

1 Diurnal and day-to-day characteristics of ambient particle mass 2 size distributions from HR-ToF-AMS measurements at an urban 3 site and a suburban site in Hong Kong

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8 **Abstract.** Mass concentration based particle size distributions measured by a high-resolution aerosol mass
9 spectrometer were systematically analyzed to assess long and short-term temporal characteristics of ambient particle
10 size distributions sampled at a typical urban environment close to emission sources and a suburban coastal site
11 representing a regional and local pollution receptor location in Hong Kong. Measured distributions were bimodal and
12 deconvoluted into submodes which were analyzed for day-to-day variations and diurnal variations.

13 Traffic and cooking emissions at the urban site contributed substantially to particle mass in both modes, while notable
14 decreases in mass median diameters were limited to the morning rush hour. Inorganic particle components displayed
15 varying diurnal behavior, including nocturnal nitrate formation and daytime photochemical formation evident in both
16 modes. Suburban particle size distributions exhibited notable seasonal disparities with differing influence of local
17 formation, particularly in spring and summer, and transport which dominated in the fall season leading to notably
18 higher sulfate and organic accumulation mode particle concentrations. Variations in particle mixing state were
19 evaluated by comparison of inter-species mass median diameter trends at both measurement sites. Internal mixing was
20 prevalent in the accumulation mode in spring at the urban site, while greater frequency of time periods with external
21 mixing of particle populations comprising different fractions of organic constituents was observed in summer. At the
22 suburban site, sulfate and nitrate in the accumulation mode more frequently exhibited differing particle size
23 distributions in all seasons signifying a greater extent of external mixing.

24 At the urban site, periods of greater submicron inorganic mass concentrations were more likely to be caused by
25 increases in both Aitken and accumulation mode particle mass in summer, while at the suburban receptor location
26 organic and nitrate Aitken mode particle mass contributed more regularly to higher total submicron species mass
27 concentrations in most seasons (spring, summer and winter).

28

29

30 1. Introduction

31 Apart from mass and chemical composition, the size distribution of fine particles represents a vital physical property
32 with important implications for human health and environmental effects of ambient aerosols (Seinfeld and Pandis,
33 2006). Particle size relates directly to the aerodynamic properties which govern the penetration and deposition of
34 particles in the airways and lungs (Davidson et al., 2005) as well as the scattering and absorption of light which affect
35 the radiative properties and hence ambient visibility (Ahlquist and Charlson, 1967;Bohren and Huffman,
36 1983;Charlson et al., 1991;Schwartz, 1996;Seinfeld and Pandis, 2006). Hygroscopic growth in response to changes in
37 ambient humidity can alter particle light scattering properties (Seinfeld and Pandis, 2006;Köhler, 1936) and activation
38 of condensation nuclei particles into cloud droplets depend on atmospheric conditions, chemical composition, mixing
39 state as well as the size and morphology of particles (Abbatt et al., 2005;Kerminen et al., 2012;Meng et al.,
40 2014;Westervelt et al., 2013).

41 Studies into the size distribution of ambient particulate matter in Hong Kong have been largely based on size-
42 segregated filter samples (Yao et al., 2007b;Zheng et al., 2008;Zhuang et al., 1999;Huang et al., 2014;Bian et al.,
43 2014) and measurements by electrostatic classifier instruments (Cheung et al., 2015;Yao et al., 2007a) and were hence
44 either limited in size resolution (offline filter samples) or chemical resolution (total particle count by classification).
45 Most measurements in Hong Kong were conducted in suburban environments. Inorganic ammonium and sulfate were
46 mainly found in fine mode particles in condensation and droplet mode size ranges, while nitrate had strong coarse
47 mode contributions (Zhuang et al., 1999). Seasonal differences were evident in solvent-extractable organics and trace
48 metals which were mainly found in PM_{0.5} particles in the wet season and winter whereas in fall a shift to larger particles
49 (0.5–2.5 µm fraction) in fall indicated a possibly stronger influence of aged particle components in the transition period
50 of the Asian monsoon (Zheng et al., 2008). Size distributions acquired by a fast mobility particle sizer at the suburban
51 HKUST supersite were investigated more recently to study the formation and accumulation of ultrafine particles under
52 different air flow regimes. Particle number concentration enhancements during the day were attributed to secondary
53 formation, while evening and nighttime peaks were thought to be related to transport of aged aerosols from upwind
54 locations. Nucleation mode particle peaks were often observed in fall and related to regional pollution influence
55 (Cheung et al., 2015). New particle formation events at the same site occurred as single and two-stage growth
56 processes with organics and sulfuric acid contributing mainly to first stage growth in the daytime while nighttime
57 second stage growth was attributed to ammonium nitrate and organics. Particle size growth into the diameter range of
58 cloud condensation nuclei (CCN) was typically only achieved with the second growth stage (Man et al., 2015).

59 Investigations into particle size distributions in urban areas of Hong Kong are even scarcer. Yao et al. (Yao et al.,
60 2007a) studied the properties and behavior of particles in vehicle plumes and reported a competing process between
61 ambient background particles and fresh soot particles in the condensation of gaseous precursors and a dependency on
62 temperature with bimodal volume size distributions observed at lower ambient temperatures and unimodal
63 distributions in the lower accumulation size range at higher ambient temperatures.

64 The Aerodyne aerosol mass spectrometer (Canagaratna et al., 2007) is widely used to determine the chemical
65 composition of major organic and inorganic components of non-refractory submicron particulate matter (NR-PM₁).
66 In contrast to most traditional aerosol sizing instruments, the AMS is capable of resolving main chemical constituents

67 within size distributions through analysis of particle flight times and particle ensemble mass spectra (Canagaratna et
68 al., 2007;Jayne et al., 2000;Jimenez et al., 2003;Rupakheti et al., 2005) and thus yields valuable additional information
69 on differences in composition of submicron particles with the gross of particle mass in the Aitken mode range ($D_p \sim$
70 10-100nm) and the accumulation mode range ($D_p \sim 100-1000\text{nm}$) covered by the AMS. Thus far most studies
71 employing ambient size distribution data from aerosol mass spectrometer measurements investigated longer time
72 period averages, i.e. campaign averages (Salcedo et al., 2006;Sun et al., 2009;Aiken et al., 2009;Huang et al.,
73 2010;Takegawa et al., 2009;Saarikoski et al., 2012;Li et al., 2015) or specific time periods of interest (Elser et al.,
74 2016;Lee et al., 2013). Mohr et al. separated organic particle mass size distributions by periods of dominant influence
75 of different PMF-resolved organic aerosol factors to study the properties of mass size distributions in relation to
76 organic aerosol composition (Mohr et al., 2012). The 3D-factorization technique is an extension of traditional AMS
77 PMF analysis on organic aerosol allowing to obtain estimates on the size distributions of organic aerosol factors,
78 however under the assumption that factor size distributions remain invariant over the measurement period (Ulbrich et
79 al., 2012).

80 The temporal evolution of species-specific size distributions, are mostly discussed qualitatively (Drewnick et al.,
81 2005) and only few studies have evaluated temporal trends in mass size distributions in greater detail.
82 Particle nucleation and subsequent growth events were investigated in Pittsburgh using size data from an AMS and
83 two SMPS as well as various gaseous pollutant instruments and meteorological information. The AMS mass size
84 distributions were evaluated quantitatively using the time series of binned particle concentrations generated from the
85 grouping of raw data into wider size bins to represent different stages in the particle growth process. (Zhang et al.,
86 2004). The same method was employed to evaluate contributions of ultrafine mode and accumulation mode particles
87 to total organic particle mass (Zhang et al., 2005) by summation of size bins in the range of 30-100 nm and 100-
88 1000nm. The authors also explored diurnal changes in size distributions of particle species by averaging over 3h
89 periods in the morning (6–9 am) and afternoon (1–4 pm). Sun et al. present a qualitative discussion of diurnal
90 variations in the mass size distributions of the m/z 44, m/z 57 and derived C_4H_9^+ ion signals from measurements at an
91 urban site in New York (Sun et al., 2011). Similarly, Setyan et al. examined diurnal changes in the mass size
92 distributions of organics and sulfate qualitatively and used binned concentrations (40–120, 120–200, and 200–800)
93 nm in their quantitative analysis to study the evolution of particle chemistry in new particle formation and growth
94 events (Setyan et al., 2012).

95 In this work, we introduce a systematic approach of assessing temporal variations in AMS mass-based particle size
96 distributions from hourly diurnal variations to 24h-average based day-to-day trends to utilize two key instrumental
97 advantages, i.e. species segregation and high time resolution, to obtain a more detailed and quantitative understanding
98 of the variabilities in ambient particle mass size distributions and to provide an additional dimension to standard AMS
99 data analysis techniques. In this context, we present a detailed discussion of particle size data from HR-ToF-AMS
100 measurements during two field campaigns in Hong Kong in both urban and suburban environments. We aim to
101 evaluate characteristic recurrent changes in size distribution as well as longer term trends in different seasons by
102 analyzing day-to-day variations and diurnal variations of size distributions of submicron organics, sulfate, and nitrate
103 particle mass. The two contrasting sites represent a typical urban source environment (inner-city, roadside station)

104 close to primary emission sources and a suburban location (coastal, HKUST supersite) which is largely a downwind
105 receptor of varying amounts of local urban, regional and long-range transported pollutants (Li et al., 2015;Huang et
106 al., 2014).

107

108 **2. Methodology**

109 **2.1. Field campaigns**

110 Sampling of ambient submicron non-refractory particulate matter (NR-PM₁) was carried out using an Aerodyne HR-
111 ToF-AMS at the HKUST air quality supersite covering four seasons between May 2011 and February 2012 (spring:
112 2011-05, summer: 2011-09, fall: 2011-11&12, winter: 2012-02). The HKUST supersite is located on the campus of
113 the Hong Kong University of Science and Technology (22°20'N, 114°16'E), on the east coast of Hong Kong in a
114 suburban area with few primary emission sources in the immediate vicinity. Sampled air was drawn from the rooftop
115 of a pump house building at an approximate height of 25m above ground level. For detailed descriptions of the
116 experimental setup, operating conditions, data treatment, and overall species composition we refer the reader to
117 previous publications (Lee et al., 2013;Li et al., 2015;Li et al., 2013). A further sampling campaign took place between
118 spring 2013 (2013-03 to 2013-05) and summer 2013 (2013-05 to 2013-07) at an inner-city urban location in the
119 densely populated and built-up Kowloon peninsula. Measurements were conducted next to the roadside air quality
120 monitoring station (AQMS) operated by the Environmental Protection Department (EPD) of the HKSAR Government
121 in the Mong Kok (MK) district on a pedestrian crossing at a major road junction. Sampled air was drawn from a height
122 of 3m above ground level. A comprehensive analysis of trends in species concentration and composition identified in
123 this urban campaign has been presented previously (Lee et al., 2015). In both campaigns, particles were sampled
124 through a PM_{2.5} cyclone at a flow rate of 16.67 L/min into a sampling port from which 0.08 L/min was drawn by the
125 AMS and the remainder drawn by co-sampling instruments and an auxiliary pump. Sample air for the AMS passed
126 through a 1m long diffusion dryer (BMI, San Francisco CA, USA) filled with silica gel to remove bulk gas- and
127 particle-phase water. Additional data from various collocated instruments including meteorological data (wind,
128 temperature, relative humidity, solar irradiation), volatile organic compounds (VOCs) and standard trace gases such
129 as NO_x, SO₂, and O₃ were available.

