



1 **Thirteen-years of observations on primary sugars and sugar alcohols over remote**

2 **Chichijima Island in the western North Pacific**

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22 **Abstract.** In order to understand the atmospheric transport of bioaerosols, we conducted long-  
23 term observations of primary sugars and sugar alcohols over remote Chichijima Island in the  
24 western North Pacific from 2001 to 2013. Our results showed that concentrations of total sugar  
25 compounds for 13 years ranged from 1.2 to 310 ng m<sup>-3</sup> (average, 46±49 ng m<sup>-3</sup>). We found that  
26 atmospheric circulations significantly affect the seasonal variations of bioaerosol distributions  
27 over the western North Pacific. The primary sugars (glucose and fructose) maximized in  
28 summer, possibly due to a decreased transport of Asian aerosols and increased local emission  
29 of vegetation products from the vascular plants in Chichijima. We also found higher  
30 concentrations of sugar components (arabitol, mannitol and trehalose) in more recent years  
31 during summer/autumn, suggesting an enhanced emission of fungal and microbial species over  
32 the island. Sucrose peaked in late winter to early spring, indicating a springtime pollen  
33 contribution by long-range atmospheric transport, while elevated concentrations of sucrose in  
34 early summer could be assumed to be long-range transport of soil dust from Southeast Asia to  
35 Chichijima. Positive matrix factorization analyses suggested the locally emitted sugar  
36 compounds as well as long-range transported air borne pollen grains, microbes and fungal  
37 spores are the major contributors to total sugar compounds in the Chichijima aerosols.  
38 Backward air mass trajectories support the atmospheric transport of continental aerosols from  
39 the Asian continent during winter/spring over Chichijima.

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41 **Keywords:** Sugar compounds, fungal and microbial tracer, pollen tracer, bioaerosols, the  
42 western North Pacific.

43



## 44 **1 Introduction**

45 East Asia has experienced rapid economic developments and population growth since last  
46 several decades (Elliot et al., 1997; Jaffe et al., 1999, 2003), whose activities emit organic and  
47 bioaerosols into the atmosphere (Xu et al., 2011). The atmospheric particles are transported to  
48 downwind region in the Pacific, associated with Asian desert dust from the Taklamakan and  
49 Gobi Deserts, and Loess plateau (Duce et al., 1980; Iwasaka et al., 1983; Jaffe et al., 1997;  
50 Prospero and Savoie, 1989; Talbot et al., 1997). The transported dust contains bacterial cells,  
51 fungal spores, and microbial cells, which fall out over the Pacific and remote islands in the  
52 Pacific Ocean (Lacey and West, 2006; Mims and Mims, 2003). The microbes associated with  
53 bioaerosols significantly affect the natural environment of marine and land ecosystem in  
54 downwind regions (Graffin et al., 2003, 2007; Prospero et al., 2005). Long-range atmospheric  
55 transport plays a key role for the global distribution of microbes from source regions to  
56 receptor site (Graffin et al., 2001). Fungi and bacteria are often attached to dust particles,  
57 which can propagate diseases to human and plants (Brown and Hovmoller, 2002). Therefore,  
58 the transported organic and bioaerosols have been the focus of extensive studies for the past  
59 years (Yamaguchi et al., 2012).

60 Organic aerosols are composed of a complex mixture of different types of molecules, in  
61 which water-soluble organic compounds (WSOCs) are enriched (Graham et al., 2002). WSOCs  
62 play an important role in climate change and global radiative forcing by scattering or absorbing  
63 light directly or indirectly (Fuzzi et al., 2007). They can act as cloud condensation nuclei  
64 (CCN) (Kanakidou et al., 2005; Martin et al., 2010). Sugar compounds (SCs) contribute 13–  
65 26% and 63% of total WSOCs identified in continental and marine aerosols, respectively  
66 (Simoneit et al., 2004a). SCs are directly emitted from biological sources such as fungi, algae,  
67 pollen, spores and bacteria (Carvalho et al., 2003; Wang et al., 2009) and transported long  
68 distances in the atmosphere (Wang et al., 2011). They are also derived from suspended soil



69 particles and associated biota (Rogge et al., 2006; Simoneit et al., 2004b; Wang et al., 2009),  
70 and biomass burning (Schmidl et al., 2008; Simoneit et al., 2002).

71 Primary sugars are emitted from biological sources (Medeiros et al., 2006). Glucose and  
72 fructose are emitted from terrestrial plant fruits, pollen, and detritus of vascular plants (Cowie  
73 and Hedges, 1984; Speranza et al., 1997). Previous studies reported that sucrose is dominant  
74 sugar component in airborne pollen grains and plays a significant role in plant blossoming  
75 activity (Bielecki, 1995; Fu et al., 2012; Pacini, 2000). Trehalose is emitted from fungal  
76 metabolic activities and resuspension of soil particles and unpaved road dust (Rogge et al.,  
77 2007; Simoneit et al., 2004). Sugar alcohols are also emitted from biological sources like fungi  
78 and microbes via metabolic activities (Bauer et al., 2008). Sugar alcohols, i.e., arabitol and  
79 mannitol, are tracers for fungal spores (Jia and Fraser; 2011; Yang et al., 2012). Di Filippo et  
80 al. (2013) reported that arabitol and mannitol are key sugar components in fungal spores.

81 Chichijima Island is located in the western North Pacific: an outflow region of Asian  
82 dust (Mochida et al., 2003a). Thus, particular attention has been paid to atmospheric chemical  
83 studies over Chichijima Island. It is one of the best remote islands to study a long-range  
84 transport of Asian aerosols, because local pollutants in Chichijima are insignificant due to low  
85 population density and no major industrial or anthropogenic activities (Chen et al., 2013).  
86 Kawamura et al. (2003) reported that lower molecular weight fatty acids derived from marine  
87 organisms showed higher concentrations in summer, while higher molecular weight fatty acids  
88 ( $C_{21}$ - $C_{34}$ ), n-alkanes ( $C_{25}$ - $C_{35}$ ), n-alcohols ( $C_{20}$ - $C_{34}$ ) and dicarboxylic acids ( $C_{20}$ - $C_{28}$ ) derived  
89 from terrestrial higher plants and soil organic matter maximized in winter to spring. Seasonal  
90 variations of low molecular weight dicarboxylic acids and levoglucosan (biomass burning  
91 tracer) have been discussed in Chichijima aerosols by Mochida et al. (2003a) and Mochida et  
92 al. (2010), respectively. Although seasonal variation of saccharides was reported in Chen et al.  
93 (2013), the observation period is rather short. Therefore, long-term observations are needed to



94 obtain a data set of SCs to better discuss the characteristics, potential sources, and possible  
95 effects of atmospheric transport over the western North Pacific.

96 Here, we report thirteen-year data set of SCs in remote Chichijima Island. The goal of  
97 this study is to characterize seasonal and annual variations of SCs and specify their possible  
98 source regions. We will also discuss a potential role of Asian dust to control the distributions  
99 of bioaerosols over the western North Pacific. The outcomes of this study will improve our  
100 understanding about a possible influence of long-range transport of bioaerosols from the  
101 continent to the clean oceanic environment. We will compare the data set of SCs for the  
102 periods 1990-1993, 2001-2003 and 2010-2013, which may provide imperative information  
103 about decadal changes in the atmospheric conditions over Chichijima. Seasonal source  
104 identifications by positive matrix factorization (PMF) analysis will also be discussed for the  
105 measured SCs.

106

## 107 **2 Materials and methods**

### 108 **2.1 Sampling site and meteorological conditions**

109 The detailed information on the sampling site was reported in Kawamura et al. (2003) and  
110 Chen et al. (2013). Briefly, Chichijima Island is located in the western North Pacific (27°04'N;  
111 142°13'E), 1000 km south of Tokyo, Japan, and 2000 km east of the Asian continent (Figure 1).  
112 Total area of the island is 24 km<sup>2</sup> with a population of 2000 (Verma et al., 2015). The climate  
113 of Chichijima is classified as subtropical; it is warm to hot (temperature, 7.8-34.1 °C) and  
114 humid (relative humidity, 66-88%) all year round.

115 Figure 2 shows monthly averaged variations in the meteorological parameters of  
116 Chichijima during 2001-2013. It receives more precipitation in between April and July,  
117 September and October during the thirteen-year periods. The sampling site is less influenced  
118 by the East Asian monsoon to receive heavy rainfall compared to Northeast Asia. The climate  
119 over Chichijima is strongly influenced by the seasonal changes in wind system. In



120 winter/spring, the westerly winds are dominant with the air masses being enriched with Asian  
121 dust, industrial pollutants, biomass burning products, organic compounds and black carbon as  
122 well as bioaerosols emitted from East Asia and Eurasia (Figure 3) (Seinfeld et al., 2004;  
123 Simoneit et al., 2004b; Wang et al., 2009). Trade winds are dominant in summer/autumn,  
124 which transport clean and pristine marine air masses from the central Pacific to Chichijima  
125 (Kawamura et al., 2003, Mochida et al., 2010).

## 126 **2.2 Aerosol sampling and chemical analysis**

127 The details on aerosol sampling and chemical analysis are reported elsewhere (Chen et al.,  
128 2013; Mochida et al., 2010). Briefly, total suspended particle (TSP) samples were collected at  
129 the Ogasawara Downrange Station of the Japan Aerospace Exploration Agency (JAXA) in  
130 Chichijima Island (254 m, above sea level, asl). The samples were collected on weekly basis  
131 (January 2001 to November 2013) using a high volume air sampler (Kimoto AS-810A) at a  
132 flow rate of  $1.0 \text{ m}^3 \text{ min}^{-1}$  and pre-combusted ( $450^\circ\text{C}$  for 6 h) quartz fiber filters (20 x 25 cm,  
133 Pallflex). Filter sample was placed in a pre-combusted glass jar with a Teflon-lined screw cap  
134 and stored in a dark freezer room at  $-20^\circ\text{C}$  prior to analysis in order to inhibit fungal growth.  
135 Due to the maintenance of the JAXA facility at sampling site, TSP samples were not collected  
136 for November–December 2004 and March–August 2005.

137 Total 590 aerosol samples were analyzed to determine the primary sugars (xylose,  
138 fructose, glucose, sucrose and trehalose) and sugar alcohols (erythritol, arabitol, mannitol and  
139 inositol) during 2001 to 2013. An aliquot ( $21 \text{ cm}^2$ ) of the filters were extracted three times with  
140 dichloromethane/methanol (2:1, v/v) mixture using ultrasonic agitation for 10 minutes. A  
141 Pasteur pipette packed with quartz wool was used to remove particles and filter debris in the  
142 extracts. Filtrates were then concentrated using a rotary evaporator under vacuum and blown  
143 down with a stream of pure nitrogen gas. The total extracts were derivatized using  $60 \mu\text{l}$  of  
144 N,O-bis-(trimethylsilyl)trifluoroacetamide (BSTFA) with 1% trimethylsilyl chloride in the  
145 presence of  $10 \mu\text{l}$  of pyridine in a sealed vial at  $70^\circ\text{C}$  for 3 hours to convert hydroxyl groups to



146 corresponding trimethylsilyl (TMS) ethers. The derivatized fractions were diluted with n-  
147 hexane containing internal standard of C<sub>13</sub> n-alkane (1.43 ng μl<sup>-1</sup>), prior to injection to gas  
148 chromatography-mass spectrometer (GC-MS).

149 Identification of SCs have been confirmed by the comparison of GC retention times and  
150 mass spectra with those of authentic standards as well as literature and library data. SCs were  
151 characterized by their common base peak at m/z 217 and 204 with specific fragment ions for  
152 individual sugars, i.e., m/z = 307 (arabitol), 205 and 319 (mannitol), 205 (erythritol), 305 and  
153 318 (inositol), 361 (sucrose and trehalose), 191 (glucose), and 437 (fructose). The selected ion  
154 peak area and relative response factors determined by injection of authentic standards have  
155 been used for the quantification of sugar compounds. Field blank filters were analyzed as a real  
156 sample, but no target compounds were detected in the field blanks. The recoveries of the target  
157 compounds were better than 90%. Therefore, the data reported here were not corrected for  
158 recoveries. Analytical errors of SCs were generally <15% based on duplicate analysis. The  
159 detection limits of primary sugars and sugar alcohols were 105-557 pg μl<sup>-1</sup>, which corresponds  
160 to ambient concentrations of 0.0015-0.0081 ng m<sup>-3</sup> under a typical sampling volume of 9000  
161 m<sup>3</sup> (Zhu et al., 2015).

162 The derivatized fractions were introduced into GC-MS using an Agilent model 7890 GC  
163 coupled to an Agilent model 5975 mass selective detector (MSD) operated in an electron  
164 impact mode at 70 eV and scanned from 40 to 650 Dalton. The GC separation was carried out  
165 on a DB-5MS fused silica capillary column (30 m long, 0.25 mm i.d., 0.25 μm film thickness),  
166 with a temperature program of 50 °C for 2 min at a rate of 15 °C min<sup>-1</sup> from 50 to 120 °C, then  
167 from 120 to 305 °C at a rate of 5 °C min<sup>-1</sup> with a final isotherm hold at 305 °C for 15 min. The  
168 sample was injected on a splitless mode at an injector temperature of 280 °C. GC-MS data  
169 were acquired and processed with the Agilent GC/MSD ChemStation software.

### 170 2.3 Backward air mass trajectory analysis



171 In order to identify the source regions of sugar compounds in Chichijima aerosols, ten-day  
172 backward trajectories were calculated at 00:00 UTC of each sampling period for thirteen-years  
173 using the NOAA Hybrid Single-Particle Lagrangian Integrated Trajectory  
174 (<http://ready.arl.noaa.gov/HYSPLIT.php>) (Figure 3). The starting height of the trajectories  
175 presented in this study is 500 m asl. We plotted thirteen-year trajectories for each sampling day  
176 but there are no significant year-to-year changes in the atmospheric circulations. Therefore, we  
177 presented seasonal trajectories for recent year (December, 2011 to November, 2012) in Figure  
178 3 to understand the seasonal aerosol mass transport from the source regions to Chichijima  
179 Island. Backward trajectories significantly supported a long-range transport of air mass under  
180 the influence of existing meteorological parameters (Figure 3). The trajectories clearly show  
181 the influences of continental air masses during mid-autumn to mid-spring and of marine air  
182 masses during mid-spring to mid-autumn.

