

28 1. Introduction

29 Apart from mass and chemical composition, the size distribution of fine particles represents a vital physical property
30 with important implications for human health and environmental effects of ambient aerosols (Seinfeld and Pandis,
31 2006). Particle size relates directly to the aerodynamic properties which govern the penetration and deposition of
32 particles in the airways and lungs (Davidson et al., 2005) as well as the scattering and absorption of light which affect
33 the radiative properties and hence ambient visibility (Ahlquist and Charlson, 1967;Bohren and Huffman,
34 1983;Charlson et al., 1991;Schwartz, 1996;Seinfeld and Pandis, 2006). Hygroscopic growth in response to changes in
35 ambient humidity can alter particle light scattering properties (Seinfeld and Pandis, 2006;Köhler, 1936) and activation
36 of condensation nuclei particles into cloud droplets depend on atmospheric conditions, chemical composition, mixing
37 state as well as the size and morphology of particles (Abbatt et al., 2005;Kerminen et al., 2012;Meng et al.,
38 2014;Westervelt et al., 2013).

39 Studies into the size distribution of ambient particulate matter in Hong Kong have been largely based on size-
40 segregated filter samples (Yao et al., 2007a;Zheng et al., 2008;Zhuang et al., 1999;Huang et al., 2014;Bian et al.,
41 2014) and measurements by electrostatic classifier instruments (Cheung et al., 2015;Yao et al., 2007b) and were hence
42 either limited in size resolution (offline filter samples) or chemical resolution (total particle count by classification).
43 Most measurements in Hong Kong were conducted in suburban environments. Inorganic ammonium and sulfate were
44 mainly found in fine mode particles in condensation and droplet mode size ranges, while nitrate had strong coarse
45 mode contributions (Zhuang et al., 1999). Seasonal differences were evident in solvent-extractable organics and trace
46 metals which were mainly found in PM_{0.5} particles in the wet season and winter whereas in fall a shift to larger particles
47 (0.5–2.5 μm fraction) in fall indicated a possibly stronger influence of aged particle components in the transition period
48 of the Asian monsoon (Zheng et al., 2008). Size distributions acquired by a fast mobility particle sizer at the suburban
49 HKUST supersite were investigated more recently to study the formation and accumulation of ultrafine particles under
50 different air flow regimes. Particle number concentration enhancements during the day were attributed to secondary
51 formation, while evening and nighttime peaks were thought to be related to transport of aged aerosols from upwind
52 locations. Nucleation mode particle peaks were often observed in fall and related to regional pollution influence
53 (Cheung et al., 2015). New particle formation events at the same site occurred as single and two-stage growth
54 processes with organics and sulfuric acid contributing mainly to first stage growth in the daytime while nighttime
55 second stage growth was attributed to ammonium nitrate and organics. Particle size growth into the diameter range of
56 cloud condensation nuclei (CCN) was typically only achieved with the second growth stage (Man et al., 2015).

57 Investigations into particle size distributions in urban areas of Hong Kong are even scarcer. Yao et al. (Yao et al.,
58 2007b) studied the properties and behavior of particles in vehicle plumes and reported a competing process between
59 ambient background particles and fresh soot particles in the condensation of gaseous precursors and a dependency on
60 temperature with bimodal volume size distributions observed at lower ambient temperatures and unimodal
61 distributions in the lower accumulation size range at higher ambient temperatures.

62 The Aerodyne aerosol mass spectrometer (Canagaratna et al., 2007) is widely used to determine the chemical
63 composition of major organic and inorganic components of non-refractory submicron particulate matter (NR-PM₁).
64 In contrast to most traditional aerosol sizing instruments, the AMS is capable of resolving main chemical constituents

102 close to primary emission sources and a suburban location (coastal, HKUST supersite) which is largely a downwind
103 receptor of varying amounts of local urban, regional and long-range transported pollutants (Li et al., 2015;Huang et
104 al., 2014).

105 **2. Methodology**

106 **2.1. Field campaigns**

107 Sampling of ambient submicron non-refractory particulate matter (NR-PM₁) was carried out using an Aerodyne HR-
108 ToF-AMS at the HKUST air quality supersite covering four seasons between May 2011 and February 2012 (spring:
109 2011-05, summer: 2011-09, fall: 2011-11&12, winter: 2012-02). The HKUST supersite is located on the campus of
110 the Hong Kong University of Science and Technology (22°20'N, 114°16'E), on the east coast of Hong Kong in a
111 suburban area with few primary emission sources in the immediate vicinity. Sampled air was drawn from the rooftop
112 of a pump house building at an approximate height of 25m above ground level. For detailed descriptions of the
113 experimental setup, operating conditions, data treatment, and overall species composition we refer the reader to
114 previous publications (Lee et al., 2013;Li et al., 2015;Li et al., 2013). A further sampling campaign took place between
115 spring 2013 (2013-03 to 2013-05) and summer 2013 (2013-05 to 2013-07) at an inner-city urban location in the
116 densely populated and built-up Kowloon peninsula. Measurements were conducted next to the roadside air quality
117 monitoring station (AQMS) operated by the Environmental Protection Department (EPD) of the HKSAR Government
118 in the Mong Kok (MK) district on a pedestrian crossing at a major road junction. Sampled air was drawn from a height
119 of 3m above ground level. A comprehensive analysis of trends in species concentration and composition identified in
120 this urban campaign has been presented previously (Lee et al., 2015). In both campaigns, particles were sampled
121 through a PM_{2.5} cyclone at a flow rate of 16.67 L/min into a sampling port from which 0.08 L/min was drawn by the
122 AMS and the remainder drawn by co-sampling instruments and an auxiliary pump. Sample air for the AMS passed
123 through a 1m long diffusion dryer (BMI, San Francisco CA, USA) filled with silica gel to remove bulk gas- and
124 particle-phase water. Additional data from various collocated instruments including meteorological data (wind,
125 temperature, relative humidity, solar irradiation), volatile organic compounds (VOCs) and standard trace gases such
126 as NO_x, SO₂, and O₃ were available.

127 AMS data were treated according to general AMS data treatment principles (DeCarlo et al., 2006;Jimenez et al., 2003)
128 with standard software packages (SQUIRREL, PIKA). Analysis of the unit-mass resolution mass spectra yielded non-
129 refractory submicron particle species concentrations of major inorganic constituents (SO₄, NO₃, NH₄, Chl) and total
130 organics at a base time resolution of 10 min. Positive Matrix Factorization (PMF) was used to deconvolute high-
131 resolution organic mass spectra acquired at 10 min time resolution following recommended PMF guidelines for AMS
132 data (Zhang et al., 2011) with the AMS PMF analysis toolkit (Ulbrich et al., 2009). At the urban Mong Kok site, six
133 organic aerosol (OA) factors were identified encompassing three secondary organic aerosol (SOA) and three primary
134 organic aerosol (POA) factors of which one was attributed to traffic emissions and two to cooking activities (Lee et
135 al., 2015). Similarly, four factors were obtained from analysis of the urban HKUST site dataset with two SOA factors

171 be found in the Supplement (Text S1, Text S2). The sequence of main data treatment and analysis steps is shown in
172 Fig. 1.

173 The transmission efficiency of the AMS aerodynamic lens is known to fall off below ~100nm and beyond ~550 nm
174 of vacuum-aerodynamic diameter (Liu et al., 2007;Takegawa et al., 2009;Zhang et al., 2004;Bahreini et al.,
175 2008;Williams et al., 2013;Knote et al., 2011) and may bias measured particle mass and mode diameters, particularly
176 in the accumulation mode towards lower values if significant particle mass fractions fall in the size region of $D_{va} >$
177 550 nm. In the Aitken mode range, the effect of limited lens transmission is expected to be less substantial as particle
178 volume (and hence particle mass) of Aitken mode particles are much smaller. We discuss the effects of lens
179 transmission briefly in section 3.4. Delayed vaporization of particle components, e.g. under high mass loadings, can
180 lead to small shifts towards larger mode diameters in AMS size distributions (Docherty et al., 2015) and enhanced
181 tails in the size distributions (Cross et al., 2009), which may lead to larger fit residuals at the trailing edges. Generally,
182 the discussion of size distributions in this work should be viewed in the context of the instrumental capabilities and
183 previously mentioned limitations of aerosol mass spectrometry. Therefore, the resolved Aitken and accumulation
184 modes in this work reflect the apparent Aitken and accumulation modes within AMS measurable particle mass size
185 distributions.

379 **Figure 4.** Diurnal variations of mode diameter (MMD), integrated mode mass concentration and width of the Aitken mode (*lighter*
380 *color*) and accumulation mode (*darker color*) from bimodal peak fits of the bin-median reconstructed size distributions at the
381 suburban HKUST site and V-mode AMS species concentrations (line with shaded background) for organics, nitrate and sulfate
382 (left to right) in (a) spring 2011, (b) summer 2011, (c) fall 2011 and (d) winter 2012; The right-most panel depicts the median
383 diurnal variations of relevant gas-phase pollutants (O₃, CO, NO_x, SO₂) measured at the same site.

384 3.1.2. Suburban coastal NR-PM₁

385 The suburban HKUST site as a downwind receptor of urban and regional pollution was generally dominated by sulfate
386 and oxygenated secondary organic aerosol (SOA) components and much lower fractions of primary organic
387 constituents, which combined typically made up less than a quarter of total organics (Li et al., 2015). Trends in the
388 species segregated particle size distributions are discussed analogously to section 3.1.1., with Fig. 4 illustrating the
389 diurnal trends of the fitting parameters (MMD, integrated mode mass, geometric standard deviation) for organics,
390 sulfate, and nitrate at the suburban HKUST site.

