

1 **Comparison of secondary organic aerosol formation from**
2 **toluene on initially wet and dry ammonium sulfate particles**
3 **at moderate relative humidity**

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20 **Abstract**

21 The formation of secondary organic aerosol (SOA) has been widely studied in the
22 presence of dry seed particles at low relative humidity (RH). At higher RH, initially dry
23 seed particles can exist as wet particles due to water uptake by the seeds as well as the
24 SOA. Here, we investigated the formation of SOA from the photooxidation of toluene
25 using an oxidation flow reactor in the absence of NO_x under a range of OH exposures
26 on initially wet or dry ammonium sulfate (AS) seed particles at an RH of 68%. The
27 ratio of the SOA yield on wet AS seeds to that on dry AS seeds, the relative SOA yield,
28 decreased from 1.31±0.02 at an OH exposure of 4.66×10¹⁰ molecules cm⁻³ s to
29 1.01±0.01 at an OH exposure of 5.28×10¹¹ molecules cm⁻³ s. This decrease may be due
30 to the early deliquescence of initially dry AS seeds after coated by highly oxidized
31 toluene-derived SOA. SOA formation lowered the deliquescence RH of AS and
32 resulted in the uptake of water by both AS and SOA. Hence the initially dry AS seeds
33 contained aerosol liquid water (ALW) soon after SOA formed and the SOA yield and
34 ALW approached those of the initially wet AS seeds as OH exposure and ALW
35 increased, especially at high OH exposure. However, a higher oxidation state of the
36 SOA on initially wet AS seeds than that on dry AS seeds was observed at all levels of
37 OH exposure. The difference in mass fractions of *m/z* 29, 43 and 44 of SOA mass
38 spectra, obtained using an aerosol mass spectrometer (AMS), indicated that SOA
39 formed on initially wet seeds may be enriched in earlier-generation products containing
40 carbonyl functional groups at low OH exposures and later-generation products
41 containing acidic functional groups at high exposures. Our results suggest that

42 inorganic dry seeds become at least partially deliquesced particles during SOA
43 formation and hence ALW is inevitably involved in the SOA formation at moderate RH.
44 More laboratory experiments conducted with a wide variety of SOA precursors and
45 inorganic seeds under different NO_x and RH conditions are warranted.

46 **1. Introduction**

47 Secondary organic aerosol (SOA) is an important component of atmospheric particulate
48 matter, which influences air quality, climate and human health (Hallquist et al., 2009).
49 SOA is mainly formed via the oxidation of volatile organic compounds (VOCs),
50 followed by partitioning to the condensed phase. Traditional atmospheric chemical
51 transport models largely underestimate the levels of SOA (de Gouw et al., 2005;
52 Volkamer et al., 2006; Hodzic et al., 2010) and the degree of oxidation (Rudich et al.,
53 2007; Ng et al., 2010). The updated models incorporating the volatility basis set (VBS)
54 formalism (Donahue et al., 2006) can better predict the observed SOA, but SOA
55 formation still remains under-constrained (Shrivastava et al., 2011; Tsigaridis et al.,
56 2014; Hayes et al., 2015; Ma et al., 2017). SOA yields in atmospheric chemical
57 transport models are obtained from smog chamber experiments using dry seed particles
58 (Barsanti et al., 2013; Mahmud and Barsanti, 2013) under dry conditions. Yet,
59 atmospheric relative humidity is often sufficiently high that aerosols often contain
60 aerosol liquid water (ALW) due to their hygroscopic properties (Liao and Seinfeld,
61 2005; Lee and Adams, 2010; Guo et al., 2015; Nguyen et al., 2016). The presence of
62 ALW in aerosols may enhance SOA formation by facilitating the partitioning of
63 semivolatile organic compounds and the uptake of water-soluble gases through
64 aqueous-phase reactions (Hennigan et al., 2008; Lim et al., 2010; Ervens et al., 2011;
65 Lee et al., 2011; Sareen et al., 2017). ALW may also promote photodegradation of
66 dissolved SOA (Romonosky et al., 2014). Therefore, SOA formation under
67 atmospherically relevant relative humidity needs to be better constrained in

68 atmospheric chemical transport models by incorporating ALW. In addition,
69 understanding water uptake of SOA is important for estimating its loss by wet
70 deposition, which is not well constrained.

71 Aromatic hydrocarbons constitute a large fraction of the total non-methane
72 hydrocarbons in the urban atmosphere (Calvert et al., 2002) and account for a
73 significant fraction of SOA in urban areas (Ding et al., 2012; Zhao et al., 2017). Toluene
74 is the most abundant aromatic hydrocarbon (Calvert et al., 2002; Zhang et al., 2016)
75 and SOA yields from the photooxidation of toluene on dry or wet ammonium sulfate
76 (AS) seeds has been studied by varying the RH in smog chambers. Kamens et al. (2011)
77 observed higher yields of SOA from toluene at higher RHs. They attributed this increase
78 to the initially wet seed particles. On the other hand, Edney et al. (2000) reported that
79 wet seeds had no effect on the SOA yields of toluene compared with dry seeds. In these
80 studies, different RHs used for dry and wet seeds experiments may influence the gas-
81 phase chemistry and complicate the comparison of SOA formation.

82 SOA formation on initially dry and wet AS seeds has been compared using
83 oxidation flow reactors at same RHs (Wong et al., 2015; Faust et al., 2017). Faust et al.
84 (2017) found a 19% enhancement in the SOA yield of toluene on wet AS seeds over
85 that on dry AS seeds at 70% RH. However, at such high RH, the initially dry and water-
86 free AS seed particles can uptake water upon SOA formation because SOA themselves
87 can be hygroscopic and they can also lower the deliquescence RH of the AS seeds
88 (Takahama et al., 2007; Smith et al., 2011, 2012, 2013). The potential influence of SOA
89 formation on the physical state of the initially dry seeds as well as and the overall water

90 uptake by the aged particles was not explicitly discussed. In addition, the hydroxyl
91 radicals (OH) exposure in Faust et al. (2017) was approximately 2×10^{11} molecules cm^{-3}
92 s, equivalent to about 1.5 days of oxidation in the atmosphere assuming an ambient
93 OH concentration of 1.5×10^6 molecules cm^{-3} (Mao et al., 2009). Atmospheric particles
94 can undergo oxidation for as long as 1-2 weeks (Balkanski et al., 1993).

95 In this study, SOA formation from the photooxidation of toluene was investigated
96 in an oxidation flow reactor at an RH of 68% under a wide range of OH exposures using
97 initially wet or dry AS seed particles. The yields and composition of SOA as well as the
98 estimated ALW contents for the initially wet and dry seeds are compared. We found
99 that as OH exposure increased, the SOA yield and ALW of the initially dry seeds
100 approached those of the initially wet seeds while the wet seeds yielded SOA of a higher
101 degree of oxidation than the dry seeds did at all exposure levels.

