

1 **Volatility and lifetime against OH heterogeneous reaction of ambient Isoprene Epoxydiols-**
2 **Derived Secondary Organic Aerosol (IEPOX-SOA)**

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31 **Abstract**

32 Isoprene epoxydiols-derived secondary organic aerosol (IEPOX-SOA) can contribute
33 substantially to organic aerosol (OA) concentrations in forested areas under low NO conditions,
34 hence significantly influencing the regional and global OA budgets, accounting for example for
35 16-36% of the submicron OA in the SE US summer. Particle evaporation measurements from a
36 thermodenuder show that the volatility of ambient IEPOX-SOA is lower than that of bulk OA
37 and also much lower than that of known monomer IEPOX-SOA tracer species, indicating that
38 IEPOX-SOA likely exists mostly as oligomers in the aerosol phase. The OH aging process of
39 ambient IEPOX-SOA was investigated with an oxidation flow reactor (OFR). New IEPOX-SOA
40 formation in the reactor was negligible, as the OFR cannot accelerate processes such as aerosol
41 uptake and reactions that do not scale with OH. Simulation results indicate that adding ~100 μg
42 m^{-3} of pure H_2SO_4 to the ambient air allows to efficiently form IEPOX-SOA in the reactor. The

43 heterogeneous reaction rate coefficient of ambient IEPOX-SOA with OH radical (k_{OH}) was
44 estimated as $4.0 \pm 2.0 \times 10^{-13} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$, which is equivalent to more than a 2-week lifetime.
45 A similar k_{OH} was found for measurements of OH oxidation of ambient Amazon forest air in an
46 OFR. At higher OH exposures in the reactor ($>1 \times 10^{12} \text{ molec. cm}^{-3} \text{ s}$), the mass loss of IEPOX-
47 SOA due to heterogeneous reaction was mainly due to revolatilization of fragmented reaction
48 products. We report for the first time OH reactive uptake coefficients ($\gamma_{OH}=0.59 \pm 0.33$ in SE US
49 and $\gamma_{OH}=0.68 \pm 0.38$ in Amazon) for SOA under ambient conditions. A relative humidity
50 dependence of k_{OH} and γ_{OH} was observed, consistent with surface area-limited OH uptake. No
51 decrease of k_{OH} was observed as OH concentrations increased. These observation of
52 physicochemical properties of IEPOX-SOA can help to constrain OA impact on air quality and
53 climate.

54

55 **1 Introduction**

56 Organic aerosol (OA), which comprises 10-90% of ambient submicron aerosol mass
57 globally, has important impacts on climate forcing and human health (Kanakidou et al., 2005;
58 Zhang et al., 2007; Hallquist et al., 2009). However, quantitative predictions of OA mass
59 concentrations often fails to match the real ambient measurements by large factors,
60 (e.g. Volkamer et al., 2006; Dzepina et al., 2011; Tsigaridis et al., 2014). Improved
61 characterization of the properties and lifetime of OA is needed to better constrain OA model
62 predictions.

63 Isoprene is the most abundant non-methane hydrocarbon (NMHC) emitted into the Earth's
64 atmosphere (Guenther et al., 2012). Many studies in the past decade have shown that the reaction
65 products of isoprene-derived epoxydiols (IEPOX), formed under low NO conditions (Paulot et
66 al., 2009), can contribute efficiently to secondary OA (SOA) via reactive uptake of gas-phase
67 IEPOX onto acidic aerosols (Eddingsaas et al., 2010; Froyd et al., 2010; Surratt et al., 2010; Lin
68 et al., 2012; Liao et al., 2015). IEPOX-SOA measurements in field studies show that it can
69 account for 6-34% of total OA over multiple forested areas across the globe, with important
70 impacts on the global and regional OA budget (Hu et al., 2015 and references therein). Although
71 the formation of IEPOX-SOA from gas-phase IEPOX has been investigated in many laboratory
72 studies (e.g. Eddingsaas et al., 2010; Lin et al., 2012; Gaston et al., 2014; Riedel et al., 2015), the
73 lifetime and aging of IEPOX-SOA in the aerosol phase is still mostly unexplored in the
74 literature.