391
392 Organics

393 There were significant seasonal differences with larger fractions (Fig. 3b) and concentrations (Fig. 5c) of Aitken mode
394 mass in total organic submicron particle mass in spring and summer compared to fall and winter, indicating greater
395 influence of closer-ranged formation sources in the warmer season. Springtime integrated Aitken mode mass
396 concentrations ($\sim 0.8 \mu\text{g}/\text{m}^3$) were twice as high as those in winter ($\sim 0.4 \mu\text{g}/\text{m}^3$). In the accumulation mode, highest
397 particle mass loadings were observed in fall ($5 \mu\text{g}/\text{m}^3$) and lowest loadings in spring ($3 \mu\text{g}/\text{m}^3$) following the frequency
398 pattern of continental air mass influence (Fig. S12 in the Supplement) in each season indicating continental transport
399 of particulate mass or gas-phase precursors. Lowest mass concentrations in the Aitken mode typically occurred in the
400 night hours (00:00 – 05:00) in a range of $0.3 - 0.5 \mu\text{g}/\text{m}^3$ in spring, summer, and winter, while in fall mass loadings
401 of $0.7 - 0.8 \mu\text{g}/\text{m}^3$ were reached. Diurnal changes were least pronounced in winter with largely constant integrated
402 Aitken mode particle concentrations. In the remaining seasons, varying degrees of daytime changes were apparent
403 with a general increase around 06:00, likely owing to citybound commuter traffic from surrounding roads to the west
404 of the sampling site at 1-2km of lateral distance. This also led to a modest increase in particle polydispersity with a
405 discernible widening of the Aitken mode size distributions (*black solid line, lowest panels in Fig. 3*). Daily maxima
406 in spring, summer and fall were reached in the early evening ($\sim 21:00$) with marked differences in absolute mass
407 concentrations depending on the respective season, from a summer time low of $0.8 \mu\text{g}/\text{m}^3$ to a fall season high of 1.4
408 $\mu\text{g}/\text{m}^3$. Mass median diameters in the Aitken mode were smaller in the night hours and displayed subtle increments
409 during the day in the range of 10-20 nm reaching their maximum typically in the late afternoon, except for the fall
410 season when mass median diameters displayed very little variation with time of day.

411 Total particle mass in the accumulation mode in spring and summer reached minima during the night hours ($2 \mu\text{g}/\text{m}^3$)
412 and maxima ($3 \mu\text{g}/\text{m}^3$) around noon, remaining stable in the daylight hours thereafter. MMDs increased notably from
413 440nm at night to 510nm during the day in spring, while in summer a morning rise by $\sim 30\text{nm}$ from 530nm to 560nm
414 was obvious between 06:00 and 10:00 and coincided with the morning rush hour and the associated early morning

415 peak of NO_x and an otherwise stable mode diameter of 530nm for the rest of the day. In fall, the increase in
416 accumulation mode organic mass occurred much earlier, starting in the dark hours at 04:00, with a corresponding
417 trend also evident for nitrate but absent for sulfate, indicating a common source of these organic and nitrate enriched
418 particles. Nighttime MMDs for organics were generally larger (540nm) and decreased to a minimum of 510nm in the
419 early afternoon accompanied by a slight widening of the distribution. In winter, mass concentrations decreased
420 appreciably in the early morning hours and started to increase only beyond 10:00. In the colder seasons (fall, winter),
421 a similar concentration pattern was also observed for gas-phase SO₂ which is considered as a largely regional pollutant
422 with few distinct local sources (Yuan et al., 2013), indicating that changes in boundary layer and mixing with regional
423 background were likely the more dominant processes in winter.

424

425 Sulfate

426 Aitken mode sulfate mass concentrations peaked in the afternoon from spring throughout fall with maximum
427 concentrations reached progressively later in the afternoon (14:00 in spring; 16:00 in fall). Nominal concentrations
428 were highest in spring and summer (0.5-0.6 µg/m³), slightly lower in fall (0.4 µg/m³) and reached the lowest levels in
429 winter (0.1 µg/m³). In addition to the afternoon peak, a conspicuous early morning peak of similar magnitude was
430 evident in spring between 02:00 and 06:00. A greater proportion of southerly winds was evident in said time period
431 compared to the overall seasonal wind frequency distribution (Fig. S13a in the Supplement) and may indicate transport
432 of sulfate from marine sources in the southern parts of Hong Kong. Diurnal variations in MMDs and GSDs were
433 generally small and without obvious regular trends. Nominal mass median diameters were significantly lower in winter
434 (~170nm) compared to spring and fall (~190nm) and summer (~210nm).

435 Trends in accumulation mode particle mass were more pronounced. In spring, a shallow concentration valley during
436 the late evening and night hours (20:00 to 03:00) with minimum concentrations of 5 µg/m³ was apparent, while
437 daytime concentrations stayed largely invariant at 6 µg/m³. The MMDs followed a similar variation with a minimum
438 mode diameter around 550nm in the early hours of the day and slightly larger daytime MMDs around 570nm. Nominal
439 concentrations were larger in summer with a nighttime valley concentration of 7 µg/m³ and a well-pronounced broad
440 day peak with a maximum of 9.5 µg/m³ in the early afternoon (14:00-15:00). A prior additional morning peak occurred
441 between 04:00 and 10:00 with particle mass concentrations reaching 8.5 µg/m³ related to a consistent north-easterly
442 morning wind pattern (Fig. S13b in the Supplement) and likely associated with transport from north-easterly coastal
443 regions or nighttime fisheries related maritime traffic. The diurnal trend in mass median diameter was similar to that
444 in spring with a night minimum of 570nm and day maximum of 590nm.

445 In fall, accumulation mode characteristics showed no significant diurnal variability, with a largely stable integrated
446 particle mass of 6 µg/m³ and only subtle MMD changes (585nm at night; 575nm during the day). In winter, two
447 concentration dips with reductions by ~0.5 µg/m³ between 06:00 and 10:00 and between 18:00 and 22:00 were evident,
448 while MMDs increased during the day between 10:00 and 15:00 from 520nm, peaking at a size of 540nm.

449

450 Nitrate

451 Nitrate particle mass in the Aitken mode was generally small from spring throughout fall amounting to 0.01 - 0.06

452 $\mu\text{g}/\text{m}^3$. Winter time concentrations were larger in a range of 0.06 - 0.08 $\mu\text{g}/\text{m}^3$ during the day and 0.10 - 0.12 $\mu\text{g}/\text{m}^3$ in
453 the late evening hours. The latter evening peak centered around 21:00 was evident in most seasons (except spring)
454 and accounted for 12-23% (0.1-0.25 $\mu\text{g}/\text{m}^3$) of total daily Aitken mode nitrate mass burden. Similar to the urban
455 roadside location, these nighttime nitrate peaks coincided with the peak period of organic cooking aerosol
456 concentrations (Fig. S14 in the Supplement), which were however significantly smaller at the suburban measurement
457 site and mainly attributed to the operation of an on-campus student canteen (Li et al., 2015). Trends in mass median
458 diameters varied between seasons with no discernible trend in winter, a subtle decreasing trend with time of day in
459 spring and broad daytime diameter increases in summer and fall. Solar irradiation in these two seasons was
460 comparatively high (Fig. S10b-c in the Supplement) indicating that photochemical nitrate production in the Aitken
461 mode may have led to this observed growth in particle size.

462 Integrated particle mass concentrations in the accumulation mode only exhibited subtle variations from spring
463 throughout fall, with essentially constant diurnal concentrations in spring, a subtle daytime peak in summer which
464 accounted for $\sim 15\%$ of total daily accumulation mode nitrate (corresponding to 0.7 $\mu\text{g}/\text{m}^3$) and a conspicuous morning
465 peak between 04:00 and 10:00 in fall accounting for $\sim 5\%$ of total daily accumulation mode nitrate (corresponding to
466 0.5 $\mu\text{g}/\text{m}^3$). Clearer seasonal differences were evident in the trends of MMDs. In spring, MMDs decreased appreciably
467 over the late evening hours (21:00-0:00) with a concurrent widening of the size distribution (increase in GSD). In
468 summer, accumulation mode diameters decreased during the day by $\sim 40\text{nm}$ with a similar trend in accumulation mode
469 organics. Winter time MMDs exhibited a more complex pattern with larger mode diameters in the early hours (04:00
470 - 10:00) and during the noon-time, and a late-afternoon dip leading to larger spread of intra-day mode diameters
471 ranging from 510nm to 570nm.

472 In comparison to the urban roadside measurements, diurnal particle size characteristics and mass concentrations in the
473 Aitken and accumulation mode were much more variable for all investigated species at the suburban HKUST site,
474 indicating that longer time scale processes and irregular events (transport patterns, local meteorology) were probably
475 more important in governing particle size distribution characteristics than diurnal processes.

