





## 34 **Abstract**

35 Decreases in surface emissions of nitrogen oxides ( $\text{NO}_x = \text{NO} + \text{NO}_2$ ) in North America have led  
36 to substantial improvements in air-quality over the last several decades. Here we show that satellite  
37 observations of tropospheric nitrogen dioxide ( $\text{NO}_2$ ) columns over the contiguous United States  
38 (US) do not decrease after about 2009, while surface  $\text{NO}_2$  concentrations continue to decline  
39 through to the present. This divergence, if it continues, could have a substantial impact on surface  
40 air quality due to mixing of free-tropospheric air into the boundary layer. Our results show only  
41 limited contributions from local effects such as fossil fuel emissions, lightning, or instrument  
42 artifacts, but we do find a possible relationship of  $\text{NO}_2$  changes to decadal climate variability. Our  
43 analysis demonstrates that the intensity of transpacific transport is stronger in El Niño years and  
44 weaker in La Niña years, and consequently, that decadal-scale climate variability impacts the  
45 contribution of Asian emissions on North American atmospheric composition. Because of the short  
46 lifetime, it is usually believed that the direct contribution of long-range transport to tropospheric  
47  $\text{NO}_x$  distribution is limited. If our hypothesis about transported Asian emissions is correct, then  
48 this observed divergence between satellite and surface  $\text{NO}_x$  could indicate mechanisms that allow  
49 for either  $\text{NO}_x$  or its reservoir species to have a larger than expected effect on North American  
50 tropospheric composition. These results therefore suggest more aircraft and satellite studies to  
51 determine the possible missing processes in our understanding of the long-range transport of  
52 tropospheric  $\text{NO}_x$ .

53

## 54 **1. Introduction**

55 Nitrogen oxides play a complex role in tropospheric chemistry and have a strong influence  
56 on air quality as precursors in the formation of ozone ( $\text{O}_3$ ) and secondary aerosols. Tropospheric



57 NO<sub>x</sub> is produced through anthropogenic combustion, biomass burning, soil (Jaegle et al., 2005),  
58 and lightning emissions (Schumann and Huntrieser, 2007), and is mainly removed by the  
59 formation of nitric acid (HNO<sub>3</sub>). Most NO<sub>x</sub> is emitted as nitric oxide (NO), however, it is most  
60 appropriate to consider the budget of the NO<sub>x</sub> as a whole, because of the rapid cycling between  
61 NO and NO<sub>2</sub> (~ 1 min). Tropospheric NO<sub>x</sub> has short lifetime, a few hours except in extratropical  
62 winter when it increases to 1-2 days (Martin et al. 2003).

63 Because of the short lifetime, state of the art chemistry/climate models suggest that the  
64 direct contribution of long-range transport to tropospheric NO<sub>x</sub> distribution is limited (e.g. Zhang  
65 et al. 2008). However, NO<sub>x</sub> can also be transported far away from the sources via the formation of  
66 long-lived reservoir species, such as peroxyacetyl nitrate (PAN, e.g., Fischer et al., 2014; Jiang et  
67 al. 2016a). Models have large uncertainties in PAN abundance (Fischer et al., 2014) and there are  
68 also potentially other missing processes in the chemical transport models used to diagnose NO<sub>x</sub>  
69 lifetime and transport. For example, a recent discovery about the rapid cycling of reactive nitrogen  
70 in the marine boundary layer (Ye et al. 2016) demonstrates processes that are not represented in  
71 modeled NO<sub>x</sub> transport, and that may help explain existing discrepancies in reactive nitrogen  
72 partitioning between models and observations.

73 Due to its critical influence in the troposphere, there are multiple space-based  
74 measurements for tropospheric NO<sub>2</sub> that are available from satellites that were launched in the past  
75 two decades. These instruments typically measure backscattered solar radiation from which the  
76 vertically integrated column abundance of NO<sub>2</sub> is retrieved. The assumption of weak long-range  
77 transport allows relatively simple applications of the space-based NO<sub>2</sub> column data to study NO<sub>x</sub>  
78 sources. For example, recent studies (e.g. Reuter et al. 2014; Itahashi et al. 2014; Duncan et al.  
79 2016; Krotkov et al. 2016) assessed the trends of surface NO<sub>x</sub> emissions by assuming a strong



80 correlation between tropospheric NO<sub>2</sub> columns with local emissions. The tropospheric NO<sub>2</sub>  
81 column data are also widely used in inverse modeling analyses to estimate NO<sub>x</sub> emissions by either  
82 scaling the surface NO<sub>x</sub> emissions with the corresponding ratio of observed over modeled  
83 tropospheric NO<sub>2</sub> column (e.g. Lamsal et al. 2011; Mijling et al. 2012; Gu et al. 2014) or through  
84 data assimilation techniques with short localization length scales (e.g. Miyazaki et al. 2017).

85 Since 1990, US regulations have required significant NO<sub>x</sub> emission reductions over many  
86 regions (US Environmental Protection Agency, 2010). The trend of decreasing local US NO<sub>x</sub>  
87 emissions has been confirmed by several studies (e.g. Lamsal et al. 2015; Tong et al. 2015; Kharol  
88 et al. 2015; Duncan et al. 2016; Krotkov et al. 2016). In contrast to the decreasing local NO<sub>x</sub>  
89 emissions, recent studies (e.g. Cooper et al. 2010; Verstraeten et al. 2015) have indicated an  
90 increase in free tropospheric O<sub>3</sub> over western North America over the past decade. The discrepancy  
91 between variations of local NO<sub>x</sub> emissions and free tropospheric O<sub>3</sub> suggests possible influences  
92 from non-local sources, and consequently, provides motivation to re-evaluate the contribution of  
93 long-range transport to the free tropospheric NO<sub>x</sub> distribution.