130 AMS data were treated according to general AMS data treatment principles (DeCarlo et al., 2006;Jimenez et al., 2003)
131 with standard software packages (SQUIRREL, PIKA). Analysis of the unit-mass resolution mass spectra yielded non-
132 refractory submicron particle species concentrations of major inorganic constituents (SO₄, NO₃, NH₄, Chl) and total
133 organics at a base time resolution of 10 min. Positive Matrix Factorization (PMF) was used to deconvolute high-
134 resolution organic mass spectra acquired at 10 min time resolution following recommended PMF guidelines for AMS
135 data (Zhang et al., 2011) with the AMS PMF analysis toolkit (Ulbrich et al., 2009). At the urban Mong Kok site, six
136 organic aerosol (OA) factors were identified encompassing three secondary organic aerosol (SOA) and three primary
137 organic aerosol (POA) factors of which one was attributed to traffic emissions and two to cooking activities (Lee et
138 al., 2015). Similarly, four factors were obtained from analysis of the urban HKUST site dataset with two SOA factors

139 and two POA factors, related to traffic and cooking respectively (Li et al., 2015). Further details on the treatment of
140 AMS size distribution data from both sampling campaigns are provided in the following section.

141 2.2. Data acquisition and treatment

142 In both campaigns, mass concentration based size distributions in terms of vacuum-aerodynamic particle diameter
143 ($dM/d\log D_{va}$) were established by joint acquisition of particle time-of-flight (PToF) measurements and unit mass
144 resolution mass spectra (V-mode) with alternation between modes every 20s for 30 cycles amounting to 5 min of total
145 sampling time. High-resolution mass spectra were acquired for the following 5 min, and thus the overall raw data time
146 resolution for each mode was equal to 10 min. The total particle mass measured in the PToF mode was normalized to
147 the V-mode mass concentration of the same time step. Daily size distributions were generated by averaging over 24h
148 periods (from 0:00 to 23:59). Hourly diurnal size distributions were reconstructed by grouping size distributions within
149 the same hour of the day and establishing representative size distributions based on average, median, 25th and 75th
150 percentile concentration values of each size bin (*referred to as size distribution sets hereinafter*).

151 At both sampling sites, the seasonally averaged AMS size distributions were bimodal (Lee et al., 2013, 2015; Li et al.,
152 2015) with similar distributions having been observed in other AMS field studies in various parts of the world. (Zhang
153 et al., 2014; Sun et al., 2011; Huang et al., 2011; Aiken et al., 2009; Zhang et al., 2005; Crippa et al., 2013; Docherty et
154 al., 2011; Mohr et al., 2012). Multimodality of size distributions is typical for environments where different sources or
155 formation processes of particles play a role and accordingly such distributions can also be represented as sums of
156 discrete lognormal distributions of the respective constituting submodes (John, 2011).

157 The measured bimodal size distributions in this work were deconvoluted by fitting two log-normal distributed modes,
158 including one closer to the Aitken size range (*mode diameter ~100nm*) and one in the accumulation size range (*mode*
159 *diameter ~500nm*) employing the Levenberg-Marquardt algorithm (Gill et al., 1981) as a non-linear least squares fit,
160 to evaluate differences in trends and formation or transformation processes in the two size regimes. An example of a
161 size distribution fit and associated parameters is depicted in Figure D1 in the Supporting Material. Additional fit
162 residual analyses were carried out in cases where the Aitken mode only accounted for small parts of (<10%) of the
163 total particle mass and uncertainties in integrated mode particle mass from the peak fitting were examined for all size
164 distributions. Details are presented in Section B of the Supporting Material. The smaller mode typically exhibited
165 mode diameters in the range of 100-200 nm (D_{va}) and is thus in the transition region between Aitken and lower
166 accumulation mode. For a clearer distinction from the larger mode which unambiguously belonged to the
167 accumulation size range, we opt to refer to the small mode as *Aitken mode* in this work. Mode diameter (*i.e.* mass
168 median diameter, MMD), curve width (*i.e.* geometric standard deviation, GSD) and curve area (*equivalent to particle*
169 *mass concentration within the mode*) are sufficient parameters to completely describe a lognormal distribution and
170 these key variables are used in the following analysis on trends in the fitted species-specific size distributions of
171 organics, nitrate, and sulfate from both HR-AMS sampling campaigns in Hong Kong. Particle diameters are discussed
172 in terms of vacuum-aerodynamic diameter, with detailed discussions on properties and relationships to other size
173 metrics available elsewhere (DeCarlo et al., 2004; Slowik et al., 2004). Further details on procedures of PToF data

174 acquisition and size distribution averaging can be found in the Supporting Material in Section A and B respectively.
175 The sequence of main data treatment and analysis steps is shown in Figure 1.
176 The transmission efficiency of the AMS aerodynamic lens is known to fall off below ~100nm and beyond ~550 nm
177 of vacuum-aerodynamic diameter (Liu et al., 2007;Takegawa et al., 2009;Zhang et al., 2004;Bahreini et al.,
178 2008;Williams et al., 2013;Knote et al., 2011) and may bias measured particle mass and mode diameters, particularly
179 in the accumulation mode towards lower values if significant particle mass fractions fall in the size region of $D_{va} >$
180 550 nm. In the Aitken mode range, the effect of limited lens transmission is expected to be less substantial as particle
181 volume (and hence particle mass) of Aitken mode particles are much smaller. We discuss the effects of lens
182 transmission briefly in section 3.4. Delayed vaporization of particle components, e.g. under high mass loadings, can
183 lead to small shifts towards larger mode diameters in AMS size distributions (Docherty et al., 2015) and enhanced
184 tails in the size distributions (Cross et al., 2009), which may lead to larger fit residuals at the trailing edges. Generally,
185 the discussion of size distributions in this work should be viewed in the context of the instrumental capabilities and
186 previously mentioned limitations of aerosol mass spectrometry. Therefore, the resolved Aitken and accumulation
187 modes in this work reflect the apparent Aitken and accumulation modes within AMS measurable particle mass size
188 distributions.
189

293 semi-volatile constituents towards the particle-phase. We expect these volatility effects to be a main contributing
294 factor, as sampling took place in direct vicinity of the emission source, i.e. next to the road and thus potential impacts
295 of physical effects such as enhanced near-ground mixing and dispersion through thermally induced convection in
296 summer are expected to be of minor influence. Considering the previously discussed estimated traffic contributions
297 during the rush hour, the seasonal difference in mass concentration was much more pronounced in the Aitken mode
298 (-67%, 0.6 $\mu\text{g}/\text{m}^3$) than the accumulation mode (-12%, 0.2 $\mu\text{g}/\text{m}^3$), consistent with the expected stronger impact of
299 reduced particle nucleation and reduced condensation of semi-volatile exhaust components on fresher, smaller
300 particles in the warmer season.

301 Comparing different size distribution sets (Figure D6 in the Supporting Material), the average concentration set in
302 summer yielded notably larger resolved mass median diameters in both modes and greater Aitken mode mass
303 compared to the median, 25th, and 75th percentile concentration sets. This indicates a strong influence of extreme
304 values (i.e. time periods with both larger particle size and larger particle mass concentrations) and thus greater
305 variability in size distributions in the warmer season caused by specific high and low concentration events such as
306 photochemical episodes and precipitation, evident in the greater relative span of organic mass concentrations in
307 summer (See Table C5 in the Supporting Material: ratio of 10th and 90th percentile to median concentration in NR-
308 PM₁). In spring, such events masked the diurnal processes to a lesser extent and with consequently greater consistency
309 across different size distribution sets.

310

311 Sulfate

312 Although variations of total submicron sulfate mass concentrations with time of day were generally subtle, distinct
313 trends were notable in MMDs and integrated mode mass concentrations in both Aitken and accumulation mode.

314 Generally, Aitken mode MMDs were 20% larger in spring (180nm) than in summer (150nm). While in spring
315 fluctuations in Aitken mode MMDs were small throughout the day within a narrow range of +/- 10nm and without
316 apparent regular features, the summertime diurnal variation exhibited a well-defined broad daytime peak with a shift
317 to ~15nm larger particle diameters. A matching trend was evident in the accumulation mode where MMDs increased
318 by ~20nm in summer. Conversely, in spring, a conspicuous nighttime peak in accumulation mode MMDs was
319 observed in the low traffic period between 01:00 and 07:00 which tracked closely with the diurnal variation of O₃
320 which peaked in the same period with the reduction of the NO_x titration effects at low nighttime traffic volumes. While
321 particulate sulfate production during the day can be achieved through both homogeneous gas-phase oxidation of SO₂
322 by the OH radical as well as heterogeneous oxidation of SO₂ by dissolved H₂O₂ or O₃ (Seinfeld and Pandis, 2006),
323 nighttime production is limited to the non-photochemical heterogeneous pathway. The apparent increase in
324 accumulation mode particle size was also associated with an increase of integrated submode particle mass by ~0.7
325 $\mu\text{g}/\text{m}^3$ and thus suggests the possibility of heterogeneous SO₂ oxidation by residual ozone in the cooler and more
326 humid spring season as a local source of particulate sulfate. In the warmer and drier summer season, no corresponding
327 trend was apparent in either accumulation mode MMD or integrated mode concentration. The small magnitude of
328 additionally produced sulfate (< 1 $\mu\text{g}/\text{m}^3$) in spring renders the nighttime production a minor source of particulate
329 sulfate however and affirms that the bulk of the accumulation mode sulfate burden at the urban roadside still originated

385 **Figure 4.** Diurnal variations of mode diameter (MMD), integrated mode mass concentration and width of the Aitken mode (*lighter*
386 *color*) and accumulation mode (*darker color*) from bimodal peak fits of the bin-median reconstructed size distributions at the
387 suburban HKUST site and V-mode AMS species concentrations (line with shaded background) for organics, nitrate and sulfate
388 (left to right) in (a) Spring 2011, (b) Summer 2011, (c) Fall 2011 and (d) Winter 2012; The right-most panel depicts the median
389 diurnal variations of relevant gas-phase pollutants (O₃, CO, NO_x, SO₂) measured at the same site.

