Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2017-1012 Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





Relationship between Chemical Composition and Oxidative Potential of Secondary Organic Aerosol from Polycyclic Aromatic Hydrocarbons

Shunyao Wang 1 , Jianhuai Ye 1 , Ronald Soong 2 , Bing Wu 2 , Legeng Yu 1 , Andre Simpson 2 , Arthur W.H. Chan 1,*

¹Department of Chemical Engineering & Applied Chemistry, University of Toronto

² Environmental NMR Centre, Department of Physical and Environmental Sciences, University of Toronto Scarborough

*Correspondence to: Arthur W.H. Chan (arthurwh.chan@utoronto.ca)

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



1



Abstract

2 Owing to the complex nature and dynamic behaviors of secondary organic aerosol (SOA), its 3 ability to cause oxidative stress (known as oxidative potential, or OP) and adverse health outcomes remain poorly understood. In this work, we probed into linkages between the chemical 4 composition of SOA and its OP, and investigated impacts from various SOA evolution pathways, 5 including atmospheric oligomerization, heterogeneous oxidation and mixing with metal. SOA 6 7 formed from photooxidation of the two most common polycyclic aromatic hydrocarbons 8 (naphthalene and phenanthrene) were studied as model systems. OP was evaluated using the 9 dithiothreitol (DTT) assay. The oligomer-rich fraction separated by liquid chromatography 10 contributed significantly to DTT activity in both SOA systems (52±10% for NSOA and 56±5% 11 for PSOA). Heterogeneous ozonolysis of NSOA was found to enhance its OP, which is consistent with the trend observed in selected individual oxidation products. DTT activities from 12 13 redox-active organic compounds and metals were found to be not additive. When mixing with highly redox-active metal (Cu), OP of the mixture decreased significantly for 1,2-14 naphthoguinone (42±7%), 2,3-dihydroxynaphthalene (35±1%), NSOA (50±6%) and PSOA 15 (43±4%). Evidence from proton nuclear magnetic resonance (¹H NMR) spectroscopy illustrates 16 that such OP reduction upon mixing can be ascribed to metal-organic binding interactions. Our 17 results highlight the role of aerosol chemical composition under atmospheric aging processes in 18 19 determining the OP of SOA, which is needed for a more accurate and explicit prediction of the 20 toxicological impacts from particulate matter.

21

22

23

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



24



1 Introduction

Exposure to particulate matter (PM) has been associated with various adverse health endpoints, 25 26 such as increased risks of myocardial infarction, ischemic heart disease, lung cancer, 27 exacerbation of asthma, and chronic obstructive pulmonary disease (de Kok et al., 2006; Li et al., 2003a; Li et al., 2003b; Nel, 2005; Risom et al., 2005; Thurston et al., 2016). As a result, 28 29 ambient PM_{2.5} exposure ranks among the top 5 global mortality risk factors (Cohen et al., 2017). Meanwhile, a decreased ambient PM level has been associated with longer life expectancies 30 31 (Pope et al., 2009). To establish causal links between aerosol exposure and health endpoints, 32 cytotoxic and carcinogenic potential has been investigated by epidemiological studies in the past 33 decades (Brunekreef and Holgate, 2002; Beelen et al., 2014; Lelieveld et al., 2015; Pope et al., 34 2002), but the underlying mechanistic pathways by which PM causes adverse health outcomes 35 still remain poorly understood. 36 Oxidative stress has been proposed as one of the main mechanisms for PM toxicity in recent years, and is often expressed as the oxidative potential (OP) (Li et al., 2003b). OP is the mass 37 38 normalized capacity of inhaled PM to induce oxidative stress, which is exhibited as redox 39 imbalance through consumption of antioxidants and generation of reactive oxygen species (ROS) 40 (Antiñolo et al., 2015; Shen et al., 2011; Shiraiwa et al., 2012). ROS include a variety of oxidants 41 such as superoxide (O₂*), hydroxyl radical (•OH) and hydrogen peroxide (HOOH), which could either be introduced into human body directly from inhaled PM or generated by targeted cells 42 (Nel et al., 1998; Pöschl and Shiraiwa, 2015; Rhee, 2006; Verma et al., 2015b). The generation 43 44 of ROS during multiphase interactions between air pollutants and human respiratory tract is 45 closely related to the chemical composition, since the combination of various pollutants may

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



46



47 nonlinear influences on its OP (Anti ñolo et al., 2015; Charrier et al., 2015; Fang et al., 2015; 48 Shiraiwa et al., 2012; Xiong et al., 2017). 49 OP of both organic and inorganic PM components have been evaluated by both cellular and 50 acellular assays. In vitro cellular assays were conducted by detecting biological endpoints of the 51 exposure, including heme oxygenase-1 (HO-1) and other cytokines as well as macrophage related biomarker expressions (Krapf et al., 2017; Li et al., 2003b). On the other hand, acellular 52 53 assays use specific chemicals, such as dithiothreitol (DTT), ascorbate (AA) and glutathione 54 (GSH), as surrogates of low-molecular weight (MW) antioxidants (Fang et al., 2016; Godri et al., 55 2011; McWhinney et al., 2013). Among acellular assays, the DTT assay quantifies OP by measuring the DTT depletion rate over a fixed time interval, which mimics the physiological 56 57 process of electron transfer from biological antioxidants to dissolved O₂ (Cho et al., 2005). DTT assay is one of the most commonly used OP evaluation methods, since DTT is a potent surrogate 58 for the total thio-pools (glutathione and protein thiols) while this assay can be conducted under 59 60 biologically relevant conditions (37 °C, pH=7.4) with relatively simple procedures (Cleland, 1964; Hansen et al., 2009; McWhinney et al., 2011). OP levels measured by this assay have been 61 found to correlate well with cellular ROS expression as well as several airway inflammation 62 63 biomarkers, such as HO-1, tumor necrosis factor- α (TNF- α) and fractional exhaled nitric oxide (FE_{NO}) (Delfino et al., 2013; Li et al., 2003b; Tuet et al., 2017). 64 65 66 Secondary organic aerosol (SOA) from atmospheric oxidation of gaseous precursors comprises a major fraction of submicron particulate matter. To date, The DTT assay has been applied in a 67 few studies to evaluate the OP of both laboratory and ambient SOA (McWhinney et al., 2013; 68

influence chemical reactivity as well as bioavailability of PM while having synergistic or

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





Tuet et al., 2017b; Verma et al., 2015a). However, owing to the complex and dynamic property 69 70 of SOA, there is limited understanding of the relationship between detailed SOA composition 71 and its OP (Charrier and Anastasio, 2012; Pöschl and Shiraiwa, 2015; Tuet et al., 2017b). Tuet et 72 al. (2017b) studied the DTT activity of chamber-generated SOA from both biogenic and 73 anthropogenic VOCs under various conditions, showing that naphthalene SOA (NSOA) has the highest OP. Previous work (Antiñolo et al., 2015; Bolton et al., 2000; Charrier and Anastasio, 74 75 2012; Cho et al., 2005; Jiang et al., 2017; Tuet et al., 2017b) indicated while polycyclic aromatic 76 hydrocarbons (PAHs) are unreactive towards DTT while their oxidation products, such as quinones, can be highly redox-active. Quinones can be directly emitted from traffic or formed 77 from secondary oxidation (Cho et al., 2004;McWhinney et al., 2013), and are able to consume 78 79 antioxidants in a catalytic cycle (Fig. 1) (Bolton et al., 2000; Valavanidis et al., 2005). McWhinney et al. (2013) found three quinones (1,2- naphthoquinone, 1,4-naphthoquinone and 5-80 81 hydroxy-1,4-naphthoquinone) in NSOA could only account 30 ±5 % for the observed OP of NSOA, and the source of the remaining DTT activity remains unknown. Peroxides, which are 82 the major contributors to the OP of isoprene SOA (Jiang et al., 2017; Lin et al., 2016; Surratt et 83 84 al., 2010), may also be abundant in NSOA (Kautzman et al., 2009), but their contribution to the OP of NSOA has not been evaluated. In addition, the composition of SOA may evolve upon 85 atmospheric aging. Previous study (Verma et al., 2009) found ambient samples collected in the 86 87 afternoon had a larger fraction of water soluble organic carbon and higher OP, suggesting the photochemical aging effects. There may also be correlations between the average carbon 88 89 oxidation state $(\overline{OS}c)$ and OP of SOA at different stages of oxidation (Tuet et al., 2017a). 90 Moreover, mixing between organics and metals was also found to change the OP of specific 91 components in SOA (Xiong et al., 2017), but the mechanisms remain the tip of iceberg.

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





92

95

96

98

99

100

101

102

103

104

105

106

Here, we focused on understanding how the composition of PAH-derived SOA is related to its

94 strong OP (Tuet et al., 2017b). Specific questions we aim to address in this work are: what are

the compounds within SOA that are important for DTT activity? How does the OP change upon

atmospheric aging processes, including oligomerization, heterogeneous oxidation, and mixing

97 with transition metals (Gao et al., 2004; Rudich et al., 2007)? In this work, SOA from

photoxidation of naphthalene and phenanthrene were studied as model systems, and compared to

SOA from ozonolysis of α -pinene and limonene. The relative OP contributions of peroxides and

high-MW oligomers were evaluated. Effects of aerosol aging on OP was evaluated by examining

the OP of individual oxidation products known to be present in NSOA, and the OP of NSOA

samples that were further oxidized in the condensed phase. Lastly, the impacts of SOA mixing

with metal was explored by mixing SOA or redox-active SOA components with Cu (II), a

transition metal which has been identified with the highest OP in ambient particles (Charrier and

Anastasio, 2012). Additivity of OP revealed by DTT assay was investigated, and the mechanisms

of Cu-organic interactions were examined in detail using proton nuclear magnetic resonance

107 spectroscopy (¹H NMR).

108

109

110

112

113

2 Methods

2.1 Flow tube experiments

111 SOA was produced in a custom-built 10 L quartz flow tube. Details about the flow tube

conditions have been described in previous work (Ye et al., 2016; Ye et al., 2017). Prior to each

experiment, the flow tube was flushed with purified compressed air at a flow rate of 20 L min⁻¹

for over 24 hours.

