1	Spectral- and size-resolved mass absorption efficiency of mineral dust aerosols in
2	the shortwave spectrum: a simulation chamber study
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### Abstract

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- 25 This paper presents new laboratory measurements of the mass absorption efficiency (MAE) between
- 26 375 and 850 nm for twelve individual samples of mineral dust from different source areas worldwide
- 27 and in of different origin in two size classes: PM<sub>10.6</sub> (mass fraction of particles of aerodynamic diameter
- lower than 10.6 μm) and PM<sub>2.5</sub> (mass fraction of particles of aerodynamic diameter lower than 2.5 μm).
- 29 EThe experiments have been were performed in the CESAM simulation chamber using generated min-
- eral dust generated from natural parent soils, and included optical and gravimetric analyses.
- 31 The Rresults show that the MAE values are lower for the PM<sub>10.6</sub> mass fraction (range 37-135 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup>
- at 375 nm) than for the PM<sub>2.5</sub> (range 95-711 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup> at 375 nm), and decrease with increasing wave-
- length as  $\lambda^{-AAE}$ , where the Angstrom Absorption Exponent (AAE) averages between 3.3-3.5, regardless
- of size. The size\_-independence of AAE suggests that, for a given size distribution, the possible variation
- of dust composition did not vary with size for this set of samples with size would not affect significantly
- 36 the spectral behavior of shortwave absorption. Because of its high atmospheric concentration, light -
- 37 absorption by mineral dust can be competitive to with black and brown carbon even during atmospheric
- transport over heavy polluted regions, when dust concentrations are significantly lower than at emission.
- 39 The AAE values of mineral dust are higher than for black carbon (~1), but in the same range as light-
- 40 absorbing organic (brown) carbon. As a result, depending on the environment, there can be some ambi-
- 41 guity in apportioning the <u>aerosol absorption optical depth (AAOD)</u> based on spectral dependence, which
- 42 is relevant to the development of remote sensing of light-absorbing aerosols from space, and their
- assimilation in climate models. We suggest that the sample-to-sample variability in our dataset of MAE
- values is related to regional differences of <u>in</u> the mineralogical composition of the parent soils. Particu-
- larly in the PM<sub>2.5</sub> fraction, we found a strong linear correlation between the dust light-absorption prop-
- erties and elemental iron rather than the iron oxide fraction, which could ease the application and the
- validation of climate models that now start to include the representation of the dust composition, as well
- as for remote sensing of dust absorption in the UV-VIS spectral region.

### 1. Introduction

- Mineral dust aerosols emitted by wind erosion of arid and semi-arid soils account for about 40% of the
- total emitted aerosol mass per year at the global scale (Knippertz and Stuut, 2014). The episodic but
- frequent transport of intense mineral dust plumes is visible from spaceborne sensors, as their high con-
- centrations, combined <u>with to-their ability of to scattering</u> and absorbing solar and thermal radiation,

54 give raise to the highest registered values of aerosol optical depth (AOD) on Earth (Chiapello, 2014). 55 The instantaneous radiative efficiency of dust particles, that is, their radiative effect per unit AOD, is of the order of tenths to hundreds of W m<sup>-2</sup> AOD<sup>-1</sup> in the solar spectrum, and of the order of tenths 56 of W m<sup>-2</sup> AOD<sup>-1</sup> in the thermal infrared (e.g., Haywood et al., 2003; di Sarra et al., 2011; Slingo et al., 57 2006 and the compilation of Highwood and Ryder, 2014). In the solar spectrum, (Boucher et al., 2013). 58 59 Albeit partially compensated by the radiative effect in the thermal infrared, the global mean radiative 60 effect of mineral dust in the shortwave is negative both at the surface and the top of the atmosphere (TOA) and produces a local warming of the atmosphere (Boucher et al., 2013). Many There are the 61 62 consequences numerous impacts of dust on the global and regional climate, that which ultimately feed 63 back on wind speed and vegetation and therefore on dust emission (Tegen and Lacis, 1996; Solmon et al., 2008; Pérez et al., 2006; Miller et al., 2014). Dust particles perturb the surface air temperature 64 65 through their radiative effect at TOA, can increase the atmospheric stability (e.g., Zhao et al. 2011) and 66 might affect precipitation at the global and regional scale (Solmon et al., 2008; Xian, 2008; Vinoj et al., 67 2014; Miller et al., 2014 and references therein). All models show indicate that the effect of mineral dust on climate has a great sensitivity to their 68 69 shortwave absorption properties of mineral dust (Miller et al., 2004; Lau et al., 2009; Loeb and Su, 2010; 70 Ming et al., 2010; Perlwitz and Miller, 2010). Absorption by mineral dust started receiving a great deal of interest-in the last ten years or so, when spaceborne and ground-based remote sensing studies (Dubo-71 vik et al., 2002; Colarco et al., 2002; Sinyuk et al., 2003) suggested that mineral dust was less absorbing 72 that than it had been suggested indicated by in situ observations (e.g., Patterson et al., 1977; Haywood et 73 al., 2001), particularly at wavelengths below 600 nm. Balkanski et al. (2007) showed that lowering the 74 75 dust absorption properties to an extent that reconciles them both with the remote-sensing observations and the state-of-knowledge of the mineralogical composition, allowed calculating the clear-sky dust 76 77 shortwave radiative effect of dust in agreement with satellite-based observations. A significant body number of observations has ve been performed in quantifyquantifieding the shortwave light-absorbing 78 79 properties of mineral dust, by direct measurements (Alfaro et al., 2004; Linke et al., 2006; Osborne et al., 2008; McConnell et al., 2008; Derimian et al., 2008; Yang et al., 2009; Müller et al., 2009; Petzold 80 et al., 2009; Formenti et al., 2011; Moosmüller et al., 2012; Wagner et al., 2012; Ryder al., 2013a; Utry 81 82 et al., 2015; Denjean et al., 2015c; 2016), and indirectly, by quantifying the amount and the speciation 83 of the light-absorbing compounds in mineral dust, principally iron oxides (Lafon et al., 2004; 2006;

84 Lazaro et al., 2008; Derimian et al., 2008; Zhang et al., 2008; Kandler et al., 2007; 2009; 2011; Formenti

85 et al., 2014a; 2014b).

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86 However, existing data are often limited to a single wavelength, which moreover are is not the same

identical for all experiments. Also, frequently they do not represent the possible regional variability of

the dust absorption, either because they are obtained from field measurements integrating the contribu-

tions of different source regions, or conversely, by laboratory investigations targeting samples from a

limited number of locations. This might lead to biases in the data. Indeed, iron oxides in mineral dust,

mostly in the form of hematite (Fe<sub>2</sub>O<sub>3</sub>) and goethite (Fe(O)OH), have specific absorption bands in the

UV-VIS spectrum (Bédidi and Cervelle, 1993), and have a variable content depending on the soil min-

eralogy of the source regions (Journet et al., 2014).

94 Henceforth, iIn this study, eExperiments on twelve aerosol samples generated from natural parent top

soils from various source regions worldwide have been were conducted with a large atmospheric simu-

lation chamber. we We present a new evaluation of the ultraviolet to near-infrared (375-850 nm) light-

absorbing properties of mineral dust by studying investigating the size-segregated mass absorption effi-

ciency (MAE, units of m<sup>2</sup> g<sup>-1</sup>) and its spectral dependence, <del>largely widely used in climate models to</del>

calculate the direct radiative effect of aerosols. Experiments on twelve aerosol samples generated from

natural parent top soils from various source regions worldwide have been conducted with a large atmos-

101 pherie simulation chamber.

## 2. Instruments and methods

At a given wavelength,  $\lambda$ , the mass absorption efficiency (MAE, units of m<sup>2</sup> g<sup>-1</sup>) is defined as the ratio

of the aerosol light-absorption coefficient  $b_{abs}(\lambda)$  (units of  $m^{-1})_{\bar{a}}$  and its mass concentration (in  $\mu g \ m^{-3}$ )

$$MAE(\lambda) = \frac{b_{abs}(\lambda)}{Mass\ Conc} \tag{1}$$

108 MAE values for mineral dust aerosol are expressed in. MAE values for mineral dust aerosol are ex-

109 pressed in  $10^{-3}$  m<sup>2</sup> g<sup>-1</sup>.