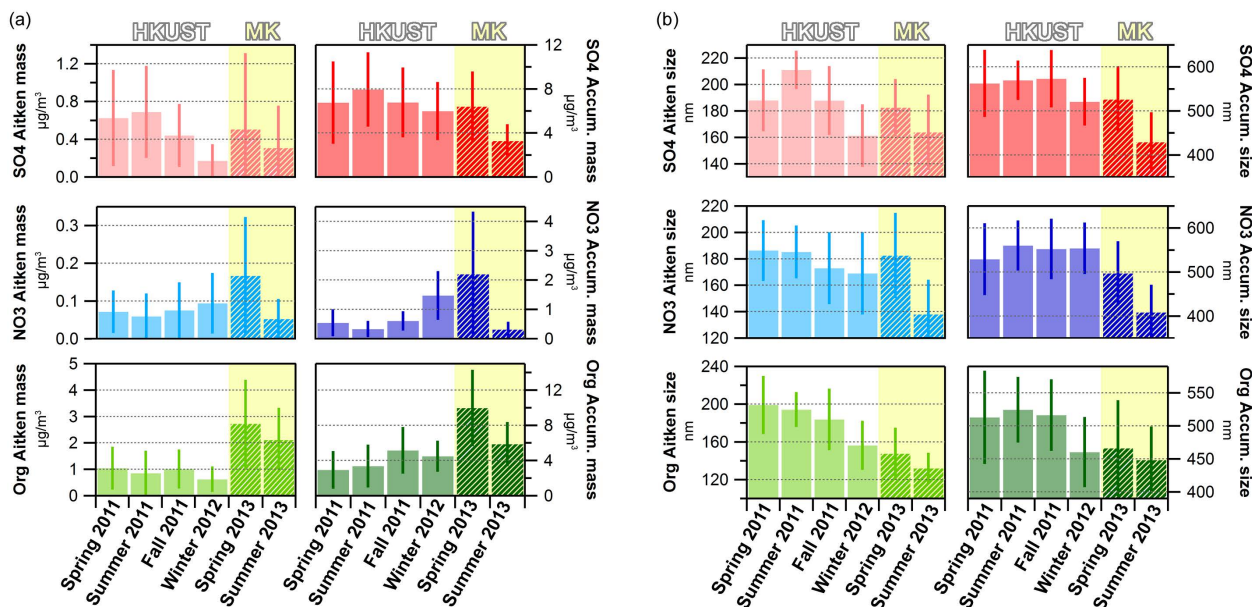
476 **3.2. Day-to-day size distributions and seasonal averages**

477 To evaluate the evolution of particle size distributions within seasons, average species-specific size distributions were
478 generated by averaging raw distributions over 24h periods (between 0:00 and 23:59). There was clear long-term
479 variability in both resolved Aitken and accumulation mode MMDs and integrated submode particle mass
480 concentrations for all species (Fig. S15-16 in the Supplement) and overall seasonal differences which have been briefly
481 addressed in the discussion of the diurnal size distribution variations between seasons. Figure 5 depicts the overall
482 average values for all daily fitted MMDs and integrated particle mass concentrations in both the Aitken and
483 accumulation mode at the suburban HKUST and urban MK sites.

484 3.2.1. Seasonal trends

485 For the MK roadside station, particle mode diameters were generally larger in spring than in summer for all three
486 investigated species, but with clear differences in the magnitude of changes among individual species. In the Aitken
487 mode, organics and sulfate displayed a moderate decrease in mode diameter from spring to summer by 7-8% each,
488 while nitrate saw a more significant decrease by 25% from spring to summer. In contrast, accumulation mode MMDs
489 for organics exhibited only a subtle decrease by 5% and more substantial decreases for sulfate and nitrate by 20-22%
490 each. Total Aitken mode particle mass decreases varied strongly: -15% for organics, -36% for sulfate and -67% for
491 nitrate. In the accumulation mode, organics and sulfate exhibited similar relative decreases by 40-46%, while nitrate
492 particle mass reduced drastically by 85%.

493 At the suburban HKUST site, Aitken mode MMDs of nitrate and organics decreased with the progression of seasons
494 from spring to winter with highest mode diameters observed in spring and summer and appreciable decreases in winter
495 by -9% for nitrate and -25% for organics compared to the warmer seasons. Sulfate displayed a similar winter time
496 decrease in MMD (-15%) and an increase of similar magnitude in the summer season (+13%) compared to spring and
497 fall. Variations in sulfate and organic accumulation mode diameters were minor between spring and fall, while
498 wintertime MMDs were 7-12% lower. Nitrate exhibited an overall higher variability in mass median diameters in the
499 accumulation mode in spring (larger standard deviation) and with on average 10% lower MMDs compared to other
500 seasons. In line with the reduction in Aitken mode MMDs in winter, the integrated Aitken mode particle mass
501 decreased as well, by -16% for organics and almost -75% for sulfate, whereas nitrate contributions remained largely
502 stable throughout the seasons. Organic accumulation mode particle mass was significantly higher in the fall and winter
503 season by factors of 1.6 – 2. Diurnal variations in the degree of oxygenation were least pronounced in these seasons
504 (Li et al., 2015) suggesting that influence of transport in autumn and winter likely dominated over local formation,
505 thus exerting greater effects on particle mass in the larger size mode. Particulate nitrate concentrations were generally
506 low in the accumulation mode from spring through fall, but increased sharply in winter by factors of 3 – 4. Sulfate
507 accumulation mode mass concentrations remained more stable but saw significant summer time enhancements by
508 ~30% likely due to photochemical activity which also led to high concentrations of Ox and a higher degree of
509 oxygenation of organic aerosol among the four seasons (Li et al., 2015).



510
 511 **Figure 5.** Average and standard deviation of daily fit values of Aitken and accumulation mode particle mass and mass median
 512 diameters at the suburban HKUST site (*solid bars*) and urban MK site (*hashed bars*). The integrated particle mass is depicted in
 513 (a) for the Aitken mode (*left panels*) as well as the accumulation mode (*right panels*) for sulfate, nitrate, and organics respectively.
 514 The mass median diameter is depicted in (b) for the Aitken mode (*left panels*) as well as the accumulation mode (*right panels*) for
 515 sulfate, nitrate and organics respectively.

516
 517 Large particles contribute more to particle volume and hence particle mass. Correspondingly, the total submicron
 518 concentration of a given species is typically governed by changes in the accumulation mode particle mass and
 519 accordingly observed correlation values between integrated accumulation mode particle mass and individual NR-PM₁
 520 species mass concentrations were generally high ($R_{pr} > 0.90$) at both measurement sites (Fig. S17 in the Supplement).
 521 This applied to both measurement sites regardless of the season. Aitken mode trends were less akin. At the urban
 522 roadside station, neither sulfate nor nitrate particle mass in the Aitken mode notably correlated with the respective
 523 total submicron species mass concentration in spring (all $R_{pr} \leq 0.20$), whereas in summer correlations were more
 524 significant with $R_{pr} = 0.51$ for sulfate and $R_{pr} = 0.80$ for nitrate. This signifies that periods of greater species mass
 525 concentrations were more likely to be caused by increases in both Aitken and accumulation mode particle mass
 526 indicating that particle formation and growth affecting smaller particles was more likely to occur in the warmer season.
 527 For organics, Aitken mode particle mass and submicron species mass correlated only weakly ($R_{pr} = 0.26$ in spring and
 528 $R_{pr} = 0.38$ in summer), i.e. each organic particle submode was governed by largely different dominant sources or
 529 formation processes in both seasons at the roadside.

530 At the suburban background site, Aitken mode particle mass for sulfate showed little correlation with total submicron
 531 sulfate concentration ($R_{pr} \leq 0.10$) apart from the spring season ($R_{pr} = 0.36$) where more frequent wet and foggy
 532 conditions may have facilitated sulfate formation in both size modes. For organics and nitrate significantly larger
 533 correlation coefficients of submode particle mass to total species concentration ($0.5 \leq R_{pr} \leq 0.7$) were observed in
 534 most seasons (spring, summer, winter) indicating significant influence of local or regional formation processes on

535 organic and nitrate Aitken mode particulate mass at the suburban receptor location. In the fall season, much weaker
536 correlations ($0.2 \leq R_{pr} \leq 0.4$) were likely caused by the dominance of continental air mass influence (Fig. S12c in the
537 Supplement) and greater influence of aged accumulation mode particles on total submicron nitrate mass
538 concentrations.

539 3.2.2. Inferred changes in mixing state

540 Shifts in mixing state of ambient particles can be inferred from the inter-species analysis of mass median diameters.
541 Close nominal agreement (i.e. diameter ratios close to 1) infer that different species were distributed similarly across
542 the particle size range which thus most likely represents a largely internally mixed particle population, while the spread
543 of data (correlation coefficient) indicates the temporal homogeneity or divergence of resolved mode diameters. A
544 hypothetically perfectly internally mixed particle population over the whole sampling period would, therefore, yield
545 MMD ratios and Pearson's R values of 1 between species, while larger or smaller values are indicative of a greater
546 frequency of heterogeneous (i.e. more externally mixed) particle populations (Fig. 6).

547 At the urban Mong Kok site, changes in accumulation mode mass median diameters for nitrate and sulfate followed
548 similar trends ($R_{pr} = 0.88-0.89$) and with diameter ratios close to 1 (0.94–0.95) Similarly, fitted accumulation mode
549 diameters of organic constituents predominantly followed that of sulfate in spring nominally (diameter ratio 0.88) and
550 temporally ($R_{pr} = 0.80$). The nominal agreement of organic and sulfate accumulation mode diameters persisted
551 (diameter ratio 1.03) overall in summer, however, there was significantly more temporal divergence ($R_{pr} = 0.65$)
552 indicating a greater frequency of time periods with external mixing of particle populations comprising different
553 fractions of organic constituents.