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



115

116

117

118

119

120

121

122

123

124

125

126

127

128

129

130

131

132

133

134

135

136

137



To produce SOA from naphthalene or phenanthrene, solid PAH was placed in a heated container (80°C), and the sublimed vapor was carried into the flow tube in a 0.2 L min⁻¹ flow of purified compressed air. O₃ and water vapor were also added into the flow tube. O₃ was produced by passing 0.5 L min⁻¹ oxygen (99.6%, Linde, Mississauga, Canada) through an UV O₃ generator (No. 97006601, UVP, Cambridge, UK). Water vapor was produced by bubbling purified air through a custom-made humidifier with a flow rate of 1.3 L min⁻¹. Residence time inside the flow tube was maintained around 5 min. The flow tube was housed inside an aluminum enclosure, which was equipped with 254 nm UV lamps (UVP, Cambridge, UK). The photolysis of O₃ produces O (¹D), which react with water vapor to produce •OH and initiate photooxidation of naphthalene/phenanthrene as well as SOA formation. During naphthalene/phenanthrene photooxidation, O₃ concentration inside the flow tube was controlled around 1 ppm. In addition, blank experiments were conducted under the same conditions, without injecting any hydrocarbons. Two types of SOA from monoterpene ozonolysis were also synthesized under similar conditions. α-pinene (Sigma-Aldrich, 98%) or limonene (Sigma-Aldrich, 97%) was pre-dissolved in cyclohexane (Sigma-Aldrich, 99.5%) with volumetric ratio of 1:500 or 1:1500, respectively. At these ratios, the reaction rates between •OH and cyclohexane are at least a hundred time higher than that of SOA precursors(Atkinson and Arey, 2003; Keywood et al., 2004). The experimental solution was injected continuously into purified air flow by a syringe (1000 mL, Hamilton) installed on a syringe pump (KDS Legato 100) to achieve an initial concentration of 588±16 ppb or 298±24 ppb of α-pinene or limonene, respectively. O₃ was produced by passing oxygen

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





through the O₃ generator at a flow rate of 0.2 L min⁻¹ or 0.3 L min⁻¹ for α-pinene or limonene, 138 respectively, such that the O_3 concentrations were at least 5 times higher than α -pinene or 139 140 limonene. Both experiments were conducted in the same flow tube without irradiation of UV 141 lights. 142 Temperature and relative humidity were monitored by an Omega HX94C RH/T transmitter. The 143 144 concentrations of SOA precursors at the inlet and outlet of the flow tube were measured by a gas chromatography-flame ionization detector (GC-FID, Model 8610C, SRI Instruments Inc., LV, 145 146 USA) equipped with a Tenax® TA trap. Size distribution and volume concentration of SOA at 147 the flow tube reactor outlet were monitored using a custom-built scanning mobility particle sizer 148 (SMPS), which was composed of a differential mobility analyzer column (DMA, Model 3081, 149 TSI, Shoreview, MN, USA) with flow controls and a condensation particle counter (CPC, Model 150 3772, TSI, Shoreview, MN, USA). The SMPS data were inverted to particle size distributions using custom code written in Igor Pro (Wavemetrics, Portland, OR, USA). By assuming a 151 particle density of 1.25 g cm⁻³ for monoterpene SOA (Kostenidou et al., 2007; Shilling et al., 152 2009) and 1.55 g cm⁻³ for PAH SOA(Chan et al., 2009; McWhinney et al., 2011), volume 153 concentrations of particle were converted into mass concentrations and integrated as a function 154 of sample collection time and flow rates to obtain the total mass of collected SOA. 155 156 157 2.2 SOA sampling and extraction 158 All SOA samples were collected on 47 mm prebaked (500 °C, 24h) quartz fiber filters (Pall, Ann 159 Arbor, MI, USA) in a stainless-steel filter holder after reaching a steady state yield, and then wrapped in prebaked aluminum foil before being stored in sterile petri dishes sealed with 160

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



183



Parafilm M[®] at -20 °C. Within 3 days of collection, the filters were extracted in methanol (HPLC 161 grade, 99.9%, Sigma Aldrich, St. Louis, MO, USA), by ultra-sonication at room temperature for 162 163 more than 3 minutes. After sonication, insoluble materials were filtered by a PTFE 164 (polytetrafluoroethylene) syringe filters (FisherbrandTM) with pore size of 0.22 μm. Chemical 165 analysis and DTT activities of the filter extracts were conducted within hours after extraction. As negative control, filter samples were also collected during blank experiments and extracted in the 166 167 same manner. 168 2.3 DTT assay 169 170 OP of SOA and selected quinone/peroxide standards were quantified by the depletion rate of 171 DTT, an antioxidant that can be consumed by oxidative components in PM (Kumagai et al., 2002). The protocols used in this work are adapted from those of McWhinney et al. (2011; 2013). 172 The SOA extracts were first evaporated to complete dryness in a 5.0 L min⁻¹ of N₂ using a blow-173 off system (N-EVAP, Organomation, USA). Phosphate buffer (0.1 M, pH 7.4) was then added to 174 175 dissolve the SOA to achieve a concentration of 0.2 mM. For quinones, copper(II) sulfate and 176 peroxides, each pure compound was weighed and dissolved in 0.1 M phosphate buffer. The concentration of quinones and copper (II) sulfate solutions was 1 µM, and the concentration of 177 each peroxide solution was 100 µM. The specific solution was then added into multiple wells in 178 179 a 96-well UV plate (Greiner Bio-One, Kremsmünster, AT), and immediately covered with 180 adhesive plate sealer (EdgeBio, Gaithersburg, USA). The plate was then placed in a UV-Vis 181 spectrophotometer (Spectramax 190, Molecular Devices Corporation, Sunnyvale, CA) for 182 incubation. The incubation temperature was maintained at 37 °C inside the spectrophotometer to

mimic human physiological conditions. 0.02 ml DTT was then added into each well to initiate

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





the redox-reactions. At each time point, 0.02 ml of 5, 5'-dithiobis (2-nitrobenzoic acid) (DTNB) 184 185 was added, which immediately consumed all the remaining DTT to form a yellow product, 2-186 nitro-5-thiobenzoic acid (TNB) (Fig.S1). TNB was quantified by the light absorption at a 187 wavelength of 412nm, which was further converted to the DTT amount by calibration curve (Fig. 188 S2b) in order to obtain the DTT decay rate. The reaction was quenched by adding DTNB at different wells at different times (all containing the same initial mixture), allowing for 189 190 quantification of the DTT decay rate over a 30 minute time interval (every 5 minutes for the first 191 10 minutes and then every 10 minutes). Blank control and the calibration curve for DTT 192 quantification are shown in Fig. S2. Here, we used the total DTT decay rate, DTT_t (µM DTT 193 min⁻¹), to report the total oxidative capacity, as well as the mass normalized DTT decay rate, DTT_m (pmol DTT min⁻¹µg⁻¹ organics), to report the OP (Charrier et al., 2016; Jiang et al., 2017; 194 195 Xiong et al., 2017). Detailed information about the chemicals used in this assay is shown in 196 section S1.

197

198

2.4 Quantification of peroxides in SOA

- 199 Quantification of total peroxides in the four types of SOA was conducted using the iodometric-
- 200 spectrophotometric method, which quantifies total aerosol peroxides in all three forms (H₂O₂,
- 201 ROOH, ROOR) (Banerjee and Budke, 1964;Docherty et al., 2005):

$$202 ROOR + 2I^- + 2H^+ \rightarrow 2ROH + I_2 (1)$$

$$I_2 + I^- \rightarrow I_3^- \tag{2}$$

where I⁻ is oxidized to I₂ by peroxides under acidic conditions, and then complexes with remaining I⁻ to form I₃⁻, a compound with brown color detected spectrophotometrically at a wavelength of 470nm. The concentration of each SOA solution was first adjusted to 5mM. 0.02

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





ml of potassium iodide solution (1 g ml⁻¹, KI dissolved in DI water), which provided the I[−] in reaction (1), and 0.02 ml of formic acid (≥95%), which maintained the acidity, were added to 0.16 ml of the SOA solution in each well of a 96-well UV plate. The plate was immediately sealed and incubated for 1 h following the same procedures as in the DTT assay, and the UV-vis absorbance was measured at 470 nm. After testing for the sensitivities of various peroxides in KI assay (Fig. S3), and following the previous work by Kautzman et al.(2009), benzoyl peroxide (≥98%) was chosen to represent peroxides in NSOA and used as standards for mass calibration in this study. All values are reported as mass fraction of peroxides in the total SOA.

2.5 Heterogeneous oxidation

Heterogeneous oxidation of NSOA was conducted by first cutting a filter with freshly collected NSOA into halves (within 3 days of flow tube synthesis and stored at -20 °C). One half of the filter was placed in a sealed container, and an O_3 stream (~3ppm, from the previously mentioned O_3 generator) was passed through the filter at a flow rate of 0.2 Lmin⁻¹. The other half of the filter was treated in parallel with a 0.2 L min⁻¹ flow of N_2 over the same time intervals, to account for evaporation and/or decomposition of SOA components at room temperature. 3 sets of experiments were conducted with exposure times of 1 h, 12 h and 24 h where the O_3 exposure can be determined by,

O₃ exposure=
$$\int_0^t [O_3] dt = \langle O_3 \rangle_t \times t$$
 (3)

where $\langle O_3 \rangle_t$ is the time averaged O_3 concentration at a total flow rate of 0.2 L min⁻¹. The DTT activity of each O_3 -exposed aerosol sample was normalized to that of the corresponding N_2 exposure group. Changes in organic carbon mass on NSOA filters exposed to O_3/N_2 for 1h, 12h

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





and 24h were also monitored with a thermal optical organic carbon/elemental carbon (OC/EC) 230 231 aerosol analyzer instrument (Sunset Laboratory Inc., Tigard, OR, USA). OC/EC content was 232 measured following the IMPROVE OC/EC protocol (Chow et al., 1993). Blanks were measured before each run and subtracted from the sample measurements. 233 234 2.6 Chromatographic Separation of NSOA 235 236 To identify the relative contributions of monomers and oligomers in N/PSOA to DTT activity, 237 the SOA extract was separated using ultra-high performance liquid chromatography (UHPLC), and analyzed using electrospray ionization/Ion Mobility-Time of Flight Mass Spectrometry 238 239 (ESI/IMS-TOF MS, TOFWERK, Switzerland, hereafter referred to as IMS-TOF). SOA methanol extract (30 g L⁻¹) was separated on a reverse phase column (ZORBAX Eclipse Plus 240 C18, Agilent, USA) with an initial mobile phase of 90% DI water and 10% HPLC methanol at a 241 flow rate of 0.15ml min⁻¹ (1290 Infinity II, Agilent, USA). The ratio of water to methanol was 242 gradually adjusted from 9:1 to 1:9 between 25 and 30 min. Separation temperature was set to 30 °C 243 with a pressure setting of 150 bar. The outlet flow was regulated using a LC-MS post column 244 245 flow splitters (Supelco, SigmaAldrich, USA) at a ratio of 30:1. The major flow was collected in two different fractions: the first fraction was collected between 6 and 14 min, and the second 246 fraction was collected between 14 and 33 min for NSOA (3-17min and 17-28min for PSOA). 247 248 DTT assay was conducted on each fraction to assess their OPs. The minor flow injected into IMS-TOF was controlled at 5ul min⁻¹ for mass spectrometric analysis in the negative mode. 249 250 251 A deactivated fused silica capillary (360μm OD, 50 μm ID, 50cm length, New Objective, Woburn, MA, USA) was used as the sample transfer line between the UHPLC and the IMS-TOF. 252

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





The ESI source was equipped with an uncoated SilicaTip Emitter (360μm OD, 50 μm ID, 50 μm 253 tip ID, New Objective, Woburn, MA, US). Charged SOA droplets generated from the tip of the 254 emitter were transferred through a desolvation region by a 1L min⁻¹ N₂ flow at room temperature, 255 256 and ions produced from the evaporated droplets were introduced into the drift tube for ion 257 mobility separation. The IMS drift voltage was set to -1.2 kV for the negative mode. The separation temperature was set to 80±1°C with an operation pressure setting of 1.2 bar for the 258 259 whole mass spectrometer. After separation in the ion mobility region, the ion m/z is measured by high-resolution time-of-flight mass spectrometry within an m/z range of 40 to 800. Resolution 260 (m/dm50) of the time-of-flight mass spectrometer is typically 3500–4000 FWHM at m/z 250 261 262 (Groessl et al., 2015; Krechmer et al., 2016). Spectra recording and data processing of the current study were performed using routines written in Igor Pro (6.37, Wavemetrics, OR, USA): 263 "Acquility" (version 2.1.0, http://www.tofwerk.com/acquility) for raw data acquisition and 264 265 "Tofware" (version 2.5.3, www.tofwerk.com/tofware) for post processing. 266 2.7 ¹H NMR spectroscopy 267 268 ¹H NMR spectroscopy was used to further investigate the mechanism behind the OP reduction upon mixing specific organics and transition metals (Simpson et al., 2011; Simpson and Simpson, 269 270 2014; Smith and van Eck, 1999). NMR measurements were performed on a Bruker Avance III NMR spectrometer (11.7 T), equipped with a 4 channel liquid state (¹H, ¹³C, ¹⁵N, ²H) inverse 271 detection probe (OXI) fitted with an actively shielded Z gradient. Typical parameters used for 272 1D ¹H experiments were: a 9.5 µs ¹H pulse, 64k acquisition points, 14 ppm spectral width and 8 273 274 transients were collected, with a total of 5.7s between scans. Before NMR analysis, each sample was dissolved into deuterium oxide (D2O, Cambridge Isotope Laboratories, 99.96 %) and 275

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





dimethyl sulfoxide (DMSO, Fisher Scientific, 99.9 %) at a ratio of 9:1. Here we used D₂O as a

lock reagent by suppressing it through pre-saturation.

