



1 2	<u>Title:</u> Fluorescent Bioaerosol Particle, Molecular Tracer, and Fungal Spore Concentrations during Dry and Rainy Periods in a Semi-Arid Forest
3 4	
4 5	Authors: Marie Ila GOSSELIN ^{1,2} , Chathurika M Rathnayake ³ , Ian Crawford ⁴ , Christopher Pöhlker ² ,
6	Janine Fröhlich-Nowoisky ² , Beatrice Schmer ² , Viviane R. Després ⁵ , Guenter Engling ⁶ , Martin
7	Gallagher ⁴ , Elizabeth Stone ³ , Ulrich Pöschl ² , and J. Alex Huffman ^{1*}
8	
9	¹ Department of Chemistry and Biochemistry, University of Denver, Denver, Colorado, USA
10	² Max Planck Institute for Chemistry, Multiphase Chemistry and Biogeochemistry Departments, Mainz,
11	Germany
12	³ Department of Chemistry, University of Iowa, Iowa City, IA 52246, USA
13	⁴ Centre for Atmospheric Science, SEAES, University of Manchester, Manchester, UK
14	⁵ Institute of General Botany, Johannes Gutenberg University, Mainz, Germany
15	⁶ Division of Atmospheric Sciences, Desert Research Institute, Reno, NV, USA
16	
17	* Correspondence to: <u>alex.huffman@DU.edu</u>
18	





19 Abstract:

- 20 Bioaerosols pose risks to human health and agriculture and may influence the evolution of mixed-phase 21 clouds and the hydrological cycle on local and regional scales. The availability and reliability of methods 22 and data on the abundance and properties of atmospheric bioaerosols, however, are rather limited. Here 23 we analyze and compare data from different real-time Ultraviolet Laser/Light Induced Fluorescence (UV-24 LIF) instruments with results from a culture-based spore sampler and offline molecular tracers for 25 airborne fungal spores in a semi-arid forest in the Southern Rocky Mountains of Colorado. Commercial 26 UV-APS (Ultraviolet Aerodynamic Particle Sizer) and WIBS-3 (Wideband Integrated Bioaerosol Sensor, 27 Version 3) instruments with different excitation and emission wavelengths were utilized to measure 28 fluorescent aerosol particles (FAP) during both dry weather conditions and periods heavily influenced by 29 rain. Seven molecular tracers of bioaerosols were quantified by analysis of total suspended particle (TSP) high-volume filter samples using High Performance Anion Exchange Chromatography with Pulsed 30 31 Amperometric Detection (HPAEC-PAD). From the same measurement campaign Huffman et al. (2013) 32 previously reported dramatic increases in total and fluorescence particle concentrations during and 33 immediately after rainfall and also showed a strong relationship between the concentrations of FAP and 34 ice nuclei (Huffman et al., 2013; Prenni et al., 2013). Here we investigate molecular tracers and show that 35 during rainy periods the atmospheric concentrations of arabitol $(35.2 \pm 10.5 \text{ ng m}^{-3})$ and mannitol $(44.9 \pm 10.5 \text{ ng})$ 36 13.8 ng m⁻³) were 3-4 times higher than during dry periods. During and after rain the correlations between FAP and tracer mass concentrations were also significantly improved. Fungal spore number 37 38 concentrations on the order of 10⁴ m⁻³, accounting for 2-4% of TSP mass during dry periods and 17-23% during rainy periods, were obtained from scaling the tracer measurements and from multiple analysis 39 methods applied to the UV-LIF data. Endotoxin concentrations were also enhanced during rainy periods, 40 41 but showed no correlation with FAP concentrations. Average mass concentrations of erythritol, 42 levoglucosan, glucose, and $(1\rightarrow 3)$ - β -D-glucan in TSP samples are reported separately for dry and rainy 43 weather conditions. Overall, the results indicate that UV-LIF measurements can be used to infer fungal 44 spore concentrations, but substantial development of instrumental and data analysis methods seems 45 required for improved quantification.
- 45 required for improved quantification.





46 1. Introduction

47 Primary biological aerosols particles (PBAP) are of keen interest within the scientific community, 48 partially because methods for their quantification and characterization are advancing rapidly (Huffman and Santarpia, 2016; Sodeau and O'Connor, 2016). The term PBAP, or equivalently bioaerosol, generally 49 50 comprises several classes of airborne biological particles including viruses, bacteria, fungal spores, pollen and their fragments (Després et al., 2012; Fröhlich-Nowoisky, 2016). Fungal spores are of particular 51 atmospheric interest because they can cause a variety of deleterious health effects in humans, animals, 52 and agriculture, and it has been shown that they can represent a significant fraction of total organic 53 54 aerosol emissions (Deguillaume et al., 2008; Gilardoni et al., 2011; Madelin, 1994), especially in tropical 55 regions (Elbert et al., 2007; Huffman et al., 2012; Pöschl et al., 2010; Zhang et al., 2010). Current 56 estimates of the atmospheric concentration of fungal spores range from 10^0 to more than $10^4 \, \text{m}^{-3}$ 57 (Frankland and Gregory, 1973; Gregory and Sreeramulu, 1958; Heald and Spracklen, 2009; Hummel et 58 al., 2015; Sesartic and Dallafior, 2011). Fungal spores may also impact the hydrological cycle as giant cloud condensation nuclei (GCCN) or as ice nuclei (IN) (Haga et al., 2013; Morris et al., 2013; Sesartic et 59 al., 2013). Additionally, several classes of bioaerosols and their constituent components, such $(1 \rightarrow 3)$ - β -60 61 D-glucan and endotoxins, have been implicated in respiratory distress and allergies (Burger, 1990; Douwes et al., 2003; Laumbach and Kipen, 2005; Pöschl and Shiraiwa, 2015). For example, asthma and 62 63 allergies have shown notable increases during thunderstorms due to elevated bioaerosol concentrations (Taylor and Jonsson, 2004) especially when attributed to fungal spores (Allitt, 2000; Dales et al., 2003). 64

65

66 Molecular tracers have long been utilized as a means of aerosol source tracking (Schauer et al., 67 1996; Simoneit and Mazurek, 1989; Simoneit et al., 2004). In recent years, analysis of molecular tracers 68 has been utilized for the quantification of PBAP in atmospheric samples and has been compared with 69 results from microscopy (Bauer et al., 2008a) and culture samples (Chow et al., 2015b; Womiloju et al., 70 2003). Three organic molecules have been predominately utilized as unique tracers of fungal spores: ergosterol, mannitol, and arabitol. The majority of atmospherically relevant fungal spores are released by 71 72 active wet discharge processes common in Ascomycota and Basidiomycota, meaning that the fungal 73 organism actively ejects spores at a time most advantageous for the spore dispersal and germination 74 processes, often when relative humidity (RH) is high (Ingold, 1971). While there are several mechanisms 75 of active spore emission (e.g. Buller's drop (Buller, 1909) and osmotic pressure canons (Ingold, 1971)), 76 they each involve the secretion of fluid containing hygroscopic compounds, such as arabitol, mannitol, 77 potassium, chloride, and other solutes (Elbert et al., 2007), released near the site of spore growth. When 78 the spores are ejected, some of the fluid adheres to the spores and becomes aerosolized. Several of these 79 secreted compounds are thought to enter the atmosphere linked uniquely with spore emission processes, 80 and so these tracers have been used to estimate atmospheric concentrations of fungal spores. Arabitol and mannitol are both sugar alcohols (polyols) that serve as energy stores for the spore (Feofilova, 2001). 81 82 Arabitol is unique to fungal spores and lichen, while mannitol is present in fungal spores, lichen, algae, and higher plants (Lewis and Smith, 1967). Ergosterol is found within the cell membranes of fungal 83 spores (Weete, 1973) and can be used as an ambient fungal spore trace (Di Filippo et al., 2013; Miller and 84 85 Young, 1997). Comparing the seasonal trends of arabitol and mannitol with ergosterol, Burshtein et al. 86 (2011) showed positive correlations between arabitol or mannitol and ergosterol only in the spring and 87 autumn suggesting that the source of these polyols is unlikely to be solely fungal in origin or that the amount of each compound emitted varies considerably between species type and season. While ergosterol 88 89 has been directly linked to fungal spores in the air, ergosterol is prone to photochemical degradation and 90 is difficult to analyze and quantify directly. Quantification of ergosterol typically requires chemical 91 derivatization by silylation before analysis via gas chromatography (Axelsson et al., 1995; Burshtein et al., 2011; Lau et al., 2006). In contrast, analysis of sugar alcohols by ion chromatography involves fewer 92 93 steps and has been successfully applied to monitor seasonal variations of atmospheric aerosol concentration at a number of sites (Bauer et al., 2008a; Caseiro et al., 2007; Yang et al., 2012; Yttri et al., 94 95 2011a; Zhang et al., 2010; Zhang et al., 2015) including pg m⁻³ levels in the Antarctic (Barbaro et al., 96 2015). By measuring spore count and tracer concentration in parallel at one urban and two suburban sites





97 in Vienna, Austria, Bauer et al. (2008a) estimated the amount of each tracer per fungal spore emitted.
98 Potassium ions have also been linked to emission of biogenic aerosol (Pöhlker et al., 2012b) and are co-

99 emitted with fungal spores, however, application of potassium as a fungal tracer is uncommon because it

is predominantly associated with biomass burning (Andreae and Crutzen, 1997). Additionally, $(1\rightarrow 3)$ - β -

D-glucan (fungal spores and pollen) and endotoxins (gram-negative bacteria) have also been widely used
 to measure other bioaerosols (Andreae and Crutzen, 1997; Cheng et al., 2012; Stone and Clarke, 1992).

102 103

104 The direct detection of PBAP has historically been limited to analysis techniques that require 105 culturing or microscopy of the samples. These systems are time-consuming, costly, and often 106 substantially undercount biological particles by an order of magnitude or more (Gonçalves et al., 2010; 107 Pyrri and Kapsanaki-Gotsi, 2007). The sampling methods associated with these measurements also offer 108 relatively low time resolution and low particle size resolution. Recently, techniques utilizing ultraviolet 109 laser/light-induced fluorescence (UV-LIF) for the real-time detection of PBAP have been developed and 110 are being utilized by the atmospheric community for bioaerosol detection. Thus far, the most widely 111 applied LIF instruments for ambient PBAP detection have been the Ultraviolet Aerosol Particle Sizer (UV-APS; TSI Inc. Model 3314, St. Paul, MN) and the Wideband Integrated Bioaerosol Sensor (WIBS; 112 113 University of Hertfordshire, Hertfordshire, UK, now licensed to Droplet Measurement Technologies, 114 Boulder, CO, USA). Both of these commercially available instruments can provide information in real-115 time about particle size and fluorescence properties of supermicron atmospheric aerosols. 116 Characterization and co-deployment of these instruments over the past ten years has expanded the 117 knowledge base regarding how to analyze and utilize the information provided from these instruments 118 (Healy et al., 2014; Huffman et al., 2013; Pohlker et al., 2013; Pöhlker et al., 2012a), though the 119 interpretation of UV-LIF results from individual particles is complicated by interfering material that is not 120 biological in nature (Gabey et al., 2010; Huffman et al., 2012; Lee et al., 2010; Saari et al., 2013; Toprak 121 and Schnaiter, 2013).

122

123 Here we present analysis of atmospheric concentrations of arabitol and mannitol in relation to 124 results from real-time, ambient particle measurements reported by UV-APS and WIBS. We interrogate 125 these relationships as they pertain to rain conditions (rainfall and RH) that have previously been shown to 126 increase fluorescent aerosol concentration (Crawford et al., 2014; Huffman et al., 2013; Prenni et al., 127 2013; Schumacher et al., 2013). Active wet discharge of ascospores and basidiospores has frequently 128 been reported to correspond with increased RH (Elbert et al., 2007), and fungal spore concentration has 129 also been shown to increase after rain events (e.g. Jones and Harrison, 2004). Here we estimate airborne 130 fungal concentrations in a semi-arid forest environment utilizing a combination of real-time fluorescence 131 methods, molecular fungal tracer methods, and direct-to-agar sampling and culturing as parallel 132 surrogates for spore analysis. This study represents the first ambient comparison of real-time aerosol UV-133 LIF instruments with results from molecular tracers or culturing.

