

# Atmospheric aerosol compositions over the South China Sea: Temporal variability and source apportionment

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**Abstract.** In order to evaluate impacts of different source emission on marine atmospheric particles over the South  
20 China Sea (SCS), major inorganic ionic concentrations ( $\text{Na}^+$ ,  $\text{Cl}^-$ ,  $\text{SO}_4^{2-}$ ,  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{K}^+$ ,  $\text{NH}_4^+$ ,  $\text{NO}_3^-$ ) were  
determined in total suspended particulates (TSP) at Yongxing Island, from March 2014 to February 2015. The  
annual average concentration of TSP was  $89.6 \pm 68.0 \mu\text{g}/\text{m}^3$ , with  $114.7 \pm 82.1$ ,  $60.4 \pm 27.0$ , and  $59.5 \pm 25.6 \mu\text{g}/\text{m}^3$   
in cool, warm, and transition seasons, respectively.  $\text{Cl}^-$  had the highest concentration, with an annual average of  
 $7.73 \pm 5.99 \mu\text{g}/\text{m}^3$ , followed by  $\text{SO}_4^{2-}$  ( $5.54 \pm 3.65 \mu\text{g}/\text{m}^3$ ),  $\text{Na}^+$  ( $4.00 \pm 1.88 \mu\text{g}/\text{m}^3$ ),  $\text{Ca}^{2+}$  ( $2.15 \pm 1.54 \mu\text{g}/\text{m}^3$ ),  $\text{NO}_3^-$   
25 ( $1.95 \pm 1.34 \mu\text{g}/\text{m}^3$ ),  $\text{Mg}^{2+}$  ( $0.44 \pm 0.33 \mu\text{g}/\text{m}^3$ ),  $\text{K}^+$  ( $0.33 \pm 0.22 \mu\text{g}/\text{m}^3$ ), and  $\text{NH}_4^+$  ( $0.07 \pm 0.07 \mu\text{g}/\text{m}^3$ ).  
Concentrations of TSP and the major ions showed seasonal variations which were higher in the cool season and  
lower in the warm and transition seasons. Factors of influence were wind speed, temperature, relatively humidity,  
rain, and air masses. Back trajectories, concentration weighted trajectories (CWT), and positive matrix factorization  
(PMF) of chemical compositions were analyzed for source apportionment, source contribution, and spatio-temporal

30 variation of major ions. Back trajectories and CWTs showed that air masses at Yongxing Island were mainly from  
the northeast, southwest, and southeast in the cool, warm, and transition seasons, respectively. The PMF results  
showed that  $\text{Na}^+$  of 77.4% and  $\text{Cl}^-$  of 99.3% were from sea salt. 60.5% of  $\text{NH}_4^+$  was of oceanic emission.  
Anthropogenic sources were very important for atmospheric aerosols over the island. Secondary inorganic aerosol  
of  $\text{SO}_2$  and  $\text{NO}_x$  from fossil fuel combustion (especially coal in Chinese coastal regions) was the dominant source  
35 of  $\text{NO}_3^-$  (69.5%) and  $\text{SO}_4^{2-}$  (57.5%).

**Keywords:** Source apportionment; crust; secondary inorganic aerosol; oceanic emission

## 1 Introduction

Aerosols or particulate matters potentially affect global atmospheric processes, chemistry, cloud formation,  
40 acid and nutrient deposition in sensitive ecosystems, affect human health as air pollution, and have effects on  
climate (Chin, 2009; Davidson, et al., 2005; Deng et al., 2010; Liu et al., 2014; Steyn and Chaumerliac, 2016;  
Zhang et al., 2011; Zhang et al., 2015). Aerosols have many sources. Primary aerosols, emitted directly from natural  
and anthropogenic sources, such as terrestrial dust from weathering, wind-blown sand and minerals, sea spray,  
combustion processes, and biological emissions (Chin, 2009; Davidson, et al., 2005; Rinaldi et al., 2010; Pavuluri  
45 et al., 2015). There are also secondary aerosols, which form from condensable atmospheric gases such as  $\text{SO}_2$ ,  $\text{NO}_x$ ,  
 $\text{NH}_3$ , and volatile organic compounds (Chin, 2009; Kolb and Worsnop, 2013; Xiao et al., 2012a, 2014; Zhang et  
al., 2011). Therefore, aerosols are composed of various chemical components (Xiao and Liu, 2004) and contain  
sulfate, nitrate, ammonium, mineral elements, black carbon and particulate organic matter (Zhang et al., 2007;  
Zhang et al., 2011; Chin, 2009; Pavuluri et al., 2015; Steyn and Chaumerliac, 2016). Because of rapid economic  
50 and industrial development in the last few decades in South and East Asia, many aerosols and their precursors  
released by human activities have become a major environmental problem (Steyn and Chaumerliac, 2016).  
Aerosols from anthropogenic emissions can be transported long distances from polluted regions to remote open  
oceans, which is well recognized as a major pathway for the supply of anthropogenic material to ocean surface  
waters (Duce et al., 2008; Lawrence and Lelieveld, 2010; Jung et al., 2012).

55 The South China Sea (SCS) is in the tropical–subtropical rim of the Northwest Pacific Ocean and is one of  
the largest marginal seas in the world (Fig. 1). The SCS is adjacent to several rapidly growing Asian economies,  
including China, the Philippines, Malaysia, Vietnam, and Indonesia (Kim et al., 2014). Consequently, the sea  
receives substantial amounts of aerosols from surrounding regions through long-range atmospheric transport (Fig.  
1; Atwood et al., 2013; Jung et al., 2012). The SCS has a monsoon climate, with a northeast monsoon in winter and

60 spring and a southwest monsoon in summer and autumn (Cui et al., 2016). Thus, aerosol optical thickness (AOT) shows spatial and seasonal variations, with higher AOT in the northern SCS during the cool season and higher AOT in the southern SCS in the warm season (Fig. 1). The SCS receives dust and pollutions from the northeast in winter and spring, e.g., from China and Japan (Liu et al., 2014; Wang et al., 2013; Xiao et al., 2015); during summer and autumn, the SCS receives particulate and gas phase pollutants from biomass burning in the southwest of SCS, e.g., from Malaysia and Indonesia (Atwood et al., 2013). However, the observational data on aerosol chemistry over the SCS are very sparse (Xiao et al., 2015).

To get better understanding of potential sources, source contributions, and spatio-temporal variations of marine aerosols over the SCS, total suspended particulate (TSP) were continuously collected at Yongxing Island from March 2014 to February 2015. The concentrations of major inorganic water-soluble ions were determined. Furthermore, back trajectories, concentration-weighted trajectory (CWT) and positive matrix factorization (PMF) analyses were also used to identify and apportion the main sources of aerosols and their chemical composition over the SCS.

## 2 Materials and Methods

### 2.1 Study site

Aerosol samples were collected from March 2014 through February 2015 on the rooftop of Xisha Deep Sea Marine Environment Observation and Research Station, South China Sea Institute of Oceanology, Chinese Academy of Sciences (SCSIO, CAS). This station is at the Yongxing Island (YXI, Fig. 1; 16.83°N, 112.33°E). This island has a tropical monsoon climate, with northeast monsoon in winter and spring, and southwest monsoon in summer and autumn (Xiao et al., 2015). The island is located in the high aerosol concentrations area of the northern SCS during winter and spring, and also at the periphery of the high concentrations area of the southern SCS during summer and autumn (Fig. 1). Therefore pollutants can arrive from Northeast Asia in winter and spring, and Southeast Asia in summer and autumn (Liu et al., 2014; Wang et al., 2013; Xiao et al., 2015; Atwood et al., 2013). It is also influenced by local SCS marine sources. More information about Yongxing Island is given elsewhere (Xiao et al., 2015 and 2016). In the present study, the sampling period was divided into three seasons based on wind direction and back trajectories. The cool season was March, April, October, November, December 2014 and January, February 2015, with principal air masses from the northeast and 25.9 °C of temperature, and the warm season was June–September 2014, with air masses primarily from the southwest and 29.7 °C of temperature. The transition season was May 2014, with local and short air masses primarily from the southeast. Yongxing Island

90 has an annual average temperature of  $27.7 \pm 2.7$  °C, relative humidity (RH) of  $80 \pm 7\%$ , and wind speeds of  $3.6 \pm 1.8$  m/s. There were strong seasonal variations between March 2014 and February 2015 (Fig. 2). Accumulated annual rainfall was 1526 mm during this period, with about 30% occurring in the cool season (Fig. 2).

## 2.2 Sample collection and chemical analyses

95 Aerosol was collected on quartz filters using a special high-flow rate ( $1.05 \pm 0.03$  m<sup>3</sup>/min) KC-1000 sampler, and aerosol water-soluble major anion concentrations ( $F^-$ ,  $Cl^-$ ,  $NO_3^-$ ,  $SO_4^{2-}$ ,  $Br^-$ ) were determined by ICS-90 ion chromatography (Dionex, California, USA) and water-soluble metal and nonmetal elemental concentrations (Al, Ca, Fe, K, Mg, Mn, Na, SiO<sub>2</sub>, Sr) were analyzed by an MPX inductively coupled plasma optical emission spectrometer (ICP-OES, Vista, CA, USA),  $NH_4^+$  concentration was determined by spectrophotometry after  
100 treatment with Nessler's reagent, as described in detail information in supplementary text S1.

## 2.3 Sea salt and non-sea salt contributions

In order to determine the source contributions to the ionic compositions in TSP, the sea salt ion (SS) and the non-sea salt ion (NSS), are calculated using Na as a reference element for marine origin by the following equations  
105 (Eq. 1 and 2; Xiao et al., 2013):

$$SS(\%) = 100(X / Na^+)_{sea} / (X / Na^+)_{TSP} \quad (1)$$

$$NSS(\%) = 100 - SS \quad (2)$$

where X represents the ionic composition and the data of  $(X/Na^+)_{sea}$  comes from Keene et al., (1986).

## 110 2.4 Back trajectories and concentration weighted trajectories analysis

Back trajectories and concentration weighted trajectories (CWT) are used to determine the long-distance transport of atmospheric pollutants and regional source areas by the program of TrajStat (Cheng et al., 2013; Wang et al., 2009; Xiao et al., 2014 and 2015). The detail methods were described in supplementary text S2.

## 115 2.5 Positive matrix factorization model

Receptor models are used to quantify the contributions of sources based on the composition or fingerprints of the sources (Norris et al., 2014). The positive matrix factorization (PMF) is an effective source apportionment receptor model that does not require source profiles prior to analysis and has no limitation on source numbers (Crippa et al., 2013; Tiwari et al., 2013; Zhang et al., 2011; Zhang et al., 2015). In our study, EPA PMF 5.0 was

120 used to determine source apportionment of each major ion. The detail methods were described in supplementary  
text S3.

### 3 Results and Discussion

#### 3.1 Aerosol chemical composition over the SCS and comparison with global marine aerosols

125 3.1.1 Aerosol characteristics over the SCS

Figs. 3, 4 and Table 1, provide information on atmospheric concentrations of TSP and major inorganic ions during the sampling period at Yongxing Island. The annual average TSP concentration at the island was  $89.6 \pm 68.0 \mu\text{g}/\text{m}^3$ , with a range of 16.4 to  $440.1 \mu\text{g}/\text{m}^3$ . The major inorganic ionic concentrations ( $\text{Na}^+$ ,  $\text{Cl}^-$ ,  $\text{SO}_4^{2-}$ ,  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{K}^+$ ,  $\text{NH}_4^+$ , and  $\text{NO}_3^-$ ) accounted for 24.8% of TSP. Total major inorganic ionic concentrations was  $22.21 \mu\text{g}/\text{m}^3$ .  $\text{Cl}^-$  had the highest concentration among these ions, from 0.39 to  $36.47 \mu\text{g}/\text{m}^3$ , with an annual average of  $7.73 \pm 5.99 \mu\text{g}/\text{m}^3$ . It was followed by  $\text{SO}_4^{2-}$  (range 0.52– $23.34 \mu\text{g}/\text{m}^3$ , average  $5.54 \pm 3.65 \mu\text{g}/\text{m}^3$ ),  $\text{Na}^+$  ( $0.90$ – $8.86 \mu\text{g}/\text{m}^3$ , average  $4.00 \pm 1.88 \mu\text{g}/\text{m}^3$ ),  $\text{Ca}^{2+}$  ( $0.17$ – $9.65 \mu\text{g}/\text{m}^3$ , average  $2.15 \pm 1.54 \mu\text{g}/\text{m}^3$ ),  $\text{NO}_3^-$  ( $0.10$ – $10.05 \mu\text{g}/\text{m}^3$ , average  $1.95 \pm 1.34 \mu\text{g}/\text{m}^3$ ),  $\text{Mg}^{2+}$  ( $0.02$ – $1.55 \mu\text{g}/\text{m}^3$ , average  $0.44 \pm 0.33 \mu\text{g}/\text{m}^3$ ),  $\text{K}^+$  ( $0.06$ – $1.13 \mu\text{g}/\text{m}^3$ , average  $0.33 \pm 0.22 \mu\text{g}/\text{m}^3$ ),  $\text{NH}_4^+$  ( $0.01$ – $0.32 \mu\text{g}/\text{m}^3$ , average  $0.07 \pm 0.07 \mu\text{g}/\text{m}^3$ ).

135 3.1.2 Aerosols over the SCS compared with global marine aerosols

The annual average TSP concentration at Yongxing Island is also lower than the annual average value over the northern Yellow Sea ( $123.2 \mu\text{g}/\text{m}^3$ ), another Chinese marginal sea (Wang et al., 2013). However, the mean TSP concentration at Yongxing Island is not lower than those at other remote islands or other seas, such as the Indian Ocean ( $21.1 \mu\text{g}/\text{m}^3$ ), Pacific Ocean ( $36.7 \mu\text{g}/\text{m}^3$ ), Mediterranean Sea ( $46.9 \mu\text{g}/\text{m}^3$ ), Southern Atlantic ( $39.1 \mu\text{g}/\text{m}^3$ ), and three islands of Okinawa ( $22.6 \mu\text{g}/\text{m}^3$  in summer and  $44.5 \mu\text{g}/\text{m}^3$  in spring) (Arakaki et al., 2014; Balasubramanian et al., 2013; Zhang et al., 2010).

