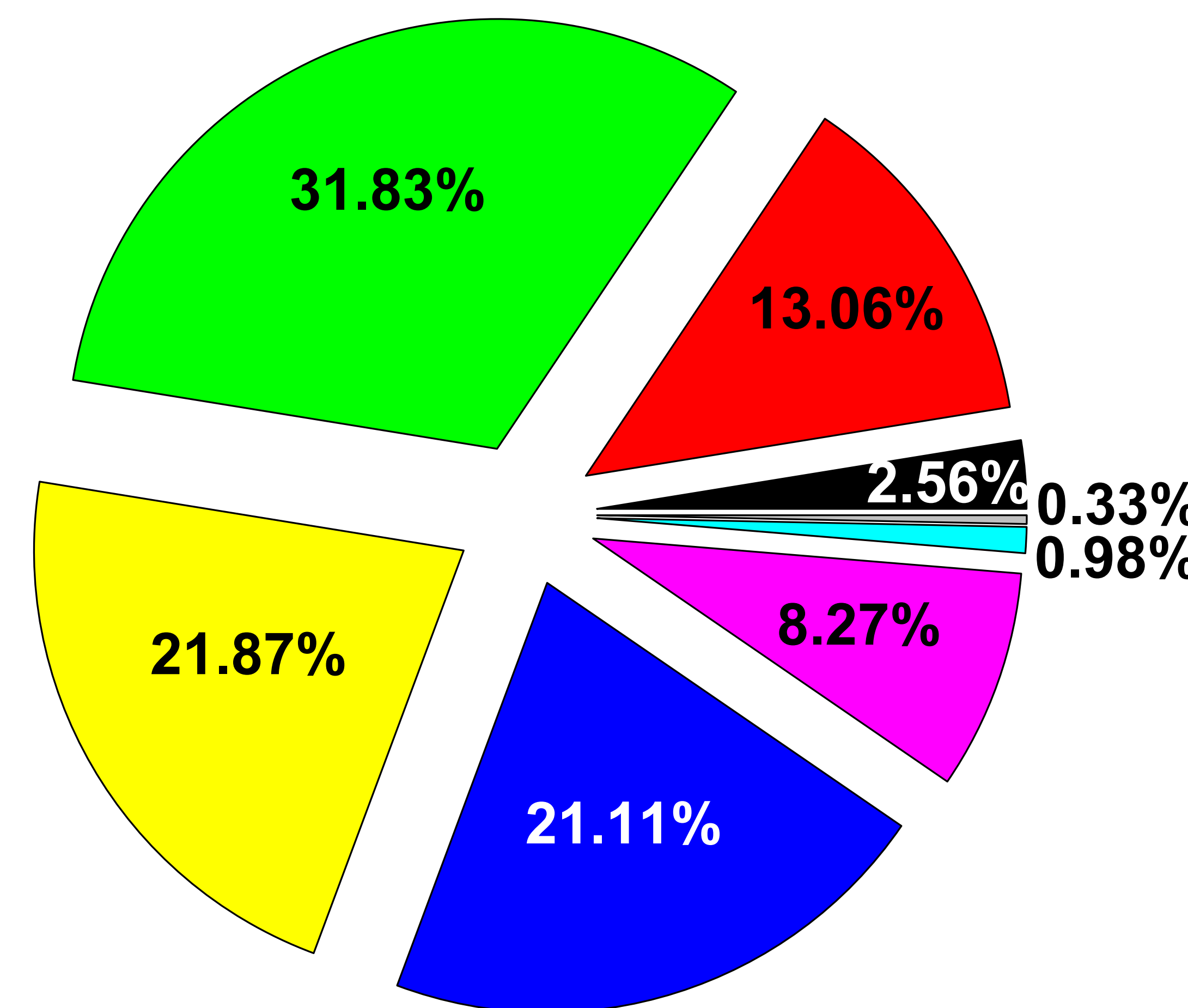
(b) Xiaoshan