#### 183 **2.4 Positive matrix factorization (PMF) analysis**

184 Positive matrix factorization (PMF 3.0, Environmental Protection Agency, USA) has  
185 been used as a powerful statistical tool that may resolve potential sources contributing to  
186 atmospheric levels of particle (as presented by %) when appropriate source profiles are not  
187 available (Paatero and Tapper, 1994). At the beginning PMF has been used in precipitation  
188 study (Juntto and Paatero, 1994) as well as air pollution and source apportionment studies  
189 (Polissar et al., 1999). Recently, it is widely using for the air quality and source apportionment  
190 (Xie and Berkowitz, 2006). In addition, PMF has been applied to the wastewater  
191 (Soonthornnonda and Christensen, 2008), lakes sediments (Bzdusek et al., 2006) and soils (Lu  
192 et al., 2008). One of the main features of PMF results is their quantitative nature; it is possible  
193 to obtain the composition of the sources determined by the model.

194 PMF uses uncertainties for each of the measured concentration ( $x_{ij}$ ). PMF minimize the  
195 residual sum of squares (Q) defined by equation:





$$Q = \sum_{i=1}^n \sum_{j=1}^m \left( \frac{e_{ij}}{s_{ij}} \right)^2$$

196  
197 Where the  $j$  is species in the  $i^{\text{th}}$  sample,  $e_{ij}$  is portion of the measurements. In PMF, only  
198 measured concentrations ( $x_{ij}$ ) are known and the goal is to estimate the contributions and the  
199 fractions. The uncertainties were computed from the measurement errors by equation:

$$s_{ij} = \sigma_{ij} + C_3 \max(|x_{ij}|, |y_{ij}|)$$

200  
201  
202 Where  $y_{ij}$  is the calculated value for  $x_{ij}$ ,  $\sigma_{ij}$  is the measurement or estimated error, and  $C_3$  is a  
203 dimensionless constant value. The estimation of the measurement errors of size distribution  
204 data were based on the combination of size bins (Zhou et al., 2004).  $C_3$  is used as the  
205 estimation of the relative uncertainties of large values (Norris et al., 2008). Using too few  
206 factors will combine sources of different nature together and using too many factors will make  
207 a real factor further dissociate into two or more non-existing factors.  $F_{\text{peak}}$  is a parameter in  
208 PMF for controlling rotations (Paatero et al., 2002). When the  $F_{\text{peak}}$  value is positive, the  
209 following additional term is included in the object function  $Q$ :

$$Q^p = \beta^2 \left( \sum_{k=1}^P \sum_{j=1}^N f_{kj} \right)^2$$

211  
212  
213 where,  $\beta^2$  corresponds to the  $F_{\text{peak}}$  value.

214 PMF analysis was performed for quantitative estimation of sources for the collected  
215 samples using tracer compounds for primary sugars, sugar alcohols, and anhydrosugars. Based  
216 on given understanding of sugar sources, 4-7 factors were examined and total five interpretable  
217 factors were characterized by the enrichment of each tracer compound, which reproduced more  
218 than 94% of SCs. Minimal robust and true  $Q$  values of the base run were 3001 and 3413,  
219 respectively. Concentrations and percentage of tracers in each factor of bootstrap run were  
220 close of those of base run results. The  $Q$  values and factor profiles of  $F_{\text{peak}}$  rotation runs showed  
221 no significant changes compared with base run, indicating stable PMF results.



222 In winter/spring, Chichijima Island receives air masses enriched with anthropogenic  
223 aerosols from the Asian continent by strong westerly winds, whereas during summer/autumn it  
224 receives clean air masses from the Pacific Ocean under the influences of trade winds. The  
225 seasonal changes in the atmospheric circulation over Chichijima may have a significant  
226 influence on the seasonal distributions of SCs. Therefore, we performed the seasonal PMF  
227 analysis on the thirteen-year sugar data set to better understand the seasonal source profile of  
228 individual sugar component. For seasonal PMF analysis, 3-5 factors were examined and 4  
229 factors were determined for each season. We included the data set of anhydrosugars from  
230 Verma et al. (2015) for PMF analysis.

231

### 232 3. Results and discussion

#### 233 3.1 Ambient concentrations of sugar compounds

234 Temporal variations of primary sugars and sugar alcohols are shown in Figure 4. Nine sugar  
235 compounds (SCs) including five primary sugars and four sugar alcohols were detected in the  
236 aerosol samples collected from Chichijima Island. The concentrations of total SCs varied from  
237 1.23 to 339 ng m<sup>-3</sup> (average, 46.7±49.5 ng m<sup>-3</sup>) during 2001 to 2013 (Table 1). Concentrations  
238 of primary sugars and sugar alcohols were in the range of 0.28 to 176 ng m<sup>-3</sup> (23.3±25.7 ng m<sup>-3</sup>)  
239 and 0.37 to 231 ng m<sup>-3</sup> (23.4±30.8 ng m<sup>-3</sup>), respectively. Average concentration of primary  
240 sugars in Chichijima aerosols is several times lower than that of primary sugars (62.0±54.9 ng  
241 m<sup>-3</sup>) reported from Cape Hedo, Okinawa, Japan (Zhu et al., 2015) while that of sugar alcohols  
242 is equivalent to or little lower than that from Cape Hedo (29.5±35.5 ng m<sup>-3</sup>).

243 Interestingly, primary sugars (49.9%) and sugar alcohols (50.1%) were found to  
244 contribute almost equal to total SCs during the entire study period. Mannitol (26.7%) and  
245 arabitol (21.4%) were the main contributors to total SCs followed by glucose (16.7%), sucrose  
246 (13.6%), fructose (10.2%), and trehalose (9.2%). Erythritol (1.6%), inositol (0.3%), and xylose  
247 (0.3%) were also present in the aerosols at lower concentration levels. Temporal plots of



248 individual sugars clearly indicate a large variation of SCs (Figure 4). This large variation in the  
249 concentrations of SCs might be involved with seasonal changes in the atmospheric circulations  
250 in the western North Pacific (Kawamura et al., 2003).

### 251 3.1.1 Concentrations of primary sugars in total SCs

252 Glucose is the dominant sugar species among the primary sugars with concentration range of  
253 0.05 to 64.3 ng m<sup>-3</sup> (average, 7.79±8.80 ng m<sup>-3</sup>). Similarly, a wide concentration range of  
254 fructose (0.03-115 ng m<sup>-3</sup>; 4.69±8.04 ng m<sup>-3</sup>) was also observed in Chichijima aerosols.  
255 Thirteen-year mean concentrations of glucose and fructose were observed to be lower than  
256 those (27.2 ng m<sup>-3</sup> and 16.4 ng m<sup>-3</sup>, respectively) reported for the aerosol samples (TSP) from  
257 Cape Hedo, Okinawa, Japan (Zhu et al., 2015). Glucose and fructose significantly contribute to  
258 total primary sugars (33.5% and 20.17%, respectively) in Chichijima aerosols. The primary  
259 sugars are abundant in the fragments of vascular plants in vegetated and forest areas (Medeiros  
260 et al., 2006). Pacini et al. (2000) reported that primary sugars are synthesized in leaves during  
261 photosynthesis and stored in root, stem, flower, pollen and fruit of growing plants. The nectars  
262 and fruits of tropical and subtropical plants also contain glucose and fructose abundantly  
263 (Backer et al., 1998). Graham et al. (2002) reported significant amounts of glucose and  
264 fructose in pollen, fern spores, and insects in aerosol samples collected from the Amazon forest.  
265 Chichijima Island is covered with endemic and vascular plants, which may emit glucose and  
266 fructose. Moreover, different sources such as soil dust (Rogge et al., 2007; Simoneit et al.,  
267 2004), lichens (Dahlman et al., 2003) and biomass burning (Medeiros et al., 2006; Nolte et al.,  
268 2001) have also been reported as dominate sources for glucose and fructose.

269 Among all the SCs detected in the Chichijima aerosols, sucrose is the second most  
270 abundant sugar species (0.002-100 ng m<sup>-3</sup>; 6.43±12.9 ng m<sup>-3</sup>), accounting for 27.3% of total  
271 primary sugars. The average sucrose concentration observed in Chichijima is twice lower than  
272 that (13.2 ng m<sup>-3</sup>) from Cape Hedo, Okinawa, Japan (Zhu et al., 2015). Sucrose is synthesized  
273 in plant leaves and circulated by phloem to different plant sections, which is accumulated in



274 root cells as well as developing flower buds (Bieleski, 1995; Jia et al., 2010). Sucrose is a  
275 dominant component in airborne pollen grains derived from flowering plants (Bieleski, 1995;  
276 Pacini, 2000). Simoneit et al. (2004a and 2004b) reported the presence of sucrose in surface  
277 soil and paved road dust. Sucrose was also observed in dry plant materials during harvesting  
278 period (Ma et al., 2009).

279 Thirteen-year mean concentration of trehalose ranged from 0.01 to 70.2 ng m<sup>-3</sup>  
280 (4.30±7.28 ng m<sup>-3</sup>), whose average concentration accounts for 18.4% of total primary sugars  
281 detected in Chichijima aerosols for 13 years. Microbes (bacterial cell), fungal spores, yeast,  
282 algae, invertebrates, suspended soil dust, as well as plant species, contribute significantly to  
283 trehalose in the atmosphere (Elbein, 1974; Graham et al., 2003; Medeiros et al., 2006; Rogge et  
284 al., 2007; Simoneit et al., 2004; Wiemken, 1990). Xylose is a less abundant primary sugar,  
285 accounting for 0.60% of total primary sugars observed in Chichijima aerosols. The  
286 concentration range of xylose was 0.001-1.35 ng m<sup>-3</sup> (0.14±0.18 ng m<sup>-3</sup>) during sampling  
287 period of thirteen years. Biomass burning activities emit xylose to the atmosphere. Cowie and  
288 Hedges (1984) reported that xylose produced by angiosperm and gymnosperm plants,  
289 phytoplankton, as well as groups of microorganisms. Simoneit et al. (2004a) have reported  
290 xylose in soil dust from various locations in the United States and Japan. Wan and Yu (2007)  
291 also observed xylose in soils and associated micro biota.

### 292 3.1.2 Concentrations of sugar alcohols in total SCs

293 Thirteen-year mean concentrations of arabitol and mannitol were found to be 9.99±13.6 ng m<sup>-3</sup>  
294 and 12.5±17.5 ng m<sup>-3</sup>, which contribute to 42.7% and 53.3% of total sugar alcohols,  
295 respectively. The concentration ranges of arabitol (0.04–106 ng m<sup>-3</sup>) and mannitol (0.10–118  
296 ng m<sup>-3</sup>) are comparable to those from the Mediterranean region, Israel (arabitol, 1.85–58.3 ng  
297 m<sup>-3</sup> and mannitol, 5.57–138 ng m<sup>-3</sup>) (Burshtein et al., 2011). Yttri et al. (2007) also reported  
298 that arabitol and mannitol were main contributors of sugar alcohols in aerosol samples  
299 collected from the different background sites in Norway. Sugar alcohols (arabitol, mannitol)



300 can be used as tracers for various fungal and algal species (Bauer et al., 2008a,b; Pashanska et  
301 al., 2002; Zhang et al., 2010). Loos et al. (1994) discussed arabitol and mannitol as potential  
302 sources of bacteria and other microbes. High levels of detritus from the spring bloom and  
303 autumn decomposition have been reported as significant sources for arabitol and mannitol in  
304 the vegetated region (Burshtein et al., 2011; Pashynska et al., 2002). Good positive correlations  
305 of arabitol ( $r = 0.63$ ) and mannitol ( $r = 0.72$ ) with glucose indicate a vegetation contribution of  
306 both sugar alcohols in Chichijima aerosols. Erythritol and inositol are less abundant sugar  
307 species, accounting for 3.29% and 0.66% of total sugar alcohols. Their concentration ranges  
308 are  $0.01\text{--}8.32\text{ ng m}^{-3}$  and  $0.01\text{--}1.81\text{ ng m}^{-3}$ , respectively. Significant positive correlations of  
309 both sugar species with arabitol and mannitol indicate similar sources for these SCs in  
310 Chichijima aerosols (Table 2).

### 311 **3.2 Seasonal variations of total sugar compounds**

312 Seasonal concentration range, mean and median values of individual sugars during the  
313 study periods of thirteen-years are presented in Table 1. The concentrations of individual  
314 sugars were extensively fluctuated from season to season in aerosol samples collected at  
315 Chichijima (Figures 4 and 5a). The seasonally averaged concentrations of total SCs are higher  
316 in summer ( $71.5\pm 70.9\text{ ng m}^{-3}$ ) and autumn ( $57.0\pm 64.2\text{ ng m}^{-3}$ ) than spring ( $39.8\pm 67.6\text{ ng m}^{-3}$ )  
317 and winter ( $18.2\pm 34.0\text{ ng m}^{-3}$ ) over Chichijima Island. Zhu et al. (2015) measured sugar  
318 components in aerosol samples collected from Cape Hedo, Okinawa, Japan and reported 2 to 3  
319 times higher concentrations in summer ( $136\text{ ng m}^{-3}$ ) and spring ( $133\text{ ng m}^{-3}$ ) than autumn ( $86$   
320  $\text{ng m}^{-3}$ ) and winter ( $40\text{ ng m}^{-3}$ ), whose seasonal trends are similar to Chichijima. Wan and Wu  
321 (2007) reported different seasonal variations with the highest concentration in autumn ( $375\text{ ng}$   
322  $\text{m}^{-3}$ ), followed by winter ( $292\text{ ng m}^{-3}$ ) and spring ( $84\text{ ng m}^{-3}$ ) for the continental urban aerosols  
323 collected from Hong Kong. These concentrations in Hong Kong are 16 and 6 times higher than  
324 those of the remote Chichijima samples for winter and autumn, respectively. Interestingly, the  
325 different seasonal trends between the continental urban sites and two islands in the western



326 North Pacific may be associated with different sources and transport pathways between the  
327 urban and marine environments.

