

# A new diagnostic for tropospheric ozone production

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## Abstract

Tropospheric ozone is important for the Earth's climate and air quality. It is produced during the oxidation of organics in the presence of nitrogen oxides. Due to the range of organic species emitted and the chain like nature of their oxidation, this chemistry is complex and understanding the role of different processes (emission, deposition, chemistry) is difficult. We demonstrate a new methodology for diagnosing ozone production based on the processing of bonds contained within emitted molecules, the fate of which is determined by the conservation of spin of the bonding electrons. Using this methodology to diagnose ozone production in the GEOS-Chem chemical transport model, we demonstrate its advantages over the standard diagnostic. We show that the number of bonds emitted, their chemistry and lifetime, and feedbacks on OH are all important in determining the ozone production within the model and its sensitivity to changes. This insight may allow future model-model comparisons to better identify the root causes of model differences.

## 1. Introduction

The chemistry of the troposphere is one of oxidation [Levy, 1973; Kroll *et al.*, 2011]. Organic compounds together with nitrogen and sulfur containing molecules are emitted into the troposphere where they are oxidised into compounds which can either be: absorbed by the biosphere; are involatile enough to form aerosols; can deposit to the surface; or be taken up by clouds and rained out. The oxidation of these compounds is significantly slower than might be expected based on the atmospheric composition of 20% molecular oxygen (O<sub>2</sub>).

31 The inefficiency of ground state O<sub>2</sub> as an atmospheric oxidant is due to its electronic  
32 structure. In quantum mechanics, all atomic particles have an intrinsic angular  
33 momentum known as spin [Atkins and De Paula, 2014]. The spin of an electron is  
34 described by the spin quantum number, *s*, and can have values of either +½ or -½ for a  
35 single electron. The Pauli exclusion principle states that if two electrons occupy the  
36 same orbital then their spins must be paired, and thus cancel. With two unpaired  
37 electrons ground state O<sub>2</sub> is a spin-triplet with a total spin quantum number  $S=½+½=1$   
38 (giving a term symbol of  ${}^3\Sigma_g^-$ ). In contrast, virtually all trace chemicals emitted into  
39 the atmosphere contain only paired electrons and are thus spin-singlets ( $S=0$ ). The  
40 quantum mechanical spin selection rule  $\Delta S=0$  means that allowed electronic  
41 transitions must not result in a change in electron spin. From a simplistic perspective  
42 (i.e. ignoring nuclear spin interactions, inter-system crossings, nuclear dipole effects  
43 etc.) this spin selection rule means that the reaction of ground state O<sub>2</sub> with most  
44 emitted compounds is effectively spin forbidden. Electronically excited O<sub>2</sub> ( ${}^1\Delta_g$  or  
45  ${}^1\Sigma_g^+$ ) is a spin singlet and is more reactive in the atmosphere but low concentrations  
46 limit its role [Larson and Marley, 1999]. Instead, atmospheric oxidation proceeds  
47 predominantly via reactions with spin-doublet oxygen-derived species ( $S=½$ ), notably  
48 the hydroxyl (OH) and peroxy radicals (RO<sub>2</sub> = HO<sub>2</sub>, CH<sub>3</sub>O<sub>2</sub>, C<sub>2</sub>H<sub>5</sub>O<sub>2</sub>, etc.), or spin-  
49 singlet species (e.g. ozone (O<sub>3</sub>)).

50 One of the few spin-triplet species in the atmosphere other than O<sub>2</sub> is the ground state  
51 of atomic oxygen (O(<sup>3</sup>P)), which readily undergoes a spin allowed reaction with O<sub>2</sub> to  
52 produce the spin-singlet O<sub>3</sub> molecule. This spin allowed reaction is responsible for the  
53 creation of O<sub>3</sub> in both the stratosphere, where it forms the protective O<sub>3</sub> layer, and the  
54 troposphere. The ability of O<sub>3</sub> to oxidise other spin-singlet species makes it a powerful  
55 oxidant, and it is thus considered a pollutant with negative health effects. Sources of  
56 O(<sup>3</sup>P) within the troposphere are limited because solar photons at sufficiently short  
57 wavelengths to directly photolyse O<sub>2</sub> to O(<sup>3</sup>P) are essentially unavailable.

58 Aside from the photolysis of O<sub>3</sub> itself, the only other significant source of  
59 tropospheric O(<sup>3</sup>P) is the photolysis of nitrogen dioxide (NO<sub>2</sub>) [Crutzen, 1971].  
60 Nitrogen oxides are emitted into the troposphere as nitrogen oxide (NO), which can be  
61 oxidised to NO<sub>2</sub> by O<sub>3</sub> and other oxidants. A large thermodynamic energy barrier  
62 prevents oxidation of NO to NO<sub>2</sub> by the OH radical [Nguyen *et al.*, 1998], and  
63 therefore NO oxidation occurs through reaction with either O<sub>3</sub> or RO<sub>2</sub>. In terms of O<sub>3</sub>

64 production, the oxidation of NO by O<sub>3</sub> forms a null cycle. Thus only the reaction of  
65 NO with RO<sub>2</sub> leads to a net production of O<sub>3</sub>.

66 Exploring the distribution, source and sinks of tropospheric O<sub>3</sub> is a central theme of  
67 atmospheric science. Chemical transport models (online and offline) are essential  
68 tools enabling this understanding but their validity needs to be continually assessed.  
69 Model-model comparison exercises are commonly performed to assess performance,  
70 and comparisons of modelled O<sub>3</sub> budgets traditionally form part of this assessment  
71 [*Stevenson et al.*, 2006; *Wu et al.*, 2007; *Wild*, 2007; *Young et al.*, 2013]. Ozone  
72 production is diagnosed from the flux of NO to NO<sub>2</sub> via reaction with each of the  
73 speciated RO<sub>2</sub> in the model's chemical schemes. This approach provides information  
74 on the relative importance of the different RO<sub>2</sub> in the fast NO + RO<sub>2</sub> reactions within  
75 the model, but gives very little detail on how the longer time scale model processes  
76 (emissions, chemistry, deposition) influence O<sub>3</sub> production. Thus exploring the  
77 reasons that models differ in their O<sub>3</sub> production is difficult and progress has been  
78 slow.

79 A new diagnostic framework that links large scale model drivers such as emission,  
80 chemistry, and deposition to O<sub>3</sub> production would allow an improved assessment of  
81 why model ozone budgets differ. We attempt to provide such a framework here.

## 82 **2. A new diagnostic framework.**

83 The rate of production of tropospheric O<sub>3</sub> is limited by the rate of oxidation of NO to  
84 NO<sub>2</sub>, which is in turn limited by the rate of production of peroxy radicals (RO<sub>2</sub>).  
85 Peroxy radicals form through association reactions of hydrogen (H) atoms or alkyl  
86 radicals (both spin-doublets, S=1/2) with O<sub>2</sub>, forming a highly reactive spin-doublet  
87 radical on an oxygen atom. This spin allowed reaction converts spin-triplet O<sub>2</sub> that  
88 cannot react with spin-singlet pollutants into a spin-doublet O<sub>2</sub> containing species that  
89 can. As such the formation of RO<sub>2</sub> is central to the atmosphere's oxidation capacity,  
90 and its production is limited by the rate of production of H atoms or alkyl radicals.  
91 Thus the maximum potential rate of tropospheric O<sub>3</sub> production is equal to the rate at  
92 which H atoms and alkyl radicals are produced.

93 Hydrogen atoms and alkyl radicals are predominantly produced via the spin allowed  
94 breaking of the spin-pairing between the two electrons in a C or H containing covalent  
95 bond (S=0), such as those in hydrocarbons. These spin-pairings can be broken in the

