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Spectral- and size-resolved mass absorption efficiency of mineral dust aerosols in the shortwave: a simulation chamber study

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Abstract

This paper presents new laboratory measurements of the mass absorption efficiency (MAE) between 375 and 850 nm for mineral dust of different origin in two size classes: $PM_{10.6}$ (mass fraction of particles of aerodynamic diameter lower than 10.6 µm) and $PM_{2.5}$ (mass fraction of particles of aerodynamic diameter lower than 2.5 µm). Experiments have been performed in the CESAM simulation chamber using generated mineral dust from natural parent soils, and optical and gravimetric analyses.

Results show that the MAE values are lower for the $PM_{10.6}$ mass fraction (range 37-135 10^{-3} m^2 g^{-1} at 375 nm) than for the $PM_{2.5}$ (range 95-711 10^{-3} m^2 g^{-1} at 375 nm), and decrease with increasing wavelength as λ^{-AAE}, where Angstrom Absorption Exponent (AAE) averages between 3.3-3.5, regardless of size. The size-independence of AAE suggests that, for a given size distribution, the possible variation of dust composition with size would not affect significantly the spectral behavior of shortwave absorption. Because of its high atmospheric concentration, light-absorption by mineral dust can be competitive to black and brown carbon even during atmospheric transport over heavy polluted regions, when dust concentrations are significantly lower than at emission. The AAE values of mineral dust are higher than for black carbon (~1), but in the same range as light-absorbing organic (brown) carbon. As a result, depending on the environment, there can be some ambiguity in apportioning the AAOD based on spectral dependence, which is relevant to the development of remote sensing of light-absorption aerosols from space, and their assimilation in climate models. We suggest that the sample-to-sample variability in our dataset of MAE values is related to regional differences of the mineralogical composition of the parent soils. Particularly in the $PM_{2.5}$ fraction, we found a strong linear correlation between the dust light-absorption properties and elemental iron rather than the iron oxide fraction, which could ease the application and the validation of climate models that now start to include the representation of the dust composition, as well as for remote sensing of dust absorption in the UV-VIS spectral region.

1. Introduction

Mineral dust aerosols emitted by wind erosion of arid and semi-arid soils account for about 40% of the total emitted aerosol mass per year at the global scale (Knippertz and Stuut, 2014). The episodic but frequent transport of intense mineral dust plumes is visible from spaceborne sensors as their high concentrations, combined to their ability of scattering and absorbing solar and thermal radiation, give raise to the highest registered values of aerosol optical depth (AOD) on Earth (Chiapello, 2014). The instantaneous radiative efficiency of dust particles, that is, their radiative effect per unit AOD, is of the order of tenths to hundreds of W m^{-2} AOD^{-1} in the solar spectrum, and of the order of order of tenths W m^{-2} AOD^{-1} in the thermal infrared (e.g., Haywood et al., 2003; di Sarra et al., 2011; Slingo et al., 2006 and the compilation of Highwood and Ryder, 2014). In the solar spectrum, (Boucher et al., 2013). Albeit partially compensated by the radiative effect in the thermal infrared, the global mean radiative effect of mineral dust in the shortwave is negative both at the surface and the top of the atmosphere (TOA) and local warming of the atmosphere. Many are the consequences on the global and regional climate, that ultimately feed back on wind speed and vegetation and therefore on dust emission (Tegen and Lacis, 1996; Solmon et al., 2008; Pérez et al., 2006; Miller et al., 2014). Dust particles perturb the surface air temperature through their radiative effect at TOA, can increase the atmospheric stability (e.g., Zhao et al. 2011) and might affect precipitation at the global and regional scale (Solmon et al., 2008; Xian, 2008; Vinoj et al., 2014; Miller et al., 2014 and references therein).

All models show that the effect of mineral dust on climate has a great sensitivity to the shortwave absorption properties of mineral dust (Lau et al., 2009; Loeb and Su, 2010; Ming et al., 2010; Perlwitz and Miller, 2010). Absorption by mineral dust started receiving a great deal of interest in the last ten years or so, when spaceborne and ground-based remote sensing studies (Dubovik et al., 2002; Colarco et al., 2002; Sinyuk et al., 2003) suggested that mineral dust was less absorbing that it had been indicated by in situ observations (e.g. Patterson et al., 1977; Haywood et al., 2001), particularly at wavelengths below 600 nm. Balkanski et al. (2007) showed that lowering the dust absorption properties to an extent that reconcile them both with the remote-sensing observations and the state-of-knowledge of the mineralogical composition, allowed calculating clear-sky dust shortwave radiative effect in agreement with satellite-based observations. A significant body of observations have been performed in quantifying the shortwave light-absorbing properties of mineral dust, by direct measurements (Alfaro et al., 2004; Linke et al., 2006; Osborne et al., 2008; McConnell et al., 2008; Derimian et al., 2008; Yang et al., 2009; Müller et al., 2009; Petzold et al., 2009; Formenti et al., 2011; Moosmüller et al., 2012; Wagner et al., 2012; Ryder al., 2013a; Utry et al., 2015; Denjean et al., 2015c; 2016), and indirectly, by quantifying the amount and the speciation of the light-absorbing compounds in mineral dust, principally iron oxides (Lafon et al., 2004; 2006; Lazaro et al., 2008; Derimian et al., 2008; Zhang et al., 2008; Kandler et al., 2007; 2009; 2011; Formenti et al., 2014a; 2014b).

However, existing data are often limited to a single wavelength, which moreover are not identical for all experiments. Also, frequently they do not represent the possible regional variability of the dust absorption, either because they are obtained from field measurements integrating the contributions of different source regions, or conversely, by laboratory investigation targeting samples from a limited number of locations. This might lead to biases. Indeed, iron oxides in mineral dust, mostly in the form of hematite (Fe_2O_3) and goethite (Fe(O)OH), have specific absorption bands in the UV-VIS spectrum (Bédidi and Cervelle, 1993), and have a variable content depending on the soil mineralogy of the source regions (Journet et al., 2014).

Henceforth, in this study we present a new evaluation of the ultraviolet to near-infrared (375-850 nm) light-absorbing properties of mineral dust by studying the size-segregated mass absorption efficiency (MAE, units of m^2 g^{-1}) and its spectral dependence, largely used in climate models to calculate the direct radiative effect of aerosols. Experiments on twelve aerosol samples generated from natural parent top soils from various source regions worldwide have been conducted with a large atmospheric simulation chamber.

2. Instruments and methods

At a given wavelength λ, the mass absorption efficiency (MAE, units of m^2 g^{-1}) is defined as the ratio of the aerosol light-absorption coefficient $b_{abs}(\lambda)$ (units of m^{-1}) and its mass concentration (in µg m^{-3})

$$MAE(\lambda) = \frac{b_{abs}(\lambda)}{Mass\ Conc} \qquad (1)$$

MAE values for mineral dust aerosol are expressed in.

The spectral dependence of the aerosol absorption coefficient $b_{abs}(\lambda)$ is described by the power-law relationship

$$b_{abs}(\lambda) \sim \lambda^{-AAE} \qquad (2)$$

where the AAE is the Ångström Absorption Exponent, representing the negative slope of $b_{abs}(\lambda)$ in a log-log plot (Moosmüller et al., 2009)

$$AAE = -\frac{dln(b_{abs}(\lambda))}{dln(\lambda)} \tag{3}$$

2.1. The CESAM simulation chamber

Experiments in this work have been performed in the 4.2 m^3 stainless-steel CESAM (French acronym for Experimental Multiphasic Atmospheric Simulation Chamber) simulation chamber (Wang et al., 2011). The CESAM chamber has been extensively used in recent years to simulate, at sub and super-saturated conditions, the formation and properties of aerosols at concentration levels comparable to those encountered in the atmosphere (Denjean et al., 2015a; 2015b; Bregonzio-Rozier et al., 2015; 2016; Di Biagio et al., 2014; 2016).