### 328 **3.2.1 Seasonal variations of primary sugars**

329 Glucose maximized in summer ( $11.0 \pm 9.02 \text{ ng m}^{-3}$ ) followed by autumn ( $9.25 \pm 8.63 \text{ ng m}^{-3}$ ),  
330 spring ( $7.68 \pm 10.3 \text{ ng m}^{-3}$ ) and winter ( $3.11 \pm 3.53 \text{ ng m}^{-3}$ ) (Table 1 and Figure 5h). Glucose is  
331 the most abundant primary sugar in Chichijima aerosols. In winter/spring, Chichijima is  
332 influenced by strong westerly winds that deliver the air masses the Asian continent including  
333 Mongolia, Russian Far East and north China, where vegetation is active. Consequently,  
334 declined concentration of glucose in winter means the depressed emission of continental  
335 bioaerosols from Asia in spite of a long-range transport of Asian dusts due to strong westerly.  
336 The local vegetation (vascular plants) in Chichijima Island might be responsible to enhanced  
337 glucose during growing season (spring and summer) and decaying periods of plant leaves  
338 (autumn). Seasonal PMF analysis also supports dominant sources of vegetation for glucose  
339 among four factors, which contributed >75% for mixed factor in summer (Figure 6c), >80%  
340 for fungal and vegetation factor in autumn (Figure 6d), and >75% for vegetation factor in  
341 spring (Figure 6b).

342 Fructose shows the highest concentrations in summer ( $7.25 \pm 7.63 \text{ ng m}^{-3}$ ) followed by  
343 spring ( $4.51 \pm 9.21 \text{ ng m}^{-3}$ ), autumn ( $3.70 \pm 2.68 \text{ ng m}^{-3}$ ) and winter ( $3.36 \pm 10.2 \text{ ng m}^{-3}$ ). As  
344 shown in Table 2, a significant correlation ( $r=0.57$ ) was obtained between glucose and fructose.  
345 Burshtein et al. (2011) reported similar correlations for both sugar species, suggesting an  
346 identical input of glucose and fructose from the local vegetation in summer (Baker et al., 1998;  
347 Pacini, 2000). Monthly mean concentrations of fructose show two prominent peaks in  
348 February-March and June-July, the latter peak may be due to the local vegetation in Chichijima  
349 (Figure 5g). The fructose peak in February-March may be influenced by air borne pollen grains  
350 in the spring bloom of flowering plants. High concentration of fructose was observed in spring  
351 followed by summer, indicating an input of this sugar compound from pollen grains (Fu et al.,



352 2012). The positive correlation of fructose with sucrose (pollen tracer) supports the similar  
353 sources. Seasonal PMF analysis further supports the identical source for fructose and sucrose;  
354 that is, among four factors, fructose contributes >70% and >60% for pollen factor in spring  
355 (Figure 6b) and winter (Figure 6a), respectively.

356 Seasonal mean concentrations of trehalose showed a maximum in summer ( $7.06 \pm 8.49$  ng  
357  $\text{m}^{-3}$ ) followed by autumn ( $6.09 \pm 8.81$  ng  $\text{m}^{-3}$ ), spring ( $2.93 \pm 6.08$  ng  $\text{m}^{-3}$ ) and winter ( $1.03 \pm 1.26$   
358 ng  $\text{m}^{-3}$ ) (Table 1). Monthly mean concentrations of SCs for 13 years show that concentrations  
359 of trehalose are higher during June to October (Figure 5j). Ma et al. (2009) reported higher  
360 concentrations of trehalose at the urban site of Guangzhou, China during summer and autumn.  
361 Similarly, Wan and Wu (2007) reported a similar autumn maximum in Hong Kong. On the  
362 other hand, different seasonal trends of trehalose were reported for the aerosol samples (TSP)  
363 collected in USA (Medeiros et al., 2006), China (Wang et al., 2011), Australia (Hackl et al.,  
364 2000), and Gosan, Jeju Island in the western North Pacific Rim (Fu et al., 2012). In the above  
365 studies, highest concentrations of trehalose were reported in early spring due to the re-  
366 suspension of soil particles during agricultural practice. Hackl et al. (2000) also obtained  
367 abundant trehalose in spring, and proposed that trehalose can be used as a tracer of soil dust  
368 emission to the atmosphere. However, we did not detect a spring peak of trehalose in  
369 Chichijima aerosols, suggesting that soil dust contribution of trehalose over Chichijima is  
370 insignificant via long-range atmospheric transport. Seasonal PMF analysis for autumn showed  
371 that more than 85% of trehalose was contributed by microbial factor among four factors  
372 (Figure 6d). An indirect contribution of trehalose from soil dust will be discussed later.

373 The seasonal mean concentrations of sucrose are almost equal during spring ( $8.80 \pm 18.0$   
374 ng  $\text{m}^{-3}$ ), summer ( $7.31 \pm 11.5$  ng  $\text{m}^{-3}$ ) and winter ( $6.60 \pm 13.1$  ng  $\text{m}^{-3}$ ), except for autumn  
375 ( $2.76 \pm 4.35$  ng  $\text{m}^{-3}$ ) (Table 1). The homogeneous seasonal distribution suggests multiple  
376 sources of sucrose in Chichijima aerosols. Monthly mean concentrations of sucrose show two  
377 peaks during February-March and June-July (Figure 5i). The March peak of sucrose was



378 reported in the forest area of Sapporo, Japan to be 3 to 7 times more abundant than other  
379 months due to springtime pollen emissions (Miyazaki et al., 2012). Fu et al. (2012) analyzed  
380 pollen samples from different plant species (white birch, Chinese willow, Peking willow) for  
381 SCs and found the highest concentrations of sucrose followed by fructose and glucose in pollen.  
382 The pollen emissions from developing buds of plants may be the reason for the increased  
383 concentration of sucrose and fructose in February and March over Chichijima Island in the  
384 western North Pacific. Seasonal PMF analysis shows that sucrose contributed 100% in spring  
385 and >80% in winter, suggesting a significant pollen contribution for sucrose in those seasons  
386 (Figures 6b and 6a).

387 However, the possibilities of pollen transport from East Asia to Chichijima cannot be  
388 excluded because pollens can travel long distances with springtime high-speed winds by  
389 westerlies (Rousseau et al., 2006). The pollen grains emitted from flowering boreal forest in  
390 China, Mongolia, Siberia and Russian Far East, could significantly be delivered to the western  
391 North Pacific during spring, which may contribute sucrose and fructose in Chichijima aerosols.  
392 Recent studies have discussed a long-range transport of airborne pollen from North America to  
393 Greenland in spring (Rousseau et al., 2008). Lorenzo et al. (2006) reported the long distance  
394 transport of airborne allergenic pollen to central Italy. Makra et al. (2010) reported a long-  
395 range transport of airborne pollen in three European cities by applying three-dimensional  
396 clustering of backward trajectories. Several studies also discussed a long-range transport of  
397 pollen to the remote arctic region (Andrews et al., 1980; Bourgeois et al., 2001; Campbell et al.,  
398 1999; Hicks et al., 2001; Hjelmroos and Franzen, 1994; Rousseau et al., 2004).

399 These observations may support that westerly winds have delivered pollen grains from  
400 the Asian continent including Mongolia, Siberia, and Russian Far East to Chichijima Island in  
401 spring. Using a box model and typical settling velocity of pollens (3 cm/sec) with the grain size  
402 of 30  $\mu\text{m}$  in diameter (Sosnoskie et al., 2009), we estimated lifetime of pollen grains to be 9.3  
403 hours in the atmospheric marine boundary layer (height of 1 km above the ocean surface). The





404 settling velocity of the pollens is ca. 20 times larger than that of typical marine aerosols (Slinn  
405 and Slinn, 1980). Because pollen grain sizes range from 10  $\mu\text{m}$  to 100  $\mu\text{m}$  in diameter, the  
406 lifetime of pollens may have a large uncertainty. If pollens could be largely transported in the  
407 free troposphere (e.g., 5 km high) to the North Pacific from the Asian continent, then lifetime  
408 of typical pollen grains would increase upto 2 days. These calculations for the lifetime of  
409 pollen grains further support their long-range atmospheric transport from the Asian continent  
410 over the western North Pacific. Based on backward air mass trajectories (Figure 3), we can  
411 roughly estimate the transport time from East Asia to Chichijima site to be 2-4 days in winter  
412 and spring. It is also of interest to note that pollens can rapture under condition of high relative  
413 humidity (RH) (Hader et al., 2014; Miguel et al., 2006; Wright et al., 2014), which leads to  
414 smaller particles with longer residence time in the atmosphere.

415 In addition, tilling process after wheat crop harvesting in farmland causes an enhanced  
416 exposure of wheat root associated with soil particles into the atmosphere. China, India, and  
417 USA are three largest countries for wheat production in the world. In China and India there are  
418 two seasons (spring and winter) for wheat crops; winter wheat is harvested from mid-May to  
419 mid-July. During those periods (early summer), Chichijima Island is highly influenced by trade  
420 winds (Figure 3). However, trajectories clearly show the occasional air mass transport during  
421 summer from the Southeast Asia to Chichijima (Pavuluri et al., 2010). PMF results of sucrose  
422 for summer (Figure 6c) and autumn (Figure 6d) account for >85% and >90%, respectively, for  
423 soil dust factor among four source factors, suggesting an additional source of soil dust for  
424 sucrose in Chichijima (Simoneit et al., 2004). The elevated sucrose concentrations in June and  
425 July (summer; non flowering seasons) suggest the transport of sucrose associated with soil  
426 particles under the influence of occasional air mass transport in summer from Southeast Asia  
427 (Figure 3).

428 Xylose was found as the least abundant sugar compound in the aerosol samples. The  
429 maximum concentration of xylose (1.35  $\text{ng m}^{-3}$ ) was found in summer whereas minimum



430 (0.001 ng m<sup>-3</sup>) in spring (Table 1). Summer mean concentration (0.18±0.26 ng m<sup>-3</sup>) was highest  
431 (Table 1). The PMF analyses showed that xylose contributed >75% for BB factor in winter  
432 (Figure 6a) and >70% in autumn (Figure 6d) for microbial factor. These results suggest  
433 different sources and seasons for xylose; i.e., biomass burning in winter (Sullivan et al., 2011)  
434 and groups of microorganisms in summer (Cowie and Hedges, 1984).

### 435 3.2.2 Seasonal variations of sugar alcohols

436 The seasonal mean concentrations of arabitol and mannitol are higher in summer/autumn than  
437 spring/winter (Table 1). The concentrations of arabitol are equally distributed between summer  
438 (15.1±12.9 ng m<sup>-3</sup>) and autumn (15.8±18.3 ng m<sup>-3</sup>) with lower levels in spring (7.13±9.50 ng  
439 m<sup>-3</sup>) and winter (1.73±2.60 ng m<sup>-3</sup>). Mannitol maximized in summer (21.7±19.7 ng m<sup>-3</sup>)  
440 followed by autumn (18.2±19.9 ng m<sup>-3</sup>), spring (7.95±13.8 ng m<sup>-3</sup>) and winter (1.89±2.81 ng  
441 m<sup>-3</sup>). Arabitol and mannitol strongly co-varied throughout the study period. As depicted in  
442 thirteen-year monthly mean concentrations of total SCs (Figure 5b,c), we found elevated  
443 concentrations of sugar alcohols from May to October. Similar seasonal trends were reported  
444 for the aerosol samples collected from Gosan, Jeju Island in the western North Pacific Rim (Fu  
445 et al. 2012) and urban aerosol samples from Ghent, Belgium (Pashynska et al., 2002). In above  
446 studies, higher relative abundances of arabitol and mannitol in total sugar alcohols were  
447 reported during late summer to autumn. The higher concentration of arabitol in autumn was  
448 also reported for aerosol samples from the Mediterranean region in Israel (Burshtein et al.,  
449 2011). Erythritol and inositol showed the similar seasonal trend, but their concentrations are  
450 lower than the former two sugar species.

451 Sugar alcohols are emitted to the atmosphere from a variety of bacteria, few green algal  
452 lichens and fungi (Dahlman et al., 2003; Filippo et al., 2013). Arabitol and mannitol are  
453 abundant in fungal spores (Lewis and Smith, 1967; Yttri et al., 2007). Arabitol (r = 0.73) and  
454 mannitol (r = 0.80) showed a strong co-variance with trehalose, suggesting identical sources of  
455 sugar species in Chichijima. The PMF analysis showed that fungal factor and mixed factor



456 (fungal, vegetation, and microbial) accounted for 25% and 54.2% of total SCs observed in  
457 summer, respectively (Figure 6c). In autumn fungal and vegetation factor contributed 71% of  
458 total SCs detected in Chichijima aerosols (Figure 6c). In winter (Figure 6a) and spring (Figure  
459 6b) fungal and vegetation factor and mixed factor account for 31.2% and 37.2% of total SCs,  
460 respectively. This is reasonable because fungal and microbial activities are lower during  
461 winter/spring as compared to summer/autumn. The meteorological factors such as RH and  
462 temperature significantly affect fungal and bacterial activity (Kim and Xiao, 2005; Malik and  
463 Singh, 2004). Higher RH and temperature are crucial in increasing fungal and bacterial growth  
464 (Sharma and Razak, 2003). Their maximum growth was observed under the condition of 92-  
465 100% RH (Ibrahim et al., 2011). Higher concentrations of arabitol and mannitol in summer and  
466 autumn may be due to the increased fungal and bacterial activities in Chichijima Island in the  
467 western North Pacific.

468 Several studies have described the occurrence of fungi in marine environment (Jones,  
469 1976; Kohlmeyer and Kohlmeyer, 1991; Moss, 1986). The fungal species eject spores from  
470 hard materials like coral and sand grains. Some fungi also eject spores from woods associated  
471 with sand in summer/autumn when higher ambient temperature and RH are available (Jones  
472 and Mitchell, 1996). Marine fungal growths are observed on several mediums of substrates  
473 such as wood, sediments, muds, soil, sand, algae, corals, decaying leaves of mangroves and  
474 living animals in marine environment (Bremer, 1995; Kohlmeyer and Kohlmeyer, 1979;  
475 Nagakiri et al., 1996). However, the above studies have claimed to the occurrence and growth  
476 of marine fungi on several mediums of substrates but still the knowledge of the role of marine  
477 fungi in sediments and decaying dead animals are insufficient due to the lack of appropriate  
478 data set. It is still unclear if these fungi are active in sediments or if inactive spores are isolated  
479 (Hyde et al., 1998). Therefore, due to the inadequate data set, we doubt the marine contribution  
480 of sugar alcohols (arabitol, mannitol) in Chichijima aerosols.



