

Reply to referee comments to “Wildfire air pollution hazard during the 21st century” by Knorr et al.

Report #1, Anonymous Referee #3

This manuscript presents a study on the effects of humans and climate on fire activity and consequences for air pollution across the 21 century. The Authors use LPJ-GUESS with SIMFIRE to produce a set of future fire emissions (based on GFED4s for present-day) and the chemical transport model CAM-Chem to study air quality impacts. The effect of climate and human intervention on future fire activity and air pollution is an important topic and results from this work are of relevant interest for the readers of the ACP. This manuscript is a revised version of a previously submitted manuscript. However, despite the substantial work done by the Authors to address the reviewers' comments, I find the manuscript still very confusing and poorly organized and cannot be published as it is.

Following the comments of Referee #3, as well as Referee #2 and the editor, we substantially revised the structure of the manuscript such that it now focuses on one point, the possible impact of wildfires on the ability to meet WHO air quality guidelines for PM_{2.5} by the WHO. The guiding question is now *“whether socio-economic developments influencing both greenhouse gas emissions as well as human population size and spatial distribution might impact wildfire emissions enough to make a difference for meeting the WHO air-quality target, provided anthropogenic PM_{2.5} emissions are reduced aggressively.”*, stated in the (new) last paragraph of Section 1. Section 3 (Results) has been completely re-written, with the assessment of current emission levels dropped (former Fig. 1 is now included in the SI as part of Fig. S1) and the number of subsections reduced to two, one for pollutant emissions and one for pollutant concentrations. A new paragraph has been added at the beginning of Section 4 (Discussion) that summarizes the impact of different scenarios on human pollutant exposure, as well as a discussion of a possible WHO target for monthly mean concentrations that would be expected to lie somewhere between the annual and daily mean value. Section 5 (Summary and conclusions) has been shortened to contain only the main three points.

The PM_{2.5} concentrations simulated with CAM-Chem using the SIMFIRE fire emissions are key in the revised manuscript. However, this additional analysis does not blend within the manuscript and is missing important information. For example, CAM-Chem does not provide directly PM_{2.5} as an output, what species did the Authors consider to calculate PM_{2.5}?

This was indeed stated in Section 2.2, 3rd paragraph of the previous revision in the following sentence: *“The computed seasonal cycle of emissions of CO, NH₃, SO₂, NO_x, black carbon and organic carbon (at monthly resolution) for current conditions and 2090 were provided as input to the Community Atmosphere Model including interactive chemistry (CAM-Chem, Lamarque et al. 2010)”*

We have now added text to Section 2.2 (line 220, after Equ. 4 and after the now revised sentence shown above) to better provide more detail on this procedure.

The Authors do not discuss how well CAM-Chem represents PM_{2.5} (without seasalt and dust). My understanding is that Tilmes et al (2016) does not evaluate PM_{2.5}, rather individual PM_{2.5} species. The simulations planned with CAM-Chem are not clear until the reader gets to the results. I understand that there were some computing time constraints. However, air quality simulations are typically performed for at least 10 years to consider any interannual climate variability. Also, the Authors could have run a simulation without fires in 2090 to determine the delta PM_{2.5} from fires.

PM_{2.5} is indeed computed as the collection of all individual aerosol components (SO₄, NO₃, NH₄, OM and BC), and limited to mostly fire and anthropogenic aerosol. As such, dust and sea-salt were not added to PM_{2.5}, as they are typically very uncertain, and limited to regions where unfortunately not many observations are available. Simulations were limited to 25 months, due to the particular model set-up, large internal variability was avoided, and small compared to the change in emissions and therefore did not warrant longer simulations. Although we did not show this in the manuscript, this information was derived from the 25-month simulations. The last paragraph of Section 2.3 has been extended substantially with a review of available validation of CAM-Chem.

The manuscript seems to be written in patches and gives the impression that the unnecessary figures from the previous version were moved to supplementary materials. For example, the paper shows first results from wildfires, peat and deforestation fires with the SSP2, SSP3 and SSP5 and RCP4.5 and RCP8.5 scenarios for 2000-2100 (Supplementary Materials). Then, it shows wildfire PM_{2.5} emissions against population for SSP3/RCP45&CLE and SSP5/RCP85&MFR for 2010, 2030, 2050 and 2090. Why aren't deforestation and peat fire emissions included in the analysis here? Finally, it presents PM_{2.5} concentrations from CAM-Chem for 2010 and 2090 for SSP3/RCP45-CLE, SSP3/RCP45-CLE and SSP5/RCP85-MFR. Again here, only wildfire emissions are considered. Can the Authors identify the key points they want to show and remove the rest?

An analysis of the magnitude of emissions from peat and deforestation fires was included in the previous revision in order to check if they are a major contribution, and if yes where. We concluded that they were not and laid this out in the manuscript. After that the manuscript proceeded without those sources for two reasons: 1) They are relevant now mostly in Indonesia because only there they occur in a region with high population densities, 2) they are difficult to project into the future. We then focussed on wildfires which are to a large extent outside the reach of policy measures, witnessed by the futility of fire suppression measures as discussed in the (new and old) text.

In the revised version, we have streamlined the presentation and focus in particular on this difference between wildfire emissions (difficult to control by policy) and anthropogenic emissions (now a routine target of policy intervention). We now only show emissions for a total of four future scenarios (summarized in a new Table 1) and only for 2090, not for 2030 or 2050 as in the previous revision. As was shown in previous publications, by 2090 we expect a substantial change in wildfire emissions and thus maximize their potential to interfere with policy targets.

Of the four future scenarios, three are fed to CAM-Chem for simulations of pollutant concentrations. Table 1 also lists two scenario differences that can be used to assess the importance of the wildfire vs. anthropogenic emissions scenarios, and these are then used to guide the analysis and discussion. Further, a guiding question is developed and explained at the end of Section 1 (Introduction) and again further explain in a new paragraph at the end of Section 2.3. The results section has been re-written to follow this guidance. We hope that this improves readability and creates are more coherent narrative for the manuscript.

One important result is the % of people that will be exposed to PM_{2.5} levels above the WHO air quality standard (10 ug/m³) as a result of fire pollution. The Authors show these results in Figure 5 and Table 2. I may be missing something in the analysis, but in Fig. 5 I do not see any population range with PM_{2.5} levels above 10 ug/m³, except for Asia. Also, over Sub-Sahara Africa and Middle East PM_{2.5} concentrations are typically dominated by dust and this emission source is not considered. Over these regions, a small influence of fire pollution may make population be exposed to PM_{2.5} levels > 10 ug/m³.

After careful consideration, we decided to drop the analysis of number of people exposed to dangerous levels of PM_{2.5} following the new focus on monthly exposure levels. Since there is no agreed air quality guidance for monthly mean PM_{2.5}, the number of people exposed becomes very strongly dependent on the finally agreed value (something that was checked but is not shown). We therefore decided to restrict the analysis to main concentrations in population-density bands, which also helped further streamlining the manuscript, but also added a new figure showing annual mean in addition to monthly PM_{2.5}.

This figure with annual mean PM_{2.5} should also reveal that averages over all areas with certain ranges of population density (shown in Fig. 3, previously Fig. 5) can mask regional differences with high values (new Fig. 4). We also added a discussion pointing out that smaller-scale urban areas with high population densities may have exposure levels that are not shown due to the fact that the highest category starts at a lower levels (line 381). We did not include a separate category for those very high population densities because wildfires are much less relevant for urban pollution, so that this is not the focus of the present study.

We do state that mineral dust or sea salt were excluded because they do not interact with the formation of aerosols from biomass burning (line 226). We also extensively discuss several limitations of the present work and recommendations for further studies, including consideration of mineral dust (e.g. line 380) in Section 4.

Some additional notes:

PM_{2.5} is PM_{2.5}.

