Wintertime Aerosol Optical and Radiative Properties in the Kathmandu Valley during the SusKat-ABC Field Campaign

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Abstract. Particulate air pollution in the Kathmandu Valley has reached severe levels that are mainly due to uncontrolled emissions and the location of the urban area in a bowl-shaped basin with associated local circulations. The AERONET measurements from December 2012 to August 2014 revealed a mean aerosol optical depth (AOD) of approximately 0.3 at 675 nm during winter, which is similar to that of the post-monsoon but half of that of the pre-monsoon AOD (0.63). The distinct seasonal variations are closely related to regional-scale monsoon circulations over South Asia and emissions in the Kathmandu Valley. During the SusKat-ABC campaign (December 2012–February 2013), a noticeable increase of both aerosol scattering (σ_s; 313→522 Mm^{-1} at 550 nm) and absorption (σ_a; 98→145 Mm^{-1} at 520 nm) coefficients occurred before and after January 4, 2013. This can be attributed to the increase of wood-burned fires due to a temperature drop and the start of firing at nearby brick kilns. The σ_s value in the Kathmandu Valley was a factor of 0.5 lower than that in polluted cities in India. The σ_a value in the Kathmandu Valley was approximately 2 times higher than that at severely polluted urban sites in India. The aerosol mass scattering efficiency of 2.6 m^2 g^{-1} in the Kathmandu Valley is similar to that reported in urban areas. However, the aerosol mass absorption efficiency was determined to be 11 m^2 g^{-1}, which is higher than that reported in the literature for pure soot particles (7.5 ± 1.2 m^2 g^{-1}). This might be due to the fact that most of the carbonaceous aerosols in the Kathmandu Valley were thought to be fresh aerosols, mostly externally mixed with other aerosols under dry conditions due to a short travel time from their sources. The σ_s and σ_a values and the equivalent black carbon (EBC) mass concentration reached up to 757 Mm^{-1}, 224 Mm^{-1}, and 29 μg m^{-3} at 08 LST, respectively but decreased dramatically during the daytime (09–18 LST), to one quarter of the morning average (06–09 LST) due to the development of valley winds and an atmospheric bounder layer. The σ_s and σ_a values and the EBC concentration remained almost constant during the night at the level of 410 Mm^{-1}, 130 Mm^{-1}, and 17 μg m^{-3}, respectively. The average aerosol direct radiative forcings over the intensive measurement period were estimated to be -6.9 ±

1.4 W m^{-2} (top of the atmosphere) and -20.8 ± 4.6 W m^{-2} (surface). Therefore, the high atmospheric forcing (i.e., 13.9 ± 3.6 W m^{-2}) and forcing efficiency (74.8 ± 24.2 W m^{-2} τ^{-1}) can be attributed to the high portion of light-absorbing aerosols in the Kathmandu Valley, as indicated by the high BC (or elemental carbon) to sulphate ratio (1.5 ± 1.1).

1. Introduction

5   The Kathmandu Valley in Nepal has been experiencing increasingly serious air pollution over recent decades because of rapid population growth and urbanization and the associated growth of economic activities and fuel consumption (Central Bureau of Statistics, 2006). Particulate air pollution in the Kathmandu Valley has reached severe levels that are mainly due to uncontrolled emissions from motor vehicles (e.g., super-emitters such as diesel trucks), brick kilns, fugitive soil dust, and biomass and garbage burning (Panday and Prinn, 2009; Kim et al., 2015; Sarkar et al., 2016). Electrical power shortages in

10   Kathmandu Valley, especially during dry cold months, have not only led to increased low-grade coal and biomass/biofuel burnings but also to the use of mostly small diesel generators between 2011 and 2016 (Chen et al., 2015). In addition, the Kathmandu Valley is located downwind of Indo–Gangetic Plain regions that experienced rapid growth in emissions and severe air quality degradation. All of this has resulted in elevated air pollution concentrations, especially during the dry season (Sapkota and Dhaubhadel, 2002; Shrestha and Shrestha, 2005; Sharma et al., 2012; Kiros et al., 2015). In addition, the bowl-

15   shaped structure of the Kathmandu Valley, with a bottom altitude of ~1.3 km and a rim of 2 to 2.8 km, and topography-induced circulations have made the air pollution more serious by limiting the transport of air pollutants out of the valley (Regmi et al., 2003; Panday and Prinn, 2009).

Many studies reported various atmospheric compounds, mainly gases and aerosols, and associated temporal patterns with meteorological conditions in the Kathmandu Valley (Sapkota and Dhaubhadel, 2002; Shrestha and Shrestha, 2005; Stone et

20   al., 2010; Chen et al., 2015; Kim et al., 2015; Kiros et al., 2015). Panday and Prinn (2009) described a very regular pattern of morning and evening peaks of the CO and PM_{10} concentrations occurring daily in the valley bottom, interspersed with low values in the afternoons and at night. Sharma et al. (2012) communicated that the black carbon (BC) concentration shows a significant temporal variation with a monthly mean concentration varying between 14.9 µg m^{-3} and 3.0 µg m^{-3}. Chen et al. (2015) observed that the total suspended particles (TSPs) and polycyclic aromatic hydrocarbons (PAHs) in the TSP were

25   relatively high in the Kathmandu Valley compared with those observed in other parts of Asia. Most previous studies performed in the Kathmandu Valley focused on adverse health effects and reduced visibility; however, limited measurement-based research was conducted to possibly estimate the radiative effects of air pollutants (e.g., Ramana et al., 2004).

In this study, we analyzed the aerosol optical and chemical properties and their direct radiative effects in the Kathmandu Valley during the intensive phase of the Sustainable Atmosphere for the Kathmandu Valley-Atmospheric Brown Clouds (SusKat-

30   ABC) campaign (December 2012–February 2013). In Section 3, we present seasonal variations of column-integrated and surface aerosol optical and radiative properties during the dry winter season. The diurnal cycle of aerosol properties and its relationship to local circulations in the valley is discussed in Section 4. Aerosol direct radiative effects estimated from columnar

aerosol property measurements are given in Section 5. The conclusions drawn based on the findings are presented in Section 6.

2. Methods

All ground-based *in situ* and column-integrated aerosol measurements were performed at the SusKat-ABC Supersite in Bode (27.68°N; 85.39°E; 1,360 m above the mean sea level), which is located approximately 3 km east of the edge of Kathmandu city. Throughout the winter season, westerly winds generally transport polluted air masses from the urban area of Kathmandu to the observational site, from late morning until sunset (Regmi et al., 2003; Panday and Prinn, 2009).

The aerosol scattering coefficient (σ_s) was measured at three wavelengths (450, 550, and 700 nm) with a nephelometer (Model 3563, TSI Inc.) from December 2012 to February 2013. The nephelometer-based scattering Ångström exponent (SÅE) was calculated from σ_s at 450 and 700 nm. The nephelometer was calibrated using CO_2 before, during, and after field deployment and all data were processed by applying the correction method described in Anderson and Ogren (1998). The aerosol absorption coefficient (σ_a) was measured with a 7-wavelength aethalometer (370, 470, 520, 590, 660, 880, and 950 nm; Model AE31, Magee Scientific) over the campaign period. The aethalometer was run at a flow rate of 3.9 L min^{-1}. However, the aethalometer data were analyzed only from December 21, 2012, to January 9, 2013, because the aethalometer data for the rest of the sampling period were not usable due to interference associated with an intermittent power supply at the observational site. The aethalometer data were corrected for filter-loading effects following the method suggested in Hansen et al. (1984) and Schmid et al. (2006).

Column-integrated spectral aerosol optical properties, such as the aerosol optical depth (hereafter, AOD or τ), single scattering albedo (ω), and asymmetric parameter (g), were obtained with the collocated 8-channel AERONET Cimel Sun/sky radiometer (340, 380, 440, 500, 675, 870, 1020, and 1640 nm) from December 2012 to August 2014. The Ångström exponent (Å) was calculated using linear regression of $\ln(\tau)$ versus $\ln(\lambda)$ (i.e., wavelength) with the data obtained at 380, 440, 500, 675, and 870 nm. These aerosol parameters were used in the column radiative model (CRM) as input data to calculate the aerosol direct radiative forcing and forcing efficiency (Won et al., 2004; Kim et al., 2010). The CRM uses a δ-Eddington approximation with 19 spectral intervals spanning from 0.2 to 5 µm and accounts for several gas absorption spectra; seven spectral bands of O_3 and H_2O, three of CO_2, and absorption bands of N_2O, CH_4, CFC-11, and CFC-12 are included (Slingo, 1989; Briegleb, 1992). In this study, a linear regression method in the log λ–log χ plane (χ denotes τ, ω, and g) was used to determine the aerosol parameters in 19 λ bands and the number of computational vertical layers was extended from 18 to 54 based on Won et al. (2004).

