

1 **Aircraft observations of water-soluble dicarboxylic acids in the aerosols over**
2 **China**

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18 **Abstract**

19 Vertical profiles of dicarboxylic acids, related organic compounds and SOA tracer
20 compounds in particle phase have not yet been simultaneously explored in East Asia, although
21 there is growing evidence that aqueous phase oxidation of volatile organic compounds may be
22 responsible for the elevated organic aerosols (OA) in the troposphere. Here, we found
23 consistently good correlation of oxalic acid, the most abundant individual organic compounds
24 in aerosols globally, with its precursors as well as biogenic-derived secondary OA (SOA)
25 compounds in Chinese tropospheric aerosols by aircraft measurements. Anthropogenically
26 derived dicarboxylic acids (i.e., C₅ and C₆ diacids) at high altitudes were 4-20 times higher than
27 those from surface measurements and even occasionally dominant over oxalic acid at altitude
28 higher than 2 km, which is in contrast to the predominance of oxalic acid previously reported
29 globally including the tropospheric and surface aerosols. This indicates an enhancement of
30 tropospheric SOA formation from anthropogenic precursors. Furthermore, oxalic acid-to-
31 sulfate ratio maximized at altitude of ~2 km, explaining aqueous-phase SOA production that
32 was supported by good correlations with predicted liquid water content, organic carbon and
33 biogenic SOA tracers. These results demonstrate that elevated oxalic acid and related SOA
34 compounds from both the anthropogenic and biogenic sources may substantially contribute to
35 tropospheric OA burden over polluted regions of China, implying aerosol-associated climate
36 effects and intercontinental transport.

37 **1 Introduction**

38 Water-soluble dicarboxylic acids (diacids), one of the most abundant organic aerosol
39 (OA) constituents, have been reported in the marine (Kawamura and Sakaguchi, 1999), remote
40 (Kawamura et al., 1996), desert (Sorooshian et al., 2012), mountainous, rural, semi-urban, and
41 urban atmosphere (Ho et al., 2007). As important components of aerosols, diacids play an
42 important role in Earth's climate by directly scattering sunlight or indirectly by enhancing the
43 ability of OA to act as cloud condensation nuclei (CCN) (Kumar et al., 2003). They can

44 contribute to wide ranges from a few percent of the water-soluble mass (Sorooshian et al.,
45 2007b), which could serve as tracers for the atmospheric processing of water-soluble OA
46 (Ervens et al., 2011), to more than 10% of organic carbon (OC) in the remote marine
47 atmosphere (Kawamura and Sakaguchi, 1999).

48 Diacids in aerosols may be directly emitted from fossil fuel combustion (e.g. vehicle
49 exhaust) (Kawamura and Kaplan, 1987), biomass burning (Narukawa et al., 1999) and cooking
50 emissions (Rogge et al., 1991). They can also be formed by degradation from unsaturated fatty
51 acids (Kawamura et al., 1996) and cyclic alkenes (Hatakeyama et al., 1987). Laboratory studies
52 have provided direct experimental evidence that stepwise aqueous oxidation of relatively
53 longer-chain (n) dicarboxylic acids can yield the corresponding short-chain (n-1) dicarboxylic
54 acids (Enami et al., 2015). In addition, in-cloud and below-cloud measurements as well as other
55 field measurements have revealed that aqueous-phase oxidation of volatile organic compounds
56 (VOCs) and intermediates such as glyoxal (Gly), methylglyoxal (MeGly) and pyruvic acid (Pyr)
57 in wet aerosols or clouds and the subsequent gas-particle partitioning are more important
58 pathway for the production of small diacids such as oxalic and malonic acids (Lim et al.,
59 2013;Carlton et al., 2007;Lim et al., 2005;Carlton et al., 2006;Yu et al., 2005;Sorooshian et al.,
60 2007b;Sorooshian et al., 2006;Zhang et al., 2016). Oxalic acid (C₂) is the most abundant diacid
61 with concentrations ranging from a few ng m⁻³ in remote locations (Kawamura et al., 1996) to
62 hundreds or even up to one thousand ng m⁻³ in urban regions (Ho et al., 2007) and highly
63 forested regions (Falkovich et al., 2005). Indeed, from a model study it has been proposed that
64 photochemical oxidation of isoprene and subsequent partitioning of water soluble
65 photooxidation products into the aqueous phase is a predominant formation pathway of oxalic
66 acid (Myriokefalitakis et al., 2011). Most of the previous studies of diacids and other organic
67 compounds in aerosols have been conducted at ground surface; however, only a few aircraft
68 observations that have been conducted over the arctic region (Talbot et al., 1992), the western
69 North Pacific (Narukawa et al., 1999), the western rim of the Pacific Ocean (Kawamura et al.,
70 2003), the coastal marine stratocumulus and cumulus clouds over USA (Crahan et al.,

71 2004;Sorooshian et al., 2013;Sorooshian et al., 2007a;Wonaschuetz et al., 2012;Prabhakar et
72 al., 2014) and over inland agricultural and urban areas in the western United States (Sorooshian
73 et al., 2015). These aircraft experiments have revealed that water-soluble diacids may be
74 produced by the photochemical oxidation of anthropogenic organic compounds in the
75 atmosphere and thus play an important role in controlling the chemical and physical properties
76 of OA in the troposphere. However, the evidence of secondary production of LWM diacids
77 (i.e., oxalic acid) in tropospheric aerosols may be not enough due to the lack of dataset from
78 simultaneous measurements of their possible precursors and/or intermediates such as particle-
79 phase glyoxal (Gly), methylglyoxal (MeGly), pyruvic acid (Pyr) and glyoxylic acid (ω C₂) as
80 well as other secondary organic aerosol (SOA) compounds from photochemical oxidation of
81 VOCs such as isoprene and monoterpenes in the aerosols collected from aircraft campaigns.

82 East Asia is one of the most important source regions of OA (Zhang et al., 2007), and
83 this is especially true for China due to the rapid urbanization, industrialization and energy
84 consumption (Huang et al., 2014;Zhang et al., 2015a;Zhang and Cao, 2015). Elevated OA were
85 found in the free troposphere from ACE-Asia aircraft observations, exceeding model-predicted
86 organic aerosols by a factor of 10-100 (Heald et al., 2011;Heald et al., 2005;Henze and Seinfeld,
87 2006). There is growing evidence that SOA formation from aqueous phase processing of VOCs,
88 especially from isoprene may partially explain the missing OA in the troposphere over East
89 Asia (Heald et al., 2011;Heald et al., 2005;Henze and Seinfeld, 2006), however, vertical
90 profiles of LWM diacids, related organic compounds (i.e., ketoacids and α -dicarbonyls) and
91 SOA tracer compounds in particle phase have not yet been simultaneously analyzed in this
92 region. Additional chemical constraints (e.g. simultaneously measuring different types of
93 organic aerosols such as diacids, ketoacids and α -dicarbonyls and SOA tracer compounds) are
94 urgently required to investigate the importance of SOA formation pathway to the
95 global/regional budget of OA. Our previous studies have reported the molecular composition
96 of primary organic aerosols (POA) including n-alkanes, fatty acids, sugars, and polycyclic
97 aromatic hydrocarbons (PAHs) as well as biogenic SOA tracers in aerosol particles over China

98 (Wang et al., 2007;Fu et al., 2014). However, chemical composition and molecular distribution
99 of diacids have never been reported by aircraft measurements over mainland and coastal China,
100 although only limited studies have been conducted on their spatial distributions based ground
101 observations (Ho et al., 2007).

102 In this study, we have collected aerosol samples over coastal to inland China using
103 aircraft during spring, summer and winter. The samples were analyzed for a series of LWM
104 diacids and related compounds as well as SOA tracer compounds to investigate the vertical
105 profiles, possible sources and formation pathways of LWM diacids and related organic
106 compounds in the polluted troposphere over China. In light of these analyses, we observed that
107 production of oxalic acid and related compounds from anthropogenic and biogenic precursors
108 is enhanced at high altitudes in the troposphere.