75 IEPOX-SOA can be measured by multiple methods. Gas chromatography/mass
76 spectrometry (GC/MS) or liquid chromatography/mass spectrometry (LC/MS) of filter extracts
77 can be used to measure some IEPOX-SOA species (accounting for 8-80% of total IEPOX-SOA
78 depending on the study, Lin et al., 2012; Budisulistiorini et al., 2015; Hu et al., 2015). Recently,
79 several studies have shown that factor analysis of real-time aerosol mass spectrometer (AMS)
80 data provides a method to obtain the total amount, overall fraction contribution, and properties of
81 IEPOX-SOA (Robinson et al., 2011; Budisulistiorini et al., 2013; Chen et al., 2015). The $C_5H_6O^+$
82 ion at m/z 82 in AMS spectra, arising from decomposition and ionization of molecular IEPOX-
83 SOA species, has also been suggested as a proxy for real-time estimation of IEPOX-SOA (Hu et
84 al., 2015).

85 Heterogeneous reaction of OA with hydroxyl radicals (OH) is a contributor to aerosol
86 aging and significantly influences aerosol lifetime (George and Abbatt, 2010; George et al.,
87 2015; Kroll et al., 2015). To describe the aging process, OA reaction rate coefficients with OH
88 radicals (k_{OH}), or alternatively uptake coefficients of OH (γ_{OH}), defined as the fraction of OH
89 collisions with a compound that result in reaction, have been reported for numerous laboratory
90 studies. Values of effective γ_{OH} (≤ 0.01 to ≥ 1) also can vary significantly under different
91 reaction conditions, such as different OA species (George and Abbatt, 2010), temperature and
92 humidity (Park et al., 2008; Liu et al., 2012; Slade and Knopf, 2014), OH concentrations (Slade
93 and Knopf, 2013; Arangio et al., 2015), and particle phase state or coatings (McNeill et al., 2008;
94 Arangio et al., 2015). Most of the studies that have reported k_{OH} and γ_{OH} are based on laboratory
95 experiments, with few experimental determinations of k_{OH} based on field measurements under
96 ambient conditions (Slowik et al., 2012; Ortega et al., 2016), while no γ_{OH} has been reported
97 based on field studies to our knowledge.

98 During the Southern Oxidant and Aerosol Study (SOAS), 17% of ambient OA was
99 estimated to be IEPOX-SOA (Hu et al., 2015). In this study, ambient gas and aerosol species
100 were sampled through an oxidation flow reactor (OFR) and a thermodenuder (TD) to investigate
101 heterogeneous oxidation and evaporation of ambient IEPOX-SOA, respectively. These systems
102 included an AMS and other on-line instruments measuring both gas and aerosol species inflow
103 and outflow. A simplified box model is used to investigate the fate of gas-phase IEPOX under
104 ambient and OFR conditions. The potential of evaporation to impact the lifetime of IEPOX-SOA
105 was evaluated. The heterogeneous reaction rate coefficient (k_{OH}) and OH uptake coefficient (γ_{OH})
106 of IEPOX-SOA with OH radicals are estimated from the OFR data. IEPOX-SOA aging during
107 the dry season of 2014 in central Amazonia as part of the Green Ocean Amazon
108 (GoAmazon2014/5, IOP2) experiment, using the same OFR experimental setup, was compared
109 to the SOAS results.

110 **2 Experimental method**

111 **2.1 Background and instrumentation**

112 The SOAS study (hereafter refer to “SE US study”) took place in the SE US in the summer
113 (June 1– July 15) of 2013. Results shown here are from the SEARCH Centreville Supersite
114 (CTR) in a mixed forest in Alabama (32.95° N, 87.13°W; Hansen et al., 2003). The average

115 (\pm standard deviation) temperature and relative humidity (RH) of ambient air were $25\pm 4^\circ\text{C}$ and
116 $83\pm 18\%$, respectively (Fig. S1). Biogenic volatile organic compounds (BVOCs) were highly
117 abundant with average isoprene and monoterpene concentrations of 3.3 ± 2.4 ppb and 0.7 ± 0.4
118 ppb, respectively, and they displayed clear diurnal variations (Fig. S1). Isoprene showed a broad
119 mid-afternoon peak (~ 5.8 ppb), and monoterpenes peaked during the nighttime and early
120 morning (~ 0.9 - 1.0 ppb). Chemically-resolved mass concentrations of submicron non-refractory
121 aerosol (PM_{10}) were measured by a high-resolution time-of-flight AMS (HR-ToF-AMS,
122 Aerodyne Research Inc., DeCarlo et al., 2006) at a time resolution of 2 min. Detailed information
123 about AMS setup, operation and data analysis is given in the supporting information and as well
124 as in Hu et al. (2015).