The aforementioned annual average TSP and ionic concentration are comparable to those reported in many remote oceans (Fig. 5), e.g., Hedo, which is at the junction of the East China Sea and Northwest Pacific. The marine ions ( $\text{Na}^+$  and  $\text{Cl}^-$ ) accounted for 53% of total major ions (Fig. 4).  $\text{Na}^+$  and  $\text{Cl}^-$  concentrations at Yongxing Island were higher than most reported values among all locations and remote sites (Fig. 5), such as the Indian Ocean ( $3.0$  and  $4.4 \mu\text{g}/\text{m}^3$ , respectively), Arabian Sea ( $1.9$  and  $2.2 \mu\text{g}/\text{m}^3$ , respectively), Oki ( $3.3$  and  $4.3 \mu\text{g}/\text{m}^3$ , respectively) and Rishiri islands ( $1.2$  and  $2.7 \mu\text{g}/\text{m}^3$ , respectively) in the Sea of Japan, Amsterdam Island ( $1.4$  and  $2.3 \mu\text{g}/\text{m}^3$ , respectively) in the Southern Ocean, Bermuda ( $3.3$  and  $4.8 \mu\text{g}/\text{m}^3$ , respectively) in the Atlantic Ocean, and Hawaii

150 (0.5 and 0.4  $\mu\text{g}/\text{m}^3$ , respectively) in the Pacific (Kumar et al., 2008; Okuda et al., 2006; Claeys et al., 2010; Moody et al., 2014; Carrillo et al., 2002). However,  $\text{Na}^+$  and  $\text{Cl}^-$  concentrations were lower than samples from cruises, such as over the northern Atlantic-2 (7.0 and 9.6  $\mu\text{g}/\text{m}^3$ , respectively) and Pacific (5.8 and 9.0  $\mu\text{g}/\text{m}^3$ , respectively) (Zhang et al., 2010). The  $\text{Ca}^{2+}$  concentration was the highest among all locations and contributes 10% of major ions (Fig. 4), followed by the Mediterranean Sea, southern Atlantic and northern Atlantic-2 (Fig. 5). The relatively high  $\text{Ca}^{2+}$  concentration may be because of Asian terrestrial dust transported to Yongxing Island. Nss- $\text{Ca}^{2+}$  accounted for 93% of total  $\text{Ca}^{2+}$ , ranging from 0.14 to 9.31  $\mu\text{g}/\text{m}^3$  with an annual average of 1.99  $\mu\text{g}/\text{m}^3$ . Large contributions of nss- $\text{Ca}^{2+}$  were also found in the Mediterranean Sea, southern Atlantic and northern Atlantic-2, being at 88.4%, 90.3% and 90.0%, respectively (Zhang et al., 2010). The relatively high nss- $\text{Ca}^{2+}$  concentrations in those oceans were potentially from the crust or dust from some deserts (Zhang et al., 2010). Comparing Yongxing Island with among all locations, average  $\text{Mg}^{2+}$  concentrations were higher at Yongxing Island than most reported values among all locations (Fig. 5).  $\text{K}^+$  was also the highest among all locations (Fig. 5). Nss- $\text{K}^+$  ranged from 0 to 0.87  $\mu\text{g}/\text{m}^3$ , with an annual average of 0.18  $\mu\text{g}/\text{m}^3$  and a contribution of 55% to total water soluble  $\text{K}^+$  at Yongxing. In general,  $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$  and  $\text{NH}_4^+$  were major in the form of secondary inorganic aerosols. They accounted for only 34.0% of total inorganic ionic concentrations, giving them an intermediate position among all locations (Fig. 5). The average  $\text{SO}_4^{2-}$  concentration at Yongxing was the highest among all locations. As shown in Fig. 4, the mean contribution of  $\text{SO}_4^{2-}$  to major inorganic ionic components was  $\sim 25\%$  at Yongxing. The nss- $\text{SO}_4^{2-}$  concentration was 3.66  $\mu\text{g}/\text{m}^3$ , with a contribution of 66.1% to total  $\text{SO}_4^{2-}$ . Similar to  $\text{SO}_4^{2-}$ , the average concentration of  $\text{NO}_3^-$  in this study was the highest among all locations. It accounted for 9% of major ions at Yongxing Island. This indicates that a large number of anthropogenic sources affected the concentrations of  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$ . It was surprising that  $\text{NH}_4^+$  had relatively low concentrations over most oceans, except for the southern Atlantic and Mediterranean Sea (Fig. 5). The average  $\text{NH}_4^+$  concentration was  $0.07 \pm 0.07 \mu\text{g}/\text{m}^3$  in aerosol at Yongxing Island (Fig. 3), representing  $< 1\%$  of total major ions (Fig. 4). Further, low concentrations of  $\text{NH}_4^+$  were also observed in rainwater on the island (Xiao et al., 2016).

### 175 3.1.3 Global marine aerosol chemical patterns

Globally, sea salt ions ( $\text{Na}^+$  and  $\text{Cl}^-$ ) were the most important components in marine atmospheric aerosol, with higher concentration of  $\text{Cl}^-$  than  $\text{Na}^+$ , except over the Mediterranean and North seas (Fig. 5; Zhang et al., 2010; Ebert et al., 2000). In the marine atmosphere, sea salt aerosol ( $\text{NaCl}$ ) can react with sulfuric acid and nitric acid to release  $\text{HCl}$ , which results in a deficit of  $\text{Cl}^-$  relative to  $\text{Na}^+$  (Zhang et al., 2010). It is also found that a deficit of  $\text{Cl}^-$

180 in transition season at Yongxing Island (Fig. 4 and Table 2), was most likely because air masses were primarily from local oceanic areas far from the continent (Fig. 1), where wind is weak (Fig. 2). The mole equivalent ratios of  $\text{Cl}^-/\text{Na}^+$  (neq/L) in aerosols were slightly larger than seawater in annual, cool and warm seasons at Yongxing Island (Table 2). This suggests that  $\text{Cl}^-$  enrichment had an anthropogenic or other natural origin (Duan et al., 2006; Jung et al., 2012; Xiao et al., 2013). For  $\text{SO}_4^{2-}$  with  $\text{ss-SO}_4^{2-}$  and  $\text{nss-SO}_4^{2-}$ ,  $\text{nss-SO}_4^{2-}$  was greatly influenced by anthropogenic sources from developed industrial areas, leading to higher concentrations of  $\text{SO}_4^{2-}$  than  $\text{Na}^+$  and  $\text{Cl}^-$ . Examples were Bermuda, Ogasawara, and the Arabian Sea (Fig. 5; Moody et al., 2014; Kumar et al., 2008), where  $\text{nss-SO}_4^{2-}$  was the preferred species for acid displacement (Zhang et al., 2010). As another important ion of anthropogenic sources,  $\text{NO}_3^-$  concentrations were often good relationships with those of  $\text{nss-SO}_4^{2-}$  (Zhang et al., 2010), with relatively high concentrations among major ions (Fig. 5). Relatively high concentrations of  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  were also found over the SCS (Figs. 3 and 4).  $\text{NH}_4^+$  had the lowest concentrations among the major ions in most marine atmospheric aerosols, suggesting little ammonia transport to the open ocean, such as Yongxing Island. However, there were some exceptions. For example, the southern Atlantic and Mediterranean Sea had the highest  $\text{NH}_4^+$  concentrations among major ions (Fig. 5). Over most seas, the order was  $\text{Ca}^{2+} > \text{Mg}^{2+} > \text{K}^+$ . However, we found that  $\text{Mg}^{2+}$  had higher concentrations than  $\text{Ca}^{2+}$  in some remote ocean areas, such as in the Pacific, Atlantic and Southern oceans (Zhang et al., 2010). This indicates that  $\text{Ca}^{2+}$  of crustal origin was difficult to transport to the remote oceans, and  $\text{Mg}^{2+}$  may mainly be from sea salt over the open ocean (Moody et al., 2014).

### 3.2 Seasonal patterns of aerosol chemical species over SCS and adjacent areas

#### 3.2.1 Seasonal characteristics at Yongxing Island

200 As illustrated in Figs. 4 and 6, seasonal and monthly TSP concentrations and major inorganic water-soluble ion concentrations had distinct features at Yongxing Island. Generally, concentrations of TSP and major inorganic ions were higher in the cool season than in the warm season (Fig. 6). Seasonal variations were the same as those in other studies, such as Okinawa, and 18 urban, rural and remote sites in various regions of China (Arakaki et al., 2014; Wang et al., 2006; Xiao and Liu, 2004; Zhang et al., 2012).

205 Average TSP concentrations were  $114.7 \pm 82.1$ ,  $60.4 \pm 27.0$  and  $59.5 \pm 25.6 \mu\text{g}/\text{m}^3$  in the cool, warm and transition seasons, respectively, with the highest monthly average in November 2014 and the lowest in April ( $39.4 \mu\text{g}/\text{m}^3$ ) and September ( $39.9 \mu\text{g}/\text{m}^3$ ) of that year (Fig. 6). There were lower concentrations in the warm season than in the cool season, because 70% of rainfall at Yongxing Island happens during the warm season (Fig. 2), being the same as other studies, such as Shanghai and over the China Sea (Wang et al., 2006; Zhao et al., 2015). The positive



210 correlation between TSP concentration and wind speed ( $p < 0.01$ ) shown in Table 3 suggests that relatively high speeds can produce many particles from both sea spray and terrigenous matter. We discovered negative correlations between TSP concentration and temperature ( $p < 0.01$ ) and relatively humidity ( $p < 0.01$ ) (Table 2), indicating that warm temperatures and high relatively humidity enhance particle activation and scavenging is happening (Liu et al., 2011).

215 As shown in Figs. 4 and 6, sea salt ions  $\text{Na}^+$  and  $\text{Cl}^-$  were characterized by a gradual increase from the transition to cool season. Their concentrations (in  $\mu\text{g}/\text{m}^3$ ) in the cool, warm and transition seasons were  $4.91 \pm 1.82$  and  $3.04 \pm 1.08$ ,  $2.28 \pm 1.35$  and  $9.93 \pm 6.78$ , and  $5.25 \pm 2.63$  and  $3.73 \pm 3.63$ , respectively, with corresponding contributions of 52%, 57% and 57% to total major ions in those seasons. The highest  $\text{Na}^+$  and  $\text{Cl}^-$  concentrations in a single sample were found in November, with the lowest concentrations in May and April, respectively. The highest average monthly concentrations were in November. Positive relationships between  $\text{Na}^+$  or  $\text{Cl}^-$  and wind speed in 220 Table 3 ( $p < 0.01$ , correlation coefficient  $R = 0.44$  and  $p < 0.01$ ,  $R = 0.43$ , respectively) at Yongxing Island suggest that sea salt concentrations were dependent on wind speed. This is consistent with results at Chichijima Island (Boreddy and Kawamura, 2015). There was a low negative relationship between  $\text{Na}^+$  and rainfall ( $p < 0.05$ ,  $R = 0.27$ ) but no relationship between  $\text{Cl}^-$  and rainfall ( $p > 0.05$ ) in Table 3, suggesting that  $\text{Na}^+$  mainly existed in coarse particles and was readily removed by rainfall. As shown in Table 3, concentrations of  $\text{Na}^+$  and  $\text{Cl}^-$  were also 225 negatively influenced by temperature and relatively humidity. Although  $\text{Mg}^{2+}$  is often treated as crustal-derived ions and elements in continental studies (Zhang et al., 2015), its highest monthly average concentrations were in November at Yongxing Island, the same as  $\text{Na}^+$  and  $\text{Cl}^-$  (Fig. 5). As shown in Fig. 5, Tables 1 and 2, similar trends and strong correlation were observed among  $\text{Na}^+$ ,  $\text{Cl}^-$  and  $\text{Mg}^{2+}$ , and the ratios of  $\text{Mg}^{2+}$  to  $\text{Na}^+$  in aerosols were close 230 to that in seawater, suggesting that  $\text{Mg}^{2+}$  may mainly derive from sea salt rather than continental sources. However, there were no relationships between  $\text{Mg}^{2+}$  and wind speed, temperature, relatively humidity, or rainfall (Table 2), in contrast to other ions, such as  $\text{Na}^+$  and  $\text{Cl}^-$ . These results reveal that  $\text{Mg}^{2+}$  has different behaviors in the marine atmosphere at Yongxing Island. The different behaviors of  $\text{Mg}^{2+}$  was also found in rainwater at Yongxing Island (Xiao et al., 2016).

235 As shown in Fig. 6, the highest monthly average concentrations of  $\text{Ca}^{2+}$  were in February. Its monthly trends were different from those of TSP,  $\text{Na}^+$  and  $\text{Cl}^-$ , and the ratios of  $\text{Ca}^{2+}$  to  $\text{Na}^+$  in aerosols were much higher than those in seawater (Table 2), suggesting that  $\text{Ca}^{2+}$  from terrestrial dust sources may be influenced by different factors.  $\text{Ca}^{2+}$  accounted for 10%, 13% and 8% of total major ions in the cool, transition and warm seasons, respectively (Fig. 4). There was no correlation between  $\text{Ca}^{2+}$  and wind speed, in contrast to TSP,  $\text{Na}^+$  and  $\text{Cl}^-$  (Table 3). However,



240 there was a negative relationship between  $\text{Ca}^{2+}$  and rainfall ( $p < 0.05$ ; Table 3). These results suggest that  $\text{Ca}^{2+}$  existed in coarse particles that can be readily removed by rainfall. Thus, a low mass concentration was observed for  $\text{Ca}^{2+}$  in the rainy (warm) season (Fig. 6), with a low percentage being in the warm season in Fig.4.

