Responses to the Comments of the Anonymous Referee #1

We very much appreciate the constructive comments and suggestions from this reviewer. Our point-by-point responses to the reviewer's comments are as follows (the reviewer's comments are marked in Italic font).

General comments:

The manuscript addresses an emerging issue for Southeast Asia which concerns the impact of biomass burning on air quality and visibility. The topic is highly relevant for publication in Atmospheric Chemistry and Physics, however major issues related to the form in which the work is structured and presented (i.e. a whole rewriting of the paper is needed), clarifications in methods and analyses need to be addressed. The overall work needs to be synthesized both in the text and in the selection of the figures presented (of the 13 figures included some of them duplicate information included in other ones. If the authors want to keep all of them, they should consider moving some of the figures to the Supplementary Materials).

Based on the reviewer's suggestion, the structure of the manuscript has been rearranged, especially in Section 2 and 3. In addition, Section 4 has been rewritten. Please note that, based on the other reviewer's suggestion, all analyses of model results and observations are now applied to the time period from 2003 to 2014.

Specific comments:

Language

A major rewriting of the paper is needed. Several sentences are not fluent and a grammar/punctuation check is needed. Below are some examples:

Line 32: remove "that"

Done.

Line 33: favorite should be "favourable"

Modified to favorable.

Line 41 and other parts: please be consistent with the tense you use.

We have checked the tense throughout the manuscript.

Line 55: "put in effect", replace with "implemented"

Done.

Line 82: please check your references (e.g. Miriam is the first name)

Corrected.

Line 118: "the great Southeast Asia" should be replaced with something similar to "over the whole Southeast Asia". Please check also elsewhere in the paper.

Modified to "the whole Southeast Asia" throughout the manuscript.

Line 135: please rephrase

The sentence is revised to "Our focus in this study is on the fire aerosol life cycle. Therefore, we chose to use WRF-Chem with a modified chemical tracer module instead of a full chemistry package, to thus model the fire $PM_{2.5}$ particles as tracers without involving much more complicated gaseous and aqueous chemical processing calculations but dry and wet depositions." in Lines 118-122 of the revised version.

Line 168: "estimations" should be always replaced with "estimates"

Modified throughout the manuscript.

Line 172: remove "with"

Done.

Line 178: "comparing" should be "compared". Please amend this everywhere in the paper.

Modified throughout the manuscript.

Line 190-202: please rephrase and summarize. This paragraph is too repetitive and needs to be more concise.

The paragraph has been rephrased to "Generally speaking, there is a strong correlation between the seasonal variation of fire emissions and that of rainfall in all fire regions as shown in Fig. 2. Because mainland Southeast Asia (s1) and northern Australia (s5) are on the edge of the seasonal migration of the ITCZ, the correlation in these two regions is even more pronounced. On the other hand, in Sumatra (s2), Borneo (s3) and the rest of Maritime Continent (s4), while inter-seasonal variations of rainfall and fire emissions are still correlated with each other in general, fire emissions do exist in some raining seasons (Fig. 2b - d), owing to the precipitation features in multiple scales over these regions (e.g., the passage of MJO events) and underground peatland burning." in Lines 172-180 in the revised version.

Line 211: units, please replace also elsewhere Line 236: "this" is missing

Done.

Line 294: "so that" is very often used incorrectly. Please check all the occurrences. Line 343: "are occurred", should be "occurred" Line 515: "reasons" should be "seasons".

Please check also other typos.

Removed "so that" in the sentence and rephrased. Done correcting typos.

Line 518-519: Please rephrase

The sentence is removed. Section 4 has been rewritten in the revised version.

Line 571-580: this section needs to be rewritten. Sentences are too long and convoluted and several grammar errors are present.

We have rewritten Section 4.

Methods:

All the introduction regarding WRF is not needed since you are using a modified version of WRF-Chem. Also you start introducing the model and have section 2.2 describing the emissions and section 2.4 discussing again the simulations. The whole method section has to be reorganized (e.g. have one section discussing the data, one on the model and one on the methods used). Please be more concise and avoid repeating the same information in different sections.

The introduction of WRF-Chem in Section 2.1 has been condensed. We have also rearranged the structure of Section 2. Besides section 2.1, the descriptions of numerical simulations and model evaluation has been moved to Section 2.2, observation data and model derivation of visibility to Section 2.3, and the "Haze Exposure Day (HED)" definition to Section 2.4.

Line 123: please refer more precisely to your "targeted science questions"

The sentence has been revised to "In this study, we have used the Weather Research and Forecasting (WRF) model coupled with a chemistry component (WRF-Chem) version 3.6 (Grell et al., 2005). Our focus in this study is on the fire aerosol life cycle. Therefore, we chose to use WRF-Chem with a modified chemical tracer module instead of a full chemistry package, to thus model the fire $PM_{2.5}$ particles as tracers without involving much more complicated gaseous and aqueous chemical processing calculations but dry and wet depositions" in Lines 117-122 of the revised version.

Line 139: you mostly focus on visibility so please also add that.

The sentence has been revised to "This configuration lowers the computational burden substantially, and thus allows us to conduct long model integrations to determine the contributions of fire aerosol to the degradation of visibility in the region over the past decade." in Lines 123-126 of the revised version.

Line 145: this is redundant information, please remove it.

Removed.

Line 146: The reported time step is for chemistry or physics?

We have made this clearly by stating: "The time step is 180 seconds for advection and physics calculation." in Line132 of the revised version.

Line 165: Did you only include fire emissions? Does WRF-Chem use other anthropogenic emissions?

We only included fire $PM_{2.5}$ particles in the model; therefore, emissions of other chemical species were excluded in the simulations. To make this clearer to the reader, we have added in the manuscript that: "Therefore, we chose to use WRF-Chem with a modified chemical tracer module instead of a full chemistry package, to thus model the fire $PM_{2.5}$ particles as tracers without involving much more complicated gaseous and aqueous chemical processing calculations but dry and wet depositions." in Lines 119-122 of the revised version.

Line 208: this should be rephrased by saying what you used for computing visibility.

The sentence has been rephrased to "In this study, the visibility is calculated by using the *Koschmeider equation:* ..." in Line 238 of the revised version.

Line 213-216: please add a reference and rephrase

The sentence has been modified to "Based on Eq. (1), a maximum visibility under an absolutely dry and pollution-free air is about 296 km owing to Rayleigh scattering, while a visibility in the order of 10 km is considered under a moderate to heavy air pollution by particulate matter (Visscher, 2013)." in Lines 242-245 of the revised version.

Reference: Visscher, A. D.: Air Dispersion Modeling: Foundations and Applications, First ed., John Wiley & Sons, Inc., pp. 50, 2013.

Line 222: please be more specific by explaining how you will use the GSOD data and to address which objectives

We have added the explanation and also rephrased the sentence to "The observational data of visibility from the Global Surface Summary of the Day (GSOD) (Smith et al., 2011) are used in our study to identify days under particulate pollution, i.e., haze events." in Lines 250-252 of the revised version.

Line 219: add "by increasing bext"

The sentence has been revised to "Similarly, fire aerosols, alone or mixed with other particulate pollutants, can degrade visibility by increasing b_{ext} and lead to occurrence of haze events too." in Lines 247-249 of the revised version.

Line 225: Here you introduce model simulations, but you have a section later discussing

that. You should reorganize the methods and be more clear on the objectives you are addressing. "In order to compare with observations", what do you mean? Are you referring to a model evaluation? If so please explain in the relevant section how you will perform it.

This paragraph describes the procedure of using observed visibility to evaluate modeled $PM_{2.5}$ concentrations in our study, and also the method of deriving modeled visibility based on the extinction coefficient of simulated fire aerosols as a function of particle size. We have modified the sentence to: "The observed visibility is also used to evaluate the modeled visibility and thus $PM_{2.5}$ concentration. The modeled visibility is derived based on the extinction coefficient of the fire aerosols as a function of particle size, by assuming a log-normal size distribution of accumulation mode with a standard deviation $\sigma = 2$ (Kim et al., 2008). Note that all these calculations are done for the wavelength of 550 nm unless otherwise indicated." in Lines 255-259. We have also added the details of particle hydroscopic growth calculation in Lines 264-270 of the revised version.

Line 227: is there a reference you can quote for these assumptions? Or some local measurements used to estimate those parameters?

We have cited Kim et al. (2008) and added this reference in the revised manuscript.

Reference:

Kim, D., Wang, C., Ekman, A. M. L., Barth, M. C., and Rasch, P. J.: Distribution and direct radiative forcing of carbonaceous and sulfate aerosols in an interactive size-resolving aerosol–climate model, Journal of Geophysical Research: Atmospheres, 113, D16309, 10.1029/2007jd009756, 2008.

Line 225-233: this paragraph should be clarified. It is not clear how you link the discussion on fire emission composition, hygroscopic growth, etc. with your work. If it is for general overview purposes, please add it to the introduction or remove it.

We have added more details of the visibility calculation, specifically the method to include the effect of particle hydroscopic growth in Section 2.4 of the revised version:

"To make the calculated visibility of the fire aerosols better match the reality, we have also considered hydroscopic growth of sulfate fraction of these mixed particles in the calculation based on the modeled relative humidity (*RH*). Based on Kiehl et al. (2000), the hydroscopic growth factor (*rhf*) is given by

$$rhf = 1.0 + exp \left(a_1 + \frac{a_2}{RH + a_3} + \frac{a_4}{RH + a_5}\right),$$
(2)

where a_1 to a5 are fitting coefficients given by 0.5532, -0.1034, -1.05, -1.957, 0.3406, respectively. The radius increase of wet particle (r_{wet}) due to hydroscopic growth will be

$$r_{wet} = r_{dry}^{rhf}, (3)$$

where r_{dry} is the radius of dry particle in micron."

Line 238-239: again this is repetition of definitions already given. Please remove this from here and elsewhere in the manuscript.

Removed.

Line 268: what is the NCAR_FNL? You have not introduced that before. Please add a reference for all datasets used.

We thank the reviewer for pointing out this typo. We have corrected "NCAR_FNL" to "NCEP_FNL".

Line 267-272: this paragraph needs to be rewritten. Is there any difference between precipitation simulated with NCAR_FNL and FNL_FINN? Otherwise synthesise this result by comparing the simulations run with FNL and ERA. What does it mean "both results appear to be higher"? Please rephrase.

We use TRMM observed precipitation to evaluate modeled rainfall in FNL_FINN and ERA-FINN. We have rewritten this paragraph. We have also added more discussions of the spatial and temporal correlations of monthly rainfall between model and observation in different seasons in Section 2.2 of the revised version.

Line 301: LVDs and VLVDs have already been defined so avoid repetitions.

Removed.

Line 332: how can you distinguish the events caused by fires? Is it because your simulations do not include other anthropogenic emissions? Otherwise please explain how you conducted your analyses.

We have revised the related descriptions. Firstly, we have emphasized that many LVDs could be induced by non-fire aerosols, therefore, modeled underestimate of $PM_{2.5}$ concentration and visibility degradation is expected. On the other hand, we used the VLVDs to specifically check the model performance because these events are known to be mainly induced by fire aerosols.

In Section 2.3 of the revised version, a largely revised paragraph now reads as: "As mentioned above, a visibility of 10 km is considered an indicator for a moderate to heavy particulate pollution. Hence a visibility of 10km in observation is used as the threshold for defining the "low visibility day (VLD)" in our study. We firstly derived the observed low visibility days in every year for a given city using the GSOD visibility data. Then, we derived the modeled low visibility days following the same procedure but using modeled visibility data that were only influenced by fire aerosols. Both the observed and modeled visibility day to occur, such a day would be attributed to fire aerosol caused LVD, regardless of whether other coexisting pollutants would have a sufficient intensity to cause low visibility or not. In addition to the LVD, we have also used a daily visibility of 7 km as the criterion to define the observed "very low visibility day (VLVD)". Such heavy haze events in the region are generally caused by severe fire aerosol pollution, thus we use their occurrence specifically to evaluate the model performance."

Line 349-362: please rephrase to remove repetitions.

We have modified the paragraph to: "The percentage of LVDs in Singapore has been rapidly increasing since 2012 (Fig. 6c). During the simulation period, this increase appears to be mostly from anthropogenic pollution other than fires, especially in 2012 and 2013. In monthly variation, similar to Kuala Lumpur, two peaks of fire aerosol influence appear in February-March and in September-October, respectively (Fig. 6g). In February and March, the trans-boundary transport of fire aerosols come from mainland Southeast Asia (s1), while in the summer monsoon season fire aerosols come from both Sumatra (s2) and Borneo (s3) (Fig. 7c). Except for the severe haze events in June 2013, VLVDs basically occur in September and October (i.e., 92%) due to both Sumatra and Borneo fires. In general, 34% of LVDs in Singapore are caused by fire aerosols in the FNL_FINN simulation and the rest by local and long-range transported pollutants (Table 3). Nevertheless, fire aerosol is still the major reason for the episodic severe haze conditions." in Lines 375-386 of the revised version.

Results

Line 374-384: this part should be moved to the methods. You need to define earlier how you will conduct your analyses. Also using LVD in equation 3 might be more appropriate than C(i).

We have moved this part to Section 2.4, the "Haze Exposure Day (HED)". We prefer to keep C(i) instead of LVD because LVD is defined as a day with visibility equal or lower than 10 km. However, C(i) represents the annual LVDs which means the sum of LVDs for each year.

Line 432: here it would be also interesting to compare with the WHO limits (i.e. the limit for annual mean $PM_{2.5}$ is 10 µg m⁻³).

The sentence has been modified to "In the FNL_FINN simulation, the seasonal mean concentration of $PM_{2.5}$ within the planetary boundary layer (PBL) can exceed 20 µg m⁻³ in this region (note that the air quality standard suggested by World Health Origination is 10 µg m⁻³ for annual mean and 25 µg m⁻³ for 24-h mean)." in Lines 430-433 of the revised version.

Line 590: Section 4 should be rewritten. The way results are presented is too repetitive and convoluted. It would be also easier for the reader to have some clear sentences summarizing the skills of different models/emissions.

Section 4 has been rewritten. The revisions are well marked in the version showing tracking results.

Figures

Thirteen figures are really too many especially since most of them have several panels. Please select the most critical ones to summarize your findings and move the others to the supplementary material. Also some figures duplicate content shown in other, so either delete them or move to the supplements. The point has been well taken. We have moved Fig. 3, 10 and 13 in the original version to the supplementary and have removed Fig. 2 and 11.

Figure 1: the number of vertical levels cannot be inferred from the figure, so please remove this part of the sentence from the caption. Also, the letters A-D are not easily readable. Please choose different colors.

We have changed the caption to "Figure 1. Model domain used for simulations. The domain has 432×148 grid points with a horizontal resolution of 36 km. Five fire source regions marked in different colors and labeled as s1, s2, s3, s4 and s5, represent mainland Southeast Asia (s1), Sumatra and Java islands (s2), Borneo (s3), the rest of Maritime Continent (s4), and northern Australia (s5). A, B, C and D indicate the location of four selected cities: Bangkok (A), Kuala Lumpur (B), Singapore (C) and Kuching (D)."

We have enlarged the font size of the letters of A-D.

Figure 2: PM2.5 on the y-axis is not as subscript 2.5. It would be easier for the reader to have the whole name of the regions on top of each panel.

Figure 2 has been removed.

Figure 3: is this the yearly average of the daily means? The units can be put after "precipitation".

The figure shows daily precipitation in 2006 only. We have added the units after "precipitation" as the reviewer suggested. This figure has been moved to the supplementary as Fig. S1.

Figure 5: From panel (a) it is clear that the model highly underestimates observations and a scaling factor is needed. This has to be commented in the text. Could you also start both the y- axes from 0? A scatter plot might also help in quantifying the underestimation or please provide some more statistics for model evaluation.