Has been corrected throughout.

Spracklen et al. (2009) and Yue et al. (2013) were the first studies to examine PM_{2.5} impacts from future wildfires. These results should be referenced. Val Martin et al. (2015) also showed future PM_{2.5} changes as a result of fires with CAM-Chem.

We have added these references, except for Val Martin et al. (2015) as this is a different focus (no prediction of future PM_{2.5} levels).

Line 200. The Authors should cite Lamarque et al. (2012), instead of (2010).

Done.

Line 240-241 Why were the simulations done at 1 deg and the analysis at 1/2 deg? It does not make sense to me to increase resolution if the bulk fire emissions are at 1 deg.

Population density detail enters the calculations at ½ degree and affects for example the exposure calculations in Table 2. We have added a statement explain that doing the final analysis at a higher resolution helps with accounting for the impact of socio-economic factors on wildfire activity (line 168).

Line 324. I don't see how SIMFIRE differs from GFED4s in Figures S5 and S6.

We are not sure what the reviewer is referring to; Figs. S5 and S6 of the previous version show SIMFIRE simulations without anomaly correction, so GFED4s does not enter these results. These figures have now been removed.

Figure 3. I find a bit odd that changes in annual PM_{2.5} emissions wrt population in Sub-Saharan Africa and South&Southeast Asia are very similar or even identical. These two regions are different in terms of fire regimes, anthropogenic emissions and population, correct?

There was an error here, thank you for spotting this! This figure has been removed to further focus the manuscript.

Report #2, Anonymous Referee #2

In this paper, the authors use a dynamic vegetation model to probe the response of global wildfires to changes in climate and the CO₂ content of the atmosphere. While warmer temperatures from climate change will likely lead to increased wildfire activity in much of the world, increasing CO₂ concentrations could affect vegetation in ways that could diminish that activity. The paper also seeks to compare future concentrations of wildfire PM_{2.5} to those of anthropogenic PM_{2.5}, and in some places the authors find that wildfire emissions will dominate pollution sources. The paper further compares the trends of future wildfire emissions and area at different population densities.

The main update from the first version of this paper seems to be the use of chemistry model to represent concentrations of PM_{2.5} in the present-day and future (2090s) atmosphere.

We agree that this is a good summary.

The key points of the paper seem to be as follows:

A. Present-day wildfires in some developing countries are greatest in regions of intermediate density (1-100 people km⁻²).

This is not a new result (e.g. Knorr et al. 2016b, Nature Climate Change), but perhaps the manuscript did leave this impression. We have reduced the number of figures showing emissions against ranges of population density from two to one, and streamlined the results and summary significantly to make this (hopefully) much clearer.

In this study, we have taken up the analysis by population density ranges again, because we believe that highlighting the linkage between wildfire on the one hand, and anthropogenic emissions on the other to population density is an interesting approach. This is because (1) the analysis clearly shows that most anthropogenic emissions are driven by population density (which is not self-evident, considering emissions from agricultural burning, transport, exploration activities etc.), but also that population density is a main driver for wildfire emissions linked to developments in rural population, but with distinctly different but often unexpected patterns; and (2) population density in polluted areas is highly relevant for policy.

B. Under some scenarios, the CO₂ fertilization effect increases shrub, and so there is less fire. This appears to matter most in Sub-Saharan Africa.

This has been shown and discussed extensively in Knorr et al. 2016a (Biogeosciences). However, it is only discussed in the introduction, so we are not sure why the referee lists it here.

C. In the slow-urbanization scenario in some regions – e.g., Australia – more people move into fire-prone areas and so more people are exposed to greater fire emissions.

This has also been discussed much more in Knorr et al. (2016a,b). Because the newly revised manuscript focuses more on concentrations and leaves out detailed comparison of urbanisation effects, it is now off-scope, but not a new result anyway.

D. In the Maximum Feasible Reduction scenario, in which anthropogenic emissions decline the most, wildfires emissions will dominate over anthropogenic emissions in many regions, especially in China and Southeast Asia, and parts of Africa and South America

This is indeed the main result. We refer to this in more general terms in a new paragraph at the start of Section 4 (Discussion) and in more detail in the following paragraph. The point made by the reviewer is also referred to by two of the remaining three bullet points of the summary (Section 5).

Main points.

1. The paper is very difficult to read, and discerning the main points is challenging. Much of the paper rambles, and the assistance of a really good editor is needed. The use of cumbersome acronyms is especially confusing – e.g., one scenario is called CLE/SSP5/RCP8.5, and another one is called MFR/SSP5/RCP8.5.

We hope that this has improved considerably by concentrating on just four scenarios, combining CLE and MFR on the one hand with SSP5/RCP8.5 and SSP3/RCP4.5 on the other (four combinations in total, which are listed in a new Table 1). We also tried to help the reader by frequently referring to SSP5/RCP8.5 as the high wildfire emissions scenario and to SSP4/RCP4.5 as the low wildfire emissions scenario. In addition, we considerably revised Section 2.2 of the methods in order to better describe the method used to calculate the final emissions fields.

2. The main goal of the paper appears to be to determine whether wildfires will affect our ability to meet WHO air quality guidelines. But it's very difficult to determine from the current set of figures where wildfires would pose such a challenge. See points #3 and #4.

We did not include a simulation with only wildfire pollutant sources, but rather assess the importance of differences between wildfire scenarios at a scenario of very low anthropogenic emissions. The regions where wildfires matter can thus be seen in Fig. 5 for peak-month, and the new Fig. 4 for annual mean concentrations by comparing different scenarios. Text has been added to the last paragraph of Section 3.2 that describes the regions that stand out as being largely influenced by wildfire emissions (line 425):
"For all scenarios, large sparsely populated regions in central South

America, northeastern Siberia and northwestern Canada experience monthly pollution levels in excess of $10\mu\text{g}/\text{m}^3$ and to a lesser extent $25\mu\text{g}/\text{m}^3$. This result is largely independent of the anthropogenic-emissions scenario (comparing Fig. 5b with 5c, d). In regions that show large reductions of annual mean pollution levels for MFR, monthly maximum concentrations can still substantially surpass $10\mu\text{g}/\text{m}^3$ even under an MFR scenario (Africa, Southeast Asia and, to some extent for SSP5/RCP8.5, south-eastern China), or even $25\mu\text{g}/\text{m}^3$ (Africa, South-East Asia)."

Besides, a simulation with only fire sources poses some theoretical issues, as we may run into non-linear responses. It is not impossible but beyond the scope of this paper.

3. The figures contain more information than is needed, obscuring the main points. For example, it appears we are meant to compare the dashed lines in Figures 3 and 4. These lines represent wildfire emissions for different timeslices at different population densities in different regions for different scenarios. But these differences across cases seem very small on the page, and are nearly nonexistent in some cases. The use of log-scales makes it even more difficult to compare fire activity across regions, and the very tiny values in some regions (e.g., the Middle East) seem not worth showing. In well-composed figures, the main messages pop out at the reader, but that doesn't happen here. The plots should be carefully designed to illustrate the main points of the paper. More detailed plots should be relegated to the Supplement.

We have streamlined the results section by removing time slices that were not used for the CAM-Chem simulations. In the newly revised version, the results section is now guided by the question of which emissions are higher as one requisite for the importance (Fig. 1), if they occur close to population for a scenario selected from Fig. 1 (Fig. 2), how emissions relate to concentrations (Fig. 3), how concentrations relate to population patterns (Fig. 3), and finally identification of regions where each source matters (Figs. 4 and 5). Fig. 5 shows peak-month concentrations, as a conclusion from the findings of Fig. 4. We believe that this new choice of figures constitutes a much better and more coherent story.

