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1 Changing trends and emissions of hydrochlorofluorocarbons and their

2 hydrofluorocarbon replacements.

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18 Abstract

High frequency, in situ global observations of HCFC-22 (CHClF₂), HCFC-141b (CH₃CCl₂F), 19 20 HCFC-142b (CH₃CClF₂) and HCFC-124 (CHClFCF₃) and their main HFC replacements HFC-134a 21 (CH₂FCF₃), HFC-125 (CHF₂CF₃), HFC-143a (CH₃CF₃), and HFC-32 (CH₂F₂) have been used to determine their changing global growth rates and emissions in response to the Montreal Protocol and its 22 recent amendments. The 2007 adjustment to the Montreal Protocol required the accelerated phase-out of 23 HCFCs with global production and consumption capped in 2013, to mitigate their environmental impact 24 25 as both ozone depleting substances and important greenhouse gases. We find that this change has coincided with a reduction in global emissions of the four HCFCs with aggregated global emissions in 26 2015 of 444 \pm 75 Gg/yr, in CO₂ equivalent units (CO₂ e) 0.75 \pm 0.1 Gt/yr, compared with 483 \pm 70 Gg/yr 27 $(0.82 \pm 0.1 \text{ Gt/yr CO}_2 \text{ e})$ in 2010. (All quoted uncertainties in this paper are 1 sigma). About 80% of the 28 total HCFC atmospheric burden in 2015 is HCFC-22, where global HCFC emissions appear to have been 29 relatively constant in spite of the 2013 cap on global production and consumption. We attribute this to a 30 probable increase in production and consumption of HCFC-22 in Montreal Protocol Article 5 31 (developing) countries and the continuing release of HCFC-22 from the large banks which dominate 32 HCFC global emissions. Conversely, the four HFCs all show increasing annual growth rates with 33 34 aggregated global HFCs emissions in 2015 of 329 ± 70 Gg/yr (0.65 ± 0.12 Gt/yr CO₂ e) compared to 2010 with 240 ± 50 Gg/yr (0.47 ± 0.08 Gt/yr CO₂ e). As HCFCs are replaced by HFCs we investigate the 35 impact of the shift to refrigerant blends which have lower global warming potentials (GWPs). We also 36





note that emissions of HFC-125 and HFC-32 appear to have increased more rapidly during the 2011-2015

- 38 5-yr period compared to 2006-2010.
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1. Introduction

42 Hydrochlorofluorocarbons (HCFCs) were introduced between the 1940s and 1980s as alternatives to chlorofluorocarbons (CFCs) in some refrigeration and air conditioning applications. Production and 43 consumption grew rapidly in "non-Article 5" developed countries until the mid-1990s. However, because 44 they are ozone depleting substances (ODS), HCFCs were included in the 1992 Montreal Protocol 45 46 amendment, with a view to eventual phase-out of production and consumption. Subsequently the 2007 adjustments to the Montreal Protocol required an accelerated phase-out of HCFCs in both "non-Article 5" 47 and "Article 5" developing countries, with a 2013 cap on global production and consumption of HCFCs. 48 Historically, HCFCs-22, -141b and -142b account for >90% of the total consumption of all HCFCs in 49 Article 5 countries. 50

A more rapid phase-out of the HCFCs should result in a faster recovery of the depleted 51 52 stratospheric ozone layer, with the additional benefit of mitigating climate change, since these compounds are also potent greenhouse gases (GHGs). Detailed studies of the rates of atmospheric 53 54 accumulation of the HCFCs indicate periods of rapid growth, temporary slowing followed by accelerated growth (Oram et al., 1995; Simmonds et al., 1998; O'Doherty et al., 2004; Reimann et al., 2004; Derwent 55 56 et al., 2007; Montzka et al., 2009; Miller et al., 2010; Saikawa et al., 2012; Fortems-Cheiney et al., 2013; Rigby et al., 2014; Graziosi et al., 2015). More recently the global growth rates of HCFC-22 and HCFC-57 58 142b have slowed significantly and Montzka et al., (2015) reported that the 2007 adjustments to the Montreal Protocol had limited HCFC emissions significantly prior to the 2013 cap on global production, 59 although the atmospheric growth rate of HCFC-141b had almost doubled between 2007 and 2012. 60

HFCs which have been introduced as replacements for the HCFCs and CFCs have grown rapidly
in abundance since their introduction (Montzka et al., 1996, 2004; Oram et al., 1996; Reimann et al.,
2004; O'Doherty et al., 2009, 2014; Carpenter et al., 2014; Rigby et al., 2014). As discussed by Velders
et al., (2009, 2015) projected HFCs emissions may make a large contribution to future climate forcing if
they are used in the transition away from ODSs.