We have changed Fig. 5 (a) and (b) (the new Fig. 3 (a) and (b)) to let the y-axes start from 0. We have accepted the reviewer's suggestion to add a new scatter plot, Fig. 4, in the revised version to show observed visibility versus modeled visibility in FNL_FINN during known fire events. We have also added discussion of this new figure as:

"The surface observational data of $PM_{2.5}$ concentration among these four cities are only available in Singapore since 2013 from the National Environment Agency (NEA) of Singapore. We thus firstly used these data along with visibility data to evaluate model's performance for fire-cause haze events reported in Singapore during 2013-2014 (Fig. 3). Note that the observed $PM_{2.5}$ level reflects the influences of both fire and non-fire aerosols, whereas the modeled $PM_{2.5}$ only includes the impact of fire aerosols. We find that the model still predicted clearly high $PM_{2.5}$ concentrations during most of the observed haze events, especially in June 2013, and in spring and fall seasons of 2014 (highlighted green areas), though with underestimates in particle concentration of up to 30-50%, likely due to the model's exclusion of non-fire aerosols, coarse model resolution, overestimated rainfall, or errors in the emission inventory. Figure 4 shows observed visibility versus modeled visibility in FNL_FINN during the fire events shown in Fig. 3. Note that all these events have an observed visibility lower than or equal to 10 km, or can be identified as LVDs. In capturing these fire-caused haze events, the model only missed about 22% of them, or reporting a visibility larger than 10 km in 40 out of 185 observed LVDs as marked with different color in Fig. 4. When observed visibility is between 7 and 10 km, model results appear to align with observations rather well. For cases with visibility lower than 7 km, the model captured all the events (by reporting a visibility lower than 10 km, or LVD) although often overestimated the visibility range. These results imply that the VLVDs only count a very small fraction in LVDs and thus are episodic events. It is very likely that the size of concentrated fire plumes in VLVDs might be constantly smaller than the 36 km model resolution, therefore, the model results could not reach the peak values of PM_{2.5} concentrations of these plumes".

Figure 6. What do you mean with "variation"? How did you compute it? Please also report the meaning of the color coding in the caption.

The caption has been changed to "Figure 6. (a) – (d) The percentage of LVDs per year derived using from GSOD visibility observations in Bangkok, Kuala Lumpur, Singapore, and Kuching, respectively. (e) – (h) The percentage of LVDs averaged over 2003-2014, derived using GSOD visibility observations in Bangkok, Kuala Lumpur, Singapore, and Kuching, respectively. Each bar presents the observed LVDs in each year or month. Red color shows the partition of fire-caused LVDs (captured by model) while green color presents non-fire LVDs (observed – modeled)."

Figure 7: Please define "variation" or rephrase. Please do the same for all other figures presenting that wording.

The caption has been changed to "Figure 7. The mean fire $PM_{2.5}$ concentrations attributed to different emission regions (s1 - s5) in: (a) Bangkok, (b) Kuala Lumpur, (c) Singapore and (d) Kuching, are all derived from FNL_FINN simulation and averaged over the period of 2003-2014."

Figure 8: (a) please rephrase saying that the size of the circles indicates the number of days and the colors refer to specific population weights. (b) Please add units on y-axes and mention in the caption the use of different scales.

Added units and days on y-axis. The caption has been changed to "Figure 8. (a) The mean low visibility days (circles) per year from 2003 to 2014 in 50 ASEAN cities. The size of the circles indicates the number of days. The colors refer to population-weighted fraction in the total Haze Exposure Days (HED). (b) Annual population-weighted HED (HED_{pw}) and arithmetic mean HED (HED_{ar}). Fire-caused HED are labeled as fHED_{pw} and fHED_{ar}. Units are in days. Note that the y-axes are in different scales."

Figure 9: region s1-s5 are not reported on the panels, so please remove them from the caption and simplify the caption as well. Also it is not clear why you report the results

separately by region instead of on one single figure. Figure 9 is essentially identical to Figure 10 averaging on a different period, so you can have just a four panels figure with on each panel a map showing different seasons and the 5 regions together and two panels with the same for wet scavenging. Otherwise you need to move one of the two figures to the supplements.

We have removed s1-s5 in the caption and removed lines in (f)-(g). We actually have moved Fig. 10 to the supplementary.

Figure 11: this is again a repetition of Figure 7. Either you condense the information in one figure or move some of the material to the supplements. It is very hard to keep in mind so many similar figures and your key message is not delivered effectively.

The reviewer's suggestion has been well taken. We have removed the Fig. 11 in the revised version.

Figure 12: Why do you have y-axes with negative numbers? You are displaying PM concentrations and precipitation, so your minimum value should be zero. This figure again contains information already presented (Figure 11, 7, 13), so please try and condense the figures or move them to the supplements. The captions of all figures should be also more informative on the message you want to deliver to the reader.

We have changed all the y-axes scales to start from 0. We have also removed the original Fig. 2. This discussed figure (i.e., original Fig. 12) now becomes Fig. 2 in the revised version.

Reference:

Grell, G. A., Peckham, S. E., Schmitz, R., McKeen, S. A., Frost, G., Skamarock, W. C., and Eder, B.: Fully coupled "online" chemistry within the WRF model, Atmospheric Environment, 39, 10.1016/j.atmosenv.2005.04.027, 2005.

Kiehl, J. T., Schneider, T. L., Rasch, P. J., Barth, M. C., and Wong, J.: Radiative forcing due to sulfate aerosols from simulations with the National Center for Atmospheric Research Community Climate Model, Version 3, Journal of Geophysical Research: Atmospheres, 105, 1441-1457, 10.1029/1999JD900495, 2000.

Kim, D., Wang, C., Ekman, A. M. L., Barth, M. C., and Rasch, P. J.: Distribution and direct radiative forcing of carbonaceous and sulfate aerosols in an interactive size-resolving aerosol–climate model, Journal of Geophysical Research: Atmospheres, 113, D16309, 10.1029/2007jd009756, 2008.

Smith, A., Lott, N., and Vose, R.: The Integrated Surface Database: Recent Developments and Partnerships, Bulletin of the American Meteorological Society, 92, 704-708, doi:10.1175/2011BAMS3015.1, 2011.

Visscher, A. D.: Air Dispersion Modeling: Foundations and Applications, First ed., John Wiley & Sons, Inc., pp. 50, 2013.

Responses to the Comments of the Anonymous Referee #2

We very much appreciate the constructive comments and suggestions from this reviewer. The following are our point-by-point responses to the reviewer's comments (the reviewer's comments are marked in Italic font).

General comments:

The paper provides a correlation of modeled particulate matter with low visibility days recorded at observation sites across South East Asia. Information is presented about the most likely source areas for biomass burning pollution for different cities and different seasons.

This is an interesting application of an alternative observation dataset for assessing the impact of biomass burning haze on the region and for validating CTM and dispersion models. However, the significant flaw in the way the results are presented is that the model is assumed to be correct and that all low visibility days that are not modeled are therefore specified to be due to other pollution contributions. The validity of this assumption is not demonstrated. It is quite possible that the model is over-estimating the biomass contribution at some sites and underestimating it at others. Fig 6 for example would suggest that the model may not be capturing up to 50% of the fire haze days, and Fig 4 would suggest that the model misses 50% of the VLVDs at Singapore. The references in the text to fire and non-fire LVD are therefore misleading. The authors need to reconsider how they interpret this data and present it in the paper.

We are fully aware of the uncertainty of our model due to factors including emissions, model resolution, and meteorological fields. The uncertainty of modeling was repeatedly indicated in the manuscript, and the additional simulations using different emission inventories and meteorological fields were all designed and conducted for the purpose of identifying, at least partially, the influences of these uncertainty factors on modeled results. Nevertheless, the reviewer's point is well taken. We have made our best effort to reiterate the model uncertainty and evaluation in the revised manuscript. In addition, we have specifically indicated in many places that the model's overestimates in visibility range (underestimates in visibility degradation) are likely due to the fact that observed visibility reflects contributions of both fire and non-fire aerosols.

We have revised the description in Section 2.3 regarding our method to attribute low visibility events to fire aerosols (such events can be induced by either fire or non-fire aerosol alone or in combination), as: "As mentioned above, a visibility of 10 km is considered an indicator for a moderate to heavy particulate pollution. Hence a visibility of 10km in observation is used as the threshold for defining the "low visibility day (VLD)" in our study. We firstly derived the observed low visibility days in every year for a given city using the GSOD visibility data. Then, we derived the modeled low visibility days following the same procedure but using modeled visibility data that were only influenced by fire aerosols. Both the observed and modeled visibilities were then used to define the fraction of low visibility days that can be caused by fire aerosols alone. It is assumed that whenever fire aerosol *alone* could cause a low visibility day to occur,

such a day would be attributed to fire aerosol caused LVD, regardless of whether other coexisting pollutants would have a sufficient intensity to cause low visibility or not. In addition to the LVD, we have also used a daily visibility of 7 km as the criterion to define the observed "very low visibility day (VLVD)". Such heavy haze events in the region are generally caused by severe fire aerosol pollution, thus we use their occurrence specifically to evaluate the model performance". In addition, we have revised statements of fire aerosol contribution to contain "up to" whenever necessary.

Furthermore, the descriptions of model evaluation based on model-observation comparison have been revised, two new or largely revised paragraphs in the revised manuscript are added in Section 3.1, they provide the procedure and present the uncertainty of the model in a greater detail and clarity:

"The surface observational data of PM_{2.5} concentration among these four cities are only available in Singapore since 2013 from the National Environment Agency (NEA) of Singapore. We thus firstly used these data along with visibility data to evaluate model's performance for fire-cause haze events reported in Singapore during 2013-2014 (Fig. 3). Note that the observed PM2.5 level reflects the influences of both fire and non-fire aerosols, whereas the modeled PM2.5 only includes the impact of fire aerosols. We find that the model still predicted clearly high PM2.5 concentrations during most of the observed haze events, especially in June 2013, and in spring and fall seasons of 2014 (highlighted green areas), though with underestimates in particle concentration of up to 30-50%, likely due to the model's exclusion of non-fire aerosols, coarse model resolution, overestimated rainfall, and errors in the emission inventory. Figure 4 shows observed visibility versus modeled visibility in FNL FINN during the fire events shown in Fig. 3. Note that all these events have an observed visibility lower than or equal to 10 km, and are identified as LVDs. In capturing these fire-caused haze events, the model only missed about 22% of them, or reporting a visibility larger than 10 km in 40 out of 185 observed LVDs as marked with purple color in Fig. 4. When observed visibility is between 7 and 10 km, model results appear to align with observations rather well. For cases with visibility lower than 7 km, the model captured all the events (by reporting a visibility lower than 10 km, or LVD) although often overestimated the visibility range. These results imply that the VLVDs only count a very small fraction in VLDs and thus are episodic events. It is very likely that the size of concentrated fire plumes in VLVDs might be constantly smaller than the 36 km model resolution; therefore, the model results could not reach the peak values of PM_{2.5} concentrations of these plumes.

Furthermore, the LVDs in the four selected near-fire-site cities during the fire seasons from 2003 to 2014 have been identified using the daily GSOD visibility database and then compared with modeled results (Fig. 5). It is difficult to identify all the fire caused haze events beyond Singapore even in recent years. However, in Southeast Asia, severe haze events equivalent to the VLVDs in visibility degradation are known to be largely caused by fire aerosol pollution. Therefore, we used the observed VLVDs in the four selected cities to evaluate the performance of the model. We find that the modeled result displays a good performance in capturing observed VLVDs despite an overestimate in visibility range during certain events compared with the observation. The model in

general only missed about 10% or fewer VLVDs observed in the past decade (Table 3; Fig. 5). In addition, the model has reasonably captured the observed LVDs despite certain biases (Fig. 5), likely due to the fact that fire aerosol might not be the only reason responsible for the degradation of visibility during many LVDs".

The paper would benefit from some reorganization of the sections and a reduction in the number of figures.

Based on the reviewer's suggestion, we have reorganized the manuscript. Specifically, Section 2 and 3. Section 4 has been rewritten.

Specific Comments:

Following on from the general comments, I am concerned that no real attempt at model validation is made within this paper. An additional source of observed data, e.g. PM10 concentrations, from a minimum of one of the sites (ideally many more) is needed to demonstrate that the WRF-Chem simulations are correctly capturing the fire component. The data shown in Fig 5(a) is misleading due to the use of different scales and a more robust analysis of this data is needed earlier in the paper. In fact this data may reveal useful information about missing "background" PM from the model. There are statements on line 320 that the model is underestimating PM2.5 concentration by up to 30-50% in this comparison. This is a significant underestimation. What impact does this then have on the visibility and hence the LVD calculations? The authors also need to discuss in more detail the impacts of uncertainty on the LVD and VLVD estimates. Without this level of validation, the model results cannot be used to the level of precision that the authors present in e.g. Table 2.

We appreciate the reviewer's comments. However, as perhaps the reviewer is well aware, observational data of aerosols in Southeast Asia are still quite limited. This is also a reason why we used surface visibility data (a proxy data of $PM_{2.5}$) in the study. Besides $PM_{2.5}$ data in Singapore, there are some PM_{10} monitoring data in Thailand and Malaysia. However, these are not the best data for visibility calculation due to a lack of knowledge of size distribution, not mentioning the sparseness of these data.

As reported in the paper, our model evaluation contains two parts: one is on modeled meteorological features and the other is on fire $PM_{2.5}$. Accepting the reviewer's suggestion, the detail discussion of meteorology evaluation including precipitation and wind field is now presented in Section 2.2 of the revised version.

Regarding the underestimate of $PM_{2.5}$ concentration by up to 30-50% compared to observation as shown in Fig. 5 (a) (new Fig. 3(a)), our response to the reviewer's general comments along with the newly added paragraphs in 3.1 should also address this specific comment. After all, observed $PM_{2.5}$ concentrations still reflect the contributions from other besides fire aerosols. We have added statements to indicate this fact in the revised manuscript.

We have also adjusted the scales of Fig. 5 (now Fig. 3).

I would also like to see some explanation as to why the modeled visibility distance for Bangkok in Fig 4 is significantly lower than that in the observations (and in comparison to the difference at other sites), and consequently what this means for the calculation of VLVDs.

Thanks for asking this interesting question. The reason why the modeled visibility in Bangkok is lower than observation in certain time period can be explained by Fig. 2 in the revised version and Fig. S5a in the supplementary section. We find that fire $PM_{2.5}$ emissions in FINNv1.5 are about a factor of 2 or 3 higher than those in GFEDv4.1s in mainland Southeast Asia (s1) during fire seasons. Note that such a difference between the two emission inventories does not show in other fire sites, i.e., $s_2 - s_4$. This implies that FINNv1.5 likely overestimated the fire emissions in mainland Southeast Asia and thus this leads to a modeled visibility in our FNL FINN lower than observation in Bangkok. We have added the discussion in Section 4 of the revised manuscript as: "Compared to FINNv1.5, fire emissions in GFEDv4.1s over mainland Southeast Asia are more than 66% lower (Fig. 2a), and this results in a 43% lower fire PM_{2.5} concentration in Bangkok The lower fire PM_{2.5} concentration in FNL GFED actually produces a (Table 4). visibility that matches better with observations in Bangkok comparing to the result of FNL FINN (Fig. S5a). This implies that the fire emissions in FINNv1.5 are perhaps overestimated in mainland Southeast Asia".

The decision that the "other pollution contribution %" is "100% minus Fire pollution contribution %" is not appropriate for the analysis that is then presented. Statements such as those on line 336-338 and line 345-347 do not hold up. The authors need to present a justification for why the reader should assume that the model data is correct. Even so, all interpretation of non-fire LVD should probably be removed.

Our analysis only implies that "by considering fire aerosol alone" how many LVDs can be attributed to fire particulate pollution. We actually emphasized this point in many places of the original manuscript. The reviewer's point is well taken. To further avoid the misunderstanding, we have made it even more clearly in the revised manuscript by: (1) laying out more details about our judgment making, (2) clarifying that other cases are those that cannot be explained by fire aerosol alone, and (3) adding "up to" in the statements when necessary when referring to fire aerosol contribution. In addition, we have made our best effort to indicate that all these implications do not need to assume a perfect model to achieve.

To aid the discussion of the changing number of LVDs further explanation of certain statements is needed. For example, Line 366-368, why is Kuching different to Singapore? Could this be because Kuching is within a fire area?

