

# 1 Phenomenology of summer ozone episodes over the Madrid 2 Metropolitan Area, central Spain

3 Xavier Querol<sup>1</sup>, Andrés Alastuey<sup>1</sup>, Gotzon Gangoiti<sup>2</sup>, Noemí Perez<sup>1</sup>, Hong K. Lee<sup>3</sup>, Heeram R.  
4 Eun<sup>3</sup>, Yonghee Park<sup>3</sup>, Enrique Mantilla<sup>4</sup>, Miguel Escudero<sup>5</sup>, Gloria Titos<sup>1</sup>, Lucio Alonso<sup>2</sup>, Brice  
5 Temime-Roussel<sup>6</sup>, Nicolas Marchand<sup>6</sup>, Juan R. Moreta, M. Arantxa Revuelta<sup>7</sup>, Pedro Salvador<sup>8</sup>,  
6 Begoña Artíñano<sup>8</sup>, Saúl García dos Santos<sup>9</sup>, Mónica Anguas<sup>10</sup>, Alberto Notario<sup>11</sup>, Alfonso Saiz-  
7 Lopez<sup>10</sup>, Roy M. Harrison<sup>12</sup>, Millán Millán<sup>4</sup>, Kang-Ho Ahn<sup>3</sup>

8  
9 <sup>1</sup>Institute of Environmental Assessment and Water Research (IDAEA-CSIC), C/Jordi Girona 18-26, Barcelona, 08034  
10 Spain

11 <sup>2</sup>Escuela Técnica Superior Ingeniería de Bilbao, Departamento Ingeniería Química y del Medio Ambiente,  
12 Universidad del País Vasco UPV/EHU, Urkixo Zumarkalea, S/N, Bilbao, 48013 Spain

13 <sup>3</sup>Department of Mechanical Engineering, Hanyang University, Ansan 425-791, Republic of Korea

14 <sup>4</sup>Centro de Estudios Ambientales del Mediterráneo, CEAM, Unidad Asociada al CSIC, Parque Tecnológico C/ Charles  
15 R. Darwin, 14 Paterna, Valencia, 46980 Spain

16 <sup>5</sup> Centro Universitario de la Defensa de Zaragoza, Academia General Militar, Ctra. de Huesca s/n, Zaragoza, 50090  
17 Spain

18 <sup>6</sup>Aix Marseille Univ, CNRS, LCE, Marseille, France

19 <sup>7</sup>Agencia Estatal de Meteorología, AEMET, C/ Leonardo Prieto Castro, 8, Madrid, 28071 Spain

20 <sup>8</sup>Department of Environment, CIEMAT, Joint Research Unit Atmospheric Pollution CIEMAT-CSIC, c/ Avenida  
21 Complutense 40, Madrid, 28040 Spain

22 <sup>9</sup>Centro Nacional de Sanidad Ambiental. Instituto de Salud Carlos III (ISCIII), Ctr Majadahonda a Pozuelo km 2,  
23 Majadahonda (Madrid), 28222 Spain

24 <sup>10</sup>Department of Atmospheric Chemistry and Climate, Institute of Physical Chemistry Rocasolano, CSIC, Madrid,  
25 28006 Spain

26 <sup>11</sup>University of Castilla-La Mancha, Physical Chemistry Department, Faculty of Chemical Science and Technologies,  
27 Ciudad Real, Spain.

28 <sup>12</sup>National Centre for Atmospheric Science, University of Birmingham, B15 2TT United Kingdom. †Also at:  
29 Department of Environmental Sciences/Centre for Excellence in Environmental Studies, King Abdulaziz University,  
30 Jeddah, Saudi Arabia

31

## 32 Abstract

33 Various studies have reported that photochemical nucleation of new ultrafine particles (UFP)  
34 in urban environments within high insolation regions occurs simultaneously with high ground  
35 ozone (O<sub>3</sub>) levels. In this work, we evaluate the atmospheric dynamics leading to summer O<sub>3</sub>  
36 episodes in the Madrid Air Basin (Central Iberia) by means of measuring a 3D distribution of  
37 concentrations for both pollutants. To this end, we obtained vertical profiles (up to 1200 m  
38 above ground level) using tethered balloons and miniaturised instrumentation at a suburban  
39 site located to the SW of the Madrid Metropolitan Area (MMA), the Majadahonda site (MJDH),  
40 in July 2016. Simultaneously, measurements of an extensive number of air quality and  
41 meteorological parameters were carried out at 3 supersites across the MMA. Furthermore,  
42 data from O<sub>3</sub>-soundings and daily radio-soundings were also used to interpret atmospheric  
43 dynamics.

44 The results demonstrate the concatenation of venting and accumulation episodes, with  
45 relative lows (venting) and peaks (accumulation) in O<sub>3</sub> surface levels. Regardless of the episode  
46 type, fumigation of high-altitude O<sub>3</sub> (arising from a variety of origins) contributes the major  
47 proportion of surface O<sub>3</sub> concentrations. Accumulation episodes are characterised by a  
48 relatively thinner planetary boundary layer (< 1500 m at midday, lower in altitude than the  
49 orographic features), light synoptic winds, and the development of mountain breezes along

50 the slopes of the Guadarrama Mountain Range (located W and NW of MMA, with a maximum  
51 elevation of >2400 m above sea level). This orographic-meteorological setting causes the  
52 vertical recirculation of air masses and enrichment of O<sub>3</sub> in the lower tropospheric layers.  
53 When the highly polluted urban plume from Madrid is affected by these dynamics, the highest  
54 O<sub>x</sub> (O<sub>3</sub>+NO<sub>2</sub>) concentrations are recorded in the MMA.

55 Vertical O<sub>3</sub> profiles during venting episodes, with strong synoptic winds and a deepening of the  
56 planetary boundary layer, reaching >2000 m above sea level, were characterised by an upward  
57 gradient in O<sub>3</sub> levels, whereas a reverse situation with O<sub>3</sub> concentration maxima at lower levels  
58 was found during the accumulation episodes due to local/regional production. The two  
59 contributions to O<sub>3</sub> surface levels (fumigation from high altitude strata and local/regional  
60 production) require very different approaches for policy actions. In contrast to O<sub>3</sub> vertical top-  
61 down transfer, UFP are formed in the lowest levels and are transferred upwards progressively  
62 with the increase of the planetary boundary layer.

63

64 **Keywords:** Ozone, ultrafine particles, photochemical pollution, air quality, vertical profiles.

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## 67 **1. Introduction**

68 The EU Directive 2008/50/EC on ambient air quality, amended by Directive 2015/1480/EC,  
69 establishes the need to comply with air quality standards to protect citizens and ecosystems. If  
70 these are not met, plans to improve air quality must be implemented by the national, regional,  
71 and local administrations. Despite the considerable improvements in air quality during the last  
72 decade, non-compliance with European air quality standards is still reported in most Europe.  
73 In particular, the limit values for nitrogen dioxide (NO<sub>2</sub>), particulate matter (PM<sub>10</sub> and PM<sub>2.5</sub>)  
74 and the tropospheric ozone (O<sub>3</sub>) target value are frequently exceeded (EEA, 2017). Therefore,  
75 in 2013, the National Plan for Air Quality and Protection of the Atmosphere (Plan AIRE), 2013-  
76 2016, was drawn up, and approved by the Spanish Council of Ministers' Agreement of 12 April  
77 2013.

78 The EEA (2017) recently reported that, in 2015, 80% of the urban EU-28 population was  
79 exposed to PM<sub>2.5</sub> levels exceeding the WHO guideline, and 90% to that of O<sub>3</sub>.

80 Measures to effectively reduce NO<sub>2</sub> and primary PM pollution are relatively easy to identify  
81 (such as abating industrial, shipping, and traffic emissions with catalytic converters for NO<sub>x</sub> and  
82 particulate controls for PM). However, defining policies for abating O<sub>3</sub>, other photochemical  
83 pollutants, and the secondary components of PM is much more complex.

84 Photochemical pollution is a subject of great environmental importance in Southern Europe  
85 due to its climatic and geographical characteristics (Ochoa-Hueso, 2017). Products of this type  
86 of pollution are many, the most noteworthy being tropospheric O<sub>3</sub>, secondary PM (nitrate,  
87 sulphate, and secondary organic compounds), and the generation of new ultra-fine particles  
88 (UFPs) by nucleation (Gomez-Moreno et al., 2011; Brines et al., 2015). In summer, the Western  
89 Mediterranean Basin (WMB), surrounded by high mountains, falls under the influence of the

90 semi-permanent Azores anticyclone. Clear skies prevail under a generalized level of subsidence  
91 aloft, and meso-meteorological processes with marked diurnal cycles dominate. Re-circulation,  
92 strong insolation, and stability in the upper layers favour the production/accumulation of O<sub>3</sub>  
93 (Millán et al., 1997, 2000, 2002; Kalabokas et al., 2008; Giannakopoulos et al., 2009; Velchev et  
94 al., 2011; Sicard et al., 2013) and the emissions of biogenic volatile organic compounds  
95 (BVOCs) (Giannakopoulos et al., 2009).

96 The abatement of tropospheric O<sub>3</sub> levels in this region is a difficult challenge due to its origin,  
97 which may be local, regional, and/or transboundary (Millán et al., 2000; Millán, 2014; Lelieveld  
98 et al., 2002; Kalabokas et al., 2008, 2013, 2015, 2017; Velchev et al., 2011; Sicard et al., 2013;  
99 Zanis et al., 2014), the complexity of the meteorological scenarios leading to severe episodes  
100 (Millán et al., 1997; Gangoiti et al., 2001; Dieguez et al., 2009, 2014; Kalabokas et al., 2017), as  
101 well as the complexity of the non-linear chemical processes that drive its formation and sinks  
102 (Monks et al., 2015, and references therein).

103 This complex context has led to a lack of 'sufficient' O<sub>3</sub> abatement in Spain (and Europe); while  
104 for primary pollutants, such as SO<sub>2</sub> and CO, and the primary fractions of PM<sub>10</sub> and PM<sub>2.5</sub>  
105 improvement has been very evident (EEA, 2017). Thus, the latest air quality assessment for  
106 Europe (EEA, 2017) shows that: i) there has been a tendency for the peak O<sub>3</sub> concentration  
107 values (those exceeding the hourly information threshold of 180 µg/m<sup>3</sup>) to decrease in recent  
108 years, although not enough to meet the WHO guidelines and EC standards; and ii) the problem  
109 of O<sub>3</sub> episodes is more pronounced in the South than in Northern and Central Europe.  
110 Likewise, O<sub>3</sub> levels are higher in rural than in urban areas, both due to i) the generation  
111 process, which requires time since the emissions of urban, industrial and biogenic precursors  
112 to the production of O<sub>3</sub>; and ii) the consumption (NO titration) of O<sub>3</sub> that takes place in urban  
113 areas.

114 Other studies, such as Sicard et al. (2013), Paoletti et al. (2014), Escudero et al. (2014), García  
115 et al. (2014), Querol et al. (2014, 2016), and EMEP (2016), also evidenced that there is a  
116 general tendency for O<sub>3</sub> to increase in urban areas, including at traffic sites, probably due to  
117 the greater reduction of NO emissions relative to NO<sub>2</sub> and, therefore, a lower NO titration  
118 effect. This trend in decreasing NO/NO<sub>2</sub> ratios from diesel vehicle emissions (the main source  
119 of NO<sub>x</sub> in urban Europe) has been widely reported (i.e., Carslaw et al., 2016). It has also been  
120 found that regional background O<sub>3</sub> levels have remained constant over the last 15 years, while  
121 acute episodes have been drastically reduced compared to the late 1990s, although these  
122 markedly increase during heat waves, such as those in the summers of 2003 and 2015 (EEA,  
123 2017; Diéguez et al., 2009, 2014; Querol et al., 2016).

124 A recent study (Saiz-Lopez et al., 2017) reported an increase of 30-40% in ambient air O<sub>3</sub> levels,  
125 along with a decrease of 20-40% in NO<sub>2</sub>, from 2007 to 2014 in Madrid, which may have led to  
126 large concentration increases of up to 70% and 90% in OH and NO<sub>3</sub>, respectively, thereby  
127 changing the oxidative capacity of this urban atmosphere. We still do not know if this increase  
128 is due to a decrease in the NO titration effect or to the fact that O<sub>3</sub> formation is dominated by  
129 VOCs since the urban areas are characterized by 'VOC-limited' conditions, and a reduction in  
130 NO<sub>x</sub> emission might yield an increase in O<sub>3</sub> formation.

131 Intensive research on O<sub>3</sub> pollution has been carried out in the Mediterranean since the late  
132 1980s and has been key in understanding the behaviour of this pollutant in Europe. It has also

133 been used to establish current European air quality standards (Millán et al., 1991, 1996a,  
134 1996b, 1996c, 2000, 2002; Millán, 2002; Lelieveld 2002; EC, 2002, 2004; Millán and Sanz, 1999;  
135 Mantilla et al., 1997; Salvador et al., 1997, 1999; Gangoiti et al., 2001; Stein et al., 2004, 2005;  
136 Chevalier et al., 2007; Kalabokas et al., 2008, 2015, 2017; Castell et al., 2008a, 2008b, 2012;  
137 Kulkarni et al., 2011; Velchev et al., 2011; Doval et al., 2012; Sicard et al., 2013; Millán et al.,  
138 2014; Escudero et al., 2014; Zanis et al., 2014; Sicard et al., 2017, among others). The EEA  
139 (2017) reports a clear increase in exceedances of the human protection 8-h O<sub>3</sub> target value in  
140 Southern and Central Europe, which are higher in the Italian Po Valley and Spain, and relatively  
141 lower in Portugal and the Eastern Mediterranean.