109 **2 Materials and Methods**

110 **2.1 Aircraft Campaigns**

111 Aircraft measurements were carried out over the coastal East China Sea in winter
112 (12/25/2002–01/06/2003, n=18) and inland China during summer (08/08/2003–09/13/2003,
113 n=14) and spring (05/19/2004–06/10/2004, n=16) using Yun-12 and Yun-5 airplanes as
114 described elsewhere(Wang et al., 2007;Fu et al., 2014) . The sampling heights were from 500
115 m to 3000 m (with uncertainty of around 10%) above ground level across many major cities
116 such as Changzhou, Nanjing, Hefei, Wuhan, Chongqing and Chengdu for inland aerosol
117 samples as well as Wenzhou, Ningbo, Shanghai, Changzhou, Qingdao and Dalian for coastal
118 aerosol samples. The detailed flight tracks and flight information are shown in Figure 1 and
119 Table S1 (see Supporting Information). PM_{2.5} aerosols were collected on pre-heated quartz
120 fiber filter (diameter of 90 mm) using a medium-volume air sampler (Beijing Geological
121 Instrument Factory, China). Aerosol samples were collected for the sampling period from ca.
122 80 min. to two hours. Air was taken via an inlet installed below the cabin of the aircraft at a
123 flow rate of 78 L/min. The lack of using organic denuders may lead to a positive artifact (e.g.,

124 10% of the organic carbon) by possible adsorption of gas-phase organic acids on the quartz
125 filters (Cheng et al., 2009), but this will not affect our conclusions. Because air conditioning
126 was not available in the aircraft, the ambient temperatures inside and outside the cabin should
127 be similar, and thus potential adsorption of gaseous organic acids on quartz filter should be
128 minimal.

129 **2.2 Measurement of diacids**

130 Diacids and related organic compounds such as ω -oxocarboxylic acids (or oxoacids),
131 pyruvic acid and α -dicarbonyls were determined as described elsewhere (Kawamura and
132 Ikushima, 1993;Ho et al., 2010). Briefly, the sample and blank filters were extracted with
133 organic free ultrapure water (10 ml \times 3) in a glass vial for 10 min. To remove insoluble particles
134 and filter debris, the extracts were passed through a Pasteur pipette packed with quartz wool.
135 The combined extracts were concentrated using a rotary evaporator under a vacuum and then
136 further concentrated using a nitrogen blow-down system. The concentrated extracts were
137 reacted with 14% BF_3 /n-butanol to derive carboxyl group to dibutyl ester and oxo group to
138 dibutoxy acetals.

139 The derivatives were extracted with n-hexane, concentrated to near dryness, dissolved
140 with n-hexane in 1.5 ml glass vials. The samples were finally determined using a gas
141 chromatography (GC, HP6980) with a split/splitless injection, fused silica capillary column
142 (HP-5, 0.2 mm i.d. \times 25 m long \times 0.52 μm film thickness) and a flame ionization detector.
143 Identification of the compounds was performed by GC retention times with those of authentic
144 standards and GC/mass spectrometry analyses. Recoveries of authentic standards spiked to a
145 pre-heated quartz fiber filter were generally better than 85% for all organic compound identified
146 except for oxalic acid (78%), pyruvic acid (72%), and methylglyoxal (55%). Duplicate analyses
147 of filter samples from other sites indicated that analytical errors are smaller than 15%. During
148 the aircraft campaigns, field blank filters were mounted onto the sampler for seconds without
149 pumping. Blank and sample filters were placed individually in a clean (pre-combusted at 450

150 °C for 6 h) glass jar sealed with a Teflon-lined screw cap, transported to the laboratory, and
151 stored at -20 °C prior to analysis (Wang et al., 2007).

152 Small peaks of oxalic, malonic and glyoxylic acids were found in the field and lab
153 blanks, but they were smaller than 10% of the real samples. The concentrations reported here
154 are all corrected for the field blanks but not for the recoveries. The method of SOA tracers
155 measurement was described elsewhere (Fu et al., 2014) .

156 **2.3 Measurements of EC and OC**

157 The concentrations of organic carbon (OC) and elemental carbon (EC) were measured
158 with thermal-optical transmittance method by OC/EC Carbon Aerosol Analyzer (Sunset
159 Laboratory Inc., USA) following the IMPROVE protocol (Chow et al., 2004). All the analyses
160 of the filter samples were completed in 2006.

161 **2.4 Estimation of liquid water content and aerosol acidity**

162 Liquid water content (LWC) and aerosol acidity were predicted by ISORROPIA II,
163 which is a thermodynamic equilibrium model with robust and rapid convergence for all aerosol
164 types (Fountoukis and Nenes, 2007). Measured concentrations of water-soluble inorganic ions,
165 ambient temperature and relative humidity were used as input.

166 **3 Results and discussion**

167 **3.1 Concentrations and molecular distributions**

168 Total concentrations of LWM diacids and related compounds quantified in the inland
169 aircraft PM_{2.5} (i.e., particulate matter with a diameter smaller than 2.5 µm) are 730±328 and
170 586±457 ng m⁻³ during summer and spring, respectively, while in the coastal aerosols, the
171 average concentration is 254±209 ng m⁻³ in winter (Table 1). It is of great interest to note that
172 the concentrations of diacids in the troposphere over the Arctic and North Pacific are generally
173 much lower than those reported at ground levels (Kawamura et al., 1996;Kawamura and

174 Sakaguchi, 1999); however, the concentrations in the troposphere over mainland China are
175 within the range of or even higher than those reported at ground levels in major cities (Ho et
176 al., 2007; Kawamura and Ikushima, 1993). Due to the enhanced anthropogenic emissions,
177 concentrations of POA (i.e., n-alkanes, fatty acids, sugars, lignin and resin products, sterols,
178 PAHs, and phthalic acids) are higher in winter than those in summer and spring (Wang et al.,
179 2007). In contrast, the concentrations of diacids in summer and spring are 2-15 times higher
180 than those in winter, indicating that these compounds are mostly of secondary origin via the
181 oxidation of their gaseous precursors such as isoprene and α/β -pinene (Kanakidou et al.,
182 2005; Carlton et al., 2006; Carlton et al., 2007; Carlton et al., 2009; Ervens et al., 2011) as
183 discussed below.

184 Molecular distributions of diacids (C_n is n-numbered carbon diacid) in our study are
185 generally characterized by the predominance of oxalic acid (C_2) followed by succinic (C_4) and
186 malonic (C_3) acids during spring and winter, being consistent with previous findings obtained
187 in Chinese megacities (Ho et al., 2007) (Figure 2a). However, in many summer samples
188 collected at height of above 2 km, we found the predominance of glutaric (C_5) and adipic (C_6)
189 acids (see Figure 2b), which are the major organic compounds produced by the oxidation of
190 anthropogenic cyclohexene and methylenecyclohexane (Hamilton et al., 2006; Muller et al.,
191 2007). Although C_5 and C_6 diacids are the ozonolysis products in smog chamber studies,
192 Pavuluri et al. (2015) recently proposed that C_5 and C_6 acids are produced by photochemical
193 processing of aqueous aerosols from a laboratory study. Such a molecular distribution has not
194 been reported for the tropospheric and ground level aerosols (Hatakeyama et al., 1987; Enami
195 et al., 2015; Kawamura and Bikkina, 2016). In our measurements, the averaged concentrations
196 of C_5 and C_6 in summer are $159 \pm 79 \text{ ng m}^{-3}$ and $93.9 \pm 23.2 \text{ ng m}^{-3}$, which are 4-20 times higher
197 than those in ground measurements in many megacities in China (Ho et al., 2007), Tokyo
198 (Kawamura and Yasui, 2005) and Los Angeles (Kawamura and Kaplan, 1987). Such high
199 abundances of C_5 and C_6 observed in summer imply an important formation pathway associated

200 with enhanced photochemical oxidation of anthropogenic precursors in the polluted
201 troposphere over China.