4. More issues with figures. The plots show annual emissions of PM_{2.5}, but of course wildfire has strong seasonal peaks. In this way, the plots are misleading. Also misleading is the apparent neglect of secondary organic aerosol (SOA) from biogenic emissions. By showing only emissions, Figure 1 implies that PM_{2.5} in the Amazon and the Southeast US is anthropogenic, but of course SOA dominates the aerosol burden in these regions. And SOA is expected to increase in the future atmosphere, due to the impact of rising temperatures on biogenic emissions. There is also inadequate treatment of dust. Figure 5 shows total PM_{2.5} and so it's unclear to what extent wildfires drive these concentrations.

Fig. 5 shows season-peak concentrations, but a newly added Fig. 4 shows annual mean concentrations, so we hope that it is now clearer that we do include seasonal variations. The impact of either anthropogenic or wildfire

emissions are seen by comparing different scenarios according to the newly added Table 1. For example, the difference between Fig. 4 or 5 b and d is due to the anthropogenic emissions scenario, while the difference between Fig. 4 or 5 c and d is due to the wildfire emissions scenario.

SOAs and biogenic volatile compounds were actually not ignored in the simulations (see line 239).

5. Perhaps the authors should just focus on a few key regions (or better subregions) and show not annual emissions but fire season concentrations for the present-day and 2090s at different population densities. The contributions of wildfires to total PM_{2.5} should be clearly denoted.

In principle, we agree with the referee that figures should convey the main message. However, we also believe that the main advantage of a full-format journal such as ACP is that readers can be guided through a narrative that presents the breadth of results, leading to intermediate conclusions that then guide further analysis. We have therefore retained some breadth with the comparison of anthropogenic vs. wildfire emissions, or the analysis of emissions vs. population density, of which not all enter the final analysis of simulated pollution levels. And we have retained the presentation of all global regions (with small exceptions). Fire-season concentrations enter Fig. 5.

6. The paper does not sufficiently reference previous papers – e.g., Pechony and Shindell (2010). This reviewer asked for this reference in the first review. Also, it is not true that wildfires will increase everywhere in the future climate: Yue et al. (2015) shows decreasing fires in the future climate in parts of Canada.

We did not include Pechony and Shindell 2010, because it simulated neither emissions nor burned area, but a measure of fire activity that is closer to number of fires, and less relevant for emissions than for example burned area. We did, however, add Yue et al. (2015) along with Yue et al. (2013).

Title:

Wildfire air pollution hazard during the 21st century

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Abstract:

Wildfires pose a significant risk to human livelihoods and are a substantial health hazard due to emissions of toxic smoke. Previous studies have shown that climate change, increasing atmospheric CO₂, as well as human demographic dynamics can lead to substantially altered wildfire risk in the future, with fire activity increasing in some regions and decreasing in others. The present study re-examines these results from the perspective of air pollution risk, focussing on emissions of airborne particulate matter (PM_{2.5}), combining an existing ensemble of simulations using a coupled fire-dynamic vegetation model with current observation-based estimates of wildfire emissions and simulations with a chemical transport model. Currently, wildfire PM_{2.5} emissions exceed those from anthropogenic sources in large parts of the world. We further analyse two extreme sets of future wildfire emissions in a

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socio-economic, demographic climate change context, and compare them to anthropogenic emission scenarios reflecting current and ambitious air pollution legislation. In most regions of the world, ambitious reductions of anthropogenic air pollutant emission have the potential to limit mean annual pollutant PM_{2.5} levels to comply with WHO air quality guidelines for PM_{2.5}. Worse-case future wildfire emissions are not likely to interfere with these annual goals, largely due to fire seasonality, as well as a tendency of wildfire sources to be situated in areas of intermediate population density, as opposed to anthropogenic sources that tend to be highest at the highest population densities. However, during the high-fire season, we find many regions where future PM_{2.5} pollution levels can reach dangerous levels even for a scenario of aggressive reduction of anthropogenic emissions.

1 Introduction

Wildland fires – or in short “wildfires” – are burning events that occur in natural or semi-natural landscapes such as (managed or un-managed) forests, shrublands, or grazing lands including savannahs. They are a major natural hazard (Bowman et al. 2009) and an important source of air pollutants (Langmann 2009), which can impact air pollution thousands of kilometres downwind (Lee et al. 2005). Wildfires also play an important role in several atmospheric chemistry–climate feedback mechanisms (Fiore et al. 2012). Emissions of fine aerosol particles, i.e. particulate matter up to an aerodynamic diameter of 2.5 micrometers (PM_{2.5}), are of particular health concern, with no known safe concentration in air, as noted by the World Health Organization (WHO 2006). Wildfires can be an important source in large, more remote areas (Granier et al. 2011, van der Werf et al. 2010), even though anthropogenic emissions are higher globally. There is a widely held view among both the general public and members of the research community that wildfire occurrence and severity have been

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94 increasing in recent decades, and will continue to increase due to climate change
 95 (Doerr and Santin 2016). Moreover, efforts to reduce anthropogenic emissions (e.g.
 96 EEA 2014) will increase the relative importance of other emission sources.
 97 Climate warming has already led to more frequent hot and dry weather in many
 98 parts of the globe, increasing the probability of wildfires (Flannigan et al. 2009), and
 99 this is expected to continue into the future. Studies based on calculated fire severity
 100 indices under climate change argue for large increases in burned area (Flannigan et al.
 101 2005 for Canada; Amatulli et al. 2013 for southern Europe) and resultant, pollutant
 102 emissions (Spracklen et al. 2009, Yue et al. 2013 for the western US), with some
 103 regional exceptions of declining emissions due to increased precipitation (Yue et al.
 104 2015 for a few sub-regions in northern Canada). However, a long-term increase in the
 105 length of the fire season or in weather conditions conducive of wildfires does not
 106 necessarily lead to increases in burned area (Doerr and Santin 2016). This is because,
 107 at longer time scales, vegetation responds not only to climate change, but also directly
 108 to rising atmospheric CO₂ levels (Buitenwerf et al. 2012, Donohue et al. 2013). While
 109 CO₂ fertilization will lead to increased fuel load, enhancing emissions, it also leads to
 110 an increase in woody as opposed to herbaceous vegetation, with on average lower
 111 emissions due to decreased fire spread in less flammable shrublands (Kelley and
 112 Harrison 2014, Knorr et al. 2016a). Indeed, simulations with coupled fire-vegetation
 113 or statistical models generally show less increase in burned area (Kloster et al. 2010,
 114 Knorr et al. 2016b) or number of fires (Krwachuk et al. 2009), when accounting not
 115 only for climate, but also for these vegetation effects.
 116 Another factor that has so far received less attention are growth and changes in
 117 human population size and distribution. Contrary to common perception, higher
 118 population density tends to be associated with lower wildfire risks when measured by

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139 burned area (Archibald et al. 2009, 2010, Lehsten et al. 2010, Knorr et al. 2014,
 140 Bistinas et al. 2014), even though higher population density in rural and remote
 141 regions tends to lead to more, but on average smaller fires (Archibald et al. 2009,
 142 2010). This can be explained by the concept of the ignition-saturated fire regime,
 143 which is reached at very low levels of population density. Above this level, human
 144 impact manifests itself less in enhancing burned area (by igniting fires), but more by
 145 creating barriers to and suppressing fire spread, thus reducing area burned (Guyette et
 146 al. 2002). Indeed, coupled vegetation–fire models that include the effects of changing
 147 human population size and spatial distribution suggest a reduced rate of increase of
 148 fire activity during the 21st century, compared to simulations not accounting for
 149 demographic changes (Kloster et al. 2010). Some studies showed even a decline in
 150 burned area (Knorr et al. 2016b) or emissions (Knorr et al. 2016a) for moderate levels
 151 of climate change, when combined with slow urbanisation and high population
 152 growth. These results are backed by observational evidence of a long-term declining
 153 trend in past fire activity or emissions from wildfires (Marlon et al. 2008, Wang et al.
 154 2010, van der Werf et al. 2013), and more recent negative trends in northern Africa
 155 that have been related to the expansion of cropland (also resulting from increasing
 156 population density; Andela and van der Werf 2014). Furthermore, the impacts of
 157 emissions from wildfires on human society are also largely determined by population
 158 growth and their spatial distribution (Knorr et al. 2016 b). It is therefore important to
 159 consider not only climate and CO₂ scenarios, but also scenarios of demographic
 160 changes.