142 Focusing on the study area, Diéguez et al. (2009, 2014) describe in detail the temporal and  
143 spatial variation of O<sub>3</sub> levels in Spain. These studies highlight the low inter-annual variability in  
144 regional background stations, as well as the existence of specific areas, such as the Madrid air  
145 basin (MAB), Northern valleys influenced by the Barcelona urban plume, Puertollano basin,  
146 and the interior of the Valencian region, where very high O<sub>3</sub> episodes are relatively frequent,  
147 and point to urban and industrial hot spots as relevant sources of precursors. Recently, Querol  
148 et al. (2016) evidenced that the highest O<sub>3</sub> episodes, with hourly exceedances of the  
149 information threshold for informing the population (180 µg/m<sup>3</sup>) during 2000-2015, occurred  
150 mostly around these densely populated or industrialised areas.

151 Querol et al. (2017) report that the high-O<sub>3</sub> plume transported from the metropolitan area of  
152 Barcelona contributed decisively to the frequent exceedances of the information threshold in  
153 the northern areas of Barcelona during the acute O<sub>3</sub> episodes in July 2015. They also  
154 demonstrate that the associated meteorology was very complex, similar to the vertical  
155 recirculation of air masses scenarios reported by Gangoiti et al. (2001), Millán (2014) and  
156 Diéguez et al. (2014) for other regions of the Western Mediterranean. Regional transport of O<sub>3</sub>  
157 is also very relevant, as well as acute O<sub>3</sub> episodes, which exceeded the information threshold  
158 and were caused largely by regional transport (with large a contribution also from local  
159 formation recirculated during prior days, on top of which an additional smaller local 'fresh'  
160 contribution was added). It is also shown that the vast majority of these exceedances are  
161 recorded in July.

162 In the Eastern Mediterranean, the regional background O<sub>3</sub> levels in the free troposphere and  
163 the boundary layer during summer might regularly exceed 60 ppb, and fumigation of these  
164 upper air masses contributes, on average, to the greatest part of the surface O<sub>3</sub> levels  
165 measured in Greece (Kalabokas et al., 2000; Kourtidis et al., 2002; Kouvarakis et al., 2002;  
166 Lelieveld et al., 2002; Kalabokas and Repapis, 2004; Gerasopoulos et al., 2005). Furthermore, a  
167 number of studies report contributions from the stratosphere to the surface O<sub>3</sub> concentrations  
168 during specific meteorological scenarios in the same region (Kalabokas et al., 2013, 2015;  
169 Zanis et al., 2014; Parrish et al., 2012; Lefohn et al., 2012; Akritidis et al., 2016, among  
170 others). In addition, recent research shows that, during springtime O<sub>3</sub> episodes (April – May)  
171 over the WMB, similar synoptic meteorological patterns might also occur, and that these are  
172 linked with regional episodes, mainly induced by large-scale tropospheric O<sub>3</sub> subsidence,  
173 influencing the boundary layer as well as the ground surface O<sub>3</sub> concentrations (Kalabokas et  
174 al., 2017). However, the most intense episodes in the WMB occur in June-July, according to the  
175 statistics for the 2000-2015 period in Spain presented by Querol et al. (2016).

176 In addition to primary emissions, nucleation or new particle formation (NPF) processes give  
177 rise to relevant contributions to the urban ambient air UFP concentrations, mostly during  
178 photochemical pollution episodes in spring and summer (Brines et al., 2015, and references  
179 therein). Ambient conditions favouring urban NPF are high insolation, low relative humidity,  
180 available SO<sub>2</sub> and VOCs, as well as a low condensation sink potential (i.e., a relatively clean  
181 atmosphere with low surface aerosol concentrations) (Kulmala et al., 2000, 2004; Kulmala and  
182 Kerminen, 2008; Sipilä, et al., 2010; Salma et al., 2016).

183 In this study, we evaluate the temporal and spatial variability of O<sub>3</sub> and UFP in the MAB (04-  
184 20/07/2016), to investigate the causes of acute summer episodes of both pollutants and  
185 possible inter-relationships. In a subsequent twin article, we will focus on the phenomenology  
186 of UFP nucleation episodes linked with these photochemical events. Data on UFPs are  
187 included in this paper only where they assist in interpreting the behaviour of O<sub>3</sub>.

188

## 189 **2. Methodology**

### 190 **2.1. The study area**

191 The MAB and the Madrid Metropolitan Area (MMA) are located in the central plain, or Meseta,  
192 of the Iberian Peninsula at around 700 m above sea level (m a.s.l.). Regarding the topographic  
193 features, the Guadarrama range, which runs in the NE-SW direction, reaches heights of up to  
194 2400 m a.s.l. and is located 40 km north from the MMA. To the S, are the Toledo Mountains  
195 which run from E to W (Figure 1). Lower mountains, located to the NE and E, are part of the  
196 Iberian range. Consequently, the Madrid plain shows a NE-SW channelling of winds, forced by  
197 the main mountain ranges, following the basin of the Tagus River and its tributaries. In  
198 particular, the MMA is located to the NE of the river basin and on its E side.

199 Climatologically, the area is characterised by continental conditions with hot summers and  
200 cold winters, with both seasons typically being dry. Mean annual precipitation of  
201 approximately 400 mm is mainly concentrated in the autumn and spring. The MMA is one of  
202 the most densely populated regions in Spain, with more than 5 million inhabitants, including  
203 Madrid City and surrounding towns. According to Salvador et al. (2015), the main  
204 anthropogenic emissions are dominated by road traffic and residential heating (in winter), with  
205 minor contributions from industry and a large airport.

206 Figure 2 shows the time series of the recorded meteorology, measured at a surface station  
207 representative of the conditions in the MMA during the field campaign of July 2016 (El Retiro,  
208 in central Madrid). In order to put the field campaign into the context of the more general  
209 meteorological situation, the time series is extended backwards to the end of June and  
210 forward to the end of July 2016. Figure 2 also shows the corresponding time series for O<sub>3</sub>, NO<sub>2</sub>,  
211 and O<sub>x</sub> concentrations in the MMA, demonstrating the occurrence of well-marked peaks  
212 alternating with relatively low O<sub>3</sub> and O<sub>x</sub> concentrations periods. The intensive field campaign  
213 (11-14/07/2016, marked with a green frame) coincides with a low O<sub>3</sub> interval preceding a  
214 higher O<sub>3</sub> period in the last two days. Red and blue frames in Figure 2 show days in which high-  
215 resolution O<sub>3</sub> free soundings were performed (red and blue indicating intervals within high and  
216 low O<sub>3</sub>, respectively).

217

## 218 2.2. Monitoring sites and instrumentation

219 To characterise acute summer episodes of O<sub>3</sub> and UFP and to investigate their possible  
220 relationships, we devised an intensive field campaign in the MMA. Three measurement  
221 supersites in and around Madrid, following a WNW direction, according the previously  
222 described dynamics, were deployed in an area where the highest levels of O<sub>3</sub> (with hourly  
223 maxima sporadically exceeding 180 µg/m<sup>3</sup>) are usually recorded (Reche et al., 2018, submitted)  
224 inside the MAB (Figure 1). Table 1 shows the equipment available at the three following  
225 supersites:

- 226 • Madrid-CSIC, located at the Spanish National Research Council headquarters. This site is  
227 located in central Madrid on the sixth floor of the building of the Instituto de Ciencias  
228 Agrarias.
- 229 • CIEMAT, located at the Centro de Investigaciones Energéticas Medioambientales y  
230 Tecnológicas headquarters, 4 km in a WNW direction from the CSIC site in a suburban  
231 area.
- 232 • MJDH-ISCI, located in the Instituto de Salud Carlos III in Majadahonda, 15 km in a NW  
233 direction from the CSIC site.

234 At MJDH-ISCI, a PTR-ToF-MS (Proton Transfer Reaction-Time of Fly-Mass Spectrometry) was  
235 deployed from 04 to 19/07/2017 and provides insights into the O<sub>3</sub> Formation Potential (OFP)  
236 of the VOC mixture over the MMA area. The operation procedure of the PTR-ToF-MS and OFP  
237 calculation are detailed in Table S1 and Figure S1.

238 Furthermore, from 11 to 14/07/2016, 28 profiles of pollutant and meteorological parameters  
239 up to 1200 m above ground level (m a.g.l.) were obtained using tethered balloons and a fast  
240 winch system (Figure S1, Tables 2). The instrumentation attached to the balloons is  
241 summarised in Table 1. The profiles were performed at the Majadahonda Rugby Course  
242 (MJDH-RC Figure 1). The balloons were equipped with a Global Position System (GPS) and a set  
243 of instruments (Figure S3), including:

- 244 • A miniaturized CPC (Condensation Particle Counter built by Hanyang University, Hy-CPC)  
245 was used to measure the number concentration of particles larger than 3 nm (PN<sub>3</sub>) with a  
246 time resolution of 1 s and a flow rate of 0.125 L/min, using butanol as a working fluid (Lee  
247 et al., 2014). Previous inter-comparison studies with conventional CPCs have yielded very  
248 good results (with r<sup>2</sup> reaching 0.65-0.98 and slopes 0.87-1.23, Minguillón et al., 2015). In  
249 this work, we will use the terms UFP and PN<sub>3</sub> as equivalents, but we measure  
250 concentrations between 3 and 1000 nm strictly while UFP is <100 nm. However, 80% of  
251 the total particle concentration falls in the range of UFPs.
- 252 • An O<sub>3</sub> monitor (PO3M, 2B Technologies) was used to determine O<sub>3</sub> concentrations. It was  
253 calibrated against an ultraviolet spectrometry reference analyser showing good  
254 agreement (n=34; PO3MO<sub>3</sub>=1.1058\*RefO<sub>3</sub>+4.41, R<sup>2</sup>=0.93). Concentrations (on 10 s basis)  
255 are reported in standard conditions (20 °C and 101.3 kPa) and corrected for the reference  
256 method.

257 In addition to the above instrumentation, we obtained the following additional meteorological  
258 and air quality data:

- 259 • Meteorological data from the CIEMAT meteorological tower (four instrumented levels  
260 between surface and 54 m a.g.l.), as well as from several AEMET (Spanish Met Office)  
261 standard meteorological stations spread out across the basin: Madrid Airport (40.46°N,  
262 3.56°W, 609 m a.s.l), Colmenar Viejo (40.69°N, 3.76°W, 994 m a.s.l), and El Retiro (in  
263 Madrid, 40.40°N, 3.67°W, 667 m a.s.l).
  - 264 • Hourly data for air pollutants (NO, NO<sub>2</sub>, SO<sub>2</sub>, O<sub>3</sub>, PM<sub>10</sub>, and PM<sub>2.5</sub>) supplied by the air  
265 quality networks of the city of Madrid, the Regional Governments of Madrid, Castilla La  
266 Mancha, Castilla y León, and the European Monitoring and Evaluation Programme (EMEP)  
267 monitoring network, all of them supplied by the National Air Quality Database of the  
268 Ministry of the Environment of Spain (MAPAMA).
  - 269 • High-resolution O<sub>3</sub>-sounding data performed by AEMET at midday each Wednesday at  
270 Madrid Airport.
  - 271 • High-resolution meteorological sounding data obtained each day at 00:00 and 12:00 h  
272 local time by AEMET, also at Madrid Airport. They were used to estimate the height of the  
273 planetary boundary layer (PBL) at 12:00 UTC by means of the simple parcel method  
274 (Pandolfi et al., 2014).
- 275 Hourly averaged wind components were calculated and used in polar plots with hourly PM<sub>1</sub>,  
276 PM<sub>2.5</sub>, NO<sub>2</sub>, O<sub>3</sub>, O<sub>x</sub> (O<sub>3</sub>+NO<sub>2</sub>), BC, and UFP concentrations by means of the OpenAir R package  
277 (Carslaw and Ropkins, 2012).

278

### 279 **3. Results**

#### 280 **3.1. Meteorological context**

281 The AEMET O<sub>3</sub>-soundings are represented in Figure 3, where it is evident that the low/high O<sub>3</sub>  
282 periods coincide with the 500 hPa gph passage of, respectively, upper level troughs/ridges  
283 over the area, associated with cold/warm deep advection of air masses. Cold advectons have  
284 usually an Atlantic origin.