134135**2. Methods**

136

137 2.1 Sampling site

138 Atmospheric sampling was conducted as a part of the BEACHON-RoMBAS (Bio-hydro-139 atmosphere interactions of Energy, Aerosols, Carbon, H2O, Organics, and Nitrogen - Rocky Mountain 140 Biogenic Aerosol Study) field campaign conducted at the Manitou Experimental Forest Observatory (MEFO) located 48 km northwest of Colorado Springs, Colorado (2370 m elevation, 39° 06' 0" N, 105° 141 142 5' 03" W) (Ortega et al., 2014). The site is located in the central Rocky Mountains and is representative of 143 semi-arid montane pine forested regions of North America. During BEACHON-RoMBAS, a large team 144 of international researchers conducted an intensive set of measurements from 20 July to 23 August 2011. 145 A summary of results from the campaign are published in the BEACHON campaign special issue of





Atmospheric Chemistry and Physics¹. All the data used in this study were gathered from instruments and sensors located within a <100 m radius (e.g. Fig. 1).

148

149 **2.2 Online fluorescent instruments**

150 A UV-APS and WIBS-3 (Model 3; University of Hertfordshire) were operated continuously as a 151 part of the study, and particle data were integrated to five-minute averages before analysis. The UV-APS 152 was operated under procedures defined in previous studies (Huffman et al., 2013; Schumacher et al., 153 2013). A total suspended particle (TSP) inlet head ~5.5 m above ground, mounted above the roof of a 154 climate-controlled, metal trailer, was used to sample aerosol directed towards the UV-APS. Bends and horizontal stretches in the 0.75 inch tubing were minimized to reduce losses of large particles (Huffman et 155 156 al., 2013). The UV-APS detects particles between 0.5-20 µm and records aerodynamic particle diameter 157 and integrated total fluorescence (420-575 nm) after pulsed excitation by a 355 nm laser (Hairston et al., 158 1997). Both UV-APS and WIBS instruments report information about particle number concentration, but 159 it is instructive here to show results in particle mass for comparison between all techniques. Total particle 160 number size distributions (irrespective of fluorescence properties) obtained from the UV-APS were 161 converted to mass distributions using unit particle mass density of as a first approximation for all direct 162 comparisons with tracer mass and unless otherwise stated. Total particle concentration values (in μ g m⁻³) 163 were obtained for each five-minute period by integrating over the size range $0.5 - 15 \,\mu$ m, and these mass 164 concentration values were averaged over the length of the filter sampling periods. Uncertainty in mass 165 concentration values reported here is influenced by utilizing a single, estimated value for particle mass 166 density and because of slight dissimilarities between UV-APS and WIBS instruments in size binning at 167 particle sizes above 10 µm that dominate particle mass.

168

169 A WIBS-3 was used to continuously sample air at a site ~50 m away from the UV-APS trailer 170 and 1.3 m above the ground. Briefly, the diameter of individual particles sampled by the WIBS is 171 estimated by the intensity of the elastic side-scatter from a continuous wave 635 nm diode laser and analyzed by a Mie scattering model (Foot et al., 2008; Kaye et al., 2005). Particles that pass through the 172 173 diode laser activate two optically-filtered Xenon flash lamps. The first lamp excites the particle at 280 nm 174 and the second at 370 nm. Emission from the 280 nm excitation is filtered separately for two PMTs, one 175 which detects in a band at 310-400 nm and the other in a band at 410-650 nm. These excitation and 176 emission wavelengths result in a total of three channels of detection: λ_{ex} 280 nm, λ_{em} 320 – 400 nm (FL 1 177 or Channel A); λ_{ex} 280 nm, λ_{em} 410 – 650 nm (FL 2 or Channel B); and λ_{ex} 370 nm, λ_{em} 410 – 650 nm (FL 178 3 or Channel C) (Foot et al., 2008; Gabey et al., 2010; Perring et al., 2015). Individual particles are 179 considered fluorescent here if they exceed fluorescent thresholds for any channel, as defined as the 180 average of a "forced trigger" baseline plus 3 standard deviations (σ) of the baseline measurement (Gabey 181 et al., 2010).

182

183 WIBS particle-type analysis is utilized to define types of particles that have specific spectral 184 patterns. As defined by Perring et al. (2015), the 3 different fluorescent channels (FL1, FL2, and FL3) can 185 be combined to produce 7 unique fluorescent categories. Observed fluorescence in channel FL1 alone, but 186 without any detectable fluorescence in Channel FL2 or FL3, categorizes a particle as type A. Similarly, 187 observed fluorescence in channels FL2 or FL3, but in no other channels, places a particle in the B or C 188 categories, respectively. Combinations of fluorescence in these channels, such as a particle that exhibits 189 fluorescence in both FL1 and FL2 categorizes a particle as type AB and so on for a possible seven particle 190 types as summarized in Figure S1.

191

As a separate tool for particle categorization, the University of Manchester has recently
 developed and applied a hierarchical agglomerative cluster analysis tool for WIBS data, which they have

¹http://www.atmos-chem-phys.net/special_issue247.html





194 applied to the BEACHON-RoMBAS campaign (Crawford et al., 2014; Crawford et al., 2015; Robinson et 195 al., 2013). Here we utilize clusters derived from WIBS-3 data as described by Crawford et al. (2015). 196 Cluster data presented here was analyzed with the Open Source Python package FastCluster (Mullner, 197 2013). Briefly, hierarchical agglomerative cluster analysis was applied to the entire data set and each 198 fluorescent particle was uniquely clustered into one of 4 groups. Cluster 1, assigned by Crawford et al. 199 (2015) as fungal spores, displayed a 1.5-2 µm mode and a daily peak in the early morning that paralleled 200 relative humidity (Schumacher et al., 2013). Clusters 2, 3, and 4 have strong, positive correlations with 201 rainfall and exhibit size modes that peak at $<1.2 \mu m$ and were initially described by Crawford et al. as 202 bacterial particles. Here we have summed Clusters 2-4 to a single group referred to as Cl_{Bact}, for 203 simplicity when comparing with molecular tracers.

204

210

The WIBS-3 utilized here has since been updated to the WIBS-4 (Univ. Hertfordshire, UK) and WIBS-4A (Droplet Measurement Technologies, Boulder, Colorado). One important difference between the models is that the WIBS-3 exhibits comparatively weak FL1 and FL2 signals with respect to the more updated models, and is thus more influenced by FL3. This results in a different break-down of channel intensity between instrument models, as will be discussed later.

211 **2.3 High volume sampler**

212 Total suspended particle samples were collected for molecular tracer and molecular genetic analyses using a high volume sampler (Digitel DHA-80) drawing 1000 L min⁻¹ through 15 cm glass fiber 213 214 filters (Macherey-Nagel GmbH, Type MN 85/90, 406015, Düren, Germany) over a variety of sampling times ranging from 4-48 h (supplemental Table S1). The sampler was located <50 m from each of the 215 216 UV-LIF instruments described here, approximately between the WIBS-3 and UV-APS. Prior to sampling 217 all filters were baked at 500 °C for 12 h to remove DNA and organic contaminants. Samples were stored 218 in pre-baked aluminum bags after sampling at -20 °C for 1-30 days and then at -80 °C after overnight, 219 international transport cooled on dry ice. Due to the low vapor pressure of the molecular tracers analyzed loss due to volatilization is considered unlikely (Zhang et al., 2010). 36 samples were collected during the 220 221 study, in addition to handling field blanks and operational field blanks. Handling blanks were acquired by 222 placing a filter into the sampler and immediately removing, without turning on the air flow control. 223 Operational blanks were placed into the sampler and exposed to 10 seconds of air flow.

224

225 2.4 Slit Sampler

226 A direct-to-agar slit sampler (Microbiological Air Sampler STA-203, New Brunswick Scientific 227 Co, Inc., Edison, NJ) was used to collect culturable airborne fungal spores. The sampler was placed $\sim 2 \text{ m}$ 228 above ground on a wooden support surface with 5 cm x 5 cm holes to allow air flow both up and down 229 through the support structure. Sampled air was drawn over the 15 cm diameter sampling plate filled with 230 growth media at a flow rate of 28 L min⁻¹ for sampling periods of 20 to 40 min. Growth media (malt 231 extract medium) was mixed with antibacterial agents (40 units streptomycin, Sigma Aldrich; 20 units 232 ampicillin, Fisher Scientific) to suppress bacterial colony growth. Plates were prepared several weeks in 233 advance and stored in a refrigerator at ca. 4 °C until used for sampling. Before each sampling period, all 234 surfaces of the samplers were sterilized by wiping with isopropyl alcohol. Handling and operational 235 blanks were collected to verify that no fungal colonies were being introduced by handling procedures. 14 236 air samples were collected over 20 days and immediately moved to an incubator (Amerex Instruments, 237 Incumax IC150R) set at 25 °C for 3 days prior to counting fungal colonies formed. Each colony, present 238 as a growing dot on the agar surface, is assumed to have originated as one colony forming unit (CFU; i.e. 239 fungal spore) deposited onto the agar by impaction during sampling. The atmospheric concentration of 240 CFU per air volume was calculated using the sampler air flow. Further discussion of methods and initial 241 results from the slit sampler were published by Huffman et al. (2013).

- 242
- 243 2.5 Offline filter analyses





245 2.5.1 Carbohydrate analysis

246 Approximately 1/8 of each frozen filter was cut for carbohydrate analysis using a sterile 247 technique, meaning that scissors were cleaned and sterilized and cutting was performed in a positive-248 pressure laminar flow hood. In order to precisely determine the fractional area of the filter to be analyzed, 249 filters were imaged from a fixed distance above using a camera and compared to a whole, intact filter. 250 Using ImageJ software (Rasband and ImageJ, 1997), the area of each filter slice showing particulate 251 matter (PM) deposit was referenced to a whole filter, and thereby the amount of each filter utilized could 252 be determined. This technique allowed for an estimate of the fraction of each sampled used for the 253 analysis, which corresponds to the fraction of PM mass deposited. The uncertainty on the filter area 254 fraction is estimated at 2%. The uncertainty was determined as the percent of variation in the area of the 255 filter edge (no PM deposit) as compared to the total filter area.

256

257 Carbohydrates were extracted from quartz filter samples and analyzed following the procedure 258 described by Rathnayake et al. (2016). A total of 36 samples were analyzed along with field and lab 259 blanks. All lab and field blanks fell below method detection limits. Extraction was performed by placing the filter slice into a centrifuge tube that had been pre-rinsed with NanopureTM water (resistance > 18.2 M Ω cm⁻¹; Barnstead EasyPure II, 7401). A volume of 8.0 mL of NanopureTM water was added to the filter 260 261 262 in the centrifuge tube to extract water-soluble carbohydrates. Samples were then exposed to rotary 263 shaking for 10 min at 125 rpm, sonication for 30 min at 60 Hz (Branson 5510, Danbury, CT, US), and 264 rotary shaking for 10 min. After shaking, the extracted solutions were filtered through a 0.45 µm 265 polypropylene syringe filter (GE Healthcare, UK) to remove insoluble particles, including disintegrated 266 filter pieces. One 1.5 mL aliquot of each extracted solution was analyzed for carbohydrates within 24 267 hours of extraction. A duplicate 1.5 mL aliquot was stored in a freezer and analyzed, if necessary due to 268 lack of instrument response and invalid calibration check, within 7 days of extraction. Analysis of 269 carbohydrates was done using a High Performance Anion Exchange Chromatography System with Pulsed 270 Amperometric Detection (HPAEC-PAD, Dionex ICS 5000, Thermo Fisher, Sunnyvale, CA, USA). 271 Details of the instrument specifications and quality standards for carbohydrate determination are available 272 in Rathnayake et al. (2016). Calibration curves for mannitol, levoglucosan, glucose (Sigma-Aldrich), 273 arabitol and erythritol (Alfa Aesar) were generated with seven points each, ranging in aqueous 274 concentration from 0.005 ppm to 5 ppm. The method detection limits for mannitol, levoglucosan, glucose, 275 arabitol and erythritol were 2.3, 2.8, 1.6, 1.0 and 0.6 ppb, respectively. Method detection limits were 276 determined as 3σ of analyte concentrations recovered from seven spiked filter samples (Rathnayake et al., 277 2016). All calibration curves were checked daily using a standard solution to ensure all concentration 278 values were within 10% of the known value. Failure to maintain a valid curve resulted in recalibration of 279 the instrument.

281 2.5.2 DNA analysis

280

291

282 Methods and initial results from DNA analysis from these high volume filters were published by 283 Huffman et al. (2013). Briefly, fungal diversity was determined by previously optimized methods for 284 DNA extraction, amplification, and sequence analysis of the internal transcribed spacer regions of 285 ribosomal genes from the high volume filter samples (Fröhlich-Nowoisky et al., 2012; Fröhlich-286 Nowoisky et al., 2009). Upon sequence determination, fungal sequences were compared with known 287 sequences using the Basic Local Alignment Search Tool (BLAST) at the National Center for 288 Biotechnology (NCBI) and identified to the lowest taxonomic rank common to the top BLAST hits after 289 chimeric sequences had been removed. When sequences displayed >97% similarity, they were grouped 290 into operational taxonomic units (OTUs).