CESAM is a multi-instrumented platform, equipped with twelve circular flanges to support its analytical environment. Basic instrumentation comprises sensors to measure the temperature, pressure and relative humidity within the chamber (two manometers MKS Baratrons (MKS, 622A and MKS, 626A) and a HMP234 Vaisala® humidity and temperature sensor). The particle size distribution is routinely measured by a combination of (i) a scanning mobility particle sizer (SMPS, mobility diameter range 0.02–0.88 μm), composed of a Differential Mobility Analyzer (DMA, TSI Inc. Model 3080) and a Condensation Particle Counter (CPC, TSI Inc. Model 3772); (ii) a SkyGrimm optical particle counter (Grimm Inc., model 1.129, optical equivalent diameter range 0.25–32 μm); and (iii) a WELAS optical particle counter (PALAS, model 2000, optical equivalent diameter range 0.5–47 μm). Full details of operations and data treatment of the particle counters are provided in Di Biagio et al. (2016).

2.2. Filter sampling

Three filter samples per top soil sample were collected on different types of substrate based on the analysis to perform. Sampling dedicated to the determination of the aerosol mass concentration by gravimetric analysis and the measurement of the absorption coefficients by optical analysis was performed on 47-mm quartz membranes (Pall Tissuquartz™, 2500 QAT-UP). Two samples were collected in parallel. The first quartz membrane sample (“total”) was collected without a dedicated size cut-off using an in-house built stainless steel sampler operated at 5 L min^{-1}. However, as detailed in Di Biagio et al. (2016), the length of the sampling line from the intake point in the chamber to the filter entrance was 50 cm, yielding with a 50% cut-off of the transmission efficiency at 10.6 μm in particle aerodynamic diameter. This fraction is therefore indicated as PM$_{10.6}$ in the forthcoming discussion. The second quartz membrane sample was collected using a 4-stage DEKATI impactor operated at the flow rate of 10 L min^{-1} to select the aerosol fraction of particles with aerodynamic diameter smaller than 2.5 µm, indicated as $PM_{2.5}$ here forth. Sampling for the analysis of the iron oxide content was performed on polycarbonate filters (47-mm Nuclepore, Whatman; pore size of 0.4 µm) using the same sample holder than used for the total quartz filters, and therefore referring to the $PM_{10.6}$ mass fraction. Samples were collected at a flow rate of 6 L min^{-1}. All flow rates were monitored by a thermal mass flow meter (TSI Inc., model 4140). These samples were also used to determine the elemental composition (including Fe) and the fraction of iron oxides in the total mass.

2.3. The Multi-Wavelength Absorbance Analyzer (MWAA)

The aerosol absorption coefficient $b_{abs}(\lambda)$ at 5 wavelengths (λ = 375, 407, 532, 635 and 850 nm) was measured by *in situ* analysis of the quartz filter samples using the Multi-Wavelength Absorbance Analyzer (MWAA), described in detail in Massabò et al. (2013; 2015).

The MWAA performs a non-destructive scan of the quartz filters on 64 different points, each ~ 1 mm^2 wide. It measures the light transmission through the filter as well as backscattering at two different angles (125° and 165°). This is necessary to constrain the multiple scattering effects occurring within the particle-filter system. Measurements are used as input of a radiative transfer model (Hänel, 1987; 1994) as implemented by Petzold and Schölinner (2004) for the Multi-Angle Absorption Photometry (MAAP) measurements. In this model, a two stream approximation is applied (Coakley and Chylek, 1975), in which the fractions of hemispherical backscattered radiation with respect to the total scattering for collimated and diffuse incident radiation are approximated on the basis of the Henyey-Greenstein scattering phase function (Hänel, 1987). This approximation assumes a wavelength-independent asymmetry parameter (g) set to 0.75, appropriate for mineral dust (Formenti et al., 2011; Ryder et al., 2013b). The total uncertainty, including the effects of photon counting and the deposit inhomogeneity, on the absorption coefficient measurement is estimated at 8%.

2.4. Gravimetric analysis

The aerosol mass deposited on filter (µg) was obtained by weighing the quartz filter before and after sampling, after a period of 48 hours of conditioning in a room with controlled atmospheric conditions (temperature, T ~ 20 ± 1 °C; relative humidity, RH ~ 50 ± 5%). Weighting is performed with an analytical balance (Sartorius model MC5, precision of 1 µg), and repeated three times to control the statistical variability of the measurement. Electrostatic effects are removed by exposing the filters, prior weighing, to a de-ionizer. The error on the measured mass is estimated at 10 µg, including the repetition variability. The aerosol mass concentration (µg m^{-3}) is obtained by dividing the mass deposited on filter to the total volume of sampled air (m^3) obtained from the mass flowmeter measurements.

2.5. Dust composition measurements

2.5.1. Elemental composition

Elemental concentrations for the major constituents of mineral dust (Na, Mg, Al, Si, P, S, Cl, K, Ca, Fe, Ti, Mn) were obtained by a Wavelength Dispersive X-ray fluorescence (WD-XRF) of the Nuclepore filters using a PW-2404 spectrometer by Panalytical. Excitation X-rays are produced by a Coolidge tube (I_{max} = 125 mA, V_{max} = 60 kV) with a Rh anode; primary X-ray spectrum can be controlled by inserting filters (Al, at different thickness) between the anode and the sample. Each element was analyzed three times, with specific conditions (voltage, tube filter, collimator, analyzing crystal and detector). Data collection was controlled by the SuperQ software provided with the instrument. The elemental mass thickness (µg cm^{-2}), that is, the analyzed elemental mass per unit surface, was obtained by comparing the elemental yields with a sensitivity curve measured in the same geometry on a set of certified mono- or bi-elemental thin layer standards by Micromatter Inc. The certified uncertainty on the standard deposit (± 5%) determines the lower limit on the uncertainty on the measured elemental concentrations, which range between 8% and 10% depending on the considered element. Thanks to the uniformity of the aerosol deposit on the filters, the atmospheric elemental concentrations (µg m^{-3}) were calculated by multiplying the analyzed elemental mass thickness by the ratio between the collection and analyzed surfaces of each sample (41 and 22 mm, respectively), then divided by the total sampled volume (m^3). Finally, concentrations of light-weight elements (atomic number $Z < 19$) were corrected for the underestimation induced by the self-absorption of the emitted soft X-rays inside aerosol particles according to Formenti et al. (2011).

Additional XRF analysis of the quartz filters has been performed both in the $PM_{10.6}$ and the $PM_{2.5}$ fractions, so to verify the absence of biases between the experiments dedicated to the determination of particle composition to those where the optical properties where measured.

2.6.2. Iron oxide content

The content and the mineralogical speciation of iron oxides, also defined as free-iron, that is, the fraction of iron which is not in the crystal lattice of silicates (Karickhoff and Bailey, 1973), was determined by

XANES (X-ray absorption near-edge structure) in the Fe K-range (K_α, 7112 eV) at the SAMBA (Spectroscopies Applied to Materials based on Absorption) beamline at the SOLEIL synchrotron facility in Saclay, France (Briois et al., 2011). The position and shape of the K pre-edge and edge peaks were analyzed as they depend on the oxidation state of iron and the atomic positions of the neighboring ions, mostly O^+ and OH^-.

As in Formenti et al. (2014b), samples were mounted in an external setup mode. A Si(220) double-crystal monochromator was used to produce a monochromatic X-ray beam, which was 3000 x 250 μm^2 in size at focal point. The energy range was scanned from 6850 eV to 7800 eV at a step resolution varying between 0.2 eV in proximity to the Fe-K absorption edge (at 7112 eV) to 2 eV in the extended range. Samples were analyzed in fluorescence mode without prior preparation. One scan acquisition lasted approximately 30 minutes, and was repeated three times to improve the signal-to-noise ratio.

