



1 **Future threat to boreal ecosystem health from wildfire air pollution**

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Abstract

27 Biomass burning is an important source of tropospheric ozone (O₃) and aerosols. These air
28 pollutants can affect vegetation photosynthesis through stomatal uptake (for O₃) and light
29 scattering (for aerosols). Climate change will significantly increase wildfire activity in boreal
30 North America by the midcentury, while little is known about the impacts of enhanced
31 emissions on the terrestrial carbon budget. Here, combining site-level and satellite
32 observations and a carbon-chemistry-climate model, we estimate the impacts of fire emitted
33 O₃ and aerosols on net primary productivity (NPP) over boreal North America. Fire
34 emissions are calculated based on an ensemble projection from 13 climate models. In the
35 present day, wildfire enhances surface O₃ by 2 ppbv (7%) and aerosol optical depth (AOD) at
36 550 nm by 0.03 (26%) in the summer. By midcentury, boreal area burned is predicted to
37 increase by 66%, contributing more O₃ (13%) and aerosols (37%). Fire O₃ causes negligible
38 impacts on NPP because ambient O₃ concentration is far below the damage threshold of 40
39 ppbv. Fire aerosols reduce surface solar radiation but enhance atmospheric absorption,
40 resulting in enhanced air stability and intensified regional drought. The domain of this drying
41 is confined to the North in the present day, but extends southward by 2050 due to increased
42 fire emissions. Consequently, wildfire aerosols enhance NPP by 72 Tg C yr⁻¹ in the present
43 day but decrease NPP by 118 Tg C yr⁻¹ in the future, mainly because of the soil moisture
44 perturbations. Our results suggest that future wildfire may accelerate boreal carbon loss, not
45 only through direct emissions, but also through the biophysical impacts of fire aerosols.

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48 **1 Introduction**

49

50 Wildfire is becoming more active in recent decades over North America boreal regions
51 (Stocks et al., 2002; Kasischke and Turetsky, 2006). Fire activity is closely related to weather
52 conditions and large-scale atmospheric oscillations (Gillett et al., 2004; Duffy et al., 2005),
53 and is projected to increase significantly in the future due to climatic changes (Flannigan et
54 al., 2005; Balshi et al., 2009). More frequent wildfires are accelerating carbon loss in boreal
55 North America (Bond-Lamberty et al., 2007; Turetsky et al., 2011). Meanwhile, fire-induced
56 air pollution, including ozone (O₃) and aerosols, is predicted to increase in boreal and
57 downwind regions by midcentury (Yue et al., 2013; Yue et al., 2015). Wildfire emissions
58 have large impacts on air quality (Wotawa and Trainer, 2000; Morris et al., 2006),
59 weather/climate conditions (Randerson et al., 2006; Zhao et al., 2014), and public health (Zu
60 et al., 2016; Liu et al., 2017). However, little is known about how these pollutants affect
61 ecosystem carbon assimilation, and how this impact will change with the increased wildfire
62 activity in the future.

63

64 Surface O₃ is detrimental to plant health because it damages photosynthesis through stomatal
65 uptake (Sitch et al., 2007). In the present climate state, fire-induced O₃ enhancements are
66 predicted to reduce net primary productivity (NPP) in the Amazon forest by 230 Tg C yr⁻¹, a
67 magnitude comparable to the direct release of CO₂ from fires in South America (Pacífico et
68 al., 2015). The aerosol effects are more uncertain because both positive and negative
69 feedbacks occur. Appearance of aerosols increases diffuse light, which is beneficial for
70 shaded leaves in the lower canopy. Consequently, photosynthesis of the whole ecosystem
71 will increase as long as the total light availability is not compromised (Kanniah et al., 2012).
72 Rap et al. (2015) estimated that biomass burning aerosols increase Amazon NPP by 78–156
73 Tg C yr⁻¹, which offsets about half of the damage caused by fire O₃ (Pacífico et al., 2015). In
74 contrast, strong light attenuation associated with high aerosol loading may decrease canopy
75 photosynthesis (Cohan et al., 2002; Oliveira et al., 2007; Cirino et al., 2014). Furthermore,
76 the aerosol radiation changes indirectly influence land carbon uptake through concomitant
77 meteorological perturbations that are only beginning to be examined (Yue et al., 2017).

78

79 Future wildfire activity is projected to increase over boreal North America but with large
80 uncertainties (Flannigan et al., 2005; Tymstra et al., 2007; Girardin and Mudelsee, 2008;
81 Nitschke and Innes, 2008; Amiro et al., 2009; Balshi et al., 2009; Bergeron et al., 2010;



82 Wotton et al., 2010; de Groot et al., 2013; Wang et al., 2016). For example, Amiro et al.
83 (2009) predicted an increase of 34% in Canadian area burned for a $2\times\text{CO}_2$ scenario (2040-
84 2060) relative to a $1\times\text{CO}_2$ condition (1975-1995), using the Canadian Fire Weather Index
85 (CFWI) and output from Canadian global climate model (CGCM) version 1. On the other
86 hand, Balshi et al. (2009) projected that area burned in boreal North America would double
87 by the year 2045-2050 relative to 1991-2000, using the Multivariate Adaptive Regression
88 Splines (MARS) approach and meteorological output from CGCM version 2. The increasing
89 rate in Balshi et al. (2009) is higher than that in Amiro et al. (2009), indicating substantial
90 uncertainties in fire projections originating from both fire models and simulated future
91 climate. Recently, Yue et al. (2015) developed stepwise regressions over boreal ecoregions
92 between area burned and multiple meteorological variables as well as CFWI indexes. They
93 used output from 13 climate models to drive these regression models and predicted an
94 average increase of 66% in boreal area burned at 2046-2065 relative to 1981-2000 under the
95 IPCC A1B scenario (Solomon et al., 2007). The multi-model ensemble projection reduces
96 uncertainties associated with single model climate projections. Yue et al. (2015) calculated
97 that the wildfire emission increase by the 2050s would increase mean summertime surface O_3
98 by 5 ppbv in Alaska and 3 ppbv in Canada. The study found regional maximum O_3
99 enhancements as high as 15 ppbv, suggesting the potential for possible vegetation damage
100 and land carbon loss due to the enhanced boreal fire-related air pollution. Wildfire aerosols
101 are also expected to increase significantly but not predicted in Yue et al. (2015).

102

103 In this study, we quantify the impacts of O_3 and aerosols emitted from boreal wildfires on the
104 land carbon uptake in North America in the present climate state and in the future world at
105 2050, taking advantage of the ensemble projection of future wildfire emissions by Yue et al.
106 (2015). We first analyze relationships between gross primary production (GPP) and aerosol
107 optical depth (AOD) at 550 nm over the boreal regions based on observations. We then
108 perform a suite of Earth system model simulations using NASA GISS ModelE2 that embeds
109 the Yale Interactive Terrestrial Biosphere model (YIBs), a framework known as ModelE2-
110 YIBs (Yue and Unger, 2015). Future projections of wildfire emissions from Yue et al. (2015)
111 are applied as input to ModelE2-YIBs model to project fire-induced O_3 and aerosol
112 concentrations in the 2010s and 2050s. The impacts of the boreal fire O_3 on forest
113 photosynthesis are predicted using the flux-based damage algorithm proposed by Sitch et al.
114 (2007). Fire aerosols induce perturbations to radiation, meteorology, and hydrology, leading
115 to multiple influences on the land carbon uptake. Sensitivity experiments are performed using



116 the YIBs model in offline mode to isolate the contributions of changes in the individual
117 meteorological drivers.

118

119

120 **2 Materials and methods**

121

122 **2.1 Observed GPP-AOD relationships**

123

124 Following the approach by Strada et al. (2015), we investigate the GPP sensitivity to diffuse
125 radiation and AOD variability in boreal regions. First, we identify study sites in Canada and
126 Alaska from the AmeriFlux (AMF) network (<http://ameriflux.lbl.gov/>). The number of
127 available boreal sites is much fewer than those in temperate regions. We select AMF sites
128 providing hourly (or half-hourly) simultaneous measurements of GPP (non gap-filled) and
129 photosynthetically active radiation (PAR, total and diffuse) for at least 3 consecutive years.
130 Only two Canadian sites meet the criteria: Groundhog River (CA-Gro, 82.2°W, 48.2°N), a
131 mixed forest (MF), and Quebec Mature Boreal Forest Site (CA-Qfo, 73.4°W, 49.7°N), an
132 evergreen needleleaf forest (ENF). At the two selected sites, we calculate the Pearson's
133 correlation coefficients between half-hourly GPP and different components of PAR. We then
134 apply instantaneous Level 2 Collection 6 of AOD pixels at 3-km resolution retrieved by the
135 Moderate Resolution Imaging Spectroradiometer (MODIS, <https://ladsweb.nascom.nasa.gov/>)
136 onboard the Aqua and Terra satellites (Levy et al., 2013). The MODIS 3-km AOD product
137 has been fully validated against ground-based sun photometers at both global (Remer et al.,
138 2013) and urban/suburban (Munchak et al., 2013) scales. Strada et al. (2015) used ground-
139 based AOD observations from the Aerosol Robotic Network (AERONET) near AMF sites to
140 validate the sampling technique of MODIS 3-km AOD product. For this study, the validation
141 against ground-based AOD observations was not possible because no AERONET stations
142 exist near to the selected AMF sites.