285 The local meteorology during the field campaign was characterized by a progressive drop in  
286 temperature (T) (-4°C in the maximal daily T) and an increase in the early morning relative  
287 humidity (RH) (+20%), with insolation remaining constant (maxima of 900-950 W/m<sup>2</sup>) (Figure  
288 4). During the nocturnal and early morning conditions of the first half of the field campaign  
289 (11-12/07/2016), relatively weak northerly winds prevailed at the main meteorological surface  
290 stations inside the basin, including CIEMAT in Figure 4, and Retiro and Colmenar in Figure S4.  
291 This is probably related to drainage (katabatic) conditions inside the MAB, with a progressive  
292 turn to a more synoptic westerly component in the central period of the day, consistent with a  
293 convective coupling with the more intense upper level wind. This coupling is also accompanied  
294 by an important increase in the wind speed at midday, up to 8 m/s (venting stage), that  
295 renewed air masses in the whole basin.

296 During the second half of the campaign, intense and persistent north-easterly winds replaced  
297 the westerlies from the evening of 12/07/2016 on, after the evolution of the upper level  
298 trough. In contrast to the previous period, during 13-14/07/2016, night-time and early  
299 morning conditions registered more intense NE winds (up to 10 m/s) than at midday, after a

300 decrease in intensity down to calm conditions (1 m/s) during the 12/07 morning, facilitating  
301 both fumigation from upper levels and local O<sub>3</sub> photochemical production. A weak wind  
302 veering to the south was also registered at the mentioned surface stations during the 13/07  
303 afternoon, which lasted for only 3 hours, and which is more characteristic of an O<sub>3</sub> enrichment  
304 episode, when the veering lasted longer (Plaza et al., 1997). A progressive decrease of the PBL  
305 height (-600 m difference) is observed in the AEMET daily radio-soundings, in particular,  
306 gradual decreases in the midday PBL height of 3400, 2200, 1900, and 1600 m a.s.l. from 11 to  
307 14/07/2016 (Figure S5) were observed. This decrease is also observed in the 12 and  
308 14/07/2017 UFP profiles (Figures 5 and 6 and S6-S8). As will be detailed later, these  
309 meteorological patterns allowed O<sub>3</sub> and UFP to smoothly and progressively accumulate in the  
310 basin (Figure 4) during the campaign.

311 In the vertical dimension, during both the high and low O<sub>3</sub> periods analysed here, all the  
312 soundings show at midday two well-defined layers separated by a temperature inversion  
313 marking the limit of the growing convection inside the PBL (Figure 3).

314 In high O<sub>3</sub> periods (6 and 27/07/2016), we found lower PBL heights (approximately 1300-1500  
315 m a.s.l.), with weak winds from the E or NE (less than 4-5 m/s) or calm conditions. This is  
316 consistent with the scheme proposed by Plaza et al. (1997), who also describe a rapid  
317 evolution of the PBL height up to 2500-3000 m a.s.l. at 15:00 UTC during their field campaigns  
318 in the area under “summer anticyclonic conditions.” They also describe a morning radiative  
319 surface inversion at around 1000 m a.s.l., which was usually “destroyed 1 hour after dawn,”  
320 containing NE winds associated with nocturnal drainage flows at lower levels (following the  
321 slope of the MAB). In this context, residual layers containing pollutants processed during the  
322 previous day(s) can develop above the stably stratified surface layer during night-time  
323 conditions. These pollutants can be transported towards the S by weak north-easterly winds,  
324 or remain stagnant under calm conditions, which leads to fumigation and mixing with fresh  
325 pollutants emitted at the surface after the destabilization of the surface layer, as we evidenced  
326 in our profiles. These residual layers are topped by the subsidence anticyclonic inversion  
327 (1000-1500 m a.s.l.), according to Plaza et al. (1997).

328 Conversely, the soundings corresponding to low O<sub>3</sub> periods have in common more elevated  
329 PBL heights (2000-2500 m a.s.l.), with more intense winds (above 6-7 m/s) that can blow from  
330 different sectors: from the NE, on 13/07/2016 (with intense N-Westerlies blowing in the free  
331 troposphere), or the S-SW, as observed on 29/06/2016 and 20/07/2016. The O<sub>3</sub>-sounding on  
332 13/07/2016, a unique day within the field campaign, presents the final stage of a low O<sub>3</sub>  
333 period, with winds in the free troposphere having a clear NW component while channelled  
334 north-easterly winds dominate below 2000 m a.s.l. The AEMET free-sounding shows low O<sub>3</sub>  
335 surface concentrations (<45 ppb) and high levels (>70 ppb) in the middle troposphere (3000-  
336 5000 m a.s.l.), associated with very low relative humidity and intense W to NW winds blowing  
337 at that height, which will be discussed in Section 4. The decrease of surface temperature  
338 observed in Figure 2 during the field campaign is also consistent with the cold advection  
339 associated with the troughing in the 500 hPa heights (13/07/2016 in Figure 3).

340

### 341 **3.2. Surface O<sub>3</sub>, O<sub>x</sub>, and UFP during the field campaign**



342 As previously stated, the field campaign was characterised by atmospheric venting conditions  
343 with the two last days marking a transitional period to a more stable anticyclonic episode of  
344 increasing O<sub>3</sub>. The lowering of the wind speed during diurnal periods, and other meteorological  
345 features mentioned above, favoured the gradual accumulation of pollutants, as indicated by  
346 the progressive increase of the O<sub>3</sub> maxima at MJDH-ISCI, where the O<sub>3</sub> maximum was reached  
347 at 15:00 UTC on 13/07/2016 and at 17:00 UTC on 14/07/2016 (Figure 4). The typical  
348 accumulation O<sub>3</sub> cycle for the zone was found only on 13 and 14/07/2016, with a maximum at  
349 14:00 UTC on 13/07/2016 and at 16:00 UTC on 14/07/2016. The two previous days presented  
350 a more irregular daily pattern, indicating unstable and atypical situations for July (perturbed  
351 conditions with the prevalence of synoptic winds). Furthermore, these meteorological  
352 conditions and the high insolation induced the concatenation of NPF episodes in the basin  
353 (with low BC and very high UFP levels at the central hours of the day), such as the one on  
354 13/07/2016 (Figure S9). Morning-midday UFP bursts were caused by nucleation and growth  
355 episodes (we will focus on the phenomenology and the vertical occurrence of these  
356 nucleation-growth events in the twin article).

357 From 11 to 12/07/2016 the highest concentrations of O<sub>3</sub> were recorded for W-SW and W  
358 winds, and peak UFP (PN<sub>3</sub>) concentrations were observed with W, SW, WNW, and NE winds.  
359 However, on 13-14/07/2016, both O<sub>3</sub> and UFP concentrations maximized during calm and NE  
360 winds (see polar plots from Figure S10). PM<sub>2.5</sub> levels were independent of the UFP and O<sub>3</sub>  
361 variation, with concentrations increasing in calm situations in the first two days, and with less  
362 pronounced variations as a function of the wind direction, but somewhat higher concentrations  
363 with NE winds in the last two days (Figure S10).

364

### 365 **3.3. Vertical O<sub>3</sub> and UFP profiles during the field campaign**

366 As shown in Figure S2 and Table 2, the vertical profiles for 14/07/2016 were the most  
367 complete of the campaign (wind speed was relatively low and this allowed extended  
368 measurements throughout the day), and, for that reason, we begin with the description of this  
369 day.

370 Figure 5 shows that there is a rapid growth of the PBL between 08:05 and 11:01 h UTC, as  
371 deduced from the vertical profile of UFP (PN<sub>3-300</sub>) concentrations. At the beginning of the  
372 measurements, the upper limit of the PBL was above 1030 m a.s.l., and in 2 h 40 min it lifted  
373 400 m (around 2.5 m/min). In this initial period, the vertical profile of O<sub>3</sub> was characterized by  
374 a succession of strata of different concentrations, but a clear tendency to increase with height  
375 (around 20 ppb of difference between surface level and 1950 m a.s.l. was observed). The  
376 discontinuity of the PBL ceiling reflected in the UFP, T, and RH profiles did not seem to affect  
377 the O<sub>3</sub> profile at all. In other words, we did not notice accumulation of O<sub>3</sub> layers in the top of  
378 the PBL, but, instead, a general increasing trend towards the highest altitudes reached with  
379 the tethered balloons.

380 Through the course of the day, the profile of concentrations of UFP and O<sub>3</sub> became  
381 homogenous in the lowest 1200 m a.g.l. (this being the maximum height reached), and a  
382 growth of O<sub>3</sub> concentrations at all altitudes was observed until 16:11 h UTC. This  
383 homogenisation and growth of O<sub>3</sub> concentrations in the PBL, caused by intense mixing by

384 convection, resulted in an uneven increase through the day with an increase of 43 ppb at  
385 surface and only 10 ppb at 1900 m a.s.l. (Figures 5 and S6).

386 Figure 6 shows the results from measurements taken at a fixed height (1400-1200 m a.s.l.)  
387 made to capture the effect of the growth of the PBL on O<sub>3</sub> and UFP levels. We started at  
388 approximately 700 m a.g.l. at 09:32 UTC with 60 ppb of O<sub>3</sub> and approximately 6000 #/cm<sup>3</sup>. At  
389 10:25 UTC, the top of the PBL reached the balloon, as deduced from the sharp increase in UFP  
390 concentrations (up to 20000 #/cm<sup>3</sup>). Meanwhile, O<sub>3</sub> concentrations experienced only a slight  
391 decrease, suggesting that O<sub>3</sub> fluxes are top down and not bottom up, as recorded for UFP.  
392 From 16:11 h UTC onwards, a reduction of O<sub>3</sub> levels at lower heights was observed (-50 ppb at  
393 surface levels from 15:55 to 17:45 h UTC, while at 1900 m a.s.l. levels remained stable, Figures  
394 5 and S6).

395 The soundings of 11 to 13/07/2016 again showed a vertical trend characterised by: i) higher O<sub>3</sub>  
396 concentrations at the highest sounding altitude in the early morning, ii) an increase in O<sub>3</sub>  
397 concentrations as the morning progressed (more pronounced at low altitudes), and iii)  
398 homogenous O<sub>3</sub> concentration along the entire vertical profile, except in the surface layers,  
399 where the deposition and titration markedly decreased O<sub>3</sub> levels reached at midday. Detailed  
400 descriptions of these soundings (Figures S7 and S8) can be found in the supplementary  
401 information.

402

#### 403 **4. Discussion**

404 Plaza et al. (1997) show, for the summer period in the study area, that the development of  
405 strong thermal convective activity, and the influence of the mountain ranges produce  
406 characteristic mesoscale re-circulations. On the other hand Crespi et al. (1995) report, also for  
407 summer and the study area, the development of a very deep mixing layer. These authors  
408 report that these re-circulations contribute markedly to the high O<sub>3</sub> episodes recorded in the  
409 region. The arrangement of the Guadarrama range favours the early heating of its S slopes,  
410 which causes a clockwise turning of wind direction, with a NE component during the night, E  
411 and S during the early morning and midday, respectively, and SW during the late afternoon,  
412 thus defining the north-western sector downwind of the city as the prone area for O<sub>3</sub>  
413 transport. Night-time downslope winds inside the basin induce the observed north-easterlies  
414 at lower levels. Influenced by these contributions, the barrier effect of the Guadarrama range  
415 against the N and NW (Atlantic) winds, as well as the repeated clockwise circulation described  
416 above, cause the sloshing of the urban plume of Madrid across the basin. Regarding the  
417 vertical scale, Plaza et al. (1997) also show that fumigation from high O<sub>3</sub>-rich layers (injected by  
418 upslope winds the previous day or days, or transported from other areas outside the MAB)  
419 could also contribute to the enhancement of the surface O<sub>3</sub> concentrations across the basin.  
420 This is attributed to the upward gradient in concentrations in the lower 1 km of the  
421 atmosphere measured in the early morning and the subsequent mixing across the PBL at  
422 midday. On the other hand, Gómez-Moreno et al. (2011) and Brines et al. (2015) report both  
423 intensive summer and winter NPF episodes in the western border of Madrid City, often  
424 simultaneously with the highest O<sub>3</sub> episodes.

425 Considering the free sounding- O<sub>3</sub> profiles in Figure 3, high O<sub>3</sub> concentrations (>70 ppb) can be  
426 observed above the PBL, between 3000 and 5000 m a.s.l., which may be related to larger-scale  
427 transport of pollutants, previously uplifted to the mid-troposphere or originated after a  
428 stratospheric intrusion and a subsequent deep subsidence into the middle troposphere, as is  
429 probably the case based on the ECMWF ERA-Interim reanalysis data. Transport of high O<sub>3</sub> air  
430 masses in the middle troposphere, as for 13/07/2016 in Figure 3, was also documented by  
431 Plaza et al. (1997) over this area in July 1994, during the final phase of a high O<sub>3</sub> period. More  
432 recently, Kalabokas et al. (2013, 2015, 2017), Zanis et al. (2014), and Akritidis et al. (2016),  
433 among others, have shown that similar transport processes of enriched O<sub>3</sub> layers at high  
434 altitude can contribute to increased surface O<sub>3</sub> concentrations during the summer in the  
435 Eastern Mediterranean. This transport has been associated with large-scale subsidence within  
436 strong northerly winds in the Eastern Mediterranean (Etesian winds), and the affected layers  
437 are dryer than average and show negative temperature anomalies. Figure S11 shows the  
438 ECMWF ERA-Interim reanalysis together with the AEMET O<sub>3</sub> free soundings at Madrid airport  
439 for 13/07/2016. The ridging at the lower troposphere over the Bay of Biscay at the rear of an  
440 upper-level trough (left panels) is accompanied by intense NW winds blowing in the middle  
441 and upper troposphere and NE winds at ground level and up to 2000 m (see the radiosonde  
442 profile in the same figure). The O<sub>3</sub> intrusion is associated with the upper-level trough (Sections  
443 A-A and B-B in the figure), and a large area of deep subsidence and extremely low relative  
444 humidity observed within the NW flows over Madrid and to the north of the Iberian Peninsula  
445 and the Bay of Biscay. High O<sub>3</sub> concentrations and low relative humidity of the ERA-Interim  
446 profiles over the airport of Madrid (green and red dotted-lines in the panel “g” of Figure S11)  
447 are in agreement with the radiosonde observations in the same panel.