The same analytical protocol was applied to five standards of Fe(III)-bearing minerals (**Table 1**), including iron oxides (hematite, goethite) and silicates (illite, montmorillonite, nontronite). The standard spectra were used to deconvolute the dust sample spectra to quantify the mineralogical status of iron. The linear deconvolution, performed the Athena IFEFFIT freeware analysis program (Ravel and Newville, 2005), provided with the proportionality factors α_i representing the mass fraction of elemental iron to be assigned to the *i*-th standard mineral. In particular, the values of α_{hem} and α_{goe} represent the mass fractions of elemental iron that can be attributed to hematite and goethite, and $\alpha_{Fe\ ox}$ ($\alpha_{hem} + \alpha_{goe}$), the mass fraction of elemental iron that can be attributed to iron oxides.

2.6.3. Calculation of the iron oxide content

The measured elemental concentrations obtained by X-ray fluorescence are expressed in the form of elemental oxides and summed to estimate the total mineral dust mass concentration MC_{dust} according to the equation from Lide (1992)

$$[MC_{dust}]=1.12\times\left\{\begin{matrix}1.658[\mathrm{Mg}]+1.889[\mathrm{Al}]+2.139[\mathrm{Si}]+1.399[\mathrm{Ca}]+1.668[\mathrm{Ti}]+1.582[\mathrm{Mn}]\\+(0.5\times1.286+0.5\times1.429+0.47\times1.204)[\mathrm{Fe}]\end{matrix}\right\} \quad (4)$$

The relative uncertainty on MC_{dust}, estimated from the analytical error on the measured concentrations, does not exceed 6%.

The fractional mass ratio (in percent) of elemental iron ($MR_{Fe\%}$) with respect to the total dust mass concentration MC_{dust} is then calculated as

$$MR_{Fe\,\%} = \frac{[Fe]}{[MC_{Dust}]} \times 100 \qquad (5)$$

The mass concentration of iron oxides or free-iron ($MC_{Fe\,ox}$), representing the fraction of elemental iron in the form of hematite and goethite (Fe_2O_3 and FeOOH, respectively), is equal to

$$MC_{Fe\,ox} = MC_{hem} + MC_{goe} \qquad (6)$$

where MC_{hem} and MC_{goe} are the total masses of hematite and goethite. These can be calculated from the values α_{hem} and α_{goe} from XANES analysis, which represent the mass fractions of elemental iron attributed to hematite and goethite, as

$$MC_{hem} = \frac{\alpha_{hem}\, x\, [Fe]}{0.70} \qquad (7.a)$$

$$MC_{goe} = \frac{\alpha_{goe}\, x\, [Fe]}{0.63} \qquad (7.b)$$

where the values of 0.70 and 0.63 represent the mass molar fractions of Fe in hematite and goethite, respectively. The relative errors on MC_{hem} and MC_{goe} are obtained from the uncertainties on values α_{hem} and α_{goe} from XANES analysis (less than 10%).

The mass ratio of iron oxides ($MR_{Fe\,ox\%}$) with respect to the total dust mass can then be calculated as

$$MR_{Fe\,ox\%} = MC_{Fe\,ox} \times MR_{Fe\,\%} \qquad (8)$$

3. Experimental protocol

At the beginning of each experiment, the chamber was evacuated by to 10^{-4}-10^{-5} hPa. Then, the reactor was filled with a mixture of 80% N_2 and 20% O_2 at a pressure slightly exceeding the current atmospheric pressure, in order to avoid contamination from ambient air. Experiments were conducted at ambient temperature and at a relative humidity <2%. As in Di Biagio et al. (2014; 2016), dust aerosols were generated by mechanical shaking of the soils, previously sieved to < 1000 µm and dried at 100 °C for about 1 h to remove any residual humidity. About 15 g of soil was placed in a Buchner flask and shaken for about 30 min at 100 Hz by means of a sieve shaker (Retsch AS200). The dust particles produced was then injected in the chamber by flushing the flask with N_2 at 10 L min^{-1} for about 10-15 min, whilst continuing shaking the soil.

Dust was injected for about 10-15 minutes, and left suspended in the chamber for approximately 120 min thanks to the 4-wheel fan located in the bottom of the chamber body. Previous measurements at the top and bottom of the chamber showed that the fan ensures a homogeneous distribution of the dust starting approximately 10 minutes after the end of the injection (Di Biagio et al., 2014).

To compensate for the air extracted from the chamber by sampling, a particle-free flow of N_2/O_2, regulated in real time as a function of the total volume of sampled air, was re-injected in the chamber. To avoid excessive dilution the flow was limited to 20 L min^{-1}. Two experiments per soil type were conducted: a first experiment for sampling on the nuclepore polycarbonate filters (determination of the elemental composition and the iron oxide fraction) and *in situ* measurements of the infrared optical constants (Di Biagio et al., 2016), and a second experiment on total quartz filter and impactor for the study of dust MAE presented in this paper.

Figure 1 illustrates as typical example the time series of the aerosol mass concentration during the two experiments conducted for the Libyan sample. The comparison demonstrates the repeatability of the dust concentrations, both in absolute values and in temporal dynamics. It also shows that the mass concentrations decreased very rapidly by gravitational settling within the first 30 minutes of the experiment (see also the discussion in Di Biagio et al. (2016)), after which concentrations only decrease by dilution. The filter sampling was started after this transient phase, and then continued through the end of the experiments, in order to collect enough dust on filter for the chemical analysis. Blank samples were collected before the start of the experiments by placing the loaded filter holders in line with the chamber and by flushing them for a few seconds with air coming from the chamber.

At the end of each experimental series with a given soil sample, the chamber was manually cleaned in order to remove carry-over caused by resuspension of particles deposited to the walls. Background concentrations of aerosols in the chamber vary between 0.5 and 2.0 µg m^{-3}, i.e. a factor of 500 to 1000 below the operating conditions.

3. Results and discussion

The geographical location of the soil collection sites is shown in **Figure 2**, whereas the coordinates are summarized in **Table 2**. As discussed in Di Biagio et al. (2016), the selection of these soils and sediments was governed by the need of representing the major arid and semi-arid regions worldwide, the need of taking into account the mineralogical diversity of the soil composition at the global scale, and finally by their availability in sufficient quantities for injection in the chamber. When doing so, we obtained a set of twelve samples distributed worldwide but mostly in Northern and Western Africa (Libya, Algeria, Mali, Bodélé) and the Middle East (Saudi Arabia and Kuwait). Individual samples from the Gobi desert in Eastern Asia, the Namib Desert, the Strzelecki desert in Australia, the Patagonian deserts in South America, and the Sonoran Desert in Arizona have also been investigated.

3.1. Elemental composition and iron oxide content

A total of 41 filters including 15 polycarbonate filters (12 samples and 3 blanks) and 25 quartz filters (12 for the total fraction, 10 for the fine fraction and 3 blanks) were collected for analysis.

The dust mass concentration found by gravimetric analysis varied between 50 µg m^{-3} and 5 mg m^{-3}, in relatively good agreement with the dust mass concentration MC_{dust} (Equation 4) based on X-ray fluorescence analysis: the slope of the linear regression between the calculated and the gravimetric values of MC_{dust} is 0.90 with $R^2 = 0.86$.