94 In this work, we investigate the variation of US tropospheric NO<sub>2</sub> in the past decade to  
95 assess the contribution of non-local sources. We will particularly explore the possible answers for  
96 the following questions: why there is good agreement between tropospheric NO<sub>2</sub> column and  
97 surface measurements over the period of 2005-2008? What is the reason for the appearance of the  
98 large and growing divergence at around 2009? What is the impact of the decreasing Chinese NO<sub>x</sub>  
99 emissions since 2013 (Liu et al. 2016) on North America? To evaluate these critical questions,  
100 multiple data sets and model are used in this work, including remotely sensed NO<sub>2</sub> column  
101 measurements from Ozone Monitoring Instrument (OMI, NASA and DOMINO products), in-situ  
102 surface NO<sub>2</sub> measurements from the Environmental Protection Agency (EPA) Air Quality System



103 (AQS) network and the Environment Canada National Air Pollution Surveillance Program (NAPS)  
104 network, flash rate density data from Lightning Imaging Sensor (LIS), and the GEOS-Chem  
105 chemical transport model.

106 This paper is organized as follows: in Section 2 we describe the observations and model  
107 used in this work. In Section 3 we demonstrate the divergence between the OMI NO<sub>2</sub> column  
108 retrievals and surface measurements over the period of 2005-2015 and focus on the evaluation of  
109 contributions from various hypotheses that could explain the divergence. Our conclusions follow  
110 in Section 4.

## 111 **2. Observations and Models**

### 112 **2.1 Tropospheric NO<sub>2</sub> column from OMI**

113 The OMI instrument was launched on NASA's Aura spacecraft. The sensor has a spatial  
114 resolution of 13 km x 24 km. OMI provides daily global coverage with measurements of both  
115 direct and atmosphere-backscattered sunlight in the ultraviolet-visible range from 270 to 500 nm;  
116 the spectral range 405-465 nm is used to retrieve tropospheric NO<sub>2</sub> columns. Two versions of the  
117 OMI retrievals (level 2) are used in this work: the NASA (version 3, Krotkov and Veefkind 2006;  
118 Bucsela et al. 2013) and DOMINO (version 2, Boersma et al. 2011) retrievals. There are significant  
119 differences in the retrieval algorithms of the two products. For example, the a priori NO<sub>2</sub> profiles  
120 of the NASA product is based on data from the Global Modeling Initiative (GMI) model with  
121 yearly varying emissions, whereas the a priori NO<sub>2</sub> profiles of the DOMINO product is from the  
122 Tracer Model 4 (TM4) without interannual variations in emissions. In addition, for the NASA  
123 product, the stratospheric contribution to the tropospheric column is estimated from the GMI  
124 model simulation. In contrast, for the DOMINO product, the stratospheric contribution is based on  
125 the assimilation of OMI data into the TM4 model.



126 Starting in 2007, anomalies were found in OMI data and diagnosed as attenuated measured  
127 radiances in certain cross-track positions. This instrument degradation has been referred to as the  
128 “row anomaly”. In order to ensure the quality and stability of the data, the following filters are  
129 applied in our analysis for both OMI products (NASA and DOMINO):

130 1) Tropospheric Column Flag = 0

131 2) Surface Albedo < 0.3

132 3) Cloud Radiance Fraction < 0.5

133 4) No edge data (rows 1-5, 56-60)

134 5) No row anomaly data (rows 27-55 for the whole period 2005-2015)

135 After the application of the filters, the number of measurements over the US is about 185,000 per  
136 month in 2010. Thus, we expect the uncertainties in the monthly/annual mean NO<sub>2</sub> columns due  
137 to random errors are small. The discrepancy between the two OMI products (see Figures 1a-b) is  
138 mainly caused by the two different retrieval algorithms.

## 139 **2.2 AQS and NAPS surface in-situ NO<sub>2</sub> concentration**

140 We use daily-averaged in-situ surface NO<sub>2</sub> measurements from the EPA AQS network,  
141 and the Environment Canada NAPS network. The AQS/NAPS networks collect ambient air  
142 pollution data from monitoring stations located in urban, suburban, and rural areas. In the analysis  
143 here, the daily data are averaged to obtain monthly mean concentration at each station.

## 144 **2.3 Flash rate density from Lightning Imaging Sensor (LIS)**

145 LIS is a component of the NASA Tropical Rain Measuring Mission (TRMM). It measures  
146 total optical pulses from cloud-to-ground and intracloud lightning flashes during both day and  
147 night with global coverage (42.5°S-42.5°N) in the period 1995-2014. Monthly flash rate density  
148 (flash/km<sup>2</sup>) with 2.5°x2.5° resolution is used in this work (Cecil et al. 2006).



#### 149 **2.4 NOAA Niño 3.4 index**

150 The Niño 3.4 index comprises sea surface temperature averaged across the region (5°S–  
151 5°N, 170°W–120°W), and their monthly anomalies relative to the 1982–2015 means to constitute  
152 the indices. Years with positive values (>0.5) are considered as El Niño, whereas years with  
153 negative values (< -0.5) are considered as La Niña.