292 2.5.3 Endotoxin and glucan analysis

293 Sample preparation for quantification of endotoxin and $(1\rightarrow 3)$ -β-D-glucan included extraction of 294 5 punches (0.5 cm² each) of the quartz filters with 5.0 mL of pyrogen-free water (Associates of Cape Cod 295 Inc., East Falmouth, MA, USA), utilizing an orbital shaker (300 rpm) at room temperature for 60 min,





296 followed by centrifuging for 15 min (1000 rpm). One-half mL of supernatant was submitted to a kinetic 297 chromogenic limulus amebocyte lysate (Chromo-LAL) endotoxin assay (Associates of Cape Cod Inc.,

East Falmouth, MA, USA) using a ELx808IU (BioTek Instrument Inc., Winooski, VT, USA) incubating

absorbance microplate reader. For $(1\rightarrow 3)$ - β -D-glucan measurement, 0.5 mL of 3 N NaOH was added to

the remaining 4.5 mL of extract and the mixture was agitated for 60 min. Subsequently, the solution was

neutralized to pH 6–8 by addition of 0.75 mL of 2 N HCl. After centrifuging for 15 min $(1\rightarrow 3)$ - β -D-

302 glucan concentration was determined in the supernatant using the Glucatell® LAL kinetic assay

303 (Associates of Cape Cod, Inc., East Falmouth, MA). The minimum detection limits (MDLs) and

reproducibility were 0.046 Endotoxin Units (EU) m⁻³ and \pm 6.4% for endotoxin and 0.029 ng m⁻³ and \pm

305 4.2% for $(1\rightarrow 3)$ -β-D-glucan, respectively. Laboratory and field blank samples were analyzed as well, 306 with lab blank values being below detection limits, while field blank values were used to subtract

with lab blank values being below detection limits, while field blank values were used to subtract
background levels from sample data. More details about the bioassays can be found elsewhere (Chow et al., 2015a).

308 309

310 2.6 Meteorology and wetness sensors

Meteorological data were recorded by a variety of sensors located at the site. Precipitation was recorded 311 312 by a laser optical disdrometer (PARticle SIze and VELocity "PARSIVEL" sensor; OTT Hydromet GmbH, Kempton, Germany) and separately by a tipping bucket rain gauge. The disdrometer provides 313 314 precipitation occurrence, rate, and physical state (rain or hail) by measuring the magnitude and duration 315 of disruption to a continuous 780 nm laser that was located in a tree clearing (Fig. 1), while the tipping 316 bucket rain gauge measures a set amount of precipitation before tipping and triggering an electrical pulse. 317 A leaf wetness sensor (LWS; Decagon Devices, Inc., Pullman, WA), provided a measurement of 318 condensed moisture by measuring the voltage drop across a leaf surface to determine a proportional 319 amount of water on or near the sensor. Additional details of these measurements can be found in Huffman 320 et al. (2013) and Ortega et al. (2014).

321

322 3. Results and Discussion

323 324 3.1 Categorization and characteristic differences of Dry and Rainy periods

325 Increases in PBAP concentration have been frequently associated with rainfall (e.g. Bigg et al., 326 2015; Faulwetter, 1917; Hirst and Stedman, 1963; Jones and Harrison, 2004; Madden, 1997). Fungal 327 polyols have also been reported to increase after rain and have been used as indicators of increased fungal 328 spore release (Liang et al., 2013; Lin and Li, 2000; Zhu et al., 2015). Recently it was shown that the 329 concentration of fluorescent aerosol particles (FAP) measured during BEACHON-RoMBAS increased 330 dramatically during and after periods of rain (Crawford et al., 2014; Huffman et al., 2013; Schumacher et 331 al., 2013) and that these particle were associated with high concentrations of ice nucleating particles that 332 could influence the formation and evolution of mixed-phase clouds (Huffman et al., 2013; Prenni et al., 333 2013; Tobo et al., 2013). It was observed that a mode of smaller fluorescent particles (2-3 µm) appeared 334 during rain episodes, and several hours after rain ceased a second mode of slightly larger fluorescent 335 particle (4-6 µm) emerged, persisting for up to 12 h (Huffman et al., 2013). The first mode was 336 hypothesized to result from mechanical ejection of particles due to rain splash on soil and vegetated 337 surfaces, and the second mode was suggested as actively emitted fungal spores (Huffman et al., 2013). 338 While the UV-APS and WIBS each provide data at high enough time resolution to see subtle changes in 339 aerosol concentration, the temporal resolution of the chemical tracer analysis was limited to 4-48 h 340 periods defined by the collection time of the high volume sampler. To compare the measurement results 341 across the sampling platforms, UV-LIF measurements were averaged to the lower time resolution of the 342 filter sampler periods, and the periods were grouped into three broad categories: Rainy, Dry, and Other, as 343 will be defined below.

Time periods were wetness-categorized in two steps: first at 15 min resolution and then averaged
for each individual filter sample. During the first stage of categorization each 15 min period was
categorized into one of four groups: rain, post-rain, dry, or other. To categorize each filter period, an





347 algorithm was established utilizing UV-APS fluorescent particle fraction and accumulated rainfall. The 348 ratio of integrated number of fluorescent particles to total particles was used as a proxy for the increased 349 emission of biological particles. Figure 2a presents a time series of the size-resolved fluorescent particle 350 concentration, showing increases during rain periods in dark red. A relatively consistent diurnal cycle of 351 increased FAP concentration in the 2-4 µm range is apparent almost every afternoon, which corresponds 352 to near daily afternoon rainfall during approximately the first half of the measurement period. 353 Disdrometer and tipping bucket rainfall measurements were each normalized to unity and summed to 354 produce a more robust measure of rainfall rate, because it was observed that often only one of the two 355 systems would record a given light rain event. If a point was described by total rainfall accumulation 356 greater than 0.201 it was flagged as rain. A point was flagged as post-rain if it immediately followed a 357 rain period and also exhibited a fluorescent particle fraction greater than 0.08. The purpose of this 358 category was to reflect the observation that sustained, elevated concentrations of FAP persisted for many 359 hours even after the rain rate, RH, and leaf wetness returned to pre-rain values. The only measurement 360 that adequately reflected this scenario was of the fluorescent particles measured by UV-APS and WIBS 361 instruments. The post-rain flag was continued until the fluorescent particle fraction fell below 0.08 or if it 362 started to rain again (with calculated rain values greater than 0.201). Points were flagged as dry periods if 363 they exhibited rainfall accumulation and fluorescent particle fraction below the thresholds stated above. 364 Several periods were not easily categorized by this system and were considered in a fourth category as 365 other. This occurred when fluorescent particle fraction above the threshold value was observed with no 366 rainfall.

367

368 Once we these categories were assigned by the algorithm at 15 min resolution, each high volume 369 filter sample was categorized by a similar nomenclature, but using only three categories. These were 370 defined as Dry, Rainy (combination of rain and post rain categories), or Other based on the relative time 371 fraction in each of the four original 15 min categories. For each sample, if the relative time fraction of a 372 given category exceeded 0.50 the sample was assigned to that category. Despite the effort to categorize 373 samples systematically, several sample periods (5 of 35) appeared mis-categorized by looking at FAP 374 concentration, rainfall, RH, and leaf wetness in more detail. In some circumstances, this was because light 375 rainfall produced observable increases in FAP, but without exceeding the rainfall threshold. Or in other 376 circumstances a period of rainfall occurred at the very end or just before the beginning of a sample, and so 377 the many-hour period was heavily influenced by aerosol triggered by a period of rain just outside of the 378 sample time window. As a result, several samples were manually re-categorized as described here. 379 Samples 20 and 21 (Table S1) were four-hour samples that displayed high relative humidity and rainfall, 380 thus samples were originally characterized as Rainy. This period was described by an extremely heavy 381 rain downpour (7.5 mm in 15 min), however, that seemingly placed the samples in a different regime of 382 rain-aerosol dynamics than the other Rainy samples and so these two samples were moved to the Other 383 category. Sample 23, originally Rainy, presented a FAP fraction marginally above the 0.08 threshold, but 384 visually displayed a trend dissimilar to other post-rain periods and so was re-categorized as Dry. Sample 385 28 showed no obvious rainfall, but the measurement team observed persistent fog in three consecutive 386 mornings (Samples 25, 27, 28), and the concentration of fluorescent (2-6 µm) particles suggested a source 387 of particles not influenced by rain, and so this Rainy sample was re-categorized as Other. Sample 38 388 displayed a fluorescent number ratio just below the threshold value, and was thus categorized as Dry, 389 however, the measurement team observed post-rain periods at the beginning and end of the sample, so the 390 sample were re-categorized as Other. For all samples other than these five, the categorization was 391 determined using the majority (> 0.50) of the 15 min periods. In no cases other than the five that were re-392 categorized was the highest category fraction less than 0.50 of the sample time. Note that we have chosen 393 to capitalize Rainy, Dry, and Other to highlight that we have rigorously defined the period using the 394 characterization scheme described above and to separate the nomenclature from the general, colloquial 395 usage of the terms. Wetness category assignment for each high volume filter sample period is shown in 396 Figure 2 as a background color (brown for Dry samples, green for Rain-influenced samples, and pink for 397 Other samples) and Table S1.





398 399 To validate the qualitative differences between wetness categories described in the last section, 400 we present observations about each of these groupings. First, we organized the WIBS data according to 401 the particle categories introduced by Perring et al. (2015). By this method, every fluorescent particle 402 detected by the WIBS can be defined uniquely into one of seven categories (i.e. A, AB, ABC and so on). 403 By plotting the relative fraction of fluorescent particles described by each particle type, temporal 404 differences between measurement periods can be observed, as shown in Figure 2e. To a first 405 approximation, this analysis style allows for coarse discrimination of particle types. For example, a given 406 population of particles would ideally exhibit a consistent fraction of particles present in the different 407 particle categories as a function of time. By this reasoning, sample periods categorized as Dry (most of 408 the latter half of the study; brown bars in Fig. 2) would be expected to have a self-consistent particle type 409 trend, whereas sample periods categorized as Rainy (most of the first half of the study; green bars in Fig. 410 2) would have a self-consistent particle type trend, but different from the Dry samples. This is broadly true. During Rainy periods as seen in Figure 3a, there is a relatively high fraction (> 65%) of ABC 411 412 particles (light blue) and a relatively low fraction (< 15%) in BC (purple) and C (vellow) type particles, suggesting heavy influence from the FL1 channel. In contrast, during Dry periods the fraction of ABC 413 414 particles (light blue) is reduced (<25%) while BC (purple) and C (yellow) type particles increase in 415 relative fraction (>30% and >40%, respectively) suggested a diminished influence of FL1 channel.

416

417 It is important to note a few important caveats here. First, the ability of the WIBS to discriminate 418 finely between PBAP types is relatively poor and it is still unclear exactly how different particle types 419 would appear by this analysis method. Particles of different kinds and from different sources are likely 420 convolved into a single WIBS particle type, which could either soften or enhance the relationships with 421 rain discussed here. Second, the assignment of particle types is heavily size-dependent and sensitive to 422 subtle instrument parameters, and so it is unclear how different instruments would present similar particle 423 types. For example, Hernandez et al. (2016) used two WIBS instruments and found differences in relative 424 fraction of particle categories for samples aerosolized in the lab. They reported fungal spores to be 425 predominately A, AB, and ABC type particles, whereas Rainy sample periods suggested to have heavy 426 fungal spore influence by Huffman et al. (2013) show predominantly C, BC, and ABC particle fraction. 427 These discrepancies may be due to the comparison of ambient particles to laboratory-grown cultures. The 428 highly controlled environment of a laboratory may not always accurately represent the humidity 429 conditions in which fungal spore release occurs in this forest setting (Saari et al., 2015). This would 430 impact the fluorescence properties of the fungal spore particles which are inhibited by increased moisture 431 level around the spore (Hill et al., 2009). More likely, however, is that the WIBS-3 used here exhibits 432 higher sensitivity in the FL3 channel with respect to the FL1 and FL2 channels (Robinson et al., 2013), as 433 compared to the WIBS-4A used one of the units reported by Hernandez et al. (2016). This would explain 434 the shift here towards particles with C-type fluorescence. One piece of evidence for this is the quantitative 435 comparison of particle measurements presented by the UV-APS and WIBS-3 instruments co-deployed 436 here (Fig. 4). The number concentration of particle exhibiting fluorescence above the FL2 baseline of the 437 WIBS-3 is approximately consistent with the number of fluorescent particles measured by the UV-APS, 438 and significantly below the concentration of FL3 particles. The UV-APS number concentration shows the 439 highest correlation with the WIBS FL2 channel: during Rainy periods, $R^2 = 0.70$; Dry, $R^2 = 0.82$; Other, R^2 440 =0.92. These observations are in stark contrast to the trends reported by Healy et al. (2014) that the UV-441 APS fluorescent particle concentration correlated most strongly with the WIBS-4 FL3 and that the 442 number concentration of FL3 was the lowest out of all three channels. Given that the FL3 channel of the 443 WIBS and the UV-APS probe cover similar excitation and emission wavelengths it is expected that these 444 two channels should correlate well. Based on these data, we suggest that the WIBS-3 utilized here may 445 present a very different particle type break-down than if a WIBS-4 had been used. So, while caution is 446 recommended when comparing the relative break-down of WIBS particle categories shown here (Fig. 3) 447 with other studies, the data are internally self-consistent, and comparing qualitative differences between, 448 e.g. Rainy and Dry periods is expected to be robust. The main point to be highlighted here is that there is





indeed a qualitative difference in particles present in the three wetness categories, as averaged and shownin Figure 3a, which generally supports the effort to segregate these samples.

