Atmos. Chem. Phys. Discuss., doi:10.5194/acp-2016-725, 2016
Manuscript under review for journal Atmos. Chem. Phys.
Published: 2 November 2016

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Deciphering the Chemical Forms of Gaseous Oxidized Mercury in Florida, USA

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Abstract

The highest mercury (Hg) wet deposition in the United States of America (USA) occurs along the Gulf of Mexico, and in the southern and central Mississippi River Valley. Gaseous oxidized Hg (GOM) is thought to be a major contributor due to high water solubility and reactivity. Therefore, it is critical to understand concentrations, potential for wet and dry deposition, and GOM compounds present in the air. Concentrations and dry deposition fluxes of GOM were measured and calculated for Outlying Landing Field (OLF), Florida, using data collected by a Tekran® 2537/1130/1135, the University of Nevada-Reno Reactive Mercury Active System (UNRRMAS) with cation exchange and nylon membranes, and Aerohead samplers that use cation-exchange membranes to determine dry deposition. Relationships with Tekran® derived data must be interpreted with caution, since GOM concentrations measured are biased low depending on the chemical compounds in air, and interferences with water vapor and ozone. Criteria air pollutants were concurrently measured.

Results from nylon membranes with thermal desorption analyses indicated five GOM compounds in this area, including $HgBr_2$, HgO, Hg-nitrogen and sulfur compounds, and an unknown compound. This indicates that the site is influenced by different gaseous phase reactions and sources. Using back trajectory analysis a high GOM event related to high CO, but average SO_2, indicated air parcels moved from the free troposphere, and across Arkansas, Mississippi, and Alabama at low elevation (<300 m). This event was initially characterized by $HgBr_2$ followed by a mixture of GOM compounds. GOM chemistry indicates reactions with local mobile source pollutants and long range transport from outside of the USA.

In order to develop methods to measure GOM concentrations and chemistry, and model dry deposition processes, the actual GOM compounds need to be known, as well as their corresponding physicochemical properties, such as Henry's Law constants.

Keywords: multiple-resistance model, dry deposition, cation exchange membrane, criteria pollutants, active samplers

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1 Introduction

Mercury (Hg) has been classified as a persistent, bioaccumulative toxin (PBT) (UNEP, 2013), and deposition from the atmosphere is considered the dominant pathway by which Hg enters remote ecosystems (Lindberg et al., 2007). In some areas, scavenging by precipitation controls atmospheric Hg removal processes, such as in the Southeastern United States of America (USA), where precipitation amounts are high (Prestbo and Gay, 2009). However, wet deposition concentrations are not necessarily correlated with precipitation amounts >81mm, and deposition did not decreased with emission reductions as coal combustion facilities in the region have implemented control technologies (Prestbo and Gay, 2009). A contributing factor to wet deposition in the Gulf Coast area may be related to high atmospheric convection during thunderstorms and scavenging of gaseous oxidized Hg (GOM) from the free troposphere (Nair et al., 2013), and down mixing of air with high GOM from the free troposphere (cf. Gustin et al., 2012).

An additional concern is that the Tekran® system measurement currently used to quantify GOM does not equally quantify all GOM forms, and has interferences with water vapor and ozone (cf. Ambrose et al., 2013; Gustin et al., 2013; Huang et al., 2013; Lyman et al., 2010; McClure et al., 2014; Lyman et al., 2016). Since GOM is considered an important form that can be rapidly removed from the atmosphere due to high water solubility (Lindberg et al., 2007); it is important to understand both atmospheric concentrations and chemistry (i.e. specific chemical compounds). Use of the University of Nevada-Reno Reactive Mercury Active System (UNRRMAS) that collects GOM on nylon membranes in tandem with cation exchange membranes has indicated that there are different chemical compounds in the air and concentrations are 2 to 13 time higher than previously thought (Huang et al., 2013; Gustin et al., 2016).

Mercury has been studied in Florida for many years, initially because of the high concentrations measured in fish and the Florida Panther (Dvonch et al., 1999; Gustin et al., 2012; Marsik et al., 2007; Pancras et al., 2011; Peterson et al., 2012). Long-term GEM and GOM concentrations as measured by the Tekran® system have declined; however, PBM concentrations increased after 2009 (Edgerton, unpublished data), suggesting the atmospheric chemistry has changed. Peterson et al. (2011) and Gustin et al.

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(2012) suggested based on detailed assessment of passive sampler and Tekran® system collected Hg data, criteria air pollutants, and meteorology that at 3 locations in Florida different GOM compounds were present and these were generated by *in situ* oxidation associated with pollutants generated by mobile sources, indirect and direct inputs of Hg from local electricity generating plants, and direct input of Hg associated with long range transport. At OLF, background deposition was equal to that associated with mobile sources, and a significant component was derived from long range transport in the spring.

In this work, GOM collected using the UNRRMAS, and the Aerohead dry deposition measurement method (Lyman et al., 2007; 2009) were analyzed, along with Tekran® Hg and criteria air pollutant data to understand GOM chemistry and dry deposition at Outlying Landing Field (OLF), located ~ 15 kilometers NW of Pensacola, Florida. Mercury was measured concurrently with various trace gases (CO, O_3, SO_2, NOx, NOy) and meteorology. Air GEM/GOM/PBM concentrations were measured by the Tekran® 2537/1130/1135 system, respectively.

