Influence of anthropogenic emissions and boundary conditions on multi-model simulations of major air pollutants over Europe and North America in the framework of AQMEII3

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Abstract

In the framework of the third phase of the Air Quality Model Evaluation International Initiative (AQMEII3), and as contribution to the second phase of the Hemispheric Transport of Air Pollution (HTAP2) activities for Europe and North America, the impacts of a 20% decrease of global and regional anthropogenic emissions on surface air pollutant levels in 2010 are simulated by an international community of regional scale air quality modeling

groups, using different state-of-the-art chemistry and transport models (CTM). The emission perturbations at the global level, as well as over the HTAP2-defined regions of Europe, North America and East Asia are first simulated by the global Composition Integrated Forecasting System (C-IFS) model from European Centre for Medium-Range Weather Forecasts (ECMWF), which provides boundary conditions to the various regional CTMs participating in AQMEII3. On top of the perturbed boundary conditions, the regional CTMs used the same set of perturbed emissions within the regional domain for the different perturbation scenarios that introduce a 20% reduction of anthropogenic emissions globally as well as over the HTAP2-defined regions of Europe, North America and East Asia.

Results show that the largest impacts over both domains are simulated in response to the global emission perturbation, mainly due to the impact of domestic emissions reductions. The responses of NO_2, SO_2 and PM concentrations to a 20% percent anthropogenic emission reductions are almost linear (~20% decrease) within the global perturbation scenario with however, large differences in the geographical distribution of the effect. NO_2, CO and SO_2 levels are strongly affected over the emission hot spots. O_3 levels generally decrease in all scenarios by up to ~1% over Europe, with increases over the hot spot regions, in particular in the Benelux region, by an increase up to ~6% due to the reduced effect of NOx-titration. O_3 daily maximum of 8-hour running average decreases in all scenarios over Europe, by up to ~1%. Over the North American domain, the central-to-eastern part and the western coast of the U.S experience the largest response to emission perturbations. Similar but slightly smaller responses are found when domestic emissions are reduced. The impact of inter-continental transport is relatively small over both domains, however, still noticeable particularly close to the boundaries. The impact is noticeable up to a few percent, for the western parts of the North American domain in response to the emission reductions over East Asia. O_3 daily maximum of 8-hour running average decreases in all scenarios over North Europe by up to ~5%. Much larger reductions are calculated over North America compared to Europe.

In addition, values of the Response to Extra-Regional Emission Reductions (RERER) metric have been calculated in order to quantify the differences in the strengths of non-local source contributions to different species among the different models. We found large RERER values for O_3 (~0.8) over both Europe and North America, indicating a large contribution from non-local sources, while for other pollutants including particles, low RERER values reflect a predominant control by local sources. A distinct seasonal variation in the local vs. non-local contributions has been found for both O_3 and $PM_{2.5}$, particularly reflecting the spring-time long-range transport to both continents.

1. Introduction

Regional air quality modeling has considerably developed during recent decades, driven by increased concern regarding the impact of air pollution on human health and ecosystems. Numerous air quality models have been developed by research groups worldwide and are being widely used for developing and testing emission control policies. Regional atmospheric chemistry and transport models (CTMs) are widely used to assess the past, present and future levels of air pollutants from continental to regional scales. There are different sources of

uncertainties in models such as emissions, meteorology, boundary conditions and chemical schemes that should be taken into account when analyzing results. These uncertainties become more critical when these models are used for regulatory applications such as impacts of emission reductions. Multi-model ensembles can help in reducing this uncertainty and provide a better estimate of impacts under different scenarios (Solazzo et al., 2013; Galmarini et al., 2013; Kioutsoukis et al., 2017).

Numerous observational and modeling studies show that long-range transport of pollutants degrade air quality over remote continents (e.g., Wilkening et al., 2000; Holloway et al., 2003; Akimoto, 2003; Fiore et al., 2009). Although the influence of foreign emissions on continental scales is seen most frequently in the free troposphere, surface levels can also be affected, in particular over locations that generally receive clean air masses (e.g. Li et al., 2002). For example, dust storms and biomass burning can influence the tropospheric composition on a hemispheric scale (e.g., Husar et al., 2001; Jaffe et al., 2004). Reducing air pollution levels in surface air would improve public health as exposure to these atmospheric constituents aggravates respiratory illness and leads to premature mortality (World Health Organization, 2013; Im et al., 2017; Liang et al., 2017). However, attributing pollution to specific source regions is complicated due to the different processes influencing intercontinental transport and by a large hemispheric background and the dominance of local emissions in contributing to high levels of particular pollutants, such as ozone (O_3) (e.g. Fiore et al., 2009). Given these difficulties, estimates of source-receptor relationships rely heavily on models.

Stjern et al. (2016), using ten models participating in the second Hemispheric Transport of Air Pollution (HTAP2) activity, showed that a 20% reduction of global anthropogenic emissions, leads to significant changes regionally. They found that for North America (NA), black carbon emissions controls in East Asia are more important than domestic mitigation. In the framework of the HTAP2 activity, UN (2007) showed that a 20% reduction of North American NOx emissions leads to a 0.22 ppb decrease in O_3 levels over Europe (EU), while a 20% decrease in East Asian NOx emissions leads to a decrease of North American surface O_3 levels by 0.12 ppb. The impacts of these emissions changes on the O_3 levels in the source regions are much higher. The impact of lateral boundary conditions (LBC) on concentration fields simulated by regional-scale air quality models can also be quite significant (Jimenez et al., 2007; Mathur, 2008; Rudich et al., 2008; Song et al., 2008; Anderrson et al., 2015; Giordano et al., 2015, Hogrefe et al., 2017; Solazzo et al., 2017a). Recently, Giordano et al. (2015) showed that the regional models can be very sensitive to the boundary conditions provided by the global models. Tang et al. (2007) showed that the simulated surface levels over polluted areas are usually not as sensitive to the variation of LBCs, but are more sensitive to the magnitude of their background concentrations. Jonson et al. (2017), in the framework of the HTAP2 activity, showed that for ozone the contributions from the rest of the world is larger than the effects from European emissions alone, with the largest contributions from North America and East Asia. The majority of these studies that address impact of emissions on regional and inter-continental transport employ global models on coarse spatial resolution or focus on just a few species, such as O_3 or carbon monoxide (CO).

On the other hand, studies using regional chemistry and transport models at finer spatial resolutions mostly focus on sub-regional scales (e.g. Im and Kanakidou, 2012; Huszar et al., 2016). Therefore, studies addressing multi-pollutant, source-receptor relationships on inter-continental and regional scales can provide valuable information on the impact of domestic and foreign emissions on regional air pollution levels. Multi-model ensembles operating on fine spatial resolutions can increase accuracy and provide an estimate of uncertainty.

The Air Quality Model Evaluation International Initiative (AQMEII), coordinated jointly by European Commission, Joint Research Centre (EC-JRC) and the U.S. Environmental Protection Agency (EPA) has brought together regional chemistry and transport modelling groups from Europe and North America since 2008 (Rao et al., 2012; Solazzo et al., 2012a,b; Im et al., 2015 a,b). AQMEII is now running its third phase as a regional sub-project of the larger Hemispheric Transport of Air Pollution (HTAP), which in turn is a taskforce of Long Range Transport of Air Pollution program (LTRAP) of United Nations Economic Commission for Europe (UNECE) (Galmarini et al., 2017). The aim of the study is to assess the impact of global and HTAP2-defined regional anthropogenic emission reductions of 20% in Europe, North America and East Asia on major air pollutant levels over Europe and North America using a multi-model ensemble approach. The study will also investigate the local vs. non-local contributions to different air pollutant levels, adopting the Response to Extra-Regional Emission Reductions (RERER) metric developed by the HTAP2 community (Galmarini et al., 2017).

2. Materials and Methods

In the framework of the AQMEII3 project, twelve groups contributed to the simulation of the air pollution levels for 2010 in Europe (EU) and three groups for North America (NA) (Table 1 and Solazzo et al., 2017b). As seen in Table 1, different groups used same CTM models, such as the CMAQ and WRF-Chem model. The main differences among these models reside in the number of vertical levels, horizontal spacing, biogenic emissions, gas/aerosol modules in the models and the model releases (Table 1). For example, regarding groups that used the CMAQ model, UK1, DE1 and US3 calculated biogenic emissions using the BEIS (Biogenic Emission Inventory System version 3) model, while TR1, UK1 and UK2 calculated biogenic emissions through the Model of Emissions of Gases and Aerosols from Nature (MEGAN) (Guenther et al., 2012). Moreover, DE1 does not include the dust module, while the other CMAQ instances use the inline calculation (Appel et al., 2013), and TR1 uses the dust calculation previously calculated for AQMEII phase 2. Finally, all runs were carried out using CMAQ version 5.0.2, except for TR1, which is based on the 4.7.1 version. The gas-phase mechanisms and the aerosol models used by each group are also presented in Table 1. IT1 used the WRF-Chem model version 3.6, with a new chemistry that includes a better representation of the secondary organic aerosol mass in the simulation of direct and indirect aerosol effects (Tuccella et al., 2015). In addition, only direct effects were included in the IT1 simulation. ES1 model also used WRF-Chem, with different gas phase chemistry. More details of the model system are provided in the supplementary material in Im et al. (2018).

The emission inventories that are used in the second phase of AQMEII for Europe and North America (Im et al., 2015a,b) and extensively described in Pouliot et al. (2015) are also used in AQMEII3. For the EU, the 2009 anthropogenic emission inventory from the Monitoring Atmospheric Composition & Climate (MACC) was used. For the NA domain, the 2008 National Emissions Inventory was used with 2010-specific adjustments for major point sources, mobile sources and wildfires (Pouliot et al., 2015). The emissions were then treated with the SMOKE emissions processing system (Mason et al., 2012). The majority of the European groups used MACC emissions over Europe, while FI1 and FRES1 supplemented the MACC emissions with HTAP emissions over North Africa (Table 1). For NA, the temporal and vertical allocation of emissions vary between the groups that used the "SMOKE" files (DE1, US1, US3) and the gridded HTAP files (DK1), however the annual total mass are exactly the same. In order to guarantee consistency between the groups using the regional scale MACC or SMOKE emissions, and the groups using the HTAPv2.2 emissions, the regional scale emission inventories were embedded in the HTAPv2.2 inventory (Janssens-Maenhout et al., 2015;) to Galmarini et al., 2017). Overall, there was a high level of harmonization of emission inputs even if there were some differences in how they were adapted by each modeling group for their system. Chemical boundary conditions for both domains were provided by the European Center for Medium Range Weather Forecasts (ECMWF) Composition – Integrated Forecast System (C-IFS) model (Flemming et al., 2015)

2.1. Emission perturbations

The perturbation scenarios feature a reduction of 20% of the anthropogenic emissions globally and in HTAP-defined regions of Europe, North America and East Asia (Table 2). The choice of 20% was motivated by the consideration that the perturbation would be large enough to produce a sizeable impact (i.e. more than numerical noise) even at long distances while small enough to be in the near-linear atmospheric chemistry regime (Galmarini et al., 2017). The emission reductions are implemented in both the global C-IFS model that provides the boundary conditions to the participating regional models, as well as in the regional models. The regional models use the corresponding set of boundary conditions extracted from the C-IFS model. Among the fourteen groups that participated to the AQMEII3 base case simulations, twelve groups from Europe and two groups from North America simulated at least one of the three emission perturbation scenarios, shown in Table 1. Two of the European groups (DE1 and DK1) also simulated the base and the three perturbation scenarios for the North American domain.

- The global perturbation scenario (GLO) reduces the global anthropogenic emissions by 20%. This change has been implemented in the C-IFS global model that provides the boundary conditions to the regional models participating in the AQMEII ensemble. Therefore, the GLO scenario introduces a change in the boundary conditions as well as a 20% decrease in the anthropogenic emissions used by the regional models. Nine groups over the EU domain and four groups over the NA domain have simulated the GLO scenario.

- The North American perturbation scenario (NAM) reduces the anthropogenic emissions in North America by 20%. This change has been implemented in the C-IFS global model that provides the boundary conditions to the regional models used in the AQMEII ensemble. Therefore, the NAM scenario introduces a change in the boundary conditions while anthropogenic emissions remain unchanged for Europe, showing the impact of long-range transport of North American pollutants to Europe while for North America, the scenario introduces a 20% reduction of anthropogenic emissions in the HTAP-defined North American region, showing the contribution from the domestic anthropogenic emissions. Seven groups over the EU domain and three groups over the NA domain have simulated the NAM scenario.
- The European perturbation scenario (EUR) reduces the anthropogenic emissions in the HTAP-defined Europe domain by 20%. The EUR scenario introduces a change in the anthropogenic emissions over the EUR region in the CTMs, showing the contribution from the domestic anthropogenic emissions. Six groups have simulated the EUR scenario over the EU domain.
- The East Asian perturbation scenario (EAS) reduces the anthropogenic emissions in East Asia by 20%. Similar to the NAM scenario for the EU domain, the EAS scenario introduces a change in the boundary conditions while anthropogenic emissions remain unchanged in the regional models, showing the impact of long-range transport from East Asia on the NA concentrations. Four groups have simulated the EAS scenario over the NA domain.

In AQMEII, all participating groups were required to upload modelled hourly surface concentrations to the ENSEMBLE system at EC-JRC, at specified monitoring stations in EU and NA, as well as surface gridded data (Galmarini et al, 2012; Im et al., 2015a, b; Solazzo et al., 2017b). This study investigates the impacts of emission perturbations and boundary conditions on O_3, NO_2, CO, SO_2, PM_{10} and $PM_{2.5}$ levels over Europe and North America.

Differences between each perturbation scenario and the base case (C-IFS global and regional models run with baseline emissions) are calculated from the gridded hourly pollutant fields, which are then monthly and annually averaged in order to estimate the impact of the perturbation of the corresponding emission or boundary condition.

To estimate the contribution of foreign emission perturbations relative to the GLO perturbation, we have also calculated the RERER metric (Galmarini et al., 2017; Huang et al., 2017; Jason et al., 2017). For Europe, RERER is calculated using the differences between the GLO vs BASE as well as the differences between EUR vs. BASE simulations for Europe (Eq. 1) while for North America; RERER is calculated using the differences between the GLO vs BASE and NAM vs. BASE simulations (Eq. 2).

$$RERER_{EUR} = \frac{R_{GLO} - R_{EUR}}{R_{GLO}} \quad \text{Eq. 1}$$

$$RERER_{NAM} = \frac{R_{GLO} - R_{NAM}}{R_{GLO}} \quad \text{Eq. 2}$$

where R_{GLO} is the response of the concentration of a given species to global emission reduction, R_{EUR} is the response of a concentration of a species to the EUR perturbation for the European domain, and R_{NAM} is the response of a concentration of a specie to the NAM perturbation for the North American domain. Therefore, a subset of modelling groups that have conducted the three simulations (BASE, GLO and EUR/NAM for Europe and North America, respectively) have been used in the metric calculations (see Table 1). The higher the local response is, the smaller the RERER metric is. The RERER value can exceed the value 1 when emission reductions lead to increasing concentrations (e.g., O_3 titration by nitrogen monoxide, NO).

3. Results

3.1. Model Evaluation

The base case simulation of each model has been evaluated on a monthly-mean basis using available surface observations from Europe and North America. The observational data used in this study are the same as the dataset used in the second phase of AQMEII (Im et al., 2015a,b). The data were provided from the surface air quality monitoring stations operating in EU and NA. In EU, surface data were provided by the European Monitoring and Evaluation Programme (EMEP, 2003; http://www.emep.int/) and the European Air Quality Database (AirBase; http://acm.eionet. europa.eu/databases/airbase/). NA observational data were obtained from the NAtChem (Canadian National Atmospheric Chemistry) database and from the Analysis Facility operated by Environment Canada (http://www.ec.gc.ca/natchem/).