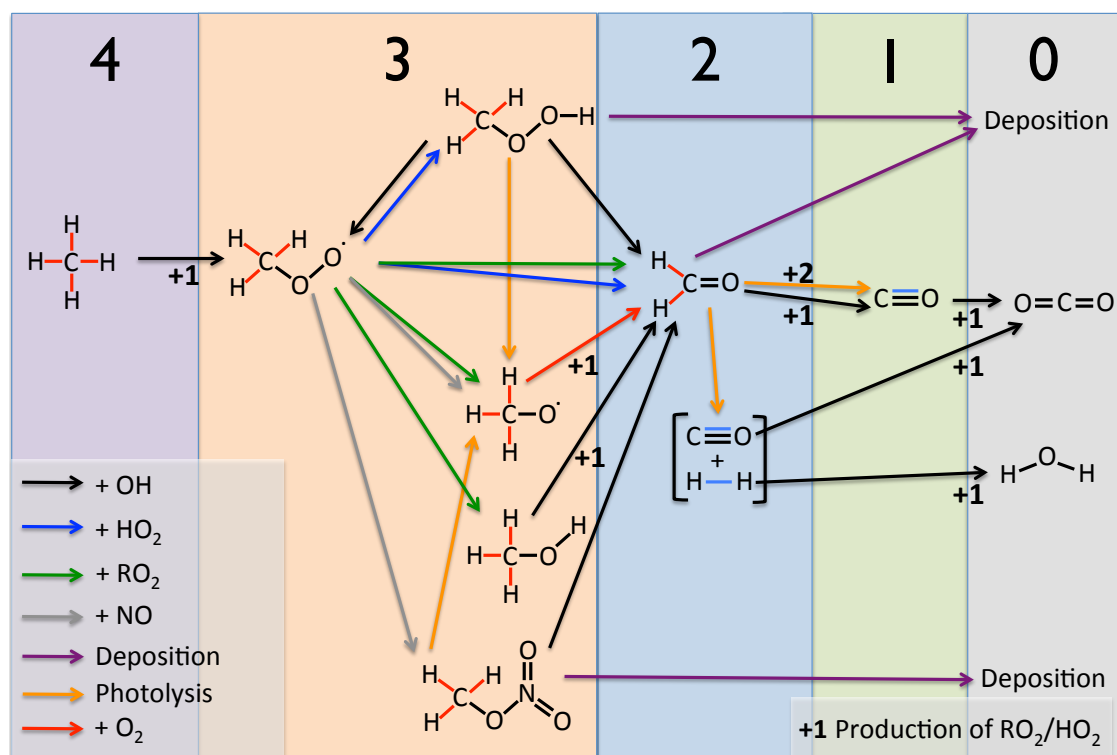
96 atmosphere either chemically or photolytically, with the products necessarily  
97 conserving spin. The breaking of a covalent bond by a photon ( $s=1$ ) can result in two  
98 products with  $S=1/2$  or two products with  $S=0$ . Likewise, oxidation by a radical ( $S = 1/2$ )  
99 will result in one product with  $S=0$  and one with  $S=1/2$ , because the unpaired electron  
100 on the radical reactant pairs with one of the covalent bond electrons to produce a spin-  
101 singlet.

102 Although the majority of  $RO_2$  is formed from emitted C or H containing covalent  
103 bonds, there are a few notable exceptions. Hydrogen atoms can also be produced  
104 through the oxidation of CO to  $CO_2$  by OH. During this reaction the coordinate bond  
105 between the C and O atom is broken and the H atom is produced via the breaking of  
106 the O-H bond. The other notable exception is the oxidation of an  $SO_2$  lone pair of  
107 electrons to  $SO_3$  by OH, where again the H atom produced comes from the OH. In  
108 both of these exceptions a spin-singlet electron pairing (CO coordinate bond or  $SO_2$   
109 lone pair) is broken during the production of the H atom, and we can therefore  
110 consider these reactions as similar to the breaking of C or H containing covalent bond.  
111 For simplicity these spin-singlet electron pairings that can be broken in the  
112 troposphere to produce either a H atom or alkyl radical will be referred to as  
113 “oxidisable bonds” (C-C, C-H, C=C, CO coordinate bond, S:).

114 Tropospheric  $O_3$  production occurs through the oxidation of NO by  $RO_2$ . Following  
115 the above rationale, these  $RO_2$  are produced during the spin allowed breaking of  
116 oxidisable bonds predominantly contained within emitted VOCs. This perspective  
117 allows us to build a new metric for the production of tropospheric  $O_3$  based around the  
118 spin conserving properties of oxidisable bond breaking. In the extreme case, all  
119 oxidisable bonds are photolysed to produce two spin-doublet  $RO_2$  products, which  
120 then react exclusively with NO to generate  $O_3$ . Thus at steady state, the maximum rate  
121 of  $O_3$  production is equal to the rate of production of  $RO_2$ , which is equal to twice the  
122 rate of destruction of the number of oxidisable bonds. This in turn is equal to twice the  
123 rate of emission of oxidisable bonds. Deviation from this maximum is determined by:

- 124 • The relative importance of processes that produce spin-singlet vs. spin-  
125 doublet products during oxidisable bond breaking;
- 126 • The fraction of spin-doublet products from oxidisable bond breaking which  
127 form  $RO_2$ ;
- 128 • The fraction of  $RO_2$  that go on to oxidize NO to  $NO_2$ .

129 To illustrate this Fig. 1 shows the tropospheric oxidation of a methane ( $\text{CH}_4$ ) molecule  
 130 through various steps to either a carbon dioxide ( $\text{CO}_2$ ) molecule or a species that is  
 131 deposited ( $\text{CH}_3\text{OOH}$ ,  $\text{CH}_2\text{O}$ ,  $\text{CH}_3\text{NO}_3$ ). Methane contains 4 x C-H oxidisable bonds  
 132 (8 paired bonding-electrons) and as the oxidation proceeds, the number of oxidisable  
 133 bonds decays to zero. Figure 1 highlights the steps in the tropospheric  $\text{CH}_4$  oxidation  
 134 mechanism that form spin-doublet products, with between 1 and 5  $\text{RO}_2$  produced  
 135 depending on the oxidation pathway. This compares with the theoretical maximum of  
 136 8 if all the original C-H bonds were photolysed to yield 2 spin-doublet products.



137  
 138 **Figure 1. Peroxy radical production during the tropospheric oxidation of  $\text{CH}_4$ .**  
 139 **Moving from left to right, the oxidisable bonds (emitted = red, produced = blue)**  
 140 **present in  $\text{CH}_4$  are removed via a range of tropospheric processes, indicated by**  
 141 **the coloured arrows. The large numbers across the top of the figure indicate the**  
 142 **number of oxidisable bonds at each stage of this oxidation. The production of**  
 143  **$\text{RO}_2$  is indicated by the +1/+2 numbers with the associated process arrows for**  
 144 **producing 1 or 2  $\text{RO}_2$  respectively.**

145 The principal atmospheric source of oxidisable bonds is the emission of C-H, C-C and  
 146 C=C bonds in hydrocarbons, with the only other significant sources being the  
 147 emission of CO and the chemical production of CO and  $\text{H}_2$  during hydrocarbon  
 148 oxidation. Over a long enough timescale, the global atmosphere can be considered to

149 be in a chemical steady state, where the rate of loss of oxidisable bonds is balanced by  
 150 the rate of production or emission. Thus the O<sub>3</sub> production rate can be described by  
 151 equation (1), where the O<sub>3</sub> production metric  $P_sO_3$  is equal to the number of spin-  
 152 paired electrons in oxidisable bonds (i.e. twice the sum of the number of oxidisable  
 153 bonds emitted ( $E_{bonds}$ ) and chemically produced ( $P_{bonds}$ )), multiplied by the number of  
 154 spin-doublet radicals produced per oxidisable bond break divided by the maximum of  
 155 2 ( $F_{Radicals}$ ), multiplied by the fraction of the radicals produced which are RO<sub>2</sub> ( $F_{RO_2}$ ),  
 156 multiplied by the fraction of RO<sub>2</sub> that goes on to react with an NO to produce an O<sub>3</sub>  
 157 molecule ( $F_{NO}$ ). A small correction ( $I$ ) for the production of RO<sub>2</sub> via reactions of spin-  
 158 doublet radicals other than those that result in the breaking of oxidisable spin-pairings  
 159 (e.g.  $O_3 + OH \rightarrow HO_2 + O_2$ ) is included.

$$160 \quad P_sO_3 = \left( (2 \times (E_{bonds} + P_{bonds}) \times F_{radicals} \times F_{RO_2}) + I \right) \times F_{NO} \quad (1)$$

### 161 **3. Implementation**

162 We use the GEOS-Chem model to evaluate this new O<sub>3</sub> production diagnostic. GEOS-  
 163 Chem is a global chemical transport model of tropospheric chemistry, aerosol and  
 164 transport ([www.geos-chem.org](http://www.geos-chem.org) version 9-02). The model is forced by assimilated  
 165 meteorological and surface fields (GEOS-5) from NASA's Global Modelling and  
 166 Assimilation Office, and was run at 4°x5° spatial resolution. The model chemistry  
 167 scheme includes O<sub>x</sub>, HO<sub>x</sub>, NO<sub>x</sub>, BrO<sub>x</sub> and VOC chemistry as described in Mao et al.  
 168 [2013] as are the emissions. The new  $P_sO_3$  diagnostic has been implemented via the  
 169 tracking of reactions by type in the GEOS-Chem chemical mechanism file (further  
 170 details given in the SI). This tracking of reactions enables the fate of all oxidisable  
 171 bonds as well as the production and loss of all RO<sub>2</sub> within the model to be determined  
 172 using the standard GEOS-Chem production and loss diagnostic tools. Model  
 173 simulations were run for 2 years (July 1<sup>st</sup> 2005 – July 1<sup>st</sup> 2007) with the first year used  
 174 as a spin up and the diagnostics performed on the second year.