278

283

284

285

286

288

279 In order to monitor the duration for the nuclear spin magnetization returning to an equilibrium

state, NMR relaxation times T1(longitudinal direction), T2 (transverse direction) were analyzed.

281 T1 of the sample was measured through the standard inversion recovery experiment. The delay

periods used for T1 measurements ramped from 0.001 s to 15s in 16 increments with a delay of

60s between scans, which represented > 5 x T1 time to permit full signal recovery. For each

delay period, 16 transients were collected. The T2 of the sample was measured through the

standard Carr-Purcell-Meiboom-Gill (CPMG) sequence. The delay periods used for T2

measurements ramped from 1.2ms to 614ms in 16 increments with a delay of 60s between scans,

which represented > 5 x T1 time to permit full signal recovery. For each delay period, 16

transients were collected. A total of 16 free induction decays were collected for each of the

289 relaxation experiments, and relaxation time calculations were done on Bruker Dynamics Centre

290 (v 2.4.5) using mono-exponential fitting functions (Eq. 3 and 4 below).

291
$$f(t) = I_0 \times [1 - 2e^{-t/T_1}]$$
 (4)

292
$$f(t) = I_0 \times e^{-t/T^2}$$
 (5)

Eq. (4) and Eq. (5) are the fitting functions for T1 and T2, respectively, where "I₀" is the thermal

294 equilibrium state of the overall proton nuclear spin magnetization; "t" is the variable delay time.

295 All the ¹H NMR spectra were collected using TopSpin (Bruker, v 3.2). Post-NMR-data

296 processing was conducted in MestReNova (Mestrelab Resesarch, v 11.0.4) and Origin

297 (OriginLab, v 9).

298

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



321



299 2.8 Statistical analysis 300 Data in this study were interpreted as mean ±standard error of the mean (SEM, n=3), and 301 significance analyses among DTT activities were performed by Student's t-test with a 95% 302 confidence interval. A statistical value of p < 0.05 was considered significant. 303 3 Results and discussion 304 305 3.1 DTT activity of laboratory generated SOA 306 Here, we chose two types of SOA derived from PAHs, naphthalene and phenanthrene, as the 307 model SOA systems. The OP of NSOA has been shown to be the highest among various types of 308 SOA previously studied (Tuet et al., 2017a; Tuet et al., 2017b). At the same time, both NSOA 309 and PSOA contain quinones which are known to be highly redox active and exhibit high OP (Cho et al., 2004; McWhinney et al., 2013). As a comparison, α-pinene and limonene SOA from 310 311 ozonolysis were chosen to represent biogenic SOA derived from monoterpenes. Experimental 312 conditions and SOA yield information are summarized in Table 1. 313 314 The mass-normalized DTT decay rate, DTTm, was applied here for OP evaluation (Fig. 2). Similar DTTm have been reported for NSOA, with values of 118 pmol min⁻¹µg⁻¹ by McWhinney 315 et al. (2013) and 110 pmol min⁻¹ µg⁻¹ by Tuet et al. (2017b) for NSOA generated from chamber 316 photooxidation under dry conditions. The DTTm of α -pinene SOA of 19.1 ± 2.5 pmol min⁻¹ μ g⁻¹ in 317 this study is also consistent with those reported by Tuet et al. (2017b) and Jiang et al. (2017). The 318 319 similar values among the different studies highlight the reproducibility of results from the DTT 320 assay. To the best of our knowledge, the OP of PSOA and limonene SOA from this study are the

first ones reported in the literature. Both SOA derived from PAHs yield a higher OP than the two

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





types of monoterpene SOA. The similarities in OP between limonene and a-pinene SOA (cyclic monoterpenes), and between naphthalene and phenanthrene SOA (PAHs) observed in this study further confirms the hypothesis proposed by Tuet et al.(2017b) that the intrinsic OP of SOA is

closely related to the molecular skeleton of the precursor.

One of the reasons for the high OP exhibited by PAH-derived SOA is the abundance of redoxactive quinone moieties in SOA compounds (Lee and Lane, 2009). The cytotoxic and carcinogenic effects from quinone-like compounds are well recognized in the field of biochemistry (Bolton et al., 2000; Valavanidis et al., 2005), and the toxicity of PM has been attributed to the presence of quinones. Charrier and Anastasio (2012) have found the OP of several quinones are comparable to transition metals in ambient particles. The importance of quinone-like components to OP was also evaluated by examining the changes in DTT activity in response to the presence of 2,4-dimethylimidazole, which has been shown to be the co-catalyst of the quinone redox cycle (Jiang et al., 2017). However, McWhinney et al. (2013) quantified three quinones in NSOA using GC/MS and found that these quinones can only account for 30% of the total NSOA DTT response. The remaining DTT activity may arise from other quinone-like compounds that have not been identified, or from other oxidation products in NSOA. Given this knowledge gap, we examine the potential roles of peroxides, oligomers and other more oxygenated products that may explain the high DTT activities of NSOA in the next sections.

3.2 OP contribution from peroxides

One of the main hypothesis in this study is that organic peroxides contribute to OP. Organic peroxides have been identified to be major components in both laboratory and ambient OA

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





(Jokinen et al., 2014; Lin et al., 2016; Surratt et al., 2010; Zhang et al., 2015; Zhang et al., 2017). 345 346 They can play important roles in forming high-MW oligomers (Docherty et al., 2005) and highly 347 oxygenated molecules (Mentel et al., 2015). Recent studies have shown that peroxides may also 348 be important for OP. Kramer et al. (2016) suggested that isoprene-derived 349 hydroxyhydroperoxide (ISOPOOH) is an essential contributor to the OP of isoprene SOA, consistent with the results of bulk peroxide measurements using 4-nitrophenylboronic acid assay 350 351 (NPBA assay) by Jiang et al. (2017). Since peroxides have been proposed to be a major 352 component in NSOA (Kautzman et al., 2010), it is essential to determine whether or not these 353 peroxides can account for the remaining OP contribution (McWhinney et al., 2013). 354 355 Here we compare the NSOA and α -pinene SOA systems to determine the role of peroxides in OP. The KI assay is known to be sensitive to all types of SOA peroxides (ROOR, ROOH and HOOH) 356 357 (Banerjee and Budke, 1964), and we confirmed its sensitivity by conducting calibrations with 4 different peroxides (Fig. S3). Similar KI response factors were observed with hydrogen peroxide, 358 cumene hydroperoxide, tert-butyl peroxide and benzoyl peroxide. Since it is likely that the 359 360 peroxides in NSOA have one aromatic ring are mostly in the form of ROOR (Kautzman et al., 2009), we used benzoyl peroxide as the mass calibration standard. The mass fraction of 361 362 peroxides and DTTm of each SOA system are shown in Table 2. A high percentage of peroxide 363 (40-100%) was observed in α-pinene SOA, which was consistent with the results (47%) in the 364 study by Docherty et al. (2005). Meanwhile, a very low percentage of peroxides (<3%) was 365 found in NSOA system, a result that is inconsistent with previous work (>20%) by Kautzman et 366 al. (2009). The difference in measured peroxide content is most likely due to the difference in the UV light source. Kautzman et al. (2009) and McWhinney et al. (2013) used H₂O₂ photolysis 367

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



368

369

370

371

372

373

374

375

376

377

378

379

380

381

382

383

384

385



under black lights (~350 nm), whereas in our study 254 nm UV lamps were used to photolyze O₃ and generate O (¹D). Organic peroxides in SOA are known to be photo-labile (Banerjee and Budke, 1964; Krapf et al., 2016; Wang et al., 2011) and had likely decomposed rapidly under the shorter UV wavelengths used in our studies. Despite the differences in light conditions and peroxide content, the DTTm measured for NSOA in this study is consistent with those measured in two separate studies (McWhinney et al., 2013; Tuet et al., 2017). Also, the DTT_m of NSOA was found to be significantly higher than that of α -pinene SOA, which contains a large fraction of peroxides. Therefore, from our work, there is no evidence showing that peroxides contribute significantly to the high DTT_m observed in NSOA. Even if organic peroxides were present at a mass fraction of around 20%, as reported by Kautzman et al., (2010), we expect these peroxides would react with DTT at a similar rate as benzoyl peroxide, which has a similar structure that the proposed peroxides. The DTTm of benzoyl peroxide (ROOR-type) is 38 pmol min⁻¹ug⁻¹, which is around 3 times lower than the DTTm of NSOA. It should also be noted that organic hydroperoxides are the major OP contributors for biogenic SOA, such as isoprene and monoterpene SOA (Jiang et al., 2017). One of the potential mechanisms is the formation of hydroxyl radicals from the decomposition of organic hydroperoxides in water (Tong et al., 2016). For the NSOA system, our results suggest that other non-peroxide species are likely to serve as major contributors to OP.

386

387

388

389

390

3.3 OP of oligomers in NSOA

Atmospheric OA also undergoes extensive oligomerization, forming high-MW compounds that have profound impacts on SOA physicochemical properties (Hallquist et al., 2009; Rudich et al., 2007; Trump and Donahue, 2014; Wang et al., 2011). Laboratory photooxidation of aromatic

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





391 compounds produces a substantial fraction of oligomers in the SOA (Kalberer et al., 2004) and these oligomers may be highly functionalized (Gao et al., 2004; Tolocka et al., 2004;). IMS-TOF 392 393 analysis reveals that a substantial fraction of the signals in NSOA and PSOA are located in the 394 high m/z range, which are associated with high-MW oligomeric products. Since previous studies 395 have largely focused on monomeric quinones (such as 1,2-naphthoquinone or 9,10phenanthrenequinone), the contribution of high-MW products to OP have not been studied and 396 397 may explain the "missing" OP contributors. 398 399 To evaluate OP of high-MW products in NSOA and PSOA, solutions of SOA extract were 400 separated in a C18 reverse phase column into two major fractions. As shown in Fig. 4, when 401 analyzed by IMS-TOF, the first fraction was found to contain relatively higher signals at m/z402 associated with monomers, and the second fraction contain products with higher signals located 403 in a higher m/z range. It should be noted that complete separation could not be achieved in this work. Other techniques, such as size exclusion chromatography (Di Lorenzo and Young, 404 405 2016; Di Lorenzo et al., 2017), may yield better separation based on molecular weights, but may 406 not be able to resolve compounds in the relative low molecular weight range in the current study. Nonetheless, the first fraction can be qualitatively described as a "monomer-rich" fraction, and 407 the second fraction can be regarded as an "oligomer-rich" fraction (Fig. 4, Fig. 5a, b). The DTT 408 409 assay was then conducted on both of the original SOA solution as well as the two separated 410 fractions. Since the amount of organic material in each fraction is not known, we use the total DTT activity (DTTt, in µM min⁻¹) to qualitatively compare the oxidative capacities of the two 411 412 fractions.