#### 154 **2.5 Passive tracer simulation using GEOS-Chem model**

155 The GEOS-Chem global chemical transport model (CTM) [[www.geos-chem.org](http://www.geos-chem.org)] is driven  
156 by assimilated meteorological fields (MERRA) from the NASA Goddard Earth Observing System  
157 at the Global Modeling and data Assimilation Office. We use version v9-01-03 of GEOS-Chem at  
158 a horizontal resolution of 4°x5°. Bertram et al. (2013) indicated the dominant role of long-lived  
159 reservoir species in the transpacific transport of reactive nitrogen using aircraft measurements from  
160 the INTEX-B campaign. Although the lifetime of tropospheric NO<sub>x</sub> is short, the lifetime of long-  
161 lived reservoir species is much longer, for example, the lifetime of free tropospheric PAN is about  
162 1 month. In order to assess the effects of physical transport processes on the long-range transport  
163 of reactive nitrogen, we performed a “passive” tracer simulation, with a constant and uniform  
164 timescale for loss of 15 days (i.e. 360 hours) over the period of 2005–2015 following the approach  
165 of Jiang et al. (2016b). The global a priori surface NO<sub>x</sub> emissions (anthropogenic, biomass burning  
166 and soil emissions) are fixed at 2005 level. For each time step (one hour), the tropospheric NO<sub>2</sub> is  
167 calculated by:  $NO_2^t = NO_2^{t-1} e^{-1/360}$ . The lightning NO<sub>x</sub> emissions are not included in the  
168 simulation. The 15-day lifetime was selected to provide an approximation for the variation of free  
169 tropospheric NO<sub>x</sub> via the formation and transport of long-lived reservoir species, due to changes  
170 in meteorology. Although actual lifetimes of long-lived reservoir species will vary, we found that  
171 15-days was a reasonable compromise to understand the influence of decadal-scale variability on



172 long-range transport patterns.

### 173 **3. Results and Discussion**

174           Figures 1a-b show the variations of mean tropospheric NO<sub>2</sub> columns from OMI (NASA  
175 and DOMINO products) over the US and east China, respectively. Although there is a significant  
176 bias in the magnitude of tropospheric NO<sub>2</sub> column between two OMI products, indicating the  
177 influence of different retrieval algorithms, this bias should not affect the trend analysis, as  
178 demonstrated by the consistent interannual variations between the two data products. Figure 1c  
179 shows percent changes, relative to 2009, of the annual mean tropospheric NO<sub>2</sub> columns over the  
180 US, and of the total US NO<sub>x</sub> emissions (anthropogenic + biomass burning) from the US EPA  
181 (<https://www.epa.gov/air-emissions-inventories/air-pollutant-emissions-trends-data>). There is  
182 good agreement between the changes in the OMI retrievals and the emissions estimates in the  
183 2005-2009 period: the annual slope of the EPA's estimates is  $-6.4\% \pm 0.03\%$  (slope of linear  
184 regression  $\pm$  uncertainty of slope), and the annual slopes of the two sets of OMI retrievals are -  
185  $6.8\% \pm 1.1\%$  (NASA) and  $-8.0\% \pm 0.8\%$  (DOMINO). Conversely, we find a large, growing  
186 separation in the 2009-2015 period: the annual slope of the EPA's estimates is  $-4.6\% \pm 0.03\%$ ,  
187 whereas the annual slopes of OMI retrievals are  $-0.5\% \pm 0.6\%$  (NASA) and  $1.6\% \pm 1.1\%$   
188 (DOMINO). Figures 1d-e show the percent changes in the seasonal mean tropospheric NO<sub>2</sub>  
189 columns from OMI retrievals, and in the EPA's estimates (annual mean). The divergence between  
190 the seasonal NO<sub>2</sub> columns and the emissions is similar to that shown in Figure 1c, suggesting there  
191 is no obvious seasonal dependence.

192           Our intention here is to understand the possible reasons for the divergence between  
193 observed changes in NO<sub>2</sub> vs. changes expected from NO<sub>x</sub> emissions. Figure 2 depicts the potential  
194 hypotheses that could explain the divergence:



- 195           • Hypothesis 1 (H1): Increasing local NO<sub>x</sub> emissions missing from the EPA inventories
- 196           • Hypothesis 2 (H2): Time dependent OMI retrieval errors
- 197           • Hypothesis 3 (H3): Non-local sources

### 198 **3.1 Increasing local NO<sub>x</sub> emissions (H1)**

199           Figure 3a shows the differences of mean surface NO<sub>2</sub> concentrations, as measured by the

200 AQS and NAPS network, from 2009-2010 to 2014-2015. These time periods were chosen to

201 determine changes in surface NO<sub>2</sub> concentrations over the period of 2009-2015 with sufficient

202 statistics. Figure 3b shows the same variations averaged with 4°x5° resolution (i.e., at the GEOS-

203 Chem grid points). The surface stations demonstrate dramatic decreases of surface NO<sub>2</sub>

204 concentrations in the period of 2009-2015. The consistent decreasing trends in the surface NO<sub>2</sub>

205 concentrations and the EPA's emission estimates during the period of 2009-2015 suggests that the

206 divergence between the OMI retrievals and the EPA's emission estimates is not caused by US

207 local emissions. We note that our analysis based on surface in-situ measurements may not provide

208 sufficient representation for emissions from oil and gas exploration and production. However,

209 though potentially important locally, these activities only contribute about 5% to total US NO<sub>x</sub>

210 emissions based on EPA's estimates. Therefore, we do not expect significant contributions to the

211 overall changes in tropospheric NO<sub>2</sub> from these sources.