175 The standard GEOS-Chem diagnostic for O<sub>3</sub> production ( $PO_3$ ) is shown on the left  
 176 side of Table 1. This emphasizes the very fast cycling between NO and NO<sub>2</sub>, but  
 177 provides little in terms of higher process level information. The right side of Table 1  
 178 shows the new budget for  $P_sO_3$ , which tracks the processing of oxidisable bonds  
 179 within the model. Both diagnostic methods give the same final answer but our new  
 180 methodology provides more process level detail. Figure 2 illustrates this new process

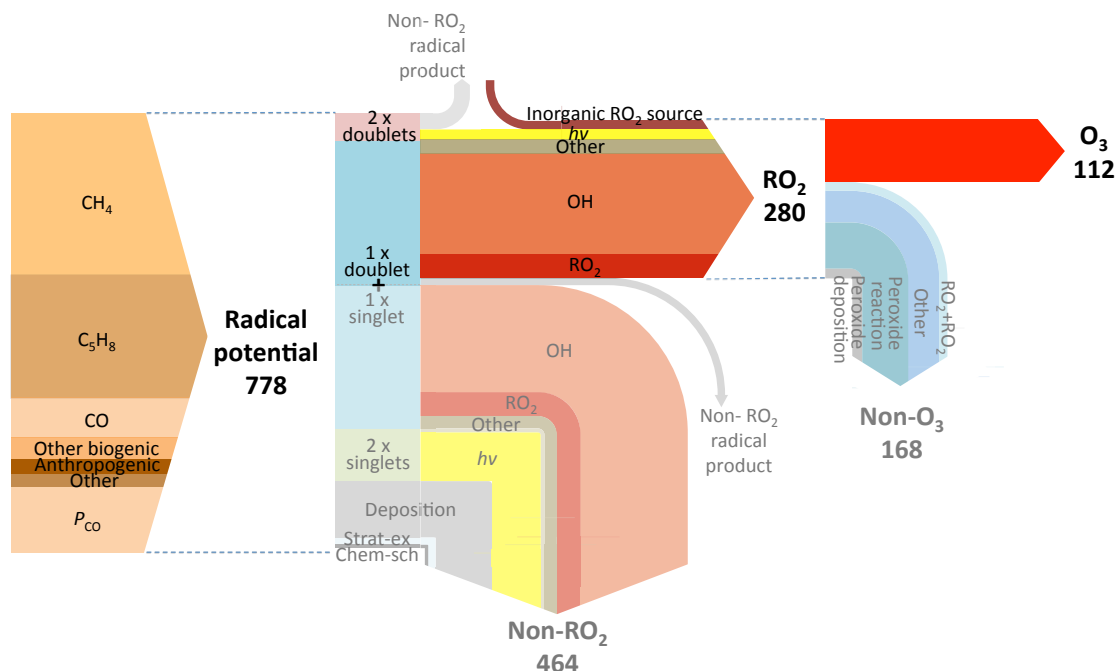
181 based approach, showing the flow of emitted oxidisable spin-paired electrons (bonds)  
 182 to O<sub>3</sub> and the magnitude of the various mechanisms that contribute to and compete  
 183 with O<sub>3</sub> production. The annual oxidisable bond emission of 389 T mol yr<sup>-1</sup> has the  
 184 potential to create 778 T mol yr<sup>-1</sup> of radicals. If all oxidisable bonds were broken by  
 185 photons to produce two radical products the RO<sub>2</sub> production would be 778 T mol yr<sup>-1</sup>.  
 186 If the oxidisable bonds were instead broken via radical reaction (e.g. OH) then RO<sub>2</sub>  
 187 production would be 389 T mol yr<sup>-1</sup>. The various oxidisable bond breaking / removal  
 188 pathways within the model result in the production of 280 T mol yr<sup>-1</sup> of RO<sub>2</sub>, with the  
 189 remainder largely producing stable spin singlet products.

190 Of the 280 T mol yr<sup>-1</sup> RO<sub>2</sub> produced, 112 T mol yr<sup>-1</sup> reacts with NO to produce O<sub>3</sub>.  
 191 The remainder is lost through the reaction or deposition of RO<sub>2</sub> reservoir species  
 192 (RO<sub>2y</sub>= RO<sub>2</sub> + peroxides + peroxy-acetyl nitrates). For example the production of  
 193 methylperoxide (CH<sub>3</sub>O<sub>2</sub> + HO<sub>2</sub> = CH<sub>3</sub>OOH) results in the loss of 2 RO<sub>2</sub>'s. However,  
 194 the reaction of methylperoxide with OH can re-release CH<sub>3</sub>O<sub>2</sub> (CH<sub>3</sub>OOH + OH =  
 195 CH<sub>3</sub>O<sub>2</sub> + H<sub>2</sub>O). Thus, the production of methylperoxide represents the loss of a HO<sub>2</sub>  
 196 and the movement of a CH<sub>3</sub>O<sub>2</sub> into a peroxide RO<sub>2y</sub> reservoir species. The deposition  
 197 of a peroxide molecule is thus the loss of a RO<sub>2y</sub> reservoir species. Notable in Fig. 2 is  
 198 that the role of PAN and nitrate removal of global RO<sub>2y</sub> is negligible, instead being  
 199 dominated by peroxide production and loss and the reaction of RO<sub>2</sub> with O<sub>3</sub>.

<b>PO<sub>3</sub> / T mol Yr<sup>-1</sup></b>		<b>PO<sub>3</sub> / T mol Yr<sup>-1</sup></b> (except F <sub>Radicals</sub> , F <sub>RO2</sub> , and F <sub>NO</sub> which are all unitless)	
NO + HO <sub>2</sub> → NO <sub>2</sub>	74	E <sub>bonds</sub>	330
NO + CH <sub>3</sub> O <sub>2</sub> → NO <sub>2</sub>	27	P <sub>bonds</sub>	58
Other RO <sub>2</sub> + NO → NO <sub>2</sub>	10	F <sub>radicals</sub>	0.40
Other	1	F <sub>RO2</sub>	0.86
		Inorganic RO <sub>2</sub> source	15
		F <sub>NO</sub>	0.40
<b>PO<sub>3</sub></b>	<b>112</b>	<b>P<sub>s</sub>O<sub>3</sub></b>	<b>112</b>

200 **Table 1. Comparison of ozone production diagnostics for GEOS-Chem base**  
 201 **simulation. Standard model PO<sub>3</sub> diagnostics (left column) show reactions**  
 202 **responsible for NO to NO<sub>2</sub> conversions but provide little process level**

203 information. The new  $P_sO_3$  (right) provides increased information on the  
 204 processes controlling  $O_3$  production within the model.



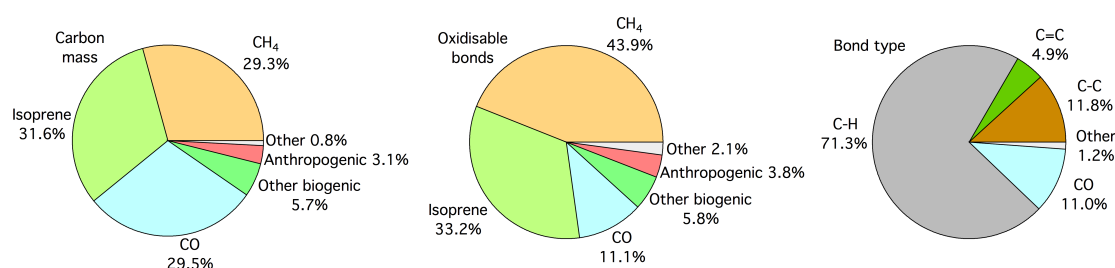
205  
 206 **Figure 2. Flow of oxidisable bonds to  $O_3$  production in the GEOS-Chem base**  
 207 **simulation. Arrows are coloured according to process and the arrow thickness is**  
 208 **proportional to the flux through that channel. Spin-paired electrons are input as**  
 209 **oxidisable bonds into the model (left arrow), with the potential to create 778 T**  
 210 **mol yr<sup>-1</sup> of radicals. The actual fate of these bonds is shown in the central arrow,**  
 211 **producing 280 T mol yr<sup>-1</sup> of  $RO_2$ , of which 112 T mol yr<sup>-1</sup> reacts with NO to**  
 212 **produce  $O_3$  (right arrow).**

### 213 3.1 Emitted oxidisable bonds

214 The fuel for tropospheric oxidation chemistry is the emission of oxidisable bonds,  
 215 predominantly in the form of hydrocarbons. The production of tropospheric  $O_3$  from  
 216 the spin-paired bonding electrons emitted into the standard GEOS-Chem model  
 217 occurs with an efficiency of 14% (112 T mol yr<sup>-1</sup> molecules of  $O_3$  produced / 778 T  
 218 mol yr<sup>-1</sup> spin-paired electrons emitted as oxidisable bonds, Fig.2). These spin-paired  
 219 bonding electrons are predominantly emitted in the form of  $CH_4$ , isoprene ( $C_5H_8$ ) and  
 220 CO (37%, 28%, and 9% respectively). Oxidisable bonds produced during chemical  
 221 reactions ( $P_{bonds}$ ), account for 15% of the net source. Figure 3 shows emissions of CO  
 222 and hydrocarbons in the standard GEOS-Chem simulation in terms of mass of carbon  
 223 per compound, number of oxidisable bonds per compound and as number of bonds in



224 different oxidisable bond types. The commonly used carbon mass approach splits  
 225 emissions approximately equally between each of the major sources (CH<sub>4</sub> (29%),  
 226 Isoprene (32%) and CO (30%)). In contrast, the oxidisable bonds accounting approach  
 227 apportions hydrocarbon emissions 44%, 33% and 11% for CH<sub>4</sub>, isoprene and CO  
 228 respectively. This highlights the high number of oxidisable bonds per carbon atom in  
 229 CH<sub>4</sub> (4) compared to isoprene (2.8) and CO (1). Thus efforts to consider emissions on  
 230 a per-bond basis may provide more insight into chemical processes, as it is these  
 231 bonds that ultimately determine the chain-like chemistry rather than the mass of  
 232 carbon atoms. This helps to emphasise the relative importance of CH<sub>4</sub> emissions on  
 233 global tropospheric chemistry compared with other emissions such as isoprene or CO.  
 234 The type of oxidisable bond emitted is overwhelmingly C-H (71%).



