

1 **Title:** Fluorescent Bioaerosol Particle, Molecular Tracer, and Fungal Spore Concentrations during Dry
2 and Rainy Periods in a Semi-Arid Forest
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19 **Abstract:**

20 Bioaerosols pose risks to human health and agriculture and may influence the evolution of mixed-phase
21 clouds and the hydrological cycle on local and regional scales. The availability and reliability of methods
22 and data on the abundance and properties of atmospheric bioaerosols, however, are rather limited. Here
23 we analyze and compare data from different real-time Ultraviolet Laser/Light Induced Fluorescence (UV-
24 LIF) instruments with results from a culture-based spore sampler and offline molecular tracers for
25 airborne fungal spores in a semi-arid forest in the Southern Rocky Mountains of Colorado. Commercial
26 UV-APS (Ultraviolet Aerodynamic Particle Sizer) and WIBS-3 (Wideband Integrated Bioaerosol Sensor,
27 Version 3) instruments with different excitation and emission wavelengths were utilized to measure
28 fluorescent aerosol particles (FAP) during both dry weather conditions and periods heavily influenced by
29 rain. Seven molecular tracers of bioaerosols were quantified by analysis of total suspended particle (TSP)
30 high-volume filter samples using High Performance Anion Exchange Chromatography with Pulsed
31 Amperometric Detection (HPAEC-PAD). From the same measurement campaign, Huffman et al. (2013)
32 previously reported dramatic increases in total and fluorescent particle concentrations during and
33 immediately after rainfall and also showed a strong relationship between the concentrations of FAP and
34 ice nuclei (Huffman et al., 2013; Prenni et al., 2013). Here we investigate molecular tracers and show that
35 during rainy periods the atmospheric concentrations of arabitol ($35.2 \pm 10.5 \text{ ng m}^{-3}$) and mannitol ($44.9 \pm$
36 13.8 ng m^{-3}) were 3-4 times higher than during dry periods. During and after rain the correlations between
37 FAP and tracer mass concentrations were also significantly improved. Fungal spore number
38 concentrations on the order of 10^4 m^{-3} , accounting for 2-4% of TSP mass during dry periods and 17-23%
39 during rainy periods, were obtained from scaling the tracer measurements and from multiple analysis
40 methods applied to the UV-LIF data. Endotoxin concentrations were also enhanced during rainy periods,
41 but showed no correlation with FAP concentrations. Average mass concentrations of erythritol,
42 levoglucosan, glucose, and (1→3)- β -D-glucan in TSP samples are reported separately for dry and rainy
43 weather conditions. Overall, the results indicate that UV-LIF measurements can be used to infer fungal
44 spore concentrations, but substantial development of instrumental and data analysis methods seems
45 required for improved quantification.

1. Introduction

Primary biological aerosols particles (PBAP) are of keen interest within the scientific community, partially because methods for their quantification and characterization are advancing rapidly (Huffman and Santarpia, 2016; Sodeau and O'Connor, 2016). The term PBAP, or equivalently bioaerosol, generally comprises several classes of airborne biological particles including viruses, bacteria, fungal spores, pollen and their fragments (Després et al., 2012; Fröhlich-Nowoisky, 2016). Fungal spores are of particular atmospheric interest because they can cause a variety of deleterious health effects in humans, animals, and agriculture, and it has been shown that they can represent a significant fraction of total organic aerosol emissions (Deguillaume et al., 2008; Gilardoni et al., 2011; Madelin, 1994), especially in tropical regions (Elbert et al., 2007; Huffman et al., 2012; Pöschl et al., 2010; Zhang et al., 2010). Current estimates of the atmospheric concentration of fungal spores range from 10^0 to more than 10^4 m^{-3} (Frankland and Gregory, 1973; Gregory and Sreeramulu, 1958; Heald and Spracklen, 2009; Hummel et al., 2015; Sesartic and Dallafior, 2011). Fungal spores may also impact the hydrological cycle as giant cloud condensation nuclei (GCCN) or as ice nuclei (IN) (Haga et al., 2013; Morris et al., 2013; Sesartic et al., 2013). Additionally, several classes of bioaerosols and their constituent components, such as (1 \rightarrow 3)- β -D-glucan and endotoxins, have been implicated in respiratory distress and allergies (Burger, 1990; Douwes et al., 2003; Laumbach and Kipen, 2005; Linneberg, 2011; Pöschl and Shiraiwa, 2015). For example, asthma and allergies have shown notable increases during thunderstorms due to elevated bioaerosol concentrations (Taylor and Jonsson, 2004) especially when attributed to fungal spores (Allitt, 2000; Dales et al., 2003).

Molecular tracers have long been utilized as a means of aerosol source tracking (Schauer et al., 1996; Simoneit and Mazurek, 1989; Simoneit et al., 2004). In recent years, analysis of molecular tracers has been utilized for the quantification of PBAP in atmospheric samples and has been compared, for example, with results from microscopy (Bauer et al., 2008a) and culture samples (Chow et al., 2015b; Womiloju et al., 2003). Three organic molecules have been predominately utilized as unique tracers of fungal spores: ergosterol, mannitol, and arabitol. The majority of atmospherically relevant fungal spores are released by active wet discharge processes common in *Ascomycota* and *Basidiomycota*, meaning that the fungal organism actively ejects spores at a time most advantageous for the spore dispersal and germination processes, often when relative humidity (RH) is high (Ingold, 1971). While there are several mechanisms of active spore emission (e.g. Buller's drop (Buller, 1909) and osmotic pressure canons (Ingold, 1971)), they each involve the secretion of fluid containing hygroscopic compounds, such as arabitol, mannitol, potassium and chloride ions, as well as other solutes (Elbert et al., 2007), released near the site of spore growth. When the spores are ejected, some of the fluid adheres to the spores and becomes aerosolized. Several of these secreted compounds are thought to enter the atmosphere linked uniquely with spore emission processes, and so these tracers have been used to estimate atmospheric concentrations of fungal spores. Arabitol and mannitol are both sugar alcohols (polyols) that serve as energy stores for the spore (Feofilova, 2001). Arabitol is unique to fungal spores and lichen, while mannitol is present in fungal spores, lichen, algae, and higher plants (Lewis and Smith, 1967). Ergosterol is found within the cell membranes of fungal spores (Weete, 1973) and has been used as an ambient fungal spore trace (Di Filippo et al., 2013; Miller and Young, 1997). Comparing the seasonal trends of arabitol and mannitol with ergosterol, Burshtein et al. (2011) showed positive correlations between arabitol or mannitol and ergosterol only in the spring and autumn suggesting that the source of these polyols is unlikely to be solely fungal in origin or that the amount of each compound emitted varies considerably between species type and season. While ergosterol has been directly linked to fungal spores in the air, ergosterol is prone to photochemical degradation and is difficult to analyze and quantify directly. Quantification of ergosterol typically requires chemical derivatization by silylation before analysis via gas chromatography (Axelsson et al., 1995; Burshtein et al., 2011; Lau et al., 2006). In contrast, analysis of sugar alcohols by ion chromatography involves fewer steps and has been successfully applied to monitor seasonal variations of atmospheric aerosol concentration at a number of sites (Bauer et al., 2008a; Caseiro et al., 2007; Yang et al., 2012; Yttri et al., 2011a; Zhang et al., 2010; Zhang et al.,

97 2015) including pg m^{-3} levels in the Antarctic (Barbaro et al., 2015). By measuring spore count and tracer
98 concentration in parallel at one urban and two suburban sites in Vienna, Austria Bauer et al. (2008a)
99 estimated the amount of each tracer per fungal spore emitted. Potassium ions have also been linked to
100 emission of biogenic aerosol (Pöhlker et al., 2012b) and are co-emitted with fungal spores, however,
101 application of potassium as a fungal tracer is uncommon because it is predominantly associated with
102 biomass burning (Andreae and Crutzen, 1997). Additionally, (1 \rightarrow 3)- β -D-glucan (fungal spores and
103 pollen) and endotoxins (gram-negative bacteria) have also been widely used to measure other bioaerosols
104 (Andreae and Crutzen, 1997; Cheng et al., 2012; Rathnayake et al., 2016b; Stone and Clarke, 1992).
105

106 The direct detection of PBAP has historically been limited to analysis techniques that require
107 culturing or microscopy of the samples. These systems are time-consuming, costly, and often
108 substantially undercount biological particles by an order of magnitude or more (Gonçalves et al., 2010;
109 Pyrri and Kapsanaki-Gotsi, 2007). The sampling methods associated with these measurements also offer
110 relatively low time resolution and low particle size resolution. Recently, techniques utilizing ultraviolet
111 laser/light-induced fluorescence (UV-LIF) for the real-time detection of PBAP have been developed and
112 are being utilized by the atmospheric community for bioaerosol detection. Thus far, the most widely
113 applied LIF instruments for ambient PBAP detection have been the Ultraviolet Aerosol Particle Sizer
114 (UV-APS; TSI Inc. Model 3314, St. Paul, MN) and the Wideband Integrated Bioaerosol Sensor (WIBS;
115 University of Hertfordshire, Hertfordshire, UK, now licensed to Droplet Measurement Technologies,
116 Boulder, CO, USA). Both of these commercially available instruments can provide information in real-
117 time about particle size and fluorescence properties of supermicron atmospheric aerosols.
118 Characterization and co-deployment of these instruments over the past ten years has expanded the
119 knowledge base regarding how to analyze and utilize the information provided from these instruments
120 (Crawford et al., 2015; Healy et al., 2014; Hernandez et al., 2016; Huffman et al., 2013; Perring et al.,
121 2015; Pöhlker et al., 2013; Pöhlker et al., 2012a; Ruske et al., 2016), though the interpretation of UV-LIF
122 results from individual particles is complicated by interfering material that is not biological in nature
123 (Gabey et al., 2010; Huffman et al., 2012; Lee et al., 2010; Saari et al., 2013; Toprak and Schnaiter,
124 2013).
125

126 Here we present analysis of atmospheric concentrations of arabitol and mannitol in relation to
127 results from real-time, ambient particle measurements reported by UV-APS and WIBS. We interrogate
128 these relationships as they pertain to rain conditions (rainfall and RH) that have previously been shown to
129 increase the concentrations of fluorescent aerosols and ice nuclei (Crawford et al., 2014; Huffman et al.,
130 2013; Prenni et al., 2013; Schumacher et al., 2013; Yue et al., 2016). Active wet discharge of ascospores
131 and basidiospores has frequently been reported to correspond with increased RH (Elbert et al., 2007), and
132 fungal spore concentration has also been shown to increase after rain events (e.g. Jones and Harrison,
133 2004). Here we estimate airborne fungal concentrations in a semi-arid forest environment utilizing a
134 combination of real-time fluorescence methods, molecular fungal tracer methods, and direct-to-agar
135 sampling and culturing as parallel surrogates for spore analysis. This study of ambient aerosol represents
136 the first quantitative comparison of real-time aerosol UV-LIF instruments with molecular tracers or
137 culturing.
138

139 **2. Methods**

141 **2.1 Sampling site**

142 Atmospheric sampling was conducted as a part of the BEACHON-RoMBAS (Bio-hydro-
143 atmosphere interactions of Energy, Aerosols, Carbon, H₂O, Organics, and Nitrogen – Rocky Mountain
144 Biogenic Aerosol Study) field campaign conducted at the Manitou Experimental Forest Observatory
145 (MEFO) located 48 km northwest of Colorado Springs, Colorado (2370 m elevation, 39° 06' 0" N, 105°
146 5' 03" W) (Ortega et al., 2014). The site is located in the central Rocky Mountains and is representative of
147 semi-arid montane pine forested regions of North America. During BEACHON-RoMBAS a large,

148 international team of researchers conducted an intensive set of measurements from 20 July to 23 August
149 2011. A summary of results from the campaign are published in the BEACHON campaign special issue
150 of Atmospheric Chemistry and Physics¹. All the data reported here were gathered from instruments and
151 sensors located within a <100 m radius (Fig. 1).