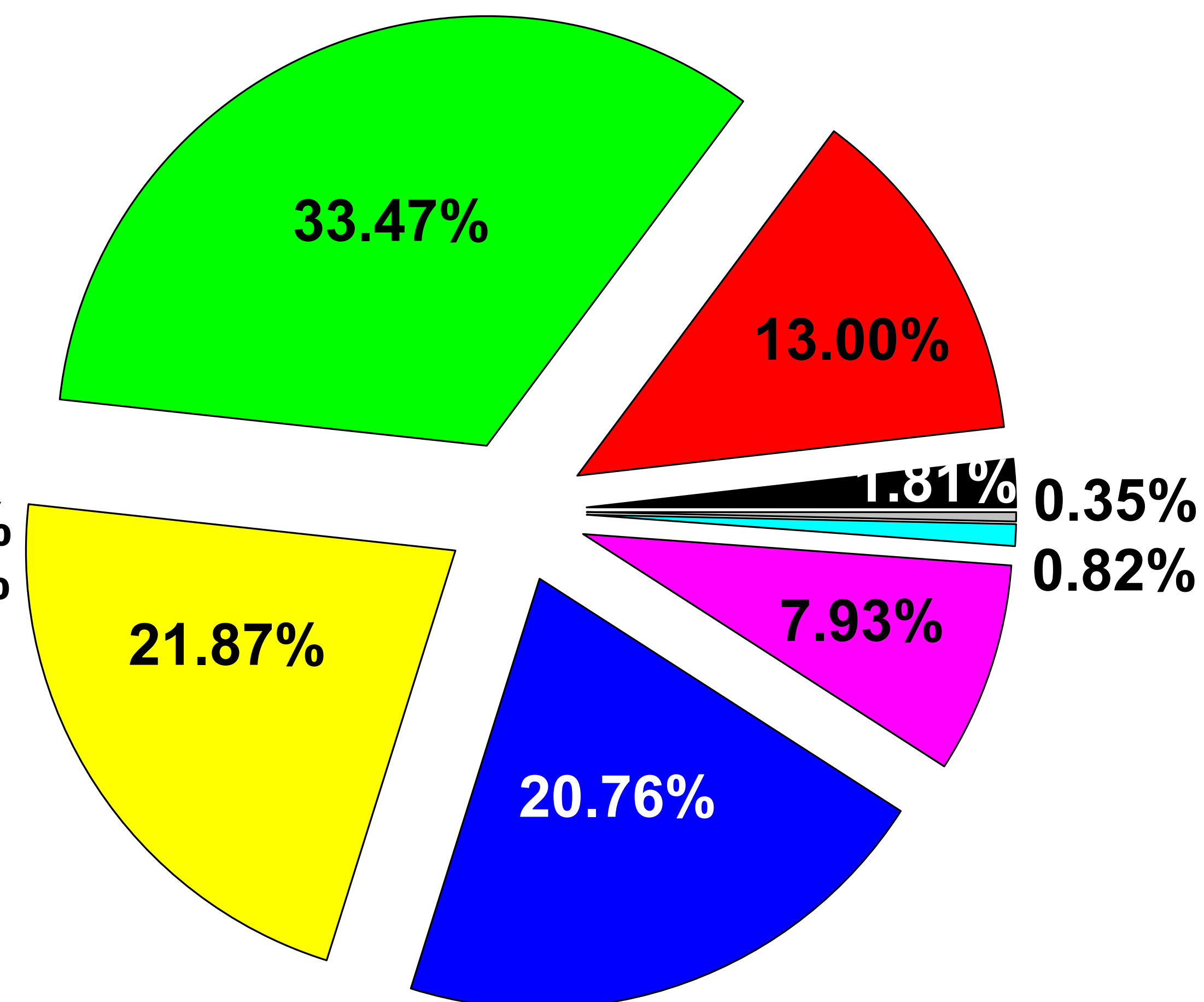
(c) Fuyang