143

144 Every day, MODIS satellite sensors pass a specific region between 10:00 and 14:00 Local
145 Time (LT), leaving patchy signals around the AmeriFlux sites. Most of MODIS AOD data at
146 high latitudes are available only in boreal summer; as a result, we narrow our explorations of
147 the GPP-AOD relationships to the noontime (10:00-14:00 LT) from June to August. The
148 chosen noontime window limits the contributions that confounding factors such as low solar
149 angles and high diffuse fraction may have on the amount of diffuse PAR and plant



150 productivity (Niyogi et al., 2004). For each summer day, we select instantaneous MODIS 3-
151 km AOD pixels that are (a) located within a distance of 0.03° (about 3 km) from the targeted
152 AMF site and (b) “quasi-coincident” with AMF data, which are available each half-hour.
153 Because of the unavoidable temporal differences between MODIS overpass and AMF data
154 availability, we name this selection “quasi-coincident”. Cloud mask, applied in the MODIS
155 retrieval procedure, conveniently filters out cloudy instants and should reduce the effect of
156 clouds in the scattering process. We calculate both the correlation and regression coefficients
157 between “quasi-coincident” GPP and AOD at the selected sites. Negative GPP is considered
158 as a missing value. To further reduce the influence of cloud cover, we discard instants (both
159 AMF and MODIS data) when precipitation is non-zero. The GPP-AOD sampling pairs are
160 much fewer than GPP-PAR, because we select instants when both instantaneous AOD and
161 GPP data are available. In addition, AOD is screened for clear instants to exclude the impacts
162 of clouds.

163

164 **2.2 Wildfire emissions**

165

166 Wildfire emissions used in climate modeling are calculated as the product of area burned,
167 fuel consumption, and emission factors. To predict area burned, we build stepwise
168 regressions for area burned in 12 boreal ecoregions (Yue et al., 2015). Observed area burned
169 aggregated from inter-agency fire reports is used as the predictand. Predictors are selected
170 from 44 ($5 \times 6 + 7 \times 2$) variables including five meteorological parameters (mean and maximum
171 temperature, relative humidity, precipitation, and geopotential height at 500 hPa) of six
172 different time intervals (winter, spring, summer, autumn, fire season (May-October), and the
173 whole year), as well as the mean and maximum values of 7 fire indexes from the CFWI
174 system during fire season. We consider the impacts of antecedent factors on current fire
175 activity by including all above variables at the same year and those in the previous two years,
176 making a total of 132 (44×3) factors. The final formats of regression are different among
177 ecoregions, depending on the selection of the factors that contribute the maximum observed
178 variance in predictand but remain the minimum collinearity among predictors. These
179 regression functions are then driven with output from 13 CMIP3 climate models (Meehl et al.,
180 2007) to predict area burned at present day (1981-2000) and midcentury (2046-2065).

181

182 Fuel consumption, the dry mass burned per fire area, is the product of fuel load and burning
183 severity. For fuel load in Alaska, we use 1-km inventory from the US Forest Service (USFS)



184 Fuel Characteristic Classification System (FCCS, McKenzie et al., 2007). For fuel load in
185 Canada, we use a 1-km fuel type map from the Canadian Fire Behavior Prediction (FBP)
186 system (Nadeau et al., 2005), combined with fuel-bed definition from the FCCS. Burning
187 severity, the fraction of fuel load burned by fires, is calculated with the USFS CONSUME
188 model 3.0 following the approach described in Val Martin et al. (2012). With both fuel load
189 and burning severity, we derive fuel consumption and further calculate biomass burned in
190 boreal North America with the predicted area burned. Fire emissions for a specific species are
191 then estimated as the product between biomass burned and the corresponding emission factor,
192 which is adopted from measurements by Andreae and Merlet (2001) except for NO_x. We use
193 the average values from six studies as NO_x emission factor, because the number reported in
194 Andreae and Merlet (2001) is much higher than field observations (Alvarado et al., 2010).
195 Based on projected area burned and observation-based fuel consumption and emission factors,
196 we derive fire emissions of NO_x, carbon monoxide (CO), non-methane volatile organic
197 compounds (NMVOCs, Alkenes and Alkanes), NH₃, SO₂, black (BC) and organic carbon
198 (OC) in the present day and midcentury.

199

200 **2.3 NASA ModelE2-YIBs model**

201

202 The NASA ModelE2-YIBs is an interactive climate-carbon-chemistry model, which couples
203 the chemistry-climate model NASA ModelE2 (Schmidt et al., 2014) and the YIBs vegetation
204 model (Yue and Unger, 2015). NASA ModelE2 is a general circulation model with
205 horizontal resolution of 2°×2.5° latitude by longitude and 40 vertical layers up to 0.1 hPa. It
206 dynamically simulates both the physical (emissions, transport, and deposition) and chemical
207 (production, conversion, and loss) processes of gas-phase chemistry (NO_x, HO_x, O_x, CO, CH₄,
208 and NMVOCs), aerosols (sulfate, nitrate, ammonium, BC, OC, dust, and sea salt), and their
209 interactions. In the model, oxidants influence the photochemical formation of secondary
210 aerosol species (e.g., sulfate, nitrate, and biogenic secondary organic aerosol), in turn,
211 aerosols alter photolysis rates and influence the online gas-phase chemistry. The model also
212 considers interactions between climate and atmospheric components. Simulated climate
213 affects formation, transport, and deposition of atmospheric components, in turn, both O₃ and
214 aerosols influence climate by altering radiation, temperature, precipitation, and other climatic
215 variables. Both observation-based evaluations and multi-model inter-comparisons indicate
216 that ModelE2 demonstrates skill in simulating climatology (Schmidt et al., 2014), radiation



217 (Wild et al., 2013), atmospheric composition (Shindell et al., 2013b), and radiative effects
218 (Shindell et al., 2013a).

219

220 YIBs is a process-based vegetation model that dynamically simulates changes in leaf area
221 index (LAI) through carbon assimilation, respiration, and allocation. Coupled photosynthesis-
222 stomatal conductance is simulated with the Farquhar-Ball-Berry scheme (Farquhar et al.,
223 1980; Ball et al., 1987). Leaf-level photosynthesis is upscaled to canopy level by separating
224 diffuse and direct light for sunlit and shaded leaves (Spitters, 1986). Plant respiration
225 considers thermal dependence as well as acclimation to temperature (Atkin and Tjoelker,
226 2003). Soil respiration is calculated based on the carbon flows among 12 biogeochemical
227 pools (Schaefer et al., 2008). Net carbon uptake is allocated among leaves, stems, and roots
228 to support leaf development and plant growth (Cox, 2001). An interactive flux-based O₃
229 damage scheme proposed by Sitch et al. (2007) is applied to quantify the photosynthetic
230 responses to ambient O₃ (Yue and Unger, 2014). The YIBs model has been benchmarked
231 against *in situ* GPP from 145 eddy covariance flux tower sites and satellite retrievals of LAI
232 and phenology (Yue and Unger, 2015).

233

234 **2.4 Simulations**

235

236 We perform 6 time-slice simulations, three for present-day (2010s) and three for midcentury
237 (2050s), with atmosphere-only configuration to explore the impacts of fire emissions on NPP
238 in boreal North America (Table 1). Simulations F10CTRL and F50CTRL turn off all fire
239 emissions as well as O₃ vegetation damage for the 2010s and 2050s, respectively. However,
240 climatic feedbacks of aerosols from other sources (both natural and anthropogenic) and
241 related photosynthetic responses are included. Simulations F10AERO and F50AERO
242 consider the responses of plant productivity to perturbations in radiation and meteorology
243 caused by aerosols, including emissions from wildfires and other sources, but do not include
244 any O₃ vegetation damage. In contrast, simulations F10O3 and F50O3 calculate offline O₃
245 damage based on the simulated O₃ from all sources including fire emissions. Vegetation NPP,
246 the net carbon uptake by biosphere, is the impact metric. The difference between AERO and
247 CTRL runs isolates the impacts of fire aerosols on NPP, and the difference between O3 and
248 CTRL runs isolates O₃ vegetation damage caused by fire and non-fire emission sources.