The spectral dependence of the aerosol absorption coefficient  $b_{abs}(\lambda)$  is described by the power-law

111 relationship

$$b_{abs}(\lambda) \sim \lambda^{-AAE} \tag{2}$$

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where the AAE is the <u>Absorption</u> Angström <u>Absorption</u> Exponent, representing the negative slope of  $b_{abs}(\lambda)$  in a log-log plot (Moosmüller et al., 2009)

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$$AAE = -\frac{dln(b_{abs}(\lambda))}{dln(\lambda)} \tag{3}$$

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#### 2.1. The CESAM simulation chamber

- 121 The eExperiments in this work have been performed in the 4.2 m<sup>3</sup> stainless-steel CESAM (French acro-
- nym for Experimental Multiphasic Atmospheric Simulation Chamber) simulation chamber (Wang et al.,
- 123 2011). The CESAM chamber has been extensively used in recent years to simulate, at sub and super-
- saturated conditions, the formation and properties of aerosols at concentration levels comparable to those
- encountered in the atmosphere (Denjean et al., 2015a; 2015b; Bréegonzio-Rozier et al., 2015; 2016; Di
- 126 Biagio et al., 2014; 20176).
- 127 CESAM is a multi-instrumented platform, equipped with twelve circular flanges to support its analytical
- environment. Basic instrumentation comprises sensors to measure the temperature, pressure and relative
- humidity within the chamber (two manometers MKS Baratrons (MKS, 622A and MKS, 626A) and a
- HMP234 Vaisala® humidity and temperature sensor). The particle size distribution is routinely meas-
- ured by a combination of (i) a scanning mobility particle sizer (SMPS, mobility diameter range 0.02–
- 0.88 μm), composed of a Differential Mobility Analyzer (DMA, TSI Inc. Model 3080) and a Conden-
- sation Particle Counter (CPC, TSI Inc. Model 3772); (ii) a SkyGrimm optical particle counter (Grimm
- Inc., model 1.129, optical equivalent diameter range 0.25–32 μm); and (iii) a WELAS optical particle
- counter (PALAS, model 2000, optical equivalent diameter range 0.5–47 µm). Full details of operations
- and data treatment of the particle counters are provided in Di Biagio et al. (20162017).

## 2.2. Filter sampling

- 138 Three filter samples per top soil sample were collected on different types of substrate based on the anal-
- ysis to be performed. Sampling dedicated to the determination of the aerosol mass concentration by

gravimetric analysis and the measurement of the absorption coefficients by optical analysis was performed on 47-mm quartz membranes (Pall Tissuquartz<sup>TM</sup>, 2500 QAT-UP). Two samples were collected in parallel. The first quartz membrane sample ("total") was collected without a dedicated size cut-off using an in-house built stainless steel sampler operated at 5 L min<sup>-1</sup>. However, as detailed in Di Biagio et al. (20162017), the length of the sampling line from the intake point in the chamber to the filter entrance was 50 cm, yielding resulting in with a 50% cut-off of the transmission efficiency at 10.6 μm in particle aerodynamic diameter. This fraction is therefore indicated as PM<sub>10.6</sub> in the forthcoming-following discussion. The second quartz membrane sample was collected using a 4-stage DEKATI impactor operated at the a flow rate of 10 L min<sup>-1</sup> to select the aerosol fraction of particles with aerodynamic diameter smaller than 2.5 μm, indicated as PM<sub>2.5</sub> here forth. Sampling for the analysis of the iron oxide content was performed on polycarbonate filters (47-mm Nuclepore, Whatman; pore size of 0.4 μm) using the same sample holder than as used for the total quartz filters, and therefore referring corresponding to the PM<sub>10.6</sub> mass fraction. Samples were collected at a flow rate of 6 L min<sup>-1</sup>. All flow rates were monitored by a thermal mass flow meter (TSI Inc., model 4140). These samples were also used to determine the elemental composition (including Fe) and the fraction of iron oxides in the total mass.

# 2.3. The Multi-Wavelength Absorbance Analyzer (MWAA)

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- The aerosol absorption coefficient,  $b_{abs}(\lambda)$ , at 5 wavelengths ( $\lambda = 375, 407, 532, 635$ , and 850 nm) was measured by *in situ* analysis of the quartz filter samples using the Multi-Wavelength Absorbance Ana-
- 158 lyzer (MWAA), described in detail in Massabò et al. (2013; 2015).
- The MWAA performs a non-destructive scan of the quartz filters on at 64 different points, each  $\sim 1 \text{ mm}^2$
- wide. It measures the light transmission through the filter as well as backscattering at two different angles
- 161 (125° and 165°). This is necessary to constrain the multiple scattering effects occurring within the par-
- ticle-filter system. The mMeasurements are used as input of to a radiative transfer model (Hänel, 1987;
- 163 1994) as implemented by Petzold and Schönlinner (2004) for the Multi-Angle Absorption Photometry
- (MAAP) measurements. In this model, a two stream approximation is applied (Coakley and Chylek,
- 165 1975), in which the fractions of hemispherical backscattered radiation with respect to the total scattering
- for collimated and diffuse incident radiation are approximated on the basis of the Henyey-Greenstein
- scattering phase function (Hänel, 1987). This approximation assumes a wavelength-independent asym-
- metry parameter (g) set to 0.75, appropriate for mineral dust (Formenti et al., 2011; Ryder et al., 2013b).

169 The total uncertainty, including the effects of photon counting and the deposit inhomogeneity, on the

absorption coefficient measurement is estimated at 8% (Massabò et al., (2013)-

## 2.4. Gravimetric analysis

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The aerosol mass deposited on the filters (µg) was obtained by weighing the quartz filter before and after

sampling, after a period of 48 hours of conditioning in a room with controlled atmospheric conditions

(temperature, T ~ 20  $\pm$  1 °C; relative humidity, RH ~ 50  $\pm$  5%). Weighting is performed with an analyt-

ical balance (Sartorius model MC5, precision of 1 µg), and repeated three times to control the statistical

variability of the measurement. Electrostatic effects are removed by exposing the filters, prior weighing,

to a de-ionizer. The error on-in the measured mass is estimated at 10 µg, including the repetition varia-

bility. The aerosol mass concentration (µg m<sup>-3</sup>) is obtained by dividing the mass deposited on the filter

to the total volume of sampled air (m<sup>3</sup>) obtained from the mass flowmeter measurements (+-5%). The

percent error on mass concentrations is estimated to 5%.

## 2.5. Dust composition measurements

## 2.5.1. Elemental composition

183 Elemental concentrations for the major constituents of mineral dust (Na, Mg, Al, Si, P, S, Cl, K, Ca, Fe,

Ti, Mn) were obtained by a Wavelength Dispersive X-ray fluorescence (WD-XRF) of the

Nuclepore filters using a PW-2404 spectrometer by Panalytical. Excitation X-rays are produced by a

Coolidge tube ( $I_{max} = 125 \text{ mA}$ ,  $V_{max} = 60 \text{ kV}$ ) with a Rh anode; the primary X-ray spectrum can be

controlled by inserting filters (Al, at different thickness) between the anode and the sample. Each ele-

ment was analyzed three times, with specific conditions (voltage, tube filter, collimator, analyzing crys-

tal, and detector). Data collection was controlled by the SuperQ software provided with the instrument.

The elemental mass thickness (µg cm<sup>-2</sup>), that is, the analyzed elemental mass per unit surface, was ob-

tained by comparing the elemental yields with a sensitivity curve measured in the same geometry on a

set of certified mono- or bi-elemental thin layer standards by Micromatter Inc. The certified uncertainty

of the standard deposit (± 5%) determines the lower limit of the uncertainty of the measured ele-

mental concentrations, which ranges between 8% and 10% depending on the element considered-ele-

195 ment. Thanks to the uniformity of the aerosol deposit on the filters, the atmospheric elemental concen-

trations (µg m<sup>-3</sup>) were calculated by multiplying the analyzed elemental mass thickness by the ratio

between the collection and analyzed surfaces of each sample (41 and 22 mm, respectively), then di-

vidinged by the total sampled volume (m<sup>3</sup>). Finally, concentrations of light-weight elements (atomic

- number Z < 19) were corrected for the underestimation induced by the self-absorption of the emitted
- soft X-rays inside aerosol particles according to Formenti et al. (20112010).
- Additional XRF analysis of the quartz filters was has been performed both in the PM<sub>10.6</sub> and the PM<sub>2.5</sub>
- fractions, so-to verify the absence of biases between the experiments dedicated to the determination of
- particle composition to and those where the optical properties where measured.