202 **3.2 C₃ (C₂) to C₄ ratio**

203 The malonic to succinic acid (C₃/C₄) ratios can provide information on source,
204 formation pathways and photochemical aging of organic aerosols. C₃/C₄ ratios in aerosols
205 derived from vehicular exhausts (i.e., 0.25–0.44, av. 0.35) (Kawamura and Kaplan, 1987) have
206 lower values than those in ambient aerosols from Tokyo (i.e., 0.56–2.9, av. 1.6) (Kawamura
207 and Ikushima, 1993) and China's megacities (i.e., 0.6–1.1, average 0.74) (Ho et al., 2007). In
208 contrast, the ratios are substantially higher for aged aerosols because C₃ is more produced by
209 photochemical processing of C₄ (Kawamura and Ikushima, 1993). Actually, higher C₃/C₄ ratios
210 are observed for remote marine aerosols from the North Pacific including tropic (range: 1–11,
211 av. 3.9) which are subjected to extensive aging during the long-range atmospheric transport
212 (Kawamura and Sakaguchi, 1999). In this study, C₃/C₄ ratios are 0.5±0.4 in summer, 0.9±0.1
213 in spring (Table 1), and 0.7±0.2 in winter, apparently smaller than those in aerosols affected by
214 atmospheric aging, but similar to that of urban aerosols in China (i.e., 0.9 in summer and 0.6 in
215 winter) (Ho et al., 2007). In addition, C₃ and C₄ diacids can be formed by OH reactions of
216 water soluble organic compounds such as glyoxal (Tan et al., 2009) and C₃/C₄ ratios gradually
217 decrease with the reaction time (Pavuluri et al., 2015). The lower C₃/C₄ ratios in summer
218 indicate that diacid aerosols are relatively fresh but will be subjected to more atmospheric aging
219 than in other seasons due to increasing OH concentrations and solar radiation.

220 C₂/C₄ ratios (2.0 in summer, 4.6 in spring and 4.6 in winter) in our study are found to
221 be much smaller than those from ground observations in China during summer (inland cities in
222 summer: av. 7.1) and winter (coastal cities in winter: av. 7.9). The lower C₂/C₄ and C₃/C₄ ratios
223 could be also resulting from degradation of C₂ and C₃ in high altitudes due to increased solar
224 radiation. However, no strong correlation ($p>0.05$) is found between C₂/C₄ (and C₃/C₄) and
225 sampling altitude although solar radiation is expected to increase with increasing altitudes.

226 These results suggest that the degradation of higher homologous diacids (i.e., C₄) is not an
227 important pathway for the production of tropospheric C₂ and C₃ in China, and primary fossil-
228 fuel emissions and/or secondary production from other precursors are major formation
229 pathways of these small diacids (i.e., C₂, C₃ and C₄).

230 3.3 Cis/trans ratio

231 It has been revealed that maleic acid (M, *cis* configuration) is produced by
232 photochemical oxidation of anthropogenic aromatic hydrocarbons such as benzene and toluene,
233 which is predominant over fumaric acid (F, *trans* configuration) (Sempere and Kawamura,
234 1996). M can be photo-isomerized to its *trans* isomer (F) in the atmosphere under solar
235 radiation. M/F ratios (4.9, 4.5 and 6.8 for summer, spring and winter, respectively) are much
236 higher in the present study than those reported in marine region (0.1-1.5) (Fu et al., 2013) and
237 Chinese megacities at ground levels (2.0 and 2.2 for summer and winter, respectively) (Ho et
238 al., 2007). This indicates that only a small fraction of maleic acid is isomerized to fumaric acid
239 by photochemical transformation and thus SOA produced from anthropogenic emissions in the
240 lower/middle troposphere in China is mostly fresh without substantially photochemical
241 processing (Cong et al., 2015).

242 3.4 C₅ (C₆) to C₉ diacid ratio

243 C₆ (or C₅) to azelaic acid (C₉) ratio (i.e., C₆/C₉ or C₅/C₉) is often used as an indicator
244 of relative contribution from anthropogenic and biogenic sources to OA (Kawamura and Yasui,
245 2005). C₆/C₉ and C₅/C₉ ratios during the summer aircraft campaign are on average 17 and 28,
246 respectively, which are >15 times higher than those reported at the ground surface from major
247 Chinese cities (Ho et al., 2007; Wang et al., 2002). This comparison further supports that
248 anthropogenic sources are an important source of OA in the high altitudes over China. Taken
249 together with other possible SOA components (e.g., oligomers with MW > 250 Da) produced
250 during the oxidation of anthropogenic VOCs, our result implies that SOA formation plays an
251 important role in OA budget in the troposphere especially during summer when atmospheric

252 oxidation capacity is significantly enhanced (Shen and Wang, 2012; Su et al., 2012). It is
253 important to note that the correlation coefficient of C₆ (or C₅) with C₂ during summer ($r^2=0.39$
254 or 0.49) is lower than that obtained in winter ($r^2=0.76$ or 0.59) and spring ($r^2=0.95$ or 0.96),
255 suggesting that C₂ has a different formation pathway and/or its precursors are mostly from
256 biogenic origins (i.e., isoprene) in summer.

257 **3.5 Correlation of diacids with SOA tracers and biomass burning tracers**

258 In the urban atmosphere, dicarboxylic acids can be emitted as primary particles from
259 motor exhausts (Kawamura and Kaplan, 1987), biomass burning (Cong et al., 2015; Falkovich
260 et al., 2005), and cooking emissions (Rogge et al., 1991). However, recent field, laboratory and
261 model studies have demonstrated that aqueous-phase SOA formation from isoprene or other
262 precursors photooxidation is a major formation pathway of small diacids (Myriokefalitakis et
263 al., 2011; Carlton et al., 2006; Ervens et al., 2011). Furthermore, significant correlations ($r>0.70$,
264 $p<0.05$) are obtained between C₂ and its possible precursors and intermediates such as glyoxylic
265 acid (ω C₂), pyruvic acid (Pyr), glyoxal (Gly) and methylglyoxal (MeGly) in all three campaigns
266 (Table 2). This demonstrates that C₂ is produced from its precursor compounds such as Pyr,
267 Gly and MeGly through the following formation pathways: (CH₃COCOOH, HCO-CHO,
268 CH₃COCHO) → HCO-COOH → HOOC-COOH (Carlton et al., 2009; Carlton et al.,
269 2007; Carlton et al., 2006; Lim et al., 2013). We found that the correlation coefficient between
270 Gly and C₂ is higher than that between MeGly and C₂ for winter and spring samples, being
271 consistent with the fact that oxalic acid is the dominant product of glyoxal with the high yield
272 but not for methylglyoxal (Tan et al., 2009; Lim et al., 2010). Indeed, a good correlation (Figure
273 3) was found between predicted liquid water content with both OC and oxalic acid, indicating
274 an important contribution from SOA formation via cloud processing and/or aqueous-phase
275 oxidation. Aqueous-phase production of oxalic acid and related compounds may increase the
276 abundances of SOA at the middle troposphere (around 2 km in this study), as discussed below.
277 With the GEOS-Chem model based on the Fu et al. (2008) scheme (Fu et al., 2008), aqueous-
278 phase SOA has a pronounced enhancement in the lower free troposphere (2-6 km) (Heald et al.,

279 2011), which may explain to some extent the elevated levels of oxalic acid around at ~2 km in
280 altitude.

281 Similarly, C₂ also shows better correlations with both anthropogenic-derived SOA such
282 as C₅ and C₆ for winter ($r^2=0.76$ or 0.59) and spring ($r^2=0.95$ or 0.96) samples than summer
283 samples ($r^2=0.39$ or 0.49), further supporting that anthropogenic VOCs play a more important
284 role in SOA formation than biogenic VOCs during winter and spring. A recent study reveals
285 that air pollutants such as SO₂, NO₂, PM_{2.5} and PM₁₀ are higher in winter than in summer and
286 spring and these air pollutants are mostly from anthropogenic emissions (Zhang and Cao, 2015).
287 No significant correlation ($r^2=0.28$, $p>0.05$) is found between EC (i.e., a primary tracer for
288 fossil fuel biomass combustion (Zhang et al., 2015b)) and C₂ in summer aerosols, whereas a
289 good correlation is found in spring. These results indicate that primary emissions are not major
290 sources of C₂ during summer, but their contribution may be more important in spring.