GOM dry deposition fluxes were calculated using deposition velocities determined using a multi-resistance model with ambient air GOM concentrations from the Tekran® system (multiplied by a factor of 3 due to bias in the Tekran® system; cf. Huang and Gustin, 2015), and compared to those obtained using Aerohead data. Results were used to estimate dry deposition velocities for the GOM compounds observed. The hypothesis for this work was that since GOM compounds can vary spatially and temporally, due to different compounds produced by different sources and processes, this will result in different dry deposition velocities and dry deposition flux.

2 Methods

2.1 Field site

The sampling site was located at OLF (30.550°N, 87.374°W, 44 m above sea level). The closest major Hg emission source is a coal-fired power plant (Plant Crist) northeast of the site (Figure 1). This area has been used for atmospheric Hg research in previous studies (Caffrey et al., 2010; Lyman et al., 2009; Gustin et al., 2012; Peterson et al., 2012; Weiss-

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Penzias et al., 2011). OLF is a coastal site (~25 km away from Gulf of Mexico) influenced by sea breezes especially during the summer (Gustin et al., 2012). Based on cluster analyses of data from one year at this location ~24% of the air comes from the marine boundary layer during the day and 60% during the night (Figure 1).

2.2 Sampling Methods

Aerohead samplers for determination of dry deposition were deployed bi-weekly from June 2012 to March 2014. UNRRAMS were deployed bi-weekly from March 2013 to March 2014. Atmospheric Hg concentrations, including GEM, GOM, and PBM, were measured using a Tekran® system (model 2357/1130/1135, Tekran® Instrument Corp., Ontario, Canada) that was operated with one-hour sampling and one-hour desorption with detection limits of 0.1 ng m^{-3}, 1.5 pg m^{-3}, and 1.5 pg m^{-3}, respectively.

Reactive Hg (GOM + PBM) concentrations were measured using the UNRRAMS with 3 sets of two in-series 47 mm cation-exchange membranes (ICE450, Pall Corp., MI, USA). Three sets of nylon membranes (0.2 µm, Cole-Parmer, IL, USA) were also deployed to assess Hg compounds in the air. Cation exchange membranes have been demonstrated to quantitatively measure specific compounds of GOM in the laboratory; however, may not measure all compounds (Gustin et al., 2015; Gustin et al., 2016). Nylon membranes do not retain GOM compounds quantitatively, and retention during transport needs to be tested (Huang et al., 2013; Gustin et al., 2015; Gustin et al., 2016). Nylon membrane retention is impacted by relative humidity that might limit uptake of specific forms. Criteria air pollutants and meteorological data, including CO, SO_2, O_3, $PM_{2.5}$, NO, NO_2, NO_y, temperature, relative humidity, wind speed, wind direction, pressure, solar radiation, and precipitation were available at this site for the sampling period. See Peterson et al. (2012) for detailed information on collection of these measurements.

Aeroheads and membranes were prepared at UNR, packed in a thermal isolated cooler, and shipped back and forth between the laboratory and site. Samples were stored in a freezer (-22°C) at UNR until analyzed. Cation-exchange membranes were digested and analyzed following EPA Method 1631 E (Peterson et al., 2012), and nylon membranes were first thermally desorbed, and then analyzed using EPA Method 1631 E (Huang et al.,

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2013). Cation-exchange membrane blanks for Aerohead and UNRRAMS were 0.40±0.18 (n=42), 0.37±0.26 (n=77) ng, respectively; and for nylon membranes used in the active system blanks were 0.03±0.03 (n=69) ng. Therefore, method detection limits (MDL, 3-sigma) for two-week sampling time (336 hrs) was 0.13 ng m^{-2} hr^{-1} for dry deposition, respectively. For the active membrane system, the Hg amount on the back-up filters and blanks were not significantly different (cation-exchange membrane: 0.4±0.3 vs 0.4±0.3 ng; nylon membrane: 0.03±0.03 vs 0.02±0.02 ng); therefore, the back-up filters were included in the calculation of the bi-weekly blanks. The bi-weekly MDL (336 hrs) for active system with cation-exchange and nylon membranes were 2-68 pg m^{-3}(mean: 24 pg m^{-3}) and 0.01-14.6 pg m^{-3}(mean: 2.1 pg m^{-3}), respectively. Bi-weekly MDL was calculated from 3 times the standard deviation of bi-weekly blanks. The MDL was calculated for each period of sampling, due to the fact this can vary based on treatment of the membranes, when preparing samples for deployment, deployment at the field site, and handling once returned to the laboratory. The membranes may also vary by material lot. All samples were corrected by subtracting the blank for the corresponding two-week period.

2.3 Data analyses

Hourly Tekran®, criteria air pollutants, and meteorological data were managed and validated by Atmospheric Research & Analysis, Inc (see Peterson et al., 2012). These were then averaged into two-week intervals to merge with the membrane measurements.

In previous studies, scaling factors similar to HNO_3 (α=β=10) were used to calculate oxidized Hg dry deposition velocity (Marsik et al., 2007; Castro et al., 2012); however, Lyman et al. (2007) used the effective Henry's Law constant, and half-redox reactions in neutral solutions of $HgCl_2$, and indicated HONO might better represent the chemical properties of oxidized Hg rather than HNO_3. Huang et al. (2015a) indicated that due to limited understanding of oxidized Hg chemical properties, no single value can be used to calculate oxidized Hg dry deposition, because α and β would change with different GOM compounds. Here dry deposition was calculated using the multiple resistance model of Lyman et al. (2007) using both α=β=2, 5, 7, and 10.