451

452 Further evidence that there is a qualitative difference in the three wetness categories is shown 453 using molecular genetic analysis (Figs. 3b, c). The analysis of fungal DNA sequences from 21 of the high 454 volume samples found 406 operational taxonomic units (OTUs), belonging to different fungal classes and 455 phyla. When organized by wetness type it was observed that 106 of these occurred only on Rainy 456 samples, 148 of these occurred on Dry samples, and 37 on Other samples, with some fraction occurring in 457 overlaps of each (Fig. 3c). This shows that the number of OTUs observed uniquely in either the Rainy or Dry periods is greater than the number of OTUs present in both wetness types, suggesting that the fungal 458 459 communities in each grouping are relatively distinct. Further, Figure 3b shows a break-down of fungal 460 taxonomic groupings for each wetness group. This analysis shows that there is a qualitative difference in 461 taxonomic break-down between periods of Rainy and Dry. Specifically, during Dry periods there is an increased fraction of Pucciniomycetes (green bar, Fig. 3c), Chytridiomycota (yellow), Sordariomyctes 462 463 (orange), and Eurotiomycetes (pink) when compared to the Rainy periods.

464

465 3.2 Atmospheric mass concentration of arabitol, mannitol, and fungal spores

To estimate fungal spore emission to the atmosphere, the concentration of arabitol and mannitol 466 467 (Fig. 5a, b, Table S2) in each aerosol sample was averaged for all samples in each of the three wetness 468 categories. The average TSP concentration of arabitol collected on Dry samples increased by a factor of 469 3.3 on Rainy samples (35.2 ± 10.5 ng m⁻³), and the average TSP mannitol concentration on Rainy samples 470 was higher by a factor of $3.7 (44.9 \pm 13.8 \text{ ng m}^{-3})$. Figures 5a, b show the concentration variability for 471 each wetness category, observed as the standard deviation from the distribution of individual samples. For 472 each polyol, there is no overlap in the ranges shown, including the outliers of the Rainy and Dry category, 473 suggesting a definitive and conceptually distinct separation between dry periods and those influenced by rain. The concentrations observed during Other periods is between those of the Dry and Rainy averages, 474 475 as expected, given the difficulty in confidently assigning these uniquely to one of these categories. The 476 observations here are roughly consistent with previous reports of polyol concentration, despite differences 477 in local fungal communities and concentrations. For example, Rathnayake et al. (2016) observed 30.2 ng 478 m⁻³ arabitol and 41.3 ng m⁻³ mannitol in PM₁₀ samples collected in rural Iowa, USA. In addition, Zhang et 479 al. (2015) reported arabitol and mannitol concentrations in PM_{10} samples of 44.0 and 71.0 ng m⁻³, 480 respectively, from a study in the mountains on Hainan Island off the coast of Southern China.

481

The square of the correlation coefficient (R^2) here between concentration values of arabitol and 482 483 mannitol during Rainy samples is very high (0.839; Table 1) suggesting that arabitol and mannitol originated from a primarily from the same source, likely active-discharge fungal spores. The correlation 484 485 is similar to the 0.87 R² reported by Bauer et al. (2008a) and the 0.93 R² reported by Graham et al. (2003). 486 In contrast, the same correlation between mannitol and arabitol concentrations, but for Dry samples is 487 relatively low (0.312). This is consistent with reports that arabitol can be used more specifically as a spore 488 tracer, but that mannitol has additional atmospheric sources besides fungal spores. The same correlation 489 was also performed between arabitol or mannitol and other molecular tracers (endotoxins and $(1 \rightarrow 3)$ - β -490 D-glucan), but all R^2 value were less than 0.43, suggesting that the endotoxins and glucans analyzed were 491 not emitted uniquely from the same sources as arabitol and mannitol.

492

Results from the two UV-LIF instruments were averaged over high volume sample periods, and a
correlation analysis was performed between tracer mass and fluorescent particle mass showing positive
correlations in all cases. The FAP mass from the UV-APS shows high correlation with the fungal polyols
during Rainy periods, with R² of 0.732 and 0.877 for arabitol and mannitol, respectively (Table 2; Figure
5c, d). The same tracers correlate poorly with the UV-LIF during Dry conditions. This is expected,
because polyols such as arabitol and mannitol are only found in *Ascomycota* and *Basidiomycota* fungal
spores which both utilize wet discharge methods for spore dispersal (Elbert et al., 2007; Feofilova, 2001;





Lewis and Smith, 1967). This high correlation suggests that the UV-APS does a good job of detecting
these wet-discharge spores, and corroborates previous statements that particles detected are often
predominately fungal spores (Healy et al., 2014; Huffman et al., 2013; Huffman et al., 2012). In contrast,
the low slope value and the poor correlation during Dry periods suggest that the UV-APS is also sensitive
to other kinds of particles, as designed. The small positive x-offset (FAP mass; Table S2, Figs. 5c,d)
during Rainy periods is likely due to particles that are too weakly fluorescent to be detected and counted
by the UV-APS, which is consistent with observations made in Brazil (Huffman et al., 2012).

507

508 Particle mass from WIBS Cl1, assigned to fungal spores (Crawford et al., 2015), also correlated 509 strongly with the same two molecular tracers. Both Rainy periods ($R^2 0.824$) and Dry periods ($R^2 0.764$) 510 correlate well with arabitol (Fig. 5e), while mannitol (Fig. 5f) only shows a strong correlation during the 511 Rainy periods (R² 0.799). Mannitol is a common polyol in higher plants while arabitol is only found in 512 fungal spores and lichen (Lewis and Smith, 1967). So the strong correlation of each polyol with UV-LIF 513 mass during Rainy periods when actively-discharged spores are expected to dominate and the similarly 514 strong correlations associated with arabitol suggests that the Cl1 cluster does a reasonably good job of 515 selecting fungal spore particles. The poor correlation between mannitol and Cl1 during dry periods 516 illustrates that the background mannitol concentration is likely not due to fungal spores alone, but has contribution from other higher plants that contain mannitol. Particle concentrations detected by individual 517 518 WIBS channels and in the other cluster were also compared with polyol concentrations, but each 519 correlation is relatively poor compared to that with respect to Cl1. As seen in Table 2 and Figures S2-S3, 520 correlations in FL1, 2, and 3 with arabitol are poor (<0.4) in the Dry category and good ($0.4 < R^2 < 0.7$) in 521 the Rainy category. For mannitol, all the UV-LIF instruments show high correlation (>0.7) in all cases. 522 This is likely due to mannitol being a non-specific tracer and suggests that the majority of UV-LIF 523 particles observed during all periods was dominated by PBAP.

524

525 **3.3 Estimated number concentration of fungal spore aerosol**

Bauer et al. (2008a) reported measurements of fungal spore number concentration in Vienna, 526 527 Austria using epifluorescence microscopy and also measured fungal tracer mass concentrations in order to 528 estimate the mass of arabitol (1.2 to 2.4 pg spore⁻¹) and mannitol (0.8 to 1.8 pg spore⁻¹) associated with 529 each emitted spore. Bauer et al. (2008a) and (Yttri et al., 2011b) reported ratios of mannitol to arabitol of 530 approximately 1.5 (\pm standard deviation of 26%) and 1.4 \pm 0.3, respectively. Our measurements show 531 slightly lower ratios of mannitol to arabitol, but that the ratio is dependent on wetness category; Rainy, 532 1.29 ± 0.17 ; Dry, 1.12 ± 0.23 ; and Other, 1.24 ± 0.54 . The mannitol to arabitol ratio would be expected to 533 vary as a function of fungal population present in the aerosol, whether between different wetness periods 534 at a given location or between different physical localities.

535

536 Using the approximate mid-point of the Bauer et al. (2008a) reported ranges, 1.7 pg mannitol per 537 spore and 1.2 pg arabitol per spore, atmospheric number concentrations of spores collected onto the high 538 volume filters were calculated from the polyol mass concentrations measured here. Based on these values, 539 and assuming all polyol mass originated with spore release, the mass concentration averages (Fig. 5) were 540 converted to fungal spore number concentrations (Fig. 6). The trends of spore concentration averages are 541 the same as with the polyol mass, because the numbers were each multiplied by the same scalar value. 542 After doing so, the analysis reveals an estimated spore concentration during Dry periods of 0.89 x 10^4 (± 0.21) spores m⁻³ using the arabitol concentration and $0.70 \times 10^4 (\pm 0.19)$ spores m⁻³ using the mannitol 543 544 concentration (Table 3). The estimated concentration of spores increased approximately three-fold during 545 Rainy periods to 2.9 x 10^4 (± 0.8) spores m⁻³ (arabitol estimate) and 2.6 x 10^4 (± 0.8) spores m⁻³ (mannitol 546 estimate) (Figure 6a, b). These estimates match well with estimates reported by Spracklen and Heald 547 (2014), who modeled the concentration of airborne fungal spores across the globe as an average of 2.5 x 10⁴ spores m⁻³, with approximately 0.5 x 10⁴ spores m⁻³ over Colorado. 548 549





550 The UV-LIF instruments discussed here are fundamentally number-counting techniques and can 551 be utilized here roughly as spore counters. As a first approximation, each particle detected by the UV-552 APS was assumed to be a fungal spore with the same properties used in the assumptions by Bauer et al. 553 (2008a). Plotting the correlation of fungal spore number concentration from polyol mass concentration 554 with respect to the fungal spore concentration assumed from the UV-LIF measurements shows 555 correlations in Figures 6c-f. The first, and most important observation is that the estimated fungal spore 556 concentration from each technique is on the same order of magnitude, 10⁴ m⁻³. Looking at individual correlations reveals a finer layer of details. These results show that the number concentration of fungal 557 558 spores estimated by the UV-APS is greater than the number of fungal spores estimated by the tracers, as evidenced by slope values of approximately 0.2 and 0.35 for Rainy and Dry conditions, respectively 559 560 (Figure 6c, d). The R² values (~0.5) during Rainy periods indicate that the additional source of particles 561 detected by the UV-APS is likely to have a similar source, such as PBAP mechanically ejected from soil 562 and vegetative surfaces with rain-splash (Huffman et al., 2013). The magnitude of the over-estimation is 563 higher during Dry periods, which would be expected if Rainy periods exhibited much higher particle 564 number fractions associated with polyol-containing spores.

565

566 The Cl1 cluster from WIBS data shows correlations with estimated fungal spores from arabitol 567 and mannitol that have slope much closer to 1.0 than correlations with UV-APS number (Figure 6e, f, 568 Table S3). For example, the slope of the Cl1 correlations with each polyol during Rainy periods is 569 approximately 0.87. This suggests only a 13% difference between the spore concentration estimates from 570 the two techniques during Rainy periods. The average number concentration of Cl1 during Rainy periods 571 is $1.6 \times 10^4 (\pm 0.8)$ spores m⁻³. In both cases the slopes with respect to Cl1 is greater than 1.0 during Dry 572 periods, suggesting that the cluster method may be missing some fraction of weakly fluorescent particles. 573 Huffman et al. (2012) similarly suggests that that particles that are weakly fluorescent may be below the 574 detection limit of the instrument, and Healy et al. (2014) suggested that both UV-APS and WIBS-4 575 instruments significantly under-count the ubiquitous *Cladosporium* spores that are most common during 576 dry weather and often peak in the afternoon when RH is low (De Groot, 1968; Oliveira et al., 2009). 577 Fundamentally, however, the results from the UV-APS, and even more so the numbers reported by the 578 clustering analysis by Crawford et al. (2015), reveal broadly similar trends with the numbers estimated 579 from polyol-to-spore values reported by Bauer et al. (2008a).