390 3.1.2. Suburban coastal NR-PM₁

391 The suburban HKUST site as a downwind receptor of urban and regional pollution was generally dominated by sulfate
392 and oxygenated secondary organic aerosol (SOA) components and much lower fractions of primary organic
393 constituents, which combined typically made up less than a quarter of total organics (Li et al., 2015). Trends in the
394 species segregated particle size distributions are discussed analogously to section 3.1.1., with Figure 4 illustrating the
395 diurnal trends of the fitting parameters (MMD, integrated mode mass, geometric standard deviation) for organics,
396 sulfate, and nitrate at the suburban HKUST site.

397 Organics

398 Organics
399 There were significant seasonal differences with larger fractions (Figure 3b) and concentrations (Figure 5c) of Aitken
400 mode mass in total organic submicron particle mass in spring and summer compared to fall and winter, indicating
401 greater influence of closer-ranged formation sources in the warmer season. Springtime integrated Aitken mode mass
402 concentrations ($\sim 0.8 \mu\text{g}/\text{m}^3$) were twice as high as those in winter ($\sim 0.4 \mu\text{g}/\text{m}^3$). In the accumulation mode, highest
403 particle mass loadings were observed in fall ($5 \mu\text{g}/\text{m}^3$) and lowest loadings in spring ($3 \mu\text{g}/\text{m}^3$) following the frequency
404 pattern of continental air mass influence (Figure D12 in the Supporting Material) in each season indicating continental
405 transport of particulate mass or gas-phase precursors. Lowest mass concentrations in the Aitken mode typically
406 occurred in the night hours (00:00 – 05:00) in a range of $0.3 - 0.5 \mu\text{g}/\text{m}^3$ in spring, summer, and winter, while in fall
407 mass loadings of $0.7 - 0.8 \mu\text{g}/\text{m}^3$ were reached. Diurnal changes were least pronounced in winter with largely constant
408 integrated Aitken mode particle concentrations. In the remaining seasons, varying degrees of daytime changes were
409 apparent with a general increase around 06:00, likely owing to citybound commuter traffic from surrounding roads to
410 the west of the sampling site at 1-2km of lateral distance. This also led to a modest increase in particle polydispersity
411 with a discernible widening of the Aitken mode size distributions (*black solid line, lowest panels in Figure 3*). Daily
412 maxima in spring, summer and fall were reached in the early evening ($\sim 21:00$) with marked differences in absolute
413 mass concentrations depending on the respective season, from a summer time low of $0.8 \mu\text{g}/\text{m}^3$ to a fall season high
414 of $1.4 \mu\text{g}/\text{m}^3$. Mass median diameters in the Aitken mode were smaller in the night hours and displayed subtle
415 increments during the day in the range of 10-20 nm reaching their maximum typically in the late afternoon, except for
416 the fall season when mass median diameters displayed very little variation with time of day.

417 Total particle mass in the accumulation mode in spring and summer reached minima during the night hours ($2 \mu\text{g}/\text{m}^3$)
418 and maxima ($3 \mu\text{g}/\text{m}^3$) around noon, remaining stable in the daylight hours thereafter. MMDs increased notably from
419 440nm at night to 510nm during the day in spring, while in summer a morning rise by $\sim 30\text{nm}$ from 530nm to 560nm
420 was obvious between 06:00 and 10:00 and coincided with the morning rush hour and the associated early morning

421 peak of NO_x and an otherwise stable mode diameter of 530nm for the rest of the day. In fall, the increase in
422 accumulation mode organic mass occurred much earlier, starting in the dark hours at 04:00, with a corresponding
423 trend also evident for nitrate but absent for sulfate, indicating a common source of these organic and nitrate enriched
424 particles. Nighttime MMDs for organics were generally larger (540nm) and decreased to a minimum of 510nm in the
425 early afternoon accompanied by a slight widening of the distribution. In winter, mass concentrations decreased
426 appreciably in the early morning hours and started to increase only beyond 10:00. In the colder seasons (fall, winter),
427 a similar concentration pattern was also observed for gas-phase SO₂ which is considered as a largely regional pollutant
428 with few distinct local sources (Yuan et al., 2013), indicating that changes in boundary layer and mixing with regional
429 background were likely the more dominant processes in winter.

430

431 Sulfate

432 Aitken mode sulfate mass concentrations peaked in the afternoon from spring throughout fall with maximum
433 concentrations reached progressively later in the afternoon (14:00 in spring; 16:00 in fall). Nominal concentrations
434 were highest in spring and summer (0.5-0.6 µg/m³), slightly lower in fall (0.4 µg/m³) and reached the lowest levels in
435 winter (0.1 µg/m³). In addition to the afternoon peak, a conspicuous early morning peak of similar magnitude was
436 evident in spring between 02:00 and 06:00. A greater proportion of southerly winds was evident in said time period
437 compared to the overall seasonal wind frequency distribution (Figure D13a in the Supporting Material) and may
438 indicate transport of sulfate from marine sources in the southern parts of Hong Kong. Diurnal variations in MMDs
439 and GSDs were generally small and without obvious regular trends. Nominal mass median diameters were
440 significantly lower in winter (~170nm) compared to spring and fall (~190nm) and summer (~210nm).

441 Trends in accumulation mode particle mass were more pronounced. In spring, a shallow concentration valley during
442 the late evening and night hours (20:00 to 03:00) with minimum concentrations of 5 µg/m³ was apparent, while
443 daytime concentrations stayed largely invariant at 6 µg/m³. The MMDs followed a similar variation with a minimum
444 mode diameter around 550nm in the early hours of the day and slightly larger daytime MMDs around 570nm. Nominal
445 concentrations were larger in summer with a nighttime valley concentration of 7 µg/m³ and a well-pronounced broad
446 day peak with a maximum of 9.5 µg/m³ in the early afternoon (14:00-15:00). A prior additional morning peak occurred
447 between 04:00 and 10:00 with particle mass concentrations reaching 8.5 µg/m³ related to a consistent north-easterly
448 morning wind pattern (Figure D13b in the Supporting Material) and likely associated with transport from north-
449 easterly coastal regions or nighttime fisheries related maritime traffic. The diurnal trend in mass median diameter was
450 similar to that in spring with a night minimum of 570nm and day maximum of 590nm.

451 In fall, accumulation mode characteristics showed no significant diurnal variability, with a largely stable integrated
452 particle mass of 6 µg/m³ and only subtle MMD changes (585nm at night; 575nm during the day). In winter, two
453 concentration dips with reductions by ~0.5 µg/m³ between 06:00 and 10:00 and between 18:00 and 22:00 were evident,
454 while MMDs increased during the day between 10:00 and 15:00 from 520nm, peaking at a size of 540nm.

455

456 Nitrate

457 Nitrate particle mass in the Aitken mode was generally small from spring throughout fall amounting to 0.01 - 0.06

458 $\mu\text{g}/\text{m}^3$. Winter time concentrations were larger in a range of 0.06 - 0.08 $\mu\text{g}/\text{m}^3$ during the day and 0.10 - 0.12 $\mu\text{g}/\text{m}^3$ in
459 the late evening hours. The latter evening peak centered around 21:00 was evident in most seasons (except spring)
460 and accounted for 12-23% (0.1-0.25 $\mu\text{g}/\text{m}^3$) of total daily Aitken mode nitrate mass burden. Similar to the urban
461 roadside location, these nighttime nitrate peaks coincided with the peak period of organic cooking aerosol
462 concentrations (Figure D14 in the Supporting Material), which were however significantly smaller at the suburban
463 measurement site and mainly attributed to the operation of an on-campus student canteen (Li et al., 2015). Trends in
464 mass median diameters varied between seasons with no discernible trend in winter, a subtle decreasing trend with time
465 of day in spring and broad daytime diameter increases in summer and fall. Solar irradiation in these two seasons was
466 comparatively high (Figure D10b-c in the Supporting Material) indicating that photochemical nitrate production in
467 the Aitken mode may have led to this observed growth in particle size.

468 Integrated particle mass concentrations in the accumulation mode only exhibited subtle variations from spring
469 throughout fall, with essentially constant diurnal concentrations in spring, a subtle daytime peak in summer which
470 accounted for ~ 15% of total daily accumulation mode nitrate (corresponding to 0.7 $\mu\text{g}/\text{m}^3$) and a conspicuous morning
471 peak between 04:00 and 10:00 in fall accounting for ~ 5% of total daily accumulation mode nitrate (corresponding to
472 0.5 $\mu\text{g}/\text{m}^3$). Clearer seasonal differences were evident in the trends of MMDs. In spring, MMDs decreased appreciably
473 over the late evening hours (21:00-0:00) with a concurrent widening of the size distribution (increase in GSD). In
474 summer, accumulation mode diameters decreased during the day by ~40nm with a similar trend in accumulation mode
475 organics. Winter time MMDs exhibited a more complex pattern with larger mode diameters in the early hours (04:00
476 – 10:00) and during the noon-time, and a late-afternoon dip leading to larger spread of intra-day mode diameters
477 ranging from 510nm to 570nm.

478 In comparison to the urban roadside measurements, diurnal particle size characteristics and mass concentrations in the
479 Aitken and accumulation mode were much more variable for all investigated species at the suburban HKUST site,
480 indicating that longer time scale processes and irregular events (transport patterns, local meteorology) were probably
481 more important in governing particle size distribution characteristics than diurnal processes.

482 **3.2. Day-to-day size distributions and seasonal averages**

483 To evaluate the evolution of particle size distributions within seasons, average species-specific size distributions were
484 generated by averaging raw distributions over 24h periods (between 0:00 and 23:59). There was clear long-term
485 variability in both resolved Aitken and accumulation mode MMDs and integrated submode particle mass
486 concentrations for all species (Figures D15-16 in the Supporting Material) and overall seasonal differences which
487 have been briefly addressed in the discussion of the diurnal size distribution variations between seasons. Figure 5
488 depicts the overall average values for all daily fitted MMDs and integrated particle mass concentrations in both the
489 Aitken and accumulation mode at the suburban HKUST and urban MK sites.