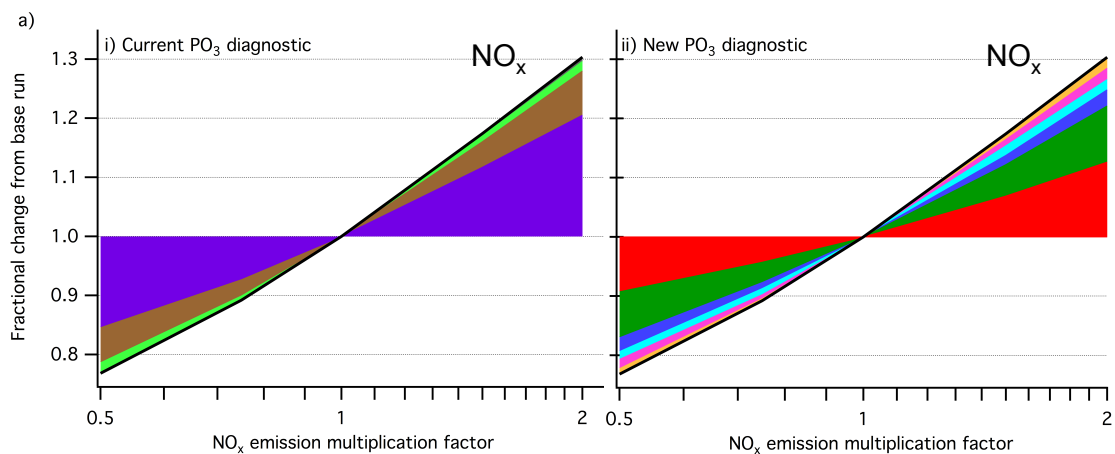
235  
 236 **Figure 3. Pie charts showing hydrocarbon emissions in the base GEOS-Chem**  
 237 **simulation. Emissions split by carbon mass (left), number of oxidisable bonds**  
 238 **(centre) and bond type (right).**

239 The total emission and production of oxidisable bonds has the potential to create 778  
 240 T mol yr<sup>-1</sup> of radicals. However, only 6% of the oxidisable spin-pairings are broken to  
 241 give the maximum 2 spin-doublet products (e.g. radical channel of CH<sub>2</sub>O photolysis).  
 242 The majority (68%) are oxidized via reaction with a spin-doublet species (OH) to  
 243 produce 1 spin-singlet and 1 spin-doublet product (e.g. OH + VOC). The remaining  
 244 26% of spin-paired electrons are removed to form two spin-singlets (e.g. the non-  
 245 radical channel of CH<sub>2</sub>O photolysis). Thus, of the 778 Tmol yr<sup>-1</sup> spin-paired electrons  
 246 emitted or produced only 265 T mol yr<sup>-1</sup> (34%) are converted into RO<sub>2</sub>, with an  
 247 additional 15 T mol yr<sup>-1</sup> produced from reactions such as O<sub>3</sub> + OH → HO<sub>2</sub> + O<sub>2</sub> (I).  
 248 The efficiency of O<sub>3</sub> production from the available oxidisable bonds is further reduced  
 249 as only 40% of the 280 T mol yr<sup>-1</sup> of RO<sub>2</sub> produced react with NO to produce NO<sub>2</sub>.  
 250 The remainder is lost either through the self-reaction of RO<sub>2</sub> or via loss through  
 251 deposition or reaction of RO<sub>2y</sub> reservoir species (e.g. peroxides). Thus overall 14% of  
 252 the emitted bonding electrons go on to make O<sub>3</sub>.

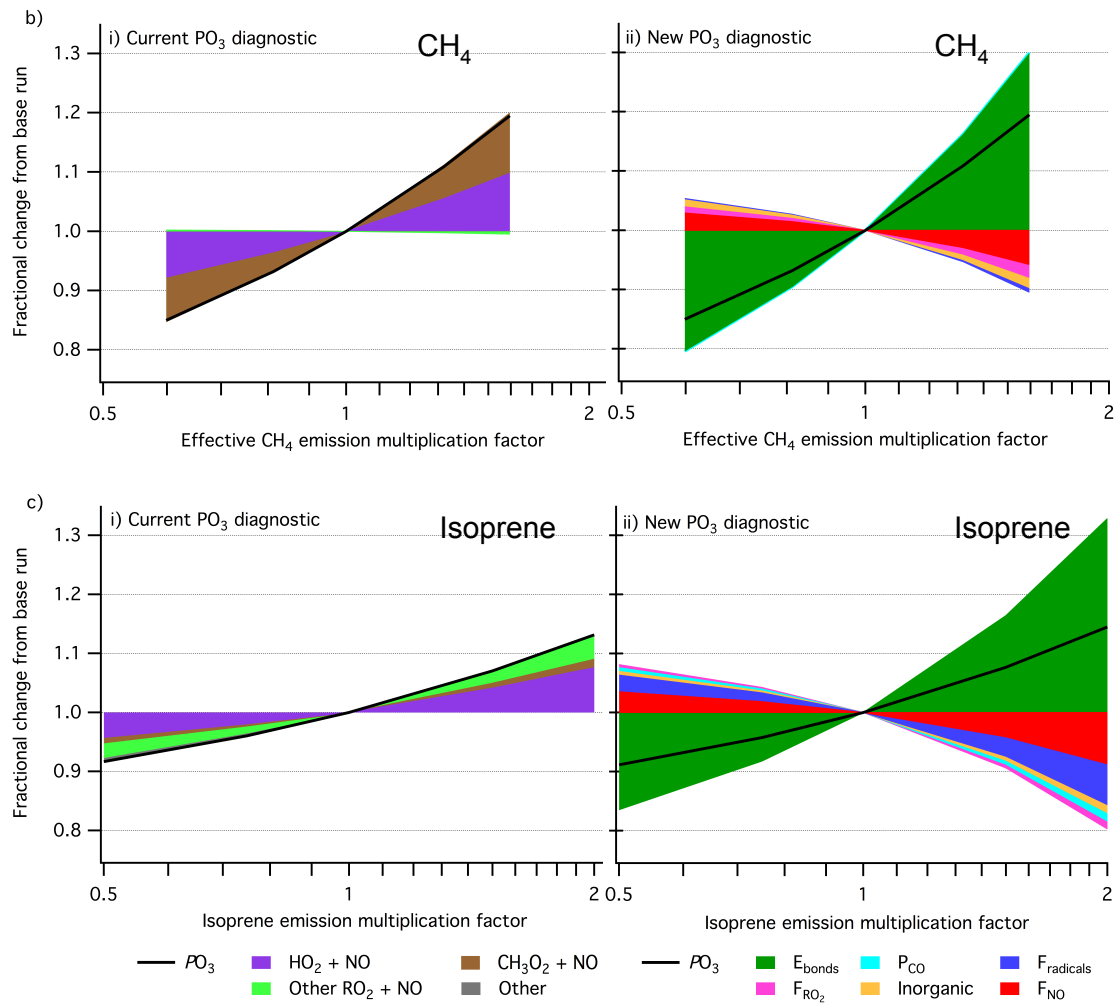
253 The new  $O_3$  production diagnostic presented here ( $P_sO_3$ ) shows the impact of  
254 processes such as emission, deposition and chemical mechanism, and provides  
255 significantly more detail than the standard  $PO_3$  diagnostic approach (Table 1). We  
256 now explore the sensitivity of model  $O_3$  production to changing emissions of  $NO_x$  and  
257 VOC from the perspective of the two diagnostic methods.

## 258 4 Model sensitivities

259 Understanding model response to changing emissions is an important tool for  
260 considering policy interventions. The major controls on  $O_3$  production are emissions  
261 of  $NO_x$  and VOCs. We show in Fig. 2 that from the perspective of global  $O_3$   
262 production, oxidisable bond emissions are dominated by  $CH_4$  and isoprene. Figure 4  
263 shows the impact of changing emissions of  $NO_x$ , isoprene and  $CH_4$  on  $O_3$  production  
264 from both the perspective of this new methodology and the conventional  $NO+RO_2$   
265 diagnostic approach. A set of 5 simulations was performed for each model sensitivity  
266 investigated ( $NO_x$ , isoprene and  $CH_4$ ), with a common base simulation, resulting in 13  
267 simulations in total. The following sections investigate these model responses and use  
268 the new diagnostic to provide insight into the processes driving the observed response  
269 in  $O_3$  production.