413

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





As shown in Fig. 5, both the monomer-rich and oligomer-rich fractions are reactive towards DTT. 414 415 For NSOA, the OP contribution from the monomer-rich fraction and the oligomer-rich fraction 416 were 16±3% and 56±10%, respectively (Fig. 5c). For PSOA, the OP contribution from the 417 monomer-rich fraction and the oligomer-rich fraction were 40±8% and 50±5%, respectively (Fig. 418 5d). In both systems, the oligomer-rich fraction caused a more rapid decay in DTT than the monomer-rich fraction even with a lower summed ion signal of low-MW constituents (Fig. 5a, b 419 420 for NSOA and PSOA, respectively). These qualitative results suggest that while the current focus 421 of health studies has been focused on monomeric quinones, other higher-MW products may be 422 important for the OP of NSOA and PSOA. Specific molecular characteristics of these high-MW 423 OP contributors are currently unknown, and understanding them will be the subject of future research. It is very likely that the oligomers also contain redox-active quinone functional groups, 424 425 such as those formed on the surface of oxidized soot (Anti ñolo et al., 2015), and are therefore 426 important for OP. 427 3.4 OP from heterogeneous oxidation 428 429 In addition to its complexity, the composition of SOA is also highly dynamic and evolves upon atmospheric oxidation (Jimenez et al., 2009). Heterogeneous oxidation in the particle phase is 430 one of the major pathways in aerosol aging (George and Abbatt, 2010; Rudich et al., 2007) and 431 432 can increase oxygen content during the functionalization processes. McWhinney et al. (2013) 433 attributed 21% of the NSOA's DTT activity to two quinone isomers (1,2-NQN and 1,4-NQN) in 434 NSOA while found a higher DTT contribution (30%) when they took 5-hydroxy-1,4-435 naphthoquinone (5-OH-1,4-NQN) into consideration. Thus, we expect that oxygenated

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



436



derivatives produced upon heterogeneous oxidation may also contribute to the OP of SOA, and 437 the OP of SOA could be enhanced by heterogeneous oxidation. 438 439 We first examined the changes in OP with additional functional groups in known organic 440 compounds. As shown in Fig. 6, two pairs of organic compounds in NSOA were chosen: 1,4-441 NQN, and 5-OH-1,4-NQN were used to study quinone-like compounds while naphthol (NPL) 442 and 1,3-dihydroxy naphthalene (1,3-DHN) were used to compare phenol-like compounds. The 443 \overline{OS} c of those four components were calculated (Kroll et al., 2011), as shown in Fig.6. Our results 444 demonstrated a higher DTTm for standards with higher oxidation states. For each addition of an OH group to the selected molecule, the OP increases. OP of an aromatic compound is therefore 445 shown here to be associated with its degree of oxygenation and is demonstrated here 446 447 fundamentally using individual organic compounds. 448 449 More broadly, oxidation also increases the degree of oxygenation in the bulk aerosol phase, and 450 increases OP. Here we conducted heterogeneous oxidation by exposing filter-collected SOA to 451 O₃/ N₂ with the same flow rate. The N₂ exposure group is used as the control group in order to 452 isolate the effects of evaporation and/or decomposition at room temperature from those of 453 heterogeneous O₃ oxidation. For each of the exposure (1h, 12h, 24h), the DTTt of O₃ exposure 454 group was normalized by the DTTt of the corresponding N₂ exposure group (Fig.7a). OC loss 455 was determined by thermal optical OC/EC analysis, and was observed to be 17% and 13% for 456 the O₃ and N₂ exposure groups after 24 hours, respectively (Fig.7b). Generally, DTTt of NSOA 457 filter under O_3 exposure was higher than that of N_2 exposure. The study of Anti ñolo et al. (2015) also showed an increased redox activity of soot accompanied by an increased amount of 458

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



459

460

461

462

463

464

465

466

467

468

469

470

471

472

473

474

475

476

477

478

479

480

481



oxygenated derivatives (quinone) under heterogeneous oxidation. However, the enhanced oxidative capacity from heterogeneous ozonolysis appeared to decrease with longer exposure to O₃ (Fig.7a), which we hypothesize may result from functionalization as well as fragmentation of organic molecules during heterogeneous oxidation (Kroll et al., 2009). This can be further confirmed with the observed changes in the different OC fractions, as shown in Fig. S4. Within the 24h exposure, the volatile fractions (OC1 and OC2) of the O₃ exposed group increased while the less volatile fractions (OC3 and OC4) decreased compared to the N₂ group, which suggested the decomposition of high-MW (low volatility) species into low-MW (high volatility) compounds. Previous work has also shown fragmentation can play a dominant role in a late stage of heterogeneous oxidation (Kroll et al., 2011). The overall increased volatility may lead to evaporation of smaller redox-active molecules and decrease the DTTt compared to the N2 exposure group. It should also be noted that the O₃ concentrations to which NSOA are exposed here are about 100 times higher than typical ambient levels (Finlayson-Pitts and Pitts Jr, 1999). Assuming heterogeneous oxidation mechanisms are linear and an ambient O₃ concentration of 30 ppb, O₃ exposure for 1h, 12h, and 24h under our experimental conditions represent 4, 50, 100 days of aging in the atmosphere. Therefore, we anticipate an overall enhancement in OP under ambient conditions. Though similar observations have been made in soot (Antiñolo et al., 2015) and diesel exhausted particles (McWhinney et al., 2013), this enhancement in OP by heterogeneous oxidation is shown here for the first time in SOA particles. 3.5 OP changes upon mixing with Cu Ambient PM forming from mixed sources is frequently composed of both organics and metals. To date, both organics (quinone) and transition metals (Cu, Fe, Mn etc.) have been shown to be

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



482

483

484

485

486

487

488

489

490

491

492

493

494

495

496

497

498

499

500

501

502

503

504



redox-active (Charrier and Anastasio, 2012; Xiong et al., 2017). Metals in ambient particles can range from insoluble substances to soluble cations, leading to various health outcomes after deposition onto the human respiratory tract (Gojova et al., 2007; Oberd örster et al., 2005). Based on the chemical composition and the assumption that DTT activities of quinones and transition metals are additive, previous studies have attempted to reconstruct the overall OP in ambient particles based on the chemical composition (Charrier and Anastasio, 2012; Charrier et al., 2015). However, addition of a transition-metal chelator did not result in significant changes in the expression of inflammatory biomarkers (Donaldson et al., 2001), suggesting that oxidative activities from different transition metals may not be additive. In our study, significant reductions in OP were observed for PSOA (43±4%) and NSOA (50±6%), when they were mixed with Cu (II) (Fig.8a). Conversely, no significant OP reduction was observed when α-pinene or limonene SOA was mixed with Cu (II). To further investigate the cause of this reduction, we examined the OP of 1,2-NQN, 1,4-NQN, 1,3-DHN and 2,3-DHN and the effects from mixing with Cu(II). Significant OP reductions for 1,2-NON (42±7%) and 2,3-DHN (35±1%)(Fig. 8b) were observed, but no such changes were observed with 1,4-NQN or 1,3-DHN upon metal mixing. It should be noted that a mixture of phenanthrenequinone and Cu(II) did not show a significant reduction in DTT activity in a study by Charrier and Anastasio (2012) while it is likely that the DTT measurements may be affected by the inefficiency of the quench regent, trichloroacetic acid (Curbo et al., 2013). Furthermore, an increased level of OP reduction was observed when an increasing amount of Cu (II) were mixed with the same amount of 1,2-NQN (Fig. S6). Based on the observation that OP reduction occurs only when there are neighboring oxygenated functional groups, we hypothesize that the OP reduction is related to formation of

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





505 covalent bonds between the electron-deficient Cu (II) and the electron-donating polar functional 506 groups. The formation of quinone-copper complexes have been demonstrated previously 507 (Dooley et al., 1990; Klinman, 1996), and may be responsible for reducing the overall OP. 508 To understand the underlying mechanism, ¹H NMR spectroscopy was used to monitor the 509 formation of the organic-Cu complex. ¹H NMR has previously been applied to study the binding 510 between metals and organics (Peana et al., 2015; Syme and Viles, 2006). The relaxometry (T2) 511 512 of NSOA illustrated in Fig. 9 shows a decreasing trend on average for T2 relaxation time when 513 Cu (II) was added to the system. Such decrease in relaxation time indicates interactions between 514 copper and SOA components. Due to the complexity in the SOA NMR spectra, it is still 515 currently challenging to specifically identify binding between individual NSOA components and Cu. To further investigate the binding reactions, ¹H NMR measurements were made for the 516 517 compounds present in NSOA, as previously mentioned, in the presence and absence of Cu, as 518 shown in Fig. 10. Significant interactions between Cu and 1,2-NQN are evidenced by the 519 broadened peak shapes (Fig. 10a), caused by the coordination of adjacent oxygen groups with 520 the copper ion, which has also been well documented in previous studies with similar compounds (Inoue and Gokel, 1990; Schmidt et al., 1990; Tolman, 1977). Protons proximal to the 521 binding sites are more significantly broadened, while protons further away are less affected. 522 523 Conversely, mixing of 1, 4-NON with Cu led to little changes in peak shape, indicating the lack 524 of any interactions with copper, as shown in Fig. 10b. Similar phenomena were also observed 525 with another pair of isomers (2, 3-DHN and 1,3-DHN). As shown in Fig.10c, 2, 3-DHN shows a 526 clear change in peak shape indicating that these hydroxyl moieties on adjacent carbons are very

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





important for copper coordination. On the other hand, the 1, 3-DHN structure shows very little 527 528 peak broadening when mixed with copper. 529 530 Such binding evidence was further supported by NMR relaxation data (Fig. 11). As copper is 531 paramagnetic, it is an effective relaxation agent, and protons brought into its proximity undergo faster T1 and T2 relaxation that manifest themselves as spectral broadening in 1D NMR (Fig. 10). 532 For the epitope maps, the largest circles indicate the least interactions with copper. For both T1 533 534 and T2 data (Fig. 11, Table S1), protons adjacent to the binding sites a and f (pronto peak 535 assignment based on Fig. 10a, Fig. S5) underwent significant changes in relaxation indicating the 536 metal coordinating to the neighboring oxygen groups. Proton e changed less as it is located 537 further away from the copper binding site. While copper has a relatively mild effect on protons b and d, a significant reduction in T1 and T2 for proton c was observed. Such interesting 538 539 phenomenon is due to the increased nuclear Overhauser effect (NOE) from the b and d sites. In the absence of binding, protons b and d would in part relax via NOE with protons a and e, which 540 relax rapidly due to the copper binding. As such protons b and d can no longer lose 541 542 magnetization via an Overhauser effect owing to a and e. Instead, they pass magnetization to position c which underwent enhanced level of relaxation as a result. The NOE effect for ring 543 544 systems with similar structure has also been demonstrated by several previous publications 545 (Kowalewski and Maler, 2006; Rehmann and Barton, 1990). It should be noted that the increased 546 level of OP depletion is accompanied by an increased ratio of Cu to 1,2-NQN (Fig. S4a), the 547 degree of broadening in the 1D NMR proton peaks becomes more significant (Fig. S4b), and the 548 relaxation times T1, T2 decreases (Fig. 11). All these observations together illustrate that the reduction of OP is proportional to the binding between Cu and organics, and further supports the 549