249



250 All simulations are conducted for 20 years and outputs for the last 15 years are used for
251 analyses. The simulations apply sea surface temperatures (SSTs) and sea ice distributions
252 from previous NASA GISS experiments under the IPCC RCP8.5 scenario (van Vuuren et al.,
253 2011) archived for the Coupled Model Intercomparison Project Phase 5 (CMIP5). Decadal
254 average monthly-varying SST and sea ice of 2006-2015 are used as boundary conditions for
255 present-day (2010s) runs while that of 2046-2055 are used for future (2050s) runs. Decadal
256 average well-mixed greenhouse gas concentrations and anthropogenic emissions of short-
257 lived species, both at present day and midcentury, are adopted from the RCP8.5 scenario
258 (Table 2). Natural emissions of soil and lightning NO_x , biogenic volatile organic compounds
259 (BVOC), dust, and sea salt are climate-sensitive and simulated interactively. The RCP8.5
260 land cover change dataset shows limited changes in land cover fractions between 2010s and
261 2050s (Oleson et al., 2010). For example, relative to the 2010s, a maximum gain of 5% is
262 predicted for grassland in the 2050s, resulting from a 1% loss in deciduous forest and another
263 1% loss in needleleaf forest over boreal North America. As a result, a land cover dataset
264 derived from satellite retrievals (Hansen et al., 2003) is applied for both the 2010s and 2050s.

265

266 To isolate the impact of individual aerosol-induced climatic perturbations on NPP, we
267 perform 10 sensitivity experiments using the YIBs vegetation model driven with offline
268 meteorology simulated by ModelE2-YIBs model (Table 3). For example, the offline run
269 Y10_CTRL is driven with variables from the online simulation of F10CTRL (Table 1). The
270 run Y10_TAS adopts the same forcing as Y10_CTRL except for temperature, which is
271 simulated by the climate simulation of F10AERO. In this case, we quantify the NPP
272 responses to individual and/or combined climate feedback (mainly in temperature, radiation,
273 and soil moisture) by fire aerosols. Each offline run is conducted for 12 years and the last 10
274 years are used for analyses.

275

276 **2.5 Observation datasets**

277

278 We use observations to evaluate GPP, AOD, and O_3 in boreal North America simulated by
279 ModelE2-YIBs. For GPP, we use a benchmark data product upscaled from FLUXNET eddy
280 covariance data using an ensemble of regression trees (Jung et al., 2009). For AOD
281 observations, we use satellite retrieval at 550 nm from Terra MODIS Level 3 data product.
282 For O_3 , gridded datasets are not available. We use site-level observations from 81 U.S. sites
283 at the Clean Air Status and Trends Network (CASTNET, <https://www.epa.gov/castnet>) and



284 202 Canadian sites at the National Air Pollution Surveillance (NAPS,
285 <http://www.ec.gc.ca/rnsps-naps/>) program. All datasets are averaged over the 2008-2012
286 period to represent present-day climatological conditions. Gridded datasets are interpolated to
287 the same $2^{\circ} \times 2.5^{\circ}$ resolution as ModelE2-YIBs model.

288

289

290 **3 Results**

291

292 **3.1 Observed GPP-AOD relationships**

293

294 Positive correlations between GPP and diffuse PAR are found at the two boreal sites (Figs
295 1b-1c). The magnitude of diffuse PAR is similar for these sites, possibly because they are
296 located at similar latitudes (Fig. 1a). GPP values at CA-Gro are generally higher than that at
297 CA-Qfo, likely because deciduous broadleaf forest (DBF) has higher photosynthetic rates.
298 Consequently, the slope of regression between GPP and PAR_{dif} is higher at CA-Gro than that
299 at CA-Qfo, suggesting that GPP of DBF (or MF) is more sensitive to changes in diffuse PAR
300 than that of ENF. We find almost zero correlation between GPP and PAR_{dir} at the two sites
301 (Table 4), indicating that photosynthesis is in general light-saturated for sunlit leaves at these
302 sites during boreal summer noontime. As a result, modest reductions in direct light by
303 aerosols will not decrease GPP of the whole canopy.

304

305 With satellite-based AOD, we find positive correlations between GPP and AOD at both sites
306 (Figs 1d-1e). However, the slope of regression between GPP and AOD is lower at CA-Gro
307 compared with that at CA-Qfo, opposite to the GPP- PAR_{dif} regressions. The cause of such
308 discrepancy might be related to the limitation of data availability. For the same reason, the
309 GPP-AOD correlation is insignificant at CA-Gro site. On average, we estimate GPP
310 sensitivity of $3.5 \pm 1.1 \mu\text{mol m}^{-2} \text{s}^{-1}$ per unit AOD at lower latitudes of boreal regions in the
311 summer.

312

313 **3.2 Evaluation of model photosynthesis and atmospheric composition**

314

315 Simulated summer GPP shows high values in mid-western Canada (Alberta and
316 Saskatchewan) and the Southeast (Ontario) (Fig. 2a). Forest GPP at high latitudes is low
317 because of the cool weather and light limitation there. Compared against the data product,



318 simulated GPP reasonably captures the spatial distribution with a high correlation coefficient
319 of 0.77 ($p \ll 0.01$) and relatively small biases within 20%. Simulated AOD reproduces the
320 observed spatial pattern including the high values in boreal forests (Fig. 2b). In contrast to the
321 MODIS observations, predicted AOD is relatively uniform over the West with a background
322 value of ~ 0.1 . This discrepancy explains the low correlation coefficient ($R = 0.25$, $p < 0.01$)
323 between the model and MODIS data. Simulated $[O_3]$ shows low values in boreal North
324 America and high values in the western and eastern U.S. (Fig. 3a). This pattern is consistent
325 with surface observations (Fig. 3b), but the model overestimates the measured surface O_3 by
326 40%. The Canadian measurement sites are located near the southern boundary, and as a result
327 do not represent the average state over the vast boreal region at higher latitudes.

328

329 3.3 Simulation of wildfire O_3 and aerosols

330

331 Compared to the present day, wildfire area burned is projected to increase by 66% in boreal
332 North America at midcentury, mainly because of the higher temperature in future fire seasons
333 (Yue et al., 2015). Consequently, enhanced fire emissions increase concentrations of surface
334 O_3 and column AOD, especially over Alaska and central Canada (Fig. 4). The maximum
335 centers of air pollutants are collocated for O_3 and AOD but with unproportional magnitudes,
336 suggesting non-linear conversion among fire emission species as well as the interactions with
337 natural emission sources (e.g., lightning/soil NO_x and BVOC). On average, wildfire
338 emissions contribute $7.1 \pm 3.1\%$ to surface O_3 and $25.7 \pm 2.4\%$ to AOD in the summer over
339 boreal North America in the present day. By midcentury, these ratios increase significantly to
340 $12.8 \pm 2.8\%$ for O_3 and $36.7 \pm 2.0\%$ for AOD.

341

342 3.4 Simulation of fire pollution impacts on NPP

343

344 Surface O_3 , including both fire and non-fire emissions (Table 2), causes limited damages to
345 summer GPP in boreal North America (Fig. 5). The most significant damage is predicted
346 over eastern U.S., where observed $[O_3]$ is high over vast forest ecosystems (Fig. 3). In the
347 western U.S., $[O_3]$ is also high but the O_3 -induced GPP reduction is trivial because low
348 stomatal conductance in the semi-arid ecosystems limits O_3 uptake there (Yue and Unger,
349 2014). Over boreal regions, the mean $[O_3]$ of 28 ppbv is much lower than the damaging
350 threshold of 40 ppbv for most tree species (Yue et al., 2016). Statistics in Yue et al. (2015)
351 show that maximum daily 8-hour average (MDA8) $[O_3]$ can be higher than 40 ppbv in Alaska



352 and Canada. However, such episodes appear at 95 percentile for present day and 90
353 percentile for midcentury, suggesting that O₃ vegetation damage is rare in boreal North
354 America and fire-induced O₃ enhancement does not exacerbate such damages. Therefore, we
355 do not consider O₃ damage effects further.