In this study we focus on high frequency atmospheric measurements (6 -12 per day) of HCFC-22,
HCFC-141b, HCFC-142b and HCFC-124 and their main replacements HFC-134a, HFC-125, HFC143a, and HFC-32 from the five core globally-distributed Advanced Global Atmospheric Gases
Experiment (AGAGE) sites with 10-20 year records (Prinn et al., 2000). We have previously estimated





- 70 global emissions of HFC-152a (CH₃CHF₂) using the same modelling methods discussed in this paper
- 71 (Simmonds et al., 2015).
- 72 The Ozone Depletion Potential (ODP), Global Warming Potential (GWP) and atmospheric
- lifetimes of these eight compounds are listed in Table 1. 73
- Table 1. Lifetimes (yr), Ozone Depletion and Global Warming Potentials (100-yr time horizon) for the 74

	-	h	
Compound	^a ODP	^b GWPs	<i>^cLIFETIME</i>
HCFC-22	0.055	1810	12
HCFC-141b	0.11	725	9.4
HCFC-142b	0.065	2310	18
HCFC-124	0.02	609	6
HFC-134a	0*	1430	14
HFC-143a	0*	4470	51
HFC-125	0*	3500	31
HFC-32	0*	675	5.4

75 HCFCs and HFCs reported in this study.

* see Hurwitz et al., 2015. 76

Notes: ^a ODPs from the Montreal Protocol, ^b GWPs from (Forster et al., 2007), 77

- ^cLifetimes, from SPARC, (2013. Report No. 6, WCRP-15/2013). 78
- 79

We combine these observations with a 2-dimensional (12-box) atmospheric chemical transport 80 model whose circulation is based on observations and on tuning to provide good agreements with global 81 distributions of reactive and stable trace gases (Cunnold et al., 1983; Rigby et al., 2013; 2014). We then 82 estimate global emissions which we relate to the global phase-out and adoption schedules of HCFCs and 83 84 HFCs, respectively. We compare these estimated global emissions with HCFC and HFC emission estimates compiled from national reports to the United Nations Environment Programme, UNEP 85 86 (consumption of HCFCs) and United Nations Framework Convention on Climate Change, UNFCCC (emissions of HFCs), respectively, and Emissions Database for Global Atmospheric Research (EDGAR 87 88 v4.2; http://edgar.jrc.ec.europa.eu/, HFCs), using the same techniques reported earlier for these HCFCs and HFCs (O'Doherty et al., 2009, 2014; Miller et al., 2010; Rigby et al., 2014). We recognise that due to 89 the historic range of use of these substances (refrigeration, foam blowing, and fire-fighting equipment), 90 91 the derivation of emissions from production and consumption data is difficult given the large and long-

92 lasting banks of these compounds.





93 We examine the evolution of the changing growth rates of the HCFCs with a view to determining 94 if the 2013 cap on their production and consumption has been reflected in an accelerated phase-out. Furthermore, we examine the rapid growth rates of the HFCs and whether these reflect manufacturers of 95 air-conditioning and refrigeration equipment switching to HFC refrigerant blends with lower GWPs. 96 97 HCFC-22 is used in commercial and domestic refrigeration, air conditioning, extruded polystyrene foams and as a feedstock in the manufacture of fluoropolymers. HCFC-141b and HCFC-98 99 142b are primarily used as foam blowing agents; in addition, HCFC-141b is used as a solvent in 100 electronics and precision cleaning applications; HCFC-142b is also used as an aerosol propellant and as a refrigerant. HCFC-124 has uses in specialized air conditioning equipment, refrigerant mixtures, fire 101 102 extinguishers and as a component of sterilant mixtures. HFC-134a has been used since the early 1990s in vehicle air conditioning systems and other 103 104 refrigeration and air conditioning largely to replace CFC-12. Other uses include plastic foam blowing, as a cleaning solvent and as a propellant. HFCs-125 and -32 have been used as a 50:50 blend (R-410A) in 105 residential air conditioning systems as well as in 3-component blends with HFC-134a. HFC-125 has also 106 found application as a fire suppressant agent. HFC-143a is predominantly a component of refrigerant 107 blends used in commercial refrigeration and in some air conditioning applications. 108 109 Although the HFCs are not significant ozone depleting substances (non-zero ODPs; Hurwitz et al., 2015), as GHGs the HFCs -143a, -125,-134a and -32 have global warming potentials (GWP 100-yr 110 horizon) of 4470, 3500, 1430, and 675, respectively. The HCFCs, in addition to their ozone depletion 111 potentials (ODPs), listed in Table 1, are also GHGs with GWPs comparable to the HFCs. This 112 combination of ozone depletion and climate forcing has provided the impetus for the accelerated phase-113 out of the HCFCs. 114

- 115 2. Materials and Methods
- 117 2.1. AGAGE in situ measurements.

The data used here are compiled from *in situ* measurements at the core AGAGE sites, listed in
Table 2, which shows the time frame when the measurements of individual HCFCs and HFCs began at
each AGAGE site.