102 **2. Materials and methods**

103 **2.1 Generation of seed particles**

104 A schematic of the experimental setup, similar to that used in Wong et al. (2015) and
105 Faust et al. (2017), is shown in Fig. 1. AS seed particles were generated from an aqueous
106 AS solution (Sigma-Aldrich) using an atomizer (TSI 3076, TSI Inc., USA). In
107 experiments using dry seeds, the atomized aqueous AS droplets passed through a silica
108 gel diffusion dryer so that the RH was reduced to less than 30% at which AS effloresced,
109 while in experiments using wet seeds, they bypassed the diffusion dryer. The dry or wet
110 seed particles then entered and mixed with a humidified $\text{N}_2/\text{O}_2/\text{O}_3$ flow in an oxidation
111 flow reactor. The RH in the flow reactor was at 68%, which lies between the

112 efflorescence and deliquescence RH of AS (Seinfeld and Pandis, 2006), so that the seed
113 particles remained in their original phase with the wet particles containing $\sim 18.6 \mu\text{g m}^{-3}$
114 ALW (see Section 2.4) and the dry particles anhydrous before reaction started.
115 Hereafter, the experiments using initially wet and dry AS seed particles are simplified
116 as wet and dry AS seeds, respectively. “Wet” and “dry” refer to the initial state of the
117 seed particles before SOA formation.

118 When atomizing a given AS solution, the diameter of wet AS droplets is much
119 larger than that of dry AS particles due to the water uptake of AS (Chan et al., 1992),
120 resulting in a larger surface area of seed particles. Previous studies have demonstrated
121 that a large surface area of seed particles may increase the SOA yields by reducing the
122 wall loss of organic vapors (Matsunaga and Ziemman, 2010, Zhang et al., 2014, 2015;
123 Huang et al., 2016; Krechmer et al., 2016). To obtain seed particles of comparable
124 surface areas, we atomized 0.013 mM and 0.015 mM of the AS solution for wet and
125 dry AS seeds, respectively. As shown in Fig. S1, the surface area distribution of initially
126 wet AS seeds was similar to that of initially dry AS seeds. Because of the difference in
127 AS concentration between the stock solutions used, wet AS seeds had a mean diameter
128 of 88 nm and were slightly smaller than dry AS seeds which had a mean diameter of
129 102 nm. The total surface area of wet AS seeds was 21% larger than that of dry AS
130 seeds. The mass loading of wet and dry AS seeds was 31.0 and $24.2 \mu\text{g m}^{-3}$, respectively.

131 **2.2 Oxidation flow reactor**

132 SOA formation from the photooxidation of toluene in the absence of NO_x on initially
133 dry or wet seeds was investigated in a potential aerosol mass (PAM) oxidation flow

134 reactor, which has been described in detail elsewhere (Kang et al., 2007, 2011; Lambe
135 et al., 2011a, 2015; Liu et al., 2017). Briefly, a PAM chamber is a continuous oxidation
136 flow reactor using high and controlled levels of oxidants to oxidize gaseous precursors
137 to produce SOA. The chamber used in this study had a volume of approximately 19 L
138 (length 60 cm, diameter 20 cm). The total flow rate in the PAM chamber was set at 3 L
139 min^{-1} using mass flow controllers, resulting in a residence time of approximately 380 s.
140 The RH and temperature of the PAM outflow were measured continuously (HMP 110,
141 Vaisala Inc, Finland) and stabilized at approximately 68% and 20 °C, respectively. High
142 OH exposures were realized through the photolysis of ozone irradiated by a UV lamp
143 ($\lambda = 254 \text{ nm}$) in the presence of water vapor. Ozone was produced by an ozone generator
144 (1000BT-12, ENALY, Japan) via the irradiation of pure O_2 . The OH concentration was
145 adjusted by varying the concentration of ozone in the PAM chamber from 0.4 ppm to
146 4.3 ppm. The corresponding upper limit of OH exposure at these operating conditions
147 ranged from $0.47 \times 10^{11} \text{ molecules cm}^{-3} \text{ s}$ to $5.28 \times 10^{11} \text{ molecules cm}^{-3} \text{ s}$, equivalent to
148 0.36 to 4.08 days of atmospheric oxidation assuming an ambient OH concentration of
149 $1.5 \times 10^6 \text{ molecules cm}^{-3}$ (Mao et al., 2009). The upper limit of OH exposure was
150 determined by measuring the decay of SO_2 (Model T100, TAPI Inc., USA) in the
151 absence of toluene, following procedures described elsewhere (Kang et al., 2007;
152 Lambe et al., 2011a). The reduction in OH exposure due to the addition of toluene was
153 estimated to range from 15% at the highest OH exposure to 25% at the lowest OH
154 exposure, using the method of Peng et al. (2016). Peng et al. (2016) found that non-OH
155 chemistry, including photolysis at $\lambda = 254 \text{ nm}$ and reactions with $\text{O}(^1\text{D})$, $\text{O}(^3\text{P})$ and O_3 ,

156 may play an important role in oxidation flow reactors. In this study, the PAM reactor
157 was operated at water vapor mixing ratios above 0.5% and external OH reactivity below
158 20 s^{-1} . Non-OH chemistry is expected to play a negligible role under these conditions
159 (Peng et al., 2016).

160 Before and after each experiment, the PAM reactor was cleaned under an OH
161 exposure of $\sim 1 \times 10^{12} \text{ molecules cm}^{-3} \text{ s}$ until the mass concentration of background
162 particles dropped below $3 \mu\text{g m}^{-3}$. After characterizing dry or wet AS seed particles for
163 half an hour, the UV lamp was turned on to oxidize the background gases at five
164 different OH levels to measure the concentrations of background organics. A toluene
165 mixture (29.6 ppm in nitrogen) with a flow rate of 0.013 L min^{-1} was then introduced
166 to initiate SOA formation. The initial concentration of toluene in the PAM reactor was
167 approximately 138 ppb. The reacted and final concentrations of toluene were calculated
168 from the OH exposure and the rate constant of the reaction between toluene and OH
169 (Atkinson and Arey, 2003) (Table 1). The flow and light conditions were the same for
170 initially wet and dry seeds. Therefore, the quantification of toluene would not introduce
171 uncertainties to the relative SOA yields described in Section 3.1 as the initial
172 concentrations of toluene and OH exposures were the same for both cases. SOA was
173 measured for at least an hour with a step-wise increase in the five OH levels.

174 **2.3 Characterization of non-refractory components**

175 The AS/SOA mixed particles were characterized for the chemical composition of non-
176 refractory components including organics, sulfate and ammonium as well as the
177 elemental ratios of organics using a high-resolution time-of-flight aerosol mass

178 spectrometer (hereafter AMS, Aerodyne Research Incorporated, USA) (DeCarlo et al.,
179 2006). The instrument was operated in the high sensitivity V-mode and the high
180 resolution W-mode alternating every one minute. The toolkit Squirrel 1.57I and Pika
181 1.16I were used to analyze the AMS data. The molar ratios of hydrogen to carbon (H:C)
182 and oxygen to carbon (O:C) were determined using the Aiken method (Aiken et al.,
183 2007, 2008). The ionization efficiency of the AMS was calibrated using 300 nm
184 ammonium nitrate particles. The particle-free matrix air, obtained by passing the air
185 flow from the PAM reactor through a HEPA filter, was measured for at least 20 min
186 before each experiment to determine the signals from major gases.