### 2.6.2. Iron oxide content

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- 205 The content and the mineralogical speciation of the iron oxides, also defined as free-iron, that isi.e., the
- fraction of iron which that is not in the crystal lattice of silicates (Karickhoff and Bailey, 1973), was
- determined by XANES (X-ray absorption near-edge structure) in the Fe K-range ( $K_{\alpha}$ , 7112 eV) at the
- SAMBA (Spectroscopies Applied to Materials based on Absorption) beamline at the SOLEIL synchro-
- tron facility in Saclay, France (Briois et al., 2011). The position and shape of the K pre-edge and edge
- 210 peaks were analyzed as they depend on the oxidation state of iron and the atomic positions of the neigh-
- 211 boring ions, mostly O<sup>+</sup> and OH<sup>-</sup>.
- As in Formenti et al. (2014b), samples were mounted in an external setup mode. A Si(220) double-
- crystal monochromator was used to produce a monochromatic X-ray beam, which was  $3000 \times 250 \mu m^2$
- in size at the focal point. The energy range was scanned from 6850 eV to 7800 eV at a step resolution
- varying between 0.2 eV in proximity to the Fe-K absorption edge (at 7112 eV) to 2 eV in the extended
- 216 range. Samples were analyzed in fluorescence mode without prior preparation. One scan acquisition
- 217 lasted approximately 30 minutes, and was repeated three times to improve the signal-to-noise ratio.
- 218 The same analytical protocol was applied to five standards of Fe(III)-bearing minerals (**Table 1**), includ-
- 219 ing iron oxides (hematite, goethite) and silicates (illite, montmorillonite, nontronite). The standard spec-
- 220 tra were used to deconvolute the dust sample spectra to quantify the mineralogical status of iron. The
- linear deconvolution was, performed with the Athena IFEFFIT freeware analysis program (Ravel and
- Newville, 2005). This, provided with the proportionality factors,  $\alpha_{i}$  representing the mass fraction of
- elemental iron to be assigned to the *i*-th standard mineral. In particular, the values of  $\alpha_{hem}$  and  $\alpha_{goe}$
- represent the mass fractions of elemental iron that can be attributed to hematite and goethite, and  $\alpha_{Fe\ ox}$
- $(\alpha_{hem} + \alpha_{goe})$ , the mass fraction of elemental iron that can be attributed to iron oxides.

### 2.6.3. Calculation of the iron oxide content

The measured elemental concentrations obtained by X-ray Fluorescence (XRF) are expressed in the

form of elemental oxides and summed to estimate the total mineral dust mass concentration  $MC_{dust}$  ac-

cording to the equation from Lide (1992)

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$$[MC_{dust}] = 1.12 \times \left\{ \begin{array}{l} 1.658 [Mg] + 1.889 [Al] + 2.139 [Si] + 1.399 [Ca] + 1.668 [Ti] + 1.582 [Mn] \\ + (0.5 \times 1.286 + 0.5 \times 1.429 + 0.47 \times 1.204) [Fe] \end{array} \right\}$$
 (4)

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233 The relative uncertainty on in  $MC_{dust}$ , estimated from the analytical error ion the measured

concentrations, does not exceed 6%. As it will be explained in the result section (paragraph 3.1), the

values of MC<sub>dust</sub> estimated from Equation 4 were found in excellent agreement with the measured

236 gravimetric mass on the filters.

The fractional mass ratio (in percent) of elemental iron ( $MR_{Fe\%}$ ) with respect to the total dust mass con-

centration,  $MC_{dust}$  is then calculated as

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$$MR_{Fe\%} = \frac{[Fe]}{[MC_{Dust}]} \times 100 \tag{5}$$

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The mass concentration of iron oxides or free-iron ( $MC_{Fe\ ox}$ ), representing the fraction of elemental iron

in the form of hematite and goethite (Fe<sub>2</sub>O<sub>3</sub> and FeOOH, respectively), is equal to

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$$MC_{Fe \ ox} = MC_{hem} + MC_{goe} \tag{6}$$

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247 where  $MC_{hem}$  and  $MC_{goe}$  are the total masses of hematite and goethite. These can be calculated from the

values  $\alpha_{hem}$  and  $\alpha_{goe}$  from XANES analysis, which represent the mass fractions of elemental iron at-

tributed to hematite and goethite, as

$$MC_{hem} = \frac{\alpha_{hem} x [Fe]}{0.70}$$
 (7.a)

$$MC_{goe} = \frac{\alpha_{goe} x [Fe]}{0.63}$$
 (7.b)

where the values of 0.70 and 0.63 represent the mass molar fractions of Fe in hematite and goethite, respectively. The relative errors of  $MC_{hem}$  and  $MC_{goe}$  are obtained from the uncertainties on of the values of  $\alpha_{hem}$  and  $\alpha_{goe}$  from XANES analysis (less than 10%).

The mass ratio of iron oxides (MR<sub>Fe ox%</sub>) with respect to the total dust mass can then be calculated as

$$MR_{Fe\ ox\%} = MC_{Fe\ ox} \times MR_{Fe\ \%} \tag{8}$$

## 3. Experimental protocol

At the beginning of each experiment, the chamber was evacuated by to 10<sup>-4</sup>-10<sup>-5</sup> hPa. Then, the reactor was filled with a mixture of 80% N<sub>2</sub> and 20% O<sub>2</sub> at a pressure slightly exceeding the current atmospheric pressure, in order to avoid contamination from ambient air. The eExperiments were conducted at ambient temperature and at a relative humidity <2%. As in Di Biagio et al. (2014; 20162017), dust aerosols were generated by mechanical shaking of the parent soils, previously sieved to < 1000 μm and dried at 100 °C for about 1 h to remove any residual humidity. About 15 g of soil was placed in a Buchner flask and shaken for about 30 min at 100 Hz by means of a sieve shaker (Retsch AS200). The dust particles produced wproduced by the mechanical shaking, mimicking the saltation processing that soils experience when eroded by strong winds, as were then injected in the chamber by flushing the flask with N<sub>2</sub> at 10 L min<sup>-1</sup> for about 10-15 min, whilst continuing shaking the soil. Di Biagio et al. (2014; 2017) have demonstrated the realism of the generation system concerning the composition and the size distribution of the generated dust with respect to the properties of mineral dust in the atmosphere.

<u>The Dd</u>ust <u>was injected for about 10-15 minutes</u>, <u>and left\_remained</u> suspended in the chamber for approximately 120 min thanks to the 4-wheel fan located in the bottom of the chamber body. Previous measurements at the top and bottom of the chamber showed that the fan ensures a homogeneous distribution of the dust starting approximately 10 minutes after the end of the injection (Di Biagio et al., 2014).

To compensate for the air extracted from the chamber by sampling, a particle-free flow of N<sub>2</sub>/O<sub>2</sub>, regulated in real time as a function of the total volume of sampled air, was re-injected in the chamber. To avoid excessive dilution the flow was limited to 20 L min<sup>-1</sup>. Two experiments per soil type were conducted: a first experiment for sampling on the nuclepore polycarbonate filters (determination of the elemental composition and the iron oxide fraction) and *in situ* measurements of the infrared optical constants (Di Biagio et al., 20162017), and a second experiment sampling on total quartz filter and impactor for the study of dust MAE presented in this paper.

**Figure 1** illustrates as typical example the time series of the aerosol mass concentration during the two experiments conducted for the Libyan sample. The comparison demonstrates the repeatability of the dust concentrations, both in absolute values and in temporal dynamics. It also shows that the mass concentrations decreased very rapidly by gravitational settling within the first 30 minutes of the experiment (see also the discussion in Di Biagio et al., (20162017)), after which concentrations only decrease by dilution. The filter sampling was started after this transient phase, and then continued through the end of the experiments, in order to collect enough dust on the filter membranes for subsequent the chemical analysis. Blank samples were collected before the start of the experiments by placing the loaded filter holders loaded with filter membranes in line with the chamber and by flushing them for a few seconds with air coming from the chamber.

At the end of each experimental series with a given soil sample, the chamber was manually cleaned in order to remove carry-over caused by resuspension of particles deposited to the walls. Background concentrations of aerosols in the chamber vary between 0.5 and 2.0 µg m<sup>-3</sup>, i.e., a factor of 500 to 1000 below the operating conditions.

### **34.** Results and discussion

The geographical location of the soil collection sites is shown in **Figure 2**, whereas and the coordinates are summarized in **Table 2**. As discussed in Di Biagio et al. (2016), the selection of these soils and sediments was governed by the need of representing the major arid and semi-arid regions worldwide, the need of taking into account the mineralogical diversity of the soil composition at the global scale, and finally by their availability in sufficient quantities for injection in the chamber. When doing so, we obtained a set of twelve samples distributed worldwide but mostly in Northern and Western Africa (Libya, Algeria, Mali, Bodélé) and the Middle East (Saudi Arabia and Kuwait). Individual samples from the Gobi desert in Eastern Asia, the Namib Desert, the Strzelecki desert in Australia, the Patagonian

309 deserts in South America, and the Sonoran Desert in Arizona have also been investigated. The selection 310 of these soils and sediments was made out of 137 individual top-soil samples collected in major arid and 311 semi-arid regions worldwide and representing the mineralogical diversity of the soil composition at the 312 global scale. As discussed in Di Biagio et al. (2017), this large sample set was reduced byto a set of 19 313 samples representing the mineralogical diversity of the soil composition at the global scale and based on their availability in sufficient quantities for injection in the chamber. Because some of the experiments 314 315 did not produce enough dust to perform good-quality optical measurements, in this paper we present a set of twelve samples distributed worldwide but mostly fromin Northern and Western Africa (Libya, 316 317 Algeria, Mali, Bodélé) and the Middle East (Saudi Arabia and Kuwait). Individual samples from the 318 Gobi desert in Eastern Asia, the Namib Desert, the Strzelecki desert in Australia, the Patagonian deserts in South America, and the Sonoran Desert in Arizona havewere also been investigated. 319