Twenty-four-hour PM_{10} filter samples, each starting at 09:00 local standard time (LST), were collected between December 2012 and February 2013 using two PM_{10} sequential samplers (Model PMS-103, APM Engineering) with a sample flow rate of 16.7 L min^{-1}, which were placed on the roof of a 5-story building at the Bode Supersite (Kim et al., 2015). The PM_{10} filter samples were analyzed for chemical composition, including organic carbon (OC), elemental carbon (EC), and ions, at the

laboratory of the National Institute of Environmental Research (NIER), Korea. The water-soluble ionic species were extracted from Zefluor™-supported PTFE filters and analyzed using ion chromatography (Dionex IC 2000) following the USEPA IO-4.1 method. The PM$_{10}$ samples collected on quartz fiber filters (pretreated at 850°C for 4 hours before sampling) were analyzed for EC and OC using a thermal optical transmittance carbon analyzer (Lab OC-EC Aerosol Analyzer Model 5, Sunset Laboratory Inc.) following the USEPA NIOSH 5040 protocol (Kim et al., 2015).

3. Aerosol optical and radiative properties

3.1 Seasonal and monthly variation of columnar aerosol optical properties

Figure 1 shows the seasonal/monthly variations of AOD, Å, and ω and the aerosol volume size distribution from AERONET Cimel Sun/sky radiometer measurements at the Bode Site during December 2012–August 2014. Here, we note that the climate seasons in Nepal are divided into winter (December–February), pre-monsoon (March–May), summer monsoon (June–September) and post-monsoon (October–November) season (Sharma et al., 2012). The AOD (675 nm) maxima were observed in the pre-monsoon season (0.63 ± 0.12). The AOD was approximately 0.3 in the summer monsoon, post-monsoon, and winter seasons. As many studies reported, these temporal variations are closely related to distinct seasonal monsoon circulations over South Asia including the Himalayan foothills (Sapkota et al., 2002; Sharma et al., 2012). Heavy precipitation in summer months (June–September) due to warm and marine air masses advected from the Indian Ocean by southerly flows (Carrico et al., 2003) efficiently removes atmospheric aerosols through the wet scavenging process (Sharma et al., 2012). After the end of the summer monsoon, on the other hand, atmospheric aerosol loadings gradually increase under the dry and stable atmospheric conditions in the Himalayan foothills and local emissions also increase (e.g., increase of the biomass burning in houses in cold months and use of low-grade coal and biomass in brick factories and kilns, which are operational from mid-December to mid-April). The peaks of AOD occurring in the pre-monsoon season can be explained by the combined effects of locally emitted pollution, wind-blown mineral dust particles, and regional forest and farm fires in the region.

The monthly mean Å ranged between 1.1 and 1.6 during December 2012 and August 2014 (Fig. 1b), which implies that fine-mode particles are predominant all year round; however, coarse soil dust particles, mainly emitted from unpaved roads and farmland in the Kathmandu Valley, also exist, especially during dry winter and pre-monsoon periods. For example, about 26% of the average PM$_{10}$ concentration in winter was attributed to soil dust during the SusKat-ABC campaign (Kim et al., 2015). This is more clearly evident in the aerosol volume size distribution (Fig. 1d). The comparison of the March–May period with other months showed an increase in coarse mode particles with a modal radius of 3 to 6 μm due to the presence of soil dust particles, as mentioned above. Enhanced fine mode aerosols with a modal radius of 0.1–0.3 μm in the pre-monsoon period can be explained by more emissions of accumulation mode particles from brick kilns (Sarkar et al., 2016), motor vehicles, diesel generator sets, and forest fires and the frequent occurrence of new particle formation and subsequent growth events (Venzac et al., 2008). The aerosol volume size distribution in the monsoon and post-monsoon periods indicated a minimal content of coarse particles.

The average ω at 675 nm in the Kathmandu Valley during winter and pre-monsoon was 0.84±0.01 (Fig. 1c), which is slightly lower than that at other locations in Asia (Xu et al., 2004; Kim et al., 2005). The low ω is likely due to a higher ratio of light-absorbing carbonaceous aerosols (i.e., black carbon) to total aerosol concentrations (Putero et al., 2015; Kim et al., 2015). This low value of ω was also observed in ground-based *in situ* measurements (Section 3.2), which results in high atmospheric radiative forcings (Section 5).

3.2 Surface aerosol optical and chemical properties during the SusKat-ABC campaign

Figure 2 shows the variations of the daily mean σ_s, SÅE, ions, and total carbon (TC: sum of organic and elemental carbon) mass concentration in PM_{10} from surface-based measurements at the Bode Supersite during the SusKat-ABC campaign. The σ_s and SÅE values averaged for the entire measurement period were estimated to be 432 ± 266 Mm^{-1} and 1.3 ± 0.2, respectively; however, the changes were significant as of January 3–4 and 18–19, 2013, respectively. To explain the variability of these properties during the experiment, we divided our dataset into three periods: Period I (December 21, 2012, to January 3, 2013), Period II (January 4–18, 2013) and Period III (January 19 to February 14, 2013). The σ_s value averaged over Period I was 313 ± 135 Mm^{-1} but distinctly increased in Period II (577 ± 310 Mm^{-1}; Fig. 2a). The main source was biogenic emission in Period I based on Sarkar et al. (2016). In Periods II and III, they observed high acetonitrile, benzene, and isocyanic acid due to emissions from biomass co-fired brick kilns and other biofuel/biomass burning activities (see the bottom of Fig. 2). We determined much higher scattering coefficients during dust and pollution events in Asia than previous studies (Kim et al., 2005; see Fig. 4 for further comparison). The σ_s value during Period III resumed similar levels as in Period I, with an arithmetic mean remaining at 426 ± 262 Mm^{-1}. The SÅE value during Period I ranged from approximately 1.3 to 1.6 but suddenly decreased to below 1.2 in Period II (Fig. 2b). We cannot explain why we observed the lowest value of SÅE (mean: 1.1 ± 0.2) in Period II because most of the fine scattering aerosols, such as SO_4^{2-} and NH_4^+, increased sharply from January 4, 2013. The Ca^{2+} concentration, which is an indicator of soil dust, did not increase much in Period II. The higher concentrations of SO_4^{2-} and NH_4^+ in Period II were mainly due to increased emissions from brick kilns. The operation of brick kilns in 2013 started in the first week of January; almost all brick kilns in the Kathmandu Valley started to operate by 20 January 2013. This contributed to the increase of these ions and that of σ_s. This means that the brick kiln activities in the valley are one of the predominant factors affecting the aerosol concentrations and aerosol optical properties in both Periods II and III compared with Period I. This can also be explained by a slight increase of the TC concentration in Periods II and III.

The cations and anions (24 ± 8 µg m^{-3}) accounted for approximately 17% of the average total PM_{10} mass concentration (153 ± 36 µg m^{-3}) during the entire measurement period. Table 1 shows the comparison of the ion mass concentration observed at polluted urban environments in Nepal, India, and China. Although the ion concentrations were measured for different sampling sizes, the ion concentrations in PM_{10} measured at the Kathmandu Valley were much lower than those in polluted urban cities in India (Hisar, Kanpur, and Allahabad) and China (Beijing). In contrast to the low ion concentrations, we found high TC concentrations in the PM_{10} filter samples. The BC (EC) to sulphate ratio for Periods I–III was ~2.9 ± 1.4, 1.1 ± 0.5, and $1.0 \pm$

0.3, respectively, which is much higher than the values reported in East and South Asia (Ramana et al., 2010). This is consistent with the low ω and high atmospheric forcing that we report in Sections 4 and 5.

Figure 3 shows the temporal variations of the aerosol properties and meteorological parameters from December 21, 2012, to January 9, 2013. Similar diurnal variations of σ_s, σ_a, EBC mass concentration, and meteorological parameters were observed during that period (see Section 4 for detailed discussion). As seen earlier in Fig. 2, Period II (January 4–9, 2013) includes values of both σ_s (522 ± 265 Mm^{-1} at 550 nm) and σ_a (145 ± 82 Mm^{-1} at 520 nm) that are ~1.5 times higher than that in Period I (December 21, 2012–January 3, 2013; σ_s: 313 ± 135 Mm^{-1}, σ_a: 98 ± 146 Mm^{-1}). The EBC mass concentration also increased from 12 ± 6 µg m^{-3} in Period I to 19 ± 10 µg m^{-3} for the period of January 4–9, 2013. However, ω did not show any visible change before and after January 4, 2013. The winds did not change much during the entire sampling period; however, the surface temperature decreased slightly after January 4, 2013 (i.e., Period II), which led to the increased usage of wood-burned fires indoors or outdoors. Additionally, as mentioned above, the emissions from biomass co-fired brick kilns increased in Period II after January 4, 2013 (Sarkar et al., 2016). Previous studies reported that the concentrations of PM$_{10}$, particulate carbon, SO$_2$, NO$_x$, and volatile organic compounds increased considerably when nearby brick kilns were operating (Raut, 2003; Pariyar et al., 2013; Kiros et al., 2015; Sarkar et al., 2016; Sarkar et al., 2017).

The comparison of the aerosol properties observed in the Kathmandu Valley with those reported in previous studies in East and South Asia is given in Fig. 4. The σ_s value in Kathmandu Valley is a factor of 0.5–0.6 lower in polluted urban cities (Hisar and Allahabad) in India, while it is similar to that of Beijing (polluted urban) and Linan (rural but affected by a large city) in China. Compared with the values measured at background sites in Asia (Himalayas, Hanimadhoo, Kaashidhoo, Anmyeon, Gosan, and Yulin) and at an urban site in Mexico City, the σ_s value in Kathmandu Valley is about 4–5 times higher.