187 The collection efficiency (CE) of an AMS is dependent on the chemical
188 composition and acidity as well as the phase state of particles (Matthew et al., 2008;
189 Middlebrook et al., 2012). Matthew et al. (2008) found that the CE for solid particles
190 thickly coated with liquid organics was 100%. In this study, experiments were
191 conducted at an RH of 68%, exceeding the RH threshold for the semisolid-to-liquid
192 phase transition for toluene-derived SOA (Bateman et al., 2015; Song et al., 2016). The
193 toluene-derived SOA in these experiments was therefore liquid-like. The unimodal size
194 distributions of particle numbers show the SOA formation on AS seed particles without
195 much nucleation mode particles (Fig. S2). A CE of 1 was used for processing all AMS
196 data since the AS seed particles were coated by liquid SOA. The adoption of this CE
197 value was supported by that the concentration of sulfate measured with the AMS varied
198 by less than 5% of the average mass of sulfate after coated by SOA for both wet and
199 dry AS seeds conditions. For the quantification of SOA, the contribution from

200 background organic aerosols was subtracted from the total organic aerosols. The ratio
201 of SOA mass to background organic mass ranged from 7 to 59, indicating that the
202 contribution from background organics was negligible. Aerosol particles typically pass
203 through a silica gel diffusion dryer to remove ALW before they are measured by AMS.
204 However, this may lead to some losses of semivolatile organics through reversible
205 partitioning (Wong et al., 2015; Faust et al., 2016). In this study, the AS/SOA mixed
206 particles stream passed through and bypassed a diffusion dryer alternately before they
207 were measured by AMS. Overall less than 8% of SOA were lost for wet and dry AS
208 seeds after passing the diffusion dryer (Fig. S3), possibly due to reversible partitioning
209 of the SVOCs. In this paper, the data reported are those bypassing the diffusion dryer.

210 A scanning mobility particle sizer (SMPS, TSI Incorporated, USA, classifier model
211 3082, CPC model 3775) was used to measure particle number concentrations and size
212 distributions. Particle size ranged from 15 nm to 661 nm.

213 To evaluate the influence of seed surface area on SOA formation, we conducted
214 another experiment at OH exposure of 0.47×10^{11} molecules cm^{-3} s with 50% of the seed
215 surface area used in the wet AS experiment. The difference in SOA concentration was
216 approximately 1% between these two experiments. Hence the 20% difference in seed
217 surface area as well as the difference in mass loadings between wet and dry AS particles
218 cannot account for the difference in SOA yield to be discussed below.

219 **2.4 Estimation of aerosol liquid water (ALW) content**

220 The ALW content of the initially dry AS was zero. However, as reactions proceed, SOA
221 themselves can uptake water and also lower the deliquescence RH of AS, leading to

222 water uptake by AS and some fractions of AS in aqueous phase. The ALW contents of
223 AS (ALW_{AS}) and toluene-derived SOA (ALW_{SOA}) were estimated from the following
224 equations (Kreidenweis et al., 2008):

$$225 \quad ALW_{AS} = V_{AS} \kappa_{AS} f \frac{\alpha_w}{1 - \alpha_w} \rho_w \quad (1)$$

$$226 \quad ALW_{SOA} = V_{SOA} \kappa_{SOA} \frac{\alpha_w}{1 - \alpha_w} \rho_w \quad (2)$$

227 where V_{AS} and V_{SOA} represent the volume concentrations of dry AS and SOA particles,
228 κ_{AS} is the hygroscopicity parameter of AS particles obtained from Kreidenweis et al.
229 (2008), κ_{SOA} is the hygroscopicity parameter of toluene-derived SOA calculated using
230 the linear correlation between κ_{SOA} and the O:C ratios of SOA proposed by Lambe et al.
231 (2011b), the term f is the fraction of AS particles that dissolved, α_w is the water activity
232 and ρ_w is the density of water (1.0 g cm^{-3}). Here, α_w was assumed to be equivalent to
233 RH/100 for simplicity. The volume concentrations of dry AS and SOA particles were
234 estimated from the measured mass concentration of AS and SOA assuming their
235 respective particle densities to be 1.77 g cm^{-3} and 1.4 g cm^{-3} (Ng et al., 2007).

236 For the initially wet AS seeds, all AS particles were completely aqueous and
237 therefore $f = 1$. For the initially dry AS seeds, before reactions, the AS particles were
238 completely dry and $f = 0$. After reactions, the AS particles became partially or entirely
239 deliquesced upon the formation of toluene-derived SOA. The dissolved fraction of AS
240 particles was regulated by the liquidus curve of the deliquescence relative humidity
241 (DRH(ϵ)) of AS particles coated with toluene-derived SOA (Smith et al., 2013):

$$f = \begin{cases} \frac{\varepsilon(1-\varepsilon_D)}{\varepsilon_D(1-\varepsilon)} & \text{for } \varepsilon < \varepsilon_D \\ 1 & \text{for } \varepsilon \geq \varepsilon_D \end{cases} \quad (3)$$

The term ε is the volume fraction of SOA (Table 1). The term ε_D , representing the volume fraction of organics at which the mixture of SOA and AS particles deliquesced at an RH of 68%, was estimated to be 0.75 based on the liquidus curve.

3. Results and discussion

3.1 SOA yields

Figure 2a shows SOA yields from the photooxidation of toluene on initially wet and dry AS seed particles as a function of OH exposure. The SOA yield was calculated as the SOA mass divided by the mass of reacted toluene. The uncertainty in the SOA yields simply reflected the standard derivation when averaging the SOA mass. In both cases, SOA yields first exhibited an increase, followed by a decrease as the level of OH exposure increased. This trend may be due to the transition of functionalization reactions to fragmentation ones (Kroll et al., 2009; Lambe et al., 2011a). Previous oxidation flow reactor studies suggest that gas-phase chemistry dominates over heterogeneous OH oxidation at OH levels below 1.0×10^{12} molecules cm^{-3} s (Ortega et al., 2016; Palm et al., 2016). In this study, the highest OH exposure was 5.28×10^{11} molecules cm^{-3} s and heterogeneous oxidation of SOA may not play an important role in reducing the mass of SOA, although we cannot exclude that it plays a role. In addition, glyoxal is an important oxidation product of toluene (Kamens et al., 2011). The reactive uptake of glyoxal has been demonstrated to enhance rather than reduce the SOA mass (Liggio et al., 2005a). The SOA yields for dry and wet AS seeds were 0.18–0.31 and

263 0.22–0.36, respectively, significantly higher than the value of 0.0059 observed in an
264 oxidation flow reactor under comparable conditions (Faust et al., 2017) and the value
265 of 0.09 obtained in another PAM chamber at 30% RH in the absence of seed particles
266 (Kang et al., 2007). Faust et al. (2017) attributed their significantly lower yields than
267 typical literature values of 0.09–0.30 (Lambe et al., 2011a; Ng et al., 2007) to the wall
268 loss of particles and the fragmentation of organics in their flow reactor. On the other
269 hand, the SOA yields we obtained are lower than the values of 0.30–0.37 from smog
270 chamber experiments conducted at a similar temperature, SOA mass loading and OH
271 exposure but a lower RH with dry AS seeds (Ng et al., 2007; Hildebrandt et al., 2009).
272 Note that the wall loss of particles was not corrected in this study, so the SOA yields
273 may be underestimated. As wet and dry AS seeds in this study had similar particle
274 number size distributions, the wall loss of particles would not affect the comparison of
275 SOA yield between wet and dry AS seeds.