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Back trajectories were calculated using the Hybrid Single Particle Lagrangian Integrated Trajectory (HYSPLIT 4.9) with EDAS 40-km, 1000-meter starting height. For day and nighttime analyses, starting times were Local Standard Time (LST) 1100-1300 h and 100-300 h, 24-hour simulations. For a high-concentration event analyses, trajectories were started for each day at 0000, 0400, 0800, 1200, 1600, and 2000 LST. Overall, the uncertainties of back trajectories calculated from HYSPLIT are 20% of the air parcel traveling distance (Draxler, 2013 ; Gebhart et al., 2005; Stohl, 1998; Stohl et al., 2003). Back trajectories for the entire sampling time were analyzed using cluster analysis (Liu et al., 2010).

Sigmaplot 14.0 (Systat Software Inc, San Jose, CA, USA), and Minitab 16.0 (Minitab Inc., PA, US) were used to do t-tests and correlation analyses. Comparisons were considered significantly different and correlations considered significant when $p < 0.05$.

3 Results and Discussion

3.1 Overall measurements

Similar to previous work at this location (Gustin et al., 2012), O_3 was highest in the spring, and CO concentrations were high in winter due to a low boundary layer and biomass burning, and low in summer (Table 1). Observations from the 3 GOM sampling methods (Tekran®, and nylon, and cation exchange membranes) showed higher GOM concentrations in spring relative to other seasons (Table 1). Concentrations of GOM measured by cation-exchange membranes in the active system were significantly (p-value < 0.05, paired-t test) higher than those measured by Tekran® KCl-coated denuder and nylon membranes, both of which have been reported to be influenced by relative humidity (Huang and Gustin, 2015b; Gustin et al., 2015). Mean cation-exchange membrane concentrations were higher than Tekran® derived GOM by 14, 48, 11, and 13 times in the spring, summer, fall and winter, respectively.

Nylon membranes collected higher GOM concentrations than those measured by the Tekran® in spring 2013 when the humidity was low. Overall, air concentrations measured by the Tekran® system in this study were similar to those measured in 2010

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(Peterson et al., 2012). Particulate-bound Hg had the same diel trend as GOM, but higher concentrations.

Understanding the oxidants present in air is important for understanding potential GOM compounds. Oxidants to consider include O_3, halogenated compounds, sulfur and nitrogen compounds (cf. Gustin et al., 2016). Since the active system is currently limited to 2-week sampling period it is difficult to use the data collected to determine specific sources; however, they are useful for understanding the specific compounds that might be present, and this in turn can be used to understand sources.

3.2 Potential GOM compounds

Standard desorption profiles for GOM compounds obtained by Huang et al. (2013) and Gustin et al. (2015) are compared to those obtained at OLF (Figure 2). Compounds in the permeation tubes included $HgBr_2$, $HgCl_2$, $HgN_2O_6•H_2O$, $HgSO_4$, and HgO. $HgCl_2$ and $HgBr_2$ have been identified as being released from permeation tubes (Lyman et al., 2016); however, the exact N and S compounds are not known. Only during 10 periods the nylon membranes (collected in triplicate) collected a significant amount of GOM based on their bi-weekly detection limit (Figure 2), and their desorption profiles varied. Although data are limited, because we have observed similar thermal desorption compounds in other studies (i.e. Huang et al, 2013 and Gustin et al. 2016), this indicates different chemical forms are being collected, and that the compounds are not being generated on the membranes. This has been shown to be the case in a limited study (Pierce and Gustin, 2016). In addition to our work, two forms of GOM in Montreal Canada air were reported (Deeds et al., 2015).

Five distinct patterns of release were observed during thermal desorption. One had a high residual tail that does not match our standard profiles; however, was also observed in Nevada (Gustin et al, 2016). These occurred on 4/2/2013, 4/9/2013 and 5/21/2013. This suggests that in spring there is a compound that is unknown based on current standard profiles. A nitrogen-based compound was found on 5/21/2013 based on the desorption profile. The second pattern occurred on 3/19/2013 and 11/19/2013, and this

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corresponds to $HgBr_2$ with some residual tail that is again some compound not accounted for.

The third pattern that occurred on 5/7/2013 and 8/27/2013, and corresponds to Hg-nitrogen based compound with a residual tail. The 4th pattern occurred on 1/14/2014, and 9/24/2013 was associated with $HgSO_4$ and the error bars are small. Lastly, the data collected on 10/22/23 was noisy and had subtle peaks that correspond with HgO, a nitrogen-based compound, and a high residual tail. It is interesting to note that the 11/19 profile was similar to $HgCl_2$.

Previous studies reported consistent desorption profiles from 3 sites in Nevada and California without significant point sources (Huang et al., 2013). Huang et al. (2013) presented desorption profiles from a highway, agriculture, and marine boundary layer site. Profiles from the marine boundary layer and agriculture impacted site did not show clear residual tails at 185ºC, but these were observed at the highway impacted site. In addition, at OLF, a significant amount of GOM (15-30%) was released after 160 ºC. This implies that we are missing one or more GOM compound(s) (Figure 2) in our permeation profiles. Interestingly, a peak was found in the 4/9/2013 sample at the GEM release temperature, and this is not due to GEM absorption as demonstrated by Huang et al.(2013), and was also observed in Nevada (Gustin et al., 2016), suggesting an additional unidentified compound. This information indicates GOM compounds at OLF varied with time, and this variation is due to complicated Hg emission sources and chemistry at this location (cf. Gustin et al., 2012).