291 Oxalic acid shows a strong positive correlation with isoprene-derived SOA tracers such
292 as 2-methylglyceric acid (2-MGA) and C₅-alkene triols (cis-2-methyl-1,3,4-trihydroxy-1-
293 butene,3-methyl-2,3,4-trihydroxy-1-butene plus trans-2-methyl-1,3,4-trihydroxy-1-butene)
294 (Figure 4), but a very weak correlation ($r^2 = 0.26$, $p>0.05$) with 2-methyltetrols (2-
295 methylthreitol and 2-methylerythritol). Previous studies have revealed that 2-methyltetrols
296 could be formed through epoxy derivatives of isoprene through acid-catalyzed hydrolysis
297 (Wang et al., 2005), whereas 2-MGA is produced by further gas-phase oxidation of its
298 intermediates such as methacrolein and methacrylic acid from isoprene (Claeys et al.,
299 2004;Surratt et al., 2006). Such a good correlation between 2-MGA and oxalic acid could
300 demonstrate that oxalic acid has a very close link with the higher-generation products of
301 isoprene in the presence of relatively high NO_x (i.e., NO_x averaged to 6.6 ± 4.0 ppbv in summer
302 and 3.9 ± 3.3 ppbv), which could serve as precursors of oxalic acid over China., which could
303 serve as precursors of oxalic acid over China. Oxalic acid also significantly correlates with α/β -
304 pinene-SOA tracers (i.e., pinonic, pinic, 3-hydroxyglutaric, and 3-methyl-1,2,3-

305 butanetricarboxylic acid) and β -caryophyllene tracer (β -caryophyllinic acid, see Figure 4)
306 (Jaoui et al., 2013).

307 Overall, both oxalic acid and SOA tracers are more abundant in summer than in spring,
308 suggesting that production of these organics is associated with higher oxidation capacity,
309 emission strength and solar radiation in summer. Based on the consistent good correlations of
310 oxalic acid with SOA tracers derived from isoprene, monoterpene and β -caryophyllene, we
311 propose that a large fraction of oxalic acid in the lower/middle troposphere over China is of
312 secondary origin, i.e., via aqueous chemistry. This study highlights that oxalic acid may serve
313 as an important tracer of SOA formation not only on the ground surface but also at high altitudes
314 within the lower/middle troposphere. Therefore, the high abundances of LWM diacids (C_2 - C_6)
315 observed in this study imply an important contribution to the OA budget from SOA production
316 from both biogenic and anthropogenic precursors emitted from the ground surface to high
317 altitudes over inland China.

318 The observed total concentrations of oxalic acid and other LWM diacids identified in
319 this study (Table 1) show the same magnitude as the levels of SOA (i.e., 299 ± 173 ng m^{-3} in
320 summer and 257 ± 210 ng m^{-3} in spring) estimated by SOA-tracer methods (Fu et al., 2014). It
321 should be noted that oxalic acid is not included in the traditional “SOA tracer method”
322 (Kleindienst et al., 2007) and therefore SOA may be underestimated if SOA is calculated by
323 this approach. Although oxalic acid has relatively high vapor pressure, it has been observed as
324 the most abundant individual organic compounds in aerosols (Kawamura and Bikkina, 2016).
325 Our study together with many previous studies have revealed that oxalic acid is mostly from
326 aqueous-phase production and is therefore of secondary origin (Myriokefalitakis et al., 2011).
327 In addition, oxalic acid may partly exist as hydrated and/or salt forms in aerosols. Therefore, it
328 is important to consider oxalic acid or oxalate as important SOA tracers.

329 Levoglucosan (1,6-Anhydro- β -D-glucopyranose) has been used as a source tracer for
330 biomass burning (Liu et al., 2013;Fu et al., 2012). Levoglucosan shows a significant correlation

331 with oxalic acid and also secondary organic aerosol (SOA) tracers (Figure 5), indicating that
332 that biomass burning is an important source of oxalic acid and SOA. It is interesting to note
333 that levoglucosan was more abundant in spring than in summer. In addition, regression slope
334 and correlation coefficient between oxalic acid and levoglucosan are higher in spring (i.e.,
335 slope: 2.7, r: 0.95) than in summer (i.e., slope: 1.7, r: 0.67), indicating that biomass-burning
336 emissions play more important roles in spring than summer. Such higher values of slope of
337 regression line and correlation coefficient were also found between levoglucosan and α/β -
338 pinene- and β -caryophyllene-oxidation products, emphasizing an importance of springtime
339 biomass burning. However, isoprene-oxidation products have a higher correlation coefficient
340 in summer than in spring, implying that biomass burning in summer is an important contributor
341 of isoprene-derived SOA.

342 **3.6 Vertical profiles of LWM-diacids**

343 As shown in Figure 6, the highest concentrations of oxalic acid and total diacids are
344 observed around at 2 km in altitude during summer with a sharp decrease toward 3 km.
345 However, during spring and winter, their concentrations decrease with increasing altitudes due
346 to the atmospheric dilution during upward transport. Much clear trends are observed in the
347 vertical profiles of oxalic acid normalized by anthropogenic tracers such as sulfate (SO_4^{2-}),
348 vanadium, and bulk OC. Interestingly, similar trends are also found in spring and winter
349 samples, but the trends are weaker, suggesting that a secondary production of oxalic acid is
350 largely enhanced during summer in the upper troposphere (~2 km). Similar vertical pattern has
351 been also reported for biogenic SOA-tracers such as 2-MGA, 3-HGA and MBTCA (3-methyl-
352 1,2,3-butanetricarboxylic acid), but not for POA such as biomass burning tracers (e.g.,
353 levoglucosan), fungal spore tracers (arabitol and mannitol), sucrose, and trehalose (Fu et al.,
354 2014). These results further demonstrate that oxalic acid is mostly likely produced by secondary
355 process in the troposphere.

356 Oxalic acid in the high altitudes could not be simply explained by uplifting transport
357 of pre-existing SOA produced on the ground surface because these SOA compounds relative
358 to anthropogenic tracers such as SO_4^{2-} , vanadium, and OC significantly increased with altitude
359 as stated above. This finding suggests that in-situ SOA production by the oxidation of VOCs
360 lifted from ground surface substantially contributes to the observed levels of oxalic acid and
361 related species. Therefore, SOA formation in cloud or wet aerosol via the oxidation of biogenic
362 and anthropogenic VOCs may increase concentrations of oxalic acid at higher altitudes in the
363 troposphere. There is growing evidence to support of in-cloud formation of oxalic acid and
364 related SOA. Many studies suggest that oxalic acid is mostly produced via aqueous-phase
365 oxidation of water-soluble organics such as glyoxal, methylglyoxal, pyruvic acid and glyoxylic
366 acid, which are oxidation intermediates of various VOCs (Ervens et al., 2004;Carlton et al.,
367 2006;Ervens et al., 2011).

368 **4 Conclusions**

369 Based on three aircraft measurements over East Asia, this study demonstrates an
370 aqueous-phase mechanism for SOA productions of diacids in the troposphere following
371 correlation analysis of oxalic acid in tropospheric aerosols with other measured chemical
372 variables including its precursors and its intermediate as well as biogenic-derived SOA from
373 isoprene, monoterpenes and β -caryophyllene. In addition to biogenic-derived SOA compounds,
374 anthropogenic-derived dicarboxylic acids (e.g. C5 and C6 diacids) are 4-20 times higher than
375 those from ground measurements and even occasionally dominant over oxalic acid at altitudes
376 higher than 2 km in summer, which is in contrast to the predominance of oxalic acid previously
377 reported globally including the tropospheric and surface aerosols. The results suggest an
378 important formation pathway associated with enhanced photochemical oxidation of
379 anthropogenic precursors in the polluted troposphere over China. Their relative contribution of
380 anthropogenic and biogenic sources is subject to future studies. The combination of radiocarbon
381 (^{14}C) measurement of water-soluble organic carbon (WSOC) and specific SOA compounds

382 (e.g., oxalic acid) may provide better insights on biogenic and fossil sources of SOA (Zhang et
383 al., 2014;Zhang et al., 2015a;Noziere et al., 2015) .