The model evaluation results for each model are presented in Fig. 1 and 2, and in Table 3, along with the results for the multi model (MM) mean and median values. The results show that the monthly variations of gaseous pollutants are well captured by all models with correlation coefficients (r) generally higher than 0.70. The biases in simulated O_3 levels are generally less than 10% with a few exceptions of up to -35%. The temporal variations of NO_2 levels are also well simulated (r>0.7), but exhibit much higher biases, with underestimations up to 75%. CO levels are underestimated by up to 45% while a majority of the models underestimated SO_2 levels by up to 68%. Few models overestimated SO_2 by up to 49%. PM_{10} and $PM_{2.5}$ levels are underestimated by 20% to 70%. Slightly higher biases are calculated for the PM_{10} levels.

The model biases can be attributed to meteorology, in particular wind speed and planetary boundary layer (PBL) height, as well as the aerosol mechanisms used in different models that can underestimate either the inorganic aerosols (e.g. IT2) or the secondary organic aerosols (e.g. DK1), leading to underestimations in simulated PM mass. As discussed in Solazzo et al. (2017), EU3 region that covers the central Europe including the Alps has the largest errors in terms of wind speed, mainly attributed to the diurnal component of the error, with some models having also large errors in the synoptic component. This region also represents the lowest correlation coefficients for all models. They further conclude that emissions nd their vertical distribution are the main source of model biases; in particular for the primary species such as CO and PM. Regarding O3, they found that the models have highest biases in the

large scale synoptic component while the diurnal variations are well-captured in general. A more comprehensive evaluation of the models is presented in Solazzo et al. (2017b), Galmarini et al. (2017) and Im et al. (2018).

C-IFS base case results have also been evaluated along with the regional CTMs, as presented in Fig. 1 and 2 and in Table 3. The seasonal variations for O_3, NO_2, CO and SO_2 are well captured with high correlation values of ~0.9. PM_{10} and $PM_{2.5}$ showed a different seasonal cycle than the observation by not reproducing the wintertime maxima (r=~-0.7). C-IFS model underestimates O_3 and CO by ~20% over Europe while NO_2 is slightly overestimated (*NMB*=7%). SO_2 is overestimated by ~10% over Europe, while PM_{10} and $PM_{2.5}$ levels are largely underestimated by ~60%, which can be attributed to the lack of secondary aerosol mechanism in the bulk C-IFS model. Over the North American domain, C-IFS well captures the seasonal variations of O_3, NO_2 and CO with correlation coefficients larger than 0.7, while the seasonal variation of SO_2 is not captured by the model (r=0.04). The seasonal variations of PM_{10} and $PM_{2.5}$ are also poorly captured (r<0.2). North American O_3 levels are slightly underestimated (*NMB*=-10%), while NO_2 and CO are overestimated by ~40% and 20%, respectively. SO_2 is overestimated by 35% while PM_{10} is largely underestimated by ~80 and $PM_{2.5}$ by ~40%. Over both Europe and North America, the wintertime PM levels are underestimated due to lack of secondary aerosols while the spring summer peaks are attributed to long range transport of desert dust from the Sahara, which affects mainly the South East of North America.

3.2. Perturbation Analyses

The annual mean relative differences of each perturbation scenario from the base case scenario, averaged over all stations, are provided in Table 4 (EU) and Table 5 (NA) for each modeling group, along with the results for the MM ensemble mean and median. The base case monthly mean time series for the participating groups are provided in Fig.1 and Fig. 2 for each pollutant, while Fig.3 and Fig. 4 shows the annual mean spatial distribution of the pollutants from the MM ensemble mean calculations over Europe and North America, respectively. As seen in the time series figures, there is a large spread among different groups, owing to the different models used and the different sets of anthropogenic emissions (Table 1). However, the temporal variation is consistent among all models, in particular for the gaseous species.

3.2.1. Impact of the global emission reduction scenario (GLO)

3.2.1.1. Europe

The monthly time series of the differences between the GLO and the BASE simulations for each pollutant are presented in Fig. 5. The annual differences are reported in Table 4. Regarding the primary gaseous pollutants, all models simulate the smallest differences during the summer months while the differences are largest in winter. For O_3, the simulated differences are positive in winter and negative in summer for all models except for DE1 that simulated a decrease in all months. Results suggest that wintertime O_3 over Europe is mainly controlled by anthropogenic emissions. For the other pollutants, results suggest that their

levels are mainly controlled by anthropogenic emission throughout the year. The annual difference is smallest for O_3, with a reduction of -0.34±1.23 ppb (-1.04±4.00%). The annual mean value of the O_3 daily maximum of 8-hour running average decreases by -0.53±1.50 ppb (-1.62±3.99%). NO_2 levels decreased by 0.97±0.45 ppb (19.34±1.59%) over Europe while CO levels decreased by 17.35±4.03 ppb (11.22±1.17%), SO_2 levels by 0.18±0.05 ppb (20.87±0.93%), PM_{10} by 2.38±0.68 μgm^{-3} (15.84±2.12%) and $PM_{2.5}$ by 2.02±0.52 μgm^{-3} (18.30±1.75%). Vivanco et al. (2017) found similar reductions regarding the deposition of sulfur and nitrogen species over Europe. Almost all models simulate an overall decrease of annual mean O_3 levels over EU (-0.94% to -4.65%), with the exception of TR1 that simulated an increase of 9.31%. Regarding other pollutants, all models simulate a decrease during the simulation period. In general, DE1 and TR1 model groups stand out for introducing the smallest and largest differences, particularly for O_3, NO_2, and PM.

The geographical distribution of the change in annual mean concentrations in the GLO scenario as simulated by the MM mean is presented in Fig. 6. Regarding O_3, most of Europe is characterized by decreased concentrations (Fig.6a). Over central Europe, where most of the primary emissions are located (e.g. NOx), O_3 levels slightly increase by ~2%. Emission hotspots, in particular the Benelux area stands out with largest increases (~6%) due to decreased NOx-titration effect, which can also be seen in Fig. 6b. In addition, O_3 levels over the northern parts of Germany and France, and southern UK are increasing in response to emission reductions. There is also a clear decrease in CO levels (Fig.6c), in particular over central Europe by up to ~16%. All primary species decrease over the whole domain, especially over the industrial hot spots such as in Poland, Po Valley and the Benelux area (Fig.6d). PM levels decrease throughout the domain by up to ~20% (Fig.6e and f).

3.2.1.2. North America

The seasonal variation of the impact of 20%-decreased global emissions on the North American pollutant levels are presented in Fig.7. All models simulated a small decrease of 3% to 5% (Table 5) in O_3 levels with the largest differences in spring to summer (Fig.7a). The mean response to the emission perturbation is estimated to be -1.39 ± 0.27 ppb (-3.52 ± 0.80%). The annual mean value of the O_3 daily maximum of 8-hour running average decreases by -1.93±0.14 ppb (-4.51±0.45%). All models simulated a largest NO_2 response in winter. Most models simulated a decrease of NO_2 levels while DK1 estimated an increase (Fig.7b). As shown is Table 5, the models simulated a NO_2 response of ~0.4 – 1.2 ppb (-17.8 ± 0.78%). Regarding CO, all models simulated very clear seasonal profile of the response to emission reductions, with maximum change in late winter/early spring and the minimum change in summer. Most models simulated a change around -15 to -25 ppb (~11%); with the exception of the DE1 model simulating a decrease of ~9 ppb (~7.9%). The MM mean response is calculated to be 19.2 ± 6.9 ppb (-11 ± 2.3%). The impact of the emission reduction on SO_2 levels was calculated to be -0.25 ppb to -0.48 ppb (-20.3 ± 0.2%).

The response of PM_{10} levels to the global emission reduction was calculated to be -2.4 ± 1.8 μgm^{-3} (-32.1 ± 26.6%) (Table 5). The largest relative change was calculated for DE1 (~63%). DK1 has almost a flat response around -1 μgm^{-3}, while DE1, which is overlapped with the

Median line, and US3 have maximum responses in early spring and mid-autumn, while they simulate a minimum response in winter and late spring. Regarding $PM_{2.5}$, the multi-model mean response was calculated to be -1.5 ± 0.9 μgm^{-3} (-17.2 ± 1.8%). DK1 (overlapped with the Median) and US3 simulated the minimum response in May (Fig.7f), while US3 has a slightly higher second minimum in September. This minimum is also simulated by DE1 as the minimum response. DE1 simulates the lowest response among the three models.

The spatial distributions of response of different pollutants to the GLO scenario are presented in Fig.8. O_3 levels are reduced over most of the domain (Fig.8a), with slight increases over the emission hotspots due to reduced effect of NOx-titration, as seen in Fig.8b, as well as decreased CO levels over the whole domain (Fig.8c). SO_2 levels are also decreased throughout the domain (Fig.8d), with the largest reductions over the Atlantic (attributable to reduction in shipping emissions). The western part of the continent is characterized by the lowest reductions. PM levels are reduced throughout the domain by up to 25% (Fig.8e and f), with the largest reductions over the eastern and central parts of the domain. A large decrease, more pronounced in the $PM_{2.5}$ response, can also be seen over California in the western coastal United States.

3.2.2. Impact of the North American emission reduction scenario (NAM)

3.2.2.1. Europe

NA emission reductions account for a reduction of European O_3 levels of -0. 22±0.07 ppb (-0.75±0.14%), with all models simulating a decrease of -0.51% to 0.86%, except for the ES1 model that simulated an increase of 1.31% (Table 4). This decrease is in agreement with previous studies, such as the HTAP2 study (UN, 2017) that calculated an O_3 reduction over Europe of 0.22 ppb in response to a 20% decrease in the North American NOx emissions, and Fiore et al. (2009) that simulated a MM mean response of -0.4 ppb in response to a 20% reduction of anthropogenic emissions in North America. NO_2 levels increase slightly by 0.16±0.01%. The annual mean value of the O_3 daily maximum of 8-hour running average decreases by -0.15±0.27 ppb (-0.45±0.77%). CO levels also decreased over the EU domain by -1.39±0.27 ppb (-0.96±0.22%), much higher than ~0.1 ppb calculated by Fiore et al. (2009). PM_{10} and $PM_{2.5}$ levels also decreased slightly by -0.03±0.03 μgm^{-3} (-0.21±0.7%) and -0.02± 0.02 μgm^{-3} (-0.18±0.25%), respectively. The models had different SO_2 responses to the NA emissions. Overall, DE1, ES1 and FRES1 simulated almost no change in the surface SO_2 levels while DK1, ES1 and TR1 simulated an increase (0.10%, 5.75% and 0.01%, respectively) and FI1 and UK1 simulated a decrease (-0.02% and -0.03%, respectively). Different responses can be due to different model setups including aqueous chemistry, vertical resolutions and aerosol modules (Solazzo et al., 2017).

All models were consistent in simulating the largest impact on O_3 during spring and a second lower peak in autumn (Fig.9a). Surface mean NO_2 concentrations (Fig.9b) increased in most models except for FRES1 that simulated a small decrease except for winter. FI1 also simulated a decrease during the winter period extending to the transition periods. All models, except for ES1, simulated a similar response of CO concentrations to perturbation to NA

emissions, with a distinct seasonality (Fig.9c). The SO_2 response in models is also consistent except for the winter period where there is a large spread in magnitude and the sign of the response (Fig.9d).

O_3 levels decreased slightly over the entire European domain by up to 3% (Fig.10a). The largest impact is simulated over the western boundary and gradually decreases eastwards. The response of O_3 levels to NAM emissions is more evident during spring where there is a clear transport from Atlantic to the western/northwestern parts of Europe such as the U.K, northern France and Scandinavia (Fig. S2a). The transport of Atlantic air masses is also shown for the springtime CO levels over Europe (Fig. S2a). The ensemble mean simulates a slight increase of up to 3% in NO_2 levels over Europe (Fig.10b). Along with the O_3 levels, CO levels show the largest decrease over northwestern Europe by up to ~2%. SO_2 levels increased over the whole domain, in particular over Eastern Europe and the Alpine region (Fig.10d), due to a decrease in the oxidative capacity of the atmosphere (see Fig.10a for O_3), leading to a decrease in the SO_2 to SO_4 conversion. This results in an increase of the SO_2 levels and a decrease in the $PM_{2.5}$ levels (Fig.10e and f).

3.2.2.2. North America

The response of North American pollutant levels to a 20% reduction of North American anthropogenic emissions (implemented in both C-IFS and the regional CTMs) are presented in Table 5. The NAM scenario led to a decrease of annual mean O_3 levels over North America by -0.36 ppb (US3) to -0.92 ppb (DE1), with *MM* ensemble mean calculated to be -0.65±0.28 ppb (-1.45±0.88%), in agreement with Fiore et al. (2009) that calculated a decrease of ~1 ppb. The annual mean value of the O_3 daily maximum of 8-hour running average decreases by -1.11±0.11 ppb (-2.60±0.36%), very similar to the change over Europe. Consequently, the largest change in NO_2 levels were simulated by US3 (-1.17 ppb) and smallest by DE1 (-0.36 ppb). The MM mean response of NO_2 is calculated to be -0.71±0.41 ppb (-17.24±0.58%). Similar to NO_2, the largest response in CO levels were simulated by US3 (-19.87 ppb) and the smallest by DE1 (-3.84 ppb), leading to a MM mean response of -12.35±8.06 ppb (-7.01±3.60%). As seen in Table 5, DE1 simulated a much lower absolute and relative change in CO response compared to DK1 and US3. SO_2 levels decreased by -0.12 ppb to -0.18 ppb, leading to a MM mean response of -0.14±0.08 ppb (-4.23±0.18%). PM_{10} levels decreased -1.78±2.08 μgm^{-3} (-15.78±3.26%). As seen in Table 5, DK1, simulated a very low response to the NAM scenario, by ~0.60 μgm^{-3}, compared to the DE1 and the US3 groups that simulated a PM_{10} response of -2.02 μgm^{-3} and -4.19 μgm^{-3}, respectively. However, the relative responses are not very different between the different groups (~16%).

The response of O_3 to the NAM scenario is largest in summer (Fig.11a): June for DK1 and US3 and August for DE1. The O_3 response clearly shows a difference from the GLO response in spring, suggesting the impact of long-range transport in spring that does not appear in the perturbation of the local emissions only. The largest NO_2 response (Fig.11b) is simulated by US3, similar to the response to the GLO scenario. The response of CO to the reductions in local emissions (Fig.11c) is different from the response to the global reduction, where DK1 and US3 has the minimum response in spring and DE1 has the minimum

response in autumn. The response of SO_2 and PM to GLO and NAM are similar, suggesting the main drivers of SO_2 and PM levels are local emissions.

Annual mean O_3 levels show large reductions (~20%) over the eastern parts of the domain, while there are slight increases or less pronounced decreases over the western parts of the domain (Fig.12a), associated with larger NOx reductions (Fig.12b). CO and SO_2 levels are mostly reduced over the central to eastern parts of the domain (Fig.12c and d, respectively), with shipping impacts over the Atlantic being more pronounced on SO_2 levels. The western parts of the U.S. experiences smaller SO_2 reductions (~5-10%) and slight increases over the southwestern U.S. The response of PM to the NAM scenario (Fig.12e and f) is very similar to the response to the GLO scenario (Fig.8e and f).

3.2.3. Impact of the European emission reduction scenario (EUR)

O_3 levels increase slightly by 0.01±0.40 ppb (0.25±1.35%) in response to the 20% reduction of the anthropogenic emissions from Europe (Table 4). This response is much lower than Fiore et al. (2009) that calculated a MM mean response of 0.8 ppb. However, as seen in Fig.13a, the positive mean response together with the large standard deviation is due to the DE1 model that simulated a decrease (-2.33%), while other groups simulated an increase (0.39% to 1.72%). There is a distinct seasonality in the response with winter levels increasing with reduced emissions and summer levels decreasing, following the emission temporal variability. The annual mean value of the O_3 daily maximum of 8-hour running average decreases by -0.21±0.10 ppb (-0.62 0.24%). NO_2 concentrations decreased by -0.75±0.26 (17.68±0.90%), with a similar seasonal response of SO_2 levels (-17.52±1.70%) and CO levels (-6.26±1.07%), consistent with the findings of Vivanco et al. (2017). An opposite seasonal variation is calculated for the O_3 response (Fig. 13.b-d)., The DE1 model also stands out in the NO_2 response together with the FRES1 model in the magnitude of the response (Fig.13b). PM_{10} and $PM_{2.5}$ levels have similar responses to the emissions reduction (-14.43±2.84% and -15.67±2.12%, respectively) with similar seasonality.