270



271

272

273 **Figure 4. Understanding the effect of  $\text{NO}_x$  and VOC emissions on ozone**  
 274 **production at the process level. Stack plots showing fractional change in model**  
 275  **$\text{PO}_3$  compared to base simulation and associated contributions from the current**  
 276  **$\text{PO}_3$  (i) and new  $\text{P}_s\text{O}_3$  (ii) diagnostic parameters under changing  $\text{NO}_x$  emissions**  
 277 **(a), effective  $\text{CH}_4$  emission (b) and isoprene emission (c). The  $\text{P}_s\text{O}_3$  diagnostic**  
 278 **parameters are derived for each model simulation using the diagnostic**  
 279 **implementation described in Sect. 3, and the fractional change in each parameter**  
 280 **from the base simulation calculated.**

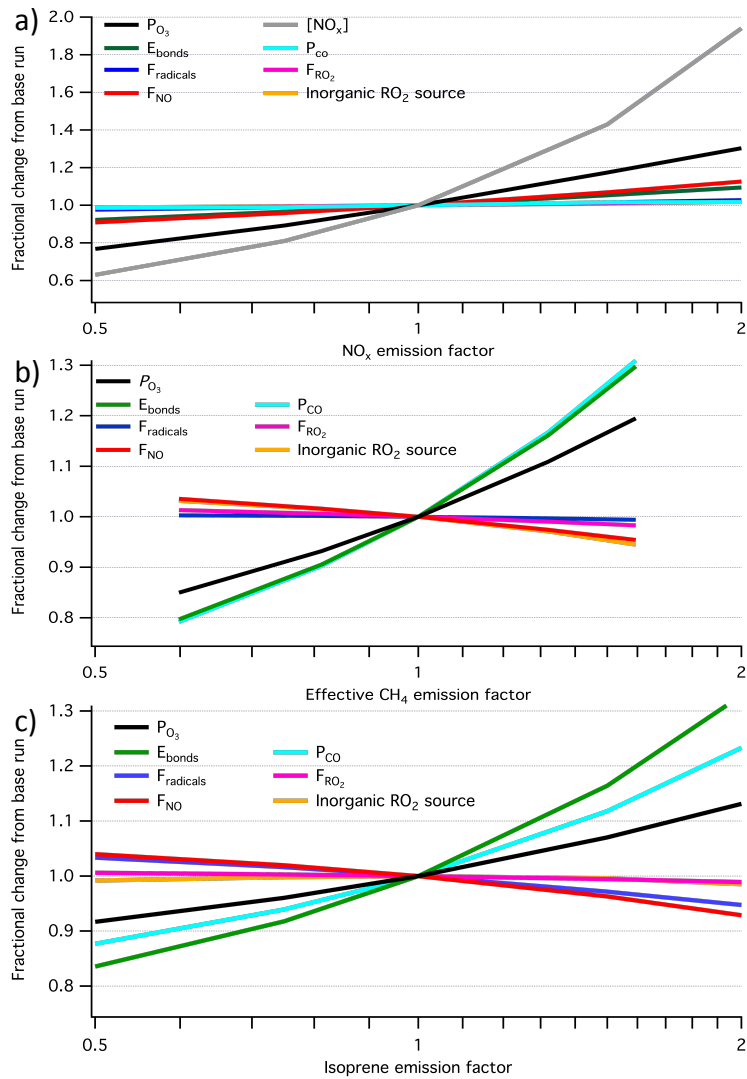
#### 281 4.1 $\text{NO}_x$ emissions

282 Figure 4a diagnoses the relative response of GEOS-Chem  $\text{O}_3$  production to changing  
 283  $\text{NO}_x$  emissions, using simulations where  $\text{NO}_x$  emissions from anthropogenic, biomass  
 284 burning, biofuels, soil and lighting sources were multiplied by factors of 0.5 - 2.  
 285 Increasing  $\text{NO}_x$  emissions increases  $\text{O}_3$  production. The standard  $\text{RO}_2+\text{NO}$  diagnostic  
 286 (Fig.4a(i)) shows that fractional contributions to the total change in  $\text{PO}_3$  from  $\text{HO}_2$

287 (67%), methyl-peroxy ( $\text{CH}_3\text{O}_2$ ) (25%), and other  $\text{RO}_2$  (8%) remain approximately  
288 constant across the  $\text{NO}_x$  emission range investigated. This diagnostic provides little  
289 detail on the processes driving the change in  $\text{O}_3$  production under changing  $\text{NO}_x$   
290 emissions. In contrast, Fig. 4a(ii) is based on the new  $P_s\text{O}_3$  diagnostic and shows a  
291 range of process level changes occurring as  $\text{NO}_x$  emissions change.

#### 292 **4.1.1 Impact of changing $\text{NO}_x$ emission on $F_{\text{NO}}$**

293 Unsurprisingly, as  $\text{NO}_x$  emissions increase the fraction of  $\text{RO}_2$  reacting with  $\text{NO}$  to  
294 produce  $\text{NO}_2$  ( $F_{\text{NO}}$ ) increases (red section in Fig. 4a(ii)). However, this impact only  
295 accounts for around 40% of the increase in  $P_s\text{O}_3$ . Figure 5a shows the fractional  
296 change in all the  $P_s\text{O}_3$  efficiency parameters and the global mean  $\text{NO}_x$  concentration  
297 as a function of the changing  $\text{NO}_x$  emission. As  $\text{NO}_x$  emissions increase the increase  
298 in  $\text{NO}_x$  concentration in the model is somewhat dampened. Halving the  $\text{NO}_x$  emission  
299 leads to  $\text{NO}_x$  burdens dropping by ~35%, and doubling leads to an increase of 95%.  
300 This dampening is due to the impact of  $\text{NO}_x$  emissions on  $\text{OH}$  (see section 4.1.2),  
301 which is the dominant sink for  $\text{NO}_x$ . Increasing  $\text{NO}_x$  increases  $\text{OH}$  concentrations,  
302 which in turn shortens the  $\text{NO}_x$  lifetime thus dampening the response of concentration  
303 to emission.

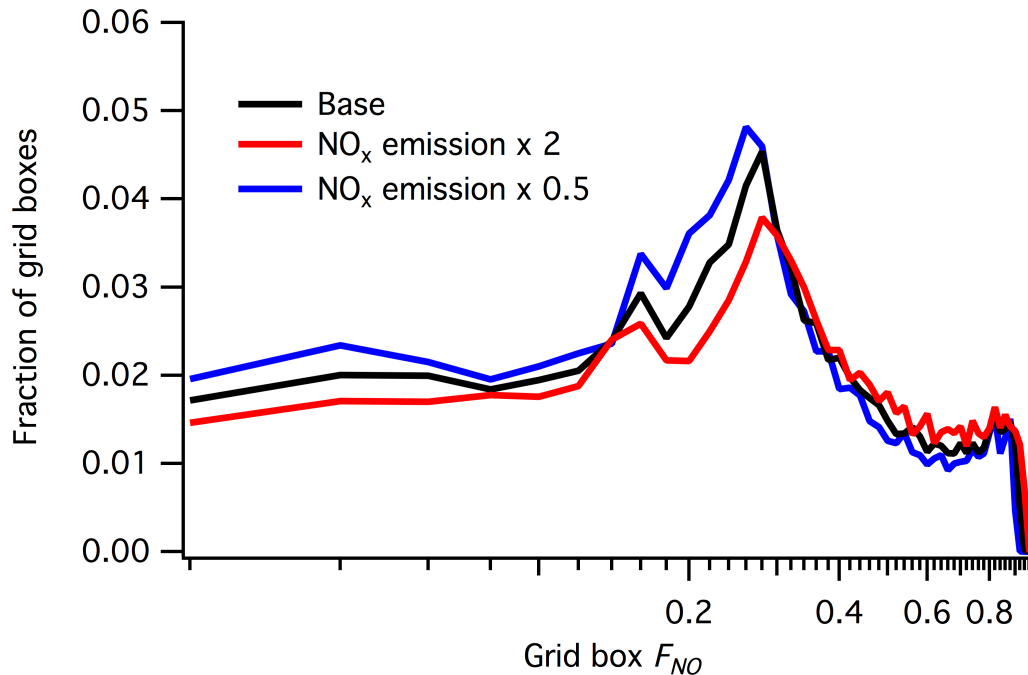


304

305 **Figure 5. Fractional change in new  $P_{O_3}$  diagnostic parameters from base run**  
 306 **against changing  $NO_x$  emission (a); effective  $CH_4$  emission (b); and isoprene**  
 307 **emission (c).**

308 The response of  $F_{NO}$  to changes in  $NO_x$  emissions is also dampened relative to the  
 309 change in  $NO_x$  emissions. This is due to spatial variability in  $F_{NO}$ , which is not  
 310 affected uniformly by changing  $NO_x$  emissions. Figure 6 shows the probability  
 311 distribution of  $F_{NO}$  values across all model grid boxes for the base simulation and the  
 312 half and doubled  $NO_x$  emission simulations (black, blue and red lines respectively).  
 313 For example, in a grid-box in the continental boundary layer where  $RO_2$  reacts  
 314 overwhelmingly with  $NO$ , doubling the  $NO_x$  emission may move  $F_{NO}$  from 0.90 to  
 315 0.95 but it can't double it. Similarly, in the remote boundary layer where  $RO_2$  reacts  
 316 overwhelmingly with other  $RO_2$  doubling  $NO_x$  emissions may move  $F_{NO}$  from 0.3 to

317 0.4 but again it doesn't double. Thus the geographical spread of  $\text{NO}_x$  chemistry limits  
 318 the change in  $F_{\text{NO}}$  caused by changing  $\text{NO}_x$  emissions. The spatial variability in the  
 319 new  $P_s\text{O}_3$  diagnostic parameters shows that this approach has significant potential in  
 320 the analysis of regional  $\text{O}_3$  budgets as well as global.