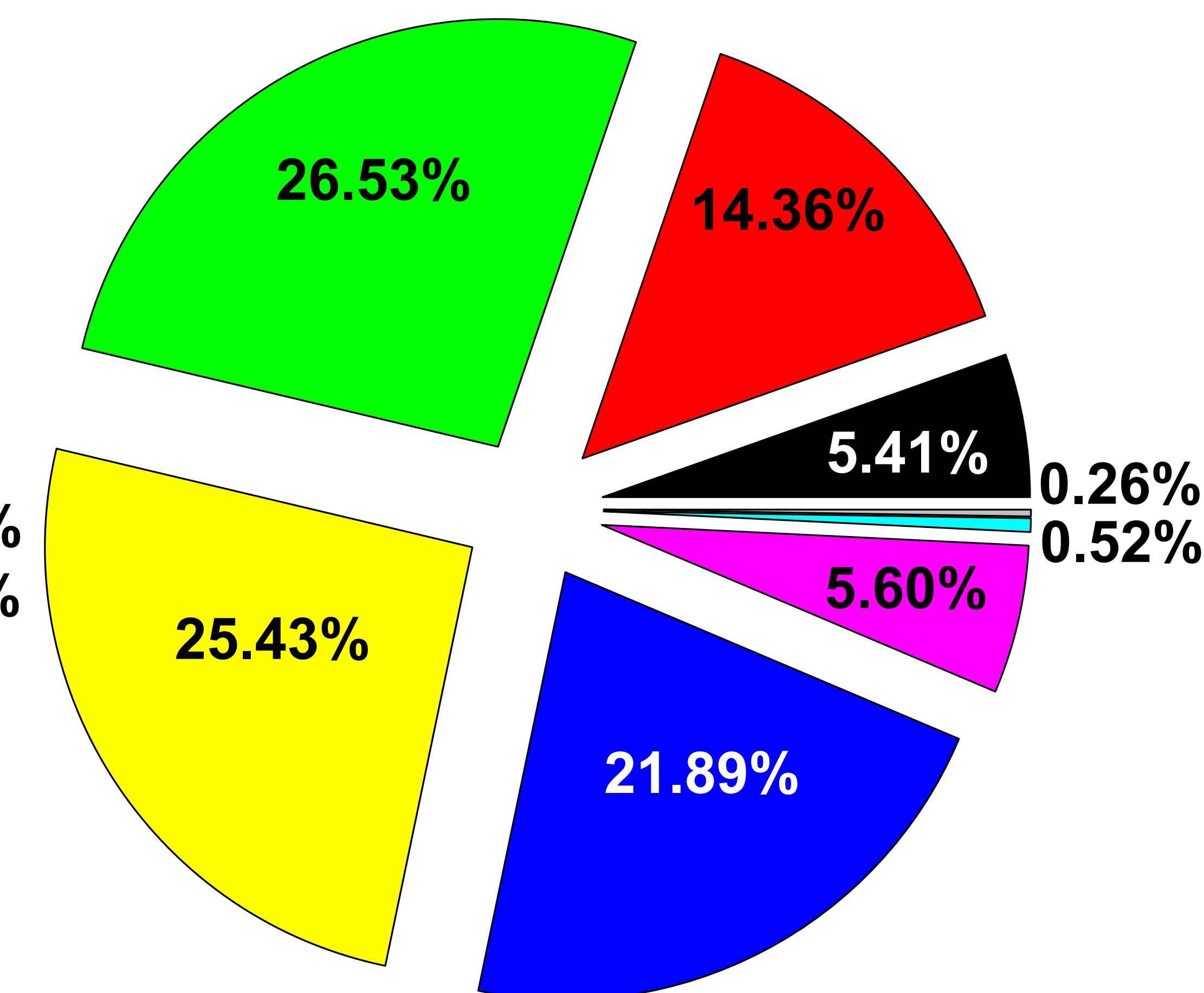
(d) LinAn



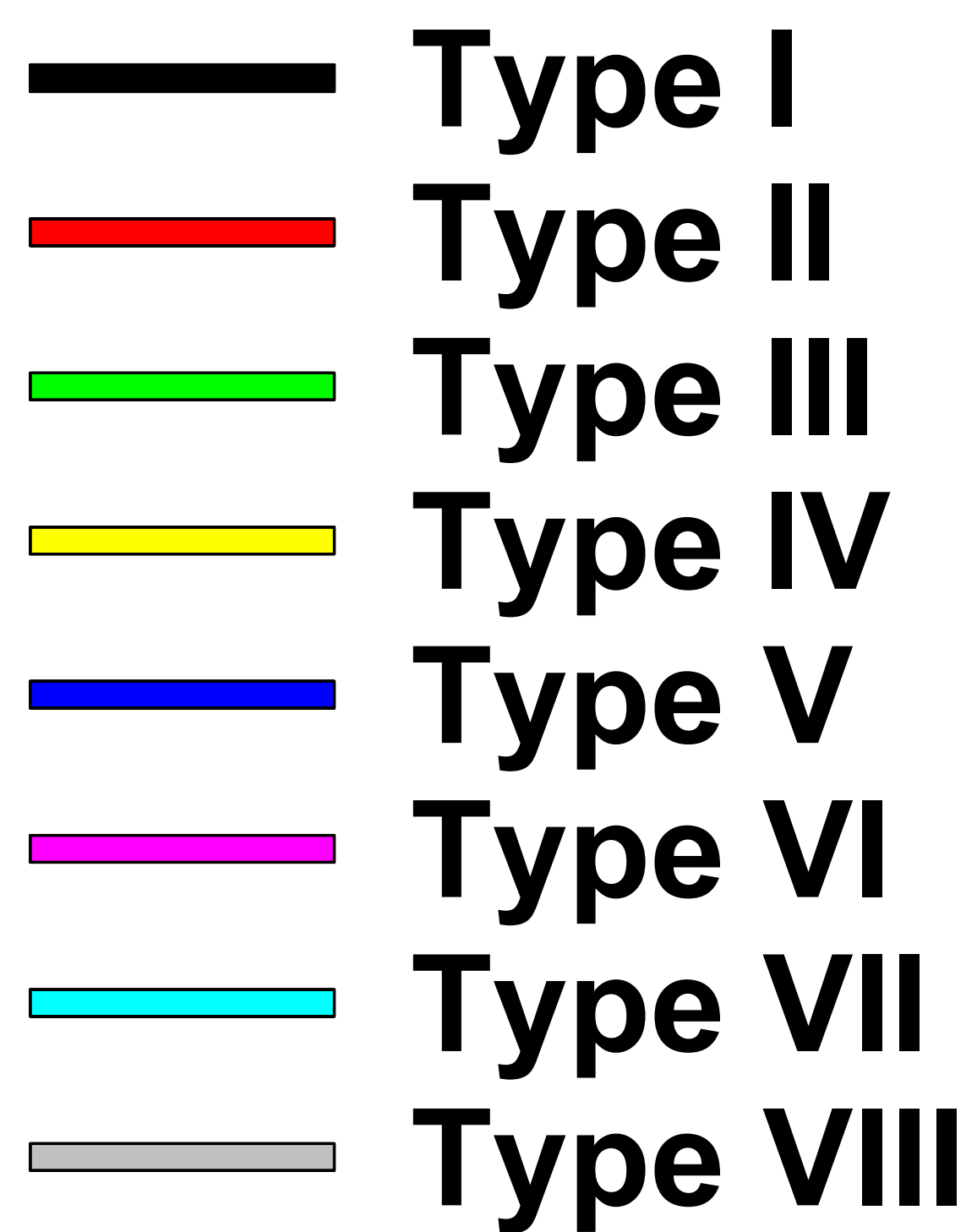