125 A “Potential Aerosol Mass” oxidation flow reactor (OFR) was used to investigate OA
126 formation/aging from ambient air over a wide range of OH exposures (10^{10} - 10^{13} molec. cm^{-3} s).
127 This field-deployable OFR provides a fast and direct way to investigate oxidation processes of
128 ambient gas and aerosol with OH radicals under low-NO chemistry (Kang et al., 2007; Lambe et
129 al., 2011; Li et al., 2015a; Peng et al., 2015; Ortega et al., 2016; Palm et al., 2016). The OFR is a
130 cylindrical vessel (~ 13 L) with an average residence time of ~ 180 - 220 s in this study, depending
131 on the flow rates of sampled ambient air (3.5 - 4.2 L min^{-1}) (Fig. S2-S3). In the “OFR185” method
132 of OH production used in this study, two low-pressure mercury lamps inside the OFR produce
133 UV radiation at 185 and 254 nm (Peng et al., 2015). OH radicals were generated when the UV
134 light initiated O_2 , H_2O , and O_3 photochemistry (Li et al., 2015a).

135 A large range of OH exposures (10^{10} - 10^{13} molec. cm^{-3} s) can be achieved by varying UV
136 light intensity, equivalent to several hours to several weeks of photochemical aging of ambient
137 air (assuming a 24-hr average $\text{OH} = 1.5 \times 10^6$ molec. cm^{-3} ; Mao et al., 2009). Thus we believe that
138 the range of OH exposures (10^{10} - 10^{13} molec. cm^{-3} s) covered by our study is the relevant range
139 for the atmosphere. We note that OH radical concentration can be calculated as the ratio of the
140 OH exposure (10^{10} - 10^{13} molec. cm^{-3} s) and the residence time (200 s). The calculated OH radical
141 concentration in our flow reactor is between 5×10^7 to 5×10^{10} molec. cm^{-3} . The lower range of
142 OH radical concentration is comparable to the higher end of observed ambient OH
143 concentrations (Mao et al., 2009).

144 OH exposures in the OFR were calculated by the real-time decay of CO added to the
145 ambient air in the OFR (1-2 ppm; OH reactivity ≈ 5 - 10 s^{-1}). The empirical estimation of OH

146 exposure based on the OFR output parameters O_3 , water, and ambient OH reactivity (15 s^{-1})
147 showed good agreement with that calculated from CO decay as shown in Fig. S4 (2015a). The
148 uncertainty of calculated OH exposures in the OFR was estimated as 35% based on regression
149 analysis (Li et al., 2015a; Peng et al., 2015).

150 The average wall loss corrections for OA in OFR during the SE US study is $2 \pm 0.7\%$. This
151 wall loss is estimated by comparing the ambient OA concentrations to those concentrations after
152 the OFR when the UV lights were off and no oxidant was present (other than ambient O_3).

153 A TD was used to investigate the volatility of ambient OA and IEPOX-SOA. The
154 temperature in the TD increased linearly during the heating period (from 30°C to 250°C over 60
155 min) and then cooled down to 30°C for 60 min. More detailed information on the TD technique
156 and instrumentation can be obtained elsewhere (Faulhaber et al., 2009; Huffman et al., 2009a;
157 Huffman et al., 2009b).

158 A typical sampling cycle during SE US study took a total of 24 min, sequentially sampling
159 ambient (4 min), TD (4 min), ambient (4 min), OFR with OH radicals as oxidant (4 min),
160 ambient (4 min), and OFR with other types of oxidation (e.g., O_3 or NO_3 as oxidants; 4 min), as
161 illustrated in the diagram in Fig. S2. Only OFR data for OH oxidation using OFR185 method is
162 presented here. A comparison of results from the OFR185 and OFR254 methods for a study at a
163 pine forest was presented by Palm et al. (2016), showing similar SOA formation by both
164 methods. Their results, together with modeling studies (Peng et al., 2015; Peng et al., 2016)
165 showed that the OFR185 method is preferable for ambient studies, and thus that was the method
166 used during SOAS. UV light intensities in the OFR were changed immediately after sampling the
167 second OFR outflow for each cycle. Thus, oxidant concentrations in the OFR had sufficient time
168 (at least 12 min, i.e. 3-4 flow e-folding times) to stabilize before the next OFR sampling interval.
169 The air from each sampling mode was sampled by the AMS, a scanning mobility particle sizer
170 (for measuring particle number size distributions; SMPS, TSI Inc.), and several other instruments
171 to measure related gas phase species, e.g., VOCs from proton-transfer-reaction mass
172 spectrometer (PTR-MS), O_3 , CO and H_2O (Table S1).