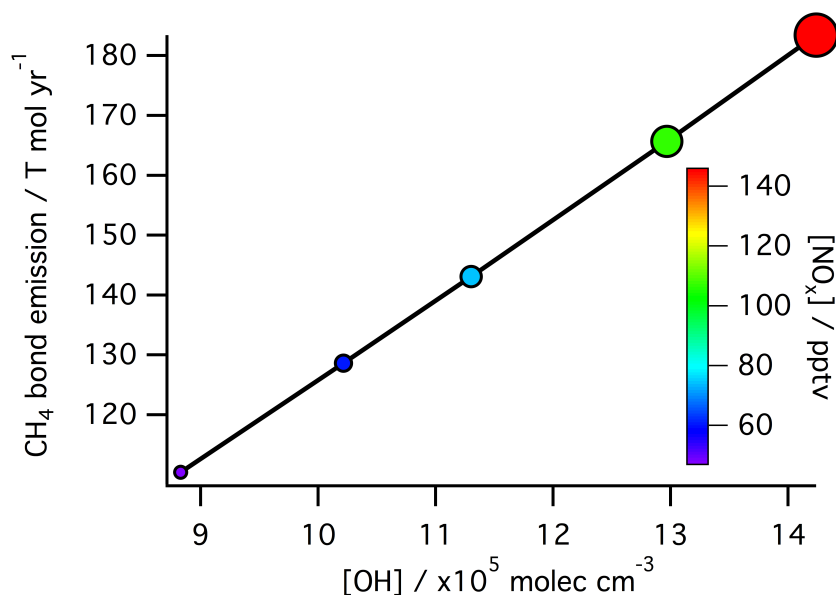


321  
 322 **Figure 6. Effect of  $\text{NO}_x$  emission on distribution of  $F_{\text{NO}}$  values (log scale).  $F_{\text{NO}}$**   
 323 **values for each model grid box in the base and  $\text{NO}_x$  emission x 0.5 and x 2**  
 324 **simulations, split into 50 x 0.02 width bins.**

#### 325 4.1.2 Impact of changing $\text{NO}_x$ emission on $E_{\text{bonds}}$

326 Figure 4a(ii) shows that 60% of the response in  $P_s\text{O}_3$  to changing  $\text{NO}_x$  emission is due  
 327 to factors other than  $F_{\text{NO}}$ , with 40% of the increase due to changes in the emissions  
 328 ( $E_{\text{bonds}}$ : 32%) and chemical production ( $P_{\text{bonds}}$ : 8%) of oxidizable bonds. This increase  
 329 in  $E_{\text{bonds}}$  is surprising given VOC emissions are unchanged in these simulations.  
 330 However, increasing  $\text{NO}_x$  emissions results in an increased OH concentration in the  
 331 model, which then leads to an increase in  $\text{CH}_4$  oxidation. Methane ( $\text{CH}_4$ )  
 332 concentrations are fixed in GEOS-Chem, resulting in an increase in the effective  $\text{CH}_4$   
 333 emission as OH concentrations increase, causing an increase in the total bond  
 334 emission ( $E_{\text{bonds}}$ ). Figure 7 shows the response of effective  $\text{CH}_4$  bond emission to  
 335 global mean OH concentration as it changes with global mean  $\text{NO}_x$  concentration.

336 More CH<sub>4</sub> oxidation also leads to more CH<sub>2</sub>O production and in turn more CO  
337 production (P<sub>CO</sub>), accounting for a significant fraction of the increase in this term.



338

339 **Figure 7. Effective CH<sub>4</sub> emissions as a function of global mean OH concentration,**  
340 **for simulations where NO<sub>x</sub> emissions were changed. Marker size and colour**  
341 **indicate global NO<sub>x</sub> concentration.**

#### 342 **4.1.3 Impact of changing NO<sub>x</sub> emission on F<sub>radicals</sub>, F<sub>RO2</sub> and I**

343 The fraction of radicals produced from bond oxidation (F<sub>radicals</sub>) and the fraction of  
344 those radicals which are RO<sub>2</sub> (F<sub>RO2</sub>) show slight positive increase with NO<sub>x</sub> emission,  
345 accounting for 9% and 6% of the change in P<sub>5</sub>O<sub>3</sub> respectively. This reflects changes in  
346 the partitioning of the fate of the oxidisable bonds, and is largely due to the changes in  
347 OH. As OH increases with NO<sub>x</sub> emission, the rate of chemical oxidation of bonds  
348 increases at the expense of other losses, in particular deposition. The inorganic RO<sub>2</sub>  
349 source term (I) also correlates with NO<sub>x</sub> emission, as it is largely determined by the  
350 concentrations of OH and O<sub>3</sub>. This change accounts for 5% of the observed change in  
351 P<sub>5</sub>O<sub>3</sub>.

352 Thus, with this new diagnostic methodology it is evident that only 40% of the model  
353 O<sub>3</sub> production response to changing NO<sub>x</sub> emission is due to the direct effect of  
354 increasing NO concentration on the rate of RO<sub>2</sub> + NO reactions. Another 40% is due  
355 to fixing the concentration of CH<sub>4</sub> within the model, with the final 20% due to the  
356 increased OH concentration competing for the available oxidisable bonds and  
357 resulting in increased RO<sub>2</sub> production.

## 358 4.2 Changing effective CH<sub>4</sub> emissions

359 As Fig. 2 shows CH<sub>4</sub> to be the largest single source of oxidisable bonds, this section  
360 investigates the response of the O<sub>3</sub> production diagnostics to changing CH<sub>4</sub> emissions.  
361 Figure 4b shows the O<sub>3</sub> production diagnostics response to varying the CH<sub>4</sub> emission  
362 rate within the model. As the model uses prescribed CH<sub>4</sub> concentrations, these were  
363 varied by factors of between 0.5 and 2 from the base simulation and the CH<sub>4</sub> emission  
364 diagnosed from the loss rate of CH<sub>4</sub> to reaction with OH, the only CH<sub>4</sub> loss in the  
365 model. We describe this as the effective CH<sub>4</sub> emission.

366 As effective CH<sub>4</sub> emission increases, O<sub>3</sub> production also increases. The standard  
367 diagnostic (Fig.4b(i)) shows that this increase occurs through an increased rate of  
368 reaction of HO<sub>2</sub> and CH<sub>3</sub>O<sub>2</sub> with NO, as would be expected as these are the RO<sub>2</sub>  
369 produced during CH<sub>4</sub> oxidation. The rate of other RO<sub>2</sub> + NO reactions actually  
370 decreases slightly as CH<sub>4</sub> emissions increase, due to lower OH concentrations and  
371 increased competition for NO from HO<sub>2</sub> and CH<sub>3</sub>O<sub>2</sub>. The new diagnostic (Fig.4b(ii)),  
372 however, shows the increase in O<sub>3</sub> production with increasing effective CH<sub>4</sub> emission  
373 is not simply a result of more HO<sub>2</sub> and CH<sub>3</sub>O<sub>2</sub>.

### 374 4.2.1 Impact of changing effective CH<sub>4</sub> emission on F<sub>NO</sub>

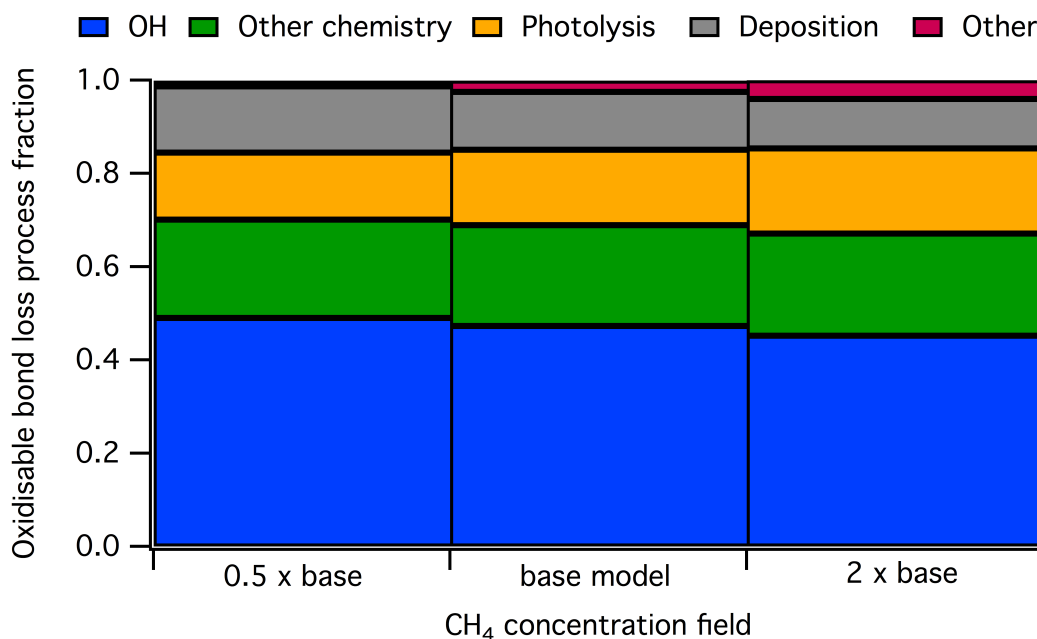
375 The observed change in  $P_s O_3$  is around one third smaller than would be expected from  
376 the increase in the oxidisable bond emission ( $E_{\text{bonds}}$ ) and bond production ( $P_{\text{bonds}}$ )  
377 terms alone. This is due to a countering decrease in the other efficiency parameters  
378 with increasing effective CH<sub>4</sub> emission. Figure 5b shows the fractional change in all  
379 the efficiency parameters as a function of the changing effective CH<sub>4</sub> emission. The  
380 decrease in the fraction of RO<sub>2</sub> reacting with NO to produce NO<sub>2</sub> ( $F_{\text{NO}}$ ) is driven by  
381 increasing O<sub>3</sub> concentrations, which push the NO/NO<sub>2</sub> ratio towards NO<sub>2</sub>. This  
382 reduces the availability of NO to react with RO<sub>2</sub> thereby reducing O<sub>3</sub> production. This  
383 shift in the NO/NO<sub>2</sub> ratio also increases NO<sub>x</sub> loss within the model with increasing  
384 CH<sub>4</sub> emission, as the increased CH<sub>4</sub> oxidation increases RO<sub>2</sub> concentrations resulting  
385 in larger losses of NO<sub>2</sub> via compounds such as peroxyacetyl nitrate (PAN) and  
386 peroxyntic acid (PNA).