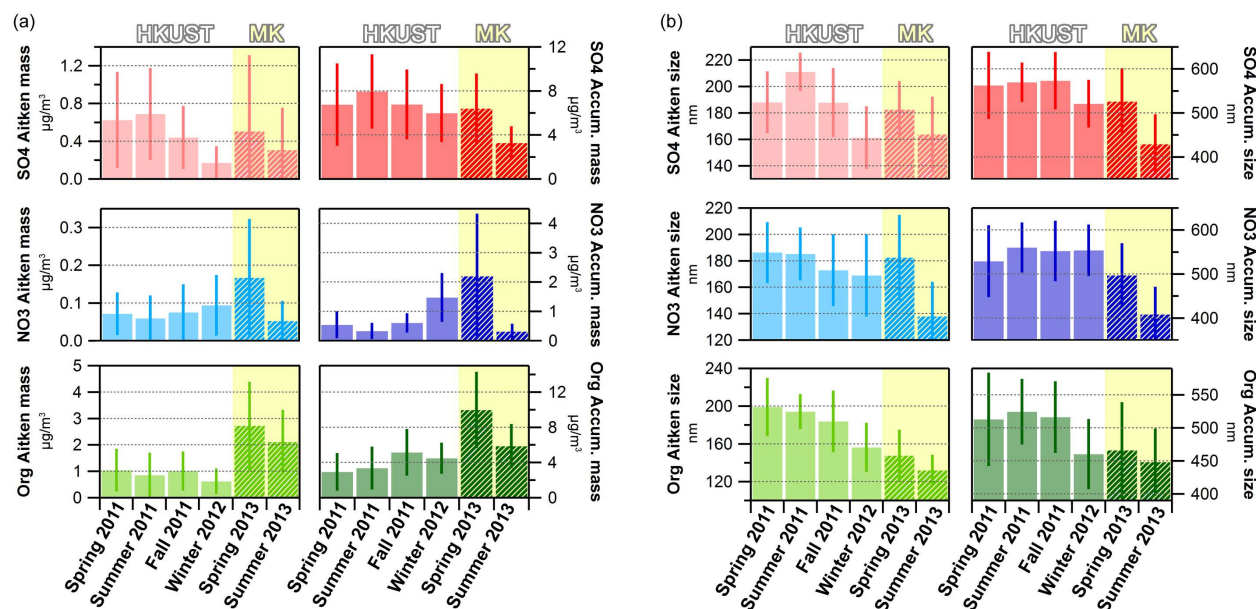
490

491

492 3.2.1. Seasonal trends

493 For the MK roadside station, particle mode diameters were generally larger in spring than in summer for all three
494 investigated species, but with clear differences in the magnitude of changes among individual species. In the Aitken
495 mode, organics and sulfate displayed a moderate decrease in mode diameter from spring to summer by 7-8% each,
496 while nitrate saw a more significant decrease by 25% from spring to summer. In contrast, accumulation mode MMDs
497 for organics exhibited only a subtle decrease by 5% and more substantial decreases for sulfate and nitrate by 20-22%
498 each. Total Aitken mode particle mass decreases varied strongly: -15% for organics, -36% for sulfate and -67% for
499 nitrate. In the accumulation mode, organics and sulfate exhibited similar relative decreases by 40-46%, while nitrate
500 particle mass reduced drastically by 85%.

501 At the suburban HKUST site, Aitken mode MMDs of nitrate and organics decreased with the progression of seasons
502 from spring to winter with highest mode diameters observed in spring and summer and appreciable decreases in winter
503 by -9% for nitrate and -25% for organics compared to the warmer seasons. Sulfate displayed a similar winter time
504 decrease in MMD (-15%) and an increase of similar magnitude in the summer season (+13%) compared to spring and
505 fall. Variations in sulfate and organic accumulation mode diameters were minor between spring and fall, while
506 wintertime MMDs were 7-12% lower. Nitrate exhibited an overall higher variability in mass median diameters in the
507 accumulation mode in spring (larger standard deviation) and with on average 10% lower MMDs compared to other
508 seasons. In line with the reduction in Aitken mode MMDs in winter, the integrated Aitken mode particle mass
509 decreased as well, by -16% for organics and almost -75% for sulfate, whereas nitrate contributions remained largely
510 stable throughout the seasons. Organic accumulation mode particle mass was significantly higher in the fall and winter
511 season by factors of 1.6 – 2. Diurnal variations in the degree of oxygenation were least pronounced in these seasons
512 (Li et al., 2015) suggesting that influence of transport in autumn and winter likely dominated over local formation,
513 thus exerting greater effects on particle mass in the larger size mode. Particulate nitrate concentrations were generally
514 low in the accumulation mode from spring through fall, but increased sharply in winter by factors of 3 – 4. Sulfate
515 accumulation mode mass concentrations remained more stable but saw significant summer time enhancements by
516 ~30% likely due to photochemical activity which also led to high concentrations of Ox and a higher degree of
517 oxygenation of organic aerosol among the four seasons (Li et al., 2015).



518
 519 **Figure 5.** Average and standard deviation of daily fit values of Aitken and accumulation mode particle mass and mass median
 520 diameters at the suburban HKUST site (*solid bars*) and urban MK site (*hashed bars*). The integrated particle mass is depicted in
 521 (a) for the Aitken mode (*left panels*) and for the accumulation mode (*right panels*) for sulfate, nitrate, and organics respectively.
 522 The mass median diameter is depicted in (b) for the Aitken mode (*left panels*) and for the accumulation mode (*right panels*) for
 523 sulfate, nitrate and organics respectively.

524
 525 Large particles contribute more to particle volume and hence particle mass. Correspondingly, the total submicron
 526 concentration of a given species is typically governed by changes in the accumulation mode particle mass and
 527 accordingly observed correlation values between integrated accumulation mode particle mass and individual NR-PM₁
 528 species mass concentrations were generally high ($R_{pr} > 0.90$) at both measurement sites (Figure D17 in the Supporting
 529 Material). This applied to both measurement sites regardless of the season. Aitken mode trends were less akin. At the
 530 urban roadside station, neither sulfate nor nitrate particle mass in the Aitken mode notably correlated with the
 531 respective total submicron species mass concentration in spring (all $R_{pr} \leq 0.20$), whereas in summer correlations were
 532 more significant with $R_{pr} = 0.51$ for sulfate and $R_{pr} = 0.80$ for nitrate. This signifies that periods of greater species mass
 533 concentrations were more likely to be caused by increases in both Aitken and accumulation mode particle mass
 534 indicating that particle formation and growth affecting smaller particles was more likely to occur in the warmer season.
 535 For organics, Aitken mode particle mass and submicron species mass correlated only weakly ($R_{pr} = 0.26$ in spring and
 536 $R_{pr} = 0.38$ in summer), i.e. each organic particle submode was governed by largely different dominant sources or
 537 formation processes in both seasons at the roadside.

538 At the suburban background site, Aitken mode particle mass for sulfate showed little correlation with total submicron
 539 sulfate concentration ($R_{pr} \leq 0.10$) apart from the spring season ($R_{pr} = 0.36$) where more frequent wet and foggy
 540 conditions may have facilitated sulfate formation in both size modes. For organics and nitrate significantly larger
 541 correlation coefficients of submode particle mass to total species concentration ($0.5 \leq R_{pr} \leq 0.7$) were observed in
 542 most seasons (spring, summer, winter) indicating significant influence of local or regional formation processes on

543 organic and nitrate Aitken mode particulate mass at the suburban receptor location. In the fall season, much weaker
544 correlations ($0.2 \leq R_{pr} \leq 0.4$) were likely caused by the dominance of continental air mass influence (Figure D12c in
545 the Supporting Material) and greater influence of aged accumulation mode particles on total submicron nitrate mass
546 concentrations.

547 **3.2.2. Inferred changes in mixing state**

548 Shifts in mixing state of ambient particles can be inferred from the inter-species analysis of mass median diameters.
549 Close nominal agreement (i.e. diameter ratios close to 1) infer that different species were distributed similarly across
550 the particle size range which thus most likely represents a largely internally mixed particle population, while the spread
551 of data (correlation coefficient) indicates the temporal homogeneity or divergence of resolved mode diameters. A
552 hypothetically perfectly internally mixed particle population over the whole sampling period would, therefore, yield
553 MMD ratios and Pearson's R values of 1 between species, while larger or smaller values are indicative of a greater
554 frequency of heterogeneous (i.e. more externally mixed) particle populations (Figure 6).

555 At the urban Mong Kok site, changes in accumulation mode mass median diameters for nitrate and sulfate followed
556 similar trends ($R_{pr} = 0.88-0.89$) and with diameter ratios close to 1 (0.94–0.95) Similarly, fitted accumulation mode
557 diameters of organic constituents predominantly followed that of sulfate in spring nominally (diameter ratio 0.88) and
558 temporally ($R_{pr} = 0.80$). The nominal agreement of organic and sulfate accumulation mode diameters persisted
559 (diameter ratio 1.03) overall in summer, however, there was significantly more temporal divergence ($R_{pr} = 0.65$)
560 indicating a greater frequency of time periods with external mixing of particle populations comprising different
561 fractions of organic constituents.

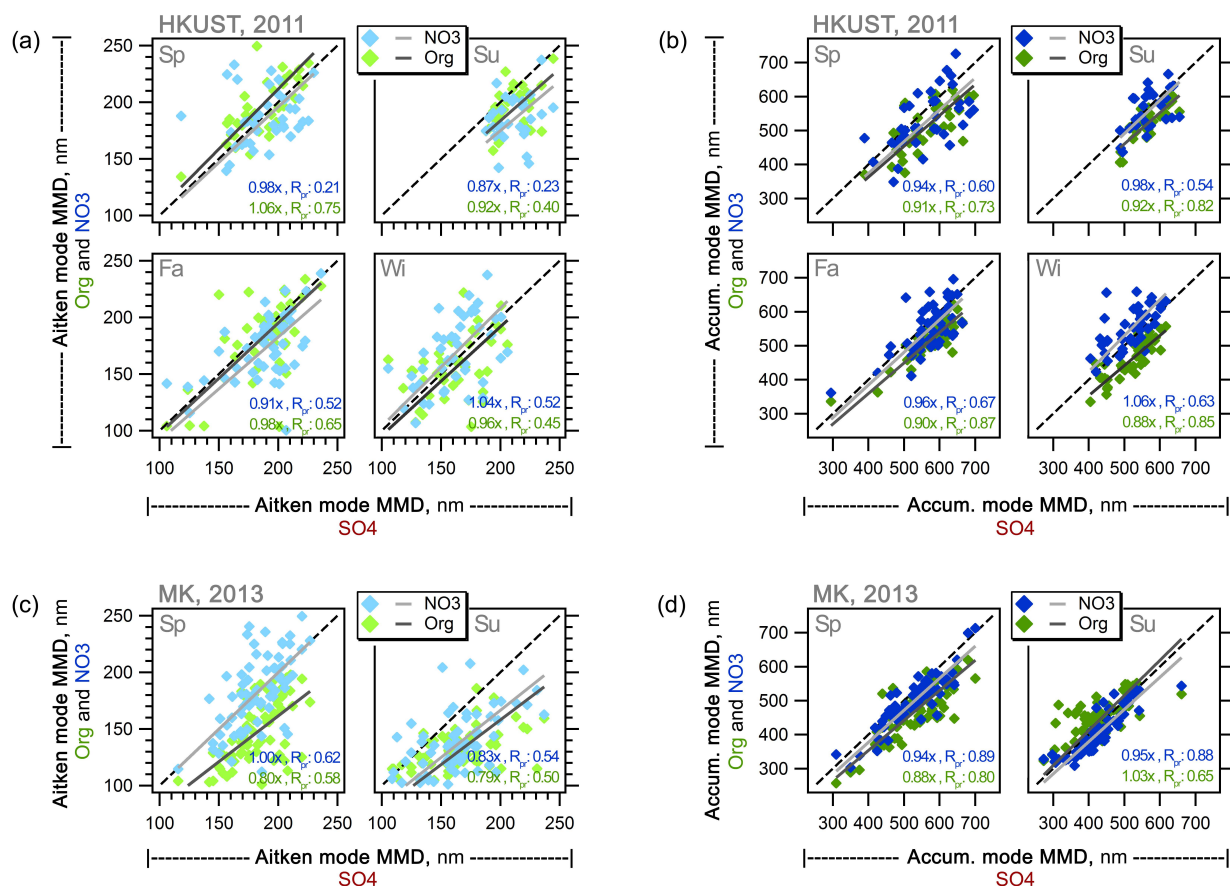
562 External mixing is more prevalent for freshly formed smaller particles which have typically undergone less
563 condensational growth, coagulation or aqueous-phase reactions. Indeed, the correlation coefficients of both nitrate
564 and organic Aitken mode MMDs with respect to sulfate were notably lower (0.50 and 0.62) indicating frequent periods
565 of particle populations with different species prevailing in different size regions within the Aitken mode.

