We thank the referees for their careful and critical review of our paper. The following are our responses to the referees' comments.

Response to Anonymous Referee #1

In the manuscript the authors measured several physicochemical properties, including size, density, chemical composition, hygroscopicity, and Cloud Condensation Nuclei (CCN) activity of coated Black Carbon (BC) particles as a function of aging (coating with condensable species from photo-oxidation of VOCs) in a quasi-atmospheric aerosol evolution study (QUALITY) chamber using a suite of instruments, such as differential mobility analyzer, aerosol particle mass analyzer, High Resolution – Time of Flight – Aerosol Mass Spectrometer (HR-ToF-AMS), Humidified Tandem Differential Mobility Analyzer (HTDMA), and a CCN Counter (CCNC). Their results show that under ambient condition in Beijing, the BC undergoes rapid growth to 77 ± 33 nm coating thickness with an average growth rate of 26 ± 11 nm h-1. The O/C ratio of the SOA coating is 0.5, lower than ambient level, indicating the lack of aqueous phase oxidation inside QUALITY chamber. The hygroscopic parameter is about 0.035 - 0.040 as measured by HTDMA and CCNC, suggesting that the initial photochemical aging of BC particles does not appreciably alter the particle hygroscopicity in Beijing. This study and the data provided are quite extensive. The most valuable addition of this manuscript is the "close to ambient (Beijing)" aging condition in the QUALITY chamber. The author showed in Figure 3 that the O3, NOx, CO, and SO2 concentrations inside and outside of the chamber is close. However, the VOCs plot inside and outside of the chamber is missing. The authors also mentioned they applied heater, drier, alumina spherules coated with potassium permanganate (KMnO4), and activated charcoal to remove the VOCs, H2S, SO2, NOx, and O3. It would be nice if the authors could also show the VOCs are efficiently removed down to sub-ppb level while injecting the BC particles into the chamber. Did the authors use proton-transfer-reaction mass spectrometer (PTR-MS) to monitor the VOCs inside and outside the chamber? Since the coating is caused by condensable species from photo-oxidation of VOCs and "close to ambient" condition is an important part of the manuscript, I would suggest the authors add the VOC plot.

We thank the reviewer very much for the insight suggestion. We have added another figure (as Figure 3 in current version), which demonstrates the comparison of VOCs concentration between chamber and ambient air before the BC aging experiment. Corresponding discussion is also added in the text as (Line 231):

"To investigate the VOCs concentration inside the reaction chamber after the injection of BC particles, both the chamber and ambient air were sampled with VOC canisters just before the BC aging experiment started. These canisters were then analyzed by a gas chromatography-mass spectrometer / flame ionization detector (GC-MS/FID, HP inc.) system (Liu et al., 2008). The concentrations of VOCs containing 4 or more carbons are illustrated in Figure 3. Slightly higher concentrations of several VOCs in the QUALITY chamber, e.g., n-butane, n-pentane, toluene, were observed compared with those in the ambient air, duo to the co-injection of a small amount of VOCs together with BC particles into the chamber. Nevertheless, the average increase of the VOC concentrations was only 16% or 0.1 ppb for all focused VOCs species, with the largest increase of 35% or 0.36 ppb, suggesting the insignificant influence of soot burner on VOCs concentrations and SOA formation in the chamber."

Besides my major point above, I also have several minor comments.

1) On page 1 in abstract (line 18), "wall loss of primary gaseous pollutants was negligible, : : : ", please add "was negligible compared with the replenish rate by gas exchange". Because other readers who don't have gas exchange might also think it is negligible in their chamber, which will cause confusion.

We thank the reviewer for this suggestion. The sentence has been modified as suggested.

2) On page 3 (line 100), is "polytetrafluoroethylene" PTFE? PFA should be "Perfluoroalkoxy alkane".

We are sorry for such mistake. The "polytetrafluoroethylene" should be "Perfluoroalkoxy alkane". It has been revised in the manuscript.

3) On page 6 (line 207), "The transmission : : : were measured using a Fourier Transform Infrared Spectroscopy (FTIR) system". However, in the following context, the authors are discussing about the UV-Vis spectrum of the Teflon film and acrylic shell. Did I miss something? Or does the authors actually mean "UV-Vis" system?

We thank the author for pointing out his mistake. The sentence has been revised as "The transmission efficiencies of each material were measured using an ultraviolet-visible (UV-Vis) spectrophotometry (PerkinElmer Inc., model 552)."

4) On page 12 (line 444), Fig. 11a and Fig. 11b should be Figure 10a and Figure 10b. Please correct that.

It has been revised. Thanks.

5) On page 13 (line 473, and equation 3), if some of SOA components are not very soluble in water, the updated formula in reference (Petters and Kreidenweis 2008) (formula 10) should be used. Otherwise, based on the kappa of 0.04, the average molecular weight of the coating material is about 450 g mol-1.

We agree with the reviewer that the updated formula in reference (Petters and Kreidenweis 2008) (formula 10) is more accurate to calculation the kappa value when coating material is not highly hygroscopic. However, the calculation requires the solubility of SOA, which is impossible to be obtained duo to the variable VOC precursors and complex photochemical reactions in the ambient condition. Nevertheless, we added discussion on the uncertainty to use of this mixing rule as "Since the SOA formed inside the chamber was not highly hygroscopic, some of the SOA components might not be able to solve in water droplets (Petters and Kreidenweis, 2008), lead to the underestimation of the κ values of the coating materials in this study."

6) This is not a comment but a discussion. In the end of "Section 4", the authors discussed about the

relationship between _ and O/C ratio. In the reference (Jimenez et al. 2009), Figure 3 shows at around O/C = 0.5, the _ is about 0.12 for ambient (including Mexico City, Jungfraujoch, Hyytilala SV-OOA, Hyytilala LV-OOA) and _-pinene SOA, but about 0.16 for Trimethylbenzene (TMB) SOA. All the cases have larger _ values than reported here (0.04). However, in another reference (Massoli et al. 2010), Figure 2 shows the _-pinene and m-xylene SOAs have _ value about 0.14 at O/C=0.5, but TMB SOA has _ value of only 0.04, consistent with the value (0.04) reported here. If we believe the latter reference is correct, does that mean the VOC source in Beijing is more TMB like? O/C could not be used as the only parameter for predicting _ as discussed. This could also suggest the VOC source in Beijing is different from other places, such as Mexico City.

The reviewer made a very interesting point to examine the application of this study. However, we don't have enough evidence to draw such conclusion yet. First, "The O/C ratio of the coating SOA is 0.5 in our experiment, corresponding to the O/C ratio of approximately 0.4 in Jimenez et al. (2009) and Massoli et al. (2010) duo the utilize of updated AMS calibration method in this study." The κ value at O/C=0.4 will be lower than that provided above by the reviewer. Second, the $\kappa_{\rm H}$ HGF of α -pinene and α -pinene/xylene mixture is consistent with our $\kappa_{\rm H}$ HGF at same O/C ratio. Third, in the measurements at other sites around the world, e.g., Mexico City, Jungfraujoch, Hyytilala, the measured κ value of organics were for the total OM in the atmosphere, including both primary and secondary OM with different aging degree (from hours to days) through both photochemical reactions with aging time scale of only a few hours. Therefore, we can't conclude that the SOA hygroscopiciy in Beijing is different from other cities at the moment. But we believe that this is an interesting aspect and we would like to focus this in our future studies.

Response to Anonymous Referee #2

The manuscript investigated the changes in the hygroscopicity of elemental carbon (EC) aerosol during aging in QUALITY chamber where the gaseous species in ambient air constantly diffused into the reaction chamber and products from their photochemical reactions interacted with mono-dispersed EC seed particles. The variations in EC particle size, mass, chemical composition, hygroscopicity and CCN properties were monitored throughout the aging processes. This study provides a unique perspective among chamber studies because the EC seed aerosol was exposed to photochemical oxidation products from an environment that closely mimics the actual ambient air. Evaluating the hygroscopicity of EC has been a challenge for the atmospheric sciences research community due to the complex morphology changes. Following the method of using mono-dispersed seed aerosol (Qiu et al. ES&T, 2012 and Khalizov et al. ES&T, 2013), this study provided new insights into this complex research problem. The current manuscript also nicely supplements a recent publication from the same authors (Peng et al. PNAS, 2016) by providing detailed evaluation of the the QUALITY chamber. The characterization of the QUALITY chamber experiments appeared to be thorough and the results appeared to be reliable. The topic is relevant to the scope of the journal of Atmospheric Chemistry and Physics and should be considered for publication. A few comments are provided to facilitate the further improvement of the manuscript:

(1) Recent studies showed (for example, Chen et al. Geophys. Res. Lett. 2016, 43, 11080) that the morphology of nascent EC particles can be highly sensitive to even a small change in the organic coating on the primary spheres. As a result, it is prudent to provide experimental data to show that all the EC seed particles started with similar morphology and chemical composition. What was the thickness (or weight percent) of volatile organics in the seed particles at the beginning of each chamber aging experiment?