448 The question now is how much of this O<sub>3</sub> could fumigate at ground level. According to the  
449 radiosonde data, the mixing height top was about 2000 m a.s.l. at midday, but could increase  
450 to about 3100 m a.s.l. after the projection of the surface temperature increase observed  
451 during the afternoon at nearby stations. This height reaches the lower part of the O<sub>3</sub> enriched  
452 layer originated in the tropopause folding. Thus, a certain impact seems likely. However, the  
453 O<sub>3</sub> concentrations were relatively low at all surface stations during that day, as it corresponds  
454 to a vented, low-O<sub>3</sub> period.

455 Thus, according to the O<sub>3</sub> soundings and radio-soundings analysed above, previous evidence  
456 described by Plaza et al. (1997), and the surface air quality measurements presented here,  
457 surface O<sub>3</sub> formation from precursor emissions within the MMA seems to develop in the core  
458 of regional processes, modulated by large-scale meteorological conditions, distinguishing two  
459 types of episodes:

- 460 • ACCUMULATION, occurring in stable, stagnant conditions and regional accumulation of  
461 pollutants (in the sense of Millan et al., 1997, 2000; Gangoiti et al., 2001; Millán, 2014),  
462 with high O<sub>3</sub> reserve strata accumulated during the previous day(s) in residual layer(s) and  
463 associated with fumigation around midday of the following day. The O<sub>3</sub> concentrations are  
464 high along the whole atmospheric column, but enriched in the lower section by additional  
465 local formation of O<sub>3</sub> within the PBL and transport-recirculation of the urban plume of  
466 Madrid around the area. This transport-recirculation is characterised by a net transport to  
467 the NW-N during daytime, after vertical mixing, and to the S and SW during night-time,  
468 inside the residual layer and decoupled from a more stable nocturnal surface layer.

469 Typically, pollutants accumulate during periods of 2-6 days, resulting in well-marked peak  
470 and valley concentration periods that affect background, peri-urban, and in-city stations.  
471 This is the case for the O<sub>3</sub>-soundings of 29/06/2016 (not shown) and, particularly,  
472 27/07/2016 (Figure 7) or the measurements with captive and free balloons by Plaza et al.  
473 (1997) in 1993 and 1994, with very high concentrations of O<sub>3</sub> in the lower atmospheric  
474 layers, usually forming a bump in the vertical profile of O<sub>3</sub> below a height of 2000 m a.s.l.,  
475 easily reachable after daytime convection (Figure 7). As illustrated for 06/07/2017, OFP  
476 (Table S1 and Figure S1) may be largely dominated by the carbonyls (mostly formaldehyde  
477 and acetaldehyde), followed by aromatic compounds (benzene, toluene, and C<sub>8</sub>,C<sub>9</sub>, and  
478 C<sub>10</sub> aromatics) when considering the VOC pool during the morning traffic peaks. The  
479 influence of aromatic VOCs on OFP rapidly decreases, while the influence of biogenic VOCs  
480 (mostly isoprene followed by monoterpenes as primary species, and methacrolein, methyl-  
481 vinyl-ketone, isoprene-derived isomers of unsaturated hydroxy hydroperoxides (ISOPOOH)  
482 and methylglyoxal, as the main secondary species) increases through the day, resulting in a  
483 similar potential influence of biogenic and aromatic VOCs on O<sub>3</sub> formation during  
484 accumulation periods, but with an OFP still dominated by carbonyls.

485 • VENTING, occurring in advective atmospheric conditions (in the sense of Millan et al., 1997,  
486 2000; Gangoiti et al., 2001; Millán, 2014), with O<sub>3</sub>-soundings characterized by (probably  
487 external) contributions from high-altitude O<sub>3</sub> strata and their fumigation on the surface  
488 (episodes 11-14/07/2016). There is no accumulation of pollutants above the stable  
489 nocturnal boundary layer because more intense and steady winds swept out the local  
490 production during the preceding day. OFP contributions of carbonyls (dominating OFP), and  
491 aromatic and biogenic VOCs did not significantly vary for 13 and 14/07/2017 from what is  
492 described above for 06/07/2017.

493 As detailed in Sections 3.1 and 3.2, with the weakening of general atmospheric circulation by  
494 the end of the campaign period, O<sub>3</sub> and UFP smoothly and progressively accumulated in the  
495 basin (Figure 5). An observed decrease of the PBL depth (up to ~1800 m at midday, according  
496 to the AEMET radio-soundings during the campaign, see Figure S5) probably also contributed  
497 to the progressive increase in pollutant concentrations through the campaign.

498 With respect to the vertical variability, the general pattern for UFP (N<sub>3</sub>) clearly showed a rapid  
499 and marked growth of the PBL in the first hours of daylight (Figure 8). In these early stages of  
500 the day, O<sub>3</sub> profiles were characterized by a succession of strata of different concentrations,  
501 but a clear increasing trend towards the higher levels (Figure 8). The discontinuity of the PBL  
502 ceiling, reflected in the UFP, temperature, and humidity profiles, was not identified as such in  
503 the O<sub>3</sub> profiles (Figures 5, 6, and S6 to S8). As the day progresses, the UFP and O<sub>3</sub>  
504 concentration profiles are homogenized and a progressive diurnal growth of O<sub>3</sub> concentrations  
505 occurs until 16:00 or 17:00 UTC (Figure 8), most clearly observed at the surface. This vertical  
506 variability points to different aspects, such as: (i) the relevance of fumigation from high  
507 altitude O<sub>3</sub>-rich strata; ii) surface titration by NO and deposition of O<sub>3</sub>; (iii) surface  
508 photochemical generation of O<sub>3</sub> from precursors (with higher concentrations close to the  
509 surface); and (iv) horizontal O<sub>3</sub> and precursor surface transport from the urban plume of  
510 Madrid towards MJDH-RC. The upper O<sub>3</sub>-rich strata might have an external (to the Madrid  
511 basin) origin or might have been injected regionally at high altitudes on the previous day(s) by  
512 the complex re-circulations of air masses already reported by Millán et al. (1997, 2000, 2002),

513 EC (2002, 2004), Gangoiti et al. (2001), Mantilla et al. (1997), Castells et al. (2008a, 2008b), and  
514 Millán (2014) for the WMB, by McKendry and Lundgren (2000) for other parts of the world,  
515 and by Plaza et al. (1997), and Diéguez et al. (2007, 2014) for the Madrid area.

516 According to the last referenced authors, due to the orientation of the Sierra de Guadarrama  
517 (Figure 1), the heating of its S slopes throughout the day forces the wind direction to veer,  
518 describing an arc that sweeps the zones to the N of Madrid clockwise, from the W to the NE.  
519 Dieguéz et al. (2014) show that the O<sub>3</sub> maxima are recorded at an intermediate point on this  
520 route (El Pardo, Colmenar V., see location in Figures 9 and S12), which is determined by the  
521 wind speed, initial composition of the urban plume, and results of photochemical processes on  
522 its route from the metropolitan area to tens of kilometres away. In addition, our results and  
523 those of Plaza et al. (1997) show that O<sub>3</sub> fumigation from high atmospheric layers decisively  
524 contributes to the increases in the surface levels, since surface concentrations during our  
525 measurements never exceeded those recorded at the highest altitude reached, and at midday  
526 homogeneous O<sub>3</sub> levels are measured across the lower 1.2 km of the PBL.

527 During the whole month of July 2016, there was a clear veering of the urban plume from  
528 Madrid, with night plume transport towards the SW (MJDH-San Martin de V., Figures 9 and  
529 S12), and towards the NW, N-NE, and, in some cases, E-SE during the morning and midday,  
530 followed by the decoupling and onset of the evening and nocturnal flow towards the SW. This  
531 veering seems to be causally associated with the high O<sub>3</sub> levels recorded in the W to E areas  
532 surrounding northern Madrid, since the peak concentrations recorded by the official air quality  
533 network follow this spatial and temporal evolution (Figure S12) for the exceedances of the O<sub>3</sub>  
534 information threshold. These plume impacts occur in periods when the O<sub>3</sub> concentration is  
535 already high because of accumulation from one day to the next in the (same) air mass, which is  
536 not completely renewed due to general circulation conditions. The relevance of the latter has  
537 been recently demonstrated by Otero et al. (2016), who report the maximum temperature as  
538 the parameter more directly related with high O<sub>3</sub> concentrations in Central Europe, whereas,  
539 in the WMB region, the O<sub>3</sub> concentrations were more related to the concentrations recorded  
540 the day before.

541 On the other hand, the differential afternoon-evening decrease of O<sub>3</sub> surface concentrations,  
542 compared with those found at the top of the soundings again demonstrates the relevance of  
543 high-altitude layers and their fumigation to the surface in the hours of maximum convection.

544 Regarding the concentrations of UFP, they were very homogeneous throughout the PBL during  
545 the vertical profiles, especially in the hours of maximum convection, showing a marked  
546 increase from 11 to 14/07/2016 for the whole depth for all profiles (Figure 8). Thus, on the  
547 12/07/2016, the upper limit of the PBL (marked by a sharp reduction in UFP levels) reached  
548 900 and 1200 m a.g.l., respectively, in the soundings conducted at 08:05 and 10:12 UTC (Figure  
549 8). In turn, on 14/07/2016, the top of the PBL exceeded 1200 m a.g.l. only in the afternoon,  
550 being constrained to 300 to 700 m a.g.l. from 08:05 and 10:45 h UTC (also shown in the  
551 progressive loss of -1800 m in the midday PBL height from 11 to 14/07/2016, revealed by  
552 AEMET radio-soundings).

553 The enhanced convection on 12/07/2016 probably favoured the dilution of UFP concentrations  
554 and reinforced the fumigation of O<sub>3</sub> from the upper levels. Conversely, the lower development  
555 of the PBL on 14/07/2016, causing less surface UFP dilution and lower top-down contributions

556 to O<sub>3</sub> surface concentrations, accounted for the opposite O<sub>3</sub> and UFP profiles. Thus, a weaker  
557 development of the PBL might result in the increase of UFP concentrations, even if UFP  
558 emission/formation rates did not vary significantly. However, we cannot discard the possibility  
559 that this UFP increase on the last day was the result of a higher intensity and duration of the  
560 nucleation episodes.

561 Consideration of the evolution of surface O<sub>3</sub> concentrations on 11 and 12/07/2016 (as shown  
562 in Figure 9), depicts a double wave: the first peak around midday (11:00-14:00 UTC on the first  
563 day, and 12:00-13:00 on the second), and the second peak in the afternoon-evening (19-22:00  
564 and 16:00-20:00 UTC, respectively), showing relative peaks (sometimes just a plateau). We  
565 interpret that the morning increase of O<sub>3</sub> concentrations is dominated by both local  
566 production and anthropogenic VOCs (Figure S1), and fumigation of upper levels, with an early  
567 maximum when layers above are rich in O<sub>3</sub>, which progressively decreases with dilution with  
568 surface concentrations. The secondary evening concentration peak corresponds to the  
569 advection of a locally enriched O<sub>3</sub> air mass (titration always causes O<sub>3</sub> depletion towards  
570 nocturnal values). When both processes (morning fumigation and evening advection) are not  
571 so strong, O<sub>3</sub> local production results in a more “typical” diurnal time evolution, with a single  
572 maximum at 15:00-16:00 UTC, as seen on 13-14/07/2016 (Figure 9).

573 The relative importance of the local contribution of the MMA to the O<sub>x</sub> concentrations  
574 registered in the monitoring stations has also been evaluated by comparing the observations  
575 at upwind and downwind locations relative to the city. In this respect, Atazar and Alcobendas  
576 (Figure 9) are located downwind for 11 and 12/07/2016, and MJDH and Fuenlabrada are  
577 upwind, while the opposite occurs for 13 and 14/07/2016. As the urban air mass is transported  
578 towards the E and NE during the first two days, a local O<sub>x</sub> contribution is superimposed on the  
579 background at Atazar and Alcobendas, where recorded O<sub>x</sub> was the highest in the basin (Figure  
580 9). The contrary holds during the next two days, when these sites show lower concentrations  
581 than the rest. MJDH and Fuenlabrada show a reversed behaviour, with lower concentrations  
582 during the first two days and higher for the last days.

