



## 1                    **Constraining the ship contribution to the aerosol of the Central** 2                    **Mediterranean**

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## 20 21 **Abstract**

22  
 23 PM<sub>10</sub> aerosol samples were collected during summer 2013 within the framework of the  
 24 Chemistry and Aerosol Mediterranean Experiment (ChArMEx) at two sites located North  
 25 (Capo Granitola, 36.6°N, 12.6°E) and South (Lampedusa Island, 35.5°N, 12.6°E),  
 26 respectively, of the main Mediterranean shipping route in the Sicily Channel.

27 The PM<sub>10</sub> samples were collected with 12 hour time resolution at both sites. Selected  
 28 metals, main anions, cations, and elemental and organic carbon were determined.

29 The evolution of soluble V and Ni concentrations (typical markers of heavy fuel oil  
 30 combustion) was related to meteorology and ship traffic intensity in the Sicily Channel,  
 31 using a high resolution regional model for back trajectories calculation. Elevated  
 32 concentration of V and Ni were associated with transport from the Sicily Channel and  
 33 coincidences between trajectories and positions of large ships, both at Capo Granitola  
 34 and Lampedusa; the vertical structure of the planetary boundary layer also appears to  
 35 play a role, with high V values associated with strong inversions and stable boundary  
 36 layer. The V concentration was generally lower at Lampedusa than at Capo Granitola,  
 37 where it reached a peak value of 40 ng/m<sup>3</sup>.

38 Concentrations of rare earth elements, La and Ce in particular, were used to identify  
 39 possible contributions from refineries, whose emissions are also characterized by  
 40 elevated V and Ni amounts; refinery emissions are expected to display high La/Ce and  
 41 La/V ratios, due to the use of La in the fluid catalytic converter systems. In general, low  
 42 La/Ce and La/V ratios were observed in the PM samples, allowing to unambiguously  
 43 identify the large role of the ship source in the Sicily Channel.



Based on the sampled aerosols, ratios of the main aerosol species arising from ship emission with respect to V were estimated with the aim of deriving a lower limit for the total ship contribution to PM<sub>10</sub>. The estimated minimum ship emission contributions to PM<sub>10</sub> was 1.9 µg/m<sup>3</sup> at Lampedusa, and 2.8 µg/m<sup>3</sup> at Capo Granitola, corresponding to 11% and 8.2% of PM<sub>10</sub>, respectively.

## 1. Introduction

Ship emissions may significantly affect atmospheric concentrations of several important pollutants, especially in maritime and coastal areas (e.g. Endresen et al., 2003). Main emitted compounds are carbon dioxide (CO<sub>2</sub>), nitrogen oxides (NO<sub>x</sub>), sulfur dioxide (SO<sub>2</sub>), carbon monoxide (CO), hydrocarbons, and primary and secondary particles. Thus, ship emissions impact the greenhouse gas budget (Stern, 2007), acid rain (through NO<sub>x</sub> and SO<sub>2</sub> oxidation products; Derwent et al., 2005), human health (CO, hydrocarbons, particles; Lloyd's Register Engineering Services, 1995; Corbett et al., 2007) and solar radiation budget through aerosol direct and indirect effects (black carbon and sulfur containing particles; Devasthale et al., 2006; Lauer et al., 2007; Coakley and Walsh, 2002).

Heavy oil fuels used by ships contain varying transition metals originating from the fuel. The aerosol emitted by ship engines is formed at high temperature (>800°C) from V, Ni, Fe compounds (Sippula et al., 2009). The thermodynamics predict that these species mainly form oxides, but when the flue gas dew point is reached, sulfuric acid (which was found to form a liquid layer on the ultra-fine particles) condenses on it leading to partial dissolution of the ultra-fine seeds, probably increasing the toxicity of the particles when inhaled.

In spite of the great amount of gas and particulate arising from ship emission, maritime transport is relatively clean if calculated per kilogram of transported material, and it is currently increasing with respect to air and road transport (Micco and Pérez, 2001; Grewal and Haugstetter, 2007). In addition, emissions from other transport sectors are decreasing due to the implementation of advanced emission reduction technologies, and the relative impact of shipping emissions is increasing.

Regulations aiming at reducing emissions based on restrictions on the fuel sulfur content (sulfur emission control areas, SECAs) have been implemented in several regions. Although the legislation is focussed on sulfur emissions, the overall health and environmental effects of the emissions depend in a complicated manner on the physical and chemical properties of the emissions (WHO, 2013). Several studies have been carried out to determine the detailed chemical composition of shipping emissions (Agrawal et al., 2008a and b, Moldanová et al., 2009, Murphy et al., 2009, Lyyränen et al. 1999, Cooper, 2003, Sippula et al. 2014); however, in comparison to on-road vehicles, the ships emissions are still poorly characterized.

A large variety of anthropic sources (refineries, power plants, intense ship traffic), also associated with a high population density, and natural emissions make the



87 Mediterranean region one of the most polluted in the world (e.g., Kouvarakis et al.,  
88 2000; Marmer and Langmann, 2005). This multiplicity of Mediterranean sources (some  
89 of which with the same markers of ship aerosol) makes difficult the quantification of  
90 ship contribution to the total aerosol amount (e.g., Becagli et al., 2012).

91 The contribution of ships and harbour emissions to local air quality, with specific focus  
92 on atmospheric aerosol, has been investigated using models (Trozzi et al., 1995;  
93 Gariazzo et al., 2007; Eyring et al., 2005; Marmer et al., 2009), experimental analyses  
94 at high temporal resolution (Ault et al., 2010; Contini et al., 2011; Jonsson et al., 2011;  
95 Diesch et al., 2013; Donato et al., 2014), receptor models based on identification of  
96 chemical tracers associated with ship emissions (Viana et al., 2009; Pandolfi et al.,  
97 2011; Cesari et al., 2014), and integrated approaches with receptor and chemical  
98 transport models (Bove et al., 2014). Few studies exist in open sea (Becagli et al.,  
99 2012; Schembari et al., 2014; Bove et al., 2016).

100 In this context, studies performed at Mediterranean sites, where it is possible to  
101 distinguish ship emission from other sources of heavy fuel oil combustion, are important  
102 to investigate the current impact of the ship emissions on primary and secondary  
103 aerosols. In a previous study (Becagli et al., 2012) we used measurements of  $PM_{10}$  and  
104 relative chemical composition carried out at Lampedusa, in the central Mediterranean,  
105 to investigate the role of ship emissions. Vanadium and Nickel were used as tracers of  
106 heavy fuel combustion together with trajectory analyses to assess the role of ship  
107 traffic. The ship source, however, could not be unequivocally separated from possible  
108 influences from refineries and power plants, which use similar fuels. In summer 2013  
109 we addressed the same topic by implementing a specific strategy to target the aerosols  
110 due to ship emissions.  $PM_{10}$  samples were collected in parallel at Lampedusa (LMP) and  
111 at Capo Granitola (CGR), i.e., respectively South and North of the main shipping route  
112 through the Mediterranean, with the aim of isolating the ship source. Figure 1 shows  
113 the map of the measurement stations in the central Mediterranean; Capo Granitola is  
114 about 230 km North of Lampedusa. The analysis is complemented with measurements  
115 of Rare Earth Elements (REEs), trajectories from a high resolution regional model, and  
116 actual observations of ship traffic. The combination of these approaches allows  
117 unambiguously identifying and providing constraints for the ship contribution to  $PM_{10}$  in  
118 the central Mediterranean.

119 The  $PM_{10}$  samples were collected in summer 2013 as a contribution to the Chemistry  
120 and Aerosol Mediterranean Experiment (ChArMEx; <http://charmex.lsce.ispl.fr>).  
121 Lampedusa is one of the supersites of the ChArMEx experiment; a list of the  
122 instruments deployed during the special observing period 1a of ChArMEx, and of the  
123 measurement strategy, meteorological conditions, and main observations is given by  
124 Mallet et al. (2016).