# 34.1. Elemental composition and iron oxide content

- A total of 41 filters including 15 polycarbonate filters (12 samples and 3 blanks) and 25 quartz filters
- (12 for the total fraction, 10 for the fine fraction and 3 blanks) were collected for analysis.
- The dust mass concentration found by gravimetric analysis varied between 50 μg m<sup>-3</sup> and 5 mg m<sup>-3</sup>, in
- relatively good agreement with the dust mass concentrations,  $MC_{dust}$ , from (Equation 4), based on  $\times$
- 325 ray Ray fluorescence Fluorescence XRF analysis Analysis: the slope of the linear regression between
- the calculated and the gravimetric values of  $MC_{dust}$  is 0.90 with  $R^2 = 0.86$ .
- Di Biagio et al. (20162017) showed that clays are the most abundant mineral phases, together with quartz
- and calcite, and that significant variability exists as function of the compositional heterogeneity of the
- parent soils. Here we use the Fe/Ca and Si/Al elemental ratios obtained from X-ray Fluorescence XRF
- analysis to discriminate the origin of used dust samples. These ratios have been extensively used in the
- past to discriminate the origin of African dust samples collected in the field (Chiapello et al., 1997;
- Formenti et al., 2011; Formenti et al., 2014a). The values obtained during our experiments are reported
- in **Table 3.** There is a very good correspondence between the values obtained for the Mali, Libya, Al-
- 334 geria, Mauritania and (to a lesser extent) Morocco experiments to values found in environmental aerosol
- samples by Chiapello et al. (1997) and Formenti et al. (2011; 2014a). These authors indicate that dust
- from local erosion of Sahelian soils, such as from Mali, have Si/Al ratios in the range of 2-2.5 and Fe/Ca
- ratios in the range 3-20, depending on the time proximity to the erosion event. Dust from sources in the
- Sahara, such as Libya and Algeria, show Si/Al ratios in the range of 2-3 and Fe/Ca ratios in the range

0.7-3, whereas dust from Morocco has Si/Al ratios around 3 and Fe/Ca ratios around 0.4. The only major difference is observed for the Bodélé experiment, for which the Fe/Ca ratio is enriched by a factor of 6 with respect to the values of 1 found during to the field observations (Formenti et al., 2011; Formenti et al., 2014a). This could reflect the fact that the Bodélé aerosol in the chamber is generated from a sediment sample and not from a soil. As a matter of fact, the Bodélé sediment sample is constituted consists of by a very fine powder which becomes very easily airborne. Henceforth, and contrary to the soil samples, this This powder is likely to be injected in the chamber with little or no size fractionation. Henceforth, the aerosol generated from it. As a consequence, should have a closer the composition to the original powder of the aerosol collected in the chamber could reflect more that of the parent sedimentary soil than the not the other samples. On the other hand, Bristow et al. (2010) and Moskowitz et al. (2016) showed that the iron content and speciation of the Bodélé sediments is very heterogeneous at the source scale. For samples from areas other than non-northern African samples, the largest variability is observed for the Fe/Ca values, ranging from 0.1 to 8, whereas the Si/Al ratio varied only between 2.5 and 4.8. In this case, values are available in the literature for comparison (e.g., Cornille et al., 1990; Reid et al., 1994; Eltayeb et al., 2001; Lafon et al., 2006; Shen et al., 2007; Radhi et al., 2010; 2011; Formenti et al., 2011; 2014a; Scheuvens et al., 2013, and references within). Values in the PM<sub>2.5</sub> fraction are very consistent with those obtained in the PM<sub>10.6</sub>: their linear correlation has a slope of 1.03 ( $\pm$  0.05) and a  $R^2$ equal to 0.97, suggesting that the elemental composition is relatively size -independent. The mass fraction of total Fe ( $MC_{Fe\%}$  from Equation 5), also reported in **Table 3**, ranged from 2.8 (Namibia) to 7.3% (Australia), values found for the Namibia and the Australia samples, respectively. This These are in the range is in good agreement with of values reported in the literature, taking into account that differences might be also due to the method (direct measurement/calculation) and/or the size fraction over which the total dust mass concentration is estimated (Chiapello et al., 1997; Reid et al., 1994; 2003; Derimian et al., 2008; Formenti et al., 2001; 2011; 2014a; Scheuvens et al., 2013). The agreement of MC<sub>Fe</sub>% values obtained by the XRF analysis of polycarbonate filters (Equation 5) and those obtained from the XRF analysis of the quartz filters, normalized to the measured gravimetric mass is well within 10% (that is, the percent error of each estimate). An the e-Exceptions are the samples of from Bodélé and Algeria, for which the values obtained from the analysis of the quartz filters are significantly lower than those obtained from the nuclepore filters (3.1% versus 4.1% for Bodélé and 4.3% versus 6.8% for Algeria). We treat that as an additional source of error in the rest of the analysis, and add it to the total uncertainty. In the PM<sub>2.5</sub> fraction<sub>.5</sub> the content of iron is more variable, ranging from 4.4% (Morocco) to

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370 33.6% (Mali), showing a size dependence. A word of caution on this conclusion as is that the two esti-371 mates are not necessarily consistent in the way that the total dust mass is estimated (from Equation 4 for 372 the PM<sub>10.6</sub> fraction and by gravimetric weighing <u>foron</u> the PM<sub>2.5</sub>). 373 Finally, between 11 and 47% of iron in the samples can be attributed to iron oxides, in variable propor-374 tions between hematite and goethite. The iron oxide fraction of total Fe in this study is on-at the lower end of the range (36-72%) estimated for field dust samples of Saharan/Sahelian origin (Formenti et al. 375 376 2014b). The highest value of Formenti et al. (2014b), obtained for a sample of locally-emitted dust col-377 lected at the Banizoumbou station in the African Sahel, is anyhow in excellent agreement with the value of 62% obtained for an experiment (not shown here) using a soil collected in the same area. Likewise, 378 379 the proportions between hematite and goethite (not shown) are reproduced, showing that goethite is more 380 abundant than hematite. The mass fraction of iron oxides ( $MR_{Fe\ ox\%}$ ), estimated from Equation 8 and 381 shown in Table 3, ranges between 0.7% (Kuwait) to and 3.6% (Australia), which is in the range of 382 available field estimates (Formenti et al., 2014a; Moskowitz et al., 2016). For China, our value of MRFe ox% is lower by almost a factor of 3 in comparison with compared to that obtained on the same dust of the 383 384 same origin sample by Alfaro et al. (2004) (0.9% against 2.8%), whereas on a sample from Niger (how-385 ever not considered in this study) our estimates and that by Alfaro et al. (2004) agree perfectly agree (5.8%). A possible underestimate of the iron oxide fraction for samples other than those from the Sahara-386 387 Sahel area could be due to the fact that - opposite to the experience of Formenti et al. (2014b) - the linear deconvolutions of the XANES spectra were not always satisfactory (see Figure S1 in the supplementary). 388 389 This resulted in a significant residual between the observed and fitted XANES spectra. Indeed In fact, the mineralogical reference for hematite is obtained from a soil from Niger (Table 1) and might not be 390 391 fully suitable for representing aerosols of different origins. Additional differences could arise from dif-392 ferences in the size distributions of the generated aerosol. As a matter of fact, the number fraction of 393 particles in the size classes above 0.5 µm in diameter is are different in the dust aerosol generated in the 394 Alfaro et al. (2004) study with respect compared to ours. In the study by Alfaro et al. (2004), the number 395 fraction of particles is lowest in the 0.5-0.7 size class and highest between 1 and 5  $\mu$ m. On the contrary In 396 contrast, in our study the number fraction is lowest in the 1-2 µm size range and highest between 0.5 397 and 0.7 µm. These differences could vield either be due to differences in the chemical composition

## 34.2. Spectral and size -variability of the mass absorption efficiency

and/or to a difference in the total mass in the denominator of Equation 8.