566 Sulfate and nitrate were still more likely to occur internally mixed in the Aitken mode in spring with similar diameters
567 (nitrate to sulfate MMD ratio = 1.00), while organic Aitken mode MMDs were consistently lower, indicating greater
568 fractions of organic dominated particles towards the lower end and more inorganic dominated particles towards the
569 upper end of the fitted Aitken mode.

570 In summer, both nitrate and organic MMDs tended to be lower than those of sulfate (diameter ratios of 0.79 – 0.83)
571 but similar to each other, thus implying a shift to externally mixed populations of more nitrate and organic enhanced
572 and internally mixed smaller Aitken mode particles and sulfate dominated larger Aitken mode particles.

573 At the suburban HKUST site, accumulation mode MMDs of both nitrate and organics were generally quite similar to
574 those of sulfate with diameter ratios of 0.88 – 1.06. Compared to the urban site, correlation coefficients of nitrate and
575 sulfate were consistently lower (0.54 – 0.67) indicating a much greater frequency of time periods where sulfate and
576 nitrate dominated particles in the accumulation exhibited significantly different particle size distributions.

577 In winter, organic MMDs were consistently lower than those of sulfate and nitrate indicating a greater proportion of
 578 externally mixed particle populations with organics enriched particles in the lower accumulation size range and
 579 inorganic dominated particles in the larger accumulation size range. The least variability in particle size was observed
 580 in the summer season where MMDs in both Aitken and accumulation mode displayed variations in relatively narrow
 581 ranges between 200-250nm and 500-700nm, whereas in the remaining seasons time periods with particle populations
 582 of lower MMD were more frequent, extending to MMDs as low as 100nm in the Aitken mode and 300nm in the
 583 accumulation mode.
 584 In the Aitken mode, mass median diameters overall were quite similar across species, with diameter ratios of organic
 585 and nitrate distributions to those of sulfate in the range of 0.87 – 1.06, indicating that they generally covered a similar
 586 size range. The temporal agreement was highly variable with correlation coefficients (R_{pt}) spanning from 0.21 to 0.75
 587 indicating that Aitken mode particle populations at the suburban site were generally more diverse and likely influenced
 588 by a greater range of particle formation and growth mechanisms compared to the urban Mong Kok site.
 589
 590



591
 592 **Figure 6.** Scatter plots of fitted mass median diameters of organics and nitrate vs. sulfate for the (a) Aitken mode and (b)
 593 accumulation mode at the HKUST suburban site, and (c) Aitken mode and (d) accumulation mode at the urban Mong Kok site

594 3.3. Comparison to previous studies

595 Particle size distribution studies in Hong Kong are generally scarce and have focused on either size segregated filter
596 samples (MOUDI) for general ambient measurements or electrostatic classification in particle formation and particle
597 growth studies (Guo et al., 2012; Cheung et al., 2015). The latter studies focus on specific and narrow time periods
598 and lack general discussions on ambient particle size distributions.

599 Two ambient studies were undertaken at the suburban coastal HKUST site using size-segregated samples from a ten-
600 stage MOUDI sampler and offline chromatographic analysis. Inorganic constituents (NH_4 , NO_3 , SO_4) in fine particles
601 (i.e. $D_p < 1.8 \mu\text{m}$) were shown to follow bimodal distributions with mode diameters in the range of 0.14–0.21 μm and
602 0.46–0.58 μm in samples collected in the winter season, while the main mode was observed in the coarse region (4–6
603 μm) for all three species (Zhuang et al., 1999). A subsequent year-long observational study also reported bimodal fine
604 particle distributions with mode diameters of 0.1–0.3 μm and 0.7–0.9 μm and 1–2 additional modes in the coarse region
605 (Bian et al., 2014), however, the main mode in the size distributions of sulfate, ammonium, potassium and oxalate
606 was observed in the droplet mode (0.7 – 0.9 μm) in this study. Vehicle exhaust plumes sampled on-road from a Mobile
607 Real-time Air Monitoring Platform (MAP) across Hong Kong's road network exhibited three distinct particle volume
608 size distributions: a unimodal distribution with an accumulation mode at 0.2 μm and two bimodal distributions with a
609 minor mode at 0.2 μm and the dominant mode at 0.5 or 0.7 μm (Yao et al., 2007a).

610 The bimodality in the fine particle range across these studies is consistent with the AMS-based results in this work.
611 Nominally, the accumulation mode diameters from filter based studies and the chase studies are larger than those from
612 AMS measurements where maximum mode diameters occurred at $D_{va} \sim 700\text{nm}$, corresponding to $D_a \sim 470$ (assuming
613 $D_{va} \sim D_a * \text{density}$; particle density $\sim 1.5 \text{ g/cm}^3$). Direct comparability is however limited due to fundamental
614 differences in sizing techniques (MOUDI: atmospheric pressure; AMS: near-vacuum), sampling times (MOUDI: 24h
615 samples, scattered time line; AMS: minute raw resolution averaged to hourly or daily, continuous time line),
616 measurement uncertainties (MOUDI: sampling artifacts such as vapor adsorption and desorption; AMS: inlet lens
617 transmission) and aerosol pretreatment (none for MOUDI with potential impacts on particle size in high humidity
618 (>80%) conditions (Fang et al., 1991); AMS: removal of water prior to introduction to instrument).

619

620 3.4. Influence of AMS lens transmission

621 The quantitative measurement of particle components in the AMS is dependent on three major factors which may lead
622 to particle loss prior to detection (Huffman et al., 2005). Irregularly shaped particles deviating from the flight path in
623 the vacuum chamber may miss the vaporizer. Particles bouncing off the vaporizer surface will not be vaporized and
624 hence may not be detected. Lastly, the aerodynamic lens which is part of the instrument's inlet system does not
625 transmit particles uniformly across all particle diameters. Small particles are lost due to insufficient focusing or
626 diffusion and large particle impact the lens apertures (Liu et al., 2007; Williams et al., 2013). Being a function of
627 particle size, the latter factor affects both total AMS quantifiable particle mass (NR-PM_{10}) and measured mass size
628 distributions. Transmission curves determined for the standard lens, which is fitted in most AMS instruments, can

629 vary but typically show efficient (i.e. close to 100%) transmission in the range of 100-550nm (Knote et al., 2011)
630 falling of significantly at either edge.

631 We examined the potential impact of lens transmission on the AMS mass size distributions on a number of 24h size
632 distributions from the fall season HKUST dataset covering both efficient and reduced lens transmission size regimes
633 (accumulation mode diameters between 400 and 600nm). Panels a-c in Figure D18 in the Supporting Material depict
634 original and lens-transmission corrected 24h mass size distributions for organic, nitrate and sulfate, assuming the
635 transmission function (subpanel d) reported by Liu et al., 2007. Impacts were generally larger in the accumulation
636 mode range with evident shifts to larger mode diameters and larger mode mass concentrations observed in all size
637 distributions. In the small diameter range, enhanced shoulders can occur which may however be artifacts due to the
638 larger uncertainties (low signal to noise ratio at small particle mass), i.e. greater noisiness at the leading end of AMS
639 size distributions. For a quantitative comparison, bimodal fitting parameters from the corrected distributions were
640 plotted against those from, the original distributions in Figure D18 in the Supporting Material (subpanels e-g refer to
641 the Aitken mode and subpanels h-j to the accumulation mode). Leading edge shoulders in the corrected size
642 distributions were not considered in the fitting. Changes in the Aitken mode mass median diameters were minor (on
643 average ~3%), while the integrated mode particle mass increased moderately (~48%). In the accumulation mode, mass
644 median diameters increased by ~28% and integrated mode particle mass doubled. The distribution widths (geometric
645 standard deviations) exhibited little change in both modes (increases of 2-3%).

646 Fitting results will therefore vary depending on whether AMS size distribution and concentration data are corrected
647 for lens transmission. While explicit lens transmission corrections can improve the accuracy of quantification of AMS
648 species concentration and size distribution measurements, few ambient studies explicitly use lens transmission
649 corrections based on individual experimental determinations or literature values e.g. (Quinn et al., 2006; Cross et al.,
650 2007). Lens transmission curves can vary between instruments (Fast et al., 2009) and are inherently difficult to
651 determine accurately experimentally. As discussed previously, scaling of size distributions by lens transmission curves
652 may introduce artifacts in noisier size distributions (e.g. low end of the Aitken mode, low concentration periods, short
653 term size distribution averages). Trailing edges from slow vaporization (e.g. under high particle mass loadings) may
654 be exacerbated and inflate mass concentrations at the upper size cut range of the AMS. The majority of ambient
655 studies employs a combined correction factor (collection efficiency, CE) considered to be the joint product of the
656 previously mentioned transmission efficiencies related to particle bounce, beam broadening and lens transmission
657 (Middlebrook et al., 2012) derived from aerosol composition and by comparison to collocated speciation or particle
658 sizing instruments. As the AMS lens transmission curve could not be determined in this study and to avoid additional
659 uncertainties from the application of non-instrument specific lens transmission values, we followed the CE correction
660 method in the analysis of the size distribution data in this study. The reported values of resolved mode diameters and
661 integrated mode should therefore be regarded as lower bound estimates in the context of the instrumental limitations
662 affecting ambient AMS measurements.