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122 2.2. AGAGE Instrumentation and Measurement Techniques

123 Two similar measurement technologies have been used at AGAGE stations over time, both based

- on gas chromatography coupled with mass spectrometry (GC-MS) and cryogenic sample pre-
- 125 concentration techniques. The earlier instrument, referred to as the GC-MS-ADS, incorporated an
 - 4





126 Adsorption-Desorption-System (ADS) based on a Peltier-cooled microtrap maintained at -50°C during the adsorption phase (Simmonds et al. 1995; Prinn et al., 2000) and was used for several years at the 127 Mace Head and Cape Grim sites. These were replaced by another GC-MS instrument, the GC-MS-128 Medusa with doubled sampling frequency and enhanced cooling to $\sim -180^{\circ}$ C, which uses the milder trap 129 130 adsorbent HayeSep D, to extend compound selection, and improve measurement precisions (Miller et al., 2008; Arnold et al., 2012). 131 132 The GC-MS-Medusa system is currently deployed at all AGAGE sites used in this study (Table 133 2). Typically for each measurement the analytes from two litres of air are collected on the sample traps and desorbed onto a single main capillary chromatography column (CP-PoraBOND Q, $0.32 \text{ mm ID} \times 25$ 134 135 m, 5 µm, Agilent Varian Chrompack, batch-made for AGAGE applications) purged with helium (grade 6.0) that is further purified using a heated getter purifier (He-purifier HP2, VICI, USA). The separation 136 137 and detection of the compounds are achieved using Agilent Technology GCs (model 6890N) and quadrupole mass spectrometers in selected ion mode (initially model 5973, progressively converted to 138 5975C over the later years). For the field GC-MS-Medusa instruments, ambient air samples are analysed 139 140 every 2 hours (c.f. 4 hours for the GC-MS-ADS) and are bracketed by measurements of quaternary 141 standards to detect and correct for short-term drift in instrument response. The quaternary standards are whole-air pressurized into 34 L internally electropolished stainless steel canisters (Essex Industries, 142 143 USA). They are filled by the groups who are in charge of the respective AGAGE stations using modified oil-free diving compressors (SA-3 and SA-6, RIX Industries, USA) to ~60 bar (older tanks to ~30 bar). 144 145 Cape Grim is an exception, where the canisters used for quaternary standard purposes are typically filled cryogenically. The on-site quaternary standards are compared weekly to tertiary standards from the 146 147 central calibration facility at the Scripps Institution of Oceanography (SIO) in order to propagate the primary calibration scales and to characterize any potential long-term drift of the measured compounds in 148 149 the quaternary standards. Importantly, all of the stations report HCFC and HFC measurements relative to the SIO (SIO-05, SIO-07 and SIO-14) and University of Bristol (UB-98) calibration scales. 150 The GC-MS-Medusa measurement precisions for the four HCFCs and four HFCs are estimated 151 based on the repeated measurements of the quaternary standards. They are typically for HCFC-22 (0.5 152 ppt, 0.2%), -141b (0.05 ppt, 0.3%), -142b (0.05 ppt, 0.25%) and -124 (0.02 ppt, 1.6%) and for HFC-134a 153 (0.2 ppt, 0.2%), -125(0.05 ppt, 0.3%), -143a (0.1 ppt, 0.4%) and -32 (0.1 ppt, 0.8%). 154 155 156 157 158 159





- 160 Table 2. Overview of the core AGAGE sites used in this study, their coordinates and periods for which
- 161 data are available.
- 162

Site	Latitude	Longitude	ADS Data*	Medusa Data**
Mace Head Ireland	53 3° N	9 9° W	1994-2004	June 2003-present
	JJ.J IN	9.9 W	1994-2004	Marsh 2005 areas
I rinidad Head, California, USA	41.0° N	124.1° W		March 2005-present
Ragged Point, Barbados	13.2° N	59.4° W		May 2005-present
Cape Matatula, American Samoa	14.2° S	170.6° W		May 2006-present
Cape Grim, Tasmania, Australia	40.7° S	144.7° E	1998-2004	Jan 2004-present

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164 * Period of HCFC and HFC data records using GC-MS-ADS.

165 ** Period of HCFC and HFC data records using GC-MS-Medusa.

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167 2.3. Calibration Scales

168	The estimated accuracy of the calibration scale for the various HCFCs and HFCs is reported
169	below and a more detailed discussion of the measurement technique and calibration procedure have been
170	reported elsewhere (Miller et al., 2008; O'Doherty et al., 2009; Mühle et al., 2010). The AGAGE
171	measurements for the HCFCs-22, -141b, -142b, and HFC-134a, are reported relative to the Scripps
172	Institution of Oceanography (SIO-05) calibration scale (as dry gas mole fractions in pmol mol ⁻¹). This
173	scale is defined through the gravimetric preparation of 13 synthetic primary standards at near-ambient
174	mole fractions (Prinn et al., 2000) at SIO in 2005. HCFC-124 is derived from the UB-98 calibration scale.
175	HFCs -143a, -32 are reported relative to the SIO-07 and HFC-125 to the SIO-14 calibration scales.
176	The overall accuracies of these primary standards sets are liberally estimated at 2% for HCFC-22,
177	-141b, and -142b, 10%, for HCFC-124 and 1.5% for HFC-134a, and 3% for HFC-125, -143a, and
178	-32 with the largest fractional uncertainty contributed from the impurities in the starting reagents.
179	
180	2.4. Selection of baseline data
181	Baseline in situ monthly mean HCFC and HFC mole fractions were calculated by excluding
182	values enhanced by local and regional pollution influences, as identified by the iterative AGAGE
183	pollution identification algorithm, (for details see Appendix in O'Doherty et al., 2001). Briefly, baseline
184	measurements are assumed to have Gaussian distributions around the local baseline value, and an
185	iterative process is used to filter out the points that do not conform to this distribution. A second-order
186	polynomial is fitted to the subset of daily minima in any 121-day period to provide a first estimate of the
187	baseline and seasonal cycle. After subtracting this polynomial from all the observations a standard
188	deviation and median are calculated for the residual values over the 121-day period. Values exceeding
189	three standard deviations above the baseline are thus identified as non-baseline (polluted) and removed