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





mechanism behind OP depletion in the DTT assay. Based on the conclusions from the individual 550 551 organic standards, the overall decrease in relaxation times for NSOA mixing with Cu (shown in 552 Fig. 9) likely indicates that Cu(II) are binding with NSOA components, limiting the redox 553 activities and OP of both the Cu(II) ions and the redox-active NSOA components. 554 4 Implications 555 556 Oxidative stress caused by ROS production and antioxidant consumption is one of the most 557 commonly studied mechanisms for PM toxicity (Nel, 2005;Rhee, 2006;Manke et al., 2013). Here 558 we performed OP evaluation of two SOA formed from PAHs (naphthalene, phenanthrene) by the 559 DTT assay, and investigated the linkages between SOA OP and chemical composition upon various atmospheric aging processes. SOA derived from ozonolysis of monoterpenes (α-pinene, 560 limonene) have a lower DTTm than that of the two PAHs derived SOA, which could be 561 562 attributed to the high redox-active quinone-like components. This is also consistent with the previous hypothesis that OP of SOA is highly dependent on the identity of its precursor (Tuet et 563 al., 2017a). To further link SOA OP to its chemical composition, this study also explored the 564 565 possible impacts from atmospheric aging processes so as to provide mechanistic understanding for ambient observations. 566 567 568 Over the span of atmospheric lifetime, the mass and chemical composition of SOA can be 569 affected by aging processes (Kroll et al., 2009; Lim et al., 2017). The aerosol aging processes 570 that we studied here include oligomerization, heterogeneous oxidation and metal mixing. Apart 571 from quinones that are well known to exhibit high OP in aerosol samples, OP contributions from peroxides in our NSOA system are likely to be insignificant. Rather, oxygenated derivatives 572

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





573

574

575

576

577

578

579

580

581

582

583

584

585

586

587

588

589

590

591

592

593

594

595

were shown here to contribute greater OP than their precursors in our study of selected organic individuals, and heterogeneous oxidation of NSOA was shown to lead to greater OP as well. Moreover, DTT activities of the monomer-rich and oligomer-rich fractions in NSOA separated by liquid chromatography showed oligomers are OP contributors in SOA. While organic peroxides have been proved to be very labile components with half-lives of minutes at room temperature (Krapf et al., 2016), SOA oligomers are relatively stable and highly oxygenated with their ratio of total organic molecular weight per organic carbon weight (OM:OC) similar to that of atmospheric humic-like substances (HULIS)(Altieri et al., 2008). Consistent with OP contributions from oxygenated components enriched HULIS fraction in ambient PM (Verma et al., 2015b), this study also shows evidence for OP contributors from atmospheric aging products of PAH-derived SOA, indicating major organic PM OP contributors could be less volatile than previously thought, and may more readily remain in the particle phase under atmospheric aging. Nevertheless, future work should focus on improving separation methods, allowing for more precise measurements of OP from SOA oligomeric constituents. A reduction in OP was observed when mixing NSOA/PSOA with Cu (II), resulting in a nonadditive effect. However, no such reduction was observed in α-pinene SOA or limonene SOA. Using ¹H NMR spectroscopy, we demonstrate that the reduction in OP is likely caused by binding between Cu (II) and redox active organic compounds. Both the peak broadening in 1D NMR spectra and shorter relaxation times are observed for compounds that exhibited OP reduction (1,2-NQN, 2,3-DHN) upon Cu(II) mixing. Additionally, the greater amount of Cu (II) mixed in, the enhanced OP reduction and the decrease in relaxation times showed up. While it is still challenging to determine which NSOA components are binding with Cu (II), the overall

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





relaxation time also decreased when NSOA was mixed with Cu(II), indicating binding between 596 597 Cu(II) and various NSOA components. Based on our work, previously recognized redox-active 598 organic and inorganic components in ambient particles (Charrier and Anastasio, 2012; 599 McWhinney et al., 2011; Monks et al., 1992; Turski and Thiele, 2009) may bind with each other 600 once mixed during atmospheric aging processes (external) or within the physiological environment of the human body (internal). The current study demonstrates that such binding 601 602 leads to a lower OP, which may be relevant to many health outcomes. In the future, a more 603 detailed understanding of SOA binding with metal components and the effects on the oxidative 604 health outcomes will be essential. It should also be noted that the DTT assay alone may not be 605 entirely representative of physiological ROS variations (Tuet et al., 2017a; Xiong et al., 2017;). 606 More in vitro and in vivo work should be performed in establishing the relationship between 607 chemical composition and the OP of aerosol. 608 ASSOCIATED CONTENT 609 **Supporting Information.** 610 611 **AUTHOR INFORMATION** 612 613 **Corresponding Author** 614 * Address: 200 College Street, Toronto, ON, M5S 3E5 615 Email: arthurwh.chan@utoronto.ca 616 Phone: +1 (416)-978-2602 617 618 **Notes** 619 The authors declare no competing financial interest. 620

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.



621



622 **ACKNOWLEDGEMENT** 623 This work was supported by Natural Sciences and Engineering Research Council Discovery Grant, Canadian Foundation for Innovation John R Evans Leaders Fund, and the Ontario Early 624 625 Researcher Award. The authors would like to thank Dr. Jon Abbatt, Dr. Greg Evans and Manpreet Takhar for helpful discussion. 626 627 **References:** 628 Altieri, K., Seitzinger, S., Carlton, A., Turpin, B., Klein, G., and Marshall, A.: Oligomers formed 629 630 through in-cloud methylglyoxal reactions: Chemical composition, properties, and mechanisms 631 investigated by ultra-high resolution FT-ICR mass spectrometry, Atmos. Environ., 42, 1476-1490, 2008. 632 633 Anti rolo, M., Willis, M. D., Zhou, S., and Abbatt, J. P. D.: Connecting the oxidation of soot to its redox cycling abilities, Nat. Commun., 6, 2015. 634 Atkinson, R., and Arey, J.: Atmospheric degradation of volatile organic compounds, Chem. Rev., 635 636 103, 4605-4638, 2003. Banerjee, D. K., and Budke, C. C.: Spectrophotometric Determination of Traces of Peroxides in 637 Organic Solvents, Anal. Chem., 36, 792-796, 1964. 638 639 Beelen, R., Raaschou-Nielsen, O., Stafoggia, M., Andersen, Z. J., Weinmayr, G., Hoffmann, B., 640 Wolf, K., Samoli, E., Fischer, P., and Nieuwenhuijsen, M.: Effects of long-term exposure to air pollution on natural-cause mortality: an analysis of 22 European cohorts within the multicentre 641 642 ESCAPE project, Lancet, 383, 785-795, 2014.

Manuscript under review for journal Atmos. Chem. Phys.





- 643 Bolton, J. L., Trush, M. A., Penning, T. M., Dryhurst, G., and Monks, T. J.: Role of quinones in
- 644 toxicology, Chem. Res. Toxicol., 13, 135-160, 2000.
- Brunekreef, B., and Holgate, S. T.: Air pollution and health, The lancet, 360, 1233-1242, 2002.
- 646 Chan, A. W. H., Kautzman, K. E., Chhabra, P. S., Surratt, J. D., Chan, M. N., Crounse, J. D.,
- 647 Kürten, A., Wennberg, P. O., Flagan, R. C., and Seinfeld, J. H.: Secondary organic aerosol
- 648 formation from photooxidation of naphthalene and alkylnaphthalenes: implications for oxidation
- of intermediate volatility organic compounds (IVOCs), Atmos. Chem. Phys., 9, 3049-3060, 2009.
- 650 Charrier, J. G., and Anastasio, C.: On dithiothreitol (DTT) as a measure of oxidative potential for
- ambient particles: evidence for the importance of soluble transition metals, Atmos. Chem. Phys.,
- 652 12, 11317, 2012.
- 653 Charrier, J. G., Richards-Henderson, N. K., Bein, K. J., McFall, A. S., Wexler, A. S., and
- 654 Anastasio, C.: Oxidant production from source-oriented particulate matter—Part 1: Oxidative
- 655 potential using the dithiothreitol (DTT) assay, Atmos. Chem. Phys., 15, 2327-2340, 2015.
- 656 Charrier, J. G., McFall, A. S., Vu, K. K., Baroi, J., Olea, C., Hasson, A., and Anastasio, C.: A
- 657 bias in the "mass-normalized" DTT response—An effect of non-linear concentration-response
- curves for copper and manganese, Atmos. Environ., 144, 325-334, 2016.
- 659 Cho, A. K., Di Stefano, E., You, Y., Rodriguez, C. E., Schmitz, D. A., Kumagai, Y., Miguel, A.
- 660 H., Eiguren-Fernandez, A., Kobayashi, T., and Avol, E.: Determination of four quinones in
- diesel exhaust particles, SRM 1649a, and atmospheric PM_{2.5} special issue of aerosol science and
- technology on findings from the fine particulate matter supersites program, Aerosol Sci. Tech.,
- 663 38, 68-81, 2004.

Manuscript under review for journal Atmos. Chem. Phys.





- 664 Cho, A. K., Sioutas, C., Miguel, A. H., Kumagai, Y., Schmitz, D. A., Singh, M., Eiguren-
- 665 Fernandez, A., and Froines, J. R.: Redox activity of airborne particulate matter at different sites
- in the Los Angeles Basin, Environ. Res., 99, 40-47, 2005.
- 667 Cleland, W. W.: Dithiothreitol, a new protective reagent for SH groups. Biochem., 3(4), 480-
- 668 482, 1964.
- 669 Cohen, A. J., Brauer, M., Burnett, R., Anderson, H. R., Frostad, J., Estep, K., Balakrishnan, K.,
- 670 Brunekreef, B., Dandona, L., and Dandona, R.: Estimates and 25-year trends of the global
- 671 burden of disease attributable to ambient air pollution: an analysis of data from the Global
- 672 Burden of Diseases Study 2015, Lancet, 389, 1907-1918, 2017.
- Curbo, S., Reiser, K., Rundlöf, A. K., Karlsson, A., & Lundberg, M.: Is trichloroacetic acid an
- insufficient sample quencher of redox reactions?, Antioxid. Redox Sign., 18(7), 795-799, 2013.
- 675 de Kok, T. M., Driece, H. A., Hogervorst, J. G., and Bried é, J. J.: Toxicological assessment of
- ambient and traffic-related particulate matter: a review of recent studies, Mutat. Res-Rev. Mutat.,
- 677 613, 103-122, 2006.
- 678 Delfino, R. J., Staimer, N., Tjoa, T., Gillen, D. L., Schauer, J. J., and Shafer, M. M.: Airway
- 679 inflammation and oxidative potential of air pollutant particles in a pediatric asthma panel, J.
- 680 Expo. Sci. Env. Epid., 23, 466-473, 2013. Di Lorenzo, R. A., and Young, C. J.: Size separation
- method for absorption characterization in brown carbon: Application to an aged biomass burning
- sample, Geophys. Res. Lett., 43, 458-465, 2016.
- 683 Di Lorenzo, R. A., Washenfelder, R. A., Attwood, A. R., Guo, H., Xu, L., Ng, N. L., Weber, R.
- 684 J., Baumann, K., Edgerton, E., and Young, C. J.: Molecular-Size-Separated Brown Carbon
- Absorption for Biomass-Burning Aerosol at Multiple Field Sites, Environ. Sci. Technol., 51,
- 686 3128-3137, 2017.

Manuscript under review for journal Atmos. Chem. Phys.