212           Similarly, we expect limited contributions from other sources which are not included in

213 EPA's inventory. The contribution from aircraft NO<sub>x</sub> is only about 3% of total US NO<sub>x</sub> emissions

214 (Skowron et al. 2014). Soil NO<sub>x</sub> emissions account for up to 40% of the tropospheric NO<sub>2</sub> columns

215 in summer over US rural areas (Hudman et al. 2010), but only contribute a few percent of the US

216 annual mean tropospheric NO<sub>2</sub> column. Similarly, NO<sub>x</sub> production by lightning is stronger in

217 summer, with an estimated annual contribution of 15% to the total emissions. Because the



218 discrepancies between the OMI retrievals and the EPA's emission estimates lack a clear seasonal  
219 dependence (see Figures 1d-e), we expect negligible contributions from soil and lightning NO<sub>x</sub>  
220 emissions to the enhanced tropospheric NO<sub>2</sub> in the period of 2009-2015. Furthermore, Figure 4  
221 indicates that the flash rate density over North America from LIS is uncorrelated with the observed  
222 NO<sub>2</sub> variation.

### 223 **3.2 Time dependent OMI retrieval errors (H2)**

224 The quality of the OMI retrievals has been evaluated with surface in-situ measurements.  
225 Lamsal et al. (2015) reported that the correlation between OMI NO<sub>2</sub> tropospheric columns (NASA)  
226 and the AQS surface in-situ NO<sub>2</sub> measurements was 0.68 for the period 2005-2010. Hoek et al.  
227 (2015) indicated that the correlation between the OMI NO<sub>2</sub> tropospheric columns (DOMINO) and  
228 surface in-situ measurements in the Netherlands was 0.74 at 2007. The stability of our analysis  
229 based on OMI retrievals (NASA and DOMINO) is ensured by the strict quality filters; these ensure  
230 that changes in OMI sampling due to detector problems (e.g. row anomaly) do not affect our  
231 conclusions.

232 Figure 5 shows the annual slopes of tropospheric NO<sub>2</sub> columns from OMI (NASA and  
233 DOMINO) over the period of 2005-2015. Both OMI products (NASA and DOMINO), with  
234 various a priori models (GMI and TM4) and algorithms, show consistent variations over the  
235 northern Pacific Ocean and the western US: insignificant changes in the period 2005-2008 (Figure  
236 5a-b), positive changes in the period 2009-2012 (Figure 5c-d) and insignificant changes in the  
237 period 2013-2015 (Figure 5e-f). Using the Berkeley High-Resolution (BEHR) NO<sub>2</sub> product for  
238 OMI, Russell et al. (2012) obtained similar positive change over the western US with the Weather  
239 Research Forecasting Chemistry model (WRF-Chem) as an a priori model in the period 2005-  
240 2011. The consistency among the various data products suggests that the variations in the retrieved



241 OMI NO<sub>2</sub> over the period of 2005-2015 are not caused by systematic biases.

242 While our adherence to published data quality filters should ensure the observed OMI  
243 based NO<sub>2</sub> values are robust, we note as a caveat that we cannot unequivocally confirm these  
244 changes with independent data due to a lack of either total column or free-tropospheric NO<sub>2</sub>  
245 measurements over the observed region. For example, Mt. Bachelor Observatory (MBO) station  
246 has free tropospheric NO<sub>2</sub> observations in the period of 2005-2009 and the CARIBIC aircraft  
247 measurements only provide free tropospheric NO<sub>2</sub> observations over the US since 2014. Using  
248 remotely sensed NO<sub>2</sub> measurements from the Global Ozone Monitoring Experiment-2 (GOME-  
249 2), Miyazaki et al. (2017) found a slight increase of tropospheric NO<sub>2</sub> columns over the US in the  
250 period of 2009-2012, consistent with our result. However, a significant sudden decrease of  
251 retrieved tropospheric NO<sub>2</sub> columns from GOME-2 has been observed since July 2013, associated  
252 with the change in the measurement mode (<http://projects.knmi.nl/atcom/news.php?id=44>).

### 253 **3.3 Non-local sources (H3)**

254 We have demonstrated that hypotheses H1 and H2 are not likely the dominant factors,  
255 which leaves hypothesis H3 (non-local sources) as possible important contributors. Figures 5g-h  
256 show the annual slope of tropospheric NO<sub>2</sub> columns (percent base) from OMI (NASA and  
257 DOMINO) over the period of 2009-2015. Our analysis demonstrates a significant positive change  
258 over the northern Pacific Ocean during this time period, and an insignificant but positive change  
259 over the western US, suggesting possible contributions from transpacific transport to tropospheric  
260 NO<sub>2</sub> over the western US. Over the period of 2005-2008, the lack of change in tropospheric NO<sub>2</sub>  
261 columns over the northern Pacific Ocean (Figure 5a-b) indicates the dominant role of local sources  
262 to the decrease of US tropospheric NO<sub>2</sub> in this period. Conversely, the increase in tropospheric  
263 NO<sub>2</sub> columns over northern Pacific Ocean over the period of 2009-2012 (Figure 5c-d) is consistent



264 with the appearance of a discrepancy between the OMI retrievals and EPA's emission estimates  
265 (Figure 1c). Accompanying with the observed decrease of Chinese NO<sub>2</sub> emissions (Figure 1b), no  
266 significant change is observed over the northern Pacific Ocean over the period of 2013-2015  
267 (Figure 5e-f).