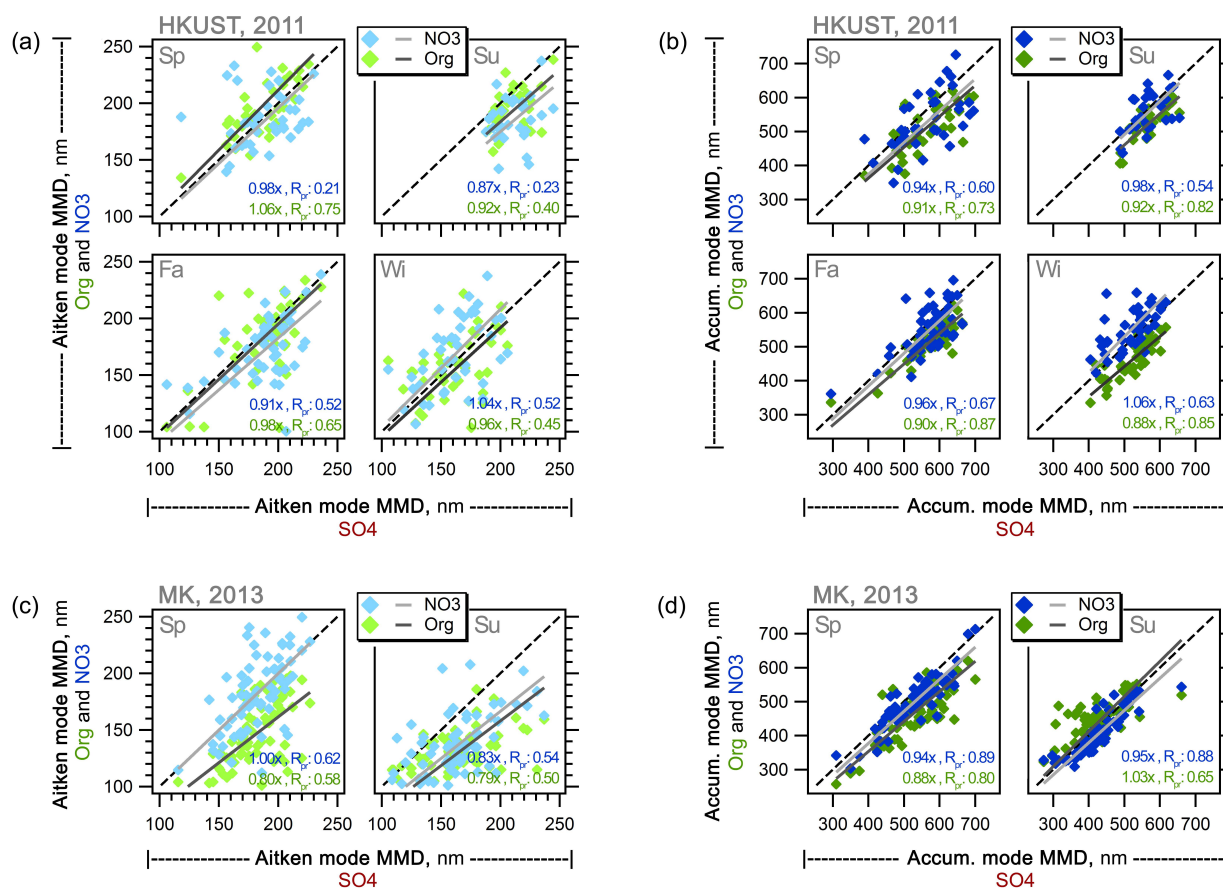
554 External mixing is more prevalent for freshly formed smaller particles which have typically undergone less
555 condensational growth, coagulation or aqueous-phase reactions. Indeed, the correlation coefficients of both nitrate
556 and organic Aitken mode MMDs with respect to sulfate were notably lower (0.50 and 0.62) indicating frequent periods
557 of particle populations with different species prevailing in different size regions within the Aitken mode.

558 Sulfate and nitrate were still more likely to occur internally mixed in the Aitken mode in spring with similar diameters
559 (nitrate to sulfate MMD ratio = 1.00), while organic Aitken mode MMDs were consistently lower, indicating greater
560 fractions of organic dominated particles towards the lower end and more inorganic dominated particles towards the
561 upper end of the fitted Aitken mode.

562 In summer, both nitrate and organic MMDs tended to be lower than those of sulfate (diameter ratios of 0.79 – 0.83)
563 but similar to each other, thus implying a shift to externally mixed populations of more nitrate and organic enhanced
564 and internally mixed smaller Aitken mode particles and sulfate dominated larger Aitken mode particles.

565 At the suburban HKUST site, accumulation mode MMDs of both nitrate and organics were generally quite similar to
566 those of sulfate with diameter ratios of 0.88 – 1.06. Compared to the urban site, correlation coefficients of nitrate and
567 sulfate were consistently lower (0.54 – 0.67) indicating a much greater frequency of time periods where sulfate and
568 nitrate dominated particles in the accumulation exhibited significantly different particle size distributions.

569 In winter, organic MMDs were consistently lower than those of sulfate and nitrate indicating a greater proportion of
 570 externally mixed particle populations with organics enriched particles in the lower accumulation size range and
 571 inorganic dominated particles in the larger accumulation size range. The least variability in particle size was observed
 572 in the summer season where MMDs in both Aitken and accumulation mode displayed variations in relatively narrow
 573 ranges between 200-250nm and 500-700nm, whereas in the remaining seasons time periods with particle populations
 574 of lower MMD were more frequent, extending to MMDs as low as 100nm in the Aitken mode and 300nm in the
 575 accumulation mode.
 576 In the Aitken mode, mass median diameters overall were quite similar across species, with diameter ratios of organic
 577 and nitrate distributions to those of sulfate in the range of 0.87 – 1.06, indicating that they generally covered a similar
 578 size range. The temporal agreement was highly variable with correlation coefficients (R_{pr}) spanning from 0.21 to 0.75
 579 indicating that Aitken mode particle populations at the suburban site were generally more diverse and likely influenced
 580 by a greater range of particle formation and growth mechanisms compared to the urban Mong Kok site.



581
 582 **Figure 6.** Scatter plots of fitted mass median diameters of organics and nitrate vs. sulfate for (a) the Aitken mode and (b) the
 583 accumulation mode at the HKUST suburban site, and (c) the Aitken mode and (d) the accumulation mode at the urban Mong Kok
 584 site

585 3.3. Comparison to previous studies

586 Particle size distribution studies in Hong Kong are generally scarce and have focused on either size segregated filter
587 samples (MOUDI) for general ambient measurements or electrostatic classification in particle formation and particle
588 growth studies (Guo et al., 2012; Cheung et al., 2015). The latter studies focus on specific and narrow time periods
589 and lack general discussions on ambient particle size distributions.

590 Two ambient studies were undertaken at the suburban coastal HKUST site using size-segregated samples from a ten-
591 stage MOUDI sampler and offline chromatographic analysis. Inorganic constituents (NH_4 , NO_3 , SO_4) in fine particles
592 (i.e. $D_p < 1.8 \mu\text{m}$) were shown to follow bimodal distributions with mode diameters in the range of 0.14–0.21 μm and
593 0.46–0.58 μm in samples collected in the winter season, while the main mode was observed in the coarse region (4–6
594 μm) for all three species (Zhuang et al., 1999). A subsequent year-long observational study also reported bimodal fine
595 particle distributions with mode diameters of 0.1–0.3 μm and 0.7–0.9 μm and 1–2 additional modes in the coarse region
596 (Bian et al., 2014), however, the main mode in the size distributions of sulfate, ammonium, potassium and oxalate
597 was observed in the droplet mode (0.7 – 0.9 μm) in this study. Vehicle exhaust plumes sampled on-road from a Mobile
598 Real-time Air Monitoring Platform (MAP) across Hong Kong's road network exhibited three distinct particle volume
599 size distributions: a unimodal distribution with an accumulation mode at 0.2 μm and two bimodal distributions with a
600 minor mode at 0.2 μm and the dominant mode at 0.5 or 0.7 μm (Yao et al., 2007b).

601 The bimodality in the fine particle range across these studies is consistent with the AMS-based results in this work.
602 Nominally, the accumulation mode diameters from filter based studies and the chase studies are larger than those from
603 AMS measurements where maximum mode diameters occurred at $D_{va} \sim 700\text{nm}$, corresponding to $D_a \sim 470$ (assuming
604 $D_{va} \sim D_a * \text{density}$; particle density $\sim 1.5 \text{ g/cm}^3$). Direct comparability is however limited due to fundamental
605 differences in sizing techniques (MOUDI: atmospheric pressure; AMS: near-vacuum), sampling times (MOUDI: 24h
606 samples, scattered time line; AMS: minute raw resolution averaged to hourly or daily, continuous time line),
607 measurement uncertainties (MOUDI: sampling artifacts such as vapor adsorption and desorption; AMS: inlet lens
608 transmission) and aerosol pretreatment (none for MOUDI with potential impacts on particle size in high humidity
609 (>80%) conditions (Fang et al., 1991); AMS: removal of water prior to introduction to instrument).

610 3.4. Influence of AMS lens transmission

611 The quantitative measurement of particle components in the AMS is dependent on three major factors which may lead
612 to particle loss prior to detection (Huffman et al., 2005). Irregularly shaped particles deviating from the flight path in
613 the vacuum chamber may miss the vaporizer. Particles bouncing off the vaporizer surface will not be vaporized and
614 hence may not be detected. Lastly, the aerodynamic lens which is part of the instrument's inlet system does not
615 transmit particles uniformly across all particle diameters. Small particles are lost due to insufficient focusing or
616 diffusion and large particle impact the lens apertures (Liu et al., 2007; Williams et al., 2013). Being a function of
617 particle size, the latter factor affects both total AMS quantifiable particle mass (NR-PM_{10}) and measured mass size
618 distributions. Transmission curves determined for the standard lens, which is fitted in most AMS instruments, can

619 vary but typically show efficient (i.e. close to 100%) transmission in the range of 100-550nm (Knote et al., 2011)
620 falling of significantly at either edge.