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400 The spectral mass absorption efficienciesy (MAE) at 375, 407, 532, 635, and 850 nm for the PM<sub>10.6</sub> and 401 the PM<sub>2.5</sub> dust fractions are summarized in **Table 4** and displayed in **Figure 3**. Regardless of particle size, the MAE values decrease with increasing wavelength (almost one order of magnitude between 375 402 403 and 850 nm), and display a larger variability at shorter wavelengths. The MAE values for the PM<sub>10.6</sub> range from 37 ( $\pm$  3) 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup> to 135 ( $\pm$  11) 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup> at 375 nm, and from 1.3 ( $\pm$  0.1) 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup> to 15 404 (± 1) 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup> at 850 nm. Maxima are found for the Australia and Algeria samples, whereas the minima 405 406 are for Bodéelé and Namibia, respectively at 375 and 850 nm. In the PM2.5 fraction, the MAE values range from 95 ( $\pm$  8)  $10^{-3}$  m<sup>2</sup> g<sup>-1</sup> to 711 ( $\pm$  70)  $10^{-3}$  m<sup>2</sup> g<sup>-1</sup> at 375 nm, and from 3.2 ( $\pm$  0.3)  $10^{-3}$  m<sup>2</sup> g<sup>-1</sup> to 36 407  $(\pm 3)$  10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup> at 850 nm. Maxima at both 375 and 850 nm are found for the Morocco sample, whereas 408 the minima are for Algeria and Namibia, respectively. The MAE values for mineral dust resulting from 409 410 this work are in-relatively in good agreement with the estimates available in the literature (Alfaro et al., 2004; Linke et al., 2006; Yang et al., 2009; Denjean et al., 2016), reported in Table 5. For the China 411 Ulah Buhn sample, Alfaro et al. (2004) reported 69.1 10<sup>-3</sup> and 9.8 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup> at 325 and 660 nm, respec-412 tively. The former is lower than the value of 99 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup> that we obtain by extrapolating our measure-413 414 ment at 375 nm. Likewise, our values for the Morocco sample are higher than reported by Linke et al. (2006) at 266 and 660 nm. Conversely, the agreement with the estimates of Yang et al. (2009) for mineral 415 416 dust locally re-suspended in Xianghe, near Beijing (China) is very good at all wavelengths between 375 and 880 nm. As expected, the MAE values for mineral dust resulting from this work are almost one 417 418 order of magnitude smaller than for other absorbing aerosols. For black carbon, MAE values are in the range of 6.5–7.5 m<sup>2</sup> g<sup>-1</sup> at 850 nm (Bond and Bergstrom, 2006; Massabò et al., 2016), and vary decrease 419 in a linear way with the logarithm of the inversely with wavelength. For brown carbon, the reported 420 MAE range between 2.3–7.0 m<sup>2</sup> g<sup>-1</sup> at 350 nm (Chen and Bond, 2010; Kirchstetter et al., 2004; Massabò 421 et al., 2016), 0.05-1.2 m<sup>2</sup> g<sup>-1</sup> at 440 nm (Wang et al., 2016) and 0.08-0.72 m<sup>2</sup> g<sup>-1</sup> at 550 nm (Chen and 422 Bond, 2010). 423 The analysis of **Table 4** indicates that, at every wavelength, the MAE values in the PM<sub>2.5</sub> fraction are 424 425 equal or higher than those for PM<sub>10.6</sub>. The PM<sub>2.5</sub>/PM<sub>10.6</sub> MAE ratios reach values of 6 for the Mali sam-426 ple, but are mostly in the range 1.5-3 for the remaining other aerosols. The v-Values decrease with wave-427 length up to 635 nm, whereas at 850 nm they have values comparable to those at 375 nm. The observed 428 size\_-dependence of the MAE values is consistent with the expected behavior of light absorption of 429 particles in the Mie and geometric optical regimes that concern are relevant for the two size fractions. Light\_-absorption of particles of sizes smaller or equivalent to the wavelength is proportional to their 430

431 bulk volume, whereas for larger particles absorption occurs on their surface only (Bohren and Huffmann, 1983). On the other hand, the size-resolved measurements of Lafon et al. (2006) show that the proportion 432 433 (by volume) of iron oxides might be higher in the coarse than in the fine fraction, which would counteract 434 the size-dependence behavior of MAE. To validate the observations, we calculated the spectrally-re-435 solved MAE values in the two size fractions using the Mie code for homogeneous spherical particles 436 (Bohren and Huffmann, 1983) and the number size distribution estimated by (Di Biagio et al., 437 2016(2017) and averaged over the duration of filter sampling. We estimated the dust complex refractive index as a volume-weighted average of a non-absorbing dust fraction having the refractive index of 438 439 kaolinite, the dominant mineral in our samples (see Di Biagio et al., 20162017), from Egan et Hilgeman 440 (1979) and an absorbing fraction estimated from the mass fraction of iron oxides and having the refractive index of hematite (Bedidi and Cervelle, 1993). The rResults of this calculation indicate that the 441 442 observed size-dependent behavior is well reproduced at all wavelengths, even in the basic hypothesis 443 that the mineralogical composition does not change with size. The only exception is 850 nm, where at 444 times, PM<sub>2.5</sub>/PM<sub>10.6</sub> MAE ratio is much higher than expected theoretically. We attribute that to the rela-445 tively high uncertainty affecting the absorbance measurements at this wavelength, where the signal-to-446 noise ratio is low. Indeed, the two sets of values (MAE in the PM<sub>2.5</sub> fraction and MAE in the PM<sub>10.6</sub> 447 fraction) are not statistically different according to a two-pair t-test (0.01 and 0.05 level of confidence), 448 confirming that any attempt of differentiation of the size -dependence at this wavelength would require 449 a stronger optical signal. 450 The analysis of the spectral dependence, using a the power-law function fit as (from Equation 2), pro-451 vides with the values of the Angstrom Absorption Exponent (AAE), also reported in **Table 4**. Contrary 452 to the MAE values, there is no statistically significant size\_dependence of the AAE values, ranging from

2.5 (± 0.2) to 4.1 (± 0.3), with an average of 3.3 (± 0.7), for the PM<sub>10.6</sub> size fraction and between 2.6 (± 0.2) and 5.1 (± 0.4), with an average of 3.5 (± 0.8), for the PM<sub>2.5</sub> fraction. Our values are in the range of those published in the open-literature (Fialho et al., 2005; Linke et al., 2006; Müller et al., 2009; Petzold et al., 2009; Yang et al., 2009; Weinzierl et al., 2011; Moosmüller et al., 2012; Denjean et al., 2016), shown in **Table 5**. AAE values close to 1.0 are found for urban aerosols where fossil fuels combustion is dominant, while AAE values for brown carbon (BrC) from incomplete combustion are in the range 3.5-4.2 (Yang et al., 2009; Chen et al., 2015; Massabò et al., 2016).

Finally, **Figure 4** shows correlations between the MAE values in the PM<sub>10.6</sub> fraction (Figure 34.a) and in the PM<sub>2.5</sub> fraction (Figure 34.b) and the estimated percent mass fraction of iron and iron oxides

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(*MCFe*% and *MCFe* ox%), respectively. Regardless of the size fraction, the correlation between the MAE values and the percent mass of total elemental iron are satisfactory. Higher higher correlations are obtained at 375, 407 and 532 nm, and in the PM<sub>2.5</sub> fraction, where a linear correlation with  $R^2$  up to 0.94 are obtained. Best correlations are obtained when forcing the intercept to zero, indicating that elemental iron fully accounts for the measured absorption. At these wavelengths, linear correlations with the mass fraction of iron oxides are loose low in the PM<sub>10.6</sub> mass fraction ( $R^2$  up to 0.38-0.62), but again satisfactory higher in the PM<sub>2.5</sub> fraction ( $R^2$  up to 0.83-0.99), where, whoeverhowever, one should keep in mind that they have been established only indirectly by considering the ratio of iron oxides to elemental iron independent of size. At 660 and 850 nm, little or no robust correlations are is obtained, often based on very few data points and with very low MAE values. It is noteworthy that, in both size fractions, the linear correlation yields a non-zero intercept—is obtained, indicating a contribution from minerals other minerals butthan iron oxides account to for the measured absorption.

## 45. Conclusive remarks

- In this paper, we reported new laboratory measurements of the shortwave mass absorption efficiency
- 476 (MAE) of mineral dust of different origins and as a function of size and wavelength in the 375-850 nm
- 477 range. Results Our results have been were obtained in the CESAM simulation chamber using generated
- 478 mineral dust generated from natural parent soils, in combination with and optical and gravimetric anal-
- 479 ysis on extracted samples.
- Our results can be summarized as follows: at 375 nm, the MAE values are lower for the  $PM_{10.6}$  mass
- fraction (range 37-135  $10^{-3}$  m<sup>2</sup> g<sup>-1</sup>) than for the PM<sub>2.5</sub> fraction (range 95-711  $10^{-3}$  m<sup>2</sup> g<sup>-1</sup>), and vary oppo-
- site to wavelength as  $\lambda^{-AAE}$ , where AAE (Angstrom Absorption Exponent) averages between 3.3-3.5
- regardless of size fraction. These results deserve some concluding comments:
  - The size\_-dependence, <u>yielding\_characterized by</u> significantly higher MAE values in the fine fraction (*PM*<sub>2.5</sub>) than <u>for the in the bulk</u> (*PM*<sub>10.6</sub>-) aerosol, indicates that light\_-absorption by mineral dust can be important even during atmospheric transport over heavily polluted regions, <u>when</u> <u>where</u> dust concentrations are significantly lower than at emission. This can be shown by comparing the aerosol absorption optical depth (AAOD) at 440 nm for China, a well-known mixing region of mineral dust and pollution (e.g., Yang et al., 2009; Laskin et al., 2014; Wang et al., 2013), <u>as well as offshore western Africa where large urban centers are downwind of dust transport areas (Petzold et al., 2011). Laskin et al. (2014) reports that the average AAOD in <u>China</u></u>

the area is of the order of  $0.1_7$  for carbonaceous absorbing aerosols (sum of black and brown carbon; Andreae and Gelencsér, 2006). This is lower or comparable to the AAOD of 0.17 and 0.11 at 407 nm (fine and total and fine mass fractions, respectively) that we obtain derive by a simple calculation (AAOD = MAE x MC<sub>dust</sub> x H), where from MAE are the values estimated in this study, (, MC<sub>dust</sub>, the dust mass concentrations typically observed in the area urban area of Beijing during dust storms (Sun et al., 20012005), and  $H_1$  a scale height factor of 1 km).