- 190 from further consideration. The process is repeated iteratively to identify and remove additional non-
- baseline values until the new and previous calculated median values agree within 0.1%.
- 192

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193 **3. Modelling studies**

195 There are several sources of information on production and emissions of HCFCs and HFCs; none of 196 which, on their own, provides a complete database of global emissions. The more geographically comprehensive source of information for HFC emissions is provided by the parties to the UNFCCC, but 197 198 only includes Annex 1 countries (developed countries). The 2014 database covers years 1990 to 2012 and emissions are reported in Table 2(II) s1 in the Common Reporting Format (CRF) available at 199 200 http://unfccc.int/national-reports/annex ighg inventories/national inventories submissions/items/8108.php. An alternative inventory estimate was also obtained from the Emissions Database for Global 201 Atmospheric Research (EDGAR v4.2; http://edgar.jrc.ec.europa.eu/), a database that estimates global 202 emission inventories of anthropogenic GHGs, including HFCs on a country, regional and gridded basis 203 204 up to 2008. Similar emission estimates are not available for HCFCs, but using HCFC consumption data 205 published by the Montreal Protocol Secretariat of the United Nations Environment Programme (UNEP, 206 2016a) we calculate HCFC emissions as described in the supplementary material (1). 207 Such bottom-up emission estimates of HFCs and HCFCs are based on industry production, 208 209 imports, distribution and usage data for these compounds, reported to national governments and thence to 210 UNEP and UNFCCC. We discuss these independent emission estimates because they are helpful as a 211 priori data constraints on our model analysis and to compare them with our observation-based top-down estimates. 212 213 3.1. Global emissions estimates using the AGAGE two-dimensional 12-box model. 214 215 To estimate global-average mole fractions and derive growth rates, a two-dimensional model of atmospheric chemistry and transport was employed. The AGAGE 12-box model simulates trace gas 216 217 transport in four equal mass latitudinal sections (divisions at 30-90°N, 0-30°N, 30-0°S and 90-30°S) and at three heights (vertical divisions at 200, 500 and 1000 hPa). The model was originally developed by 218 Cunnold et al. (1983) (nine-box version), with subsequent improvements by Cunnold et al. (1994) and 219 Rigby et al. (2014). Emissions were estimated between 1998 and 2015 using a Bayesian method in which 220 221 an a priori constraint (EDGAR v4.2) on the emissions growth rate was adjusted using the baseline-222 filtered AGAGE observations (Rigby et al., 2011, 2014). Global emissions were derived that included





- estimates of the uncertainties due to the observations, the prior and the current best-estimate lifetimes of these compounds from SPARC (2013); as detailed in the supplementary material in Rigby et al. (2014).
- 225

226 4. Results and Discussion

227 4.1. Atmospheric Mole Fractions

Based on the output from the 12-box model, into which AGAGE observations had been assimilated, Figure 1 illustrates the global mean mole fractions for the four HCFCs and the four HFCs, (the model output was used for "gap-filling" purposes). Figure 2 shows the average annual growth rates. Global mean mole fractions of HCFC-22, -141b, and -142b have increased throughout the observation period reaching 234, 24.3 and 22.4 pmol mol⁻¹, respectively in 2015. HCFC-124 reached a maximum global mean mole fraction of 1.48 pmol mol⁻¹ in 2007 and has since declined by 23% to 1.14 pmol mol⁻¹ in 2015.

- The HFCs all show increasing global mean mole fractions and growth rates over the entire period
- of observations. In 2015 the global mean mole fractions (pmol mol⁻¹) in descending order of abundance
- 237 are HFC-134a (83.3), HFC-125 (18.4), HFC-143a (17.7) and HFC-32 (10.5) with growth rates (pmol
- 238 mol⁻¹ yr⁻¹) for HFC-134a (5.6), HFC-125 (2.3), HFC-143a (1.5) and HFC-32 (1.6).







Figure 1. Global mean mole fractions for the four HCFCs and the four HFCs, (the model output was used
for "gap-filling" purposes). Shading in the figure reflects the uncertainty on the mole fractions derived in
the inversion and includes a contribution from random and scale-related measurement errors and
modelling uncertainties (further details are provided in Rigby et al., 2014). Note that HCFC-124, HFC143a and HFC-32 use only GC-MS-Medusa data for these calculations; all others use combined GC-MSADS and GC-MS-Medusa data. HFCs shown with individual annual mole fractions.