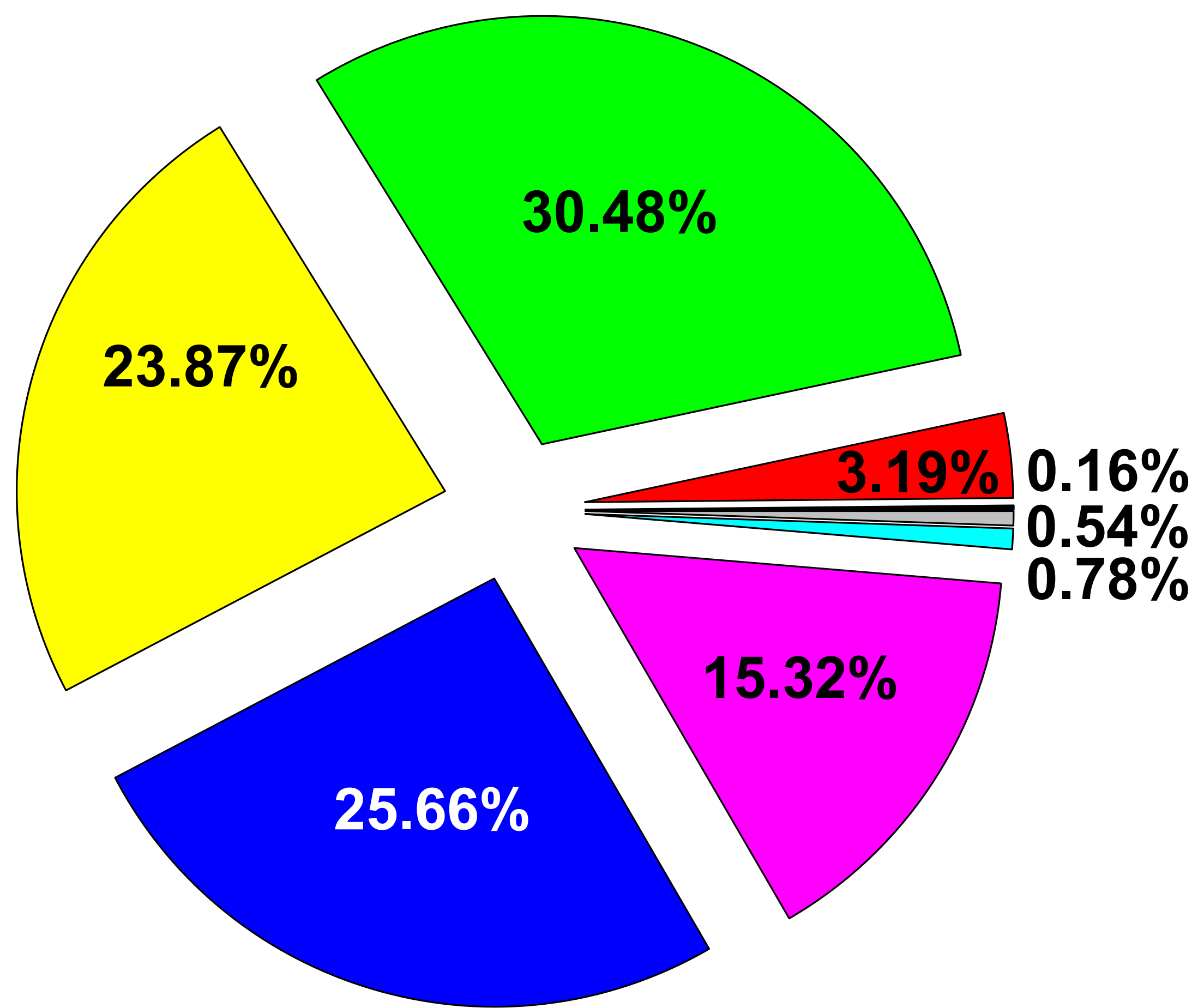
(e) Tonglu