We appreciate the reviewer's suggestion. We have stated "Kuching is in the coastal area of Borneo so Kuching is directly affected by Borneo fire events (s3)", and also "Because of its geographic location, Kuching is affected heavily by local fire events during the fire season (Fig. 7d). Fire aerosols can often degrade the visibility to below 7 km and can even reach 2 km (Fig. 3d)" in the revised version.

More information and explanation on the model set-up and analysis approach are needed to help the reader understand what has been done. Including (a) in section 2, further explanation about the "chemistry tracer module" is required – is there any chemistry at all? It doesn't appear so, so this is a bit misleading. It would be better to sav "chemical tracer module" and be clear that the pollutants are being modeled as tracers only. The lines on p8 (163-164) describing the deposition processes could usefully be moved to this earlier point in the text. An explanation for why the domain extends so far west would also be helpful. (b) p9 line 180 – the authors need to clarify whether emissions have been injected at just 700 m or from the surface to 700 m. Is this asl or agl? (c) More detail (ideally the equations used) is needed as to how the hydroscopic growth is calculated on p11 line 232 and how this relates to the visibility calculation. Also where has the environmental relative humidity data that is used come from? This is fundamental part of the model data processing, and will introduced it's own uncertainties, but is rushed over (d) There is currently no information on how the model output has been produced for each site, so this needs to be added. For example, is it based on the modeled concentration in the lowest WRF-Chem layer for the grid box corresponding to each observation site? (e) A brief explanation as to how the runs have been conducted to identify the different source sectors is needed. Did these use labeled tracers?

(a) The sentence has been changed to "to thus model the fire $PM_{2.5}$ particles as tracers without involving much more complicated gaseous and aqueous chemical processing calculations but dry and wet depositions." We have also moved the description of deposition calculation to this place in Lines 120-122 of the revised version.

(b) We have changed the sentence to: "Therefore, we have limited the plume injection height of peat fire by a ceiling of 700 m above the ground in this study based on Tosca et al. (2011). The vertical distribution of emitted aerosols is calculated using the plume model." in Lines 160-162 of the revised version.

(c) We have added the calculation of hydroscopic growth factor and the radius increase adjustment after hydroscopic growth in Eq. (2) and (3) in the revised version. The data of relative humidity for the hydroscopic growth calculation are from the model results.

"We also consider hydroscopic growth of sulfate fraction of these mixed particles in the calculation based on the modeled relative humidity (*RH*). Based on Kiehl et al. (2000), the hydroscopic growth factor (*rhf*) is given by

$$rhf = 1.0 + exp \left(a_1 + \frac{a_2}{RH + a_3} + \frac{a_4}{RH + a_5}\right),$$
(2)

where a_1 to a5 are fitting coefficients given by 0.5532, -0.1034, -1.05, -1.957, 0.3406, respectively. The radius increase of wet particle (r_{wet}) due to hydroscopic growth will be $r_{wet} = r_{dry}^{rhf}$, (3) where r_{dry} is the radius of dry particle in micron." has been added in Section 2.4 in the revised version.

(d) The fire $PM_{2.5}$ concentration presented in the paper is averaged within the PBL for the grid box corresponding to each observation site. This information has been added in the caption of Fig. 7 and 9.

(e) Yes, we labeled tracers from each source region when we created fire emission in WRF-Chem inputs. This is actually described in the emissions section, Section 2.1.

The use of two different time periods for the analysis of the results for the FINN data vs. the GFED data introduces differences in the outputs, which could be misinterpreted. It makes Table 3 particularly complicated to interpret. I would recommend that throughout the paper the authors only present data for the same period for all 3 model simulations (i.e. 2003-2014) to avoid introducing additional uncertainty and confusion in their results and analysis.

The reviewer's suggestion is well taken. All discussion and data in the revised manuscript are now presented from 2003 to 2014.

I would also recommend that Table 3 is modified to present the total number of days in the 12 year period rather than an annual average, as the latter significantly distorts the true year to year variability and introduces false precision.

We believe the reviewer's comment applies to Table 2 not Table 3 in the original version. Actually, the percentage values used in current Tables (i.e., mean LVDs/365 x 100%) serve the same purpose to describe the haze situation in any given year as suggested by the reviewer. The standard deviation shows year to year variation.

The language needs some improvement particularly in the abstract and the introduction. The use of "particulate matters" rather than "matter" is somewhat unconventional.

We thank the reviewer's comment and we have tried our best to polish the language of the manuscript.

The discussion of the role of precipitation jumps around the sections, so the authors are encouraged to see if this could be pulled together into one, shorter overview section. Some of the text regarding the precipitation in section 2.4 needs further explanation. For example on line 275 more detail and/or a citation is needed for the FDDA grid nudging. The use of mean monthly rainfall to compare the models and observations (lines 269-274) seems strange given that the authors have nicely demonstrated the large annual variation in rainfall timing and magnitude across the region. It would be useful to explore whether the models are better in some seasons than others in this region? On Line 281 the authors mention the temporal correlation, but also need to state over what averaging period this is, e.g. is this based on daily, weekly, monthly mean or total ppt data? Figure 3 is particularly hard to interpret. Difference plots would be more useful here, but this figure is a candidate for removal.

The reviewer's point is well taken. We have added the discussion about the evaluation of simulated rainfall and wind field and moved them all to Section 2.2. We have also added Table 2 in the revised version to present the spatial and temporal correlation of monthly rainfall between model and observation in different season.

The original Fig. 3 has been moved to the supplementary.

Section 4 would benefit from a broader discussion of the NWP datasets, for example there is currently no discussion of the wind fields, which are of higher order relevance than the precipitation, particularly for the source area identification. I also find it slightly surprising that given that the LBCs are a long way from Sumatra that WRF develops such a discrepancy in precipitation over the central region of the domain in the different runs. Is there a similar difference in the winds, which would therefore impact the transport? Has any verification of the WRF wind data been conducted? This section would benefit from being merged with the other sections on meteorology.

We have added a discussion of the surface wind difference in Section 2.2 along with related figures (Fig. S2 and S3) in the supplementary. Figure S2 and S3 show the surface wind of reanalysis data of FNL and ERA in the summer and winter monsoon seasons and the difference between FNL_FINN and ERA_FINN modeled winds. In responding to the reviewer's suggestion, we have also added discussions of the mesoscale wind pattern change in Section 2.2 besides rainfall evaluation. The discussion about the impacts of different meteorology inputs on modeled $PM_{2.5}$ concentration and LVDs are presented in Section 4 of the revised manuscript.

The attempt by the authors to use the data to assess the impact of the haze on populations in SE Asia is to be commended, but the approach taken is needlessly complicated. The units of the HED metrics are unclear and the dominance of population size on the HED_{pw} metric needs more careful explanation. What the results are showing are that the total number of LVDs in the region (based on observations at 50 cities) has increased over the analysis period. This conclusion could be reached without the HED and is easier to explain and understand for the reader. As explained previously the statements in this section about non-fire pollution are not justified by the approach.

Haze Exposure Day (HED) can be defined by the population weighted or arithmetic mean over the included cities. The latter perhaps is the format suggested by the reviewer. As shown in the paper, we have provided results of both. The population weighted exposure is commonly used in health and policy analyses because it clearly indicates the impact correlated to population distribution. The meanings of both types of HED have been described along with their definition. The reviewer's point is well taken and we have made our best effort to clarify the implication of our results relating to fire aerosols.

The manuscript would benefit from fewer figures and I am not sure the supplementary material adds anything. The line thickness in many of the line graphs means that the bottom lines are often hidden, this is always a problem with this sort of graph, but a reduction in the line thickness would be beneficial.

We thank the reviewer's suggestion. We have moved the Fig. 3, 10 and 13 in the original version to the supplementary and have removed Fig. 2 and 11. All y-axes in the figures have been set to start from zero in the revised version.

Technical Corrections

P2 line 45 – 99.1% is over stating the precision here. I would suggest using only 99%

which is in line with the precision of other numbers given in the abstract.

Modified.

P4 line 66-73 – The discussion of radiative impact isn't relevant to the rest of this work, so seems unnecessary. Recommend deleting these lines.

We have shortened the discussion of radiative impact of fire aerosols in the Introduction.

Line 325-327 - it would be more helpful to the reader if these percentages were expressed as a number of days. The language at the end of this sentence could also be improved.

The sentence has been modified to "We find that the annual mean LVDs in Bangkok has increased from 47% (172 days per year) in the first 5-year period of the simulation (2003-2007) to 74% (272 days per year) in the last 5-year period (2010-2014). The LVDs caused by fire aerosols has increased as well (Fig. 6a)." in Lines 352-355 of the revised version.

Line 237 – *Is the total population figure here correct? It is not clear if this the combined total, or if each city has more than 2 million?*

There is no population figure presented in the paper. We are not sure to which figure the reviewer was referred. The population information of 50 ASEAN cities has been added in the supplementary (Table S1) in the revised version.

Table 2 – The table would benefit from explanation that the VLD and VLVD for FNL_FINN and ERA_FINN are identical as they are based on observations, and that the data for FNL_GFED is different as it covers a shorter time period. However see comments regarding making the time period consistent.

The caption of Table 3 in the revised version has been changed to "Annual mean low visibility days (LVDs; observed visibility ≤ 10 km) and very low visibility days (VLVDs; observed visibility ≤ 7 km) per year in Bangkok, Kuala Lumpur, Singapore and Kuching during 2003-2014 are presented in the second column. Parentheses show the percentage of year. The third and fourth columns show the percentage contributions along with standard deviations of fire and non-fire (other) pollutions for total low visibility days."

Table 2 - The FNL_FINN LVD line for Singapore does not add up to 100%.

In the revised version, the data have been changed to 36% and 64% based on the analysis from 2003 to 2014.

In Table 3, the caption states that "parentheses show the fire aerosol fraction in total PM2.5" – this is very unclear and confusing. It could be taken to imply that the model also contains non-fire PM2.5, but I don't think this is the case. I think the table would be more informative and cleaner if all of the parentheses data were removed.

We would like to keep the information of the percentage of fire aerosol contribution from each source region in the table. We have modified the caption to "Parentheses show the percentage of fire $PM_{2.5}$ contribution originating from each source region." to clarify the meaning in the parentheses.

Figure 2 - it would be useful to highlight in the caption that all of the plots have different axes scales.

Highlighted as suggested. Figure 2 has been removed to reduce the number of figures in the manuscript.

Figure 5 – the use of different axis scales in (a) is very misleading. Both data sets should be presented with the same scale and starting from 0. Where is the data that gives the green areas from? This data could usefully contribute to the discussion in the text and the validation of the model.

We now use the same scales starting from zero. The haze events highlighted in green are manually selected based on observed $PM_{2.5}$ concentration and visibility. A detailed discussion has been added in Section 3.1.

Figure 6 - A better way to present this data would be to have the green data as the GSOD observed LVDs and the red data as the modeled fire LVDs. This would be a more robust comparison of model vs. observations and start to address issues in the comments above.

We very much appreciate the reviewer's suggestion. However, since the observations actually contain both fire and non-fire contributions, therefore, we believe the current column charts present the results rather well. In this figure, each column presents the observed LVDs in each year or month. For example, in Fig. 6a of the revised version, column 2003 shows 40% observed LVDs (greed + red), which includes 10% fire LVDs (red) and 30% other LVDs (green).

Figure 7 – the S1 and S5 line colors are too similar in my copy, so can one of these be changed please.

Changed the s5 line color to orange.

Figure 9 – Need to specify that these are "fire" concentrations in the caption. In this and Fig 10, the purple contours on the right hand plots prevent the underlying colors from being seen and are so small that they are unreadable, so recommend that these are removed.

We have modified the caption to contain "fire $PM_{2.5}$ concentration". We have also removed the contour lines in Fig. 9 (f) – (g) and Fig. S4 (f) – (g).

Figure 11 - To ensure that there is no unintentional bias, the plot would be better if it depicted data for only 2003-2014 for all of the data sources.

We have removed this figure in the revised version.

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3	Biomass Burning Aerosols and the Low Visibility Events in Southeast Asia	
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5	Hsiang-He Lee ^{1@} , Rotem Z Bar-Or ² , and Chien Wang ^{1,2}	
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16 17	Atmospheric Chemistry and Physics	
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22 23 24 25 26 27 28 29	[@] Corresponding author address: Dr. Hsiang-He Lee, 1 <u>GreateCREATE</u> Way, #09-03 CREATE Tower, Singapore, 138602 E-mail: hsiang-he@smart.mit.edu	Formatted: Font: (Default) Times New Roman

30 Abstract

31 Fires including peatland burning in Southeast Asia have become a major concern ofto 32 the general public as well as governments in the region. This is because that aerosols 33 emitted from such fires can cause persistent haze events under favoritecertain weather 34 conditions in downwind locations, degrading visibility and causing human health issues. 35 In order to improve our understanding of the spatial-temporal coverage and influence of 36 biomass burning aerosols in Southeast Asia, we have used surface visibility and particulate 37 matter concentration observations, addedsupplemented by decadal long (20022003, to 38 2014) simulations using the Weather Research and Forecasting (WRF) model with a fire 39 aerosol module, driven by high-resolution biomass burning emission inventories. We find 40 that in the past decade, fire aerosols are responsible for nearly all the events with very low 41 visibility (< 7km), and). Fire aerosols alone are also responsible for a substantial fraction 42 of the low visibility events (visibility < 10 km) in the major metropolitan areas of Southeast 43 Asia: 38up to 39% in Bangkok, 3536% in Kuala Lumpur, and 34% in Singapore. Biomass 44 burningsburning in Mainlandmainland Southeast Asia account for the largest contributorcontribution to total fire-produced PM2.5 in Bangkok (99.1%), while biomass 45 burning in Sumatra is thea major contributor to fire-produced PM2.5 in Kuala Lumpur 46 47 (4950%) and Singapore (41%). To examine the general situation across the region, we 48 have further defined and derived a new integrated metric for 50 cities of the Association of 49 Southeast Asian Nations, (ASEAN); i.e., the Haze Exposure Days (HEDs) that measures 50 the annual exposure days of these cities to low visibility (< 10 km) caused by particulate 51 matter pollution. It is shown that HEDs have increased steadily in the past decade across cities with both high and low populations. Fire events alone are found to be responsible 52

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53	for <u>up to about half of the total HEDs</u> . Therefore, our <u>Our</u> result suggests that in order to		Formatted: Font: (Default) Times New Roman
54	improve the overall air quality in Southeast Asia, mitigation policies targeting at both		Formatted: Font: (Default) Times New Roman Formatted: Font: (Default) Times New Roman
55	biomass <u>burning</u> and fossil fuel burning sources need to be <u>put in effectimplemented</u> .	<	Formatted: Font: (Default) Times New Roman Formatted: Font: (Default) Times New Roman

56 1 Introduction

57	In recent decades, biomass burning has become frequent and widely spread across the
58	mainland of Southeast Asia to and the islands of Sumatra and Borneo (Langner et al., 2007;
59	Carlson et al., 2012; Page et al., 2002; van der Werf et al., 2010). Abundant particulate
60	mattersaerosols emitted from such fires cause the haze events to occur in the downwind
61	locations such as Singapore (Koe et al., 2001; Heil et al., 2007; See et al., 2006), degrading
62	visibility and threatening on-human health (Emmanuel, 2000; Kunii et al., 2002; Johnston
63	et al., 2012; Mauderly and Chow, 2008). Besides causing air quality issues, the fire
64	aerosols contain rich carbonaceous compounds such as black carbon (BC) (Fujii et al.,
65	2014) and thus can reduce sunlight through both absorption and scattering. Based on
66	satellite data and numerical simulations, Tosca et al. (2010) found that tropospheric
67	heating from BC absorption in the Maritime Continent (MC) is 20.5 \pm 9.3 W m ⁻² , and
68	the reduction of both surface net shortwave radiation and regional precipitation can
69	be as high as 10% due to the direct and semi-direct effects of fire aerosols.
70	Nevertheless, indirectIndirect effects of fire aerosols are even more complicated due to
71	various cloud types and meteorological conditions in the MCMaritime Continent (MC)
72	(Sekiguchi et al., 2003; Lin et al., 2013; Wu et al., 2013).
73	MajorityThe majority of present day fires in Southeast Asia occursoccur due to
74	human interferences: oil palm plantation related<u>interference such as</u> land clearing<u>, for</u>
75	oil palm plantations, other causes of deforestation, and poor peatland management, and
76	burning of agriculture wasteswaste (Dennis et al., 2005; MiriamMarlier et al.,
77	2015b2015a). Certain policies and regulations, such as those regarding, e.g., migration,
78	also affect the occurrence of burning events. For example, largeLarge fires have occurred