621 We examined the potential impact of lens transmission on the AMS mass size distributions on a number of 24h size
622 distributions from the fall season HKUST dataset covering both efficient and reduced lens transmission size regimes
623 (accumulation mode diameters between 400 and 600nm). Panels a-c in Fig. S18 in the Supplement depict original and
624 lens-transmission corrected 24h mass size distributions for organic, nitrate and sulfate, assuming the transmission
625 function (subpanel d) reported by Liu et al., 2007. Impacts were generally larger in the accumulation mode range with
626 evident shifts to larger mode diameters and larger mode mass concentrations observed in all size distributions. In the
627 small diameter range, enhanced shoulders can occur which may however be artifacts due to the larger uncertainties
628 (low signal to noise ratio at small particle mass), i.e. greater noisiness at the leading end of AMS size distributions.
629 For a quantitative comparison, bimodal fitting parameters from the corrected distributions were plotted against those
630 from, the original distributions in Fig. S18 in the Supplement (subpanels e-g refer to the Aitken mode and subpanels
631 h-j to the accumulation mode). Leading edge shoulders in the corrected size distributions were not considered in the
632 fitting. Changes in the Aitken mode mass median diameters were minor (on average ~3%), while the integrated mode
633 particle mass increased moderately (~48%). In the accumulation mode, mass median diameters increased by ~28%
634 and integrated mode particle mass doubled. The distribution widths (geometric standard deviations) exhibited little
635 change in both modes (increases of 2-3%).

636 Fitting results will therefore vary depending on whether AMS size distribution and concentration data are corrected
637 for lens transmission. While explicit lens transmission corrections can improve the accuracy of quantification of AMS
638 species concentration and size distribution measurements, few ambient studies explicitly use lens transmission
639 corrections based on individual experimental determinations or literature values e.g. (Quinn et al., 2006; Cross et al.,
640 2007). Lens transmission curves can vary between instruments (Fast et al., 2009) and are inherently difficult to
641 determine accurately experimentally. As discussed previously, scaling of size distributions by lens transmission curves
642 may introduce artifacts in noisier size distributions (e.g. low end of the Aitken mode, low concentration periods, short
643 term size distribution averages). Trailing edges from slow vaporization (e.g. under high particle mass loadings) may
644 be exacerbated and inflate mass concentrations at the upper size cut range of the AMS. The majority of ambient
645 studies employs a combined correction factor (collection efficiency, CE) considered to be the joint product of the
646 previously mentioned transmission efficiencies related to particle bounce, beam broadening and lens transmission
647 (Middlebrook et al., 2012) derived from aerosol composition and by comparison to collocated speciation or particle
648 sizing instruments. As the AMS lens transmission curve could not be determined in this study and to avoid additional
649 uncertainties from the application of non-instrument specific lens transmission values, we followed the CE correction
650 method in the analysis of the size distribution data in this study. The reported values of resolved mode diameters and
651 integrated mode should therefore be regarded as lower bound estimates in the context of the instrumental limitations
652 affecting ambient AMS measurements.

653 4. Conclusion

654 A detailed analysis of AMS mass-based particle size distributions of sulfate, nitrate, and organics in submicron
655 particulate matter measured at two contrasting locations in Hong Kong during two field campaigns has been
656 undertaken. Deconvolution of size distributions into Aitken and accumulation submodes was accomplished by log-
657 normal peak fitting and trends in particle size (mass median diameters), dispersity (geometric standard deviation) and
658 overall particle mass (integrated mode area) were discussed on a diurnal time scale and on a daily basis to evaluate
659 longer-term changes in size distribution characteristics. At the urban roadside location, clear diurnal influences of
660 primary particle and gas-phase species were evident affecting both inorganic and organic component size distributions.
661 Traffic and cooking contributed an estimated $0.3 - 0.9 \mu\text{g}/\text{m}^3$ and $0.5 - 1.8 \mu\text{g}/\text{m}^3$ of organic component particle mass
662 in the Aitken mode, and $1.6 - 1.8 \mu\text{g}/\text{m}^3$ and $1.0 - 2.7 \mu\text{g}/\text{m}^3$ respectively in the accumulation mode with concentrations
663 level varying with seasons. Notable changes in Aitken mode mass median diameters of organics were limited to the
664 morning rush hour. Daytime particle concentration maxima of sulfate and nitrate in summer indicated substantial
665 influence of photochemical processes, which also led to increments in mass median diameters in the accumulation
666 mode thus inferring associated particle growth. Nocturnal nitrate formation was apparent in the accumulation mode
667 in spring concurring with the nighttime peak of ozone at the roadside, while in the Aitken mode nitrate particle
668 concentrations were significantly elevated during the dinner hours. Organics-related size distributions were mostly
669 governed by intra-day changes at the urban site with very similar trends across different size distribution sets (i.e.
670 concentration regimes), while disparities in diurnal variations among different size distribution sets were evident for
671 nitrate and sulfate, particularly affecting the average sets, indicating stronger influence of irregular external factors
672 which were not associated with diurnal time scale processes.

673 Suburban particle size distributions exhibited variable diurnal characteristics, suggesting that irregular processes such
674 as transport and seasonal meteorological conditions were the more dominant processes influencing particle size
675 characteristics. Aitken mode particle mass of organics was significantly larger in spring and summer indicating greater
676 influence of more local formation sources in the warm season. In the accumulation mode, organic particle mass
677 concentrations were highest in fall and lowest in spring, following the frequency pattern of continental air mass
678 influence. For sulfate, Aitken mode mass concentrations mass concentrations peaked in the afternoon from spring
679 throughout fall with highest nominal concentrations in spring and summer and lowest levels in winter, while
680 accumulation mode particle mass was highest in summer and fall and lowest in winter, similar to the trend observed
681 among organic constituents.

682 Nitrate particle mass in the Aitken mode was generally small in most seasons ($0.01 - 0.06 \mu\text{g}/\text{m}^3$), except winter where
683 daytime concentrations reached $\sim 0.1 \mu\text{g}/\text{m}^3$. In both modes, changes in mass median diameters varied temporally and
684 in magnitude with seasons, indicating a stronger influence of specific meteorological conditions on the properties of
685 nitrate-containing particles at the suburban site. At the urban site, periods of greater inorganic species mass
686 concentrations were more likely to be caused by increases in both Aitken and accumulation mode particle mass in
687 summer, indicating that particle formation and growth affecting smaller particles was more likely to occur in the
688 warmer season. At the suburban receptor location, significant correlation of submode particle mass to total species
689 concentration ($0.5 \leq R_{pr} \leq 0.7$) was observed for organics and nitrate in most seasons (spring, summer, winter)

690 suggesting notable influence of local or regional formation processes on organic and nitrate Aitken mode particulate
691 mass. Variations in particle mixing state were examined by evaluation of inter-species mass median diameter trends
692 at both measurement sites. In the accumulation mode at the urban site, internal mixing appeared to be prevalent in
693 spring, while greater frequency of time periods with external mixing of particle populations comprising different
694 fractions of organic constituents was observed in summer. External mixing was predominant in the Aitken mode at
695 the urban location in both seasons. At the suburban site, sulfate and nitrate in the accumulation mode more frequently
696 exhibited differing particle size distributions in all seasons signifying a greater extent of external mixing. In winter,
697 external mixing of more organics enriched particles in the lower accumulation size range was evident.

698 **Data availability**

699 The data is available upon request. To obtain the data, please contact Chak K. Chan (chak.k.chan@cityu.edu.hk).

700 **Competing interests**

701 The authors declare that they have no conflict of interest.

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705 **References**

- 706 Abbatt, J. P. D., Broekhuizen, K., and Kumal, P. P.: Cloud condensation nucleus activity of internally mixed
707 ammonium sulfate/organic acid aerosol particles, *Atmos. Environ.*, 39, 4767-4778,
708 <https://doi.org/10.1016/j.atmosenv.2005.04.029>, 2005.
- 709 Ahlquist, N. C., and Charlson, R. J.: A New Instrument for Evaluating the Visual Quality of Air, *J. Air Pollut.*
710 *Control Assoc.*, 17, 467-469, <https://doi.org/10.1080/00022470.1967.10469006>, 1967.
- 711 Aiken, A. C., Salcedo, D., Cubison, M. J., Huffman, J. A., DeCarlo, P. F., Ulbrich, I. M., Docherty, K. S., Sueper,
712 D., Kimmel, J. R., Worsnop, D. R., Trimborn, A., Northway, M., Stone, E. A., Schauer, J. J., Volkamer, R.
713 M., Fortner, E., de Foy, B., Wang, J., Laskin, A., Shutthanandan, V., Zheng, J., Zhang, R., Gaffney, J.,
714 Marley, N. A., Paredes-Miranda, G., Arnott, W. P., Molina, L. T., Sosa, G., and Jimenez, J. L.: Mexico
715 City aerosol analysis during MILAGRO using high resolution aerosol mass spectrometry at the urban
716 supersite (T0) - Part 1: Fine particle composition and organic source apportionment, *Atmos. Chem. Phys.*,
717 9, 6633-6653, <https://doi.org/10.5194/acp-9-6633-2009>, 2009.

718 Bahreini, R., Dunlea, E. J., Matthew, B. M., Simons, C., Docherty, K. S., DeCarlo, P. F., Jimenez, J. L., Brock, C.
719 A., and Middlebrook, A. M.: Design and operation of a pressure-controlled inlet for airborne sampling with
720 an aerodynamic aerosol lens, *Aerosol Sci. Technol.*, 42, 465-471,
721 <https://doi.org/10.1080/02786820802178514>, 2008.