161 The overarching research question addressed in this paper is whether socio-
 162 economic developments influencing both greenhouse gas emissions as well as human
 163 population size and spatial distribution might impact wildfire emissions enough to

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183 make a difference for meeting the WHO air quality target, provided anthropogenic
 184 PM_{2.5} emissions are reduced aggressively. The work in this study takes the following
 185 steps towards this end: Building on a similar study for Europe (Knorr et al. 2016 c)
 186 for computing emissions, we use a state-of-the art chemical transport model to
 187 compute future levels of human exposure to PM_{2.5}, using observation-based wildfire
 188 emissions, combined with relative changes in emissions from vegetation–fire model
 189 simulations. Our analysis focuses on the relative importance of changes in global
 190 wildfire emissions for air quality and atmospheric pollutant load, as compared to
 191 anthropogenic sources.

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192 **2 Methods**

193 **2.1 Vegetation–fire model and driving data**

194 We use the LPJ-GUESS global dynamic vegetation model (Smith et al. 2001,
 195 Ahlström et al. 2012) coupled to the global semi-empirical fire model SIMFIRE
 196 (Knorr et al. 2014), with details given by Knorr et al. (2016a). Shortly, LPJ-GUESS is
 197 a patch-scale dynamic vegetation model that represents age cohorts of perennial
 198 vegetation and computes vegetation establishment and growth, allocation of carbon
 199 pools in living plants, and turnover of carbon in plant litter and soils. SIMFIRE
 200 provides burned area to LPJ-GUESS on an annual basis, which then evokes plant
 201 mortality according to a plant-functional-type (PFT) dependent probability. Specified
 202 fractions of plant litter and live leaf biomass are burnt and emitted into the air in a
 203 fire, while the remaining biomass of the killed vegetation is transferred to the litter
 204 pool of LPJ-GUESS (see Knorr et al. 2012). Population data needed to drive
 205 SIMFIRE are based on gridded data from HYDE 3.1 (Klein-Goldewijk et al. 2010) up
 206 to 2005, and then re-scaled using per-country relative growth in rural and urban
 207 population, retaining the urban masks of the HYDE data. Grid cells with more than

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233 50% past or future cropland area in either the RCP6.0 or 4.5 land use scenarios (Hurt
234 et al. 2011; see Section 2.2) were excluded from the calculations (see Knorr et al.
235 2016a,b for details).

236 In order to compute emissions of chemical species, we use the emission factors of
237 the Global Fire Emissions Database version 4 (GFED 4, van der Werf et al. 2010,
238 based mainly on Akagi et al. 2011, see <http://www.falw.vu/~gwerf/GFED/GFED4>),
239 which are fixed ratios between emission rates of various pollutant species and rates of
240 combustion of dry biomass differentiated by where the fire occurs: (1) savannahs and
241 grasslands, (2) tropical, (3) boreal and (4) temperate forests. In order to select the
242 appropriate emissions factor, we assign a grid cell to (1) if the PFT with the largest
243 leaf area index at full leaf development is a grass, to (2) if it is a tropical tree, and to
244 (3) or (4) if it is a boreal or temperate tree, respectively (see Knorr et al. 2012 for a list
245 of PFTs used).

246 2.2 Simulations and scenarios

247 Climate simulations were driven by output from an ensemble of eight global climate
248 models from the Climate Model Intercomparison Project 5 (CMIP5, Taylor et al.
249 2012) for two climate scenarios based on the Representative Concentration Pathways
250 (van Vuuren et al. 2011) RCP4.5 with moderate, and RCP8.5 with high degree of
251 climate change. Simulations for 1901 to 2100 are carried out on a global equal-area
252 grid with 1° x 1° spatial resolution at the equator, but constant east-west spacing of
253 the grid cells when moving towards the poles in order to keep the grid cell area
254 constant (Knorr et al. 2016a). These climate scenarios were combined with population
255 and urbanisation projections following the Shared Socioeconomic Pathways (SSPs,
256 Jiang 2014). The SSPs are based on qualitative narratives of five different
257 development pathways, which have been translated into quantitative projections of a

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range of socio-economic, demographic and biophysical factors. Here, we focus on
two combinations that represent, respectively, the highest and lowest wildfire
combusted carbon emissions globally (Knorr et al. 2016a). These are RCP8.5 (high
greenhouse-gas emissions, strong warming) combined with SSP5 (rapid fossil-fuel
driven economic growth with globally low population growth and fast urbanisation
leading to land abandonment and increased wildfire activity), and RCP4.5 (moderate
greenhouse-gas emissions and climate warming) combined with SSP3 (globally high
population growth and slow urbanisation leading to increased fire suppression). Note,
however, that in contrast to developing countries and the world as a whole, high-
income countries have high population growth for SSP5, but low population growth
for SSP3.

In addition to the emissions simulated by LPJ-GUESS-SIMFIRE, we also use the
GFED4.1s observation-based emissions fields for wildfires (van der Werf et al. 2010,
updated using Randerson et al. 2012 and Giglio et al. 2013) aggregated to 0.5° x 0.5°
resolution and then re-scaled in time by country or region (following the methodology
of Knorr et al. 2016c). For larger countries, scaling is done by sub-national regions,
which were chosen in such a way as to isolate major fire areas found in GFED4.1s
(For a list of regions/countries, see Table S1 in the Supplementary Information). The
use of country boundaries and performing the analysis at 0.5x0.5° instead of the 1°-
resolution used by LPJ-GUESS-SIMFIRE better accounts for the high degree of fire-
policy or cultural impact on fire regimes (Bowman et al. 2009).

For wildfires, we use the sum of boreal forest fires, temperate forest fires and
savannah fires from GFED4.1s. Agricultural waste burning from GFED4.1s has been
excluded from the calculations. Instead, we use anthropogenic emissions that include
agricultural burning from the ECLIPSE data set (Granier et al. 2011). Deforestation

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331 fires (caused by deforestation activities) and peat fires (fires occurring in forested or
 332 non-forested peatlands, see van der Werf et al. 2010) were found to be minor sources
 333 globally, despite of their regional importance mainly in Southeast-Asia (Fig. S1,
 334 Table S2)

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335 In contrast to Knorr et al. (2016c), where changes in the spatial distribution of
 336 population within a country did not affect predicted wildfire emissions, here we also
 337 account for demographic effects at the grid-cell scale. To do so, we combine a scalar
 338 accounting for climate and vegetation effects, f_{cv} , which is uniform in space across
 339 each region/country, with a scalar accounting for population effects, f_p , which is
 340 applied at each grid cell separately:

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341
$$E^S(x,y,m) = f_{cv}^S(R(x),y) * f_p(\Delta p'(x,y)) * E_{GFED}^S(x,m) \quad (1)$$

 342 with E^S the re-scaled per-grid-cell emissions of chemical species S , x the geographic
 343 location on the $0.5^\circ \times 0.5^\circ$ grid used for the analysis, y year, m month of the year, $R(x)$
 344 the country or region found at location x , and E_{GFED}^S the (per-grid-cell) emissions
 345 climatology of species S from GFED 4.1s averaged over 1997 to 2014. The
 346 population effect, f_p , is equal to the population multiplier of SIMFIRE (Knorr et al.
 347 2016a):