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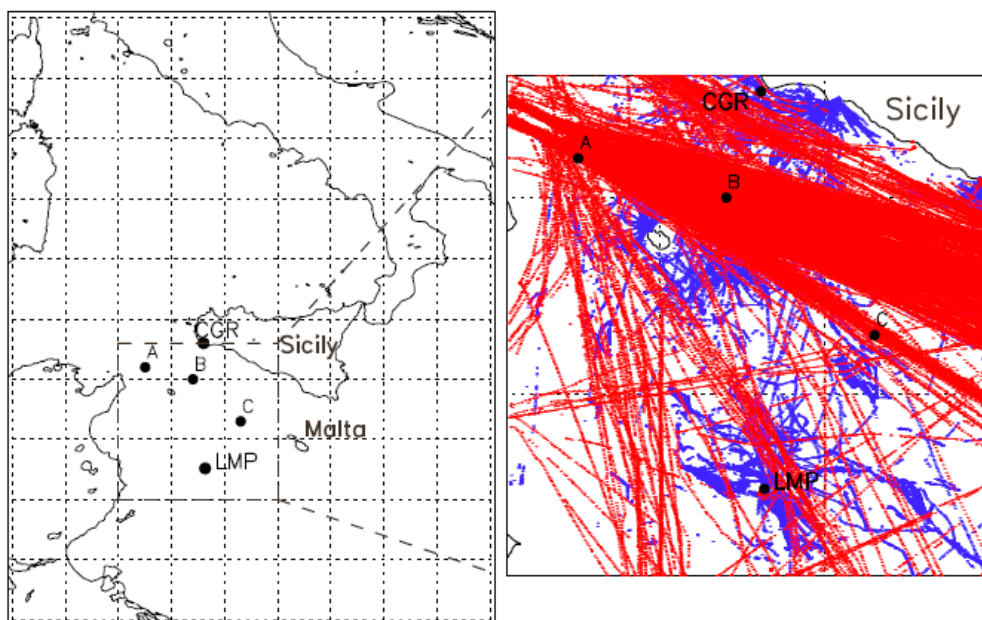


Figure 1. Map of the study area with the sites of Lampedusa (LMP) and Capo Granitola (CGR) (left panel). A, B and C indicate the three sites selected to study the stability of the boundary layer in the Sicily Channel (see section 3.2.2). The ship routes in the study area during the first 10 days of June 2013 are displayed in the right panel. Red and blue dots show the routes of merchant and fishing vessels, respectively.

## 2. Measurements and methods

### 2.1. Aerosol sampling and chemical analyses

PM<sub>10</sub> was sampled at two sites: at the Station for Climate Observations, maintained by ENEA (the Italian Agency for New Technologies, Energy, and Sustainable Economic Development) on the island of Lampedusa (35.5°N, 12.6°E), and at the Italian CNR (National Research Council) Research Centre at Capo Granitola (36.6°N, 12.6° E ). Lampedusa is a small island in the Central Mediterranean sea, more than 100 km far from the nearest Tunisian coast. At the Station for Climate Observations, which is located on a 45 m a.s.l. plateau on the North-Eastern coast of Lampedusa, continuous observations of greenhouse gases concentration (Artuso et al., 2007, 2009), aerosol properties (di Sarra et al., 2011, 2015; Becagli et al., 2013; Marconi et al., 2014; Calzolari et al., 2015), total ozone, ultraviolet irradiance (Meloni et al., 2005), solar and infrared radiation (di Sarra et al., 2011; Meloni et al., 2012, 2015), and other climatic parameters are carried out.



154 PM<sub>10</sub> is routinely sampled on a daily basis at LMP (Becagli et al., 2013; Marconi et al.,  
155 2014; Calzolari et al., 2015). For the intensive ChArMEx campaign, samples were  
156 collected from 1 June to 3 August at 12-hour resolution by using a low volume dual  
157 channel sequential sampler (HYDRA FAI Instruments) equipped with two PM<sub>10</sub> sampling  
158 heads operating in accord with UNI EN12341.

159 The two channels operated in parallel and were loaded with different types of filters:  
160 the first one with 47 mm diameter, 2 µm nominal porosity Teflon filters, for ion  
161 chromatographic analysis of soluble ions, atomic emission spectroscopy for soluble  
162 metals, and proton-induced X-ray emission (PIXE) for the total (soluble+insoluble)  
163 elemental composition; the second one with 47 mm pre-fired, 2 µm nominal porosity,  
164 quartz filters for elemental (EC) and organic carbon (OC) determinations.

165 The sampling site at CGR is located at Torretta Granitola, a Research Center of the  
166 Italian National Research Council, in South-Western Sicily (12 km from Mazara del  
167 Vallo). The sampler was installed on the roof of one of the research centre buildings at  
168 about 20 m a.s.l., directly on the coastline, facing the strait of Sicily.

169 At CGR PM<sub>10</sub> samples were collected at 12 hour resolution with a TECORA Skypost  
170 sequential sampler on 47 mm pre-fired, 2 µm nominal porosity, quartz filters allowing  
171 the determination of ions, metals, EC and OC on different fractions of the filter. Due to  
172 technical problems, some diurnal samplings were lost at CGR.

173

174 The PM<sub>10</sub> mass was determined by weighting the filters before and after sampling with  
175 an analytical balance in controlled conditions of temperature (20±1 °C) and relative  
176 humidity (50±5 %). The estimated error on PM<sub>10</sub> mass is around 1% at 30 µg/m<sup>3</sup> in the  
177 applied sampling conditions.

178 A quarter of each Teflon filter from LMP and a 1.5x1 cm punch of the quartz filter from  
179 CGR were analysed by Ion Chromatography (IC) in the analytical conditions described  
180 in Marconi et al. (2014). The estimated uncertainty for IC measurements is 5% for all  
181 the considered ions.

182 Blank values were negligible with respect to the concentration in the samples for Teflon  
183 filters. Blank values for quartz filters were negligible for most of the analyzed species,  
184 and when not negligible, anyway lower than 25<sup>th</sup> percentile, they were subtracted from  
185 the measured concentrations.

186 Another quarter of the Teflon filter from LMP, and another 1.5x1 cm punch of the  
187 quartz filter from CGR were extracted in ultrasonic bath for 15 min with MilliQ water  
188 acidified at pH 1.5–2 with ultrapure HNO<sub>3</sub> obtained by sub-boiling distillation. This  
189 extract was used for the determination of the metals soluble part by means of an  
190 Inductively Coupled Plasma Atomic Emission Spectrometer (ICP-AES, Varian 720-ES)  
191 equipped with an ultrasonic nebulizer (U5000 AT+, Cetac Technologies Inc.). The  
192 chosen value of pH is the lowest found in rainwater (Li and Aneja, 1992) and leads to  
193 the determination of the metals fraction available to biological organisms and, for some  
194 metals (e.g. V and Ni), related to the anthropic source (Becagli et al., 2012).

195 The remaining half Teflon filter from Lampedusa was analysed by proton induced X ray  
196 emission (PIXE) technique (Lucarelli et al., 2011). PIXE analysis is a non-destructive





method for metals. Thus, after the PIXE analysis, this part of the Teflon filter and another punch of the quartz filter from CGR were used for the determination of metals by ICP-AES through the solubilisation procedure according with the method reported in the EU EN14902 (2005) rule, by using concentrated sub-boiling distilled  $\text{HNO}_3$  and 30% ultrapure  $\text{H}_2\text{O}_2$  in a microwave oven at  $220^\circ\text{C}$  for 25 min ( $P = 55$  bar). Although this solubilisation procedure is not able to completely dissolve the silicate species, is able to recover at least 70% of the same elements measured by PIXE also for elements having dominant crustal source (unpublished data) due to the low crustal aerosol load in these sampling period (e.g., Mailler et al., 2016).

The OC and EC measurements were carried out on a  $1.5 \times 1$  cm punch of the quartz filters from Lampedusa and Capo Granitola by means of a Sunset thermo-optical transmittance analyser, following the NIOSH protocol (Wu et al, 2016).

### 2.3. Atmospheric model and trajectory calculations

Numerical simulations with a non-hydrostatic mesoscale atmospheric model were used to characterize the meteorological conditions in the Sicily Channel during the campaign and to support the interpretation of the experimental results. The Weather Research and Forecasting (WRF) model (Skamarock et al., 2008) outputs, provided by the Department of Physics of the University of Genoa, Italy, were used, covering the entire Mediterranean with a grid spacing of 10 km and hourly temporal resolution. Initial and boundary conditions to drive WRF simulations are obtained from the Global Forecast System operational global model (Environmental Modeling Center, 2003) outputs ( $0.5 \times 0.5$  square degree). Some recent applications of the modelling chain are described in Mentaschi et al. (2015) and Cassola et al. (2016), where full details on the model configuration can also be found.

In particular, the WRF 3-D hourly meteorological fields were used to perform a backward trajectory analysis with the NOAA HYbrid Single-Particle Lagrangian Integrated Trajectory Model (HYSPLIT; Stein et al. 2015), aimed at assessing the origin of air masses impacting the monitoring sites and at supporting the source attribution suggested by the analysis of specific markers (see, in particular, Section 3.2.2). The use of a high-resolution regional atmospheric model for trajectory calculations allows a better representation of boundary layer properties and mesoscale phenomena such as land/sea breezes, which can have a relevant impact especially in complex topography coastal sites like CGR.