384 The present study demonstrates that secondary formation of oxalic acid in aqueous
385 phase plays an important role in the SOA budget from the near surface to the lower FT (i.e., 2
386 km) over inland China, whereas dilution of pre-existing particles and VOCs, photochemical
387 decomposition and aerosol processing may decrease the levels of oxalic acid and related
388 compounds at higher altitudes (>2 km). Our findings also highlight that water-soluble diacids
389 and other SOA components may have important impacts on the chemical compositions,
390 physical properties and budget of OA in the polluted troposphere over China, and thus
391 significantly affect the regional/global climate and intercontinental transport especially over
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633

634 **Table 1.** Average concentrations (ng m^{-3}) and concentration ranges (ng m^{-3}) for straight chain
 635 diacids ($\text{C}_2\text{--C}_{11}$), branched chain diacids ($\text{iC}_4\text{--iC}_6$), unsaturated diacids (M, F and mM),
 636 multifunctional diacids (hC_4 , kC_3 and kC_7), oxoacids ($\omega\text{C}_2\text{--}\omega\text{C}_4$, ωC_9 , and Pyr), and α -
 637 dicarbonyls (Gly and MeGly) of aerosol samples collected by aircraft campaigns over China.

	2003 Summer (inland China)		2004 Spring (inland China)		2002/2003 Winter (coastal China)	
	Range	Mean (SD) ^a	Range	Mean (SD)	Range	Mean (SD)
Oxalic, C_2	36.4-401	183 (111)	76.5-918	286 (216)	13.3-425	92.6 (94.5)
Malonic, C_3	6.3-131	54 (35)	12.2-216	57.5 (47.6)	1.4-79	15 (17.9)
Succinic, C_4	9.4-277	117 (68.2)	16-319	69.1 (69.9)	2.3-87.9	21.4 (19.8)
Glutaric, C_5	20.7-289	159 (79.1)	6.5-73.8	18.1 (16.3)	0.9-26.1	9.7 (7.3)
Adipic, C_6	52.3-135	93.9 (23.2)	5.7-69	19.4 (18.1)	3.7-34.5	13.4 (8.1)
Pimelic, C_7	0-2.5	0.8 (0.7)	n.d.-8.1	1.7 (2)	0.3-7.5	1.9 (1.7)
Suberic, C_8	n.d. ^b	n.d.	n.d.-2.7	0.2 (0.7)	n.d.-11	3.2 (2.9)
Azelaic, C_9	2-13.5	5.6 (3)	2.4-18.2	6.3 (4.3)	2.9-20.6	8.5 (4.8)
Sebacic, C_{10}	0.3-3.6	1.3 (1.0)	n.d.-8.4	3.6 (2.5)	0-6.9	1.1 (1.8)
Undecanedioic, C_{11}	0.8-4.3	2.1 (0.9)	n.d.-4.3	1.5 (1.3)	n.d.	n.d.
Methylmalonic, iC_4	1.2-5.9	3.7 (1.7)	1.1-12.5	4.9 (3)	0-3.9	0.9 (0.9)
Methylsuccinic, iC_5	0.6-10.8	4.4 (3.1)	1.3-27.5	5.9 (6.1)	0.7-23.2	5.9 (5.8)
Methylglutaric, iC_6	n.d.-1.3	0.4 (0.4)	0.4-5.9	1.2 (1.3)	n.d.-2.8	0.7 (0.8)
Maleic, M	1.8-12.5	6.5 (3.6)	3.3-22.3	9.4 (5)	1.6-11.1	5.7 (2.9)
Fumaric, F	0.1-3.9	1.7 (1.1)	0.5-8.4	3.0 (2.1)	0.1-6.2	1.5 (1.5)
Methylmaleic, mM	2.3-15.1	6.3 (3.6)	2.2-18.0	7.4 (3.8)	1.3-8.2	4 (2)
Hydroxysuccinic, hC_4	1.7-12.5	5.3 (3.4)	n.d.-9.3	1.9 (2.2)	n.d.-13.7	1.9 (3.2)
Ketomalonic, kC_3	0.4-9.2	4.2 (2.5)	n.d.-22.8	5.6 (5.3)	n.d.-26	5.1 (5.6)
Ketopimelic, kC_7	0.4-8.2	3.0 (2.3)	n.d.-18.7	4.0 (4.5)	n.d.-3.9	0.6 (0.9)
Total diacids	139-1230	653 (290)	148-1780	507 (402)	40.4-757	193 (164)
n.s. $\text{C}_2\text{--C}_{11}$	128-1160	615(272)	128-1630	464 (371)	31.3-678	167 (149)
Pyruvic acid, Pyr	n.d.-9.6	2.9 (3.0)	0.1-11.4	2.1 (2.8)	0.7-36.5	10 (8.6)
Glyoxylic, ωC_2	8.1-89.6	37.7 (25.2)	8.3-146	46.0 (38.0)	6.7-129	30.6 (28.9)
3-oxopropanoic, ωC_3	0.1-9.7	3.3 (2.4)	0.1-1.1	0.5 (0.3)	n.d.-1.9	0.5 (0.5)
4-oxobutanoic, ωC_4	0-23.1	8.0 (7.0)	6.8-38.9	14.9 (8.5)	0.6-35.5	7.5 (8.6)
9-oxononoic, ωC_9	3.4-36.2	11.6 (8.9)	0.3-20.5	5.8 (5.2)	0.2-5.5	1.8 (1.4)
Total Keto acid	18.7-131	63.5 (36.0)	23.7-178	69.3 (48.5)	12.2-176	50.0 (36.9)
Glyoxal, Gly	0.7-14.8	4.0 (3.6)	0.2-9.5	2.3 (2.5)	0.6-23	4.3 (5.1)
Methylglyoxal, MeGly	0.6-28.2	10.8 (7.7)	0.8-27.3	7.4 (8.3)	2.5-24.3	7.6 (5.6)
Total dicabonyls	1.3-42.9	14.8 (11.1)	1.7-36.8	9.8 (10.7)	3.1-47.3	11.8 (10.5)
Total	170-1390	731 (329)	174-1990	586 (457)	68.5-980	255 (209)
Gly/MeGly	0.1-1.2	0.4 (0.2)	0.04-1.0	0.4 (0.3)	0.2-0.9	0.5 (0.2)
M/F	1.7-13.9	4.9 (2.9)	1.1-13.8	4.5 (3.1)	1.6-27.4	6.8 (6.2)
$\omega\text{C}_2/\text{C}_2$	0.1-0.2	0.2 (0.03)	0.1-0.3	0.2 (0.04)	0.2-0.8	0.4 (0.1)
C_2/C_4	0.6-6.7	2.0 (1.6)	2.9-5.9	4.6 (0.9)	2.8-13.5	4.6 (2.3)
C_3/C_4	0.2-1.9	0.5 (0.42)	0.7-1.1	0.9 (0.1)	0.4-1.7	0.7 (0.3)
C_5/C_9	5.2-64.6	31.3 (15.6)	1.6-4.3	2.8 (0.7)	0.1-2.7	1.1 (0.6)
C_6/C_9	10.0-41.0	19.8 (7.21)	1.3-4.6	2.8 (0.9)	0.5-2.8	1.6 (0.6)

638 ^aSD denotes standard deviation (1σ); ^bn.d. denotes not detected.