The σ_a value in the Kathmandu Valley was far higher than in any of the other studies listed in Fig. 4 throughout the measurement period. The σ_a value in the Kathmandu Valley was more than 100 times higher than that at the high-altitude site in the Himalaya (Nepal Climate Observatory–Pyramid, 5079 a.m.s.l.), roughly a factor of 12 higher than that at other background sites, and more than 2 times higher than that at severely polluted urban sites in India. The observed high σ_a is consistent with the large portion of light-absorbing aerosols (i.e., black carbon or EC), as represented by the low ω with a range of 0.6–0.85 at 520 nm (Section 3.1).

3.3 Aerosol mass scattering and absorption efficiencies

Table 2 shows the aerosol mass scattering efficiency (Q_s, m^2 g^{-1}) and aerosol mass absorption efficiency (Q_a, m^2 g^{-1}) estimated from the simultaneous measurements of aerosol optical and chemical properties. The Q_s, calculated as the ratio of the averaged σ_s at 550 nm to the PM$_{10}$ mass concentration over the period of January 4–9, 2013 (Period II), was a factor of 1.3 higher than that in Period I. The enhanced formation of secondary scattering aerosols from increased emissions of aerosol precursors (e.g., SO$_2$), mainly from brick factories, as evident in the abundance of SO$_4^{2-}$, NO$_3^-$, and NH$_4^+$ concentrations in this period, is thought to be a possible factor for a higher Q_s value in Period II, in addition to direct emissions from brick kilns, motor vehicles, and

house heating. Similar estimates of Q_s (2.2–4.4 m^2 g^{-1}) have been found in other studies for both urban and rural areas (Bergin et al., 2001; Xu et al., 2002; Kim et al., 2005).

The Q_a value was calculated as the ratio of the averaged σ_a at 520 nm to the EC mass concentration from PM$_{10}$ measurements. It was ~11 m^2 g^{-1} over the entire measurement period. This value is within the range of values reported for polluted regions of

5   5–25 m^2 g^{-1} in previous studies (Liousse et al., 1993; Xu et al., 2002; Kim et al., 2005; Schwarz et al., 2008). As reported by Bond and Bergstrom (2006), the Q_a value for purely emitted soot particles at 550 nm is 7.5 ± 1.2 m^2 g^{-1}; however, Q_a could double because the particles are mixed with other aerosol constituents (Jacobson et al., 2012). We can conclude that most of the carbonaceous aerosols in Bode were fresh aerosols that were mostly externally mixed with other aerosols under dry conditions during the transport over short distances from their sources, as discussed in Kim et al. (2015).

10   **4. Diurnal variations of aerosol and meteorological properties**

Figure 5 shows the diurnal cycles of aerosol and meteorological parameters, averaged over the study period. A distinct diurnal variation of the σ_s and σ_a values and EBC mass concentration was observed, with two peaks occurring at 08:00 and 19:00 LST. On the other hand, the wind speed, temperature, and mixing layer height showed a peak around 15:00–16:00 LST. As already mentioned above, ω did not show much visible change during a day. This suggests that the proportion of light-absorbing

15   aerosols of the total aerosol composition remained unchanged, likely because these were emitted from the same sources in the valley.

The distinct diurnal variations of aerosol optical properties were closely linked with unique local circulations in the Kathmandu Basin, as explained in several previous studies (e.g., Panday and Prinn, 2009; Regmi et al. 2003). Figure 6 illustrates 3-h averaged near-surface winds simulated by the Weather Research and Forecast (WRF) model over the Kathmandu Valley and

20   Fig. 7 shows the coupling of winds and aerosol scattering and absorption coefficients for three time periods in a day, separated by color into morning (06–09 LST), daytime (09–18 LST), and night (18–06 LST). A detailed discussion of diurnal variations of aerosol optical and meteorological properties is provided below:

1) 06:00 – 09:00 LST (morning)

25   The meteorological conditions at the valley basin in the morning were almost stagnant flows with calm winds (near surface wind speed ≤ 1 m s^{-1}). The weakening of the downslope winds as the cold pool of air kept building through the night was also conducive to the accumulation of aerosols emitted within the valley basin throughout the night and early morning. The σ_s and σ_a values and EBC mass concentration at 520 nm were rapidly elevated from the value of 488 Mm^{-1}, 146 Mm^{-1}, and 19 µg m^{-3} at 06:00 LST to the maximum value of 757 Mm^{-1}, 224 Mm^{-1}, and 29 µg m^{-3} at 08:00 LST, respectively, because considerable

30   amounts of aerosols, emitted from increased vehicles during the traffic rush hour and cooking activities, were added to pre-existing aerosols of the shallow nighttime boundary layer. The average value of σ_s and σ_a during 3 hours was 628 ± 252 Mm^{-1}

and 194 ± 68 Mm^{-1}, respectively. The SÅE value, which increased from 1.4 at 06:00 LST to 1.8 at 08:00 LST (not shown), suggests that tiny particles were largely emitted from automobile and biofuel/biomass burning in the morning hours.

2) 09:00 – 18:00 LST (daytime)

The development of the mixing layer height (MLH) and valley winds induced by solar heating in the high-altitude Kathmandu Valley played a crucial role in the reduction of the aerosol concentration near the surface by enhancing the vertical and horizontal ventilation (Panday and Prinn, 2009). The MLH estimated from ceilometer measurements varied from about 100 m in the nighttime to more than 1,500 m above the ground during the day, reaching about 1,000 m at 15:00 LST (Fig. 5; Panday and Prinn, 2009; Mues et al., 2017). Within the developed MLH, aerosols were both vertically mixed and ventilated out of the Kathmandu Valley by winds sweeping through. The up-valley winds entering along the Bagmati River Gorge from the south and westerly winds passing through the Mudku, Bhimdhunga, and Nagdhunga passes merged into a strong westerly winds in the Kathmandu Valley. The aerosols were carried by the westerly winds to the low mountain outflow passes at Sanga and Nala (Regmi et al., 2003). The westerly winds recorded speeds up to 5 m s^{-1} during the daytime, efficiently diluted polluted air masses within the valley basin and led to decreased σ_s (250 ± 164 Mm^{-1} at 550 nm) and σ_a (70 ± 50 Mm^{-1} at 520 nm; Fig. 7) values. The σ_s value at 550 nm, σ_a at 520 nm, and EBC mass concentration at 520 nm decreased to minimum values of 130 Mm^{-1}, 37 Mm^{-1}, and 5 µg m^{-3} at 15:00 LST, respectively (Fig. 5).

3) 18:00 – 06:00 LST (nighttime)

When the inflows from southern and western passes retreated (wind speed ≤ 3 m s^{-1}) after the sunset, second peaks of the σ_s and σ_a values and EBC mass concentration appeared approximately 19:00 LST. These high values remained stable throughout the night. In addition to the abrupt collapse of the boundary layer height, the aerosol concentrations of the night were mostly contributed by emissions from vehicles, cooking activities, and house heating in the early evening. The use of diesel-fired power generators might also contribute to the elevated σ_s, σ_a, and EBC mass concentration. Based on Chen et al. (2015), Nepal experienced acute power shortages, particularly during winter (e.g., 10–18-h power cuts in a day), which resulted in the use of nearly 250,000 small diesel generators. Because of the emission of these aerosols, the residual aerosols during the daytime were also trapped in the collapsed boundary layer due to the radiative cooling at the surface. Earlier studies found evidence of nighttime lifting of pollutants over the valley as a result of converging katabatic winds. In the early morning, the elevated pollutants returned to the surface when the mixing layer grew and entrained the lifted layer, as explained before by Panday and Prinn (2009), adding to the growing morning emissions that caused the morning peak described above. The average value of σ_s and σ_a during the night was estimated to be 410 ± 163 Mm^{-1} at 550 nm and 130 ± 47 Mm^{-1} at 520 nm, respectively, which is a factor of 0.6 lower than the value of the maximum peak in the morning.

5. Aerosol radiative forcing effects

Figure 8 shows the daily mean AOD, ω, and aerosol direct radiative forcing (ΔDRF) estimated from the AERONET sun/sky radiometer measurements at the Bode Supersite from December 21 to January 9, 2013. The daily mean AOD at 675 nm ranged from 0.2 to 0.4, with an average AOD of 0.29 ± 0.07. Contrary to *in situ* surface aerosol measurements, which showed a significant change of the aerosol properties between periods I and II (Sections 3 and 4), AOD did not show such a big change between the two periods. This discrepancy can be attributed to the fact that the sun/sky radiometer measurements were made only for the daytime under cloud-free conditions when aerosols were naturally diluted due to the development of a mixing layer and local circulations. Similar to AOD, the aerosol extinction coefficient (σ_e) at 500 nm at the surface during the daytime (09:00– 18:00 LST) was 37% lower in Period I (285 Mm^{-1}) than in Period II (392 Mm^{-1}). In the nighttime, σ_e was about 75% lower in Period I (468 Mm^{-1}) than in Period II (817 Mm^{-1}). This means that the increase of the surface aerosols due to increased emissions in Period II was more concentrated under calm and stable conditions during the night. The daily mean ω (at 675 nm) ranged from 0.84 to 0.90 and the average ω was 0.87 ± 0.02.