We thank the reviewer for the suggestion. In our study, the fresh BC particles were highly fractal. For three initial mobility diameters (Dm) of about 100 nm, 150 nm, and 200 nm, the ranges of effective density for fresh BC particles in each experiment are 0.43-0.50 g cm⁻³, 0.34-0.34 g cm⁻³, and 0.24-0.32 g cm⁻³, respectively (Table 2 of the manuscript). Accordingly, the dynamic shape factors (DSF) of 100 nm, 150 nm, and 200 nm BC particle are 2.11-2.30, 2.54-2.55 and 2.44-2.85, respectively. This suggests the highly fractal morphology of the fresh BC particles, and the consistency of fresh BC properties in different experiment. Also, as the BC particles formed from the burner were heated to 300 °C before introduced into the chamber, most of the organic coatings were removed. The combined measurement of particle size distribution, density and chemical composition (AMS) shows that organics accounted for less than 10% of fresh BC mass concentration. The single scattering albedo (SSA) of fresh BC particles was only 0.1, further confirming the purity of BC particles. We modified two places in the manuscript to make this clear:

Line 132, "The combined measurement of particle size distribution, density and chemical composition exhibited that organics accounted for less than 10% of fresh BC mass concentration. Moreover, the single scattering albedo (SSA) of fresh BC particles was only 0.1, further confirming that few organic coatings exited on fresh BC particles."

Line 238, "For three initial D_m of 100 nm, 150 nm, and 200 nm, the ranges of effective density of fresh BC particles in each experiment were 0.43-0.50 g cm⁻³, 0.34-0.34 g cm⁻³, and 0.24-0.32 g cm⁻³, respectively, indicating highly fractal BC aggregates (Zhang et al., 2008).

The small variation for particles with the same D_m also demonstrates the consistency of fresh BC morphology in different experiments."

(2) Evaluating the hygroscopic changes of EC has been a challenge. It has been proposed (for example, Qiu et al. ES&T, 2012) that the hygroscopicity of the coating materials can be measured or estimated to reflect the real hygroscopic growth of the EC particles. The authors may consider incorporating such approach to covert the apparent HGF into the real HGF. Furthermore, I wonder if similar approach can be developed for the CCN data to estimate the real _ values of aged EC particles.

The reviewer made a very good point here. As discussed in the manuscript, the apparent HGF may underestimate the hygroscopicity of BC particles when particles are not spherical shape. We attribute this the major reason of the difference between $\kappa_{\rm CCN}$ and $\kappa_{\rm HTDMA}$ in the beginning of experiment. However, duo to the fast growth of BC particles in our study, the morphology of BC particles shifted from fractal to spherical shape within a very short time. Therefore, the apparent HGF at the end of experiments can represent real HGF of particles.

Besides, in the CCN data processing, the particle mass equivalent diameter was used as the particle dry diameter in the calculation of κ _CCN value by Equ (10). Therefore, the κ _CCN should reflect the real hygroscopicity of BC particles from this point of view.

(3) What was the final size of the new particles from nucleation in the chamber? It could be a concern if the nucleation in the chamber was fast and the new particles grew too fast into the size range of the aged soot particles and hence interfere with the property measurements.

Nucleation occurred in the chamber during several experiments of this study. Moreover, the newly formed particles grew along with BC particles inside the chamber. However, as shown in the Figure below, the difference in the peak diameter between BC particles and nucleation particles was always larger than 80 nm for all experiments. Therefore, the influence of nucleation particles to the measurement of BC properties was insignificant.

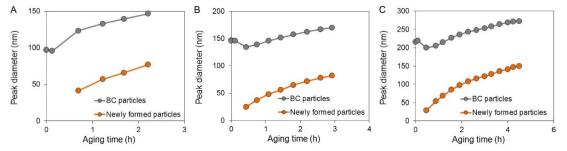


Figure 1. Peak diameter of BC and nucleation particles inside the chamber during experiments with initial BC diameter of 100 nm (A), 150 nm (B), and 200 nm, respectively.

Aging and hygroscopicity variation of black carbon particles in Beijing

2 measured by a quasi-atmospheric aerosol evolution study (QUALITY) chamber

Jianfei Peng^{1, 2*}, Min Hu^{1, 3*}, Song Guo^{1,2}, Zhuofei Du¹, Dongjie Shang¹, Jing Zheng¹, Jun
 Zheng², Limin Zeng¹, Min Shao¹, Yusheng Wu¹, Don Collins², Renyi Zhang^{1,2*}

⁵ ¹State Key Joint Laboratory of Environmental Simulation and Pollution Control, College of

6 Environmental Sciences and Engineering, Peking University, Beijing, China 100871

²Department of Atmospheric Sciences, Texas A&M University, College Station, Texas, USA
 77843

³Beijing Innovation Center for Engineering Sciences and Advanced Technology, Peking
 University, Beijing, China 100871

11 *To whom correspondence should be addressed. E-mail: pengjianfeipku@gmail.com (Jianfei

12 Peng); minhu@pku.edu.cn (Min Hu); renyi-zhang@geos.tamu.edu (Renyi Zhang)

Abstract. Measurements of aging and hygroscopicity variation of black carbon (BC) particles 13 in Beijing were conducted using a 1.2 m³ quasi-atmospheric aerosol evolution study 14 (QUALITY) chamber, which consisted of a bottom flow chamber where ambient air was pulled 15 16 through continuously and an upper reaction chamber where aging of BC particles occurred. Within the reaction chamber, transmission of the solar ultraviolet irradiation was approximately 17 18 50% - 60%, wall loss of primary gaseous pollutants was negligible compared with the replenish rate by gas exchange, and BC exhibited a half-lifetime about 3-7 hours. Typically, equilibrium 19 for the primary gases, temperature, and relative humidity between the reaction chamber and 20 21 ambient air was established within 1 hour. Rapid growth of BC particles was observed, with an average total growth of 77 ± 33 nm and average growth rate of 26 ± 11 nm h⁻¹. Secondary 22 organic aerosolaerosols (SOA) accounted for more than 90% of the coating mass. The O/C 23 ratio of SOA was 0.5, lower than the ambient level. The hygroscopic growth factor of BC 24 25 particles decreased slightly with an initial thin coating layer because of BC reconstruction, but subsequently increased to 1.06-1.08 upon further aging. The κ (kappa) values for BC particles 26 and coating materials were calculated as 0.035 and 0.040 at the subsaturation and 27 supersaturation conditions, respectively, indicating low hygroscopicity of coated SOA on BC 28 particles. Hence, our results indicate that initial photochemical aging of BC particles does not 29 30 appreciably alter the particle hygroscopicity in Beijing.

31

1

32 **1 Introduction**

Atmospheric aerosols undergo continuous and complicated transformation during their 33 residence time in the atmosphere. The aging of aerosols is likely resulted from both physical 34 (i.e., coagulation, condensation, equilibrium partitioning, and evaporation) and chemical (i.e., 35 photochemical gas-phase oxidation and multi-phase reactions) processes (Zhao et al., 2006; 36 Qiu et al., 2013; Zhang et al., 2015). Also, there are typically large variations in the particle 37 properties (i.e., size, mass, chemical composition, morphology, and optical and hygroscopic 38 parameters) during aging, significantly influencing the aerosol impacts on visibility, human 39 40 health, weather, and climate (Jacobson, 2001; Guo et al., 2014). A better understanding of the aging process of aerosols in the atmosphere is critical in atmospheric and climate research. 41