245  $\text{K}^+$  concentrations were  $0.42 \pm 0.23 \mu\text{g}/\text{m}^3$  in the cool season,  $0.22 \pm 0.18 \mu\text{g}/\text{m}^3$  in warm season, and  $0.15 \pm 0.07 \mu\text{g}/\text{m}^3$  in the transition season at Yongxing, with the maximum monthly average concentrations in February and the minimum in July (Fig. 6). However, the lowest nss- $\text{K}^+$  monthly average concentration was in August. The results suggest that nss- $\text{K}^+$  is derived from Chinese biomass/biofuel burning in the cool season (Lawrence and Lelieveld, 2010). Streets et al. (2003) computed that China contributes 25% of total biomass burning in Asia. Sites in Chinese coastal regions had higher  $\text{K}^+$  and nss- $\text{K}^+$  concentrations than those at Yongxing Island (Wang et al. 2006) were observed, further indicating that Chinese and other Northeast Asian regions' biomass/biofuel burning have a strong influence on atmospheric composition over the SCS.

250 Similar to  $\text{K}^+$ , the highest monthly concentrations of  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  were observed in February, being at  $13.08 \pm 9.04$  and  $4.99 \pm 4.33 \mu\text{g}/\text{m}^3$ , respectively (Fig. 6). As shown in the figure,  $\text{SO}_4^{2-}$  concentrations in the cool and warm seasons were  $7.22 \pm 3.92$  and  $3.26 \pm 1.26 \mu\text{g}/\text{m}^3$ , respectively, accounting for 26% and 22% of total major ions, and  $\text{NO}_3^-$  concentrations were  $2.43 \pm 1.54$  and  $1.30 \pm 0.64 \mu\text{g}/\text{m}^3$ , accounting for 9% and 9%. The  $\text{NH}_4^+$  showed maxima in the cool season and minima in the warm season, being  $0.08 \pm 0.08$  and  $0.04 \pm 0.03 \mu\text{g}/\text{m}^3$ , respectively.

### 3.2.2 Seasonal patterns over SCS and adjacent areas

260 The spatial variability in seasonal patterns of the major inorganic ionic components at Yongxing Island and adjacent sites of the Acid Deposition Monitoring Network in East Asia (EANET) in 2011 is portrayed in Fig. 7. In general, total major inorganic ionic concentrations tended to be higher in cool seasons and lower in warm seasons to the north of Phnom Penh, including Phnom Penh, Hoa Binh, Hanoi, Hongwen, Hedo, Ogasawara, and Yongxing Island, consistent with previous studies (Boreddy and Kawamura, 2015; Wang et al., 2006; Xiao and Liu, 2004). There was no substantial seasonal variation at other sites of EANET, and there was no strong seasonal variation of rainfall there either. These results suggest that rainfall and wind patterns, anthropogenic activities influence the ionic seasonal variations (Lawrence and Lelieveld, 2010; Wang et al., 2006; Xiao et al., 2013; Xiao and Liu, 2004). Additionally, total major ionic concentrations were higher in the north than in the south, indicating more anthropogenic pollutants in the north, such as  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  (Lawrence and Lelieveld, 2010). As it is well known, the most densely populated regions in the north, including Hanoi, northeastern China, Pearl River Delta of China,

270 Korea, and Japan release large amounts of pollutants (Lawrence and Lelieveld, 2010), which then transport to the  
SCS in the cool seasons (Fig. 1).

The total ionic concentrations were higher at the three islands than sites to the south of Phnom Penh. As shown  
in Fig. 7, relatively high concentrations of  $\text{Na}^+$  and  $\text{Cl}^-$  were found at those islands, suggesting that ions from sea  
salt had large contributions to total major ions, i.e., 52.8%, 62.5% and 55.6% at Yongxing, Ogasawara and Hedo,  
275 respectively. This represents high mass concentrations of sea salt in the marine atmospheric aerosol. The highest  
concentrations of both  $\text{Na}^+$  and  $\text{Cl}^-$  appeared in November at Yongxing and Hedo islands, which were influenced  
by a strong northeast monsoon. The highest concentrations of both  $\text{Na}^+$  and  $\text{Cl}^-$  were in September at Ogasawara  
Island, which were influenced by a strong southeast monsoon from the Pacific. The relationship between  $\text{Na}^+$ ,  $\text{Cl}^-$   
and wind speed at Yongxing ( $p < 0.01$ ) is shown in Table 3. Other sites in Fig. 7 were also influenced by wind  
280 speed and winds directly from the ocean. However, the highest  $\text{Na}^+$  and  $\text{Cl}^-$  concentrations at some sites did not  
appear in the same month, e.g., at Hongwen, the highest concentrations of  $\text{Na}^+$  were in April and the highest of  $\text{Cl}^-$   
were in January. Excess  $\text{Cl}^-$  in January has been observed by anthropogenic sources in China (Duan et al., 2006).

The highest concentrations of  $\text{Mg}^{2+}$  were in the same months as  $\text{Na}^+$  at most sites, indicating that  $\text{Mg}^{2+}$  may  
be from sea salt with  $\text{Na}^+$  for almost all stations. The exceptions were at Hoa Binh and Tanah Rata with the  
285 maximum  $\text{Mg}^{2+}$  concentrations being in December and July, respectively. This suggests that  $\text{Mg}^{2+}$  originates from  
the crust rather than oceans (Xiao and Liu, 2004), or from both crust and ocean at these sites. Hoa Binh was  
influenced by the northeast monsoon, which carries strongly weathering crustal matters from China Yunnan-  
Guizhou Plateau karst (Hien et al., 2004; Xiao et al., 2013), and there was a strong relationship between  $\text{Mg}^{2+}$  and  
 $\text{Ca}^{2+}$  ( $R = 0.7$ ,  $p < 0.05$ ).  $\text{Ca}^{2+}$  had its highest concentrations in July at Phnom Penh, Tanah Rata, Petaling Jaya,  
290 Serpong, and Danum Valley, all of which are located in the south of the SCS. In these regions, relatively little  
rainfall (rainfall data from EANET) and strong sunlight were observed in that month, leading to strong weathering  
that generated  $\text{Ca}^{2+}$ . However, the highest  $\text{Ca}^{2+}$  concentrations were found at other sites in the cool season, during  
which there was much dust from Northeast Asia (Fig. S1; Boreddy and Kawamura, 2015; Liu et al., 2014; Wang  
et al., 2011). This result is consistent with earlier studies (Boreddy and Kawamura, 2015; Cheng et al., 2000; Liu  
295 et al., 2014; Zhao et al., 2015). The  $\text{Ca}^{2+}$  data also proved that Asian dust can affect the northern SCS, but it is  
difficult for Asian dust to be transported to the southern SCS (Figs. 1).

Figure 8 shows fire spot data from MODIS global fire mapping around the SCS during March 2014 through  
February 2015. Additionally, smoke surface concentrations every day in that period and region is shown in Fig. S1.  
The fire spot and smoke data give information on seasonal variations of biomass burning around the SCS. This

300 activity was strong from January to April in the west of the SCS, including Vietnam, Thailand and Laos, and  
between July and October in the south of the SCS, including Malaysia and Indonesia (Figs. 8 and S1). These data  
are consistent with other studies showing substantial monthly CO emissions from biomass burning during  
February–April and August–October in Southeast Asia, and February–May in southern China and Taiwan (Streets  
et al., 2003).  $K^+$  is commonly used as a tracer of biomass and biofuel burning (Deng et al., 2010). As shown in Fig.  
305 7, we found that the maximum  $K^+$  was in the aforementioned months at most sites, suggesting that Asian biomass  
burning influenced the SCS region.

Fossil fuel combustion, industrial processes, biofuel burning, agricultural and waste handling often generate  
large quantities of  $SO_2$ ,  $NO_x$ ,  $NH_3$  in Asia (Lawrence and Lelieveld, 2010; Liu et al., 2013; Xiao et al., 2012a; Xiao  
et al., 2014; Xiao et al., 2015), although natural emissions of  $SO_2$  and  $NH_3$  include from biomass burning, marine  
310 and soil biological processes (Streets et al., 2003; Altieri et al., 2014; Boreddy and Kawamura, 2015; Xiao et al.,  
2012a), and  $NO_x$  from those processes and lightning (Price et al., 1997; Xiao et al., 2015). In general, the three  
marine sites (Yongxing, Hedo, and Ogasawara islands) had smaller proportions of  $SO_4^{2-}$ ,  $NO_3^-$  and  $NH_4^+$  than inland  
sites, with the three ions accounting for ~35% at the three marine sites and up to 65% at the other sites. This  
indicates that anthropogenic contributions are smaller over remote open oceans than at continental sites. Figure 6  
315 shows that the highest  $SO_4^{2-}$ ,  $NO_3^-$  and  $NH_4^+$  concentrations were found during the cool season in the north of  
Phnom Penh, including Phnom Penh, Hoa Binh, Hanoi, Hongwen, Hedo, Ogasawara, and Yongxing, consistent  
with total inorganic major ions. This indicates that the pollutants from Northeast Asia have a great impact on the  
Northwest Pacific. **Figure S1 confirms these findings that pollutants from nature and anthropogenic activities effect  
the Northwest Pacific.** We also found that most sites in the south of Phnom Penh had maximum  $SO_4^{2-}$  and  $NO_3^-$   
320 concentrations in the same months as the highest  $K^+$  concentrations occurred, suggesting that biomass and biofuel  
burning are important sources for  $SO_4^{2-}$  and  $NO_3^-$  in those regions. Lawrence and Lelieveld (2010) found that such  
burning was important in the emissions of  $SO_4^{2-}$  and  $NO_3^-$  in southern Asia, whereas fossil fuel combustion and  
industrial processes tended to be dominant in northern Asia (Xiao et al., 2015). However, maximum  $NH_4^+$   
concentrations at some sites (e.g., Petaling Jaya, Serpong, Danum Valley) were inconsistent with  $SO_4^{2-}$  and  $NO_3^-$ .  
325 Moreover, there was no relationship between  $SO_4^{2-}$  or  $NO_3^-$  and  $NH_4^+$  at these sites in the south of Phnom Penh,  
including Phnom Penh and Yongxing Island (both  $p > 0.05$ ). The results are inconsistent with previous studies  
(Boreddy and Kawamura, 2015; Hsu et al., 2007; Wang et al., 2006; Xiao et al., 2013; Xiao and Liu, 2004). In the  
marine atmosphere,  $SO_4^{2-}$  and  $NO_3^-$  are predominant in coarse particles (Boreddy and Kawamura, 2015; Xiao et al.,

2015); while  $\text{NH}_4^+$  is often predominant in fine particles and may exist in the form of  $(\text{NH}_4)_2\text{SO}_4$  in their  
330 **accumulation mode** (Ooki et al., 2007; Ottley and Harrison, 1992).

### **3.3 Source identification, apportionment, and region**

Based on the PMF 5.0 model, five potential sources of atmospheric chemical components at Yongxing Island were identified: sea salt (two species), crust, SIA, and oceanic emission. Table 4 summarizes source apportionment  
335 of the relative contributions of each identified source to major ions. Figs. 9 and 10 show the modeled source profiles and the time series of modeled concentrations for each identified main source.

Further, CWTs were plotted for TSP and major ions (Fig. 11) to explore likely regional sources and transport pathways for the island. Air masses at Yongxing Island had obvious unique and seasonal variations, from northeast of the island in the cool season, southwest in the warm season, and southeast in the transition season (Fig. 11). This  
340 reveals that aerosol or chemical compositions at the island originated from different regions in different seasons (Figs. 1, 8, 11 and S1). The air masses with high TSP concentrations were from China coastal regions bordering the Yellow and East China Seas and northern South China (Fig. 11). This is consistent with the seasonal variations of TSP concentrations in Fig. 6. The average aerosol optical thickness (AOT) over the Northwest Pacific (Fig. 1) confirmed this result. A relatively large average AOT was found over the northern SCS and East China Sea in the  
345 cool season, and Karimata Strait in the warm season (Fig. 1). But there was a relatively low average AOT over the entire SCS in the transition season (Fig. 1).

The first source, sea salt, generally has strong marine elements, such as  $\text{Na}^+(\text{Cl}^-)$  and  $\text{Mg}^{2+}$ , which  $\text{Na}^+$  exists in super-micron size aerosols, whereas  $\text{Mg}^{2+}$  in sub-micron size aerosols (Moody et al., 2014). They contributed 77.4% ( $\text{Na}^+$ ), 93.9% ( $\text{Cl}^-$ ) and 70.4% ( $\text{Mg}^{2+}$ ) from sea salt at Yongxing Island (Table 4). Although the CWT for  
350  $\text{Mg}^{2+}$  was larger in the cool season and lower in the warm season, air masses with relatively high concentrations of  $\text{Mg}^{2+}$  originated offshore of China (Fig. 10). This further indicates that  $\text{Mg}^{2+}$  was mainly from sea salt. Other significant sources were crust, SIA and oceanic emission, with contributions  $< 10\%$  for  $\text{Na}^+$  and  $\text{Cl}^-$ , and  $< 16\%$  for  $\text{Mg}^{2+}$ . According to data of rainwater at Yongxing Island, a part of  $\text{Na}^+$  and  $\text{Cl}^-$  could be from crust and produced by burning (Xiao et al., 2016). Coal combustion and biomass burning also produce  $\text{Na}^+$  and  $\text{Cl}^-$  (Liu et al., 2000;  
355 Tiwari et al., 2013; Zhang et al., 2015). Zhang et al. (2015) found that coal combustion was the most likely dominant source of  $\text{Cl}^-$  in Beijing. The mole equivalent  $\text{Cl}^-/\text{Na}^+$  ratios were larger in the cool season than in the transition and warm seasons at Yongxing, indicating that the crust, fossil combustion, and biofuel and biomass burning affected  $\text{Na}^+$  and  $\text{Cl}^-$  concentrations over the Northwest Pacific (Figs. 8 and S1). As shown in Table 3, there were strong

relationships between  $\text{Na}^+$  ( $\text{Cl}^-$ ) and  $\text{SO}_4^{2-}$  ( $\text{Ca}^{2+}$ ,  $\text{K}^+$ ), further proving that crust, fossil combustion and biomass  
360 burning can generate  $\text{Na}^+$  and  $\text{Cl}^-$ . Moreover,  $\text{NaCl}$  can react with acids such as  $\text{H}_2\text{SO}_4$ ,  $\text{HNO}_3$  and  $\text{H}_2\text{C}_2\text{O}_4$ , altering  
 $\text{Na}^+$  and  $\text{Cl}^-$  concentrations in the marine atmosphere and producing secondary chlorine-containing salt (Boreddy  
and Kawamura, 2015). Sea salt provided  $\text{K}^+$ ,  $\text{Ca}^{2+}$  and  $\text{SO}_4^{2-}$ , constituting 53.2%, 33.9% and 24.0%, respectively.  
The ratios of  $\text{K}^+/\text{Na}^+$ ,  $\text{Ca}^{2+}/\text{Na}^+$  and  $\text{SO}_4^{2-}/\text{Na}^+$  in Table 2 also indicate that part of them are from sea salt. The  
results are consistent with other studies (Boreddy and Kawamura, 2015).