Figure 9 shows the relationship between AOD and ω at 675 nm from 1.75-year measurements at the Bode Site including the SusKat-ABC campaign period in the Kathmandu Valley. The AOD and ω (in parenthesis) values at 675 nm for each season were 0.14 ± 0.07 (0.88 ± 0.05) in the monsoon, 0.19 ± 0.13 (0.87 ± 0.05) in the post-monsoon, 0.21 ± 0.1 (0.85 ± 0.04) in the winter, and 0.43 ± 0.2 (0.84 ± 0.06) in the pre-monsoon seasons. Contrary to Kim et al. (2005) who reported that ω increases as σ_s or σ_e increased under polluted conditions, the relationship between the AOD and ω across the four seasons in the Kathmandu Valley showed that ω decreased with increasing AOD. Here, we also compared the AOD and ω at Kathmandu Valley with that from different environments at several sites in East and South Asia. The higher AOD values were observed in polluted regions, such as Beijing (0.39 ± 0.43) and Kanpur (0.48 ± 0.25); however, the AOD was lower in background regions such as Shirahama (0.17 ± 0.13), Maldives (0.22 ± 0.11), Gosan (0.25 ± 0.19), and Anmyeon (0.24 ± 0.22). The AOD (0.28 ± 0.19) in the Kathmandu Valley was similar to that in background regions. However, ω (0.85 ± 0.06) in the Kathmandu Valley was much lower than that in both the background (0.94 in Shirahama, Maldives, and Anmyeon; 0.95 in Gosan) and polluted regions (0.91 in Beijing; 0.90 in Kanpur). The low value of ω in the Kathmandu Valley suggests that the ratio of light-absorbing aerosols to total aerosols was much higher than in other regions, as we discussed in Section 3.

The high portion of light-absorbing aerosols resulted in atmospheric warming and cooling at the top of the atmosphere (TOA) and at the surface. The averaged ΔDRF over the measurement period (December 21,2012–January 9, 2013) was -6.9 ± 1.4 W m^{-2} at TOA and -20.8 ± 4.6 W m^{-2} at the surface and thereby 13.9 ± 3.6 W m^{-2} in the atmosphere (Fig. 8c). The high atmospheric forcing efficiency (74.8 ± 24.2 W m^{-2} τ^{-1}) can be attributed to the high portion of light-absorbing aerosols in the Kathmandu Valley, as indicated by the high BC (or elemental carbon) to sulphate ratio (overall mean: 1.5 ± 1.1; see Section 3.2).

Figure 10 shows the comparison of the ΔDRF and aerosol direct radiative forcing efficiency (ΔDRFe) at various environmental locations. There is a strong seasonal variation of ΔDRF in the Kathmandu Valley, as we discussed for AOD. The ΔDRF dramatically increased in the pre-monsoon season, with the highest values of -5.0 ± 5.7 W m^{-2} (TOA), 39.7 ± 23.9 W m^{-2} (atmosphere), and -44.7 ± 22.1 W m^{-2} (surface). After the dry season, the summer monsoon depressions extended to Nepal

from the southeastern Arabian or northern bay of Bengal, producing heavy rain events (Ueno and Toyotsu, 2008). The substantial daily rainfall in the monsoon season scavenged the aerosols from the atmosphere, resulting in lower or similar values of ΔDRF in the monsoon and post-monsoon seasons with those at other background sites (Maldives, Gosan, Anmyeon, and Shirahama). The average ΔDRF in the Kathmandu Valley over the period of December 2012–November 2015 was estimated to be -6.1 ± 4.3 W m^{-2} (TOA), 22.5 ± 20 W m^{-2} (atmosphere), and -28.6 ± 19.1 W m^{-2} (surface). The ΔDRF value of the atmosphere was roughly a factor of 2.5–3 higher than that at background stations: Maldives (5.8 ± 4.8 W m^{-2}), Shirahama (8.6 ± 6.0 W m^{-2}), Gosan (10.2 ± 8.5 W m^{-2}), and Anmyeon (14 ± 8.7 W m^{-2}). The ΔDRF value of the atmosphere in the Kathmandu Valley was similar to that in polluted sites such as Beijing (34.9 ± 27.4 W m^{-2}) and Kanpur (27.7 ± 15.4 W m^{-2}). As we discussed in Fig. 9, the AOD and ω in Beijing and Kanpur were much higher than those in the Kathmandu Valley. Although the aerosol loadings over the Kathmandu Valley were lower than those in Beijing and Kanpur, this implies that atmospheric warming can be enhanced over the Kathmandu Valley due to the high portion of light-absorbing aerosols. This can be clearly seen in Fig. 10b. The ΔDRF of -26.7 ± 15 W m^{-2} τ^{-1} (TOA), 76.9 ± 32.7 W m^{-2} τ^{-1} (atmosphere), and -103.6 ± 25.9 W m^{-2} τ^{-1} (surface) in the Katmandu Valley was a factor of 1.5 higher than that at two urban sites mentioned above (Beijing and Kanpur) and 3 times higher than that at other background sites.

6. Summary and conclusions

In this study, we investigated the optical and chemical properties of aerosols and their direct radiative effects in the Kathmandu Valley during the SusKat-ABC campaign in the winter of 2012–2013. The principal findings of our analysis are summarized below:

- The 1.75-year AERONET Cimel sun/sky radiometer measurements at the Bode Site in the Kathmandu Valley showed that the AOD in the winter and post-monsoon seasons was about 0.3. The AOD maxima were observed in the pre-monsoon season (0.63 ± 0.09 at 675 nm), whereas AOD minima occurred in the summer monsoon season (0.19 ± 0.1 at 675 nm). The seasonal variations are closely related to distinct seasonal monsoon circulations over South Asia including the Himalayan foothills.

- Both σ_s (average: 522 ± 265 Mm^{-1} at 550 nm) and σ_a (145 ± 82 Mm^{-1} at 520 nm) values increased distinctly after January 4, 2013 compared with the period from December 21, 2012–January 3, 2013 (σ_s: 313 ± 135 Mm^{-1}, σ_a: 98 ± 146 Mm^{-1}). This noticeable increase can be attributed to emissions from brick kilns (operations started January 4, 2013). The σ_s value in the Kathmandu Valley was a factor of 0.5–0.6 lower than those observed in polluted urban cities (Hisar and Allahabad) in India, while it was similar to those in Beijing in China. The σ_a value at Kathmandu Valley was roughly a factor of 12 higher than that at other background sites and more than two times higher than those at severely polluted urban sites in India.

- The aerosol mass scattering efficiency of 2.6 m^2 g^{-1} in Kathmandu Valley is similar to those reported in other urban areas. However, the aerosol absorption efficiency of 11 m^2 g^{-1} is 47% higher than that of pure soot particles (7.5 ± 1.2 m^2 g^{-1}; Bond

and Bergstrom, 2006). This is thought to be due to the fact that most of the carbonaceous aerosols in Kathmandu Valley are fresh aerosols and are mostly externally mixed with other aerosols under dry conditions because of the short travel time/distance from their sources to the observation site at Bode.

- The aerosol optical properties from *in situ* surface measurements showed distinct diurnal variations, coupled with the development of local circulations and the boundary layer height in the Kathmandu Basin. The σ_s, σ_a, and EBC mass concentration values reached up to 757 Mm^{-1}, 224 Mm^{-1}, and 29 µg m^{-3} at 08:00 LST, mainly due to high emissions from increased motor vehicles in the traffic rush hour and cooking activities in the morning (06:00-09:00 LST) and pre-existing aerosols under calm and stable conditions. Because of the development of valley winds and the boundary layer during the daytime (09:00 – 18:00 LST), these values reduced to one quarter of the morning average. After sunset, the σ_s, σ_a, and EBC mass concentration values were back at 410 ± 163 Mm^{-1}, 130 ± 47 Mm^{-1}, and 17 ± 6 µg m^{-3}, respectively, and remained constant throughout the night.

- The ΔDRF in Kathmandu was estimated to be -6.1 ± 4.3 W m^{-2} (TOA), 22.5 ± 20 W m^{-2} (atmosphere), and -28.6 ± 19.1 W m^{-2} (surface). The high atmospheric forcing (i.e., atmospheric warming) and forcing efficiency (76.9 ± 32.7 W m^{-2} τ^{-1}) can be attributed to the high proportion of light-absorbing aerosols in the aerosol composition, as indicated by the high BC (EC) to sulphate ratio (1.5 ± 1.1).

The mitigation of high aerosol concentrations in the Kathmandu Valley is of particular importance, especially that of particulate carbon, to improve the health, visibility, and climate effects. The continuous comprehensive measurements of aerosol chemical compositions and associated optical and radiative properties and detailed air quality modeling in this region are necessary to develop most effective mitigation strategies and control measures. This study only covers a short period of the dry winter season; however, but it provides unprecedented insights into aerosol optical properties and their radiative forcings in the Kathmandu Valley. Further measurements, particularly at the surface level, and investigations during both winter and pre-monsoon periods (November to June) are required to understand the seasonal characteristics at both the surface and columnar levels.

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Table 1: Comparison of ion mass concentrations observed in polluted urban environments. The values reflect the mean and standard deviation (unit: µg m^{-3}).