583 In addition of the local O<sub>3</sub>, the background contribution can also be very relevant. At high  
584 elevation, changes in the background tropospheric O<sub>3</sub> can be attributed to: (i) hemispheric  
585 background concentrations, (ii) exchange between the free troposphere and boundary layer,  
586 and iii) stratospheric inputs (Chevalier et al., 2007; Kulkarni et al., 2011; Parrish et al., 2012;  
587 Lefohn et al., 2012; Kalabokas et al., 2013, 2015, 2017; Zanis et al., 2014; Akritidis et al., 2016;  
588 Sicard et al., 2017).

589

## 590 **5. Conclusions**

591 The phenomenology of O<sub>3</sub> episodes in the Madrid Metropolitan Area (MMA, Central Iberia)  
592 has been characterised. We found that O<sub>3</sub> episodes linked with precursors emitted in the  
593 Madrid conurbation are modulated by the complex regional atmospheric dynamics.

594 Vertical profiles (up to 1200 m a.g.l.), obtained using tethered balloons and miniaturised  
595 instrumentation at Majadahonda (MJDH), a sub-urban site located on the southwestern flank  
596 of the Madrid Metropolitan Area (MMA) during 11-14/07/2016, showed how complex is O<sub>3</sub>  
597 with altitude and time. Simultaneously, measurements of air quality and meteorological

598 parameters were carried out at 3 supersites within the MMA, where spatial differences  
599 highlight the influence of atmospheric dynamics on different scales.

600 The results presented here confirm prior findings regarding the concatenation of relatively low  
601 (venting) and high (accumulation) O<sub>3</sub> episodes in summer. In the Madrid Air Basin (MAB),  
602 during both types of episodes, fumigation of high altitude O<sub>3</sub>-rich layers (from a remote or  
603 regional origin) contributes a relevant fraction to surface O<sub>3</sub> concentrations. Moreover, we  
604 propose here a conceptual model (shown in Figure 10). To be specific:

605 • Accumulation episodes are activated by a relatively thinner PBL (< 1500 m a.g.l. at  
606 midday), light synoptic winds, and the development of anabatic winds along the slope of  
607 the Sierra de Guadarrama (W and NW of MAB, with >2400 m a.g.l. peaks). This PBL height,  
608 lower than the mountain range, and the development of the mountain breezes cause the  
609 vertical recirculation of air masses, enrichment of O<sub>3</sub> in the lower troposphere, as well as  
610 the formation of reservoir layers that fumigate to the surface as the diurnal convective  
611 circulation develops. This dynamics accounts for the occurrence of the high O<sub>x</sub> (O<sub>3</sub>+NO<sub>2</sub>)  
612 surface concentrations.

613 • During venting episodes, with more intense synoptic winds, and the top of the PBL usually  
614 reaching >2000 m a.g.l., vertical O<sub>3</sub> profiles were characterised by an upward increase in  
615 concentrations (whereas lower-altitude O<sub>3</sub> maxima were observed in the accumulation  
616 periods). Interestingly, vertical profiles demonstrated that, during the study period, O<sub>3</sub>  
617 fumigation (top-down) from upper layers prevailed as a contribution to surface O<sub>3</sub>  
618 concentrations, whereas the increase of UFP takes place bottom-up, progressing with the  
619 development of the PBL and the occurrence of nucleation and growth episodes occurring  
620 within the PBL. Thus, crossing the boundary of the PBL from the free troposphere  
621 increases of UFP concentrations by an order of magnitude, and slight decreases in O<sub>3</sub>  
622 levels were registered. This O<sub>3</sub> and UFP vertical distribution through the day is consistent  
623 with the existence of an efficient venting mechanism which is able to sweep out the local  
624 production of the day. Thus, there is no accumulation of pollutants above the observed  
625 stable nocturnal boundary layer from one day to the next, and new UFP production is  
626 added from below the following day. The presence of O<sub>3</sub>-enriched layers well above the  
627 stable nocturnal boundary layer, transported by sustained intense westerly winds,  
628 suggests a remote origin of this pollutant, after photochemical reactions and uplift  
629 processes developed at least the day before away from the MAB, or stratospheric  
630 intrusions, such as the one documented on 13/07/2016, during the field campaign.  
631 However, surface O<sub>3</sub> concentrations at all stations of the MAB were low during this day;  
632 consequently, even when fumigation from this intrusion was very likely, its air quality  
633 affection was irrelevant for the same days, but the effect in the forthcoming days of  
634 subsided O<sub>3</sub> cannot be evaluated with our tools. The high-O<sub>3</sub> period in the area initiated  
635 the day after, 14/07/2016, and attained its highest concentration on 16/07/2016.

636 The results obtained in this intensive field campaign can be summarized in the following  
637 conclusions and recommendations concerning O<sub>3</sub> abatement policies:

638 • The O<sub>3</sub> source apportionment is very complex, with contributions from local/regional and  
639 remote sources, including the stratosphere. The relative contributions of these might vary  
640 in time and space (e.g. Lefhon et al., 2014).

- 641 • Climate change might reduce the benefits of the O<sub>3</sub> abatement policies (since heat waves  
642 increase O<sub>3</sub> episodes). This, as well as the measures and policies in Northern America and  
643 Asia, will need to be considered in future Europe policies for O<sub>3</sub> mitigation (Lefohn and  
644 Cooper, 2015; Sicard et al., 2017).
- 645 • The phenomenology of O<sub>3</sub> episodes in the WMB is extremely complex, mainly due to the  
646 close coupling between photochemistry processes and mesoscale atmospheric dynamics.  
647 This requires, consequently, abatement policies very different from the ones useful for  
648 Central and Northern Europe, as intensive research has demonstrated in the last decades.
- 649 • In the MAB, during the highest O<sub>3</sub> (accumulation) episodes, in addition to the contribution  
650 (to surface concentrations) by fumigation of upper O<sub>3</sub> (from regional transport,  
651 hemispheric free troposphere O<sub>3</sub>, and intruded stratospheric O<sub>3</sub>, X in Figure 10), there is  
652 an added fraction produced locally and transported-recirculated within the MAB, which  
653 accumulates from one day to the next (Y in Figure 10). If sensitivity analyses demonstrate  
654 that abatement of specific precursors would have an effect on reducing O<sub>3</sub> peaks, then  
655 the reduction strategies (geographic extension, timing, and so on) for decreasing the X  
656 and Y components are very different, and, in most cases, the X component will dominate  
657 the relative contributions. Thus, probably, structural measures over wider regions would  
658 be more effective than local episodic measures (which might have a larger effect on the Y  
659 component). In terms of precursors, the OFP analysis carried out at the ISCIII site shows  
660 that, even if anthropogenic emissions may dominate the O<sub>3</sub> formation through the  
661 potential impact of alkenes and alkanes (not measured) and the high contribution of  
662 carbonyls (formaldehyde and acetaldehyde), biogenic emissions must be considered.  
663 Biogenic VOC (primary and secondary) and aromatic compounds (C<sub>6</sub> to C<sub>10</sub>) contribute to  
664 the same extent to the OFP, according our calculations (Table S1 and Figure S1).
- 665 • The meteorological scenarios causing the summer accumulation episodes in the MAB  
666 (high temperatures, low synoptic winds, and relatively thinner PBL) should be forecast in  
667 order to drive an effective alert system.
- 668 • A more detailed characterisation of O<sub>3</sub> precursors (VOC and BVOCs) in the MAB is  
669 necessary, especially in the source areas, to effectively predict the photochemical  
670 evolution of the plumes, the main impact areas where O<sub>3</sub> from high-altitude reservoir  
671 layers formed the previous day(s) fumigates to the surface levels enriched in O<sub>3</sub> and other  
672 precursors.
- 673 • Modelling techniques and sensitivity analyses will allow the simulation of real conditions  
674 concerning O<sub>3</sub> abatement potential only if the following is achieved in advance: i) the  
675 recirculation cells and other local/regional meteorological processes, such as the  
676 fumigation timing and regional plume transport, are reproduced; ii) a geographically  
677 resolved and accurate emission inventory of O<sub>3</sub> precursors in the source areas and their  
678 temporal modulation is included; and iii) the origin of the high- altitude O<sub>3</sub> strata from  
679 external origins is reproduced.
- 680 • A good combination of regional/local scale modelling, able to reproduce  
681 horizontal/vertical re-circulations of air masses and the behaviour of urban/industrial  
682 plumes in complex topography/meteorology, with modelling able to calculate



683 contributions from long-range transport, free troposphere, and stratospheric O<sub>3</sub> will be  
684 needed to efficiently support policy (see, for example, the ACP special issue on the  
685 Atmospheric Chemistry and Climate Model Inter-comparison Project, ACCMIP  
686 [https://www.atmos-chem-phys.net/special\\_issue296.html](https://www.atmos-chem-phys.net/special_issue296.html); the FAIRMODE initiative,  
687 Thunis et al., 2015; or the Monitoring Atmospheric Composition & Climate (MACC)).

688 The conceptual model described in this study for O<sub>3</sub> episodes in the MMA confirms the  
689 relevance of the vertical re-circulations (on top of the high atmospheric multi-source O<sub>3</sub>  
690 background) that Millan et al. (1997, 2000), Gangoiti et al. (2001), and Millán (2014)  
691 highlighted, controlled in this case by specific synoptic conditions and the PBL depth, may also  
692 be applicable to most of the Western Mediterranean Basin (WMB). Thus, Otero et al. (2016)  
693 demonstrate that, in Central Europe, the highest temperature is the most statistically related  
694 parameter for O<sub>3</sub> episodes, whereas in the WMB it is the O<sub>3</sub> level recorded the day before  
695 (reflecting re-circulation).

696

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953 **FIGURES AND TABLES**

954

955 **Figure Captions**

956 Figure 1. Location of the study area, profiles showing the major orographic patterns and  
957 location of three supersites (CSIC, CIEMAT, ISCIII) and the site where vertical profile  
958 measurements were carried out (MJDH).

959 Figure 2. Top: Hourly meteorological parameters recorded at El Retiro air quality monitoring  
960 station in central Madrid (from 28/06/2016 to 01/08/2016). Middle: Hourly concentrations of  
961 O<sub>3</sub> and O<sub>x</sub> (O<sub>3</sub>+NO<sub>2</sub>) recorded at a selection of air quality monitoring station representing the  
962 Greater Madrid area, together with those from the remote background station of  
963 Campisábalos. Bottom: Hourly NO<sub>2</sub> concentrations recorded at the same sites for the same  
964 period. Periods with available AEMET free-soundings of O<sub>3</sub> are bracketed with red  
965 (accumulation) or blue (venting) squares. The vertical O<sub>3</sub> and UFP profiling campaign is marked  
966 with a green square.

967 Figure 3: Left: Climate Forecast System Reanalysis (CFSR) for the 500 hPa geopotential heights  
968 (gpdams) and mean sea level pressure (MSLP) contours (hPa) at 12:00 UTC (obtained from the  
969 Climate Forecast System reanalysis, Saha et al., 2014) in July 2016 (Wetterzentrale,  
970 <http://www.wetterzentrale.de/>), simultaneous with, Right: AEMET O<sub>3</sub>-free soundings at  
971 Madrid airport.

972 Figure 4. Variation of meteorological parameters (temperature, relative humidity, solar  
973 radiation and wind speed and direction), and levels of NO<sub>2</sub>, NO, O<sub>3</sub>, PM2.5, PM1, BC and UFP  
974 (with lower detection limits of 1, 3 and 7 nm, PN<sub>1</sub>, PN<sub>3</sub> and PN<sub>7</sub>) measured at Madrid-CSIC,  
975 Madrid-CIEMAT and ISCIII, as well as in MJDH-RC from 11 to 14/07/2016.

976 Figure 5. Vertical profiles of levels of O<sub>3</sub>, UFP (PN<sub>3</sub>), temperature and relative humidity  
977 obtained on 14/07/2016 (8:05 to 17:45 UTC). A: Ascending; D: Descending.

978 Figure 6. UFP (PN<sub>3</sub>) concentrations for different vertical profiles obtained on 14/07/2016, as  
979 well as O<sub>3</sub> and UFP during two periods focusing to evaluate changes produced in a fixed height  
980 when reached by the growth of the PBL.

981 Figure 7. Top: Vertical profiles of O<sub>3</sub> levels, and temperature obtained on 12/07/1994 (with  
982 free sounding) and 15/07/1993 (with tethered balloons). Data obtained from Plaza et al  
983 (1997). Bottom: Vertical profiles of O<sub>3</sub> levels of the free soundings by AEMET at Madrid airport  
984 (26.6 km east of MJDH-RC) in 06-07/2017.

985 Figure 8. 11-14/07/2017 profiles of O<sub>3</sub> and UFP (PN<sub>3</sub>) grouped by hourly stretches from  
986 morning to afternoon.

987 Figure 9. Time evolution of hourly O<sub>x</sub> (O<sub>3</sub>+NO<sub>2</sub>) and O<sub>3</sub> concentrations from 11 to 14/07/2016  
988 at selected air quality monitoring sites of the Madrid Basin and an external reference site  
989 (Campisábalos), as well as the locations of these monitoring sites.

990 Figure 10. Conceptual model of the venting and accumulation O<sub>3</sub> episodes in the Madrid Air  
991 Basin, their associated vertical O<sub>3</sub> profiles and the X (fumigation from upper layers, and flows  
992 from free troposphere and stratosphere) and Y (local/regional) contributions to surface O<sub>3</sub>  
993 concentrations in the accumulation episodes.