The MM mean geographical distribution of the O_3 response is very similar with that of the GLO perturbation (Fig.14a), with relatively smaller decreases by up to ~3%. O_3 levels increase over the central and in particular over northwestern Europe by up to ~6%. NO_2 levels decrease uniformly over the entire domain by up to ~20% (Fig.14b). CO levels decrease over the emission sources, mainly over central and Eastern Europe (Fig.14c). PM levels also decrease over the entire domain, especially over central and Eastern Europe (Fig.14e and f).

3.2.4. Impact of the East Asian emission reduction scenario (EAS)

As seen in Table 5, the impacts of East Asian emissions on North American O_3 levels are much lower than the impacts from the reductions in global and local emissions. The largest impact is simulated by DE1 as -0.99 ppb (-0.35%), while other models give similar responses (~0.60 ppb; -0.20%). The O_3 response as calculated by the MM mean ensemble is -0.25±0.07 ppb, in agreement with the HTAP2 findings and Fiore et al. (2009). The annual mean value of the O_3 daily maximum of 8-hour running average decreases by -0.28±0.07 ppb (-

0.65±0.20%). NO_2 and SO_2 response to reductions in EAS emissions were simulated to be very small (-0.04±0.08% and 0.01±0.02%, respectively). The CO response to EAS was simulated to be -2.60 ppb (DE1) to -4.16 ppb (DK1), with the MM mean response of -3.37±0.68 ppb (-2±0.29%). Regarding PM_{10}, DE1 simulated a very large response (~-0.56 μgm^{-3}) compared to DK1 and US3 (~-0.05 μgm^{-3}), leading to a MM mean response of -0.21±0.30 μgm^{-3} (-5.63±8.50%). However, the $PM_{2.5}$ response was much lower (-0.02±0.03 μgm^{-3}; -0.20±0.35%), suggesting that the $PM_{2.5}$ levels are largely driven by local emissions.

The O_3 response to EAS emission reductions was highest in spring and autumn, suggesting that long-range transport is important in these seasons (Fig.15a). The NO_2 response was negative, being maximum in winter and minimum in summer, except for DK1 showing an increase in NO_2 levels in all seasons (Fig.15b). The impact of EAS emissions on North American CO levels showed a distinct seasonality (Fig.15c), similar to the impact of the global emission reductions (Fig.5c), suggesting that regional CO levels over North America are driven by both local emissions and long-range transport. The response of SO_2 to East Asian emission reductions varied largely from model to model with US3 showing an overall reduction while DE1 and DK1 simulated increases in winter, spring, and autumn, and decreases in summer (Fig.15d). The PM_{10} response simulated by DK1 (overlapped with the median) and US3 were simulated to be small, being largest in spring (Fig.15e). However, DE1 simulated a large and opposite response, with spring having the smallest response and winter with the largest response. DE1 also simulated a different $PM_{2.5}$ response in terms of the sign of the change and thus, seasonality in response to DK1 and US3 (Fig.15f). Largest differences were simulated in spring, similar to PM_{10} by DK1 and US3, while DE1 simulated the largest response in winter and summer and the spring response was minimum.

The impact of the East Asian emissions over the western parts of North America is clearly seen for all pollutants in Fig.16. The impacts are low for all pollutants, being up to 5%. The impacts are particularly pronounced for CO (Fig.16c), SO_2 (Fig.16d) and PM (Fig.16e and f). The largest O_3 response was simulated over the northwestern parts of North America (Fig.16a). The springtime transport of O3 from East Asia is more evident compared to the annual average of the perturbation response (Fig. S3a), where the western NA O_3 levels decrease by up to ~1.5%. The springtime CO levels also decrease by up to 6% (Fig. S3b), showing the importance of long-range transport from East Asia.

3.2.5. RERER analyses

As discussed in Section 2, the RERER metric (Galmarini et al., 2017; Hang et al., 2017; Jason et al., 2017) is designed to quantify the relative impact of local vs. non-local emission sources on pollutant levels in the receptor regions EU and NA., The RERER metrics for the EU have been calculated using gridded annual mean pollutant concentrations from the BASE, GLO and EUR simulations for the individual groups as well as for the ensemble mean. For the NA domain. The RERER metrics have been calculated using the annual mean concentrations from the BASE, GLO and NAM simulations. Table 6 presents the RERER metric calculated for the European domain. The table shows differences in the strengths of non-local source contributions to different species among the different models. Regarding the

RERER metric for O_3 in Europe, most values calculated are below one, except for the IT1 model, which shows a significant increase of O_3 levels in Europe in response to emission reductions compared with the other models. A RERER value of 0.8-0.9 is calculated for the majority of models, implying the dominance of non-local sources in Europe, except for the DE1 model, where the RERER value is lower (~0.5), giving an equal contribution of local vs. non-local sources in Europe. The MM mean RERER value for O_3 is ~0.8, showing a much larger contribution of non-local sources compared to local sources in Europe. This result is in agreement with, however slightly smaller, Jonson et al. (2017) that calculated a MM mean RERER value of 0.89.

Regarding NO_2, the RERER metrics (< 0.4) show that NO_2 is controlled by local sources. In addition, the RERER metrics calculated for DE1 and FI1 are slightly negative, implying that the signal is not sensitive to non-local emissions. RERER calculated for the ensemble mean for NO_2 (~0.2) also shows the high sensitivity of NO_2 concentrations to local sources. The RERER metric calculations for CO shows similar contributions from local vs. non-local sources, with RERER values of 0.4-06, except for IT1. IT1 has a RERER metric value of ~0.9 suggesting a large contribution of non-local sources, leading to the higher sensitivity of CO to non-local sources compared to other model groups. The RERER values calculated for the ensemble mean (~0.6) shows a slightly larger contribution of non-local sources compared to local sources. The MM mean RERER value of 0.55 for CO from this study is in very good agreement with Jonson et al. (2017) that calculated a MM mean RERER of 0.51. RERER metrics calculated for SO_2 are also in the low range (0-0.4). While DE1 and FI1 show almost no signal for the non-local contribution, DK1, IT1 and UK1 are in the higher end of the range. The CO MM mean RERER value of ~0.3 shows that CO levels are largely controlled by local emissions. Finally, the metrics calculated for PM_{10} and $PM_{2.5}$ shows that local sources are the main contributor to the PM levels in Europe (RERER = ~0 - 0.3), leading to an ensemble mean contribution of local sources (RERER = ~0.2).

Regarding the local vs. non-local contributions to different pollutants over the North American domain, three groups out of four simulated the GLO and NAM scenarios needed to calculate the RERER metrics. RERER metrics show that O_3 is largely controlled by non-local sources. European model groups DE1 and DK1 simulate a larger influence of non-local sources (~0.8 - ~0.9) compared to the US3 group, which simulated lower RERER metric values of ~0.5, indicating that O_3 levels are driven equally by local and non-local sources. This lower value is also consistent with the findings of Huang et al. (2017), who simulated the largest impacts on O_3 in May and June with RERER values around ~0.5. The ensemble mean shows that O_3 responses are largely attributable to non-local sources (RERER = ~0.8), which are similar to those found for Europe. RERER metric values calculated for NO_2 by different models (RERER = ~0 – ~0.2) and the ensemble mean (RERER = 0.05) clearly shows that NO_2 is controlled by local sources, similar to the Europe case. The sensitivity of CO to local and non-local sources are similar to those for O_3, with DE1 and DK1 simulating a large contribution from non-local sources while US1 shows that CO is controlled equally by local and non-local sources (RERER = 0.5). Similar to NO_2, all models show that SO_2 is largely driven by local sources with RERER values between ~0.1 and ~0.2. Regarding the

particles, models simulate very similar responses to changes in the local and non-local sources. RERER values are calculated to be ~0.08 and ~0.11 for PM_{10} and $PM_{2.5}$, respectively, showing the large local contribution compared to non-local sources.

Fig. 17 shows the spatial distributions of the MMM RERER values for O_3 and $PM_{2.5}$, as constructed from the annual mean responses to perturbation scenarios over Europe and North America. Fig. 17a shows that O_3 is dominantly controlled by non-local sources with RERER values higher than 0.5 throughout the domain. Higher values are calculated over the north western Europe, in particular over UK and the north western part of the domain covering the Atlantic. In contrary, $PM_{2.5}$ levels are controlled by local sources with RERER values around 0.2 (Fig. 17b). North American O_3 levels are largely controlled by non-local sources over the western part of the domain, with RERER values above 0.5 (Fig. 17c). Local sources play a more important role in controlling O_3 levels over the eastern part of the U.S. where much lower RERER values are calculated. $PM_{2.5}$ levels are dominantly controlled by the local sources, similar to the case in Europe, with low RERER values throughout the domain (Fig. 17d). $PM_{2.5}$ levels over the western part of the domain has however a relatively larger contribution from non-local sources. It is important to note that the sharp gradients in the $PM_{2.5}$ RERER values over both the eastern part of the Europe domain and the Mexican part of the NA domain is due to HTAP2-definition of source regions where the perturbations are introduced. Therefore, due to the large contribution of the local sources to $PM_{2.5}$ levels, large gradients are calculated across the HTAP2 borders. As O_3 is largely controlled by non-local sources, these gradients do not exist.

In order to further analyze the impact of local vs non-local sources, the monthly variations of RERER values for O_3 and $PM_{2.5}$ over both domains are presented in Fig. 18. All models simulate a larger non-local source contribution during the spring period for both domains and pollutants. For both pollutants and domains, the local sources have relatively larger contribution in winter periods, reflected by the lower RERER values compared to other parts of the year. Regarding European O_3, majority of the models show a RERER value of between 0.5 and 1, while DE1 shows much lower and IT1 much higher values (see also Table 6). DE1 and FI1 simulates the lowest RERER values for $PM_{2.5}$ (< 0.1), while other models calculate RERER values between 0.1 and 0.5. Regarding O_3 over North America, US3 shows that in winter months, O_3 is controlled more by local emission with RERER values much lower than 0.5, while DE1 shows the highest non-local contributions throughout the year.

CONCLUSIONS

In the framework of the third phase of the Air Quality Model Evaluation International Initiative (AQMEII3), the impacts of local vs. foreign emissions over the European and North American receptor regions are simulated by introducing a 20% decrease of global and regional emissions by research groups, using different state-of-the-art chemistry and transport models. The emission perturbations were introduced globally, as well as over the HTAP2-defined regions of Europe, North America and East Asia. Base case and the perturbation

scenarios are first simulated using the global C-IFS global model, which provides the boundary conditions to the regional CTMs.

The base case simulation of each model has been evaluated against surface observations from Europe and North America. The temporal variabilities of all pollutants are well captured by all models with correlations generally higher than 0.70. O_3 levels are generally simulated with a *MNB* less than 10% with few exceptions of *MNB* values up to -35%. NO_2, CO and SO2 levels are simulated with underestimations up to 75%, 45% and 68%, respectively. PM_{10} and $PM_{2.5}$ levels are underestimated by 20% to 70%, with slightly higher biases in PM_{10} levels.

Results from the perturbation simulations show that the largest impacts over both Europe and North American domains are simulated in response to the global emission perturbation (GLO). These responses are similar, however slightly lower, as compared to the local emission perturbation scenarios for Europe (EUR) and North America (NAM). In contrast to the GLO scenario, O_3 levels over Europe slightly increase by 0.13 ppb (0.02%). The annual mean value of the O_3 daily maximum of 8-hour running average decreases in all scenarios over Europe, highest in the GLO scenario by ~1% and lowest in the NAM scenario by ~0.3%. Over North America, the annual mean value of the O_3 daily maximum of 8-hour running average decreased by ~5% in the GLO scenario, 3% in the NAM scenario and 0.7% in the EAS scenario. The impact of foreign emissions simulated by the NAM scenario for Europe and EAS scenario for North America were found to be lowest, however still noticeable, particularly close to the boundaries. This impact is especially noticeable (up to only a few percent) for the western parts of the North American domain in response to the emission reductions over East Asia. The response is almost linear (~20% decrease) to the change in emissions for NO_2, SO_2 and PM in the global perturbation scenario (GLO), while O_3 levels decrease slightly (~1%).

Despite these small differences, there are large geographical differences. NO_2, CO and SO_2 levels are mainly affected over emission hot spots in the GLO scenario as well as in the EUR scenario for Europe and the NAM scenario for North America. O_3 levels increase over the hot spot regions, in particular the Benelux region in Europe, by up to ~6% due to the reduced effect of NOx-titration. Over the North American domain, the central-to-eastern part and the western coast of the U.S experience the largest response to the global emission perturbation. For most of the pollutants, there is distinct seasonality in the responses particularly to the global and local emission perturbations. The largest responses are calculated during winter months, where anthropogenic emission are highest, except for O_3, where largest responses are seen during spring/summer months, suggesting photochemistry still plays an important role in O_3 levels.

The RERER metrics have been calculated to examine the differences in the strengths of non-local source contributions to different species among the different models. The large RERER values over Europe and North America for O_3 (~0.8), show a larger contribution of non-local sources, while for other gaseous pollutants (NO_2, CO and SO_2) and particles (PM_{10} and $PM_{2.5}$), low RERER values (< 0.5) indicate that these pollutants are largely controlled by

local sources. Results show that the contribution of local sources on NO_2, SO_2 and PM levels are larger in North America compared to Europe, while for CO, local sources have a larger share in Europe in comparison with North America. In addition, RERER analyses shows that European O_3 is largely controlled by non-local sources (RERER > 0.5) throughout the domain. $PM_{2.5}$ levels are largely controlled by local sources with RERER values around 0.2 throughout the domain. Local sources play a more important role in controlling O_3 levels over the eastern part of the U.S. $PM_{2.5}$ levels over the western part of NA has a relatively larger contribution from non-local sources compared to the rest of the domain. A larger non-local source contribution during the spring period for both domains and pollutants has been calculated, suggesting long-range transport from non-local sources. For both pollutants and domains, the local sources have relatively larger contribution in winter periods, reflected by the lower RERER values compared to other parts of the year.

Overall results show that there is a large spread among the models, although the majority of the models simulate a similar seasonal variation. These differences suggest that despite the harmonization of inputs, such as emissions and boundary conditions, to regional models, there are still large differences between models, such as different gas phase and aerosol modules, deposition schemes, meteorological drivers and spatial and vertical resolutions. Therefore, the use of multi model ensembles can help to reduce the uncertainties inherent in individual models.

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Table 1. Key features (meteorological/chemistry and transport models, emissions, horizontal and vertical grids) of the regional models participating to the AQMEII3 health impact study and the perturbation scenarios they performed.

Group Code	Model	Emissions	Horizontal Resolution	Vertical Resolution	Gas Phase	Aerosol Model	Europe				North America			
							BASE	GLO	NAM	EUR	BASE	GLO	EAS	NAM
DE1	COSMO-CLM/CMAQ	HTAP	24 km × 24 km	30 layers, 50 hPa	CB5-TUCL	3 modes	×	×	×	×	×	×	×	×
DK1	WRF/DEHM	HTAP	50 km × 50 km	29 layers, 100 hPa	Brandt et al. (2012)	2 modes	×	×	×	×	×	×	×	×
ES1	WRF/CHEM	MACC	23 km × 23 km	33 layers, 50 hPa	RADM2	3 modes, MADE/SORGAM	×		×					
FI1	ECMWF/SILAM	MACC+HTAP	0.25° × 0.25°	12 layers, 13 km	CB4	1-5 bins, VBS	×	×	×	×				
FRES1	ECMWF/CHIMERE	HTAP+HTAP	0.25° × 0.25°	9 layers, 50 hPa	MELCHIOR2	8 bins	×	×	×	×				
IT1	WRF/CHEM	MACC	23 km × 23 km	33 layers, 50 hPa	RACM-ESRL	3 modes, MADE/VBS	×	×		×				
IT2	WRF/CAMx	MACC	23 km × 23 km	14 layers, 8 km	CB5	3 modes	×	×						
NL1	LOTOS/EUROS	MACC	0.50° × 0.25°	4 layers, 3.5 km	CB4	2 modes, VBS	×							
TR1	WRF/CMAQ	MACC	30 km × 30 km	24 layers, 10hPa	CB5	3 modes	×	×	×					
UK1	WRF/CMAQ	MACC	15 km × 15 km	23 layers, 100 hPa	CB5-TUCL	3 modes	×	×	×	×				
UK2	WRF/CMAQ	HTAP	30 km × 30 km	23 layers, 100 hPa	CB5-TUCL	3 modes	×	×						
UK3	WRF/CMAQ	MACC	18 km × 18 km	35 layers, 16 km	CB5	3 modes	×	×	×					
US3	WRF/CMAQ	SMOKE	12 km × 12 km	35 layers, 50 hPa	CB5-TUCL	3 modes					×	×	×	×

[1] MACC: Modelling group used only the MACC emissions, MACC+HTAP: Modelling group used MACC emissions for Europe and HTAP emissions over North Africa.