Di Biagio et al. (2016) show that clays are the most abundant mineral phases, together with quartz and calcite, and that significant variability exists as function of the compositional heterogeneity of the parent soils. Here we use the Fe/Ca and Si/Al elemental ratios obtained from X-ray Fluorescence analysis to discriminate the origin of used dust samples. These ratios have been extensively used in the past to discriminate the origin of African dust samples collected in the field (Chiapello et al., 1997; Formenti et al., 2011; Formenti et al., 2014a). The values obtained during our experiments are reported in **Table 3.** There is a very good correspondence between the values obtained for the Mali, Libya, Mauritania and (to a lesser extent) Morocco experiments to values found in environmental aerosol samples. The only

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major difference is observed for the Bodélé experiment, for which the Fe/Ca ratio is enriched by a factor of 6 with respect to the field observations (Formenti et al., 2011; Formenti et al., 2014a). This could reflect the fact that the Bodélé aerosol in the chamber is generated from a sediment sample and not from a soil. As a matter of fact, the Bodélé sediment sample is constituted by a very fine powder which becomes very easily airborne. Henceforth, and contrary to the soil samples, this powder is likely to be injected in the chamber with little or no size fractionation. As a consequence, the composition of the aerosol collected in the chamber could reflect more that of the parent sedimentary soil than not the other samples. On the other hand, Bristow et al. (2010) and Moskowitz et al. (2016) show that the iron content and speciation of the Bodélé sediments is very heterogeneous at the source scale. For non-northern African samples, the largest variability is observed for the Fe/Ca values, ranging from 0.1 to 8, whereas the Si/Al ratio varied between 2.5 and 4.8. In this case, values are available in the literature for comparison (e.g., Cornille et al., 1990; Reid et al., 1994; Eltayeb et al., 2001; Lafon et al., 2006; Shen et al., 2007; Radhi et al., 2010; 2011; Formenti et al., 2011; 2014a; Scheuvens et al., 2013 and references within). Values in the $PM_{2.5}$ fraction are very consistent with those obtained in the $PM_{10.6}$: their linear correlation has a slope of 1.03 (± 0.05) and a R^2 equal to 0.97, suggesting that the elemental composition is relatively size-independent.

The mass fraction of total Fe ($MC_{Fe\%}$ from Equation 5), also reported in **Table 3**, ranged from 2.8 to 7.3%, values found for the Namibia and the Australia samples, respectively. This range is in good agreement with values reported in the literature, taking into account that differences might be also due to the method (direct measurement/calculation) and/or the size fraction over which the total dust mass concentration is estimated (Chiapello et al., 1997; Reid et al., 1994; 2003; Derimian et al., 2008; Formenti et al., 2001; 2011; 2014a; Scheuvens et al., 2013). The agreement of $MC_{Fe\%}$ values obtained by the XRF analysis of polycarbonate filters (Equation 5) and those obtained from the XRF analysis of the quartz filters, normalized to the measured gravimetric mass is well within 10% (that is, the percent error of each estimate). An the exception are the samples of Bodélé and Algeria, for which the values obtained from the analysis of the quartz filters are significantly lower than those obtained from the nuclepore filters. We treat that as an additional source of error in the rest of the analysis, and add it to the total uncertainty. In the $PM_{2.5}$, the content of iron is more variable, ranging from 4.4% (Morocco) to 33.6% (Mali), showing a size dependence. A word of caution on this conclusion as the two estimates are not necessarily consistent in the way that the total dust mass is estimated (from Equation 4 for the $PM_{10.6}$ fraction and by gravimetric weighing on the $PM_{2.5}$).

Finally, between 11 and 47% of iron in the samples can be attributed to iron oxides, in variable proportions between hematite and goethite. The iron oxide fraction of total Fe in this study is on the lower end of the range (36-72%) estimated for field dust samples of Saharan/Sahelian origin (Formenti et al. 2014b). The highest value of Formenti et al. (2014b), obtained for a sample of locally-emitted dust collected at the Banizoumbou station in the African Sahel, is anyhow in excellent agreement with the value of 62% obtained for an experiment (not shown here) using a soil collected in the same area. Likewise, the proportions between hematite and goethite (not shown) are reproduced, showing that goethite is more abundant than hematite. The mass fraction of iron oxides ($MR_{Fe\ ox\%}$), estimated from Equation 8 and shown in Table 3, ranges between 0.7% (Kuwait) to 3.6% (Australia), which is in the range of available field estimates (Formenti et al., 2014a; Moskowitz et al., 2016). For China, our value of $MR_{Fe\ ox\%}$ is lower by almost a factor of 3 in comparison with that obtained on the same dust sample by Alfaro et al. (2004) (0.9% against 2.8%), whereas on a sample from Niger (however not considered in this study) our estimates and that by Alfaro et al. (2004) perfectly agree (5.8%). A possible underestimate of the iron oxide fraction for samples other than those from the Sahara-Sahel area could be due to the fact that - opposite to the experience of Formenti et al. (2014b) - the linear deconvolutions of the XANES spectra were not always satisfactory (see Figure S1 in the supplementary). This resulted in a significant residual between the observed and fitted XANES spectra. Indeed, the mineralogical reference for hematite is obtained from a soil from Niger (Table 1) and might not be fully suitable for representing aerosols of different origins. Additional differences could arise from differences in the size distributions of the generated aerosol. As a matter of fact, the number fraction of particles in the size classes above 0.5 µm in diameter are different in the dust aerosol generated in the Alfaro et al. (2004) study with respect to ours. In the study by Alfaro et al. (2004), the number fraction is lowest in the 0.5-0.7 size class and highest between 1 and 5 µm. On the contrary, in our study the number fraction is lowest in the 1-2 µm size range and highest between 0.5 and 0.7 µm. These differences could yield either to difference in the chemical composition and/or to a difference in the total mass in the denominator of Equation 8.

3.2. Spectral and size-variability of the mass absorption efficiency

The spectral mass absorption efficiency (MAE) at 375, 407, 532, 635 and 850 nm for the $PM_{10.6}$ and the $PM_{2.5}$ dust fractions are summarized in **Table 4** and displayed in **Figure 3**. Regardless of particle size, the MAE values decrease with increasing wavelength (almost one order of magnitude between 375 and 850 nm), and display a larger variability at shorter wavelengths. The MAE values for the $PM_{10.6}$ range from 37 (± 3) 10^{-3} m^2 g^{-1} to 135 (± 11) 10^{-3} m^2 g^{-1} at 375 nm, and from 1.3 (± 0.1) 10^{-3} m^2 g^{-1} to 15 (± 1)

10^{-3} m^2 g^{-1} at 850 nm. Maxima are found for the Australia and Algeria samples, whereas the minima are for Bodelé and Namibia, respectively at 375 and 850 nm. In the PM$_{2.5}$ fraction, the MAE values range from 95 (± 8) 10^{-3} m^2 g^{-1} to 711 (± 70) 10^{-3} m^2 g^{-1} at 375 nm, and from 3.2 (± 0.3) 10^{-3} m^2 g^{-1} to 36 (± 3) 10^{-3} m^2 g^{-1} at 850 nm. Maxima at both 375 and 850 nm are found for the Morocco sample, whereas the minima are for Algeria and Namibia, respectively. The MAE values for mineral dust resulting from this work are in relative good agreement with the estimates available in literature (Alfaro et al., 2004; Linke et al., 2006; Yang et al., 2009; Denjean et al., 2016), reported in **Table 5**. For the China Ulah Buhn sample, Alfaro et al. (2004) reported 69.1 10^{-3} and 9.8 10^{-3} m^2 g^{-1} at 325 and 660 nm, respectively. The former is lower than the value of 99 10^{-3} m^2 g^{-1} that we obtain by extrapolating our measurement at 375 nm. Likewise, our values for the Morocco sample are higher than reported by Linke et al. (2006) at 266 and 660 nm. Conversely, the agreement with the estimates of Yang et al. (2009) for mineral dust locally re-suspended in Xianghe, near Beijing (China) is very good at all wavelengths between 375 and 880 nm. As expected, the MAE values for mineral dust resulting from this work are almost one order of magnitude smaller than for other absorbing aerosols. For black carbon, MAE values are in the range of 6.5–7.5 m^2 g^{-1} at 850 nm (Bond and Bergstrom, 2006; Massabò et al., 2016), and vary in a linear way inversely with wavelength. For brown carbon, the reported MAE range between 2.3–7.0 m^2 g^{-1} at 350 nm (Chen and Bond, 2010; Kirchstetter et al., 2004; Massabò et al., 2016), 0.05–1.2 m^2 g^{-1} at 440 nm (Wang et al., 2016) and 0.08–0.72 m^2 g^{-1} at 550 nm (Chen and Bond, 2010).