Country	Site	Sample	SO$_4^{2-}$	NO$_3^-$	NH$_4^+$	K$^+$	Ca^{2+}	Na$^+$	Mg^{2+}	Cl$^-$	Period	Reference
Nepal	Kathmandu	PM$_{10}$	6.1 ± 4.6	4.6 ± 1.3	2.8 ± 2.1	0.9 ± 0.3	1.1 ± 0.4	0.2 ± 0.1	0.1 ± 0.0	2.6 ± 1.1	Dec 2012 –Jan 2013	This Study
India	Hisar	TSP[*]	12.7 ± 1.8	16.7 ± 4.2	6.8 ± 1.9	2.6 ± 0.6	4.4 ± 0.8	0.8 ± 0.2	0.4 ± 0.1	0.4 ± 0.1	Dec 2004	Ram et al. 2012
	Kanpur	PM$_{10}$	12.7 ± 5.1	17.1 ± 1.9	8.9 ± 4.5	1.9 ± 0.9	1.4 ± 1.1	0.5 ± 0.4	0.2 ± 0.2	3.1 ± 2.3	Dec 2007 –Feb 2008	Ram et al. 2012
	Allahabad	TSP[*]	18.6 ± 2.6	9.9 ± 3.7	7.3 ± 2.3	2.5 ± 0.4	5.2 ± 2.2	0.8 ± 0.2	0.6 ± 0.1	0.1 ± 0.1	Dec 2004	Ram et al. 2012
China	Beijing	PM$_{2.5}$	21.3	13.8	10.0	4.6	1.9	1.0	0.3	5.5	30 Nov–09 Dec 2004	Sun et al. 2006

[*]Total suspended particulate

Table 2: Aerosol mass scattering efficiency (at 550 nm) and aerosol mass absorption efficiency (at 520 nm) for each period and all periods combined from December 21, 2012–January 9, 2013 (unit: $m^2 g^{-1}$).