356

357 Fire aerosols cause significant perturbations in shortwave radiation at surface (Fig. 6). The
358 direct light is largely attenuated especially over Alaska and central Canada, where fire
359 aerosols are most abundant (Fig. 4). In contrast, diffuse light widely increases due to particle
360 scattering. In the present day, the average reduction of 5.6 W m⁻² in the direct light
361 component is in part offset by the enhancement of 2.6 W m⁻² in the diffuse light component,
362 leading to a net reduction of 3.0 W m⁻² in solar radiation over boreal North America. By the
363 midcentury, a stronger reduction of 9.5 W m⁻² in direct light is accompanied by an increase of
364 4.0 W m⁻² in diffuse light, resulting in a net reduction of 5.5 W m⁻² in solar radiation. Fire-
365 induced BC aerosols strongly absorb solar radiation in the atmospheric column (Figs 7a-7b).
366 On average, fire aerosols absorb 1.5 W m⁻² in the present day and 2.6 W m⁻² by the
367 midcentury.

368

369 Atmospheric circulation patterns respond to the aerosol-induced radiative perturbations (Figs
370 7c-7d). Surface radiative cooling and atmospheric heating together increase air stability and
371 induce anomalous subsidence. In the present day, such descending motion is confined to 55-
372 68°N, accompanied by a rising motion at 52-55°N (Fig. 7c). As a result, fire aerosols induce
373 surface warming at higher latitudes but cooling at lower latitudes in boreal regions (Fig. 8a).
374 Meanwhile, precipitation is inhibited by the subsidence in northwestern Canada but is
375 promoted by the rising motion in the Southwest (Fig. 8c). By the midcentury, the range of
376 subsidence expands southward to 42°N (Fig. 7d) due to strengthened atmospheric heating
377 (Fig. 7b). The downward convection of warm air offsets surface radiative cooling (Fig. 6b),
378 leading to a significant warming in the Southwest (Fig. 8b). The expanded subsidence further
379 inhibits precipitation in vast domain of Canada (Fig. 8d). Soil moisture is closely related to
380 rainfall and as a result exhibits dipole changes (drier north and wetter south) in the present
381 day (Fig. 8e) but widespread reductions (Fig. 8f) by the midcentury.

382

383 In response to the climatic effects of fire aerosols, boreal NPP shows distinct changes
384 between the present day and midcentury (Fig. 9). In the 2010s, forest NPP increases by 5-15%
385 in Alaska and southern Canada, but decreases by 5-10% in northern and eastern Canada. This



386 pattern of NPP changes (Δ NPP) is connected to the climatic effects of aerosols, especially
387 changes in soil moisture (Fig. 8). The correlation between Δ NPP (Fig. 9a) and changes in soil
388 moisture (Fig. 8e) reaches $R = 0.56$ ($n = 356$), much higher than the values of $R = -0.11$ for
389 temperature change (Fig. 8a) and $R = 0.22$ for precipitation change (Fig. 8c). At the
390 continental scale, the patchy responses of NPP offset each other. Since the dominant fraction
391 of carbon uptake occurs in southern Canada (Fig. 2a), where positive NPP change is
392 predicted (Fig. 9a), wildfire aerosols enhance the total NPP by 72 Tg C yr^{-1} in the present day
393 (Table 5). In contrast, increased wildfire emissions in the 2050s inhibit precipitation and
394 decrease soil moisture in boreal North America (Fig. 8), leading to widespread NPP
395 reductions and a total NPP loss of 118 Tg C yr^{-1} (Fig. 9b, Table 5).

396

397

398 4 Discussion

399

400 4.1 Roles of aerosol climatic feedback

401

402 The contrasting sign of NPP responses in the present day and midcentury are closely related
403 to the aerosol-induced surface climatic feedback. Sensitivity experiments using offline YIBs
404 model (Table 3) allowed assessment of the impacts of individual changes in the major
405 meteorological drivers, including temperature, radiation (diffuse and direct), and soil
406 moisture (Table 5). The offline simulations driven with changes in all three variables yield
407 Δ NPP of 126 Tg C yr^{-1} for the 2010s and -97 Tg C yr^{-1} for the 2050s. These values are
408 different from the online simulations, which predict Δ NPP of 72 Tg C yr^{-1} for the 2010s and
409 $-118 \text{ Tg C yr}^{-1}$ for the 2050s. Missing of other aerosol climatic feedbacks in the offline model,
410 for example, changes in relative humidity, surface pressure, soil temperature, and turbulence
411 momentum, may cause such discrepancy between the online and offline simulations.
412 Seasonal analyses show that summertime Δ NPP is 99 Tg C at present day and -95 Tg C at
413 midcentury, dominating the NPP changes all through the year, because both wildfire
414 emissions and ecosystem photosynthesis maximize in boreal summer.

415

416 Observations show that aerosols can promote plant photosynthesis through increasing diffuse
417 radiation (Niyogi et al., 2004; Cirino et al., 2014; Strada et al., 2015). Our analyses with
418 ground and satellite data also show positive correlations between GPP and AOD (Fig. 1).
419 Wildfire aerosols enhance diffuse radiation by 2.6 W m^{-2} (1.7%) at present day and 4.0 W m^{-2}



420 (2.3%) at midcentury in boreal North America (Fig. 6). With these changes, simulated NPP
421 increases by 8 Tg C yr⁻¹ at the 2010s and 14 Tg C yr⁻¹ at the 2050s (Table 5). Near the two
422 AmeriFlux sites (Fig. 1a), wildfires increase local AOD by 0.03 (Fig. 4c). Meanwhile, we
423 estimate that summer average (00:00-24:00) GPP increases by 0.04 μmol m⁻² s⁻¹ in the same
424 region due to aerosol diffuse fertilization effects (DFE) based on the results of (Y10_PAR –
425 Y10_CTRL). This change suggests a simulated GPP sensitivity of 1.2 μmol m⁻² s⁻¹ (22%) per
426 unit AOD. Observed GPP sensitivity to AOD at the two sites are 2.3 (19%) and 4.5 μmol m⁻²
427 s⁻¹ (58%) per unit AOD, respectively (Figs 1d-e). The absolute value of GPP sensitivity from
428 simulations is much smaller than that of observations, because the former is for 24-h average
429 while the latter is only for noontime (10:00-14:00). The relative change of 22% in YIBs
430 model falls within the observed range of 19-58%.

431

432 The estimated NPP changes of 8 Tg C yr⁻¹ by the radiative effects of boreal fire aerosols are
433 much weaker than the enhancement of 78-156 Tg C yr⁻¹ by fires in Amazon basin (Rap et al.,
434 2015). There are at least two reasons for such discrepancies in the DFE between boreal and
435 Amazon fire aerosols. First, wildfire emissions and associated impacts on radiation are much
436 smaller in boreal regions. Wildfires in Alaska and Canada directly emit 68 Tg C yr⁻¹ at the
437 2010s, resulting in enhancement of summer AOD by 35% and diffuse radiation by 1.7%.
438 These boreal emissions are much smaller than the ~240 Tg C yr⁻¹ in Amazon basin (van der
439 Werf et al., 2010), where fires enhances regional PM_{2.5} concentrations by 85% and diffuse
440 radiation by 6.2% in dry seasons (Rap et al., 2015). Second, larger solar insolation in lower
441 latitudes allows stronger DFE for the same unit change of diffuse radiation. In our prediction,
442 most of NPP changes occur at high latitudes of boreal regions (Fig. 9), where total insolation
443 is not so abundant as that at the tropical areas. Consequently, decline of direct radiation in
444 boreal regions more likely converts the light availability of sunlit leaves from light-saturation
445 to light-limitation, offsetting the benefit from enhanced diffuse radiation for shaded leaves.
446 For this study, we do not find GPP reduction by the decline of direct light at the two
447 AmeriFlux sites (Table 4), possibly because these sites are located at middle latitudes (<50°N).
448 In the future, more observations at higher latitudes (> 55°N) are required to explore the
449 sensitivity of GPP to AOD at the light-limited conditions.

450

451 Simulations have shown that absorbing aerosols can cause regional drought by increasing air
452 stability (Liu, 2005; Cook et al., 2009; Tosca et al., 2010). Our results confirm such tendency
453 but with varied range of hydrological responses depending on the magnitude of wildfire



454 emissions (Figs 8c-8f). Observations suggest that precipitation (and the associated soil
455 moisture) is the dominant driver of the changes in GPP over North America, especially for
456 the domain of cropland (Beer et al., 2010). Sensitivity experiments with offline YIBs model
457 show that changes in soil moisture account for 82.5% of Δ NPP at present day and 70.5% of
458 Δ NPP at midcentury (Table 5). These results suggest that aerosol-induced changes in soil
459 water availability, instead of temperature and radiation, dominantly contribute to the changes
460 of boreal NPP, consistent with observational and experimental results (Ma et al., 2012;
461 Girardin et al., 2016; Chen et al., 2017).