481 Thirteen-year monthly mean concentrations of SCs clearly show slightly decreased  
482 concentrations of arabitol, mannitol, and erythritol in July and August; a similar trend was  
483 observed for trehalose (Figure 5b,c,e,j). These sugar compounds are derived from the microbial  
484 activities in source regions. The thirteen-year precipitation record over Chichijima Island  
485 shows that precipitations were lowered in July and August (Figure 2). The lower precipitation  
486 amount decreases the RH (Figure 2) and thus depresses the fungal and microbial activities. The  
487 lower precipitation also suppresses the moisture contents in the surface soil of Chichijima,  
488 which should cause a significant decline of local fungal and other microbial activities on the  
489 ground of Chichijima Island. Decreased precipitation might be a possible reason for the lower  
490 concentrations of arabitol, mannitol, and trehalose in July and August.

### 491 3.3 Annual variations and decadal comparisons of SCs

492 The annual variations in the concentrations of primary sugars and sugar alcohols are shown in  
493 Figure 7a. The annual mean concentrations of total SCs varied randomly during 2001 to 2013.  
494 As shown in Figure 7 (i, j, f and d), concentrations of sucrose, trehalose, xylose, and inositol  
495 increase from 2001 to 2013 in Chichijima aerosols. Similarly, arabitol (Figure 7b), glucose  
496 (Figure 7h), and fructose (Figure 7g) show clear increasing trends from 2006 to 2013 whereas  
497 mannitol (Figure 7c) and erythritol (Figure 7e) show a random trend. Thirteen-year data set  
498 provides significant information regarding the decadal variation in the atmospheric conditions  
499 over Chichijima Island in the western North Pacific. Here, we compare the data set of SCs for  
500 the periods of 1990-1993 (Period, P-I) with the current observations from 2001 to 2003 (P-II)  
501 and 2010 to 2013 (P-III) (Table 3) (Figures 8a, b, c).

502 The comparison for three periods indicates that concentrations of anhydrosugars are  
503 highest in winter followed by autumn. Their concentrations significantly increased from P-I to  
504 P-II/P-III (Figure 8a). The detailed discussions on anhydrosugars were reported in Verma et al.  
505 (2015). Here, we refer the data set of anhydrosugars for the decadal comparison with SCs.  
506 Interestingly, biomass-burning tracers (BB tracers; levoglucosan, mannosan and galactosan)



507 showed a significant difference in the decadal trends among three periods (i.e., P-I, P-II, and P-  
508 III) during winter and autumn. In winter, BB tracers showed an increasing trend from P-I to P-  
509 III (Figure 8a). Biomass burning is common in winter for house heating (Simoneit et al.,  
510 2004a), thus it is obvious that the lower ambient temperature is the more biomass burning  
511 activities are. Westerly winds abundantly transport biomass-burning products over Chichijima  
512 Island in the western North Pacific with air masses derived from East Asia, Siberia, Mongolia,  
513 and Russian Far East during winter (Bendle et al., 2007; Simoneit and Elias, 2010; Verma et al.,  
514 2015). In contrast, BB-tracers in autumn show an opposite trend (i.e., higher concentrations in  
515 P-I followed by P-II and P-III) compared to those in winter (Figure 8a).

516 The difference in the concentrations of anhydrosugars in winter and autumn during P-I is  
517 insignificant while the concentrations are 3 and 5 times higher in winter than autumn for P-II  
518 and P-III, respectively. This seasonal shifting in the concentrations of anhydrosugars may be  
519 attributable to the changes in the strength of both westerly and trade wind systems from mid  
520 autumn to early winter among three periods (Chen et al., 2013). In contrast, the concentrations  
521 of primary sugars were 2 to 7 times higher during P-II and P-III than P-I period in summer and  
522 autumn (Figure 8b). PMF analysis showed that local emissions from vegetation are important  
523 contributor for primary sugars (glucose, fructose and sucrose). Therefore, a drastic increase in  
524 concentration of primary sugars in summer/autumn for P-II and P-III than P-I may be caused  
525 by an increase in the emission of primary sugars by local vegetation under the influence of  
526 meteorological conditions in the western North Pacific. However, a possible soil dust  
527 contribution of primary sugar (sucrose) associated with an occasional air mass transport from  
528 Southeast Asia cannot be excluded.

529 Similarly to primary sugars, a drastic increase in the concentrations of sugar alcohols was  
530 observed for P-II and P-III compared to P-I period (Figure 8c). The concentrations of sugar  
531 alcohols in P-II and P-III are 6 to 19 times higher than those of P-I in summer/autumn (Table  
532 3; Figure 8c). Arabitol and mannitol are key sugar alcohols and reported as fungal and



533 microbial tracers, which contribute significantly to total SCs (Bauer et al., 2008b; Lewis and  
534 Smith, 1967; Zhang et al., 2010). Microbes such as fungi and bacteria are significantly  
535 increasing in the Asian and European countries (Yamaguchi et al., 2012). They are largely  
536 transported towards downwind regions in the Pacific Ocean from the Asian Continent in  
537 winter/spring under the influence of strong westerlies (Griffin et al., 2001, 2003, 2007; Hua et  
538 al., 2007; Uno et al., 2009) and settled down by wet and dry deposition in the western North  
539 Pacific according to air mass trajectories (Figure 3). Hirst and Stedman (1967) reported the life  
540 time of the fungal spores, they studied long-range spores transport and discussed that the  
541 fungus spores are between 2 and 200  $\mu\text{m}$  in diameter (mostly 5 to 30  $\mu\text{m}$ ); they have settling  
542 velocity between 0.005 and 3.0 cm/sec with a mode of less than 1 cm/sec, and different fungal  
543 spores vary significantly in shape and ornamentation. Most fungus spores are usually nearly  
544 spherical and larger, have a specific gravity close to 1.0. Because of their larger mass, settle  
545 velocity is at 1 to 40 cm/sec, with a mode of 3 cm/sec. This study suggested the possibilities of  
546 the fungal spores long-range fungal spores. Jeon et al. (2011) and Yamaguchi et al. (2012)  
547 have analyzed aerosol samples collected during the Asian dust event over the Sea of Japan, and  
548 identified similar groups of microbes (bacterial cells) transported from the source regions of  
549 Asian dust. Consequently, bacteria and fungi grow extensively during summer/autumn, when  
550 the climate conditions (i.e., higher RH and temperature) are favorable for their metabolic  
551 activities (Morris et al., 2004). Accordingly, an increased transport of bioaerosols since the last  
552 decade may have caused a drastic increase in the concentrations of sugar alcohols during P-II  
553 and P-III compared to P-I period over the western North Pacific.

#### 554 **3.4 Source apportionment of SCs**

555 To investigate the source apportionment of sugar components, the data sets of Chichijima  
556 aerosols were subjected to positive matrix factorization (PMF) analysis. Based on the PMF  
557 analysis, total five factors were determined to be significant to classify the sources of sugar



558 compounds (SCs). Five factors successfully explored the source profile for the individual sugar  
559 component. Factor profiles resolved by PMF analysis are shown in Figures 6 and 9-11, where  
560 percentages of each component summed for factors 1 to 5 are calculated to be 100%.

561 Vegetation factor (Figure 9) was dominated by xylose (75%), glucose (48%), and  
562 fructose (36%). Xylose is significantly produced by gymnosperm and angiosperm (Cowie and  
563 Hegdes, 1984; Sjoström, 1981). Fructose and glucose are highly water-soluble sugar species,  
564 and present in the bark and leaves of plants (Fu et al., 2012). Glucose is the second most  
565 abundant sugar that contributed to this factor. Cowie and Hegdes (1984) reported higher  
566 concentration of glucose in vascular plants and phytoplankton in the marine environment. The  
567 SCs emitted by the vegetation during growing season significantly contribute to vegetation  
568 factor. In Chichijima aerosols, glucose and fructose are significantly contributed in spring  
569 (Figure 6b), summer (Figure 6c) and autumn (Figure 6d), Therefore, the respective factors in  
570 Figure 6 are termed as a vegetation source for the both sugar species. This is reasonable  
571 because plants started growing in spring and summer seasons. In autumn, leaf senescence and  
572 decay result in the emission of glucose and fructose to the atmosphere.

573 Fungal and microbial factor (Figure 9) was characterized by trehalose (88%), mannitol  
574 (64%), and arabitol (54%). These sugars that contribute to fungal and microbial factor are  
575 associated with fungal spores, bacteria and yeast (Bauer et al., 2008a; Medeiros et al., 2006;  
576 Wiemken, 1990). The three sugars are good tracers of fungal spores and microbes (Loos et al.,  
577 1994; Rogge et al., 2007). Arabitol and mannitol are produced by a large variety of fungal  
578 species (Ion et al., 2005; Medeiros et al., 2006), and considered as a suitable tracer for fungal  
579 and bacterial metabolic activities (Bauer et al., 2008b; Elbein et al., 1974; Rogge et al., 2007).  
580 Arabitol is strongly correlated with mannitol ( $r=0.88$ ), suggesting similar sources for both  
581 species (Table 2) (Elbert et al., 2007). Fungi, bacteria, and other microbes in soils are the main  
582 sources for trehalose (Graham et al., 2003; Rogge et al., 2007; Simoneit et al., 2004). An



583 excellent correlation of trehalose with arabitol and mannitol suggested similar sources in the  
584 marine environment (Table 2) (Lewis and Smith, 1967).

585 Sugar alcohols have been proposed as tracers for microbes and fungal spores (Bauer et  
586 al., 2008b; Ion et al., 2005; Medeiros et al., 2006; Rogge et al., 2007). The fungal and  
587 microbial activities are considered higher during summer and autumn due to higher  
588 temperature and RH. The above discussions for the sources of arabitol, mannitol and trehalose  
589 are well supported by the seasonal PMF analysis. Arabitol and mannitol are well contributed in  
590 summer (Figure 6c, fungal factor and mixed factor) and autumn (Figure 6d, fungal and  
591 vegetation factor). Correspondingly, trehalose also contributed in summer (Figure 6c, mixed  
592 factor) and autumn (Figure 6d, microbial factor). Therefore, significant contributions of  
593 arabitol, mannitol and trehalose are observed during the respective seasons in Chichijima  
594 aerosols.

595 Mixed factor (Figure 9) is associated with erythritol (94%), arabitol (44%), mannitol  
596 (34%), inositol (32%), glucose (24%), and fructose (31%). Due to the highly miscellany  
597 characteristics of fungi, other microbes, and plant debris, it is quite difficult to specify the  
598 particular source for individual sugar species (Percival, 1970). Arabitol and mannitol are also  
599 attributed to the vegetation (photosynthesized by mature leaves) (Burshtein et al., 2011;  
600 Pashynska et al., 2002). The contributions of arabitol and mannitol in winter (Figure 6a, mixed  
601 factor) and spring (Figure 6b, mixed factor) indicate other sources than fungal spores in the  
602 Chichijima aerosols. Because the fungal and microbial growth is less important in winter and  
603 spring compared to summer/autumn. Therefore, this factor (Figure 9) should be associated with  
604 mixed sources from microbial and vegetational activities. Contemplating of mixed sources is  
605 very likely because sugar species that are highly apportioned to vegetation factor contribute, to  
606 some extent, to the same sources that are responsible to fungal and microbial factor (Figure 9).

607 BB (biomass burning) factor (Figure 9) is loaded significantly with levoglucosan (94%),  
608 mannosan (91%), and galactosan (90%), and moderately with xylose (20%). These species are





609 associated with biomass burning (Fraser and Lakshmanan, 2000; Graham et al., 2002;  
610 Simoneit, 2002). Nolte et al. (2001) and Medeiros et al. (2006) also reported that biomass  
611 burning-influenced aerosols are enriched with levoglucosan, mannosan, and galactosan.  
612 Kawamura et al. (2003) and Mochida et al. (2010) reported that biomass-burning products are  
613 abundantly transported to Chichijima under the influence of westerly winds during  
614 winter/spring (Figure 3). These results are well supported by the fact that BB factor is  
615 associated with SCs, which are derived from biomass burning in East Asia. The seasonal PMF  
616 analysis (Figure 6) also supports the above explanation; BB products are contributed highest in  
617 winter/spring (18.8% and 6.3%) under the influence of westerly winds, followed by  
618 summer/autumn (3.8% and 4.1%) for total sugars in the aerosol samples collected from  
619 Chichijima Island.

620 Pollen factor (Figure 9) is characterized by high loading of sucrose (91%), fructose  
621 (35%), and inositol (33%); these sugar species are associated with airborne pollen sources.  
622 Previous studies have also reported that sucrose is an excellent tracer for airborne pollen grains  
623 of flowering plants (Pacini, 2000; Graham et al., 2003; Wang et al., 2008; Medeiros et al.,  
624 2006). Fructose is well correlated with inositol ( $r=0.57$ ), indicating a similar origin for both  
625 sugar species (Table 2). Two prominent peaks of sucrose, fructose, and inositol, which  
626 appeared in late winter to early spring or summer, also indicate a similar source of those sugars.  
627 Contributions of sucrose (pollen factor) in winter (39.5%; Figure 6a) and spring (36%; Figure  
628 6b) are supported the sources of airborne pollen for sucrose in Chichijima Island. The sucrose  
629 contribution (soil dust factor) in non-flowering seasons, i.e., summer (17.1%; Figure 6c) and  
630 autumn (8.4%; Figure 6d) indicates different sources for sucrose in Chichijima aerosols.  
631 According to the seasonal PMF analysis (Figure 6), we termed additional sources of sucrose in  
632 Chichijima Island as soil dust factor as well as pollen factor.

633 In Figure 6a, b, c and d, PMF analysis for seasonal source identification indicated  
634 variable contributions of individual SCs in different seasons according to their seasonal source



635 origin. In winter (Figure 6a), airborne pollen (39.5%) contributed highest followed by  
636 vegetation, microbial, fungal (mixed) (31.2%), biomass burning (18.8%) and microbial  
637 (10.5%) sources. However, in spring, vegetation, microbial, and fungal (mixed) sources  
638 (37.2%) contributed almost equal to the airborne pollen (36.0%) followed by vegetation  
639 (21.2%) and biomass burning (6.3%) sources. The vegetation, microbial, fungal (mixed)  
640 (54.2%), fungal (25.2%), soil particles (17.1%) and biomass burning (3.8%) sources are  
641 characterized to maximize in summer season. Similar to summer, vegetation and fungal  
642 (mixed) sources (71.2%) are also leading to the contribution of total SCs followed by microbial  
643 (16.4%), soil particles (8.4%) and biomass burning (4.1%) in total SCs observed in autumn  
644 season for Chichijima aerosols.