The analysis of **Table 4** indicates that, at every wavelength, the MAE values in the PM$_{2.5}$ fraction are equal or higher than those for PM$_{10.6}$. The PM$_{2.5}$/PM$_{10.6}$ MAE ratios reach values of 6 for the Mali sample, but are mostly in the range 1.5-3 for the remaining aerosols. Values decrease with wavelength up to 635 nm, whereas at 850 nm they have values comparable to those at 375 nm. The observed size-dependence of the MAE values is consistent with the expected behavior of light absorption of particles in the Mie and geometric optical regimes that concern the two size fractions. Light-absorption of particles of size smaller or equivalent to wavelength is proportional to their bulk volume, whereas for larger particles absorption occurs on their surface only (Bohren and Huffmann, 1983). On the other hand, the size-resolved measurements of Lafon et al. (2006) show that the proportion (by volume) of iron oxides might be higher in the coarse than in the fine fraction, which would counteract the size-dependence behavior of MAE. To validate the observations, we calculated the spectrally-resolved MAE values in the two size fractions using the Mie code for homogeneous spherical particles (Bohren and Huffmann, 1983) and the number size distribution estimated by (Di Biagio et al., 2016) and averaged over the duration of filter sampling. We estimated the dust complex refractive index as a volume-weighted average of a non-absorbing dust fraction having the refractive index of kaolinite, dominant mineral in our samples (see Di Biagio et al., 2016), from Egan et Hilgeman (1979) and an absorbing fraction estimated from the mass fraction of iron oxides and having the refractive index of hematite (Bedidi and Cervelle, 1993). Results of this calculation indicate that the observed size-behavior is well reproduced at all wavelengths, even in the basic hypothesis that the mineralogical composition does not change with size. The only exception is 850 nm, where at times, $PM_{2.5}/PM_{10.6}$ MAE ratio is much higher than expected theoretically. We attribute that to the relatively high uncertainty affecting the absorbance measurements at this wavelength where the signal-to-noise ratio is low. Indeed, the two sets of values (MAE in the $PM_{2.5}$ fraction and MAE in the $PM_{10.6}$ fraction) are not statistically different according to a two-pair t-test (0.01 and 0.05 level of confidence), confirming that any attempt of differentiation the size-dependence at this wavelength would require a stronger optical signal.

The analysis of the spectral dependence, using a power-law function fit as from Equation 2, provides with the values of the Angstrom Absorption Exponent (AAE), also reported in **Table 4**. Contrary to the MAE values, there is no statistically significant size-dependence of the AAE values, ranging from 2.5 (± 0.2) to 4.1 (± 0.3), with an average of 3.3 (± 0.7), for the $PM_{10.6}$ size fraction and between 2.6 (± 0.2) and 5.1 (± 0.4), with an average of 3.5 (± 0.8), for the $PM_{2.5}$ fraction. Our values are in the range of those published in the open literature (Fialho et al., 2005; Linke et al., 2006; Müller et al., 2009; Petzold et al., 2009; Yang et al., 2009; Weinzierl et al., 2011; Moosmüller et al., 2012; Denjean et al., 2016), shown in **Table 5**. AAE values close to 1.0 are found for urban aerosols where fossil fuels combustion is dominant, while AAE values for brown carbon (BrC) from incomplete combustion are in the range 3.5-4.2 (Yang et al., 2009; Chen et al., 2015; Massabò et al., 2016).

Finally, **Figure 4** shows correlations between MAE values in the $PM_{10.6}$ fraction (Figure 3.a) and in the $PM_{2.5}$ fraction (Figure 3.b) and the estimated percent mass fraction of iron and iron oxides ($MC_{Fe\%}$ and $MC_{Fe\ ox\%}$), respectively. Regardless of the size fraction, the correlation between the MAE values and the percent mass of total elemental iron are satisfactory. Higher correlations are obtained at 375, 407 and 532 nm, and in the $PM_{2.5}$ fraction, where a linear correlation with R^2 up to 0.94 are obtained. Best correlations are obtained when forcing the intercept to zero, indicating that elemental iron fully accounts for the measured absorption. At these wavelengths, linear correlations with the mass fraction of iron oxides are loose in the $PM_{10.6}$ mass fraction (R^2 up to 0.38-0.62), but again satisfactory in the $PM_{2.5}$ fraction (R^2 up to 0.83-0.99), where, whoever, one should keep in mind that they have been established only indirectly by considering the ratio of iron oxides to elemental iron independent of size. At 660 and 850 nm, little or no robust correlation is obtained, often on very few data points and with very low MAE values. It is noteworthy that, in both size fractions, the linear correlation yields a non-zero intercept is obtained, indicating other minerals but iron oxides account for the measured absorption.

4. Conclusive remarks

In this paper, we reported new laboratory measurements of the shortwave mass absorption efficiency (MAE) of mineral dust of different origin and as a function of size and wavelength in the 375-850 nm range. Results have been obtained in the CESAM simulation chamber using generated mineral dust from natural parent soils, and optical and gravimetric analysis on extracted samples.

Our results can be summarized as follows: at 375 nm, the MAE values are lower for the $PM_{10.6}$ mass fraction (range 37-135 10^{-3} m^2 g^{-1}) than for the $PM_{2.5}$ (range 95-711 10^{-3} m^2 g^{-1}), and vary opposite to wavelength as λ^{-AAE}, where AAE (Angstrom Absorption Exponent) averages between 3.3-3.5 regardless of size fraction. These results deserve some conclusive comments:

- The size-dependence, yielding significantly higher MAE values in the fine fraction ($PM_{2.5}$) than for the $PM_{10.6}$ aerosol, indicates that light-absorption by mineral dust can be important even during atmospheric transport over heavy polluted regions, when dust concentrations are significantly lower than at emission. This can be shown by comparing the aerosol absorption optical depth (AAOD) at 440 nm for China, a well-known mixing region of mineral dust and pollution (e.g., Yang et al., 2009; Laskin et al., 2014; Wang et al., 2013). Laskin et al. (2014) reports that the average AAOD in the area is of the order of 0.1, for carbonaceous absorbing aerosols (sum of black and brown carbon). This is lower or comparable to the AAOD of 0.17 and 0.11 at 407 nm (fine and total mass fractions, respectively) that we obtain by a simple calculation (AAOD = MAE x MC_{dust} x H), where MAE are the values estimated in this study, MC the dust mass concentrations typically observed in the area during dust storms (Sun et al., 2001), and H a scale height factor of 1 km.
- The spectral variability of the dust MAE values, represented by the AAE parameter, is equal in the $PM_{2.5}$ and $PM_{10.6}$ mass fractions. This suggests that, for a given size distribution, the possible variation of dust composition with size do not affect in a significant way the spectral behavior of the absorption properties. Our average value for AAE is 3.3 ± 0.7, higher than for black carbon, but in the same range than light-absorbing organic (brown) carbon. As a result, depending on the

environment, there can be some ambiguity in apportioning the AAOD based on spectral dependence. Bahadur et al. (2012) and Chung et al. (2012) couple the AAE and the spectral dependence of the total AOD and/or its scattering fraction only to overcome this problem. Still, Bahadur et al. (2012) show that there is an overlap in the scatterplots of the spectral dependence of the scattering and absorption fractions of the AOD based on analysis of ground-based remote sensing data for mineral dust, urban and non-urban fossil fuel over California. A closer look to observations in mixing areas where biomass burning have different chemical composition and/or mineral dust has heavy loadings should be given in order to generalize the clear separation observed in the spectral dependences of mineral dust and biomass burning (Bahadur et al., 2012). This aspect is relevant to the development of remote sensing of light-absorption aerosols from space, and their assimilation in climate models (Torres et al., 2007; Buchard et al., 2015; Hammer et al., 2016).