- 687 Docherty, K. S., Wu, W., Lim, Y. B., and Ziemann, P. J.: Contributions of organic peroxides to
- 688 secondary aerosol formed from reactions of monoterpenes with O3, Environ. Sci. Technol., 39,
- 689 4049-4059, 2005.
- 690 Donaldson, K., Stone, V., Seaton, A., and MacNee, W.: Ambient particle inhalation and the
- 691 cardiovascular system: potential mechanisms, Environ. Health Persp., 109, 523, 2001.
- 692 Dooley, D. M., McIntire, W. S., McGuirl, M. A., Cote, C. E., and Bates, J. L.: Characterization
- 693 of the active site of Arthrobacter P1 methylamine oxidase: evidence for copper-quinone
- 694 interactions, J. Am. Chem. Soc., 112, 2782-2789, 10.1021/ja00163a047, 1990.
- Fang, T., Guo, H., Verma, V., Peltier, R. E., and Weber, R. J.: PM 2.5 water-soluble elements in
- the southeastern United States: automated analytical method development, spatiotemporal
- distributions, source apportionment, and implications for heath studies, Atmos. Chem. Phys., 15,
- 698 11667-11682, 2015.
- 699 Fang, T., Verma, V., Bates, J. T., Abrams, J., Klein, M., Strickland, M. J., Sarnat, S. E., Chang,
- 700 H. H., Mulholland, J. A., and Tolbert, P. E.: Oxidative potential of ambient water-soluble PM 2.5
- 701 in the southeastern United States: contrasts in sources and health associations between ascorbic
- acid (AA) and dithiothreitol (DTT) assays, Atmos. Chem. Phys., 16, 3865-3879, 2016.
- Finlayson-Pitts, B. J., and Pitts Jr, J. N.: Chemistry of the upper and lower atmosphere: theory,
- 704 experiments, and applications, Academic press, 1999.
- Gao, S., Ng, N. L., Keywood, M., Varutbangkul, V., Bahreini, R., Nenes, A., He, J., Yoo, K. Y.,
- 706 Beauchamp, J. L., and Hodyss, R. P.: Particle phase acidity and oligomer formation in secondary
- 707 organic aerosol, Environ. Sci. Technol., 38, 6582-6589, 2004.
- 708 George, I., and Abbatt, J.: Heterogeneous oxidation of atmospheric aerosol particles by gas-
- 709 phase radicals, Nat. Chem., 2, 713-722, 2010.

Manuscript under review for journal Atmos. Chem. Phys.





- 710 Godri, K. J., Harrison, R. M., Evans, T., Baker, T., Dunster, C., Mudway, I. S., and Kelly, F. J.:
- 711 Increased oxidative burden associated with traffic component of ambient particulate matter at
- roadside and urban background schools sites in London, PloS one, 6, e21961, 2011.
- Gojova, A., Guo, B., Kota, R. S., Rutledge, J. C., Kennedy, I. M., and Barakat, A. I.: Induction
- of inflammation in vascular endothelial cells by metal oxide nanoparticles: effect of particle
- 715 composition, Environ. Health Persp., 403-409, 2007.
- 716 Groessl, M., Graf, S., and Knochenmuss, R.: High resolution ion mobility-mass spectrometry for
- separation and identification of isomeric lipids, Analyst, 140, 6904-6911, 2015.
- Hallquist, M., Wenger, J., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M., Dommen, J.,
- 719 Donahue, N., George, C., and Goldstein, A.: The formation, properties and impact of secondary
- organic aerosol: current and emerging issues, Atmos. Chem. Phys., 9, 5155-5236, 2009.
- 721 Hansen, R. E., Roth, D., and Winther, J. R.: Quantifying the global cellular thiol-disulfide status.
- 722 P. Natl. Acad. Sci., 106(2), 422-427, 2009.
- 723 Inoue, Y., and Gokel, W.G.: Cation Binding by Macrocycles. Complexation of cationic species
- by crown ethers, Marcel Dekker, New York, 1990.
- 725 Jiang, H., Jang, M., and Yu, Z.: Dithiothreitol Activity by Particulate Oxidizers of SOA
- 726 Produced from Photooxidation of Hydrocarbons under Varied NOx Levels, Atmos. Chem. Phys.
- 727 Discuss., 2017, 1-24, doi:10.5194/acp-2017-214, 2017.
- Jimenez, J. L., Canagaratna, M. R., Donahue, N. M., Prevot, A. S. H., Zhang, Q., Kroll, J. H.,
- 729 DeCarlo, P. F., Allan, J. D., Coe, H., and Ng, N. L.: Evolution of organic aerosols in the
- 730 atmosphere, Science, 326, 1525-1529, 2009.

Manuscript under review for journal Atmos. Chem. Phys.





- Jokinen, T., Sipil ä, M., Richters, S., Kerminen, V. M., Paasonen, P., Stratmann, F., Worsnop, D.,
- 732 Kulmala, M., Ehn, M., and Herrmann, H.: Rapid autoxidation forms highly oxidized RO2
- radicals in the atmosphere, Angew. Chem. Int. Edit., 53, 14596-14600, 2014.
- Kalberer, M., Paulsen, D., Sax, M., Steinbacher, M., Dommen, J., Prevot, A., Fisseha, R.,
- 735 Weingartner, E., Frankevich, V., and Zenobi, R.: Identification of polymers as major components
- of atmospheric organic aerosols, Science, 303, 1659-1662, 2004.
- 737 Kautzman, K. E., Surratt, J. D., Chan, M. N., Chan, A. W. H., Hersey, S. P., Chhabra, P. S.,
- 738 Dalleska, N. F., Wennberg, P. O., Flagan, R. C., and Seinfeld, J. H.: Chemical composition of
- 739 gas-and aerosol-phase products from the photooxidation of naphthalene, J. Phys. Chem. A, 114,
- 740 913-934, 2009.
- Keywood, M. D., Kroll, J. H., Varutbangkul, V., Bahreini, R., Flagan, R. C., and Seinfeld, J. H.:
- 742 Secondary organic aerosol formation from cyclohexene ozonolysis: Effect of OH scavenger and
- the role of radical chemistry, Environ. Sci. Technol., 38, 3343-3350, 2004.
- 744 Klinman, J. P.: Mechanisms whereby mononuclear copper proteins functionalize organic
- 745 substrates, Chem. Rev., 96, 2541-2562, 1996.
- 746 Kostenidou, E., Pathak, R. K., and Pandis, S. N.: An algorithm for the calculation of secondary
- 747 organic aerosol density combining AMS and SMPS data, Aerosol Sci. Technol., 41, 1002-1010,
- 748 2007.
- 749 Kowalewski, J., and Maler, L.: Nuclear spin relaxation in liquids: theory, experiments, and
- applications, CRC press, 2006.
- 751 Kramer, A. J., Rattanavaraha, W., Zhang, Z., Gold, A., Surratt, J. D., and Lin, Y.-H.: Assessing
- 752 the oxidative potential of isoprene-derived epoxides and secondary organic aerosol, Atmos.
- 753 Environ., 130, 211-218, 2016.

Manuscript under review for journal Atmos. Chem. Phys.





- 754 Krapf, M., El Haddad, I., Bruns, Emily A., Molteni, U., Daellenbach, Kaspar R., Pr év ôt, Andr é S.
- 755 H., Baltensperger, U., and Dommen, J.: Labile Peroxides in Secondary Organic Aerosol, Chem,
- 756 1, 603-616, 2016.
- 757 Krapf, M., Künzi, L., Allenbach, S., Bruns, E. A., Gavarini, I., El-Haddad, I., Slowik, J. G.,
- 758 Prévôt, A. S. H., Drinovec, L., and Močnik, G.: Wood combustion particles induce adverse
- effects to normal and diseased airway epithelia, Environ. Sci. Proc. Impacts, 2017.
- 760 Krechmer, J. E., Lambe, A. T., Kimmel, J. R., Cubison, M. J., Budisulistiorini, S. H., Surratt, J.
- D., Jayne, J. T., Worsnop, D. R., and Canagaratna, M. R.: Ion mobility spectrometry-mass
- 762 spectrometry (IMS-MS) for on-and offline analysis of atmospheric gas and aerosol species,
- 763 Atmos. Meas. Tech., 9, 3245, 2016.
- Kroll, J. H., Smith, J. D., Che, D. L., Kessler, S. H., Worsnop, D. R., and Wilson, K. R.:
- Measurement of fragmentation and functionalization pathways in the heterogeneous oxidation of
- oxidized organic aerosol, Phys. Chem. Chem. Phys., 11, 8005-8014, 2009.
- Kroll, J. H., Donahue, N. M., Jimenez, J. L., Kessler, S. H., Canagaratna, M. R., Wilson, K. R.,
- 768 Altieri, K. E., Mazzoleni, L. R., Wozniak, A. S., and Bluhm, H.: Carbon oxidation state as a
- metric for describing the chemistry of atmospheric organic aerosol, Nat. Chem., 3, 133-139,
- 770 2011.
- 771 Kumagai, Y., Koide, S., Taguchi, K., Endo, A., Nakai, Y., Yoshikawa, T., and Shimojo, N.:
- Oxidation of proximal protein sulfhydryls by phenanthraquinone, a component of diesel exhaust
- particles, Chem. Res. Toxicol., 15, 483-489, 2002.
- 774 Lee, J. Y., and Lane, D. A.: Unique products from the reaction of naphthalene with the hydroxyl
- radical, Atmos. Environ., 43, 4886-4893, 2009.

Manuscript under review for journal Atmos. Chem. Phys.





- 776 Lelieveld, J., Evans, J. S., Fnais, M., Giannadaki, D., and Pozzer, A.: The contribution of outdoor
- air pollution sources to premature mortality on a global scale, Nature, 525, 367-371, 2015.
- Li, N., Hao, M., Phalen, R. F., Hinds, W. C., and Nel, A. E.: Particulate air pollutants and asthma:
- 779 a paradigm for the role of oxidative stress in PM-induced adverse health effects, Clin. Immunol.,
- 780 109, 250-265, 2003a.
- 781 Li, N., Sioutas, C., Cho, A., Schmitz, D., Misra, C., Sempf, J., Wang, M., Oberley, T., Froines, J.,
- 782 and Nel, A.: Ultrafine particulate pollutants induce oxidative stress and mitochondrial damage,
- 783 Environ. Health Persp., 111, 455, 2003b.
- Lim, C. Y., Browne, E. C., Sugrue, R. A., and Kroll, J. H.: Rapid heterogeneous oxidation of
- organic coatings on submicron aerosols. Geophys. Res. Lett., 44(6), 2949-2957,2017.
- Lin, Y., Arashiro, M., Martin, E., Chen, Y., Zhang, Z., Sexton, K. G., Gold, A., Jaspers, I., Fry,
- 787 R. C., and Surratt, J. D.: Isoprene-Derived Secondary Organic Aerosol Induces the Expression of
- 788 Oxidative Stress Response Genes in Human Lung Cells, Environ. Sci. Technol. L., 3, 250-254,
- 789 2016.
- 790 Manke, A., Wang, L., and Rojanasakul, Y.: Mechanisms of nanoparticle-induced oxidative stress
- 791 and toxicity, Biomed. Res. Int., 2013, 1-15, 2013.
- 792 McWhinney, R. D., Gao, S. S., Zhou, S., and Abbatt, J. P. D.: Evaluation of the effects of ozone
- oxidation on redox-cycling activity of two-stroke engine exhaust particles, Environ. Sci.
- 794 Technol., 45, 2131-2136, 2011.
- 795 McWhinney, R. D., Zhou, S., and Abbatt, J. P. D.: Naphthalene SOA: redox activity and
- naphthoquinone gas-particle partitioning, Atmos. Chem. Phys., 13, 9731-9744, 2013.
- 797 Monks, T. J., Hanzlik, R. P., Cohen, G. M., Ross, D., and Graham, D. G.: Quinone chemistry
- 798 and toxicity, Toxicol. Appl. Pharm., 112, 2-16, 1992.

Manuscript under review for journal Atmos. Chem. Phys.