At OLF, GOM composition on the nylon membrane was more complicated than that collected at rural sites in the Western US (cf. Huang et al.,2013; Gustin et al., 2016); however, similar complexity was observed at a highway location in Reno, Nevada (Gustin et al., 2016). Desorption curves from the nylon filters collected at rural locations in Nevada were in the range of the standard GOM compounds that have been investigated (Huang et al., 2013: Gustin et al., 2016). Curves with multiple peaks in this study imply that there were at least 5 GOM compounds collected on the nylon membranes.

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3.3 Dry deposition measurements

Dry deposition of GOM measured by the Aerohead sampler ranged from 0 to 0.5 ng m^{-2} hr^{-1}, and 83% of GOM dry deposition was higher than the detection limit (0.12 ng m^{-2} hr^{-1}). Higher GOM dry deposition was observed in spring relative to winter (ANOVA one-way rank, p-value < 0.01); GOM dry deposition was slightly lower in summer and fall (not statistically different) relative to the spring due to high wet deposition and scavenging processes during these seasons. The pattern in GOM seasonal dry deposition was similar to that reported by Peterson et al. (2012). However, GOM dry deposition rates were significantly higher in this study than 2010 values (0.2 vs 0.05 ng m^{-2} h^{-1}). This is due to the correction of 0.2 ng m^{-2} h^{-1} applied in Peterson et al. (2012) to account for contamination of the Aerohead that has been demonstrated to be unnecessary (Huang et al., 2014). Although, highest GOM dry deposition measured using the Aerohead sampler and GOM concentrations measured using the UNRRAMS were observed in spring 2013, the value in March 2014 was relatively low. In March 2014, atmospheric conditions were more similar to winter than spring, with low temperatures and high CO concentrations. These results are different from those calculated using Tekran® measurements that suggest low GOM concentrations and high deposition velocities, and this is because the denuder measurements are biased low.

Modeled GOM dry deposition fluxes were calculated using GOM concentrations measured by the Tekran® system that were multiplied by a factor of three (cf. Huang et al., 2014). In general, measured Hg dry deposition fluxes were similar to those modeled simulations with modeled GOM dry deposition $\alpha=\beta=2$ during winter, spring, and fall (see below; Figure 3). However, measured Hg dry deposition was significantly higher than modeled results (both $\alpha=\beta=2$ and 10) in summer and early fall (Figure 3). This indicates there are compounds of GOM in the summer that are poorly collected by the denuder, and this also can help explain the higher wet deposition measured during this season (Prestbo and Gay, 2009). Highest deposition was measured during the spring, when the input from long range transport is greatest (Gustin et al., 2012). Figure 3 shows the disparity that occurs by season, and comparing model and measured vales. For example

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in spring a=b=10 significantly overestimates deposition, while in the summer and early fall measured deposition is greater than modeled values.

Because of the low GOM concentrations and influence of humidity on the nylon membrane measurements (Huang and Gustin, 2015b), GOM compounds were identified only in one summertime sample as $HgN_2O_6•H_2O$. During this time, measured GOM dry deposition was ~6 times higher than both modeled results, and considering the Tekran® correction factor of 3, membrane-based $HgN_2O_6•H_2O$ dry deposition flux was ~18 times higher than the Tekran®-model-based value. Gustin et al. (2015) indicated $HgN_2O_6•H_2O$ collection efficiency on cation-exchange membrane in charcoal scrubbed air was ~ 12.6 times higher than on Tekran (KCl-coated denuder.

However, in May 2013, two samples were dominated by a profile similar the Hg nitrogen-based compound with lower measured/modeled ratios (2.1-6.0 with Tekran® correction factor). This might be due to ambient air GOM chemistry being dominated by a compound with a different dry deposition velocity, less interference on the denuder surface, or parameters in the dry deposition scheme. In May, GOM concentrations measured by the Tekran® were higher than in summer due to lower wet deposition and mean humidity (Table 1). Therefore, despite the fact that GOM collection efficiency associated with the Tekran and nylon membranes are impacted by environmental conditions, this demonstrates the presence of different compounds in the air. The dry deposition scheme needs Henry's Law constants for determining the scaling factors for specific resistances for different compounds (Lyman et al., 2007; Zhang et al., 2002).

Lin et al. (2006) stated that the dry deposition velocity of HgO is two times higher than that for $HgCl_2$, due to the different Henry's Law constant. The Henry's Law constants for $HgCl_2$, $HgBr_2$, and HgO presented in previous literature (Schroeder and Munthe, 1998) have high uncertainty, for how these calculations were done is not clear (S. Lyman, Utah State University, personal communication, 2015), and the constants for $HgN_2O_6•H_2O$ and $HgSO_4$ are unknown. Some researchers considered that GOM is similar to HNO_3 ($\alpha=\beta=10$), and some treated GOM as HONO ($\alpha=\beta=2$) (Castro et al., 2012; Lyman et al., 2007; Marsik et al., 2007); however, using the parameters of HNO_3 could overestimate

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GOM dry deposition velocities due to the differences of effective Henry's law constants ($HgCl_2$: ~10^6 HNO_3: ~10^{13} M atm^{-1}).