- There is an important sample-to-sample variability in our dataset of MAE values for mineral dust aerosols. At 532 nm, our estimated MAE average at 34 ± 14 $m^2 g^{-1}$ and 78 ± 70 $m^2 g^{-1}$ in the $PM_{10.6}$ and $PM_{2.5}$ mass fractions, respectively. Figure 3, showing the correlation with the estimated mass fraction of elemental iron and iron oxides, suggests that this variability could be related to the regional differences of the mineralogical composition of the parent soils. These observations lead to different considerations. To start with, our study reinforces the need for regionally-resolved representation of the light-absorption properties of mineral dust in order to improve the representation of its effect on climate. As a matter of fact, the natural variability of the absorption properties that we obtain from our study is in the range 50-100%, even when we limit ourselves to smaller spatial scales, for example those of north Africa (samples from Libya, Algeria, Mali and Bodélé). This is far above the ± 5% sensitivity factor used by Solmon et al. (2008) to vary the single scattering albedo (as a proxy of absorption) of mineral dust over western Africa, and to show how this could drastically change the climate response in the region.

The question is then “how to represent this regional variability?” As Moosmüller et al. (2012), we found that elemental iron is a very good proxy for the MAE, especially in the $PM_{2.5}$ fraction, where iron-bearing absorbing minerals (hematite, goethite, illite, smectite clays) would be more concentrated. In the coarse fraction, Ca-rich minerals, quartz and feldspars could also play a role, and that could result in the observed lowered correlation (although adding a term proportional to elemental Ca does not ameliorate the result in the present study). The correlation of the spectral

MAE values with the iron oxide fraction is satisfactory but rather noisy, also owing to some uncertainty in the quantification of iron oxides from X-Ray Absorption measurements. In this case, the intercept is significantly different from zero, again indicating that a small but clear fraction of absorption is due to minerals other than iron oxides. There are contrasting results on this topic: Alfaro et al. (2004) found an excellent correlation between MAE and the iron oxide content, whereas Klaver et al. (2011) found that the single scattering albedo (representing the capacity of an aerosol population to absorb light with respect to extinction) was almost independent on the mass fraction of iron oxides. Moosmüller et al. (2012) disagreed, pointing out to the uncertainty in the correction procedure of the measurement of absorption by Klaver et al. (2011). As a matter of fact, Klaver et al. (2011) and Alfaro et al. (2004) used the same correction procedure. It is more likely that the lack of correlation found in Klaver et al. (2011) is due to the fact that other minerals than iron oxides contribute to absorption, in particular at their working wavelength (567 nm), where the absorption efficiency of iron oxides starts to weaken. Clearly, the linear correlation between elemental iron in mineral dust and its light-absorption properties could ease the application and validation of climate models that now starting including the representation of the mineralogy (Perlwitz et al., 2015a; 2015b; Scanza et al., 2015). Also, they would facilitate detecting source regions based on remote sensing of dust absorption in the UV-VIS spectral region (e.g., Hsu et al., 2004). However, such a quantitative relationship cannot uniquely determined from these studies, including the present one, which use different ways of estimating elemental iron, iron oxides and the total dust mass. A more robust estimate should be obtained by estimating the imaginary parts of the complex refractive indices associated to these measurements of absorption, and their dependence on the mineralogical composition.

Author contributions

L. Caponi, P. Formenti, D. Massabò, P. Prati, C. Di Biagio, and J. F. Doussin designed the chamber experiments and discussed the results. L. Caponi and C. Di Biagio realized the experiments with contributions by M. Cazaunau, E. Pangui, P. Formenti, and J.F. Doussin. L. Caponi, D. Massabò and P. Formenti performed the full data analysis with contributions by C. Di Biagio, P. Prati and J.F. Doussin. L. Caponi, P. Formenti and S. Chevaillier performed the XRF measurements. P. Formenti and G. Landrot performed the XAS measurements. D. Massabò performed the MWAA and the gravimetric measurements. M. O. Andreae, K. Kandler, T. Saeed, S. Piketh, D. Seibert, and E. Williams collected the soil

samples used for experiments. L. Caponi, P. Formenti, D. Massabò and P. Prati wrote the manuscript with comments from all co-authors.

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Table captions

Table 1. Characteristics of the standards used for the quantification of the iron oxides in the XAS analysis.

Table 2. Summary of information on the soil samples used in this work.

Table 3. Chemical characterisation of the dust aerosols in $PM_{10.6}$ and $PM_{2.5}$ (in parentheses) size fractions. Columns 3 and 4 report the Si/Al and Fe/Ca elemental ratios obtained from X-Ray Fluorescence analysis. The uncertainty on each individual value is estimated to be 10%. Column 5 reports $MR_{Fe\%}$, the fractional mass of elemental iron with respect to the total dust mass concentration (uncertainty 10%). Column 5 reports $MR_{Fe\%}$, the mass fraction of iron oxides with respect to the total dust mass concentration (uncertainty 15%). For $PM_{2.5}$ the determination of the Si/Al ratio is impossible due to the composition of the filter medium.

Table 4. Mass absorption efficiency (MAE, 10^{-3} m^2 g^{-1}) and Ångström Absorption Exponent (AAE) in the $PM_{10.6}$ and $PM_{2.5}$ size fractions. Absolute errors are in brackets.

Table 5. Mass absorption efficiency (MAE, 10^{-3} m^2 g^{-1}) and Ångström Absorption Exponent (AAE) of literature data discussed in the paper

Figure captions

Figure 1. Time series of aerosol mass concentration in the chamber for the two companion experiments (Libya sample). Experiment 1 (top panel) was dedicated to the determination of the chemical composition (including iron oxides) by sampling on polycarbonate filters. Experiment 2 (bottom panel) was dedicated to the determination of the absorption optical properties by sampling on quartz filters.

Figure 2. Location (red star) of the soil and sediment samples used to generate dust aerosols.

Figure 3. Spectral dependence of the MAE values for the samples investigated in this study in the $PM_{10.6}$ (left) and in the $PM_{2.5}$ (right) mass fractions.

Figure 4. Illustration of the links between the MAE values and the dust chemical composition found in this study. Left column, from top to bottom: linear regression between MAE values between 375 and 850 nm and the fraction of elemental iron with respect to the total dust mass ($MR_{Fe\%}$) in the $PM_{10.6}$ fraction; Middle column: same as left column but respect to the mass fraction of iron oxides to the total

dust mass ($MR_{Fe\ ox\%}$) in the $PM_{10.6}$ size fraction; Right column: same as left column but in the $PM_{2.5}$ size fraction.

Table 1. Characteristics of the standards used for the quantification of the iron oxides in the XAS analysis.

Standard	Stoechiometric Formula	Origin
Illite of Puy	$(Si_{3.55}Al_{0.45})(Al_{1.27}Fe_{0.36}Mg_{0.44})O_{10}(OH)_2(Ca_{0.01}Na_{0.01}K_{0.53}X(I)_{0.12})$	Puy, France
Goethite	FeO OH	Minnesota
Hematite	Fe_2O_3	Niger
Montmorillonite	$(Na,Ca)_{0,3}(Al,Mg)_2Si_4O10(OH)_2\cdot n(H_2O)$	Wyoming
Nontronite	$Na_{0.3}Fe_2(Si,Al)_4O10(OH)_2\cdot nH2O$	Pennsylvania

Table 2. Geographical information on the soil samples used in this work.