645 Overall, average contributions of each factor to measured SCs as resolved by the PMF  
646 analyses are shown in Figure 10. Fungal and microbial factor accounts for 41% of total SCs  
647 measured. The emission from microbes including fungal spores was found as a dominant  
648 contributor to total SCs. Mixed factor (27%) indicates a common involvement of fungal,  
649 microbial and vegetation sources. Figure 11 shows annual trends in % contributions of five  
650 source factors to SCs in Chichijima aerosols. Fungal and microbial factor, and mixed factor  
651 contributed higher than other sources for SCs during 2001 to 2013. However, no clear trends in  
652 annual % contributions were observed for both source factors during thirteen-years study  
653 period. The sugar species assigned as pollen tracers were found to contribute 18% in pollen  
654 factor. Vegetation accounts for 11% of total SCs, indicating less emission from vegetation as  
655 compared to fungi and microbes. As indicated by BB factor, biomass-burning source  
656 contributes only 3% of total SCs.

657 Interestingly, we found an increasing trend in % contribution of vegetation, pollen, and  
658 BB factors to SCs in 2006 to 2013 (see Figure 11). Sugar components, which are contributed  
659 from pollen (sucrose, fructose, and inositol), vegetation (glucose and fructose), and biomass  
660 burning sources (levoglucosan, galactosan, and mannosan), also show a similar increasing



661 trend for the period of 2006 to 2013 (Figure 7). The increased annual trends of BB and pollen  
662 factors might be due to an enhanced long-range transport of airborne pollens and biomass  
663 burning products from the Asian continent to the western North Pacific under the influence of  
664 strong westerly winds. The increasing annual trends in % contribution of vegetation factor to  
665 SCs may denote a significantly increased activity of local vegetation in Chichijima Island from  
666 2006 to 2013, which could be involved with a recent global warming especially in the western  
667 North Pacific region (<http://climate.nasa.gov/vital-signs/global-temperature/>).

668

#### 669 **4 Summary and conclusions**

670 In this study we reported thirteen-years of temporal, seasonal and decadal observations on  
671 sugar compounds (SCs) in the aerosol samples collected at Chichijima Island in the western  
672 North Pacific, an outflow region of Asian aerosols. We observed the highest abundances of  
673 total SCs and primary sugars in summer, while sugar alcohols are almost equally distributed  
674 during summer and autumn. Thus, seasonal variations are well regulated by the atmospheric  
675 circulation and meteorological parameters of Chichijima Island. The seasonal distributions of  
676 arabitol, mannitol and trehalose are strongly influenced by long-range transport of microbe-,  
677 fungi- and bacteria-associated bioaerosols and their metabolic activities under the influences of  
678 westerly winds and other meteorological parameters (high RH and temperature) in  
679 summer/autumn.

680 Seasonal variation of sucrose is controlled by locally emitted and long-range transported  
681 pollen from East Asia to Chichijima during spring bloom periods. On the other hand, the  
682 increased concentrations of sucrose and fructose in summer may be caused by the local activity  
683 of vegetation and possibly by the atmospheric transport of plant root-associated soil dust  
684 particles potentially delivered from East and Southeast Asia with the occasional transport of air  
685 mass from the respective regions.



686 PMF analysis of long-term observations clearly indicated specific sources for individual  
687 SCs during different seasons. The results separate biogenic emissions into two parts, i.e.,  
688 vegetation and microbes including fungal species. The emissions from vegetation, pollen as  
689 well as microbial activities contributed almost 97% of total SCs determined, with the  
690 remaining fraction being derived from biomass burning activities. In the present decadal study,  
691 we found a drastic increase in the concentrations of sugar alcohols and primary sugars during  
692 2001-2003 and 2010-2013, which may be explained by a significant transport of bioaerosols in  
693 last decades from East Asia to Chichijima Island in the western North Pacific.

694

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700 request to the corresponding author (Kimitaka Kawamura, [kkawamura@isc.chubu.ac.jp](mailto:kkawamura@isc.chubu.ac.jp)).

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702 **References**

- 703 Andrews, J. T., W. N. Mode, and P. T. Davis (1980), Holocene climate based on pollen transfer-  
704 functions, Eastern Canadian Arctic, *Arct. Alp. Res.*, 12, 41-64.
- 705 Baker, H. G., I. Baker, S. A. Hodges (1998), Sugar composition of nectars and fruits consumed  
706 by birds and bats in the tropics and subtropics, *Biotropica*. 30, 559-586.
- 707 Bauer, H., E. Schueller, G. Weinke, A. Berger, R. Hitzemberger, I. L. Marr, and H. Puxbaum  
708 (2008a), Significant contributions of fungal spores to the organic carbon and to the  
709 aerosol mass balance of the urban atmospheric aerosol, *Atmos. Environ.*, 42(22), 5542-  
710 5549.
- 711 Bauer, H., M. Claeys, R. Vermeylen, E. Schueller, G. Weinke, A. Berger, and H. Puxbaum  
712 (2008b), Arabitol and mannitol as tracers for the quantification of airborne fungal spores,  
713 *Atmos. Environ.*, 42(3), 588-593.
- 714 Bendle, J., K. Kawamura, K. Yamazaki, and T. Niwai (2007), Latitudinal distribution of  
715 terrestrial lipid biomarkers and n-alkane compound-specific stable carbon isotope ratios  
716 in the atmosphere over the western Pacific and Southern Ocean, *Geochim. Cosmochim.*  
717 *Acta*, 71(24), 5934-5955.
- 718 Bieleski, R. L. (1995), Onset of phloem export from senescent petals of Daylily, *Plant Physiol.*,  
719 109(2), 557-565.
- 720 Bourgeois, J. C., K. Gajewski, and R. M. Koerner (2001), Spatial patterns of pollen deposition  
721 in arctic snow, *J. Geophys. Res.*, 106, 5255-5265.
- 722 Bremer, G. B. (1995), Lower marine fungi (Labyrinthulomycetes) and the decay of mangrove  
723 leaf-litter, *Hydrobiologia*, 304(3), 243-245.
- 724 Brown, J. K. M., and M. S. Hovmoller (2002), Epidemiology—Aerial dispersal of pathogens on  
725 the global and continental scales and its impact on plant disease, *Science*, 297(5581),  
726 537-541.
- 727 Burshtein, N., N. Lang-Yona, and Y. Rudich (2011), Ergosterol, arabitol and mannitol as tracers  
728 for biogenic aerosols in the eastern Mediterranean, *Atmos. Chem. Phys.*, 11(2), 829-839.
- 729 Bzdusek P.A., E. R. Christensen, C. M. Lee, U. Pakadeesusuk, D. C. Freedman, (2006). PCB  
730 congeners and dechlorination in sediments of Lake Hartwell, South Carolina, determined  
731 from cores collected in 1987 and 1988. *Environ. Sci. Technol.*, 40, 109-119.
- 732 Campbell, I. D., K. McDonald, M. Flannigan, and J. Kringayark (1999), Long-distance transport  
733 of pollen into the Arctic, *Nature*, 399, 29- 30.
- 734 Carvalho, A., C. Pio, and C. Santos (2003), Water-soluble hydroxylated organic compounds in  
735 German and Finnish aerosols, *Atmos. Environ.*, 37(13), 1775-1783.
- 736 Chen, J., K. Kawamura, C. Q. Liu, and P. Q. Fu (2013), Long-term observations of saccharides  
737 in remote marine aerosols from the western North Pacific: A comparison between 1990-  
738 1993 and 2006-2009 periods, *Atmos. Environ.*, 67, 448-458.
- 739 Cowie, G. L., and J. I. Hedges (1984), Carbohydrate sources in a coastal marine-environment,  
740 *Geochim. Cosmochim. Acta*, 48(10), 2075-2087.
- 741 Dahlman, L., J. Persson, T. Nasholm, and K. Palmqvist (2003), Carbon and nitrogen distribution  
742 in the green algal lichens *Hypogymnia physodes* and *Platismatia glauca* in relation to  
743 nutrient supply, *Planta*, 217(1), 41-48.
- 744 Duce, R. A., et al. (1980), Long-range atmospheric transport of soil dust from Asia to the



- 745 tropical North Pacific: temporal variability, *Science*, 209, 1522–1524.
- 746 Elbein, A. D. (1974), The metabolism of  $\alpha$ ,  $\alpha$ -trehalose, *Adv. Carboh. Chem. and Biochem.*, 30,  
747 227–256.
- 748 Elbert, W., P. E. Taylor, M. O. Andreae, and U. Poschl (2007), Contribution of fungi to primary  
749 biogenic aerosols in the atmosphere: wet and dry discharged spores, carbohydrates, and  
750 inorganic ions, *Atmos. Chem. Phys.*, 7(17), 4569–4588.
- 751 Elliot S., Blake D.R., Duce R.A., Lai C.A., McCreary I., McNair L.A., Rowland F.S., Russell  
752 A.G., Streit G.E., and Turco R.P. Motorization of China implies changes in pacific air  
753 chemistry and primary production. *Geophys. Res. Letts.* 24, 2671 –2674, 1997.
- 754 Filippo, P. D., D. Pomata, C. Riccardi, F. Buiarelli, and C. Perrino (2013), Fungal contribution  
755 to size-segregated aerosol measured through biomarkers, *Atmos. Environ.*, 64,132-140.
- 756 Fraser, M. P., and K. Lakshmanan (2000), Using levoglucosan as a molecular marker for the  
757 long-range transport of biomass combustion aerosols, *Environ. Sci. Technol.*, 34(21),  
758 4560-4564.
- 759 Fu, P. Q., K. Kawamura, M. Kobayashi, and B. R. T. Simoneit (2012), Seasonal variations of  
760 sugars in atmospheric particulate matter from Gosan, Jeju Island: Significant  
761 contributions of airborne pollen and Asian dust in spring, *Atmos. Environ.*, 55, 234-239.
- 762 Fuzzi, S., et al. (2007), Overview of the inorganic and organic composition of size-segregated  
763 aerosol in Rondonia, Brazil, from the biomass-burning period to the onset of the wet  
764 season, *J. Geophys. Res.-Atmos.*, 112(D1), doi:10.1029/2005jd006741.
- 765 Graham, B., O. L. Mayol-Bracero, P. Guyon, G. C. Roberts, S. Decesari, M. C. Facchini, P.  
766 Artaxo, W. Maenhaut, P. Koll, and M. O. Andreae (2002), Water-soluble organic  
767 compounds in biomass burning aerosols over Amazonia - 1. Characterization by NMR  
768 and GC-MS, *J. Geophys. Res.-Atmos.*, 107(D20), doi:10.1029/2001jd000336.
- 769 Graham, B., P. Guyon, P. E. Taylor, P. Artaxo, W. Maenhaut, M. M. Glovsky, R. C. Flagan, and  
770 M. O. Andreae (2003), Organic compounds present in the natural Amazonian aerosol:  
771 Characterization by gas chromatography-mass spectrometry, *J. Geophys. Res.-Atmos.*,  
772 108(D24), doi:10.1029/2003jd003990.
- 773 Griffin, D. W., et al. (2001), African desert dust in the Caribbean atmosphere: Microbiology and  
774 public health, *Aerobiologia*, 17, 203–213.
- 775 Griffin, D. W., et al. (2003), Atmospheric microbiology in the northern Caribbean during  
776 African dust events, *Aerobiologia*, 19, 143–157.
- 777 Griffin, D. W., (2007), Atmospheric movement of microorganisms in clouds of desert dust and  
778 implications for human health, *Clin. Microbiol. Rev.* 20, 459–477.
- 779 Hackl, E., G. Bachmann, and S. Zechmeister-Boltenstern (2000), Soil microbial biomass and  
780 rhizosphere effects in natural forest stands, *Phyton.-Ann. Rei. Bot. A*, 40(4), 83-90.
- 781 Hader, J. D., T. P. Wright, and M. D. Petters (2014), Contribution of pollen to atmospheric ice  
782 nuclei concentrations, *Atmos. Chem. Phys.*, 14, 5433-5449.
- 783 Hicks, S., H. Tinsley, A. Huusko, C. Jensen, M. Hattestrand, A. Gerasimides, and E. Kvavadze  
784 (2001), Some comments on spatial variation in arboreal pollen deposition: First records  
785 from the Pollen Monitoring Programme (PMP), *Rev. Palaeobot. Palynol.*, 117, 183-194.
- 786 Hirst J. M. and O. J. Stedman (1967), Long-distance spore transport: Methods of measurement,  
787 vertical spores profile and the detection of immigrant spores, *J. Gen. Microbiol.* 48, 329-  
788 355.