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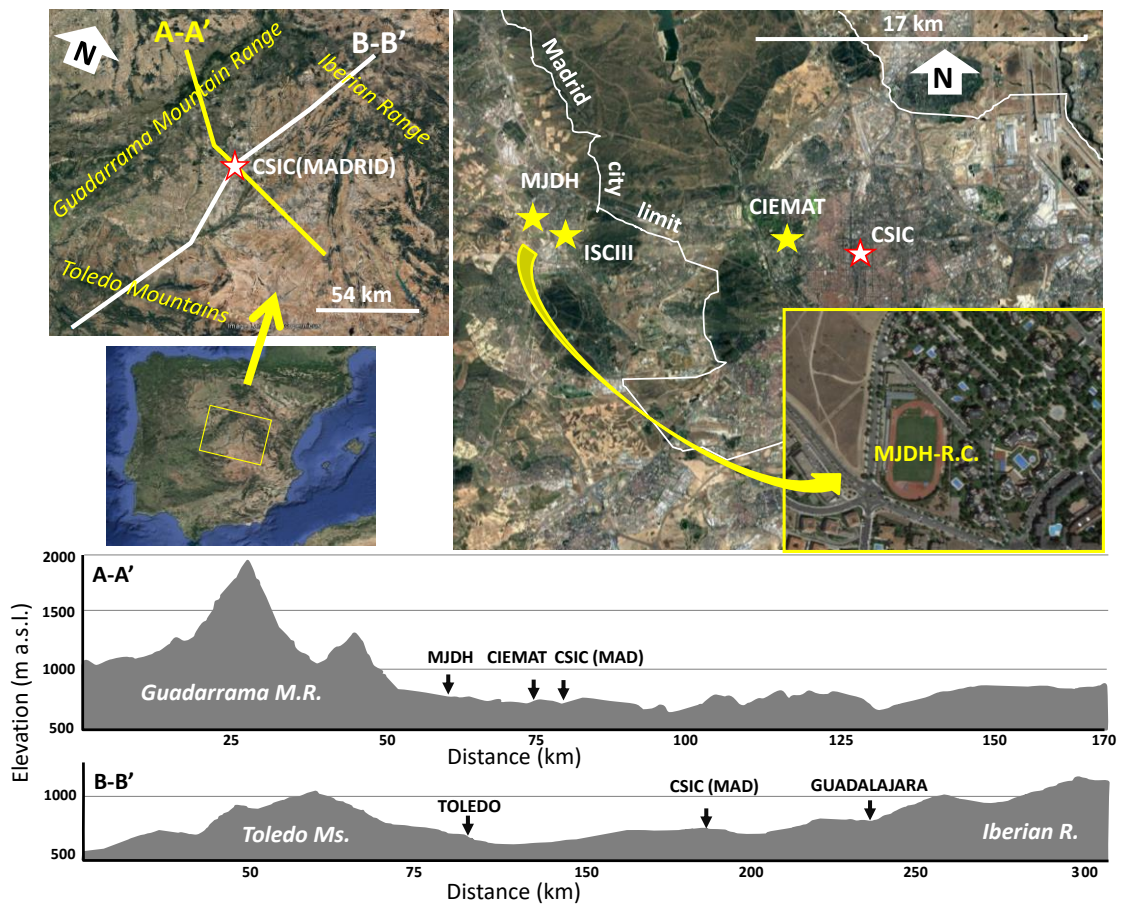


Figure 1

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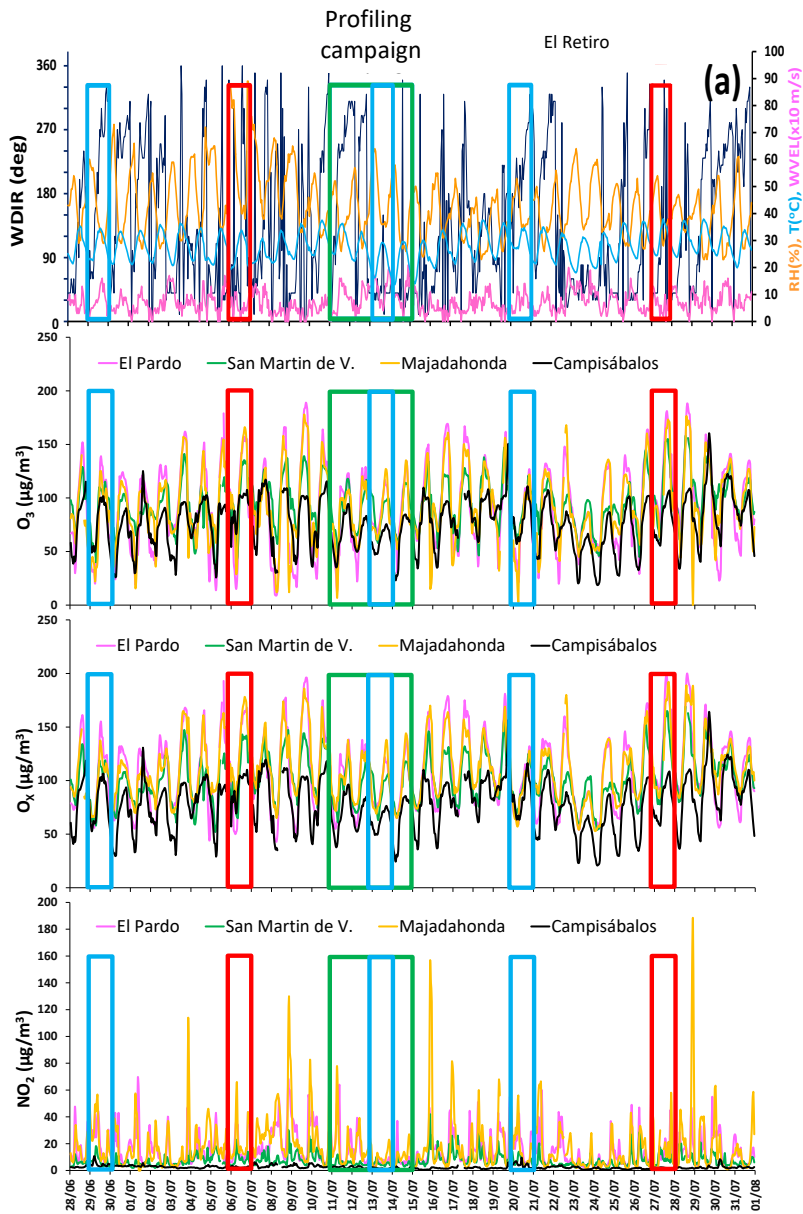
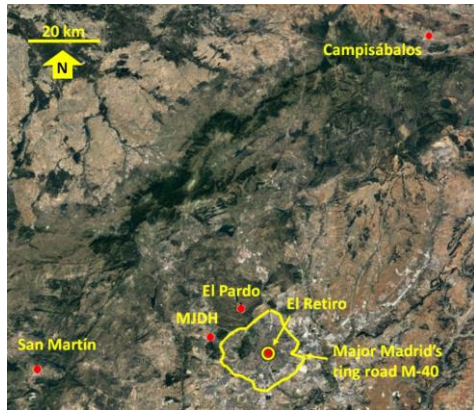


Figure 2

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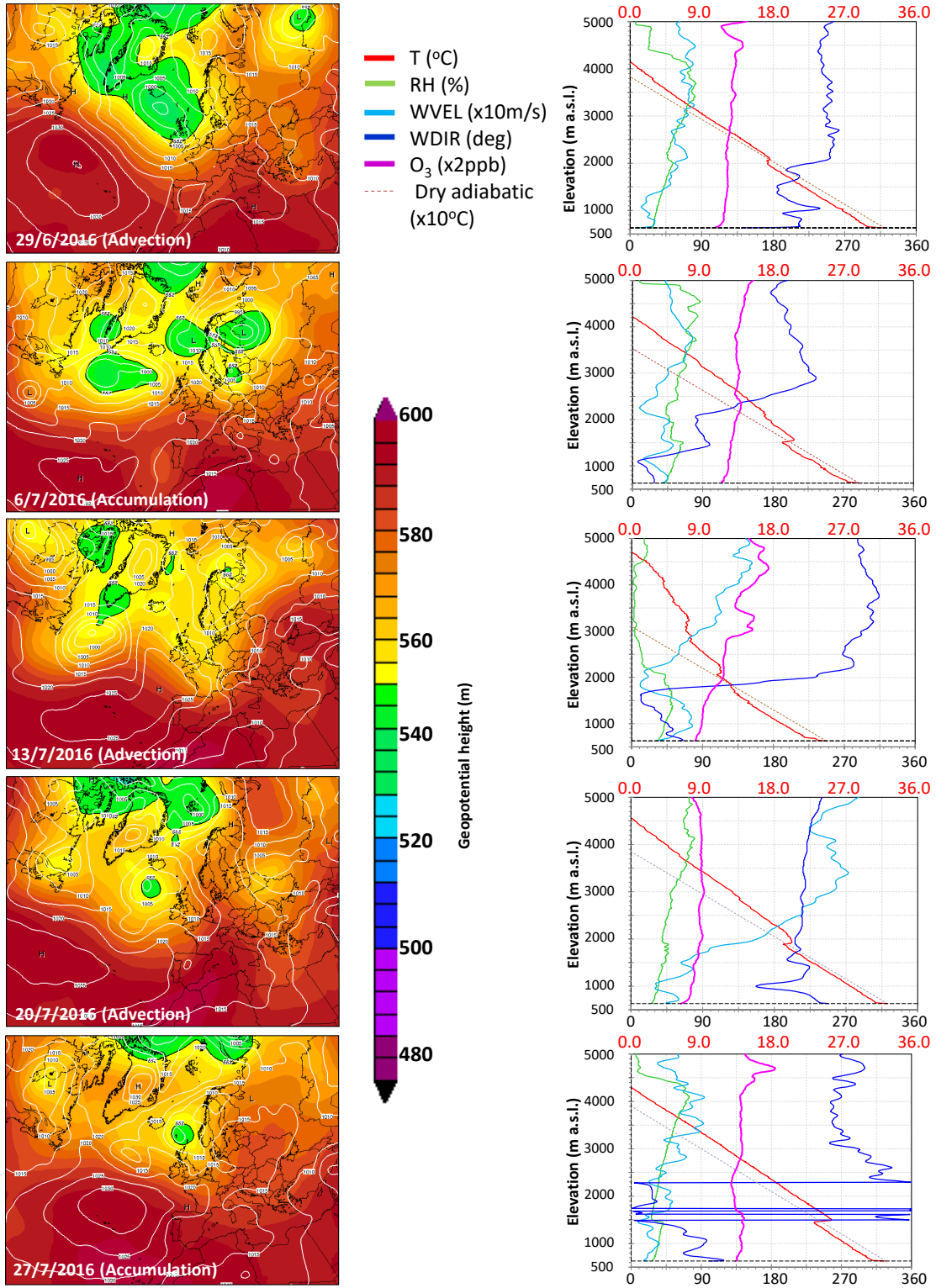


Figure 3

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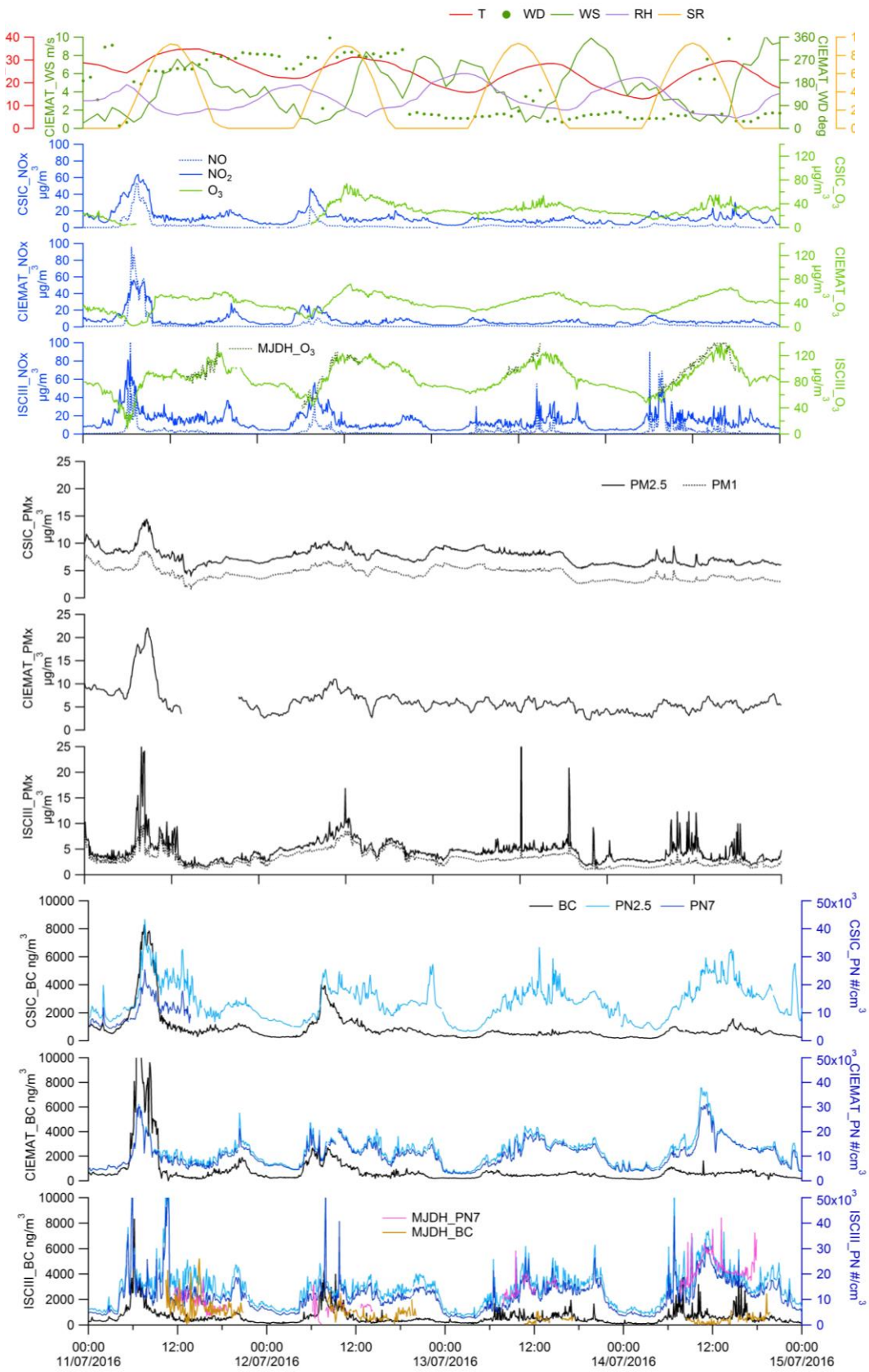