268 Decadal climate variability has non-negligible influences on tropospheric compositions by  
269 affecting the physical and chemical processes. For example, Lin et al. (2014) indicated that  
270 transpacific transport of O<sub>3</sub> is modulated by decadal variability of El Niño–Southern Oscillation  
271 (ENSO). El Niño is defined as the appearance of anomalously warm water off northern Peru and  
272 Ecuador in December. The atmospheric component tied to El Niño is called the Southern  
273 Oscillation (Trenberth 1997). To the best of our understanding, ENSO is the dominant climate  
274 phenomenon linked to extreme weather conditions globally (Cai et al. 2015), and it also exerts a  
275 major influence on the interannual variability of O<sub>3</sub> in the troposphere (Doherty et al., 2006).  
276 Following Jiang et al. (2016b), we conducted an analysis using an idealized passive tracer to assess  
277 the possible influences of transport patterns. We performed a GEOS-Chem model simulation for  
278 tropospheric NO<sub>2</sub> over the period of 2005-2015 with an NO<sub>2</sub>-like tracer with a constant 15-day  
279 lifetime and fixed (2005 level) surface NO<sub>x</sub> emissions. The passive tracer simulation with constant  
280 lifetime avoids the possible influences from uncertainties in the modeled nonlinear NO<sub>x</sub> chemistry,  
281 particularly, the conversion between NO<sub>x</sub> and its longer-lived reservoir species.

282 Figures 6a-c show that even with emissions held constant, interannual variations in  
283 transport produce differences in NO<sub>2</sub> (or the passive tracer) columns over the eastern Pacific.  
284 Based on the passive tracer simulation, transpacific transport decreased over the period of 2005-  
285 2008 (Figure 5a). During this four-year period, declining transport efficiency appears to have offset  
286 the increase of Asian emissions, resulting in insignificant changes of tropospheric NO<sub>2</sub> columns



287 over the northern Pacific Ocean (Figures 5a-b), and consequently, good agreement between the  
288 OMI retrievals and the EPA's emission estimates (Figure 1c). The efficiency of transpacific  
289 transport is more stable over the period of 2009-2012 (Figure 6b), which allows stronger  
290 transpacific transport of rising Asian emissions. It leads to positive changes of tropospheric NO<sub>2</sub>  
291 columns over the northern Pacific Ocean (Figure 5c-d), and consequent growing discrepancy  
292 between the OMI retrievals and the EPA's emission estimates (Figure 1c). Increasing efficiency  
293 in transpacific transport over the period of 2013-2015 (Figure 6c) counteracts the decrease of  
294 Chinese NO<sub>2</sub> (Figure 1b), again resulting in no change in the tropospheric NO<sub>2</sub> column over  
295 northern Pacific Ocean (Figure 5e-f) and relatively flat changes in US tropospheric NO<sub>2</sub> column.  
296 Figure 6d shows the comparison between regional mean of passive tracer columns over the  
297 northern Pacific Ocean with the NOAA Niño 3.4 index. There is strong correlation between  
298 transpacific transport and ENSO: the transpacific transport is stronger in El Niño years and weaker  
299 in La Niña years, demonstrating strong influence of decadal climate variability on the transpacific  
300 transport.

301 Brown-Steiner et al. (2011) showed that Asian O<sub>3</sub> is transported across Pacific Ocean  
302 primarily in the lower/middle troposphere in winter and spring, and primarily in the middle/upper  
303 troposphere in summer and fall, but that differences in tropospheric column remain small for  
304 different seasons. Figure 7 shows the seasonal mean tropospheric columns for the NO<sub>2</sub>-like tracer  
305 between 2005-2015. Our analysis indicates the transpacific transport, with constant 15-day lifetime,  
306 is strongest in summer and spring. However, the lifetime of PAN approximately doubles for every  
307 4°C decrease in temperature. Consequently, due to the temperature effect associated with the  
308 change in transport pathways from the lower troposphere (in spring) to the upper troposphere (in  
309 fall), transport in fall, relative to that in spring, should be greater than that shown in Figure 7c.



310 Similarly, the actual transpacific transport of NO<sub>2</sub> in winter, relative to that in spring, should be  
311 greater than shown in Figure 7d because of the temperature differences between winter and spring.  
312 Thus, in agreement with Brown-Steiner et al. (2011), our analysis would suggest weak seasonality  
313 of transpacific transport of reactive nitrogen, consistent with the observed changes in OMI NO<sub>2</sub>  
314 over the US (Figures 1d-e). It should be noticed that our conclusion about weak seasonality of  
315 transpacific transport is different from previous studies (e.g. Liang et al. 2004) based on carbon  
316 monoxide (CO) simulations. The discrepancy in the seasonality is associated with the strong  
317 springtime biomass burning CO emissions in southeast and boreal Asia, whereas the contribution  
318 of biomass burning to tropospheric NO<sub>x</sub> is relatively small.