Geographical area	Sample	Desert area	Geographical coordinates
Sahara	Morocco	East of Ksar Sahli	31.97°N, 3.28°W
	Libya	Sebha	27.01°N, 14.50°E
	Algeria	Ti-n-Tekraouit	23.95°N, 5.47°E
Sahel	Mali	Dar el Beida	17.62°N, 4.29°W
	Bodélé	Bodélé depression	17.23°N, 19.03°E
Middle East	Saudi Arabia	Nefud	27.49°N, 41.98°E
	Kuwait	Kuwaiti	29.42°N, 47.69°E
Southern Africa	Namibia	Namib	21.24°S, 14.99°E
Eastern Asia	China	Gobi	39.43°N, 105.67°E
North America	Arizona	Sonoran	33.15 °N, 112.08°W
South America	Patagonia	Patagonia	50.26°S, 71.50°W
Australia	Australia	Strzelecki	31.33°S, 140.33°E

Table 3. Chemical characterisation of the dust aerosols in $PM_{10.6}$ and $PM_{2.5}$ (in parentheses) size fractions. Columns 3 and 4 report the Si/Al and Fe/Ca elemental ratios obtained from X-Ray Fluorescence analysis. The uncertainty on each individual value is estimated to be 10%. Column 5 reports $MR_{Fe\%}$, the fractional mass of elemental iron with respect to the total dust mass concentration (uncertainty 10%). Column 5 reports $MR_{Fe\ ox\%}$, the mass fraction of iron oxides with respect to the total dust mass concentration (uncertainty 15%). For $PM_{2.5}$ the determination of the Si/Al ratio is impossible due to the composition of the filter medium.

Geographical area	Sample	Si/Al	Fe/Ca	$MC_{Fe\%}$	$MC_{Fe-ox\%}$
Sahara	Morocco	3.12 (---)	0.24 (0.28)	3.6 (4.4)	1.4 (1.8)
	Libya	2.11 (---)	1.19 (1.12)	5.2 (5.6)	3.1 (3.4)
	Algeria	2.51 (---)	3.14 (4.19)	6.6 (5.4)	2.7 (2.2)
Sahel	Mali	3.03 (---)	2.99 (3.67)	6.6 (33.6)	3.7 (18.7)
	Bodélé	5.65 (---)	12.35 (----)	4.1 (----)	0.7 (----)
Middle East	Saudi Arabia	2.95 (---)	0.29 (0.27)	3.8 (5.1)	2.6 (3.5)
	Kuwait	3.15 (---)	0.89 (1.0)	5.0 (13.6)	1.5 (4.2)
Southern Africa	Namibia	3.41 (---)	0.11 (0.10)	2.4 (6.9)	1.1 (3.1)
Eastern Asia	China	2.68 (---)	0.77 (0.71)	5.8 (13.6)	0.9 (2.5)
North America	Arizona	3.30 (---)	0.95 (----)	5.3 (----)	1.5 (----)
South America	Patagonia	4.80 (---)	4.68 (4.64)	5.1 (----)	1.5 (---)
Australia	Australia	2.65 (---)	5.46 (4.86)	7.2 (11.8)	3.6 (5.9)

Table 4. Mass absorption efficiency (MAE, 10^{-3} m^2 g^{-1}) and Ångström Absorption Exponent (AAE) in the PM$_{10.6}$ and PM$_{2.5}$ size fractions. Absolute errors are in brackets.

		PM$_{10.6}$					
Geographical area	**Sample**	**375 nm**	**407 nm**	**532 nm**	**635 nm**	**850 nm**	**AAE**
Sahara	Morocco	--- (---)	--- (---)	--- (---)	--- (---)	--- (---)	--- (---)
	Libya	89 (11)	75 (9)	30 (5)	--- (---)	--- (---)	3.2 (0.3)
	Algeria	99 (10)	80 (10)	46 (7)	16 (3)	15 (3)	2.5 (0.3)
Sahel	Mali	--- (---)	103 (18)	46 (12)	--- (---)	--- (---)	--- (---)
	Bodélé	37 (4)	25 (3)	13 (2)	6 (1)	3 (1)	3.3 (0.3)
Middle East	Saudi Arabia	90 (9)	79 (8)	28 (3)	6 (1)	4 (1)	4.1 (0.4)
	Kuwait	--- (---)	--- (---)	--- (---)	--- (---)	--- (---)	2.8 (0.3)
Southern Africa	Namibia	52 (7)	49 (7)	13 (3)	5 (2)	1 (2)	4.7 (0.5)
Eastern Asia	China	65 (8)	58 (7)	32 (4)	8 (2)	7 (2)	3 (0.3)
North America	Arizona	130 (15)	99 (12)	47 (7)	21 (4)	13 (4)	3.1 (0.3)
South America	Patagonia	102 (11)	80 (9)	29 (4)	17 (2)	10 (2)	2.9 (0.3)
Australia	Australia	135 (15)	121 (13)	55 (7)	26 (4)	14 (3)	2.9 (0.3)

		$PM_{2.5}$					
Geographical area	**Sample**	**375 nm**	**407 nm**	**532 nm**	**635 nm**	**850 nm**	**AAE**
Sahara	Morocco	107 (13)	88 (11)	34 (6)	14 (3)	15 (4)	2.6 (0.3)
	Libya	132(17)	103 (14)	33 (7)	--- (---)	--- (---)	4.1 (0.4)
	Algeria	95(8)	71 (11)	37 (7)	12 (5)	12 (5)	2.8 (0.3)
Sahel	Mali	711 (141)	621 (124)	227 (78)	--- (---)	--- (---)	3.4 (0.3)
	Bodelé	--- (---)	--- (---)	--- (---)	--- (---)	--- (---)	--- (---)
Middle East	Saudi Arabia	153 (18)	127 (15)	42 (7)	8 (4)	6 (4)	4.5 (0.5)
	Kuwait	270 (100)	324 (96)	--- (---)	54 (52)	--- (---)	3.4 (0.3)
Southern Africa	Namibia	147 (36)	131 (32)	31 (21)	6 (16)	3 (15)	5.1 (0.5)
Eastern Asia	China	201 (30)	176 (26)	89 (17)	14 (10)	23 (10)	3.2 (0.3)
North America	Arizona	--- (---)	--- (---)	--- (---)	--- (---)	--- (---)	--- (---)
South America	Patagonia	--- (---)	--- (---)	--- (---)	--- (---)	--- (---)	2.9 (0.3)
Australia	Australia	335 (39)	288 (33)	130 (19)	57 (11)	36 (9)	2.9 (0.3)

Table 5. Mass absorption efficiency (MAE, 10^{-3} m^2 g^{-1}) and Ångström Absorption Exponent (AAE) of literature data discussed in the paper

Geographical area	**Sample**	**266 nm**	**325 nm**	**428 nm**	**532 nm**	**660 nm**	**880 nm**	**1064 nm**	**AAE**
Sahara	Morocco[*]								2.25–5.13
	Morocco, $PM_{2.5}$[£]								2.0–6.5
	Morocco, submicron[#]	1100			60			30	4.2
	Egypt, submicron[#]	810			20				5.3
	Tunisia[$]		83			11			
	Saharan, transported[µ]								2.9 ± 0.2
	Saharan, transported (PM_{10})[%]			37	27[%%]	15[%%%]			2.9
	Saharan, transported (PM_1)[%]			60	40[%%]	30[%%%]			2.0
Sahel	Niger[$]		124			19			
Eastern Asia	China[$]		69			10			
	China[&]		87[&&]	50[&&&]	27[&&&&]	13	1		3.8
Arabian Peninsula, N/NE Africa, Central Asia	Various locations[@]								2.5-3.9

[*] Müller et al. (2008)
[£] Petzold et al. (2008)
[#] Linke et al. (2006)
[$] Alfaro et al. (2004)
[µ] Fialho et al. (2005)
[%] Denjean et al. (2016); [%%] at 528 nm, [%%%] at 652 nm
[&] Yang et al. (2009); [&&] at 375 nm, [&&&] at 470 nm, [&&&&] at 590 nm
[@] Mossmüller et al. (2012)

Figure 1. Time series of aerosol mass concentration in the chamber for the two companion experiments (Libya sample). Experiment 1 (top panel) was dedicated to the determination of the chemical composition (including iron oxides) by sampling on polycarbonate filters. Experiment 2 (bottom panel) was dedicated to the determination of the absorption optical properties by sampling on quartz filters.