152

153 **2.2 Online fluorescent instruments**

154 UV-APS and WIBS-3 (Model 3; University of Hertfordshire) instruments were operated
155 continuously as a part of the study, and particle data were integrated to five-minute averages before
156 further analysis. The UV-APS was operated under procedures defined in previous studies (Huffman et al.,
157 2013; Schumacher et al., 2013). A total suspended particle (TSP) inlet head ~5.5 m above ground,
158 mounted above the roof of a climate-controlled, metal trailer, was used to sample aerosol directed towards
159 the UV-APS. Bends and horizontal stretches in the 0.75 inch tubing were minimized to reduce losses of
160 large particles (Huffman et al., 2013). The UV-APS detects particles between 0.5-20 μm and records
161 aerodynamic particle diameter and integrated total fluorescence (420-575 nm) after pulsed excitation by a
162 355 nm laser (Hairston et al., 1997). Both UV-APS and WIBS instruments report information about
163 particle number concentration, but it is instructive here to show results in particle mass for comparison
164 between all techniques. Total particle number size distributions (irrespective of fluorescence properties)
165 obtained from the UV-APS and WIBS were converted to mass distributions assuming spherical particles
166 of unit particle mass density, unless otherwise stated, as a first approximation. Total particle concentration
167 values (in $\mu\text{g m}^{-3}$) were obtained for each five-minute period by integrating over the size range 0.5 – 15
168 μm , and these mass concentration values were averaged over the length of the filter sampling periods.
169 Uncertainty in mass concentration values reported here is influenced by assuming a single value for
170 particle mass density and because of slight dissimilarities between size bins of UV-APS and WIBS
171 instruments at particle sizes above 10 μm that dominate particle mass.

172

173 A WIBS-3 was used to continuously sample air at a site ~50 m from the UV-APS trailer and 1.3
174 m above the ground. Briefly, the diameter of individual particles sampled by the WIBS is estimated by
175 the intensity of the elastic side-scatter from a continuous wave 635 nm diode laser and analyzed by a Mie
176 scattering model (Foot et al., 2008; Kaye et al., 2005). Particles that pass through the diode laser activate
177 two optically-filtered Xenon flash lamps. The first lamp excites the particle at 280 nm and the second at
178 370 nm. Emission from the 280 nm excitation is filtered separately for two PMTs, one which detects in a
179 band at 320-400 nm and the other in a band at 410-650 nm. These excitation and emission wavelengths
180 result in a total of three channels of detection: λ_{ex} 280 nm, λ_{em} 320 – 400 nm (FL 1 or Channel A); λ_{ex} 280
181 nm, λ_{em} 410 – 650 nm (FL 2 or Channel B); and λ_{ex} 370 nm, λ_{em} 410– 650 nm (FL 3 or Channel C)
182 (Crawford et al., 2014). Individual particles are considered fluorescent here if they exceed fluorescent
183 thresholds for any channel, as defined as the average of a “forced trigger” baseline plus 3 standard
184 deviations (σ) of the baseline measurement (Gabey et al., 2010).

185

186 WIBS particle-type analysis is utilized to define types of particles that have specific spectral
187 patterns. As defined by Perring et al. (2015), the 3 different fluorescent channels (FL1, FL2, and FL3) can
188 be combined to produce 7 unique fluorescent categories. Observed fluorescence in channel FL1 alone, but
189 without any detectable fluorescence in Channel FL2 or FL3, categorizes a particle as type A. Similarly,
190 observed fluorescence in channels FL2 or FL3, but in no other channels, places a particle in the B or C
191 categories, respectively. Combinations of fluorescence in these channels, such as a particle that exhibits
192 fluorescence in both FL1 and FL2 categorizes a particle as type AB and so on for a possible seven particle
193 types as summarized in Figure S1.

194

¹http://www.atmos-chem-phys.net/special_issue247.html

195 As a separate tool for particle categorization, the University of Manchester has recently developed and
196 applied a hierarchical agglomerative cluster analysis tool for WIBS data, which they have previously
197 applied to the BEACHON-RoMBAS campaign (Crawford et al., 2014; Crawford et al., 2015; Robinson et
198 al., 2013). Here we utilize clusters derived from WIBS-3 data as described by Crawford et al. (2015).
199 Cluster data presented here was analyzed with the Open Source Python package FastCluster (Mullner,
200 2013). Briefly, hierarchical agglomerative cluster analysis was applied to the entire data set and each
201 fluorescent particle was uniquely clustered into one of 4 groups. Cluster 1, assigned by Crawford et al.
202 (2015) as fungal spores, displayed a 1.5-2 μm mode and a daily peak in the early morning that paralleled
203 relative humidity (Schumacher et al., 2013). Clusters 2, 3, and 4 have strong, positive correlations with
204 rainfall and exhibit size modes that peak at $<1.2 \mu\text{m}$ and were initially described by Crawford et al. as
205 bacterial particles. Here we have summed Clusters 2-4 to a single group referred to as Cl_{Bact} , for
206 simplicity when comparing with molecular tracers. It should be noted that assignment of name and origin
207 (e.g. fungal spores or bacteria) to clusters is approximate and does not imply naming accuracy or particle
208 homogeneity. Each cluster likely contains an unknown fraction of contaminating particles, but the clusters
209 are beneficial to group particles more selectively than using fluorescent intensity alone. For more details
210 see Robinson et al. (2013) and Crawford et al. (2015).

211
212 The WIBS-3 utilized here has since been superseded by the WIBS-4 (Univ. Hertfordshire, UK)
213 and WIBS-4A (Droplet Measurement Technologies, Boulder, Colorado). One important difference
214 between the models is that the optical chamber design and filters of the WIBS-4 models were updated to
215 enhance the overall sensitivity of the instrument (Crawford et al., 2014). Additionally, slight differences
216 in detector gain between models and individual units can impact the relative sensitivity of the
217 fluorescence channels. This may result in differences in fluorescent channel intensity between instrument
218 models, as will be discussed later.

219 220 **2.3 High volume sampler**

221 Total suspended particle samples were collected for molecular tracer and molecular genetic
222 analyses using a high volume sampler (Digitel DHA-80) drawing 1000 L min^{-1} through 15 cm glass fiber
223 filters (Macherey-Nagel GmbH, Type MN 85/90, 406015, Düren, Germany) over a variety of sampling
224 times ranging from 4-48 h (supplemental Table S1). The sampler was located $<50 \text{ m}$ from each of the
225 UV-LIF instruments described here, approximately between the WIBS-3 and UV-APS. Prior to sampling
226 all filters were baked at $500 \text{ }^\circ\text{C}$ for 12 h to remove DNA and organic contaminants. Samples were stored
227 in pre-baked aluminum bags after sampling at $-20 \text{ }^\circ\text{C}$ for 1-30 days and then at $-80 \text{ }^\circ\text{C}$ after overnight,
228 international transport cooled on dry ice. Due to the low vapor pressure of the molecular tracers analyzed
229 loss due to volatilization is considered unlikely (Zhang et al., 2010). 36 samples were collected during the
230 study, in addition to handling field blanks and operational field blanks. Handling blanks were acquired by
231 placing a filter into the sampler and immediately removing, without turning on the air flow control.
232 Operational blanks were placed into the sampler and exposed to 10 seconds of air flow.

233 234 **2.4 Slit Sampler**

235 A direct-to-agar slit sampler (Microbiological Air Sampler STA-203, New Brunswick Scientific
236 Co, Inc., Edison, NJ) was used to collect culturable airborne fungal spores. The sampler was placed $\sim 2 \text{ m}$
237 above ground on a wooden support surface with 5 cm x 5 cm holes to allow air flow both up and down
238 through the support structure. Sampled air was drawn over the 15 cm diameter sampling plate filled with
239 growth media at a flow rate of 28 L min^{-1} for sampling periods of 20 to 40 min. Growth media (malt
240 extract medium) was mixed with antibacterial agents (40 units streptomycin, Sigma Aldrich; 20 units
241 ampicillin, Fisher Scientific) to suppress bacterial colony growth. Plates were prepared several weeks in
242 advance and stored in a refrigerator at ca. $4 \text{ }^\circ\text{C}$ until used for sampling. Before each sampling period, all
243 surfaces of the samplers were sterilized by wiping with isopropyl alcohol. Handling and operational
244 blanks were collected to verify that no fungal colonies were being introduced by handling procedures. 14
245 air samples were collected over 20 days and immediately moved to an incubator (Amerex Instruments,

246 Incumax IC150R) set at 25 °C for 3 days prior to counting fungal colonies formed. Each colony, present
247 as a growing dot on the agar surface, is assumed to have originated as one colony forming unit (CFU; i.e.
248 fungal spore) deposited onto the agar by impaction during sampling. The atmospheric concentration of
249 CFU per air volume was calculated using the sampler air flow. Further discussion of methods and initial
250 results from the slit sampler were published by Huffman et al. (2013).

251

252 **2.5 Offline filter analyses**

253

254 *2.5.1 Carbohydrate analysis*

255 Approximately 1/8 of each frozen filter was cut for carbohydrate analysis using a sterile
256 technique, meaning that scissors were cleaned and sterilized and cutting was performed in a positive-
257 pressure laminar flow hood. In order to precisely determine the fractional area of the filter to be analyzed,
258 filters were imaged from a fixed distance above using a camera and compared to a whole, intact filter.
259 Using ImageJ software (Rasband and ImageJ, 1997), the area of each filter slice showing particulate
260 matter (PM) deposit was referenced to a whole filter, and thereby the amount of each filter utilized could
261 be determined. The total PM mass was not measured and so this technique allowed for an estimation of
262 the fraction of each sample used for the analysis, which corresponds to the fraction of PM mass deposited.
263 The uncertainty on the filter area fraction is estimated at 2%, determined as the percent of variation in the
264 area of the filter edge (no PM deposit) as compared to the total filter area.