663

664

665 4. Conclusion

666 A detailed analysis of AMS mass-based particle size distributions of sulfate, nitrate, and organics in submicron
667 particulate matter measured at two contrasting locations in Hong Kong during two field campaigns has been
668 undertaken. Deconvolution of size distributions into Aitken and accumulation submodes was accomplished by log-
669 normal peak fitting and trends in particle size (mass median diameters), dispersity (geometric standard deviation) and
670 overall particle mass (integrated mode area) were discussed on a diurnal time scale and on a daily basis to evaluate
671 longer-term changes in size distribution characteristics. At the urban roadside location, clear diurnal influences of
672 primary particle and gas-phase species were evident affecting both inorganic and organic component size distributions.
673 Traffic and cooking contributed an estimated $0.3 - 0.9 \mu\text{g}/\text{m}^3$ and $0.5 - 1.8 \mu\text{g}/\text{m}^3$ of organic component particle mass
674 in the Aitken mode, and $1.6 - 1.8 \mu\text{g}/\text{m}^3$ and $1.0 - 2.7 \mu\text{g}/\text{m}^3$ respectively in the accumulation mode with concentrations
675 level varying with seasons. Notable changes in Aitken mode mass median diameters of organics were limited to the
676 morning rush hour. Daytime particle concentration maxima of sulfate and nitrate in summer indicated substantial
677 influence of photochemical processes, which also led to increments in mass median diameters in the accumulation
678 mode thus inferring associated particle growth. Nocturnal nitrate formation was apparent in the accumulation mode
679 in spring concurring with the nighttime peak of ozone at the roadside, while in the Aitken mode nitrate particle
680 concentrations were significantly elevated during the dinner hours. Organics-related size distributions were mostly
681 governed by intra-day changes at the urban site with very similar trends across different size distribution sets (i.e.
682 concentration regimes), while disparities in diurnal variations among different size distribution sets were evident for
683 nitrate and sulfate, particularly affecting the average sets, indicating stronger influence of irregular external factors
684 which were not associated with diurnal time scale processes.

685 Suburban particle size distributions exhibited variable diurnal characteristics, suggesting that irregular processes such
686 as transport and seasonal meteorological conditions were the more dominant processes influencing particle size
687 characteristics. Aitken mode particle mass of organics was significantly larger in spring and summer indicating greater
688 influence of more local formation sources in the warm season. In the accumulation mode, organic particle mass
689 concentrations were highest in fall and lowest in spring, following the frequency pattern of continental air mass
690 influence. For sulfate, Aitken mode mass concentrations mass concentrations peaked in the afternoon from spring
691 throughout fall with highest nominal concentrations in spring and summer and lowest levels in winter, while
692 accumulation mode particle mass was highest in summer and fall and lowest in winter, similar to the trend observed
693 among organic constituents.

694 Nitrate particle mass in the Aitken mode was generally small in most seasons ($0.01 - 0.06 \mu\text{g}/\text{m}^3$), except winter where
695 daytime concentrations reached $\sim 0.1 \mu\text{g}/\text{m}^3$. In both modes, changes in mass median diameters varied temporally and
696 in magnitude with seasons, indicating a stronger influence of specific meteorological conditions on the properties of
697 nitrate-containing particles at the suburban site. At the urban site, periods of greater inorganic species mass
698 concentrations were more likely to be caused by increases in both Aitken and accumulation mode particle mass in
699 summer, indicating that particle formation and growth affecting smaller particles was more likely to occur in the
700 warmer season. At the suburban receptor location, significant correlation of submode particle mass to total species
701 concentration ($0.5 \leq R_{\text{pr}} \leq 0.7$) was observed for organics and nitrate in most seasons (spring, summer, winter)

702 suggesting notable influence of local or regional formation processes on organic and nitrate Aitken mode particulate
703 mass. Variations in particle mixing state were examined by evaluation of inter-species mass median diameter trends
704 at both measurement sites. In the accumulation mode at the urban site, internal mixing appeared to be prevalent in
705 spring, while greater frequency of time periods with external mixing of particle populations comprising different
706 fractions of organic constituents was observed in summer. External mixing was predominant in the Aitken mode at
707 the urban location in both seasons. At the suburban site, sulfate and nitrate in the accumulation mode more frequently
708 exhibited differing particle size distributions in all seasons signifying a greater extent of external mixing. In winter,
709 external mixing of more organics enriched particles in the lower accumulation size range was evident.

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713 **References**

- 714 Abbatt, J. P. D., Broekhuizen, K., and Pradeep Kumar, P.: Cloud condensation nucleus activity of internally mixed
715 ammonium sulfate/organic acid aerosol particles, *Atmospheric Environment*, 39, 4767-4778,
716 10.1016/j.atmosenv.2005.04.029, 2005.
- 717 Ahlquist, N. C., and Charlson, R. J.: A New Instrument for Evaluating the Visual Quality of Air, *Journal of the Air
718 Pollution Control Association*, 17, 467-469, 10.1080/00022470.1967.10469006, 1967.
- 719 Aiken, A. C., Salcedo, D., Cubison, M. J., Huffman, J. A., DeCarlo, P. F., Ulbrich, I. M., Docherty, K. S., Sueper,
720 D., Kimmel, J. R., Worsnop, D. R., Trimborn, A., Northway, M., Stone, E. A., Schauer, J. J., Volkamer, R. M.,
721 Fortner, E., de Foy, B., Wang, J., Laskin, A., Shutthanandan, V., Zheng, J., Zhang, R., Gaffney, J., Marley, N.
722 A., Paredes-Miranda, G., Arnott, W. P., Molina, L. T., Sosa, G., and Jimenez, J. L.: Mexico City aerosol
723 analysis during MILAGRO using high resolution aerosol mass spectrometry at the urban supersite (T0) - Part 1:
724 Fine particle composition and organic source apportionment, *Atmospheric Chemistry and Physics*, 9, 6633-
725 6653, 2009.
- 726 Bahreini, R., Dunlea, E. J., Matthew, B. M., Simons, C., Docherty, K. S., DeCarlo, P. F., Jimenez, J. L., Brock, C.
727 A., and Middlebrook, A. M.: Design and Operation of a Pressure-Controlled Inlet for Airborne Sampling with
728 an Aerodynamic Aerosol Lens, *Aerosol Science and Technology*, 42, 465-471, 10.1080/02786820802178514,
729 2008.
- 730 Bian, Q., Huang, X. H. H., and Yu, J. Z.: One-year observations of size distribution characteristics of major aerosol
731 constituents at a coastal receptor site in Hong Kong – Part 1: Inorganic ions and oxalate, *Atmos. Chem.
732 Phys.*, 14, 9013-9027, 10.5194/acp-14-9013-2014, 2014.
- 733 Bohren, C. F., and Huffman, D. R.: Absorption and scattering of light by small particles, in: *Absorption and
734 Scattering of Light by Small Particles*, Wiley-VCH Verlag GmbH, 1-11, 1983.

735 Canagaratna, M. R., Jayne, J. T., Jimenez, J. L., Allan, J. D., Alfarra, M. R., Zhang, Q., Onasch, T. B., Drewnick, F.,
736 Coe, H., Middlebrook, A., Delia, A., Williams, L. R., Trimborn, A. M., Northway, M. J., DeCarlo, P. F., Kolb,
737 C. E., Davidovits, P., and Worsnop, D. R.: Chemical and microphysical characterization of ambient aerosols
738 with the aerodyne aerosol mass spectrometer, *Mass Spectrometry Reviews*, 26, 185-222, 10.1002/mas.20115,
739 2007.

740 Charlson, R. J., Langner, J., Rodhe, H., Leovy, C. B., and Warren, S. G.: Perturbation of the northern hemisphere
741 radiative balance by backscattering from anthropogenic sulfate aerosols*, *Tellus B*, 43, 152-163,
742 10.1034/j.1600-0889.1991.t01-1-00013.x, 1991.

743 Cheung, K., Ling, Z. H., Wang, D. W., Wang, Y., Guo, H., Lee, B., Li, Y. J., and Chan, C. K.: Characterization and
744 source identification of sub-micron particles at the HKUST Supersite in Hong Kong, *Science of The Total
745 Environment*, 527-528, 287-296, <http://dx.doi.org/10.1016/j.scitotenv.2015.04.087>, 2015.

746 Crippa, M., DeCarlo, P. F., Slowik, J. G., Mohr, C., Heringa, M. F., Chirico, R., Poulain, L., Freutel, F., Sciare, J.,
747 Cozic, J., Di Marco, C. F., Elsasser, M., Nicolas, J. B., Marchand, N., Abidi, E., Wiedensohler, A., Drewnick,
748 F., Schneider, J., Borrmann, S., Nemitz, E., Zimmermann, R., Jaffrezo, J. L., Prévôt, A. S. H., and
749 Baltensperger, U.: Wintertime aerosol chemical composition and source apportionment of the organic fraction
750 in the metropolitan area of Paris, *Atmos. Chem. Phys.*, 13, 961-981, 10.5194/acp-13-961-2013, 2013.

751 Cross, E. S., Slowik, J. G., Davidovits, P., Allan, J. D., Worsnop, D. R., Jayne, J. T., Lewis †, D. K., Canagaratna,
752 M., and Onasch, T. B.: Laboratory and Ambient Particle Density Determinations using Light Scattering in
753 Conjunction with Aerosol Mass Spectrometry, *Aerosol Science and Technology*, 41, 343-359,
754 10.1080/02786820701199736, 2007.

755 Cross, E. S., Onasch, T. B., Canagaratna, M., Jayne, J. T., Kimmel, J., Yu, X. Y., Alexander, M. L., Worsnop, D. R.,
756 and Davidovits, P.: Single particle characterization using a light scattering module coupled to a time-of-flight
757 aerosol mass spectrometer, *Atmos. Chem. Phys.*, 9, 7769-7793, 10.5194/acp-9-7769-2009, 2009.

758 Davidson, C. I., Phalen, R. F., and Solomon, P. A.: Airborne Particulate Matter and Human Health: A Review,
759 *Aerosol Science and Technology*, 39, 737-749, 10.1080/02786820500191348, 2005.

760 DeCarlo, P. F., Slowik, J. G., Worsnop, D. R., Davidovits, P., and Jimenez, J. L.: Particle morphology and density
761 characterization by combined mobility and aerodynamic diameter measurements. Part 1: Theory, *Aerosol
762 Science and Technology*, 38, 1185-1205, 10.1080/027868290903907, 2004.

763 Docherty, K. S., Aiken, A. C., Huffman, J. A., Ulbrich, I. M., DeCarlo, P. F., Sueper, D., Worsnop, D. R., Snyder,
764 D. C., Peltier, R. E., Weber, R. J., Grover, B. D., Eatough, D. J., Williams, B. J., Goldstein, A. H., Ziemann, P.
765 J., and Jimenez, J. L.: The 2005 Study of Organic Aerosols at Riverside (SOAR-1): instrumental
766 intercomparisons and fine particle composition, *Atmospheric Chemistry and Physics*, 11, 12387-12420,
767 10.5194/acp-11-12387-2011, 2011.

768 Docherty, K. S., Lewandowski, M., and Jimenez, J. L.: Effect of Vaporizer Temperature on Ambient Non-
769 Refractory Submicron Aerosol Composition and Mass Spectra Measured by the Aerosol Mass Spectrometer,
770 *Aerosol Science and Technology*, 49, 485-494, 10.1080/02786826.2015.1042100, 2015.