If the ratios ($HgBr_2$: 1.6, $HgCl_2$:2.4, $HgSO_4$: 2.3, HgO: 3.7, and HgN_2O_6•H_2O: 12.6) of GOM concentrations measured by the Tekran® versus cation-exchange membranes for different GOM permeated compounds (Gustin et al., 2015; Huang et al., 2013) are used to correct Tekran® GOM data in this study, modeled GOM dry deposition (Figure 3) are not correlated with measurements. For example, on 3/9/2013 and 11/19/2013 (Figure 3), GOM was dominated by $HgBr_2$ and $HgCl_2$. Dry deposition of $HgBr_2$ from Aerohead measurements and modeling were close to α=β=10; however, modeled and measured $HgCl_2$ dry deposition were matched as α=β=2. Average deposition velocity for α=β=2 was 0.78 cm s^{-1}, and for α=β=10 is 1.59 cm s^{-1}, if we assume the model is right. There were three samples that were identified as Hg-nitrogen based compounds using nylon membranes; however, the ratios of measurement and modeling HgN_2O_6•H_2O dry deposition were inconsistent over time. In spring, all modeled HgN_2O_6•H_2O dry deposition values were much higher than measured values; however, in summer, measured and modeled HgN_2O_6•H_2O dry deposition were similar as α=β=5 (Table 2). If you assume the dry deposition measurements made by the surrogate surfaces are accurate then this demonstrates there are different forms that occur over time, and these will have different deposition velocities as suggested by Peterson et al. (2012).

3.4 Elevated Pollution Event

In spring 2013, there was a time period when high concentrations of O_3, CO, and all Hg measured (Figure 4). Figure 5 shows that during this time air masses traveled west to east across the continent. The air movement pattern is similar to that found in Gustin et al. (2012) for OLF Class 2 events which had low SO_2 concentrations. During this 4-week period, air parcels traveling to OLF were in the free troposphere and descended to the surface (Figure 5). Although there are coal-fired power plants in the upwind area within a 500 km range (Figure 1), the low SO_2 concentrations, and elevated CO, O_3, and GOM values were not from fossil fuel combustion. Gustin et al. (2012) also indicated that free troposphere air impacted OLF. The first few endpoints for these trajectories indicate air

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parcels entered North America at > 1000 m agl; therefore, there was transport of some air measured during this time from the free troposphere. Ozone concentrations were also similar to those measured in Nevada in the free troposphere at this time (Gustin et al., 2014). It is important to note that the back trajectories are only for 72 hours and the ones that subsided to surface levels in the Midwest were traveling fast. We hypothesize this is a common event in the spring that represents free troposphere transport. The chemical composition of this event suggests potential input from Asia as previously suggested for numerous locations in Florida in the spring by Gustin et al. (2012). During this time $HgBr_2$ was an important compound initially and then the profile was a gradual increase with a high residual tail. This would suggest initial subsidence of air from the stratosphere/troposphere (cf. Lyman et al., 2012) followed by a mixture of polluted air as observed in the Western United States (c.f. VanCuren and Gustin, 2015)

4 Conclusions

The chemical forms of GOM in the atmosphere at OLF varied by season as suggested by Gustin et al. (2012). Five potential different GOM compounds were identified at OLF using nylon membranes with thermal desorption analysis, including $HgBr_2$, $HgCl_2$, HgO, Hg-nitrogen and sulfur compounds, and 2 unknown compounds. Given the long sampling time detailed assessment of specific sources is difficult, but the presence of different compounds indicate multiple sources and different GOM chemistry. Comparing modeled and measured Hg dry deposition fluxes also demonstrate there are different forms in air and this will affect dry deposition velocities. In order to improve our understanding of Hg air-surface exchange, and measure GOM physiochemical properties of different GOM compounds need to be understood.

5 Acknowledgements

The authors thank The Southern Company (project manager-John Jansen) for their support, and Bud Beghtel for deploying and collecting our membranes and passive samplers at OLF and managing this site in general. This work was also supported by EPRI and a National Science Foundation Grant 1326074. We thank the following students for coordinating shipment of membranes and passive samplers, analyses of the

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membranes in the lab, and keeping the glassware clean (Keith Heidecorn, Douglas Yan, Matt Peckham, Jennifer Arnold, Jen Schoener, and Addie Luippold).

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Table 1 – Overall seasonal average of air species, GEM, PBM, GOM (measured using three different methods) concentration, GOM dry deposition (DD), and meteorological data at OLF.