#### 319 **4. Conclusions**

320 In this work, we investigated the variation of US tropospheric NO<sub>2</sub> in the past decade, to  
321 evaluate the contribution of non-local sources to the tropospheric NO<sub>x</sub> budget. We demonstrated  
322 significant divergence between the time variation in tropospheric NO<sub>2</sub> columns from the OMI  
323 retrievals and the EPA's NO<sub>x</sub> emission estimates. Our analysis suggests limited contributions from  
324 local effects such as fossil fuel emissions, lightning, or instrument artifacts, and indicates possible  
325 important contributions from long-range transport of Asian emissions that are modulated by ENSO.  
326 Passive tracer simulation with fixed emissions demonstrates that the intensity of transpacific  
327 transport is stronger in El Niño years and weaker in La Niña years. The unexpected important  
328 contributions from long-range transport contradict assumptions of weak long-range transport for  
329 NO<sub>x</sub>, suggesting potential underestimation of transported reactive nitrogen in the state of the art  
330 models.

331 In related studies, long-term free tropospheric O<sub>3</sub> observations over Europe demonstrated  
332 a significant increase in the past three decades (e.g. Logan et al. 2012; Parrish et al. 2012), whereas



333 state of the art models cannot reproduce this variation (Parrish et al. 2014). Enhanced long-range  
334 transport of NO<sub>x</sub> could potentially reconcile the large discrepancy between modeled and observed  
335 free tropospheric O<sub>3</sub> over Europe. Because of uncertain processes in the long-range transport of  
336 reactive nitrogen (Ye et al. 2016) and the dominant role of long lifetime reservoirs (Bertram et al.  
337 2013), we are not able to quantify the different contributions to the observed tropospheric NO<sub>x</sub> in  
338 this study. This quantification will require comprehensive observations and modeling efforts to  
339 understand the formation and transport of long lifetime reservoirs of reactive nitrogen.

340

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351

352 **Reference:**

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495 **Tables and Figures**

496 **Figure 1.** (a-b): monthly mean (dash lines) and annual mean (solid lines) tropospheric NO<sub>2</sub> column  
497 over contiguous United States and East China from OMI (NASA and DOMINO) products; (c-e):  
498 percent changes of annual mean (c) and seasonal mean (d-e) tropospheric NO<sub>2</sub> column over the  
499 US from the OMI and EPA's emission estimates, normalized at 2009.

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501 **Figure 2.** Schematic figure showing the sources of tropospheric NO<sub>x</sub>.

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503 **Figure 3.** (a) difference of mean NO<sub>2</sub> concentrations of surface in-situ measurements (AQS and  
504 NAPS stations) from 2009-2010 to 2014-2015; Blue (red) means decrease (increase) of NO<sub>2</sub>  
505 concentrations. (b) same as panel a, but averaged with 4°x5° resolution.

506

507 **Figure 4.** Flash rate density (1x10<sup>-3</sup> flash/km<sup>2</sup>/month) over North America (15°N-42.5°N, 130°W-  
508 60°W) from Lightning Imaging Sensor (LIS).

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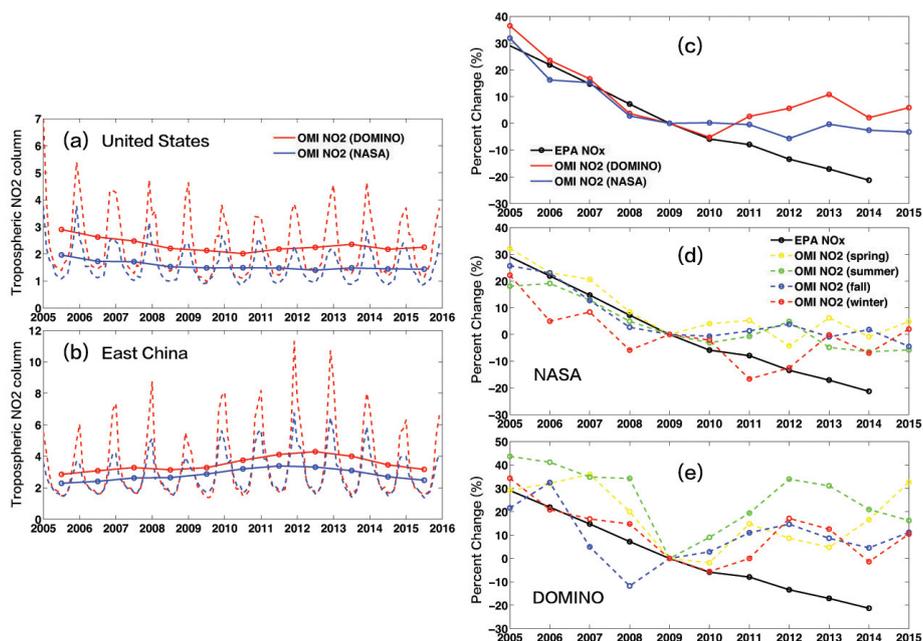
510 **Figure 5.** (a-f) annual slope of tropospheric NO<sub>2</sub> column (unit 1x10<sup>15</sup> molec/cm<sup>2</sup>) from OMI  
511 (NASA and DOMINO products); (g-h) same as panels a-f, with percent (%) as unit.