`	Period I (December 21, 2012 – January 3, 2013)	Period II (January 4-9, 2013)	All Periods (December 21, 2012 – January 9, 2013)
Aerosol Scattering Efficiency (α_s)	2.4 ± 0.4	3.2 ± 1.2	2.6 ± 0.8
Aerosol Absorption Efficiency (α_a)	11.9 ± 3.7	9.6 ± 0.6	11.2 ± 0.8
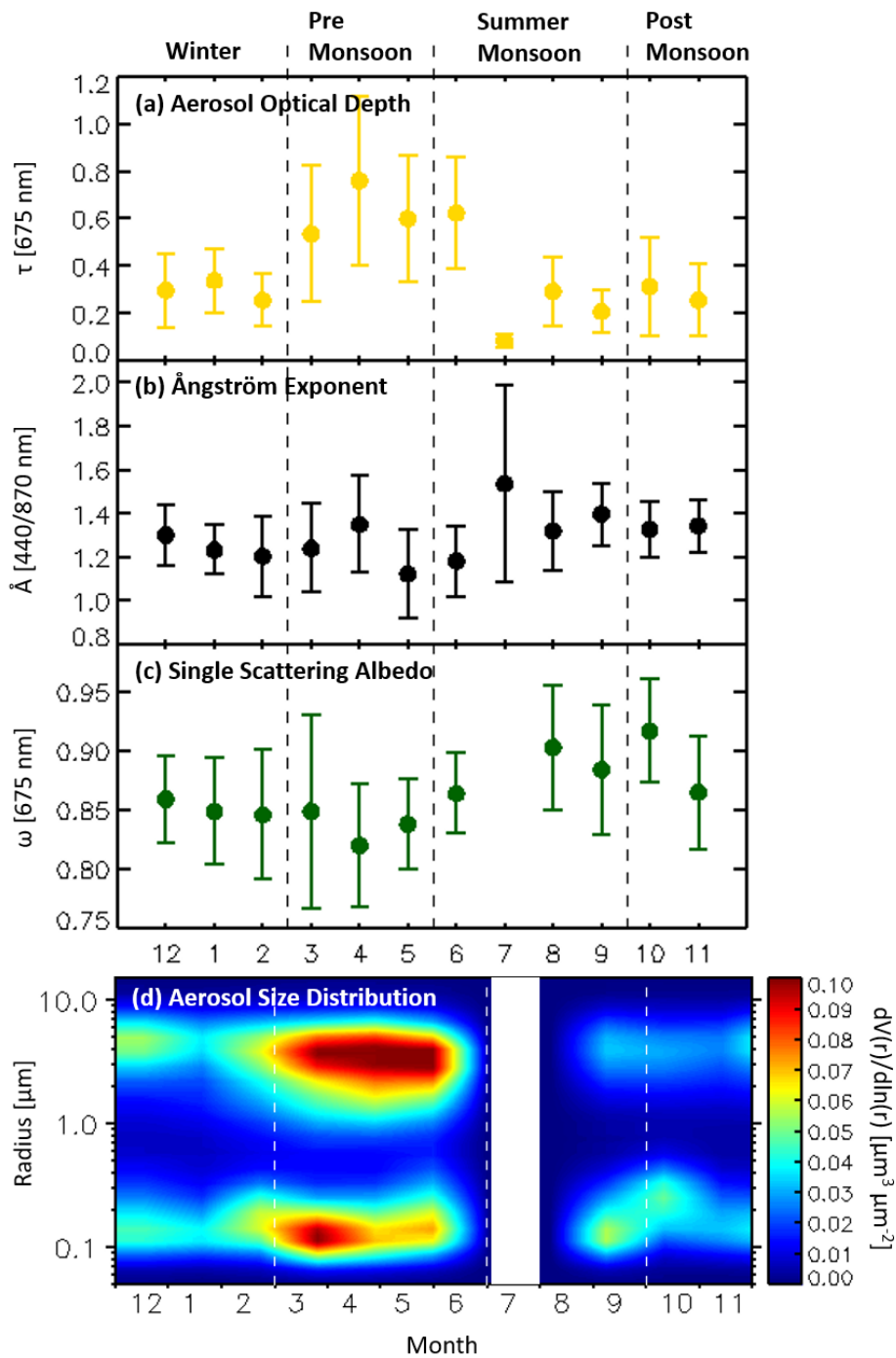
Figure 1: Monthly variation of (a) aerosol optical depth (b) Ångström exponent (c) single scattering albedo, and (d) aerosol volume size distribution from 1.75-year AERONET sun/sky radiometer measurements (December 2012 to August 2014) at Bode, Kathmandu Valley. The four seasons based on Sharma et al. (2012) are noted on top of the figure.
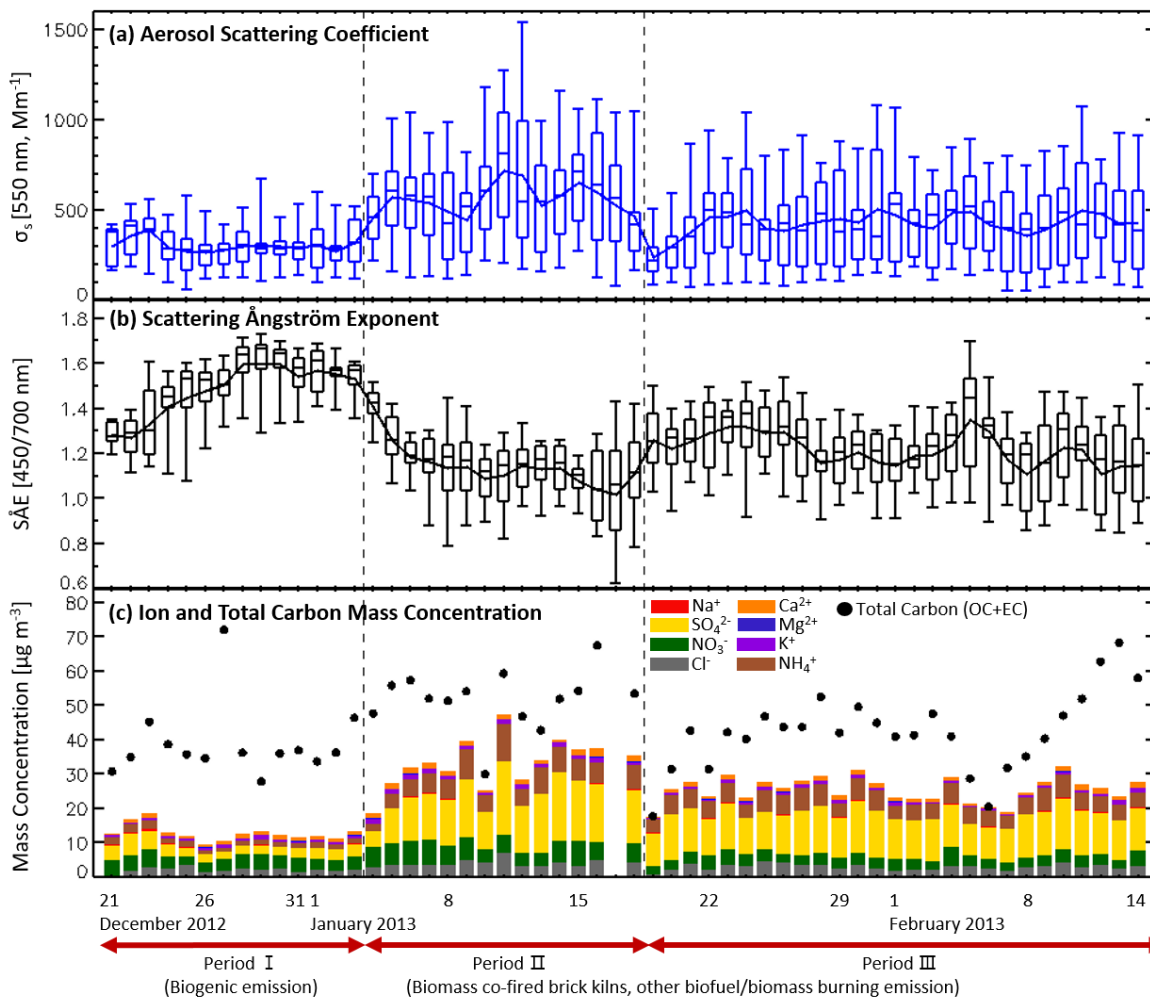
Figure 2: Daily variation of (a) aerosol scattering coefficient, (b) scattering Ångström exponent, and (c) chemical composition of PM₁₀ at the Bode Supersite during the SusKat-ABC campaign. Based on Sarkar et al. (2016), the main source is biogenic emission in Period I (December 21, 2012, to January 3, 2013) and biomass co-fired brick kilns and other biofuel/biomass burning emissions in Period II (January 4–18, 2013). Period III is from January 19 to February 14, 2013.
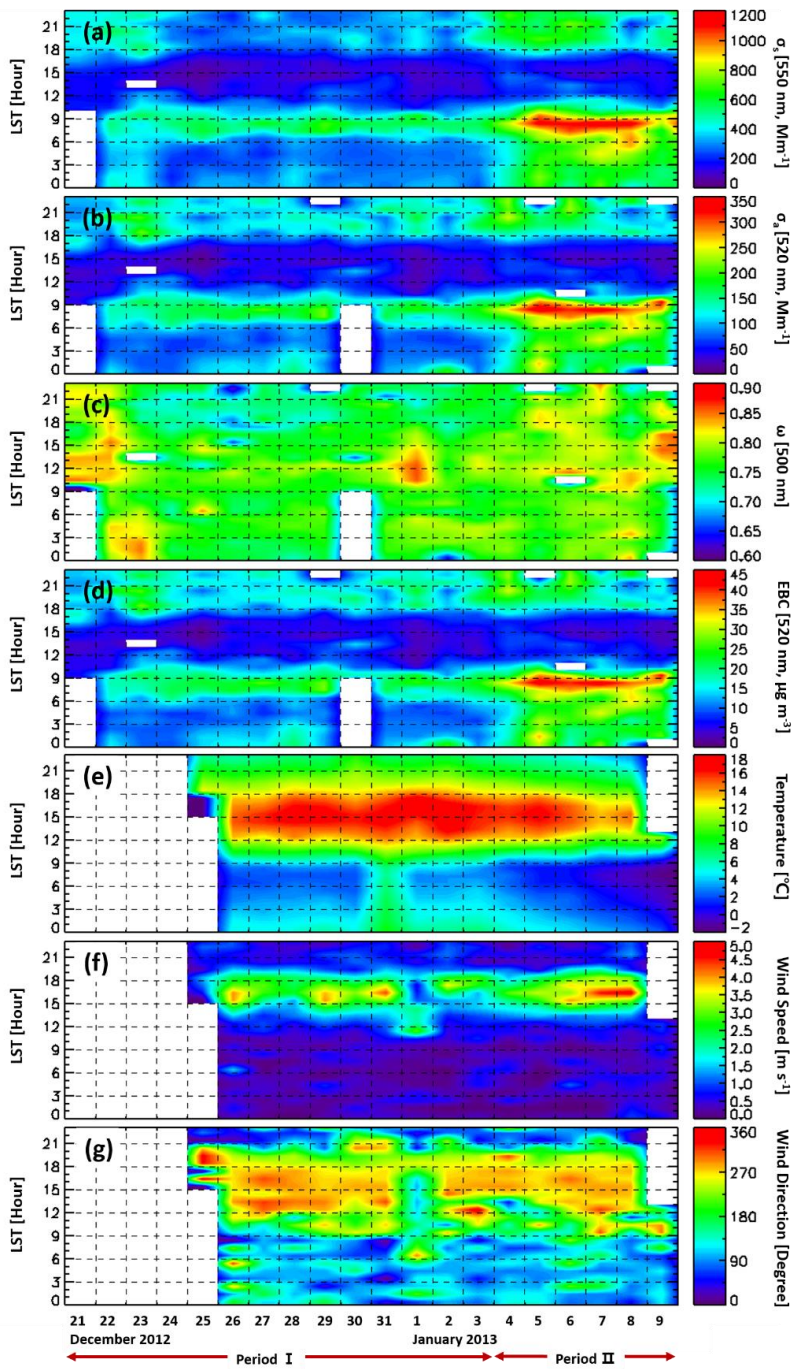