Figure 2. Annual average global mean mole fraction growth rates (pmol mol⁻¹ yr⁻¹) determined from the
12-box model for the HCFCs and HFCs. Note that HCFC-124, HFC-143a and HFC-32 use only Medusa
data for these calculations; all others use combined GC-MS-ADS and GC-MS-Medusa data. HFCs shown
with individual annual growth rates.

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307	Global mean HCFC-22 reached a maximum annual rate of increase of 8.2 pmol mol ⁻¹ yr ⁻¹ in 2007
308	and slowed by 54% to 3.7 pmol mol ⁻¹ yr ⁻¹ in 2015. The global mean annual growth rates of HCFC-141b
309	reached a first maximum of 1.9 pmol mol ⁻¹ yr ⁻¹ in 1998, followed by a second maximum in 2011 of 1.0
310	pmol mol ⁻¹ yr ⁻¹ and then slowed to ~0.3 pmol mol ⁻¹ yr ⁻¹ in 2015, a 70% decline. Similarly, HCFC-142b
311	reached a maximum in 2008 of 1.1 pmol mol ⁻¹ yr ⁻¹ , followed by a steep 90% decline to just 0.11 pmol
312	mol ⁻¹ yr ⁻¹ . These observations reflect substantial changes in the quantity of HCFCs emitted to the
313	atmosphere over time with the rates of increase in 2015 considerably slower than their historical highs
314	and broadly in response to the MP and its amendments.
315	
316 317	4.2. Top-down Emission Estimates
318	4.2.1. Global estimates of HCFCs
319	Estimated annual global emissions (Gg/yr) of the HCFCs using the 12-box model (emissions
320	listed in supplementary material 1) and those calculated from consumption reported by UNEP (2016a, see
321	supplementary material 2) and EDGAR emission inventories are shown in Figure 3 a-d. The blue solid
322	line represents our model-derived emissions, with the uncertainties shown by the shaded areas.
323	
324	4.2.1.1. HCFC-22
325	Model derived global emissions of HCFC-22 increased from 234 ± 35 Gg/yr in 1995 to a
326	maximum of 383 ± 54 Gg/yr in 2010 increasing by ~10 Gg/yr. Since 2010 global HCFC-22 emissions
327	have declined by perhaps 6.6% to 357 \pm 58 Gg/yr in 2015. Figure 3a includes estimated emissions from
328	UNEP and other reported global HCFC-22 emissions estimates (Saikawa et al., 2012; Xiang et al., 2014;
329	Montzka et al., 2015) which all agree within the uncertainties of our estimates. Fortems-Cheiney et al.,
330	(2013) using observations from multiple networks, an inversion and a new gridded bottom-up inventory
331	estimated global emissions of 387 ± 9 Gg/yr in 2010, very close to the estimated HCFC-22 emissions
332	derived in this study.
333 334	4.2.1.2. HCEC-141b and HCEC-142b
335	These two HCFCs have exhibited similar but fluctuating emissions with maxima in 2000 of $63 \pm$
336	6 Gg/vr (HCFC-141b) and 31 + 6 Gg/vr (HCFC-142b) followed by a decline during 2004-2005 to 46 + 7
337	Gg/yr (HCFC-141b) and $29 + 5 Gg/yr$ (HCFC-142b) HCFC-141b and HCFC-142b emissions then grew
338	rapidly to new maxima of $68 + 8$ Gg/yr (2012) and $39 + 5$ Gg/yr (2008) respectively. These trends were
339	again reversed with subsequent declines to 59 ± 10 Gg/yr (HCFC-141b) and 25 ± 6 Gg/yr (HCFC-142b)
340	in 2015. Montzka et al. (2015), using an independent sampling network, also provided emissions
341	estimates for HCFCs-141b and -142b in 2012 which are included in Figures 3 b.c and agree within the
J f I	11





- 342 uncertainties of our estimates. Inventory emissions estimates reported from EDGAR v4.2 are
- 343 considerably lower than all other reported emissions post-2005. Global emissions of HCFC-141b and
- HCFC-142b have declined by 2.0% and 30%, respectively, from 2010 to 2015.
- 345

346 4.2.1.3. HCFC-124

Global emissions of this less abundant HCFC had a maximum in 2005 of 6.1 ± 1.8 Gg/yr, followed by a steady decline to 3.0 ± 0.9 Gg/yr in 2015, a decrease in emissions over this 12-year time frame of 51%. There are no bottom-up estimates of HCFC-124 emissions or top-down global estimates to compare with our results. However, we note that HCFC-124 emissions make only a minor contribution to climate change in CO₂ equivalent units (CO₂ e) of 0.002 Gt/yr in 2015.

The combined model derived aggregated global emissions of these four HCFCs in 2015 were 444 \pm 75 Gg/yr (0.75 \pm 0.1 Gt/yr CO₂ e). We estimate that these four HCFCs contribute 55 \pm 1 mW/m² to climate forcing in 2015 with HCFC-22 accounting for 79% of this forcing.