365 The second source, crust, has a substantial crustal elements  $\text{Ca}^{2+}$  and  $\text{Mn}$  (Fig. 9), which is a tracer of crust  
(Suzuki and Tsunogai, 1988; Xiao et al., 2013; Xiao and Liu, 2004; Norris et al., 2014). However,  $\text{Ca}^{2+}$  from the  
crust, only had a contribution of 9.8% (Table 4). The result indicates that it is difficult for  $\text{Ca}^{2+}$  directly derived  
from crust to transport to open ocean. But CWTs for  $\text{Ca}^{2+}$  were larger in the cool season and lower in the warm  
season, indicating that dust from Northeast Asia influenced aerosol chemistry in the remote marine areas (Figs. 1,  
370 11 and S1). The result suggests that when it reacts with  $\text{H}_2\text{SO}_4$  and  $\text{HNO}_3$  to generate secondary inorganic aerosol,  
e.g.  $\text{CaSO}_4$  and  $\text{Ca}(\text{NO}_3)_2$ , it may transport to a longer distance.

The third source is relevant to secondary inorganic aerosol, which are typically characterized by remarkable  
 $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$ . They contributed 57.5% and 69.5% from SIA at Yongxing Island (Table 4) and good relationship  
between them was observed (Table 2). Fossil fuel (especially coal) combustion releases large amounts of  $\text{SO}_2$  and  
375  $\text{NO}_x$  (Xiao et al., 2012b, 2014 and 2015). Lawrence and Lelieveld (2010) attributed 61% of total  $\text{NO}_x$  and 77% of  
total  $\text{SO}_2$  emissions from fossil fuel combustion in southern Asia, 76% of total  $\text{NO}_x$  and 75% of total  $\text{SO}_2$  in northern  
Asia, resulting in the transport of substantial secondary inorganic aerosols containing  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  to the SCS  
(Figs. 1, 11 and S1). As with TSP and  $\text{Ca}^{2+}$ , air masses from China coastal regions had high  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$   
concentrations in the cool season (Fig. 11), owing to rapid economic development and great coal demand in the  
380 country, especially in its coastal regions (Lawrence and Lelieveld, 2010). Figure S1 also shows that  $\text{SO}_4^{2-}$  from  
central and eastern China reached coastal regions in the cool season. In Chinese coastal provinces, emission  
intensities of  $\text{SO}_2$  and  $\text{NO}_x$  were about 10 and 15 tons/ $\text{km}^2$ , respectively, much higher than other Chinese provinces  
(China Environment Statistical Yearbook, 2014). Rapidly growing economies and high population densities in  
these regions (Kim et al., 2014) release pollutants that are transported to the Northwest Pacific. In the atmosphere,  
385 the products ( $\text{H}_2\text{SO}_4$  and  $\text{HNO}_3$ ) of  $\text{SO}_2$  and  $\text{NO}_x$  can easily combine alkaline ions, such as  $\text{Ca}^{2+}$  and  $\text{K}^+$  (Xiao et  
al., 2013), which  $\text{K}^+$  is characterized as an effective tracer of biomass and biofuel burning aerosols (Zhang et al.,  
2015). As shown in Table 4, SIA had much bigger contributions of 56.3% to  $\text{Ca}^{2+}$  and 27.7% to  $\text{K}^+$ . We found that  
there were strong relationships between  $\text{SO}_4^{2-}$  ( $\text{NO}_3^-$ ) and  $\text{Ca}^{2+}$  ( $\text{K}^+$ ) (Table 3). These suggest that  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$

changed source attribution of some alkaline ions to transport to open ocean. In this case, it is difficult to distinguish  
390 the  $\text{Ca}^{2+}$  and  $\text{K}^+$  sources using PMF model.

The fourth source is oceanic emission, which released  $\text{NO}_x$ ,  $\text{NH}_3$  and DMS (Altieri et al., 2014; Jickells et al.,  
2003; Boreddy and Kawamura, 2015; Phinney et al., 2006), with respective contributions of 1.6% to  $\text{NO}_3^-$ , 60.5%  
to  $\text{NH}_4^+$ , and 6.9% to  $\text{SO}_4^{2-}$  at Yongxing Island (Table 4). As shown in Fig. 11, air masses with high  $\text{NH}_4^+$   
395 concentrations were from remote open oceans such as the southeastern and northeastern SCS. There were relatively  
higher  $\text{NH}_4^+$  concentrations in the cool season and lower values in the warm season at Yongxing Island (Fig. 6).  
These suggest that it is feasible for the ocean to be a  $\text{NH}_4^+$  source. Substantial  $\text{NH}_x$  may be released from degraded  
organic nitrogen-containing compounds and excretion from zooplankton in the ocean (Norman and Leck, 2005).  
Altieri et al. (2014) suggested that the efficient kinetics of ammonia evasion from surface seawater causes  $\text{NH}_3$  to  
accumulate in the marine atmosphere. The contribution of oceanic emission to  $\text{NH}_4^+$  was much larger at Yongxing  
400 than at global marine atmospheric  $\text{NH}_x$  sources in the review of Duce et al. (2008), which showed 87.5% from  
anthropogenic sources. However, Altieri et al. (2014) found that the anthropogenic contribution was  $< 87.5\%$  at  
Bermuda, an island in the North Atlantic Ocean (Fig. 5). Atmospheric  $\text{NH}_x$  is usually rapidly deposited near source  
regions and has a short residence time, about several hours in the marine boundary layer (Boreddy and Kawamura,  
2015; Xiao et al., 2012a; Xiao and Liu, 2002). Thus,  $\text{NH}_x$  transportation from continental to remote sea sites is  
405 limited. Therefore,  $\text{NH}_4^+$  in aerosol at Yongxing Island was possibly from oceanic emission, as being reported at  
other marine sites (Altieri et al., 2014; Jickells et al., 2003). DMS is the most abundant marine biogenic volatile  
sulfur emitted from the ocean surface to atmosphere, and can be oxidized to  $\text{SO}_4^{2-}$  in the marine atmosphere  
(Phinney et al., 2006). Yang et al. (2015) reported that biogenic  $\text{SO}_4^{2-}$  from the Bohai and northern Yellow seas  
near China was 0.114–0.551  $\mu\text{g}/\text{m}^3$ , with an average of 0.247  $\mu\text{g}/\text{m}^3$ , accounting for 1.4% of nss- $\text{SO}_4^{2-}$ . Biogenic  
410  $\text{SO}_4^{2-}$  in the northern SCS was  $\sim 1.2$  and  $0.6 \mu\text{g}/\text{m}^3$  in summer and winter, respectively (Zhang et al., 2007),  
constituting  $\sim 8\%$  and  $12\%$  of nss- $\text{SO}_4^{2-}$ . CWTs for  $\text{SO}_4^{2-}$  in Fig. 11, also show some  $\text{SO}_4^{2-}$  were from marine  
biogenic source. Thus, natural sources had large contributions to marine atmospheric aerosols over the Northwest  
Pacific.

Figure 10 illustrates the time series of daily concentrations contributed by each identified source. In order to  
415 examine if the results are reasonable, we compared the modeled results of each source with the observed seasonal  
variations of the specific chemical species (Figs. 6 and 11). As shown in Fig. 9, the highest and lowest contributions  
of sea salt were in November-December and April-May, respectively, which are consistent with the seasonal  
variations of aerosol  $\text{Na}^+$ ,  $\text{Cl}^-$  and  $\text{Mg}^{2+}$ . For crust, the resulting time series show that it has a higher concentration

in March, 2014 and February, 2015, relatively close to the observed values of  $\text{Ca}^{2+}$  and  $\text{K}^+$  (Figs. 6 and 11). SIA  
420 has higher contributions from November, 2014 to February, 2015 and lower other months, which also consistent  
with the observed and CWTs data of  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  shown in Figs. 6 and 11. This result is definitely related to the  
photochemistry that accounts for SIA formation (Zhang et al., 2013). The formed SIA species may not appear in  
their original emission sources (Zhang et al., 2013), such as coal combustion, biomass burning and crust. It is  
obvious that oceanic emission has highest contribution in March, and secondary highest in October, consistent with  
425 the seasonal change of  $\text{NH}_4^+$ . The PMF and CWTs modeled results seem to be promising because the corresponding  
time series of each source's contribution are very consistent with the observations.

#### 4 Conclusions

Chemical compositions of 1-year aerosols at Yongxing Island were investigated to help better understand their  
430 chemical characteristics, sources, and transport pathways over the SCS. Sea salt ( $\text{Na}^+$  and  $\text{Cl}^-$ ) had the greatest  
contribution to total major inorganic ions in aerosols at the island, followed by  $\text{SO}_4^{2-}$ ,  $\text{Ca}^{2+}$ ,  $\text{NO}_3^-$ ,  $\text{Mg}^{2+}$ ,  $\text{K}^+$  and  
 $\text{NH}_4^+$ . The concentrations of TSP and all major inorganic ions showed seasonal variations, with higher  
concentrations in the cool season and lower in the warm season, which was influenced by meteorological  
parameters (e.g., wind speed, temperature, relative humidity and rainfall) and air masses. Using PMF and CWT  
435 models, fire spot and AOT, we found that  $\text{Na}^+$ ,  $\text{Cl}^-$ , and  $\text{Mg}^{2+}$  were mainly derived from sea salt,  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$  of  
secondary inorganic aerosol mainly came from fossil fuel combustion (especially coal combustion in the Northern  
Asia).  $\text{NH}_4^+$  was mainly from oceanic emission in the remote ocean. In summary, fossil fuel combustion seriously  
affected marine atmospheric aerosol chemical compositions over the Northwest Pacific.

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640 **Table Captions**

**Table 1** Annual average, minimum and maximum mass concentrations ( $\mu\text{g}/\text{m}^3$ ) of TSP and aerosol chemical species at Yongxing Island.

**Table 2** Mole equivalent ratios for major ionic species in aerosols at Yongxing Island (annual, cool, transition and warm seasons), together with seawater ratios for comparison.

645 **Table 3** Correlation coefficients among major ions in aerosol and meteorological parameters.

**Table 4** Relative contributions (%) for different major ions from potential five sources of TSP at Yongxing Island over the year, based on PMF 5.0 model.

## Figure Captions

650 **Figure 1** Distribution of seasonal average aerosol optical thickness (AOT) at 550 nm (T550) over Northwest Pacific in cool, warm and transition seasons during sampling period. Monthly AOT products (from Moderate Resolution Imaging Spectrometer, MODIS) with 4-km resolution were downloaded from Globcolour (<http://hermes.acri.fr/>). The GlobColour project began in 2005 as an ESA Data User Element project to provide a continuous dataset of merged Level 3 Ocean Colour products.

655 **Figure 2** Three-hour temperature, relative humidity, wind speed and precipitation at Yongxing Island during sampling period (March 2014 through February 2015).

**Figure 3** Annual average mass concentrations of aerosol chemical species at Yongxing Island. Box boundary indicates 25<sup>th</sup> and 75<sup>th</sup> percentile. Lines within the box show the mean. Whiskers above and below the box indicate 90<sup>th</sup> and 10<sup>th</sup> percentiles.

660 **Figure 4** Mean contributions of each major ionic component to total ionic mass concentration of (a) Yongxing Island (YXI) annual, (b) YXI cool season, (c) YXI warm season, and (d) YXI transition season.

**Figure 5** Comparisons of major ions in aerosol at Yongxing Island with global ocean. Data of Oki, Ogasawara and Hedo are from EANET ([www.eanet.asia](http://www.eanet.asia)); those of Rishiri Island are from Okuda et al. (2006); those of Hawaii are from Carrillo et al. (2002); those of Bermuda are from Moody et al. (2014); those of Amsterdam Island are from Claeys et al. (2010); those of the Arabian Sea and Indian Ocean are from Kumar et al. (2008); those of Helgoland are from Ebert et al. (2000); those of the Mediterranean Sea, northern Atlantic-1 and 2, Pacific and southern Atlantic are from Zhang et al. (2010). Pentagrams represent sampling sites on islands; others represent cruises. N.A. indicates no data.

670 **Figure 6** Seasonal variations of TSP mass concentration and associated species, including Na<sup>+</sup>, Cl<sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, Mg<sup>2+</sup>, K<sup>+</sup>, Ca<sup>2+</sup>, NH<sub>4</sub><sup>+</sup>, and NO<sub>3</sub><sup>-</sup> at Yongxing Island (cool season: C; warm season: W; annual: A). Shown are the mean and standard deviation for each bar.

**Figure 7** Comparison of aerosol chemical species between Yongxing Island and around the South China Sea (data from EANET).