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79	since the 1960s in Sumatra; however, the first fire event in Kalimantan happened in the Formatted: Font: (Default) Times New Roman
80	1980s (Field et al., 2009). Based on economic incentives and population growth in
81	Southeast Asia, future land-use management will play an important role in determining the
82	coverageoccurrence of fires across the region (Carlson et al., 2012; <u>MiriamMarlier</u> et al., Formatted : Font: (Default) Times New Roman
83	2015a 2015b).
84	Besides human interventions, meteorological factors, such as rainfall, can also Formatted: Font: (Default) Times New Roman
85	influence fire initiation, intensity, and duration (Reid et al., 2012; Reid et al., 2015). Of
86	particular importance is rainfall. Reid et al. (2012) investigated relationships between fire Formatted: Font: (Default) Times New Roman
87	hotspot appearance and various climate variabilities as well as meteorological phenomena
88	in different temporal scales over the MC, including: (1) the El Nino and Southern Formatted: Font: (Default) Times New Roman
89	Oscillation (ENSO) (Rasmusson and Wallace, 1983: McBride et al., 2003) and the Indian
90	Ocean Dipole (IOD)(Saji et al., 1999); (2) Seasonal seasonal migration of the Inter Formatted: Font: (Default) Times New Roman
91	tropical Convergence Zone (ITCZ) and associated Southeast Asia monsoons (Chang et al.,
92	2005); (3) Intraintra-seasonal variabilities such as variability associated with the Madden-
93	Julian Oscillation (MJO) (Madden and Julian, 1971: <u>Zhang, 2005</u>) and the west Sumatran
94	low (Wu and Hsu, 2009); (4) Waveequatorial waves, mesoscale features, and tropical Formatted: Font: (Default) Times New Roman
95	cyclones; and (5) Convections convection. One interesting finding is that the influence of Formatted: Font: (Default) Times New Roman
96	these factors on fire events varies over different parts of the MC. For example, the fire
97	signal in aone part of Kalimantan is strongly related to both the monsoons and ENSO. In Formatted: Font: (Default) Times New Roman
98	contrast, fire activity in Central Sumatra is not as-closely tied to the monsoons and ENSO Formatted: Font: (Default) Times New Roman
99	but MJO-signal, Formatted: Font: (Default) Times New Roman
100	Above climate variabilities or <u>Climate variability of</u> meteorological phenomena
101	affectaffects not only biomass burning emissions but also fire aerosol_transport_of fire Formatted: Font: (Default) Times New Roman Formatted: Font: (Default) Times New Roman

102	aerosols (Reid et al., 2012). Seasonal The seasonal migration of the ITCZ and the
103	associated monsoonal circulation dominate seasonal wind flows, whereas sea breeze,
104	typhoon, orbreezes, tropical cyclones, and topography determine air flow inon smaller
105	spatial scales or shorter<u>and</u> temporal scales<u>, –</u> all of themthese phenomena play
106	significant roles in determining the transport pathway of fire aerosols (Wang et al., 2013).
107	For example, during the intense haze episode of June 2013, thea long lasting
108	situationevent with a "very unhealthy" air pollution level in Singapore, was actually
109	caused by an enhanced fire aerosol transport from Sumatra to West Malaysia owing to a
110	tropical stormcyclone located in South China Sea. Recently, using a global chemistry
111	transport model combiningcombined with a back-trajectory tracer model, Reddington et
112	al. (2014) attempted to attribute particulate pollutionspollution in Singapore over a short
113	time period of 5 years to different burning sites in surrounding regions over a short time
114	period of 5 years. The coarse 2.8-degree resolution model used in the study, however, has
115	left many open questions.
116	In this study, we aim to examine and quantify the impact of fire aerosols on the
117	visibility and air quality of Southeast Asia inover the past decade. Analyses of
118	observational data and comprehensive regional model simulationsresults have both been
119	performed in order to improve our understanding of this issue. We firstly describe
120	methodologies adopted in the study, followed by the results and findings from our
121	assessment of the fire aerosol on the degradation of visibility in several selected cities and
122	also inover the great whole Southeast Asia. We then discuss the sensitivity of our findings
123	to the use of different meteorological datasets as well as fire emission inventories. The last
124	section summarizes and concludes our work.

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125 2 Methodology

126 **2.1 The model**

127 In order to address the targeted science question, we have used the Weather 128 Research and Forecasting (WRF) model coupled with chemistry component (WRF-129 Chem). The WRF model is a compressible, non-hydrostatic regional meteorology 130 model that uses the Arakawa C grid and terrain following hydrostatic pressure 131 coordinates, and includes various dynamic cores and physical parameterizations for 132 different scientific purposes (Skamarock et al., 2008). The WRF-Chem model is a 133 version of the standard WRF with an additional interactively coupled model of 134 atmospheric chemistry. WRF-Chem simulates atmospheric evolutions of chemical 135 species including particulate matters concurrently with meteorological fields, using 136 the same grid structure, advection scheme, and physics schemes for sub-grid scale 137 transport as in the standard WRF model (Grell et al., 2005). In this study, we use 138 WRF-Chem version 3.6 with a modified chemistry tracer module instead of a full 139 chemistry package. This is for the purpose to focus on the fire aerosol life cycle as the 140 first step, without involving a much more complicated gaseous and aqueous chemical 141 processing calculations. This configuration also lowers the computational burden 142 substantially, and thus enablesIn this study, we have used the Weather Research and 43 Forecasting (WRF) model coupled with a chemistry component (WRF-Chem) version 3.6 44 (Grell et al., 2005). Our focus in this study is on the fire aerosol life cycle. Therefore, we 145 chose to use WRF-Chem with a modified chemical tracer module instead of a full chemistry package, to thus model the fire PM2.5 particles as tracers without involving much 146 147 more complicated gaseous and aqueous chemical processing calculations but dry and wet

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148	depositions. Emissions of other chemical species were excluded in the simulations. This
149	configuration lowers the computational burden substantially, and thus allows us to conduct
150	long model integrations to determine the contributions of fire aerosol to the degradation of
151	air quality visibility in the region over the past decade. In WRF-Chem, the sinks of $PM_{2.5}$
152	particles include dry deposition and wet scavenging calculated at every time step. The
153	numerical simulations are employed within a model domain with a horizontal resolution
154	of 36 km, including 432 \times 148 horizontal grid points (Fig. 1), and 31 vertically staggered
155	layers based on a terrain-following pressure coordinate system. The vertical
156	layersthat are stretched with to have a higher resolution near the surface (an average depth
157	of ~30 m in the first model half layer). Variables other than vertical velocity and
158	geopotential are stored at the half model layers) based on a terrain-following pressure
159	coordinate system. The time step is 180 seconds for advection and physics calculation,
160	The physics schemes included in the simulations are listed in Table 1. The initial and
161	boundary meteorological conditions are taken from reanalysis meteorological datasetdata.
162	In order to examine the potential influence of different reanalysis products on simulation
163	results, we have used two such datasets: (1) the National Center for Environment Prediction
164	FiNaL (NCEP-FNL) reanalysis data (National Centers for Environmental Prediction,
165	2000), which has a spatial resolution of 1 degree and a temporal resolution of 6 hours; and
166	(2) ERA-Interim, which is a global atmospheric reanalysis from European Centre for
167	Medium-Range Weather Forecasts (ECMWF) (European Centre for Medium-Range
168	Weather, 2009), providing 6-hourly atmospheric fields on sixty pressure levels from
169	surface to 0.1 hPa with a horizontal resolution of approximately 80 km. Sea surface
170	temperature is updated every 6 hours in both NCEP-FNL and ERA-Interim. All

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171	simulations used four-dimensional data assimilation (FDDA) to nudge NCEP-FNL or	
172	ERA-Interim temperature, water vapor, and zonal andas well as meridional wind speeds	Formatted: Font: (Default) Times New Roman
173	above the planetary boundary layer (PBL). This approach has been shown to provide	Formatted: Font: (Default) Times New Roman
174	realistic temperature, moisture, and wind fields in a long simulation (Stauffer and Seaman,	Formatted: Font: (Default) Times New Roman
175	1994),	Formatted: Font: (Default) Times New Roman
176	In WRF Chem, the sinks of PM2.5 particles include dry deposition and wet seavenging	
177	calculated at every time step.	
178	2.2—Biomass burning emissions	
179	Two biomass burning emission inventories are also used in this study to investigate	Formatted: Font: (Default) Times New Roman
180	the sensitivity of modeled fire aerosol concentration to different emission	Formatted: Font: (Default) Times New Roman
181	estimationsestimates. The first emission inventory is the Fire INventory from NCAR	Formatted: Font: (Default) Times New Roman
182	version 1.5 (FINNv1.5) (Wiedinmyer et al., 2011), which classifies burnings of extra	
183	tropical forest, topical forest (including peatland), savanna, and grassland. It is	 Formatted: Font: (Default) Times New Roman
184	used in this study to provide daily, 36 km resolution $PM_{2.5}$ emissions. The second emission	
185	inventory is the Global Fire Emission Database with version 4.1 with small firefires	Formatted: Font: (Default) Times New Roman

included (GFEDv4.1s) (van der Werf et al., 2010; Randerson et al., 2012; Giglio et al.,

2013). GFEDv4.1s provides PM_{2.5} emissions with the same spatiotemporal resolution as
FINNv1.5.

A plume rise algorithm for fire emissions was implemented in WRF-Chem by Grell et al. (2011) to estimate fire injection height. This algorithm, however, often derives an injection height for tropical peat fire that is too high <u>comparingcompared</u> to the estimated value based on remote sensing retrievals (Tosca et al., 2011). Therefore, we have limited the plume injection height of peat fire <u>withinby a ceiling of</u> 700 m <u>above the ground</u> in this

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194	study based on Tosca et al. (2011). The vertical distribution of emitted aerosols is	
195	calculated using the plume model. This modification has clearly improved the modeled	_
196	surface PM _{2.5} concentration comparing when compared to observations in Singapore.	_
197	In order to distinguish the spatial-temporal coverage and influence of biomass burning	
198	aerosols from different regions in Southeast Asia and nearby northern Australia, we have	
199	created five tracers to represent fire aerosols respectively from Mainlandmainland	_
200	Southeast Asia (s1), Sumatra and Java islands (s2), Borneo (s3), the rest of the Maritime	
201	Continent (s4), and northern Australia (s5) as illustrated in Fig. 1. The major fire season	
202	in Mainland Mainland Southeast Asia (s1) is from February to April. In the other four	
203	regions (s2-s5), it is from August to October.	
204	Generally speaking, there areis a strong correlation between the seasonal	_
205	variationsvariation of fire emissions coordinating with those and that of rainfall in all fire	
206	regions as shown in Fig. 2. Because Mainlandmainland Southeast Asia (s1) and northern	
207	Australia (s5) are on the edge of <u>the seasonal migration of the ITCZ</u> , seasonal variations	
208	of rainfallthe correlation in these two regions areis even more pronounced. On the other	
209	hand, in Sumatra (s2), Borneo (s3), and the rest of the Maritime Continent (s4) are all	/
210	influenced by similar meteorological regimes, i.e., seasonal migration of the ITCZ.	
211	However, the passage of MJO events adds more intra-season variability of rainfall and	$\langle \rangle$
212	fire emissions in these three regions. Therefore, the seasonal variations of rainfall	
213	and fire emissions in s2, s3, and s4 are not as apparent as in the s1 and s5 regions (Fig.	
214	2b – d), owing to the influences of multiple scales of precipitation features over these	
215	areas. Nevertheless,), while inter-seasonal variations of rainfall and fire emissions are	
216	still highly correlated with each other in general, however, fire emissions do exist in some	

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217	raining seasons (Fig. $2b - d$), owing to the precipitation features in multiple scales over	
218	these regions (e.g., the passage of MJO events) and underground peatland burning.	
219	2.2 Numerical simulations and model evaluation	
220	Our simulations cover a time period slightly longer than a decade from 2003 to 2014	
221	based on available biomass burning emission estimates. The simulation of each year	Formatted: Font: (Default) Times New Roman
222	started on 1 November of the previous year and lasted for 14 months. The first two months	
223	were used for spin-up.	
224	Three sets of decadal long simulations have been conducted. The first simulation used	Formatted: Font: (Default) Times New Roman
225	NCEP-FNL reanalysis data and the FINNv1.5 fire emission inventory. This simulation is	
226	hereafter referred to as FNL_FINN and is discussed as the base simulation. In order to	
227	examine the influence of different meteorological inputs on fire aerosol life cycle, the	
228	second simulation was conducted using the same FINNv1.5 fire emission inventory as in	
229	FNL FINN but different reanalysis dataset, the ERA-Interim, and is referred to as	
230	ERA_FINN. In addition, to investigate the variability of fire aerosol concentration brought	
231	by the use of different estimates of fire emissions, the third simulation, FNL_GFED, was	
232	driven by the same NCEP-FNL meteorological input as in FNL_FINN but with a different	
233	fire emission inventory, the GFEDv4.1s. Note that the simulation period from 2003 to	
234	2014 of all these simulations was solely decided based on the temporal coverage of	
235	GFEDv4.1s.	
236	Precipitation and wind are two key factors in determining the transport and scavenging	
237	of fire aerosols. They are also the variables we use to evaluate the model's performance in	
238	simulating meteorological features. The WRF simulation driven by NCEP-FNL reanalysis	
239	data, the FNL_FINN run, produced a monthly mean precipitation of 6.80±0.55 mm day-1	