Figure 3 (a-d). Global HCFCs emissions (Gg/yr, blue line) from the 12-box model (solid blue line);
shading, representing 1 sigma uncertainties; emissions derived from UNEP* consumption data (- see
supplementary material 2). EDGAR v4.2 and other works are also shown. Note that HCFC-124 uses only
GC-MS-Medusa data for these calculations; all others use combined GC-MS-ADS and GC-MS-Medusa
data.

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469

470 Figure 4 (a-d). Global HFCs emissions (Gg/yr, blue line) from the 12-box model; shading represents 1

471 sigma uncertainties; UNFCCC* values are the global aggregate of national data reported in UNFCCC

472 (2016). Note that HFC-143a and HFC-32 use only Medusa data for these calculations; all others use

- 473 combined GC-MS-ADS and GC-MS-Medusa data.
- 474
- 475
- 476





477 4.2.2. Global estimates of HFC

478	In contrast to the HCFCs, estimates of HFCs emissions shown in Figure 4 (a-d) (see
479	supplementary material 1 for actual values) have increased annually over the entire observational record,
480	reaching maxima in 2015 of 60 \pm 10 Gg/yr (HFC-125), 209 \pm 43 Gg/yr (HFC-134a), 27 \pm 3.0 Gg/yr
481	(HFC-143a) and 32 ± 14 Gg/yr (HFC-32). UNFCCC emission estimates are consistently lower than our
482	estimates, even for HFC-32 which barely agrees within our uncertainties. EDGAR v4.2 inventory
483	emissions of HFC-143a post- 2000 are substantially larger than our estimates, but the other three HFCs
484	are in reasonable agreement (within the uncertainties of our estimates). Recently published HFC
485	emissions estimates by Velders et al., (2015) are in close agreement with the results from this work and
486	within the uncertainties of our estimates, except for HFC-143a after 2014.
487	The combined model derived aggregated emissions of these four HFCs in 2015 were
488	329 ± 70 Gg/yr (0.65 \pm 0.14 Gt/yr CO_2 e). We estimate that these four HFCs contribute 21.0 ± 0.5
489	mW/m^2 to radiative forcing in 2015, less than half the combined forcing of the 4 HCFCs treated in this
490	study.
491	
492	4.3. Overall emissions trends
493	
494	In Figure 5 we plot individual HCFC and HFC in terms of CO2-e emissions, noting that HCFC-22
495	is the largest contributor to these emissions and they have declined relatively slowly since 2010. HCFC-
496	141b and HCFC-142b exhibit declines in CO2-e emissions after 2010, in spite of substantial changes in
497	emissions over time and dramatic declines in their global mean mole fraction growth rates (see Figure 2).
498	All of the HFCs have increasing CO_2 -e emissions over time with HFC-125 and HFC-32 showing the
499	most rapid increases in CO ₂ -e emissions.







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Figure 5. Individual HCFC and HFC carbon dioxide equivalent (CO₂-e) emissions (Gg/yr) derived from
the 12-box model. Shading, representing 1 sigma uncertainties. Note that HCFC-124, HFC-143a and
HFC-32 use only GC-MS-Medusa data for these calculations; all others use combined GC-MS-ADS and
GC-MS-Medusa data.

Figure 6 shows the trends in CO2-e emissions when the HCFCs and HFCs are aggregated
together from 2005 to 2015. The figure indicates an increase in aggregate HCFC emissions until around
2010, and subsequent reduction in emissions through 2015. HFC emissions were observed to increase
throughout this period. Using 2010 as a "reference" year, we aimed to compare the relative trends in
HCFC and HFC emissions and determine broadly whether a decline in HCFC emissions was being
matched by acceleration in HFC usage. In order to determine a potential "business as usual" trajectory for
HCFC and HFC emissions post-2010, we first assumed that emissions would continue to follow the 2005