276 As shown in Fig. 2a, a higher SOA yield was observed for wet AS seeds than for
277 dry AS seeds at the same OH exposure and the difference in SOA yield decreased as
278 the OH exposure increased. The ratio of SOA yields on wet AS seeds to those on dry
279 AS seeds, the relative SOA yield, was 1.31 ± 0.02 at an OH exposure of 0.47×10^{11}
280 molecules cm^{-3} s but decreased to 1.01 ± 0.01 when the OH exposure was increased to
281 5.28×10^{11} molecules cm^{-3} s (Fig. 2b). These ratios are comparable to the 1.19 ± 0.05
282 observed by Faust et al. (2017) at an OH exposure of approximately 2.0×10^{11} molecules
283 cm^{-3} s.

284 The formation of SOA on initially dry AS particles may alter the deliquescence

285 relative humidity (DRH) of AS particles. Smith et al. (2013) found that when coated
286 with toluene-derived SOA, the DRH of AS particles decreased from 80% to 58% as the
287 organic volume fraction increased from 0 to 0.8. Therefore, coating AS particles with
288 toluene-derived SOA can change the physical state of initially dry AS seeds and
289 increase the content of $ALW_{AS, dry}$. As shown in Fig. 3a, after reactions, the mass
290 concentrations of ALW_{tot} ($= ALW_{SOA} + ALW_{AS}$) and ALW_{SOA} increased for both wet
291 and dry seeds as the OH exposure increased. The uncertainties for ALW_{SOA} and ALW_{AS}
292 were 22% and less than 3%, respectively. They reflect the uncertainties in κ and volume
293 concentrations of AS and SOA. The increase in $ALW_{tot, wet}$ was due to the increase in
294 $ALW_{SOA, wet}$ while the increase in $ALW_{tot, dry}$ was driven by the increase in $ALW_{AS, dry}$ at
295 lower OH exposure and by $ALW_{SOA, dry}$ at higher OH exposures. At OH exposure of
296 0.47×10^{11} molecules cm^{-3} s, $ALW_{AS, dry}$ increased from 0 to $6.2 \mu g m^{-3}$ after reactions
297 due to the partial deliquescence ($f=0.43$) of the originally dry AS particles after SOA
298 formation. The difference in $ALW_{AS, dry}$ and $ALW_{AS, wet}$ narrowed and the ALW_{total} of
299 initially dry AS seeds partially resembled those of the wet ones. At OH exposure
300 between 1.66×10^{11} and 5.28×10^{11} molecules cm^{-3} s, the total final organic volume
301 fraction increased to approximately 0.8 and the initially dry AS particles entirely
302 deliquesced after reactions. Based on the reported SOA yield, initial toluene
303 concentration, OH exposure and assumed concentrations of AS seeds (~ 10 - $40 \mu g m^{-3}$)
304 in Faust et al. (2017), we estimated that an upper limit of 48% of the initially dry AS
305 seeds has deliquesced in their study. Similar to this study, SOA coatings on seed
306 particles may change the physical state of initially dry seeds and lower the difference

307 of SOA yields between initially dry and wet seeds experiments.

308 The hydrophilic products can partition more readily into initially wet AS seeds than
309 dry seeds and partially account for the difference in SOA yields. For example, as one
310 of the important oxidation products, glyoxal was estimated to have an effective Henry's
311 law constant of $4.52 \times 10^8 \text{ m atm}^{-1}$ for our initially wet AS seeds due to the "salting-in"
312 effect (Kampf et al., 2013), approximately 3 orders of magnitude higher than that in
313 pure water (Ip et al., 2009). The uptake rate constant of glyoxal can be calculated as
314 $(\gamma v A)/4$, where γ is the uptake coefficient, v is the gas-phase velocity of glyoxal, and A
315 is the total surface area of AS seeds. The uptake rate constant is $4.5 \times 10^{-4} \text{ s}^{-1}$ for initially
316 wet seeds with $\gamma = 2.4 \times 10^{-3}$ estimated from glyoxal uptake in AS seeds at 68% RH
317 (Liggio et al., 2005b). The average gas-phase glyoxal concentration was modeled to be
318 4.3 ppb at OH exposure of $0.47 \times 10^{11} \text{ molecules cm}^{-3} \text{ s}$ using the Master Chemical
319 Mechanism v 3.3.1 (Jenkin et al., 2003; Bloss et al., 2005), which would result in
320 approximately $1.6 \mu\text{g m}^{-3}$ of glyoxal in particle phase for initially wet AS seeds. If the
321 particle-phase concentration of glyoxal was assumed to be 0 for initially dry AS seeds,
322 the enhanced partitioning of glyoxal alone would account for 24.5% of the mass
323 difference of SOA. Note that other hydrophilic products were not included in this
324 calculation. This analysis suggests that the enhanced partitioning of hydrophilic
325 products may play an important role in the difference of SOA yields at low OH
326 exposures. As discussed above, the initially dry AS seeds approached wet seeds and
327 reduce the differences between wet and dry SOA yields at high OH exposures.

328 **3.2 Chemical composition of SOA**

329 Figure 4 shows the high-resolution mass spectra of SOA for initially wet and dry AS
330 seeds at OH exposures of 0.47×10^{11} molecules cm^{-3} s and 5.28×10^{11} molecules cm^{-3} s.
331 For both types of AS seeds, at an OH exposure of 0.47×10^{11} molecules cm^{-3} s, the most
332 prominent peaks were m/z 29 and 43, followed by m/z 28 and 44. m/z 29 was dominated
333 by ion CHO^+ , a tracer for alcohols and aldehydes (Lee et al., 2012). The m/z 28 and m/z
334 44 signals, respectively dominated by CO^+ and CO_2^+ , are tracers for organic acids (Ng
335 et al., 2010). At the OH exposure of 5.28×10^{11} molecules cm^{-3} s, the dominant peaks
336 were m/z 28 and 44, followed by m/z 29 and 43. The increase of mass fractions of the
337 oxygen-containing ions in the SOA mass spectra at a relatively high OH exposure
338 suggests the formation of more oxidized organic aerosols. On the basis of the mass
339 fraction of ions, Fig. S4 shows that as OH exposure increased, the difference (wet minus
340 dry) in the spectra of toluene-derived SOA changed from positive in m/z 29 (CHO^+)
341 and m/z 43 ($\text{C}_2\text{H}_3\text{O}^+$) to m/z 28 (CO^+) and m/z 44 (CO_2^+). The increase in OH exposure
342 resulted in a change from more alcohols or aldehydes to more organic acids in the wet
343 seeded case when compared to the dry seeded case.