240	over the modeled domain for the period from 2003 to 2014, very close to the value of	
241	6.30±0.43 mm day-1 produced in another simulation driven by ERA-Interim, the	
242	ERA_FINN run. However, the average rainfall in both runs appears to be higher than the	
243	monthly mean of 4.71±0.37 mm day-1 from the satellite-retrieved precipitation of the	
244	Tropical Rainfall Measuring Mission (TRMM) 3B43 (V7) dataset (Huffman et al., 2007),	
245	Based on the sensitivity tests for FDDA grid nudging, the wet bias in both experiments	
246	mainly comes from water vapor nudging. Figure S1a – c are the Hovmöller plots of daily	
247	TRMM, FNL FINN, and ERA FINN precipitation in 2006, respectively. Compared to	
248	the satellite-retrieved data, both FNL_FINN and ERA_FINN have produced more light	
249	rain events, and this appears to be the reason behind the model precipitation bias. Despite	
250	the model overestimate in average total precipitation, the temporal correlation of monthly	
251	rainfall between FNL_FINN and TRMM is 0.68 and the spatial correlation is 0.85 during	
252	2003-2014 (Table 2). For ERA_FINN, the temporal correlation with TRMM is 0.90, while	
253	the spatial correlation is 0.85. In the summer monsoon season (i.e., May, June and July),	
254	both runs show the highest temporal correlations with observation but the lowest in the	
255	spatial correlations. The comparisons show that simulated rainfall generally agrees with	
256	the observation in space and time, especially when ERA-Interim reanalysis is used (i.e., in	
257	<u>ERA_FINN)</u> .these three	
258	The representative wind pattern in Southeast Asia is the monsoon wind flow. In the	
259	winter monsoon season (i.e., February, March and April), mean surface winds are from	
260	northeast in the Northern Hemisphere and turn to the northwesterly once past the Equator	
261	(Fig. S2a). On the other hand, the wind directions are reversed in the summer monsoon	
262	season (i.e., August, September and October) (Fig. S2b). We use the wind data from	
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263	NCEP-FNL and ERA-Interim reanalysis to evaluate model simulated winds. We find that	
264	both runs overestimated the u component (stronger easterly) in South China Sea (Fig. S3a	
265	and c) in the winter monsoon season, and overestimated the v component (stronger	
266	southerly) in Java Sea in the summer monsoon season (Fig. S3b and d). These regions are	
267	the entrances of monsoon wind flow into the MC. In general, model has well captured the	
268	general wind flows in Southeast Asia during both monsoon seasons but overestimated	
269	about 1 m sec ⁻¹ in wind speed in some regions (see additional discussion in Section	Formatted: Font:
270	4). likely due to terrain effect and model resolution limitation.	Formatted: Font:
271	2.3 Observational data and model derivation of visibility	
272	The definition of "visibility" is the farthest distance at which one can see a large, black	
273	object against a bright background at the horizon (Seinfeld and Pandis, 2006), There are	Formatted: Font:
274	several factors to determinedetermining visibility, but in this studyhere we mainly	Formatted: Font: Formatted: Font:
275	consider the absorption and scattering of light by gases and <u>aerosol particles</u> excluding fog	Formatted: Font:
276	or misty days. One of In this study, the most widely used equations, visibility is calculated	Formatted: Font:
277	by using the Koschmeider equation, is given by ;	Formatted: Font: Formatted: Font:
278	$VIS = 3.912 / b_{ext}, \tag{1}$	Formatted: Font: Formatted: Font:
279	where VIS is visibility with a unit in meter and b_{ext} is the extinction coefficient with a unit	
280	of m ⁻¹ . Visibility Excluding fog, visibility degradation is most readily observed from the	Formatted: Font:
281	impact of particulate pollution besides fog. Based on Eq. (1), a maximum visibility under	Formatted: Font:
282	an absolutely dry and pollution-free air is about 296 km owing to Rayleigh scattering, while	Formatted: Font:
283	a visibility onin the order of 10 km is considered asunder a moderatelymoderate to	Formatted: Font: Formatted: Font:
284	heavily pollutedheavy air pollution by particulate matters.matter (Visscher, 2013),	Formatted: Font:
285	Abnormal and persistent low visibility situations are also referred to as "haze" events.	Formatted: Font: Formatted: Font:
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286	Urban air pollutionsAir pollution sources such as fossil fuel burning can cause low	Format
287	visibility and haze events to occur. Similarly, fire aerosols, alone or mixed with other	Format Format
288	particulate pollutants, can degrade visibility by increasing best and lead to occurrence of	Format
289	haze events too.	Format
290	The observational data of visibility from the Global Surface Summary of the Day	Format
291	(GSOD) (Smith et al., 2011) are used in our study, as to identify days under particulate	Format
292	pollution, i.e., haze events. The GSOD is derived from the Integrated Surface Hourly (ISH)	Format color: B
293	dataset and archived at the National Climatic Data Center (NCDC). The daily visibility in	Format Format
294	the dataset is available from 1973 to the present.	Format
295	In order to compare with observations, we also calculate the The observed	
296	visibility using is also used to evaluate the modeled fire aerosol data, visibility and thus	Format
297	<u>PM_{2.5} concentration</u> . The modeled visibility is derived based on the extinction coefficient	Format Format
298	of thesethe fire aerosols as functions function of particle size (, by assuming a log-normal	Format
299	size distribution of accumulation mode _x with a standard deviation $\sigma = 2$), the complex	Format Format
300	refractive index of the particles, and a wavelength of 550 nm of the incident light. As	Format
301	fire plumes contain both sulfur compounds and carbonaceous aerosols, we assume	
302	the fire aerosols are aged internal mixtures with black carbon as core and sulfate as	
303	shell (Kim et al., 2008). We also consider hydroscopic growth of sulfate fraction of	Format
304	these mixed particles in the calculation based on environmental relative humidity.	Format
305	. Note that all these calculations are done for the wavelength of 550 nm unless	
306	otherwise indicated. As fire plumes contain both sulfur compounds and carbonaceous	
307	aerosols, we assume the fire aerosols are aged internal mixtures with black carbon as the	
308	core and sulfate as the shell (Kim et al., 2008). To make the calculated visibility of the fire	

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fraction of these mixed particles in the calculation based on the modeled relative humidity	
(RH). Based on Kiehl et al. (2000), the hydroscopic growth factor (rhf) is given by	
$rhf = 1.0 + exp(a_1 + \frac{a_2}{RH + a_3} + \frac{a_4}{RH + a_5}), $ (2)	
where a1 to a5 are fitting coefficients given by 0.5532, -0.1034, -1.05, -1.957, 0.3406,	
respectively. The radius increase of wet particle (r_{wet}) due to hydroscopic growth will be	
$r_{wet} = r_{dry}^{rhf},$ (3)	
where r_{dry} is the radius of dry particle in micron.	
As mentioned above, a visibility of 10 km is considered as under moderately to	
heavilyan indicator for a moderate to heavy particulate pollution so that this quantity.	
Hence a visibility of 10km in observation is used as the threshold for deriving defining the	<
"low visibility day (VLD)" in our study. In analysis, we derived We firstly derived the	<
observed low visibility days in every year for a given city using the GSOD visibility data.	
Such day is identified when the daily averaged visibility in the observation site is	
lower or equal to 10 km. Then, we derived the modeled low visibility days infollowing	\leq
the same procedure, but using modeled visibility data that were only influenced by fire	
aerosols. Both the observed and modeled visibilities were then used to define the fraction	
of low visibility days that can be caused by fire aerosols alone. It is assumed that whenever	<
fire aerosol <i>alone</i> could cause a low visibility day to occur, such a day would be attributed	<
to fire aerosol caused LVD, regardless of whether other coexisting pollutants would have	
ana sufficient intensity to cause low visibility or not. WeIn addition to the LVD, we have	<
also used a daily visibility of 7 km as the criterion to define the observed "very low	
visibility day (VLVD)". Such heavy haze events in the region are generally caused by	
	$(RH). Based on Kiehl et al. (2000), the hydroscopic growth factor (rhf) is given by rhf = 1.0 + exp(a_1 + \frac{a_2}{RH + a_3} + \frac{a_4}{RH + a_5}). (2)where at to as are fitting coefficients given by 0.5532, -0.1034, -1.05, -1.957, 0.3406,respectively. The radius increase of wet particle (rwei) due to hydroscopic growth will ber_{wet} = r_{dry}rhf_{\perp}. (3)where rwet is the radius of dry particle in micron.As mentioned above, a visibility of 10 km is considered as under moderately toheavilyan indicator for a moderate to heavy, particulate pollution-so that this quantity.Hence a visibility of 10km in observation is used as the threshold for derivingdefining the"low visibility day (VLD)" in our study. In analysis, we derived We firstly derived theobserved low visibility days in every year for a given city using the GSOD visibility data.Such day is identified when the daily averaged visibility in the observation site islower or equal to 10 km. Then, we derived the modeled low visibility days infollowingthe same procedure, but using modeled visibilities were then used to define the fractionof low visibility days that can be caused by fire aerosols along. It is assumed that wheneverfire aerosol caused LVD, regardless of whether other coexisting pollutants would haveana sufficient intensity to cause low visibility or not. WeIn addition to the LVD, we havealso used a daily visibility of 7 km as the criterion to define the observed "very lowvisibility day (VLDD)". Such heavy haze events in the region are generally caused by$

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332	severe fire aerosol pollution, thus we use their occurrence specifically to evaluate the model	
333	performance.	Formatted: Font: (Default) Times New Roman
334	2.4 The "Haze Exposure Day (HED)"	
335	We have derived a metric, the Haze Exposure Day (HED), to measure the exposure of	
336	the whole Southeast Asia, represented by 50 cities of the Association of Southeast Asian	
337	Nations (ASEAN), to low visibility events. HED can be defined in a population weighted	
338	format for the analyzed 50 cities, indicating the relative exposure of the populations in	
339	these cities to the low visibility events caused by particulate pollution;	Formatted: Font: (Default) Times New Roman
340	$HED_{pw} = \sum_{i=1}^{N} C_{pw}(i) $	Formatted: Font: (Default) Times New Roman
341	where	Formatted: Font: (Default) Times New Roman
342	$C_{pw}(i) = pop(i) \cdot C(i) / \sum_{i=1}^{N} pop(i) $ (5)	Formatted: Font: (Default) Times New Roman
343	is the population-weighted fraction of the total Haze Exposure Days, N equals to the total	
344	number of cities (50), <i>i</i> is the index for the 50 analyzed cities, <i>pop(i)</i> is the population for	
345	a given city (Table S1), and C(i) represents the annual LVDs for that city calculated from	
346	the GSOD dataset. Note that we assume that the population of each city stays constant	
347	throughout the analyzed period. Another assumption of HED_{pw} is that everyone in a given	
348	city would be equally exposed to the particulate pollution,	Formatted: Font: (Default) Times New Roman
349	In addition, HED can be also defined in an arithmetic mean format, assuming each city	
350	weights equally regardless of its population. Its value hence emphasizes on the relative	
351	exposure of each area within the analyzed region:	
352	$HED_{ar} = \sum_{i=1}^{N} C(i)/N.$ (6)	
354	absolute and relative contributions of fire aerosols to the total low visibility events in the	Forr
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355	region. We will label the fire-caused HED as <i>fHED_{pw}</i> and <i>fHED_{ar}</i> thereafter.	Forr
356	2.4—Numerical simulations	
357	Our simulations cover a time period slightly longer than a decade from 2002 to	
358	2014 based on availability of biomass burning emission estimations. The simulation	Forr
359	of each year started on 1 November of the previous year and lasted for 14 months. The	
360	first two months are used for spin-up.	
361	Three sets of decadal long simulations have been conducted. The first simulation	Forr
362	used reanalysis data of NCEP-FNL and fire emission inventory of FINNv1.5. This	
363	simulation is hereafter referred to as FNL_FINN and discussed as the base simulation.	
364	In order to examine the influence of different meteorological inputs on fire aerosol	
365	life cycle, the second simulation was conducted using the same FINNv1.5 fire	
366	emission inventory as in FNL_FINN but a different reanalysis data of ERA_Interim,	
367	referring to as ERA_FINN. In addition, to investigate the variability of fire aerosol	
368	concentration brought by the use of different estimations of fire emissions, the third	
369	simulation, FNL_GFED, was driven by the same NCEP-FNL meteorological input as in	
370	FNL_FINN but a different fire emission inventory, the widely used GFEDv4.1s. Since	
371	the daily emission of GFEDv1.1s is only available after 2003, the period of the	
372	FNL_GFED simulation is from 2003 to 2014.	
373	Precipitation is one of the key factors in determining the transport and	
374	scavenging of fire aerosols. WRF simulation driven by NCAR_FNL reanalysis data, or	
375	the FNL_FINN run, produced a monthly mean precipitation of 6.81 ± 0.55 mm day $^{-1}$	
	17	

Both HED_{pw} and HED_{ar} can be also calculated using fire-caused LVDs to define the

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376	over the modeled domain for the period from 2002 to 2014, very close to the value of		
377	6.29±0.43 mm day ⁻¹ produced in another simulation driven by ERA_Interim, or the		
378	ERA_FINN run. Comparing to the monthly mean of 4.69±0.38 mm day ⁻¹ from the		
379	satellite retrieved precipitation in the Tropical Rainfall Measuring Mission (TRMM)		
380	3B43 (V7) dataset (Huffman et al., 2007), however, both results appear to be higher,		Formatted: Font: (Default) Times New Roman
381	Based on the sensitivity tests for FDDA grid nudging, the wet bias in both experiments		
382	mainly comes from water vapor nudging. Figure 3a - c are the Hovmöller plot of daily		
383	TRMM, FNL_FINN, and ERA_FINN precipitation in 2006, respectively. Comparing to		
384	the observations, both FNL_FINN and ERA_FINN have produced more light rain		
385	events, and this appears to be the reason behind the model precipitation bias. Despite		
386	the model overestimation in averaged total precipitation, the temporal correlation of		
387	normalized rainfall anomaly between FNL_FINN (ERA_FINN) and TRMM is 0.69		
388	(0.90) and the spatial correlation is 0.86 (0.85) during 2002–2014. The comparisons		
389	show that simulated rainfall generally agrees with the observation in space and time,		
390	especial when ERA-Interim reanalysis is used (i.e., in ERA_FINN).		Formatted: Font: (Default) Times New Roman
891	3 Assessment of the impact of fire aerosols on the visibility in Southeast Asia		Formatted: Font: (Default) Times New Roman
392	visibility	\sum	Formatted: Indent: Left: 0 cm, Hanging: 0.76 of Line spacing: 1.5 lines
		$\backslash \rangle$	Formatted: Font: (Default) Times New Roman
393	3.1 Impact of fire aerosols on the visibility in four selected cities		Formatted: Font: (Default) Times New Roman
394	We first to focus our analysis on four selected cities in the region, Bangkok (Thailand),		
395	Kuala Lumpur (Malaysia), Singapore (Singapore), and Kuching (Malaysia), all located		
396	close to the major Southeast fire sites ranging from the mainland to the islands. of		Formatted: Font: (Default) Times New Roman
897	Southeast Asia Specifically Bangkok is a smoke recentor city of the fire events in the		Formatted: Font: (Default) Times New Roman
	<u>Sourcest risk</u> sponteury, Bungkok is a sinoke receptor ery of the fire events in the	\leq	Formatted: Font: (Default) Times New Roman

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398	mainland of Southeast Asia (s1) while Kuala Lumpur and Singapore are two cities
399	frequently under the influence of Sumatra (s2) as well as Borneo fires (s3). Kuching is in
400	the coastcoastal area of Borneo so that and directly affected by Borneo fire events (s3).
401	The low visibility events in these four near-fire-site cities during the fire seasons
402	from 2002 to 2014, defined as days with daily averaged visibility lower or equal to 10
403	km, or Low Visibility Days (LVDs), have been identified using the daily GSOD visibility
404	database and then compared with modeled results (Fig. 4). We find that the model
405	has reasonably captured the LVDs despite certain biases. Specifically, for the Very
406	Lower Visibility Days (VLVDs), here defined as events with daily averaged visibility
407	lower or equal to 7 km, the modeled and observed results display a good correlation
408	despite a model overestimate in visibility value or underestimate in degrading
409	visibility in certain events. In Southeast Asia, severe haze events equivalent to the
410	VLVDs in visibility degradation are largely caused by fire aerosol pollutions.
411	Assuming this is true, the performance of our model in reproducing the major fire
412	events is very good since only 10% or fewer VLVDs observed in the past decade were
413	not captured by the model (Table 2; Fig. 4). Note that other than these VLVDs, for
414	many LVDs fire aerosol might not be the only reason responsible for the degradation
415	of visibility.
416	In addition to the visibility data, we have also obtained the ground based*
417	observations of PM _{2.5} concentration in recent years from the National Environment
418	Agency (NEA) of Singapore. Figure 5a shows the comparison of time series of
419	observed and FNL_FINN simulated daily PM25-during 2013-2014. The surface
420	observational data of PM _{2.5} concentration among these four cities are only available in
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421	Singapore since 2013 from the National Environment Agency (NEA) of Singapore. We
422	thus firstly used these data along with visibility data to evaluate model's performance for
423	fire-caused haze events reported in Singapore during 2013-2014 (Fig. 3), Note that the
424	observed $PM_{2.5}$ level reflects the influences of both fire and non-fire aerosols, whereas the
425	modeled PM _{2.5} only includes the impact of fire aerosols. However, We find that the model
426	still predicted clearly high $PM_{2.5}$ concentrations during most of the observed haze events,
427	especially in June 2013, and in spring and fall seasons of 2014 (highlighted green areas),
428	though with underestimates in particle concentration of up to 30-50%, likely due to the
429	model resolution, a model overestimation of rainfall, and the errors in emission
430	inventory. Once again, the model has shown a solid performance in capturing all the
431	major known haze events caused by fire PM in Singapore (Fig. 5b). Specifically to the
432	observed VLVDs, we evidence that fire aerosol is the main reason behind these
433	eventsmodel's exclusion of non-fire aerosols, coarse model resolution, overestimated
434	rainfall, or errors in the emission inventory. Figure 4 shows observed visibility versus
435	modeled visibility in FNL_FINN during the fire events shown in Fig. 3. Note that all these
436	events have an observed visibility lower than or equal to 10 km, or can be identified as
437	LVDs. In capturing these fire-caused haze events, the model only missed about 22% of
438	them, or reporting a visibility larger than 10 km in 40 out of 185 observed LVDs as marked
439	with purple color in Fig. 4. When observed visibility is between 7 and 10 km, model results
440	appear to align with observations rather well. For cases with visibility lower than 7 km,
441	the model captured all the events (by reporting a visibility lower than 10 km, or LVD)
442	although often overestimated the visibility range. These results imply that the VLVDs only
443	count a very small fraction in LVDs and thus are episodic events. It is very likely that the
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4	size of concentrated fire plumes in VLVDs might be constantly smaller than the 36 km	
5	model resolution; therefore, the model results could not reach the peak values of PM _{2.5}	
6	concentrations of these plumes.	
7	Furthermore, the LVDs in the four selected near-fire-site cities during the fire seasons	
8	from 2003 to 2014 have been identified using the daily GSOD visibility database and then	
9	compared with modeled results (Fig. 5). It is difficult to identify all the fire caused haze	
0	events beyond Singapore even in recent years. However, in Southeast Asia, severe haze	
1	events equivalent to the VLVDs in visibility degradation are known to be largely caused	
2	by fire aerosol pollution. Therefore, we used the observed VLVDs in the four selected	
3	cities to evaluate the performance of the model. We find that the modeled result displays	
4	a good performance in capturing VLVDs despite an overestimate in visibility range during	
5	certain events compared with the observation. The model in general only missed about	
5	10% or fewer VLVDs observed in the past decade (Table 3; Fig. 5). In addition, the model	
7	has reasonably captured the observed LVDs despite certain biases (Fig. 5), likely due to	
•	the fact that fire aerosol might not be the only reason responsible for the degradation of	
	visibility during many LVDs.	
)	We find that the annual mean LVDs in Bangkok has increased from 46%47% (172	
	days) in the first 5-year period of the simulation duration ($\frac{2002-20092003-2007}{2003-2007}$) to 74%	
	(272 days) in the last 5-year period (2010-2014), so does the). The LVDs caused by fire	
	aerosols has increased as well (Fig. 6a). Overall, fire aerosols are responsible for more	
	than one third of these LVDs (i.e. $38_{}39$ % in average; Table $23_{}$. The largest source of	
	fire aerosols affecting Bangkok is burning of agriculture waste and other biomass burning	
	in s1 during the dry season of spring (Fig. 7a; Table 34). During the fire season, abundant	