Specifically, 48-h long back trajectories were computed at each site from a reference height of 10 m above ground level, starting every six hours for the whole period of the campaign, from 10<sup>th</sup> June to 31<sup>st</sup> July 2013.

### 2.4. Ships/marine traffic



The position and the main characteristics of the ships travelling in the central Mediterranean were derived from the MarineTraffic database (<http://www.marinetraffic.com/>), which provides the position of the ships with a high temporal resolution (about 3-5 minutes) by means of the Automatic Identification System (AIS).

Three classes of ships defined by the AIS classification were considered: all the ships, the merchant (i.e. cargo and tanker), and the fishing vessels. The merchant and fishing vessels are the most frequent ships in the Sicily Channel and the former are expected to produce the highest impact on the V concentration due to their higher emissions ([http://ec.europa.eu/environment/archives/air/pdf/chapter2\\_ship\\_emissions.pdf](http://ec.europa.eu/environment/archives/air/pdf/chapter2_ship_emissions.pdf)).

### 3. Results

#### 3.1. PM<sub>10</sub> chemical composition at the two sites

The sea salt aerosol (SSA) component of PM<sub>10</sub> was estimated as the sum of the sea salt (ss) fractions of Na<sup>+</sup>, Mg<sup>2+</sup>, Ca<sup>2+</sup>, K<sup>+</sup>, sulfate and chloride. Details on the calculation of sea salt Na<sup>+</sup> and Ca<sup>2+</sup>, and non-sea salt (nss) fractions are reported in Marconi et al. (2014). The sea salt fractions of Mg<sup>2+</sup>, Ca<sup>2+</sup>, K<sup>+</sup>, and sulphate were calculated from sea salt Na<sup>+</sup> (ssNa<sup>+</sup>) by using the ratio of each component to Na<sup>+</sup> in bulk sea water: Mg<sup>2+</sup>/Na<sup>+</sup> = 0.129, Ca<sup>2+</sup>/Na<sup>+</sup> = 0.038, K<sup>+</sup>/Na<sup>+</sup> = 0.036, SO<sub>4</sub><sup>2-</sup>/Na<sup>+</sup> = 0.253 (Bowen, 1979). Chloride undergoes depletion processes during aging of sea spray, mainly due to exchange reactions with anthropic H<sub>2</sub>SO<sub>4</sub> and HNO<sub>3</sub>, leading to re-emission of HCl in the atmosphere. Thus, for chloride we use the measured chloride concentration instead of the one calculated from ssNa<sup>+</sup>. Thus,

$$\text{SSA} = 1.46 * [\text{ssNa}^+] + [\text{Cl}^-]$$

The crustal component is calculated from Al, which represents 8.2% of the upper continental crust, UCC (Henderson and Henderson 2009). A previous study using an extensive data set at Lampedusa showed that the crustal content determined from the total Al was in very good agreement with calculations made from the sum of the metal oxides (Marconi et al., 2014). However, in this study we use measurements of the soluble Al concentration obtained by ICP-AES on the solution obtained with H<sub>2</sub>O<sub>2</sub> and HNO<sub>3</sub> in microwave oven, instead of the total Al content. Therefore, in this work we underestimate the crustal contribution by about 30% (unpublished results). However, it must be emphasized that the crustal aerosol contribution was very low throughout the measurement campaign.

Figure 2 shows the time series of the main PM<sub>10</sub> components at LMP and CGR. It must be noticed that an intense Mistral event occurred from 22<sup>nd</sup> June to 1<sup>st</sup> July. Mistral



events are characterized by strong winds from the north-westerly sector, and often by subsiding air masses originating from the free troposphere. Thus, elevated values of SSA and low concentrations of other compounds are generally found during Mistral. Average concentrations of  $PM_{10}$  and of the different aerosol components for the whole measurement campaign and for the non-Mistral conditions are reported in Table 1. The averages were calculated over a homogeneous dataset, i.e., only for the 12-hour intervals with observations at both sites.

The largest  $PM_{10}$  values were linked to elevated SSA during the Mistral event at both sites.  $PM_{10}$  is about two times larger at Capo Granitola than at Lampedusa. The  $PM_{10}$  measured during the campaign at Lampedusa was significantly smaller than its long-term average ( $31.5 \mu\text{g}/\text{m}^3$ ; Marconi et al., 2014). No Saharan dust transport events occurred at low altitude in this period (e.g., Mailler et al., 2015), and the crustal aerosol contribution remained very low and almost constant at both sites (average  $< 1 \mu\text{g}/\text{m}^3$  at LMP and around  $3 \mu\text{g}/\text{m}^3$  at CGR).

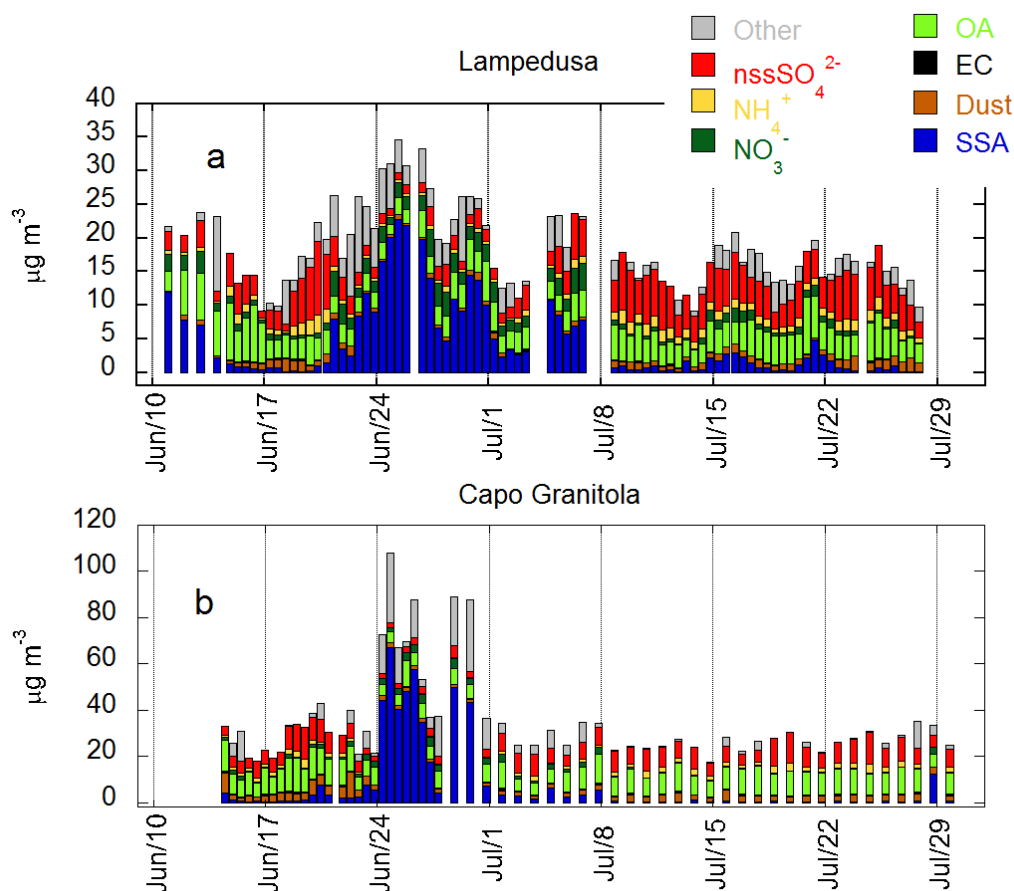






Figure 2. Time series of the main aerosol components at LMP (plot a) and CGR (plot b). Note the different vertical scales of the graphs. For calculation of Organic Aerosol (OA) see text.

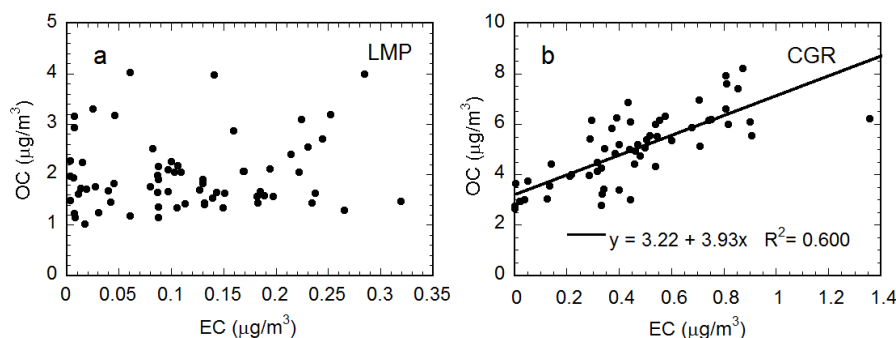


Figure 3. Scatter plot of OC vs. EC at LMP (plot a) and CGR (plot b). Note the different scales of the graphs.