- The spectral variability of the dust MAE values, represented by the AAE parameter, is equal in the PM<sub>2.5</sub> and PM<sub>10.6</sub> mass fractions. This suggests that, for a given size distribution, the possible variation of dust composition with size does not affect in a significant way the spectral behavior of the absorption properties. Our average value for AAE is  $3.3 \pm 0.7$ , higher than for black carbon, but in the same range than as light-absorbing organic (brown) carbon. As a result, depending on the environment, there can be some ambiguity in apportioning the AAOD based on spectral dependence. Bahadur et al. (2012) and Chung et al. (2012) couple the AAE and the spectral dependence of the total AOD (and/or its scattering fraction only) to overcome this problem. Still, Bahadur et al. (2012) show that there is an overlap in the scatterplots of the spectral dependence of the scattering and absorption fractions of the AOD based on an analysis of ground-based remote sensing data for mineral dust, urban, and non-urban fossil fuel over California. A closer look should be taken at\_to-observations in mixing areas where biomass burning aerosols may have different chemical composition and/or mineral dust has heavy loadings should be given in order to generalize the clear separation observed in the spectral dependences of mineral dust and biomass burning (Bahadur et al., 2012). This aspect is relevant to the development of remote sensing retrievals of light-absorption by aerosols from space, and their assimilation in climate models (Torres et al., 2007; Buchard et al., 2015; Hammer et al., 2016).
  - There is an important sample-to-sample variability in our dataset of MAE values for mineral dust aerosols. At 532 nm, our <u>estimated\_average\_MAE values\_average\_toare\_at</u> 34 ± 14 m<sup>2</sup> g<sup>-1</sup> and 78 ± 70 m<sup>2</sup> g<sup>-1</sup> in the PM<sub>10.6</sub> and PM<sub>2.5</sub> mass fractions, respectively. Figure 3, showing the correlation with the estimated mass fraction of elemental iron and iron oxides, suggests that this variability could be related to the regional differences of the mineralogical composition of the parent soils. These observations lead to <u>different considerations\_further conclusions</u>. To start with, our study reinforces the need for regionally-resolved representation of the light\_-absorption properties of

mineral dust in order to improve the representation of its effect on climate. As a matter of fact, the natural variability of the absorption properties that we obtain from our study is in the range 50-100%, even when we limit ourselves to smaller spatial scales, for example those of from north Africa (samples from Libya, Algeria, Mali and Bodélé). This is far above the  $\pm$  5% sensitivity factor used by Solmon et al. (2008) to vary the single scattering albedo (as a proxy of absorption) of mineral dust over western Africa, and to show how this could drastically change the climate response in the region. As a comparison, Solmon et al. (2008) showed that varying the single scattering albedo of mineral dust over western Africa by  $\pm$  5%, that is, varying the co-albedo (or absorption) by 45% (0.1 $\pm$  0.045) could drastically change the climate response in the region.

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The question is then "how to represent this regional variability?" As-Like Moosmüller et al. (2012) and Engelbrecht et al. (2016), we found that elemental iron is a very good proxy for the MAE, especially in the PM<sub>2.5</sub> fraction, where iron-bearing absorbing minerals (hematite, goethite, illite, smectite clays) would beare more concentrated. In the coarse fraction, Ca-rich minerals, quartz, and feldspars could also play a role, and that could result in the observed lowered correlation (although adding a term proportional to elemental Ca does not ameliorate improve the <u>correlation-result</u> in the present study). The correlation of the spectral MAE values with the iron oxide fraction is satisfactory but rather noisy, also owing to some uncertainty in the quantification of iron oxides from X-Ray Absorption absorption measurements. In this case, the intercept is significantly different from zero, again indicating that a small but elear-distinct fraction of absorption is due to minerals other than iron oxides. There are contrasting results on this topic: Alfaro et al. (2004) found an excellent correlation between MAE and the iron oxide content, whereas Klaver et al. (2011) found that the single scattering albedo (representing the capacity of an aerosol population to absorb light with respectin relation to extinction) was almost independent on the mass fraction of iron oxides. Moosmüller et al. (2012) disagreed, pointing out to the uncertainty in the correction procedure of the measurement of absorption by Klaver et al. (2011). As a matter of fact, Klaver et al. (2011) and Alfaro et al. (2004) used the same correction procedure. It is more likely that the lack of correlation found in Klaver et al. (2011) is due to the fact that other minerals other than iron oxides contribute to absorption, in particular at their working wavelength (567 nm), where the absorption efficiency of iron oxides starts to weaken. Clearly, the linear correlation between elemental iron in mineral dust and its light-absorption properties could ease the application and validation of climate models that are now starting to include including the representation of the mineralogy (Perlwitz et al., 2015a; 2015b; Scanza et al., 2015). Also, they this would facilitate detecting source regions based on remote sensing of dust absorption in the UV-VIS spectral region (e.g., Hsu et al., 2004). However, such a quantitative relationship cannot be uniquely determined from these studies, including the present one, which use different ways of estimating elemental iron, iron oxides, and the total dust mass. A more robust estimate should be obtained from these measurements of absorption, and their dependence on the mineralogical composition.

### **Author contributions**

L. Caponi, P. Formenti, D. Massabò, P. Prati, C. Di Biagio, and J. F. Doussin designed the chamber experiments and discussed the results. L. Caponi and C. Di Biagio realized conducted the experiments with contributions by M. Cazaunau, E. Pangui, P. Formenti, and J.F. Doussin. L. Caponi, D. Massabò and P. Formenti performed the full data analysis with contributions by C. Di Biagio, P. Prati and J.F. Doussin. L. Caponi, P. Formenti and S. Chevaillier performed the XRF measurements. P. Formenti and G. Landrot performed the XAS measurements. D. Massabò performed the MWAA and the gravimetric measurements.-M. O. Andreae, K. Kandler, T. Saeed, S. Piketh, D. Seibert, and E. Williams collected the soil samples used for experiments. L. Caponi, P. Formenti, D. Massabò and P. Prati wrote the manuscript with comments from all co-authors.

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## 932 **Table captions**

- Table 1. Characteristics of the standards used for the quantification of the iron oxides in the XAS anal-
- 934 ysis.
- Table 2. Summary Geographical of information on the soil samples used in this work.
- Table 3. Chemical characterisation of the dust aerosols in PM<sub>10.6</sub> and PM<sub>2.5</sub> (in parentheses) size frac-
- tions. Columns 3 and 4 report give the Si/Al and Fe/Ca elemental ratios obtained from X-Ray Fluores-
- cence analysis. The uncertainty of each individual value is estimated to be 10%. Column 5 reports
- $MR_{Fe\%}$ , the fractional mass of elemental iron with respect to the total dust mass concentration
- (uncertainty 10%). Column 5 reports  $MR_{Fe\%}$ , the mass fraction of iron oxides with respect to the total
- dust mass concentration (uncertainty 15%). For PM<sub>2.5</sub> the determination of the Si/Al ratio is impossible
- due to the composition of the filter medium\_membranes (quartz).
- Table 4. Mass absorption efficiency (MAE, 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup>) and Ångström Absorption Exponent (AAE) in
- 944 the PM<sub>10.6</sub> and PM<sub>2.5</sub> size fractions. Absolute errors are in brackets.
- Table 5. Mass absorption efficiency (MAE, 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup>) and Ångström Absorption Exponent (AAE) of
- 946 <u>from the literature data discussed in the paper</u>

# 948 Figure captions

- Figure 1. Time series of aerosol mass concentration in the chamber for the two companion experiments
- 950 (Libyan dust sample). Experiment 1 (top panel) was dedicated to the determination of the chemical
- composition (including iron oxides) by sampling on polycarbonate filters. Experiment 2 (bottom panel)
- was dedicated to the determination of the absorption optical properties by sampling on quartz filters.
- Figure 2. Locations (red stars) of the soil and sediment samples used to generate dust aerosols.
- Figure 3. Spectral dependence of the MAE values for the samples investigated in this study in the PM<sub>10.6</sub>
- 955 (left) and in the PM<sub>2.5</sub> (right) mass fractions.
- 956 **Figure 4.** Illustration of the links between the MAE values and the dust chemical composition found in
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- mass  $(MR_{Fe}\%)$  in the PM<sub>10.6</sub> fraction; Middle column: same as left column but respect to for the mass

- fraction of iron oxides <u>relative</u> to the total dust mass ( $MR_{Fe\ ox\%}$ ) in the PM<sub>10.6</sub> size fraction; Right column:
- same as left column but in the  $PM_{2.5}$  size fraction.