- 789 Hjelmroos, M., and L. G. Franzen (1994), Implications or recent long distance pollen transport  
790 events for the interpretation of fossil pollen records in Fennoscandia, *Rev. Palaeobot.*  
791 *Palynol.*, 82, 175-189.
- 792 Hua, N.-P. et al. (2007), Detailed identification of desert-originated bacteria carried by Asian  
793 dust storms to Japan, *Aerobiologia*, 23, 291-298.
- 794 Hyde, K. D., E. B. G. Jones, E. Leano, S. B. Pointing, A. D. Poonyth, and L. L. P. Vrijmoed  
795 (1998), Role of fungi in marine ecosystems, *Biodivers. Conserv.*, 7(9), 1147-1161.
- 796 Ibrahim, M., A. B. Rabah, B. Liman, and N. T. Ibrahim (2011), Effect of temperature and  
797 relative humidity on the growth of *Helminthosporium fulvum*, *Nigerian J. Basic Appli.*  
798 *Sci.*, 19(1) 127- 129.
- 799 Ion, A. C., R. Vermeylen, I. Kourtchev, J. Cafmeyer, X. Chi, A. Gelencser, W. Maenhaut, and  
800 M. Claeys (2005), Polar organic compounds in rural PM(2.5) aerosols from K-pusztá,  
801 Hungary, during a 2003 summer field campaign: Sources and diel variations, *Atmos.*  
802 *Chem. Phys.*, 5, 1805-1814.
- 803 Iwasaka, Y., et al. (1983), The transport and spacial scale of Asian dust-storm clouds: a case  
804 study of the dust-storm event of April 1979, *Tellus*, 35B, 189–196.
- 805 Jaffe, D., et al. (1999), Transport of Asian air pollution to North America, *Geophys. Res. Lett.*,  
806 26(6), 711–714.
- 807 Jaffe, D., I. McKendry, T. Anderson, and H. Price (2003), Six “new” episodes of trans-Pacific  
808 transport of air pollutants, *Atmos. Environ.*, 37(3), 391–404.
- 809 Jaffe, D.A., Mahura A., Kelley, J., Atkins J., Novelli P.C. and Merrill J. Impact of Asian  
810 Emissions on the Remote North Pacific Atmosphere: Interpretation of CO Data from  
811 Shemya, Guam, Midway and Mauna Loa. *J. Geophys. Res.* 23, 28,627-28,636, 1997.
- 812 Jeon, E.-M. et al. (2011), Impact of Asian dust events on airborne bacterial community assessed  
813 by molecular analyses, *Atmos. Environ.* 45, 4313–4321.
- 814 Jia, Y. L., A. L. Clements, and M. P. Fraser (2010), Saccharide composition in atmospheric  
815 particulate matter in the southwest US and estimates of source contributions, *J. Aerosol*  
816 *Sci.*, 41(1), 62-73.
- 817 Jones, E.B.G. (ed.) (1976), *Recent Advances in Aquatic Mycology*, London: Elek Science.
- 818 Jones, E.B.G., and J. I. Mitchell (1996), Biodiversity of marine fungi, In *Biodiversity:*  
819 *International Biodiversity Seminar* (A. Cimerman and N. Gunde-Cimerman, eds), pp. 31-  
820 42. Ljubljana: National Inst. Chemistry and Slovenia National Commission for UNESCO.
- 821 Juntto S and P. Paatero (1994), Analysis of daily precipitation data by positive matrix  
822 factorization. *Environmetrics*, 5, 127–144.
- 823 Kanakidou, M., et al. (2005), Organic aerosol and global climate modelling: a review, *Atmos.*  
824 *Chem. Phys.*, 5, 1053-1123.
- 825 Kawamura, K., Y. Ishimura, and K. Yamazaki (2003), Four years' observations of terrestrial  
826 lipid class compounds in marine aerosols from the western North Pacific, *Global*  
827 *Biogeochem. Cy.*, 17(1), doi:10.1029/2001gb001810.
- 828 Kim, V. K., and C.L. Xiao (2005), Influence of culture media and environmental factors on  
829 mycelial growth and picnidial production of *Sphaeropsis pyriputrescens*, *Mycologia*,  
830 97(1): 25-32.
- 831 Kohlmeyer, J., and B. Volkmann-Kohlmeyer (1991), *Illustrated key to the filamentous fungi*,





- 832 Bot. Mar. 34, 1-61.
- 833 Kohlmeyer, J., and E. Kohlmeyer (1979), *Marine Mycology: The Higher Fungi*, London:  
834 Academic Press.
- 835 Lacey, M. E., and J. S. West (2006), *The Air Spora: a Manual for Catching and Identifying*  
836 *Airborne Biological Particles*, pp. 32-33.
- 837 Lewis, D. H., and D. C. Smith (1967), Sugar alcohols (polyols) in fungi and green plants. I.  
838 Distribution physiology and metabolism, *New Phytologist*, 66(2), 143.
- 839 Loos, H., R. Kramer, H. Sahm, and G. A. Sprenger (1994), Sorbitol promotes growth of  
840 *Zymomonas mobilis* in environments with high concentrations of sugar: evidence for a  
841 physiological function of glucose-fructose oxidoreductase in osmoprotection, *J.*  
842 *Bacteriol.*, 176(24), 7688-7693.
- 843 Lu J., P. Jiang, et al. (2008), Assessing soil quality data by positive matrix factorization.  
844 *Geoderma*, 145, 259-266.
- 845 Ma, S. X., Z. Z. Wang, X. H. Bi, et al. (2009), Composition and source of saccharides in  
846 aerosols in Guangzhou, China, *Chinese Sci. Bull.*, 54, 4500-4506.
- 847 Makra, L., T. Santa, I. Matyasovszky, A. Damialis, K. Karatzas, K. C. Bergmann, and D. Vokou  
848 (2010), Airborne pollen in three European cities: Detection of atmospheric circulation  
849 pathways by applying three-dimensional clustering of backward trajectories, *J. Geophys.*  
850 *Res.-Atmos.*, 115, doi:10.1029/2010jd014743.
- 851 Malik, V. K., and S. Singh (2004), Effect of temperature and relative humidity on teliospore  
852 germination in *Ustilago hordei*, *J. Mycol. Plant Pathol.*, 34, 410-411.
- 853 Martin, S. T., et al. (2010), Sources and Properties of Amazonian Aerosol Particles, *Rev.*  
854 *Geophys.*, 48, doi:10.1029/2008rg000280.
- 855 Medeiros, P. M., M. H. Conte, J. C. Weber, and B. R. T. Simoneit (2006), Sugars as source  
856 indicators of biogenic organic carbon in aerosols collected above the Howland  
857 Experimental Forest, Maine, *Atmos. Environ.*, 40(9), 1694-1705.
- 858 Medeiros, P. M., and B. R. T. Simoneit, (2008), Source profiles of organic compounds emitted  
859 upon combustion of green vegetation from temperate climate forests, *Environ. Sci.*  
860 *Technol.*, 42, 8310-8316.
- 861 Miguel, A. G., Taylor P. E., House J., Glovsky M. M., and Flagan R. C. (2006), Meteorological  
862 influences on respirable fragment release from Chinese elm pollen, *Aerosol Sci. Technol.*  
863 40, 690-696.
- 864 Mims, S. A., and F. M. Mims (2003), Fungal spores are transported long distances in smoke  
865 from biomass fires, *Atmos. Environ.*, 38, 651-655.
- 866 Miyazaki, Y., J. Jung, P. Fu, Y. Mizoguchi, K. Yamanoi, and K. Kawamura (2012), Evidence of  
867 formation of submicrometer water-soluble organic aerosols at a deciduous forest site in  
868 northern Japan in summer, *J. Geophys. Res.*, 117, D19213, doi:10.1029/2012JD018250.
- 869 Mochida, M., A. Kawabata, K. Kawamura, H. Hatsushika, and K. Yamazaki (2003a), Seasonal  
870 variation and origins of dicarboxylic acids in the marine atmosphere over the western  
871 North Pacific, *J. Geophys. Res.-Atmos.*, 108(D6), doi:10.1029/2002jd002355.
- 872 Mochida, M., K. Kawamura, N. Umemoto, M. Kobayashi, S. Matsunaga, H. J. Lim, B. J. Turpin,  
873 T. S. Bates, and B. R. T. Simoneit (2003b), Spatial distributions of oxygenated organic  
874 compounds (dicarboxylic acids, fatty acids, and levoglucosan) in marine aerosols over  
875 the western Pacific and off the coast of East Asia: Continental outflow of organic





- 876 aerosols during the ACE-Asia campaign, *J. Geophys. Res.-Atmos.*, 108(D23),  
877 doi:10.1029/2002jd003249.
- 878 Mochida, M., K. Kawamura, P. Q. Fu, and T. Takemura (2010), Seasonal variation of  
879 levoglucosan in aerosols over the western North Pacific and its assessment as a biomass-  
880 burning tracer, *Atmos. Environ.*, 44(29), 3511-3518.
- 881 Morris, C. E., D. G. Georgakopoulos, and D. C. Sands (2004), Ice nucleation active bacteria and  
882 their potential role in precipitation, *J. Phys. Chem. A*, 108, 87-103.
- 883 Moss, S.T. (ed.) (1986), *The Biology of Marine Fungi*, Cambridge: Cambridge University Press.
- 884 Nakagiri, A., S. Y. Newell, T. Ito, and T. K. Tan (1996), Biodiversity and ecology of the  
885 oomycetous fungus, *Halophytophthora*, In *Biodiversity and the Dynamics of Ecosystems*.  
886 (DIWPA series. Vol.1) (I.M. Turner, C.H. Diong, S.S.L. Lim, and P.K.L. Ng, eds), pp.  
887 273-80. Tokyo: International Network for DIVERSITAS in Western Pacific and Asia.
- 888 Nolte, C. G., J. J. Schauer, G. R. Cass, and B. R. T. Simoneit (2001), Highly polar organic  
889 compounds present in wood smoke and in the ambient atmosphere, *Environ. Sci.*  
890 *Technol.*, 35(10), 1912-1919.
- 891 Paatero, P., and U. Tapper (1994), Positive Matrix Factorization - a Nonnegative Factor Model  
892 with Optimal Utilization of Error-Estimates of Data Values, *Environmetrics*, 5(2), 111-  
893 126.
- 894 Paatero, P. et al. (2002), Understanding and controlling rotations in factor analytic models,  
895 *Chemom. Intell. Lab. Syst.*, 60, 253-264.
- 896 Pacini, E. (2000), From anther and pollen ripening to pollen presentation, *Plant Syst. Evol.* 222,  
897 19-43.
- 898 Pashynska, V., R. Vermeylen, G. Vas, W. Maenhaut, and M. Claeys (2002), Development of a  
899 gas chromatographic/ion trap mass spectrometric method for the determination of  
900 levoglucosan and saccharidic compounds in atmospheric aerosols. Application to urban  
901 aerosols, *J. Mass Spectrom.*, 37(12), 1249-1257.
- 902 Pavuluri, C. M., K. Kawamura, E. Tachibana, and T. Swaminathan (2010), Elevated nitrogen  
903 isotope ratios of tropical Indian aerosols from Chennai: Implication for the origins of  
904 aerosol nitrogen in South and Southeast Asia, *Atmos. Environ.*, 44(29), 3597-3604.
- 905 Polissar A.V., P. K. Hopke, P. Paatero, Y. J. Kaufmann, D. K. Hall, B.A. Bodhaine, E.G. Dutton,  
906 J. M. Harris, (1999). The aerosol at Barrow, Alaska: long-term trends and source  
907 locations. *Atmos. Environ.*, 33, 2441-2458.
- 908 Prospero, J.M. and D.L. Savoie (1989), Effect of continental sources of nitrate concentrations  
909 over the Pacific Ocean, *Nature*, 339, 687- 689.
- 910 Prospero, J. M. et al. (2005), Interhemispheric transport of viable fungi and bacteria from Africa  
911 to the Caribbean with soil dust, *Aerobiologia*, 21, 1-19.
- 912 Rogge, W. F., P. M. Medeiros, and B. R. T. Simoneit (2006), Organic marker compounds for  
913 surface soil and fugitive dust from open lot dairies and cattle feedlots, *Atmos. Environ.*,  
914 40(1), 27-49.
- 915 Rogge, W. F., P. M. Medeiros, and B. R. T. Simoneit (2007), Organic marker compounds in  
916 surface soils of crop fields from the San Joaquin Valley fugitive dust characterization  
917 study, *Atmos. Environ.*, 41, 8183-8204.
- 918 Rousseau, D.-D., P. Schevin, J. Ferrier, D. Jolly, T. Andreasen, S. E. Ascanius, S. E. Hendriksen,  
919 and U. Poulsen (2008), Long-distance pollen transport from North America to Greenland



- 920 in spring, *J. Geophys. Res.-Biogeo.*, 113(G2), doi:10.1029/2007jg000456.
- 921 Rousseau, D.-D., D. Duzer, J.-L. Etienne, G. Cambon, D. Jolly, J. Ferrier, and P. Schevin (2004),  
922 Pollen record of rapidly changing air trajectories to the North Pole, *J. Geophys. Res.*, 109,  
923 D06116, doi:10.1029/2003JD003985.
- 924 Rousseau, D.-D., P. Schevin, D. Duzer, G. Cambon, J. Ferrier, D. Jolly, and U. Poulsen (2006),  
925 New evidence of long distance pollen transport to southern Greenland in late Spring, *Rev.*  
926 *Palaeobot. Palynol.*, 141, 277–286.
- 927 Schmidl, C., H. Bauer, A. Dattler, R. Hitzemberger, G. Weissenboeck, I. L. Marr, and H.  
928 Puxbaum (2008), Chemical characterisation of particle emissions from burning leaves,  
929 *Atmos. Environ.*, 42(40), 9070-9079.
- 930 Seinfeld, J. H., et al., (2004), ACE-ASIA - Regional climatic and atmospheric chemical effects  
931 of Asian dust and pollution, *Bull. Amer. Meteorol. Soc.*, 85(3), 367-380.
- 932 Sharma, R. and R. C. Rajak (2003), Keratinophilic fungi: Nature's keratin degrading machines,  
933 *Resonance*, 8, 28-30.
- 934 Simoneit, B. R. T. (2002), Biomass burning - A review of organic tracers for smoke from  
935 incomplete combustion, *Appl. Geochem.*, 17(3), 129-162.
- 936 Simoneit, B. R. T. (2004), Biomarkers (molecular fossils) as geochemical indicators of life, *Adv.*  
937 *Space. Res.*, 33(8), 1255-1261.
- 938 Simoneit, B. R. T., and V. O. Elias (2000), Organic tracers from biomass burning in  
939 atmospheric particulate matter over the ocean, *Mar. Chem.*, 69(3-4), 301-312.
- 940 Simoneit, B. R. T., V. O. Elias, M. Kobayashi, K. Kawamura, A. I. Rushdi, P. M. Medeiros, W.  
941 F. Rogge, and B. M. Didyk (2004a), Sugars-dominant water-soluble organic compounds  
942 in soils and characterization as tracers in atmospheric particulate matter, *Environ. Sci.*  
943 *Technol.*, 38(22), 5939-5949.
- 944 Simoneit, B. R. T., M. Kobayashi, M. Mochida, K. Kawamura, M. Lee, H. J. Lim, B. J. Turpin,  
945 and Y. Komazaki (2004b), Composition and major sources of organic compounds of  
946 aerosol particulate matter sampled during the ACE-Asia campaign, *J. Geophys. Res.-*  
947 *Atmos.*, 109(D19), doi:10.1029/2004jd004598.
- 948 Slinn, S. A., and W. G. N. Slinn (1980), Predictions for particle deposition on natural waters,  
949 *Atmos. Environ.*, 14, 1013-1016.
- 950 Soonthornnonda P. and E. R. Christensen (2008). Source apportionment of pollutants and flows  
951 of combined sewer wastewater. *Water Res.*, 42, 1989–1998.
- 952 Sosnoskie, L. M., T. M. Webster, D. Dales, G. C. Rains, T. L. Grey, and A. S. Culpepper (2009),  
953 Pollen grain size, density, and settling velocity for Palmer amaranth (*Amaranthus*  
954 *palmeri*), *Weed Science*, 57, 404-409. <http://dx.doi.org/10.1614/WS-08-157.1>.
- 955 Speranza, A., G. L. Calzoni, and E. Pacini (1997), Occurrence of mono- or disaccharides and  
956 polysaccharide reserves in mature pollen grains, *Sex. Plant Reprod.*, 10(2), 110–115.
- 957 Sullivan, A. P., N. Frank, D. M. Kenski, Jr., and J. L. Collett (2011), Application of high-  
958 performance anion-exchange chromatography-pulsed amperometric detection for  
959 measuring carbohydrates in routine daily filter samples collected by a national network: 2.  
960 Examination of sugar alcohols/polyols, sugars, and anhydrosugars in the upper Midwest,  
961 *J. Geophys. Res.* 116. <http://dx.doi.org/10.1029/2010JD014169>.
- 962 Talbot R.W. et al. (1997), Chemical characteristics of continental outflow from Asia to the  
963 troposphere over the western Pacific Ocean during February-March 1994: Results from



