

1 Volatility of mixed atmospheric Humic-like Substances and ammonium sulfate particles

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21 Abstract

22 The volatility of organic aerosols remains poorly understood due to the complexity of speciation and
23 multi-phase processes. In this study, we extracted HUMIC-Like Substances (HULIS) from four
24 atmospheric aerosol samples collected at the SORPES station in Nanjing, eastern China, and
25 investigated the volatility behavior of particles at different sizes using a Volatility Tandem Differential

26 Mobility Analyzer (VTDMA). In spite of the large differences in particle mass concentrations, the
27 extracted HULIS from the four samples all revealed very high oxidation states ($O : C > 0.95$),
28 indicating secondary formation as the major source of HULIS in Yangtze River Delta (YRD). An
29 overall low volatility was identified for the HULIS samples, with the volume fraction remaining (VFR)
30 higher than 55% for all the re-generated HULIS particles at the temperature of 280 °C. A kinetic mass
31 transfer model was applied to the thermodenuder (TD) data to interpret the observed evaporation
32 pattern of HULIS, and to derive the mass fractions of semi-volatile (SVOC), low-volatility (LVOC)
33 and extremely low-volatility components (ELVOC). The results showed that LVOC and ELVOC
34 dominated (more than 80%) the total volume of HULIS. Atomizing processes led to a size dependent
35 evaporation of regenerated HULIS particles, and resulted in more ELVOCs in smaller particles. In
36 order to understand the role of interaction between inorganic salts and atmospheric organic mixtures in
37 the volatility of an organic aerosol, the evaporation of mixed samples of ammonium sulfate (AS) and
38 HULIS was measured. The results showed a significant but nonlinear influence of ammonium sulfate
39 on the volatility of HULIS. The estimated fraction of ELVOCs in the organic part of largest particles
40 (145 nm) increased from 26% in pure HULIS samples to 93% in 1:3 (mass ratio of HULIS:AS) mixed
41 samples, to 45% in 2:2 mixed samples, and to 70% in 3:1 mixed samples, [suggesting that the](#)
42 [interaction with ammonium sulfate tends to decrease the volatility of atmospheric organic compounds.](#)
43 Our results demonstrate that HULIS are important low volatile, or even extremely low volatile,
44 compounds in the organic aerosol phase. As important formation pathways of atmospheric HULIS,
45 multi-phase processes, including oxidation, oligomerization, polymerization and interaction with
46 inorganic salts, are indicated to be important sources of low volatile and extremely low volatility
47 species of organic aerosols.

48 1. Introduction

49 Atmospheric organic aerosol (OA) comprises 20-90% of the total submicron aerosol mass depending
50 on location (Kanakidou et al., 2005; Zhang et al., 2007; Jimenez et al., 2009), and play a critical role in
51 air quality and global climate change. Given the large variety of organic species, OA is typically
52 grouped in different ways according to its sources and physicochemical properties. These include the
53 classifications based on aerosol optical properties (brown carbon and non-light absorption OA),
54 formation pathways (primary (POA) and secondary (SOA) organic aerosol) and solubility (water
55 soluble OA (WSOA) and water insoluble OA (WISOA)). Humic-Like Substances (HULIS), according

56 to their operational definition, are the hydrophobic part of WSOA, and contribute to more than half of
57 the WSOA (e.g. Krivácsy et al., 2008). Secondary formation (Lin et al., 2010b) and primary emission
58 from biomass burning (Lukács et al., 2007; Lin et al., 2010a) have been identified as the two major
59 sources of atmospheric HULIS. Because they are abundantly present, water-soluble, light-absorbing
60 and surface-active, HULIS in atmospheric particles have been demonstrated to play important roles in
61 several processes, including cloud droplet formation, light [absorption](#) and heterogeneous redox
62 activities (Kiss et al., 2005; Graber and Rudich, 2006; Hoffer et al., 2006; Lukács et al., 2007; Lin and
63 Yu, 2011; Verma et al., 2012; Kristensen et al., 2012).

64 Volatility of atmospheric organic compounds is one of their key physical properties determining their
65 partitioning between the gas and aerosol phases, thereby strongly influencing their lifetimes and
66 concentrations. Atmospheric OA can be divided into semi-volatile organic compounds (SVOC), low
67 volatility organic compounds (LVOC) and extremely low volatility organic compounds (ELVOC)
68 (Donahue et al., 2012; Murphy et al., 2014). LVOC and ELVOC are predominantly in the aerosol
69 phase and contribute largely to the new particle formation and growth (Ehn et al., 2014), while SVOC
70 have considerable mass fractions in both phases and usually dominate the mass concentration of OA.
71 As far as we know, volatility studies on OA have mostly focused on laboratory-generated organic
72 particles or ambient particles (Kroll and Seinfeld, 2008; Bilde et al., 2015). Laboratory-generated
73 organic particles contain only a small fraction of compounds present in atmospheric OA, whereas
74 ambient particles are usually complex mixtures of thousands of organic and several inorganic
75 compounds. One way to interlink laboratory and ambient studies, and to understand the volatility of
76 ambient OA systematically, might be to isolate some classes of OA from ambient particles before
77 investigating their volatility separately. As an important sub-group of organic aerosols in the real
78 ambient aerosols, the physicochemical properties of HULIS have been studied widely, including their
79 mass concentrations (Lin et al., 2010b), chemical composition (Lin et al., 2012; Kristensen et al., 2015;
80 Chen et al., 2016), density (Dinar et al., 2006) and hygroscopicity (Wex et al., 2007; Kristensen et al.,
81 2014). However, to the best of our knowledge, the volatility of atmospheric HULIS has never been
82 reported so far.

83 In the ambient aerosol, organic aerosol (OA, including HULIS) mostly co-exist with inorganic
84 compounds, such as ammonium sulfate. The volatility of OA has been demonstrated to be affected by
85 aerosol-phase reactions when mixed with inorganic compounds (Bilde et al., 2015). The most typical

86 examples of these are interactions between particulate inorganic salts with organic acids to form
87 organic salts, which evidently can enhance the partitioning of organic acids onto the aerosol phase
88 (Zardini et al., 2010; Laskin et al., 2012; Häkkinen et al., 2014; Yli-Juuti et al., 2013;). Recent studies
89 have reported that the saturation vapor pressure (p_{sat}) of ammonium oxalate is significantly lower than
90 that of pure oxalic acid, with p_{sat} being around 10^{-6} Pa for ammonium oxalate (Ortiz-Montalvo et al.,
91 2014; Paciga et al., 2014). However, this has not shown to be the case for adipic acid vs. ammonium
92 adipate, indicating that not all dicarboxylic acids react with ammonium to form low-volatility organic
93 salts (Paciga et al., 2014). Given that HULIS contain acidic species (Paglione et al., 2014; Chen et al.,
94 2016), their interaction with inorganic salts would plausibly influence their volatility.

95 In this study, HULIS were extracted from PM_{2.5} filter samples collected at the SORPES station (Station
96 for observing Regional Processes of the Earth System) in western Yangtze River delta (YRD) during
97 the winter of 2014 to 2015. A Volatility-Hygroscopicity Tandem Differential Mobility Analyzer
98 (VHTDMA) was then used to measure the volatility properties of extracted HULIS and their mixtures
99 with ammonium sulfate. A kinetic mass transfer model was deployed to re-build the measured
100 thermograms, and to separate the mixture into three volatility fractions having an extremely low
101 volatility, low volatility and semi-volatility. Our main goals were (1) to characterize the volatility of
102 size-dependent, re-generated HULIS particles and to get insight into the relationship between
103 atmospheric HULIS and ELVOC, and (2) to understand how the interaction between HULIS and
104 inorganic salts affect their volatilities.