265

266 Water soluble carbohydrates were extracted from quartz filter samples and analyzed following the
267 procedure described by Rathnayake et al. (2016a). A total of 36 samples were analyzed along with field
268 and lab blanks. All lab and field blanks fell below method detection limits. Extraction was performed by
269 placing the filter slice into a centrifuge tube that had been pre-rinsed with Nanopure™ water (resistance >
270 18.2 MΩ cm⁻¹; Barnstead EasyPure II, 7401). A volume of 8.0 mL of Nanopure™ water was added to the
271 filter in the centrifuge tube to extract water-soluble carbohydrates. Samples were then exposed to rotary
272 shaking for 10 min at 125 rpm, sonication for 30 min at 60 Hz (Branson 5510, Danbury, CT, US), and
273 rotary shaking for another 10 min. After shaking, the extracted solutions were filtered through a 0.45 μm
274 polypropylene syringe filter (GE Healthcare, UK) to remove insoluble particles, including disintegrated
275 filter pieces. One 1.5 mL aliquot of each extracted solution was analyzed for carbohydrates within 24
276 hours of extraction. A duplicate 1.5 mL aliquot was stored in a freezer and analyzed, if necessary due to
277 lack of instrument response or invalid calibration check, within 7 days of extraction. Analysis of
278 carbohydrates was done using a High Performance Anion Exchange Chromatography System with Pulsed
279 Amperometric Detection (HPAEC-PAD, Dionex ICS 5000, Thermo Fisher, Sunnyvale, CA, USA).
280 Details of the instrument specifications and quality standards for carbohydrate determination are available
281 in Rathnayake et al. (2016). Calibration curves for mannitol, levoglucosan, glucose (Sigma-Aldrich),
282 arabitol and erythritol (Alfa Aesar) were generated with seven points each, ranging in aqueous
283 concentration from 0.005 ppm to 5 ppm. The method detection limits for mannitol, levoglucosan, glucose,
284 arabitol, and erythritol were 2.3, 2.8, 1.6, 1.0, and 0.6 ppb, respectively. Method detection limits were
285 determined as 3σ of analyte concentrations recovered from seven spiked filter samples (Rathnayake et al.,
286 2016a). All calibration curves were checked daily using a standard solution to ensure all concentration
287 values were within 10% of the known value. Failure to maintain a valid curve resulted in recalibration of
288 the instrument.

289

290 *2.5.2 DNA analysis*

291 Methods and initial results from DNA analysis from these high volume filters were published by
292 Huffman et al. (2013). Briefly, fungal diversity was determined by previously optimized methods for
293 DNA extraction, amplification, and sequence analysis of the internal transcribed spacer regions of
294 ribosomal genes from the high volume filter samples (Fröhlich-Nowoisky et al., 2012; Fröhlich-
295 Nowoisky et al., 2009). Upon sequence determination, fungal sequences were compared with known
296 sequences using the Basic Local Alignment Search Tool (BLAST) at the National Center for

297 Biotechnology (NCBI) and identified to the lowest taxonomic rank common to the top BLAST hits after
298 chimeric sequences had been removed. When sequences displayed >97% similarity, they were grouped
299 into operational taxonomic units (OTUs).

300

301 *2.5.3 Endotoxin and glucan analysis*

302 Sample preparation for quantification of endotoxin and (1→3)-β-D-glucan included extraction of
303 5 punches (0.5 cm² each) of the quartz filters with 5.0 mL of pyrogen-free water (Associates of Cape Cod
304 Inc., East Falmouth, MA, USA), utilizing an orbital shaker (300 rpm) at room temperature for 60 min,
305 followed by centrifuging for 15 min (1000 rpm). One-half mL of supernatant was submitted to a kinetic
306 chromogenic limulus amoebocyte lysate (Chromo-LAL) endotoxin assay (Associates of Cape Cod Inc.,
307 East Falmouth, MA, USA) using a ELx808IU (BioTek Instrument Inc., Winooski, VT, USA) incubating
308 absorbance microplate reader. For (1→3)-β-D-glucan measurement, 0.5 mL of 3 N NaOH was added to
309 the remaining 4.5 mL of extract and the mixture was agitated for 60 min. Subsequently, the solution was
310 neutralized to pH 6–8 by addition of 0.75 mL of 2 N HCl. After centrifuging for 15 min (1→3)-β-D-
311 glucan concentration was determined in the supernatant using the GlucateLL® LAL kinetic assay
312 (Associates of Cape Cod, Inc., East Falmouth, MA). The minimum detection limits (MDLs) and
313 reproducibility were 0.046 Endotoxin Units (EU) m⁻³ ± 6.4% for endotoxin and 0.029 ng m⁻³ ± 4.2% for
314 (1→3)-β-D-glucan, respectively. Laboratory and field blank samples were analyzed as well, with lab
315 blank values being below detection limits, while field blank values were used to subtract background
316 levels from sample data. More details about the bioassays can be found elsewhere (Chow et al., 2015a).

317

318 **2.6 Meteorology and wetness sensors**

319 Meteorological data were recorded by a variety of sensors located at the site. Precipitation was recorded
320 by a laser optical disdrometer (PARTICLE SIZE and VELOCITY “PARSIVEL” sensor; OTT Hydromet
321 GmbH, Kempten, Germany) and separately by a tipping bucket rain gauge. The disdrometer provides
322 precipitation occurrence, rate, and physical state (rain or hail) by measuring the magnitude and duration
323 of disruption to a continuous 780 nm laser that was located in a tree clearing (Fig. 1), while the tipping
324 bucket rain gauge measures a set amount of precipitation before tipping and triggering an electrical pulse.
325 A leaf wetness sensor (LWS; Decagon Devices, Inc., Pullman, WA), provided a measurement of
326 condensed moisture by measuring the voltage drop across a leaf surface to determine a proportional
327 amount of water on or near the sensor. Additional details of these measurements can be found in Huffman
328 et al. (2013) and Ortega et al. (2014).

329

330 **3. Results and Discussion**

331

332 **3.1 Categorization and characteristic differences of Dry and Rainy periods**

333 Increases in PBAP concentration have been frequently associated with rainfall (e.g. Bigg et al.,
334 2015; Faulwetter, 1917; Hirst and Stedman, 1963; Jones and Harrison, 2004; Madden, 1997). Fungal
335 polyols have also been reported to increase after rain and have been used as indicators of increased fungal
336 spore release (Liang et al., 2013; Lin and Li, 2000; Zhu et al., 2015). Recently it was shown that the
337 concentration of fluorescent aerosol particles (FAP) measured during BEACHON-RoMBAS increased
338 dramatically during and after periods of rain (Crawford et al., 2014; Huffman et al., 2013; Schumacher et
339 al., 2013) and that these particles were associated with high concentrations of ice nucleating particles that
340 could influence the formation and evolution of mixed-phase clouds (Huffman et al., 2013; Prenni et al.,
341 2013; Tobo et al., 2013). It was observed that a mode of smaller fluorescent particles (2-3 μm) appeared
342 during rain episodes, and several hours after rain ceased a second mode of slightly larger fluorescent
343 particle (4-6 μm) emerged, persisting for up to 12 h (Huffman et al., 2013). The first mode was
344 hypothesized to result from mechanical ejection of particles due to rain splash on soil and vegetated
345 surfaces, and the second mode was suggested as actively emitted fungal spores (Huffman et al., 2013).
346 While the UV-APS and WIBS each provide data at high enough time resolution to see subtle changes in
347 aerosol concentration, the temporal resolution of the chemical tracer analysis was limited to 4-48 h

348 periods defined by the collection time of the high volume sampler. To compare the measurement results
349 across the sampling platforms, UV-LIF measurements were averaged to the lower time resolution of the
350 filter sampler periods, and the periods were grouped into three broad categories: Rainy, Dry, and Other, as
351 will be defined below.

352 Time periods were wetness-categorized in two steps: first at 15 min resolution and then averaged
353 for each individual filter sample. During the first stage of categorization each 15 min period was
354 categorized into one of four groups: rain, post-rain, dry, or other. To categorize each filter period, an
355 algorithm was established utilizing UV-APS fluorescent particle fraction and accumulated rainfall. The
356 ratio of integrated number of fluorescent particles to total particles was used as a proxy for the increased
357 emission of biological particles. Figure 2a presents a time series of the size-resolved fluorescent particle
358 concentration, showing increases during rain periods in dark red. A relatively consistent diurnal cycle of
359 increased FAP concentration in the 2-4 μm range is apparent almost every afternoon, which corresponds
360 to near daily afternoon rainfall during approximately the first half of the measurement period.
361 Disdrometer and tipping bucket rainfall measurements were each normalized to unity and summed to
362 produce a more robust, unitless measure of rainfall rate, because it was observed that often only one of the
363 two systems would record a given light rain event. If a point was described by total rainfall accumulation
364 greater than 0.201 it was flagged as rain. A point was flagged as post-rain if it immediately followed a
365 rain period and also exhibited a fluorescent particle fraction greater than 0.08. The purpose of this
366 category was to reflect the observation that sustained, elevated concentrations of FAP persisted for many
367 hours even after the rain rate, RH, and leaf wetness returned to pre-rain values. The only measurement
368 that adequately reflected this scenario was of the fluorescent particles measured by UV-APS and WBS
369 instruments. The post-rain flag was continued until the fluorescent particle fraction fell below 0.08 or if it
370 started to rain again (with calculated rain values greater than 0.201). Points were flagged as dry periods if
371 they exhibited rainfall accumulation and fluorescent particle fraction below the thresholds stated above.
372 Several periods were not easily categorized by this system and were considered in a fourth category as
373 other. This occurred when fluorescent particle fraction above the threshold value was observed with no
374 discernable rainfall.

375
376 Once wetness categories were assigned by the algorithm at 15 min resolution, each high volume
377 filter sample was categorized by a similar nomenclature, but using only three categories. These were
378 defined as Dry, Rainy (combination of rain and post rain categories), or Other based on the relative time
379 fraction in each of the four original 15 min categories. For each sample, if a given category represented
380 more than 50% of the 15 min periods, the sample was assigned to that category. Despite the effort to
381 categorize samples systematically, several sample periods (5 of 35) appeared mis-categorized by looking
382 at FAP concentration, rainfall, RH, and leaf wetness in more detail. In some circumstances, this was
383 because light rainfall produced observable increases in FAP, but without exceeding the rainfall threshold.
384 Or in other circumstances a period of rainfall occurred at the very end or just before the beginning of a
385 sample, and so the many-hour period was heavily influenced by aerosol triggered by a period of rain just
386 outside of the sample time window. As a result, several samples were manually re-categorized as
387 described here. Samples 20 and 21 (Table S1) were four-hour samples that displayed high relative
388 humidity and rainfall, thus samples were originally characterized as Rainy. This period was described by
389 an extremely heavy rain downpour (7.5 mm in 15 min), however, that seemingly placed the samples in a
390 different regime of rain-aerosol dynamics than the other Rainy samples and so these two samples were
391 moved to the Other category. Sample 23, originally Rainy, presented a FAP fraction marginally above the
392 0.08 threshold, but visually displayed a trend dissimilar to other post-rain periods and so was re-
393 categorized as Dry. Sample 28 showed no obvious rainfall, but the measurement team observed persistent
394 fog in three consecutive mornings (Samples 25, 27, 28), and the concentration of fluorescent particles (2-
395 6 μm) suggested a source of particles not influenced by rain, and so this Rainy sample was re-categorized
396 as Other. Sample 38 displayed a fluorescent number ratio just below the threshold value, and was first
397 categorized as Dry, however, the measurement team observed post-rain periods at the beginning and end
398 of the sample, so the sample were re-categorized as Other. For all samples other than these five, the

399 categorization was determined using the majority (> 0.50) of the 15 min periods. In no cases other than
400 the five that were re-categorized was the highest category fraction less than 0.50 of the sample time. Note
401 that we have chosen to capitalize Rainy, Dry, and Other to highlight that we have rigorously defined the
402 period using the characterization scheme described above and to separate the nomenclature from the
403 general, colloquial usage of the terms. Wetness category assignment for each high volume filter sample
404 period is shown in Figure 2 as a background color (brown for Dry samples, green for Rain-influenced
405 samples, and pink for Other samples) and Table S1.