722 Bian, Q., Huang, X. H. H., and Yu, J. Z.: One-year observations of size distribution characteristics of major aerosol
723 constituents at a coastal receptor site in Hong Kong – Part 1: Inorganic ions and oxalate, *Atmos.*
724 *Chem. Phys.*, 14, 9013-9027, <https://doi.org/10.5194/acp-14-9013-2014>, 2014.

725 Bohren, C. F., and Huffman, D. R.: Absorption and scattering of light by small particles, in: *Absorption and*
726 *Scattering of Light by Small Particles*, Wiley-VCH Verlag GmbH, 1-11, 1983.

727 Canagaratna, M. R., Jayne, J. T., Jimenez, J. L., Allan, J. D., Alfarra, M. R., Zhang, Q., Onasch, T. B., Drewnick, F.,
728 Coe, H., Middlebrook, A., Delia, A., Williams, L. R., Trimborn, A. M., Northway, M. J., DeCarlo, P. F.,
729 Kolb, C. E., Davidovits, P., and Worsnop, D. R.: Chemical and microphysical characterization of ambient
730 aerosols with the aerodyne aerosol mass spectrometer, *Mass Spectrom. Rev.*, 26, 185-222,
731 <https://doi.org/10.1002/mas.20115>, 2007.

732 Charlson, R. J., Langner, J., Rodhe, H., Leovy, C. B., and Warren, S. G.: Perturbation of the Northern-Hemisphere
733 Radiative Balance by Backscattering from Anthropogenic Sulfate Aerosols, *Tellus A*, 43, 152-163,
734 <https://doi.org/DOI.10.1034/j.1600-0870.1991.00013.x>, 1991.

735 Cheung, K., Ling, Z. H., Wang, D. W., Wang, Y., Guo, H., Lee, B., Li, Y. J., and Chan, C. K.: Characterization and
736 source identification of sub-micron particles at the HKUST Supersite in Hong Kong, *Sci. Total Environ.*,
737 527-528, 287-296, <https://doi.org/10.1016/j.scitotenv.2015.04.087>, 2015.

738 Crippa, M., DeCarlo, P. F., Slowik, J. G., Mohr, C., Heringa, M. F., Chirico, R., Poulain, L., Freutel, F., Sciare, J.,
739 Cozic, J., Di Marco, C. F., Elsassser, M., Nicolas, J. B., Marchand, N., Abidi, E., Wiedensohler, A.,
740 Drewnick, F., Schneider, J., Borrmann, S., Nemitz, E., Zimmermann, R., Jaffrezo, J. L., Prevot, A. S. H.,
741 and Baltensperger, U.: Wintertime aerosol chemical composition and source apportionment of the organic
742 fraction in the metropolitan area of Paris, *Atmos. Chem. Phys.*, 13, 961-981, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-13-961-2013)
743 13-961-2013, 2013.

744 Cross, E. S., Slowik, J. G., Davidovits, P., Allan, J. D., Worsnop, D. R., Jayne, J. T., Lewis, D. K., Canagaratna, M.,
745 and Onasch, T. B.: Laboratory and ambient particle density determinations using light scattering in
746 conjunction with aerosol mass spectrometry, *Aerosol Sci. Technol.*, 41, 343-359,
747 <https://doi.org/10.1080/02786820701199736>, 2007.

748 Cross, E. S., Onasch, T. B., Canagaratna, M., Jayne, J. T., Kimmel, J., Yu, X. Y., Alexander, M. L., Worsnop, D. R.,
749 and Davidovits, P.: Single particle characterization using a light scattering module coupled to a time-of-
750 flight aerosol mass spectrometer, *Atmos. Chem. Phys.*, 9, 7769-7793, [https://doi.org/10.5194/acp-9-7769-](https://doi.org/10.5194/acp-9-7769-2009)
751 2009, 2009.

752 Davidson, C. I., Phalen, R. F., and Solomon, P. A.: Airborne particulate matter and human health: A review, *Aerosol*
753 *Sci. Technol.*, 39, 737-749, <https://doi.org/10.1080/02786820500191348>, 2005.

754 DeCarlo, P. F., Slowik, J. G., Worsnop, D. R., Davidovits, P., and Jimenez, J. L.: Particle morphology and density
755 characterization by combined mobility and aerodynamic diameter measurements. Part 1: Theory, *Aerosol*
756 *Sci. Technol.*, 38, 1185-1205, <https://doi.org/10.1080/027868290903907>, 2004.

757 Docherty, K. S., Aiken, A. C., Huffman, J. A., Ulbrich, I. M., DeCarlo, P. F., Sueper, D., Worsnop, D. R., Snyder,
758 D. C., Peltier, R. E., Weber, R. J., Grover, B. D., Eatough, D. J., Williams, B. J., Goldstein, A. H.,
759 Ziemann, P. J., and Jimenez, J. L.: The 2005 Study of Organic Aerosols at Riverside (SOAR-1):
760 instrumental intercomparisons and fine particle composition, *Atmos. Chem. Phys.*, 11, 12387-12420,
761 <https://doi.org/10.5194/acp-11-12387-2011>, 2011.

762 Docherty, K. S., Lewandowski, M., and Jimenez, J. L.: Effect of Vaporizer Temperature on Ambient Non-
763 Refractory Submicron Aerosol Composition and Mass Spectra Measured by the Aerosol Mass
764 Spectrometer, *Aerosol Sci. Technol.*, 49, 485-494, <https://doi.org/10.1080/02786826.2015.1042100>, 2015.

765 Drewnick, F., Hings, S. S., DeCarlo, P., Jayne, J. T., Gonin, M., Fuhrer, K., Weimer, S., Jimenez, J. L., Demerjian,
766 K. L., Borrmann, S., and Worsnop, D. R.: A new time-of-flight aerosol mass spectrometer (TOF-AMS) -
767 Instrument description and first field deployment, *Aerosol Sci. Technol.*, 39, 637-658,
768 <https://doi.org/10.1080/02786820500182040>, 2005.

769 Elser, M., Huang, R. J., Wolf, R., Slowik, J. G., Wang, Q. Y., Canonaco, F., Li, G. H., Bozzetti, C., Daellenbach, K.
770 R., Huang, Y., Zhang, R. J., Li, Z. Q., Cao, J. J., Baltensperger, U., El-Haddad, I., and Prevot, A. S. H.:
771 New insights into PM_{2.5} chemical composition and sources in two major cities in China during extreme
772 haze events using aerosol mass spectrometry, *Atmos. Chem. Phys.*, 16, 3207-3225,
773 <https://doi.org/10.5194/acp-16-3207-2016>, 2016.

774 Farmer, D. K., Matsunaga, A., Docherty, K. S., Surratt, J. D., Seinfeld, J. H., Ziemann, P. J., and Jimenez, J. L.:
775 Response of an aerosol mass spectrometer to organonitrates and organosulfates and implications for
776 atmospheric chemistry, *Proceedings of the National Academy of Sciences of the United States of America*,
777 107, 6670-6675, <https://doi.org/10.1073/pnas.0912340107>, 2010.

778 Fast, J., Aiken, A. C., Allan, J., Alexander, L., Campos, T., Canagaratna, M. R., Chapman, E., DeCarlo, P. F., de
779 Foy, B., Gaffney, J., de Gouw, J., Doran, J. C., Emmons, L., Hodzic, A., Herndon, S. C., Huey, G., Jayne,
780 J. T., Jimenez, J. L., Kleinman, L., Kuster, W., Marley, N., Russell, L., Ochoa, C., Onasch, T. B., Pekour,
781 M., Song, C., Ulbrich, I. M., Warneke, C., Welsh-Bon, D., Wiedinmyer, C., Worsnop, D. R., Yu, X. Y.,
782 and Zaveri, R.: Evaluating simulated primary anthropogenic and biomass burning organic aerosols during
783 MILAGRO: implications for assessing treatments of secondary organic aerosols, *Atmos. Chem. Phys.*, 9,
784 6191-6215, 2009.

785 Gill, P. E., Murray, W., and Wright, M. H.: The Levenberg-Marquardt method, in: *Practical optimization*, Academic
786 Press, London, 1981.

787 Griffith, S. M., Huang, X. H. H., Louie, P. K. K., and Yu, J. Z.: Characterizing the thermodynamic and chemical
788 composition factors controlling PM_{2.5} nitrate: Insights gained from two years of online measurements in
789 Hong Kong, *Atmos. Environ.*, 122, 864-875, <https://doi.org/10.1016/j.atmosenv.2015.02.009>, 2015.

790 Guo, H., Wang, D. W., Cheung, K., Ling, Z. H., Chan, C. K., and Yao, X. H.: Observation of aerosol size
791 distribution and new particle formation at a mountain site in subtropical Hong Kong, *Atmos. Chem. Phys.*,
792 12, 9923-9939, <https://doi.org/10.5194/acp-12-9923-2012>, 2012.

793 Huang, X. F., He, L. Y., Hu, M., Canagaratna, M. R., Sun, Y., Zhang, Q., Zhu, T., Xue, L., Zeng, L. W., Liu, X. G.,
794 Zhang, Y. H., Jayne, J. T., Ng, N. L., and Worsnop, D. R.: Highly time-resolved chemical characterization
795 of atmospheric submicron particles during 2008 Beijing Olympic Games using an Aerodyne High-
796 Resolution Aerosol Mass Spectrometer, *Atmos. Chem. Phys.*, 10, 8933-8945, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-10-8933-2010)
797 10-8933-2010, 2010.