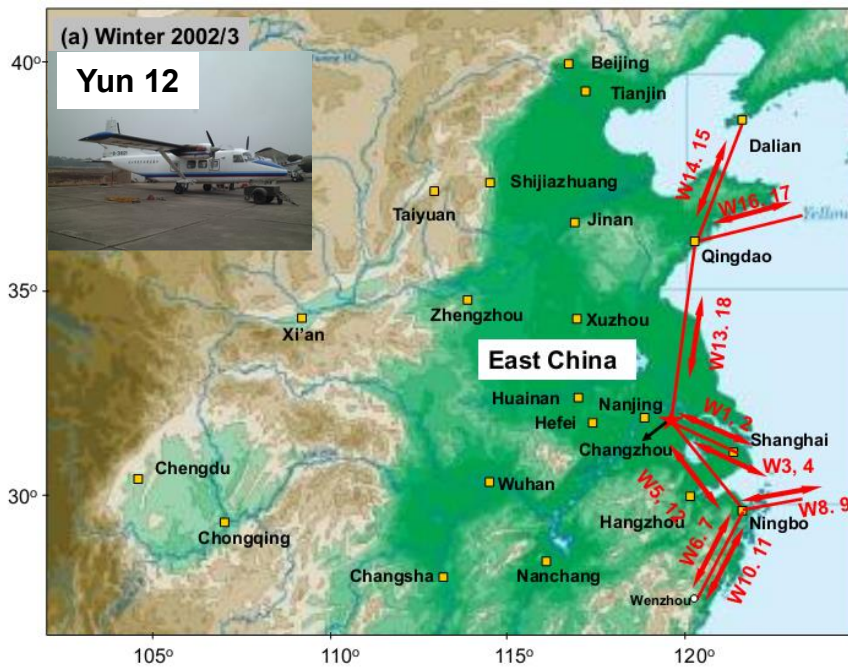
639 **Table 2.** Correlation coefficients (r^2) among oxalic acid (C_2), pyruvic acid (Pyr), glyoxylic acid
 640 (ωC_2), glyoxal (Gly) and methylglyoxal (MeGly) detected in aerosol samples from aircraft
 641 campaigns during summer 2003 (n=14), spring 2004 (n=16) and winter 2002/2003 (n=18) over
 642 China. See Table 1 for abbreviations.

Summer 2003					
	C_2	Pyr	ωC_2	Gly	MeGly
C_2	1.00				
Pyr	0.89	1.00			
ωC_2	0.97	0.95	1.00		
Gly	0.75	0.86	0.85	1.00	
MeGly	0.91	0.97	0.97	0.90	1.00

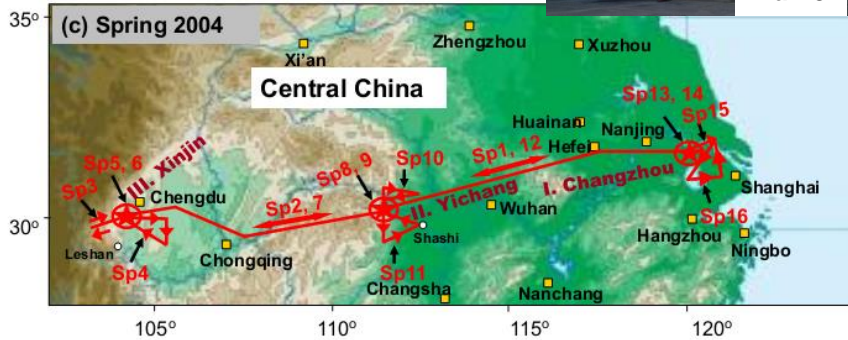
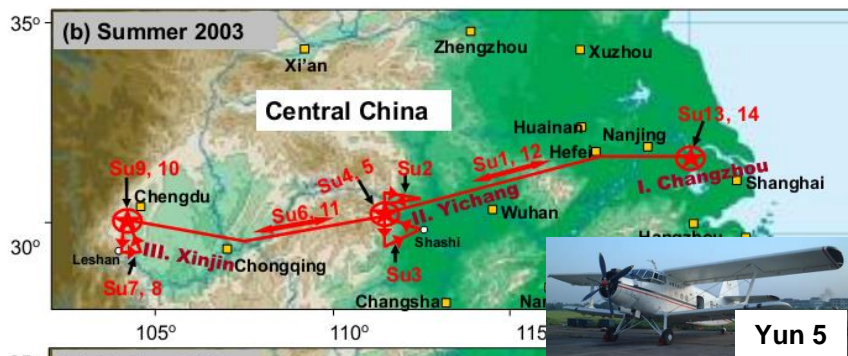
Spring 2004					
	C_2	Pyr	ωC_2	Gly	MeGly
C_2	1.00				
Pyr	0.95	1.00			
ωC_2	0.97	0.93	1.00		
Gly	0.96	0.95	0.97	1.00	
MeGly	0.93	0.92	0.94	0.93	1.00

Winter 2002/2003					
	C_2	Pyr	ωC_2	Gly	MeGly
C_2	1.00				
Pyr	0.70	1.00			
ωC_2	0.98	0.70	1.00		
Gly	0.92	0.69	0.90	1.00	
MeGly	0.85	0.63	0.83	0.94	1.00