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467 fire aerosols degrade visibility and even cause VLVDs to occur (Fig., mainly 60)- Ninety-468 eight percent of VLVDs in Bangkok occurred from December to April (Fig. <u>6e)</u>. Based 469 on our model results, 8987% of VLVDs can be identified as fire caused. 470 In Kuala Lumpur, the percentage of LVDs also gradually increases since 2006 to reach 471 a peak in 2011 and again in 2014 (Fig. 6b). During 2005-2010 the frequency of total LVDs 472 have increased 10-15% each year, mainly attributing to the pollution sources other than 473 fires. However, fire-caused LVDs arebecome more evident after 2009. Seasonal wise, 474 there are two peaks of fire aerosol influence, one in February-March and another in August (Fig. 6f), corresponding to the trans-boundary transport of fire aerosols from 475 476 Mainland mainland Southeast Asia (s1) in the winter monsoon season and from Sumatra 477 (s2) in the summer monsoon season, respectively (Fig. 7b). Three quarter of VLVDs are 478 occurred in the summer monsoon season due to Sumatra fires. NotedNote that in 479 November and December the percentage of LVDs is over 50% and dominated by-the 480 pollutants other than fire aerosols. These non-fire aerosols come from either local sources 481 or the areas further inland riding on the winter monsoon circulation. Overall, fire pollution 482 is responsible for 35% or 36%, a substantial fraction of total low visibility events in Kuala 483 Lumpur during 20022003-2014 (Table 23). 484 The percentage of LVDs in Singapore has been rapidly increasing since 2012 (Fig. 485 6c). Except for 2014During the simulation period, this increase isappears to be mostly 486 from anthropogenic pollution other than fires, especially in 2012 and 2013. High 487 percentage of LVDs in November and December could be induced by aerosols from 488 further inland of Mainland Southeast Asia through long range transport driven by the 489 monsoon circulation (Fig. 6g). Similar In monthly variation, similar to Kuala Lumpur,

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490	there are two peaks of fire aerosol influence, one appear in February-March and another
491	in September-October, respectively (Fig. 6g). TheIn February and March, the trans-
492	boundary transported fire aerosols can transport of fire aerosols come from mainland
493	Southeast Asia (s1), while in the summer monsoon season fire aerosols come from both
494	Sumatra (s2) and Borneo (s3) in the summer monsoon season (Fig. 7c). Except for the
495	severe haze events in June 2013, VLVDs basically occur in September and October (i.e.
496	92%) due to both Sumatra and Borneo fires. In general, up to 34% of LVDs in Singapore
497	are caused by fire aerosols inbased on the FNL_FINN simulation and the rest by local and
498	long-range transported pollutants (Table 2). Fire3). Nevertheless, fire aerosol is still the
499	major reason for the episodic severe haze conditions.
500	Because of its geographic location, Kuching is affected heavily by local fire events
501	during the fire season (Fig. 7d). Fire aerosols can often degrade the visibility easily to
502	lower than <u>below</u> , 7 km and even reach <u>reaching</u> , 2 km (Fig. 445d). The LVDs mainly
503	occur in August and September during the fire season (Fig. 6d and h). The frequency of
504	LVDs in Kuching is similar to Singapore; however, 25% of those LVDs are considered to
505	be VLVDs in Kuching while only 4% are in Singapore in comparison (Table $\frac{23}{2}$).
506	3.2 Impact of fire aerosols on the visibility inover , the greater whole. Southeast
507	Asia
508	Air quality degradation caused by fires apparently occurs in regions beyond the above-
509	analyzed four cities. To examine such degradation inover the greater whole Southeast
510	Asia, we have extended our analysis to cover 50 cities of the Association of Southeast
511	Asian Nations (ASEAN)., The impact of particulate pollution on the greater whole
512	Southeast Asia is measured by a metric of the "Haze Exposure Day" (HED). HED can be
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532 Apparently, both HED_{pw} and HED_{ar} can be also calculated using
533 fire-caused LVDs (here using the results of FNL_FINN) to define the
534 absolute and relative contributions of fire acrosols to the total low visibility events in the

535 region. We will label the fire-caused HED as *fHED*_{pv} and *fHED*_{ar} thereafter.

536 <u>Table S1).</u>

537 We find that both HEDpw and HEDar increase rather steadily over the past decade (Fig. 538 8b), demonstrating that the exposure to haze events either weighted by population or not 539 has become worse in the region. Generally speaking, the fire aerosols are responsible for up to 40-60% of the total exposures exposure to low visibility across the region. In both 540 measures, the increase of fire-caused HED (2.64 and 3.37 days per year for population-541 542 weighted and arithmetic mean, respectively) is similar to that of overall HED (2.61 and 543 3.59 days per year for population-weighted and arithmetic mean, respectively) (Fig. 8b), 544 suggesting that fire aerosol has taken the major role in causing the degradation of air quality 545 in Southeast Asia comparingcompared to the non-fire particulate pollution. The result that HED_{pw} is higher than HED_{ar} in most of the years indicates that the particulate pollution is 546 547 on average worse over more populous cities than the others. Interestingly, the discrepancy 548 of these two variables, however, has become smaller in recent years and even reversed in 549 2014, implying an equally worsening of haze event occurrence across from the smaller to 550 the bigger cities in terms of population in the region. The reason behind this result could 551 be a widely wider spread of fire events in the region, particularly causing acute haze events 552 in the cities even with relatively low populations. Regarding the increase of fire-caused HED, because biomass burning, especially peatland burning, usually occurs in the rural 553 554 areas, higher fire emissions would extend low visibility conditionconditions to a larger

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area regardless of its population. On the other hand, <u>air pollution caused bydue to</u> industrialization, urbanization, and other factors such as population growth<u>increases</u> rapidly, <u>air pollution has become worse</u> across the region so that-even cities with lower populationpopulations now increasingly suffer from low visibility from fossil fuel burning and other sources of particulate pollution. Therefore, the mitigation of air quality degradation needs to consider both fire and non-fire sources.

561 **3.3** The influence of wind and precipitation on fire aerosol life cycle

Seasonal migrations of the ITCZ and associated summer and winter monsoons dominate seasonal wind flows that drive fire aerosol transport. Additionally, as discussed previously, certain small—<u>scale</u> or short-term phenomena such as sea <u>breeze</u>, <u>typhoonbreezes</u>, <u>typhoons</u>, and topography-<u>forced</u> circulations also play important roles in distributing fire aerosols. Nevertheless, we focus our <u>discussionsdiscussion</u> here on the former.

From February to April is the main fire season in Mainland Southeast Asia 568 569 (s1). In the FNL FINN simulation, the seasonal mean concentration of PM2.5 within the 570 planetary boundary layer (PBL) can exceed 20 μ g m⁻³ in this region. (note that the air 571 quality standard suggested by World Health Origination is 10 µg m⁻³ for annual mean and 572 25 μg m⁻³ for 24-h mean). During this fire season, the most common wind direction is 573 from northeast to southwest across the region (Fig. 9a). Fire aerosol plumes with 574 concentration<u>concentrations</u> higher than 0.1 μ g m⁻³ can transport with the main 575 windbe transported westward as far as 7000 km from the burning sites. In contrast, 576 February to April is not the typical burning season in the islands. Low fire emissions added

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577	by in combination with a lack of long-range transport of fire aerosols from the mainland
578	due to the seasonal circulation result in a low PM _{2.5} level over these regions (Fig. 9b - d).
579	Wet scavenging is a major factor to determinedetermining the lifetime and thus
580	abundance of suspended fire aerosols in the air. The effect of wet scavenging of fire
581	aerosols is reflected from the wet scavenging time calculated using the modeled results-
582	The wet scavenging time, which is a ratio of the aerosol mass concentration and to the
583	scavenging rate, the latter is (a function of precipitation rate). Thus, short scavenging
584	timetimes often indicates indicate high scavenging raterates except for the sites with
585	extremely low aerosol concentrationDuring February-April, at the ITCZ's furthest
586	southern extent, the short scavenging time < 1 day around 10°S shows a quick removal of
587	fire aerosols by heavy precipitation that has prevented, preventing the southward
588	transport of aerosols (Fig. 9f). Whereas On the other hand, the long scavenging time (> 5
589	days) in the Western Pacific warm pool, South China Sea, the Indochina peninsula, Bay of
590	Bengal, and Arabian Sea leads to a long suspending time of aerosols transported to these
591	regions. During the same season, over the islands of Sumatra and Borneo, the abundance
592	along with the likelihood of being transported to other places of fire aerosols, either
593	emitted locally or trans-boundary transported, are greatly limited by the high scavenging
594	rate (short scavenging time) over this these regions (Fig. 9g and h). The South China Sea
595	is in a dry condition has little precipitation during this time period, therefore, fire aerosols
596	from the northern part of Philippinethe Philippines can be transported to this region and
597	stay longer than 5 days (Fig. 9i).

The months of August to October, when the ITCZ reaches its furthest northern extent, mark the major fire season of Sumatra, Borneo, and some other islands in the Maritime Formatted: Font: (Default) Times New Roman

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600	ContinentMC (Fig. 10bS5b - d). Australia fires also mainly occur in this season (Fig.
601	10eS5e). Mean wind flows are from southeast to northwest in the Southern Hemisphere,
602	and turn to the northeast direction once passpast the Equator. Within the MC the seasonal
603	variation of rainfall is small during this time, with heavy precipitation and thus short
604	scavenging timetimes (< 3 days) mostly existentiating along the MJO path (Fig. 10fS5f -
605	i) (Wu and Hsu, 2009). The high scavenging rate in the regions close to the fire sites in
606	the islands shortens the transport distance of fire aerosol plumes with PM2.5 concentration
607	> 0.1 μ g m ⁻³ to less than 3000 km (Fig. 10bS5b - d). Long scavenging timetimes (> 5
608	days) primarily existsexist in the Banda Sea and northern Australia due to the ITCZ
609	location. Fire aerosols from Java Island (s2) (Fig. 10gS5g), Papua New Guinea (s4) (Fig.
610	10iS5i), and northern Australia (s5) (Fig. 10jS5j) can thus suspended in the air
611	for a relatively long time over these regions.
612	The above-discussed seasonal features of precipitation and aerosol scavenging
613	strengthrate help us to better understand the variability of haze occurrence and also to
614	identify the major source regions of fire aerosols influencing selected Southeast Asian
615	cities (Fig. 7). For example, the geographic location of Bangkok, which is inside the s1
616	emission region, determines that about 99% nearly all the fire aerosols is (99%) are from
617	sources within the region from December to April (Fig. 7a and Table 34). Fire aerosols
618	from all the other burning sites stay at very low level levels even during the burning seasons
619	there due to circulation and precipitation scavenging. For Kuala Lumpur and Singapore,
620	over 90% of totalthe fire aerosols reachedreaching both cities come from
621	Mainlandmainland Southeast Asia (s1) in January-April due to the dominant winter
622	monsoon circulation. During May-October, however, the major sources of fire aerosols
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623	shift to Sumatra (s2) and Borneo (s3) aidingaided by northward wind (Fig. 10bS5b and c).
624	The monthly variations of PM _{2.5} concentration in Kuala Lumpur and Singapore also have
625	a largely similar pattern (Fig. 8b7b and d). The annual mean contribution of different
626	emission regions in Kuala Lumpur are 43% from Mainlandmainland Southeast Asia (s1),
627	4950% from Sumatra (s2), 4% from Borneo (s3), 3% from the rest of Maritime Continent
628	(s4), and 0.43% from northern Australia (s5) in FINL_FINN (Table 34). Similar to Kuala
629	Lumpur, there are two peak seasons of the monthly low visibility days contributed by fire
630	aerosols in Singapore (Fig. 6g), well correlated with modeled high fire PM2.5 concentration
631	(Fig. 7c). The low visibility days in March and April mainly are caused by fire aerosols
632	from Mainlandmainland Southeast Asia (s1) under southward wind pattern (Fig. 9a), and
633	those in May to October are affected by Sumatra (s2) first in May to June, and then by both
634	s2 and s3 (Borneo) during August to October due to north- or northwest-ward monsoonal
635	circulation (Fig. 10b <u>S5b</u> and c; also Table 34). Kuching, similar to Bangkok, is strongly
636	affected by local fire aerosols (s3) during the fire season (July – October). The annual
637	mean contribution from Borneo (s3) is 85% while%, with only 78% from
638	Mainlandmainland Southeast Asia (s1) and 5% from Sumatra (s2) (Table 34).
639	Reddington et al. (2014) applied two different models, a 3D global chemical transport
640	model and a Lagrangian atmospheric transport <u>tracer</u> model to examine the long-term
641	mean contributions of fire emissions to PM2.5 from different regions to PM2.5 in several
642	cities in Southeast Asia. The Their estimated contribution from Mainland mainland
643	Southeast Asia to the above-discussed four selected cities was lower than our result during
644	January-May, likely due to their use of a different emission inventory and the coarse
645	resolution of their global model. The FINNv1.5 dataset used in our study specifically



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646	provides	higher	PM2.5	emissions	from	agriculture	fires	(the	major	fire	type	ın
647	Mainland	mainlar	nd Sout	heast Asia)	than G	FED4.1s doe	s , – the	latte	r is an u	pdate	d vers	ion

648 of the dataset (GFEDv3) used in Reddington et al. (2014) (Fig. 2). The detaildetailed

649 comparison of FNL_FINN and FNL_GFED will be discussed in the following section.

inventories on modeled fire aerosol abundance

650 4 Influence of different reanalysis<u>meteorological</u> datasets and emission