	2012			2013				2014
	Summer	Fall	Winter	Spring	Summer	Fall	Winter	March
Ozone [ppb]	30±15	30±12	29±11	38±12	24±12	26±11	27±10	35±12
CO [ppb]	143±38	161±35	167±41	165±36	139±35	156±33	167±35	183±33
SO_2 [ppb]	0.3±0.4	0.6±1.5	0.4±0.5	0.3±0.5	0.2±0.3	0.4±0.5	0.7±1.2	0.3±0.4
NO [ppb]	0.3±0.7	0.3±0.7	0.3±0.8	0.2±0.5	0.3±0.7	0.3±0.8	0.4±0.8	0.2±0.5
NO_2 [ppb]	2.4±2.4	3.0±2.7	3.0±3.1	2.0±2.3	2.2±2.1	3.1±2.9	3.2±3.0	2.3±2.8
NO_y [ppb]	3.6±2.9	4.3±3.1	4.3±3.6	3.1±2.8	3.2±2.5	4.4±3.3	4.2±3.4	3.6±3.1
GEM [ng m^{-3}][a]	1.2±0.1	1.2±0.1	1.3±0.1	1.2±0.2	1.1±0.1	1.0±0.1	1.2±0.3	1.2±0.1
GOM [pg m^{-3}][a]	0.6±1.3	1.1±2.8	1.0±2.2	2.9±5.1	0.5±1.0	1.1±2.1	1.3±2.5	2.0±3.6
PBM [pg m^{-3}][a]	2.4±2.6	3.6±3.8	7.3±8.7	5.9±6.8	2.3±2.0	2.9±2.3	4.9±5.3	4.0±3.4
GOM [pg m^{-3}][b]	-	-	-	43±110	24±57	14±18	17±23	24±15
GOM [pg m^{-3}][c]	-	-	-	4±10	0.4±1.3	1.2±1.1	0.6±0.6	0.6±0.5
GOM DD [ng m^{-2} hr^{-1}]	0.24±0.20	0.17±0.12	0.15±0.06	0.40±0.23	0.20±0.13	0.13±0.18	0.20±0.50	0.14±0.04
WS [m s-1]	2.1±1.2	2.1±1.0	2.8±1.7	2.9±1.8	2.0±1.1	2.1±1.1	2.5±1.3	2.5±1.5
TEMP [°C]	26±3	19±6	14±6	18±6	26±3	20±7	11±7	14±5
RH [%]	83±14	76±18	79±19	73±21	84±13	77±17	76±23	78±21
SR [w m^{-2}]	230±302	193±271	121±199	266±304	210±278	175±255	129±212	182±278
Precipitation [mm]	637	186	385	223	1010	254	357	183

[a]: Tekran data
[b]: cation-exchange membrane data
[c]: nylon membrane data

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Table 2 – Modeled (multiple-resistance model) and measured (surrogate surfaces) GOM dry deposition (ng m^{-2} hr^{-1}), GOM concentrations used to calculate for modeled results are from the Tekran® data and corrected by compounds' corresponding ratios from Gustin et al. (submitted). The sample with unknown compound is used the Tekran® data with correction factor of three (average ratio). The tentative GOM compounds are identified from nylon membrane results.

Start date	Tentative GOM compound	Measured GOM dry deposition flux	Modeled GOM dry deposition α=β=2	Modeled GOM dry deposition α=β=5	Modeled GOM dry deposition α=β=7	Modeled GOM dry deposition α=β=10
3/12/2013	$HgBr_2$	0.50±0.06	0.34	0.49	0.54	0.58
3/26/2013	unknown	0.40±0.11	0.34	0.47	0.52	0.56
4/30/2013	$Hg(NO_3)_2$	0.50±0.13	1.21	1.67	1.81	1.95
5/14/2013	$Hg(NO_3)_2$	0.40±0.09	1.19	1.69	1.88	2.07
8/20/2013	$Hg(NO_3)_2$	0.15±0.07	0.10	0.14	0.16	0.17
11/12/2013	$HgCl_2$	0.08±0.03	0.11	0.16	0.17	0.19
1/7/2014	$HgSO_4$	0.19±0.03	0.18	0.24	0.27	0.29

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Figure Caption

Figure 1 – Sampling site and point sources (NEI 2011) map. Cluster trajectories for daytime (11:00-13:00) and nighttime (1:00-3:00).

Figure 2 – Desorption profiles from nylon membranes with standard materials in laboratory investigation (top) and field measurements. Whisker is 1 standard variation, and only present in the desorption peak. Note the Hg-nitrogen compound in the permeation tube was $HgN_2O_6 \bullet H_2O$.

Figure 3 – Measured and modeled GOM dry deposition fluxes, Tekran® data (correction factor of three) were used with multiple resistance models ($\alpha=\beta=2$ and 10). Tentative GOM compounds were determined using the results from nylon membranes desorption.

Figure 4 – Temporal variation of GOM concentrations (mean ± standard deviation, bi-week average), outlined rectangle indicates a polluted event with high Hg, CO, and ozone concentrations. Data are missing for 3 weeks because it was not collected. Tekran data is presented when >75% of the data were available and membrane data are shown when above the method detection limit.

Figure5 – Results of gridded frequency distribution (left), light color indicates less endpoints in a grid. Altitude of 72-hr trajectories during the polluted event (3/12/2013-4/2/2013), light color of dots on left panel represents low altitude.

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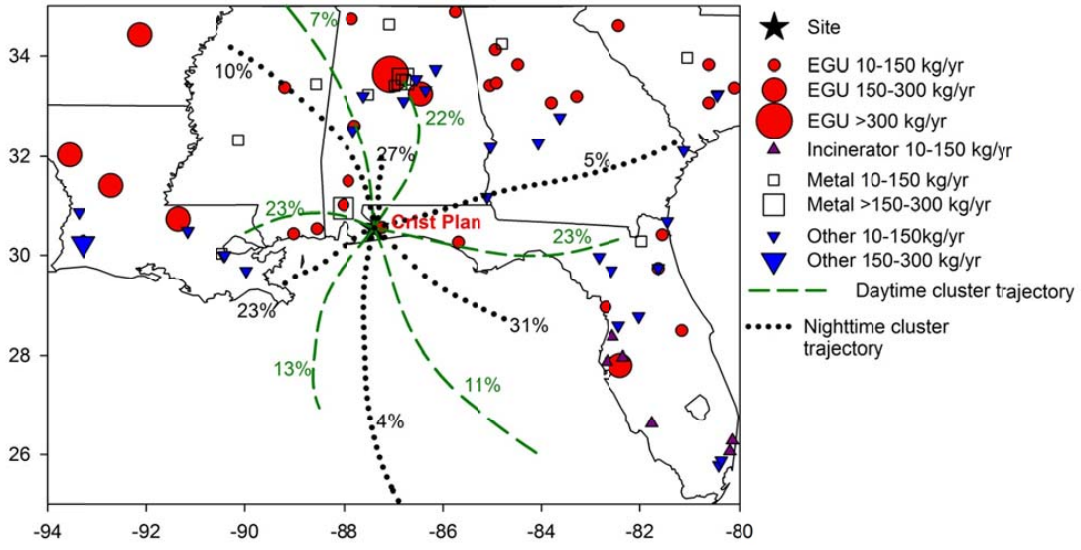