SSA accounted for about 26% and 24% of  $PM_{10}$  at LMP and CGR, respectively. During the periods non influenced by the Mistral the SSA contribution was about 14% at LMP and 8% at CGR. Non-sea salt  $SO_4^{2-}$  was the most abundant among the secondary inorganic species.

Organic aerosol was the most abundant component at CGR, where its mean concentration was  $> 9 \mu g/m^3$  and represented 35% of  $PM_{10}$  in the days not characterized by Mistral. EC was about 4 times higher at CGR than at LMP.

Elemental carbon (EC) and organic carbon (OC) displayed a quite different behavior at the two sites. Figure 3 shows the behavior of OC versus EC at LMP and CGR. OC is correlated with EC ( $R^2=0.60$ ;  $n=59$ ) at CGR, suggesting a strong influence from carbon species primary sources, which are characterized by the simultaneous emission of EC and OC. At LMP, on the contrary, OC was not correlated with EC, indicating a strong impact of OC secondary and/or natural sources.



Thus, we used a conversion factor of 1.8 (typical for urban background sites, Turpin and Lin, 2001) at CGR, and a conversion factor of 2.1 (typical for remote sites characterized by high impact of secondary sources, Turpin and Lin, 2001) at LMP to estimate the total organic aerosol (OA) amount from the measured values of OC. Once estimated OA with this method, the sum of the various species accounted to more than 85% of the measured mass at both sites. The unreconstructed mass could be due to an underestimation of OA from OC, or to the presence of bound water not removed by the desiccation procedure at 50% relative humidity (Tsyro, 2005; Canepari et al., 2013).

Figure 4 shows the combined evolution of  $\text{nssSO}_4^{2-}$ , OA, and V at CGR between 14 and 21 June, based on the 12-hour resolution data. As discussed above, OA was mainly due to inland primary sources, while V was expected to be mainly associated with ship emissions (Becagli et al., 2012). Non-sea salt  $\text{SO}_4^{2-}$  sources are expected to be present both on land and on sea. It is interesting to note the diurnal cycle at CGR in the period 14-18 June (Figure 4). The daily cycle is very likely related to the sea breeze regime which is expected to play a significant role at CGR, and a negligible role at LMP, due to the island small size.

In the days with dominant sea breeze regime, air masses are advected from the sea during daytime, and displayed low values of OA and elevated values of V. In nighttime, when land air masses were driven to the coast, OA was larger, and V lower than in daytime.

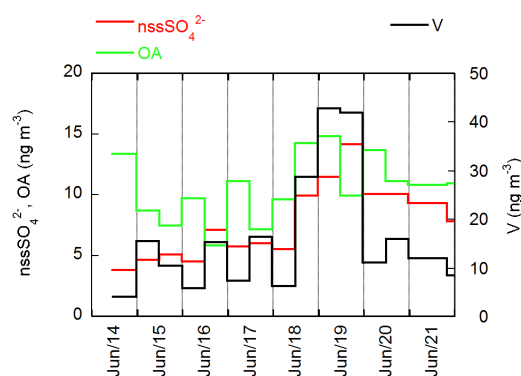


Figure 4. Time series of  $\text{nssSO}_4^{2-}$ , organic aerosol (OA), and  $V_{\text{sol}}$  at CGR.

### 3.1.1. Ship emission markers: V and Ni

Several studies focussed on the identification of specific tracers of shipping emissions (Viana et al., 2008; Becagli et al., 2012, Isakson et al., 2001, Hellebust et al., 2010).



385 Vanadium and Nickel are generally considered the best markers for this source because,  
386 after sulfur, they are the main impurities in heavy fuel oil (Agrawal et al., 2008a and b).  
387 The soluble fraction of these metals is even more representative for the ship source  
388 (Becagli et al., 2012).

389 Following Becagli et al. (2012), we used measurements of V and Ni soluble fractions  
390 ( $V_{sol}$  and  $Ni_{sol}$ , respectively). In the data set here considered the  $V_{sol}$  and  $Ni_{sol}$  ratio with  
391 respect to Al were always more than 10 times larger than for UCC, as expected for  
392 cases dominated by heavy oil combustions sources (ships, refineries, power plants,  
393 stainless steel production plants).

394 Figure 5 shows the time series of the V soluble fraction at LMP and CGR. Table 2  
395 reports slope, correlation coefficient and number of samples of the linear correlation  
396 between  $V_{sol}$  and  $Ni_{sol}$ .

397  $V_{sol}$  and  $Ni_{sol}$  are highly correlated, suggesting a common source. The obtained slope of  
398 the regression line (2.8-2.9, that increases to 3.0 for samples with  $V_{sol} > 6 \text{ ng/m}^3$ ) is  
399 typical for heavy fuel oil combustion sources (Mazzei et al., 2008; Agrawal et al., 2008a  
400 and b, Viana et al. 2009; Pandolfi et al., 2011). The same value was found at  
401 Lampedusa by Becagli et al. (2012), considering data from 2004 to 2008. The behaviour  
402 of V, Ni, and their ratio are then representative of heavy fuel oil combustion. It is  
403 however difficult to distinguish V and Ni originating from power plants, refineries, or  
404 ship engines. Moreover, several refineries are present in Sicily (Siracusa, Gela, Milazzo)  
405 and in Sardinia (Cagliari) which may potentially influence the sampling sites.

406  
407 A combination of methods is thus used in this study to unequivocally identify the ship  
408 source. The analysis is based on: additional chemical tracers, like the Rare Earth  
409 Elements, whose behaviour is specific for the refinery and the ship sources; high  
410 resolution back-trajectories, based on data from the high resolution regional model;  
411 information on the vertical mixing in the atmospheric boundary layer; coincidences  
412 between the high resolution back-trajectories and the position of different types of  
413 ships in the Sicily Channel.

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415

### 416 3.1.2. Rare Earth elements

417

418 As discussed above, anthropic V and Ni originate from heavy oil combustion, and may  
419 be considered markers of the ship source only when other sources can be excluded.  
420 Few studies propose the use of lanthanoid elements (La to Lu) to distinguish refinery  
421 from ship emissions (Moreno et al. 2008). In particular, the ratio between the La and  
422 Ce concentrations (La/Ce ratio, hereafter LCR) has been used to identify specific  
423 sources. Shipping emissions are characterised by values of LCR between 0.6 and 0.8,  
424 similarly to the earth crust. Conversely, elevated values of LCR (from 1 to 5) are  
425 associated with emissions from refinery zeolitic fluid catalytic converter plants (Moreno  
426 et al., 2008).



LCR at LMP and CGR was generally around the value expected for UCC (Handerson and Handerson, 2009), also in events with high  $V_{sol}$  concentration (fig. 5). At Capo Granitola LCR was  $>1$  in 10% of the samples, and  $>2$  in 3% of samples. At Lampedusa 24% of samples displayed  $LCR>1$ , and 8%  $LCR>2$ . Although during these days with  $LCR>1$  the concentration of  $V_{sol}$  is usually low, a contribution of aerosol from refinery cannot be excluded. The behaviour of V, La, and Ce is shown in a 3-component plot in figure 6. La and Ce were scaled in order to have the typical UCC composition in the central part of the plot. By comparison, the typical UCC composition and that of uncontaminated and La-contaminated (Refinery) Asian dust collected at Mauna Loa, Hawai'i by Olmez and Gordon (Olmez and Gordon, 1985) are also displayed in figure 6. The data points from LMP and CGR are grouped in a region with elevated values of V, and marked different amounts of La and Ce with respect to V, differently from UCC and refineries.