344 Fragments derived from the AMS data have been extensively used to infer the bulk
345 compositions and evolution of organic aerosols (Zhang et al., 2005; Ng et al., 2010;
346 Heald et al., 2010). Here we used the approach of Ng et al. (2010) and plotted the
347 fractions of the total organic signal at m/z 43 (f_{43}) vs. m/z 44 (f_{44}) as well as the triangle
348 based on the analysis of ambient AMS data (Fig. 5). Ng et al. (2010) proposed that
349 aging would cause f_{43} and f_{44} to converge toward the triangle apex ($f_{43} = 0.02$, $f_{44} = 0.30$).
350 For both wet and dry AS seeds, f_{43} first increased and then decreased with the increase

351 of OH exposure, while f_{44} increased all the time. This reversing trend of f_{43} was the
352 result of the increase and subsequent decrease in $C_2H_3O^+$ (Fig. S5), an indicator of
353 products containing carbonyl functional groups. The f_{43} - f_{44} plot supports our earlier
354 assertion that as OH exposure increased, the reaction products changed from earlier-
355 generation products containing carbonyl functional groups dominated to later-
356 generation products containing acidic functional groups dominated. It was also
357 observed for SOA formed from other precursors such as alkanes and naphthalene
358 (Lambe et al., 2011b). Before the decrease in f_{43} , SOA formed on wet AS seeds had
359 higher f_{43} and similar f_{44} to SOA formed on dry AS seeds at the same OH exposure. As
360 OH exposure increased, SOA formed on wet AS seeds had higher f_{44} and lower f_{43} than
361 SOA formed on dry AS seeds. In addition, as OH exposure increased, SOA formed on
362 wet AS seeds initially had more earlier-generation products but later had more acidic
363 later-generation products than SOA formed on dry AS seeds, likely due to the enhanced
364 partitioning of these products on initially wet AS seeds and/or enhanced uptake of
365 water-soluble gases through aqueous phase reactions.

366 Figure 6 shows the changes in H:C and O:C ratios as a function of OH exposure in
367 a Van Krevelen diagram (Heald et al., 2010). The standard deviations for H:C and O:C
368 values, determined for the steady-state periods, were all less than 0.01. The O:C ratios
369 for dry and wet AS seeds were in the ranges of 0.59–0.89 and 0.63–0.95, respectively.
370 At the same OH exposure, SOA on wet AS seeds had both higher O:C ratios and
371 estimated average carbon oxidation state (OS_C) ($OS_C \approx 2 \times O:C - H:C$) (Kroll et al., 2011)
372 than dry AS seeds had. Fig. 6 also shows some of the identified SOA products from the

373 photooxidation of toluene (Bloss et al., 2005; Hamilton et al., 2005; Sato et al., 2007).
374 The elevated OS_C (exceeding 0.5) could only be due to the formation of highly
375 oxygenated small acids such as pyruvic acid ($OS_C = 0.67$), glycolic acid ($OS_C = 1$),
376 formic acid ($OS_C = 2$), oxalic acid ($OS_C = 3$), malonic acid ($OS_C = 1.33$) and glyoxylic
377 acid ($OS_C = 2$). Small acids may be important products of toluene-derived SOA at high
378 OH exposures. Fisseha et al. (2004) found that small organic acids accounted for 20–
379 45% of SOA from the photooxidation of 1,3,5-trimethylbenzene. The higher OS_C at
380 high OH exposures for wet AS seeds might suggest that these small acids were more
381 abundant, likely due to their enhanced retention in the presence of ALW and/or the
382 more efficient uptake of OH radicals by wet AS seeds and further oxidation reactions
383 in aqueous phase (Ruehl et al., 2013).

384 We evaluate whether enhanced uptake of OH radicals on initially wet AS seeds
385 could explain the difference in oxygen contents, following the method of DeCarlo et al.
386 (2008). We calculated R , the ratio of the difference in oxygen of OA between the
387 initially wet and dry AS seed particles to the difference in the total number of OH
388 collisions with OA at different OH exposures. To obtain R , the uptake coefficient (γ) of
389 OH radicals was assumed to be 1 and 0.1/0.8 (lower/upper limit) for initially wet and
390 dry AS seed particles, respectively (George and Abbatt, 2000). Note that as SOA
391 formation takes place, the initially dry AS can become wet and the difference in γ
392 between initially wet and dry seeds is reduced, especially at higher OH exposures. We
393 also assumed that each collision of OH with OA resulted in the addition of one oxygen
394 atom to SOA. A value of R smaller than unity qualitatively indicates that the uptake of

395 OH radicals can potentially explain the differences in oxygen contents in the dry and
396 wet experiments. Fig. S6 shows that R is larger than unity at low OH exposures and
397 smaller than unity at high OH exposures. This analysis suggests that the enhanced OH
398 uptake may contribute to the difference in oxygen contents between dry and wet cases
399 at higher OH exposures. At low OH exposures, the enhanced gas-particle partitioning
400 may dominate the difference.

401 The change in the slope of H:C vs O:C is consistent with the earlier analysis that
402 the mechanism of SOA formation changed from functionalization dominated by the
403 addition of alcohol/peroxide (Heald et al., 2010; Ng et al., 2011) at low exposures to
404 the addition of both acid and alcohol/peroxide functional groups without fragmentation,
405 and/or the addition of acid groups with fragmentation at high exposures.

406 **3.3 Atmospheric implications**

407 In this work, yields and composition of SOA formed from the photooxidation of toluene
408 on initially wet and dry AS seeds were compared over a wide range of OH exposures,
409 covering the transition from functionalization reactions to fragmentation reactions. We
410 found that the ratio of SOA yield on wet AS seeds to that on dry AS seeds decreased
411 from 1.31 to 1.01 as the OH exposure increased from 0.47×10^{11} to 5.28×10^{11} molecules
412 cm^{-3} s. This decrease coincides with the decrease of differences in ALW between the
413 wet and dry cases, which may be due to water uptake by SOA as well as the early
414 deliquescence of dry AS particles as a result of SOA formation. Hence, the SOA yield
415 and ALW of the initially dry AS seeds approached those of the initially wet AS seeds
416 as OH exposure and ALW increased.

417 In addition to relatively higher SOA yields, higher O:C and OS_c of SOA derived
418 from the photooxidation of toluene were also observed on initially wet AS seeds.
419 Particularly, the O:C in the presence of initially wet AS seeds could be as high as 0.95.
420 Chen et al. (2015) observed large gaps between laboratory and ambient measured O:C
421 of OA and suggested that OA having a high O:C (> 0.6) was required to bridge these
422 gaps. The multiphase oxidation of toluene in the presence of wet aerosols may be a
423 pathway to contribute to this gap. However, the relative importance of such chemistry
424 to the evolution of ambient OA remains unclear.