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513 **Figure 6.** (a-c) Annual slope of passive tracer column (percent base). The passive tracer is  
514 simulated with GEOS-Chem model with constant 15-day lifetime. The surface NO<sub>x</sub> emissions are  
515 fixed at 2005 level. The lightning NO<sub>x</sub> emissions are not included in the simulation. (d) Blue line:  
516 regional mean (box in panel a) of passive tracer column, normalized by the 11-year mean (2005-  
517 2015). The black line shows the NOAA Niño 3.4 index.

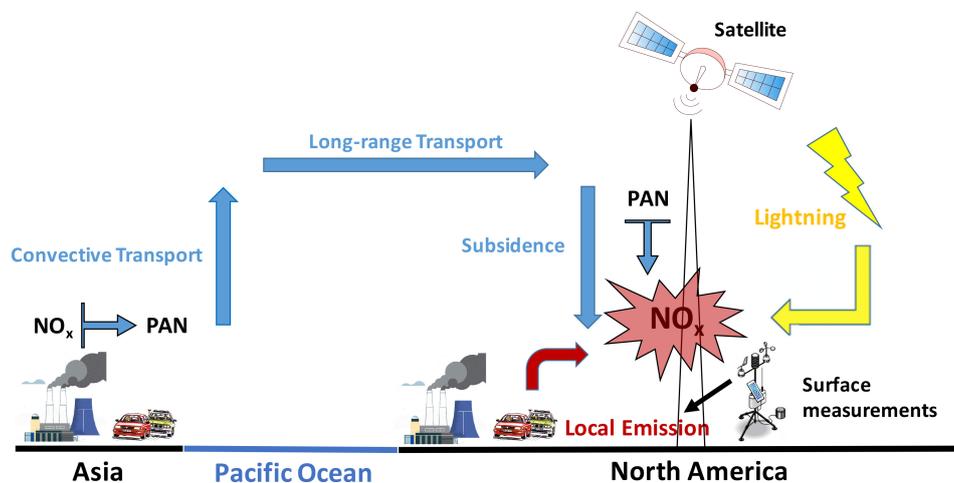
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519 **Figure 7.** Seasonal mean passive tracer column (2005-2015), with 10<sup>16</sup> molec cm<sup>-2</sup>. The passive  
520 tracer is simulated with GEOS-Chem model with constant 15-day lifetime. The surface NO<sub>x</sub>  
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524 **Figure 1.** (a-b): monthly mean (dash lines) and annual mean (solid lines) tropospheric  $\text{NO}_2$   
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 527 column over the US from the OMI and EPA's emission estimates, normalized at 2009.  
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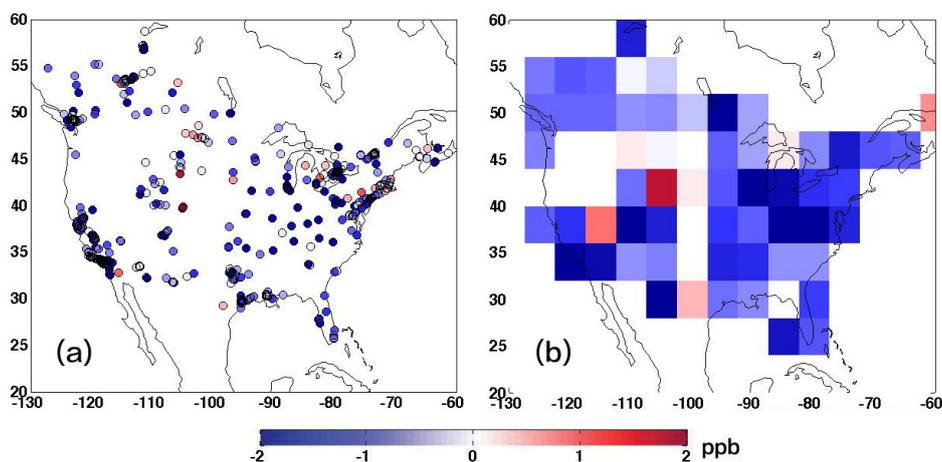


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530 **Figure 2.** Schematic figure showing the sources of tropospheric  $\text{NO}_x$ .  
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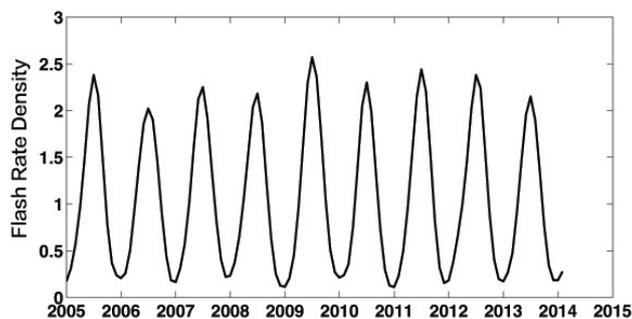