580

581 The fungal culture samples show similar division during Rainy and Dry periods as arabitol and 582 mannitol concentrations (Figure 6c), with an increase of approx. 1.6 during Rainy periods. The trend of a positive slope with respect to the UV-LIF measurements is also similar between the tracer and culturing 583 584 methods. In general, however, the R² value correlating CFU to fungal spore number calculated from UV-585 LIF number is lower than between tracers and UV-LIF numbers (Tables 2, S4). This is not unexpected for 586 several reasons. First, the short sampling time of the culture samples (20 min) leads to poor counting statistics and high number concentration variability, whereas each data point from the high volume air 587 588 samples represents a period of 4 - 48 hours. Second, culture samplers, by their nature, only account for 589 cultural fungal spores. It has been estimated that as low as 17% of aerosolized fungal spores are 590 culturable, and so it is expected that the CFU concentration observed is significantly less than the total 591 airborne concentration of spores (Bridge and Spooner, 2001; Després et al., 2012). Nonetheless, the 592 culturing analysis here supports the tracer and UV-LIF analyses and the most important trends are 593 consistent between all analysis methods. The concentration of fungal spores is higher during the Rainy 594 periods, and there is a positive correlation between both tracer and CFU concentration and UV-LIF 595 number.

596

597 In pristine environment, such as the Amazon, supermicron particle mass has been found to consist 598 of up to 85% biological material (Pöschl et al., 2010). Total particulate matter mass was calculated here 599 from the UV-APS number concentrations (m^{-3}) and converted to mass for particles of aerodynamic 600 diameter 0.5 – 15 µm. In only this case a density of 1.5 g cm⁻³ was utilized to calculate a first





601 approximation of total particle mass to which all other mass measurements were compared. An average 602 TSP mass density of 1.5 g cm⁻³ was utilized, because organic aerosol is typically estimated with density <603 1.0 g cm⁻³, biological particles are often assumed to have ca. 1.0 g cm⁻³ density, and mineral dust particles have densities of up to ca. 3.5 g cm⁻³ (Dexter, 2004; Tegen and Fung, 1994). Fungal spore mass 604 605 was estimated here using the fungal spore concentrations calculated from arabitol and mannitol mass (Fig. 606 6) and then using an estimated 33 pg reported by Bauer et al. (2008b) as an average mass per spore. 607 Dividing the resultant fungal spore mass by total particulate mass provides a relative mass fraction for 608 each high volume sample period. These calculations suggest that fungal spores represent approximately 609 $23\% \pm 9$ (using arabitol) or $21\% \pm 8$ (using mannitol) of total particulate mass during Rainy periods (Table 3, Figure 7). This represents a nearly 6 fold increase in percentage compared to Dry periods (4.8% 610 611 \pm 1.4 and 3.7% \pm 1.1, respectively). A similar increase during Rainy periods was also seen in the mass 612 fraction of fungal cluster Cl1, which represented $17\% \pm 10$ of the particle mass during Rainy and $2\% \pm 1$ 613 during Dry periods (Table S4).

614

632

615 **3.5 Variations in endotoxin and glucan concentrations**

Endotoxins are components of gram-negative bacteria (Andreae and Crutzen, 1997). Here, we 616 617 show correlations between total endotoxin mass and WIBS Cl_{Bact}, which were assigned by Crawford et al. (2015) to be bacteria due to the small particle size ($< 1 \mu m$) and high correlation with rain. These cluster 618 619 assignments are quite uncertain, however, and should be treated loosely. The correlation between 620 endotoxin mass and UV-APS and the WIBS clusters was very poor, in most cases $R^2 < 0.1$ (Table 2, 621 Figure 8), suggesting no apparent relationship. Analysis of bacteria by both UV-LIF techniques is 622 hampered by the fact that bacteria can be $< 1 \mu m$ in size and because both instruments detect particles 623 with decreased efficiency at sizes below 0.8 µm. So weak correlations may not have been apparent due to 624 reduced overlap in particle size. Despite the lack of apparent correlation between the techniques, the 625 relatively variable endotoxin concentrations were elevated during Rainy periods, consistent with Jones 626 and Harrison (2004), who showed that bacteria concentration were elevated after rainy periods.

627 Glucans, such as $(1\rightarrow 3)$ -β-D-glucan, are components of the cell walls of pollen, fungal spores, 628 plant detritus, and bacteria (Chow et al., 2015b; Lee et al., 2006; Stone and Clarke, 1992). In contrast to 629 the observed difference in endotoxin concentration during the different wetness periods, however, $(1\rightarrow 3)$ -630 β-D-glucan showed no correlations with UV-LIF concentrations (Table 2) and no differentiation during 631 the different wetness periods.

633 4. Conclusions

634 Increased concentrations of fluorescent aerosol particles and ice nuclei attributed to having 635 biological origin were observed during and immediately after rain events throughout the BEACHON-636 RoMBAS study in 2011 (Huffman et al., 2013; Prenni et al., 2013; Schumacher et al., 2013). Here we 637 expand upon the previous reports by utilizing measurements from two commercially available UV-LIF 638 instruments, of several molecular tracers extracted from high volume filter samples, and from a culture-639 based sampler in order to compare three very different methods of atmospheric fungal spore analysis. 640 This study represents the first reported correlation of UV-LIF and molecular tracer measurements and 641 provided an opportunity to understand how an important class of PBAP might be influenced by periods of rainy and dry weather. We found clear patterns in the fungal molecular tracers, arabitol and mannitol, 642 643 associated with Rainy conditions that are consistent with previous findings (Bauer et al., 2008a; Elbert et 644 al., 2007; Feofilova, 2001). Fungal polyols increased 3-fold over Dry conditions during Rainy weather 645 samples, with arabitol concentration of 35.2 ± 10.5 ng m⁻³ and mannitol concentration of 44.9 ± 13.8 ng 646 m^{-3} . Additionally, the very high correlation of the fungal tracers with WIBS Cl1 ($R^2 > 0.8$ in many cases) 647 provides support for its assignment by Crawford et al. (2015) to fungal spores. Similarly, the UV-APS 648 correlates well with fungal tracers, however over-counts the number concentration estimated from the 649 tracers, confirming that the UV-APS is sensitive also to other types of particles beyond fungal spores, as 650 expected. The estimated spore count from the WIBS Cl1 concentration was within ~13% of the spore 651 count estimated by the tracer method, with concentrations ranging from $1.6 - 2.9 \times 10^4$ spores m⁻³. These





652 values are broadly consistent with concentrations modeled by, e.g. Spracklen and Heald (2014), Hoose et 653 al. (2010), and Hummel et al. (2015). These spore counts represent 17-23% of the total particle mass 654 during Rainy conditions and 2-4% during Dry conditions. Culture-based sampling also shows a similar 655 relationship between CFU and UV-LIF concentrations and an increase of ~1.6 between Dry and Rainy conditions. Despite the fact that the tracer and UV-LIF approaches to estimating atmospheric fungal spore 656 concentration are fundamentally different, they provide remarkably similar estimates and temporal trends. 657 With further improvements in instrumentation and analysis methods (e.g. advanced clustering algorithms 658 659 applied to UV-LIF data), the ability to reliably discriminate between PBAP types is improving. As we 660 have shown here, this technology represents a potential for monitoring approximate fungal spore mass and for contributing improved information on fungal spore concentration to global and regional models 661 662 that to this point has been lacking (Spracklen and Heald, 2014).

663

664

665 5. Acknowledgements

666 The BEACHON-RoMBAS campaign was partially supported by an ETBC (Emerging Topics in 667 Biogeochemical Cycles) grant to the National Center for Atmospheric Research (NCAR), the University 668 of Colorado, Colorado State University, and Penn State University (NSF ATM-0919189). The authors wish to thank Jose Jimenez, Douglas Day (Univ. Colorado-Boulder); Anthony Prenni, Paul DeMott, 669 670 Sonia Kreidenweis, and Jessica Prenni (Colorado St. Univ.); Alex Guenther, and Jim Smith (NCAR) for 671 BEACHON-RoMBAS project organization and logistical support and the USFS, NCAR, and Richard 672 Oakes for access to the Manitou Experimental Forest Observatory field site. Measurements of 673 temperature, relative humidity, wind speed, and wind direction were provided by Andrew Turnipseed 674 (NCAR) and leaf wetness and disdrometer data were provided by Dave Gochis (NCAR). Marie I. 675 Gosselin thanks the Max Planck Society for financial support. J. Alex Huffman thanks the University of 676 Denver for intramural funding for faculty support. The Mainz team acknowledges financial support from the Max Planck Society (MPG), the Max Planck Graduate Center with the Johannes Gutenberg 677 678 University Mainz (MPGC), the Geocycles Cluster Mainz (LEC Rheinland-Pfalz), and the German 679 Research Foundation (DFG PO1013/5-1 and FR3641/1-2, FOR 1525 INUIT). The Manchester team 680 acknowledges funding from the UK NERC (UK-BEACHON, Grant # NE/H019049/1) to participate in 681 the BEACHON experiment, and development support of the WIBS instruments. Manchester would also 682 like to thank Prof. Paul Kaye, the developer of the WIBS instruments and his team at the University of 683 Hertfordshire, for their technical support. The authors thank Cristina Ruzene, Isabell Müller-Germann, 684 Petya Yordanova, Tobias Könemann (Max Planck Inst. For Chem.), and Nicole Savage (Univ. Denver)

685 for technical assistance.





686 6. References

- 687
- Allitt, U.: Airborne fungal spores and the thunderstorm of 24 June 1994, Aerobiologia, 16, 397-406,
 2000.
- 690 Andreae, M. O. and Crutzen, P. J.: Atmospheric Aerosols: Biogeochemical Sources and Role in
- 691 Atmospheric Chemistry, Science, 276, 1052-1058, 1997.
- 692 Axelsson, B.-O., Saraf, A., and Larsson, L.: Determination of ergosterol in organic dust by gas
- 693 chromatography-mass spectrometry, Journal of Chromatography B: Biomedical Sciences and694 Applications, 666, 77-84, 1995.
- Barbaro, E., Kirchgeorg, T., Zangrando, R., Vecchiato, M., Piazza, R., Barbante, C., and Gambaro, A.:
 Sugars in Antarctic aerosol, Atmos Environ, 118, 135-144, 2015.
- Bauer, H., Claeys, M., Vermeylen, R., Schueller, E., Weinke, G., Berger, A., and Puxbaum, H.: Arabitol
 and mannitol as tracers for the quantification of airborne fungal spores, Atmos Environ, 42, 588-593,
 2008a.
- 700 Bauer, H., Schueller, E., Weinke, G., Berger, A., Hitzenberger, R., Marr, I. L., and Puxbaum, H.:
- Significant contributions of fungal spores to the organic carbon and to the aerosol mass balance of the
 urban atmospheric aerosol, Atmos Environ, 42, 5542-5549, 2008b.
- Bigg, E. K., Soubeyrand, S., and Morris, C. E.: Persistent after-effects of heavy rain on concentrations of
 ice nuclei and rainfall suggest a biological cause, Atmos Chem Phys, 15, 2313-2326, 2015.
- 705 Bridge, P. and Spooner, B.: Soil fungi: diversity and detection, Plant and soil, 232, 147-154, 2001.
- 706 Buller, A.: Spore deposits—the number of spores, Researches on fungi, 1, 79-88, 1909.
- 707 Burger, H.: Official Publication of American Academy of Allergy and ImmunologyBioaerosols:
- Prevalence and health effects in the indoor environment, Journal of Allergy and Clinical Immunology, 86,687-701, 1990.
- Burshtein, N., Lang-Yona, N., and Rudich, Y.: Ergosterol, arabitol and mannitol as tracers for biogenic
 aerosols in the eastern Mediterranean, Atmos Chem Phys, 11, 829-839, 2011.
- 712 Caseiro, A., Marr, I. L., Claeys, M., Kasper-Giebl, A., Puxbaum, H., and Pio, C. A.: Determination of
- saccharides in atmospheric aerosol using anion-exchange high-performance liquid chromatography and
- pulsed-amperometric detection, Journal of Chromatography A, 1171, 37-45, 2007.
- Cheng, J. Y. W., Hui, E. L. C., and Lau, A. P. S.: Bioactive and total endotoxins in atmospheric aerosols
 in the Pearl River Delta region, China, Atmos Environ, 47, 3-11, 2012.
- Chow, J. C., Lowenthal, D. H., Chen, L.-W. A., Wang, X., and Watson, J. G.: Mass reconstruction
 methods for PM2. 5: a review, Air Quality, Atmosphere & Health, 8, 243-263, 2015a.
- 719 Chow, J. C., Yang, X., Wang, X., Kohl, S. D., Hurbain, P. R., Chen, L. A., and Watson, J. G.:
- 720 Characterization of Ambient PM10 Bioaerosols in a California Agricultural Town, Aerosol Air Qual Res,
- 721 15, 1433-1447, 2015b.