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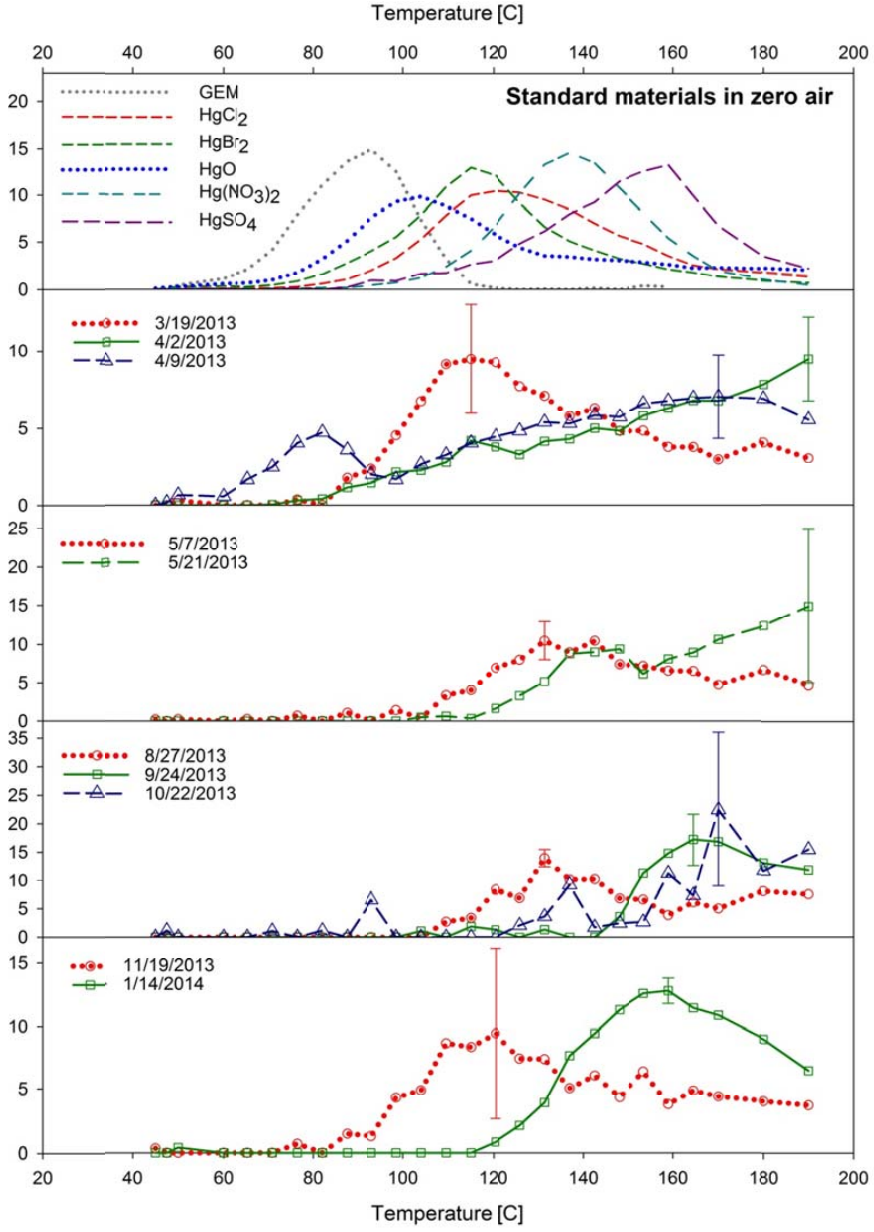

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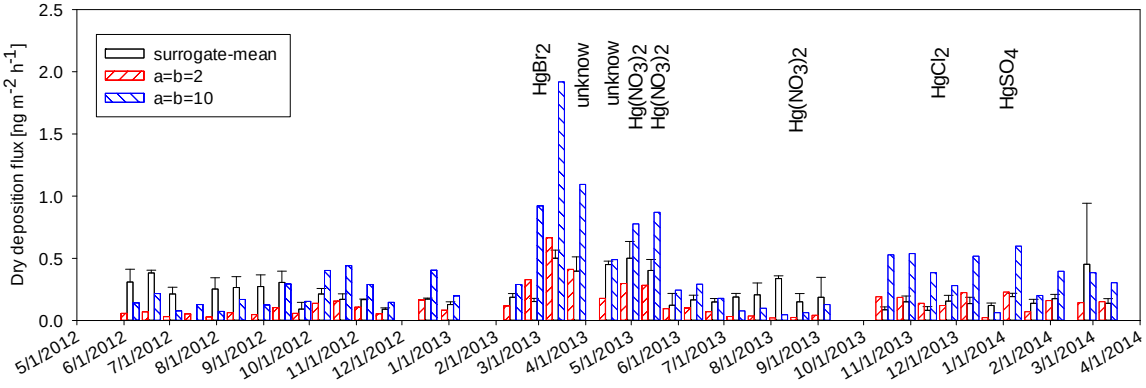