42 For example, the scientific interest in the climate effects of black carbon (BC) has 43 remained, since BC is the strongest absorber of visible solar radiation (Wang et al., 2013). BC solar absorption represents a central issue in climate change research, since the synthesis of 44 satellite, in situ, and ground observations shows that the global solar absorption (direct 45 radiative forcing or DRF) by BC is as much as $0.9 \text{ W} \cdot \text{m}^{-2}$, second only to that of CO₂ (Jacobson, 46 2001; Bond et al., 2013; IPCC, 2013). BC also represents an important component of air 47 48 pollution for large parts of the world (Zhang et al., 2015). The properties of BC are considerably modified during aging, including the size, mass, morphology, and optical and hygroscopic 49 50 parameters (Khalizov et al., 2009; Xue et al., 2009a). Enhanced light absorption of BC particles during aging not only contributes to atmospheric stabilization and exacerbation of haze 51 formation, but also imposes large positive radiative forcing on climate (Peng et al., 2016). 52 Furthermore, the variation in hygroscopicity during aging also regulates the lifetimes of BC 53 particles. Hygroscopic particles serve efficiently as cloud condensed nuclei (CCN), affecting 54 55 the formation, longevity and albedo of clouds (Yuan et al., 2008; Wang et al., 2011). Also, deposition of BC particles, via in-cloud scavenging and wet deposition, depends highly on the 56 57 particle hygroscopicity (Bond et al., 2013). In addition, the hygroscopicity also affects the aqueous-phase reactions of atmospheric pollutants (Ervens et al., 2011; Wang et al., 2016). 58 Previous studies using hygroscopic tandem differential mobility analyzer (H-TDMA) 59 instruments have shown that coating of hydrophilic materials significantly increases the 60 hygroscopic growth factor of BC particles (Saathoff et al., 2003; Khalizov et al., 2009; Guo et 61 al., 2016). The ability of BC particles to form CCN also enhanced after coating of hydrophilic 62 materials (Kuwata et al., 2007; Tritscher et al., 2011; Ma et al., 2013; Wittbom et al., 2014). 63 64 The activation supersaturation depends on the particle size, hygroscopicity of coating materials, and the coating thickness (Ma et al., 2013). The coating materials in the previous experiments 65 include sulfuric acid (Zhang and Zhang, 2005; Khalizov et al., 2009), oxidation products from 66 biogenic and anthropogenic hydrocarbon species (Saathoff et al., 2003; Ma et al., 2013; 67 68 Khalizov et al., 2013), or secondary organic aerosols (SOA) from single emission source (Tritscher et al., 2011). However, there still exist uncertainties for parametrization of the BC 69 lifetime in atmospheric models, because of insufficient constraints on the hydrophobic to 70 hydrophilic conversion of BC particles under variable ambient conditions. 71

Atmospheric field measurements have been performed to evaluate aging of particles on
different platforms, e.g., ground, aircraft, and cruise (Moffet and Prather, 2009; DeCarlo et al.,
2010; Peng et al., 2014; Liu et al., 2015) and over different spatial scales (intensive campaigns)

or long-term measurements). Typically, a wide variety of state-of art instruments are employed to characterize the changes of the chemical and physical properties of aerosols. On the other hand, field measurements at fixed sites are affected by transport, local emissions, and chemistry, and quantification of the particle parameters during aging involves complex decoupling of the various processes (Peng et al., 2016). In particular, it is challenging to isolate the chemical processes from those related to meteorology (i.e., transport and mixing) and emissions.

81 The methods of environmental chambers or reactors have been widely employed in atmospheric chemistry research, including photochemical oxidation of volatile organic 82 compounds (VOCs) (Zhang et al., 2000), formation and growth of aerosols (Claeys, 2004; 83 Kalberer, 2004), nucleation of nanoparticles (Zhang et al., 2004; Wang et al., 2010; Zhang et 84 85 al., 2012), aging of BC particles (Zhang et al., 2008), and cloud formation (Ruehl et al., 2016). Dependent of the scientific objectives, the designs of environmental chambers and reactors 86 vary considerably (Zhang et al., 2015). However, few of the previous experimental methods 87 88 are able to characterize have characterized the evolution of aerosols under the ambient conditions. 89

In this study, we present measurements of aging and hygroscopicity variation of BC particles in Beijing using a quasi-atmospheric aerosol evolution study (QUALITY) chamber (Reed, 2010; Peng et al., 2016). The performance of the QUALITY chamber for mimicking the ambient gaseous concentrations (i.e., the wall loss, and gas mixing rate), ultraviolet transmission, and meteorology parameters (i.e., temperature and relative humidity, RH) has been evaluated.

96 2 Experimental method

97 The 1.2 m³ QUALITY chamber was employed to study BC aging under ambient 98 conditions (Fig. 1). The two-layer chamber was comprised of an inner layer of 0.13 mm 99 polytetrafluoro ethylenePerfluoroalkoxy alkane (PFA) Teflon and an outer rigid 5.6 mm thick 100 acrylic shell (Cyro Industries Acrylite, OP-4). Both Acrylite OP-4 Acrylic and PFA Teflon 101 allowed for efficient ultraviolet (UV) transmission in UV-B (280-315 nm) and UV-A (315-400 102 nm) ranges. When exposed to sunlight, the UV light transmitted through the chamber wall and 103 initiated photochemical reactions inside the chamber.

The two individual subdivisions of the QUALITY chamber included a bottom flow 104 chamber, where ambient air was pulled through continuously over each experiment, and an 105 upper reaction chamber, where aging of BC particles occurred (Fig. 1). The two chambers were 106 separated by a 5 µm thick semi-permeable expanded polytetrafluoroethylene (e-PTFE) 107 membrane with high degrees of chemical resistivity, microporosity, nonpolarity, and thermal 108 stability. Gaseous species penetrated the membrane by either bulk gas flow or diffusion. The 109 permeability of the e-PTFE was greater than 90 % for nearly all the tested volatile organic 110 components (Fig. S1) and other gaseous pollutants, i.e., SO₂, NO_x, O₃ and CO. Particles, on 111 the other hand, were blocked from penetration into the reaction chamber. The filtration 112 efficiency of the e-PTFE membrane was measured to be over 99.5% for particles larger than 113 114 15 nm. During each chamber experiment, ambient air was pulled through the flow chamber continuously and gases in lower chamber permeated through the membrane into the reaction 115 chamber. Hence, an environment that continuously captured the ambient gas concentrations 116

117 without the presence of ambient particles was created inside the reaction chamber. Since the 118 chamber was continuously exposed to ambient gas concentrations during experiments, gases 119 lost due to reaction, deposition or adsorption to the seed aerosols within the reaction chamber 120 were steadily replenished by the exchange with the flow chamber. Several sampling ports were 121 set at the side of the reaction chamber for injection of seed particles or sampling during 122 experiment.

Seed particles were introduced in the reaction chamber via an injection line (Fig. S2). To 123 investigate the growth of BC particles under ambient condition, monodisperse BC particles 124 were injected into the QUALITY chamber. BC particles were generated by incomplete 125 combustion of propane fuel in a custom-made laminar diffusion burner (Santoro et al., 1983; 126 Qiu et al., 2012). The aerosol stream sequentially passed through a 300°C heater to evaporate 127 the semi-volatile organic compounds in the particle phase, a Nafion dryer to remove excess 128 moisture in the flow, and four one-meter-long cylinder cylinders containing both alumina 129 spherules coated with potassium hypermanganate and activated carbon to remove all the 130 gaseous pollutants (i.e., VOCs, H₂S, SO₂, NO_x, O₃). The combined measurement of particle 131 132 size distribution, density and chemical composition exhibited that organics accounted for less 133 than 10% of fresh BC mass concentration. Moreover, the single scattering albedo (SSA) of fresh BC particles was only 0.1, further confirming that few organic coatings exited on fresh 134 BC particles after treatment. The measured removal efficiency in the cylinders for SO₂, NO 135 and NO₂ were 99.2%, 100%, and 99.9%, respectively. The aerosol stream was then introduced 136 through an ionizer and into a differential mobility analyzer (DMA, model 3081, TSI, Inc.) with 137 138 stable voltage to create a monodisperse BC particles flow.

A suite of high time resolution state-of-the-art aerosol instruments simultaneously 139 140 measured a comprehensive set of BC properties throughout the BC aging process- (Table S1). The particle diameter, mass, chemical composition, hygroscopicity and ability as cloud 141 142 condensation nuclei (CCN) were measurement by a scanning mobility particle sizer (SMPS), a differential mobility analyzer-aerosol particle mass analyzer (DMA-APM) system, a high-143 resolution time-of-flight aerosol mass spectrometer (HR-ToF-AMS), a humidified tandem 144 differential mobility analyzer (HTDMA) system, and a could condensation nuclei counter 145 (CCNC), respectively. Detail information of the instruments is provided in the supplementary 146 material and previous publications (DeCarlo et al., 2006; Khalizov et al., 2009). Specific 147 measurement procedures in this study are discussed below. 148

DMA-APM measurement. The DMA-APM was used to measure the effective density of BC 149 particles (Pagels et al., 2009). Before any DMA-APM measurement, a SMPS scan was made 150 to obtain the size distribution of particles inside the chamber. The particle size distribution was 151 then fitted with a lognormal Gaussian distribution to derive the peak diameter. During a DMA-152 APM measurement, the aerosol flow passed through DMA with a fixed voltage to select 153 particles with a fixed diameter. The APM then measured the mass distribution of the selected 154 particles with the same diameter, and the effective density of these particles was obtained by 155 fitting the mass distribution with a normal Gaussian distribution. 156

157 **DMA-CCN measurement.** Similar to the density measurements, CCN activation was also 158 measured on the basis of monodisperse particles with a peak diameter. A DMA with a fixed

- voltage selected particles with a peak diameter. Both a CPC and a CCN counter were placed in
- 160 parallel after the DMA to simultaneously measure the total particle number concentration (N_{cn})
- 161 as well as the activated number concentration (N_{ccn}) at a fixed supersaturation. The activation
- 162 fraction of the BC particles with peak diameter is calculated as:

163
$$f_{CCN} = \frac{N_{CN}}{N_{CCN}}$$
(1)

164 Several gradients of supersaturation were set for the CCN counter, with each one being 6 165 min. This method yielded a steeper curve of the CCN activation rate, which was employed to 166 estimate the particle diameter with 50% activation fraction (D_{50}) and the kappa value.