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348
$$f_p(\Delta p') = \exp(-0.0168 * \Delta p'). \quad (2)$$

 349 with $\Delta p'(x,y) = p'(p(x,y)) - p'(p(x,y_0))$, where $p'(p)$ is the minimum of population
 350 density p and 100 inhabitants per km^2 . $y_0=2010$ is the reference year relative to which
 351 future emission levels are computed. The cap at $100 / \text{km}^2$, which is only used for
 352 scaling observation-based inventories by LPJ-GUESS-SIMFIRE output (not by
 353 SIMFIRE itself) is used to prevent unrealistically large relative increases in emissions
 354 resulting from the scaling procedure, when population density decreases from present
 355 values that are much above $100 / \text{km}^2$. In other words, we consider areas with a

432 population density [above](#) 100 /km² to be essentially wildfire free. The combined
 433 effect of climate and vegetation on emissions is defined as:
 434
$$f_{cv}(R,y) = \{ \sum_{x' \in R} E_{SIM}^S(x',y) / \sum_{x' \in R} E_{SIM}^S(x',y_0) \} \cdot \dots$$

 435
$$\dots \{ \sum_{x' \in R} f_p(p(x',y)-p(x',y_0)) E_{GFED}^S(x',m) / \sum_{x' \in R} E_{GFED}^S(x',m) \} \quad (3)$$

 436 Here, E_{SIM}^S are LPJ-GUESS-SIMFIRE emissions of species S computed on the 1°
 437 equal-area grid. The sums are over all grid cells x' of the 1° equal-area grid that
 438 belong to region/country R . The first term in curly brackets is the SIMFIRE simulated
 439 relative change in emissions or region R by year y (shown in Fig. S2), which is
 440 divided by the projected change in emissions only due to changes in population
 441 density (second term in curly brackets; see Fig. S3 for a map of projected changes
 442 only due to population density, i.e. $f_p(p'(x,y)-p'(x,y_0))$ for $y=2090$). Finally, we
 443 compute future projected emissions of species S as the 21-year climatological mean
 444 around 2090:
 445
$$E_{2090}^S(x,m) = \sum_{y \in [2080, 2100]} E^S(x,y,m) \quad (4)$$

 446 Countries where 90% or more of the grid cells of the LPJ-GUESS grid have
 447 either a current or future cropland fraction of $\geq 50\%$ (highly agricultural regions:
 448 Moldavia and Bangladesh), or for which LPJ-GUESS simulates zero current
 449 emissions in at least one simulation (Greenland) were excluded from this scaling
 450 procedure by setting $f_{cv}(R,y)=f_p(p)=1$. Note that the procedure retains the seasonal
 451 cycle of the GFED4.1 emissions, $E_{GFED}(m)$, by scaling each month by the same factor.
 452 The computed seasonal cycle of anthropogenic emissions of CO, NH₃, SO₂, NO_x,
 453 black carbon and organic carbon (at monthly resolution) for current conditions and
 454 2090 were provided as input to the Community Atmosphere Model including
 455 interactive chemistry (CAM-Chem, Lamarque et al. 2012). In the present
 456 configuration, CAM-Chem simulates the aerosol distributions for all types (i.e.,

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sulfate, nitrate, black carbon, organic carbon, sea-salt and mineral dust). From those, the PM_{2.5} distribution is calculated using all components except dust or sea-salt, since those are not affected by biomass burning, and assuming all component mass to be present below 2.5 µm. For current emissions we take $E_{GFED}^S(x,m)$, and for 2090 emissions $E_{2090}^S(x,m)$ (Eq. 4). Anthropogenic emissions are constant over the year.

The configuration of CAM-Chem is identical to the one used in the recent Chemistry-Climate Model Initiative simulations discussed in Tilmes et al. (2016) under REF-C1 (specified sea-surface temperatures and sea-ice distribution), except for using a higher horizontal resolution of 0.9° latitude x 1.25° longitude. The model has 26 vertical layers from the surface to approximately 40 km. CAM-Chem is here used as a chemical transport model, with the meteorology being the same between simulations with different emissions fields, thus excluding the effect of changing atmospheric composition on meteorology. This is done so that short simulations are sufficient to identify the chemical impacts of different emission scenarios. Emissions of sea-salt, dust and biogenic volatile organic compounds (VOCs, i.e. isoprene and mono-terpenes, which are precursors to secondary-organic aerosols) are also identical between the different simulations, and are computed as in (Tilmes et al. 2016). For all species, emissions (including biomass burning) are included as flux boundary conditions to the vertical diffusion module, and are therefore quickly (within hours) redistributed within the boundary layer.

The current representation of aerosols in CAM-Chem has been extensively tested and compared with observations. In particular, Lamarque et al. (2012) provide a comparison of present-day observations of the IMPROVE network over the conterminous United States, indicating an overall good representation of the probability density function for all species considered in the present study. In

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541 addition, Shindell et al. (2013) have shown that CAM-Chem indicated the lowest bias
 542 of the models when compared to observed aerosol optical depth. Finally, CAM-Chem
 543 results have been used in several health impacts studies (West et al., 2013; Silva et al.,
 544 2013, 2016) and is usually very consistent with other chemistry-climate models used
 545 for those analyses. Four simulations are carried out for a period of 25 months each,
 546 and a mean annual cycle is computed from months 2 to 25. While there is some
 547 residual interannual variability in the two meteorological years simulated by the
 548 model, this signal is small compared to the impact associated with the changes in
 549 emissions (not shown).

550 2.3 Analytical Framework

551 Our analysis focuses on two time windows, current and 2090. To assess the relevance
 552 of wildfire PM_{2.5} emission rates, we compare them to those from anthropogenic
 553 sources, and also judge simulated surface concentrations by their proximity to the
 554 World Health Organization (WHO) air quality guidelines of 10 µg/m³ on an annual
 555 average, keeping in mind that there is no established safe upper limit and that the
 556 target was set considering background concentrations of 3–5 µg m⁻³ in North America
 557 and Western Europe.

558 For anthropogenic emissions, we use the ECLIPSE-GAINS-v4a data (Amann et
 559 al., 2011) developed as part of the ECLIPSE project (Granier et al. 2011, Klimont et
 560 al. 2013, Stohl et al., 2015). This dataset provides two scenarios: current legislation
 561 (CLE), and maximum feasible reductions (MFR) on top of business-as-usual
 562 projections until 2050 from the Energy Technology Projections study by the
 563 International Energy Agency (IEA, 2012), which are considered roughly equivalent to
 564 RCP6.0. MFR corresponds to a policy and technology driven abatement scenario,
 565 implementing all currently known technologies at a reasonable cost, with the aim,

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among others, to lower PM_{2.5} emissions to levels of limited health impacts (Amann et al. 2011). For the present, we use emissions for 2010 under the CLE scenario. Following Knorr et al. (2016~~c~~), we estimate 2090 CLE emissions assuming constant *per capita* emissions after 2050, but changing population densities according to the SSP3 scenario, which directly affects magnitude and spatial patterns of emissions. For MFR, 2090 emissions are assumed half of the corresponding 2050 levels (i.e. somewhat optimistic compared with e.g. Braspenning-Radu et al., 2016). In principle the emissions under CLE and MFR conditions may differ also for different SSP and climate mitigation assumptions (Rao et al., 2016; Braspenning-Radu et al, 2016). However, the assumptions above can be seen as a schematic approximation of a wide range of possible air pollution emission futures that have become available in the recent literature,

To facilitate the regional aspects of our analysis, we use a global map of nine major world regions (see Fig. S4). Of these, three belong to the high-income group of countries of the SSP scenarios (see Jiang and O'Neill 2015): High-income Europe, Australia & New Zealand, and North America. Countries of Europe belonging to the middle-incoming group were assigned to the region Eastern Europe and Central Asia, which also includes Russia. High-income countries of the Middle East (Israel, oil-rich states of the Persian Gulf) or East Asia (Japan, South Korea), which only account for a very small fraction of wildfire emissions in their respective regions, have been excluded from the analysis.