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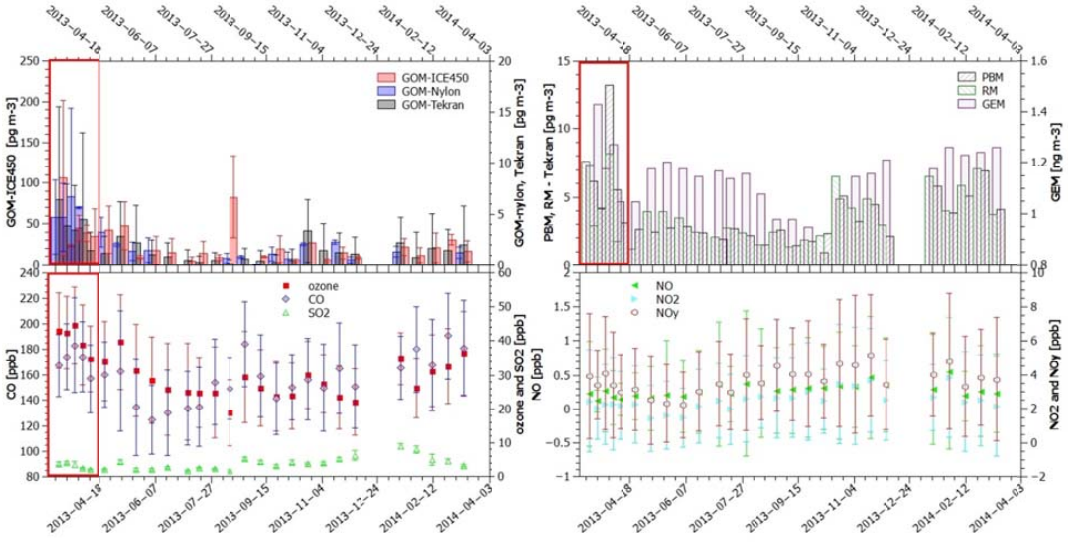

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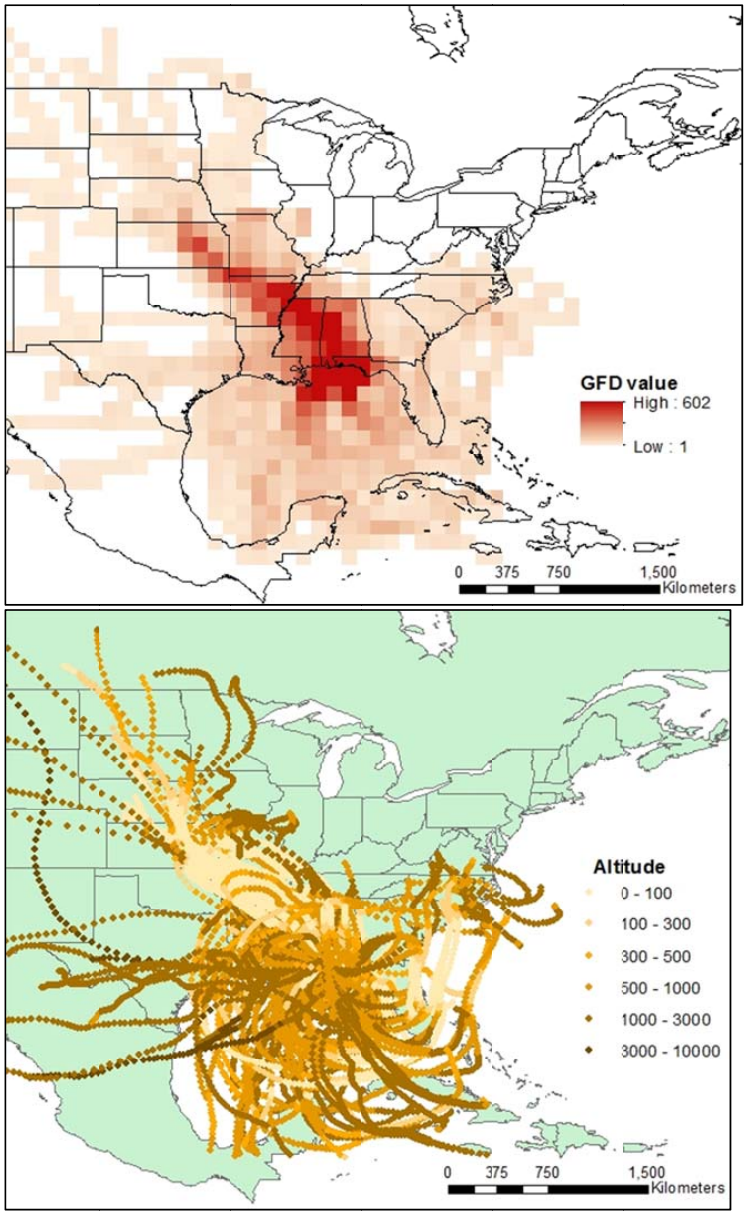

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