Table 2. Perturbations of global/regional anthropogenic emissions and boundary conditions in the perturbation scenarios.

	GLO	Europe		North America	
		NAM	EUR	NAM	EAS
Emissions	-20%	-	-20%	-20%	-
Boundary conditions (Emissions in the IFS model)	-20%	-20%	-20%	-20%	-20%

900 Table 3. Monthly statistics of Pearson's Correlation (*r*), Normalized Mean Bias (*NMB*), Normalized Mean Gross Error (*NMGE*) and Root Mean
901 Square Error (*RMSE*: μg m^{-3} for Europe, while ppb for gases and μg m^{-3} for particles for North America) calculated for each model group.

		EUROPE													**NORTH AMERICA**						
		DE1	DK1	ES1	FI1	FRES1	IT1	IT2	TR1	UK1	UK2	MEAN	MEDIAN	C-IFS	DE1	DK1	US1	US3	MEAN	MEDIAN	C-IFS
O_3	r	0.63	0.90	0.82	0.83	0.91	0.92	0.93	0.87	0.92	0.90	0.93	0.92	0.89	0.78	0.59	0.89	0.87	0.84	0.83	0.71
	NMB	0.10	0.07	-0.14	-0.36	-0.10	0.04	-0.14	0.09	0.08	-0.03	-0.04	-0.04	-0.20	0.12	0.22	0.14	-0.02	0.09	0.11	-0.10
	NMGE	0.17	0.12	0.15	0.36	0.12	0.13	0.15	0.26	0.11	0.09	0.08	0.08	0.20	0.17	0.23	0.14	0.08	0.12	0.13	0.19
	RMSE	12.68	8.81	11.58	23.13	9.01	8.54	10.94	17.66	8.05	6.79	5.91	6.31	14.63	6.16	9.81	5.72	3.23	4.63	5.28	7.31
NO_2	r	0.80	0.88	0.89	0.95	0.74	0.90	0.92	0.90	0.85	0.85	0.95	0.93	0.92	0.99	0.92	0.94	0.93	0.98	0.99	0.91
	NMB	-0.75	-0.38	-0.47	0.00	0.05	-0.29	-0.30	0.58	-0.32	-0.06	-0.17	-0.24	0.07	-0.18	-0.35	0.05	0.31	-0.03	-0.02	0.41
	NMGE	0.75	0.38	0.47	0.20	0.23	0.29	0.30	0.58	0.32	0.17	0.18	0.24	0.20	0.18	0.35	0.10	0.31	0.06	0.02	0.41
	RMSE	9.38	5.41	6.00	2.89	3.44	4.43	4.15	7.39	4.65	2.74	2.70	3.49	2.59	1.01	2.05	0.62	1.77	0.40	0.26	2.30
CO	r	0.83	0.76	0.74	0.88	0.82	0.84	0.79	0.87	0.63	0.72	0.92	0.84	0.91	0.79	0.74	0.74	0.73	0.88	0.82	0.80
	NMB	-0.42	-0.42	-0.44	-0.27	-0.32	-0.38	-0.44	-0.20	-0.41	-0.43	-0.33	-0.38	-0.25	-0.19	-0.07	-0.06	-0.04	-0.07	-0.07	0.17
	NMGE	0.42	0.42	0.44	0.27	0.32	0.38	0.44	0.21	0.41	0.43	0.33	0.38	0.25	0.19	0.11	0.08	0.08	0.08	0.07	0.17
	RMSE	128.62	134.31	132.78	89.99	107.81	128.14	135.83	70.04	130.21	135.82	106.98	123.61	84.73	40.27	24.90	22.44	20.51	19.94	20.41	37.30
SO_2	r	0.85	0.90	0.88	0.86	0.87	0.86	0.86	0.54	0.83	0.83	0.93	0.92	0.70	0.79	0.81	0.80	0.78	0.87	0.78	0.04
	NMB	-0.01	-0.47	-0.65	-0.20	-0.16	-0.30	-0.55	0.04	-0.13	0.20	-0.19	-0.10	0.41	-0.46	-0.42	0.07	-0.13	-0.19	-0.13	0.35
	NMGE	0.24	0.48	0.65	0.28	0.22	0.31	0.55	0.28	0.19	0.28	0.21	0.12	0.45	0.46	0.42	0.11	0.13	0.19	0.13	0.35
	RMSE	0.92	1.47	2.03	0.95	0.80	1.23	1.71	1.14	0.86	1.05	0.76	0.58	1.39	1.27	1.18	0.32	0.40	0.53	0.40	1.02
PM_{10}	r	0.86	0.82	0.17	0.41	0.82	0.60	0.10	0.52	0.71	0.71	0.87	0.73	-0.74	-0.31	-0.47	*NA*	0.07	0.47	-0.07	0.02
	NMB	-0.71	-0.59	-0.47	-0.42	-0.51	-0.20	-0.48	-0.25	-0.47	-0.42	-0.41	-0.45	-0.62	-0.67	-0.84	*NA*	-0.25	-0.44	-0.46	-0.86
	NMGE	0.71	0.59	0.47	0.42	0.51	0.25	0.48	0.26	0.47	0.42	0.41	0.45	0.62	0.67	0.84	*NA*	0.27	0.44	0.46	0.86
	RMSE	20.43	18.25	16.16	14.67	15.74	9.78	16.48	10.45	14.78	13.72	13.15	14.63	19.87	20.42	25.09	*NA*	9.85	13.51	14.74	25.58
$PM_{2.5}$	r	0.89	0.86	0.24	0.58	0.84	0.75	0.11	0.62	0.77	0.77	0.89	0.82	-0.73	0.52	0.02	*NA*	0.54	0.61	0.56	0.18
	NMB	-0.64	-0.47	-0.27	-0.27	-0.36	-0.19	-0.48	-0.17	-0.40	-0.28	-0.32	-0.33	-0.59	-0.63	-0.14	*NA*	0.17	-0.15	-0.08	-0.39
	NMGE	0.64	0.47	0.35	0.30	0.36	0.24	0.49	0.24	0.41	0.30	0.32	0.33	0.59	0.63	0.20	*NA*	0.22	0.15	0.11	0.40
	RMSE	11.95	9.92	9.20	8.02	8.06	6.57	11.65	6.82	8.65	7.15	7.51	7.99	12.97	6.79	2.40	*NA*	2.78	1.92	1.41	5.04

902 Table 4. Annual mean absolute differences (ppb for gases and μg m^{-3} for particles) between the base case and the different emission perturbation
903 scenarios as calculated by the different model groups over the European domain.

Pollutant	Scenario	DE1	DK1	ES1	FI1	IT1	IT2	TR1	UK1	UK2	FRES1	All Mean	Common Mean
O_3	GLO	-1.54	-0.71		-0.40	-0.37	-0.63	2.83	-0.83	-0.79	-0.63	-0.34	-0.82
	NAM	-0.28	-0.24	0.77	-0.13			-0.30	-0.22		-0.22	-0.09	-0.22
	EUR	-0.77	0.14		0.09	0.43			0.06		0.12	0.01	-0.07
NO_2	GLO	-0.28	-0.72		-1.20	-0.93	-0.95	-1.93	-0.75	-1.10	-0.89	-0.97	-0.77
	NAM	0.00	0.01	0.17	0.00	0.00		0.01				0.03	0.00
	EUR	-0.30	-0.69		-1.05	-0.85			-0.70		-0.89	-0.75	-0.73
CO	GLO	-15.97	-14.03		-21.10	-18.13	-15.04	-26.01	-12.83	-16.94	-16.11	-17.35	-16.01
	NAM	-1.50	-1.71	3.26	-1.41			-1.35	-1.33		-1.55	-0.80	-1.50
	EUR	-10.49	-6.91		-14.63	-10.11			-7.87		-9.51	-9.92	-9.88
SO_2	GLO	-0.23	-0.12		-0.17	-0.17	-0.11	-0.23	-0.20	-0.28	-0.15	-0.18	-0.17
	NAM	0.00	0.00	0.03	0.00			0.00	0.00		0.00	0.00	0.00
	EUR	-0.23	-0.10		-0.14	-0.13			-0.16		-0.15	-0.15	-0.16
PM_{10}	GLO	-1.47	-1.90		-2.52	-2.97	-1.58	-3.58	-2.32	-2.81	-2.27	-2.38	-2.10
	NAM	-0.01	-0.09	0.00	-0.02			-0.04	-0.03		-0.04	-0.03	-0.04
	EUR	-2.03	-1.53		-2.20	-2.46			-1.96		-2.07	-2.04	-1.96
$PM_{2.5}$	GLO	-1.30	-1.76		-2.15	-2.56	-1.33	-2.79	-1.78	-2.44	-2.10	-2.02	-1.82
	NAM	0.01	-0.05	0.00	-0.02			-0.03	-0.02		-0.04	-0.02	-0.02
	EUR	-1.29	-1.42		-1.82	-2.05			-1.47		-1.89	-1.66	-1.58

904

905 Table 5. Annual mean absolute differences (ppb for gases and μg m^{-3} for particles) between the base case and the different emission perturbation
906 scenarios as calculated by the different model groups over the North American domain.

Pollutant	Scenario	DE1	DK1	US1	US3	All Mean	Common Mean
O_3	GLO	-1.70	-1.42	-1.41	-1.03	-1.39	-1.39
	NAM	-0.92	-0.66		-0.36	-0.65	-0.65
	EAS	-0.35	-0.24	-0.23	-0.19	-0.25	-0.26
NO_2	GLO	-0.35	-0.63	-1.07	-1.20	-0.81	-0.73
	NAM	-0.36	-0.62		-1.17	-0.71	-0.71
	EAS	0.00	0.00	0.00	-0.01	0.00	0.00
CO	GLO	-9.31	-20.48	-22.12	-25.01	-19.23	-18.27
	NAM	-3.84	-13.35		-19.87	-12.35	-12.35
	EAS	-2.60	-4.16	-3.64	-3.07	-3.37	-3.28
SO_2	GLO	-0.33	-0.32	-0.48	-0.25	-0.34	-0.30
	NAM	-0.33	-0.32		-0.48	-0.37	-0.37
	EAS	0.00	0.00		0.00	0.00	0.00
PM_{10}	GLO	-2.26	-0.66		-4.24	-2.39	-2.39
	NAM	-2.02	-0.59		-4.19	-2.27	-2.27
	EAS	-0.56	-0.05		-0.03	-0.21	-0.21
$PM_{2.5}$	GLO	-0.60	-1.67		-2.29	-1.52	-1.52
	NAM	-0.62	-1.56		-2.24	-1.47	-1.47
	EAS	0.01	-0.04		-0.03	-0.02	-0.02

907

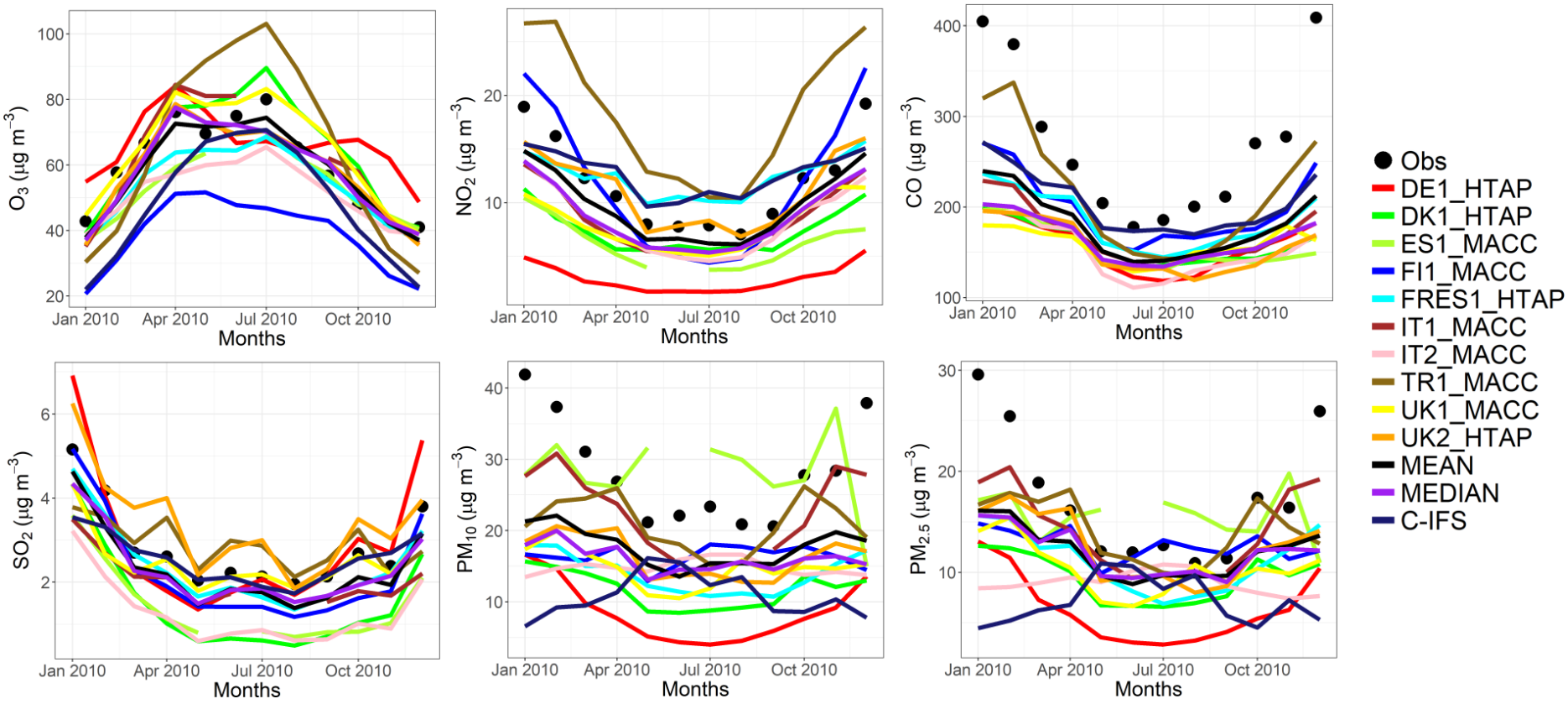
908

909 Table 6. Annual mean RERER values calculated for the multi-model mean ensembles over Europe and North America.

910

	O_3	NO_2	CO	SO_2	PM_{10}	$PM_{2.5}$
	EUROPE					
DE1	0.44	-0.09	0.44	0.02	0.01	0.01
DK1	0.85	0.23	0.63	0.37	0.17	0.28
FI1	0.76	-0.01	0.40	0.01	0.02	0.02
FRES1	0.78	0.15	0.56	0.30	0.20	0.20
IT1	1.10	0.34	0.93	0.42	0.27	0.26
UK1	0.92	0.35	0.52	0.43	0.33	0.34
MMM	0.77	0.18	0.55	0.27	0.18	0.19
MEDIAN	0.81	0.19	0.54	0.34	0.18	0.23
	NORTH AMERICA					
DE1	0.77	0.12	0.73	0.07	0.09	0.12
DK1	0.93	0.06	0.90	0.15	0.07	0.12
US3	0.54	0.02	0.47	0.11	0.08	0.10
MMM	0.75	0.05	0.71	0.11	0.08	0.11
MEDIAN	0.77	0.06	0.73	0.11	0.08	0.12