In this study, we address the question whether there is a risk that the combined direct (as estimated by changing population patterns) and indirect impacts (through climate change) of human activities on wildfire pollutant emissions will compromise meeting the WHO guideline value of 10 $\mu\text{g} / \text{m}^3$, under a scenario in which

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anthropogenic PM_{2.5} emissions are reduced aggressively. The range of plausible future air quality policy scenarios is spanned by the difference between our projections using MFR and CLE assumptions. For a range of plausible wildfire emissions scenarios, we consider the combinations SSP5/RCP8.5 (highest) and SSP3/RCP4.5 (lowest global wildfire emissions). Of the possible four emission combinations, we select three in order to assess the impact of either the plausible range of wildfire, or of air quality policy scenarios (see Table 1). These three scenarios plus current emission fields are used for simulating present and future (2090) PM_{2.5} surface concentrations with CAM-Chem.

3 Results

3.1 *Comparison of wildfire and anthropogenic pollutant emissions*

In order for wildfire emissions to be relevant for atmospheric pollution levels, at least two conditions should be met: (1) they need to be of a similar or greater magnitude compared to anthropogenic sources, and (2) they should be in or close to populated areas. Therefore we first compare emission levels of both types of sources for the current and future time slices. Fig. 1 shows areas where wildfire emissions currently exceed those from anthropogenic sources (in either red or light blue; shown in all panels). At first sight, these regions appear to lie in relatively remote areas of wildfire-prone regions (cf. Fig. S5, e.g. the boreal forest zones of Canada and Alaska, eastern Siberia, western US, the Brazilian interior, and in Africa away from the main population centres of Nigeria, Ethiopia and Kenya). A breakdown of emissions by population density shown for the nine world regions (cf. Fig. S4) shows that current (or future) anthropogenic emissions display a universal increase with increasing population density (solid blue lines in Fig. 2). Wildfire emissions also often increase with increasing population density, and only for North America per-area emissions

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decrease with increasing population density across the entire range of population densities (dashed blue lines in Fig. 2, i.e. current wildfire emissions). Most of the remaining regions have peaked distributions with highest per-area emissions in some intermediate population-density category, and it is not the most remote and sparsely inhabited areas that have the highest emissions. Therefore, despite of a dominance of anthropogenic emissions in the more densely populated areas, wildfires emissions are often found to be an important pollution source in regions with intermediate population density, meaning that they approach or even exceed anthropogenic sources for population densities in the range 10 to 100 / km² (Sub-Saharan Africa, Latin America & Caribbean, South & Southeast Asia, Australia & New Zealand). This 10 to 100 / km² category is significant not only because of its relatively high population density, but also because it represents the wildland-urban interface, which is often particularly prone to wildfires (Syphard et al. 2007).

For the simulated future, we either expect anthropogenic emissions to surpass those from wildfires in many regions (mainly in Africa, light blue areas in Fig. 1a, b) for the CLE scenario, or the reverse for the MFR scenario, where wildfire emissions surpass anthropogenic sources in a wide range of regions (yellow areas in Fig. 1c, d: South America, Central America, Africa, Eastern and Southern Europe, Central Asia, Southeast Asia, and southern China). This dominant role of the anthropogenic-emissions scenario is largely independent of the wildfire emissions scenario (i.e. the difference between Fig. 1a and b or between Fig. 1b and d is much small than the difference between Fig. 1a, b on the one hand and Fig. 1 c, d on the other). The strongest impact of the wildfire emissions scenario is found for Sub-Saharan Africa for the CLE scenario, where SSP3/RCP4.5 shows many more regions newly dominated by anthropogenic emissions than SSP5/RCP8.5 (Fig. 1a vs. Fig. 1b). This

is due to declining wildfire emissions in that scenario (see Fig. S2a). The scenario in which wildfire emissions are most likely to become a relevant source of PM_{2.5} emissions is represented by the combination MFR/SSP5/RCP8.5 (Fig. 1d and Fig. 2, red lines; cf. also Fig. S2b).

The MFR scenario assumes a decline in anthropogenic emissions by approximately one order of magnitude in areas with at least 10 people / km², with somewhat less decrease for the more sparsely populated categories (Fig. 2). When compared to this magnitude of change, the simulated increase in wildfire emissions for SSP5/RCP8.5 is much smaller throughout. The highest absolute increase for wildfire emissions is found for Sub-Saharan Africa (in the 1 to 100 / km² categories), Latin America (0.1 to 1 / km²), South & Southeast Asia (below 10 / km²) and Developing East Asia (1 to 100 / km²). As a consequence mainly of the reduction implicit in the MFR scenario, wildfire emissions approach or exceed anthropogenic emissions even in the most densely populated category (>100 / km²) for Sub-Saharan Africa, Latin America & Caribbean, Eastern Europe-Russia-Central Asia, South & Southeast Asia and Australia & New Zealand. Note that the areas of the population density categories shift as well (see dotted lines in Fig. 2 and Fig. S5), and that the underlying population patterns used to compute the MFR scenario do not necessarily match those from SSP5.

3.3 Changing patterns of PM_{2.5} pollutant exposure

Simulated changes of PM_{2.5} concentrations between current and 2090 for the high-wildfire emissions scenario combined with MFR (Fig. 3 blue to red line, corresponding to emissions of Fig. 2, same colours) indicate a reduction in PM_{2.5} pollutant concentration by between one half and three-quarters in the most densely

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populated category ($>100 / \text{km}^2$) in each world region, resulting from the large reductions in anthropogenic emissions. Both the current and future pollution levels are substantially lower in the most sparsely populated categories. For current emissions, only two regions show average anthropogenic concentrations above the WHO air-quality target of $10 \mu\text{g} / \text{m}^3$ (Developing East Asia, South & Southeast Asia; note the different axis scales), and none for 2090. We emphasize that the results presented here do not include sea salt and dust derived particulate matter, and are not downscaled using information on the emissions attributable to urban regions. In fact, in Figs. 2 and 3 small areas with typical urban population densities e.g. above $1000 / \text{km}^2$ are included in same category in with the much larger areas in the 100 to 1000 km^{-2} range. Therefore these results are likely to represent a lower limit of population-weighted $\text{PM}_{2.5}$ concentrations, which are consequently closer to the WHO guidelines. In general, relative changes in per-area emissions (Fig. 2) are much larger than relative changes in concentrations (Fig. 3), mainly due to the effect of long-range atmospheric transport of aerosols and precursor gases.

In contrast to the current situation with a steady increase in pollutant concentrations with population density (dark blue in Fig. 3), for 2090 concentrations peak at an intermediate level of population density in most regions (red line: Sub-Saharan Africa, Latin America & Caribbean, Eastern Europe-Russia-Central Asia, Australia & New Zealand, High-Income Europe), reflecting the increased importance of wildfires for the spatial distribution of pollutant concentrations. This change over time is most pronounced for Sub-Saharan Africa, where emissions from wildfires far outweigh anthropogenic $\text{PM}_{2.5}$ emissions by 2090 (Fig. 2, red lines).

The effect of the anthropogenic-emissions scenario can be seen by comparing the MFR and CLE scenarios for the case of the high-wildfire emissions scenario (red and

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797 green lines in Fig. 3, cf. Table 1). CLE/SSP5/RCP8.5 compared to
 798 MFR/SSP5/RCP8.5 leads to large reductions in concentrations in pollutant
 799 concentrations in all regions by 2090, in particular for the most densely populated
 800 categories. Compared to this large effect of the anthropogenic emissions scenario, the
 801 effect of the wildfire scenario (difference between high and low wildfire-emissions
 802 scenario, red and light-blue lines in Fig. 3) is relatively small and only visible in Sub-
 803 Saharan Africa, South & South-East Asia and Developing East Asia.