- 964 PEM -West B. J. Geophys. Res., 102, 28,255-28,274, 1997.
- 965 Uno, I., K. Eguchi, K. Yumimoto, T. Takemura, A. Shimizu, M. Uematsu, Z. Y. Liu, Z. F.  
966 Wang, Y. Hara, and N. Sugimoto (2009), Asian dust transported one full circuit around  
967 the globe, *Nat. Geosci.*, 2(8), 557-560.
- 968 Verma, S. K., K. Kawamura, J. Chen, P. Q. Fu and C. Zhu (2015), Thirteen years observation of  
969 biomass-burning organic tracers over Chichijima Island in the western North Pacific: An  
970 outflow region of Asian aerosols, *J. Geophys. Res. Atmos.*, 120,  
971 doi:10.1002/2014JD022224. .
- 972 Wan, E. C. H., and J. Z. Yu (2007), Analysis of sugars and sugar polyols in atmospheric  
973 aerosols by chloride attachment in liquid chromatography/negative ion electrospray mass  
974 spectrometry, *Environ. Sci. Technol.*, 41(7), 2459-2466.
- 975 Wang, G. H., C. L. Chen, J. J. Li, B. H. Zhou, M. J. Xie, S. Y. Hu, K. Kawamura, and Y. Chen  
976 (2011), Molecular composition and size distribution of sugars, sugar-alcohols and  
977 carboxylic acids in airborne particles during a severe urban haze event caused by wheat  
978 straw burning, *Atmos. Environ.*, 45(15), 2473-2479,
- 979 Wang, G. H., K. Kawamura, and M. Lee (2009), Comparison of organic compositions in dust  
980 storm and normal aerosol samples collected at Gosan, Jeju Island, during spring 2005,  
981 *Atmos. Environ.*, 43(2), 219-227.
- 982 Wiemken, A. (1990), Trehalose in Yeast, Stress Protectant Rather Than Reserve Carbohydrate,  
983 *Anton. Leeuw. Int. J. G.*, 58(3), 209-217.
- 984 Wright, T. P., J. D. Hader, G. R. McMeeking, and M. D. Peters (2014), High relative humidity  
985 as a trigger for widespread release of ice Nuclei, *Aerosol Sci. Technol.*, 48, 11, i-v, DOI:  
986 10.1080/02786826.2014.968244.
- 987 Xie Y. and C. M. Berkowitz (2006), The use of positive matrix factorization with conditional  
988 probability functions in air quality studies: an application to hydrocarbon emissions in  
989 Houston, Texas. *Atmos. Environ.*, 40, 3070– 3091.
- 990 Xu, W. Y., et al. (2011), Characteristics of pollutants and their correlation to meteorological  
991 conditions at a suburban site in the North China Plain, *Atmos. Chem. Phys.*, 11(9), 4353-  
992 4369.
- 993 Yamaguchi, N., T. Ichijo, A. Sakotani, T. Baba, and M. Nasu (2012), Global dispersion of  
994 bacterial cells on Asian dust, *Sci. Rep.*, 2, DOI: 10.1038/srep00525.
- 995 Yang, Y. H., C. Y. Chan, J. Tao, M. Lin, G. Engling, Z. S. Zhang, T. Zhang, and L. Su (2012),  
996 Observation of elevated fungal tracers due to biomass burning in the Sichuan Basin at  
997 Chengdu City, China, *Sci. Total Environ.*, 431, 68–77.
- 998 Yttri, K. E., C. Dye, and G. Kiss (2007), Ambient aerosol concentrations of sugars and sugar-  
999 alcohols at four different sites in Norway, *Atmos. Chem. Phys.*, 7(16), 4267-4279.
- 1000 Zhang, T., Engling, G., Chan, C.-Y., Zhang, Y.-N., Zhang, Z.-S., Lin, M., Sang, X.-F., Li, Y.D.,  
1001 Li, Y.-S., 2010. Contribution of fungal spores to particulate matter in a tropical rainforest.  
1002 *Environ. Res. Lett.* 5. doi:10.1088/1748-9326/1085/1082/ 024010.
- 1003 Zhou, L., E. Kim et al., (2004), Advanced factor analysis on Pittsburgh particle size distribution  
1004 data, *Aerosol Sci. Technol.*, 38, 118-132.
- 1005 Zhu, C., K. Kawamura, and B. Kunwar (2015), Organic tracers of primary biological aerosol  
1006 particles at subtropical Okinawa Island in the western North Pacific Rim, *J. Geophys.*  
1007 *Res. Atmos.*, 120, 5504–5523.



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1009 **Figure Captions**

1010 **Figure 1.** Geographical location of Chichijima Island (27°04'N; 142°13'E; 254 m, asl) in the  
1011 western North Pacific.

1012 **Figure 2.** The monthly variation of the meteorological parameters over Chichijima Island  
1013 during 2001-2013 (The error bars denote the standard deviations).

1014 **Figure 3.** The seasonal ten-day air mass backward trajectories over Chichijima for 2012  
1015 (winter: Dec-Feb, spring: Mar-May, summer: Jun-Aug, autumn: Sep-Nov). The trajectory  
1016 calculations were performed everyday starting at Chichijima Island.

1017 **Figure 4.** Temporal plots for the concentrations ( $\text{ng m}^{-3}$ ) of sugar compounds in Chichijima  
1018 aerosol samples collected for 2001-2013 in the western North Pacific.

1019 **Figure 5.** Monthly mean concentrations ( $\text{ng m}^{-3}$ ) of sugar compounds in aerosol samples from  
1020 Chichijima Island in the western North Pacific during 2001-2013.

1021 **Figure 6.** Seasonal source contributions to sugar compounds from various sources based on  
1022 PMF analyses. (BB – biomass-burning; Mixed – vegetation, fungal and microbial sources).

1023 **Figure 7.** Annual mean concentrations ( $\text{ng m}^{-3}$ ) of sugar compounds in aerosol samples  
1024 collected from Chichijima Island in the western North Pacific during 2001-2013.

1025 **Figure 8.** The seasonal concentrations of anhydrosugars (biomass burning tracers), primary  
1026 sugars and sugar alcohols measured in Chichijima aerosols during three periods, i.e., P-I  
1027 (1990-1993), P-II (2001-2003) and P-III (2010-2013).

1028 **Figure 9.** PMF analyses of sugar compounds in Chichijima aerosols based on the 2001-2013  
1029 data set. (BB – biomass-burning; Mixed – vegetation, fungal and microbial sources).

1030 **Figure 10.** Source contributions to sugar compounds from various sources based on PMF  
1031 analyses. (BB – biomass-burning; Mixed – vegetation, fungal and microbial sources).

1032 **Figure 11.** Annual trends in % contributions of five source factors: (a) vegetation, (b) fungal  
1033 and microbial, (c) mixed, (d) biomass burning (BB), and (e) pollen factors to SCs in  
1034 Chichijima aerosols. The data of 2005 are not plotted due to limited data points.

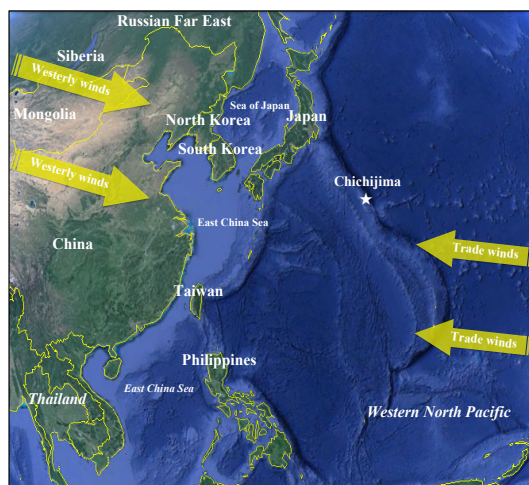
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**Figure 1.**



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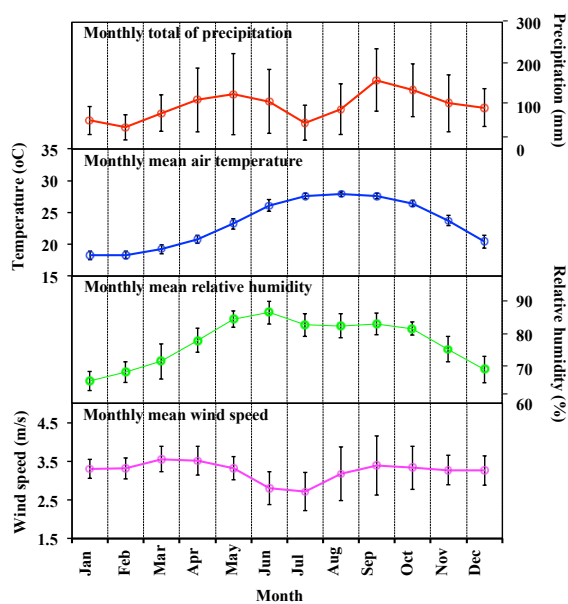
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1071 **Figure 2.**



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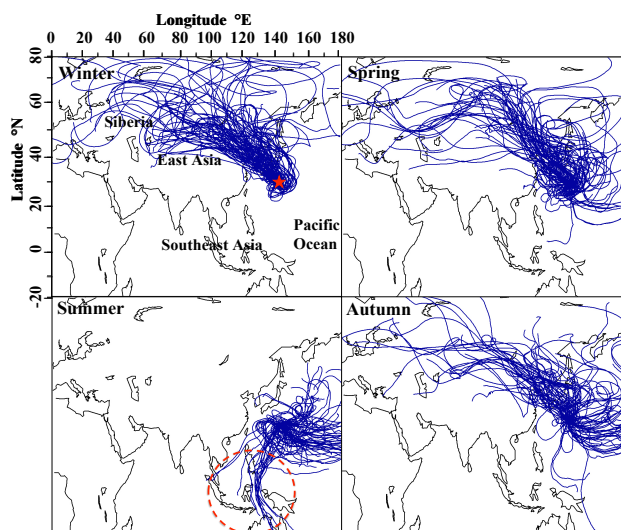
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1089 **Figure 3.**



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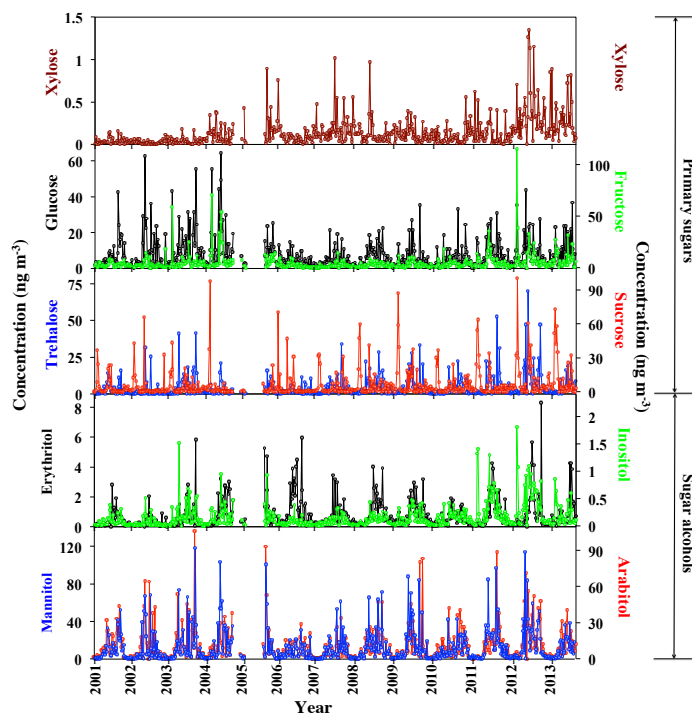
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Figure 4.



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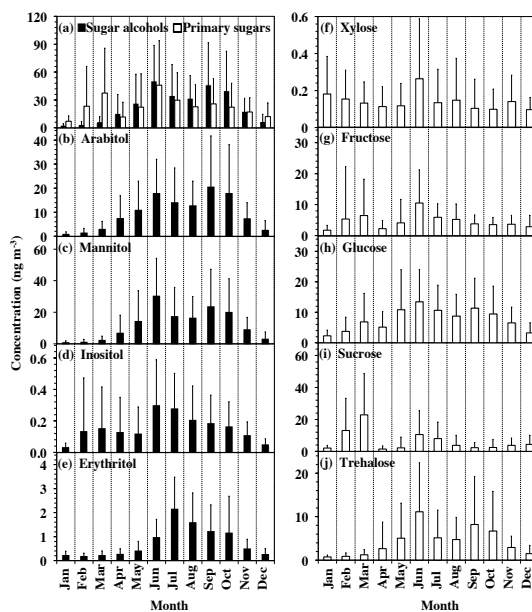
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1127 **Figure 5.**



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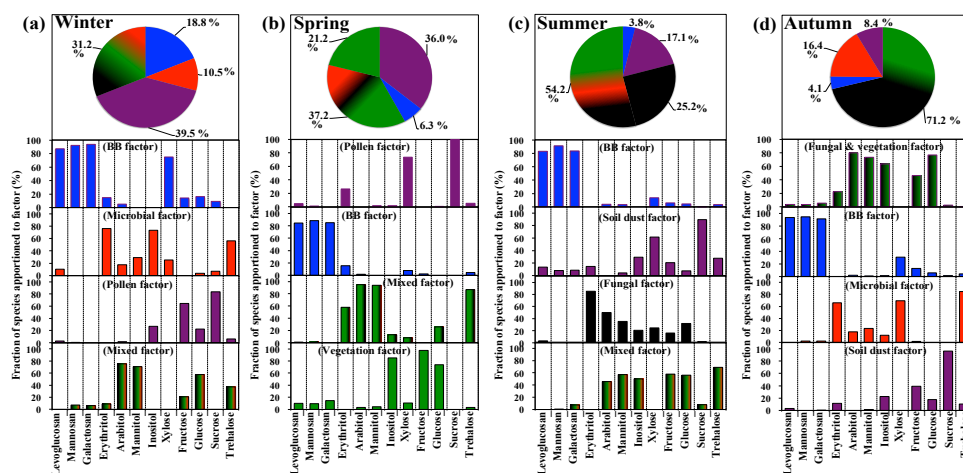
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Figure 6.