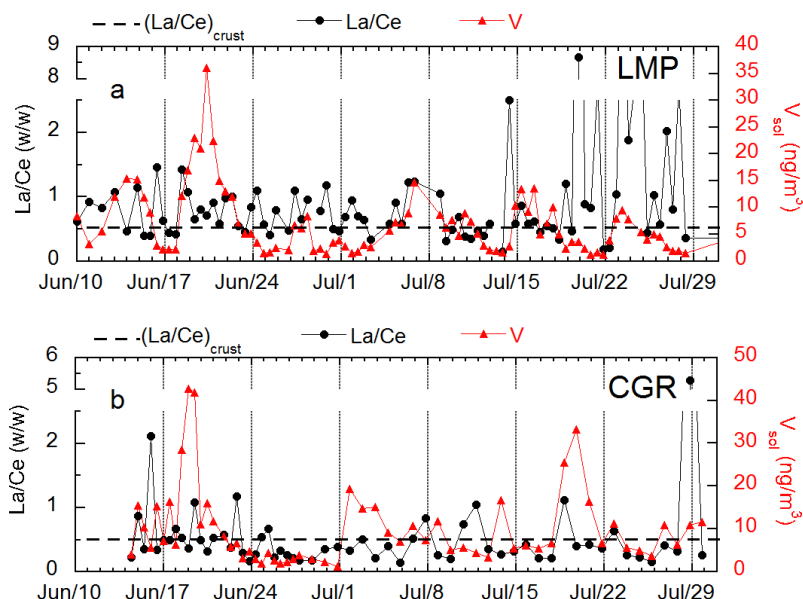


Figure 5. Time series of LCR and V at a), Lampedusa, and b), Capo Granitola. The horizontal black lines in each plot represent the LCR in the upper continental crust (0.5 w/w).

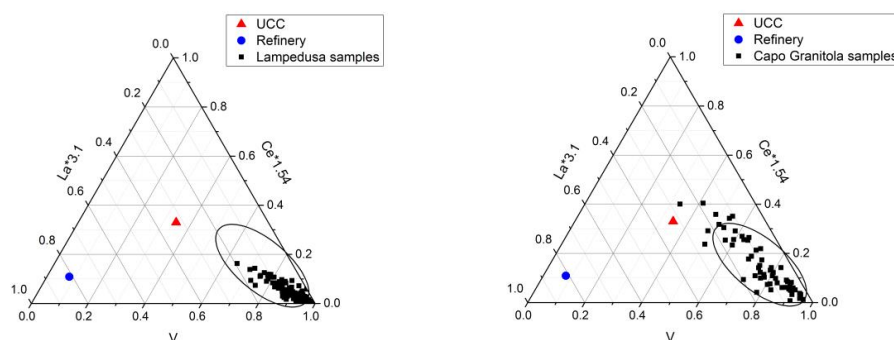


Figure 6. Three-component Ce-La-V plots for LMP (left) and CGR (right). The UCC composition is marked with a red triangle in the centre of the plots. The blue dot represents the composition of refinery-contaminated Asian dust (Olmez and Gordon, 1985). The black encircled area represents the ship emission composition.

The behaviour of the different chemical tracers support the conclusion that the V due to ship emissions is largely dominant in the  $PM_{10}$  measured at LMP and CGR during the measurement campaign. Thus, cases with elevated V can be used to identify cases with a large contribution from the ship source.

## 3.2. Trajectories and ship traffic

### 3.2.1 Origin of air masses during the campaign

All the trajectories arriving at LMP and CGR, calculated with the HYSPLIT model driven by the WRF meteorological fields (see Section 2.3), are shown in an aggregated way in Figure 7, where the trajectory frequency at each point of the computing grid is shown for the whole period (upper panels) and for the June 10<sup>th</sup> – June 30<sup>th</sup> interval (lower panels). While at LMP the trajectory frequency pattern is quite elongated in the NW-SE



direction, at CGR trajectories are distributed over a wider range of directions, despite a general prevalence of northerly sectors. The predominance of air masses coming from the northwest is particularly evident in June (lower panels), when areas with trajectory frequencies exceeding 10% are found farther to the north, up to the Gulf of Lion. During the first part of the campaign (June 2013), indeed, the synoptic situation was characterized by a "dipolar" sea level pressure anomaly pattern, with positive anomalies in the western Mediterranean and negative ones in the eastern part of the basin (Denjean et al., 2016). This situation induced stronger and more frequent than usual north-westerly winds (i.e. Mistral episodes, see Section 3.1) over the Sardinia and Sicily Channels.

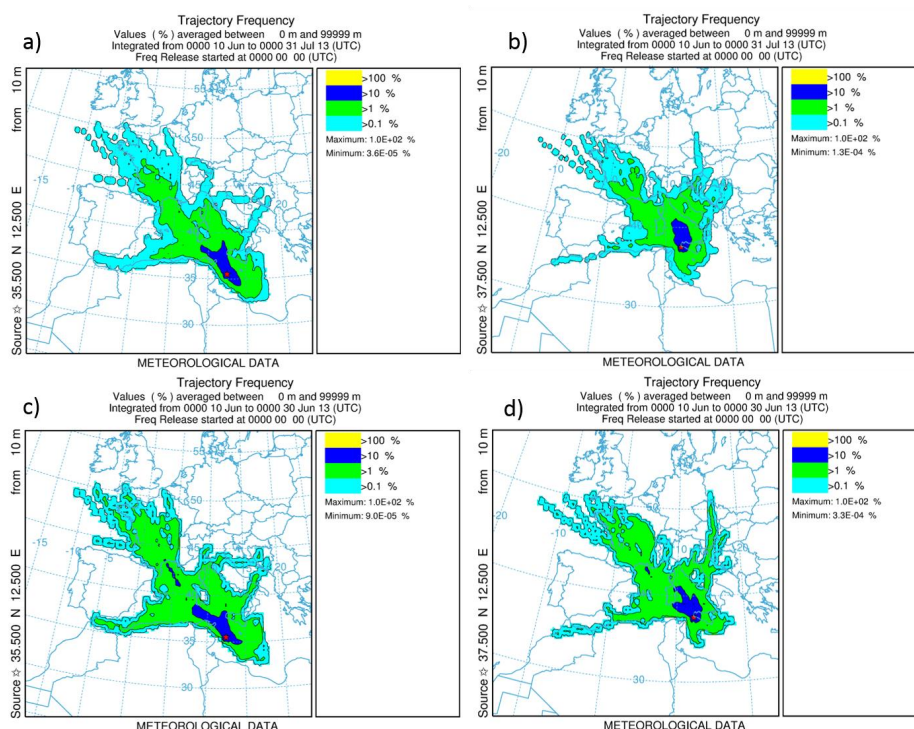


Figure 7. Trajectory frequency computed at each grid cell with starting points at LMP (panels a,c) and CGR (b,d). Upper panels show values averaged over the whole period of the campaign (10<sup>th</sup> June – 31<sup>st</sup> July 2013), while lower panels are relative to the June 10<sup>th</sup> – June 30<sup>th</sup> interval.

### 3.2.2 Ship traffic





To further investigate the mechanisms determining the V enhancement at the two sites we investigated the relationships among the amount of V, the back-trajectory pattern, the effective number of ships influencing the air mass, and the stability of the boundary layer in the ship source region (i.e., the Sicily Channel).

All back-trajectories arriving at LMP and CGR were considered and all trajectory-ship coincidences occurring within the last 36 hours before sampling were taken into account

541

It was assumed that the ship plume influenced the sampled air mass if:

- the trajectory passed at less than 15 km from the position of the ship
- the altitude of the air mass was lower than 500 m.

The total number of ships fulfilling these criteria was associated to each trajectory. The analysis was based on the available 1-hour time resolution meteorological fields (a ship influencing a trajectory was counted once every hour).

To further explore the impact of different types of ships, the analysis was carried out considering the following three ship categories: all the ships, the merchant (i.e. cargo and tanker), and the fishing vessels.

551

The atmospheric stability is also expected to play a large role in modulating the V amounts (Becagli et al. 2012). A temperature inversion, TI, index, was calculated based on the 3D atmospheric fields from the WRF model at three sites in the Sicily Channel. The temperature inversions have been used as a proxy to identify the periods characterized by a stable boundary layer. The three sites, A (37.2°N, 11.5°E), B (37.0°N, 12.4°E), and C (36.3°N, 13.3°E), were selected in the regions of most frequent ship passage and crossing with the trajectories from LMP and CGR. The TI index was calculated as the difference between the temperature at the altitude of the maximum T, and the surface T. A positive TI indicates an inversion, and the TI value provides an indication of the intensity of the inversion. Only positive values are considered in this analysis.

563

Figure 8 summarizes the results of this analysis. It shows the times series of the number of the ships influencing the trajectories arriving at LMP and CGR, respectively, and the corresponding measured values of V. Results are shown for the three classes of ships. The TI intensity is also shown.

568

In general, there is a rather good correspondence between the measured values of V and the number of ships encountered along the associated air mass trajectory at CGR. The correspondence is somewhat less evident at LMP. As discussed above, the V concentration is generally higher at CGR than at LMP. Part of this difference may be ascribed to the shorter distance between CGR and the main shipping route crossing the Sicily Channel with respect to Lampedusa, and the consequent larger number of encountered ships.