105 2. Methods

106 2.1 Sample collection and HULIS extraction

107 The SORPES station is located on the top of a hill in the Xianlin campus of Nanjing University, which
108 is about 20 km east from the downtown Nanjing and can be regarded as a regional background site of
109 Yangtze River delta (YRD) (Ding et al., 2013; Ding et al., 2016). 24-hour PM_{2.5} samples were collected
110 on quartz filters using a middle-volume PM_{2.5} sampler during the winter of 2014 to 2015. HULIS were
111 extracted from four aerosol samples for the following volatility measurements.

112 Water-soluble inorganic ions, organic carbon (OC) and elemental carbon (EC) were measured online
113 using a Monitor for Aerosols and Gases in Air (MARGA) and a sunset OC/EC analyzer during the
114 sampling periods. WSOC were extracted from portions of the sampled filters using sonication in

115 ultrapure water with the ratio of 1 mL water per 1 cm² filter. Insoluble materials were removed by
116 filtering the extracts with a 0.45 µm Teflon filter (Millipore, Billerica, MA, USA). A TOC analyzer
117 with a non-dispersive infrared (NDIR) detector (Shimadzu TOC-VCPH, Japan) was used to determine
118 WSOC concentrations. The aerosol water extracts were then acidified to pH = 2 by HCl and loaded
119 onto a SPE cartridge (Oasis HLB, 30 µm, 60 mg / cartridge, Waters, USA) to isolate the HULIS
120 following the procedure described in Lin et al (2010b). Most of the inorganic ions, low-molecular-
121 weight organic acids and sugars were removed, with HULIS retaining on the SPE cartridge. A total 20
122 ml of methanol was then used to elute the HULIS. The eluate was evaporated to dryness under a gentle
123 stream of nitrogen gas. A part of the HULIS eluate was re-dissolved in 1.0 mL water to be quantified
124 with an evaporative light scattering detector (ELSD). It should be noted here that HULIS extracted in
125 this work refers to the part of water-soluble organic compounds that are hydrophobic. In case that the
126 compounds removed by the isolation processes, especially the inorganic species, may influence the
127 evaporation behavior of HULIS, we also re-induce ammonium sulfate, the most important inorganic
128 salt to the extracted HULIS and investigate the volatility of the mixed samples.

129 2.2. Volatility measurements by VTDMA measurements

130 The evaporation behavior of HULIS and their mixtures with AS was measured using a Volatility
131 Tandem Mobility Analyzer, which is part of a Volatility-Hygroscopicity Tandem Differential Mobility
132 Analyzer (VH-TDMA) system (Hong et al., 2014). During the measurements, the hygroscopicity mode
133 was deactivated, so that only the volatility mode of this instrument was functioning. Briefly, aerosol
134 particles were generated by atomizing aqueous solutions consisting of HULIS and their mixtures with
135 AS by using an atomizer (TOPAS, ATM 220). Then, a monodisperse aerosol with particle sizes of 30,
136 60, 100 and 145 nm were selected by a Hauke-type Differential Mobility Analyzer (DMA, Winklmayr
137 et al., 1991). The monodisperse aerosol flow was then heated by a thermodenuder at a certain
138 temperature, after which the number size distribution of the particles remaining was determined by a
139 second DMA and a condensation particle counter (CPC, TSI 3010). The thermodenuder was a 50-cm-
140 long stainless steel tube with an average residence time of around 5 s.

141 The VTDMA measures the shrinkage of the particle diameter after heating particles of some selected
142 initial size at different temperatures. Conventionally, the volume fraction remaining (VFR), i.e. the
143 fraction of aerosol mass left after heating particles of diameter D_p , is used to describe the evaporation

144 quantitatively. $D_p(T_{room})$ is the initial particle diameter at room temperature. $D_p(T)$ is the particle
145 diameter after passing through the thermodenuder at the temperature T .

146 The VFR can be defined as:

$$147 \text{VFR}(D_p) = \frac{D_p^3(T)}{D_p^3(T_{room})} \quad (1)$$

148 In this work, we totally analyzer 8 samples collected during both winter and summer, and covering a
149 wide range of PM concentration from less than $40 \mu\text{g}/\text{m}^3$ to higher than $150 \mu\text{g}/\text{m}^3$. All these 8 samples
150 showed similar evaporation behavior with some small differences in details (figures not shown).
151 Therefore, in terms of volatility, we believe that there were no large differences between the collected
152 HULIS samples. We finally selected 4 samples with different PM concentrations to represent the
153 HULIS samples at the SORPES station and made the argument clear.

154 2.3. Kinetic mass transfer model

155 A kinetic mass transfer model (Riipinen et al., 2010) was applied to help interpreting the HULIS
156 evaporation data. The size distribution, chemical composition and physicochemical properties of the re-
157 generated HULIS particles, as well as the residence time of the particles traveling through the
158 thermodenuder, were predefined in the model. As an output, the model provided the particle mass
159 change as a function of the residence time, which can either increase or decrease depending on the
160 particle composition, volatility of compounds and concentrations of surrounding vapors. With the aim
161 to reproduce the observed evaporation pattern of HULIS particles measured by the VTDMA, the model
162 applied an optimization procedure to minimize the difference between the measured and modeled
163 evaporation curves of the HULIS particles.

164 In the model, particles were assumed to consist of compounds that can be grouped into three volatility
165 bins: semi-volatile, low-volatility and extremely low-volatility components. These three “bins” were
166 quantified by assuming that they had fixed volatilities with $p_{\text{sat}}(298 \text{ K}) = [10^{-3} \ 10^{-6} \ 10^{-9}] \text{ Pa}$. Modeling
167 was performed for each experiment / sample separately, with 4 samples and 4 different initial particle
168 sizes ($D_p = 30, 60, 100$ and 145 nm), leading to 16 different model runs, each providing information on
169 how much semi-volatile, low-volatile and extremely low-volatility matter (X_i) was present in the
170 investigated particles. The initial particle size refers to the particle diameter prior to heating. The values
171 for $p_{\text{sat}}(298 \text{ K})$ and ΔH_{vap} (see Table 1 and text above) were selected by doing a preliminary test model

172 runs. With ΔH_{vap} of around [40 40 40] kJ mol⁻¹ and p_{sat} (298 K) of [10⁻³ 10⁻⁶ 10⁻⁹] Pa the model was
173 best able to reproduce the observed evaporation curves of the HULIS aerosol. Such low vaporization
174 enthalpies (referred often as effective vaporization enthalpies) for aerosol mixtures, for example for
175 SOA from α -pinene oxidation, have been reported also in previous studies (Häkkinen et al.,
176 2014;Donahue et al., 2005;Offenberg et al., 2006;Riipinen et al., 2010). The molecular weight and
177 density of HULIS were assumed to be 280 g mol⁻¹ (Kiss et al., 2003; Lin et al., 2012) and 1.55 kg m⁻³
178 (Dinar et al., 2006), respectively. Sensitivity of this kinetic evaporation model was tested towards different
179 values of mass accommodation coefficient (MAC) (i.e. MAC=1, 0.1, 0.01) for both pure HULIS sample and
180 mixed samples (figure not shown). The results suggested that 1 is the proper MAC value to best reproduce the
181 measured evaporation behavior (Table 1).