### 387 4.2.2 Impact of changing effective CH<sub>4</sub> emission on $E_{\text{bonds}}$

388 Increasing the effective CH<sub>4</sub> emission results in an increase in  $E_{\text{bonds}}$ . Changing the  
389 fraction of total emitted oxidisable bonds from CH<sub>4</sub> does however have significant



390 consequences on the loss mechanisms of these bonds, which influences the other  
 391 efficiency parameters. Figure 8 show the split of oxidisable bond loss mechanisms in  
 392 the base simulation and those with the CH<sub>4</sub> concentration fields multiplied by 0.5 and  
 393 2. As the effective CH<sub>4</sub> emission increases the fraction of bonds lost via OH  
 394 decreases, despite the actual number of oxidisable bonds lost to OH increasing. A  
 395 larger fraction of bonds are therefore lost via the other mechanisms shown in Fig. 8  
 396 rather than reaction with OH. As CH<sub>4</sub> removal occurs predominantly in the free  
 397 troposphere, increasing the effective CH<sub>4</sub> emission also results in a reduction in the  
 398 fraction of oxidisable bonds lost via deposition. The largest fractional increase in bond  
 399 loss mechanism with increasing effective CH<sub>4</sub> emission is for photolysis, with the  
 400 increase in the “other” fraction due to increased loss of bonds to the stratosphere with  
 401 increasing CH<sub>4</sub>.



402

403 **Figure 8. Oxidisable bond loss mechanism fractions under changing effective**  
 404 **CH<sub>4</sub> emissions (0.5 x CH<sub>4</sub> concentration field, base simulation and 2 x CH<sub>4</sub>**  
 405 **concentration field).**

406 **4.2.3 Impact of changing effective CH<sub>4</sub> emission on F<sub>radicals</sub>, F<sub>RO2</sub> and I**

407 The fraction of oxidisable bonds that goes on to produce radicals (F<sub>radicals</sub>) and the  
 408 fraction of these that are RO<sub>2</sub> (F<sub>RO2</sub>) also decrease with increasing effective CH<sub>4</sub>  
 409 emissions. This is due to decreasing global OH concentration resulting from increased  
 410 loss by reaction with CH<sub>4</sub> and a decreasing NO concentration. This favours bond loss

411 via pathways that produce less RO<sub>2</sub> (e.g. CH<sub>2</sub>O photolysis). The long lifetime of CH<sub>4</sub>  
412 compared with the majority of other sources of oxidisable bonds, also results in a  
413 decrease in the fraction of bonds lost to deposition as total bond oxidation increases  
414 fractionally in the free troposphere where deposition is a less significant loss  
415 mechanism than in the boundary layer.

### 416 **4.3 Changing isoprene emission**

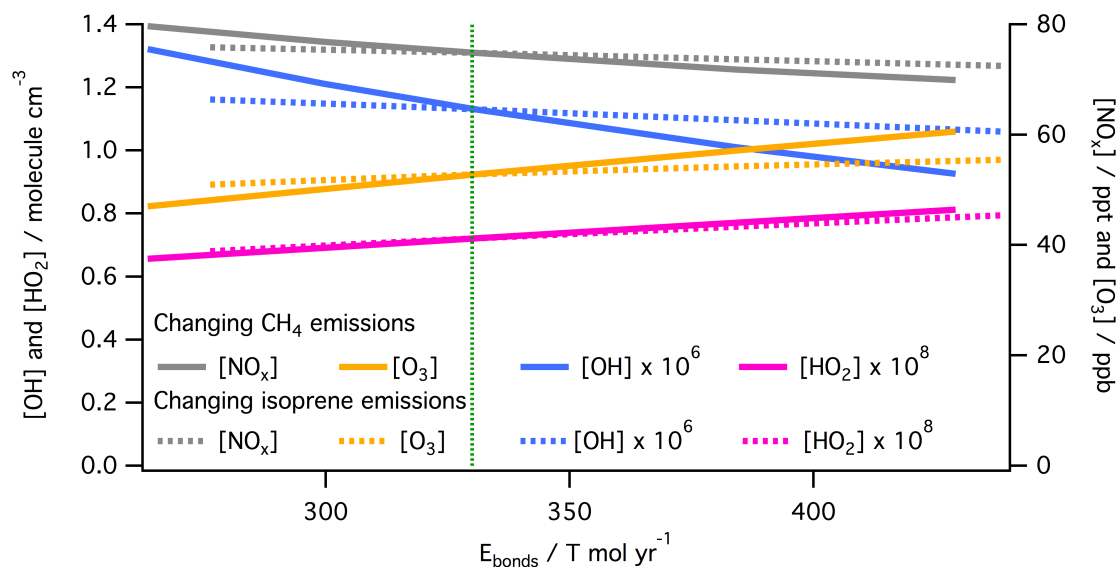
417 The species through which the oxidisable bonds are emitted has a significant impact  
418 on O<sub>3</sub> production, due to their subsequent removal mechanisms. For example, in a  
419 simulation where the only emission of oxidisable bonds is CO, F<sub>radicals</sub> is 0.5 and F<sub>RO<sub>2</sub></sub>  
420 is 1 as the only CO sink is reaction with OH to produce one HO<sub>2</sub> (OH + CO → HO<sub>2</sub> +  
421 CO<sub>2</sub>). The CO coordinate bond, which in theory has the potential to produce 2  
422 radicals, only produces 1 radical, which is an RO<sub>2</sub>.

423 Isoprene has the most complex chemistry in the model and is the second largest  
424 source of bonds into the atmosphere after CH<sub>4</sub> (Fig. 3). Figure 4c shows the response  
425 of the two O<sub>3</sub> production diagnostics to varying the isoprene emission within the  
426 model. The standard diagnostic (Fig.4c(i)) shows that the most significant increase in  
427 *PO*<sub>3</sub> from increasing isoprene emissions is from NO + HO<sub>2</sub> and non-CH<sub>3</sub>O<sub>2</sub> peroxy  
428 radicals, with a smaller increase from CH<sub>3</sub>O<sub>2</sub>. The new *P<sub>s</sub>O*<sub>3</sub> diagnostic (Fig.4c(ii))  
429 again provides more insight, showing significant offsetting of around a half between  
430 the terms.

#### 431 **4.3.1 Impact of changing isoprene emission on F<sub>NO</sub>**

432 The increased isoprene emission leads to a similar change in the magnitude of the  
433 total number of oxidisable bonds emitted (E<sub>bonds</sub>) as the simulations in which effective  
434 CH<sub>4</sub> emission were varied. However, the countering decrease in all of the efficiency  
435 parameters is much larger for isoprene than for CH<sub>4</sub>. Figure 5c shows the fractional  
436 change in the new *P<sub>s</sub>O*<sub>3</sub> ozone production diagnostic parameters as a function of  
437 isoprene emissions compared to the base simulation. The change in F<sub>NO</sub> is due to both  
438 a decrease in global mean NO<sub>x</sub> concentrations with increasing isoprene and the spatial  
439 distribution of isoprene emissions. The majority of global isoprene emissions are in  
440 regions with low NO<sub>x</sub> emissions, and thus low values of F<sub>NO</sub>. Figure 9 shows a  
441 decrease in global mean NO<sub>x</sub>, and global mean OH concentrations with increasing  
442 isoprene emissions, however, the effect is less than that seen when CH<sub>4</sub> is responsible

443 for the same increase in oxidisable bond emission. This is due in a large part to the  
 444 spatial scales over which the two compounds impact.