675 **Figure 8** Fire spot data from MODIS global fire mapping from March 2014 to February 2015 around South China Sea (<https://firms.modaps.eosdis.nasa.gov/firemap/>).

**Figure 9** Profiles of five sources identified from the PMF 5.0 model, including sea salt (two species), crust, secondary inorganic aerosol and oceanic emission.



**Figure 10** Time series contributions from each identified sources, including sea salt (two species), crust, secondary inorganic aerosol, and oceanic emission.

680 **Figure 11** Ten-day back trajectories of warm (black, June through September 2014), cool (blue, March and April 2014 and October 2014 through February 2015) seasons, and transition season (red, May 2014) at Yongxing Island. Additionally, CWT (concentration weighted trajectory) plots for daily weighted-average concentrations of TSP,  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{K}^+$ ,  $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$  and  $\text{NH}_4^+$  at Yongxing Island.

685 **Supporting Information**

**Figure S1.** Modeled NAAPS total aerosol optical depth (AOD) for every month of March 2014 through February 2015, for total, sulfate, dust and smoke (data from <http://www.nrlmry.navy.mil/aerosol/#aerosolobservations>).

**Table 1** Annual average, minimum and maximum mass concentrations ( $\mu\text{g}/\text{m}^3$ ) of TSP and aerosol chemical species at Yongxing Island.

	TSP	Na <sup>+</sup>	Cl <sup>-</sup>	K <sup>+</sup>	Ca <sup>2+</sup>	Mg <sup>2+</sup>	SO <sub>4</sub> <sup>2-</sup>	NO <sub>3</sub> <sup>-</sup>	NH <sub>4</sub> <sup>+</sup>
Annual	89.6 ± 68.0	4.00 ± 1.88	7.73 ± 5.99	0.33 ± 0.22	2.15 ± 1.54	0.44 ± 0.33	5.54 ± 3.65	1.95 ± 1.34	0.07 ± 0.07
Minimum	16.4	0.90	0.39	0.06	0.17	0.02	0.52	0.10	0.01
Maximum	440.1	8.86	36.47	1.13	9.65	1.55	23.34	10.05	0.32

**Table 2** Mole equivalent ratios for major ionic species in aerosols at Yongxing Island (annual, cool, transition and warm seasons), together with seawater ratios for comparison.

	Yongxing Island				Seawater <sup>a</sup>
	annual	cool	transition	warm	
Cl <sup>-</sup> /Na <sup>+</sup>	1.25	1.31	1.06	1.12	1.17
Mg <sup>2+</sup> /Na <sup>+</sup>	0.21	0.21	0.19	0.23	0.22
K <sup>+</sup> /Na <sup>+</sup>	0.048	0.051	0.040	0.042	0.021
Ca <sup>2+</sup> /Na <sup>+</sup>	0.62	0.64	0.83	0.47	0.044
SO <sub>4</sub> <sup>2-</sup> /Na <sup>+</sup>	0.66	0.71	0.73	0.51	0.12
nss-SO <sub>4</sub> <sup>2-</sup> /Na <sup>+</sup>	0.54	0.58	0.61	0.39	-
NO <sub>3</sub> <sup>-</sup> /Na <sup>+</sup>	0.18	0.18	0.22	0.16	-
NH <sub>4</sub> <sup>+</sup> /Na <sup>+</sup>	0.022	0.021	0.044	0.016	-
NO <sub>3</sub> <sup>-</sup> /nss-SO <sub>4</sub> <sup>2-</sup>	0.34	0.32	0.36	0.41	-
NH <sub>4</sub> <sup>+</sup> /nss-Ca <sup>2+</sup>	0.038	0.035	0.056	0.038	-

<sup>a</sup>Seawater ratios from Keene et al. (1986).

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696 **Table 3** Correlation coefficients among major ions in aerosol and meteorological parameters.

	TSP	Na <sup>+</sup>	Cl <sup>-</sup>	SO <sub>4</sub> <sup>2-</sup>	Ca <sup>2+</sup>	Mg <sup>2+</sup>	K <sup>+</sup>	NH <sub>4</sub> <sup>+</sup>	NO <sub>3</sub> <sup>-</sup>	WS	T	RH	R
TSP	1	0.77**	0.92**	0.77**	0.92**	0.32**	0.75**	-0.05	0.52**	0.36**	-0.47**	-0.44**	-0.20
Na <sup>+</sup>		1	0.91**	0.69**	0.72**	0.57**	0.78**	-0.03	0.48**	0.44**	-0.51**	-0.46**	-0.27*
Cl <sup>-</sup>			1	0.71**	0.83**	0.49**	0.77**	-0.04	0.45**	0.43**	-0.37**	-0.36**	-0.19
SO <sub>4</sub> <sup>2-</sup>				1	0.86**	0.56**	0.85**	0.26*	0.87**	0.04	-0.56**	-0.58**	-0.29*
Ca <sup>2+</sup>					1	0.36**	0.81**	-0.03	0.69**	0.24	-0.51**	-0.53**	-0.27*
Mg <sup>2+</sup>						1	0.63**	0.45**	0.59**	0.04	-0.12	-0.08	-0.18
K <sup>+</sup>							1	-0.18	0.72**	0.15	-0.51**	-0.45**	-0.27*
NH <sub>4</sub> <sup>+</sup>								1	0.36**	-0.13	-0.18	0.11	-0.18
NO <sub>3</sub> <sup>-</sup>									1	-0.05	-0.50**	-0.50**	-0.31**

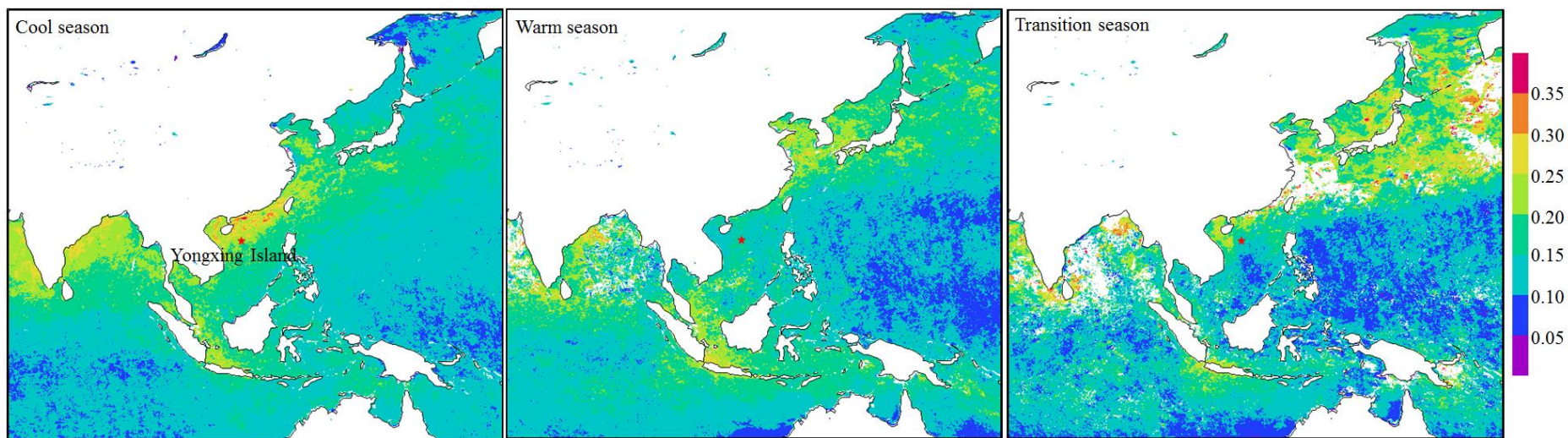
697 \*\*Correlation significant at 0.01 level (2-tailed), \* significant at 0.05 level (2-tailed). WS: wind speed (m/s); T: temperature (°C); RH: relative humidity (%); R: rainfall (mm/h)

698 **Table 4** Relative contributions (%) for different major ions from potential five sources of TSP  
 699 at Yongxing Island over the year, based on PMF 5.0 model.

Source	Na <sup>+</sup>	Cl <sup>-</sup>	K <sup>+</sup>	Ca <sup>2+</sup>	Mg <sup>2+</sup>	SO <sub>4</sub> <sup>2-</sup>	NO <sub>3</sub> <sup>-</sup>	NH <sub>4</sub> <sup>+</sup>
Sea salt (two species)	77.4	93.9	53.2	33.9	70.4	24.0	13.1	6.7
Crust	6.1	5.5	10.8	9.8	11.9	11.5	15.9	23.6
SIA	8.8	0.0	27.7	56.3	15.8	57.5	69.5	9.1
Oceanic emission	7.7	0.6	8.3	0.0	2.0	6.9	1.6	60.5

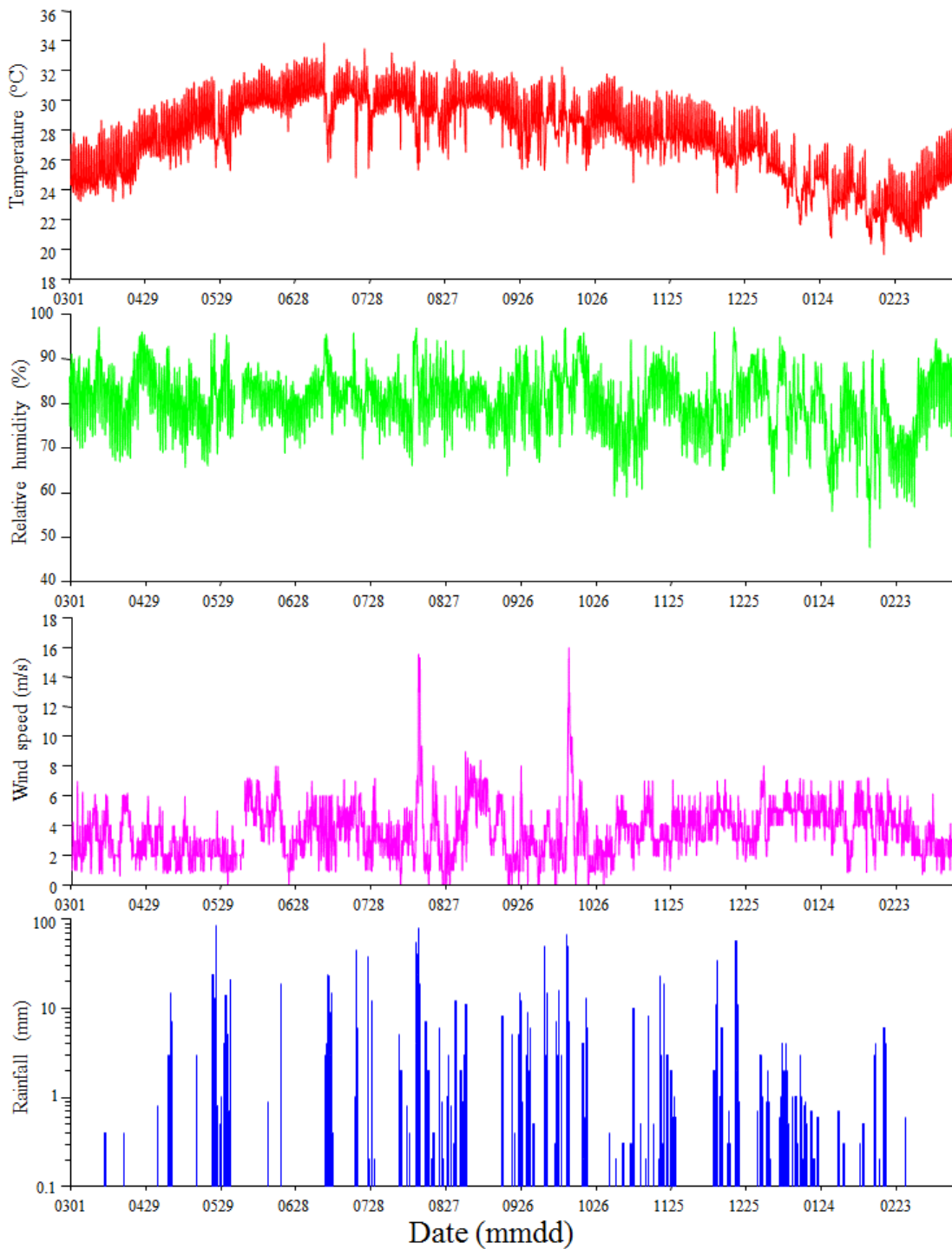
700

701 **Figure 1** Distribution of seasonal average aerosol optical thickness (AOT) at 550 nm (T550) over Northwest Pacific in cool, warm and transition seasons during  
702 sampling period. Monthly AOT products (from Moderate Resolution Imaging Spectrometer, MODIS) with 4-km resolution were downloaded from Globcolour  
703 (<http://hermes.acri.fr/>). The GlobColour project began in 2005 as an ESA Data User Element project to provide a continuous dataset of merged Level 3 Ocean  
704 Colour products.