- 722 Crawford, I., Robinson, N. H., Flynn, M. J., Foot, V. E., Gallagher, M. W., Huffman, J. A., Stanley, W.
- 723 R., and Kaye, P. H.: Characterisation of bioaerosol emissions from a Colorado pine forest: results from
- the BEACHON-RoMBAS experiment, Atmos Chem Phys, 14, 8559-8578, 2014.
- 725 Crawford, I., Ruske, S., Topping, D., and Gallagher, M.: Evaluation of hierarchical agglomerative cluster
- 726 analysis methods for discrimination of primary biological aerosol, Atmos Meas Tech, 8, 4979-4991,
 727 2015.
- Dales, R. E., Cakmak, S., Judek, S., Dann, T., Coates, F., Brook, J. R., and Burnett, R. T.: The role of
 fungal spores in thunderstorm asthma, Chest, 123, 745-750, 2003.
- De Groot, R.: Diurnal cycles of air-borne spores produced by forest fungi, Phytopathology, 58, 12231229, 1968.
- 732 Deguillaume, L., Leriche, M., Amato, P., Ariya, P. A., Delort, A. M., Pöschl, U., Chaumerliac, N., Bauer,
- H., Flossmann, A. I., and Morris, C. E.: Microbiology and atmospheric processes: chemical interactions
- of primary biological aerosols, Biogeosciences, 5, 1073-1084, 2008.
- 735 Després, V. R., Huffman, J. A., Burrows, S. M., Hoose, C., Safatov, A. S., Buryak, G., Frohlich-
- Nowoisky, J., Elbert, W., Andreae, M. O., Poschl, U., and Jaenicke, R.: Primary biological aerosol
 particles in the atmosphere: a review, Tellus B, 64, 58, 2012.
- Dexter, A.: Soil physical quality: Part I. Theory, effects of soil texture, density, and organic matter, and
 effects on root growth, Geoderma, 120, 201-214, 2004.
- Di Filippo, P., Pomata, D., Riccardi, C., Buiarelli, F., and Perrino, C.: Fungal contribution to sizesegregated aerosol measured through biomarkers, Atmos Environ, 64, 132-140, 2013.
- Douwes, J., Thorne, P., Pearce, N., and Heederik, D.: Bioaerosol health effects and exposure assessment:
 progress and prospects, Annals of Occupational Hygiene, 47, 187-200, 2003.
- Elbert, W., Taylor, P. E., Andreae, M. O., and Poschl, U.: Contribution of fungi to primary biogenic
 aerosols in the atmosphere: wet and dry discharged spores, carbohydrates, and inorganic ions, Atmos
 Chem Phys, 7, 4569-4588, 2007.
- Faulwetter, R.: Wind-blown rain, a factor in disease dissemination, J. agric. Res, 10, 639-648, 1917.
- Feofilova, E. P.: The Kingdom Fungi: Heterogeneity of Physiological and Biochemical Properties and
 Relationships with Plants, Animals, and Prokaryotes (Review), Applied Biochemistry and Microbiology,
 37, 124-137, 2001.
- Foot, V. E., Kaye, P. H., Stanley, W. R., Barrington, S. J., Gallagher, M., and Gabey, A.: Low-cost realtime multiparameter bio-aerosol sensors, 2008, 71160I-71160I-71112.
- Frankland, A. and Gregory, P.: Allergenic and agricultural implications of airborne ascosporeconcentrations from a fungus, Didymella exitialis, 1973. 1973.
- 755 Fröhlich-Nowoisky, J., Burrows, S., Xie, Z., Engling, G., Solomon, P., Fraser, M., Mayol-Bracero, O.,
- 756 Artaxo, P., Begerow, D., and Conrad, R.: Biogeography in the air: fungal diversity over land and oceans,
- 757 Biogeosciences, 9, 1125-1136, 2012.





- 758 Fröhlich-Nowoisky, J., Kampf, C. J., Weber, B., Huffman, J. A., Pöhlker, C., Andreae, M. O., Lang-
- 759 Yona, N., Burrows, S. M., Gunthe, S. S., Elbert, W., Su, H., Hoor, P., Thines, E., Hoffmann, T., Després,
- V. R., and Pöschl, U.: Bioaerosols in the Earth system: Climate, health, and ecosystem interactions,
- 761 Atmospheric Research, 182, 346-376, doi: 10.1016/j.atmosres.2016.07.018, 2016.
- Fröhlich-Nowoisky, J., Pickersgill, D. A., Després, V. R., and Pöschl, U.: High diversity of fungi in air
 particulate matter, Proceedings of the National Academy of Sciences, 106, 12814-12819, 2009.
- Gabey, A., Gallagher, M., Whitehead, J., Dorsey, J., Kaye, P. H., and Stanley, W.: Measurements and
 comparison of primary biological aerosol above and below a tropical forest canopy using a dual channel
- fluorescence spectrometer, Atmos Chem Phys, 10, 4453-4466, 2010.
- Gilardoni, S., Vignati, E., Marmer, E., Cavalli, F., Belis, C., Gianelle, V., Loureiro, A., and Artaxo, P.:
 Sources of carbonaceous aerosol in the Amazon basin, Atmos Chem Phys, 11, 2747-2764, 2011.
- 769 Gonçalves, F. L. T., Bauer, H., Cardoso, M. R. A., Pukinskas, S., Matos, D., Melhem, M., and Puxbaum,
- H.: Indoor and outdoor atmospheric fungal spores in the São Paulo metropolitan area (Brazil): species and
- numeric concentrations, International journal of biometeorology, 54, 347-355, 2010.
- 772 Graham, B., Guyon, P., Taylor, P. E., Artaxo, P., Maenhaut, W., Glovsky, M. M., Flagan, R. C., and
- Andreae, M. O.: Organic compounds present in the natural Amazonian aerosol: Characterization by gas
 chromatography-mass spectrometry, J Geophys Res-Atmos, 108, 4766-4766, 2003.
- 774 chromatography-mass spectrometry, f deophys Res-Atmos, 100, 4700-4700, 2005.
- 775 Gregory, P. H. and Sreeramulu, T.: Air spora of an estuary, T Brit Mycol Soc, 41, 145-156, 1958.
- Haga, D., Iannone, R., Wheeler, M., Mason, R., Polishchuk, E., Fetch, T., Kamp, B., McKendry, I., and
- 777 Bertram, A.: Ice nucleation properties of rust and bunt fungal spores and their transport to high altitudes,
- where they can cause heterogeneous freezing, Journal of Geophysical Research: Atmospheres, 118, 7260 7272, 2013.
- 780 Hairston, P. P., Ho, J., and Quant, F. R.: Design of an instrument for real-time detection of bioaerosols
- vising simultaneous measurement of particle aerodynamic size and intrinsic fluorescence, Journal of
 Aerosol Science, 28, 471-482, 1997.
- Heald, C. L. and Spracklen, D. V.: Atmospheric budget of primary biological aerosol particles from
 fungal spores, Geophysical Research Letters, 36, L09806/09801-L09806/09805, 2009.
- Healy, D., Huffman, J., O'Connor, D., Pöhlker, C., Pöschl, U., and Sodeau, J.: Ambient measurements of
 biological aerosol particles near Killarney, Ireland: a comparison between real-time fluorescence and
 microscopy techniques, Atmos Chem Phys. 14, 2055 2060, 2014
- microscopy techniques, Atmos Chem Phys, 14, 8055-8069, 2014.
- Hernandez, M., Perring, A. E., McCabe, K., Kok, G., Granger, G., and Baumgardner, D.: Chamber
 catalogues of optical and fluorescent signatures distinguish bioaerosol classes, Atmos Meas Tech, 9,
 3283-3292, 2016.
- Hill, S. C., Mayo, M. W., and Chang, R. K.: Fluorescence of bacteria, pollens, and naturally occurring
 airborne particles: excitation/emission spectra, DTIC Document, 2009.
- Hirst, J. and Stedman, O.: Dry liberation of fungus spores by raindrops, Microbiology, 33, 335-344, 1963.





- Hoose, C., Kristjánsson, J. E., Chen, J.-P., and Hazra, A.: A Classical-Theory-Based Parameterization of
 Heterogeneous Ice Nucleation by Mineral Dust, Soot, and Biological Particles in a Global Climate Model,
- Journal of the Atmospheric Sciences, 67, 2483-2503, 2010.
- 797 Huffman, J. A., Prenni, A. J., DeMott, P. J., Pöhlker, C., Mason, R. H., Robinson, N. H., Fröhlich-
- 798 Nowoisky, J., Tobo, Y., Després, V. R., Garcia, E., Gochis, D. J., Harris, E., Müller-Germann, I., Ruzene,
- 799 C., Schmer, B., Sinha, B., Day, D. A., Andreae, M. O., Jimenez, J. L., Gallagher, M., Kreidenweis, S. M.,
- 800 Bertram, A. K., and Pöschl, U.: High concentrations of biological aerosol particles and ice nuclei during
- and after rain, Atmos. Chem. Phys., 13, 6151-6164, 2013.
- Huffman, J. A. and Santarpia, J.: Online techniques for quantification and characterization of biological
 aerosol. In: Microbiology of aerosols, Delort, A.-M. and Amato, P. (Eds.), Wiley, Hoboken, NJ, 2016.
- Huffman, J. A., Sinha, B., Garland, R. M., Snee-Pollmann, A., Gunthe, S. S., Artaxo, P., Martin, S. T.,
- 805 Andreae, M. O., and Pöschl, U.: Size distributions and temporal variations of biological aerosol particles
- 806 in the Amazon rainforest characterized by microscopy and real-time UV-APS fluorescence techniques
- during AMAZE-08, Atmos. Chem. Phys., 12, 11997-12019, 2012.
- 808 Hummel, M., Hoose, C., Gallagher, M., Healy, D. A., Huffman, J. A., O'Connor, D., Poeschl, U.,
- 809 Poehlker, C., Robinson, N. H., Schnaiter, M., Sodeau, J. R., Stengel, M., Toprak, E., and Vogel, H.:
- 810 Regional-scale simulations of fungal spore aerosols using an emission parameterization adapted to local
- 811 measurements of fluorescent biological aerosol particles, Atmos Chem Phys, 15, 6127-6146, 2015.
- Ingold, C. T.: Fungal spores. Their liberation and dispersal, Fungal spores. Their liberation and dispersal.,
 1971. 1971.
- Jones, A. M. and Harrison, R. M.: The effects of meteorological factors on atmospheric bioaerosol
- concentrations—a review, Science of the Total Environment, 326, 151-180, 2004.
- Kaye, P., Stanley, W., Hirst, E., Foot, E., Baxter, K., and Barrington, S.: Single particle multichannel bio aerosol fluorescence sensor, Optics Express, 13, 3583-3593, 2005.
- Lau, A. P. S., Lee, A. K. Y., Chan, C. K., and Fang, M.: Ergosterol as a biomarker for the quantification
 of the fungal biomass in atmospheric aerosols, Atmos Environ, 40, 249-259, 2006.
- Laumbach, R. J. and Kipen, H. M.: Bioaerosols and sick building syndrome: particles, inflammation, and
 allergy, Current opinion in allergy and clinical immunology, 5, 135-139, 2005.
- 822 Lee, T., Grinshpun, S. A., Kim, K. Y., Iossifova, Y., Adhikari, A., and Reponen, T.: Relationship between 823 indoor and outdoor airborne fungal spores, pollen, and $(1 \rightarrow 3)$ -β-D-glucan in homes without visible mold
- growth, Aerobiologia, 22, 227-235, 2006. $(1 \rightarrow 3)$ -p-D-grucan in nones without visible more
- 825 Lee, T., Sullivan, A. P., Mack, L., Jimenez, J. L., Kreidenweis, S. M., Onasch, T. B., Worsnop, D. R.,
- Malm, W., Wold, C. E., Hao, W. M., and Collett, J. L., Jr.: Chemical Smoke Marker Emissions During
- Flaming and Smoldering Phases of Laboratory Open Burning of Wildland Fuels, Aerosol Science and
 Technology, 44, I-V, 2010.
- Lewis, D. H. and Smith, D. C.: Sugar alcohols (polyols) in fungi and green plants, New Phytol, 66, 185204, 1967.