Figure 4

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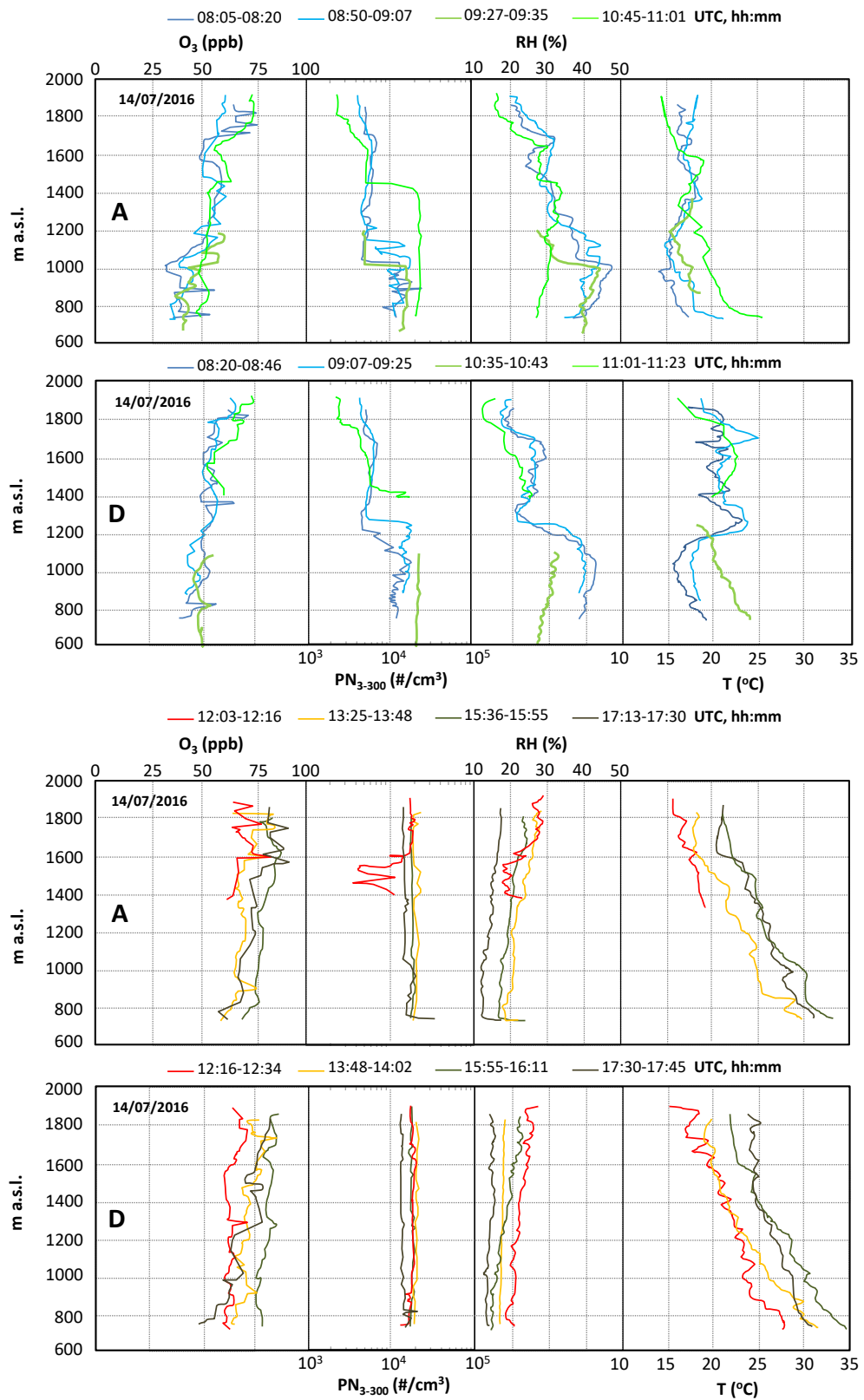
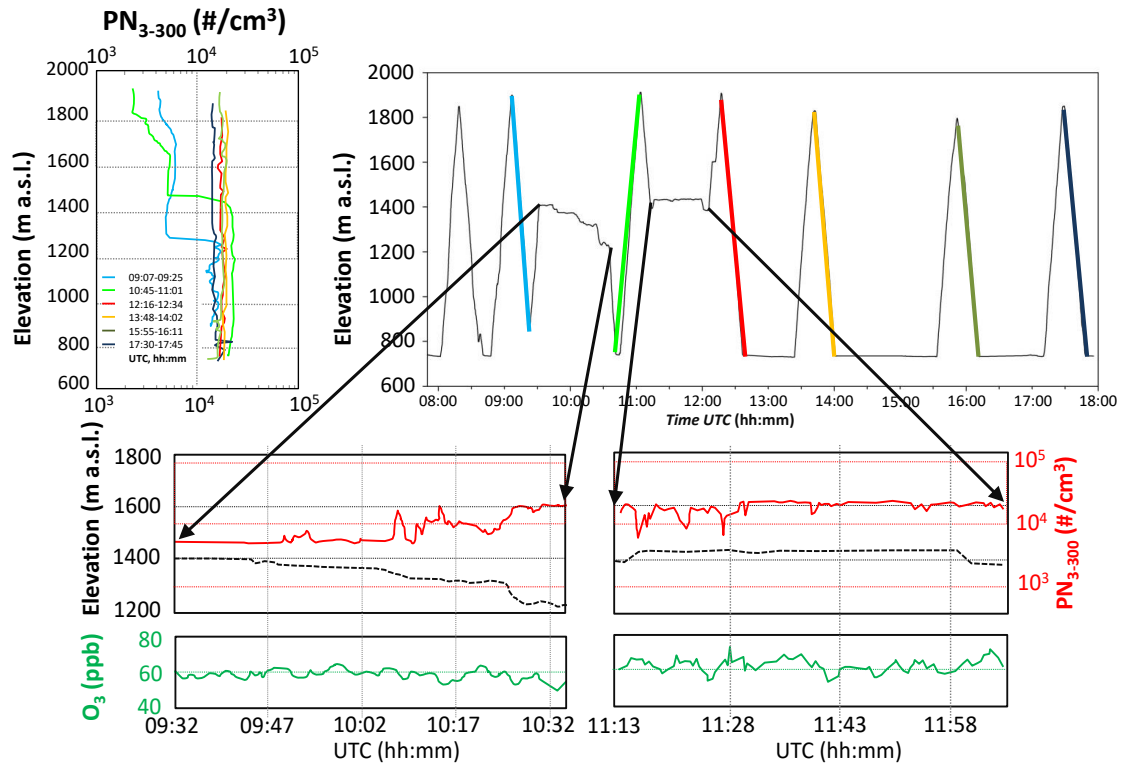


Figure 5



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Figure 6

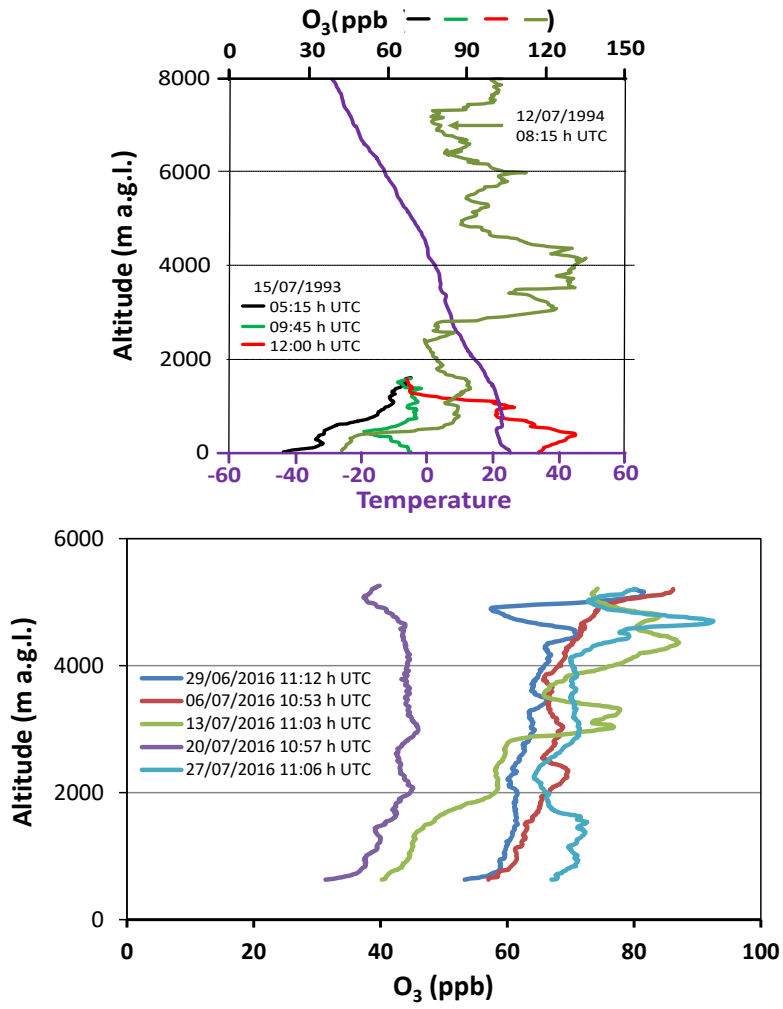
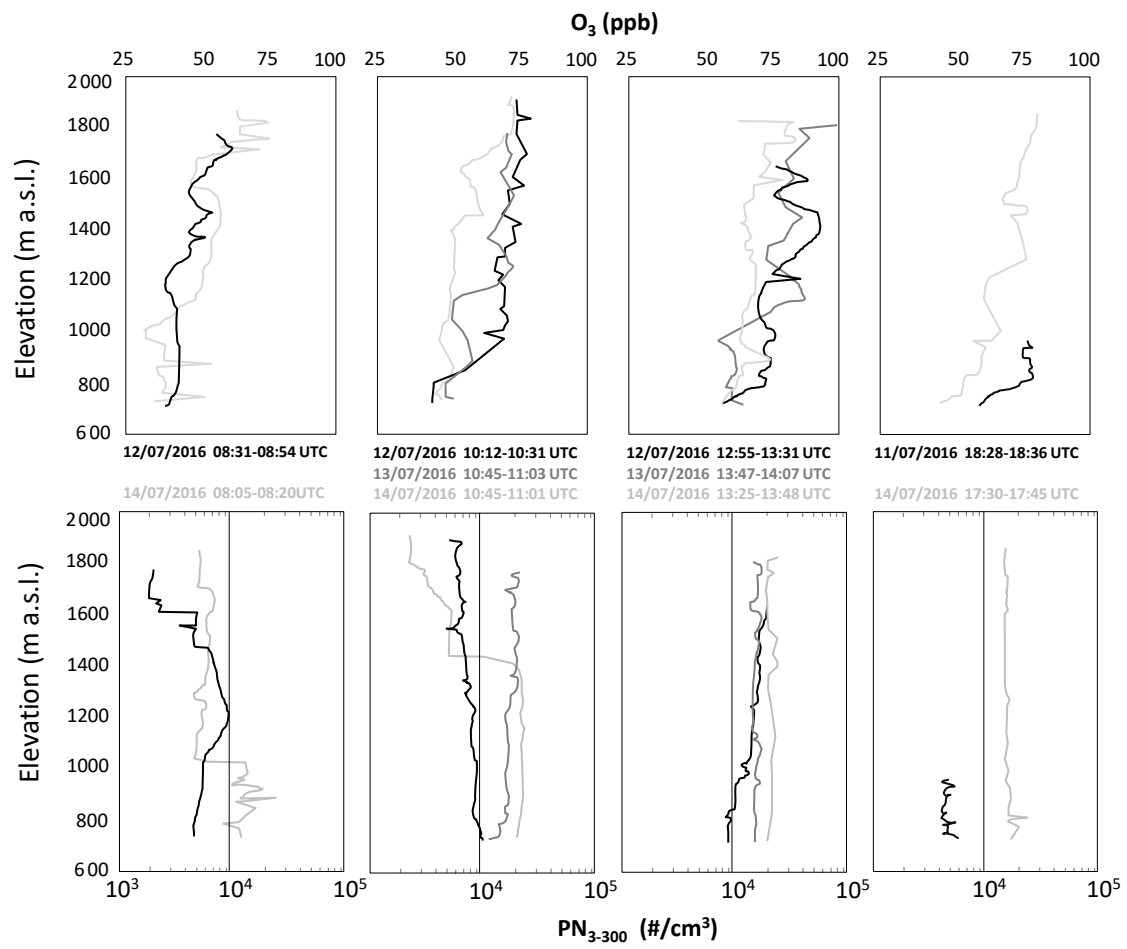