771 Drewnick, F., Hings, S. S., DeCarlo, P., Jayne, J. T., Gonin, M., Fuhrer, K., Weimer, S., Jimenez, J. L., Demerjian,
772 K. L., Borrmann, S., and Worsnop, D. R.: A new time-of-flight aerosol mass spectrometer (TOF-AMS) -
773 Instrument description and first field deployment, *Aerosol Science and Technology*, 39, 637-658,
774 10.1080/02786820500182040, 2005.

775 Elser, M., Huang, R. J., Wolf, R., Slowik, J. G., Wang, Q., Canonaco, F., Li, G., Bozzetti, C., Daellenbach, K. R.,
776 Huang, Y., Zhang, R., Li, Z., Cao, J., Baltensperger, U., El-Haddad, I., and Prévôt, A. S. H.: New insights into
777 PM_{2.5} chemical composition and sources in two major cities in China during extreme haze events using aerosol
778 mass spectrometry, *Atmos. Chem. Phys.*, 16, 3207-3225, 10.5194/acp-16-3207-2016, 2016.

779 Farmer, D. K., Matsunaga, A., Docherty, K. S., Surratt, J. D., Seinfeld, J. H., Ziemann, P. J., and Jimenez, J. L.:
780 Response of an aerosol mass spectrometer to organonitrates and organosulfates and implications for
781 atmospheric chemistry, *Proceedings of the National Academy of Sciences of the United States of America*, 107,
782 6670-6675, 10.1073/pnas.0912340107, 2010.

783 Fast, J., Aiken, A. C., Allan, J., Alexander, L., Campos, T., Canagaratna, M. R., Chapman, E., DeCarlo, P. F., de
784 Foy, B., Gaffney, J., de Gouw, J., Doran, J. C., Emmons, L., Hodzic, A., Herndon, S. C., Huey, G., Jayne, J. T.,
785 Jimenez, J. L., Kleinman, L., Kuster, W., Marley, N., Russell, L., Ochoa, C., Onasch, T. B., Pekour, M., Song,
786 C., Ulbrich, I. M., Warneke, C., Welsh-Bon, D., Wiedinmyer, C., Worsnop, D. R., Yu, X. Y., and Zaveri, R.:
787 Evaluating simulated primary anthropogenic and biomass burning organic aerosols during MILAGRO:
788 implications for assessing treatments of secondary organic aerosols, *Atmospheric Chemistry and Physics*, 9,
789 6191-6215, 2009.

790 Gill, P. E., Murray, W., and Wright, M. H.: The Levenberg-Marquardt method, in: *Practical optimization*, Academic
791 Press, London, 1981.

792 Griffith, S. M., Huang, X. H. H., Louie, P. K. K., and Yu, J. Z.: Characterizing the thermodynamic and chemical
793 composition factors controlling PM_{2.5} nitrate: Insights gained from two years of online measurements in Hong
794 Kong, *Atmospheric Environment*, 122, 864-875, <http://dx.doi.org/10.1016/j.atmosenv.2015.02.009>, 2015.

795 Guo, H., Wang, D. W., Cheung, K., Ling, Z. H., Chan, C. K., and Yao, X. H.: Observation of aerosol size
796 distribution and new particle formation at a mountain site in subtropical Hong Kong, *Atmos. Chem. Phys.*, 12,
797 9923-9939, 10.5194/acp-12-9923-2012, 2012.

798 Huang, X. F., He, L. Y., Hu, M., Canagaratna, M. R., Sun, Y., Zhang, Q., Zhu, T., Xue, L., Zeng, L. W., Liu, X. G.,
799 Zhang, Y. H., Jayne, J. T., Ng, N. L., and Worsnop, D. R.: Highly time-resolved chemical characterization of
800 atmospheric submicron particles during 2008 Beijing Olympic Games using an Aerodyne High-Resolution
801 Aerosol Mass Spectrometer, *Atmospheric Chemistry and Physics*, 10, 8933-8945, 10.5194/acp-10-8933-2010,
802 2010.

803 Huang, X. F., He, L. Y., Hu, M., Canagaratna, M. R., Kroll, J. H., Ng, N. L., Zhang, Y. H., Lin, Y., Xue, L., Sun, T.
804 L., Liu, X. G., Shao, M., Jayne, J. T., and Worsnop, D. R.: Characterization of submicron aerosols at a rural site
805 in Pearl River Delta of China using an Aerodyne High-Resolution Aerosol Mass Spectrometer, *Atmospheric
806 Chemistry and Physics*, 11, 1865-1877, 10.5194/acp-11-1865-2011, 2011.

807 Huang, X. H. H., Bian, Q. J., Ng, W. M., Louie, P. K. K., and Yu, J. Z.: Characterization of PM_{2.5} Major
808 Components and Source Investigation in Suburban Hong Kong: A One Year Monitoring Study, *Aerosol and Air*
809 *Quality Research* 14, 237-250, 2014.

810 Huffman, J. A., Jayne, J. T., Drewnick, F., Aiken, A. C., Onasch, T., Worsnop, D. R., and Jimenez, J. L.: Design,
811 modeling, optimization, and experimental tests of a particle beam width probe for the aerodyne aerosol mass
812 spectrometer, *Aerosol Science and Technology*, 39, 1143-1163, 10.1080/02786820500423782, 2005.

813 Jayne, J. T., Leard, D. C., Zhang, X. F., Davidovits, P., Smith, K. A., Kolb, C. E., and Worsnop, D. R.:
814 Development of an aerosol mass spectrometer for size and composition analysis of submicron particles, *Aerosol*
815 *Science and Technology*, 33, 49-70, 2000.

816 Jimenez, J. L., Jayne, J. T., Shi, Q., Kolb, C. E., Worsnop, D. R., Yourshaw, I., Seinfeld, J. H., Flagan, R. C., Zhang,
817 X., Smith, K. A., Morris, J. W., and Davidovits, P.: Ambient aerosol sampling using the Aerodyne Aerosol
818 Mass Spectrometer, *J. Geophys. Res.*, 108, 8425, 10.1029/2001jd001213, 2003.

819 John, W.: Size Distribution Characteristics of Aerosols, in: *Aerosol Measurement*, John Wiley & Sons, Inc., 41-54,
820 2011.

821 Kerminen, V. M., Paramonov, M., Anttila, T., Riipinen, I., Fountoukis, C., Korhonen, H., Asmi, E., Laakso, L.,
822 Lihavainen, H., Swietlicki, E., Svenningsson, B., Asmi, A., Pandis, S. N., Kulmala, M., and Petäjä, T.: Cloud
823 condensation nuclei production associated with atmospheric nucleation: a synthesis based on existing literature
824 and new results, *Atmos. Chem. Phys.*, 12, 12037-12059, 10.5194/acp-12-12037-2012, 2012.

825 Knote, C., Brunner, D., Vogel, H., Allan, J., Asmi, A., Äijälä, M., Carbone, S., van der Gon, H. D., Jimenez, J. L.,
826 Kiendler-Scharr, A., Mohr, C., Poulain, L., Prévôt, A. S. H., Swietlicki, E., and Vogel, B.: Towards an online-
827 coupled chemistry-climate model: evaluation of trace gases and aerosols in COSMO-ART, *Geosci. Model Dev.*,
828 4, 1077-1102, 10.5194/gmd-4-1077-2011, 2011.

829 Köhler, H.: The nucleus in and the growth of hygroscopic droplets, *Transactions of the Faraday Society*, 32, 1152-
830 1161, 10.1039/tf9363201152, 1936.

831 Lee, B. P., Li, Y. J., Yu, J. Z., Louie, P. K. K., and Chan, C. K.: Physical and chemical characterization of ambient
832 aerosol by HR-ToF-AMS at a suburban site in Hong Kong during springtime 2011, *Journal of Geophysical*
833 *Research: Atmospheres*, 118, 8625-8639, 10.1002/jgrd.50658, 2013.

834 Lee, B. P., Li, Y. J., Yu, J. Z., Louie, P. K. K., and Chan, C. K.: Characteristics of submicron particulate matter at
835 the urban roadside in downtown Hong Kong—Overview of 4 months of continuous high-resolution aerosol
836 mass spectrometer measurements, *Journal of Geophysical Research: Atmospheres*, 120, 7040-7058,
837 10.1002/2015JD023311, 2015.

838 Li, Y. J., Lee, B. Y. L., Yu, J. Z., Ng, N. L., and Chan, C. K.: Evaluating the degree of oxygenation of organic
839 aerosol during foggy and hazy days in Hong Kong using high-resolution time-of-flight aerosol mass
840 spectrometry (HR-ToF-AMS), *Atmos. Chem. Phys.*, 13, 8739-8753, 10.5194/acp-13-8739-2013, 2013.

841 Li, Y. J., Lee, B. P., Su, L., Fung, J. C. H., and Chan, C. K.: Seasonal characteristics of fine particulate matter (PM)
842 based on high resolution time-of-flight aerosol mass spectrometric (HR-ToF-AMS) measurements at the
843 HKUST Supersite in Hong Kong, *Atmos. Chem. Phys.*, 15, 37-53, doi:10.5194/acp-15-37-2015, 2015.

844 Liu, P. S. K., Deng, R., Smith, K. A., Williams, L. R., Jayne, J. T., Canagaratna, M. R., Moore, K., Onasch, T. B.,
845 Worsnop, D. R., and Deshler, T.: Transmission efficiency of an aerodynamic focusing lens system: Comparison
846 of model calculations and laboratory measurements for the Aerodyne Aerosol Mass Spectrometer, *Aerosol*
847 *Science and Technology*, 41, 721-733, 10.1080/02786820701422278, 2007.

848 Man, H., Zhu, Y., Ji, F., Yao, X., Lau, N. T., Li, Y., Lee, B. P., and Chan, C. K.: Comparison of Daytime and
849 Nighttime New Particle Growth at the HKUST Supersite in Hong Kong, *Environmental Science & Technology*,
850 49, 7170-7178, 10.1021/acs.est.5b02143, 2015.

851 Meng, J. W., Yeung, M. C., Li, Y. J., Lee, B. Y. L., and Chan, C. K.: Size-resolved cloud condensation nuclei
852 (CCN) activity and closure analysis at the HKUST Supersite in Hong Kong, *Atmos. Chem. Phys.*, 14, 10267-
853 10282, 10.5194/acp-14-10267-2014, 2014.

854 Middlebrook, A. M., Bahreini, R., Jimenez, J. L., and Canagaratna, M. R.: Evaluation of Composition-Dependent
855 Collection Efficiencies for the Aerodyne Aerosol Mass Spectrometer using Field Data, *Aerosol Science and*
856 *Technology*, 46, 258-271, 10.1080/02786826.2011.620041, 2012.