**Table 1.** Characteristics of the standards used for the quantification of the iron oxides in the XAS analysis.

Standard	Stoechiometric Stoichiometric Formula	Origin
Illite of Puy	$(Si_{3.55}Al_{0.45})(Al_{1.27}Fe_{0.36}Mg_{0.44})O_{10}(OH)_2(Ca_{0.01}Na_{0.01}K_{0.53}X(I)_{0.12})$	Puy, France
Goethite	FeO OH	Minnesota
Hematite	Fe <sub>2</sub> O <sub>3</sub>	Niger
Montmorillonite	$(Na,Ca)_{0,3}(Al,Mg)_2Si_4O10(OH)_2 \cdot n(H_2O)$	Wyoming
Nontronite	Na <sub>0.3</sub> Fe <sub>2</sub> (Si,Al) <sub>4</sub> O <sub>1</sub> O(OH) <sub>2</sub> ·nH2O	Pennsylvania

**Table 2.** Geographical information on the soil samples used in this work.

Geographical area	Sample	Desert area	Geographical coordinates			
	Morocco	East of Ksar Sahli	31.97°N, 3.28°W			
Sahara	Libya	Sebha	27.01°N, 14.50°E			
	Algeria	Ti-n-Tekraouit	23.95°N, 5.47°E			
Sahel	Mali	Dar el Beida	17.62°N, 4.29°W			
Saller	Bodélé	Bodélé depression	17.23°N, 19.03°E			
Middle East	Saudi Arabia	Nefud	27.49°N, 41.98°E			
Wilddle East	Kuwait	Kuwaiti	29.42°N, 47.69°E			
Southern Africa	Namibia	Namib	21.24°S, 14.99°E			
Eastern Asia	China	Gobi	39.43°N, 105.67°E			
North America	Arizona	Sonoran	33.15 °N, 112.08°W			
South America	Patagonia	Patagonia	50.26°S, 71.50°W			
Australia	Australia	Strzelecki	31.33°S, 140.33°E			

**Table 3.** Chemical characterisation of the dust aerosols in PM<sub>10.6</sub> and PM<sub>2.5</sub> (in parentheses) size fractions. Columns 3 and 4 give the Si/Al and Fe/Ca elemental ratios obtained from X-Ray Fluorescence analysis. The uncertainty of each individual value is estimated to be 10%. Column 5 shows  $MR_{Fe\%}$ , the fractional mass of elemental iron with respect to the total dust mass concentration (uncertainty 10%). Column 5 reports  $MR_{Fe\%}$ , the mass fraction of iron oxides with respect to the total dust mass concentration (uncertainty 15%). For PM<sub>2.5</sub> the determination of the Si/Al ratio is impossible due to the composition of the filter- membranes (quartz)

Chemical characterisation of the dust aerosols in PM<sub>10.6</sub> and PM<sub>2.5</sub> (in parentheses) size fractions. Columns 3 and 4 report the Si/Al and Fe/Ca elemental ratios obtained from X-Ray Fluorescence analysis. The uncertainty on each individual value is estimated to be 10%. Column 5 reports *MR<sub>Fe/Si</sub>*, the fractional mass of elemental iron with respect to the total dust mass concentration (uncertainty 10%). Column 5 reports *MR<sub>Fe-ox/Si</sub>*, the mass fraction of iron oxides with respect to the total dust mass concentration (uncertainty 15%). For PM<sub>2.5</sub> the determination of the Si/Al ratio is impossible due to the composition of the filter medium.

Geographical area	Sample	Si/Al	Fe/Ca	MC <sub>Fe</sub> %	MC <sub>Fe-ox%</sub>
	Morocco	3.12 ()	0.24 (0.28)	3.6 (4.4)	1.4 (1.8)
Sahara	Libya	2.11 ()	1.19 (1.12)	5.2 (5.6)	3.1 (3.4)
	Algeria	2.51 ()	3.14 (4.19)	6.6 (5.4)	2.7 (2.2)
Sahel	Mali	3.03 ()	2.99 (3.67)	6.6 (33.6)	3.7 (18.7)
Sallel	Bodélé	5.65 ()	12.35 ()	4.1 ()	0.7 ()
Middle East	Saudi Arabia	2.95 ()	0.29 (0.27)	3.8 (5.1)	2.6 (3.5)
Middle East	Kuwait	3.15 ()	0.89 (1.0)	5.0 (13.6)	1.5 (4.2)
Southern Africa	Namibia	3.41 ()	0.11 (0.10)	2.4 (6.9)	1.1 (3.1)
Eastern Asia	China	2.68 ()	0.77 (0.71)	5.8 (13.6)	0.9 (2.5)
North America	Arizona	3.30 ()	0.95 ()	5.3 ()	1.5 ()
South America	Patagonia	4.80 ()	4.68 (4.64)	5.1 ()	1.5 ()
Australia	Australia	2.65 ()	5.46 (4.86)	7.2 (11.8)	3.6 (5.9)

**Table 4.** Mass absorption efficiency (MAE,  $10^{-3}$  m<sup>2</sup> g<sup>-1</sup>) and Ångström Absorption Exponent (AAE) in the PM<sub>10.6</sub> and PM<sub>2.5</sub> size fractions. Absolute errors are in brackets.

		$\mathrm{PM}_{10.6}$					_
Geographical area	Sample	375 nm	407 nm	532 nm	635 nm	850 nm	AAE
	Morocco	()	()	()	()	()	()
Sahara	Libya	89 (11)	75 (9)	30 (5)	()	()	3.2 (0.3)
	Algeria	99 (10)	80 (10)	46 (7)	16 (3)	15 (3)	2.5 (0.3)
Cahal	Mali	()	103 (18)	46 (12)	()	()	()
Sahel	Bodélé	37 (4)	25 (3)	13 (2)	6 (1)	3 (1)	3.3 (0.3)
Middle East	Saudi Arabia	90 (9)	79 (8)	28 (3)	6 (1)	4 (1)	4.1 (0.4)
Middle East	Kuwait	()	()	()	()	()	2.8 (0.3)
Southern Africa	Namibia	52 (7)	49 (7)	13 (3)	5 (2)	1 (2)	4.7 (0.5)
Eastern Asia	China	65 (8)	58 (7)	32 (4)	8 (2)	7 (2)	3 (0.3)
North America	Arizona	130 (15)	99 (12)	47 (7)	21 (4)	13 (4)	3.1 (0.3)
South America	Patagonia	102 (11)	80 (9)	29 (4)	17 (2)	10 (2)	2.9 (0.3)
Australia	Australia	135 (15)	121 (13)	55 (7)	26 (4)	14 (3)	2.9 (0.3)

	PM <sub>2.5</sub>						
Geographical area	Sample	375 nm	407 nm	532 nm	635 nm	850 nm	AAE
	Morocco	107 (13)	88 (11)	34 (6)	14 (3)	15 (4)	2.6 (0.3)
Sahara	Libya	132(17)	103 (14)	33 (7)	()	()	4.1 (0.4)
	Algeria	95(8)	71 (11)	37 (7)	12 (5)	12 (5)	2.8 (0.3)
Sahel	Mali	711 (141)	621 (124)	227 (78)	()	()	3.4 (0.3)
	Bodelé	()	()	()	()	()	()
	Saudi Arabia	153 (18)	127 (15)	42 (7)	8 (4)	6 (4)	4.5 (0.5)
Middle East	Kuwait	270 (100)	324 (96)	()	54 (52)	()	3.4 (0.3)
Southern Africa	Namibia	147 (36)	131 (32)	31 (21)	6 (16)	3 (15)	5.1 (0.5)
Eastern Asia	China	201 (30)	176 (26)	89 (17)	14 (10)	23 (10)	3.2 (0.3)
North America	Arizona	()	()	()	()	()	()
South America	Patagonia	()	()	()	()	()	2.9 (0.3)
Australia	Australia	335 (39)	288 (33)	130 (19)	57 (11)	36 (9)	2.9 (0.3)

Table 5. Mass absorption efficiency (MAE, 10<sup>-3</sup> m<sup>2</sup> g<sup>-1</sup>) and Ångström Absorption Exponent (AAE) from the of literature data discussed in the paper