798 Huang, X. F., He, L. Y., Hu, M., Canagaratna, M. R., Kroll, J. H., Ng, N. L., Zhang, Y. H., Lin, Y., Xue, L., Sun, T.
799 L., Liu, X. G., Shao, M., Jayne, J. T., and Worsnop, D. R.: Characterization of submicron aerosols at a rural
800 site in Pearl River Delta of China using an Aerodyne High-Resolution Aerosol Mass Spectrometer, *Atmos.*
801 *Chem. Phys.*, 11, 1865-1877, <https://doi.org/10.5194/acp-11-1865-2011>, 2011.

802 Huang, X. H. H., Bian, Q. J., Ng, W. M., Louie, P. K. K., and Yu, J. Z.: Characterization of PM2.5 Major
803 Components and Source Investigation in Suburban Hong Kong: A One Year Monitoring Study, *Aerosol*
804 *Air Qual. Res.*, 14, 237-250, <https://doi.org/10.4209/aaqr.2013.01.0020>, 2014.

805 Huffman, J. A., Jayne, J. T., Drewnick, F., Aiken, A. C., Onasch, T., Worsnop, D. R., and Jimenez, J. L.: Design,
806 modeling, optimization, and experimental tests of a particle beam width probe for the aerodyne aerosol
807 mass spectrometer, *Aerosol Sci. Technol.*, 39, 1143-1163, <https://doi.org/10.1080/02786820500423782>,
808 2005.

809 Jayne, J. T., Leard, D. C., Zhang, X. F., Davidovits, P., Smith, K. A., Kolb, C. E., and Worsnop, D. R.:
810 Development of an aerosol mass spectrometer for size and composition analysis of submicron particles,
811 *Aerosol Sci. Technol.*, 33, 49-70, <https://doi.org/10.1080/027868200410840>, 2000.

812 Jimenez, J. L., Jayne, J. T., Shi, Q., Kolb, C. E., Worsnop, D. R., Yourshaw, I., Seinfeld, J. H., Flagan, R. C., Zhang,
813 X. F., Smith, K. A., Morris, J. W., and Davidovits, P.: Ambient aerosol sampling using the Aerodyne
814 Aerosol Mass Spectrometer, *J. Geophys. Res. Atmos.*, 108, 8425, <https://doi.org/10.1029/2001jd001213>,
815 2003.

816 John, W.: Size Distribution Characteristics of Aerosols, in: *Aerosol Measurement*, John Wiley & Sons, Inc., 41-54,
817 2011.

818 Kerminen, V. M., Paramonov, M., Anttila, T., Riipinen, I., Fountoukis, C., Korhonen, H., Asmi, E., Laakso, L.,
819 Lihavainen, H., Swietlicki, E., Svenningsson, B., Asmi, A., Pandis, S. N., Kulmala, M., and Petaja, T.:
820 Cloud condensation nuclei production associated with atmospheric nucleation: a synthesis based on
821 existing literature and new results, *Atmos. Chem. Phys.*, 12, 12037-12059, [https://doi.org/10.5194/acp-12-](https://doi.org/10.5194/acp-12-12037-2012)
822 12037-2012, 2012.

823 Knote, C., Brunner, D., Vogel, H., Allan, J., Asmi, A., Aijala, M., Carbone, S., van der Gon, H. D., Jimenez, J. L.,
824 Kiendler-Scharr, A., Mohr, C., Poulain, L., Prevot, A. S. H., Swietlicki, E., and Vogel, B.: Towards an
825 online-coupled chemistry-climate model: evaluation of trace gases and aerosols in COSMO-ART, *Geosci.*
826 *Model Dev.*, 4, 1077-1102, <https://doi.org/10.5194/gmd-4-1077-2011>, 2011.

827 Köhler, H.: The nucleus in and the growth of hygroscopic droplets, *Trans. Faraday Soc.*, 32, 1152-1161,
828 <https://doi.org/10.1039/tf9363201152>, 1936.

829 Lee, B. P., Li, Y. J., Yu, J. Z., Louie, P. K. K., and Chan, C. K.: Physical and chemical characterization of ambient
830 aerosol by HR-ToF-AMS at a suburban site in Hong Kong during springtime 2011, *J. Geophys. Res.*
831 *Atmos.*, 118, 8625-8639, <https://doi.org/10.1002/jgrd.50658>, 2013.

832 Lee, B. P., Li, Y. J., Yu, J. Z., Louie, P. K. K., and Chan, C. K.: Characteristics of submicron particulate matter at
833 the urban roadside in downtown Hong Kong-Overview of 4 months of continuous high-resolution aerosol
834 mass spectrometer measurements, *J. Geophys. Res. Atmos.*, 120, 7040-7058,
835 <https://doi.org/10.1002/2015jd023311>, 2015.

836 Li, Y. J., Lee, B. Y. L., Yu, J. Z., Ng, N. L., and Chan, C. K.: Evaluating the degree of oxygenation of organic
837 aerosol during foggy and hazy days in Hong Kong using high-resolution time-of-flight aerosol mass
838 spectrometry (HR-ToF-AMS), *Atmos. Chem. Phys.*, 13, 8739-8753, [https://doi.org/10.5194/acp-13-8739-](https://doi.org/10.5194/acp-13-8739-2013)
839 [2013](https://doi.org/10.5194/acp-13-8739-2013), 2013.

840 Li, Y. J., Lee, B. P., Su, L., Fung, J. C. H., and Chan, C. K.: Seasonal characteristics of fine particulate matter (PM)
841 based on high-resolution time-of-flight aerosol mass spectrometric (HR-ToF-AMS) measurements at the
842 HKUST Supersite in Hong Kong, *Atmos. Chem. Phys.*, 15, 37-53, <https://doi.org/10.5194/acp-15-37-2015>,
843 2015.

844 Liu, P. S. K., Deng, R., Smith, K. A., Williams, L. R., Jayne, J. T., Canagaratna, M. R., Moore, K., Onasch, T. B.,
845 Worsnop, D. R., and Deshler, T.: Transmission efficiency of an aerodynamic focusing lens system:
846 Comparison of model calculations and laboratory measurements for the Aerodyne Aerosol Mass
847 Spectrometer, *Aerosol Sci. Technol.*, 41, 721-733, <https://doi.org/10.1080/02786820701422278>, 2007.

848 Man, H., Zhu, Y., Ji, F., Yao, X., Lau, N. T., Li, Y., Lee, B. P., and Chan, C. K.: Comparison of daytime and
849 nighttime new particle growth at the HKUST supersite in Hong Kong, *Environ. Sci. Technol.*, 49, 7170-
850 7178, <https://doi.org/10.1021/acs.est.5b02143>, 2015.

851 Meng, J. W., Yeung, M. C., Li, Y. J., Lee, B. Y. L., and Chan, C. K.: Size-resolved cloud condensation nuclei
852 (CCN) activity and closure analysis at the HKUST Supersite in Hong Kong, *Atmos. Chem. Phys.*, 14,
853 10267-10282, <https://doi.org/10.5194/acp-14-10267-2014>, 2014.

854 Middlebrook, A. M., Bahreini, R., Jimenez, J. L., and Canagaratna, M. R.: Evaluation of Composition-Dependent
855 Collection Efficiencies for the Aerodyne Aerosol Mass Spectrometer using Field Data, *Aerosol Sci.*
856 *Technol.*, 46, 258-271, <https://doi.org/10.1080/02786826.2011.620041>, 2012.

857 Mohr, C., DeCarlo, P. F., Heringa, M. F., Chirico, R., Slowik, J. G., Richter, R., Reche, C., Alastuey, A., Querol, X.,
858 Seco, R., Penuelas, J., Jimenez, J. L., Crippa, M., Zimmermann, R., Baltensperger, U., and Prevot, A. S. H.:
859 Identification and quantification of organic aerosol from cooking and other sources in Barcelona using
860 aerosol mass spectrometer data, *Atmos. Chem. Phys.*, 12, 1649-1665, [https://doi.org/10.5194/acp-12-1649-](https://doi.org/10.5194/acp-12-1649-2012)
861 [2012](https://doi.org/10.5194/acp-12-1649-2012), 2012.

862 Quinn, P. K., Bates, T. S., Coffman, D., Onasch, T. B., Worsnop, D., Baynard, T., de Gouw, J. A., Goldan, P. D.,
863 Kuster, W. C., Williams, E., Roberts, J. M., Lerner, B., Stohl, A., Pettersson, A., and Lovejoy, E. R.:

864 Impacts of sources and aging on submicrometer aerosol properties in the marine boundary layer across the
865 Gulf of Maine, *J. Geophys. Res. Atmos.*, 111, D23S36, <https://doi.org/10.1029/2006jd007582>, 2006.

866 Rupakheti, M., Leaitch, W. R., Lohmann, U., Hayden, K., Brickell, P., Lu, G., Li, S. M., Toom-Saunty, D.,
867 Bottenheim, J. W., Brook, J. R., Vet, R., Jayne, J. T., and Worsnop, D. R.: An intensive study of the size
868 and composition of submicron atmospheric aerosols at a rural site in Ontario, Canada, *Aerosol Sci.*
869 *Technol.*, 39, 722-736, <https://doi.org/10.1080/02786820500182420>, 2005.

870 Saarikoski, S., Carbone, S., Decesari, S., Giulianelli, L., Angelini, F., Canagaratna, M., Ng, N. L., Trimborn, A.,
871 Facchini, M. C., Fuzzi, S., Hillamo, R., and Worsnop, D.: Chemical characterization of springtime
872 submicrometer aerosol in Po Valley, Italy, *Atmos. Chem. Phys.*, 12, 8401-8421,
873 <https://doi.org/10.5194/acp-12-8401-2012>, 2012.