- 831 Liang, L., Engling, G., He, K., Du, Z., Cheng, Y., and Duan, F.: Evaluation of fungal spore characteristics
- in Beijing, China, based on molecular tracer measurements, Environmental Research Letters, 8, 014005,
 2013.
- 834 Lin, W.-H. and Li, C.-S.: Associations of fungal aerosols, air pollutants, and meteorological factors,
- 835 Aerosol Science & Technology, 32, 359-368, 2000.
- Madden, L.: Effects of rain on splash dispersal of fungal pathogens, Canadian Journal of Plant Pathology,
 19, 225-230, 1997.
- 838 Madelin, T.: Fungal aerosols: a review, Journal of Aerosol Science, 25, 1405-1412, 1994.
- Miller, J. D. and Young, J. C.: The use of ergosterol to measure exposure to fungal propagules in indoor
 air, American Industrial Hygiene Association Journal, 58, 39-43, 1997.
- 841 Morris, C., Sands, D., Glaux, C., Samsatly, J., Asaad, S., Moukahel, A., Goncalves, F. L. T., and Bigg, E.:
- Urediospores of rust fungi are ice nucleation active at>- 10 C and harbor ice nucleation active bacteria,
 Atmos Chem Phys, 13, 4223-4233, 2013.
- 844 Oliveira, M., Ribeiro, H., Delgado, J., and Abreu, I.: The effects of meteorological factors on airborne
- fungal spore concentration in two areas differing in urbanisation level, International journal of
 biometeorology, 53, 61-73, 2009.
- Ortega, J., Turnipseed, A., Guenther, A. B., Karl, T. G., Day, D. A., Gochis, D., Huffman, J. A., Prenni,
 A. J., Levin, E. J. T., Kreidenweis, S. M., DeMott, P. J., Tobo, Y., Patton, E. G., Hodzic, A., Cui, Y. Y.,
 Harley, P. C., Hornbrook, R. S., Apel, E. C., Monson, R. K., Eller, A. S. D., Greenberg, J. P., Barth, M.
- 850 C., Campuzano-Jost, P., Palm, B. B., Jimenez, J. L., Aiken, A. C., Dubey, M. K., Geron, C., Offenberg,
- J., Ryan, M. G., Fornwalt, P. J., Pryor, S. C., Keutsch, F. N., DiGangi, J. P., Chan, A. W. H., Goldstein,
- A. H., Wolfe, G. M., Kim, S., Kaser, L., Schnitzhofer, R., Hansel, A., Cantrell, C. A., Mauldin, R. L., and
- 853 Smith, J. N.: Overview of the Manitou Experimental Forest Observatory: site description and selected
- science results from 2008 to 2013, Atmos Chem Phys, 14, 6345-6367, 2014.
- Perring, A., Schwarz, J., Baumgardner, D., Hernandez, M., Spracklen, D., Heald, C., Gao, R., Kok, G.,
 McMeeking, G., and McQuaid, J.: Airborne observations of regional variation in fluorescent aerosol
- across the United States, Journal of Geophysical Research: Atmospheres, 120, 1153-1170, 2015.
- Pohlker, C., Huffman, J. A., Forster, J. D., and Poschl, U.: Autofluorescence of atmospheric bioaerosols:
 spectral fingerprints and taxonomic trends of pollen, Atmos Meas Tech, 6, 3369-3392, 2013.
- Pöhlker, C., Huffman, J. A., and Poeschl, U.: Autofluorescence of atmospheric bioaerosols-fluorescent
 biomolecules and potential interferences, Atmos Meas Tech, 5, 37-71, 2012a.
- 862 Pöhlker, C., Wiedemann, K. T., Sinha, B., Shiraiwa, M., Gunthe, S. S., Smith, M., Su, H., Artaxo, P.,
- Chen, Q., and Cheng, Y.: Biogenic potassium salt particles as seeds for secondary organic aerosol in the
 Amazon, Science, 337, 1075-1078, 2012b.
- 865 Pöschl, U., Martin, S., Sinha, B., Chen, Q., Gunthe, S., Huffman, J., Borrmann, S., Farmer, D., Garland,
- 866 R., and Helas, G.: Rainforest aerosols as biogenic nuclei of clouds and precipitation in the Amazon,
- 867 Science, 329, 1513-1516, 2010.





- Pöschl, U. and Shiraiwa, M.: Multiphase chemistry at the atmosphere–biosphere interface influencing
 climate and public health in the Anthropocene, Chemical reviews, 115, 4440-4475, 2015.
- 870 Prenni, A. J., Tobo, Y., Garcia, E., DeMott, P. J., Huffman, J. A., McCluskey, C. S., Kreidenweis, S. M.,
- Prenni, J. E., Pöhlker, C., and Pöschl, U.: The impact of rain on ice nuclei populations at a forested site in
 Colorado, Geophysical Research Letters, 40, 227-231, 2013.
- Pyrri, I. and Kapsanaki-Gotsi, E.: A comparative study on the airborne fungi in Athens, Greece, by viable
 and non-viable sampling methods, Aerobiologia, 23, 3-15, 2007.
- 875 Rasband, W. and ImageJ, U.: Bethesda, Md, USA. ImageJ, 1997.
- 876 Rathnayake, C. M., Metwali, N., Baker, Z., Jayarathne, T., Kostle, P. A., Thorne, P. S., O'Shaughnessy, P.
- 877 T., and Stone, E. A.: Urban enhancement of PM10 bioaerosol tracers relative to background locations in
- the Midwestern United States, Journal of Geophysical Research: Atmospheres, 121, 5071-5089, 2016.
- Robinson, N. H., Allan, J. D., Huffman, J. A., Kaye, P. H., Foot, V. E., and Gallagher, M.: Cluster
 analysis of WIBS single-particle bioaerosol data, Atmos Meas Tech, 6, 337-347, 2013.
- 881 Saari, S., Niemi, J., Rönkkö, T., Kuuluvainen, H., Järvinen, A., Pirjola, L., Aurela, M., Hillamo, R., and
- 882 Keskinen, J.: Seasonal and diurnal variations of fluorescent bioaerosol concentration and size distribution
- in the urban environment, Aerosol and Air Quality Research, 15, 572-581, 2015.
- Saari, S., Putkiranta, M., and Keskinen, J.: Fluorescence spectroscopy of atmospherically relevant
 bacterial and fungal spores and potential interferences, Atmos Environ, 71, 202-209, 2013.
- 886 Schauer, J. J., Rogge, W. F., Hildemann, L. M., Mazurek, M. A., Cass, G. R., and Simoneit, B. R. T.:
- Source apportionment of airborne particulate matter using organic compounds as tracers, Atmos Environ,
 30, 3837-3855, 1996.
- 889 Schumacher, C. J., Pöhlker, C., Aalto, P., Hiltunen, V., Petäjä, T., Kulmala, M., Pöschl, U., and Huffman,
- J. A.: Seasonal cycles of fluorescent biological aerosol particles in boreal and semi-arid forests of Finland
 and Colorado, Atmos. Chem. Phys., 13, 11987-12001, 2013.
- Sesartic, A. and Dallafior, T. N.: Global fungal spore emissions, review and synthesis of literature data,
 Biogeosciences, 8, 1181-1192, 2011.
- Sesartic, A., Lohmann, U., and Storelvmo, T.: Modelling the impact of fungal spore ice nuclei on clouds
 and precipitation, Environmental Research Letters, 8, 014029, 2013.
- Simoneit, B. R. and Mazurek, M.: Organic tracers in ambient aerosols and rain, Aerosol Science and
 Technology, 10, 267-291, 1989.
- 898 Simoneit, B. R. T., Kobayashi, M., Mochida, M., Kawamura, K., Lee, M., Lim, H.-J., Turpin, B. J., and
- 899 Komazaki, Y.: Composition and major sources of organic compounds of aerosol particulate matter
- sampled during the ACE-Asia campaign, Journal of Geophysical Research, [Atmospheres], 109,
 D19S10/11-D19S10/22, 2004.
- 902 Sodeau, J. and O'Connor, D.: Bioaerosol Monitoring of the Atmosphere for Occupational and
- 903 Environmental Purposes, Comprehensive Analytical Chemistry, 2016. 2016.





- 904 Spracklen, D. and Heald, C. L.: The contribution of fungal spores and bacteria to regional and global 905 aerosol number and ice nucleation immersion freezing rates, Atmos Chem Phys, 14, 9051-9059, 2014.
- 906 Stone, B. and Clarke, A.: Chemistry and biology of (1, 3)-D-glucans, Victoria, Australia.: La Trobe 907 University Press, 1992. 236-239, 1992.
- 908 Taylor, P. E. and Jonsson, H.: Thunderstorm asthma, Current allergy and asthma reports, 4, 409-413, 909 2004.
- 910 Tegen, I. and Fung, I.: Modeling of mineral dust in the atmosphere: Sources, transport, and optical 911 thickness, Journal of Geophysical Research: Atmospheres, 99, 22897-22914, 1994.
- 912 Tobo, Y., Prenni, A. J., DeMott, P. J., Huffman, J. A., McCluskey, C. S., Tian, G., Pöhlker, C., Pöschl,
- 913 U., and Kreidenweis, S. M.: Biological aerosol particles as a key determinant of ice nuclei populations in
- 914 a forest ecosystem, Journal of Geophysical Research: Atmospheres, 118, 10,100-110,110, 2013.
- 915 Toprak, E. and Schnaiter, M.: Fluorescent biological aerosol particles measured with the Waveband
- 916 Integrated Bioaerosol Sensor WIBS-4: laboratory tests combined with a one year field study, Atmos. 917 Chem. Phys, 13, 225-243, 2013.
- 918 Weete, J. D.: Sterols of the fungi: distribution and biosynthesis, Phytochemistry, 12, 1843-1864, 1973.
- 919 Womiloju, T. O., Miller, J. D., Mayer, P. M., and Brook, J. R.: Methods to determine the biological 920 composition of particulate matter collected from outdoor air, Atmos Environ, 37, 4335-4344, 2003.
- Yang, Y., Chan, C.-y., Tao, J., Lin, M., Engling, G., Zhang, Z., Zhang, T., and Su, L.: Observation of 921 922 elevated fungal tracers due to biomass burning in the Sichuan Basin at Chengdu City, China, Science of 923
- the Total Environment, 431, 68-77, 2012.
- 924 Yttri, K. E., Simpson, D., Noejgaard, J. K., Kristensen, K., Genberg, J., Stenstrom, K., Swietlicki, E.,
- 925 Hillamo, R., Aurela, M., Bauer, H., Offenberg, J. H., Jaoui, M., Dye, C., Eckhardt, S., Burkhart, J. F.,
- Stohl, A., and Glasius, M.: Source apportionment of the summer time carbonaceous aerosol at Nordic 926 927 rural background sites, Atmos Chem Phys, 11, 13339-13357, 2011a.
- 928 Yttri, K. E., Simpson, D., Stenstrom, K., Puxbaum, H., and Svendby, T.: Source apportionment of the 929 carbonaceous aerosol in Norway - quantitative estimates based on 14C, thermal-optical and organic tracer 930 analysis, Atmos Chem Phys, 11, 9375-9394, 2011b.
- 931 Zhang, T., Engling, G., Chan, C.-Y., Zhang, Y.-N., Zhang, Z.-S., Lin, M., Sang, X.-F., Li, Y. D., and Li, 932 Y.-S.: Contribution of fungal spores to particulate matter in a tropical rainforest, Environmental Research
- 933 Letters, 5, No pp. given, 2010.
- 934 Zhang, Z., Engling, G., Zhang, L., Kawamura, K., Yang, Y., Tao, J., Zhang, R., Chan, C.-y., and Li, Y.:
- 935 Significant influence of fungi on coarse carbonaceous and potassium aerosols in a tropical rainforest, 936 Environmental Research Letters, 10, 1-9, 2015.
- 937 Zhu, C., Kawamura, K., and Kunwar, B.: Organic tracers of primary biological aerosol particles at
- subtropical Okinawa Island in the western North Pacific Rim, Journal of Geophysical Research: 938
- Atmospheres, 120, 5504-5523, 2015. 939
- 940





941 <u>Tables and Figures</u>:

942

					Mass Con	centration			
			Aral (ng		Man	nitol m ⁻³)	$(1 \rightarrow 3)$ - β -D-glucan (pg m ⁻³)		
			Rainy	Dry	Rainy	Dry	Rainy	Dry	
	nitol n ⁻³)	Rainy	<u>0.839</u>						
	Mannitol (ng m ⁻³)	Dry		0.312					
Mass Concentration	$(1 \rightarrow 3)$ - β -D-glucan (pg m ⁻³)	Rainy	0.000		0.003				
Mass Con		Dry		0.000		0.327			
	oxins m ⁻³)	Rainy	0.116		0.126		0.427		
	Endotoxins (EU m ⁻³)	Dry		0.012		0.113		0.103	

943

944 <u>**Table 1**</u>: Square of correlation coefficients (\mathbb{R}^2) comparing total mass concentration of molecular tracers

945 to each other. EU: endotoxin units. Boxes colored by coefficient value (<u>Bold Underline</u>> 0.7; 0.7 > Bold
946 > 0.4).