406
407 To validate the qualitative differences between wetness categories described in the last section,
408 we present observations about each of these groupings. First, we organized the WIBS data according to
409 the particle categories introduced by Perring et al. (2015). By this method, every fluorescent particle
410 detected by the WIBS can be defined uniquely into one of seven categories (i.e. A, AB, ABC and so on).
411 By plotting the relative fraction of fluorescent particles described by each particle type, temporal
412 differences between measurement periods can be observed, as shown in Figure 2e. To a first
413 approximation, this analysis style allows for coarse discrimination of particle types. For example, a given
414 population of particles would ideally exhibit a consistent fraction of particles present in the different
415 particle categories as a function of time. By this reasoning, sample periods categorized as Dry (most of
416 the latter half of the study; brown bars in Fig. 2) would be expected to have a self-consistent particle type
417 trend, whereas sample periods categorized as Rainy (most of the first half of the study; green bars in Fig.
418 2) would have a self-consistent particle type trend, but different from the Dry samples. This is broadly
419 true. During Rainy periods as seen in Figure 3a, there is a relatively high fraction ($> 65\%$) of ABC
420 particles (light blue) and a relatively low fraction ($< 15\%$) in BC (purple) and C (yellow) type particles,
421 suggesting heavy influence from the FL1 channel. In contrast, during Dry periods the fraction of ABC
422 particles (light blue) is reduced ($< 25\%$) while BC (purple) and C (yellow) type particles increase in
423 relative fraction ($> 30\%$ and $> 40\%$, respectively) suggested a diminished influence of FL1 channel.

424
425 It is important to note a few caveats here. First, the ability of the WIBS to discriminate finely
426 between PBAP types is relatively poor and it is still unclear exactly how different particle types would
427 appear by this analysis method. Particles of different kinds and from different sources are likely
428 convolved into a single WIBS particle type, which could either soften or enhance the relationships with
429 rain discussed here. Second, the assignment of particle types is heavily size-dependent and sensitive to
430 subtle instrument parameters, and so it is unclear how different instruments would present similar particle
431 types. For example, Hernandez et al. (2016) used two WIBS instruments and found differences in relative
432 fraction of particle categories for samples aerosolized in the lab. They reported fungal spores to be
433 predominately A, AB, and ABC type particles, whereas Rainy sample periods suggested to have heavy
434 fungal spore influence by Huffman et al. (2013) show predominantly C, BC, and ABC particle fraction.
435 These discrepancies may be due to the comparison of ambient particles to laboratory-grown cultures. The
436 highly controlled environment of a laboratory may not always accurately represent the humidity
437 conditions in which fungal spore release occurs in this forest setting (Saari et al., 2015). This could impact
438 the fluorescence properties of fungal spore particles that have different amounts of adsorbed or associated
439 water (Hill et al., 2009; 2013; 2015). More likely, however, is that the WIBS-3 used here exhibits
440 differences in fluorescence sensitivity from the WIBS-4A used by Hernandez et al. (2016). Even a slight
441 increase in sensitivity in the FL3 channel with respect to the FL1 or FL2 channels could explain the shift
442 here towards particles with C-type fluorescence. One piece of evidence for this is the quantitative
443 comparison of particle measurements presented by the UV-APS and WIBS-3 instruments co-deployed
444 here (Fig. 4). The number concentration of particle exhibiting fluorescence above the FL2 baseline of the
445 WIBS-3 is approximately consistent with the number of fluorescent particles measured by the UV-APS,
446 and significantly below the concentration of FL3 particles. The UV-APS number concentration shows the
447 highest correlation with the WIBS-3 FL2 channel: during Rainy periods, $R^2 = 0.70$; Dry, $R^2 = 0.82$; Other,
448 $R^2 = 0.92$. These observations are in stark contrast to the trends reported by Healy et al. (2014) that the
449 UV-APS fluorescent particle concentration correlated most strongly with the WIBS-4 FL3 and that the

450 number concentration of FL3 was the lowest out of all three channels. Given that the FL3 channel of the
451 WIBS and the UV-APS probe cover similar excitation and emission wavelengths it is expected that these
452 two channels should correlate well. Based on these data, we suggest that the WIBS-3 utilized here may
453 present a very different particle type break-down than if a WIBS-4 had been used. So, while caution is
454 recommended when comparing the relative break-down of WIBS particle categories shown here (Fig. 3)
455 with other studies, the data are internally self-consistent, and comparing qualitative differences between,
456 e.g. Rainy and Dry periods is expected to be robust. The main point to be highlighted here is that there is
457 indeed a qualitative difference in particles present in the three wetness categories, as averaged and shown
458 in Figure 3a, which generally supports the effort to segregate these samples.

459
460 Further evidence that there is a qualitative difference in the three wetness categories is shown
461 using molecular genetic analysis (Figs. 3b, c). The analysis of fungal DNA sequences from 21 of the high
462 volume samples found 406 operational taxonomic units (OTUs), belonging to different fungal classes and
463 phyla. When organized by wetness type it was observed that 106 of these occurred only on Rainy
464 samples, 148 of these occurred on Dry samples, and 37 on Other samples, with some fraction occurring in
465 overlaps of each (Fig. 3c). This shows that the number of OTUs observed uniquely in either the Rainy or
466 Dry periods is greater than the number of OTUs present in both wetness types, suggesting that the fungal
467 communities in each grouping are relatively distinct. Further, Figure 3b shows a break-down of fungal
468 taxonomic groupings for each wetness group. This analysis shows that there is a qualitative difference in
469 taxonomic break-down between periods of Rainy and Dry. Specifically, during Dry periods there is an
470 increased fraction of *Pucciniomycetes* (green bar, Fig. 3c), *Chytridiomycetes* (yellow), *Sordariomycetes*
471 (orange), and *Eurotiomycetes* (pink) when compared to the Rainy periods.

472

473 **3.2 Atmospheric mass concentration of arabitol, mannitol, and fungal spores**

474 To estimate fungal spore emission to the atmosphere, the concentration of arabitol and mannitol
475 (Fig. 5a, b, Table S2) in each aerosol sample was averaged for all samples in each of the three wetness
476 categories. The average concentration of arabitol collected on Rainy TSP samples ($35.2 \pm 10.5 \text{ ng m}^{-3}$)
477 increased by a factor of 3.3 with respect to Dry samples, and the average mannitol concentration on Rainy
478 samples was higher by a factor of 3.7 ($44.9 \pm 13.8 \text{ ng m}^{-3}$). Figures 5a, b show the concentration
479 variability for each wetness category, observed as the standard deviation from the distribution of
480 individual samples. For each polyol, there is no overlap in the ranges shown, including the outliers of the
481 Rainy and Dry category, suggesting a definitive and conceptually distinct separation between dry periods
482 and those influenced by rain. The concentrations observed during Other periods is between those of the
483 Dry and Rainy averages, as expected, given the difficulty in confidently assigning these uniquely to one
484 of these categories. The observations here are roughly consistent with previous reports of polyol
485 concentration, despite differences in local fungal communities and concentrations. For example,
486 Rathnayake et al. (2016a) observed 30.2 ng m^{-3} arabitol and 41.3 ng m^{-3} mannitol in PM_{10} samples
487 collected in rural Iowa, USA. In addition, Zhang et al. (2015) reported arabitol and mannitol
488 concentrations in PM_{10} samples of 44.0 and 71.0 ng m^{-3} , respectively, from a study in the mountains on
489 Hainan Island off the coast of Southern China. More recently, Yue et al. (2016) studied a rain event in
490 Beijing and observed increased polyol concentrations at the onset of the rain. The observed mannitol
491 concentration (45 ng m^{-3}) was approximately consistent with observations reported here and with
492 previous reports, while the arabitol concentration values observed were approximately an order of
493 magnitude lower (0.3 ng m^{-3}).

494

495 The square of the correlation coefficient (R^2) here between concentration values of arabitol and
496 mannitol during Rainy samples is very high (0.839; Table 1) suggesting that arabitol and mannitol
497 originated primarily from the same source, likely active-discharge fungal spores. The correlation is
498 similar to the $0.87 R^2$ reported by Bauer et al. (2008a) and the $0.93 R^2$ reported by Graham et al. (2003).
499 In contrast, the same correlation between mannitol and arabitol concentrations, but for Dry samples is
500 relatively low (0.312). This is consistent with reports that arabitol can be used more specifically as a spore

501 tracer, but that mannitol has additional atmospheric sources besides fungal spores. The same correlation
502 was also performed between arabitol or mannitol and other molecular tracers (endotoxins and (1→3)-β-
503 D-glucan), but all R^2 value were less than 0.43, suggesting that the endotoxins and glucans analyzed were
504 not emitted uniquely from the same sources as arabitol and mannitol.

505
506 Results from the two UV-LIF instruments were averaged over high volume sample periods, and a
507 correlation analysis was performed between tracer mass and fluorescent particle mass showing positive
508 correlations in all cases. The FAP mass from the UV-APS shows high correlation with the fungal polyols
509 during Rainy periods, with R^2 of 0.732 and 0.877 for arabitol and mannitol, respectively (Table 2; Figure
510 5c, d). The same tracers correlate poorly with the UV-APS during Dry conditions. This is expected,
511 because *ascomycetes* and *basidiomycetes* emitted by wet discharge methods are the only fungal spores
512 reported to be associated with arabitol and mannitol (Elbert et al., 2007; Feofilova, 2001; Lewis and
513 Smith, 1967). This high correlation suggests that the UV-APS does a good job of detecting these wet-
514 discharge spores, and corroborates previous statements that particles detected in ambient air by the UV-
515 APS are often predominately fungal spores (Healy et al., 2014; Huffman et al., 2013; Huffman et al.,
516 2012). In contrast, the low slope value and the poor correlation during Dry periods suggest that the UV-
517 APS is also sensitive to other kinds of particles, as designed. The small positive x-offset (FAP mass;
518 Table S2, Figs. 5c,d) during Rainy periods is likely due to particles that are too weakly fluorescent to be
519 detected and counted by the UV-APS, which is consistent with observations made in Brazil (Huffman et
520 al., 2012).

521
522 Particle mass from WIBS C11, assigned to fungal spores (Crawford et al., 2015), also correlate
523 strongly with the same two molecular tracers. Both Rainy periods (R^2 0.824) and Dry periods (R^2 0.764)
524 correlate well with arabitol (Fig. 5e), while mannitol (Fig. 5f) only shows a strong correlation during the
525 Rainy periods (R^2 0.799). Mannitol is a common polyol in higher plants while arabitol is only found in
526 fungal spores and lichen (Lewis and Smith, 1967). So the strong correlation of each polyol with UV-LIF
527 mass during Rainy periods when actively-discharged spores are expected to dominate and the similarly
528 strong correlations associated with arabitol suggest that the C11 cluster does a reasonably good job of
529 selecting fungal spore particles. The poor correlation between mannitol and C11 during dry periods
530 illustrates that the background mannitol concentration is likely not due to fungal spores alone, but has
531 contribution from other higher plants that contain mannitol. Particle concentrations detected by individual
532 WIBS channels and in the other cluster were also compared with polyol concentrations, but each
533 correlation is relatively poor compared to that with respect to C11. As seen in Table 2 and Figures S2-S3,
534 correlations in FL1, 2, and 3 with arabitol are poor (<0.4) in the Dry category and good ($0.4 < R^2 < 0.7$) in
535 the Rainy category. For mannitol, all the UV-LIF instruments show high correlation (>0.7) in all cases.
536 This is likely due to mannitol being a non-specific tracer and suggests that the majority of UV-LIF
537 particles observed during all periods was dominated by PBAP.