462

463 **4.2 Limitations and uncertainties**

464

465 For this study, we use the ensemble projected fire emissions from Yue et al. (2015). Area
466 burned is predicted based on the simulated meteorology from multiple climate models. Such
467 an approach may help reduce model uncertainties in climatic responses to CO₂ changes, but
468 cannot remove the possible biases in the selection of climate scenarios and fire models. All
469 predictions in Yue et al. (2015) are performed under the IPCC A1B scenario. With two
470 different scenarios, A2 of high emissions and B2 of low emissions, Balshi et al. (2009)
471 showed that future area burned in boreal North America increases at a similar rate until the
472 2050s, after which area burned in A2 scenario increases much faster than that in B2 scenario.
473 On average, boreal area burned in Balshi et al. (2009) increases by ~160% at 2051-2060
474 compared with 2001-2010, much higher than the change of 66% in Yue et al. (2015). In
475 contrast, Amiro et al. (2009) predicted that boreal area burned at the 2×CO₂ scenario
476 increases only by 34% relative to the 1×CO₂ scenario. This ratio is only half of the estimate
477 in Yue et al. (2015), which compared results between periods with 1.44×CO₂ and 1×CO₂.
478 The discrepancies among these studies are more likely attributed to the differences in fire
479 models. Although both Amiro et al. (2009) and Yue et al. (2015) developed fire-weather
480 regressions in boreal ecoregions, the former study did not include geopotential height at 500
481 hPa and surface relative humidity as predictors, which make dominant contributions to area
482 burned changes in the latter study. On the other hand, Balshi et al. (2009) developed
483 nonlinear regressions between area burned and climate at grid scale, which helps retain
484 extreme values at both the temporal and spatial domain. Compared to previous estimates,
485 Yue et al. (2015) predicted median increases in future fire emissions over boreal North
486 America.

487



488 Predicted surface $[O_3]$ is much higher than observations over boreal North America (Fig. 3).
489 This bias does not affect main conclusions of this study, because predicted O_3 causes limited
490 damages to boreal GPP even with the overestimated $[O_3]$ (Fig. 5). The result confirms that
491 fire-induced O_3 vegetation damage is negligible in boreal North America. For aerosols, the
492 model captures reasonable spatial pattern of AOD but with a background value of ~ 0.1
493 outside fire-prone regions, where the actual AOD is usually 0.1-0.2 (Fig. 2). This discrepancy
494 may be related to the insufficient representations of physical and chemical processes in the
495 model, but may also result from the retrieval biases in MODIS data due to the poor surface
496 conditions (Liu et al., 2005) and small AOD variations (Vachon et al., 2004) at high latitudes.

497

498 Simulated aerosol climatic effects depend on radiative and physical processes implemented in
499 the climate model. We find that present-day boreal fire aerosols on average absorb 1.5 W m^{-2}
500 in the atmosphere (Fig. 6), which is much smaller than the value of $20.5 \pm 9.3 \text{ W m}^{-2}$ for fires
501 in equatorial Asia (Tosca et al., 2010). This is because boreal fires enhance AOD only by
502 0.03 while tropical fires increase AOD by ~ 0.4 . Previous modeling studies showed that fire
503 plumes induce regional and downwind drought through enhanced atmospheric stability
504 (Feingold et al., 2005; Tosca et al., 2010; Liu et al., 2014). Most of these results were based
505 on the direct and/or semi-direct radiative effects of fire aerosols. Inclusion of the indirect
506 aerosol effect may further inhibit precipitation and amplify drought, but may also introduce
507 additional uncertainties for the simulations. The fire-drought interaction may promote fire
508 activity, especially in a warmer climate. Ignoring this interaction may underestimate future
509 area burned and the consequent emissions.

510

511 4.3 Implications

512

513 Inverse modeling studies have shown that the land ecosystems of boreal North America are
514 carbon neutral in the present day, with the estimated land-to-air carbon flux from -270 ± 130
515 Tg C yr^{-1} to $300 \pm 500 \text{ Tg C yr}^{-1}$ (Gurney et al., 2002; Rodenbeck et al., 2003; Baker et al.,
516 2006; Jacobson et al., 2007; Deng et al., 2014). Here, we reveal a missing land carbon source
517 due to future wildfire pollution, taking into account full coupling among fire activity, climate
518 change, air pollution, and the carbon cycle. Fire pollution aerosol increases boreal NPP by 72
519 Tg C yr^{-1} in the present day, comparable to the direct carbon loss of 68 Tg C yr^{-1} from
520 wildfire CO_2 emissions. By midcentury, increasing fire emissions instead cause a NPP
521 reduction of 118 Tg C yr^{-1} due to the amplified drought. Although NPP is not a direct



522 indicator of the land carbon sink, reduction of NPP is always accompanied with the decline
523 of net ecosystem exchange (NEE) and the enhanced carbon loss. In combination with the
524 enhanced carbon emission of 130 Tg C yr^{-1} , future boreal wildfire presents an increasing
525 threat to the regional carbon balance and global warming mitigation. Furthermore, the NPP
526 reductions are mostly located in southern Canada, where cropland is the dominant ecosystem,
527 newly exposing the future wildfire-related air pollution risk to food production.

528

529

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533

534

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837 **Table 1.** Online simulations with ModelE2-YIBs climate model ^a

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Simulations	SST	[CO ₂]	Emissions	Fires	O ₃ effect	Aerosol effect
F10O3	2010s	2010s	2010s	2010s	Yes	No
F10AERO	2010s	2010s	2010s	2010s	No	Yes
F10CTRL	2010s	2010s	2010s	No	No	Yes
F50O3	2050s	2050s	2050s	2050s	Yes	No
F50AERO	2050s	2050s	2050s	2050s	No	Yes
F50CTRL	2050s	2050s	2050s	No	No	Yes

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840 ^a Values of SST, [CO₂], and emissions are adopted from RCP8.5 scenario, with the average

841 of 2006-2015 for the 2010s and that of 2046-2055 for the 2050s. For fire emissions, values at

842 the 2010s are predicted based on meteorology for 1981-2000 and those at the 2050s are for

843 2046-2065.

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848 **Table 2.** Emissions from wildfires and non-fire sources over boreal North America

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Species	Fire emissions (Tg yr ⁻¹)		Non-fire emissions (Tg yr ⁻¹)	
	2010s	2050s	2010s	2050s
NO _x ^a	0.39	0.74	2.43	2.08
CO	15.7	28.8	5.9	4.0
SO ₂ ^a	0.12	0.22	1.95	1.28
NH ₃	0.22	0.40	0.80	1.15
BC	0.08	0.16	0.03	0.01
OC	1.10	2.04	0.04	0.02
NM VOC	0.39	1.34	0.49	0.30
BVOC ^b	N/A	N/A	15.3	15.1

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851 ^a Natural emissions are included for NO_x (lightning and soil) and SO₂ (volcano).852 ^b ModelE2-YIBs calculates BVOC emissions using photosynthesis-dependent scheme
853 implemented by Unger et al. (2013).

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859 **Table 3.** Simulations with YIBs vegetation model driven by offline meteorology from

860 ModelE2-YIBs climate model

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Simulations	Base forcing	Temperature	PAR	Soil moisture
Y10_CTRL	F10CTRL			
Y10_ALL	F10CTRL	F10AERO	F10AERO	F10AERO
Y10_TAS	F10CTRL	F10AERO		
Y10_PAR	F10CTRL		F10AERO	
Y10_SLM	F10CTRL			F10AERO
Y50_CTRL	F50CTRL			
Y50_ALL	F50CTRL	F50AERO	F50AERO	F50AERO
Y50_TAS	F50CTRL	F50AERO		
Y50_PAR	F50CTRL		F50AERO	
Y50_SLM	F50CTRL			F50AERO

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869 **Table 4.** Pearson's correlation coefficients for GPP-PAR and GPP-AOD relationships at870 Ameriflux (AMF) sites ^a

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Site	Period ^b	Pearson's <i>R</i>					
		GPP-PAR	GPP-PAR _{dir}	GPP-PAR _{dif}	GPP-AOD	AOD-PAR _{dif}	AOD-PAR _{dir}
CA-Gro	2004-2013	0.19 (2432)	-0.01 (2432)	0.42 (2432)	0.15 (65)	0.60 (65)	-0.52 (65)
CA-Qfo	2003-2014	0.16 (3201)	-0.04 (3201)	0.45 (3201)	0.36 (59)	0.91 (34)	-0.80 (34)

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873 ^a Both GPP and PAR (direct PAR_{dir} and diffuse PAR_{dif}) data are adopted from site-level AMF
 874 measurements. AOD data are adopted from instantaneous MODIS Aqua and Terra 3-km
 875 retrievals. Correlations are calculated for quasi-coincident AMF and MODIS data over
 876 summer noontime (June-August, 10:00-14:00 Local Time). The sampling number for each
 877 correlation is denoted in brackets. Significant ($p < 0.05$) correlation coefficients are bolded.