Prior to each experiment, the QUALITY chamber was flushed by zero air for more than 167 40 hours to ensure a clean condition and covered with two layers of anti-UV cloth to shield it 168 from sunlight. In the beginning of each experiment, monodisperse BC particles were 169 introduced into the chamber. The injection of BC particles typically lasted for 1 to 2 hours. 170 During the injection period, zero air passed through the bottom chamber continually to remove 171 any possible remaining gaseous pollutant that were removed by the activated carbon. After the 172 injection, ambient air was pulled through the bottom chamber at a flow rate of about 50 L min⁻ 173 ¹ for at least half an hour in order to produce a quasi-ambient condition inside the chamber. 174 Finally, the anti-UV cloth was removed, and BC particles underwent aging inside the reaction 175 chamber. A charged zero air stream continuously passed through the space between the two 176 chamber layers to reduce particle wall loss. Various properties of BC particles, including the 177 particle diameter, mass, chemical composition, hygroscopicity, and optical coefficients, were 178 179 simultaneously measured by a suite of state-of-art aerosol instruments in every 0.5 - 1 hour. 180 Ambient particles and chamber particles were measured alternately every 30 min. The aging experiments lasted for about 2-6 hours depending on the initial BC concentrations and ambient 181 conditions. 182

The BC aging experiments were conducted from August 18th to October 17th, 2013 at an urban site (<u>PKUERS</u>) located on the campus of Peking University in the northwestern Beijing (39.99°N, 116.31°E) (Hu et al., 2012).).

186 **3.** Characterization and Validation of the QUALITY chamber

187 **3.1 Wall loss of gases and aerosols**

188 To evaluate the wall loss of both particles and gases in the QUALITY chamber, different 189 gaseous pollutants and particles were introduced into the chamber separately and the decay of 190 their concentration inside the chamber was measured by gas analyzers and SMPS, respectively. 191 All of the ports connected to the ambient air were closed to ensure an enclosed system in the 192 reaction chamber.

Particles with different chemical composition exhibited different wall loss rates. Monodispersed BC particles with different diameters showed a small half-lifetime (τ) of about 4-7 hours (Reed, 2010). Aerosol nucleation also occurred inside the chamber likely from organic species (Zhao et al., 2009), which corresponded to a half time of about 3.5 hours (Fig. S3), because the nucleated particles inside the chamber was neutral with a slow electrostatic 198 loss to the wall.

Gas species had a longer residence time inside the chamber. Toluene and isoprene did not show obvious wall loss during a two-day experiment (Reed, 2010). O_3 , SO_2 and NO_x decreased by 50% inside the chamber after more than 20 hours, suggesting slow loss. Since the loss rate of these primary gaseous pollutants was much slower than the gas exchange rate between the chamber, the loss of gases was replenished by the exchange with the flow chamber.

204 **3.2** UV transmission

The QUALITY chamber contained two layers of walls, an acrylic shell layer and a PFA Teflon layer. Since the QUALITY chamber used sunlight as the photochemical origin, the transmission spectra of the two-layer walls was of great important for the photochemical reactions inside the reaction chamber.

The transmission efficiencies of each material were measured using a Fourier transform 209 infrared spectroscopy (FTIR) system.an ultraviolet-visible (UV-Vis) spectrophotometry 210 (PerkinElmer Inc., model 552). As shown in Figure 2, the Teflon film exhibited stable 211 212 transmission efficiency of about 60% in the focused wavelength range. The Acrylic shell, however, showed very low transmission efficiency when the wavelength range was shorter 213 than 270 nm, and high transmission efficiency (nearly 90%) when the wavelength range was 214 longer than 300 nm. In general, approximately 60% of the UVA irradiation (315-400 nm range) 215 and 50% of the UVB irradiation (280-315 nm range) penetrated through the chamber walls, 216 allowing photochemical processes to take place in the upper reaction chamber. 217

A NO₂-photolysis experiment was also conducted to characterize the UV transmission of QUALITY chamber. NO₂ was introduced inside the chamber at a clean and sunny day. Reactions among NO₂, NO and O₃ will occur as the following equations:

(2)

(3)

(4)

- 221 $NO + O_3 \rightarrow NO_2 + O_2$
- 222 $NO_2 \xrightarrow{hv} NO + O$
- $223 \qquad 0 + 0_2 \rightarrow 0_3$

By simultaneously monitoring the concentrations of NO_2 , NO and O_3 inside the chamber, the photolysis constant of NO_2 , $J(NO_2)$, inside the chamber was estimated. The photolysis constant of NO_2 inside the chamber was on average 55% of that in the ambient air, in agreement with the transmission spectra measurement, further confirming that the two-layer chamber walls allowed 50% - 60% of solar irradiation in the UV range into the reaction chamber.

229 **3.3** Exchange of gases between Gas concentrations in the reaction chamber

To investigate the VOCs concentration inside the reaction chamber after the injection of
 BC particles, both the chamber and ambient air was sampled with VOC canisters just before
 the BC aging experiment started. These canisters were then analyzed by a gas chromatography mass spectrometer / flame ionization detector (GC-MS/FID, HP inc.) system (Liu et al., 2008).
 The concentrations of VOCs containing 4 or more carbons are illustrated in Figure 3. Slightly
 higher concentrations of several VOCs in the QUALITY chamber, e.g., n-butane, n-pentane,

toluene, were observed compared with those in the ambient air, duo to the co-injection of a
small amount of VOCs together with BC particles into the chamber. Nevertheless, the average
increase of the VOC concentrations was only 16% or 0.1 ppb for all focused VOCs species,
with the largest increase of 35% or 0.36 ppb, suggesting the insignificant influence of soot
burner on VOCs concentrations and SOA formation in the chamber.

Additional experiments were conducted to characterize the exchange time scale for gases 241 in the QUALITY chamber. The chamber was firstly cleaned and flushed with zero air for 40 242 243 hours. Ambient air was then pulled through the bottom chamber (Fig. 34), and the concentrations of gaseous pollutants, including O₃, NO_x, CO, and SO₂, were measured 244 alternatively in the ambient air and in the upper chamber. In the beginning of the experiment 245 the concentrations of all gaseous species were lower than those in the ambient air. After ambient 246 247 air was pulled through the bottom chamber (labeled as the black dots in Fig. 34), the concentrations of the gaseous pollutants in the reaction chamber increased sharply. For 248 example, the CO concentration inside the chamber was approximate 70% of that the ambient 249 250 concentration after 30-min mixing. The concentrations inside the chamber and ambient air 251 exhibited little difference after 1-hour mixing (Fig. <u>3c4c</u>). Hence, the QUALITY chamber well 252 replicated the ambient gas concentrations. The gas exchange rate between the bottom and reaction chambers was calculated to be approximately 0.06 min⁻¹. 253

254 **3.4** Temperature and RH

The greenhouse effect for an outdoor chamber typically increases the temperature and decreases the RH inside the chamber. For the QUALITY chamber, however, heat produced by the greenhouse effect inside the chamber was effectively taken away as the ambient air continuously passed through the bottom chamber and exchanged with air in the upper chamber. As illustrated in Figure 4<u>5</u>, there was little difference in temperature or RH inside and outside the chamber, when the chamber experiments lasted for more than 1 hour, suggesting that the QUALITY chamber effectively captured the ambient temperature and RH.

262 **3.5** Sulfuric acid production

Though gaseous pollutants such as SO₂, O₃, NO_x and VOCs penetrated into the reaction 263 chamber from the bottom chamber through the semi-permeable membrane, low volatile and 264 sticky gases, i.e., sulfuric acid, were unlikely to penetrate through the membrane (Fortner et al., 265 2004). To characterize sulfuric acid production inside the chamber, we conducted a special 266 experiment by pulling ambient air into the bottom chamber, while injecting SO₂ directly into 267 the reaction chamber. The experiment was conducted around noon on a clean day when the O₃ 268 concentration was around 50 ppb. A custom-made atmospheric pressure-ion drift chemical 269 ionization mass spectrometry (AP-ID-CIMS) (Fortner et al., 2004; Zheng et al., 2010) was used 270 271 to directly measure the concentration of gaseous sulfuric acid. Fig. 56 shows a good correlation between SO₂ and gaseous H₂SO₄ inside the chamber, suggesting that the QUALITY chamber 272 well simulated the formation of low volatile gaseous species and hence the photochemical 273 274 processes.