538 539 **3.3 Estimated number concentration of fungal spore aerosol**

540 Bauer et al. (2008a) reported measurements of fungal spore number concentration in Vienna,
541 Austria using epifluorescence microscopy and also measured fungal tracer mass collected onto filters in
542 order to estimate the mass of arabitol (1.2 to 2.4 pg spore⁻¹) and mannitol (0.8 to 1.8 pg spore⁻¹) associated
543 with each emitted spore. Bauer et al. (2008a) and Yttri et al. (2011b) reported ratios of mannitol to
544 arabitol of approximately 1.5 (\pm standard deviation of 26%) and 1.4 ± 0.3 , respectively. Our
545 measurements show slightly lower ratios of mannitol to arabitol, but that the ratio is dependent on
546 wetness category; Rainy, 1.29 ± 0.17 ; Dry, 1.12 ± 0.23 ; and Other, 1.24 ± 0.54 . The mannitol to arabitol
547 ratio would be expected to vary as a function of fungal population present in the aerosol, whether between
548 different wetness periods at a given location or between different physical localities.

549
550 Using the approximate mid-point of the Bauer et al. (2008a) reported ranges, 1.7 pg mannitol per
551 spore and 1.2 pg arabitol per spore, atmospheric number concentrations of spores collected onto the high

552 volume filters were calculated from the polyol mass concentrations measured here. Based on these values,
553 and assuming all polyol mass originated with spore release, the mass concentration averages (Fig. 5) were
554 converted to fungal spore number concentrations (Fig. 6). The trends of spore concentration averages are
555 the same as with the polyol mass, because the numbers were each multiplied by the same scalar value.
556 After doing so, the analysis reveals an estimated spore concentration during Dry periods of 0.89×10^4 (\pm
557 0.21) spores m^{-3} using the arabitol concentration and 0.70×10^4 (± 0.19) spores m^{-3} using the mannitol
558 concentration (Table 3). The estimated concentration of spores increased approximately three-fold during
559 Rainy periods to 2.9×10^4 (± 0.8) spores m^{-3} (arabitol estimate) and 2.6×10^4 (± 0.8) spores m^{-3} (mannitol
560 estimate) (Figure 6a, b). These estimates match well with estimates reported by Spracklen and Heald
561 (2014), who modeled the concentration of airborne fungal spores across the globe as an average of $2.5 \times$
562 10^4 spores m^{-3} , with approximately 0.5×10^4 spores m^{-3} over Colorado.

563
564 The UV-LIF instruments discussed here are number-counting techniques and in this instance have
565 been applied as spore counters. As a first approximation, each particle detected by the UV-APS was
566 assumed to be a fungal spore with the same properties used in the assumptions by Bauer et al. (2008a).
567 Figures 6d, e, g, h show correlations of fungal spore number concentration estimated from polyol mass on
568 the y-axes and from UV-LIF measurements on the x-axes. The first, and most important observation is
569 that the estimated fungal spore concentration from each technique is on the same order of magnitude, 10^4
570 m^{-3} . Looking at individual correlations reveals a finer layer of detail. These results show that the number
571 concentration of fungal spores estimated by the UV-APS is greater than the number of fungal spores
572 estimated by the tracers, as evidenced by slope values of approximately 0.2 and 0.35 for Rainy and Dry
573 conditions, respectively (Tables S3, Figs. 6d, e). Again, this suggests that the UV-APS detects fungal
574 spores as well as other types of fluorescent particles. The R^2 values (~ 0.5) during Rainy periods indicate
575 that the additional source of particles detected by the UV-APS is likely to have a similar source, such as
576 PBAP mechanically ejected from soil and vegetative surfaces with rain-splash (Huffman et al., 2013).
577 The magnitude of the over-estimation is higher during Dry periods, which would be expected because
578 Rainy periods exhibited much higher particle number fractions associated with polyol-containing spores.

579
580 The C11 cluster from WIBS data shows correlations with estimated fungal spores from arabitol
581 and mannitol that have slope much closer to 1.0 than correlations with UV-APS number (Figs. 6g, h,
582 Table S3). For example, the slope of the C11 correlations with each polyol during Rainy periods is
583 approximately 0.87. This suggests only a 13% difference between the spore concentration estimates from
584 the two techniques during Rainy periods. The average number concentration of C11 during Rainy periods
585 is 1.6×10^4 (± 0.8) spores m^{-3} . In both cases the slopes with respect to C11 is greater than 1.0 during Dry
586 periods, suggesting that the cluster method may be missing some fraction of weakly fluorescent particles.
587 Huffman et al. (2012) similarly suggests that that particles that are weakly fluorescent may be below the
588 detection limit of the instrument, and Healy et al. (2014) suggested that both UV-APS and WIBS-4
589 instruments significantly under-count the ubiquitous *Cladosporium* spores that are most common during
590 dry weather and often peak in the afternoon when RH is low (De Groot, 1968; Oliveira et al., 2009).
591 Fundamentally, however, the results from the UV-APS, and even more so the numbers reported by the
592 clustering analysis by Crawford et al. (2015), reveal broadly similar trends with the numbers estimated
593 from polyol-to-spore values reported by Bauer et al. (2008a).

594
595 The fungal culture samples show similar division during Rainy and Dry periods as arabitol and
596 mannitol concentrations (Figure 6c), with an increase of approx. 1.6 during Rainy periods. The trend of a
597 positive slope with respect to the UV-LIF measurements is also similar between the tracer and culturing
598 methods. In general, however, the R^2 value correlating CFU to fungal spore number calculated from UV-
599 LIF number is lower than between tracers and UV-LIF numbers (Table 2). This is not unexpected for
600 several reasons. First, the short sampling time of the culture samples (20 min) leads to poor counting
601 statistics and high number concentration variability, whereas each data point from the high volume air
602 samples represents a period of 4 – 48 hours. Second, culture samplers, by their nature, only account for

603 culturable fungal spores. It has been estimated that as low as 17% of aerosolized fungal species are
604 culturable, and so it is expected that the CFU concentration observed is significantly less than the total
605 airborne concentration of spores (Bridge and Spooner, 2001; Després et al., 2012). Nonetheless, the
606 culturing analysis here supports the tracer and UV-LIF analyses and the most important trends are
607 consistent between all analysis methods. The concentration of fungal spores is higher during the Rainy
608 periods, and there is a positive correlation between both tracer and CFU concentration and UV-LIF
609 number.

610
611 In pristine environment, such as the Amazon, supermicron particle mass has been found to consist
612 of up to 85% biological material (Pöschl et al., 2010). Total particulate matter mass was calculated here
613 from the UV-APS number concentrations (m^{-3}) and converted to mass for particles of aerodynamic
614 diameter 0.5 – 15 μm . In only this case a density of 1.5 g cm^{-3} was utilized to calculate a first
615 approximation of total particle mass to which all other mass measurements were compared. An average
616 TSP mass density of 1.5 g cm^{-3} was utilized, because organic aerosol is typically estimated with density <
617 1.0 g cm^{-3} , biological particles are often assumed to have ca. 1.0 g cm^{-3} density, and mineral dust
618 particles have densities of up to ca. 3.5 g cm^{-3} (Dexter, 2004; Tegen and Fung, 1994). Fungal spore mass
619 was estimated here using the fungal spore concentrations calculated from arabitol and mannitol mass (Fig.
620 6) and then using an estimated 33 pg reported by Bauer et al. (2008b) as an average mass per spore.
621 Dividing the resultant fungal spore mass by total particulate mass provides a relative mass fraction for
622 each high volume sample period. These calculations suggest that fungal spores represent approximately
623 23% \pm 9 (using arabitol) or 21% \pm 8 (using mannitol) of total particulate mass during Rainy periods
624 (Table 3, Figure 7). This represents a nearly 6 fold increase in percentage compared to Dry periods (4.8%
625 \pm 1.4 and 3.7% \pm 1.1, respectively). A similar increase during Rainy periods was also seen in the mass
626 fraction of fungal cluster C11, which represented 17% \pm 10 of the particle mass during Rainy and 2% \pm 1
627 during Dry periods (Table S4).

628 **3.5 Variations in endotoxin and glucan concentrations**

629 Endotoxins measured in the atmosphere are uniquely associated with gram-negative bacteria
630 (Andreae and Crutzen, 1997). Here, we show correlations between total endotoxin mass and WIBS Cl_{Bact},
631 which was assigned by Crawford et al. (2015) to be bacteria due to the small particle size (< 1 μm) and
632 high correlation with rain. These assignment of particle type to this set of clusters is quite uncertain,
633 however, and should be treated loosely. The correlation between endotoxin mass and UV-APS and the
634 WIBS clusters was very poor, in most cases $R^2 < 0.1$ (Table 2, Figure 8), suggesting no apparent
635 relationship. Analysis of bacteria by both UV-LIF techniques is hampered by the fact that bacteria can be
636 < 1 μm in size and because both instruments detect particles with decreased efficiency at sizes below 0.8
637 μm . So weak correlations may not have been apparent due to reduced overlap in particle size. Despite the
638 lack of apparent correlation between the techniques, the relatively variable endotoxin concentrations were
639 elevated during Rainy periods, consistent with Jones and Harrison (2004), who showed that bacteria
640 concentration were elevated after rainy periods.

641
642 Glucans, such as (1 \rightarrow 3)- β -D-glucan, are components of the cell walls of pollen, fungal spores,
643 plant detritus, and bacteria (Chow et al., 2015b; Lee et al., 2006; Stone and Clarke, 1992). In contrast to
644 the observed difference in endotoxin concentration during the different wetness periods, however, (1 \rightarrow 3)-
645 β -D-glucan showed no correlations with UV-LIF concentrations (Table 2) and no differentiation during
646 the different wetness periods.