576



577 The analysis of the event with elevated V concentration at both sites, between 18<sup>th</sup> and  
578 21<sup>st</sup> June, provides further information to understand the link between the ship traffic  
579 and measured V concentration. This is the only event observed almost simultaneously  
580 at both stations. Maxima of V occurred between 19<sup>th</sup> and 20<sup>th</sup> June at CGR (about 42  
581 ng/m<sup>3</sup>), and on 21<sup>st</sup> June at LMP (36.1 ng/m<sup>3</sup>). It is worth noting that similar  
582 concentrations were measured at CGR also around 18<sup>th</sup>-19<sup>th</sup> July, in conjunction with an  
583 increase in the number of merchant vessels.  
584 On the other hand, the episode of 19<sup>th</sup>-21<sup>st</sup> June is the largest occurring at LMP, both  
585 for duration and V concentration. Especially at the beginning of the event, large values  
586 of V do not correspond with an increase of the number of ships along the air mass  
587 trajectories.

588  
589 A possible explanation for this behavior is provided by the temporal evolution of TI in  
590 the Sicily Channel. The temperature inversion started to develop on 14 June, and  
591 gradually increased in intensity until 22 June; the TI persistence and progressive  
592 increase in intensity provided suitable conditions for the trapping of the ship plumes in  
593 the boundary layer, with a consequent build-up of the ship aerosol and V concentration.  
594 This process appears particularly efficient at CGR between 21 and 25 June.

595  
596 A similar combined dependency on number of ships and TI appears also at LMP around  
597 7 July. It is interesting to note that V seems to depend more directly on the number of  
598 merchant ships (see, e.g., the lack of V peaks on 17 June, 12 and 29 July at LMP, when  
599 the number of fishing vessels was high and the number of merchant ships was low)  
600 than on the total or the fishing ships.

601 Thus, the trajectory analysis carried out in combination with the available information  
602 on the ship tracks confirms that the ship emissions are the main responsible for the  
603 moderate and elevated values of V measured at LMP and CGR during the campaign.  
604 This analysis also clearly suggests that the boundary layer structure plays a very  
605 important role in determining the impact produced by the emissions. This simplified  
606 approach confirms the importance to carefully characterize the emission scenario and  
607 the meteorological conditions in studies on the impact of ships emissions on the air  
608 quality.

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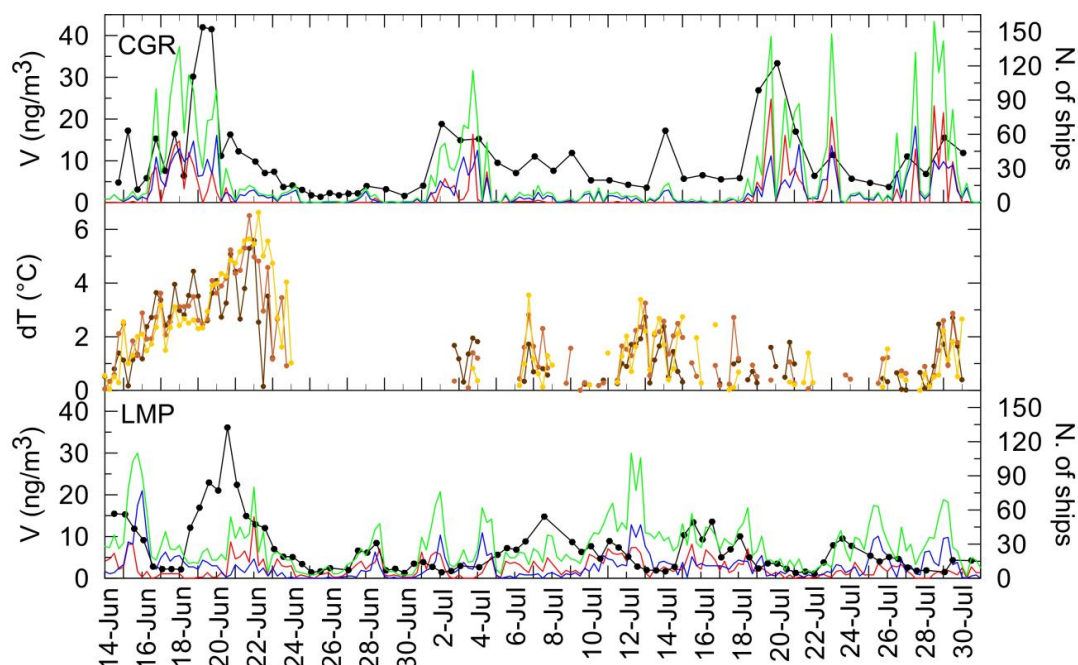


Figure 8. Time series of Vanadium concentration (black line with dots) and number of ships affecting the air masses sampled at CGR (upper panel) and LMP (lower panel). Green, red and blue lines indicate respectively the total number of ships, the number of merchant (i.e. cargo and tanker), and of fishing vessels. The time evolution of the temperature inversion index ( $dT$  in the figure) at three different locations in the Sicily Channels is shown in the middle panel; brown, red, and yellow curves show the behavior at sites A, B, and C (see text).

### 3.3. Sulfate, nitrate, and organic carbon from ships

$\text{SO}_2$  is one of the main species emitted in the ship plume in the gas phase (Agrawal et al., 2008a, b).  $\text{SO}_2$  is produced through oxidation of the S contained as impurity in heavy fuel oil, and is an aerosol precursor.

A previous study performed at Lampedusa over 5 years (Becagli et al., 2012) showed that the behavior of non-sea salt sulfate is not directly correlated with V and Ni because several other  $\text{SO}_4^{2-}$  sources (anthropic, marine biogenic, crustal, volcanic) contribute to the non-sea salt sulfate in the Central Mediterranean Sea.

The same study suggests a lower limit of about 200 for the  $\text{nssSO}_4^{2-}/\text{V}$  ratio for particles originating from heavy oil combustion at Lampedusa.



633 Figure 9 shows  $\text{nssSO}_4^{2-}/V$  versus  $V$  at LMP and CGR. At both sites  $\text{nssSO}_4^{2-}/V$  decreases  
634 for increasing  $V$  and reaches a lower limit of about 200 at elevated values of  $V$  ( $> 15$   
635  $\text{ng/m}^3$ ). We assume that the ship emission is the dominant source of the sampled  
636 particles for these cases with elevated  $V$ . This implies that in these cases virtually all  
637 sulfate originated from the ship source, and the observed lower limit for  $\text{nssSO}_4^{2-}/V$  can  
638 be considered the lower limit for the sulfate to  $V$  ratio in the ship plume.  
639 We use a value of 200 in this work as a rough estimate of the sulfate to  $V$  ratio, based  
640 on the values obtained in the previous study and confirmed by the data set used in this  
641 study for two sites and reported in figure 9a and b. The  $\text{nssSO}_4^{2-}/V$  limit values appears  
642 similar at LMP and CGR confirming the reliability of such values for the central  
643 Mediterranean Sea.

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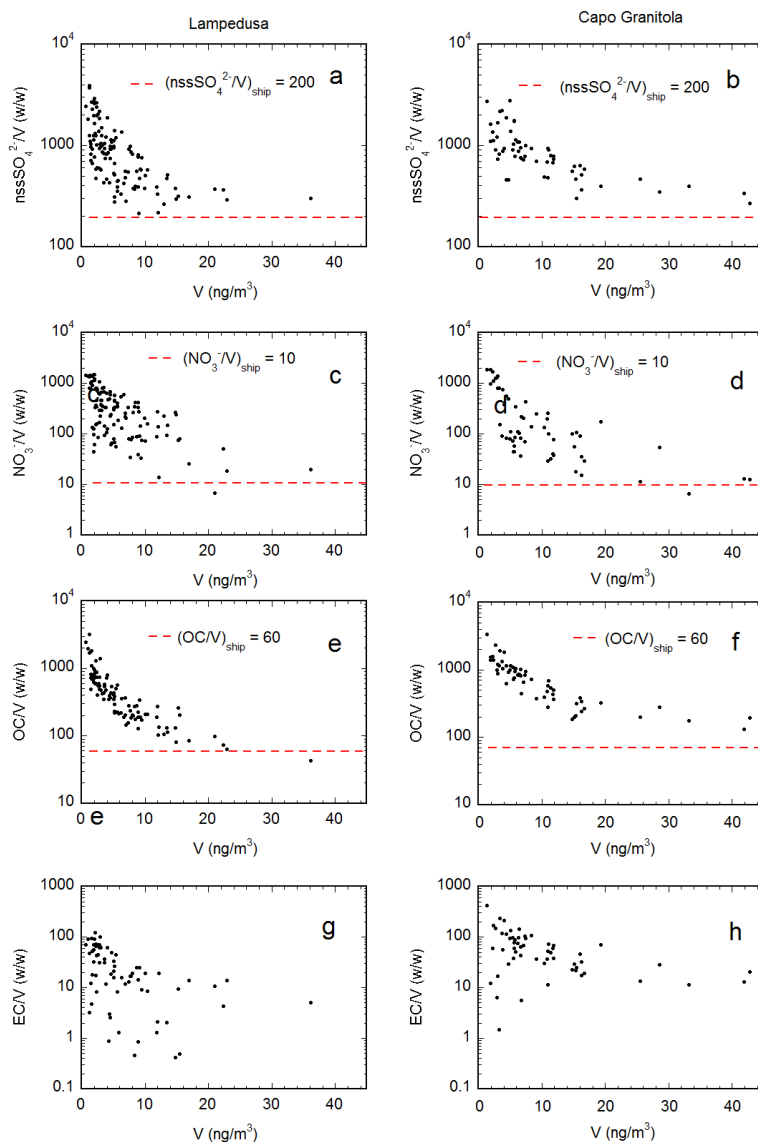