651

652 As discussed in the previous section, meteorological conditions, particularly wind field 653 and precipitation, could substantially influence the life cycle and transport path of fire 654 aerosols during the fire reasons; thereforeseasons. Therefore, it is necessary to examine 655 any potential discrepancies discrepancy in modeled particulate matter abundance 656 attributing toarising from the use of different meteorological datasets. 657 In When comparing the two of our simulations, one was driven by the NCAR_NCEP-658 FNL (i.e., FNL FINN), another) and the other, by the ERA_Interim (i.e., ERA FINN) 659 meteorological input, we find that the ERA FINN run consistently produces less 660 precipitation than the FNL FINN run during the rainingrainy seasons over the past decade (Fig. 2) (; also see the comparison results of both runs with observations in Section 2.4).2.). 661 662 Regarding fire aerosol life cycle, less rainfall in ERA FINN results in a-weaker wet

scavenging condition and thus higher abundance of fire aerosol concentrationaerosols
than in FNL FINN. We find that the annual mean concentration of fire PM_{2.5} produced in

- 665 the ERA_FINN run in Bangkok, Kuala Lumpur, Singapore, and Kuching is 8.89.2, 5.48
- 666 3.4, and 7.97, μ g m⁻³, respectively, clearly higher than the corresponding results of the
- 667 FNL_FINN run of 8.0, 4.95, 5.3, 3.0, and 7.16.9, μ g m⁻³ (Table 34). In Mainland

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668	Southeast Asia, a twenty-one percent lower rainfall in ERA_FINN causes the	
669	significantly different general, fire PM2.5 concentration comparing to in ERA_FINN is	Formatted: Font: (Default) Time
670	about 10% higher than in FNL FINN-result in. However, the occurrence of low visibility	Formatted: Font: (Default) Time
671	events is less sensitive to the fire season (February - April) (Fig. 2a and 11a). In Kuala	Formatted: Font: (Default) Time
672	Lumpurdifferences in rainfall in places near the burning areas such as Bangkok and	
673	Kuching, as indicated by a nearly negligible enhancement of VLVDs in the ERA_FINN	
674	run in Bangkok and Kuching (~1%) (Table 3). In comparison, the difference in fire PM _{2.5}	Formatted: Font: (Default) Time
675	concentrationwind field between thesethe two runs mainly comes from Sumatra (s2)	Formatted: Font: (Default) Time
676	during June to September; however, in Singapore and Kuching the concentration	Formatted: Font: (Default) Time
677	difference comes from both Sumatra (s2) and Borneo (s3) in August to October (Fig.	
678	11b - d and Table 3), all corresponding to the discrepancy of rainfall between	
679	FNL_FINN and ERA_FINN in these regions (Fig. 2b and c).has a much smaller impact	
	the sector of the sector of the sector field as a first sector of the sector of the sector of the sector of the	
680	than that of precipitation on modeled particulate matter abundance.	Formatted: Font: (Default) Time
680 681	The difference in aerosol scavenging between ERA_FINN and FNL_FINN extends	Formatted: Font: (Default) Time
680 681 682	than that of precipitation on modeled particulate matter abundance. The difference in aerosol scavenging between ERA_FINN and FNL_FINN extends to a difference as high as 7% and 12% in the resulted LVDs of Bangkok and Kuching,	Formatted: Font: (Default) Time
680 681 682 683	than that of precipitation on modeled particulate matter abundance. The difference in aerosol scavenging between ERA_FINN and FNL_FINN extends to a difference as high as 7% and 12% in the resulted LVDs of Bangkok and Kuching, respectively, (Table 2), though its influence on the results of Kuala Lumpur and	Formatted: Font: (Default) Time
680 681 682 683 684	than that of precipitation on modeled particulate matter abundance. The difference in aerosol scavenging between ERA_FINN and FNL_FINN extends to a difference as high as 7% and 12% in the resulted LVDs of Bangkok and Kuching, respectively, (Table 2), though its influence on the results of Kuala Lumpur and Singapore is much smaller (3~4%). In general, fire PM _{2.5} concentration in ERA_FINN	Formatted: Font: (Default) Time
680 681 682 683 684 685	than that of precipitation on modeled particulate matter abundance. The difference in aerosol scavenging between ERA_FINN and FNL_FINN extends to a difference as high as 7% and 12% in the resulted LVDs of Bangkok and Kuching, respectively, (Table 2), though its influence on the results of Kuala Lumpur and Singapore is much smaller (3~4%). In general, fire PM _{2.5} concentration in ERA_FINN is about 10% higher than in FNL_FINN; however, the substantial impact of fire	Formatted: Font: (Default) Time
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680 681 682 683 684 685 686 687	The difference in aerosol scavenging between ERA_FINN and FNL_FINN extends to a difference as high as 7% and 12% in the resulted LVDs of Bangkok and Kuching, respectively, (Table 2), though its influence on the results of Kuala Lumpur and Singapore is much smaller (3~4%). In general, fire PM _{2.5} concentration in ERA_FINN is about 10% higher than in FNL_FINN; however, the substantial impact of fire aerosols on LVDs is more sensitive in places near the burning areas, i.e., Bangkok and Kuching. Interestingly, a mild increase of VLVDs in the ERA_FINN run in Bangkok and	Formatted: Font: (Default) Time
680 681 682 683 684 685 686 687 688	The difference in aerosol scavenging between ERA_FINN and FNL_FINN extends to a difference as high as 7% and 12% in the resulted LVDs of Bangkok and Kuching, respectively, (Table 2), though its influence on the results of Kuala Lumpur and Singapore is much smaller (3~4%). In general, fire PM _{2.5} concentration in ERA_FINN is about 10% higher than in FNL_FINN; however, the substantial impact of fire aerosols on LVDs is more sensitive in places near the burning areas, i.e., Bangkok and Kuching. Interestingly, a mild increase of VLVDs in the ERA_FINN run in Bangkok and Kuching (~1%) (Table 2) implies that the occurrence of severe haze events is less	Formatted: Font: (Default) Time
680 681 682 683 684 685 686 687 688 688 689	than that of precipitation on modeled particulate matter abundance. The difference in aerosol scavenging between ERA_FINN and FNL_FINN extends to a difference as high as 7% and 12% in the resulted LVDs of Bangkok and Kuching, respectively, (Table 2), though its influence on the results of Kuala Lumpur and Singapore is much smaller (3~4%). In general, fire PM _{2.5} concentration in ERA_FINN is about 10% higher than in FNL_FINN; however, the substantial impact of fire aerosols on LVDs is more sensitive in places near the burning areas, i.e., Bangkok and Kuching. Interestingly, a mild increase of VLVDs in the ERA_FINN run in Bangkok and Kuching (~1%) (Table 2) implies that the occurrence of severe haze events is less affected by the rainfall difference in the burning areas.	Formatted: Font: (Default) Time

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690	In addition to meteorological inputs, differences varioususing different fire emission	$\langle \langle$
691	estimationsestimates could also affect the modeled results. To examine such an influence,	
692	we have compared two simulations with the same meteorological input but different fire	
693	emission inventories, the FNL_FINN using FINNv1.5 and FNL_GFED using GFEDv4.1s.	
694	The main differences between the two emission inventories appear mostly in	
695	Mainlandmainland Southeast Asia (s1) and northern Australia (s5) (Fig. 2a and e ; Fig. 12a	
696	and e). For instance, the peak month of fire $PM_{2.5}$ concentration in Bangkok shifts	
697	from March in FNL_FINN to January in FNL_GFED (Fig. 11a), owing to the difference	
698	in temporal pattern between the two fire emission inventories (Fig. 2a). Comparing).	
699	Compared to FINNv1.5, fire emissions in GFEDv4.1s over Mainland Southeast	\langle
700	Asia are more than 66% lower (Fig. 2a), and this results in a 4043% lower fire PM _{2.5}	
701	<u>concentration</u> in Bangkok (Fig. 11a and Table 34). The lower fire $PM_{2.5}$ concentration in	\leq
702	FNL_GFED actually produced produces a visibility that matches better with	\mathbb{Z}
703	observationobservations in Bangkok comparing to the result of FNL_FINN (Fig. S1a).	
704	The difference in monthly S5a). This implies that the fire emissions over the islands	
705	between the two emission inventories is small, with the fire emission in FINNv1.5	
706	generally higher than that in GFEDv4.1s (Fig. 2b – d). However, fire emissions in	
707	GFEDv4.1s are much higher during the fire season in the dry years (i.e. 2004, 2006	(
708	and 2009) over s2 and s3 (Fig. 12b and c), leading to a modeled mean $PM_{2.5}$	
709	concentration by FNL_GFED in Kuala Lumpur and Singapore that is higher than that	
710	by FNL_FINN during the fire season (Fig. 11b and c). On the other hand, the higher	
711	PM _{2.5} concentration simulated in FNL_GFED during the June 2013 severe haze event	
712	in Kuala Lumpur and Singapore is due to the spatiotemporal distribution of fire spots	
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713	rather than absolute perhaps overestimated in mainland Southeast Asia. In northern	
714	Australia, fire aerosol emissions. Based on our simulations, fire aerosols from Sumatra	Formatted: Font: (Default) Times New Roman
715	(s2) are mainly responsible for the severe haze event in June 2013 (Fig. 7b – c and	
716	Fig. S2b – c). During this event, the total amount of fire emissions in Sumatra (s2) is	
717	lower in GFEDv4.1s than FINNv1.5, however, distributed rather more densely over a	
718	smaller area (Fig. 13c and d). As a result, under the same meteorological condition,	
719	the simulated $PM_{2.5}$ in the FNL_GFED simulation reaches Singapore in a higher	
720	concentration that also matches better with observation than the result of FNL_FINN	
721	(Fig. 13b). A similar result also appears in Kuching, where the difference in modeled	
722	PM _{2.5} concentration between the two model runs is likely related to the difference in	
723	spatial or temporal distributions rather than the mean quantities of PM _{2.5} emissions	
724	since the latter suggested by FINNv1.5 are almost the same in both fire emissions	Formatted: Font: (Default) Times New Roman
725	inventories.	
726	The most evident difference between the two emission inventories occurs in	
727	northern Australia, where FINNv1.5 suggests an almost negligible fire aerosol	Formatted: Font: (Default) Times New Roman
728	emission comparingcompared to GFEDv4.1s (Fig. 2e). Therefore, in the FNL_GFED	Formatted: Font: (Default) Times New Roman
729	simulation, Australia fire aerosols play an important role in Singapore air quality,	
730	contributing to about 22% of the modeled PM2.5 concentration in Singapore. In contrast,	Formatted: Font: (Default) Times New Roman
731	Australia fires have nearly no effect on Singapore air quality in the FNL_FINN run (Table	
732	3). Our results raise the important issue of the sensitivity of modeled aerosol	
733	concentration in downwind areas to the spatiotemporal distribution, besides the	
734	absolute emission amount from the fire spots. A further study regarding this topic	
735	would be much needed. <u>4).</u>	Formatted: Font: (Default) Times New Roman

736	We would also like to point out the importance of spatiotemporal distribution of fire		
737	emission to the modeled results. For example, during the June 2013 severe haze event in		
738	Kuala Lumpur and Singapore, the total amount of fire emissions from Sumatra (s2) in		
739	GFEDv4.1s are lower than those of FINNv1.5 (Fig. S6a) but distributed rather more		
740	densely over a smaller area (Fig. S6c and d). As a result, under the same meteorological		
741	conditions, the simulated PM2.5 in the FNL_GFED simulation reaches Singapore in a		
742	higher concentration that also matches better with observations than the result of		
743	FNL FINN (Fig. S7b).		
744	5 Summary and Conclusions	(For
745	We have examined the extent of the biomass burning aerosol's impact on the air		
746	quality of Southeast Asia in the past decade using surface visibility and surface PM2.5	_(Foi
747	measurements along with the WRF model with a modified fire tracer module. The model	1	Foi
748	has shown a good performance in capturing 90% of the observed severe haze events		
749	(visibility < 7 km) <u>caused by fire aerosols occurred inover</u> past decade in several cities <u>that</u>	(Foi
750	are close to the major burning sitesSuch events are known to be induced mainly by		For For
751	biomass burning. On the more general cases of particulate pollution, ourOur study	(Foi
752	also suggests that fire aerosols are responsible for a substantial fraction of the low visibility	(Foi
753	days (visibility < 10 km) in several<u>these</u> cities: <u>38up to 39</u>% in Bangkok, <u>3536</u>% in Kuala_	(Foi
754	Lumpur, 34% in Singapore, and 3233% in Kuching.	\triangleleft	Foi Foi
755	The life cycle and transport path, and thus spatial and temporal distributions of	\geq	For
756	fire aerosols are all influenced by meteorological conditions, especially the seasonal		
757	precipitation distribution and atmospheric circulations. These impacts are well		
	prospiration and annospirate encandrons. These impacts are wen		

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758 reflected from the variations of abundance of fire aerosols in the selected cities in 759 analysis. In general, Mainland In attributing the low visibility events to fire emissions 760 from different sites, we find that mainland Southeast Asia is the major contributor during 761 the Northeast or winter monsoon season in Southeast Asia. In the Southwest or summer 762 monsoon season, however, most fire aerosols come from Sumatra and Borneo. 763 Specifically, fires in Mainlandmainland Southeast Asia are accounted for the largest 764 percentage of the total fire PM_{2.5} in Bangkok (99.2%), and fires from Sumatra are the major 765 contributor in Kuala Lumpur (5150%) and Singapore (4241%). Kuching receives 8885% 766 of fire aerosols from local Borneo fires.

By comparing the results from two modeled runs with the same fire emissions but driven by different meteorological inputs, we have examined the potential sensitivity of modeled results to meteorological datasets. The discrepancy in modeling themodeled low visibility events due to arising from the use of different meteorological datasets is clearly evident, especially in the results of Bangkok and Kuching. However, using different meteorological input datasets does not appear to have influenced the modeled very low visibility events, or the severe haze events in the cities close to burning sites.

We have also examined the sensitivity of modeled results to the use of different emission inventories. We find that significant discrepancies of fire emissions in <u>Mainlandmainland</u> Southeast Asia and northern Australia between the two emission inventories used in theour study have caused significant substantial difference in modeled fire aerosol concentration and visibility, particularlyespecially in Bangkok and Singapore. For instance, the contribution to fire aerosol in Singapore from northern Australia changes from nearly zero in the simulation driven by FINNv1.5 to about 22% in Formatted: Font: (Default) Times New Roman

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781 another simulation driven by GFEDv4.1s. We have also identified the influence of the 782 discrepancydifference, in spatiotemporal distribution rather than total emitted quantities Formatted: Font: (Default) Times New Roman 783 from the fire hotspots on modeled PM2.5 concentration. Further analysis on this direction 784 is much needed, 785 To further assess the impacts of fire eventsparticulate pollution on the air 786 guality surface visibility of the great whole Southeast Asia and to estimate the fire aerosol's 787 contribution, we have defined and derived a metric of "Haze Exposure Days" (HEDs), by 788 integrating annual low visibility days of 50 cities of the Association of Southeast Asian 789 Nations and weighted by population or averaged arithmetically. We find that a very large 790 population of Southeast Asia has been exposed to relatively persistent hazy 791 condition conditions. The top four cities in the HED ranking, Jakarta, Bangkok, Hanoi, 792 and Yangon, with a total population exceeding two millions, all have experienced more 793 than 200 days per year of low visibility due to particulate pollution over the past decade. 794 Even worse is that the number of annual low visibility days have been increasing steadily 795 not only in high population cities but also those with relatively low populations, suggesting 796 a widelywide spread of particulate pollutions into the great across Southeast Asian 797 region. Generally speaking, the fire aerosols are found to be responsible for up to about 798 half of the total exposes to low visibility acrossin the region. Our result suggests that in 799 order to improve the air quality in Southeast Asia, besides reducing or even prohibiting 800 planned or unplanned fires, mitigation policies targeting at pollution sources other than 801 fires need to be put in effect implemented as well. 802

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812	for providing the numerical model for this study. We thank the National Supercomputing	
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¥97		

998 999	Table 1. WRF ph	vsics scheme configuration		
-	Physics Processes	Scheme		
_	microphysics	Morrison (2 moments) scheme		Formatted: Font: (Default) Times New Roman
	longwave radiation	rrtmg scheme		Formetted, Font: (Default) Times New Doman
	shortwave radiation	rrtmg scheme		Formatted: Font. (Delauit) Times New Roman
	surface-layer	MYNN surface layer		Formatted: Font: (Default) Times New Roman
	land surface	Unified Noah land-surface model		Formattad: Font: (Default) Times New Reman
	planetary boundary layer	WYNN 2.5 level TKE scheme		Formatteu. Form. (Derault) Times New Roman
-	cumulus parameterization	Grell-Freitas ensemble scheme		Formatted: Font: (Default) Times New Roman
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Table 2. The spatial and temporal correlation of monthly rainfall between model and observation during 2003-2014. FMA, MJJ, ASO, NDJ and All indicate February-April, May-July, August-October, November-January and whole year, respectively. 1005

	FNL FINN	vs. TRMM	ERA_FINN	vs. TRMM
	Spatial cor.	Temporal cor.	Spatial cor.	Temporal cor.
FMA	<u>0.89</u>	<u>0.61</u>	<u>0.89</u>	<u>0.89</u>
MJJ	<u>0.83</u>	0.69	<u>0.81</u>	<u>0.90</u>
ASO	<u>0.86</u>	0.59	<u>0.84</u>	0.89
<u>NDJ</u>	0.88	0.60	0.88	0.85
All	0.86	0.68	0.85	0.90

1009 1010

1011 Table 23. Annual mean low visibility days (LVDs: <u>observed visibility ≤ 10 km</u>) and very /

1012 low visibility days (VLVDs; observed visibility ≤ 7 km) per year, in Bangkok, Kuala

1013 Lumpur, Singapore and Kuching during 2003-2014 are presented in the second column,

1014 <u>Parentheses show the percentage of year. The third and fourth columns show the</u>/ percentage contributions along with standard deviations of fire and non-fire (other)

percentage contributions along with standard deviations of fire and non-fire (other)
 pollutions for total low visibility days-in Bangkok, Kuala Lumpur, Singapore and
 Kuching during 2002-2014 (FNL_GFED is from 2003 to 2014), Parentheses show the

1018 percentage of year.