													E	1.0	N	C
				bitol m ³)	Man		$(1 \rightarrow 3)$ - β -(pg		Endotox	·	Aral (spore			al Spore 1 (spores	Colony	Forming CFU m ⁻³)
			Rainy	Dry	Rainy	Dry	Rainy	Dry	Rainy	Dry	Rainy	Dry	Rainy	Dry	Rainy	Dry
	UVAPS		<u>0.732</u>	0.127	<u>0.877</u>	0.160	0.006	0.012	0.153	0.067	0.483	0.278	0.504	0.571	0.469	0.491
ion		FL	0.554	0.250	<u>0.810</u>	0.255	0.128	0.010	0.068	0.066	0.159	0.200	0.088	0.314	0.330	<u>0.737</u>
Concentration		FL1	0.602	0.445	<u>0.819</u>	0.412	0.042	0.001	0.090	0.012	0.667	0.339	<u>0.863</u>	0.621	0.470	0.546
Conc		FL2	0.617	0.248	<u>0.843</u>	0.342	0.092	0.001	0.039	0.094	0.485	0.302	0.442	0.340	0.560	0.543
mber		FL3	0.561	0.222	<u>0.818</u>	0.251	0.124	0.008	0.071	0.065	0.178	0.181	0.104	0.306	0.367	<u>0.736</u>
UV-LIF Mass or Number	WIBS	Cl1	<u>0.824</u>	<u>0.764</u>	<u>0.799</u>	0.109	0.000	0.134	0.229	0.011	0.679	0.543	<u>0.775</u>	0.423	0.128	0.690
Mass	ŕ	Cl2	0.005	0.002	0.004	0.006	0.002	0.047	0.006	0.017	0.052	0.056	0.001	0.075	0.081	<u>0.930</u>
-LIF		Cl3	0.267	0.164	0.261	0.198	0.003	0.011	0.016	0.066	0.052	0.116	0.087	0.439	0.262	0.383
Ν		Cl4	0.048	0.046	0.172	0.118	0.115	0.011	0.179	0.145	0.062	0.089	0.001	0.065	0.120	0.000
		Cl _{Bact}							0.041	0.081						

947 <u>**Table 2**</u>: Square of correlation coefficients (R²) comparing fluorescent particle measurements from UV-LIF instruments to measurements from

molecular tracers. Columns marking tracer mass (top line) indicate correlations between time-averaged UV-LIF and tracer mass concentrations
 (left side), and columns marking fungal spore number indicate correlations between fungal spore number concentrations estimated from time-

averaged UV-LIF and tracer or culture measurements (right side). FL1, FL2, FL3 represent individual channels from the WIBS. FL represents all

951 particle exhibiting fluorescence in any channel. Cl1, Cl2, Cl3, Cl4 are clusters that estimate particle concentrations as a mixture of various

channels (Crawford et al., 2015). Cl_{Bact} is a sum of the "bacteria" clusters Cl2-4. Boxes colored by coefficient value (**Bold Underline** > 0.7; 0.7 >

953 **Bold**> 0.4).





			М	ass Concentratio	n							
	Arabitol	Mannitol	Erythritol	Levoglucosa	Glucose	Endotoxins	(1→3)-β-					
	$(ng m^{-3})$	$(ng m^{-3})$	$(ng m^{-3})$	n	$(ng m^{-3})$	(EU m ⁻³)	D-glucan					
				(ng m ⁻³)			(pg m ⁻³)					
Dry	10.6	11.9	0.840	14.2	38.7	0.192	8.8 5					
5	± 2.5	± 3.2	± 0.610	± 10.7	± 21.3	± 0.0970	± 7.68					
	n = 18	n=18	n=16	n=15	n=18	n=18	n=18					
Rainy	35.2	44.9	1.12	12.4	73.2	1.43	10.6					
	± 10.5	± 13.8	± 0.38	± 19.1	± 50.5	± 1.22	± 8.2					
	n=11	n=11	n=3	n=8	n=11	n=10	n=11					
Other	20.2	22.7	0.664	9.21	56.5	0.311	6.08					
	± 8.9	± 8.3	± 0.515	± 1.66	± 39.2	± 0.159	± 6.08					
	n=6	n=6	n=6	n=5	n=6	n=6	n=6					
	Mass Contribution (%)											
Dry	0.18 %	0.202 %	0.0.14 %	0.21 %	0.67 %		0.16 %					
5	± 0.05	± 0.073	± 0.011	±0.17	±0.49		±0.16					
	n=18	n=18	n=16	n=15	n=18		n=18					
Rainy	0.83 %	1.07 %	0.032 %	0.27 %	1.60 %		0.25 %					
5	± 0.32	±0.44	±0.009	±0.41	±1.09		±0.21					
	n=11	n=11	n=3	n=8	n=11		n=11					
Other	0.25 %	0.37 %	0.013 %	0.15 %	0.83 %		0.12 %					
	± 0.28	± 0.29	±0.015	±0.11	±0.64		±0.19					
	n=6	n=6	n=6	n=5	n=6		n=6					
		•	Fungal Spore	Number Concer	ntration (m ⁻³)							
Dry	8870	6890										
2	± 2060	± 1870										
	n=18	n=18										
Rainy	29310	26430										
2	± 8727	± 8139										
	n=11	n=11										
Other	16850	13350										
	± 7415	± 4863										
	n=6	n=6										
	Fungal Spore Mass Contribution (%)											
Dry	4.81 %	3.72 %										
-	± 1.36	± 1.12										
	n=18	n=18										
Rainy	22.88 %	20.66 %										
-	± 8.84	±8.49										
	n=11	n=11										
Other	9.80 %	7.31 %										
	± 7.67	± 5.60										
	n=6	n=6										

954

Table 3: Campaign-average concentrations of molecular tracers (measured) and fungal spores (number

955 concentration estimated from arabitol and mannitol mass). Each set of data broken into wetness

categories. Values are mean \pm standard deviation; *n* shows the number of samples used for averaging. 956

Fungal spore mass contribution was based on the assumption by Bauer et al. (2008b) of 33 pg spore⁻¹. 957

958 Total particulate matter mass calculated from UV-APS number concentration (m⁻³) and converted to mass

959 over aerodynamic particle diameter range $0.5 - 15 \,\mu m$ using density of 1.5 g cm⁻³.





960





963 **Figure 1:** Aerial overview of BEACHON-RoMBAS field site at the Manitou Experimental Forest

964 Observatory located northwest of Colorado Springs, CO. Locations of all instruments and sensors

965 discussed here are marked and were located within a 50 m radius. Figure adapted from Figure 1a of

966 Huffman et al. (2013)







967

968 **Figure 2**: Time series of key species concentrations and meteorological data over entire campaign. (a)

969 Fluorescent particle number size distribution measured with UV-APS instrument. Color scale indicates

970 fluorescent particle number concentration (L⁻¹). (b) Meteorological data: relative humidity (RH),
971 disdrometer rainfall (mm per 15 min), leaf wetness (mV). (c) Wetness category indicated as colored bars;

green, Rainy; brown, Dry; pink, Other. Bar width corresponds to filter sampling periods. Lightened

green, Kanty, brown, Dry, pink, other. Bar with corresponds to inter sampling periods. Eightercolored bars extend vertically to highlight categorization. (d) Colored traces show fungal spore

974 concentrations estimated from molecular tracers (circles) and WIBS Cl1 data (squares). I Stacked bars

show relative fraction of fluorescent particle type corresponding to each WIBS category.







976

977 Figure 3: Characteristic differences between different wetness periods (Dry, Rainy, Other). (a) Relative 978 fraction of fluorescent particle number corresponding to each WIBS category. Bars show relative standard 979 deviation of category fraction in each wetness group (Dry, 19 samples; Rainy, 11 samples; Other, 6 980 samples). (b, c) Distribution of fungal OTU (operational taxonomic unit) values. (b) Fungal community 981 composition at phylum and class level with Agaricomycetes (dominant class with consistently ~60% of diversity) removed. Relative proportion of OTUs assigned to different fungal classes and phyla for each 982 983 sample category shown. (c) Venn diagram showing the number of unique (wetness category specific) and 984 shared OTUs (represented by numbers in overlapping areas) among the sample categories (Dry, 11 985 samples; Rainy, 7 samples; Other, 3 samples). OTUs classified as cluster of sequences with $\geq 97\%$ 986 similarity. Taxonomic assignments were performed using BLAST against NCBI database. In total, 3902 987 sequences, representing 406 fungal OTUs from 3 phyla and 12 classes were detected. Despite differences 988 in community structure across the sample categories, phylogenetic representation appears largely similar.







989 UV-APS FAPs (10°m°) UV-APS FAPs (10°m°)
 990 Figure 4: Number concentration of fluorescent particles as a function of instrument channel, averaged
 991 over entire measurement period. (a) Box-whisker plot of fluorescent particle number concentration for
 992 WIBS FL1, FL2, FL3, and UVAPS. Circle markers shows mean values, internal horizontal line shows
 993 median, top and bottom of box show inner quartile, and whiskers show 5th and 95th percentiles. (b) WIBS
 994 FL1 versus UV-APS (c) WIBS FL2 versus UV-APS (d) WIBS FL3 versus UV-APS. Crosses represent 5 995 minute average points. Linear fits assigned for data in each wetness category.







996 997

Figure 5: Mass concentrations of molecular tracers and fluorescent particles (assuming unit density particle mass): arabitol – top row, and mannitol – bottom row. Average mass concentration of arabitol (a) and mannitol (b) in each wetness category. Central marker shows mean value of individual filter
concentration values, bars represent standard deviation (*s*) range of filter values, and individual points show outliers beyond mean ± *s*. Correlation of arabitol (c) and mannitol (d) with fluorescent particle mass from UV-APS. Correlation of arabitol I and mannitol (f) with fluorescent particle mass from WIBS
Cluster 1. R² values shown for each fit in c, d, e, f. Linear fit parameters are shown in Table S2.







1005 Figure 6: Estimated fungal spore number concentration, calculated using mass of arabitol and mannitol 1006 1007 per spore reported by Bauer et al. (2008a). Estimates from arabitol (top row) and mannitol (middle row). 1008 Average fungal spore concentration, calculated using arabitol mass (a), mannitol mass (b), and colony 1009 forming units (c) in each wetness category. Central marker shows mean value of individual filter 1010 concentration values, bars represent standard deviation (s) range of filter values, and individual points 1011 show outliers beyond mean $\pm s$. Correlation of fungal spore number calculated from arabitol (d), mannitol 1012 (e), and colony forming units (f) concentration with estimated fluorescent particle mass from UV-APS. 1013 Correlation of fungal spore number calculated from arabitol (g), mannitol (h), and colony forming unit (i) 1014 concentration with fluorescent particle concentration from WIBS Cluster 1. R² value shown for each fit 1015 (right two columns). Linear fit parameters are shown in Table S3.







1016

1017 **Figure 7:** Estimated fraction of total aerosol mass contributed by fungal spores. Fungal spore mass 1018 concentration (μ g/m³) calculated separately from mannitol and arabitol concentration and using average

1019 mass per spore reported by Bauer et al. (2008b). Total particulate matter mass calculated from UV-APS

1020 number concentration (m⁻³) and converted to mass over aerodynamic particle diameter range $0.5 - 15 \ \mu m$

using density of 1.5 g cm⁻³. Central marker shows mean value of individual filter concentration values,

bars represent standard deviation (*s*) range of filter values, and individual points show outliers beyond

1023 mean $\pm s$.









Figure 8: Endotoxin mass concentration as an approximate indicator of gram-negative bacteria1028concentration. (a) Averaged concentration in each wetness category. Central marker shows mean value of1029individual filter concentration values, bars represent standard deviation (s) range of filter values, and1030individual points show outliers beyond mean $\pm s$. (b) Correlation of endotoxin mass concentration with1031estimated fluorescent particle mass from UV-APS. (c) Correlation of endotoxin mass concentration with1032estimated fluorescent particle mass summed from Clusters 2, 3, and 4 from Crawford et al. (2015).