182 Volatility information, specifically described as the saturation vapor pressure and vaporization enthalpy
183 here, of ammonium sulfate was determined by interpreting the evaporation behavior of laboratory-
184 generated AS particles using the kinetic evaporation model. By setting the saturation vapor pressures
185 and enthalpy of vaporization of AS as fitting parameters, the optimum solution was obtained by
186 minimizing the difference between the measured and model-interpreted thermograms of AS particles.
187 Hence, p_{sat} (298 K) of 1.9·10⁻⁸ Pa and ΔH_{vap} of 97 kJ mol⁻¹ for AS were determined and used in the
188 following analysis.

189 2.4 AMS measurement for oxygen to carbon ratio

190 The O : C (Oxygen to carbon) ratios of re-generated HULIS particles were measured using a high-
191 resolution time-of-flight aerosol mass spectrometer (HR-Tof-AMS, Aerodyne Research Inc., Billerica,
192 USA). Detailed descriptions of the instrument and data processing can be found in previous
193 publications (DeCarlo et al., 2006; Canagaratna et al., 2007). The HULIS solution was atomized to
194 generate poly-dispersed aerosol particles and introduced into AMS. The AMS was operated in V mode
195 and the data was acquired at 5-min saving intervals. The AMS data were analyzed using standard Tof-
196 AMS data analysis toolkits (SQUIRREL version 1.57H and PIKA version 1.16H in Igor Pro software
197 (version 6.22A, WaveMetrics Inc.). For mass calculations, the default relative ionization efficiency
198 (RIE) values 1.1, 1.2, 1.3 and 1.4 for nitrate, sulfate, chloride and organic were applied, respectively.
199 The RIE for ammonium was 2.6, determined from the ionization efficiency calibration. In elemental
200 analysis, the “Improved- Ambient” method was applied to calculate O:C ratios by considering the
201 CHO⁺ ion correction (Canagaratna et al., 2015).

202 3. Results and discussions

203 Figure 1 shows the chemical compositions of the four PM_{2.5} samples, and the oxygen to carbon ratio
204 (O : C) of the extracted HULIS in related samples. The four samples can be classified into two groups
205 based on their PM_{2.5} concentrations (the sum of all measured chemical compositions), with one group
206 (samples 1 and 2) having the PM_{2.5} higher than 110 µg/m³ and the other one (samples 3 and 4) having
207 the PM_{2.5} lower than 40 µg m⁻³. The concentrations of inorganic compounds (sulfate, nitrate and
208 ammonium) were significantly higher in samples 1 and 2 than in samples 3 and 4. The HULIS
209 concentrations were also higher in samples 1 and 2 (about 9 µg/m³, ratio of HULIS-carbon to OC were
210 about 0.3) than in samples 3 and 4 (about 6 µg/m³, ratio of HULIS-carbon to OC were about 0.4). The
211 oxidation states of the HULIS, however, did not show any notable differences, showing very high
212 values for all the four samples (O:C > 0.95), indicating that the HULIS in YRD could be mostly
213 secondarily formed even during the relatively clean days. Such high oxidation states suggest further
214 that the extracted HULIS were very likely highly-oxidized, multifunctional compounds (HOMs)
215 originating from multi-phase oxidation (Graber and Rudich, 2006). Hydroxyl radical (OH) initiated
216 oxidation of aromatics followed by auto-oxidation was recently demonstrated to be able to form HOMs
217 in a laboratory work (Molteni et al., 2016).

218 3.1 Volatility of atmospheric HULIS

219 The volume fraction remaining (VFR) of the HULIS particles as a function of the heating temperature
220 obtained from VTDMA is illustrated in Fig. 2. An overall low volatility was identified for the HULIS
221 particles, with the VFR higher than 55% for the particles of all 4 sizes at the heating temperature of
222 280 °C and residence time of 5 s. Small differences in the volatility could be observed between the
223 samples of high mass concentrations and low mass concentrations in that the evaporation of HULIS in
224 samples 1 and 2 was in general weaker than that in samples 3 and 4. In addition, all the samples started
225 to evaporate from the very beginning of the heating program (around 20 °C to 25 °C) and the
226 evaporation curves varied smoothly, suggesting that the HULIS particles were mixtures of compounds
227 having wide range of saturation vapor pressures.

228 A kinetic mass transfer model was applied to reproduce the observed evaporation of the HULIS, and to
229 estimate the mass fractions of semi-volatile (SVOC, $p_{\text{sat}}(298\text{K}) = 10^{-3}$ Pa), low-volatility (LVOC, p_{sat}
230 (298K) = 10^{-6} Pa) and extremely low-volatility organic components (ELVOC, $p_{\text{sat}}(298\text{K}) = 10^{-9}$ Pa).

231 As shown in Fig. 3, the model performed reasonably well in simulating the “pure” HULIS particles
232 (example for sample 1). Noting that the HULIS mixtures were represented with only three model
233 compounds of different volatilities, the modeled evaporation curves of the HULIS in all samples
234 showed a relatively good agreement with the measured evaporation curves for all the four particle sizes.
235 The shape of the modeled thermograms is not as smooth as that of the measured ones suggesting lower
236 number of volatilities in simulations compared with in the real samples. The model-simulated
237 distributions of SVOC, LVOC and ELVOC of each HULIS sample gave indication on the volatility of
238 HULIS. As shown in Fig. 4, all the HULIS samples consisted of compounds from all the 3 volatility
239 “bins”, further confirming HULIS to be mixtures of compounds with wide range of volatilities. SVOC
240 was estimated to account for only small proportion (less than 20% of the particle mass) of the HULIS
241 samples, while LVOC and ELVOC dominated these samples (78% - 97% of the particle mass),
242 suggesting an overall low volatility of the extracted HULIS. Given that the heating program has the
243 potential to raise the evaporation of HULIS by decomposing large molecules, the real volatility of
244 atmospheric HULIS could be even lower than obtained here.

245 In spite of their overall low values, the volatilities of the HULIS varied between the different samples.
246 The HULIS extracted from the samples of higher particle mass loadings (samples 1 and 2) had, in
247 general, lower volatilities than those extracted from the samples of lower particle mass concentrations
248 (samples 3 and 4). By taking 30 nm particles as an example, sample 2 had the largest mass fraction of
249 ELVOC, up to 72%, followed by sample 1 (66%) and sample 3 (64%), while sample 4 had the least
250 amount of ELVOC (58%). Correspondingly, the mass fraction of SVOC in 30 nm particles was the
251 highest in sample 4 (9%) and the lowest in sample 2 (6%). Several factors, including the molecular
252 weight, oxidation state and molecular structure of the compounds, as well as their interaction with other
253 compounds, can influence the volatility of HULIS. Although there is not enough information to support
254 the final conclusion, we excluded the oxidation state as a key factor here because its variation did not
255 match the volatility changes of the HULIS samples. As can be seen from Figs.1 and 4, sample 2
256 showed the lowest volatility but the third highest oxidation state of the four samples. Instead of the
257 oxidation state, the interaction between HULIS and inorganic species is a more likely candidate for
258 influencing the observed variation of the HULIS volatility, especially as the lower-volatility samples
259 (sample 1 and sample 2) had higher concentrations and fractions of inorganic species (Fig. 1).