911



912

913 Fig.1. Observed and simulated monthly mean air pollutant levels, averaged over the monitoring stations over Europe.

914

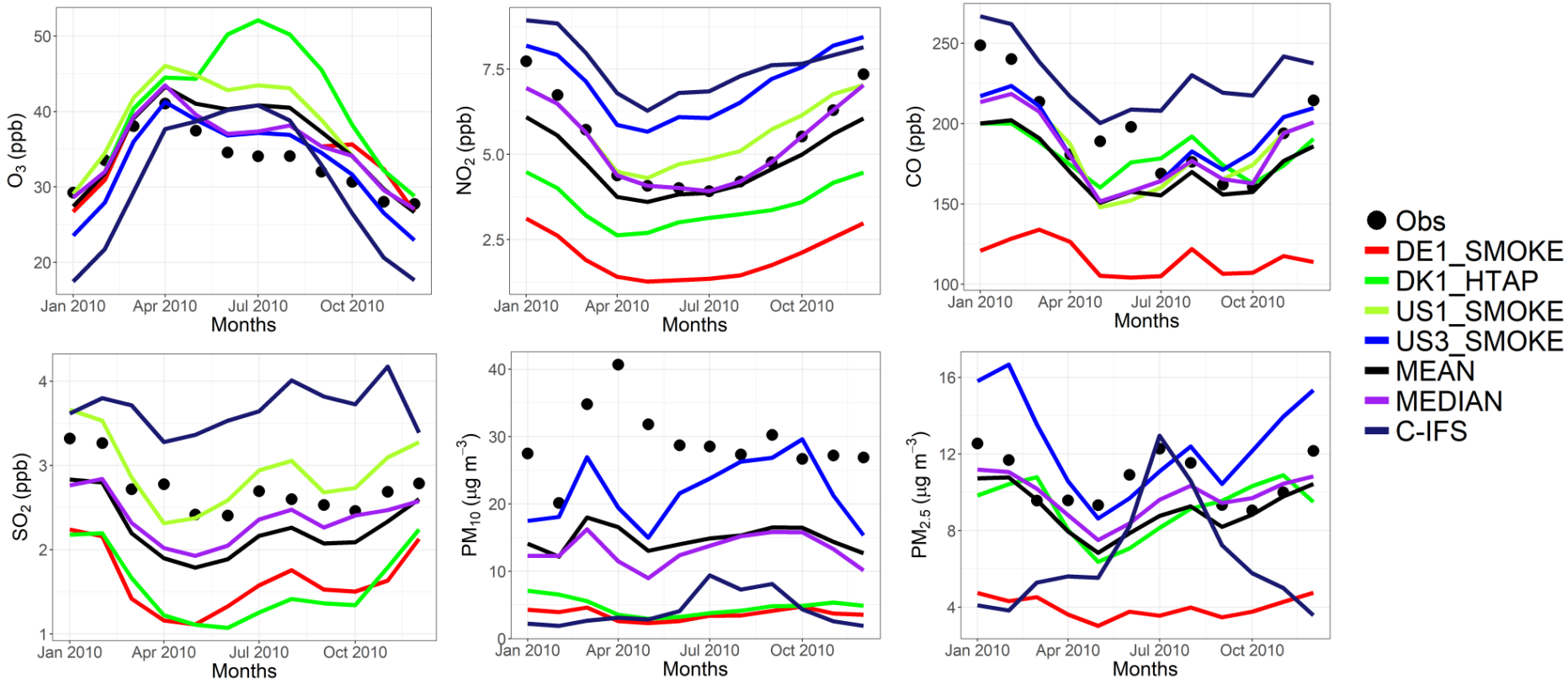


Fig.2. Observed and simulated monthly mean air pollutant levels, averaged over the monitoring stations over North America.

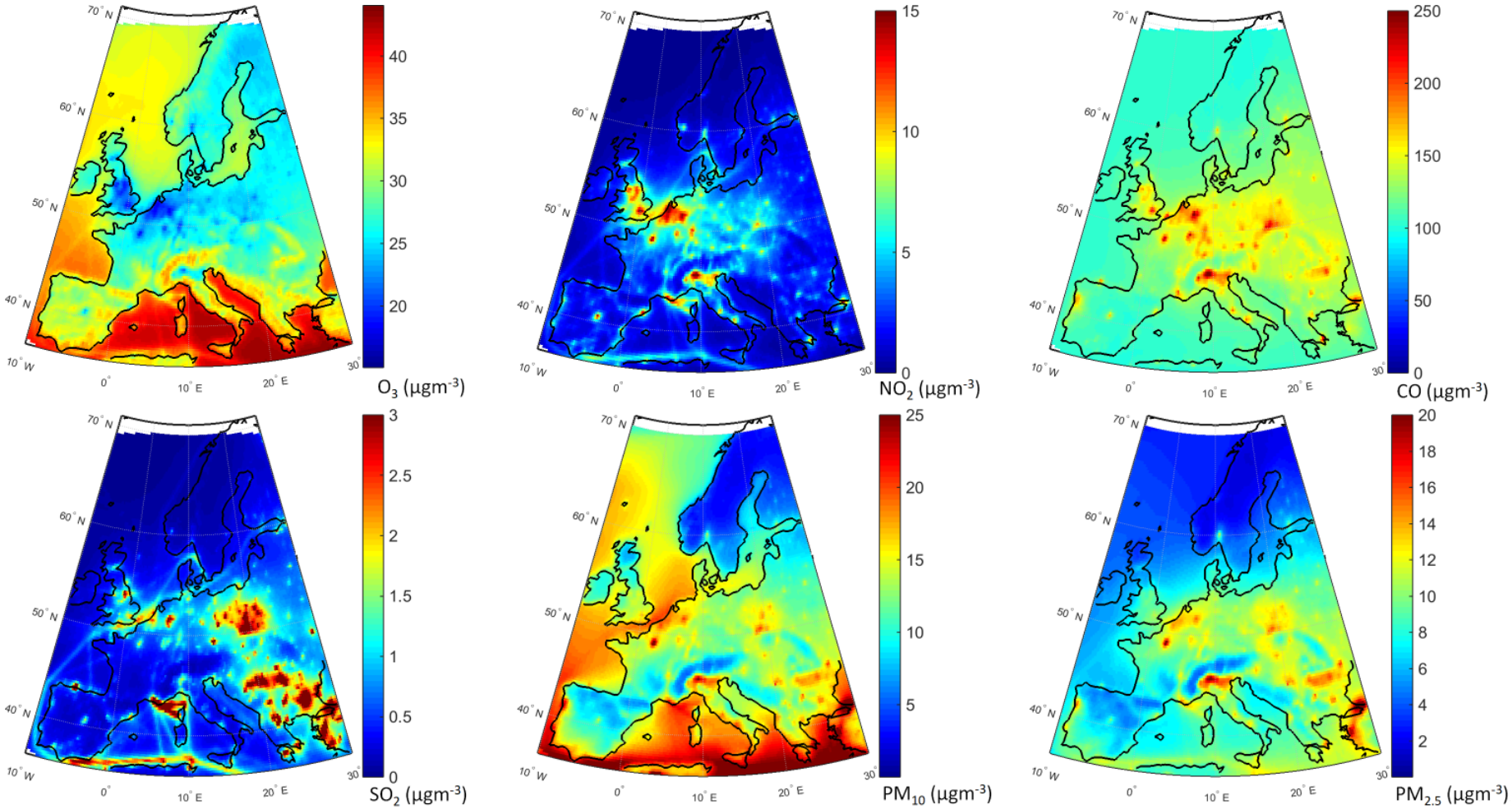


Fig.3. Multi-model mean air pollutant levels over Europe as simulated in the base case.

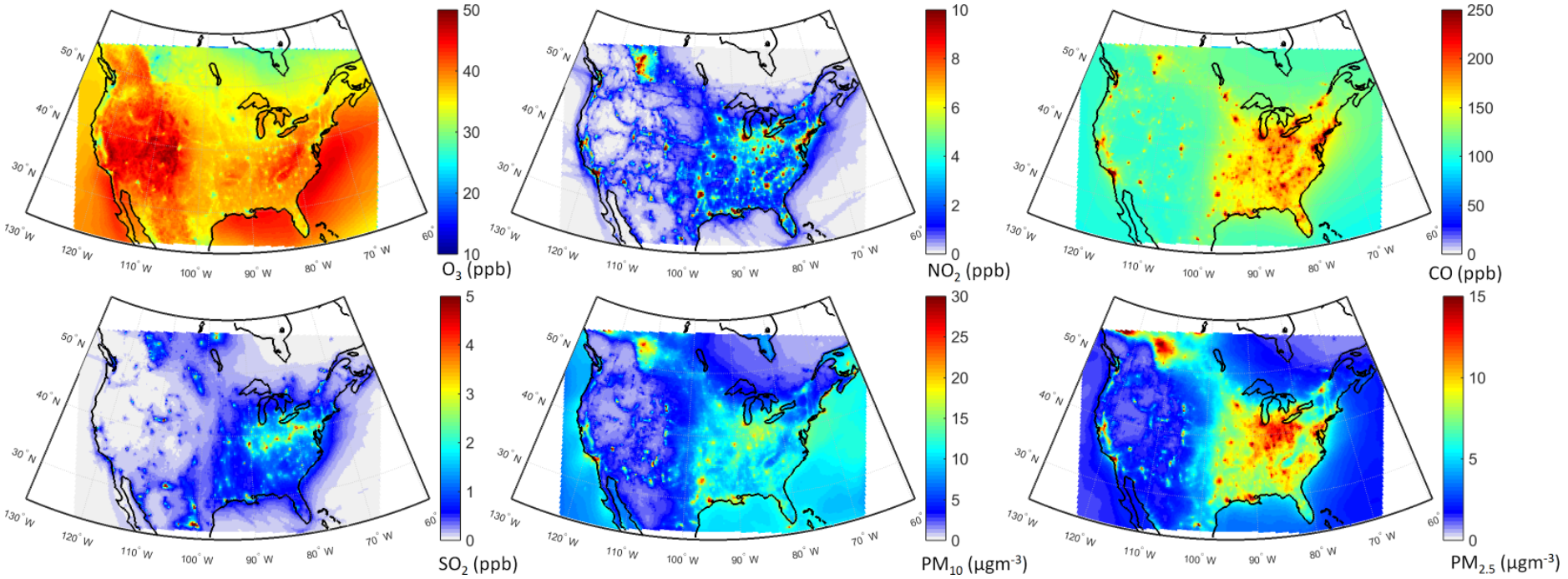


Fig.4. Multi-model mean air pollutant levels over North America as simulated in the base case.

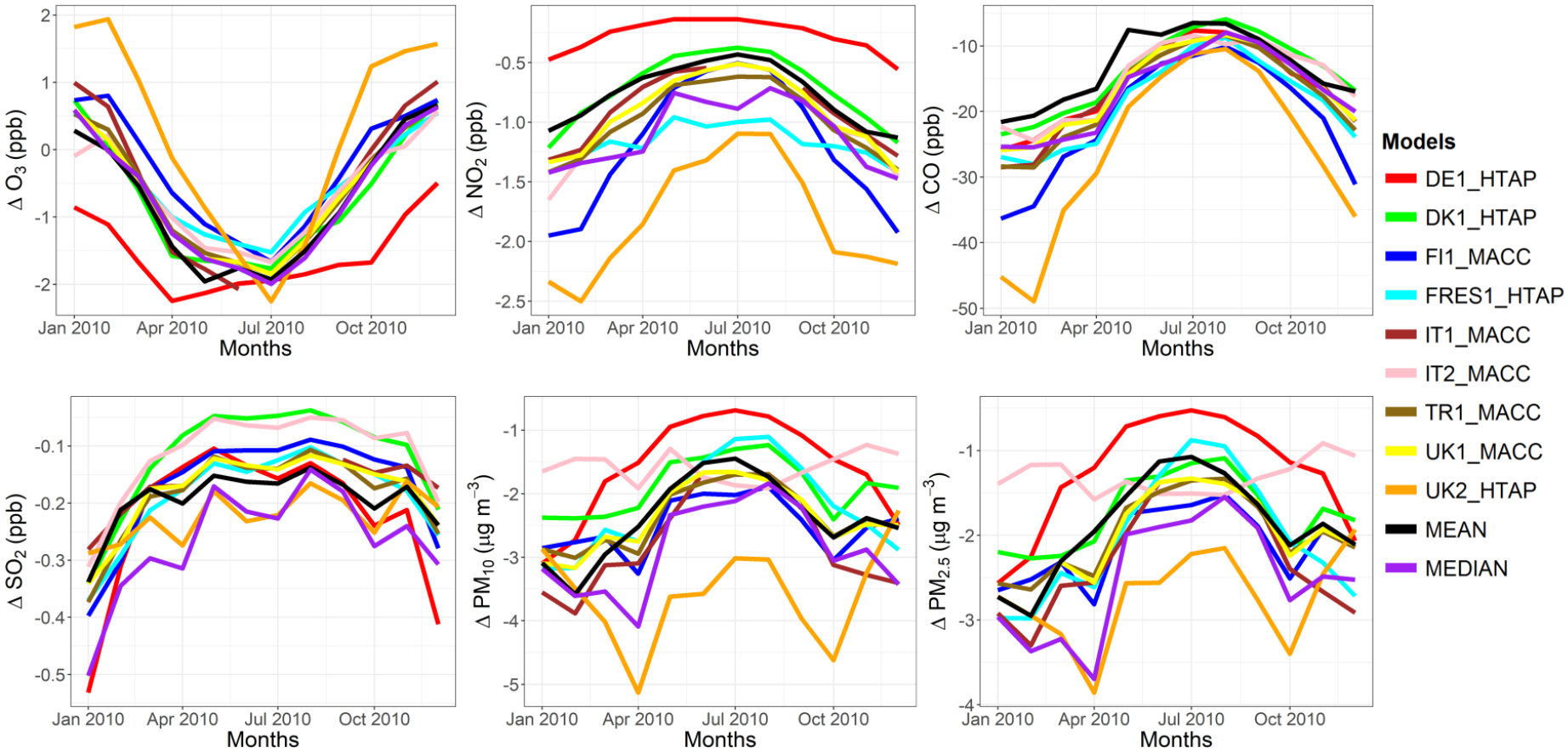
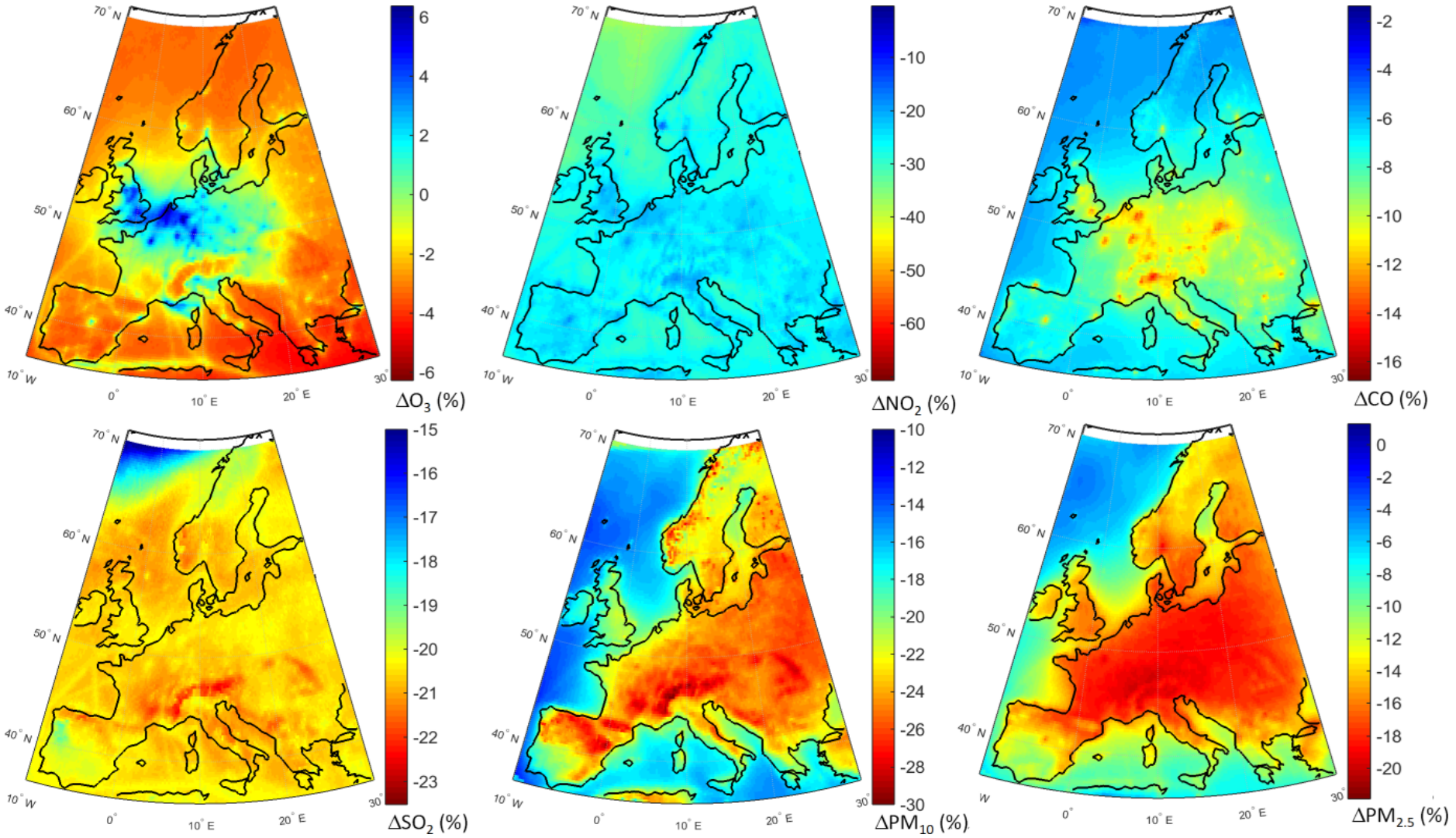