878 ^b For CA-Gro site, diffuse PAR observations of 2005-2009 have been discarded because of
 879 poor calibration, as documented on the AMF website.

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885 **Table 5.** Changes in NPP (Tg C yr^{-1}) caused by composite and individual climatic effects of
886 fire aerosols

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	2010s	2050s
Online ^a	72	-118
Offline total ^b	126	-97
Temperature	11	-22
Radiation	8	14
Soil moisture	104	-86

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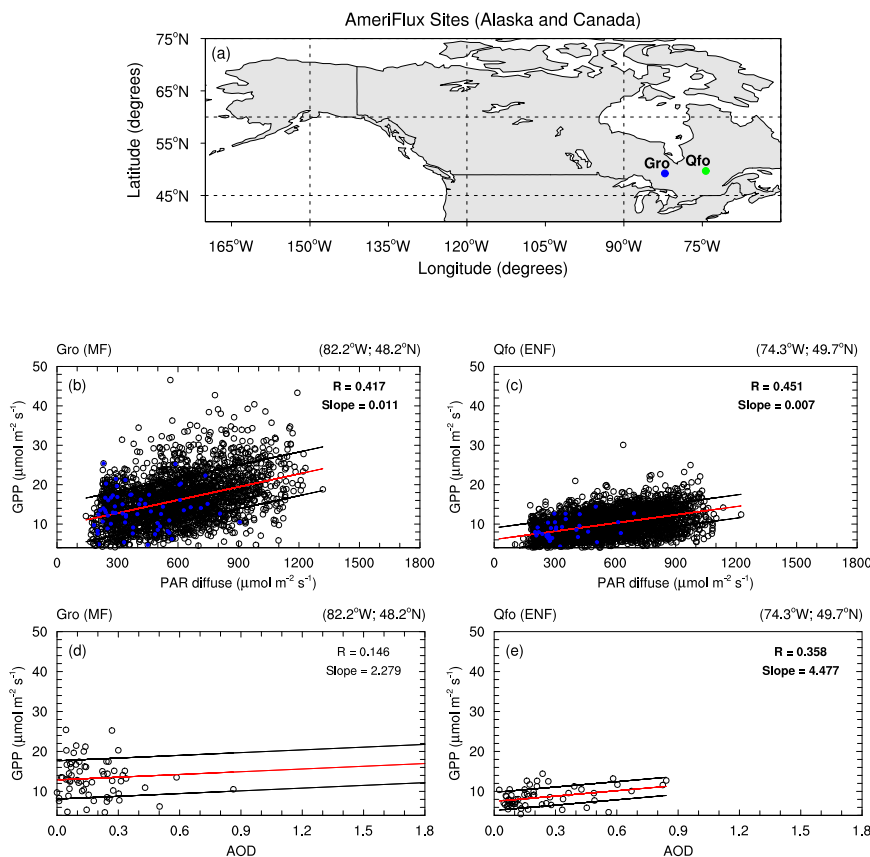
889 ^a Online results are calculated using the ModelE2-YIBs model with (F10AERO – F10CTRL)
890 for the 2010s and (F50AERO – F50CTRL) for the 2050s.

891 ^b Offline results are calculated with the YIBs model driven with individual or combined
892 changes in temperature, radiation, and soil moisture.

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Figure 1. Relationships between (b, c) GPP and diffuse PAR and (d, e) GPP and MODIS AOD at (a) two boreal sites: Groundhog River (Gro) and Quebec Mature Boreal Forest Site (Qfo). The two sites are from the AmeriFlux network in Canada and are dominated by mixed forest (MF at Gro) and evergreen needleleaf forest (ENF at Qfo) (Table 1). Data cover summer days (June-August). AmeriFlux diffuse PAR and GPP (in $\mu\text{mol m}^{-2} \text{s}^{-1}$) are half-hourly observations (10:00-14:00 LT). Instantaneous MODIS Aqua and Terra 3-km AOD are selected in a time span centered on AmeriFlux record time. For each plot: the red line indicates the regression line, black lines depict the 1- σ interval; the regression slope and correlation coefficient are both included for each site (in bold if statistically significant at 95% confidence level). Blue dots in (b, c) show instants when MODIS Aqua and Terra 3-km AODs overlap AmeriFlux data.

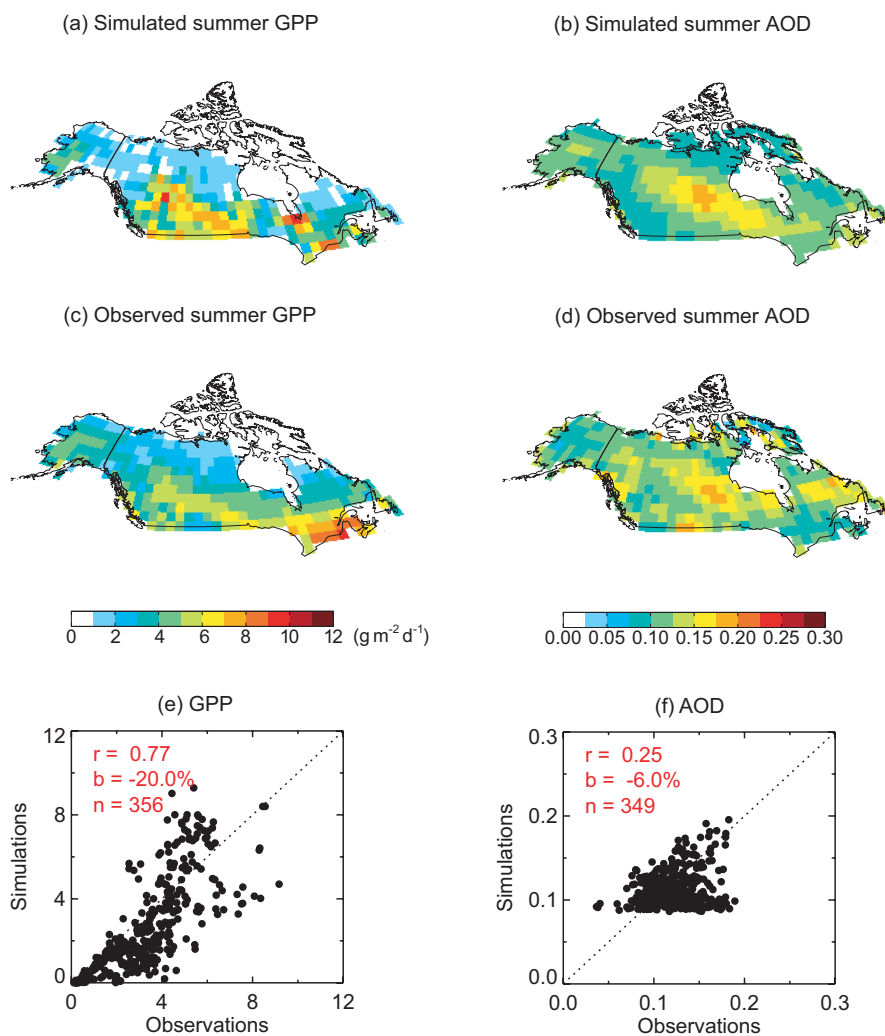
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914 **Figure 2.** Evaluation of simulated summer (a) GPP and (b) AOD at 550 nm with (c, d)
915 observations. Simulation results are from F10AERO (Table 1). Each point on the (e, f) scatter
916 plot represents one grid square in boreal North America. The number of points (n),
917 correlation coefficient (r), and relative bias (b) for the evaluation are presented on the plot.

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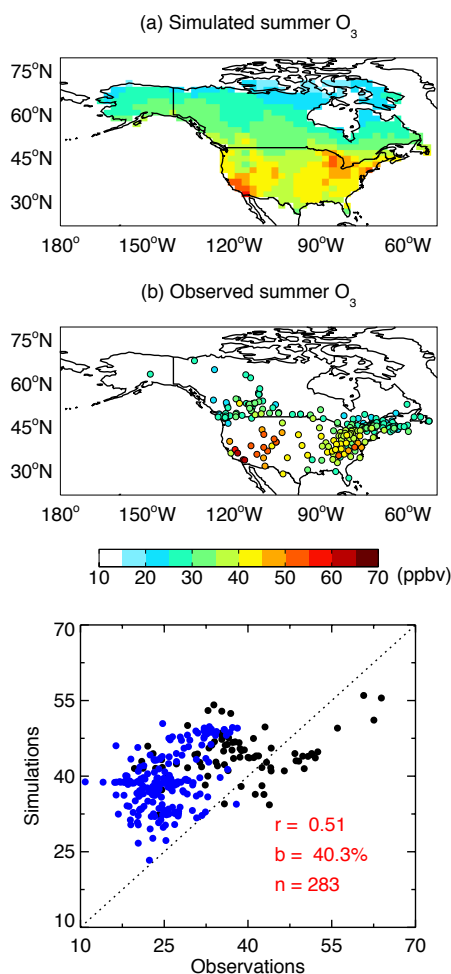


Figure 3. Evaluation of simulated summer surface $[O_3]$ with observations for 2008-2012.