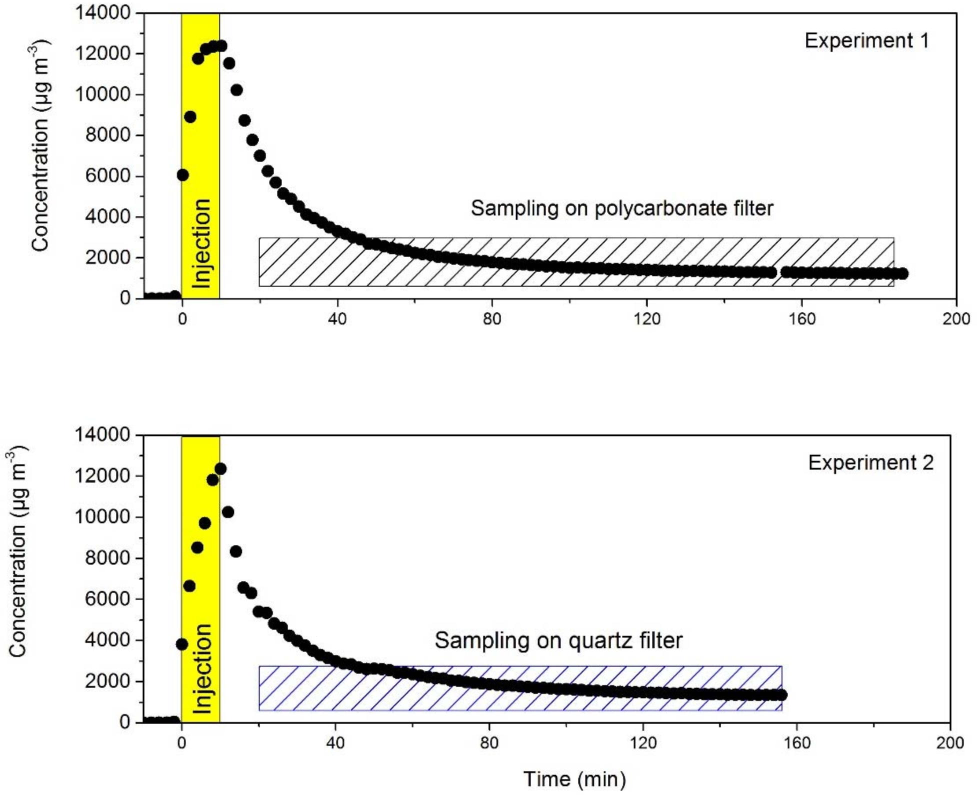

Figure 2. Location (red star) of the soil and sediment samples used to generate dust aerosols.

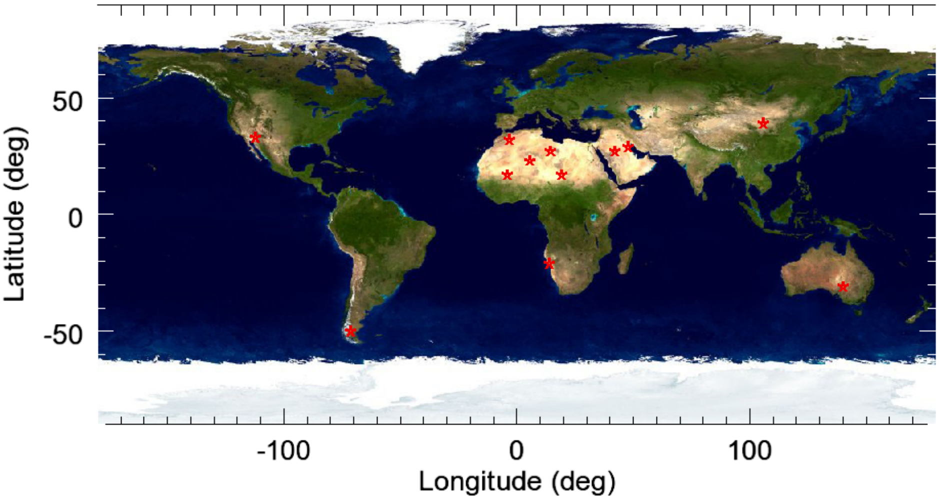

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Figure 3. Spectral dependence of the MAE values for the samples investigated in this study in the $PM_{10.6}$ (left) and in the $PM_{2.5}$ (right) mass fractions.

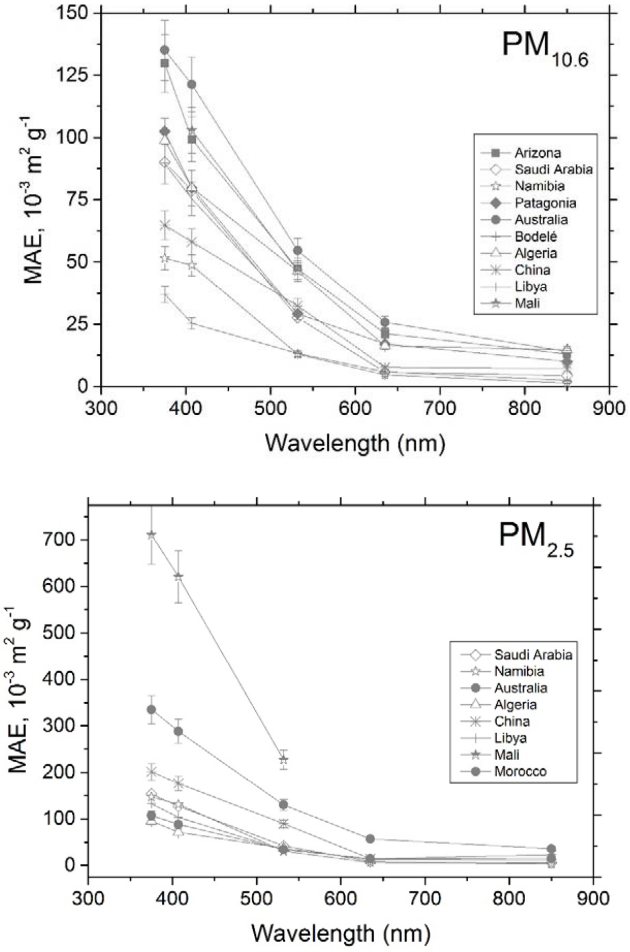

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Figure 4. Illustration of the links between the MAE values and the dust chemical composition found in this study. Left column, from top to bottom: MAE values between 375 and 850 nm versus the fraction of elemental iron with respect to the total dust mass ($MR_{Fe\%}$) in the $PM_{10.6}$ fraction; Middle column: same as left column but versus the mass fraction of iron oxides to the total dust mass ($MR_{Fe\ ox\%}$) in the $PM_{10.6}$ size fraction; Right column: same as left column but in the $PM_{2.5}$ size fraction. The linear regression lines between MAE and $MR_{Fe\%}$ and MAE and $MR_{Fe\ Ox\%}$ are reported in each plot.