804 Simulated spatial patterns of PM_{2.5} concentrations for current conditions as well
 805 as for the CLE scenario (Fig. 4 a, b) show large parts of central Africa as well as most
 806 of South, East and Southeast Asia with concentrations above the WHO air quality
 807 guidance value. For MFR (Fig. 4c, d), however, only limited areas in central Africa
 808 show levels exceeding 10 µg / m³. For the high-wildfire scenario (SSP5/RCP8.5, Fig.
 809 4d), concentrations are higher than for low-wildfire emissions (SSP3/RCP4.5, Fig.
 810 4c), but the areas are much more sparsely populated (see Fig. S4). It appears therefore
 811 that under the MFR scenario, annual mean anthropogenic and wildfire-induced
 812 concentrations will remain below the WHO air-quality guidance value for most areas
 813 of the globe. Even though the previous analysis (Fig. 3) showed wildfire emissions
 814 becoming the dominant source for the high-wildfire scenario, it appears the levels of
 815 emissions are not high enough to bring concentrations above the annual WHO air-
 816 quality value.

817 Further analysis hereafter focuses on monthly mean PM_{2.5} concentrations,
 818 assuming this to be a health-related variable more appropriately reflecting the
 819 seasonality of fire emissions. We note that there is currently no guideline for monthly
 820 concentrations, but the WHO does provide guidelines for 24-hour means that amount
 821 to 25 µg m⁻³. We here assume that formally declared safe monthly PM_{2.5}

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concentrations would be somewhere between 10 and 25 $\mu\text{g m}^{-3}$. When investigating monthly $\text{PM}_{2.5}$ concentrations, the picture changes considerably. For all scenarios, large but sparsely populated regions in central South America, northeastern Siberia and northwestern Canada (cf. Fig. S5) experience monthly pollution levels in excess of 10 $\mu\text{g m}^{-3}$ (Fig. 5) and to a lesser extent 25 $\mu\text{g m}^{-3}$. This result is largely independent of the anthropogenic-emissions scenario (comparing Fig. 5b with 5c, d). In regions that show large reductions of annual mean pollution levels for MFR, monthly maximum concentrations can still substantially surpass 10 $\mu\text{g m}^{-3}$ even under an MFR scenario (Africa, Southeast Asia and, to some extent for SSP5/RCP8.5, south-eastern China), or even 25 $\mu\text{g m}^{-3}$ (Africa, South-East Asia). The wildfire scenario is thus found to be of globally limited, but regionally important significance for air pollution-related health impacts.

4 Discussion

In most regions of the world, decisive and effective reductions of anthropogenic air pollutant emissions will likely be able to limit pollutant levels below the WHO threshold of 10 $\mu\text{g m}^{-3}$ annual mean $\text{PM}_{2.5}$ concentrations irrespective of the evolution of future wildfire emissions. This is mainly the result of the highly seasonal character of wildfire emissions, as well as a tendency of wildfire sources to be situated in areas of intermediate population density, as opposed to anthropogenic sources that tend to be highest at the highest population densities. We therefore see a shift of the highest pollution levels away from the most densely populated areas, which reduces the impacts of future human exposure to $\text{PM}_{2.5}$. However, as the example of the Russian wildfires in 2010 shows (Kaiser et al. 2012), high pollution levels even from single wildfires can persist over many weeks and thus pose a significant health threat over extended lengths of time. On a seasonal basis, we

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974 therefore find many regions where pollution levels can reach dangerous levels even
 975 for a scenario of aggressive reduction of anthropogenic emissions.

976 Previous simulations with chemistry-climate models using RCP emission
 977 projections have already shown a strong future downward trend in East and South
 978 Asia, driven by reduced anthropogenic emissions, but no notable trend in Africa
 979 (Fiore et al. 2012). Knorr et al. (2016a) have shown a general picture of climate-
 980 driven fire emission increases, both for an RCP 4.5 and 8.5 scenario, that may be
 981 overridden by demographic changes only in Sub-Saharan Africa. This simulated
 982 future decline in Africa is in agreement with observations of currently already
 983 declining burned area that has been linked to demographic trends of increasing rural
 984 population for the northern part of Sub-Saharan Africa (Andela and van der Werf,
 985 2014). In the present study, southern China is identified as a new area of possible high
 986 human exposure to wildfire-generated air pollution under a scenario of rapid
 987 urbanisation and population decline in rural areas. While forest fires in China may
 988 have received comparatively little attention, they can still be substantial, with over
 989 670,000 ha area burnt annually between 1950 and 2010 (Shu et al. 2003, Su et al.
 990 2015).

991 While any additional emission source of PM_{2.5} poses a health risk (WHO 2006),
 992 wildfires are likely to be ignored by air quality policy if they emit considerably less
 993 than anthropogenic sources, in particular as their occurrence tends to be sporadic and
 994 of short-term nature. One factor is that wildfire emissions are much more difficult to
 995 legislate given the difficulties of long-term fire suppression (Donovan and Brown
 996 2007). In this study, we find that in large parts of the world, wildfires are the main air
 997 pollutant source. While on an annual-mean basis, they are unlikely to lift pollution

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1061 levels above the WHO air-quality guidance levels, they can become important sources
 1062 on a seasonal basis.

1063 In our simulations, air pollutant concentrations follow similar, but more dispersed
 1064 patterns compared to emissions, with highest levels in densely populated and lowest
 1065 in sparsely populated areas. In the future, in the case of strong reduction in
 1066 anthropogenic sources, this pattern is predicted to shift to one where the highest
 1067 pollution levels are found in regions of intermediate population density for most
 1068 regions (Sub-Saharan Africa, Latin America & Caribbean, and Eastern Europe-
 1069 Russia-Central Asia), resembling more the pattern of wildfire emissions. This means
 1070 that also due to their geographical distribution, wildfires pose a smaller risk to humans
 1071 than anthropogenic emissions. Nevertheless, in our simulations under strong emission
 1072 reduction from anthropogenic sources, the future trajectory of wildfire emissions has a
 1073 discernible impact on air pollution in certain regions (Sub-Saharan Africa, South &
 1074 Southeast Asia, and Developing East Asia), and is particularly relevant if we consider
 1075 seasonal maxima in pollution levels. Even though the WHO recommendations are
 1076 based on annual mean concentrations of PM_{2.5}, the WHO also states that health effects
 1077 persist below these values. Fire emissions typically occur on the timescale of a week
 1078 up to one or several months – the length of the fire season – and the WHO annual
 1079 guidelines do not account for such monthly timescales. We hypothesise that, if a
 1080 guidance level for monthly means were to be established, it would lie somewhere
 1081 between 10 (annual) and 25 µg / m³ (daily WHO air-quality guidance). These levels
 1082 of PM_{2.5} concentration are indeed frequently exceeded in our simulations even with
 1083 the extremely low emission levels taken here for the MFR scenario. It also needs to be
 1084 taken into account that the PM_{2.5} levels that are reported and compared to WHO or
 1085 other air-quality guidelines usually include some water content, as opposed to the