425 Our results suggest that dry seeds would quickly turn to at least partially
426 deliquesced particles upon SOA formation under moderate RH conditions. We only
427 studied the photooxidation of toluene in the absence of NO_x as it is still a challenge to
428 study high-NO chemistry in oxidation flow reactors without using atmospherically
429 irrelevant high concentrations of NO_x (Peng and Jimenez, 2017). However, the ALW
430 may also be important to SOA formation under high NO_x conditions that preferentially
431 form highly water-soluble products (Ervens et al., 2011). Since ambient RH is rarely at
432 such low values that inorganic particles remain dry even after SOA formation, more
433 laboratory and field studies are needed to elucidate the formation and evolution of OA
434 under various NO_x conditions at moderate RH.

435 **Acknowledgments**

436 The work described in this paper was sponsored by the Science Technology and
437 Innovation Committee of Shenzhen Municipality (project no.
438 JCYJ20160401095857424). Zijun Li and ManNin Chan are supported by a Direct
439 Grant for Research (4053159), The Chinese University of Hong Kong and a Research
440 Grants Council grant (RGC 2191111). Chak K. Chan would like to thank the Hong
441 Kong University of Science and Technology for the use of the AMS.

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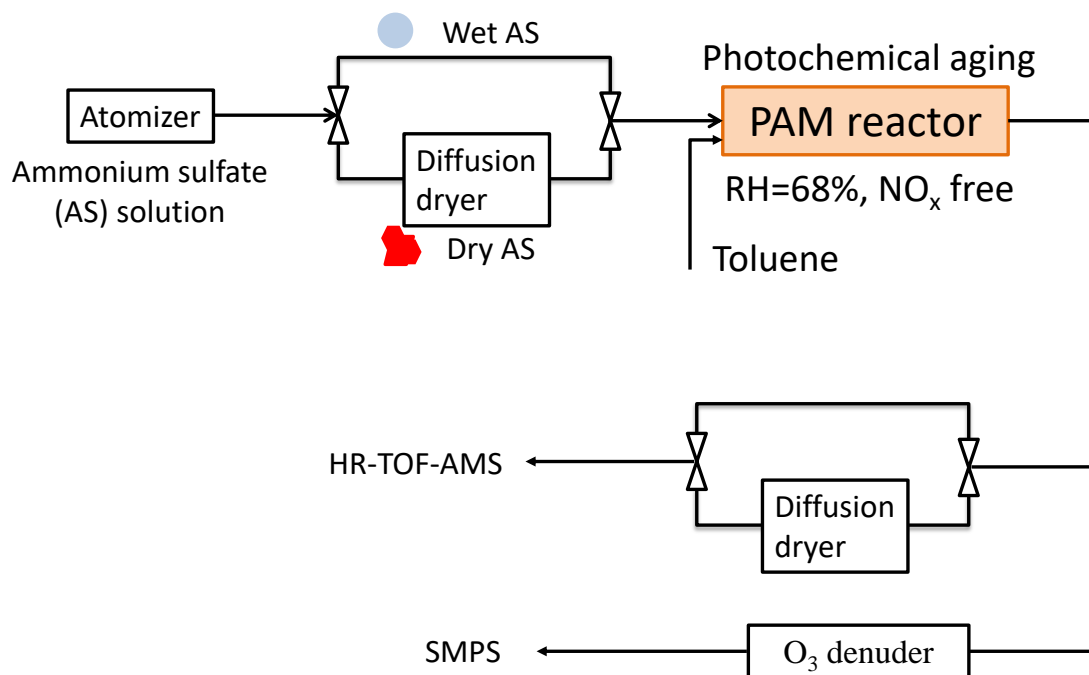
779

780 **Table 1.** Summary of the results for the initially dry and wet AS seeds experiments.

OH exposure ($\times 10^{11}$ molecules cm^{-3} s)	[toluene] _{reacted} (ppb)	[toluene] _{final} (ppb)	ε^a	
			wet AS	dry AS
0.47	32.4	106.0	0.57	0.56
1.66	84.9	53.5	0.82	0.82
2.97	113.1	25.3	0.83	0.85
4.34	126.9	11.5	0.83	0.85
5.28	131.7	6.7	0.83	0.85

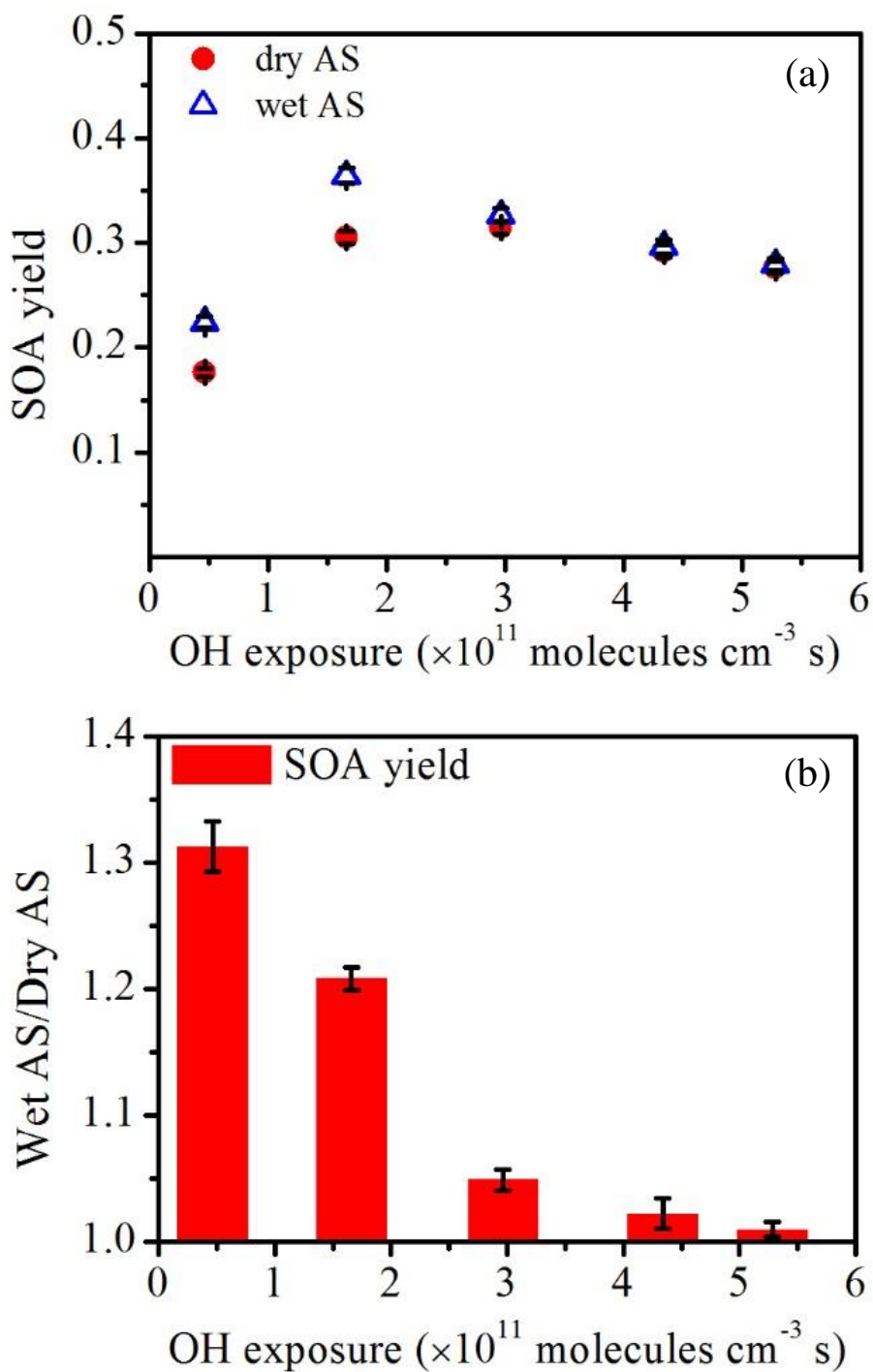
781 ^a The volume fraction of organics.