Figure 3: Hourly variation of the (a) aerosol scattering coefficient, (b) aerosol absorption coefficient, (c) single scattering albedo, (d) EBC mass concentration, (e) ambient temperature (f) wind speed, and (g) wind direction at the Bode Site in the Kathmandu Valley over the study period of December 21, 2012–January 9, 2013.
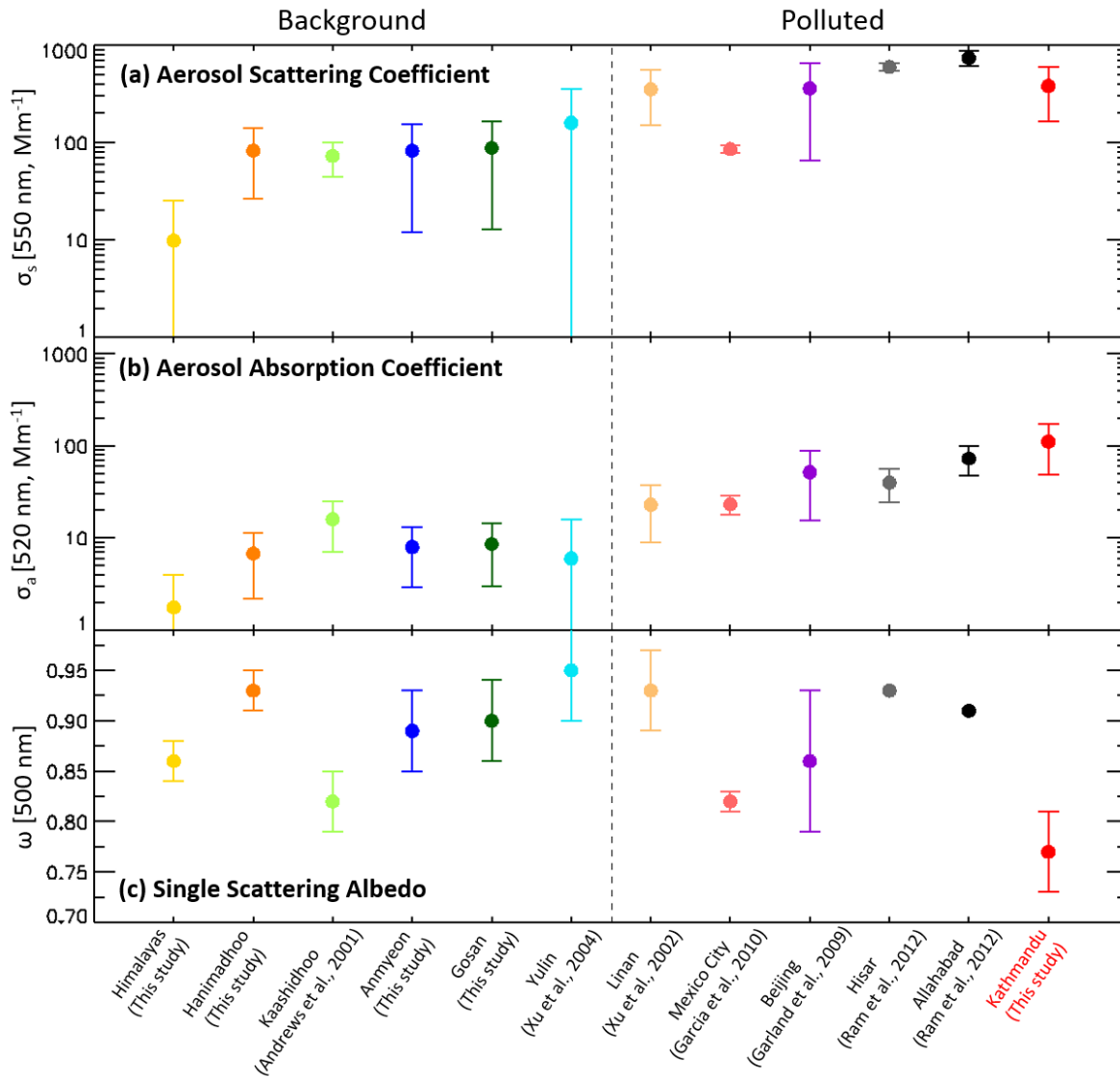
Figure 4: Comparison of the (a) aerosol scattering coefficient, (b) aerosol absorption coefficient, and (c) single scattering albedo measured at several locations worldwide, segregated by the background and polluted area. Note that the scale of the scattering and absorption coefficients is logarithmic (log).
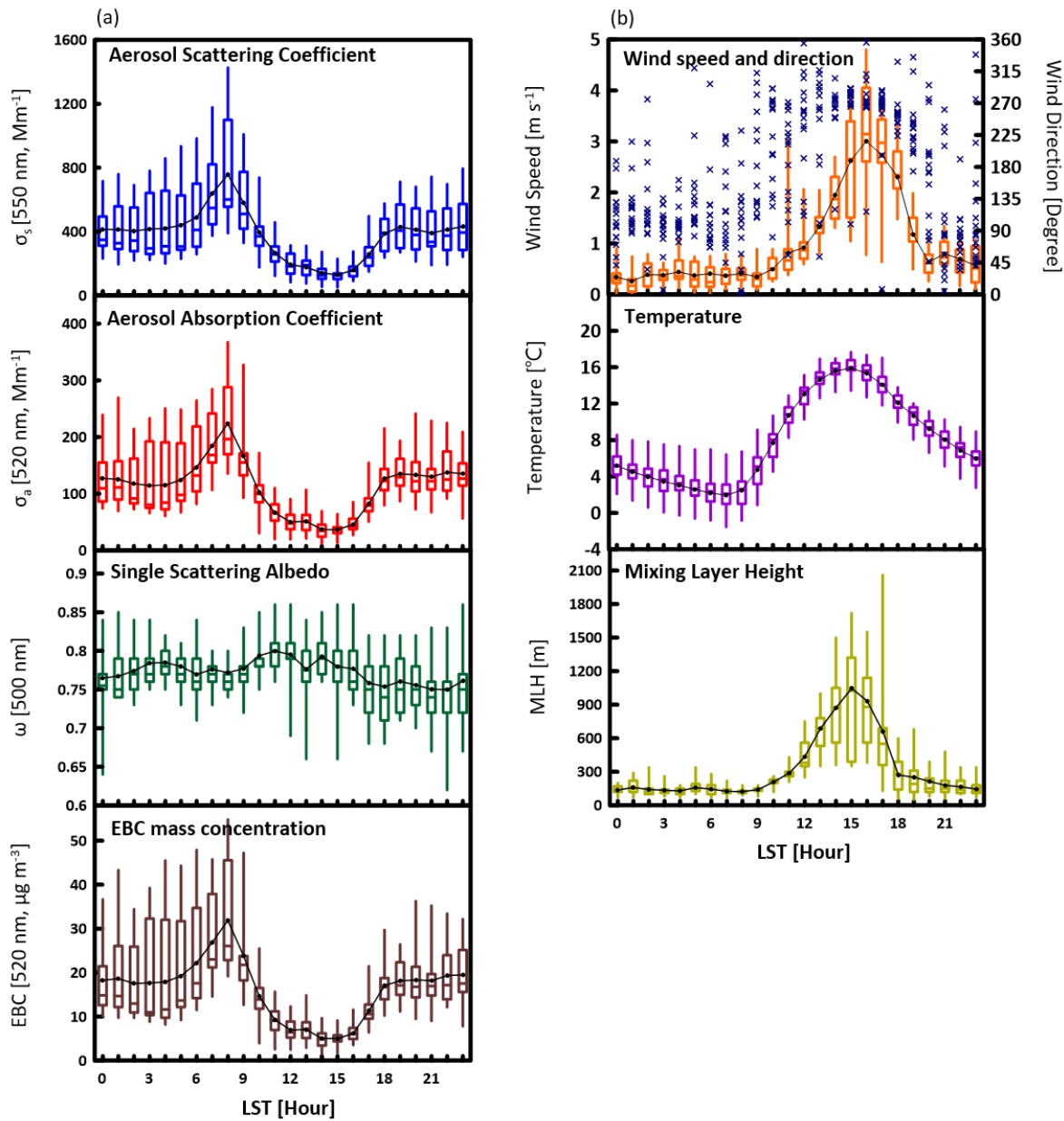
Figure 5: Diurnal cycle of the (a) aerosol optical properties and EBC mass concentration (blue box-whisker; scattering coefficient, red box-whisker; absorption coefficient, green box-whisker; single scattering albedo, brown box-whisker; EBC mass concentration; black lines are mean values) and (b) meteorological parameters (orange box-whisker; wind speed, x symbol; wind direction, purple box-whisker; temperature, yellow box-whisker; mixing layer height; black lines are mean values) at Bode in the Kathmandu Valley. Note that the mixing layer height was measured by a ceilometer from January 27, 2013–February 14, 2013 [see Mues et al. (2017) for more information on ceilometer measurements at Bode].
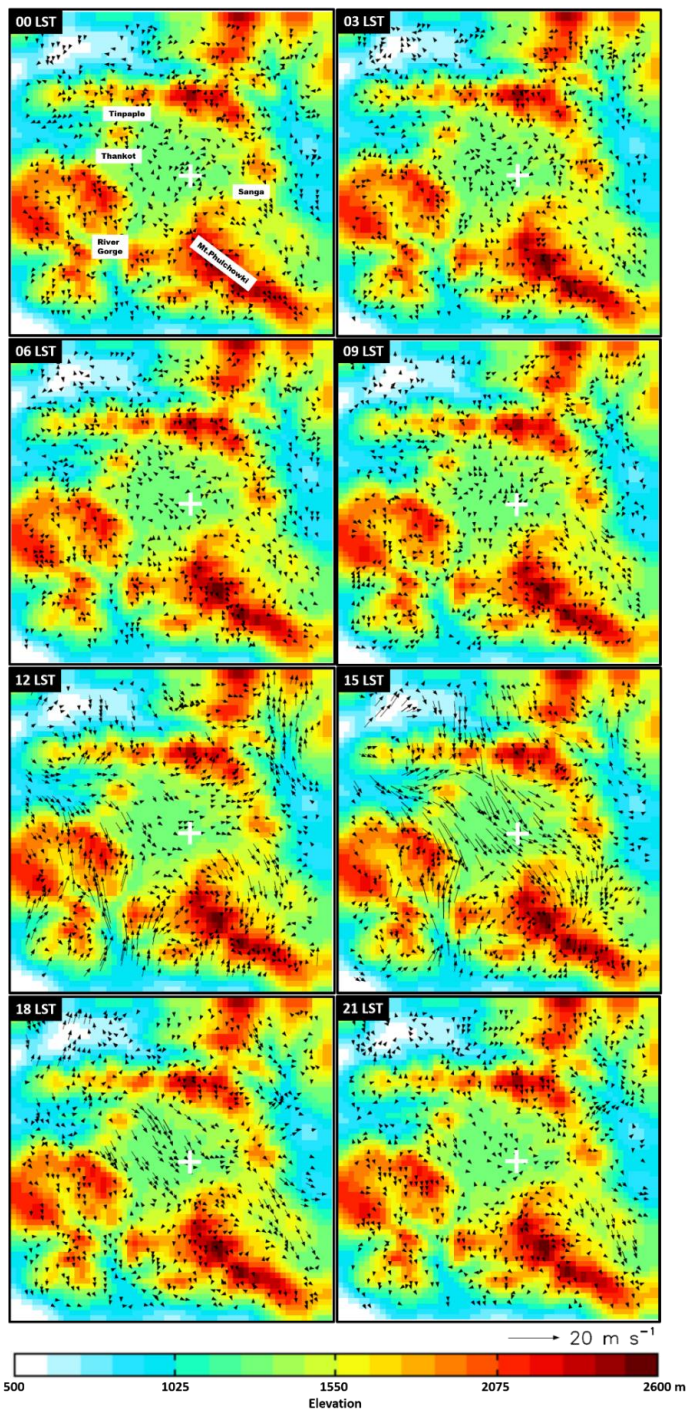