Observations are collected from 81 U.S. sites at the Clean Air Status and Trends Network

(CASTNET) and 202 Canadian sites at the National Air Pollution Surveillance (NAPS)

program. The number of points (n), correlation coefficient (r), and mean bias (b) for the

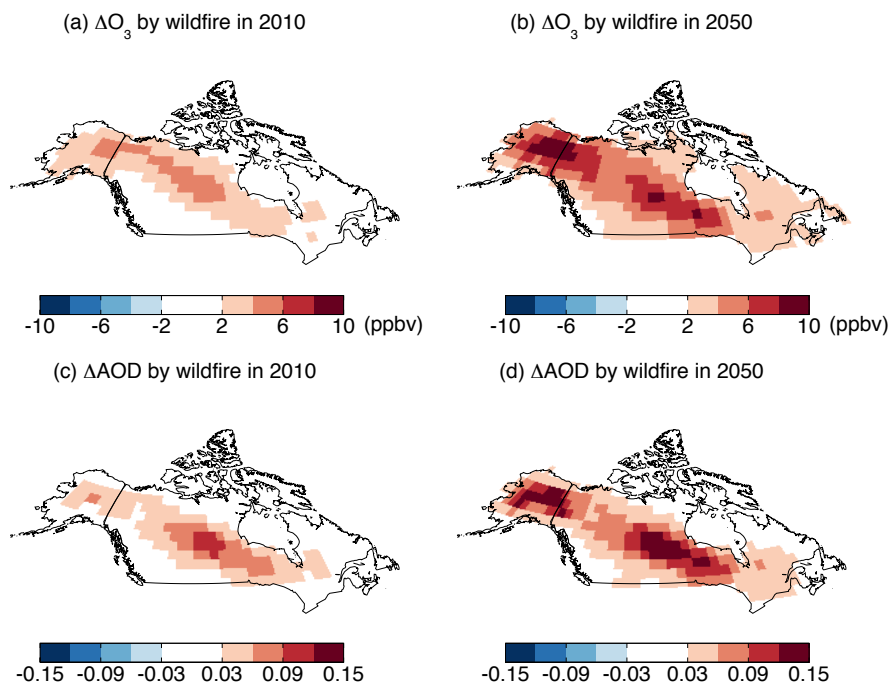
evaluation are presented on the plot. Values over Canada and Alaska are denoted with blue

points.



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967 **Figure 4.** Changes in summer (a, b) $[O_3]$ and (c, d) AOD at 550 nm induced by wildfire
 968 emissions in (a, c) the 2010s and (b, d) the 2050s over boreal North America. Only
 969 significant changes ($p < 0.05$) are shown.

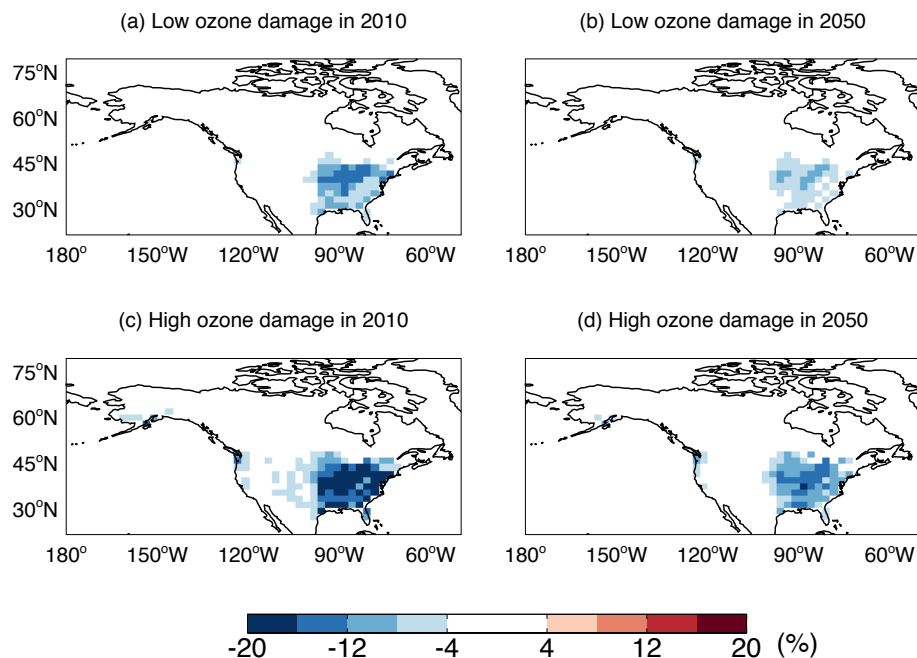
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975 **Figure 5.** Simulated O₃ damages to summer GPP in North America. Results shown are from

976 simulations with (a, b) low and (c, d) high O₃ sensitivities for (a, c) 2010 and (b, d) 2050.

977 Simulated [O₃] includes contributions from both wildfire and non-fire emissions. Results for

978 2010 are derived as $(F_{10O3}/F_{10CTRL}-1)\times 100\%$. Results for 2050 are derived as

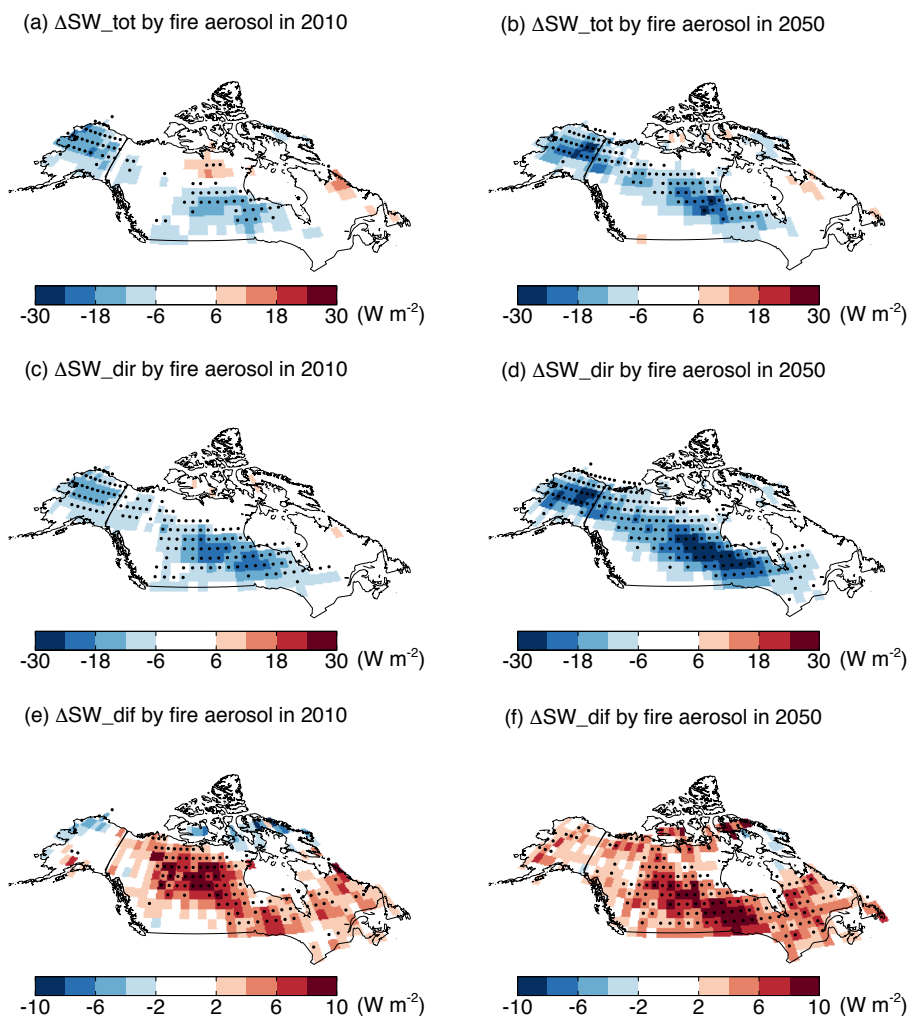
979 $(F_{50O3}/F_{50CTRL}-1)\times 100\%$.