782



783

784 **Fig. 1.** Schematic of the experimental setup. The aqueous ammonium sulfate (AS) seed
785 particles either passed through a diffusion dryer so that the phase of the seed particles
786 could be altered or bypassed the diffusion dryer. Either wet or dry AS served as seed
787 particles for the experiments.



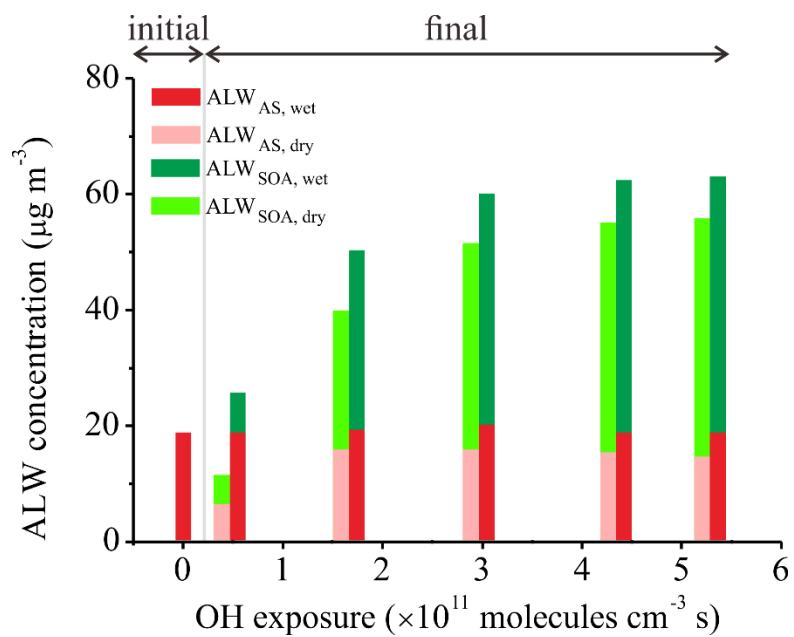
788

789 **Fig. 2.** (a) Yield of toluene-derived SOA formed on initially wet and dry AS as a

790 function of OH exposure. (b) Ratio of SOA yields on initially wet AS to those on

791 initially dry AS as a function of OH exposure.

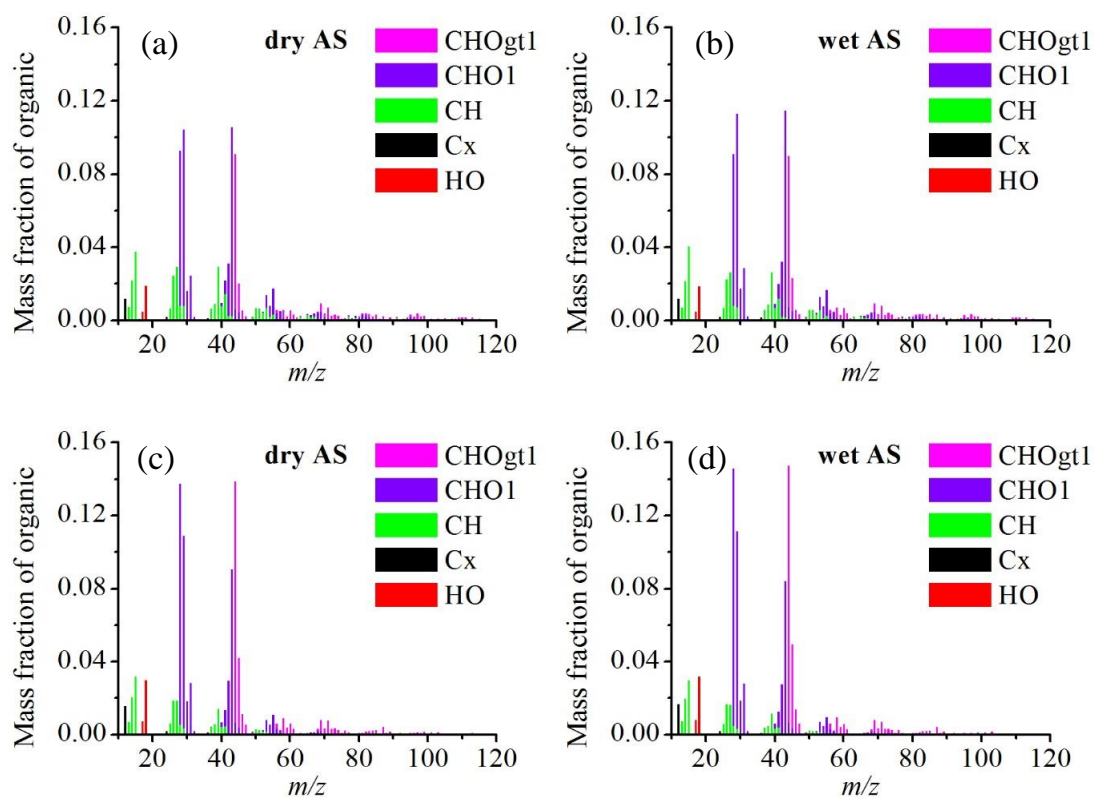
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793

794 **Fig. 3.** Mass concentration of ALW uptake by AS and toluene-derived SOA before
 795 (initial) and after reactions (final) for both initially wet and dry AS seeds. Adjoining
 796 bars for initially wet and dry seeds have same OH exposures.