275 4. BC evolution in the QUALITY chamber

276 Time series of the ambient PM, gas, and meteorology parameters of ambient air during

chamber experiment period are illustrated in Figure 67. Except for the last experiment conducted on October 17th, all experiments were conducted between August 18th, 2013 and September 21st, 2013. During this period in Beijing, strong photochemical reactions and frequent heavy pollution events occurred (Huang et al., 2010; Zhao et al., 2013). Temperature and RH during this period (August 18th to September 21st) varied from 24°C to 38°C and from 20 % to 90 %, respectively. The average concentration of PM_{2.5}, SO₂ and NO_x were $60 \pm 45 \mu g$ m⁻³, 3.2 ± 2.6 ppb and 33.9 ± 20.9 ppb, respectively.

The red shaded areas in Figure 67 represent the period of the nine chamber experiments. 284 In general, chamber experiments were conducted in the afternoon of relatively clean and sunny 285 days, when strong solar radiation led to fast photochemical reactions. There were two 286 experiments conducted under polluted days, e.g., the experiments on October 22rd and 287 September 11th. Table 1 summaries the conditions of the experiments. Totally, 10 BC aging 288 experiments were conducted, including four experiments using BC particles with initial 289 mobility diameter of 100 nm, three experiments using 150 nm particles, and three using 220 290 nm BC particles. The average concentrations of PM_{2.5} and NO_x over each chamber experiment 291 were only 9 to 69 μ g m⁻³ and 9 to 41 ppb, respectively. The concentrations of VOCs, such as 292 toluene and m/p-xylene were relatively low during most of the experiments compared with 293 severe pollution episodes in Beijing (Guo et al., 2014). $J(O^{1}D)$ and O_{3} exhibited higher values 294 during the chamber experiments. The average $J(O^{1}D)$ values and O_{3} concentrations ranged 295 from 3.2×10^{-6} s⁻¹ to 21.1×10^{-6} s⁻¹ and 26 to 92 ppb, respectively. 296

297 **4.1 BC growth**

To quantify the growth of BC particles, several parameters were used to describe the properties of BC particle, including the effective density, mobility diameter and mass equivalent diameter. Material density (ρ_m) is the average density of the solid and liquid material in the particle. Assuming that the volume of the species coexisting in an individual particle does not change upon mixing, the density of an internally mixed particle is calculated using the material densities and mass concentrations of particulate constituents (DeCarlo et al., 2004; Pagels et al., 2009),

305
$$\rho_m = \frac{\sum_{species} m_i}{\sum_{species} V_i} = \frac{\sum_{species} m_i}{\sum_{species} \frac{m_i}{\rho_i}} = \frac{\sum_{species} MC_i}{\sum_{species} \frac{MC_i}{\rho_i}}$$
(5)

306

where ρ_i is the material density of species *i*, V_i is its volume, m_i is its mass, and MC_i is its mass concentration. This approach is based on the assumption that there is no void space enclosed within the particle envelope. Hence, the material density is larger than the true particle density with internal voids in particles. In this study, ρ_m is calculated from the chemical composition of coating materials measured by AMS. A value of 1.35 for the material density of SOA formed during chamber experiment was obtained by directly measuring the density of newly form particles inside the chamber via DMA-APM system.

Effective density (ρ_{eff}) is defined as the ratio of the measured particle mass (m_p) to the particle volume calculated assuming a spherical particle with a diameter equal to the measured mobility diameter (D_m) (DeCarlo et al., 2004; Pagels et al., 2009; Xue et al., 2009b):

317
$$\rho_{eff} = \frac{6m_p}{\pi D_m^3} \tag{6}$$

318

In this study, m_p of BC particles was measured by the APM and mobility diameter (D_m) was measured by the DMA. The effective density reflects the information on both particle density and shape. If particles are spherical in the absence of internal void, the effective density equals the material density. If particles are non-spherical, the calculated volume and volume concentration are larger than the true values, and the effective density is less than true particle and the material density.

325 Figure $\frac{78}{2}$ exhibits the change of particle density and diameter in three typical BC aging experiments using BC particles with the initial diameter of 100 nm, 150 nm, and 220 nm 326 (experiment #4, #5, and # 9 in Table 1, respectively). The average $PM_{2.5}$ 327 concentration concentrations in these experiments was were 40, 27, and 12 µg m⁻³, respectively, 328 suggesting relatively clean conditions during the experiments. In all three experiments, aging 329 of BC particles occurred between 13:00-14:00 and 17:00 in the afternoon. The highest J(O¹D) 330 value varied from 1.7 to 2.4×10^{-5} s⁻¹ and decreased generally over the experiment period. 331 Average O₃ concentrations during the three experiments were 68, 83 and 54 ppb, respectively, 332 indicating strong oxidation during the experiment periods. 333

The For three initial ρ_{eff} values of BC particles were 0.46, 0.34 and 0.25 g cm⁻³ for particles 334 335 with the initial diameter D_m of 100 nm, 150 nm, and 220 nm, the ranges of effective density of fresh BC particles in each experiment were 0.43-0.50 g cm⁻³, 0.34-0.34 g cm⁻³, and 0.24-0.32 336 <u>g cm⁻³</u>, respectively, indicating <u>highly</u> fractal BC aggregates with chain-like branches (Zhang 337 et al., 2008). The small variation for particles with the same D_m also demonstrated the 338 consistency of fresh BC morphology in different experiments. After aging of 1 hour, ρ_{eff} of BC 339 particles in the three experiments increased to approximately 1.2 g cm⁻³, suggesting that 340 formation of the secondary components changed the morphology from chain-like BC particles 341 342 into a more compact shape (Zhang et al., 2008; Peng et al., 2016). The morphology change was further confirmed by a decrease of D_m , particularly for larger BC particles that were more 343 fractal than smaller BC particles. 344

As the D_m is largely influenced by the particle morphology, we utilized the parameter of mass equivalent diameter (D_{me}) to describe the growth of particles. Based on the mobility diameter (D_m), material density (ρ_m) and effective density (ρ_{eff}), the D_{me} is calculated assuming that particles are compact and with a spherical morphology (DeCarlo et al., 2004):

349
$$D_{me} = \sqrt[3]{\frac{\rho_{eff}}{\rho_m}} D_m \tag{7}$$

350

The change in the mass equivalent diameter (ΔD_{me}) during BC aging is defined as the total coating thickness, and the ratio of the total coating thickness to the initial mass equivalent diameter ($\Delta D_{me}/D_{me,0}$) is defined as the coating fraction.

The initial D_{me} of fresh BC particles with initial D_m of 100, 150 and 220 nm were 61, 84 354 and 114 nm, respectively. In contrast to the mobility diameter, D_{me} increased continuously 355 during the entire experiment. After 3-4 hours, D_{me} in the three experiments increased to 133, 356 169 and 197 nm, respectively (Fig. 78), with the average growth rates of 19, 29 and 31 nm h⁻¹. 357 Higher growth rates in D_{me} occurred around noontime, when the J(O¹D) value was higher and 358 the photochemical reaction was stronger. On the other hand, much less growth rate was 359 observed during late afternoon or with cloud coverage (As shown in Fig. 7e8c at 15:00), 360 361 indicating that the growth was driven by photochemical reactions.

The increases of the particle density and diameter in all the experiments are summarized in Table 2. Fast aging of BC particles occurred in all experiments. The total growth of D_{me} ranged from 40 nm to 152 nm within 3-6 hours, with an average growth of 73 nm. The average growth rate was 26±11 nm h⁻¹, demonstrating large secondary production under the ambient conditions in Beijing. The largest growth rate ($\Delta D_{me} = 152$ nm) was observed in experiment #8, when solar irradiation was the strongest among all experiments (Table 2).

Correlation analysis was made between the average growth rate of BC particles $(\Delta D_{me}/\Delta t)$ with O₃, PM_{2.5}, *J*(O¹D), and temperature during the different experiments (Fig. S4). The growth rate of BC particles exhibits no correlation with O₃ concentration (R²=0.00), weak negative correlation with PM_{2.5} concentration (*R*² = 0.25), and strong positive correlations with *J*(O¹D) (*R*² = 0.80) and temperature (*R*² = 0.67), indicating the importance of photochemical production on the BC coating materials.

4.2 Chemical composition of coating materials

Particle composition measurements by AMS during chamber experiments reveal a majority of coating materials (above 90%) as SOA (Fig. 78). The concentration of SOA inside the chamber reached up to 9 μ g m⁻³ in several experiments, suggesting fast formation of SOA via gas phase oxidation of VOCs. The SOA formation in Beijing is likely attributed to a large amount of anthropogenic aromatic VOCs (Peng et al., 2017).