FIGURE 7

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**Figure 8**

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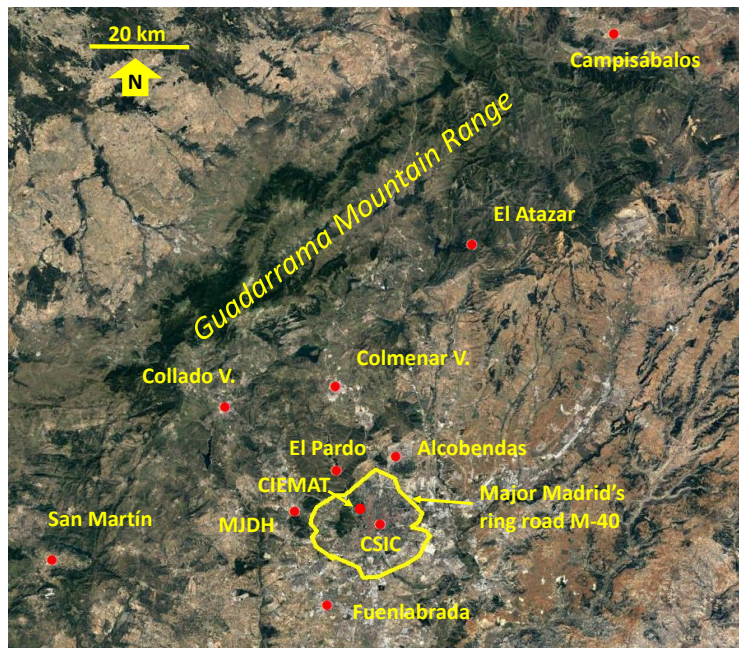
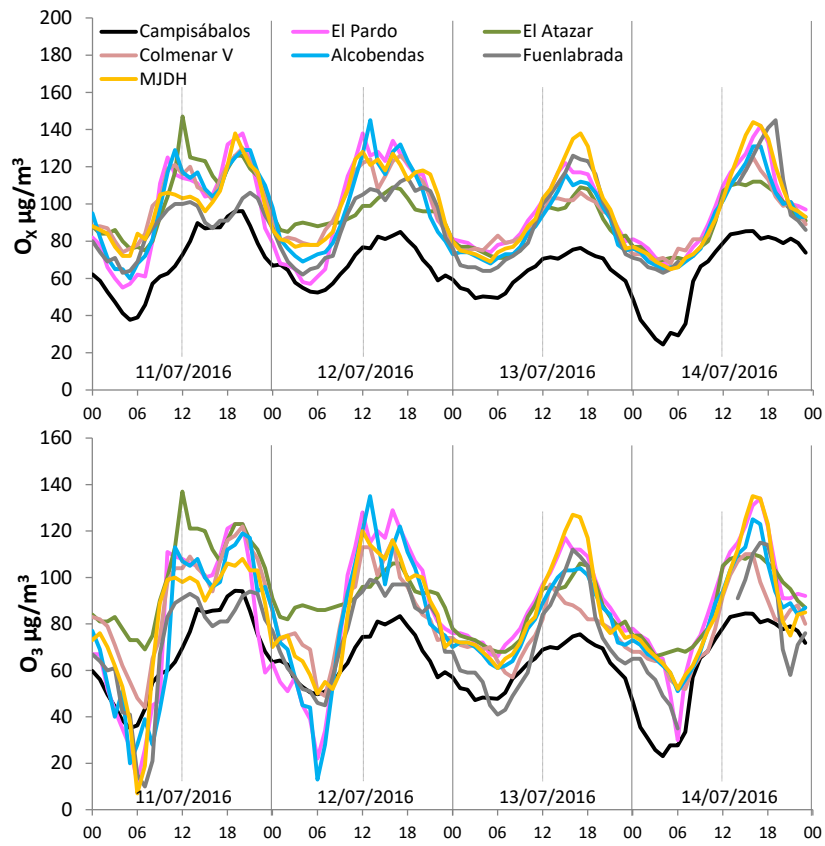
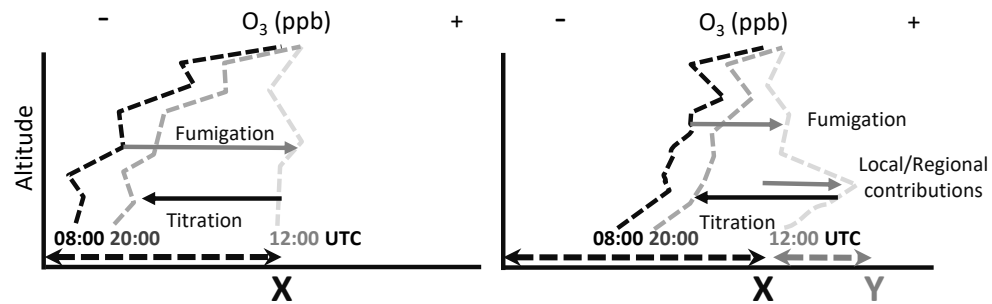


FIGURE 9

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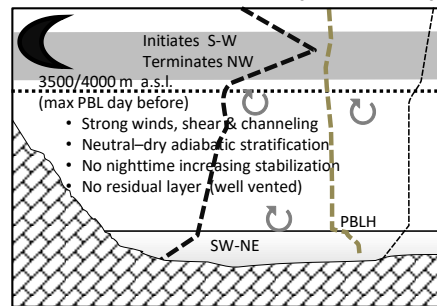


**VENTING/TROUGHING (VT EVENT)      ACCUMULATING/RIDGING (AR EVENT)**

Free troposphere. Strong winds  
Occasional high O<sub>3</sub> peaks (free troposphere O<sub>3</sub>, stratosphere O<sub>3</sub>, long range transport & regional layers)

**Low O<sub>3</sub> (mixed, more external)**  
**Intense ventilation**, no accumulation from the day before  
Mechanical Turbulence

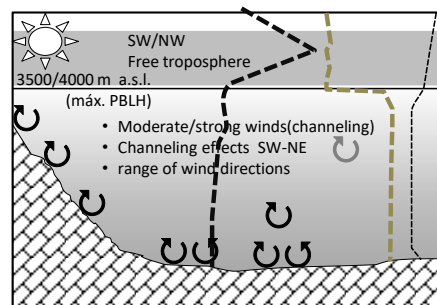
Surface layer, occasionally stably stratified  
**Low O<sub>3</sub> concentrations, titration**  
Strong winds SW-NE



Free troposphere. Strong winds  
Occasional high O<sub>3</sub> peaks (free troposphere O<sub>3</sub>, stratosphere O<sub>3</sub>, long range transport & regional layers)

**No O<sub>3</sub> accumulation in the PBL**  
No re-circulatory winds  
**New O<sub>3</sub> /UFP formation**  
**O<sub>3</sub> fumigation**

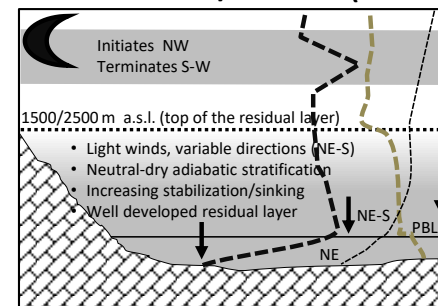
Thicker PBL: > 2000-2500 m a.s.l. at 12:00 UTC  
Rapid growing up to 3500/4000 m  
Intense mechanical & convective turbulence; Intense convection



Free troposphere. Light winds  
Occasional high O<sub>3</sub> peaks (free troposphere O<sub>3</sub>, stratosphere O<sub>3</sub>, long range transport & regional layers)

**Higher O<sub>3</sub> (mixed external + Local)**  
**Low ventilation**, re-circulatory winds, accumulation from the day before  
No Turbulence

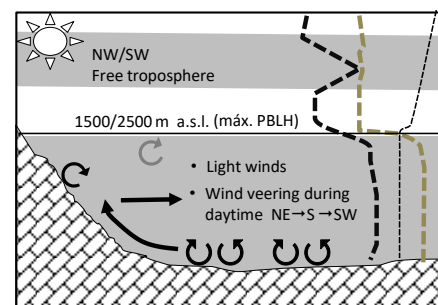
NE'ly jet over stably stratified surface layer  
**Low O<sub>3</sub> concentrations, titration**  
Light winds (NE).



Free troposphere. Light winds  
Occasional high O<sub>3</sub> peaks (free troposphere O<sub>3</sub>, stratosphere O<sub>3</sub>, long range transport & regional layers)

**O<sub>3</sub> accumulation in the PBL**  
Re-circulation over the MMA basin  
**New ozone/UFP formation, O<sub>3</sub> fumigation**

Thinner PBL: < 1500 m a.s.l. at 12:00 UTC  
Slower deepening to 1500/2500 m  
Intense convective turbulence  
**Additional O<sub>3</sub> formation of local origin**  
Thermally driven wind veering NE→S→SW  
Intense convection



**FIGURE 10**

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**TABLES**

Table 1. Details of the instrumentation used in the three supersites and the platform mounted on tethered balloons. BC, black carbon; UFP, ultrafine particles; CPC, condensation particle counter; OPC Optical particle counter; MAAP, Multi-angle Absorption Photometre; PTR-ToF-MS, Proton Transfer Reaction-Time of Fly-Mass Spectrometer.

Site	Latitude (N)	Longitude (W)	Elevation (m a.s.l.)	Parameter (Device-Model)	Operation period
CSIC	40°26'25"	03°41'17"	713	NO <sub>x</sub> (Teledyne API 200EU) O <sub>3</sub> (2B Technologies 202) UFP>2.5nm (CPC-TSI 3775) BC (Aethalometer-AE33) PM <sub>1</sub> (OPC-GRIMM 1107)	09-20/07/2016
CIEMAT	40°27'23"	03°43'32"	669	NO <sub>x</sub> (THERMO 17i) O <sub>3</sub> (THERMO 49i) UFP>7nm (CPC-TSI 3772) UFP>2.5nm (CPC-TSI 3776) BC (Aethalometer-AE33) PM <sub>2.5</sub> (TEOM®) Meteorological tower	04-20/07/2016
ISCIH	40°27'27"	03°51'54"	739	NO <sub>x</sub> (THERMO 17i) O <sub>3</sub> (THERMO 49i) UFP>7nm (CPC-TSI 3783) UFP>2.5nm (CPC-TSI 3776) BC (MAAP-THERMO) PM <sub>1</sub> (OPC-GRIMM 1108) PTR-ToF-MS (HR 8000, Ionicon)(operating procedures described in SI)	04-20/07/2016
MJDH-RC (vertical profiles)	40°28'30"	03°52'55"	729	UFP>3nm (CPC Hy-CPC) O <sub>3</sub> (PO3M™ 2B Technologies) Meteorology (temperature., relative humidty, pressure., wind speed and direction)	11-14/07/2016
El Retiro	40°24'55"	3°41 04°	667	Meteorological parameters	04-20/07/2016

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Table 2. Vertical measurement profiles obtained during 11-14/07/2016 at Majadahonda (MJDH-RC).

<b>Day</b>	<b>Starting</b>	<b>Final</b>	<b>Number</b>	<b>Maximum</b>
	<b>hour (UTC)</b>	<b>hour (UTC)</b>	<b>of profiles</b>	<b>altitude (m a.g.l.)</b>
11/07/2016	18:30	18:45	2	200
12/07/2012	07:02	07:40	2	850
	08:30	09:10	2	1000
	10:10	10:56	2	1100
	11:55	13:43	2	900
13/07/2008	10:45	11:25	2	1000
	11:25	12:00	2	1000
	13:47	14:29	2	1000
	14:29	15:12	2	1100
14/07/2004	08:03	08:44	2	1150
	08:48	10:37	2	1100
	10:46	12:45	2	1200
	13:22	14:02	2	1100
	15:23	16:13	2	1025
	17:12	17:31	2	1100

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