445

446 **Figure 9. The effect of oxidisable bond parent species on OH, HO<sub>2</sub>, O<sub>3</sub> and NO<sub>x</sub>**  
 447 **concentrations. Global mean [OH], [HO<sub>2</sub>], [O<sub>3</sub>] and [NO<sub>x</sub>] for simulations where**  
 448 **the effective CH<sub>4</sub> emission (solid lines) and isoprene emission (dashed lines) were**  
 449 **changed, against model E<sub>bonds</sub>. The dashed vertical green line indicates E<sub>bonds</sub> in**  
 450 **the base simulation (330 T mol yr<sup>-1</sup>).**

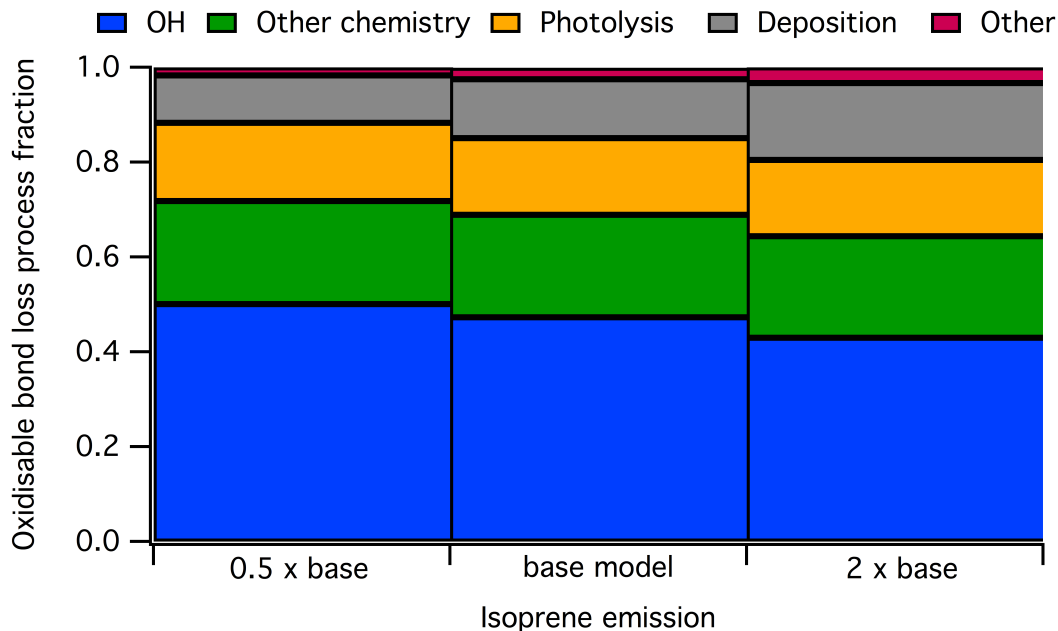
#### 451 4.2.2 Impact of changing isoprene emission on E<sub>bonds</sub>

452 As isoprene is the second largest source of oxidisable bonds (Fig. 3), increasing the  
 453 isoprene emission results in a significant increase in E<sub>bonds</sub>. Differences in both the  
 454 spatial distribution of emissions and the oxidation chemistry of isoprene and CH<sub>4</sub>,  
 455 however, means that the impact of the increases in E<sub>bonds</sub> on O<sub>3</sub> production are  
 456 significantly different for the two compounds. This is predominantly because the  
 457 fraction of oxidisable bonds that are physically deposited for isoprene is high  
 458 compared to those emitted as CH<sub>4</sub>. This increase is due to i) the higher solubility of  
 459 isoprene oxidation products compared to those of CH<sub>4</sub>, and ii) the higher reactivity of  
 460 isoprene means its oxidation occurs in the boundary layer where both dry and wet  
 461 deposition is most effective.

462 Figure 10 shows the fate of oxidisable bonds in the base simulation and those with the  
 463 isoprene emissions multiplied by 0.5 and 2. The complex myriad of products formed  
 464 during the isoprene oxidation mechanism also results in the production of many

465 highly oxygenated multifunctional compounds with high Henry's law solubility  
466 constants, meaning they are more readily lost to deposition.

467



468

469 **Figure 10. Oxidisable bond loss mechanism fractions under changing isoprene**  
470 **emissions.**

471 Increasing the isoprene emission also has a slight offsetting impact on the effective  
472 CH<sub>4</sub> emission, as increased isoprene concentrations decrease OH concentrations, and  
473 thus decrease the effective CH<sub>4</sub> emission. A doubling in isoprene emission causes a  
474 6% reduction in the effective emission of CH<sub>4</sub>.

475

### 476 4.3.3 Impact of changing isoprene emission on $F_{\text{radicals}}$ , $F_{\text{RO2}}$ and $I$

477 As shown in Fig. 3c(ii), increasing the isoprene emission results in a reduction in all  
478  $P_sO_3$  efficiency parameters. The reductions in  $F_{\text{radicals}}$  is due to the higher fraction of  
479 oxidisable bonds that are lost via non-radical forming pathways (e.g. deposition) for  
480 isoprene relative to the other main oxidisable bond emission sources  $CH_4$  and  $CO$ .  
481 The slight decreases of  $F_{\text{RO2}}$  and  $I$  with increasing isoprene emission are  
482 predominantly due to changes in  $OH$  and  $NO_x$  (Fig. 9).

483 The complex chemistry of isoprene oxidation combined with the spatial distribution of  
484 isoprene emissions means the increase in  $O_3$  production due to increases in isoprene  
485 emissions is roughly half what might be expected from the increase in oxidisable bond  
486 emission alone (i.e. if the increase was *via*  $CO$  instead of isoprene).

## 487 5. Conclusions

488 We have shown that this bond-focussed approach to  $O_3$  production provides a  
489 significantly more detailed understanding of the processes involved. The role of  
490 modelled VOC emissions and  $O_3$  burden has been reported previously [Wild, 2007;  
491 Young *et al.*, 2013]. However previous efforts extending this to a general process led  
492 approach has not been successful. This new approach provides a tool with which the  
493 processes controlling  $O_3$  production can be investigated, and a metric by which  
494 different emissions can be compared. For example, the differing chemistry of isoprene  
495 and  $CH_4$  shows that even though their emissions of carbon mass are comparable, the  
496 atmosphere responds in different ways, with the isoprene bonds being less effective in  
497 producing  $O_3$  than  $CH_4$  bonds. By quantifying multiple steps in the  $O_3$  production  
498 process, competing changes in the system become apparent (as shown in Fig. 4b(ii)  
499 and c(ii)) and are thus testable. This enables the effect of model approximations on  $O_3$   
500 production to be quantified (e.g. the effect of  $NO_x$  on  $CH_4$  emissions when using  $CH_4$   
501 concentration fields).

502 This new diagnostic also points towards the importance of observational datasets for  
503 assessing our understanding of tropospheric chemistry. Although the budget presented  
504 in Fig. 2 provides an annually integrated global estimate it points towards local  
505 comparisons that can be made to assess model fidelity. Comparisons, both their  
506 magnitude and their ratios, between observed and modelled bond concentration, bond  
507 emission and loss fluxes (e.g.  $OH$  reactivity [Yang *et al.*, 2016] or depositional fluxes

508 [*Wesely and Hicks, 2000*]), and O<sub>3</sub> production [*Cazorla and Brune, 2010*] would all  
509 provide comparisons for outputs from the *P<sub>s</sub>O<sub>3</sub>* diagnostic and help assess model  
510 performance.

511 Future work is necessary to identify the usefulness of this approach on smaller spatial  
512 and temporal scales. For regional modelling scale, the transport flux of bonds into the  
513 domain would need to be considered alongside the emissions of bonds. However, this  
514 might help to disentangle O<sub>3</sub> production due to local VOC emissions from that due to  
515 VOC emissions outside of the domain. This bond focussed approach may also have  
516 usefulness on shorter timescales. For example, when considering vertical fluxes in and  
517 out of the boundary layer, a bond centred approach could help. What fraction of the  
518 bonds emitted at the surface are exported to the free troposphere. If a measurement of  
519 reactivity flux could be made this could be tested experimentally.

520 Another potentially important application is in model-model comparisons. Increases  
521 in our understanding of why different models calculate different O<sub>3</sub> production and  
522 burdens has been slow [*Stevenson et al., 2006; Wu et al., 2007; Young et al., 2013*].  
523 Although a complete tagging like that described here is unlikely to occur for all of the  
524 models involved in the comparison, a small number of additional diagnostics is likely  
525 to produce a significantly better understanding of the models. Diagnosing (1) the total  
526 bond flux (direct emissions plus the flux for those species kept constant), (2) the rate  
527 of production of RO<sub>2</sub> and (3) the rate of production of O<sub>3</sub>, could help differentiate why  
528 certain models produce more or less O<sub>3</sub> than others. The ratios between these fluxes  
529 would help identify what aspect of the emissions of chemistry differs between the  
530 models.

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572

### 573 **Author contributions**

574 All work presented here was conceived by P.M.E. and M.J.E. The implementation,  
575 model simulations and analysis were carried out by P.M.E., and the manuscript was  
576 written by P.M.E. with substantial input from M.J.E..

### 577 **Additional information**

578 The authors declare no competing financial interests.

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