1019

FNL_FINN	LVD per year (days)	Fire pollution contribution (%)	Other pollution contribution (%)	
Bangkok, Thailand	211±49 (58215±50	38<u>39</u>± 8	62 61±8 •	ĺ
-	(59±14%)			
Kuala Lumpur, Malaysia	166±80 (45±22<u>174</u>±78	35±1836±17	65±1864±17	
	<u>(48±21%)</u>		/	
Singapore, Singapore	92±84 (25±23<u>96</u>±87	34±16 <u>17</u>	67±1666±17	h
	<u>(26±24%)</u>			ļ
Kuching, Malaysia	95±54 <u>57 (26±1517%)</u>	32±14<u>33</u>±15	68±14 <u>67±15</u>	
FNL_FINN	VLVD per year (days)	Fire pollution contribution (%)	Other pollution contribution (%)	[
Bangkok, Thailand	17±10 (5±3<u>15</u>±8	89±19 87±20	<u>11±1987±20</u> ◀	ļ
	<u>(4±2%)</u>			Î
Kuala Lumpur, Malaysia	18<u>19</u>±18 (5±5%)	85±17	15±17	/
Singapore, Singapore	4±4 (1±1%)	92±32 91±33	8±32 9±33	/
Kuching, Malaysia	24±19 (7<u>22±18 (6</u>±5%)	94±12<u>9</u>3±11	6±12 <u>7±11</u>	
ERA_FINN	VLD per year (days)	Fire pollution contribution (%)	Other pollution contribution (%)	1
Bangkok, Thailand	211±49 (58215±50	45±846±7	55±854±7 ◀	1
	(59±14%)			į
Kuala Lumpur, Malaysia	166±80 (45±22174±78	39<u>40</u>±16	61 <u>60</u> ±16	1
	$(48\pm21\%)$		/	(
Singapore, Singapore	92±84 (25±23<u>96</u>±87	37±18	63±18	ų
	<u>(26±24%)</u>			/
Kuching, Malaysia	95±54 <u>57 (26±1517%)</u>	44 <u>45</u> ±17	56<u>55</u>±17	1
ERA FINN	VLVD ner vear (days)	Fire pollution	Other pollution	
	· 1. · 2 per year (augs)	contribution (%)	contribution (%)	
Bangkok, Thailand	$\frac{17\pm10(5\pm315\pm8)}{17\pm10(5\pm315\pm8)}$	<u>9088</u> ±20	<u>1012</u> ±20	×
	<u>(4±2%)</u>	00.10	10.10	
Kuala Lumpur, Malaysia	$\frac{1819\pm18(5\pm5\%)}{14(1+10(1))}$	90±18	10±18	
Singapore, Singapore	$4\pm4(1\pm1\%)$	<u>98±56</u>	<u>2±56</u>	
Kuching, Malaysia	$\frac{24\pm19(722\pm18(0\pm5\%))}{22\pm18(0\pm5\%)}$	9594±11	11± <u>0</u> €	A.
FNL_GFED	VLD per year (days)	Fire pollution contribution (%)	Other pollution contribution (%)	- AND
Bangkok, Thailand	215±50 (59±14%)	36±8	64±8	Ì
Kuala Lumpur, Malaysia	174±78 (48±21%)	28±17	72±17	
Singapore, Singapore	96±87 (26±24%)	29±21	71±21	1
Kuching, Malaysia	95±57 (26±1517%)	26±18	74±18	1
FNL_GFED	VLVD per year (days)	Fire pollution	Other pollution	A DESCRIPTION OF THE OWNER OWNER OF THE OWNER OWNER OF THE OWNER OWNE
- Denskalt Theiland	15+9(4+29/)	contribution (%)	contribution (%)	
Dangkok, Inaliand	$13\pm 8 (4\pm 2\%)$ 1910+18 (5+59/)	90±19 92±29	10±19 17±28	I
Singapora Singapora	$\frac{1019\pm10}{4\pm4}$ (3±3%)	0J=20 80+27	11+27	۱
Kuching Malaysic	$4\pm4(1\pm170)$ 22+18(6+5%)	80±28	11±3/	1
Kuuning, malaysia	22±10 (0±370)	07=20	11±20	í

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1022	Table <u>34</u> .	Annual 1	mean a	and standa	rd deviation	of fi	ire PM _{2.5}	concentration	(μ g	g m ⁻³	3)
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1023 1024 1025 1026 contributed by each source region in Bangkok, Kuala Lumpur, Singapore, and Kuching during 20022003-2014 (FNL_GFED is from 2003 to 2014). Parentheses show the

percentage of fire aerosol fraction in total $PM_{2.5-}$ contribution originating from each source region. The same regions, s1-s5, are explained in Fig. 1.

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N	s1	\$2	\$3	s4	\$5
Bangkok	8 04+2 63	0.0+0.0	0.0+0.0	0.1+0+0-	0.0+0.0
Bunghon	(99.12+0.5%)	$(0.1\pm0.1\%)$	$(0.1\pm0.1\%)$	0. <u>1±</u> 0± <u>.</u> 0.	$(0.0\pm0.0\%)$
	$()).12 \pm 0.570)$	(0.1±0.170)	(0.1±0.170)	$\frac{10}{(0.7.6\pm0.5\%)}$	(0.0±0.070)
Kuala	2. 13 ±1.2	2. 5 7±1.4	0.2±0. 1 2	0.1±0.1	0.0 ± 0.0
Lumpur	$(43.3 \pm 14.68\%)$	(49. 36± 14. 3 9%	$(4.1\pm43.3\pm3.4\%)$	(2. 9 5±2. 6 3%)	(0.43±0.2%
	()	(· · · · · · · · · · · · · · · · · · ·	())
Singapore	1. <mark>01</mark> ±0.7	1.2±0.8	0. 5 4±0.4	0.2±0.1	0.1±0.0
	(34.3±16.4<u>36.7</u>±14.	(40.7±15. <u>39</u> %)	(16<u>14.3±10</u>.0±11.	(6. 7±4.2<u>1±3.8</u>	(2.2±1.1%)
	<u>7</u> %)		3 %)	%)	
Kuching	0.4 <u>5</u> ±0.4	0.3±0.1	6. 30 ±3.2	0.1±0.1	0.0 ± 0.0
	(7. <u>38</u> ±6. <u>65</u> %)	(4. <u>67</u> ±2.4 <u>5</u> %)	(85.3<u>84.6</u>±9.<u>67</u>%)	(2.3±2. <u>35</u> %)	(0.6±0. <u>23</u> %
)
ERA_FIN				4	
N	sl	s2	\$3	s4	\$5
Bangkok	$\frac{8.79.1}{2.73}$	0.0±0.0	0.0±0.0	0.1±0.0	0.0±0.0
	$(99.\frac{12}{2}\pm 0.4\%)$	$(0.1\pm0.1\%)$	$(0.1\pm 0.1\%)$	$(0.\frac{46}{6} \pm 0.4\%)$	$(0.0\pm0.0\%)$
Kuala	2. <u>+3</u> ±1.2	3. <u>02</u> ±1. <u>54</u>	0.2±0.2	0.1 ± 0.0	0.0 ± 0.0
Lumpur	(38.6<u>39.7</u>±12.7%)	$(53.7 \pm 11.9 \times 12.3)$	$(4.7\pm4.23.9\pm3.3\%)$	$(2.6\pm 2.3\pm 1.8)$	$(0.4\pm0.2\%)$
		%)		%)	
~.	1. <u>01</u> ±0.6	1.4±0.9	0. <u>4</u> <u>6</u> ±0.6		0.1±0.0
Singapore	$(31.9\pm15.334.2\pm13.)$	(40.45 ± 13.17)	$(\frac{18.9\pm12}{17.2\pm11})$	0.2±0.1	(1.9± <u>1.</u> 0 <u>.9</u>
	<u>5</u> %))	8%)	$(6.82 \pm 3.71\%)$	%)
			6. <u>97</u> ±3. <u>89</u>		0.0 ± 0.0
Kuching	0.5±0.4	0.4±0.2	$(\frac{83.482.5}{10.10})$	0.1±0.1	$(0.6\pm 0.23\%)$
	$(7.8.1\pm5\pm5.7.6\%)$	(5.9<u>6.1</u>±3.9%))	$(2.7\pm \frac{2.9}{3.0}\%)$)
FNL_GFE	c1	s?	c ²	сA	.5
Bangkok	1 8+1 2	0.0+0.0	0.0+0.0	0.0+0.0	0.0±0.0
Duligkok	$(00.6\pm0.2\%)$	$(0.1\pm0.0\%)$	$(0.1\pm0.1\%)$	$(0.2\pm0.2\%)$	$(0.1\pm0.0\%)$
Kuala	1 2+0.6	2 7+1 0	0.1+0.2		0.1 ± 0.070
Lumpur	$(38.6\pm 20.8\%)$	$(52.8 \pm 21.10\%)$	(28+250/)	$(0.8\pm0.8\%)$	$(2 0 \pm 2 40/2)$
Singanore	(30.0-20.070)	(33.0-21.170)	(2.0-3.370)	(0.0±0.070)	0.4+0.2
Singapore	0 3+0 2	1 5+1 8	0.4+0.5	0.1+0.0	(223+132)
	(22 1+17 3%)	(40.2+23.6%)	(125+0.5)	(2.0+2.4%)	(22.5±15.2
Kuching	(22.1±17.370)	$(+0.2\pm23.0\%)$	(12.3±9.370)	(2.9=2.470)	0 3+0 2
Kuching					0.5±0.2
Ruening	0.1+0.1	0.1 ± 0.1	2 2+2 2	0.0+0.0	(11.6 ± 6.70)

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C and D indicate the location of four selected cities: Bangkok, (A), Kuala Lumpur, (B),

Singapore (C) and Kuching, respectively. (D).



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Figure 2. <u>MonthlyTime series of monthly PM2.5 emissionsemission</u> (Tg year⁻¹) in FINNv1.5 (redpink solid lines) and GFEDv4.1s <u>(pink(red dashed lines)</u>. Also shown are precipitation raterates (mm day⁻¹) simulated in FNL FINN (light blue <u>solid lines</u>) and ERA_FINN (blue <u>dashed lines</u>). All data are averaged) during 20022003-2014 forin; (a) Mainland<u>mainland</u> Southeast Asia (s1), (b) Sumatra and Java islands (s2), (c) Borneo (s3), (d) the rest of the Maritime Continent (s4), and (e) northern Australia (s5). Note that GFEDv4.1s PM_{2.5} emission is averaged from 2003 to 2014.

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1085 1086 1087 Figure 5. (a) Time series of daily surface PM2.5 from the ground-based observations (black 1088 line) and FNL_FINN simulated results (red line) in Singapore during 2013-2014. (b) Time 1089 beries of daily visibility of GSOD observation (black line) and calculated result from 1090 FNL_FINN (red line) in Singapore during 2013-2014. Highlighted green areas are known 1091 haze events caused by fire aerosols. Two gray lines mark the visibility of 7 and 10 km, 1092 respectively. 1093

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Figure 4. A scatter plot of observed visibility and FNL_FINN visibility during known fire events as labeled in Fig. 4b. Black dash line refers 1:1 line and red line is the threshold of

1095 1096 1097 1098 1099 1100 VLVD (7 km). Data points marked with purple color are the events that model failed to produce a visibility qualified for LVD.



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- 1124 1125 1126 1127 2014.2003-2014, derived using GSOD visibility observations in Bangkok, Kuala Lumpur, Singapore, and Kuching, respectively. Each bar presents the observed LVDs in each year or month. Red color shows the partition of fire-caused LVDs (captured by model) while green color presents non-fire LVDs (observed – modeled).

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1142	Bangkok, (b) Kuala Lumpur, (c) Singapore and (d) Kuching, all derived from FNL_FINN	Formatted: Font: (Default) Times New Roman
1143	simulation and averaged over the period 2002 of 2003 2014.	 Formatted: Font: (Default) Times New Roman
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148Figure 8. (a) The mean low visibility days (circles) per year from 20022003 to 2014 in 50149ASEAN cities and their. The size of the circles indicates the number of days. The colors150refer to population-weighted fraction in the total Haze Exposure Days (HED; colors). (b)151Annual variation of population-weighted HED (HED_{pw}) and arithmetic mean HED152(HEDar). Fire-caused HED are labeled as fHED_{pw} and fHEDar. Units are in days. -Note that153the y-axes are in different scales.

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PBL modeled in FNL_FINN during February to April, 2002-2003-2014 in: Mainland for
fire PM_{2.5} source region from (a) mainland Southeast Asia (s1), (b) Sumatra and Java
island (s2),islands, (c) Borneo (s3), (d) the rest of the Maritime Continent (s4), and (e)
northern Australia (s5), respectively. (f) (g) Same as (a)-(e) but for seasonal mean wet
scavenging time (days; shaded) and column intergraded PM_{2.5} concentration (µg m⁻²;
contours) within the PBL height.)

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1189 Kuching over the period 2002-2014 (FNL_GFED is from 2003 to 2014).





Figure 12. Temporal variation of monthly PM2.5 emission (Tg year-1) in FINNv1.5

1195 1196 (pink solid lines) and GFEDv4.1s (red dashed lines). Also shown are precipitation

1197 rates (mm day⁻¹) simulated in FNL_FINN (light blue solid lines) and ERA_FINN (blue

dashed lines) during 2002-2014 in: (a) Mainland Southeast Asia (s1), (b) Sumatra 1198

1199 (s2), (c) Borneo (s3), (d) the rest of the Maritime Continent (s4), and (e) northern 1200 1201 Australia (s5).



Figure 13. (a) Time series of daily mean PM2.5 emissions (Tg year*) in Sumatra (S2)
 from FINNv1.5 (red line) and GFEDv4.1s (green line). (b) Time series of daily mean
 PM2.5 concentration (µg m⁻³) in Singapore from observation (black line), and modeled
 results from FNL_FINN (red line) and FNL_GFED (green line). (c) Monthly mean PM2.5
 emissions (Tg year⁻¹) from FINNv1.5 in June 2013. (d) same as (c) but from
 GFEDv4.1s.