647 **4. Conclusions**

648 Increased concentrations of fluorescent aerosol particles and ice nuclei attributed to having
649 biological origin were observed during and immediately after rain events throughout the BEACHON-
650 RoMBAS study in 2011 (Huffman et al., 2013; Prenni et al., 2013; Schumacher et al., 2013). Here we
651 expand upon the previous reports by utilizing measurements from two commercially available UV-LIF
652 instruments, of several molecular tracers extracted from high volume filter samples, and from a culture-

654 based sampler in order to compare three very different methods of atmospheric fungal spore analysis.
655 This study represents the first reported correlation of UV-LIF and molecular tracer measurements and
656 provided an opportunity to understand how an important class of PBAP might be influenced by periods of
657 rainy and dry weather. We found clear patterns in the fungal molecular tracers, arabitol and mannitol,
658 associated with Rainy conditions that are consistent with previous findings (Bauer et al., 2008a; Elbert et
659 al., 2007; Feofilova, 2001). Fungal polyols increased 3-fold over Dry conditions during Rainy weather
660 samples, with arabitol concentration of $35.2 \pm 10.5 \text{ ng m}^{-3}$ and mannitol concentration of $44.9 \pm 13.8 \text{ ng}$
661 m^{-3} . Additionally, the very high correlation of the fungal tracers with WIBS C11 ($R^2 > 0.8$ in many cases)
662 provides support for its assignment by Crawford et al. (2015) to fungal spores. Similarly, the UV-APS
663 correlates well with fungal tracers, however over-counts the number concentration estimated from the
664 tracers, confirming that the UV-APS is sensitive also to other types of particles beyond fungal spores, as
665 expected. The estimated spore count from the WIBS C11 concentration was within ~13% of the spore
666 count estimated by the tracer method, with concentrations ranging from $1.6 - 2.9 \times 10^4 \text{ spores m}^{-3}$. These
667 values are broadly consistent with concentrations modeled by, e.g. Spracklen and Heald (2014), Hoose et
668 al. (2010), and Hummel et al. (2015). These spore counts represent 17-23% of the total particle mass
669 during Rainy conditions and 2-4% during Dry conditions. Culture-based sampling also shows a similar
670 relationship between CFU and UV-LIF concentrations and an increase of ~1.6 between Dry and Rainy
671 conditions. Despite the fact that the tracer and UV-LIF approaches to estimating atmospheric fungal spore
672 concentration are fundamentally different, they provide remarkably similar estimates and temporal trends.
673 With further improvements in instrumentation and analysis methods (e.g. advanced clustering algorithms
674 applied to UV-LIF data), the ability to reliably discriminate between PBAP types is improving. As we
675 have shown here, this technology represents a potential for monitoring approximate fungal spore mass
676 and for contributing improved information on fungal spore concentration to global and regional models
677 that to this point has been lacking (Spracklen and Heald, 2014).

678
679

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701

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984 **Tables and Figures:**

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			Mass Concentration					
			Arabitol (ng m ⁻³)		Mannitol (ng m ⁻³)		(1→3)-β-D-glucan (pg m ⁻³)	
			Rainy	Dry	Rainy	Dry	Rainy	Dry
Mass Concentration	Mannitol (ng m ⁻³)	Rainy	<u>0.839</u>					
		Dry		0.312				
	(1→3)-β-D-glucan (pg m ⁻³)	Rainy	0.000		0.003			
		Dry		0.000		0.327		
	Endotoxins (EU m ⁻³)	Rainy	0.116		0.126		0.427	
		Dry		0.012		0.113		0.103

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987 **Table 1:** Square of correlation coefficients (R^2) comparing total mass concentration of molecular tracers

988 to each other. EU: endotoxin units. Boxes colored by coefficient value (**Bold Underline** > 0.7; 0.7 > **Bold**

989 > 0.4).

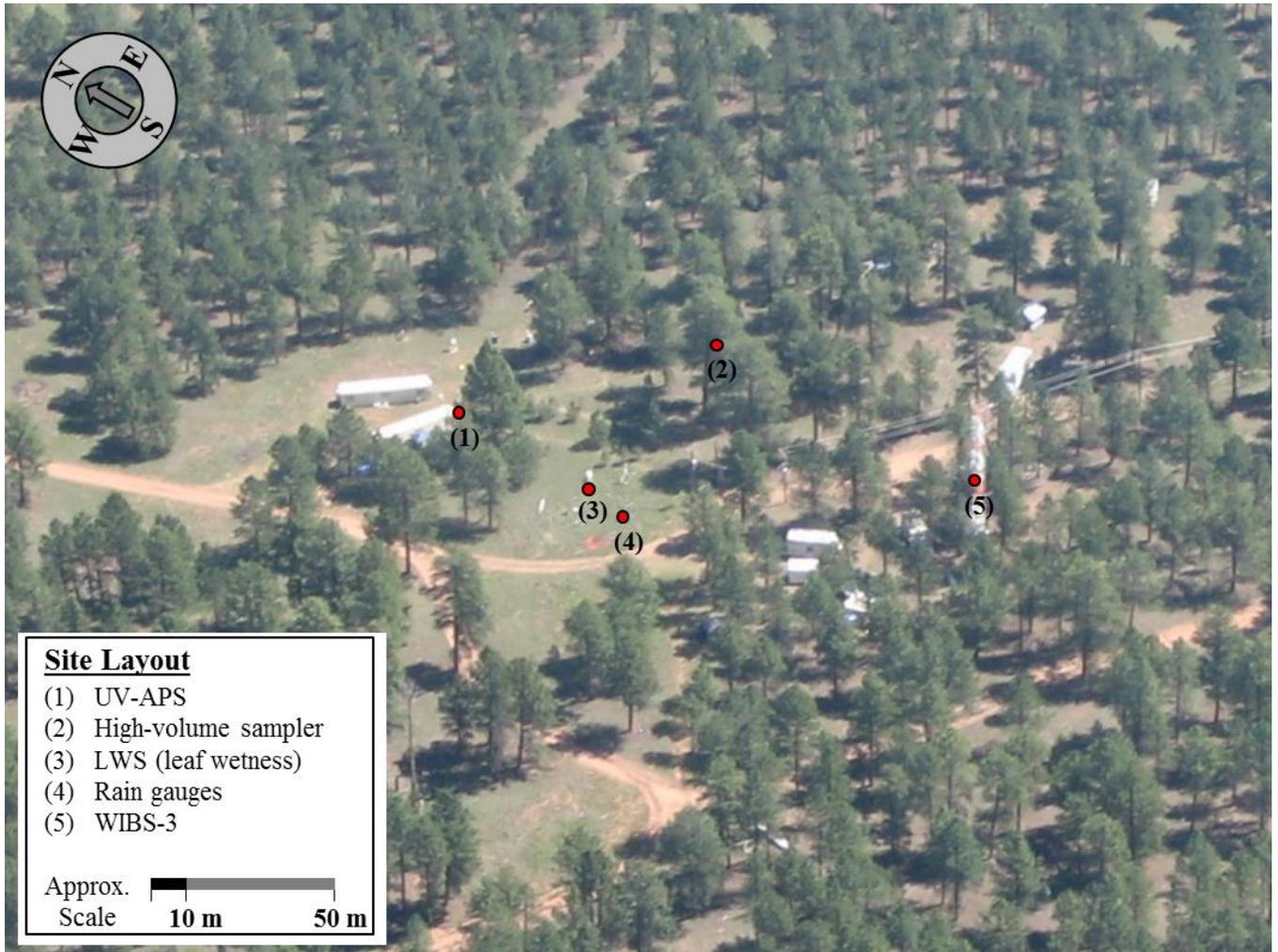
			Mass Concentration								Fungal Spore Number Concentration						
			Arabitol (ng m ⁻³)		Mannitol (ng m ⁻³)		(1→3)-β-D-glucan (pg m ⁻³)		Endotoxins (EU m ⁻³)		Arabitol (spores m ⁻³)		Mannitol (spores m ⁻³)		Colony Forming Units (CFU m ⁻³)		
			Rainy	Dry	Rainy	Dry	Rainy	Dry	Rainy	Dry	Rainy	Dry	Rainy	Dry	Rainy	Dry	
UV-LIF Mass or Number Concentration	UVAPS		<u>0.732</u>	0.127	<u>0.877</u>	0.160	0.006	0.012	0.153	0.067	0.483	0.278	0.504	0.571	0.469	0.491	
	WIBS	FL		0.554	0.250	<u>0.810</u>	0.255	0.128	0.010	0.068	0.066	0.159	0.200	0.088	0.314	0.330	<u>0.737</u>
		FL1		0.602	0.445	<u>0.819</u>	0.412	0.042	0.001	0.090	0.012	0.667	0.339	<u>0.863</u>	0.621	0.470	0.546
		FL2		0.617	0.248	<u>0.843</u>	0.342	0.092	0.001	0.039	0.094	0.485	0.302	0.442	0.340	0.560	0.543
		FL3		0.561	0.222	<u>0.818</u>	0.251	0.124	0.008	0.071	0.065	0.178	0.181	0.104	0.306	0.367	<u>0.736</u>
		Cl1		<u>0.824</u>	<u>0.764</u>	<u>0.799</u>	0.109	0.000	0.134	0.229	0.011	0.679	0.543	<u>0.775</u>	0.423	0.128	0.690
		Cl2		0.005	0.002	0.004	0.006	0.002	0.047	0.006	0.017	0.052	0.056	0.001	0.075	0.081	<u>0.930</u>
		Cl3		0.267	0.164	0.261	0.198	0.003	0.011	0.016	0.066	0.052	0.116	0.087	0.439	0.262	0.383
		Cl4		0.048	0.046	0.172	0.118	0.115	0.011	0.179	0.145	0.062	0.089	0.001	0.065	0.120	0.000
Cl _{Bact}								0.041	0.081								

990 **Table 2:** Square of correlation coefficients (R^2) comparing fluorescent particle measurements from UV-LIF instruments to measurements from
991 molecular tracers and direct-to-agar sampler. Columns marking tracer mass (top line) indicate correlations between time-averaged UV-LIF and
992 tracer mass concentrations (left side), and columns marking fungal spore number indicate correlations between fungal spore number
993 concentrations estimated from time-averaged UV-LIF and tracer or culture measurements (right side). FL1, FL2, FL3 represent individual
994 channels from the WIBS. FL represents particles exhibiting fluorescence in any channel. Cl1, Cl2, Cl3, Cl4 are clusters that estimate particle
995 concentrations as a mixture of various channels (Crawford et al., 2015). Cl_{Bact} is a sum of the “bacteria” clusters Cl2-4. Boxes colored by
996 coefficient value (**Bold Underline** > 0.7; 0.7 > **Bold** > 0.4).

	Mass Concentration						
	Arabitol (ng m ⁻³)	Mannitol (ng m ⁻³)	Erythritol (ng m ⁻³)	Levoglucozan (ng m ⁻³)	Glucose (ng m ⁻³)	Endotoxins (EU m ⁻³)	(1→3)-β-D- glucan (pg m ⁻³)
Dry	10.6 ± 2.5 n = 18	11.9 ± 3.2 n=18	0.840 ± 0.610 n=16	14.2 ± 10.7 n=15	38.7 ± 21.3 n=18	0.192 ± 0.0970 n=18	8.85 ± 7.68 n=18
Rainy	35.2 ± 10.5 n=11	44.9 ± 13.8 n=11	1.12 ± 0.38 n=3	12.4 ± 19.1 n=8	73.2 ± 50.5 n=11	1.43 ± 1.22 n=10	10.6 ± 8.2 n=11
Other	20.2 ± 8.9 n=6	22.7 ± 8.3 n=6	0.664 ± 0.515 n=6	9.21 ± 1.66 n=5	56.5 ± 39.2 n=6	0.311 ± 0.159 n=6	6.08 ± 6.08 n=6
	Mass Contribution (%)						
Dry	0.18 % ± 0.05 n=18	0.20 % ± 0.073 n=18	0.014 % ± 0.011 n=16	0.21 % ±0.17 n=15	0.67 % ±0.49 n=18		0.16 % ±0.16 n=18
Rainy	0.83 % ± 0.32 n=11	1.07 % ±0.44 n=11	0.032 % ±0.009 n=3	0.27 % ±0.41 n=8	1.60 % ±1.09 n=11		0.25 % ±0.21 n=11
Other	0.25 % ± 0.28 n=6	0.37 % ± 0.29 n=6	0.013 % ±0.015 n=6	0.15 % ±0.11 n=5	0.83 % ±0.64 n=6		0.12 % ±0.19 n=6
	Fungal Spore Number Concentration (m ⁻³)						
Dry	8900 ± 2100 n=18	6900 ± 1900 n=18					