260 Within individual HULIS samples, the estimated amount of ELVOC, LVOC and SVOC varied with the
261 particle size (Fig. 4). The mass fraction of ELVOC was in the range of 58–72% for the smallest
262 particles (30 nm in diameter) and decreased to the range of 47–60% for the 60 nm, to the range of
263 35–53% for the 100 nm particles, and to the range of 20–39% for the 145 nm particles. The amount of
264 LVOC increased correspondingly with an increasing particle size, from 23–33% for the 30 nm particles
265 to 52–65% for the 145 nm particles. The amount of SVOC slightly increased with an increasing
266 particle size, on average from 7.5% (30 nm) to 14.5% (145 nm). The most likely explanation for this
267 behavior is that, due to the Kelvin effect, compounds with higher volatilities are likely to evaporate
268 more from smaller particles. This result indicates that size-resolved chemical compositions of
269 laboratory-generated particles from aqueous solutions of mixtures should be examined more carefully
270 to support their size-dependent physical properties from lab studies.

271 3.2 Interaction between HULIS and ammonium sulfate

272 Interactions between inorganic and organic matter have been shown to influence the volatility of the
273 organic matter. However, recent work has focused on the interaction between one specific organic
274 compound and some inorganic salt(s). For example, Laskin et al. (2012) observed the formation of
275 sodium organic salt in a submicron organic acid-NaCl aerosol. Ma et al. (2013) reported that the
276 formation of sodium oxalate can occur in particles containing oxalic acid and sodium chloride.
277 Häkkinen et al. (2014) demonstrated that low-volatility material, such as organic salts, were formed
278 within aerosol mixtures of inorganic compounds with organic acids. Zardini et al. (2010) and Yli-Juuti
279 et al. (2013b) suggested that interactions between inorganic salts and organic acids in the particle phase
280 might further enhance the partitioning of organic acids onto the particle phase. Given the complex
281 nature of organic aerosols in the real atmosphere, large uncertainties will be induced when using
282 simplified laboratory results for explaining observations in the real atmosphere. In this study, we
283 investigated the volatility of mixed samples of HULIS and ammonium sulfate in different ratios in
284 order to get better understand organic-inorganic interactions under atmospherically relevant conditions.

285 Three samples were prepared by mixing HULIS (extracted from sample 1) and pure ammonium sulfate
286 (AS) with the mass ratios (HULIS to AS) of 0.25:0.75, 0.5:0.5 and 0.75:0.25 (actually 0.29:0.71,
287 0.55:0.45 and 0.79:0.21). As shown by Fig. 5, pure ammonium sulfate particles started to evaporate at
288 100°C, and were almost entirely evaporated at 180 °C, whereas HULIS aerosol started to evaporate at
289 the very beginning (about 20 °C) and more than 80% of its volume still remained at 180 °C. The

290 evaporation curves for the three mixed samples (Fig. 6) showed generally slow evaporation rates within
291 the temperature windows from 20 °C to 100 °C and from 180 °C to 280 °C, and much faster
292 evaporation rates between 100 °C and 180 °C. Interactions between HULIS and ammonium sulfate
293 obviously influenced the observed volatility. For example, the VFRs of 0.25:0.75 samples (Fig. 6a) at
294 the temperature of 180 °C were around 0.4 (varied from 0.397 to 0.428 for different size particles),
295 which is significantly higher than the calculated VFR ($0.29 \times 0.8 + 0.71 \times 0.06 = 0.275$) by assuming
296 HULIS and ammonium sulfate independently separated. This indicates that mixing of ammonium
297 sulfate to a HULIS solution decreases the volatility of the organic group or, alternatively, forms new
298 compounds of low volatility. For the 0.5:0.5 and 0.75:0.25 samples (Fig. 6b and 6c), the VFRs at
299 180 °C were around 0.43 (0.395 to 0.460 for different size particles) and 0.64 (0.595 to 0.655), which
300 are comparable to the calculated VFR (0.467 for the 0.5:0.5 samples and 0.645 for the 0.75:0.25
301 samples). These results indicate that the role of HULIS-AS interactions in the volatility of their
302 mixtures is complex and nonlinear.

303 In order to quantify the volatility changes of HULIS induced by its interaction with ammonium sulfate,
304 the kinetic mass transfer model was again applied to estimate the mass fractions of SVOC, LVOC and
305 ELVOC for the HULIS part in the mixed samples. As shown in Fig. 7, the model's performance in
306 simulating mixed HULIS-AS samples was fairly good, yet poorer than in simulating the "pure" HULIS
307 sample. The poorest agreement between the simulated and measured evaporation curves was found for
308 the 1:3 mixed samples (mass ratio of HULIS to AS), indicating relatively high uncertainties in the
309 calculated mass fractions of compounds with different volatility bins for this mixture. These visible
310 differences between modeled and measured results indicate that interactions between HULIS and AS
311 indeed influence their volatility distribution. As can be seen from Fig. 8, the estimated fraction of
312 ELVOC in the HULIS part of the 0:25:0.75 (Fig. 8b) and 0.75:0.25 (Fig. 8d) samples was much higher
313 than in the pure HULIS sample (Fig. 8a), while the ELVOC fraction in the 0.5:0.5 sample was
314 comparable to that in the pure HULIS sample. By taking 30 nm and 145 nm particles as an example,
315 the corresponding estimated ELVOC fractions were 0.66 and 0.26 in the pure HULIS sample, 1.0 and
316 0.93 in the 0.25:0.75 sample, 0.53 and 0.45 in the 0.5:0.5 sample, and 0.83 and 0.71 in the 0.75:0.25
317 sample, respectively. In spite of the possible overestimation of ELVOCs fraction in 1:3 mixed samples,
318 these results suggest that the interaction between HULIS and ammonium sulfate tend to decrease the
319 volatility of HULIS, and that this effect is nonlinear. [It should be emphasized here in case HULIS are](#)
320 [always mixed with ammonium sulfate, which accounted for 30% of the mass of PM_{2.5} \(Xie et al., 2015\),](#)

321 in ambient aerosols of YRD region, it is possible that these mixed samples are more representative of
322 the real volatility of HULIS in ambient aerosols.

323 4. Conclusion and implication

324 In this study, we analyzed the volatility of atmospheric HULIS extracted from four PM_{2.5} samples
325 collected at the SORPES station in the western YRD of eastern China, and investigated how the
326 interactions between HULIS and ammonium sulfate affected the volatility of HULIS aerosol fraction.
327 Overall, low volatilities and high oxidation states were identified for all the four samples, with VFRs at
328 280°C being higher than 55 % and O to C ratio being higher than 0.95 for all the re-generated HULIS
329 particles. A kinetic mass transfer model was deployed to divide the HULIS mixture into SVOC, LVOC
330 and ELVOC groups. We found that HULIS were dominated by LVOC and ELVOC (more than 80%)
331 compounds. Given the possible thermo-decomposition of large molecules during the heating program,
332 an even lower volatility than found here is possible for atmospheric HULIS in eastern China. The
333 Kelvin effect was supposedly taking place in atomizing the solutions of the HULIS mixtures, which
334 resulted in a size dependent distribution of the relative fractions of SVOC, LVOC and ELVOC in the
335 generated particles. The interaction between HULIS and ammonium sulfate was found to decrease the
336 volatility of the HULIS part in the mixed samples. However, these volatility changes were not linearly
337 correlated with the mass fractions of ammonium sulfate, indicating a complex interaction between the
338 HULIS mixture and inorganic salts.