The elemental compositions of OA inside the chamber, i.e., the oxygen to carbon (O/C)380 ratio and the hydrogen to carbon (H/C) ratio, were calculated based on the updated ambient 381 calibrations (He et al., 2011; Canagaratna et al., 2015). The H/C and O/C ratios of organics for 382 383 coating on BC particles exhibit notable trends during the aging process. Figure 8A9A shows an example of the evolution of H/C and O/C ratios in experiment #8. The data were corrected 384 for the CO₂ concentration in the chamber, which were introduced into the chamber with BC 385 particles and influenced the abundance of m/z = 28 and 44 in the AMS mass spectra. The H/C 386 ratio decreased from 1.73 to 1.45 over six hours. Accordingly, the O/C ratio increased from 387 0.32 to 0.50 during the same time, revealing that further oxidation of SOA occurred in the latter 388 part of the experiment. The lower final O/C ratio (0.5) in the chamber experiment (0.5) than 389 that under the ambient conditions (Hu et al., 2016) implies that there areis oxidation on a longer 390 391 timescale or by the aqueous pathway for the formation of highly oxidized SOA in the ambient air (Zhang et al., 2015). 392

393 Furthermore, the mass spectra of OA inside the chamber shows strong correlation with the

less-oxidized oxygenated organic aerosols (LO-OOA) derived from field measurements in Beijing (Hu et al., 2016), which likely arose from oxidation of aromatic VOCs emitted from vehicles (Peng et a., 2017). The correlation coefficient (\mathbb{R}^2) initially was 0.88 and raised to 0.99 sharply (Fig. **\$B9B**), indicating that the chamber well simulated the formation of LO-OOA.

398 In our study, the secondary inorganic aerosols, i.e., sulfate, nitrate and ammonium, only accounted for less than 10% of the coating materials on BC particles. This is consistent with 399 the previous studies showing that the concentration of organics is much larger than those of 400 sulfate and nitrate during the early stage of haze development in Beijing (Guo et al., 2014). 401 402 The low observed sulfate concentration in this study suggests that the gas phase formation of 403 sulfuric acid was unimportant under our experimental conditions. On the other hand, it has 404 been shown that the aqueous-phase reactions represent the dominate pathway for sulfate formation in Beijing (Guo et al., 2010; Wang et al., 2016). 405

406 NO₂ has a higher reaction coefficient with the OH radical ($8 \times 10^{-12} \text{ cm}^3$ molecule⁻¹ s⁻¹) 407 than SO₂ (Zhang et al., 2015). Nitrate acid formed in the gas phase is transformed into nitrate 408 salts by the reaction with ammonia in the equilibrium process:

(8)

409
$$HNO_3(g) + NH_3(g) \leftrightarrow NH_4NO_3(s)$$

The equilibrium of this reaction is highly depended on ambient temperature and RH (Zheng et al., 2008). In this study, chamber experiments were conducted in the afternoon with high temperature and low RH (Table 1), which shifted the thermodynamic equilibrium to the gas phase.

414 4.3 Hygroscopicity evolution

415 HTDMA measurement

The hygroscopic growth factors (HGF) of particles in each experiment were continuously
 measured by the HTDMA system and corrected for the reference "dry" diameters,

418
$$HGF = \frac{D_{wet,t}/D_{m,t}}{D_{dry,0}/D_{m,0}}$$
(9)

419 where D_m is the mobility diameter of fresh or coated particles at dry condition, D_{dry} is the 420 mobility diameter of particles after experiencing a low humidity (below 30%) cycle in 421 HTDMA, and D_{wet} is the mobility diameter of particles after experiencing a high humidity 422 cycle (87%) in HTDMA.

Figure 910 shows the hygroscopicity variation of BC particles with the initial mobility 423 diameter $(D_{m,0})$ of 100 nm and 150 nm. The measured HGF of 0.999 - 1.004 for fresh BC 424 particles suggests high hydrophobicity, consistent with the previous studies (Khalizov et al., 425 2009; Weingartner et al., 1997). After exposed to sunlight and ambient gaseous pollutants for 426 several hours, the HGF of these BC particles increased to 1.02-1.08 at the end of each 427 428 experiment. The HGF value varied with the total growth (ΔD_{me}) of BC particles, but was 429 constant at the same ΔD_{me} for different experiments (Fig.9<u>10</u>). The final HGF values shown 430 in Figure 910 (1.02-1.08) were much lower than those in previous laboratory studies (Khalizov

et al., 2009;Tritscher et al., 2011) but similar to the low hygroscopic fraction in field
observations (Swietlicki et al., 2008), even for growth of particle size up to 90 nm in our
experiments.

The HGF is affected by many factors, e.g., the particle chemical composition and 434 morphology as well as RH (Qiu et al., 2012). The hygroscopicity of BC particles coated with 435 inorganic components, i.e., sulfuric acid (Khalizov et al., 2009), is significantly higher than 436 that coated by organic compounds (Tritscher et al., 2011). In this study, the major component 437 of the coating substance was LO-OOA with a O/C ratio about 0.5. The low oxygen content of 438 SOA coated on BC particles explains the low hygroscopicity (Jimenez et al., 2009), indicating 439 that coating of BC particles during the early stage haze development in Beijing does not 440 441 considerably increase the particle hygroscopicity.

442 The morphology of BC particles directly affects the HGF. As illustrated in Figure 1011, when the ΔD_{me} was 18 nm and 22 nm for 100 nm and 150 nm BC particles, respectively, the 443 HGF decreased slightly to about 0.99, suggesting that a thin layer of coatings on BC particles 444 decreased the particle diameter, even though a certain amount of water absorbed by BC 445 particles increased the particle mass. The surface tension of the water layer produced an inward 446 447 force on the "chain-like" branches of BC particles, leading to particle reconstruction, and a more compact morphology. Such change was also identified in laboratory studies (Weingartner 448 449 et al., 1997; Tritscher et al., 2011; Qiu et al., 2012). In this study, the BC particles became spherical when ΔD_{me} was 30 nm and 40 nm for particles with initial D_m of 100 nm and 150 nm, 450 respectively (Peng et al., 2016). Therefore, when ΔD_{me} was large, the HGF value was not 451 influenced by reconstruction. 452

453 *CCN measurements and* κ *closure*

The CCN activation faction (f_{CCN}) of BC particles at different supersaturation during two 454 typical experiments is illustrated in Figure 1011. Fresh BC particles were not activated even at 455 very high supersaturation conditions (0.7%). With aging, f_{CCN} rapidly raised to nearly 100% at 456 high supersaturation (0.7% for experiment #4 and 0.6% for experiment #6). After several hours, 457 BC particles became CCN at lower supersaturation. The f_{CCN} at 0.4 supersaturation (Fig. 11a 458 in experiment #4) and 0.3 supersaturation (Fig. 11b in experiment #6) exceeded 50% before 459 the end of these two experiments, suggesting that aging increases the ability of BC particles to 460 become CCN (Wittbom et al., 2014) and a large amount of coatings results in activation at 461 462 lower supersaturation.

To further investigate the hygroscopicity of BC particles and combine the measurements using HTDMA and CCN, we evaluated the hygroscopicity parameter, kappa (κ) (Petters and Kreidenweis, 2007). The approximate relationship between the dry particle mass equivalent diameter (D_{me}), the critical saturation ratio (Sc) and the apparent κ value of particles is describe as:

468
$$\kappa = \frac{4A^3}{27D_{me} \ln^2 S_c}$$
 (10)

$$470 A = \frac{4\sigma_{s/a}M_w}{RT\rho_w} (11)$$

471 M_w is the molecular weight of water, ρ_w is the density of water, $\sigma_{s/a}$ is the surface tension of the 472 solution/air interface, R is the universal gas constant, and T is temperature.

In addition to the supersaturated condition, the κ theory also adopts the form for the subsaturated condition, using the HGF from HTDMA measurement and RH:

$$\frac{RH}{exp\left(\frac{A}{D_{wet}}\right)} = \frac{HGF^3 - 1}{HGF^3 - (1 - \kappa)}$$
(12)

476 where D_{wet} is the wet diameter of particles.

The apparent κ values of BC particles calculated by HTDMA (κ HTDMA) and CCN 477 (κ CCN) are shown in Figure <u>1011</u>. The κ of fresh BC particles was near zero. With aging, 478 SOA coated on BC particles increased the κ HTDMA and κ CCN to approximately 0.04, 479 although the κ HTDMA and κ CCN exhibited difference features. A slightly higher κ CCN 480 481 than κ HTDMA at the beginning of aging was identified, attributed to reconstruction of BC particles after humidified and underestimation of HGF and thus the K HTDMA value. Such a 482 difference between K CCN and K HTDMA was also observed in previous studies (Tritscher et 483 al., 2011; Martin et al., 2013). Nevertheless, the apparent κ values from two both methods were 484 comparable at the end of both experiments. 485

486

487 Assuming that a simple mixing rule is applicable to coated BC particles, the κ for coating 488 materials can be calculated based on the volume fraction of BC and SOA:

489
$$\kappa = \sum_{i} \varepsilon_{i} \kappa_{i} \tag{13}$$

490 where ε_i represents the volume fraction of species *i*. Since the SOA formed inside the 491 chamber was not highly hygroscopic, some of the SOA components might not be able to solve 492 in water droplets (Petters and Kreidenweis, 2008), leading to the underestimation of the κ 493 values of the coating materials in this study.