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1103 simulated concentrations of dry aerosol mass, leading to a possible low bias of our
 1104 simulations of the order of 30 to 50% (Tsyro 2005). Also some emissions and
 1105 components are not considered in our simulation (e.g. the ash component of fire
 1106 emissions), or have a very high uncertainty (e.g. secondary aerosol formation
 1107 associated with burning emissions, but also in general with vegetation emissions).
 1108 Significant health impacts are therefore very likely, even if this limit is exceeded
 1109 only on a seasonal basis. For certain regions, it will be of critical importance whether
 1110 future air quality policy objectives will converge to the current WHO guidelines, in
 1111 which case fire management will become increasingly important. If anthropogenic
 1112 emissions are aggressively curtailed (MFR scenario), wildfires in Sub-Saharan Africa
 1113 are predicted to decline less than anthropogenic sources, and in parts of Southeast
 1114 Asia, southern China and central South America climate change may even lead to new
 1115 areas with wildfire emission levels relevant for air quality and the associated health
 1116 impacts. In many boreal regions wildfires will also increase to levels where they
 1117 become pollution sources with relevant health impacts. Because past efforts aimed at
 1118 a lasting reduction in wildfire activity have largely failed (Donovan et al. 2007, Doerr
 1119 and Santin 2016), it is questionable whether it is even possible to devise policy
 1120 measure aimed at bringing down wildfire emissions to avoid adverse health effects.
 1121 LPJ-GUESS-SIMFIRE only simulates wildfires. The predictions presented in this
 1122 study therefore leave out the possibility of significant increases or decreases in
 1123 deforestation or peat fire sources. Peat fires can be associated with considerable
 1124 emissions (Page et al. 2002, Kajii et al. 2002), and forest conversion into cropland or
 1125 pasture is often accompanied by burning (van der Werf et al. 2010). Both are of minor
 1126 importance for air pollution except for Southeast Asia (see Fig. S1, Table S2), mainly
 1127 in Indonesia (Field et al. 2009), where they are the dominant pollution source and

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1143 occur even in more densely populated areas. In other regions, including Russia, peat
 1144 fires are of minor importance. Whether or not future land-use change will lead to an
 1145 increase or a decrease in deforestation is unknown. Based on integrated-assessment
 1146 model realisations of the four RCPs, Hurtt et al. (2011) projected little increase or
 1147 even a decline in future crop and pasture areas. However, in studies that assessed
 1148 future land-use change from a broader perspective, a much larger range of crop and
 1149 pasture changes emerged (Eitelberg et al. 2015, Prestele et al. 2016), which makes the
 1150 relative change of deforestation vs. wildfires highly uncertain. In the present analysis,
 1151 declining wildfire emissions are only predicted for Sub-Saharan Africa, where it
 1152 appears to be related to conversion of savanna to cropland (Andela and van der Werf
 1153 2014).
 1154 Apart from the issue of dry vs. wet aerosols and the omission of some difficult to
 1155 model components, there are additional limitations of this study. We expect that
 1156 results will be affected by the presence of natural aerosols, such as mineral dust and
 1157 sea salt, which – depending on location and time of year – could be significant
 1158 fractions of the PM_{2.5} concentrations (Monks et al. 2009). We do not evaluate
 1159 changes in natural emissions of mineral dust, sea salt or other naturally occurring
 1160 emissions other than biomass burning. Also, the anthropogenic emission scenarios do
 1161 not consider the benefits of climate mitigation scenarios, and are based on rather
 1162 crude assumptions regarding the development of emission controls beyond 2050, as
 1163 the original scenarios do not consider developments beyond that year. The scenarios
 1164 are thus study are only a subset of a large range identified in recent studies
 1165 (Braspenning-Radu et al. 2016, Rao et al. 2016) and therefore call for a further
 1166 evaluation of consistency in drivers across anthropogenic emissions, climate, and fire
 1167 scenarios. The study furthermore only considers climatological monthly emissions

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1182 during specified time windows, even though wildfire impacts on air quality can have
 1183 large interannual (Jaffe et al. 2008) and intra-seasonal variations. ~~We also did not~~
 1184 account for relevant secondary emission products, such as ozone from wildfires,
 1185 which can reach policy-relevant levels (Jaffe and Wigder 2012). ~~Other~~ studies on the
 1186 possible impact of climate change on wildfire-related air pollution hazards have
 1187 concentrated on changes in meteorological conditions (Jacob and Winner 2009, Tai et
 1188 al. 2010) instead of emissions. There is also recent progress in the incorporation of
 1189 injection height (Sofiev et al. 2012) into chemistry-enabled atmospheric general
 1190 circulation models (Veira et al. 2015), which is not considered.

1191 ~~Nevertheless, to our knowledge,~~ this is the first global-scale **air quality** study to
 1192 consider changes in both climate and demographic drivers of air pollutant emissions
 1193 from wildfires. Future work should aim at using general circulation models with
 1194 realistic plume heights for a series of dedicated present and future time slices at
 1195 combining observed plume height information, fire radiative energy data (for their
 1196 finer temporal resolution), satellite-derived burned area (for better spatial coverage),
 1197 projected emission changes from coupled dynamic vegetation-fire models (as the
 1198 present study), and improved demographic scenarios accounting for changes in urban
 1199 population density. Such studies would then also simulate the temporal statistics of
 1200 pollution events on a daily time scale. Wildfire episodes can elevate **PM_{2.5}** pollution
 1201 levels to dangerous levels with serious health impacts (Haikerwal et al. 2015). Such
 1202 results could then be used, for example, to assess for how many days the WHO 24-
 1203 hour **PM_{2.5}** limit of 25 µg/m³ ~~from the WHO air quality guidelines~~ (WHO 2006) is
 1204 exceeded as a result of wildfire emissions.

1205 5 Summary and conclusions

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- 1216 • Globally, wildfire emissions are unlikely to thwart aggressive measures to reduce
 1217 anthropogenic pollutant emissions enough to stay under the ambitious 10 $\mu\text{g} / \text{m}^3$
 1218 annual mean limit of the WHO.
- 1219 • In a number of regions, wildfire emissions will remain or could rise above critical
 1220 thresholds relevant for health policy, in particular when pollution levels during the
 1221 fire season are considered. So far, there is no generally accepted method for
 1222 wildfire management that has been shown to lead to lasting reductions in fire
 1223 activity or emissions.
- 1224 • Demographic changes appear to be the main driver for the expected changes in
 1225 wildfire emissions in Sub-Saharan Africa. For a scenario of high population
 1226 growth and slow urbanisation, anthropogenic sources could surpass air pollutant
 1227 emissions from wildfires in most populated areas. Exposure of humans to $\text{PM}_{2.5}$ in
 1228 Sub-Saharan Africa is expected to drop if measures are put in place to reduce,
 1229 anthropogenic emissions, but wildfires may remain as a health relevant pollution
 1230 source in a scenario of fast urbanisation and low population growth.

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1238 **Author contributions:** WK conceived of the study, carried out the analysis and wrote
 1239 the first draft of the manuscript. JFL performed the CAM-Chem simulations. All
 1240 authors contributed to discussions and writing.

Wolfgang Knorr 3/5/2017 09:16

Deleted: <#>Globally, the percentage of world population exposed to dangerously high $\text{PM}_{2.5}$ emissions from wildfires is expected to decrease in all scenarios considered. The future of anthropogenic emissions has the largest impact, and for a scenario of current legislation, the predicted decrease is very small. -

Wolfgang Knorr 3/5/2017 09:16

Moved down [3]: <#>Demographic changes appear to be the main driver for the expected changes in wildfire emissions in Sub-Saharan Africa. For a scenario of high population growth and slow urbanisation, there will be large decreases in emissions in many parts of the continent. Exposure of humans to $\text{PM}_{2.5}$ in Sub-Saharan Africa is expected to drop if measures are put in place to reduced anthropogenic emissions, but wildfires may remain a health relevant pollution sources in a scenario of fast urbanisation and slow population growth. -

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Moved (insertion) [3]

Wolfgang Knorr 3/5/2017 09:52

Deleted: there will be large decreases in emissions in many parts of the continent

Wolfgang Knorr 3/5/2017 12:11

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Wolfgang Knorr 25/5/2017 22:05

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Wolfgang Knorr 15/5/2017 16:18

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Wolfgang Knorr 27/4/2017 12:40

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