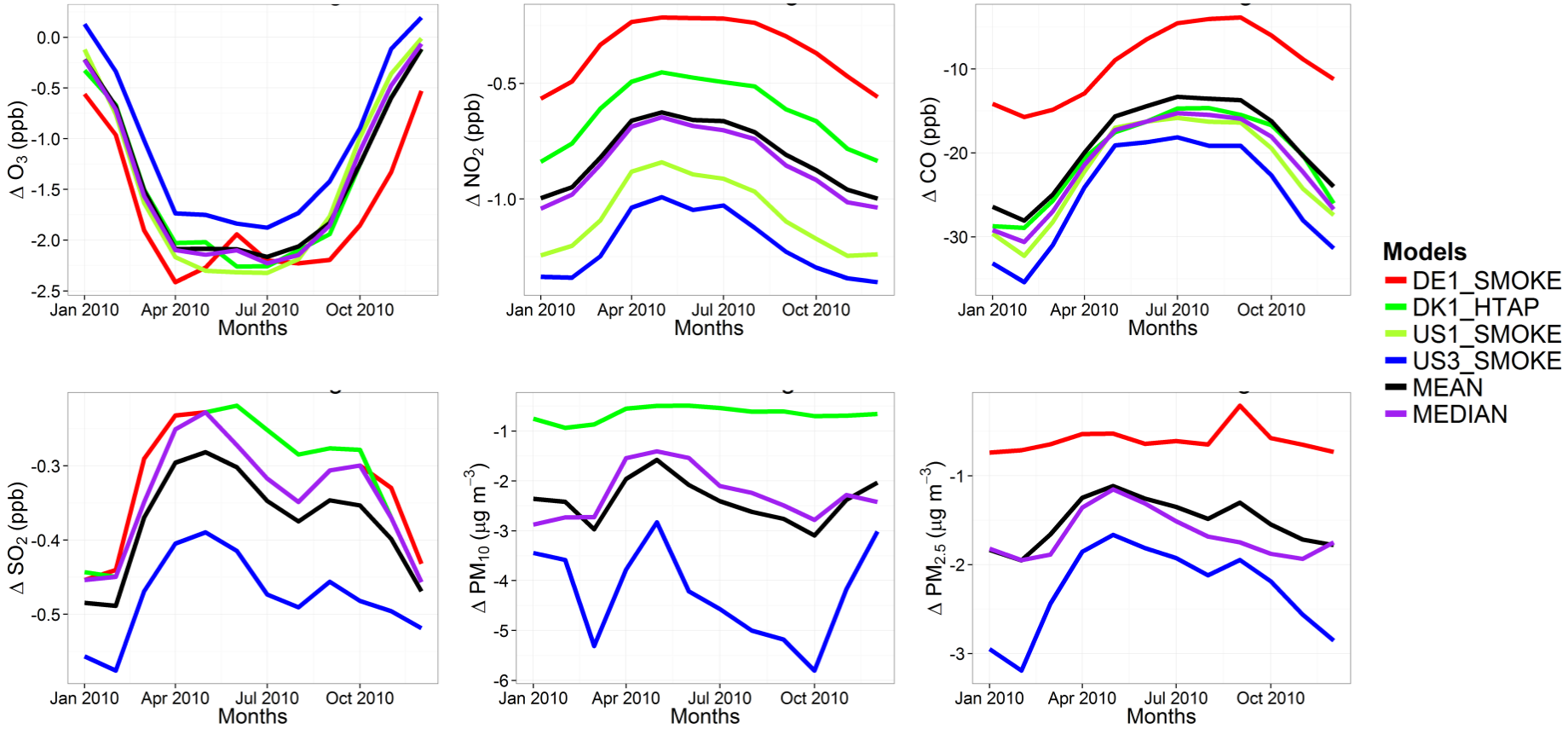
Fig.5. Absolute impact of the 20% reduction of the global anthropogenic emissions over Europe (GLO_{EUR}-$BASE_{EUR}$).



930

931 Fig.6. Spatial distribution of the annual mean relative differences between the global perturbation scenario and the base case over Europe as
932 simulated by the multi-model mean ensemble.

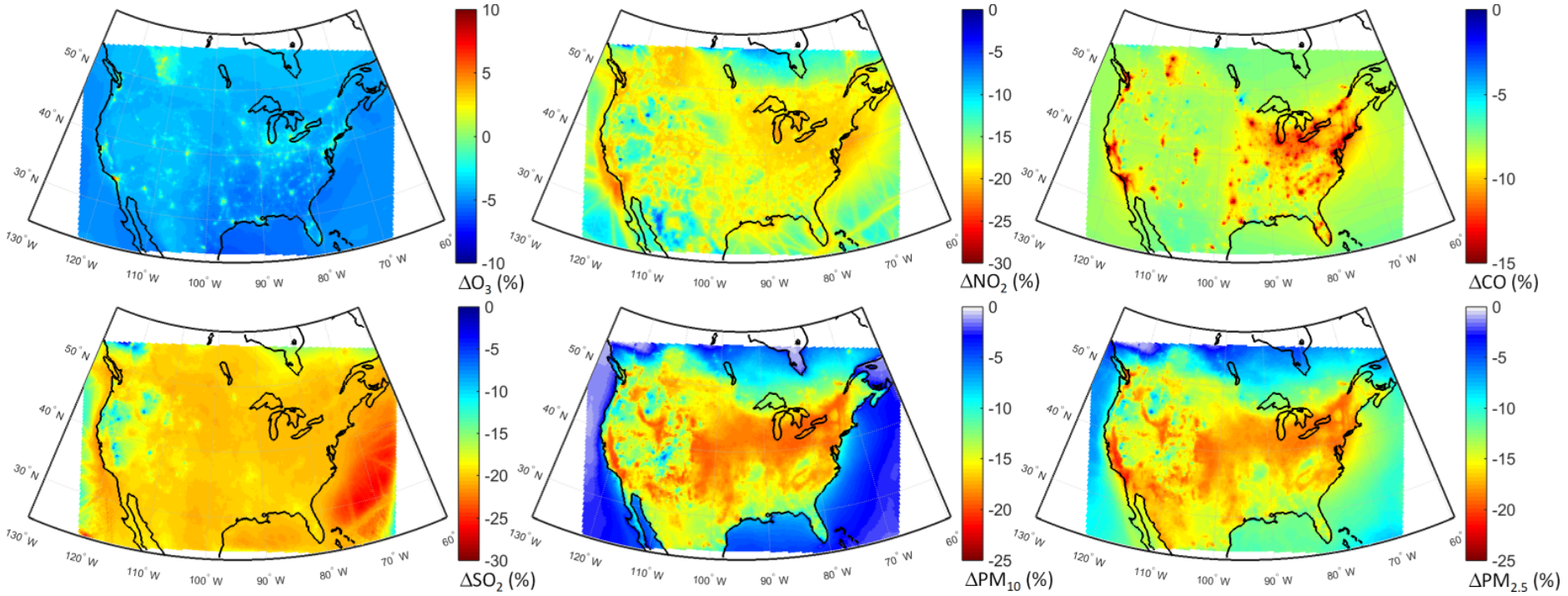
933



934

935 Fig.7. Absolute impact of the 20% reduction of the global anthropogenic emissions over North America (GLO_{NAM}-$BASE_{NAM}$).

936



937

938 Fig.8. Spatial distribution of the annual mean relative differences between the global perturbation scenario and the base case over North America
939 as simulated by the multi-model mean ensemble.

940

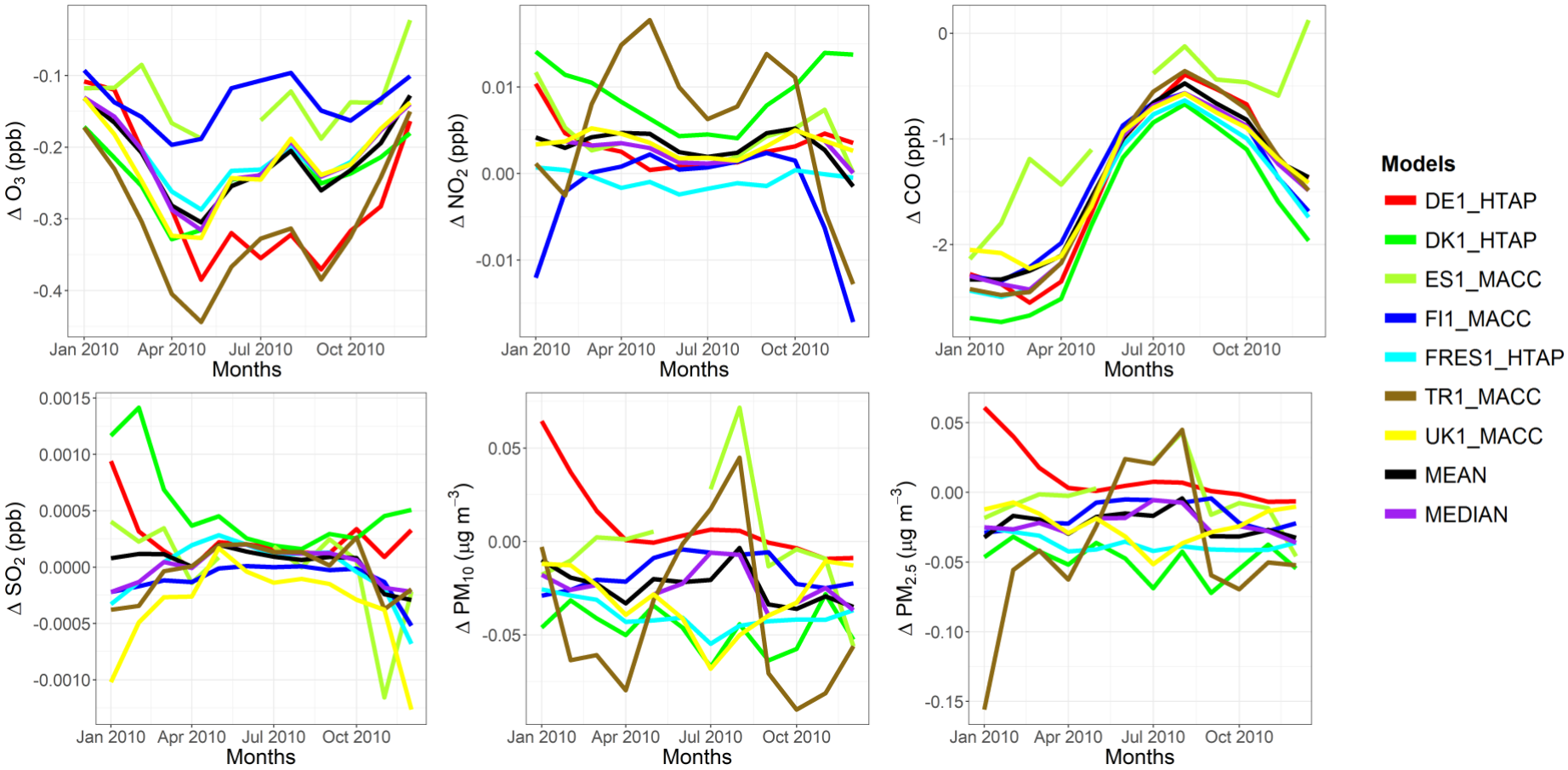
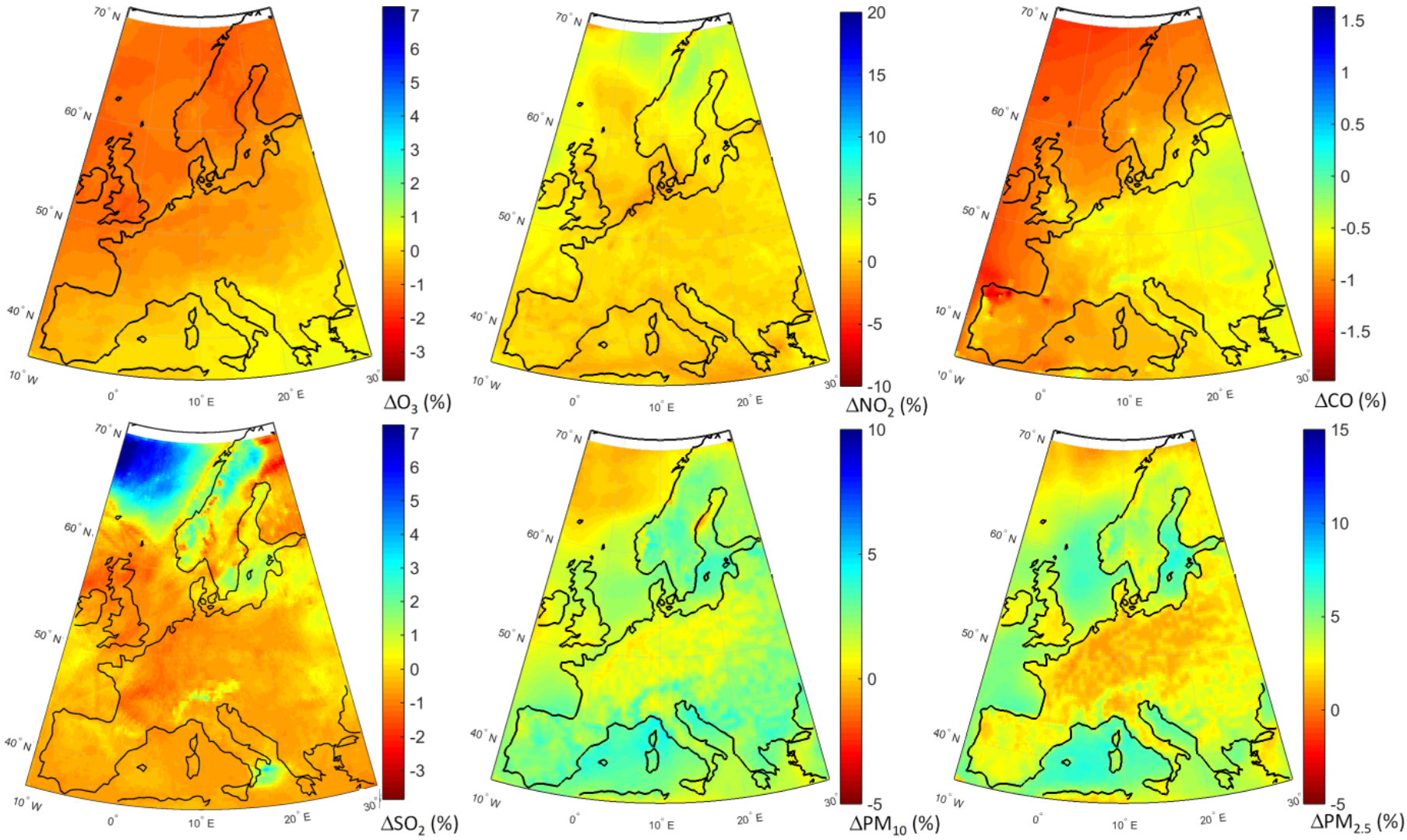


Fig.9. Absolute impact of the 20% reduction of the North American anthropogenic emissions over Europe (NAM_{EUR}-$BASE_{EUR}$).