Figure 9. Scatter plots of  $\text{nssSO}_4^{2-}/V$  (a and b),  $\text{NO}_3^-/V$  (c and d),  $\text{OC}/V$  (e and f) and  $\text{EC}/V$  (g and h) vs. V concentration at LMP (plots on the left) and CGR (plots on the right) sites. The dashed lines in the plot represent lower limits for the characteristic ratio in the ship plume.



NO<sub>x</sub> are among the main compounds emitted in gas phase acting as aerosol precursors. The photochemistry of NO<sub>x</sub> leading to NO<sub>3</sub><sup>-</sup> formation in the particulate phase is complex, especially in summer due to the presence of high amounts of OH radical (see e.g., Chen et al., 2005), and the NO<sub>x</sub> contribution to the particulate phase is not easy to be quantified.

Here we try to use the same approach used for sulfate for the determination of a lower limit for the NO<sub>3</sub><sup>-</sup>/V ratio in the ship plume.

Figure 8 (plot c and d) shows the NO<sub>3</sub><sup>-</sup>/V ratio versus V at the two sites. Similarly to sulfates, the NO<sub>3</sub><sup>-</sup>/V ratio tends to a lower limit value (around 10 for V higher than 15 ng/m<sup>3</sup>) at both sites. The NO<sub>x</sub> concentration is about two times larger than that of SO<sub>2</sub> in the ship plume close to the source (Agrawal et al 2008b) and lifetime of NO<sub>x</sub> is extremely low (1.8 hour during day and 6.5 hour during night, Chen et al., 2005). However, the NO<sub>3</sub><sup>-</sup>/V limit ratio values is low compared to the limit ratio for SO<sub>4</sub><sup>2-</sup>. It has to be considered that NO<sub>3</sub><sup>-</sup> takes part in other photochemical atmospheric reactions that lead to its removal. Besides, the presence of HNO<sub>3</sub> in gas phase not neutralized by NH<sub>3</sub> or by sea salt could explain the low NO<sub>3</sub><sup>-</sup>/nssSO<sub>4</sub><sup>2-</sup> ratio in the aerosol. Indeed, the NO<sub>3</sub><sup>-</sup> concentration measured at LMP and CGR is 4-6 times lower than that of nssSO<sub>4</sub><sup>2-</sup> (table 1). Low amount of NO<sub>3</sub><sup>-</sup> with respect to SO<sub>4</sub><sup>2-</sup> from ship emissions are found in model simulations in Southern California (Dabdub, 2008). Indeed, Dabdub (2008) shows that the contribution to aerosol from ship emissions is 0.05% for NO<sub>3</sub><sup>-</sup>, and 44% for SO<sub>4</sub><sup>2-</sup>.

Elemental and Organic Carbon are also present in the ship plume (Shah et al., 2004). In particular, OC constitutes about 15-25% and EC is generally lower than 1% of the PM sampled at the plume of main ship engine powered by heavy fuel oil (Agrawal et al., 2008b).

Figure 9 shows EC/V and OC/V versus V at LMP and CGR. Similarly to sulfate and nitrate, OC/V decreases with increasing V and appears to reach a minimum value for V > 15 ng/m<sup>3</sup>. As discussed in section 3.1, other OC sources in addition to ships are probably present at CGR even at high values of V. Thus, we assume that the OC/V value obtained at Lampedusa for V > 15 ng/m<sup>3</sup> is representative of cases dominated by ship emissions, and this ratio is used to estimate the OC contribution due to ships at both sites.

The pattern of the ratio EC/V versus V is less clear; in particular, several very low values of EC/V appear also at small values of V. This result is unexpected because V and EC are both markers of the primary ship aerosol, but the data here presented seem to suggest that non negligible EC contributions from other sources were present, or that different fractionating effects acted during the transport.

### 3.4 Contribution of the ship aerosol to PM<sub>10</sub>





699 With all the limitations above described, by using the lower limits for the ratios  
700 ( $\text{nssSO}_4^{2-}/V$ ), ( $\text{NO}_3^-/V$ ), and ( $\text{OC}/V$ ) representative for ship aerosol it is possible to  
701 estimate the minimum contribution of  $\text{nssSO}_4^{2-}$ ,  $\text{NO}_3^-$  and OC emitted by ships to the  
702 total budget of these component, and also to the total  $\text{PM}_{10}$  mass. It has to be noticed  
703 that the aerosol quantification obtained by this method is a rough estimate useful to  
704 constrain the ship aerosol contribution. In addition, due to possibly different  
705 meteorological conditions and photochemical activity, such values cannot be applied in  
706 general as they can vary spatially and seasonally.

707 The minimum ratio of each specie with respect to  $V$ , the minimum estimated  
708 contribution of ship emissions, for the average amount and for the maxima, to the total  
709 concentration of these species and to  $\text{PM}_{10}$ , are reported in Table 3. As previously  
710 discussed, the measured OC contribution is multiplied by 2.1 at LMP and by 1.8 at CGR  
711 to obtain the total organic aerosol contribution.

712

713 At LMP, the estimated minimum concentration of non-sea-salt sulfate from ship  
714 emissions was  $1.3 \mu\text{g}/\text{m}^3$ , on average during this campaign. This value is lower than in  
715 the previous study by Becagli et al. (2012) obtained over a longer period (2004-2008).  
716 The relative contribution to the total sulfate is however similar here and in Becagli et al.  
717 (2012), suggesting a similar role of  $\text{nssSO}_4^{2-}$  from ship emissions to the total  $\text{nssSO}_4^{2-}$   
718 budget. At CGR the minimum ship contribution to sulfate, averaged over the same time  
719 period, is higher than at LMP ( $2.0 \mu\text{g}/\text{m}^3$ ), but this higher value corresponds to a lower  
720 contribution to the total  $\text{nssSO}_4^{2-}$ , confirming that other  $\text{nssSO}_4^{2-}$  sources are important  
721 at CGR.

722 Marmer and Langmann (2005) estimate that ship emissions contribute by 50% to the  
723 total amount of  $\text{nssSO}_4^{2-}$  in the Mediterranean. This value is larger than the one we  
724 derive (about 30%); the reader is however reminded that we estimate a lower limit for  
725 the ship contribution, useful to constrain the ship impact.

726 However, our data show that in cases with largest ship impact the  $\text{nssSO}_4^{2-}$  from ship  
727 contributes at least by 66% and 75% to the total  $\text{nssSO}_4^{2-}$  at LMP and CGR,  
728 respectively.

729 Ships appear to contribute by small fractions to the total budget of  $\text{NO}_3^-$ . As previously  
730 mentioned, the atmospheric chemistry of  $\text{NO}_3^-$  is complex and the contribution of nitrate  
731 from ship emission could be highly variable especially in the Mediterranean region  
732 where high amount of UV radiation and highly reactive radical species are present.

733 Organic aerosol from ships also contributes significantly to the total OA amount and to  
734 the total PM; in particular, at LMP at least about 92% of the total OA may be attributed  
735 to the ship source in the case with maximum ship impact.