- 799 Nel, A.: Air pollution-related illness: effects of particles, Science, 308, 804-806, 2005.
- 800 Nel, A. E., Diaz-Sanchez, D., Ng, D., Hiura, T., and Saxon, A.: Enhancement of allergic
- inflammation by the interaction between diesel exhaust particles and the immune system, J.
- 802 Allergy. Clin. Immun., 102, 539-554, 1998.
- 803 Oberd örster, G., Oberd örster, E., and Oberd örster, J.: Nanotoxicology: an emerging discipline
- evolving from studies of ultrafine particles, Environ. Health Persp., 823-839, 2005.
- 805 Peana, M., Medici, S., Nurchi, V. M., Lachowicz, J. I., Crisponi, G., Crespo-Alonso, M., Santos,
- 806 M. A., and Zoroddu, M. A.: An NMR study on the 6,6' -(2-
- 807 (diethylamino)ethylazanediyl)bis(methylene)bis(5-hydroxy-2-hydroxymethyl-4H-pyran-4-one)
- interaction with AlIII and ZnII ions, J. Inorg. Biochem., 148, 69-77,
- 809 https://doi.org/10.1016/j.jinorgbio.2015.01.016, 2015.
- 810 Pope, C. A., Burnett, R. T., Thun, M. J., Calle, E. E., Krewski, D., Ito, K., and Thurston, G. D.:
- Lung cancer, cardiopulmonary mortality, and long-term exposure to fine particulate air pollution,
- 812 Jama, 287, 1132-1141, 2002.
- Pope, C. A., Ezzati, M., and Dockery, D. W.: Fine-particulate air pollution and life expectancy in
- the United States, N. Engl. J. Med., 2009, 376-386, 2009.
- 815 Pöschl, U., and Shiraiwa, M.: Multiphase chemistry at the atmosphere–biosphere interface
- influencing climate and public health in the anthropocene, Chem. Rev., 115, 4440-4475, 2015.
- 817 Rehmann, J. P., and Barton, J. K.: Proton NMR studies of tris (phenanthroline) metal complexes
- bound to oligonucleotides: characterization of binding modes, Biochem., 29, 1701-1709, 1990.
- 819 Rhee, S. G.: H₂O₂, a necessary evil for cell signaling, Science, 312, 1882-1883, 2006.
- 820 Risom, L., Møller, P., and Loft, S.: Oxidative stress-induced DNA damage by particulate air
- 821 pollution, Mutat.Res-Fund. Mol. M., 592, 119-137, 2005.

Manuscript under review for journal Atmos. Chem. Phys.





- 822 Rudich, Y., Donahue, N. M., and Mentel, T. F.: Aging of organic aerosol: Bridging the gap
- between laboratory and field studies, Annu. Rev. Phys. Chem., 58, 321-352, 2007.
- 824 Schmidt, M. H., Miskelly, G. M., and Lewis, N. S.: Effects of redox potential, steric
- 825 configuration, solvent, and alkali metal cations on the binding of carbon dioxide to cobalt (I) and
- nickel (I) macrocycles, J. Am. Chem. Soc., 112, 3420-3426, 1990.
- 827 Shen, H., Barakat, A. I., and Anastasio, C.: Generation of hydrogen peroxide from San Joaquin
- Valley particles in a cell-free solution, Atmos. Chem. Phys., 11, 753-765, 2011.
- Shilling, J. E., Chen, Q., King, S. M., Rosenoern, T., Kroll, J. H., Worsnop, D. R., DeCarlo, P. F.,
- 830 Aiken, A. C., Sueper, D., and Jimenez, J. L.: Loading-dependent elemental composition of α-
- pinene SOA particles, Atmos. Chem. Phys., 9, 771-782, 2009.
- 832 Shiraiwa, M., Selzle, K., and Pöschl, U.: Hazardous components and health effects of
- atmospheric aerosol particles: reactive oxygen species, soot, polycyclic aromatic compounds and
- allergenic proteins, Free. Radical. Res., 46, 927-939, 2012.
- 835 Simpson, A. J., McNally, D. J., and Simpson, M. J.: NMR spectroscopy in environmental
- research: from molecular interactions to global processes, Prog. Nucl. Mag. Res. Sp., 58, 97-175,
- 837 2011.
- 838 Simpson, M. J., and Simpson, A. J.: NMR Spectroscopy: A Versatile Tool for Environmental
- Research, John Wiley & Sons, 2014.
- 840 Smith, M. E., and van Eck, E. R.: Recent advances in experimental solid state NMR
- methodology for half-integer spin quadrupolar nuclei, Prog. Nucl. Mag. Res. Sp., 34, 159-201,
- 842 1999.
- 843 Surratt, J. D., Chan, A. W. H., Eddingsaas, N. C., Chan, M., Loza, C. L., Kwan, A. J., Hersey, S.
- 844 P., Flagan, R. C., Wennberg, P. O., and Seinfeld, J. H.: Reactive intermediates revealed in

Manuscript under review for journal Atmos. Chem. Phys.





- secondary organic aerosol formation from isoprene, P. Natl. Acad. Sci., 107, 6640-6645,
- 846 10.1073/pnas.0911114107, 2010.
- 847 Syme, C. D., and Viles, J. H.: Solution 1 H NMR investigation of Zn 2+ and Cd 2+ binding to
- 848 amyloid-beta peptide (Aβ) of Alzheimer's disease, BBA-Proteins and Proteom., 1764, 246-256,
- 849 2006.
- Thurston, G. D., Burnett, R. T., Turner, M. C., Shi, Y., Krewski, D., Lall, R., Ito, K., Jerrett, M.,
- 851 Gapstur, S. M., and Diver, W. R.: Ischemic heart disease mortality and long-term exposure to
- source-related components of US fine particle air pollution, Environ. Health Persp., 124, 785,
- 853 2016.
- 854 Tolman, C. A.: Steric effects of phosphorus ligands in organometallic chemistry and
- 855 homogeneous catalysis, Chem. Rev., 77, 313-348, 1977.
- Tolocka, M. P., Jang, M., Ginter, J. M., Cox, F. J., Kamens, R. M., and Johnston, M. V.:
- 857 Formation of oligomers in secondary organic aerosol, Environ. Sci. Technol., 38, 1428-1434,
- 858 2004.
- 859 Trump, E. R., and Donahue, N. M.: Oligomer formation within secondary organic aerosols:
- equilibrium and dynamic considerations, Atmos. Chem. Phys., 14, 3691-3701, 2014.
- Tuet, W. Y., Fok, S., Verma, V., Rodriguez, M. S. T., Grosberg, A., Champion, J. A., and Ng, N.
- L.: Dose-dependent intracellular reactive oxygen and nitrogen species (ROS/RNS) production
- from particulate matter exposure: comparison to oxidative potential and chemical composition,
- 864 Atmos. Environ., 144, 335-344, 2016.
- Tuet, W. Y., Chen, Y., Fok, S., Champion, J. A., and Ng, N. L.: Inflammatory responses to
- secondary organic aerosols (SOA) generated from biogenic and anthropogenic precursors,
- 867 Atmos. Chem. Phys. Discuss., 2017, 1-42, 10.5194/acp-2017-262, 2017a.

Manuscript under review for journal Atmos. Chem. Phys.





- Tuet, W. Y., Chen, Y., Xu, L., Fok, S., Gao, D., Weber, R. J., and Ng, N. L.: Chemical oxidative
- 869 potential of secondary organic aerosol (SOA) generated from the photooxidation of biogenic and
- anthropogenic volatile organic compounds, Atmos. Chem. Phys., 17, 839-853, 2017b.
- Turski, M. L., and Thiele, D. J.: New roles for copper metabolism in cell proliferation, signaling,
- and disease, J. Biol. Chem., 284, 717-721, 2009.
- Valavanidis, A., Fiotakis, K., Bakeas, E., and Vlahogianni, T.: Electron paramagnetic resonance
- 874 study of the generation of reactive oxygen species catalysed by transition metals and quinoid
- redox cycling by inhalable ambient particulate matter, Redox. Rep., 10, 37-51, 2005.
- Verma, V., Ning, Z., Cho, A. K., Schauer, J. J., Shafer, M. M., and Sioutas, C.: Redox activity of
- urban quasi-ultrafine particles from primary and secondary sources, Atmos. Environ., 43, 6360-
- 878 6368, 2009.
- Verma, V., Fang, T., Xu, L., Peltier, R. E., Russell, A. G., Ng, N. L., and Weber, R. J.: Organic
- aerosols associated with the generation of reactive oxygen species (ROS) by water-soluble PM_{2.5},
- 881 Environ. Sci. Technol., 49, 4646-4656, 2015a.
- 882 Verma, V., Wang, Y., El-Afifi, R., Fang, T., Rowland, J., Russell, A. G., and Weber, R. J.:
- Fractionating ambient humic-like substances (HULIS) for their reactive oxygen species activity—
- Assessing the importance of quinones and atmospheric aging, Atmos. Environ., 120, 351-359,
- 885 2015b.
- Wang, L., Xu, W., Khalizov, A. F., Zheng, J., Qiu, C., and Zhang, R.: Laboratory investigation
- on the role of organics in atmospheric nanoparticle growth, J. Phys. Chem. A, 115, 8940-8947,
- 888 2011.

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





- 889 Xiong, Q., Yu, H., Wang, R., Wei, J., and Verma, V.: Rethinking The Dithiothreitol (DTT)
- 890 Based PM Oxidative Potential: Measuring DTT Consumption versus ROS Generation, Environ.
- 891 Sci. Technol., 2017.
- Ye, J., Gordon, C. A., and Chan, A. W. H.: Enhancement in secondary organic aerosol formation
- in the presence of preexisting organic particle, Environ. Sci. Technol., 50, 3572-3579, 2016.
- 894 Ye, J., Salehi, S., North, M. L., Portelli, A. M., Chow, C.-W., and Chan, A. W. H.: Development
- of a Novel Simulation Reactor for Chronic Exposure to Atmospheric Particulate Matter, Sci.
- 896 Rep., 7, 2017.
- Zhang, R., Wang, G., Guo, S., Zamora, M. L., Ying, Q., Lin, Y., Wang, W., Hu, M., and Wang,
- Y.: Formation of urban fine particulate matter, Chem. Rev., 115, 3803-3855, 2015.
- Zhang, X., McVay, R. C., Huang, D. D., Dalleska, N. F., Aumont, B., Flagan, R. C., and
- 900 Seinfeld, J. H.: Formation and evolution of molecular products in α-pinene secondary organic
- 901 aerosol, P. Natl. Acad. Sci., 112, 14168-14173, 2015.
- 902 Zhang, X., Lambe, A. T., Upshur, M. A., Brooks, W. A., Gray Be, A., Thomson, R. J., Geiger, F.
- 903 M., Surratt, J. D., Zhang, Z., and Gold, A.: Highly Oxygenated Multifunctional Compounds in α-
- pinene Secondary Organic Aerosol, Environ. Sci. Technol., 2017.

905

906

907

908

909

Manuscript under review for journal Atmos. Chem. Phys.





Table 1. Flow tube experimental conditions

Compund	Reaction ^a	ΔHC ppb	Y ^{b,c} %	RH
limonene	ozonolysis	251±23	25±3.9	14±1%
α-pinene	ozonolysis	304±17	19±4.2	14±1%
naphthalene	photooxidation	6436±402	28±6.7	57 ±5%
phenanthrene	photooxidation	4050±578	12±2.6	57 ±5%

a. Temperature in all experiments is around room temperature (22-25°C).

Table 2. SOA peroxide content and OP

Organics	Peroxide percentage %	DTTm pmol min ⁻¹ µg ⁻¹
Naphthalene SOA	<3%	100-129
α-pinene SOA	40-100%	10-20
Benzoyl peroxide	100%	160

b. SOA mass yields were calculated without particle wall loss correction.

c. SOA density in this study was assumed to be 1.25 g cm⁻³ for monoterpene SOA (Shilling et al., 2009), and 1.55 g cm⁻³ for PAHs SOA(Chan et al., 2009).

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 10 November 2017 © Author(s) 2017. CC BY 4.0 License.