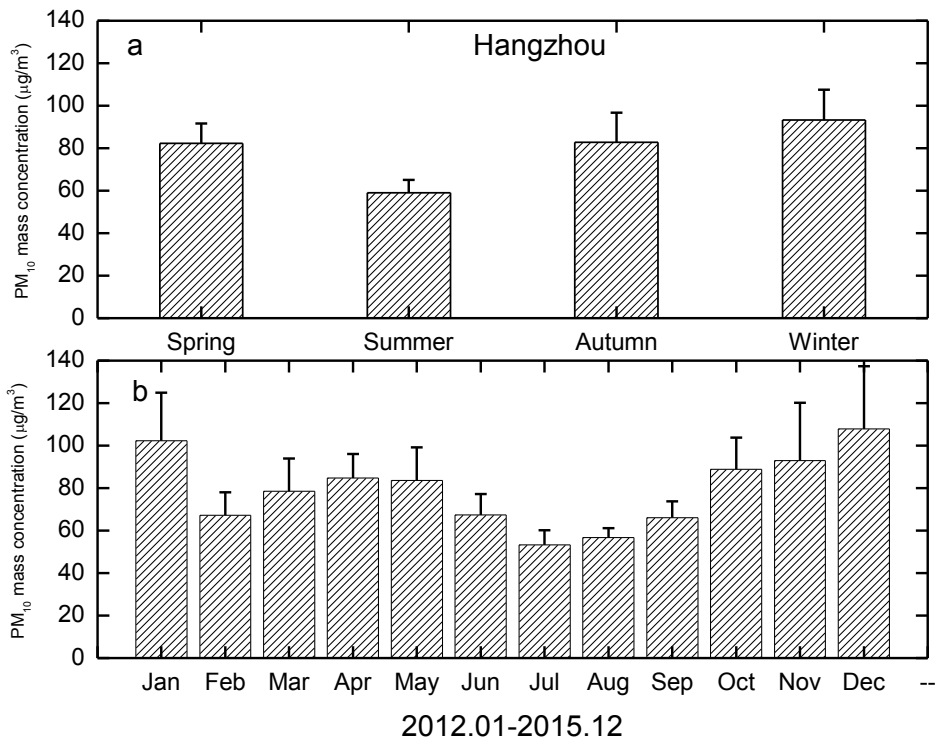


(f) Jiande



(g) ChunAn





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1035 Seasonal and monthly variations of PM₁₀ mass concentrations (µg/m³) in Hangzhou

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during 2012-2015

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