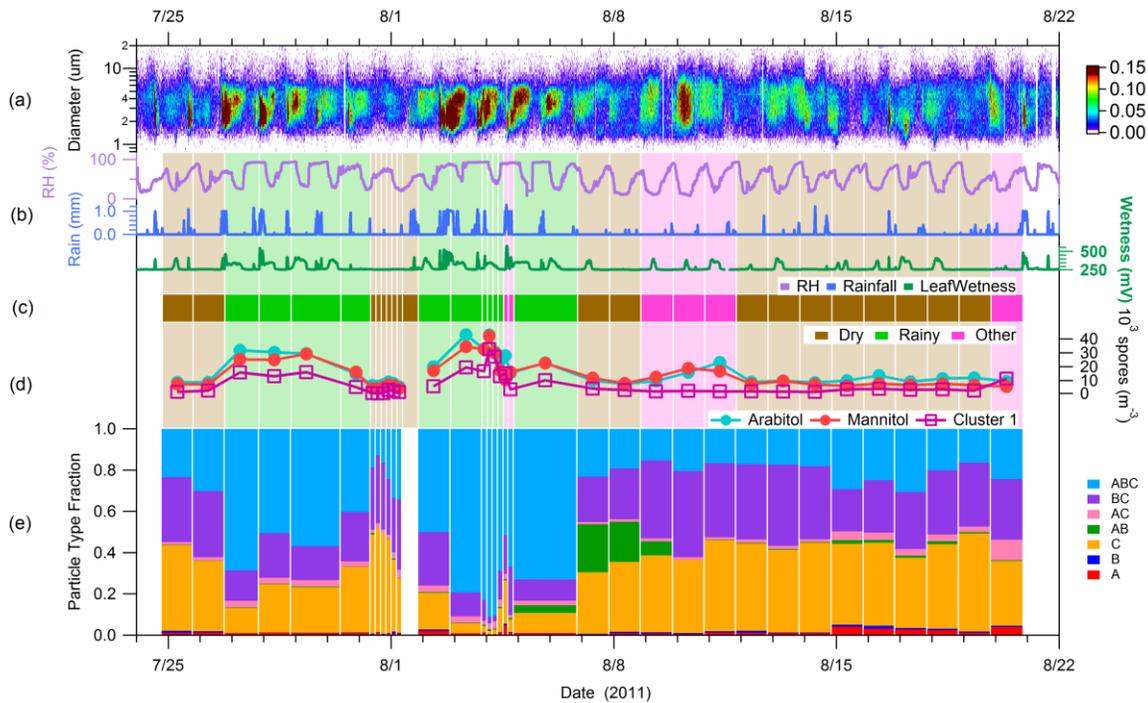
Rainy	29300 ± 8700 n=11	26400 ± 8100 n=11					
Other	16900 ± 7400 n=6	13400 ± 4900 n=6					
Fungal Spore Mass Contribution (%)							
Dry	4.8 % ± 1.43 n=18	3.7 % ± 1.1 n=18					
Rainy	22.9 % ±8.8 n=11	20.7 % ±8.5 n=11					
Other	9.8 % ± 7.7 n=6	7.3 % ± 5.6 n=6					

998 **Table 3:** Campaign-average concentrations of molecular tracers (measured) and fungal spores (number
999 concentration estimated from arabitol and mannitol mass). Each set of data broken into wetness
1000 categories. Values are mean ± standard deviation; *n* shows the number of samples used for averaging.
1001 Fungal spore mass contribution was based on the assumption by Bauer et al. (2008b) of 33 pg spore⁻¹.
1002 Total particulate matter mass calculated from UV-APS number concentration (m⁻³) and converted to mass
1003 over aerodynamic particle diameter range 0.5 – 15 µm using density of 1.5 g cm⁻³.

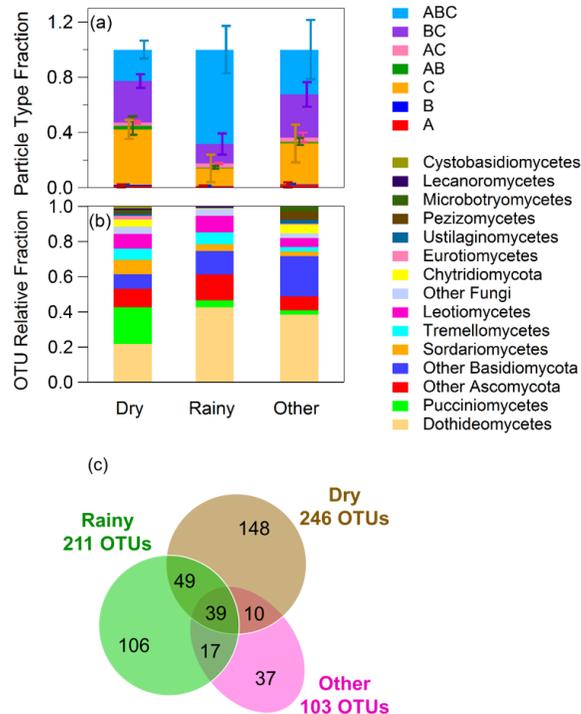


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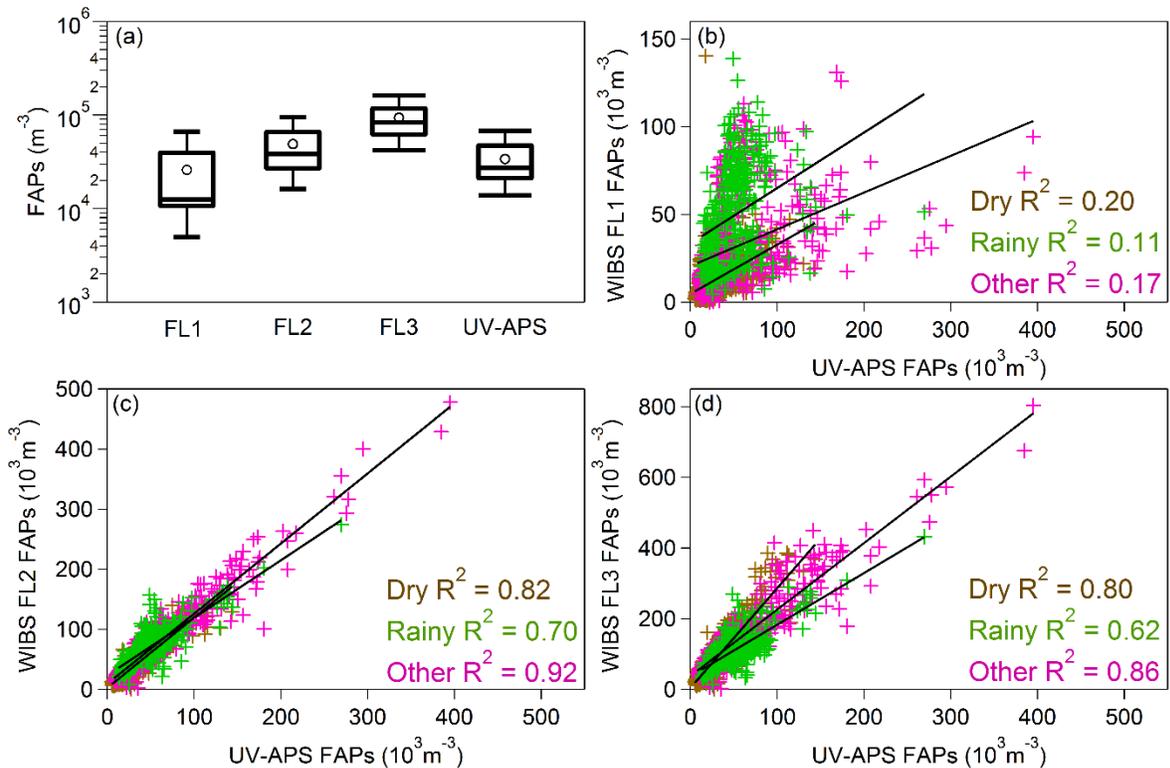
Figure 1: Aerial overview of BEACHON-RoMBAS field site at the Manitou Experimental Forest Observatory located northwest of Colorado Springs, CO. Locations of all instruments and sensors discussed here are marked and were located within a 50 m radius. Figure adapted from Figure 1a of Huffman et al. (2013)



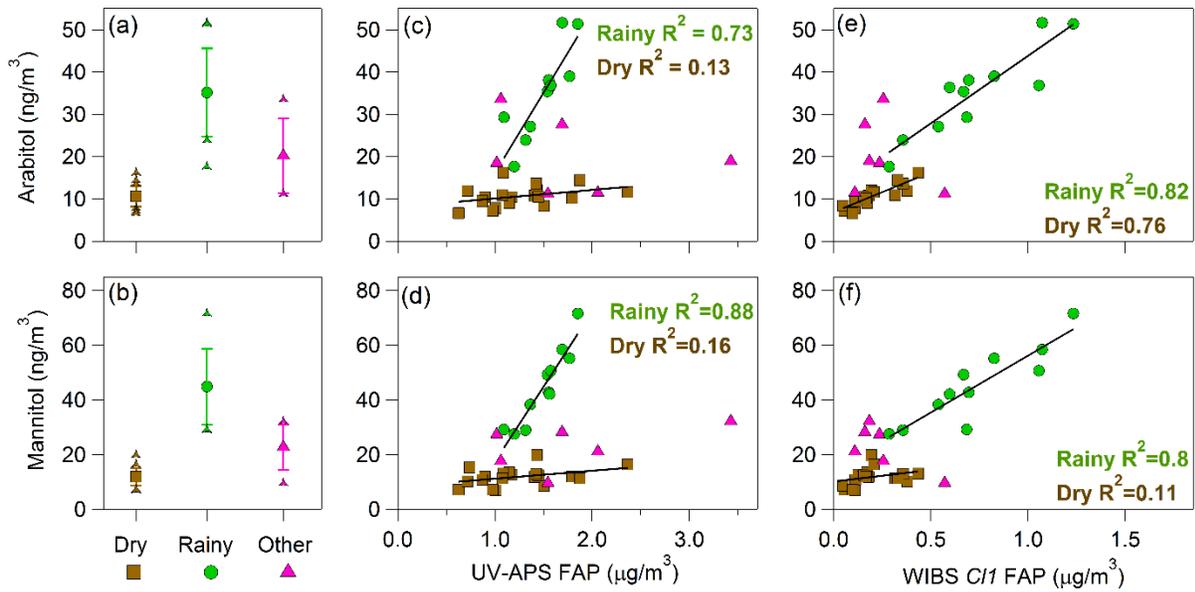
1011
 1012 **Figure 2:** Time series of key species concentrations and meteorological data over entire campaign. (a)
 1013 Fluorescent particle number size distribution measured with UV-APS instrument. Color scale indicates
 1014 fluorescent particle number concentration (L^{-1}). (b) Meteorological data: relative humidity (RH),
 1015 disdrometer rainfall (mm per 15 min), leaf wetness (mV). (c) Wetness category indicated as colored bars;
 1016 green, Rainy; brown, Dry; pink, Other. Bar width corresponds to filter sampling periods. Lightened
 1017 colored bars extend vertically to highlight categorization. (d) Colored traces show fungal spore
 1018 concentrations estimated from molecular tracers (circles) and WIBS C11 data (squares). (e) Stacked bars
 1019 show relative fraction of fluorescent particle type corresponding to each WIBS category.