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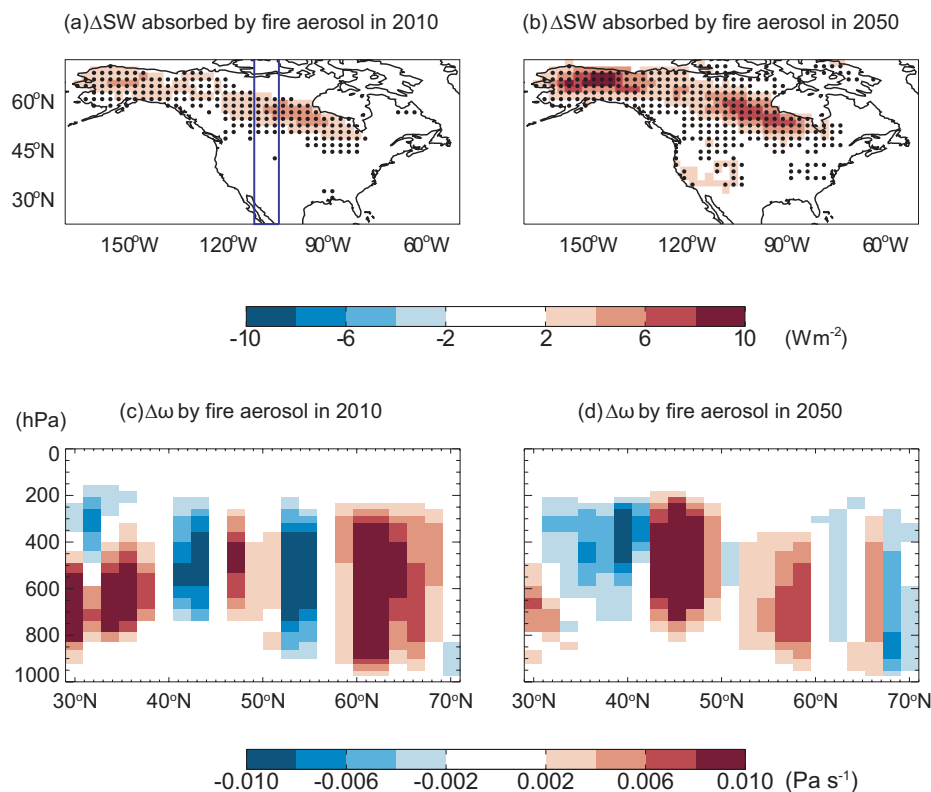
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Figure 6. Changes in surface radiative fluxes induced by wildfire aerosols in boreal North America. Results shown are for the changes in summertime (June-August) (a, b) total, (c, d) direct, and (e, f) diffuse solar radiation at surface caused by aerosols from wildfire emissions at (a, c, e) present day and (b, d, f) midcentury. Significant changes ($p < 0.05$) are marked with black dots. Results for 2010 are calculated as (F10AERO - F10CTRL). Results for 2050 are calculated as (F50AERO - F50CTRL).



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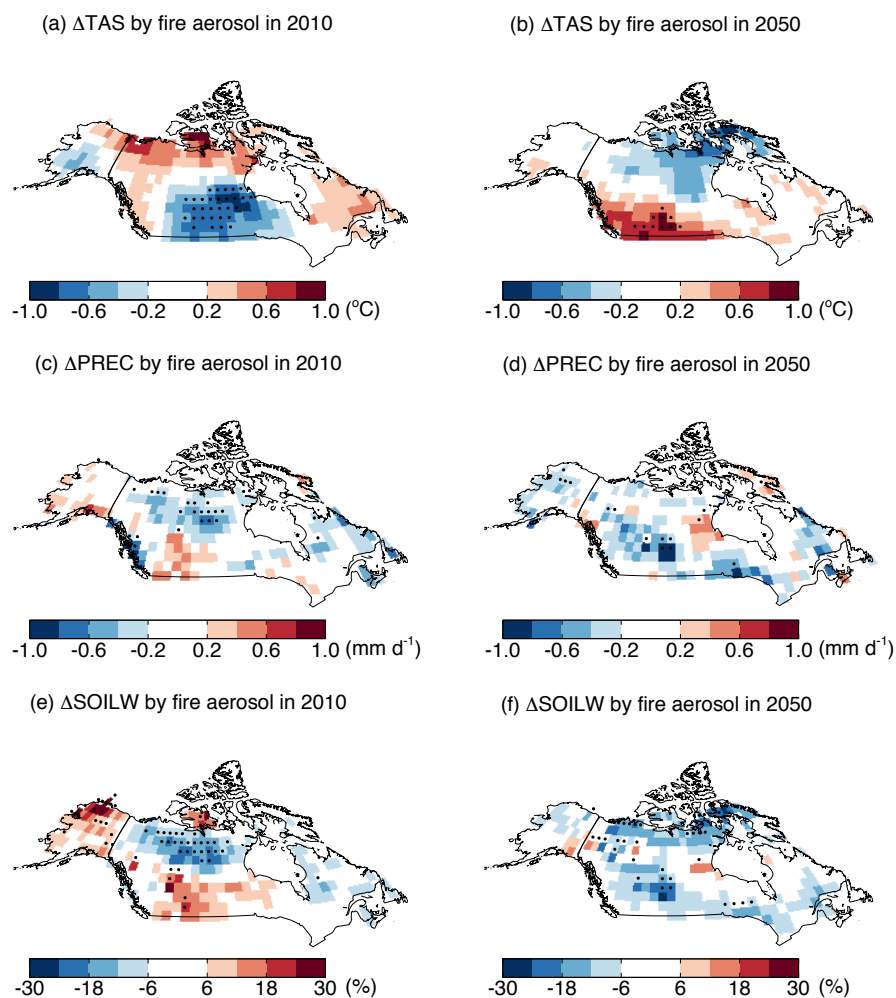
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998 **Figure 7.** Predicted (a, b) absorption of shortwave radiation and (c, d) perturbations in
999 vertical velocity by wildfire aerosols at (a, c) present day and (b, d) midcentury. The
1000 absorption of shortwave radiation is calculated as the differences of radiative perturbations
1001 between top of atmosphere and surface. Vertical velocity is calculated as the longitudinal
1002 average between 105°W and 112.5°W (two blue lines in a). Positive (negative) values
1003 indicate descending (rising) motion. Results for the 2010s are calculated as (F10AERO -
1004 F10CTRL). Results for the 2050s are calculated as (F50AERO - F50CTRL). Significant
1005 changes ($p < 0.05$) in (a, b) are indicated as black points.

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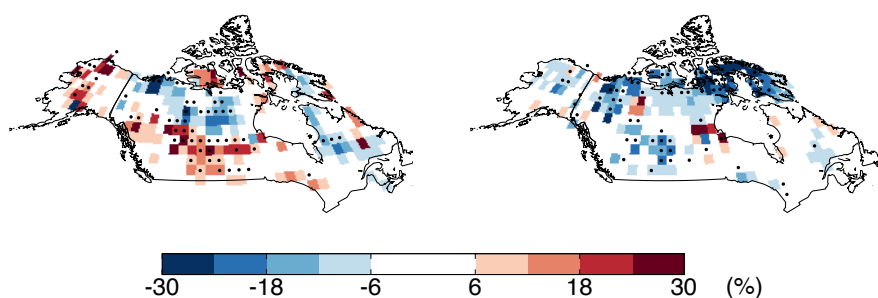
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1009 **Figure 8.** Predicted changes in summertime (a, b) surface air temperature, (c, d) precipitation,
1010 and (e, f) soil water content at surface caused by aerosols from wildfire emissions at (a, c, e)
1011 present day and (b, d, f) midcentury. Results for temperature and precipitation are shown as
1012 absolute changes. Results for soil water are shown as relative changes. Results for the 2010s
1013 are calculated as (F10AERO - F10CTRL). Results for the 2050s are calculated as (F50AERO
1014 - F50CTRL). Significant changes ($p < 0.05$) are marked with black dots.
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(a) Δ NPP by fire aerosol in 2010

(b) Δ NPP by fire aerosol in 2050



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1023 **Figure 9.** Predicted percentage changes in summer NPP caused by wildfire aerosols at (a)
1024 present day and (b) midcentury. Results for the 2010s are calculated as $(F10AERO/F10CTRL$
1025 $- 1) \times 100\%$. Results for the 2050s are calculated as $(F50AERO/F50CTRL - 1) \times 100\%$.
1026 Significant changes ($p < 0.05$) are marked with black dots.

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