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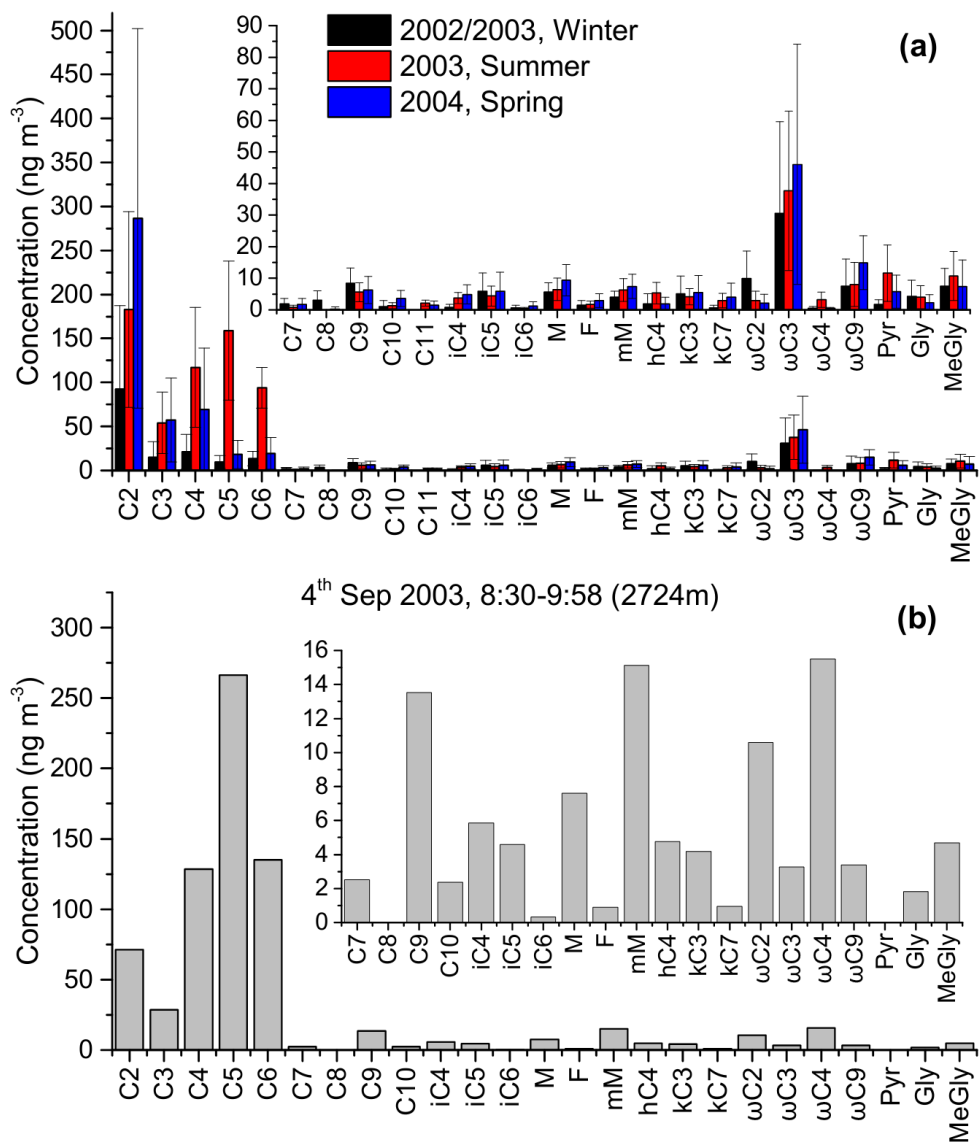
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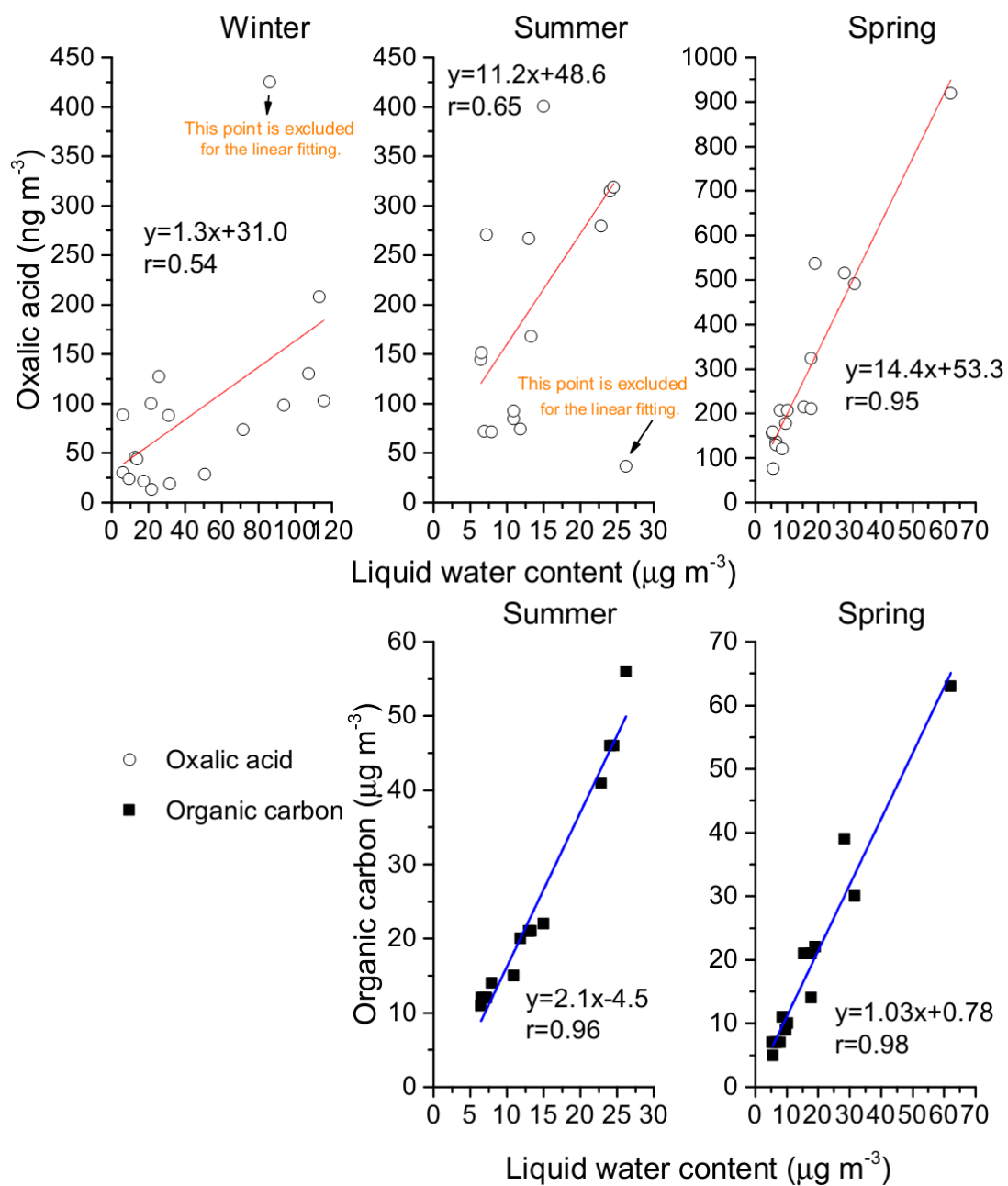
646 **Figure 1.** Tracks of research flights during aircraft measurements over China during (a) winter
 647 2002, (b) summer 2003 and (c) spring 2004. The detailed sampling information with flight No.
 648 is listed in the Supplementary Table S1. The maps with the flight tracks were drawn by the
 649 software of PowerPoint 2010, <https://products.office.com/>.

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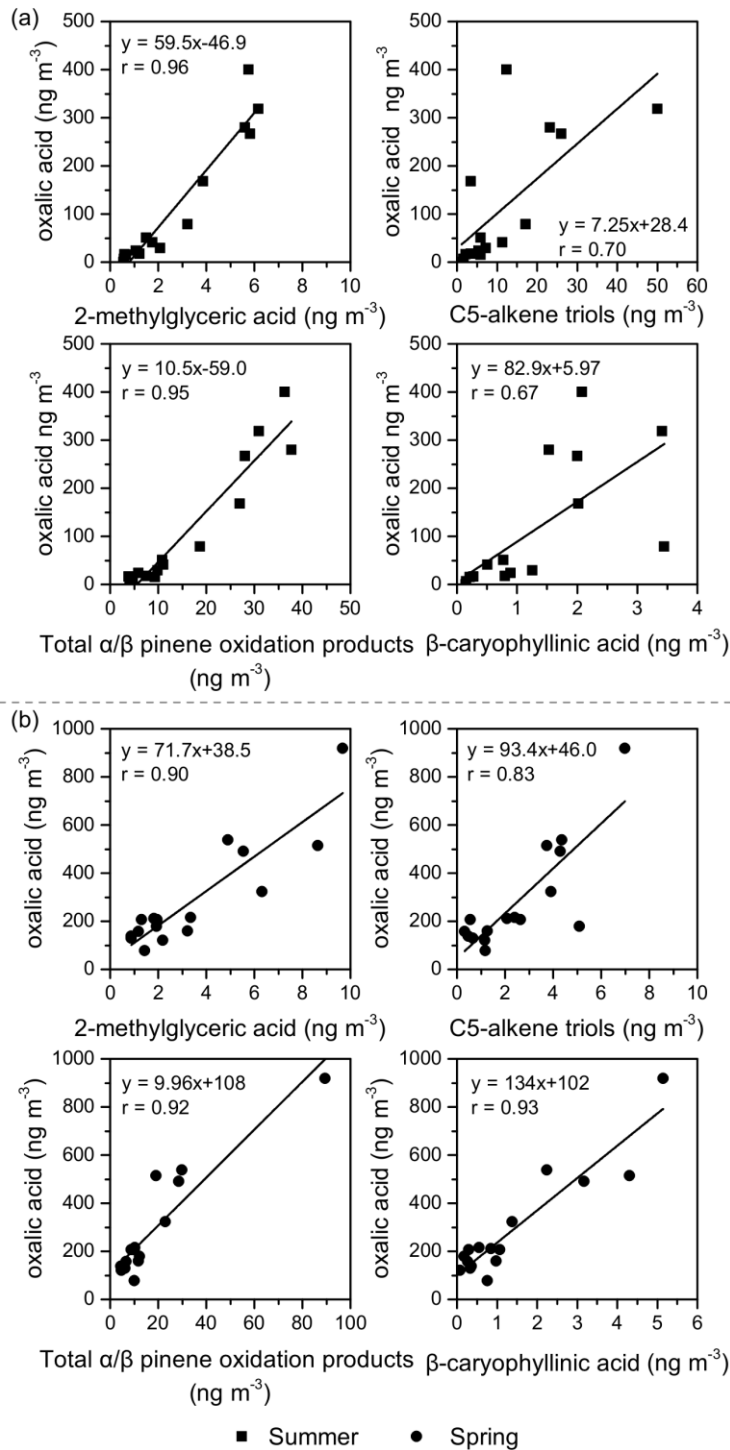
652 **Figure 2.** Molecular distributions of straight chain diacids (C₂–C₁₁), branched chain diacids
 653 (iC₄–iC₆), unsaturated diacids (M, F and mM), multifunctional diacids (hC₄, kC₃ and kC₇),
 654 ketoacids (ωC₂–ωC₄, ωC₉, and Pyr), and α-dicarbonyls (Gly and MeGly) in aircraft
 655 measurement over China; (a) averaged concentrations with bars of standard deviation during
 656 winter 2002/2003, summer 2003 and spring 2004; (b) molecular distributions of the measured
 657 compounds in the sample collected on 4th Sep 2003. See Table 1 for abbreviations.