874 Salcedo, D., Onasch, T. B., Dzepina, K., Canagaratna, M. R., Zhang, Q., Huffman, J. A., DeCarlo, P. F., Jayne, J. T.,
875 Mortimer, P., Worsnop, D. R., Kolb, C. E., Johnson, K. S., Zuberi, B., Marr, L. C., Volkamer, R., Molina,
876 L. T., Molina, M. J., Cardenas, B., Bernabe, R. M., Marquez, C., Gaffney, J. S., Marley, N. A., Laskin, A.,
877 Shutthanandan, V., Xie, Y., Brune, W., Leshner, R., Shirley, T., and Jimenez, J. L.: Characterization of
878 ambient aerosols in Mexico City during the MCMA-2003 campaign with Aerosol Mass Spectrometry:
879 results from the CENICA Supersite, *Atmos. Chem. Phys.*, 6, 925-946, 2006.

880 Schwartz, S. E.: The Whitehouse effect - Shortwave radiative forcing of climate by anthropogenic aerosols: An
881 overview, *J. Aerosol Sci*, 27, 359-382, [https://doi.org/10.1016/0021-8502\(95\)00533-1](https://doi.org/10.1016/0021-8502(95)00533-1), 1996.

882 Seinfeld, J. H., and Pandis, S. N.: *Atmospheric Chemistry and Physics - From Air Pollution to Climate Change*, 2
883 ed., John Wiley & Sons, 2006.

884 Setyan, A., Zhang, Q., Merkel, M., Knighton, W. B., Sun, Y., Song, C., Shilling, J. E., Onasch, T. B., Herndon, S.
885 C., Worsnop, D. R., Fast, J. D., Zaveri, R. A., Berg, L. K., Wiedensohler, A., Flowers, B. A., Dubey, M. K.,
886 and Subramanian, R.: Characterization of submicron particles influenced by mixed biogenic and
887 anthropogenic emissions using high-resolution aerosol mass spectrometry: results from CARES, *Atmos.*
888 *Chem. Phys.*, 12, 8131-8156, <https://doi.org/10.5194/acp-12-8131-2012>, 2012.

889 Slowik, J. G., Stankin, K., Davidovits, P., Williams, L. R., Jayne, J. T., Kolb, C. E., Worsnop, D. R., Rudich, Y.,
890 DeCarlo, P. F., and Jimenez, J. L.: Particle morphology and density characterization by combined mobility
891 and aerodynamic diameter measurements. Part 2: Application to combustion-generated soot aerosols as a
892 function of fuel equivalence ratio, *Aerosol Sci. Technol.*, 38, 1206-1222,
893 <https://doi.org/10.1080/027868290903916>, 2004.

894 Sun, C., Lee, B. P., Huang, D., Li, Y. J., Schurman, M. I., Louie, P. K. K., Luk, C., and Chan, C. K.: Continuous
895 measurements at the urban roadside in an Asian megacity by Aerosol Chemical Speciation Monitor
896 (ACSM): particulate matter characteristics during fall and winter seasons in Hong Kong, *Atmos. Chem.*
897 *Phys.*, 16, 1713-1728, <https://doi.org/10.5194/acp-16-1713-2016>, 2016.

898 Sun, Y., Zhang, Q., Macdonald, A. M., Hayden, K., Li, S. M., Liggio, J., Liu, P. S. K., Anlauf, K. G., Leaitch, W.
899 R., Steffen, A., Cubison, M., Worsnop, D. R., van Donkelaar, A., and Martin, R. V.: Size-resolved aerosol

900 chemistry on Whistler Mountain, Canada with a high-resolution aerosol mass spectrometer during INTEX-
901 B, *Atmos. Chem. Phys.*, 9, 3095-3111, <https://doi.org/10.5194/acp-9-3095-2009>, 2009.

902 Sun, Y. L., Zhang, Q., Schwab, J. J., Demerjian, K. L., Chen, W. N., Bae, M. S., Hung, H. M., Hogrefe, O., Frank,
903 B., Rattigan, O. V., and Lin, Y. C.: Characterization of the sources and processes of organic and inorganic
904 aerosols in New York city with a high-resolution time-of-flight aerosol mass spectrometer, *Atmos. Chem.*
905 *Phys.*, 11, 1581-1602, <https://doi.org/10.5194/acp-11-1581-2011>, 2011.

906 Takegawa, N., Miyakawa, T., Watanabe, M., Kondo, Y., Miyazaki, Y., Han, S., Zhao, Y., van Pinxteren, D.,
907 Bruggemann, E., Gnauk, T., Herrmann, H., Xiao, R., Deng, Z., Hu, M., Zhu, T., and Zhang, Y.:
908 Performance of an Aerodyne Aerosol Mass Spectrometer (AMS) during Intensive Campaigns in China in
909 the Summer of 2006, *Aerosol Sci. Technol.*, 43, 189-204, <https://doi.org/10.1080/02786820802582251>,
910 2009.

911 Ulbrich, I. M., Canagaratna, M. R., Cubison, M. J., Zhang, Q., Ng, N. L., Aiken, A. C., and Jimenez, J. L.: Three-
912 dimensional factorization of size-resolved organic aerosol mass spectra from Mexico City, *Atmos. Meas.*
913 *Tech.*, 5, 195-224, <https://doi.org/10.5194/amt-5-195-2012>, 2012.

914 Westervelt, D. M., Pierce, J. R., Riipinen, I., Trivitanurak, W., Hamed, A., Kulmala, M., Laaksonen, A., Decesari,
915 S., and Adams, P. J.: Formation and growth of nucleated particles into cloud condensation nuclei: model-
916 measurement comparison, *Atmos. Chem. Phys.*, 13, 7645-7663, <https://doi.org/10.5194/acp-13-7645-2013>,
917 2013.

918 Williams, L. R., Gonzalez, L. A., Peck, J., Trimborn, D., McInnis, J., Farrar, M. R., Moore, K. D., Jayne, J. T.,
919 Robinson, W. A., Lewis, D. K., Onasch, T. B., Canagaratna, M. R., Trimborn, A., Timko, M. T., Magoon,
920 G., Deng, R., Tang, D., Blanco, E. D. L. R., Prevot, A. S. H., Smith, K. A., and Worsnop, D. R.:
921 Characterization of an aerodynamic lens for transmitting particles greater than 1 micrometer in diameter
922 into the Aerodyne aerosol mass spectrometer, *Atmos. Meas. Tech.*, 6, 3271-3280,
923 <https://doi.org/10.5194/amt-6-3271-2013>, 2013.

924 Yao, X., Ling, T. Y., Fang, M., and Chan, C. K.: Size dependence of in situ pH in submicron atmospheric particles
925 in Hong Kong, *Atmos. Environ.*, 41, 382-393, <https://doi.org/10.1016/j.atmosenv.2006.07.037>, 2007a.

926 Yao, X. H., Lau, N. T., Chan, C. K., and Fang, M.: Size distributions and condensation growth of submicron
927 particles in on-road vehicle plumes in Hong Kong, *Atmos. Environ.*, 41, 3328-3338,
928 <https://doi.org/10.1016/j.atmosenv.2006.12.044>, 2007b.

929 Yuan, Z. B., Yadav, V., Turner, J. R., Louie, P. K. K., and Lau, A. K. H.: Long-term trends of ambient particulate
930 matter emission source contributions and the accountability of control strategies in Hong Kong over 1998-
931 2008, *Atmos. Environ.*, 76, 21-31, <https://doi.org/10.1016/j.atmosenv.2012.09.026>, 2013.

932 Zhang, J. K., Sun, Y., Liu, Z. R., Ji, D. S., Hu, B., Liu, Q., and Wang, Y. S.: Characterization of submicron aerosols
933 during a month of serious pollution in Beijing, 2013, *Atmos. Chem. Phys.*, 14, 2887-2903,
934 <https://doi.org/10.5194/acp-14-2887-2014>, 2014.

935 Zhang, Q., Stanier, C. O., Canagaratna, M. R., Jayne, J. T., Worsnop, D. R., Pandis, S. N., and Jimenez, J. L.:
936 Insights into the chemistry of new particle formation and growth events in Pittsburgh based on aerosol
937 mass spectrometry, *Environ. Sci. Technol.*, 38, 4797-4809, <https://doi.org/10.1021/es035417u>, 2004.
938 Zhang, Q., Canagaratna, M. R., Jayne, J. T., Worsnop, D. R., and Jimenez, J. L.: Time- and size-resolved chemical
939 composition of submicron particles in Pittsburgh: Implications for aerosol sources and processes, *J.*
940 *Geophys. Res. Atmos.*, 110, <https://doi.org/10.1029/2004jd004649>, 2005.
941 Zheng, M., Kester, D. R., Wang, F., Shi, X. M., and Guo, Z. G.: Size distribution of organic and inorganic species in
942 Hong Kong aerosols during the wet and dry seasons, *J. Geophys. Res. Atmos.*, 113, D16303,
943 <https://doi.org/10.1029/2007jd009494>, 2008.
944 Zhuang, H., Chan, C. K., Fang, M., and Wexler, A. S.: Size distributions of particulate sulfate, nitrate, and
945 ammonium at a coastal site in Hong Kong, *Atmos. Environ.*, 33, 843-853, [https://doi.org/10.1016/S1352-](https://doi.org/10.1016/S1352-2310(98)00305-7)
946 [2310\(98\)00305-7](https://doi.org/10.1016/S1352-2310(98)00305-7), 1999.