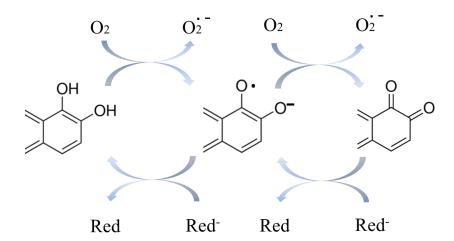


Figure 1. A simplified mechanism of redox cycling of quinone-like substances and formation of superoxide anion radicals. *Red* refers to a general reductant. In the redox cycle, regenerated quinone serves as a chemical intermediate to transfer electrons from reductants to oxygen to form superoxide $(O_2$. $\overline{}$).

43





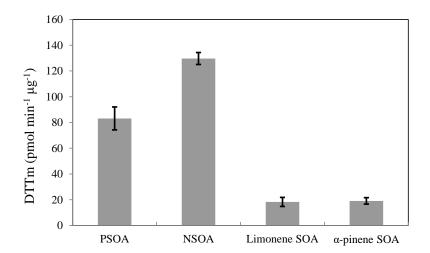


Figure 2. DTTm (pmol min⁻¹ μ g⁻¹) for SOA formed from various types of hydrocarbons (phenanthrene, naphthalene, limonene and α -pinene). Each measurement was conducted in triplicates, and the error bar represents the standard error of the mean (SEM).





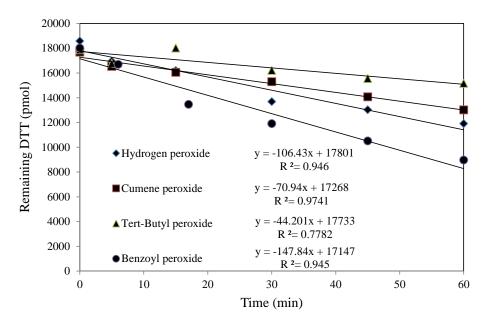
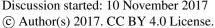


Figure 3. DTT activity of various types of peroxides (hydrogen peroxide, cumene peroxide, tert-Butyl peroxide, benzoyl peroxide). With the same initial concentration of peroxide (0.1mM), benzoyl peroxide has the highest DTT activity (147.8 pmol min⁻¹, which can be converted to DTTm of 38 pmol min⁻¹ ug⁻¹).

Discussion started: 10 November 2017







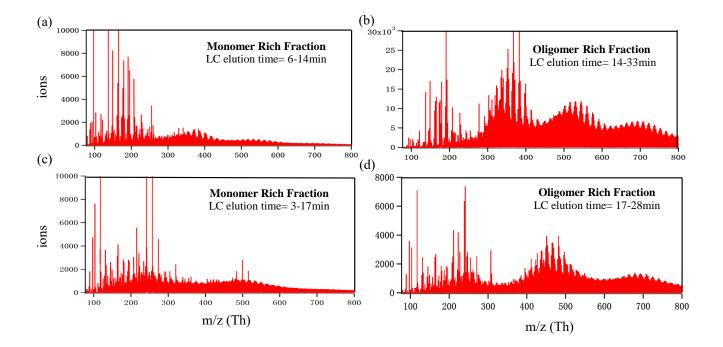


Figure 4. IMS-TOF mass spectra for (a) monomer-rich fraction (b) oligomer-rich fraction in NSOA and (c) monomer-rich fraction (d) oligomer-rich fraction in PSOA. During a total of 46min elution, the majority of NSOA monomers eluted at 6-14 min, and most of the oligomers eluted at 14-33min. For PSOA system, the majority of monomers eluted at 3-17 min, and most of the oligomer- rich fraction eluted at 17-28 min.





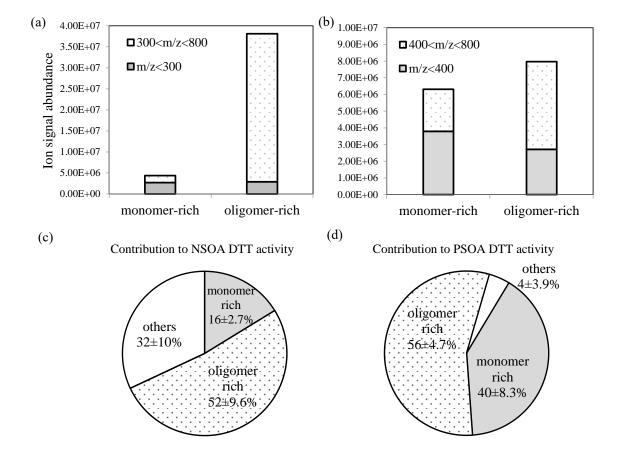


Figure 5. Sum of ion signals for monomers and oligomers in monomer-rich fraction and oligomer-rich fraction for NSOA (a) and PSOA (b) systems. OP (DTTt) contributions from monomer-rich fraction and oligomer-rich fraction in NSOA (c) and PSOA (d) systems. The remaining DTT activity (others, white) is attributed to residual SOA fractions that did not clearly elute out.





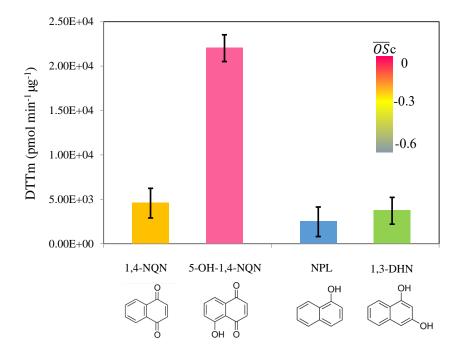


Figure 6. DTTm for two selected pairs of oxygenated derivatives in NSOA system (1,4-NQN vs. 5-OH 1,4-NQN, NPL vs. 1,3-DHN). Averaged carbon oxidation state (\overline{OS} c) of each component is shown in color (color scale shown on top-right). Each measurement was conducted in triplicates, and the error bar here represents the SEM. The asterisk indicates significant difference between each pair of measurements at the 95% confidence level.





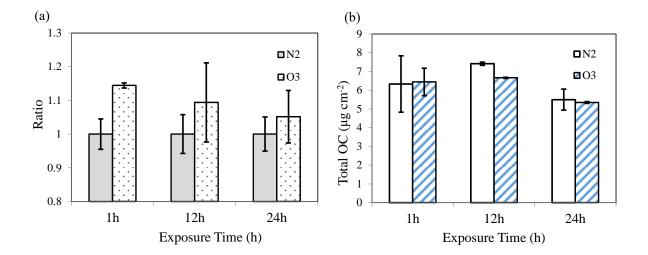


Figure 7. (a) Relative DTTt for NSOA after O_3 and N_2 exposure. For each of the exposure duration (1h, 12h, 24h), the DTTt of O_3 exposure group was normalized by the DTTt of the corresponding N_2 exposure group. Generally, DTTt of NSOA that underwent heterogeneous ozonolysis was higher than that of the N_2 control. (b) OC/EC measurement results show total OC mass loss (17% and 13% for O_3 and N_2 exposure, respectively) after 24-hour exposures. Each measurement was conducted in triplicates, and the error bar represents the SEM.





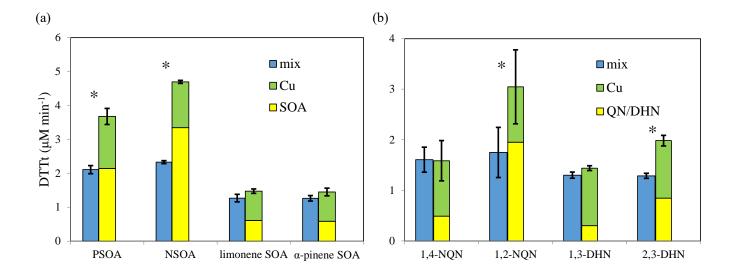


Figure 8. Significant OP depletions were observed when PSOA($43\pm4\%$), NSOA($50\pm6\%$), 1,2-NQN($42\pm7\%$) and 2,3-DHN($35\pm1\%$) mixed with Cu (II). The asterisk indicates significant difference between a pair of bars at a 95% confidence level.





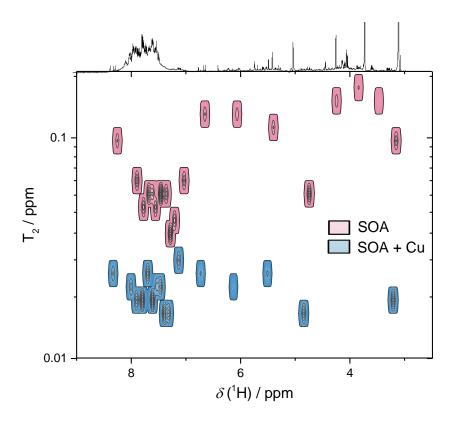


Figure 9. 2D ¹H-NMR T2 relaxation contour map for NSOA with (blue) and without copper (red) with 1D NMR projections from the top. A general decreasing trend in T2 is observed here, which indicates interactions (binding) of Cu (II) with many NSOA components.

Manuscript under review for journal Atmos. Chem. Phys.





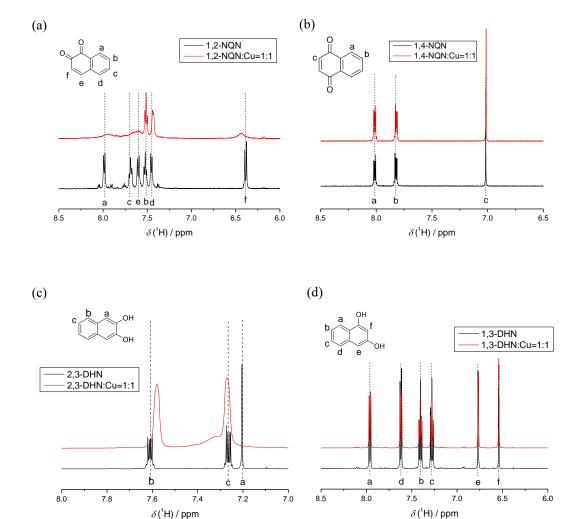


Figure 10. 1 D ¹H-NMR spectra of (a) 1,2-NQN (b) 1,4-NQN (c) 2,3-DHN (d)1,3-DHN and their mixture with 1:1 ratio of Cu (II). Both 1,2-NQN and 2,3-DHN show the broadening of ¹H-NMR peaks (protons at a, c, e, f and a, b, c for 1,2-NQN and 2,3-DHN, respectively) after mixing with Cu(II).





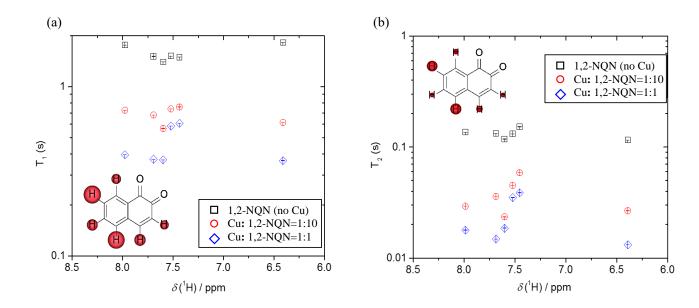


Figure 11. ¹H NMR relaxometry analyses for (a) T1, (b) T2 of 1,2-NQN mixed with Cu (II) at different ratios: 1,2-NQN with no Cu (**black**),Cu(II):1,2-NQN = 1:10 (**red**), Cu(II):1,2-NQN = 1:1(**blue**). Both T1, T2 decreased when Cu was introduced into the system, indicating a smaller scale of nuclear spin dynamic resulted from organic-metal binding. The molecular epitopes illustrate the influence of Cu (II) binding on individual proton. A smaller sphere shadow on a proton denotes a larger relaxation influence from Cu (II)-organic binding.