The κ values of the coating materials were 0.04 at the end of our experiments for both 494 CCN and HTDMA method, much lower than that of ambient aerosols in Beijing (Gunthe et al., 495 2011; Wu et al., 2016) and those of SOA in previous chamber studies (Jimenez et al., 2009; 496 Tritscher et al., 2011; Martin et al., 2013). As discussed above, the coating substances on BC 497 498 particles were mainly SOA formed from photochemical oxidation. The k of SOA depends on the oxidation degree, which is correlated to the O/C ratio (Jimenez et al., 2009). The O/C ratio 499 of the coating SOA was 0.5 in our experiment, likely explaining the low κ with LO-OOA.; 500 Massoli et al. 2010). The O/C ratio of the coating SOA was 0.5 in our experiment, 501 502 corresponding to the O/C ratio of approximately 0.4 in Jimenez et al. (2009) and Massoli et al. 503 (2010) due the utilize of updated AMS calibration method in this study. The κ value of coating materials here is in general slightly lower than those of SOA with similar oxidation degree 504 (O/C ratio) formed from different VOCs precursors (Jimenez et al., 2009; Massoli et al. 2010). 505

As motion above, the κ values of the coating materials might be underestimated duo to the
 ignore of solubility of the SOA products. Also, our experiments represent the aging of BC
 under typical urban condition, where variety of anthropogenic VOCs, i.e., toluene, xylene,
 TMB and small molecular PAHs, contribute significantly to the SOA formation. Some products
 of these VOC precursors may exhibit low hygroscopicity.

511 5 Conclusions

512 In this paper, we present measurements of aging and hygroscopicity of BC particles in Beijing using the QUALITY chamber. The unique two sub-chamber design facilitates the 513 evaluation of aging of BC particles under ambient conditions, by mimicking the ambient 514 gaseous concentrations without the presence of ambient aerosols. High UV transmission 515 efficiency (50-60%) and negligible wall loss of primary gaseous pollutants are shown for the 516 517 chamber performance. The validation experiments demonstrate little differences in the primary gas concentrations, temperature, and RH between the chamber and the atmosphere, suggesting 518 that the chamber captures the evolution of ambient conditions. In addition, our results show 519 sulfuric acid production correlated with SO₂, indicating that the chamber well simulates 520 photochemical-driven formation of low volatile gaseous species by the hydroxyl radical. 521

BC aging experiments were performed using the QUALITY chamber in Beijing. Fast growth of BC particles (on average 26 ± 11 nm h⁻¹) was observed, and SOA was identified as the dominate component of the coating materials on BC particles, while inorganic species, such as sulfate and nitrate, were unimportant under our experimental condition and timescale.

526 The HGF of BC particles exhibited a very low value (1.02-1.08) after several hours aging. 527 A slight decrease of HGF with a thin coating layer indicated reconstruction of BC particles 528 after humidified. Also, a very low kappa value (0.035) for BC particles at both subsaturation 529 and supersaturation conditions were found, with HTDMA and CCN measurements. Hence, our 530 results indicate that initial photochemical aging of BC particles does not appreciably alter the 531 particle hygroscopicity in Beijing.—

532

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Table 1. Summary of ambient conditions for BC aging experiments conducted in Beijing. The PM _{2.5} , gas concentrations and meteorological condition were
averaged from the entire experimental time. D _m and T represent the mobility diameter and temperature, respectively.

No.	Date	Time		Initial Dm	PM _{2.5}	J(O ¹ D)	Gas Concentration (ppb)					Meteorological Conditions	
		Start	End	(nm)	(µg m ⁻³)	(10-6)	Toluene	Xylene	SO ₂	NO _x	O ₃	Τ (°C)	RH (%)
#1	Aug. 18 th	12:41	14:47	95	43	19	0.49	0.13	1.7	8.8	56	35	27
#2	Aug. 22 nd	13:32	16:51	96	69	3.7	3.41	0.78	2	36.1	26	26	69
#3	Sep. 7 th	12:40	14:52	97	12	17.5	0.71	0.17	2.7	10.2	75	30	35
#4	Sep. 9 th	13:13	17:06	97	40	6.3	0.77	0.21	4	11	68	26	50
#5	Sep. 1 st	13:19	16:13	147	27	11.6	0.76	0.19	4.6	19.9	83	31	33
#6	Sep. 11 st	13:50	17:25	147	57	6.1	1.57	0.39	6.7	17.1	92	29	42
#7	Sep. 21 st	15:31	17:41	146	30	2.1	0.75	0.29	2	10.6	90	28	37
#8	Aug. 24 th	11:37	16:06	216	8.8	21.1	0.98	0.3	1.7	15.6	57	36	25
#9	Sep. 5 th	14:06	16:44	220	12	8.3	0.45	0.2	2.2	14.6	54	29	35
#10	Oct. 17 th	12:54	17:13	224	57	3.2	-	-	13.8	41	34	18	30

		D _m		<u> </u>		D _{me}					
No.	Date	Initial (nm)	Final (nm)	Initial (g cm ⁻³)	Final (g cm⁻ ³)	Initial (nm)	Final (nm)	∆D _{me} (nm)	Growth Rate ^a (nm h ⁻¹)		
#1	Aug. 18 th	95	157	0.50	1.35	62	162	100	47		
#2	Aug. 22 nd	96	129	0.46	1.31	61	126	65	20		
#3	Sep. 7 th	97	147	0.45	1.25	62	142	80	36		
#4	Sep. 9 th	97	136	0.43	1.30	61	133	62	19		
#5	Sep. 1 st	147	170	0.34	1.36	85	168	83	29		
#6	Sep. 11 st	147	162	0.34	1.34	84	159	75	20		
#7	Sep. 21 st	146	132	0.34	1.05	84	116	32	15		
#8	Aug. 24 th	216	272	0.32	1.37	123	275	152	34		
#9	Sep. 5 th	220	202	0.25	1.33	114	197	83	31		
#10	Oct. 17 th	224	224	0.24	0.52	117	157	40	11		
Average									26±11		

Table 2. Summary of particle properties for BC aging experiments conducted in Beijing. D_m , ρ_{eff} and D_{me} represent the mobility diameter, effective density and mass equivalent diameter, respectively.

^a The growth rate is calculated using the data between 12:00 and 17:00 for each experiment.

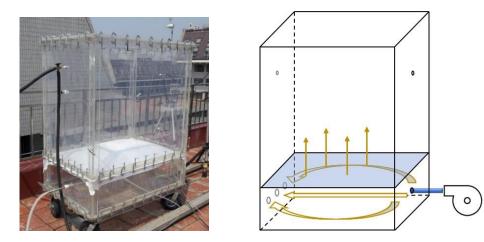


Figure 1. Photo (left) and schematic (right) of the quasi-atmospheric aerosol evolution study (QUALITY) chamber

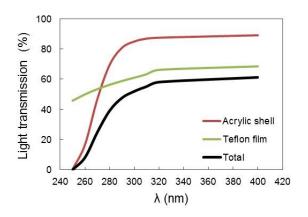


Figure 2. Light transmission spectra of the PFA Teflon film (yellow line), the acrylite shell (red line), and their total transmission in the UV range (black line).

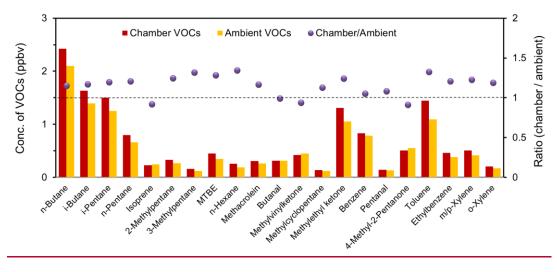


Figure 3. Comparison of VOCs concentrations in the QUALITY chamber (red bar) and in the ambient air (yellow bar) before the start of BC aging experiment. Purple cycles represent the ratio of each species in the reaction chamber to in the ambient air. Only the VOCs species that contains 4 or more carbons with the concentration higher than 0.1 ppb are shown in the figure.