945


70°N
60°N
50°N
40°N
10°W
0°
10°E
20°E
30°
7 6 5 4 3 2 1 0 -1 -2 -3
ΔO3 (%)
20 15 10 5 0 -5 -10
ΔNO2 (%)
1.5 1 0.5 0 -0.5 -1 -1.5
ΔCO (%)
ΔSO2 (%)
10 5 0 -5
ΔPM10 (%)
15 10 5 0 -5
ΔPM2.5 (%)

946

Fig.10. Spatial distribution of the annual mean relative differences between the North American emissions perturbation scenario and the base case over Europe as simulated by the multi-model mean ensemble.

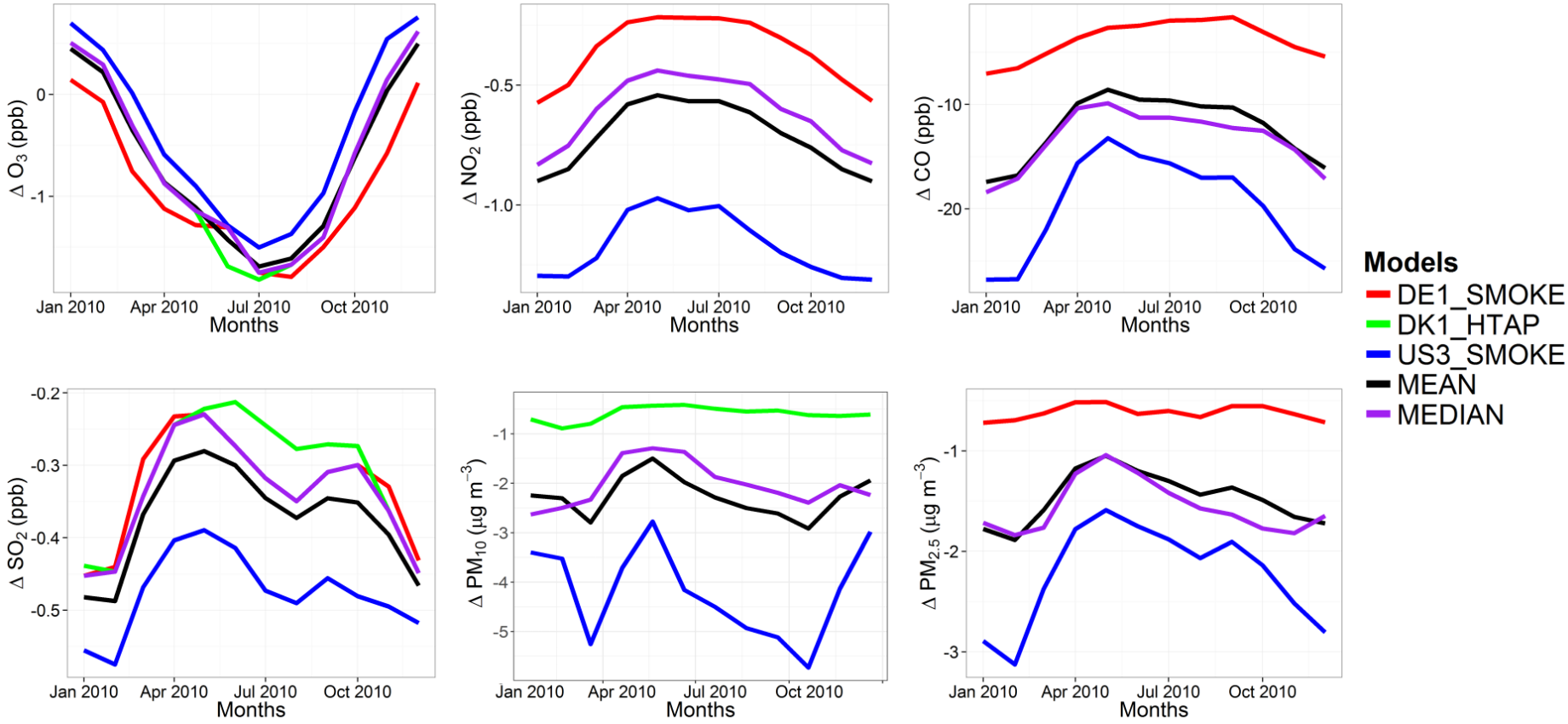
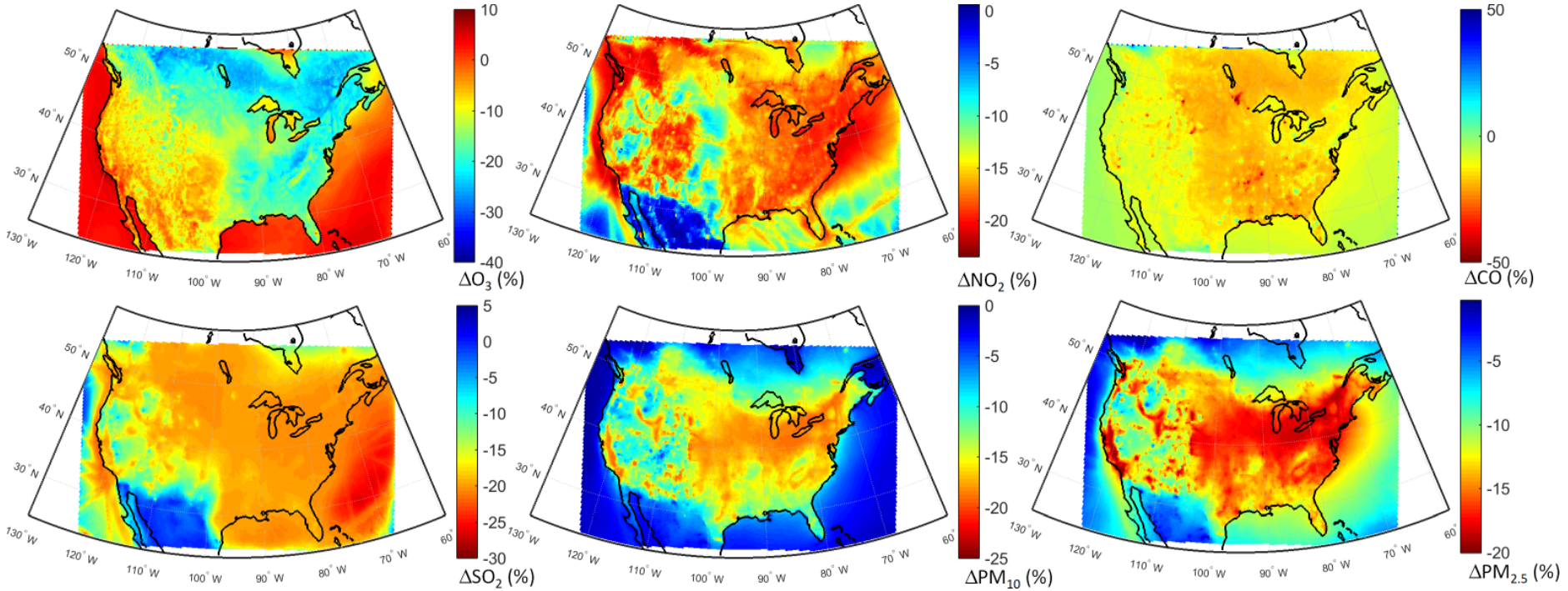


Fig.11. Absolute impact of the 20% reduction of the North American anthropogenic emissions over North America (GLO_{NAM}-$BASE_{NAM}$).

953



954

955 Fig.12. Spatial distribution of the annual mean relative differences between the North American emissions perturbation scenario and the base
956 case over North America as simulated by the multi-model mean ensemble.

957

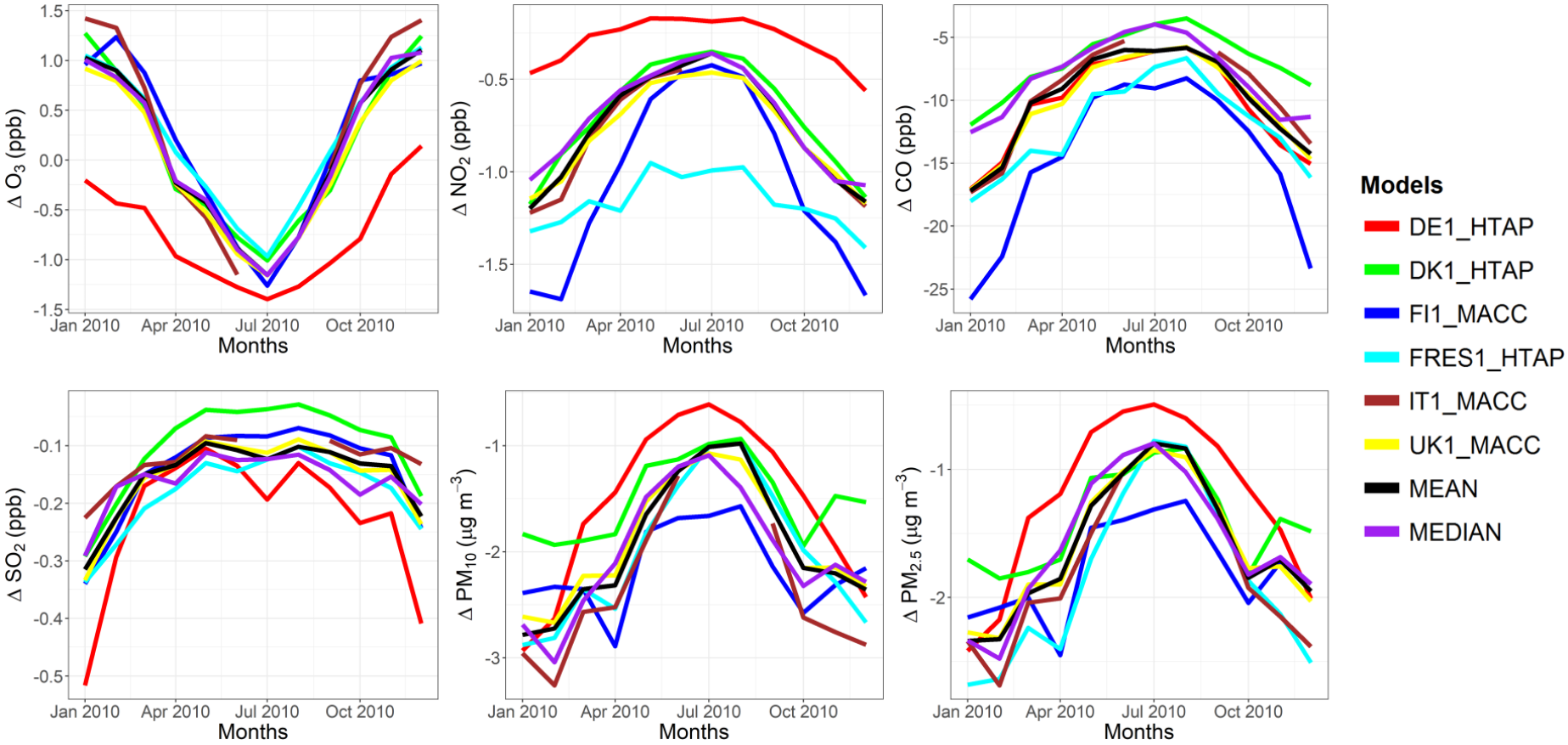
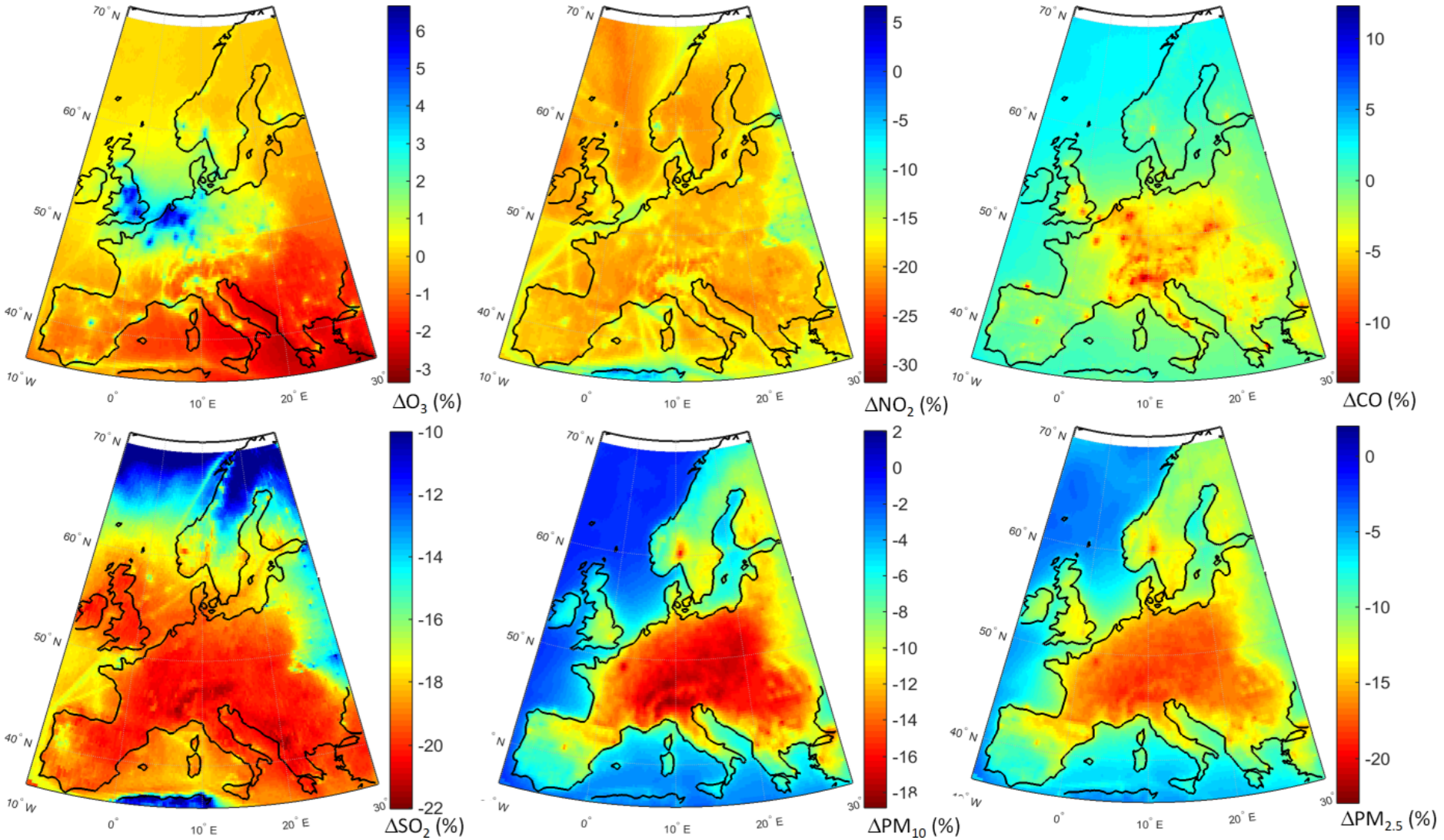


Fig.13. Absolute impact of the 20% reduction of the European anthropogenic emissions over Europe (EUR_{EUR}-$BASE_{EUR}$).