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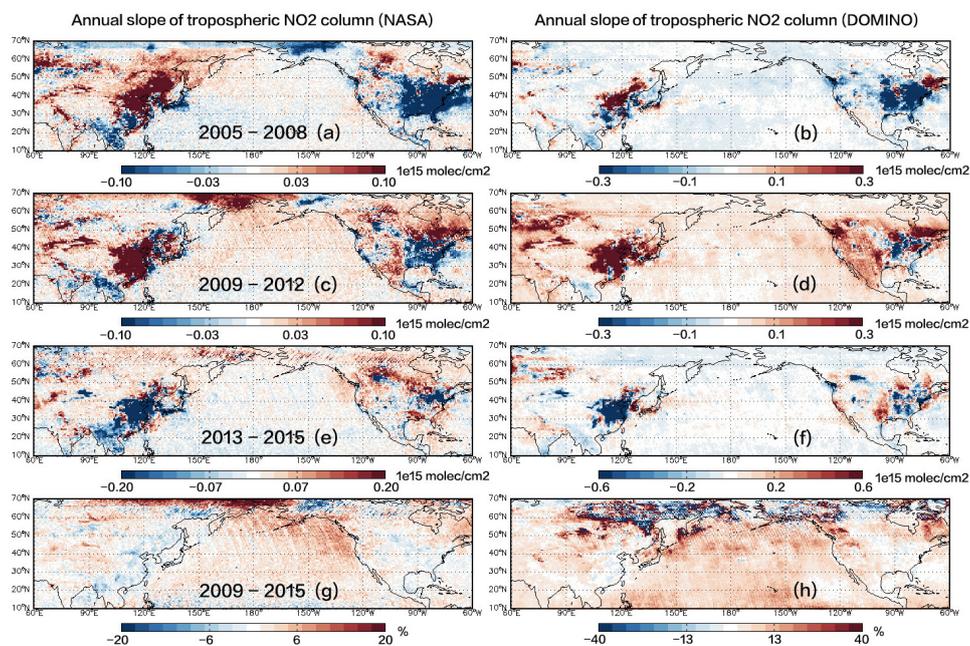


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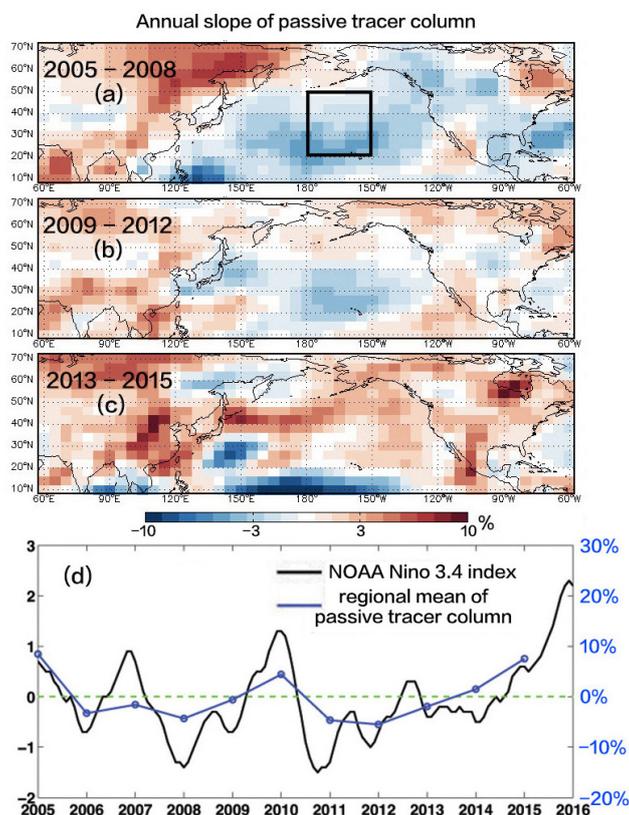


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**Figure 5.** (a-f) annual slope of tropospheric NO<sub>2</sub> column (unit  $1 \times 10^{15} \text{ molec/cm}^2$ ) from OMI (NASA and DOMINO products); (g-h) same as panels a-f, with percent (%) as unit.

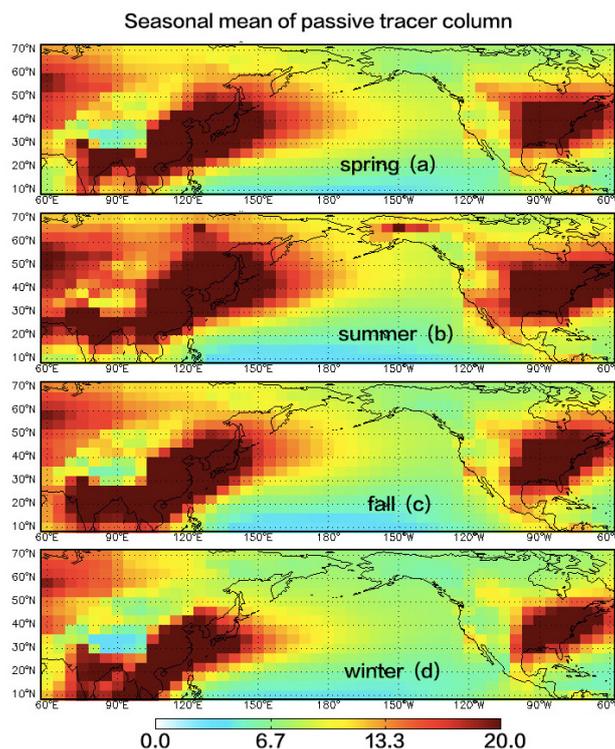


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**Figure 6.** (a-c) Annual slope of passive tracer column (percent base). The passive tracer is simulated with GEOS-Chem model with constant 15-day lifetime. The surface  $\text{NO}_x$  emissions are fixed at 2005 level. The lightning  $\text{NO}_x$  emissions are not included in the simulation. (d) Blue line: regional mean (box in panel a) of passive tracer column, normalized by the 11-year mean (2005-2015). The black line shows the NOAA Niño 3.4 index.



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