339 This study demonstrates that HULIS are important low volatility and extremely low volatility
340 compounds in the aerosol phase, and sheds new light on the connection between atmospheric HULIS
341 and ELVOCs. In a view of the important sources of HULIS, multi-phase processes, including multi-
342 phase oxidation, oligomerization, polymerization and interaction with inorganic salts, have the
343 potential to lower the volatility of organic compounds in the aerosol phase, and to influence their gas-
344 aerosol partitioning. Multiphase processes could be one of the important reasons that most models tend
345 to underestimate the formation of SOA.

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Table 1 Kinetic model input settings for three-component HULIS aerosol.

Model input parameter	Unit	HULIS
Molar mass, M	g mol^{-1}	[280 280 280]
Density, ρ	kg m^{-3}	[1550 1550 1550]
Surface tension, σ	N m^{-1}	[0.05 0.05 0.05]
Diffusion coefficient, D	$10^{-6} \text{ m}^2 \text{ s}^{-1}$	[5 5 5]
Parameter for the calculation of T -dependence of D , μ	-	[1.75 1.75 1.75]
Saturation vapor pressure, p_{sat} (298 K)	Pa	[10^{-3} 10^{-6} 10^{-9}]
Saturation vapor concentration, c_{sat} (298 K)	$\mu\text{g m}^{-3}$	[10^2 10^{-1} 10^{-4}]
Enthalpy of vaporization, ΔH_{vap}	kJ mol^{-1}	[40 40 40]
Mass accommodation coefficient, α_m	-	[1 1 1]
Activity coefficient, γ	-	[1 1 1]
Particle initial diameter, d_p	nm	30, 60, 100, 145
Particle total mass, $m_{p,\text{tot}}$	$\mu\text{g m}^{-3}$	1
		Thermodenuder
Length of the flow tube	m	0.50 (i.d of 6 mm)
Residence time	s	5

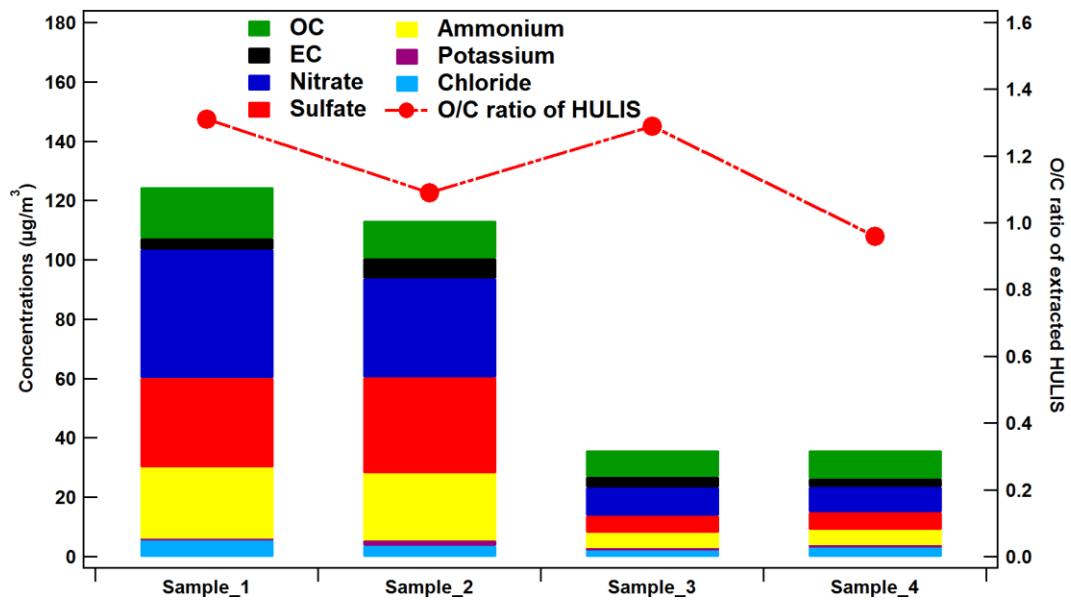


Figure 1 Chemical composition of the four PM_{2.5} samples collected at SORPES station and oxygen to carbon ratio of extracted HULIS from related samples

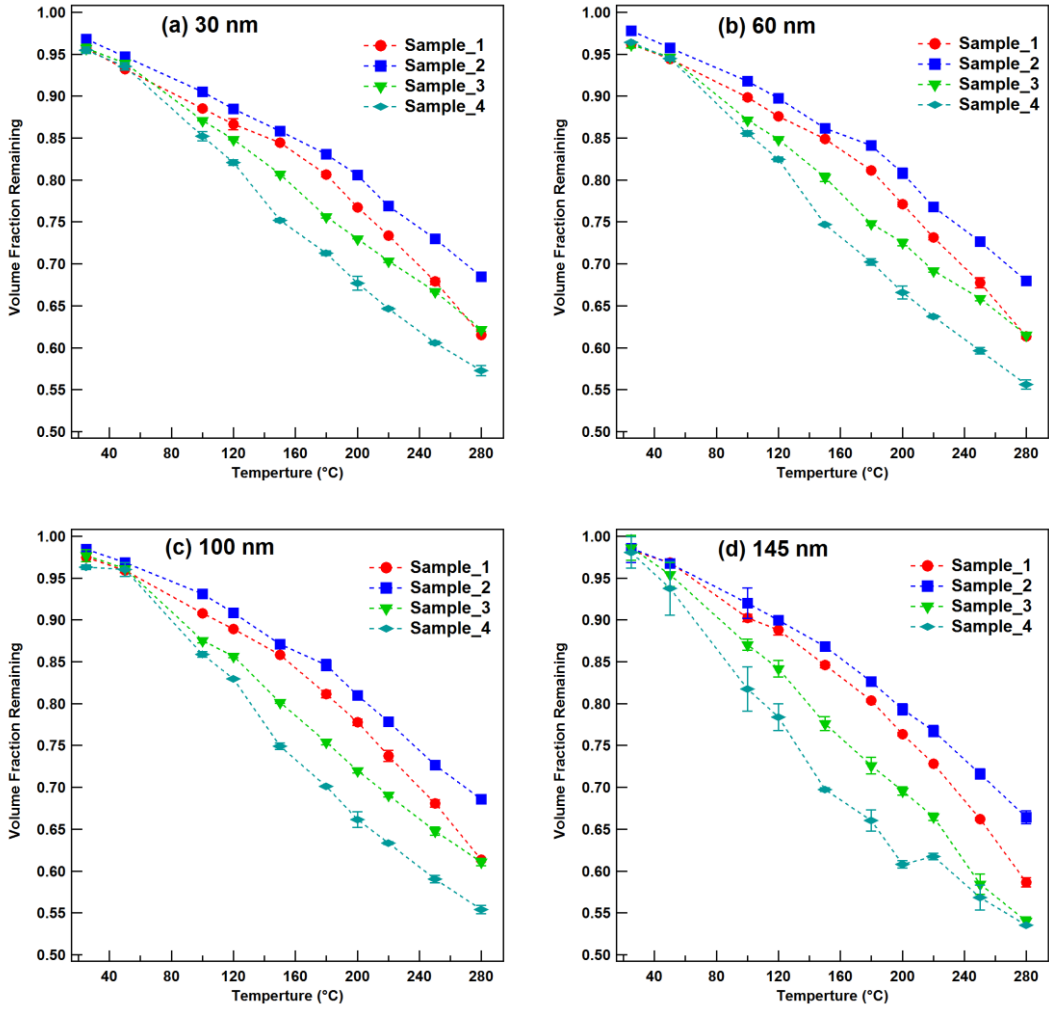


Figure 2 Volume fraction remaining (VFR) as a function of heating temperature for 4 samples at four different sizes of (a) 30 nm, (b) 60 nm, (c) 100 nm, and (d) 145 nm