962

963 Fig.14. Spatial distribution of the annual mean relative differences between the European emissions perturbation scenario and the base case over
964 Europe as simulated by the multi-model mean ensemble.

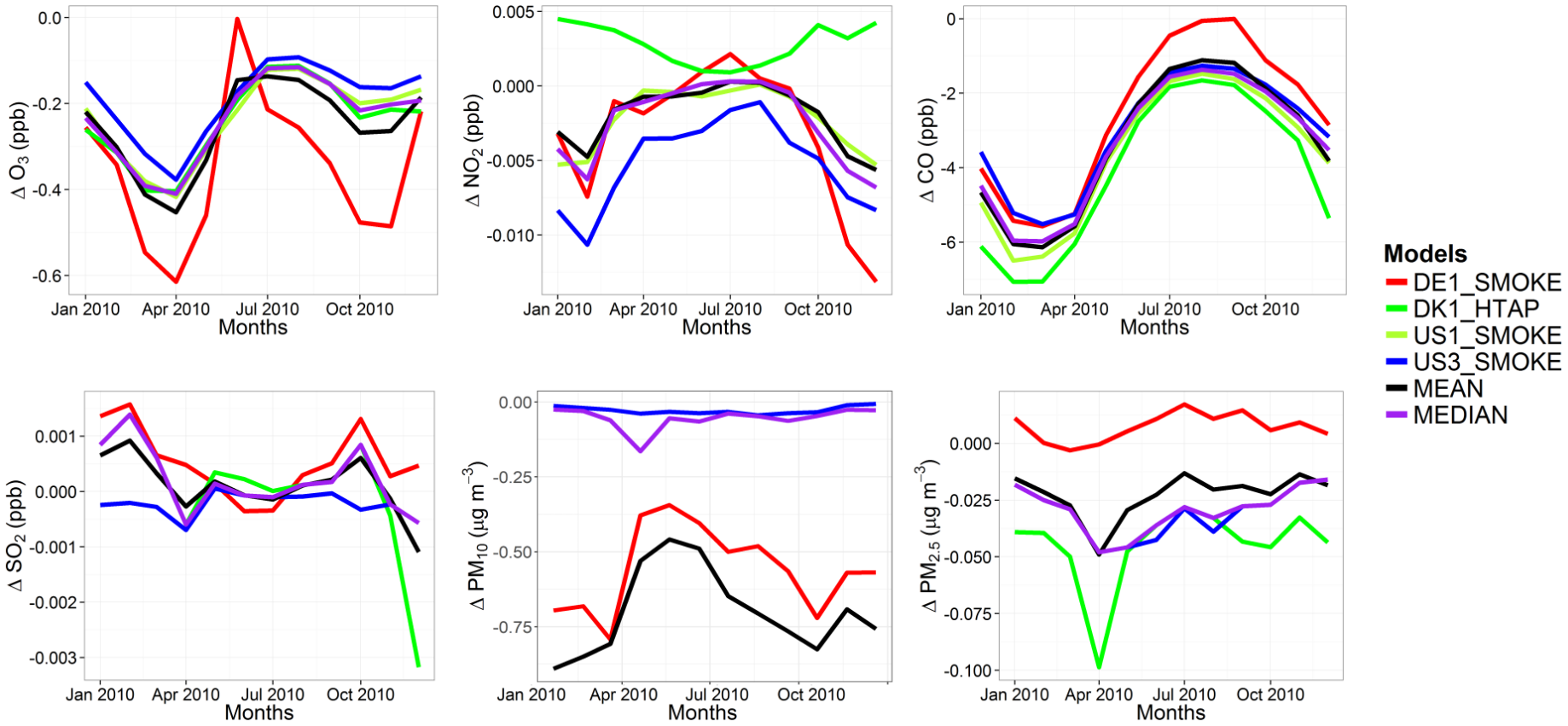
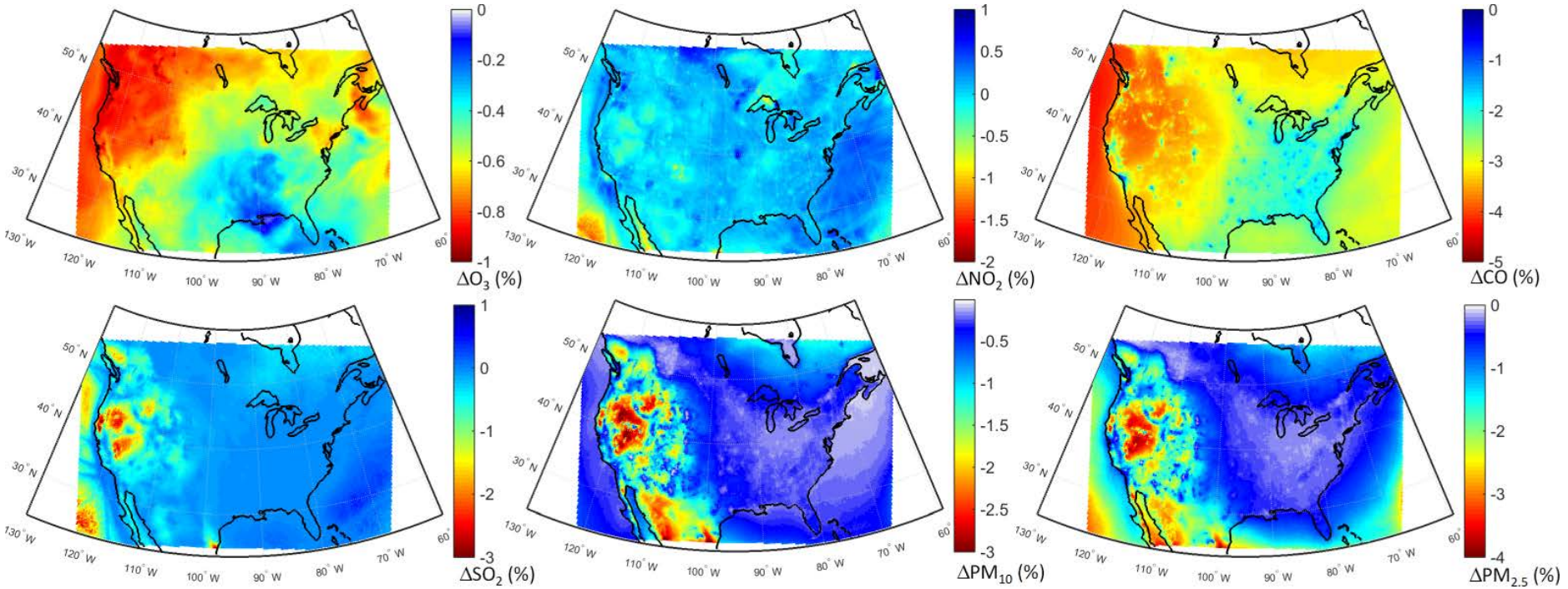


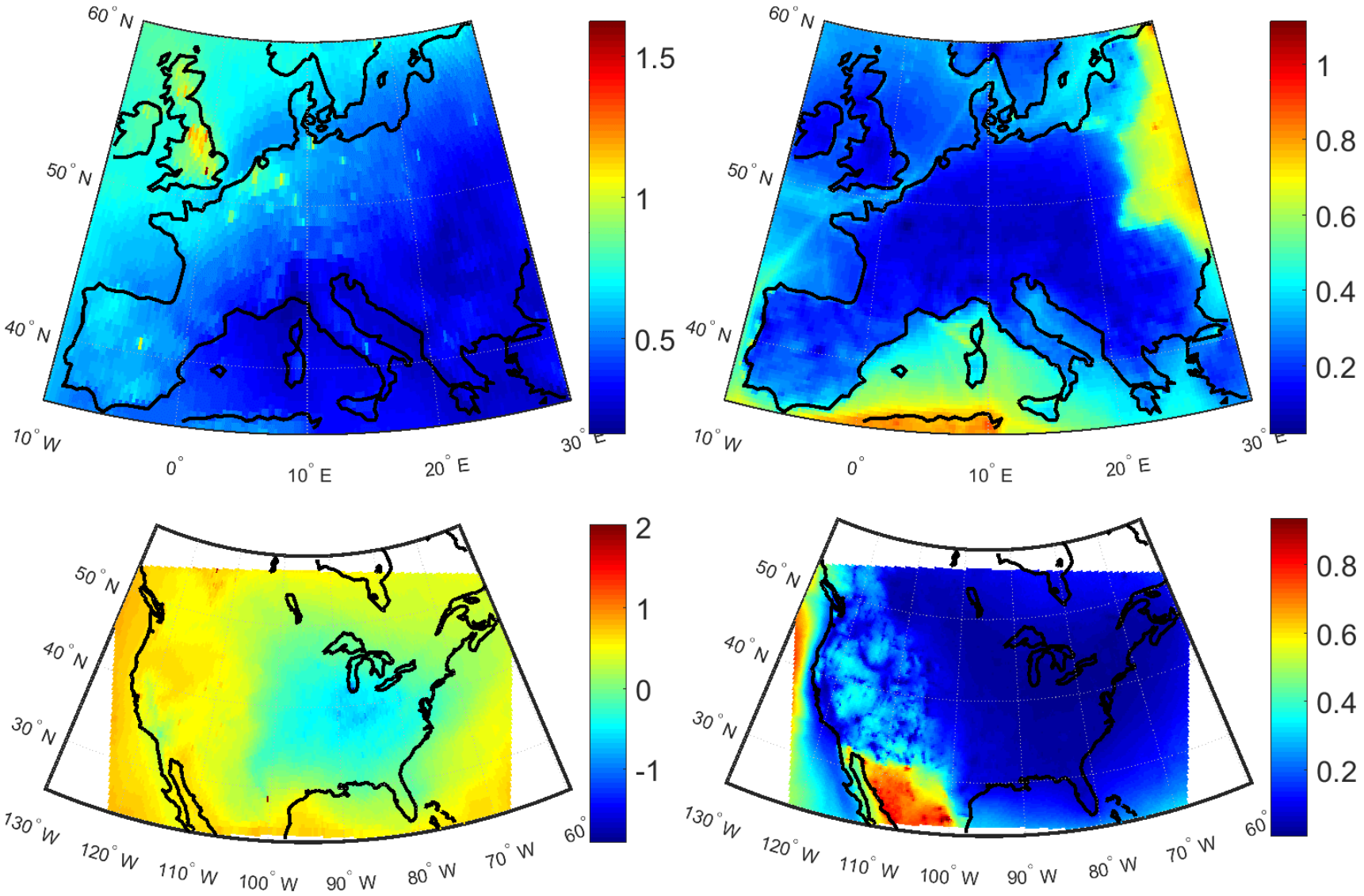
Fig.15. Absolute impact of the 20% reduction of the East Asian anthropogenic emissions over North America (GLO_{NAM}-$BASE_{NAM}$).



971

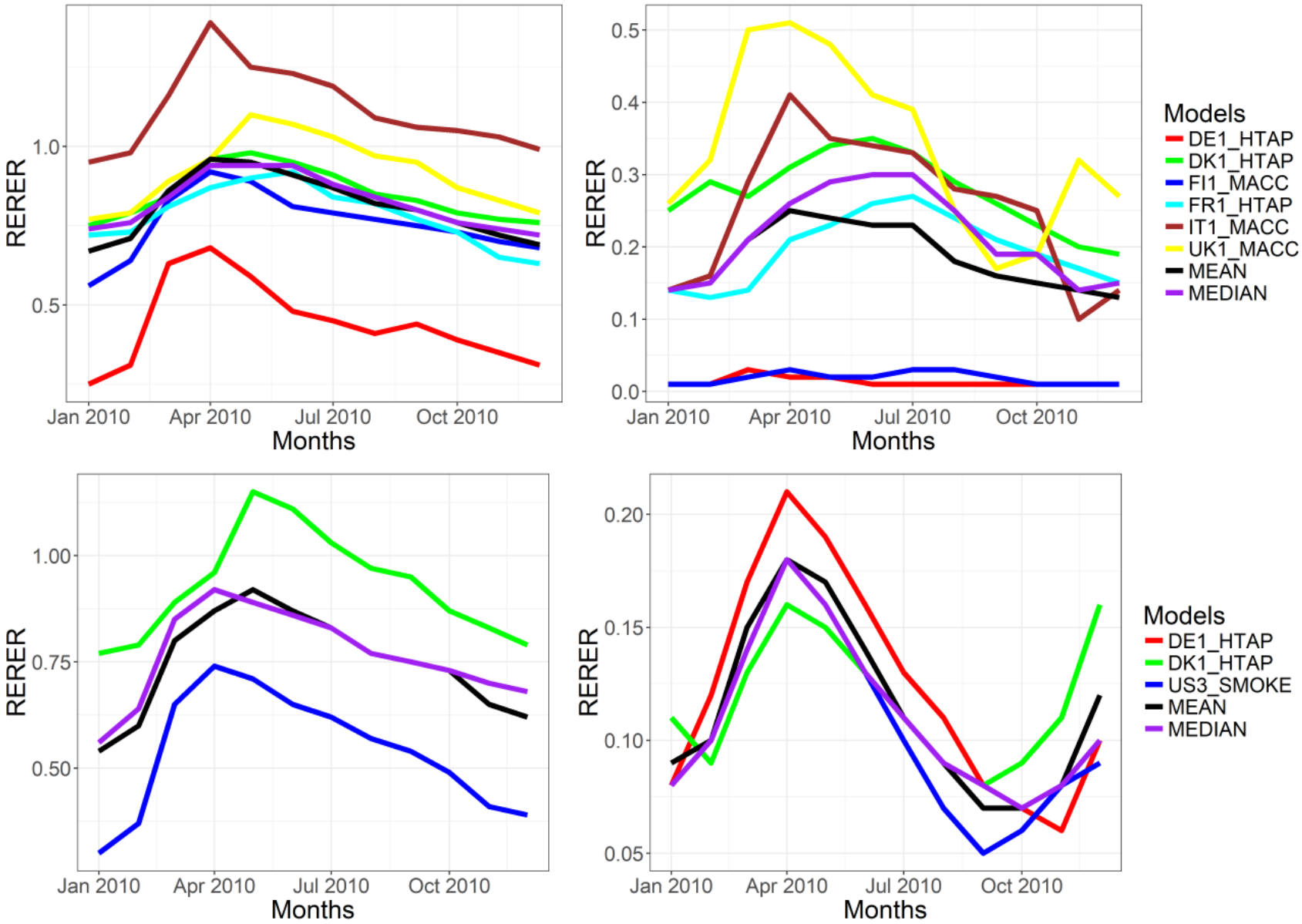
972 Fig.16. Spatial distribution of the annual mean relative differences between the East Asian emissions perturbation scenario and the base case
973 over North America as simulated by the multi-model mean ensemble.

974



975

976 Fig. 17. Spatial distribution of RERER values constructed from the annual mean responses of O_3 and $PM_{2.5}$ over Europe and North America.


RERER
1.0
0.5
Jan 2010
Apr 2010
Jul 2010
Oct 2010
Months
RERER
0.5
0.4
0.3
0.2
0.1
0.0
Jan 2010
Apr 2010
Jul 2010
Oct 2010
Months
Models
DE1_HTAP
DK1_HTAP
FI1_MACC
FR1_HTAP
IT1_MACC
UK1_MACC
MEAN
MEDIAN
RERER
1.00
0.75
0.50
Jan 2010
Apr 2010
Jul 2010
Oct 2010
Months
RERER
0.20
0.15
0.10
0.05
Jan 2010
Apr 2010
Jul 2010
Oct 2010
Months
Models
DE1_HTAP
DK1_HTAP
US3_SMOKE
MEAN
MEDIAN

977

978 Fig. 18. Seasonal variations of RERER values of O_3 and $PM_{2.5}$ over Europe and North America.