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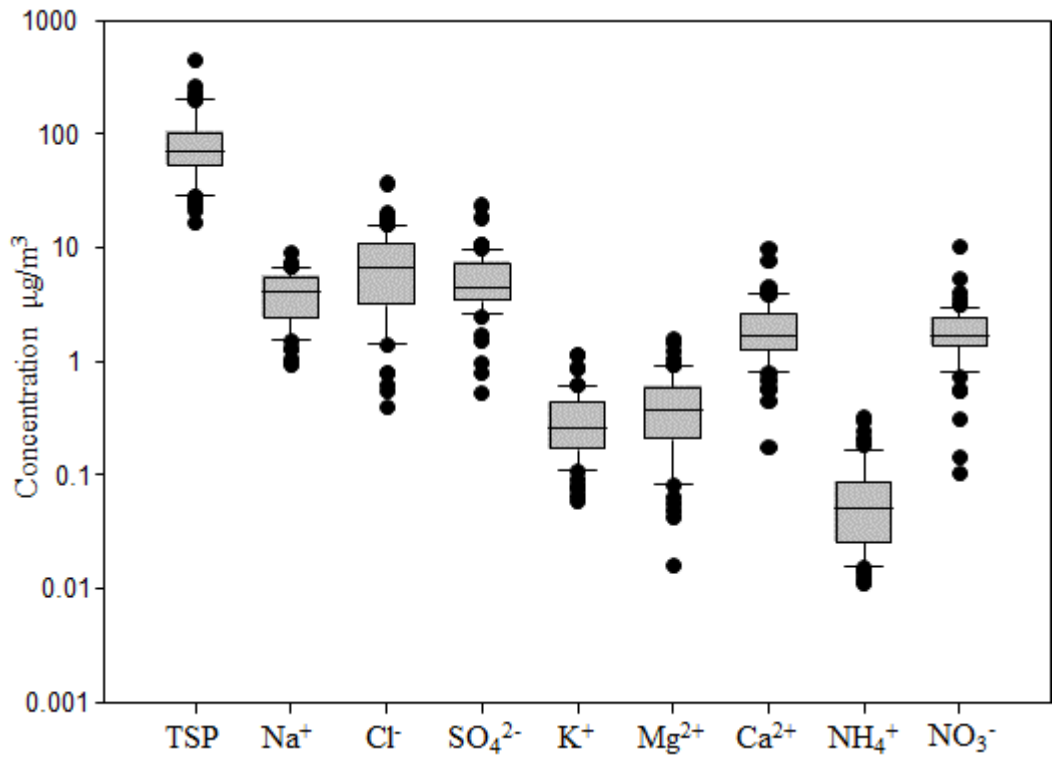
706 **Figure 2** Three-hour temperature, relative humidity, wind speed and precipitation at Yongxing  
707 Island during sampling period (March 2014 through February 2015).



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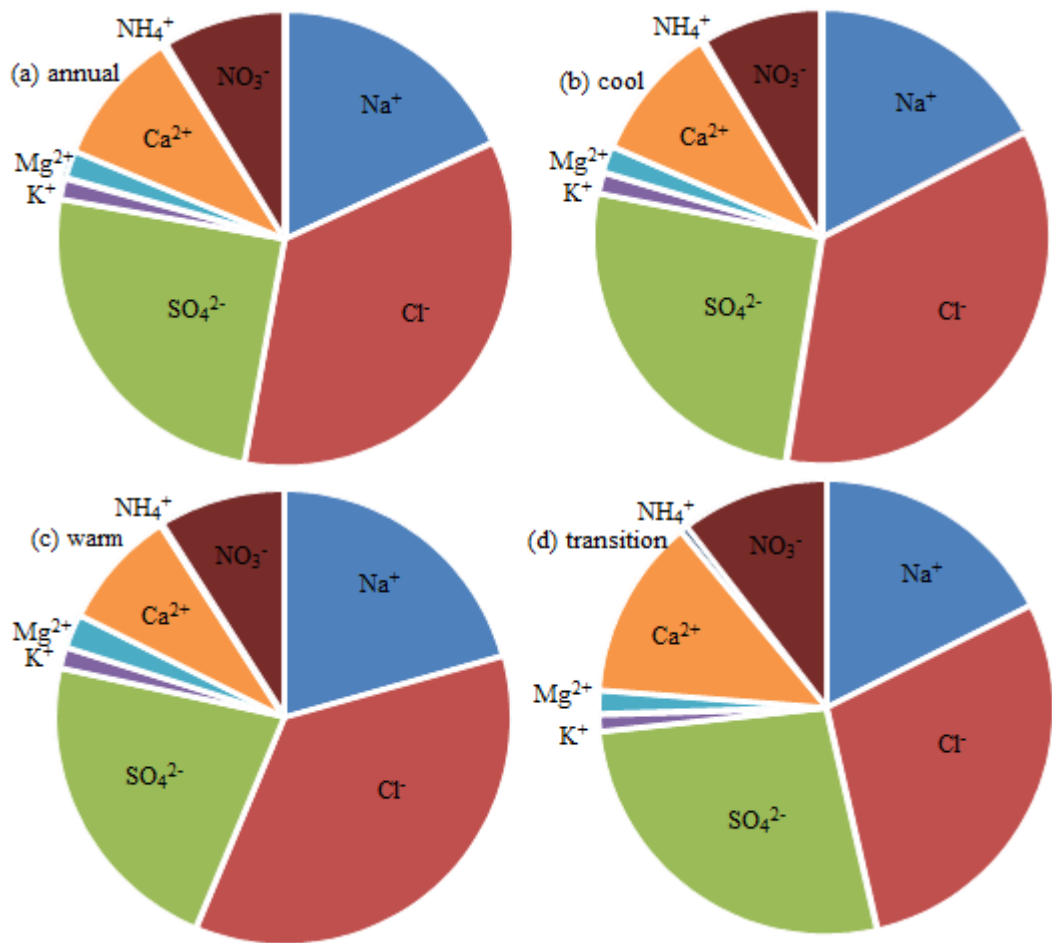


709 **Figure 3** Annual average mass concentrations of aerosol chemical species at Yongxing Island.  
710 Box boundary indicates 25<sup>th</sup> and 75<sup>th</sup> percentile. Lines within the box show the mean. Whiskers  
711 above and below the box indicate 90<sup>th</sup> and 10<sup>th</sup> percentiles.



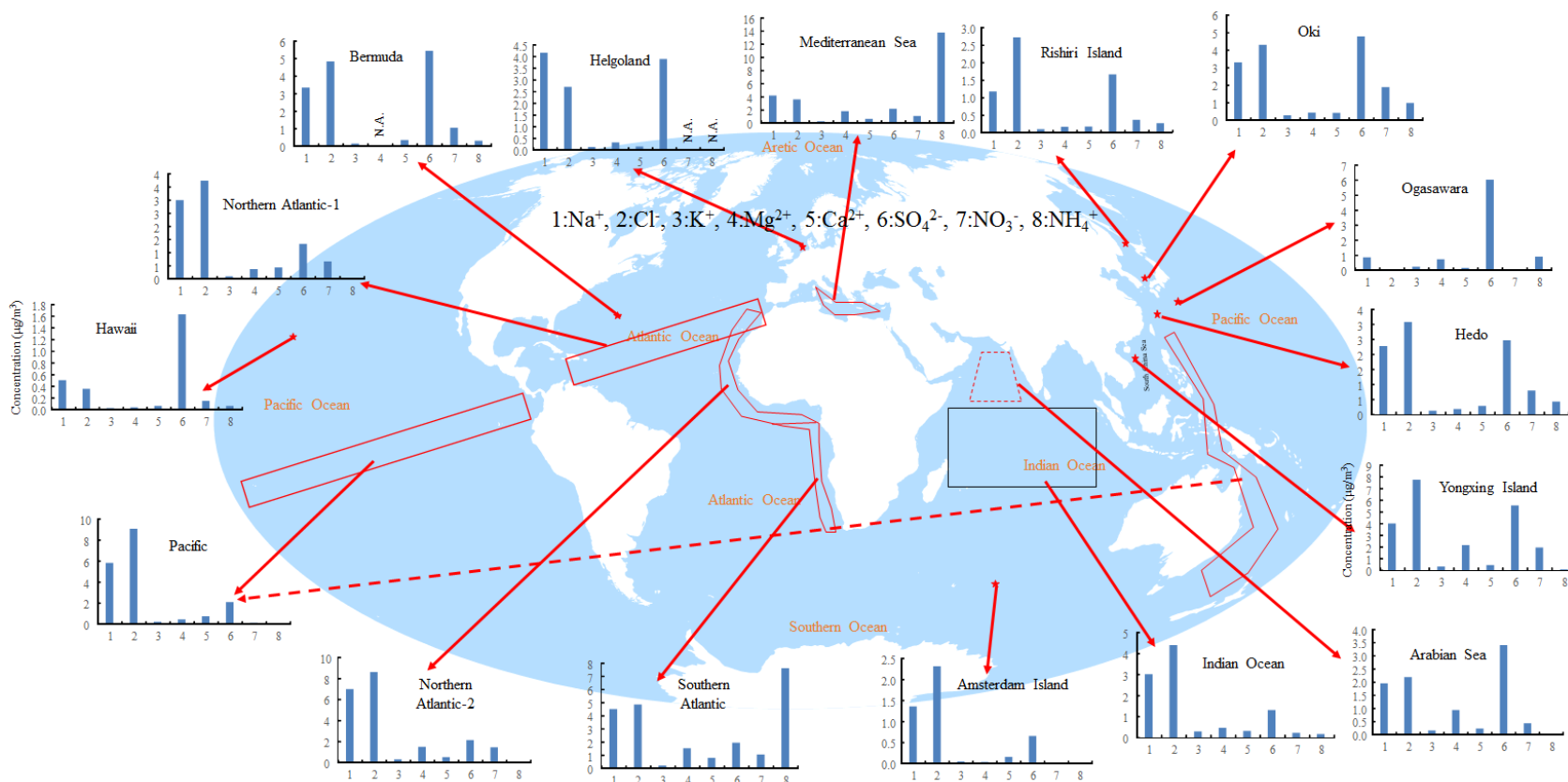
712

713 **Figure 4** Mean contributions of each major ionic component to total ionic mass concentration  
714 of (a) Yongxing Island (YXI) annual, (b) YXI cool season, (c) YXI warm season, and (d) YXI  
715 transition season.



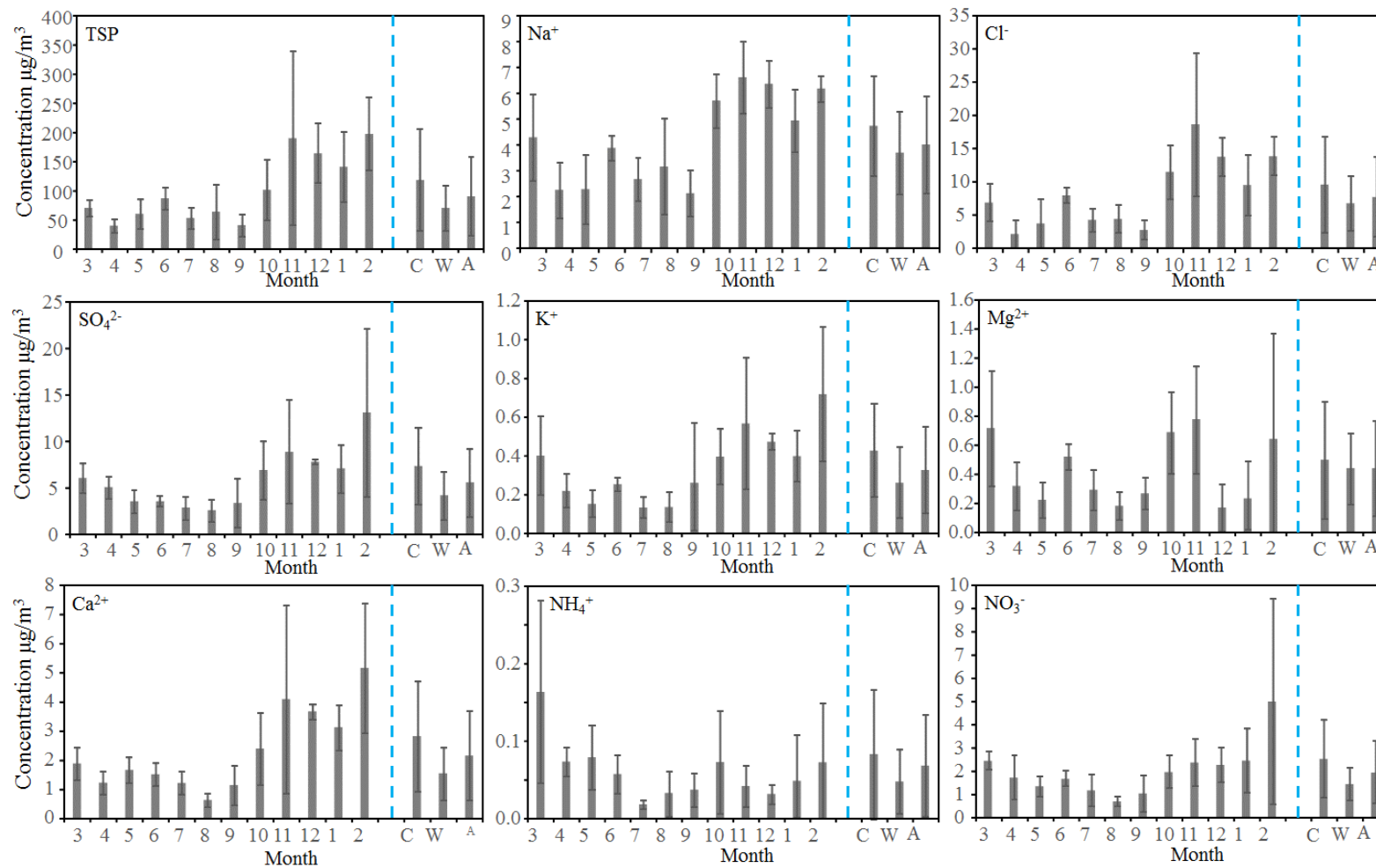
716

717 **Figure 5** Comparisons of major ions in aerosol at Yongxing Island with global ocean. Data of Oki, Ogasawara and Hedo are from EANET (www.eanet.asia);  
 718 those of Rishiri Island are from Okuda et al. (2006); those of Hawaii are from Carrillo et al. (2002); those of Bermuda are from Moody et al. (2014); those of  
 719 Amsterdam Island are from Claeys et al. (2010); those of the Arabian Sea and Indian Ocean are from Kumar et al. (2008); those of Helgoland are from Ebert et  
 720 al. (2000); those of the Mediterranean Sea, northern Atlantic-1 and 2, Pacific and southern Atlantic are from Zhang et al. (2010). Pentagrams represent sampling  
 721 sites on islands; others represent cruises. N.A. indicates no data.