736 By summing these three contributions, it is possible to estimate the total aerosol mass  
737 due to ship emissions, and its contribution to the total mass of  $\text{PM}_{10}$ . The lower limit for  
738 the ship contribution was  $1.9 \mu\text{g}/\text{m}^3$  and  $2.8 \mu\text{g}/\text{m}^3$ , corresponding to 11% and 8.2% of  
739  $\text{PM}_{10}$  at LMP and CGR, respectively.

740



741 These percent contributions are higher than the annual average for the Mediterranean  
742 Region estimated by Viana et al. (2014). It has to be considered that these authors  
743 used data from harbour or coastal sites, which are highly affected by other sources in  
744 addition to ships, and where gas-to-particle conversion is still at its initial phase.  
745 Moreover, the percentage reported in this study refers to the summer season, when the  
746 ship contribution in the Mediterranean region is higher (Becagli et al., 2012).

747  
748 In cases with maximum ship impact, the estimated lower limit for the ship contribution  
749 was between 40% and 48% of the total  $PM_{10}$ .

## 750 751 **Summary and conclusions**

752  
753 In this study, we investigate the impact of the ship emissions to  $PM_{10}$  on measurements  
754 made at two sites in the central Mediterranean. The main objectives of the study were  
755 to unambiguously identify the tracers of ship emissions in the sampled aerosol, and to  
756 obtain a lower limit for the produced impact.

757 The  $PM_{10}$  samples were collected in summer 2013, as a contribution to the Chemistry  
758 and Aerosol Mediterranean Experiment, in parallel at Lampedusa and at Capo Granitola,  
759 respectively South and North of the main shipping route through the Mediterranean.

760  
761 The identification of aerosol originating from ships was based on an integrated analysis  
762 combining chemical analyses, calculations of backward trajectories using a high  
763 resolution regional model, and on tracking of ship traffic in the Mediterranean through  
764 the Automatic Identification System.

765  
766 The main results of this study may be summarized as follows:

767  
768 1. moderate and elevated values of V and Ni in the aerosol were unambiguously  
769 associated with the ship source; this attribution was based on:

- 770 - the V to Ni ratio, which corresponds to what expected for heavy fuel oil  
771 combustion;
- 772 - low amounts of La and Ce with respect to V, and La/Ce ratio similar to those in  
773 the UCC, which allowed to exclude power plants or refineries as sources  
774 significantly contributing to the observed aerosol;
- 775 - coincidences between air mass trajectories and travelling ships;

776 2. in addition to travelling ships, also the planetary boundary layer vertical structure  
777 played an important role in determining the dispersion of aerosols from the ship  
778 source; temperature inversions appeared to be associated with elevated amounts of  
779 ship emissions tracers, suggesting that they favoured the build-up of aerosol  
780 concentration in the lowest atmospheric layers;



3. merchant ships (cargo and tankers) appeared to produce a larger impact on the measured aerosol than fishing vessels;
4. lower limits for the ratios  $\text{nssSO}_4^{2-}/V$ ,  $\text{NO}_3^-/V$ , and  $\text{OC}/V$ , identifying the ship-dominated emission cases, were derived from the observations. The lower limits are respectively 200, 10, and 40. These lower limits are expected to be season and site-dependent;
5. by using these ratios, the lower limits to the contribution of the ship source to  $\text{nssSO}_4^{2-}$ ,  $\text{NO}_3^-$ , OA, and to  $\text{PM}_{10}$  during the measurement campaign were estimated. Ship emissions contributed by at least 30% to the total amount of sulfate, by at least 4-7% to the total amount of  $\text{NO}_3^-$ , and by at least 8-14% to the total amount of organic aerosol. All these contributions correspond at least to 11% of  $\text{PM}_{10}$  at LMP ( $1.9 \mu\text{g}/\text{m}^3$ ), and about 8% of  $\text{PM}_{10}$  at CGR ( $2.8 \mu\text{g}/\text{m}^3$ ). In cases with largest ship impact, ships contributed up to  $12 \mu\text{g}/\text{m}^3$  to  $\text{PM}_{10}$ , and by about 48% of  $\text{PM}_{10}$  at LMP and 40% at CGR.

## Acknowledgements

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**Table 1.** Mean PM<sub>10</sub> load and composition with the related standard deviation and percentage with respect to PM<sub>10</sub> (in bracket) at Lampedusa and Capo Granitola. Mean, standard deviation and percentage are calculated on homogeneous data sets for both sites considering all the common sampling ("all data" columns) and excluding the mistral events ("Mistral excluded" columns).

	Lampedusa		Capo Granitola	
	All data	Mistral excluded	All data	Mistral excluded
<b>PM<sub>10</sub> (μg/m<sup>3</sup>)</b>	18.0±6.6	16.3±5.2	34.1±18.9	27.2±6.5
<b>Sea Salt Aerosol (μg/m<sup>3</sup>)</b>	4.63±6.30 (25.7%)	2.33±3.21 (14.3%)	8.14±15.50 (23.9%)	2.12±6.51 (7.8%)
<b>Crustal Aerosol (μg/m<sup>3</sup>)</b>	0.82±0.44 (4.6%)	0.90±0.43 (5.5%)	2.80±1.7 (8.2%)	3.02±1.75 (11.1%)
<b>nssSO<sub>4</sub><sup>2-</sup> (μg/m<sup>3</sup>)</b>	3.95±2.28 (21.9%)	4.40±2.22 (27.0%)	6.78±3.08 (19.9%)	7.53±2.78 (27.7%)
<b>NH<sub>4</sub><sup>+</sup> (μg/m<sup>3</sup>)</b>	0.98±0.56 (5.5%)	1.09±0.55 (6.7%)	1.48±0.94 (4.3%)	1.66±0.87 (6.1%)
<b>NO<sub>3</sub><sup>-</sup> (μg/m<sup>3</sup>)</b>	1.25±1.00 (7.0%)	1.02±0.02 (6.2%)	1.35±1.11 (4.0%)	1.01±0.82 (3.7%)
<b>OA (μg/m<sup>3</sup>)</b>	3.86±1.56 (21.4%)	4.04±1.59 (24.8%)	9.02±2.52 (26.5%)	9.53±2.29 (35.0%)
<b>EC (μg/m<sup>3</sup>)</b>	0.15±0.08 (0.8%)	0.15±0.08 (0.9%)	0.44±0.28 (1.3%)	0.51±0.26 (1.9%)
<b>Unknown (μg/m<sup>3</sup>)</b>	2.52±3.26 (14.0%)	2.20±3.40 (13.5%)	4.11±7.78 (12.1%)	1.82±4.48 (6.7%)

838



**Table 2.** Correlation parameters between V and Ni at LMP and CGR PM<sub>10</sub> samples for all the samples and for samples with V concentration higher than 6 ng/m<sup>3</sup>.

		Slope ( $\pm$ uncertainty)	R <sup>2</sup>	n.
<b>LMP</b>	All data	2.94 $\pm$ 0.03	0.986	124
	V <sub>sol</sub> > 6ng/m <sup>3</sup>	2.99 $\pm$ 0.03	0.994	44
<b>CGR</b>	All data	2.82 $\pm$ 0.08	0.950	59
	V <sub>sol</sub> > 6ng/m <sup>3</sup>	3.00 $\pm$ 0.05	0.989	34

**Table 3.** Estimated minimum ratio of nssSO<sub>4</sub><sup>2-</sup>, NO<sub>3</sub><sup>-</sup>, OA with respect to V, minimum contribution of each species from ship emissions averaged over the considered time period and for the cases with highest ship impact of nssSO<sub>4</sub><sup>2-</sup>, NO<sub>3</sub><sup>-</sup>, OA and PM<sub>10</sub> at LMP and CGR.

	<b>nssSO<sub>4</sub><sup>2-</sup></b> (nssSO <sub>4</sub> <sup>2-</sup> /V) <sub>min</sub> =200		<b>NO<sub>3</sub><sup>-</sup></b> (NO <sub>3</sub> <sup>-</sup> /V) <sub>min</sub> =10		<b>OA</b> (OC/V) <sub>min</sub> =40		<b>PM<sub>10</sub></b>	
	LMP	CGR	LMP	CGR	LMP	CGR	LMP	CGR
<b>Average contribution</b> <b>μg/m<sup>3</sup></b> <b>(%)</b>	1.3 (33%)	2.0 (30%)	0.065 (4.5%)	0.10 (7.2%)	0.55 (14%)	0.72 (8.1%)	1.9 (11%)	2.8 (8.2%)
<b>Maximum contribution</b> <b>μg/m<sup>3</sup></b> <b>(%)</b>	7.2 (66%)	8.6 (75%)	0.36 (50%)	0.43 (80%)	3.0 (92%)	3.1 (21%)	10.6 (48%)	12.1 (40%)

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