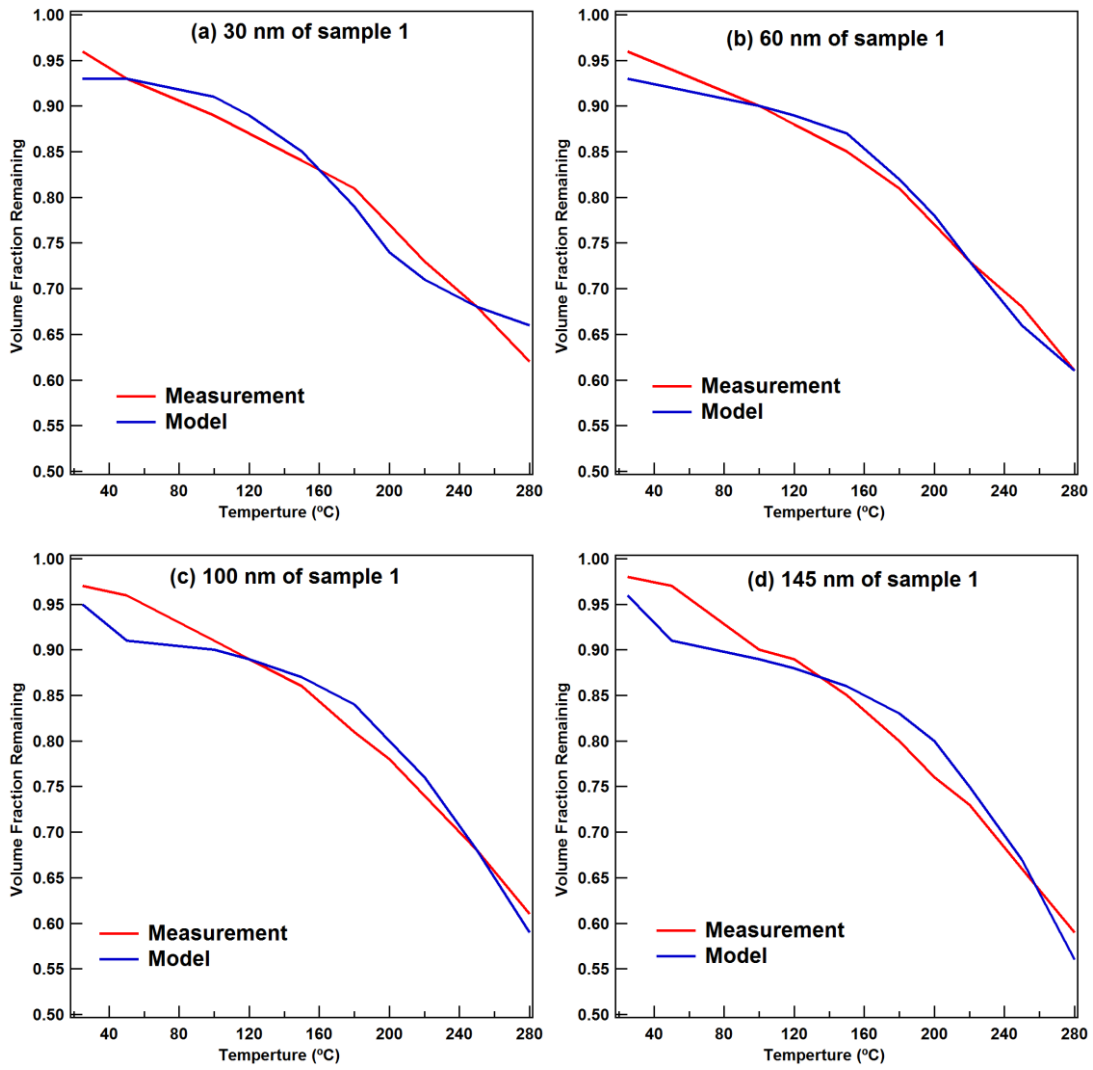


Figure 3 Measured and modeled volume fraction remaining (VFR) as a function of temperature for HULIS of sample 1 at four different particle sizes of (a) 30 nm, (b) 60 nm, (c) 100 nm and (d) 145 nm