Sample	266 nm	325 nm	428 nm	532 nm	660 nm	880 nm	106 4 nm	AAE
Morocco*								2.25– 5.13
Morocco, PM <sub>2.5</sub> <sup>£</sup>								2.0–6.5
Morocco, submicron#	1100			60			30	4.2
Egypt, submicron#	810			20				5.3
Tunisia <sup>\$</sup>		83			11			
Saharan, transported <sup>µ</sup>								$2.9 \pm 0.2$
Saharan, transported $(PM_{10})^{9/6}$			37	27%%	15%%%			2.9
Saharan, transported (PM <sub>1</sub> )%			60	40%%	30%%%			2.0
Niger <sup>\$</sup>		124			19			
China <sup>§</sup>		69			10			
China&		87 <sup>&amp;</sup> &	50&&&	27***	13	1		3.8
Various locations <sup>@</sup>								2.5-3.9
	Morocco*  Morocco, PM <sub>2.5</sub> <sup>£</sup> Morocco, submicron <sup>#</sup> Egypt, submicron <sup>#</sup> Tunisia <sup>\$</sup> Saharan, transported  (PM <sub>10</sub> ) <sup>%</sup> Saharan, transported  (PM <sub>1</sub> ) <sup>%</sup> Niger <sup>\$</sup> China <sup>\$</sup> China <sup>&amp;</sup>	Morocco*  Morocco, PM <sub>2.5</sub> <sup>£</sup> Morocco, submicron#  Egypt, submicron#  Tunisia <sup>\$</sup> Saharan, transported  (PM <sub>10</sub> )%  Saharan, transported  (PM <sub>10</sub> )%  Niger <sup>\$</sup> China <sup>\$</sup> China <sup>&amp;</sup>	Morocco*  Morocco, $PM_{2.5}^{£}$ Morocco, submicron#  Egypt, submicron#  Tunisia\$  Saharan, transported $^{\mu}$ Saharan, transported $(PM_{10})^{\%}$ Saharan, transported $(PM_1)^{\%}$ Niger\$  Niger\$  124  China\$  69  China\$	Morocco*  Morocco, $PM_{2.5}^{£}$ Morocco, submicron#  Egypt, submicron#  Saharan, transported $(PM_{10})^{\%}$ Saharan, transported $(PM_1)^{\%}$ Niger\$ $124$ China\$ $69$ China\$ $87^{\&}$ $80^{\&\&\&}$	Morocco*         Morocco, PM $_{2.5}^{\pounds}$ Morocco, submicron#         1100         60           Egypt, submicron#         810         20           Tunisia\$         83         Saharan, transported#           Saharan, transported (PM $_{10}$ )%         37 $27$ %%           Saharan, transported (PM $_{10}$ )%         60 $40$ %%           Niger\$         124           China\$ $87$ % $87$	Morocco*         Morocco, PM <sub>2.5</sub> <sup>£</sup> Morocco, submicron#         1100         60           Egypt, submicron#         810         20           Tunisia\$         83         11           Saharan, transported (PM <sub>10</sub> )%         37         27%%         15%%%           Saharan, transported (PM <sub>1</sub> )%         60         40%%         30%%%           Niger\$         124         19           China\$         87% & 50&&& & & & & & 13	Morocco*         Morocco, PM2.5 <sup>£</sup> Morocco, submicron#         1100         60         11           Egypt, submicron#         810         20         11           Saharan, transported#         83         11         11           Saharan, transported(PM10)%         37 $27\%$ $15\%\%$ 15           Saharan, transported(PM10)%         60 $40\%\%$ $30\%\%\%$ 10           Niger\$         124         19         10           China\$ $87\%$ $50\%\%$ $27\%\%$ $30\%\%$ China\$ $87\%$ $50\%\%$ $27\%\%$ $30\%\%$	Sample       266 nm       325 nm       428 nm       532 nm       660 nm       880 nm       4 nm         Morocco*       Morocco, PM2.5 <sup>£</sup> 428 nm       420 n

<sup>\*</sup> Müller et al. (2008)

<sup>£</sup> Petzold et al. (2008)

<sup>‡</sup> Linke et al. (2006)

<sup>§</sup> Alfaro et al. (2004) 

Fialho et al. (2005)

Fialho et al. (2016); % at 528 nm, % at 652 nm

Yang et al. (2009); & at 375 nm, & at 470 nm, & at 590 nm

Mossmüller et al. (2012)

**Figure 1.** Time series of aerosol mass concentration in the chamber for two companion experiments (Libyan dust). Time series of aerosol mass concentration in the chamber for the two companion experiments (Libya sample). Experiment 1 (top panel) was dedicated to the determination of the chemical composition (including iron oxides) by sampling on polycarbonate filters. Experiment 2 (bottom panel) was dedicated to the determination of the absorption optical properties by sampling on quartz filters.

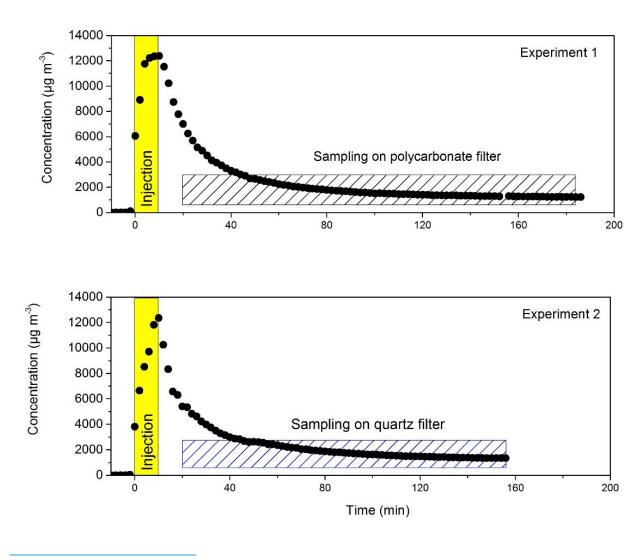
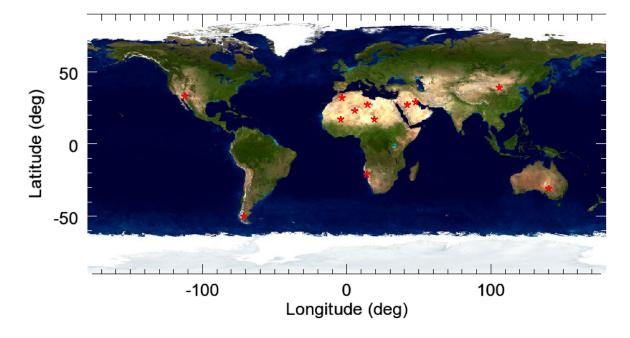


Figure 2. Locations (red stars) of the soil and sediment samples used to generate dust aerosols.



**Figure 3.** Spectral dependence of the MAE values for the samples investigated in this study in the PM<sub>10.6</sub> (left) and in the PM<sub>2.5</sub> (right) mass fractions.

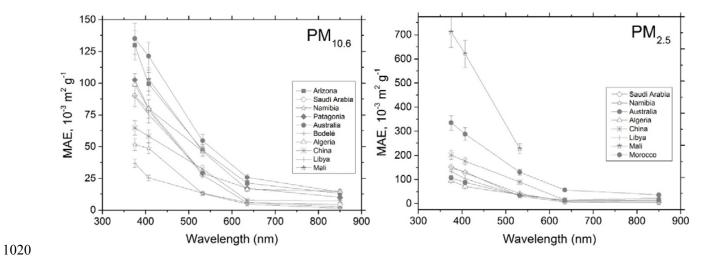


Figure 4. Illustration of the links between the MAE values and the dust chemical composition found in this study. Left column, from top to bottom: linear regression between the MAE values in the range from 375 to 850 nm and the fraction of elemental iron relative to the total dust mass (MR<sub>Fe</sub>%) in the PM<sub>10.6</sub> fraction; Middle column: same as left column but for the mass fraction of iron oxides relative to the total dust mass ( $MR_{Fe\ ox\%}$ ) in the PM<sub>10.6</sub> size fraction; Right column: same as left column but in the PM<sub>2.5</sub> size fraction. Illustration of the links between the MAE values and the dust chemical composition found in this study. Left column, from top to bottom: MAE values between 375 and 850 nm versus the fraction of elemental iron with respect to the total dust mass (MR<sub>Fe%</sub>) in the PM<sub>10.6</sub> fraction; Middle column: same as left column but versus the mass fraction of iron oxides to the total dust mass (MR<sub>Fe ox/6</sub>) in the PM<sub>10.6</sub> size fraction; Right column: same as left column but in the PM2.5 size fraction. The linear regression lines between MAE and MR<sub>Fe%</sub> and MAE and MR<sub>Fe Ox%</sub> are reported in each plot. 