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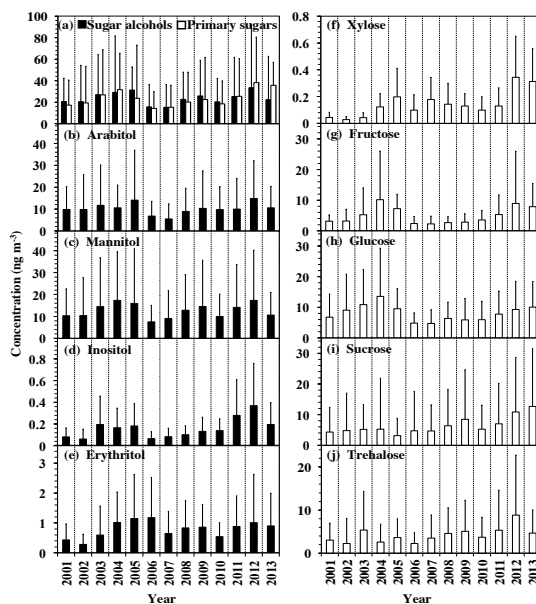
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1171 **Figure 7.**



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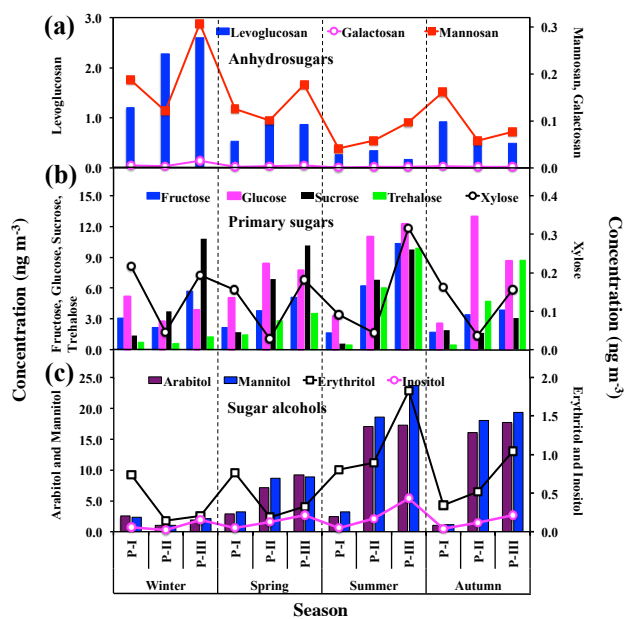
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Figure 8.



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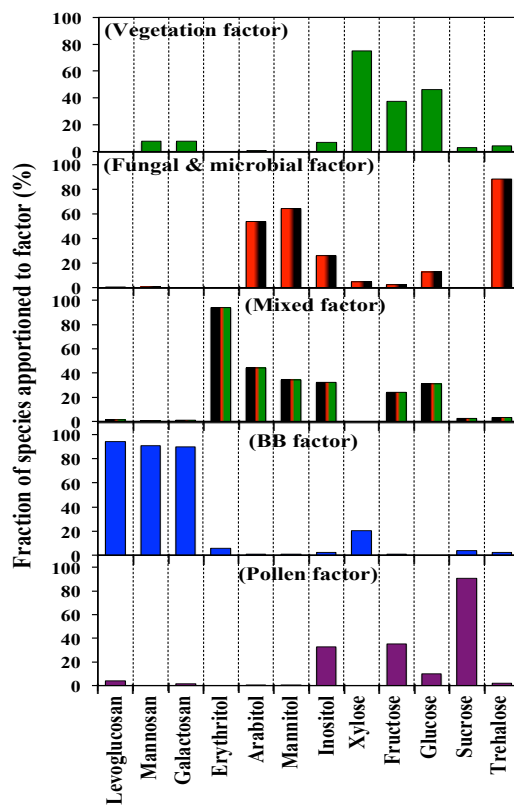
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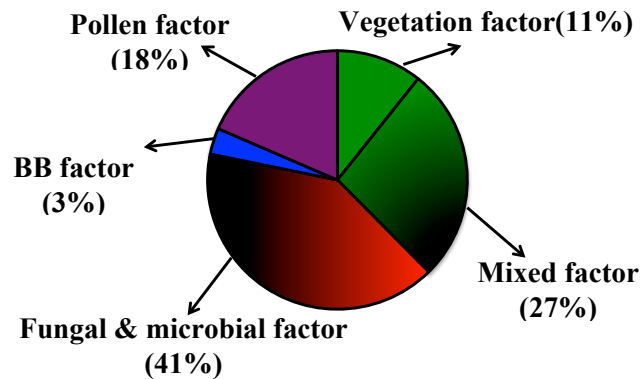
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**Figure 10.**



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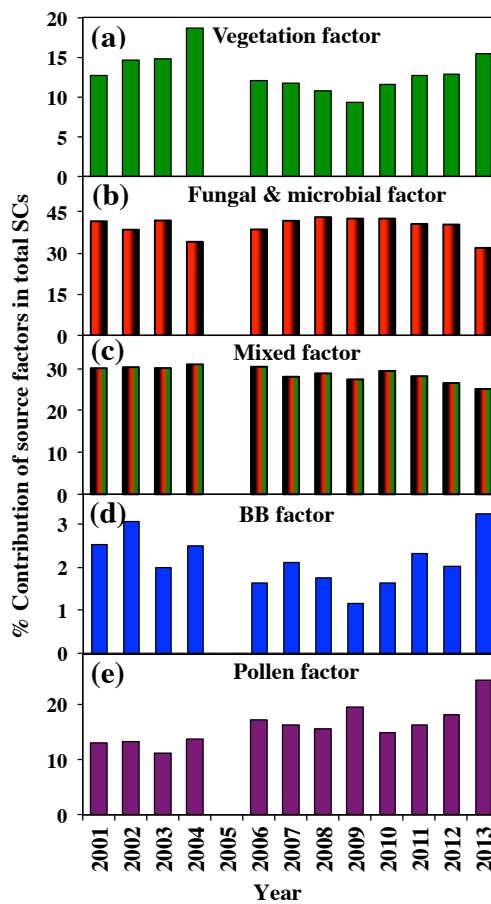
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1251 **Figure 11.**



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**Table 1.** Seasonal concentrations ( $\text{ng m}^{-3}$ ) of sugar compounds (SCs) in the aerosol samples collected at ChichiJima Island in the western North Pacific during 2001–2013.

Sugars	Winter <sup>a</sup> (n=139)			Spring <sup>b</sup> (n=155)			Summer <sup>c</sup> (n=146)			Autumn <sup>d</sup> (n=150)			2001-2013 (n=590)		
	Range	Mean±SD	Med <sup>e</sup>	Range	Mean±SD	Med <sup>e</sup>	Range	Mean±SD	Med <sup>e</sup>	Range	Mean±SD	Med <sup>e</sup>	Range	Mean±SD	Med <sup>e</sup>
<b>Primary sugars</b>															
Xylose	0.05–0.89	0.14±0.15	0.12	0.001–0.56	0.12±0.11	0.08	0.002–1.35	0.18±0.26	0.08	0.004–0.89	0.11±0.14	0.07	0.001–1.35	0.14±0.18	0.08
Fructose	0.17–115	3.36±10.2	1.38	0.03–70.4	4.51±9.21	1.90	0.24–54.0	7.25±7.63	5.23	0.22–13.0	3.70±2.68	2.82	0.03–115	4.69±8.04	2.53
Glucose	0.27–23.3	3.11±3.53	2.13	0.05–62.6	7.68±10.3	4.36	0.23–64.3	11.0±9.02	7.99	0.72–55.5	9.25±8.63	5.77	0.05–64.3	7.79±8.80	4.73
Sucrose	0.02–73.4	6.60±13.1	2.08	0.005–100	8.80±18.0	1.30	0.003–66.0	7.31±11.5	2.58	0.002–31.1	2.76±4.35	1.14	0.002–100	6.43±12.9	1.71
Trehalose	0.03–10.5	1.03±1.26	0.72	0.006–47.2	2.93±6.08	1.23	0.03–70.2	7.06±8.49	4.46	0.04–52.5	6.09±8.81	2.46	0.01–70.2	4.30±7.28	1.61
$\Sigma$ Primary sugars	<b>0.49–223</b>	<b>14.2±28.2</b>		<b>0.09±281</b>	<b>24.2±43.8</b>		<b>0.51–256</b>	<b>32.8±36.9</b>		<b>0.99–153</b>	<b>22.0±24.6</b>		<b>0.28–176</b>	<b>23.3±25.7</b>	
<b>Sugar alcohols</b>															
Erythritol	0.03–1.17	0.23±0.18	0.18	0.008–2.25	0.31±0.28	0.23	0.07–5.70	1.55±1.21	1.18	0.05–8.32	0.99±1.16	0.56	0.01–8.32	0.77±1.01	0.37
Arabitol	0.12–21.2	1.73±2.60	0.96	0.04–70.8	7.13±9.50	4.44	0.24–64.5	15.1±12.9	12.0	0.57–106	15.8±18.3	8.83	0.04–106	9.99±13.6	4.97
Mannitol	0.10–23.9	1.89±2.81	1.11	0.16–114	7.95±13.8	4.07	0.25–104	21.7±19.7	16.9	0.59–118	18.2±19.9	9.15	0.10–118	12.5±17.5	5.54
Inositol	0.01–1.81	0.07±0.20	0.03	0.008–1.51	0.13±0.22	0.06	0.01–1.29	0.26±0.25	0.17	0.01–0.93	0.16±0.15	0.10	0.01–1.81	0.16±0.22	0.08
$\Sigma$ Sugar alcohols	<b>0.26–48.2</b>	<b>3.93±5.79</b>		<b>0.22–188</b>	<b>15.5±23.8</b>		<b>0.56–175</b>	<b>38.6±34.1</b>		<b>1.22–234</b>	<b>35.1±39.5</b>		<b>0.37–231</b>	<b>23.4±30.8</b>	
$\Sigma$ SCs	<b>0.75–272</b>	<b>18.2±34.0</b>	<b>10.6</b>	<b>0.31–469</b>	<b>39.8±67.6</b>	<b>18.2</b>	<b>1.07–431</b>	<b>71.5±70.9</b>	<b>50.7</b>	<b>2.21–387</b>	<b>57.0±64.2</b>	<b>31.1</b>	<b>1.23–339</b>	<b>46.7±49.5</b>	<b>30.9</b>

<sup>a</sup> Winter (December–February), <sup>b</sup> Spring (March–May), <sup>c</sup> Summer (June–August) and <sup>d</sup> Autumn (September–November); Med. = Median

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**Table 2.** Pearson correlation coefficients ( $r$ ) for the dataset of sugars in Chichijima aerosols during 2001–2013 ( $n = 590$ ).

	Levogluco- sac	Mannosan <sup>a</sup>	Galactosan <sup>a</sup>	Erythritol	Arabitol	Mannitol	Inositol	Xylose	Fructose	Glucose	Sucrose	Trehalose
Levogluco- sac	1.00											
Mannosan	0.79	1.00										
Galactosan	0.55	0.58	1.00									
Erythritol	-0.18	-0.12	-0.16	1.00								
Arabitol	-0.16	-0.09	-0.16	0.48	1.00							
Mannitol	-0.18	-0.12	-0.17	0.49	0.88	1.00						
Inositol	-0.06	-0.02	0.10	0.35	0.49	0.58	1.00					
Xylose	0.20	0.32	0.34	0.15	0.18	0.23	0.42	1.00				
Fructose	0.02	0.08	0.26	0.17	0.16	0.28	0.57	0.31	1.00			
Glucose	-0.11	0.00	-0.06	0.32	0.63	0.72	0.53	0.19	0.57	1.00		
Sucrose	0.05	0.07	0.18	-0.02	-0.06	0.01	0.40	0.26	0.30	0.14	1.00	
Trehalose	-0.10	-0.05	-0.10	0.33	0.73	0.80	0.55	0.33	0.22	0.54	0.13	1.00

<sup>a</sup> data from Verma et al. (2015)

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**Table 3.** Comparisons of seasonal concentrations ( $\text{ng m}^{-3}$ ) of primary sugars and relative contributions (%) of sugar compounds (SCs) in total SCs in Chichijima aerosols among 1990-1993<sup>a</sup>, 2001-2003 and 2010-2013.

Season	Anhydrosugars			Sugar alcohols			Primary sugars			Total Sugars		
	1990-93 <sup>a</sup>	2001-03 <sup>b</sup>	2010-13 <sup>b</sup>	1990-93 <sup>a</sup>	2001-03	2010-13	1990-93 <sup>a</sup>	2001-03	2010-13	1990-93 <sup>a</sup>	2001-03	2010-13
Winter	1.44	2.44	3.05	5.64	2.27	4.37	10.6	9.34	21.8	17.7	14.0	29.2
%	11.4	23.7	16.6	31.6	16.5	19.8	57.0	59.9	63.6			
Spring	0.68	1.12	1.10	6.91	16.2	18.6	10.5	21.3	26.7	18.1	38.6	46.4
%	5.08	5.17	4.04	39.2	39.7	42.0	55.7	55.1	53.9			
Summer	0.31	0.43	0.29	6.50	36.7	43.1	6.07	30.0	42.5	12.9	67.1	86.0
%	2.92	0.91	0.43	55.6	51.3	51.1	41.4	47.8	48.5			
Autumn	1.12	0.65	0.60	2.57	34.7	38.3	6.79	22.6	24.5	10.5	57.9	63.4
%	12.9	3.45	1.73	29.2	50.4	57.2	57.9	46.2	41.1			

<sup>a</sup> data from Chen et al. (2013), <sup>b</sup> data from Verma et al. (2015).

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