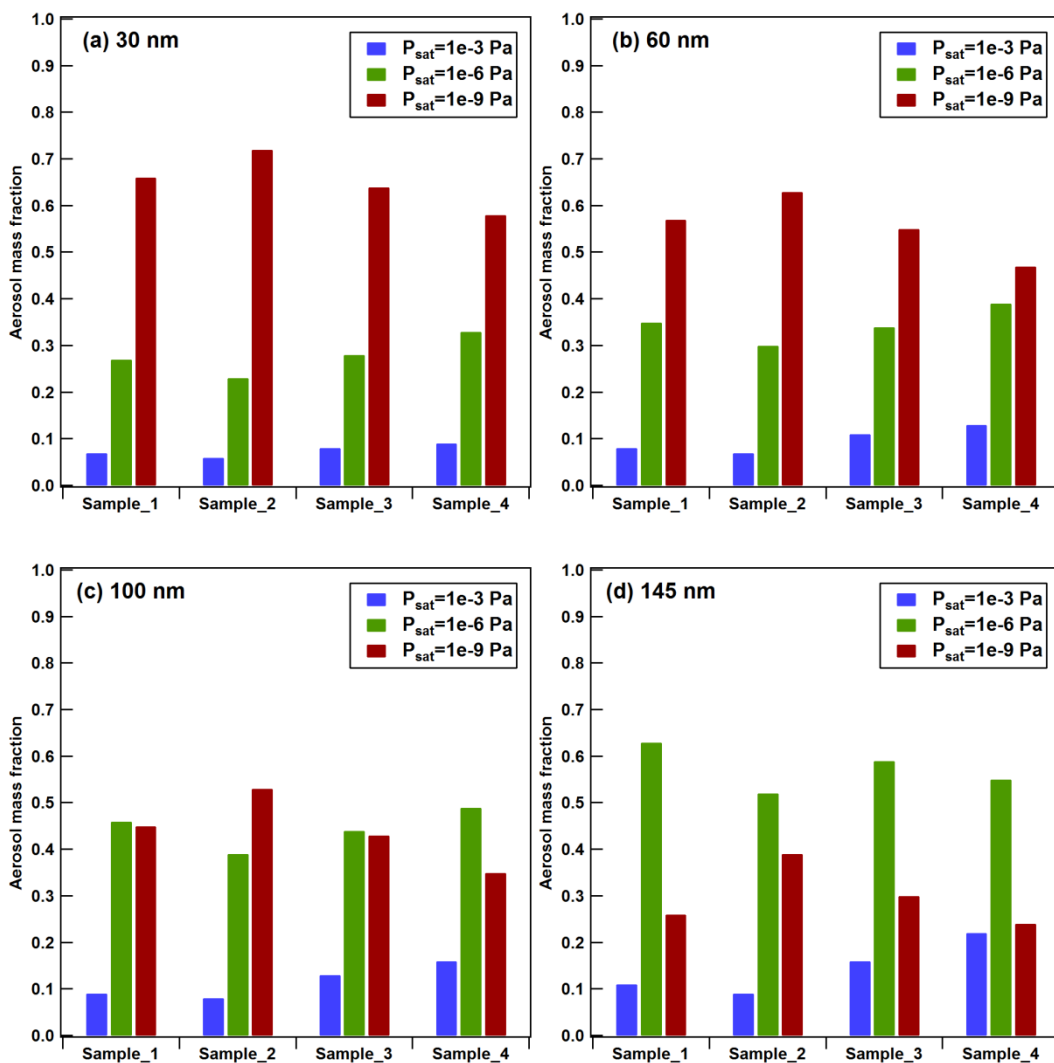


Figure 4 Mass fractions of compounds of SVOC ($p_{sat}=10^{-3}$ Pa), LVOC ($p_{sat}=10^{-6}$ Pa) and ELVOC, ($p_{sat}=10^{-9}$ Pa) in four aerosol samples with different particle sizes of (a) 30 nm, (b) 60 nm, (c) 100 nm, and (d) 145 nm

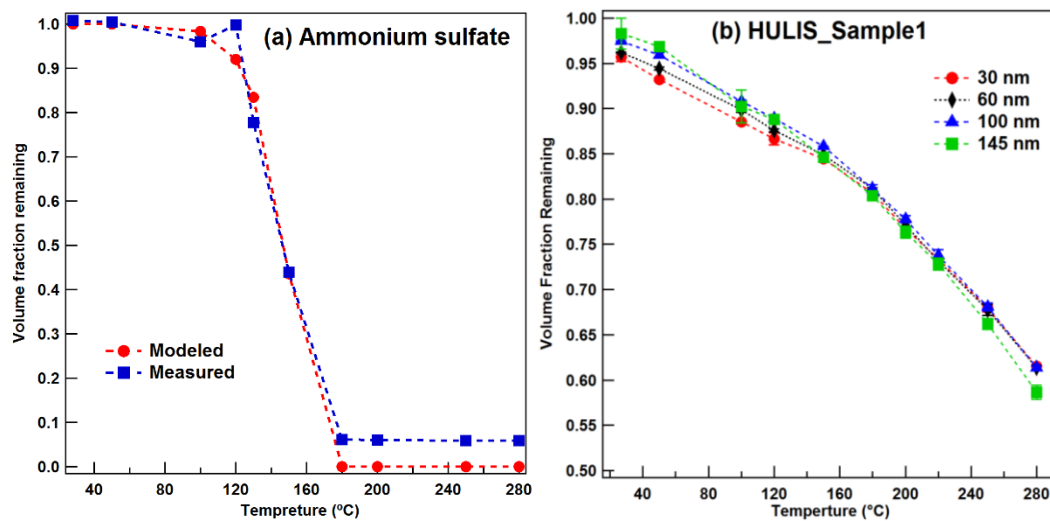


Figure 5 Volume fraction remaining (VFR) as a function of heating temperature for (a) measured and modeled pure ammonium sulfate particles at 100 nm, and (b) HULIS sample 1 at four different sizes of 30 nm, 60 nm, 100 nm, and 145 nm

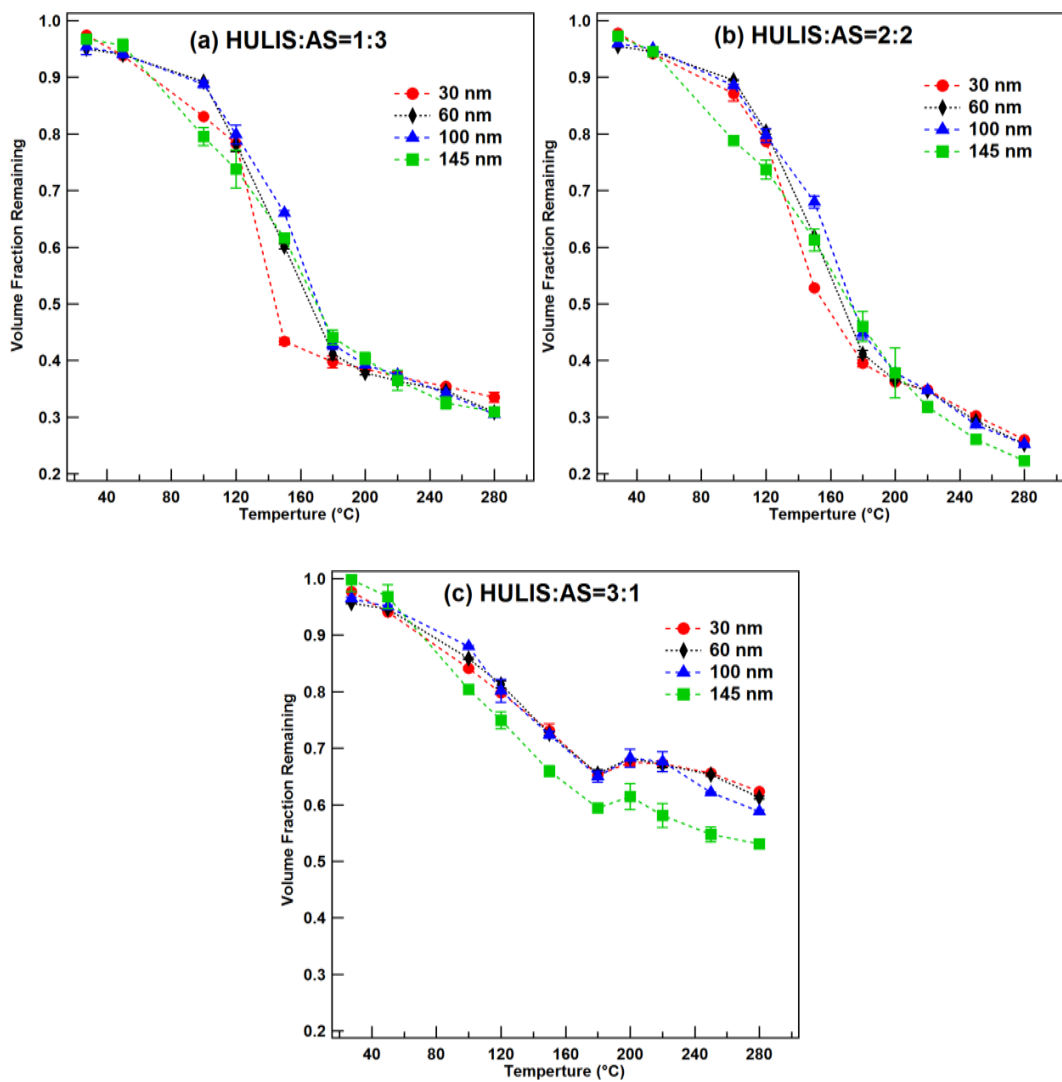


Figure 6 Volume fraction remaining (VFR) as a function of heating temperature for (a) 1:3 HULIS-AS mixed sample, (b) 2:2 HULIS-AS mixed samples, and (c) 3:1 HULIS-AS mixed samples at four different sizes of 30 nm, 60 nm, 100 nm, and 145 nm