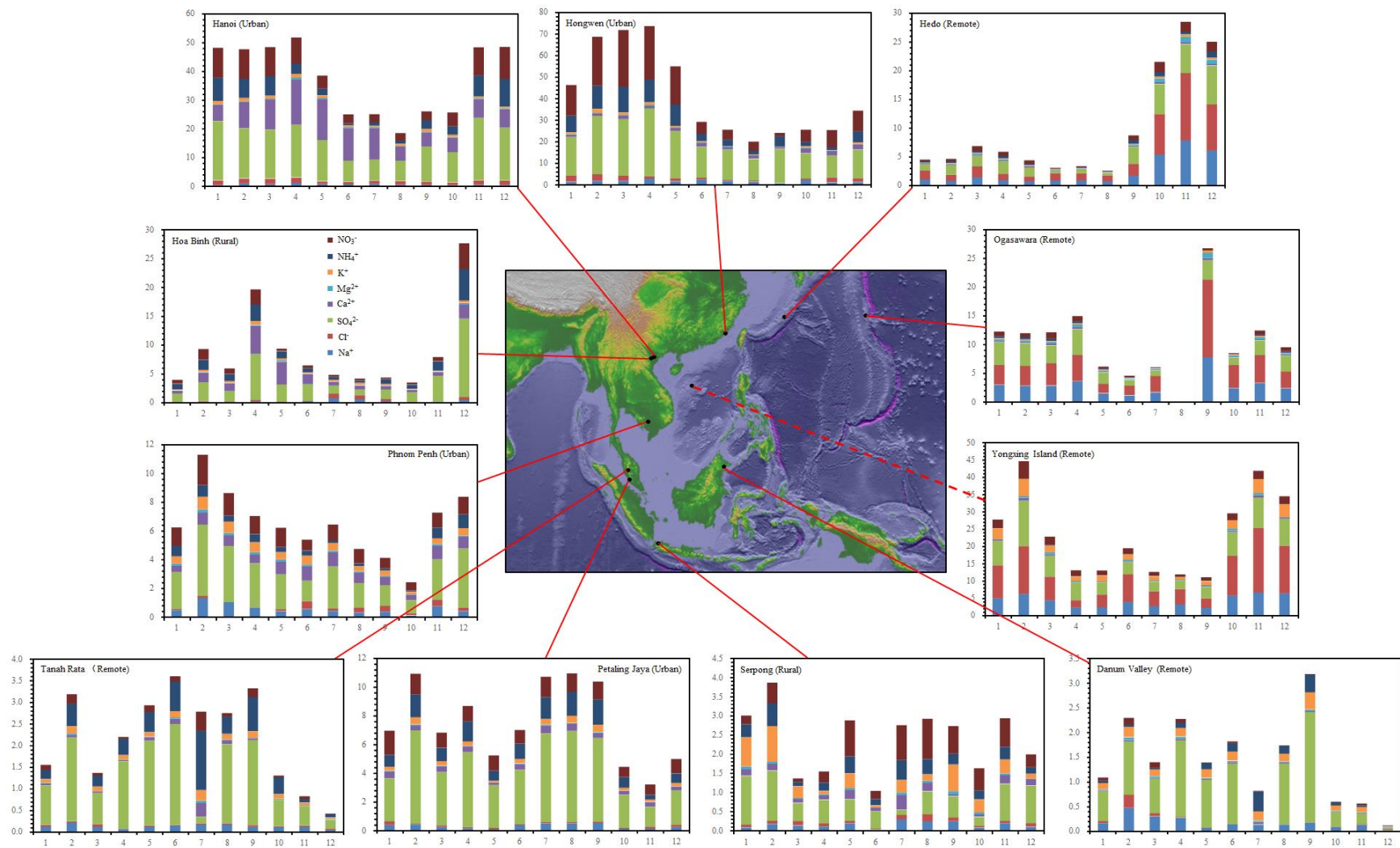
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723 **Figure 6** Seasonal variations of TSP mass concentration and associated species, including  $\text{Na}^+$ ,  $\text{Cl}^-$ ,  $\text{SO}_4^{2-}$ ,  $\text{Mg}^{2+}$ ,  $\text{K}^+$ ,  $\text{Ca}^{2+}$ ,  $\text{NH}_4^+$ , and  $\text{NO}_3^-$  at Yongxing Island  
 724 (cool season: C; warm season: W; annual: A). Shown are the mean and standard deviation for each bar.



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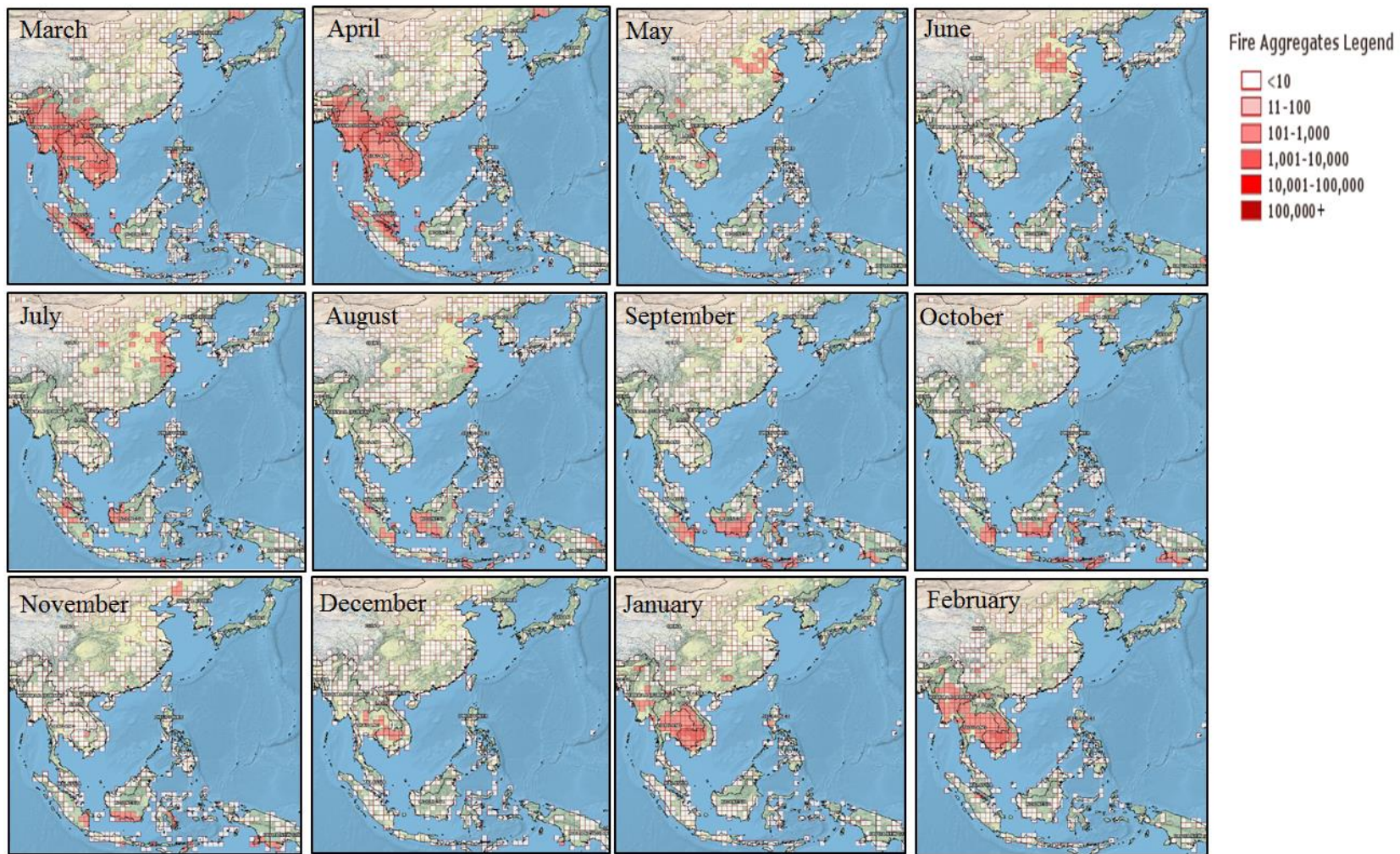
726 **Figure 7** Comparison of aerosol chemical species between Yongxing Island and around the South China Sea (data from EANET).



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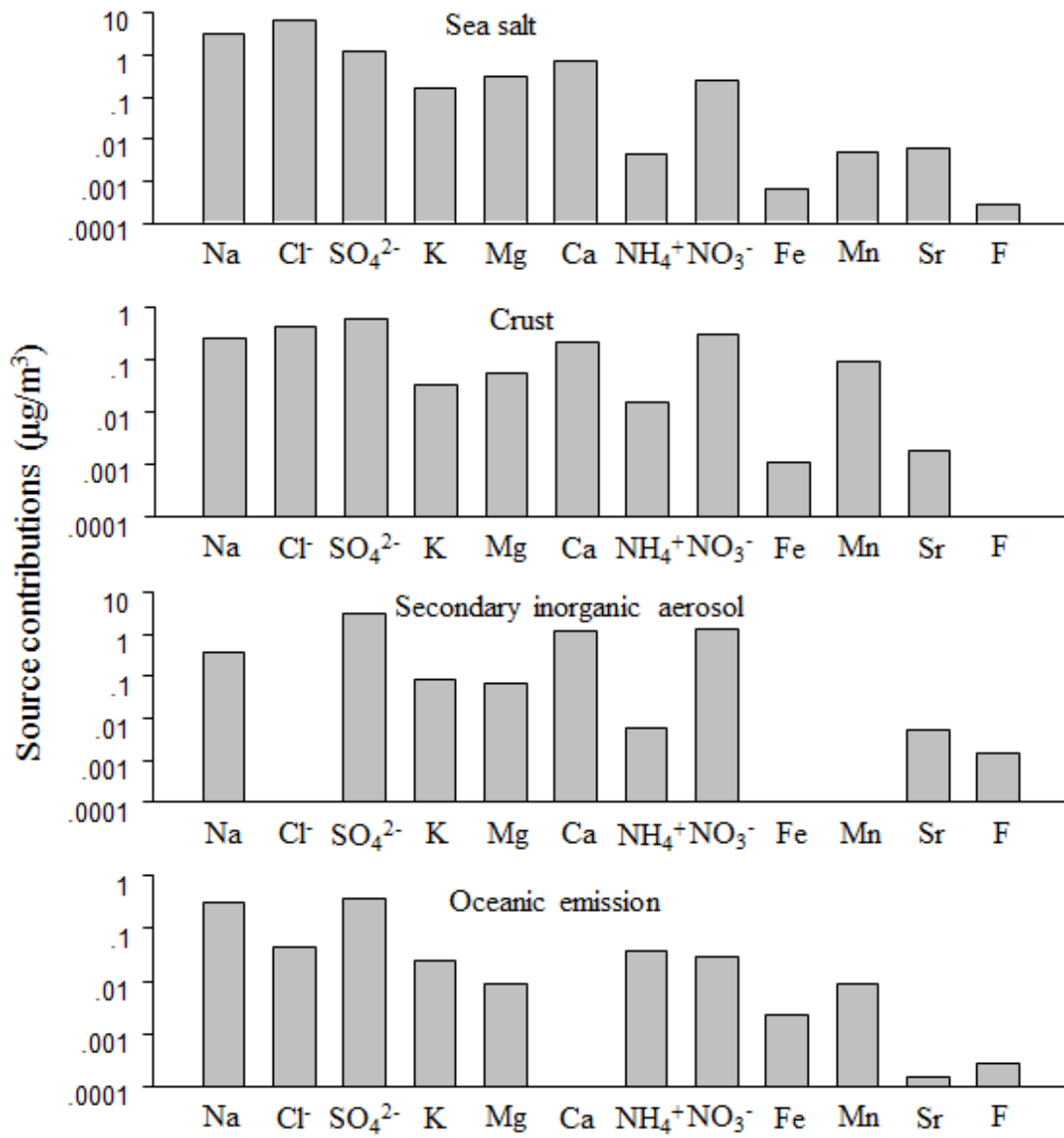


728 **Figure 8** Fire spot data from MODIS global fire mapping from March 2014 to February 2015 around South China Sea  
729 (<https://firms.modaps.eosdis.nasa.gov/firemap/>).