Figure 6: WRF simulation of near surface winds for every 3-h averaged during the study period in Kathmandu. The white crosshair is the location of the Bode Site; the color shade represents the elevation (m).
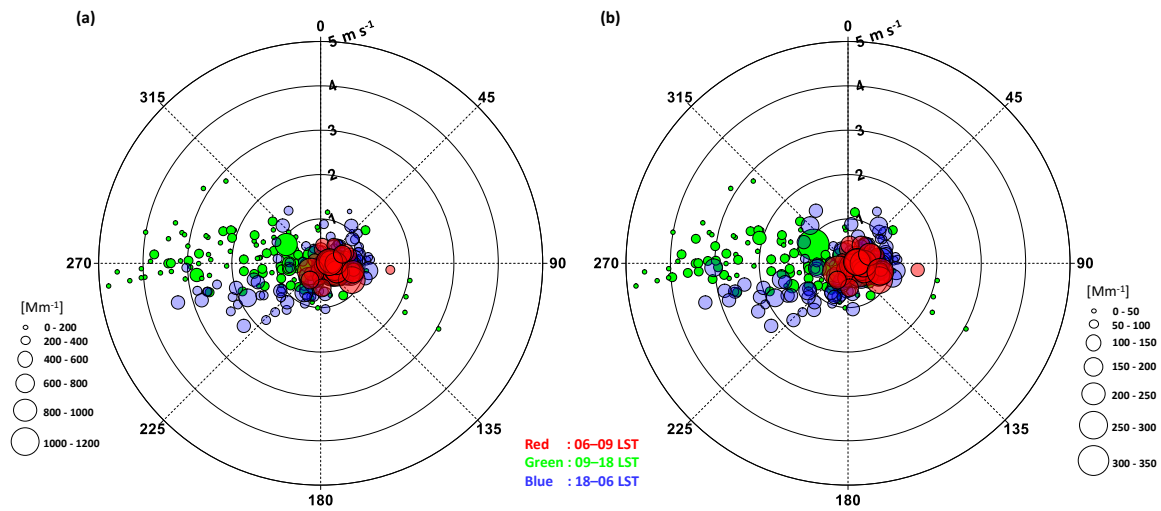
Figure 7: (a) Aerosol scattering coefficient and (b) aerosol absorption coefficient for three time periods of the day: morning (red dot; 06–09 LST), day (green dot; 09–18 LST), and night (blue dot; 18–06 LST) for different wind speeds (m s⁻¹) and wind directions (degree).
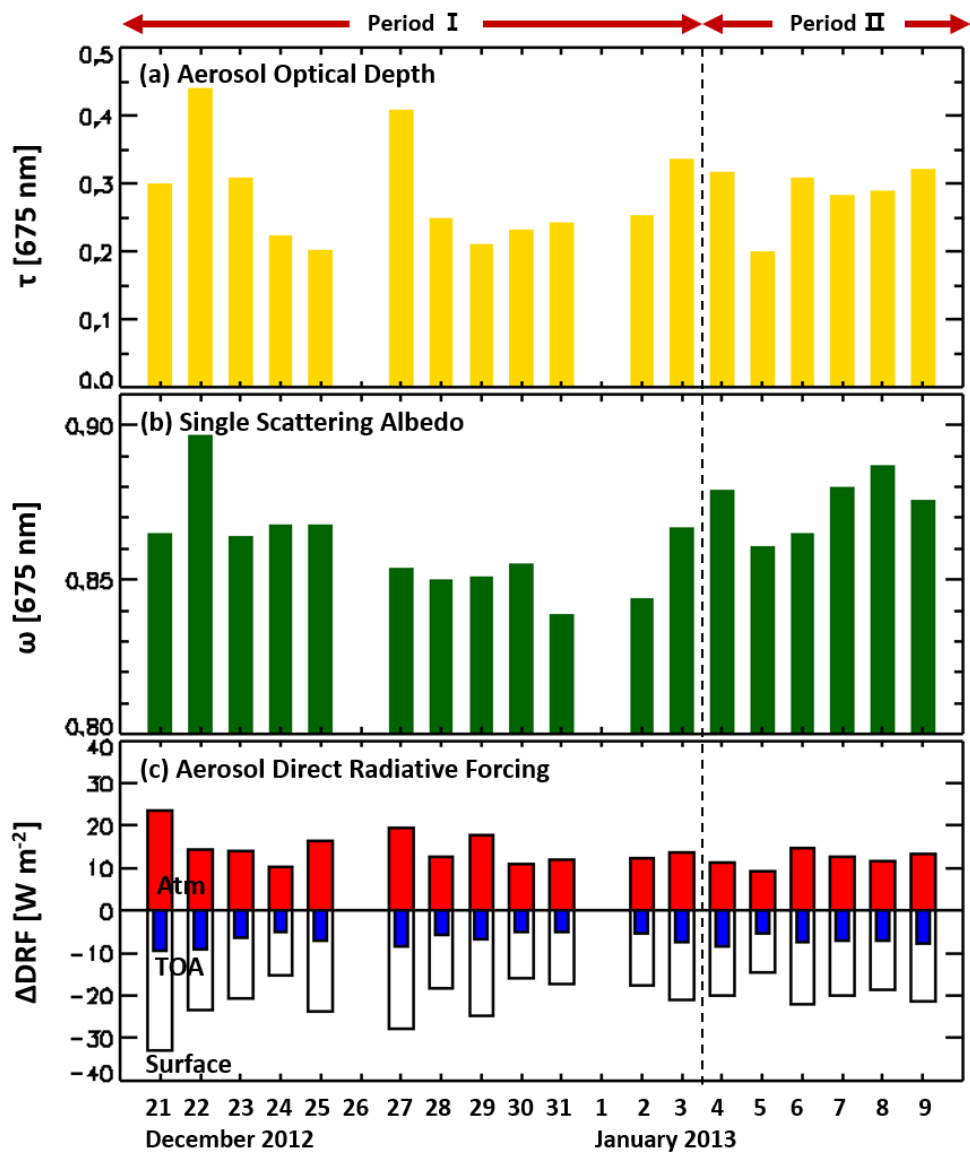
Figure 8: Daily variation of the (a) aerosol optical depth, (b) single scattering albedo, and (c) estimated aerosol radiative forcing at the surface, atmosphere (Atm), and top of atmosphere (TOA) in the Kathmandu Valley from December 21, 2012– January 9, 2013.
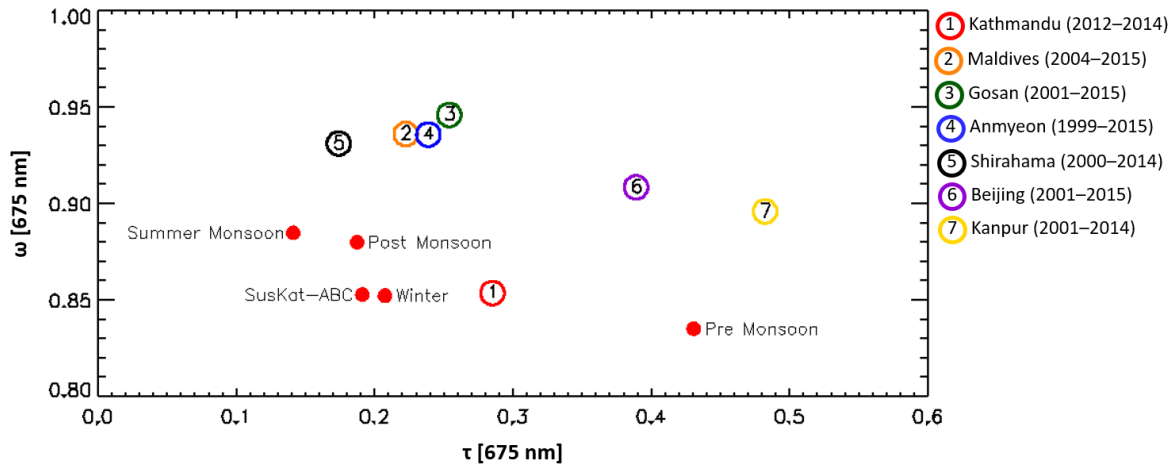
Figure 9: Relationship between th eaerosol optical depth and single scattering albedo measured at several sites in East and South Asia (open circle with number, data period in parentheses) and in the Kathmandu Valley, segregated by season (closed circle).
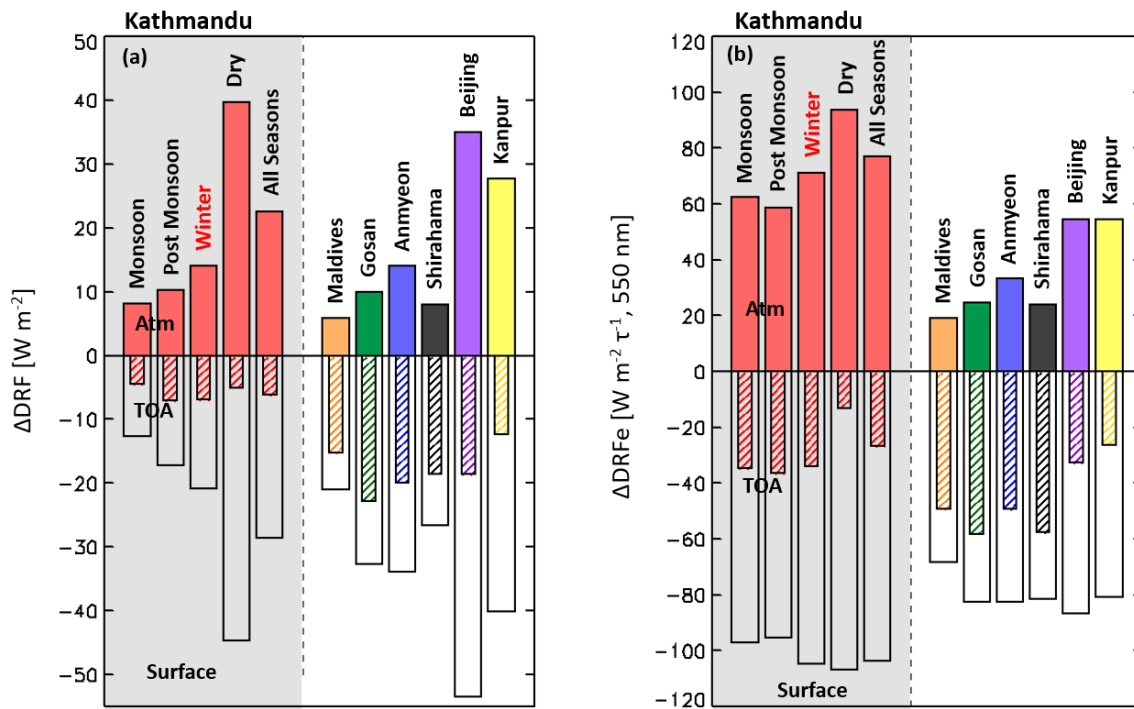
Figure 10: Comparison of the (a) aerosol radiative forcing and (b) aerosol radiative forcing efficiency in the Kathmandu Valley with that of other locations in Asia. Data period is the same as that given in Figure 9.