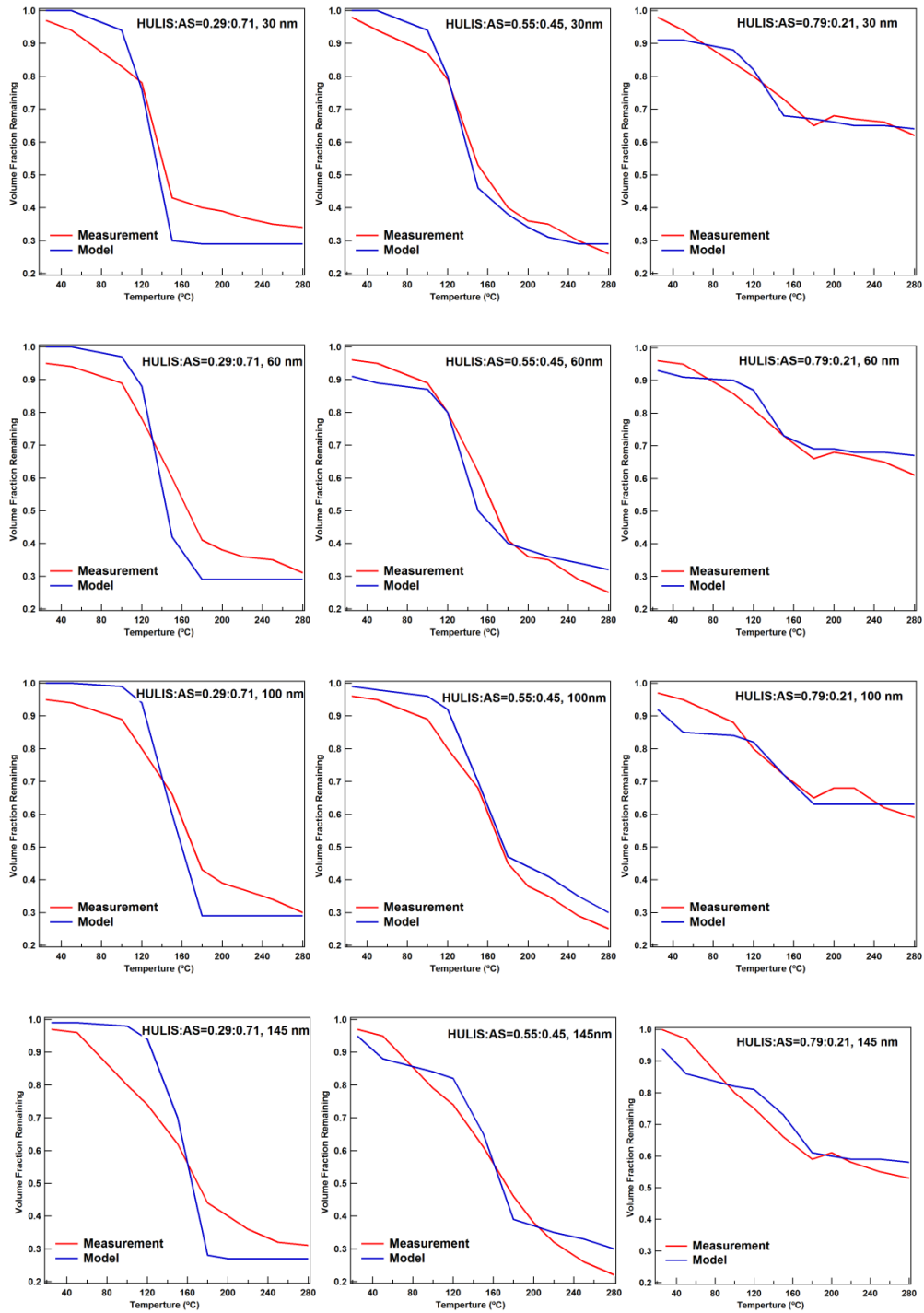


Figure 7 Measured and modeled volume fraction remaining (VFR) as a function of temperature for HULIS-AS mixed samples of 3 different mixing ratios at four different particle sizes of 30 nm, 60 nm, 100 nm and 145 nm

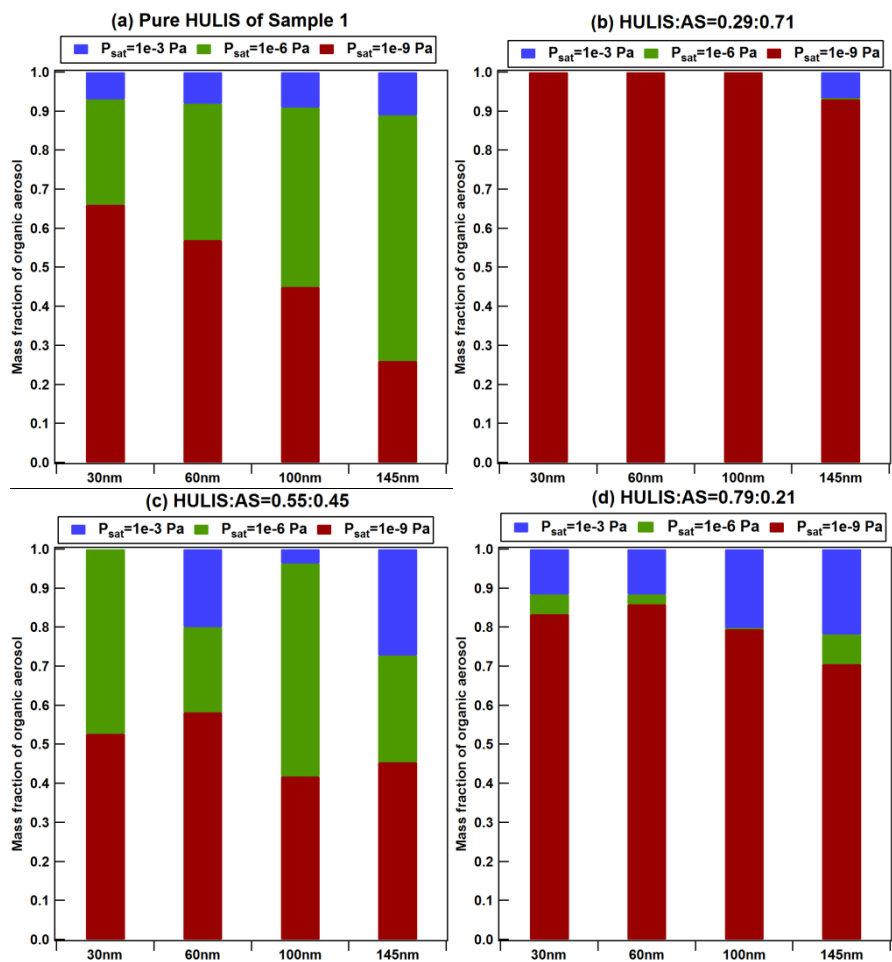


Figure 8 Model-derived mass fractions of organic compounds with different volatilities in four aerosol samples with different particle sizes of (a) 30 nm, (b) 60 nm, (c) 100 nm, and (d) 145 nm