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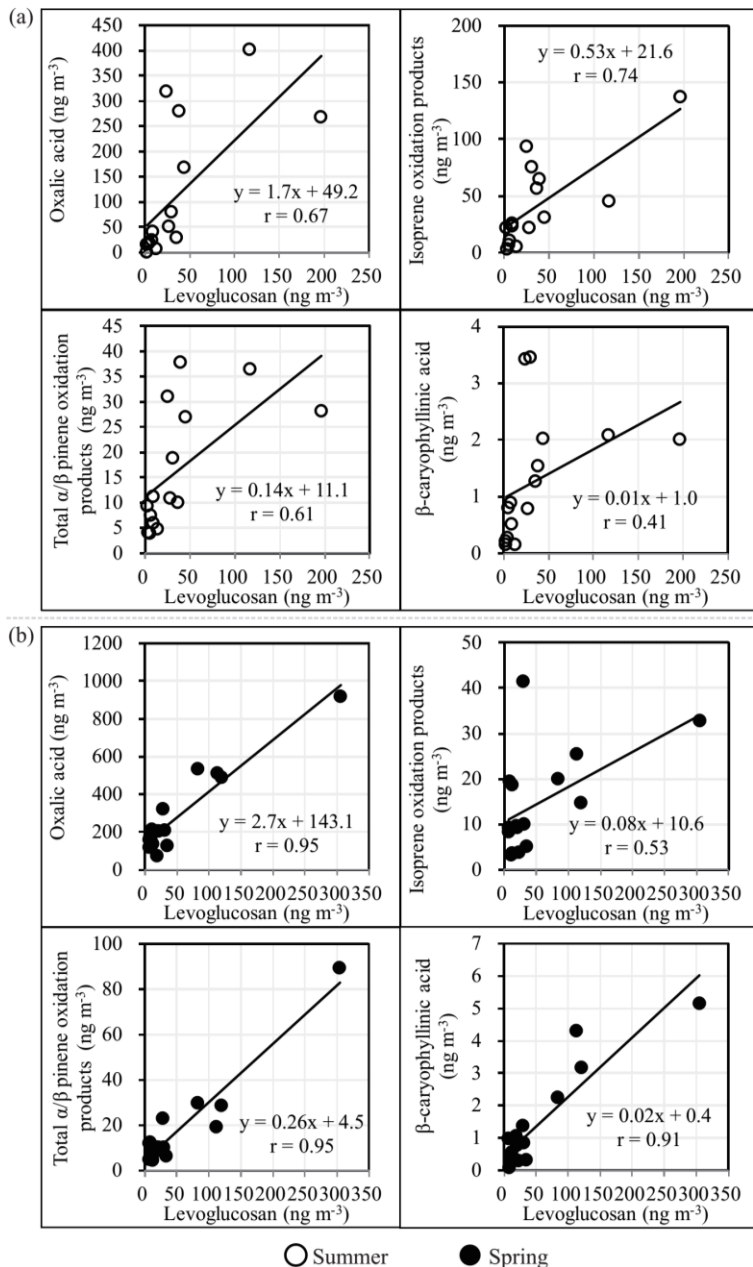
659 **Figure 3.** The linear fitting of the predicted liquid water content with organic carbon (OC) and
 660 oxalic acid aerosols during winter, summer, and spring aircraft measurements over China.

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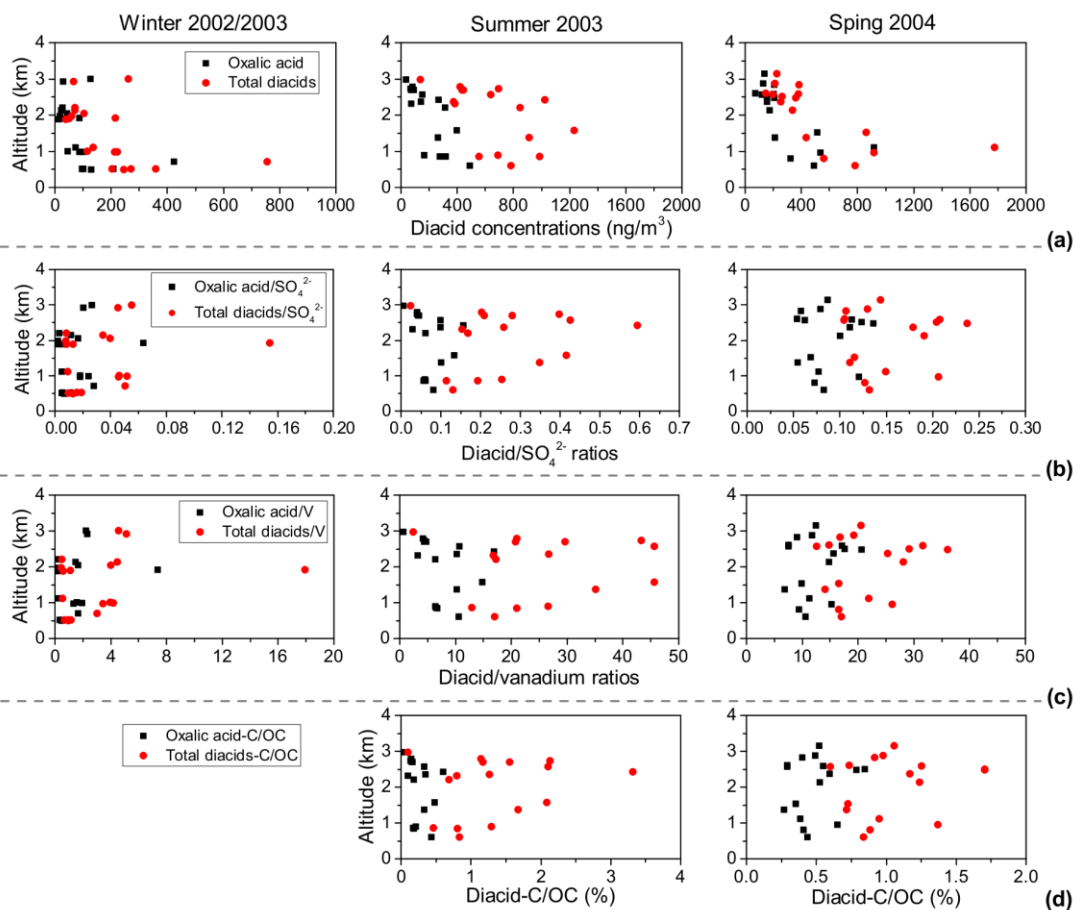
663 **Figure 4.** Linear relationship of the concentrations of oxalic acid with the detected tracer
 664 compounds for secondary organic aerosols (SOA) such as isoprene SOA tracers such as 2-
 665 methylglyceric acid and C5-alkene triols (sub-total of cis-2-methyl-1,3,4-trihydroxy-1-butene,
 666 3-methyl-2,3,4-trihydroxy-1-butene, trans-2-methyl-1,3,4-trihydroxy-1-butene), α/β -pinene
 667 SOA tracers (subtotal of 3-hydroxyglutaric acid, pinonic acid, pinic acid, 3-methyl-1,2,3-
 668 butanetricarboxylic acid), and β -caryophyllene SOA tracer (i.e., β -caryophyllinic acid) from
 669 aircraft measurements over China during (a) summer and (b) spring.



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671 **Figure 5.** Linear relationship for the concentrations of levoglucosan with oxalic acid, isoprene
 672 SOA tracers, α/β -pinene SOA tracers and β -caryophyllene SOA tracer from the aircraft
 673 measurements over China during (a) summer and (b) spring.

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676 **Figure 6.** Vertical profiles of (a) concentrations of oxalic acid and total diacids, and their
 677 relative abundance to (b) sulfate (SO_4^{2-}), (c) vanadium (V), and (d) organic carbon (OC, %) in
 678 aerosol samples collected during winter 2002 (left), summer 2003 (middle) and spring 2004
 679 (right) aircraft campaigns over China.