546	Figure 6. Aggregated HCFCs and HFCs emissions as CO ₂ -e (Gt yr ⁻¹) solid blue and green lines,
547	respectively; shading representing 1 sigma uncertainties. The grey line and shading represents the total
548	HCFC and HFC CO ₂ -e emissions. The dashed line is a linear projection from 2010 to 2015, based on the
549	2005 – 2010 emissions growth rate. The dashed lines with the circular data points are the Velders et al.
550	(2009) projection, rescaled to the 2010 emissions rate, and the dashed line with the diamonds show the
551	same but for HCFC emissions projections WMO (2010), Chapter 1. S.A. Montzka and S. Reimann.
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553	In Table 3 we compare the cumulative emissions over two 5-years periods, 2006-2010 when
554	HCFCs were still increasing (with the exception of HCFC-124) and 2011-2015 when all four HCFCs
555	show declining emissions. From the percentage change in emissions between the two periods we note
556	that HCFC-141b emissions have increased by 18% and HCFC-22 by 1.7%. In comparison HCFC-142b
557	and -124 emissions have decreased by 23% and 30%, respectively. Aggregating the four HCFCs we
558	observe a 1.4% increase in the combined emissions between the two 5-yr periods. However, over both 5-
559	yr periods there is an equal contribution of 3.9 Gt CO_2 -e.
560	Conversely, global emissions of the HFCs have grown continuously throughout the period of
561	observations with substantial increases between the two 5-yr periods. The largest increases were observed
562	for HFC-32 (143%) and HFC-125 (92%) with smaller increases for HFC-143a (37%) and HFC-134a
563	(28%). In terms of the aggregated HFC emissions we see a 43% increase representing a rise from 1.95 to
564	2.87 Gt CO ₂ -e, or an additional 0.92 Gt CO ₂ -e between the two periods.
565	It is also apparent that emissions of HCFC-22 represent 79.3% of the global cumulative HCFC
566	burden during 2011-2015 and HCFC-22 alone contributes 0.45 Gt CO2-e (13.5%) more than all HFCs
567	cumulative emissions during 2011-2015.
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- 579 Table 3. Global emissions of HCFCs and HFCs in 2015 and a comparison of the cumulative emissions
- and the percentage change in emissions over two 5-year periods (2006-2010 and 2011-2015).
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HCFCs	2015	2006-2010 HCFCs	2011-2015 HCFCs	% Change in
	Emissions	Cumulative Emissions	Cumulative Emissions	cumulative emissions
	(Gg)	(Gg) and CO ₂ - e	(Gg) and CO ₂ - e	between 2006-2010
		Emissions (Gt CO ₂ - e)	Emissions (Gt CO ₂ - e)	and 2011-2015
22	357.4 ± 58.1	1802.0 (3.26)	1833.2 (3.32)	1.7
141b	59.3 ± 9.8	272.5 (0.20)	322.4 (0.23)	18.3
142b	24.6 ± 6.3	180.8 (0.42)	138.9 (0.32)	-23.2
124	3.0 ± 0.89	25.4 (0.02)	17.8 (0.01)	-29.9
Total		2281 (3.9)	2312 (3.9)	1.4
HFCs	2015	2006-2010 HFCs	2011-2015 HFCs	% change in
	Emissions	Cumulative Emissions	Cumulative Emissions	cumulative emissions
	(Gg)	(Gg) and CO ₂ - e	(Gg) and CO ₂ - e	between 2006-2010
		Emissions (Gt CO ₂ - e)	Emissions (Gt CO ₂ - e)	and 2011-2015
125	60.3 ± 9.5	131.3 (0.46)	251.2 (0.88)	91.7
134a	209.4 ± 42.9	736.6 (1.05)	943.8 (1.35)	28.1
143a	27.3 ± 3.0	89.2 (0.4)	122.0 (0.55)	36.9
32	31.9 ± 14.4	51.6 (0.04)	125.2 (0.09)	142.7
Total		1009 (1.9)	1442 (2.9)	43

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Even though global production and consumption of HCFCs in Article 5 countries was capped in 584 585 2013 these developing countries have substantially increased their usage of HCFCs (Montzka et al., 2009) and are not required to phase-out potentially emissive consumption until 2040. As noted in Figure 586 587 2 the HCFCs show decreasing annual rates of growth with HCFC-142b declining by ~93% from 2007-588 2015, HCFC-141b by 53% and HCFC-22 by 54%. This is reflected in the aggregated HCFCs global 589 CO₂-e emissions (shown as in Figure 6) exhibiting a small but steady decline from 2010-2015. In spite of the 2013 cap on global production and consumption, the pace of decline has to some extent been 590 591 moderated by the increased emissions of HCFC-22 and HCFC-141b in article 5 countries. HCFC-22 592 remains the dominant refrigerant used in Article 5 countries and it has been estimated that approximately 1 million tonnes of HCFC-22 are currently in use in air conditioners operating worldwide (UNEP, 593 594 2016b). As the HFCs have replaced HCFCs, their aggregated emissions in 2015 have risen to a level in CO_2 -e that is 13% less than the aggregated emissions of the HCFCs. 595 596 597 Although there has been a shift in developed countries (non-Article 5) from HCFCs to high-GWP HFCs (Lunt at al., 2015; Montzka et al., 2015) the consumption of both HCFCs and HFCs in Article 5 598

countries has substantially increased, most notably in China (Fang et al., 2012; Zhang et al., 2014; Su et

al., 2015; Velders et al., 2015). There has been a trend in recent years to move to refrigerant blends with





601 lower GWPs and in Japan in 2014 residential air-conditioners were switched to using HFC-32 (UNEP,

602 2016b).

HFCs-134a, -125, -143a and -32 are the principal components of all alternative substitutes for the
HCFCs and in the supplementary material (3) we examine the composition of the main refrigerant blends
to determine if there is evidence for a significant use of single component refrigerants. In general, we find
that the atmospheric mole fractions of HFCs -32, -125, and -143a are consistent with their release
predominantly as blends.