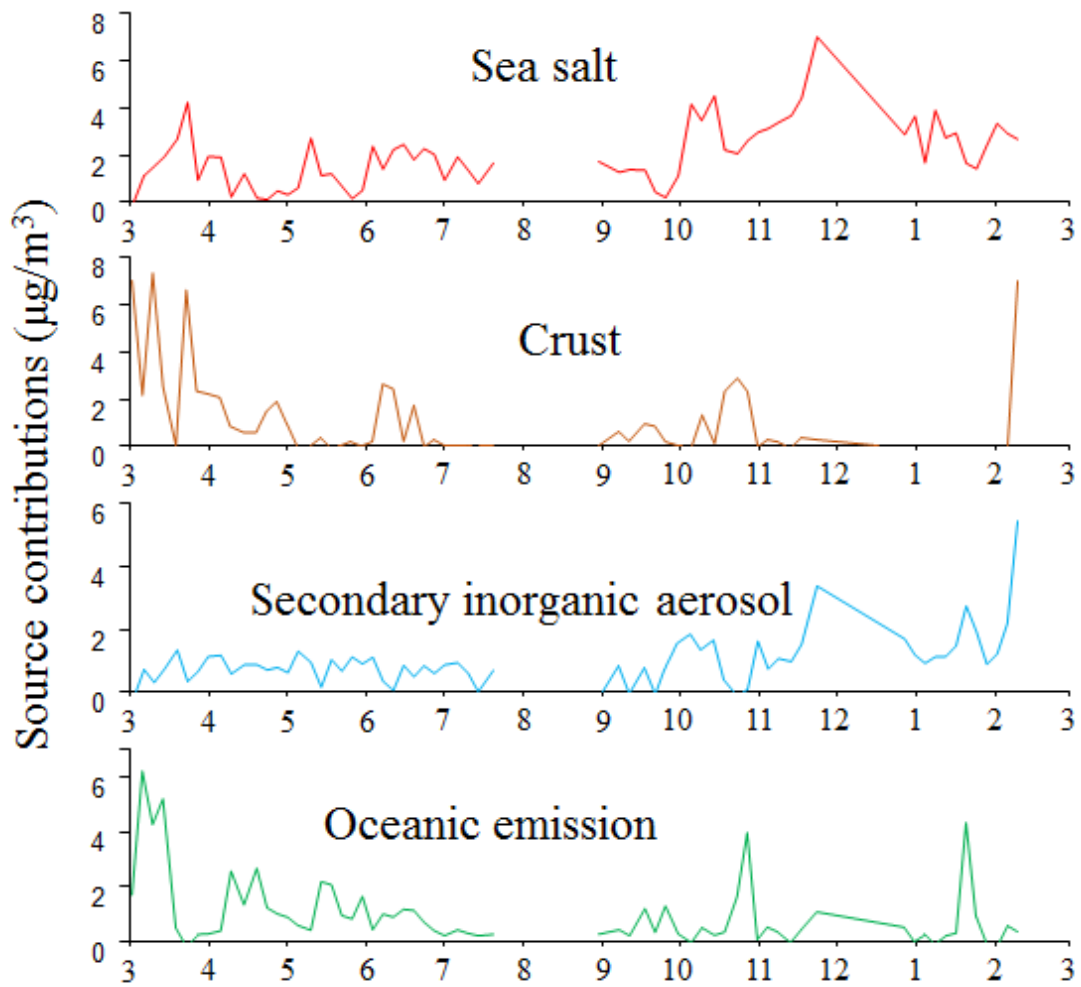
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731 **Figure 9** Profiles of five sources identified from the PMF 5.0 model, including sea salt (two  
 732 species), crust, secondary inorganic aerosol and oceanic emission.  
 733



734

735 **Figure 10** Time series contributions from each identified sources, including sea salt (two  
736 species), crust, secondary inorganic aerosol, and oceanic emission.f

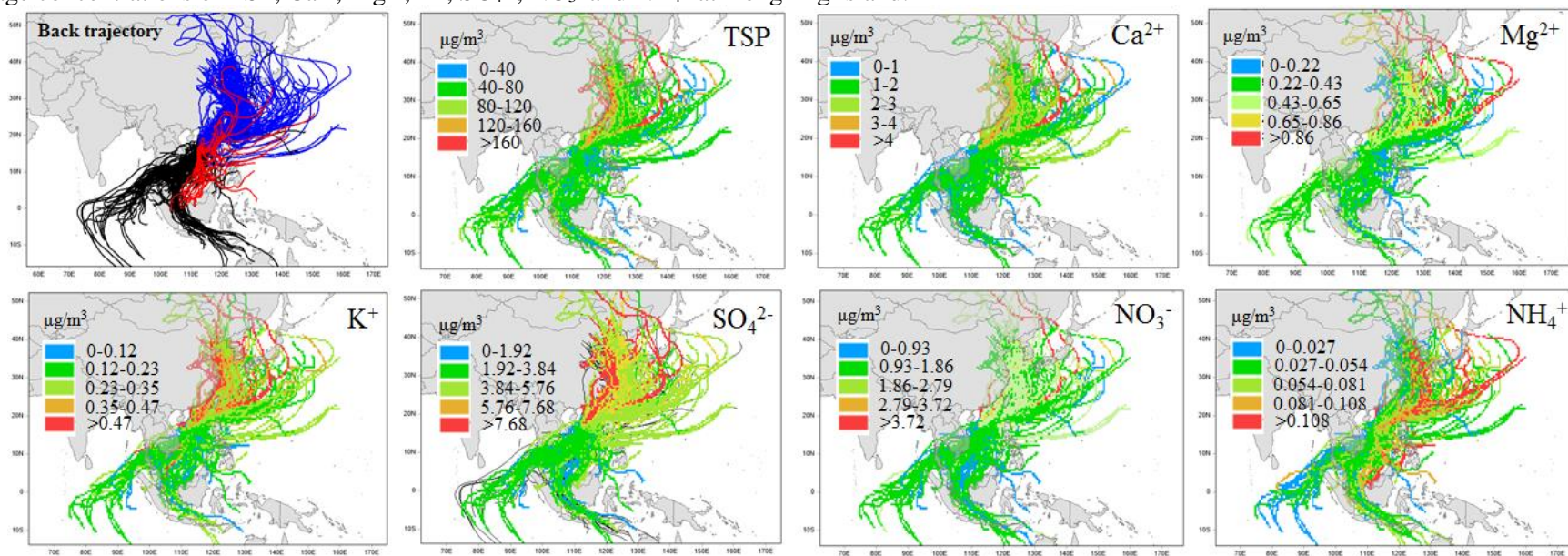


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739 **Figure 11** Ten-day back trajectories of warm (black, June through September 2014), cool (blue, March and April 2014 and October 2014 through February  
 740 2015) seasons, and transition season (red, May 2014) at Yongxing Island. Additionally, CWT (concentration weighted trajectory) plots for daily weighted-  
 741 average concentrations of TSP, Ca<sup>2+</sup>, Mg<sup>2+</sup>, K<sup>+</sup>, SO<sub>4</sub><sup>2-</sup>, NO<sub>3</sub><sup>-</sup> and NH<sub>4</sub><sup>+</sup> at Yongxing Island.



742

Supplementary Materials for  
**Atmospheric aerosol compositions over the South China Sea:**  
**Temporal variability and source apportionment**

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5 Zhen-Hua Long<sup>5</sup>, Da-Ning Li<sup>5</sup>

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This PDF file includes:

Supplementary Text: S1, S2 and S3

10 Reference List

**S1 Sample collection and chemical analyses**

Aerosol was collected on quartz filters (8 × 10 inch, Tissuquartz™ Filters, 2500 QAT-UP, Pallflex, Washington, USA) using a special high-flow rate (1.05 ± 0.03 m<sup>3</sup>/min) KC-1000  
15 sampler (Laoshan Institute for Electronic Equipment, Qingdao, China), which were installed 1m above the building (about 15m) roof's surface of the station of SCSIO, CAS, and there was not obvious pollution around this station. The sampling time was nominally 96 hours (4 days one sample). Three blank filters were taken from each package (25 filters). All samples and blank filters were stored in a refrigerator at -20°C until analysis in the laboratory.

20 In the laboratory, one eighth filters were cut and placed in a clean 50-ml Nalgene tube with additional 35-ml ultrapure water. These tubes were washed for 30 minutes using ultrasonic vibration. They were then shaken for 30 minutes on a horizontal shaker at a rate of ~300 rpm and left to rest for another 30 minutes at room temperature. The extract was filtered using pinhole filters, which were then rinsed twice with 5-ml ultrapure water. The extract and rinse  
25 were put into 50-ml tubes together and stored in a refrigerator at -20 °C until chemical analyses.

Major anion concentrations (F<sup>-</sup>, Cl<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, Br<sup>-</sup>) were determined by ICS-90 ion chromatography (Dionex, California, USA). Water-soluble metal and nonmetal elemental concentrations (Al, Ca, Fe, K, Mg, Mn, Na, SiO<sub>2</sub>, Sr) were analyzed by an MPX inductively coupled plasma optical emission spectrometer (ICP-OES, Vista, CA, USA). NH<sub>4</sub><sup>+</sup>

30 concentration was determined by spectrophotometry after treatment with Nessler's reagent. The  
detection limits of F<sup>-</sup>, Cl<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, Br<sup>-</sup> were 0.03, 0.03, 0.08, 0.075 and 0.1 mg/L,  
respectively, and the relative standard deviation of these ions of standard samples were 0.57%,  
2.55%, 1.16%, 1.36% and 11.36%, respectively (Xiao et al., 2013 and 2016). The detection  
limits of Al, Ca, Fe, K, Mg, Mn, Na, SiO<sub>2</sub>, Sr were 0.025, 0.003, 0.002, 0.06, 0.0005, 0.0005,  
35 0.02, 0.015 and 0.00008 mg/L, respectively, and the relative standard deviation of these ions of  
standard samples were less than 1.5% (Xiao et al., 2013 and 2016). The detection limit of NH<sub>4</sub><sup>+</sup>  
was 0.1 mg/L and its relative standard deviation was less than 5.0% (Xiao et al., 2013 and 2016).  
In this study, Al and Br<sup>-</sup> in most of samples was less than the detection limit.

#### 40 **S2 Back trajectories and concentration weighted trajectories analysis**

Back trajectories and concentration weighted trajectories (CWT) are used to determine the  
long-distance transport of atmospheric pollutants and regional source areas (Cheng et al., 2013;  
Xiao et al., 2014 and 2015). The CWT is a good model to estimate potential sources areas,  
when grid cells are more than 2 trajectory segment endpoint (Cheng et al., 2013). For each day,  
45 10-day (240 hours) back trajectories of air masses (Pavuluri et al., 2015) arriving at Yongxing  
Island were computed by the program of TrajStat (version 1.2.26) (Wang et al., 2009). 10-day  
back trajectories are used in this study since trajectories of a short duration are not long enough  
to indicate possible distant sources regions (Harris and Kahl, 1990). We also used the program  
of TrajStat to model CWT of TSP, and Ca<sup>2+</sup>, Mg<sup>2+</sup>, K<sup>+</sup>, SO<sub>4</sub><sup>2-</sup>, NO<sub>3</sub><sup>-</sup> and NH<sub>4</sub><sup>+</sup> concentrations at  
50 the island. In CWT model, each grid cell receives a weighted concentration obtained by  
averaging sample concentrations that have associated trajectories crossing the grid cell as  
follows (Eq. S1; Wang et al., 2009; Xiao et al., 2014 and 2015).

$$C_{ij} = \frac{1}{\sum_{l=1}^M \tau_{ijl}} \sum_{l=1}^M C_l \tau_{ijl} \quad (\text{S1})$$

where  $C_{ij}$  is the average weighted concentration in grid cell  $(i, j)$ ;  $l$  is the index of the trajectory;  
55  $C_l$  is the measured ionic concentration, corresponding with the arrival of back-trajectory  $l$  at the  
sampling site;  $\tau_{ijl}$  is the time spent in the grid cell  $(i, j)$  by trajectory  $l$ ;  $M$  is the total number of  
back trajectories. The region from 70°E to 160°E and from 20°S to 60°N was defined as the

source domain based on back trajectories during the sampling period, containing 14,400 grid cells of  $0.5^\circ \times 0.5^\circ$ .

60

### S3 Positive matrix factorization model

Receptor models are used to quantify the contributions of sources to samples based on the composition or fingerprints of the sources (Norris et al., 2014). The positive matrix factorization (PMF) is an effective source apportionment receptor model that does not require source profiles  
65 prior to analysis and has no limitation on source numbers (Crippa et al., 2013; Tiwari et al., 2013; Zhang et al., 2011; Zhang et al., 2015). The PMF model describes the observation ( $x_{ij}$ ) as a linear combination of a number of factors  $p$  for each time step  $i$  and  $j$ , whose contribution over time is always positive ( $g_k$ ) and whose mass spectra ( $f_k$ ) are static (see Eq. S2; Crippa et al., 2013; Paatero et al., 2014).

$$70 \quad x_{ij} = \sum_{k=1}^p g_{ik} f_{kj} + e_{ij} \quad (\text{S2})$$

where  $e_{ij}$  is the residual. PMF decomposes the matrix of speciated sample data into two matrices: factor contributions (G) and factor profiles (F) (Norris et al., 2014). G and F are derived by the PMF model minimizing the objective function Q (Eq. S3):

$$Q = \sum_{i=1}^n \sum_{j=1}^m \left[ \frac{e_{ij}}{u_{ij}} \right]^2 \quad (\text{S3})$$

75 where  $u_{ij}$  is the measurement uncertainty. In our study, PMF 5.0 (United States Environmental Protection Agency) was used to determine source apportionment of each major ion based on  $\text{F}^-$ ,  $\text{Cl}^-$ ,  $\text{NO}_3^-$ ,  $\text{SO}_4^{2-}$ ,  $\text{NH}_4^+$ , Ca, K, Mg, Mn, Fe, Na, and Sr by sampling time, with uncertainties by species provided by the analytical library. However, we downweighted Fe and  $\text{F}^-$  since they had low signal-to-noise ratios (S/N), and there were no excluded species and samples. We run PMF  
80 for the range of number factors from 2 to 10, and examine the  $Q(\text{Robust})/Q_{\text{exp}}$  to choose the best model number of factors ( $P$ ). Five physically realistic sources major ions were identified, i.e., sea salt (two species, Na and Mg), secondary inorganic aerosol (SIA;  $\text{F}^-$ ,  $\text{SO}_4^{2-}$  and  $\text{NO}_3^-$ ), oceanic emission ( $\text{NH}_4^+$ ), and crust (Mn). The model results show that there are good relationships between observed (input data) values and predicted (modeled) values of each  
85 major ion except Sr and  $\text{F}^-$  (correlation coefficients  $R^2$  were 0.94, 0.93, 0.94, 0.80, 1.00, 0.93,

0.95, 0.91, 0.59, 1.00, 0.40 and 0.40 for Na, Cl<sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, K, Mg, Ca, NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, Fe, Mn, Sr, and F<sup>-</sup>, respectively), and the modeled values of time series are fitting the observed values well. For the best solution chosen above, we run PMF for F<sub>peak</sub> in Rotational Tools (-1.0, -0.8, -0.6, -0.4, -0.2, -0.01, 0.01, 0.2, 0.4, 0.6, 0.8 and 1.0). The results show that the best solution is the  
90 F<sub>peak</sub> near to zero (-0.01 and 0.01) for P 5.

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