1020
 1021 **Figure 3:** Characteristic differences between different wetness periods (Dry, Rainy, Other). (a) Relative
 1022 fraction of fluorescent particle number corresponding to each WIBS category. Bars show relative standard
 1023 deviation of category fraction in each wetness group (Dry, 19 samples; Rainy, 11 samples; Other, 6
 1024 samples). (b, c) Distribution of fungal OTU (operational taxonomic unit) values. (b) Fungal community
 1025 composition at phylum and class level with *Agaricomycetes* (dominant class with consistently ~60% of
 1026 diversity) removed. Relative proportion of OTUs assigned to different fungal classes and phyla for each
 1027 sample category shown. (c) Venn diagram showing the number of unique (wetness category specific) and
 1028 shared OTUs (represented by numbers in overlapping areas) among the sample categories (Dry, 11
 1029 samples; Rainy, 7 samples; Other, 3 samples). OTUs classified as cluster of sequences with $\geq 97\%$
 1030 similarity. Taxonomic assignments were performed using BLAST against NCBI database. In total, 3902
 1031 sequences, representing 406 fungal OTUs from 3 phyla and 12 classes were detected. Despite differences
 1032 in community structure across the sample categories, phylogenetic representation appears largely similar.

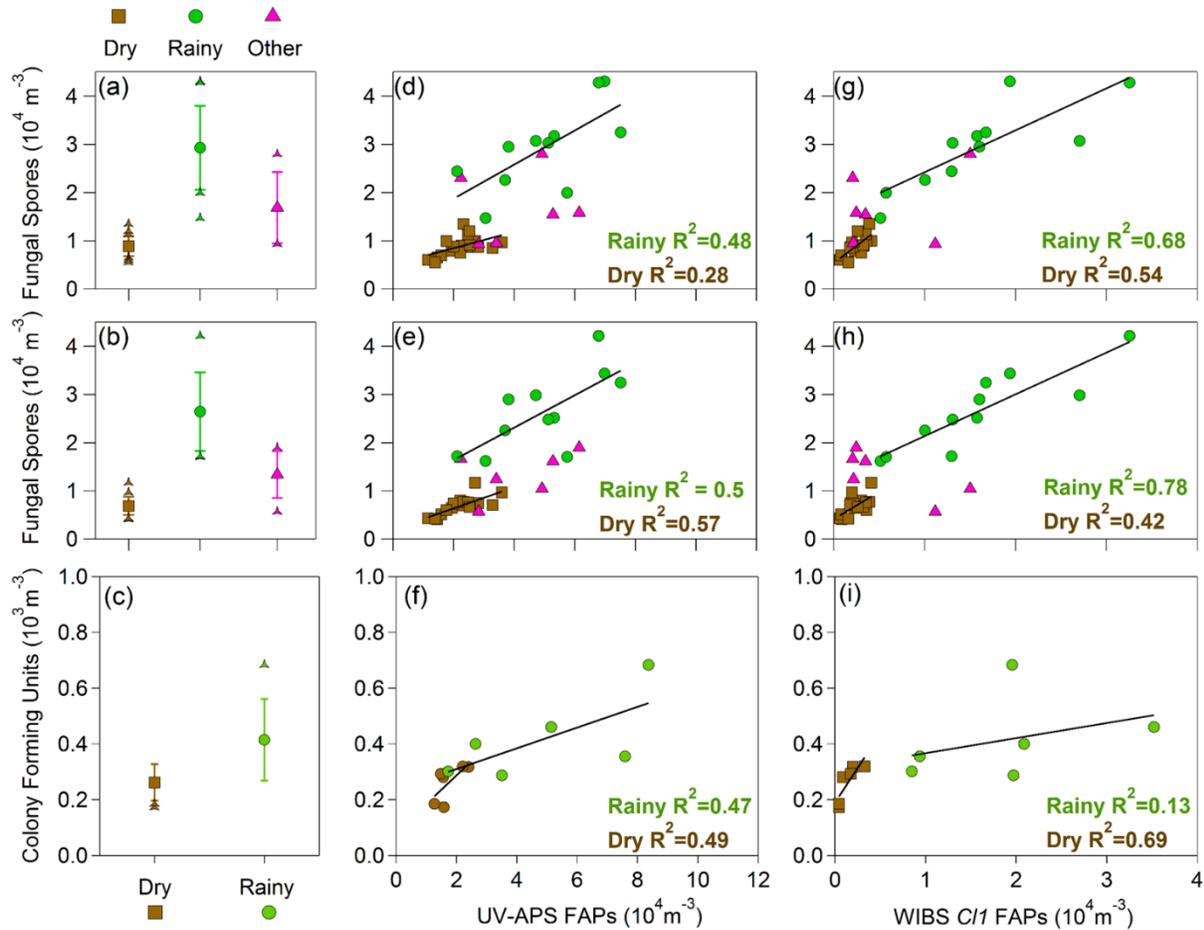


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 1034 **Figure 4:** Number concentration of fluorescent particles as a function of instrument channel, averaged
 1035 over entire measurement period. (a) Box-whisker plot of fluorescent particle number concentration for
 1036 WIBS FL1, FL2, FL3, and UVAPS. Circle markers shows mean values, internal horizontal line shows
 1037 median, top and bottom of box show inner quartile, and whiskers show 5th and 95th percentiles. (b) WIBS
 1038 FL1 versus UV-APS (c) WIBS FL2 versus UV-APS (d) WIBS FL3 versus UV-APS. Crosses represent 5-
 1039 minute average points. Linear fits assigned for data in each wetness category.

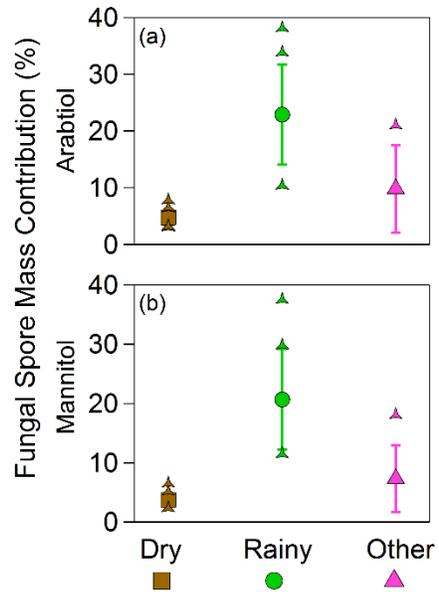


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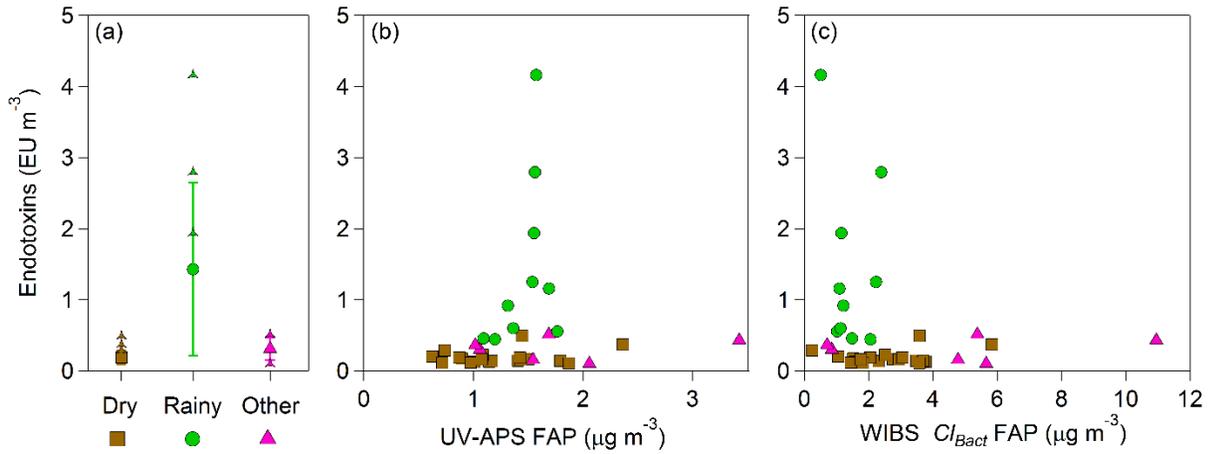
Figure 5: Mass concentrations of molecular tracers and fluorescent particles (assuming unit density particle mass and spherical particles): arabitol – top row, and mannitol – bottom row. Average mass concentration of arabitol (a) and mannitol (b) in each wetness category. Central marker shows mean value of individual filter concentration values, bars represent standard deviation (*s*) range of filter values, and individual points show outliers beyond mean ± *s*. Correlation of arabitol (c) and mannitol (d) with fluorescent particle mass from UV-APS. Correlation of arabitol (e) and mannitol (f) with fluorescent particle mass from WIBS Cluster 1. R² values shown for each fit in c-f. Linear fit parameters are shown in Table S2.



1050
 1051 **Figure 6:** Estimated fungal spore number concentration, calculated using mass of arabitol and mannitol
 1052 per spore reported by Bauer et al. (2008a). Estimates from arabitol (top row) and mannitol (bottom row).
 1053 Average fungal spore concentration, calculated using arabitol mass (a), mannitol mass (b), and colony
 1054 forming units (c) in each wetness category. Central marker shows mean value of individual filter
 1055 concentration values, bars represent standard deviation (s) range of filter values, and individual points
 1056 show outliers beyond mean $\pm s$. Correlation of fungal spore number calculated from arabitol (d) mannitol
 1057 (e), and colony forming units (f) concentration with estimated fluorescent particle mass from UV-APS.
 1058 Correlation of fungal spore number calculated from arabitol (g), mannitol (h), and colony forming unit (i)
 1059 concentration with fluorescent particle concentration from WIBS Cluster 1. R^2 value shown for each fit
 1060 (right two columns). Linear fit parameters are shown in Table S3.



1061
 1062 **Figure 7:** Estimated fraction of total aerosol mass contributed by fungal spores. Fungal spore mass
 1063 concentration ($\mu\text{g}/\text{m}^3$) calculated separately from mannitol and arabinol concentration and using average
 1064 mass per spore reported by Bauer et al. (2008b). Total particulate matter mass calculated from UV-APS
 1065 number concentration (m^{-3}) and converted to mass over aerodynamic particle diameter range $0.5 - 15 \mu\text{m}$
 1066 using density of 1.5 g cm^{-3} . Central marker shows mean value of individual filter concentration values,
 1067 bars represent standard deviation (s) range of filter values, and individual points show outliers beyond
 1068 $\text{mean} \pm s$.



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Figure 8: Endotoxin mass concentration as an approximate indicator of gram-negative bacteria concentration. (a) Averaged concentration in each wetness category. Central marker shows mean value of individual filter concentration values, bars represent standard deviation (*s*) range of filter values, and individual points show outliers beyond mean ± *s*. (b) Correlation of endotoxin mass concentration with estimated fluorescent particle mass from UV-APS. (c) Correlation of endotoxin mass concentration with estimated fluorescent particle mass summed from Clusters 2, 3, and 4 from Crawford et al. (2015).