797

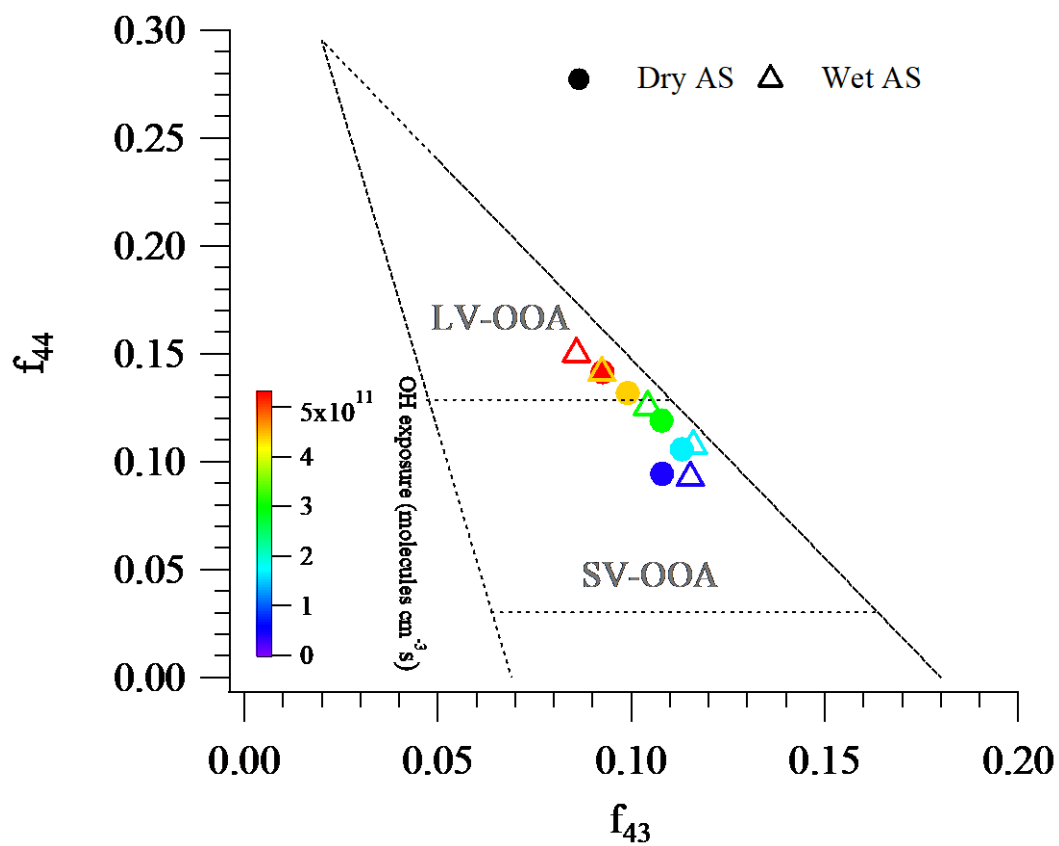


798

799 **Fig. 4.** High-resolution mass spectra of toluene-derived SOA on initially wet and dry

800 AS at an OH exposure of (a, b) 0.47×10^{11} molecules cm^{-3} s and (c, d) 5.28×10^{11}

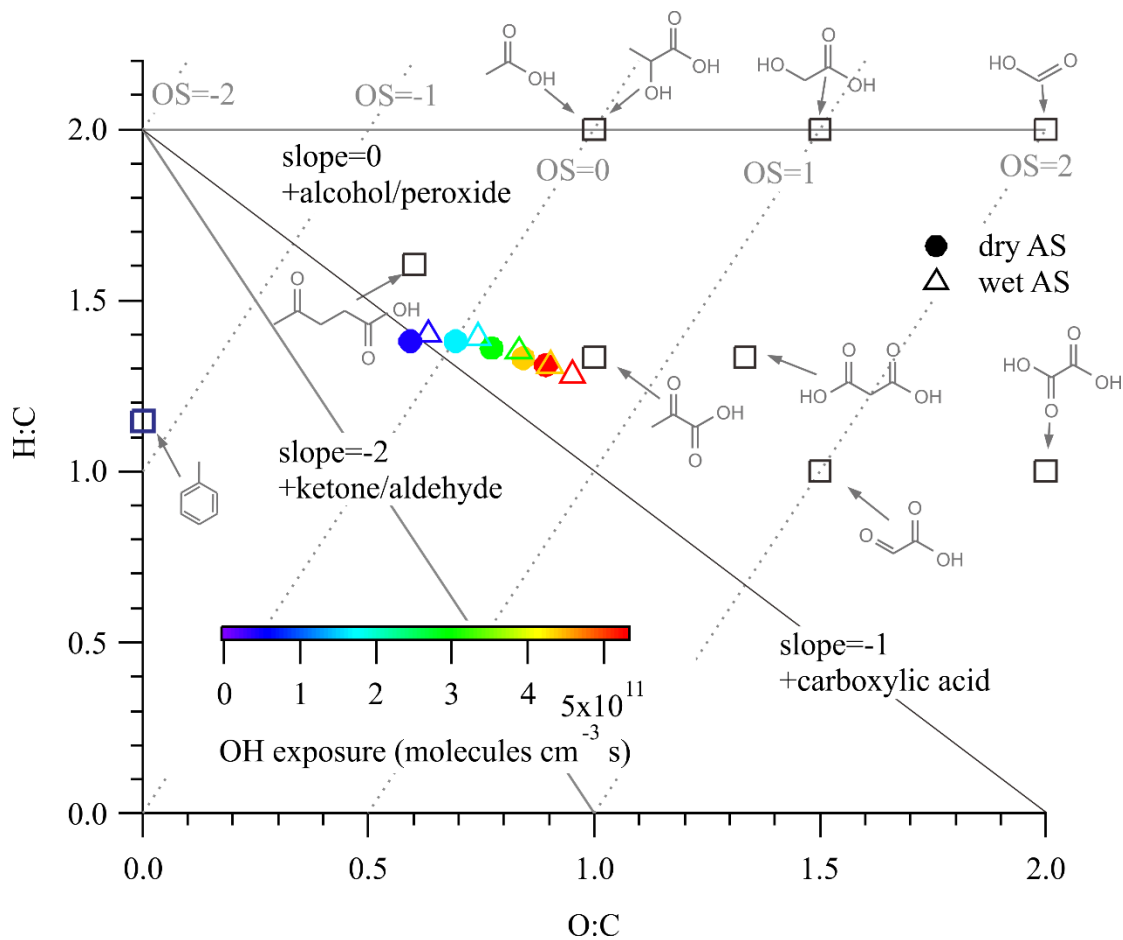
801 molecules cm^{-3} s.



802

803 **Fig. 5.** Fractions of total organic signal at m/z 43 (f_{43}) vs. m/z 44 (f_{44}) from SOA data
 804 obtained in this study together with the triangle plot of Ng et al. (2010). Ambient SV-
 805 OOA and LV-OOA regions are adapted from Ng et al. (2010). Data are colored
 806 according to the OH exposure.

807



808

809 **Fig. 6.** Van Krevelen diagram of SOA derived from the photooxidation of toluene on
 810 initially wet and dry AS seed particles. SOA data are colored according to the OH
 811 exposure. Products identified in toluene-derived SOA are shown in boxes (Bloss et al.,
 812 2005; Hamilton et al., 2005; Sato et al., 2007). Average carbon oxidation states from
 813 Kroll et al. (2011) and functionalization slopes from Heald et al. (2010) are shown for
 814 reference.

815