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609 5. Conclusions

610 This study confirms that the Montreal Protocol and its amendments have been effective in slowing the atmospheric accumulation of the HCFCs. If there had been no change in the emissions 611 612 growth rate, we find that based on our linear projection there would have been an additional 0.67 ± 0.24 Gt CO₂-e of aggregated HCFCs -22, -141b, -142b emitted to the atmosphere from 2010-2015. This 613 compares with the forward projections of 'business as usual' in the Velders et al., 2009 and WMO (2010) 614 scenario of an additional 0.71 ± 0.25 Gt CO₂-e and 0.66 ± 0.24 Gt CO₂-e, respectively, of total HCFCs 615 616 that would have been emitted during this same period. However, we also calculate that the aggregated cumulative emissions of HCFCs-22, -141b, -142b, and -124 during the most recent 5-yr period (2011-617 2015) are slightly larger (1.4%) than in the previous five years (2006-2010). This increase has likely been 618 619 driven by the substantial emissions of HCFCs-22 and -141b in Article 5 countries. HCFC-22 represents 620 about 79% of total global HCFC emissions. As shown in Figure 2, the annual HCFC-22 growth rate has steadily declined since the introduction of the 2007 adjustment to the Montreal Protocol, yet global 621 emissions have tended to remain approximately constant with only a modest decline in emissions post-622 623 2007 (see Figure 5). We also note the linkage between HCFC-22 and HFC-23, also a potent GHG, which 624 is an unavoidable by-product of HCFC-22 production, (Miller et al., 2010; Rigby et al., 2014). Therefore further reductions in HCFC-22 production and consumption will benefit the efforts of the UNFCCC 625 Clean Development Mechanism (CDM) that mitigates HFC-23 emissions by voluntary incineration. 626 Since we are only two years beyond the 2013 cap on global production and consumption of HCFCs, it is 627 probably too early for our current observations through 2015 to show an accelerating phase-out for all 628 HCFCs, although HCFCs-142b and -124 have both recently undergone substantial declines in global 629 emissions. 630 Although global emissions of HFCs have increased throughout the course of this study, 631 Montzka et al., (2014) suggested that there may have been a shift to lower GWP refrigerant blends which 632

- 633 can account for the observed emissions. It is also noteworthy that the two HFCs with the largest
- percentage changes in emissions in recent years are HFCs-125 and -32 (see Table 3), which implies a
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trend towards blends containing these two refrigerants. However, we are unable to confirm the extent to which other lower GWP blends (e.g. R404A and R507A) have been substituted with the possibility that there has simply been a switch from one blend to another with similar GWPs. In terms of CO₂-e, surprisingly HCFC-22 emissions alone contribute 0.45 Gt CO₂-e, about 14% more to climate change than the four aggregated HFCs during 2011-2015. This could potentially be attributed to the continuing use of HCFC-22 in existing refrigeration equipment with the consequence of large slowly leaking banks.

641 With regard to inventory-based HFC emissions estimates it is important to acknowledge that attempting to quantify the release of individual HFCs to the atmosphere is complicated by the continuous 642 introduction of new blends many of which contain hydrocarbons and lower GWP refrigerants and their 643 substitution into older existing refrigeration equipment. This problem is further compounded by the lack 644 of information on the actual usage of the various blends in commercial and residential refrigeration, 645 646 coupled with the difficulty of quantifying emission magnitudes from the many banks and immediate release as solvents and in foam blowing applications. Nevertheless, the atmospheric mole fractions 647 observed are consistent with emissions of HFCs in refrigerant blends, rather than substantial emissions 648 from single component refrigerants. 649

650 We find that the increase in HFC emissions from 2010 to 2015 has been more rapid than a linear growth would imply. However, compared to this linear trend, the cumulative excess of emissions during 651 this period is 0.12 ± 0.15 Gt CO₂-e, which is smaller than the deficit in HCFCs during the same time 652 frame. This suggests that the phase-down in HCFCs post-2010 has not coincided with an equivalent 653 654 increase in HFC emissions in CO₂-e terms. Compared to alternative projections by Velders et al. (2009), our top-down estimates show a much more rapid decline in emissions of HCFCs, and slower increase in 655 656 HFC emissions. Therefore, similarly to our linear projection, where Velders et al. (2009) predicted a relatively steady growth in CO_2 -e emissions due to HCFCs and HFCs from 2005 - 2015, we find an 657 658 overall slowing of emissions post-2010.

Finally we note that national regulations to limit HFC use are already in place in the European
Union, Japan and the USA, and recently there has been an agreement to amend the Montreal Protocol to
further restrict HFC use beginning in 2019, (28th meeting of the parties to the Montreal Protocol, Kigali,
Rwanda, October 2016). By including HFCs, which have been shown to have a small, but non-zero
ozone-depletion potential (Hurwitz et al., 2015) into the Montreal Protocol, this has the benefit of further

- 664 regulating production and sales.
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- 683

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684 Data availability

- 686 The entire ALE/GAGE/AGAGE data base comprising every calibrated measurement
- 687 including pollution events is archived on the Carbon Dioxide Information and Analysis
- 688 Center (CDIAC) at the U.S. Department of Energy, Oak Ridge National Laboratory.

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