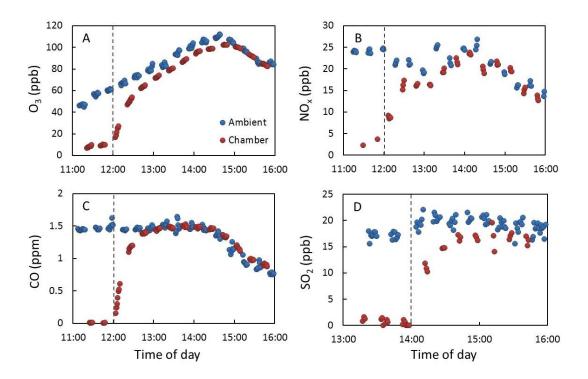


Figure 34. Concentrations of O_3 (A), NO_x (B), CO (C), and SO_2 (D) measured inside the QUALITY chamber (Red circles) and in the ambient air (blue circles). The vertical dashed lines denote the time when the ambient air started to be pulled through the bottom flow chamber and the ambient gases began to exchange into the upper reaction chamber.

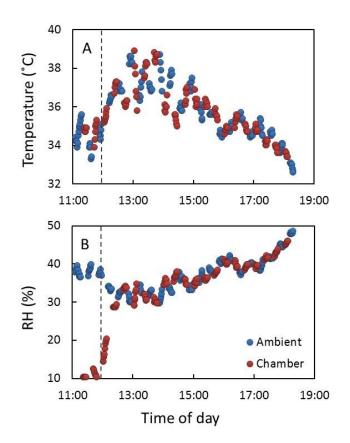


Figure 45. Temperature (A) and RH (B) measured inside the QUALITY chamber (Red circles) and in the ambient air (blue circles). The temperature and RH were measured by probes placed in the sampling tube adjacent to the chamber. The vertical dashed lines denote the time when the ambient air started to be pulled through the bottom flow chamber and the ambient gases began to exchange into the upper reaction chamber.

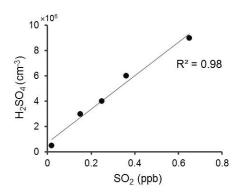
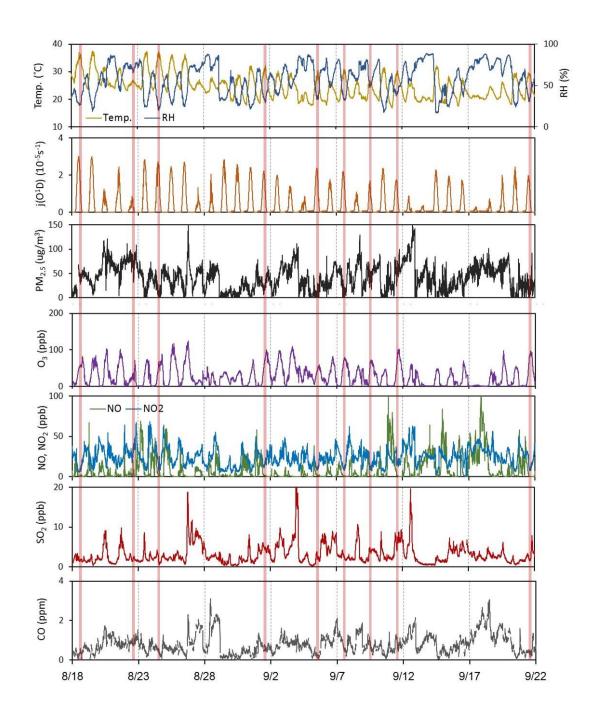


Figure 56. Sulfuric acid concentration as function of SO₂ concentration inside the chamber



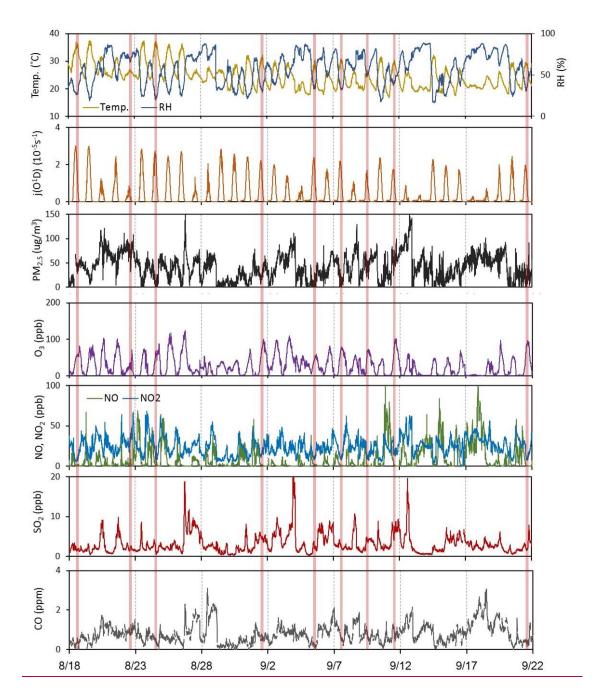


Figure 67. Time series of ambient pollutant concentrations and meteorological parameters during the experimental period in Beijing. Red bars indicate periods when BC aging experiments were conducted.

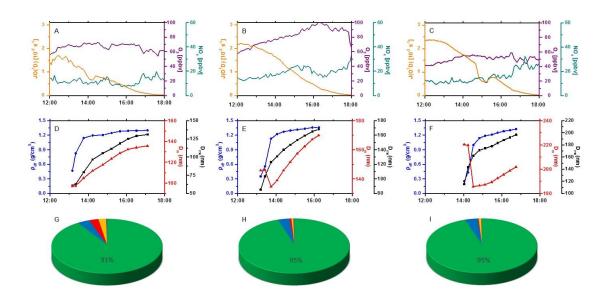


Figure 78. Ambient condition (A, B, C), changes of diameter and density of BC particle (D, E, F), and the chemical composition of coating materials (G, H, I) during three typical aging experiments. A and D correspond to experiment #4; B and E correspond to experiment #5, and C and E correspond to experiment #9. D_m is the peak mobility diameter of BC particles, ρ_{eff} is the best-fit effective density of BC particles with the mobility diameter D_m , D_{me} is the mass equivalent diameter of BC particles, and $J(O^1D)$ represents the measured photolysis rate constant for $O(^1D)$. The colors of green, blue, red, yellow and purple in the pie charts represent organics, nitrate, sulfate, ammonium, and chlorine, respectively. The numbers in figure G, H and I are the mass fraction of organics.

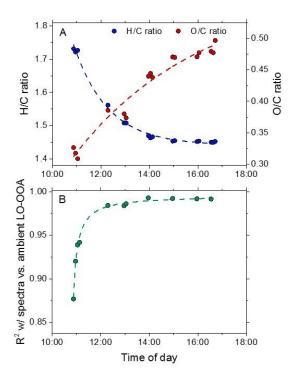


Figure 89. The evolution of organic aerosols inside the chamber during an aging experiment (#8) in Beijing. (A) the H/C and O/C ratios of organics on aged BC particles; (b) the correlation coefficients (R²) between the evolving total OA spectra in chamber experiment and the LO-OOA spectra derived from the Beijing field data set.

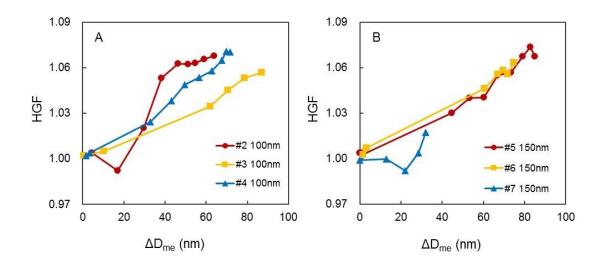


Figure 910. Evolution of hygroscopic growth factors (HGF) of BC particles during aging as a function of ΔD_{me} . (A) three experiments with 100 nm BC particle; (B) three experiments with 150 nm BC particle. Different colors in each figure represent different experiments. Hygroscopicity measurement is not available for experiment #1.

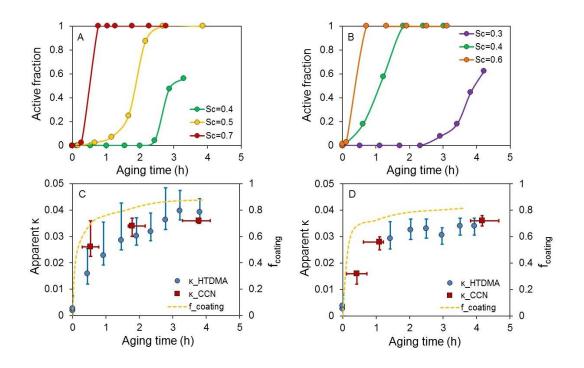


Figure 1011. The active fraction of BC particles under diversified supersaturation (A, B) and the closure of apparent κ for BC particles with initial diameter (C, D) during aging in two typical experiments. A, C are the results from experiment #4 with 100nm BC particles and B, D represents experiment #6 with 150nm BC particles. Red squre and blue circle in C and D represent the apparent κ calculated using CCN counter data and HTDMA data, respectively. Yellow slash line represents the fraction of coating materials on BC particles. The error bars of κ _CCN and κ _HTDMA represent the uncertainty in the calculation.