857 Mohr, C., DeCarlo, P. F., Heringa, M. F., Chirico, R., Slowik, J. G., Richter, R., Reche, C., Alastuey, A., Querol, X.,
858 Seco, R., Peñuelas, J., Jiménez, J. L., Crippa, M., Zimmermann, R., Baltensperger, U., and Prévôt, A. S. H.:
859 Identification and quantification of organic aerosol from cooking and other sources in Barcelona using aerosol
860 mass spectrometer data, *Atmos. Chem. Phys.*, 12, 1649-1665, 10.5194/acp-12-1649-2012, 2012.

861 Quinn, P. K., Bates, T. S., Coffman, D., Onasch, T. B., Worsnop, D., Baynard, T., de Gouw, J. A., Goldan, P. D.,
862 Kuster, W. C., Williams, E., Roberts, J. M., Lerner, B., Stohl, A., Pettersson, A., and Lovejoy, E. R.: Impacts of
863 sources and aging on submicrometer aerosol properties in the marine boundary layer across the Gulf of Maine,
864 *J. Geophys. Res.*, 111, D23S36, 10.1029/2006jd007582, 2006.

865 Rupakheti, M., Leaitch, W. R., Lohmann, U., Hayden, K., Brickell, P., Lu, G., Li, S. M., Toom-Sauntry, D.,
866 Bottenheim, J. W., Brook, J. R., Vet, R., Jayne, J. T., and Worsnop, D. R.: An intensive study of the size and
867 composition of submicron atmospheric aerosols at a rural site in Ontario, Canada, *Aerosol Science and*
868 *Technology*, 39, 722-736, 10.1080/02786820500182420, 2005.

869 Saarikoski, S., Carbone, S., Decesari, S., Giulianelli, L., Angelini, F., Canagaratna, M., Ng, N. L., Trimborn, A.,
870 Facchini, M. C., Fuzzi, S., Hillamo, R., and Worsnop, D.: Chemical characterization of springtime
871 submicrometer aerosol in Po Valley, Italy, *Atmos. Chem. Phys.*, 12, 8401-8421, 10.5194/acp-12-8401-2012,
872 2012.

873 Salcedo, D., Onasch, T. B., Dzepina, K., Canagaratna, M. R., Zhang, Q., Huffman, J. A., DeCarlo, P. F., Jayne, J. T.,
874 Mortimer, P., Worsnop, D. R., Kolb, C. E., Johnson, K. S., Zuberi, B., Marr, L. C., Volkamer, R., Molina, L. T.,
875 Molina, M. J., Cardenas, B., Bernabe, R. M., Marquez, C., Gaffney, J. S., Marley, N. A., Laskin, A.,
876 Shutthanandan, V., Xie, Y., Brune, W., Leshner, R., Shirley, T., and Jimenez, J. L.: Characterization of ambient
877 aerosols in Mexico City during the MCMA-2003 campaign with Aerosol Mass Spectrometry: results from the
878 CENICA Supersite, *Atmospheric Chemistry and Physics*, 6, 925-946, 2006.

879 Schwartz, S. E.: The whitehouse effect—Shortwave radiative forcing of climate by anthropogenic aerosols: an
880 overview, *Journal of Aerosol Science*, 27, 359-382, [http://dx.doi.org/10.1016/0021-8502\(95\)00533-1](http://dx.doi.org/10.1016/0021-8502(95)00533-1), 1996.

881 Seinfeld, J. H., and Pandis, S. N.: Atmospheric Chemistry and Physics - From Air Pollution to Climate Change (2nd
882 Edition). John Wiley & Sons, 2006.

883 Setyan, A., Zhang, Q., Merkel, M., Knighton, W. B., Sun, Y., Song, C., Shilling, J. E., Onasch, T. B., Herndon, S.
884 C., Worsnop, D. R., Fast, J. D., Zaveri, R. A., Berg, L. K., Wiedensohler, A., Flowers, B. A., Dubey, M. K., and
885 Subramanian, R.: Characterization of submicron particles influenced by mixed biogenic and anthropogenic
886 emissions using high-resolution aerosol mass spectrometry: results from CARES, *Atmos. Chem. Phys.*, 12,
887 8131-8156, 10.5194/acp-12-8131-2012, 2012.

888 Slowik, J. G., Stanken, K., Davidovits, P., Williams, L. R., Jayne, J. T., Kolb, C. E., Worsnop, D. R., Rudich, Y.,
889 DeCarlo, P. F., and Jimenez, J. L.: Particle morphology and density characterization by combined mobility and
890 aerodynamic diameter measurements. Part 2: Application to combustion-generated soot aerosols as a function of
891 fuel equivalence ratio, *Aerosol Science and Technology*, 38, 1206-1222, 10.1080/027868290903916, 2004.

892 Sun, C., Lee, B. P., Huang, D., Jie Li, Y., Schurman, M. I., Louie, P. K. K., Luk, C., and Chan, C. K.: Continuous
893 measurements at the urban roadside in an Asian megacity by Aerosol Chemical Speciation Monitor (ACSM):
894 particulate matter characteristics during fall and winter seasons in Hong Kong, *Atmos. Chem. Phys.*, 16, 1713-
895 1728, 10.5194/acp-16-1713-2016, 2016.

896 Sun, Y., Zhang, Q., Macdonald, A. M., Hayden, K., Li, S. M., Liggio, J., Liu, P. S. K., Anlauf, K. G., Leaitch, W.
897 R., Steffen, A., Cubison, M., Worsnop, D. R., van Donkelaar, A., and Martin, R. V.: Size-resolved aerosol
898 chemistry on Whistler Mountain, Canada with a high-resolution aerosol mass spectrometer during INTEX-B,
899 *Atmos. Chem. Phys.*, 9, 3095-3111, 10.5194/acp-9-3095-2009, 2009.

900 Sun, Y. L., Zhang, Q., Schwab, J. J., Demerjian, K. L., Chen, W. N., Bae, M. S., Hung, H. M., Hogrefe, O., Frank,
901 B., Rattigan, O. V., and Lin, Y. C.: Characterization of the sources and processes of organic and inorganic
902 aerosols in New York city with a high-resolution time-of-flight aerosol mass spectrometer, *Atmospheric
903 Chemistry and Physics*, 11, 1581-1602, 10.5194/acp-11-1581-2011, 2011.

904 Takegawa, N., Miyakawa, T., Watanabe, M., Kondo, Y., Miyazaki, Y., Han, S., Zhao, Y., van Pinxteren, D.,
905 Bruggemann, E., Gnauk, T., Herrmann, H., Xiao, R., Deng, Z., Hu, M., Zhu, T., and Zhang, Y.: Performance of
906 an Aerodyne Aerosol Mass Spectrometer (AMS) during Intensive Campaigns in China in the Summer of 2006,
907 *Aerosol Science and Technology*, 43, 189-204, 10.1080/02786820802582251, 2009.

908 Ulbrich, I. M., Canagaratna, M. R., Cubison, M. J., Zhang, Q., Ng, N. L., Aiken, A. C., and Jimenez, J. L.: Three-
909 dimensional factorization of size-resolved organic aerosol mass spectra from Mexico City, *Atmos. Meas. Tech.*,
910 5, 195-224, 10.5194/amt-5-195-2012, 2012.

911 Westervelt, D. M., Pierce, J. R., Riipinen, I., Trivittayanurak, W., Hamed, A., Kulmala, M., Laaksonen, A., Decesari,
912 S., and Adams, P. J.: Formation and growth of nucleated particles into cloud condensation nuclei: model-
913 measurement comparison, *Atmos. Chem. Phys.*, 13, 7645-7663, 10.5194/acp-13-7645-2013, 2013.

914 Williams, L. R., Gonzalez, L. A., Peck, J., Trimborn, D., McInnis, J., Farrar, M. R., Moore, K. D., Jayne, J. T.,
915 Robinson, W. A., Lewis, D. K., Onasch, T. B., Canagaratna, M. R., Trimborn, A., Timko, M. T., Magoon, G.,
916 Deng, R., Tang, D., de la Rosa Blanco, E., Prévôt, A. S. H., Smith, K. A., and Worsnop, D. R.: Characterization

917 of an aerodynamic lens for transmitting particles greater than 1 micrometer in diameter into the Aerodyne
918 aerosol mass spectrometer, *Atmos. Meas. Tech.*, 6, 3271-3280, 10.5194/amt-6-3271-2013, 2013.

919 Yao, X., Lau, N. T., Chan, C. K., and Fang, M.: Size distributions and condensation growth of submicron particles
920 in on-road vehicle plumes in Hong Kong, *Atmospheric Environment*, 41, 3328-3338,
921 10.1016/j.atmosenv.2006.12.044, 2007a.

922 Yao, X., Ling, T. Y., Fang, M., and Chan, C. K.: Size dependence of in situ pH in submicron atmospheric particles
923 in Hong Kong, *Atmospheric Environment*, 41, 382-393, 10.1016/j.atmosenv.2006.07.037, 2007b.

924 Yuan, Z., Yadav, V., Turner, J. R., Louie, P. K. K., and Lau, A. K. H.: Long-term trends of ambient particulate
925 matter emission source contributions and the accountability of control strategies in Hong Kong over 1998–2008,
926 *Atmospheric Environment*, 76, 21-31, <http://dx.doi.org/10.1016/j.atmosenv.2012.09.026>, 2013.

927 Zhang, J. K., Sun, Y., Liu, Z. R., Ji, D. S., Hu, B., Liu, Q., and Wang, Y. S.: Characterization of submicron aerosols
928 during a month of serious pollution in Beijing, 2013, *Atmos. Chem. Phys.*, 14, 2887-2903, 10.5194/acp-14-
929 2887-2014, 2014.

930 Zhang, Q., Stanier, C. O., Canagaratna, M. R., Jayne, J. T., Worsnop, D. R., Pandis, S. N., and Jimenez, J. L.:
931 Insights into the chemistry of new particle formation and growth events in Pittsburgh based on aerosol mass
932 spectrometry, *Environmental Science & Technology*, 38, 4797-4809, 10.1021/es035417u, 2004.

933 Zhang, Q., Canagaratna, M. R., Jayne, J. T., Worsnop, D. R., and Jimenez, J. L.: Time- and size-resolved chemical
934 composition of submicron particles in Pittsburgh: Implications for aerosol sources and processes, *Journal of*
935 *Geophysical Research-Atmospheres*, 110, D07s09 Artn d07s09, 2005.

936 Zheng, M., Kester, D. R., Wang, F., Shi, X., and Guo, Z.: Size distribution of organic and inorganic species in Hong
937 Kong aerosols during the wet and dry seasons, *J. Geophys. Res.*, 113, D16303, 10.1029/2007jd009494, 2008.

938 Zhuang, H., Chan, C. K., Fang, M., and Wexler, A. S.: Size distributions of particulate sulfate, nitrate, and
939 ammonium at a coastal site in Hong Kong, *Atmospheric Environment*, 33, 843-853, 10.1016/s1352-
940 2310(98)00305-7, 1999.