173 Measurements collected during the second Intensive Operating Period (IOP2) of the Green
174 Ocean Amazon (GoAmazon2014/5, hereinafter "Amazon study") Experiment (Martin et al.,
175 2016), which took place in the dry season of central Amazonia, are also presented here. The
176 region has high isoprene and monoterpene emissions (Karl et al., 2007; Martin et al., 2010). In

177 this analysis, data from the “T3” ground site (3.213 S, 60.599 W), a rural location 60 km west of
178 Manaus (Pop. 2 million) in the dry season (Aug. 15 to Oct. 15, 2014) are also shown. Unlike SE
179 US study, the aerosols in dry season of Amazon study were heavily influenced by biomass
180 burning (Martin et al., 2016), thus providing a difference dataset to investigate IEPOX-SOA
181 heterogeneous reaction. The instrument setup, OFR settings, sampling schemes and data
182 processing were similar to those for SE US study.

183 **2.2 IEPOX-SOA identification**

184 We classified ambient OA using positive matrix factorization (PMF) on the time series of
185 peak-fitted, high-resolution organic spectra measured by the AMS (Ulbrich et al., 2009). A factor
186 corresponding to ambient IEPOX-SOA was assigned based on its spectral features (e.g.
187 prominent $C_5H_6O^+$ ion at m/z 82), and strong correlation with hourly or daily-measured 2-
188 methyltetrols ($R=0.79$), an oxidation product of isoprene oxidation via the IEPOX pathway
189 (Surratt et al., 2010; Hu et al., 2015), as well as with sulfate ($R=0.75$), which facilitates IEPOX-
190 SOA formation through direct reactions or nucleophilic effects (Nguyen et al., 2014a; Liao et al.,
191 2015). Unconstrained PMF analysis often fails when the factor fractions become too small
192 ($<5\%$), e.g., as is for the IEPOX-SOA at higher OH exposures in the OFR in this study (Ulbrich
193 et al., 2009). To overcome this, a more advanced algorithm, the Multilinear Engine (ME-2)
194 (Paatero, 2007; Canonaco et al., 2013), was applied through the recently implemented Source
195 Finder (SoFi, Canonaco et al., 2013). In SoFi, the mass spectrum of the IEPOX-SOA factor was
196 constrained based on the ambient spectrum of IEPOX-SOA from conventional PMF, and the
197 concentrations of IEPOX-SOA factors were retrievable even at low concentrations. More
198 information can be found in Supp. Info. (Sect. 2 and Fig. S5-S9). Here after we will call IEPOX-
199 SOA PMF factor to be IEPOX-SOA for abbreviation.

200 In this study, $C_5H_6O^+$ data directly measured from AMS is used as a complementary tool to
201 examine/interpret the analysis results from IEPOX-SOA PMF factor, since both lab and ambient
202 results have shown $C_5H_6O^+$ is a very good tracer for IEPOX-SOA (Robinson et al., 2011; Lin et
203 al., 2012; Allan et al., 2014; Hu et al., 2015). Analyzing $C_5H_6O^+$ is an easy alternative method to
204 evaluate the physicochemical evolution of IEPOX-SOA, that avoids the uncertainties related to
205 PMF analysis, and thus provides further confidence in the results. This is especially true when
206 periods where the OA is dominated by IEPOX-SOA are analyzed.

207 **2.3 Box model to simulate gas-phase IEPOX**

208 The chemistry of OH oxidation in the OFR is typical of low-NO conditions with HO₂ being
209 the dominant reaction partner of RO₂ radicals due to the greatly elevated HO₂ concentrations and
210 the very short lifetime of NO and NO_x in OFR (Li et al., 2015a; Peng et al., 2015). A box model
211 (KinSim 3.2 in Igor Pro. 6.37) was used to simulate the fate of gas-phase IEPOX under both
212 ambient and OFR conditions, as shown in Fig. 3 (Paulot et al., 2009; Xie et al., 2013; Bates et
213 al., 2014; Krechmer et al., 2015). A detailed description, including reactions and parameters in
214 the model, pH-dependent uptake coefficient of IEPOX onto aerosols (γ_{IEPOX}), aerosol surface
215 area calculations and estimated photolysis of IEPOX, can be found in Supp. Info. Section 3
216 (Table S2-3 and Fig. S10-14).

217 **3 Results and discussion**

218 **3.1 Low Volatility of IEPOX-SOA**

219 TDs are widely used to investigate the volatility distribution of OA in ambient air (e.g.
220 Faulhaber et al., 2009; Cappa and Jimenez, 2010). IEPOX-SOA evaporates more slowly upon
221 heating (Fig. 1a) than total OA over a very wide range of TD temperatures (<170°C), indicating
222 that IEPOX-SOA has a lower volatility than bulk OA. Consistent with that result, a lower
223 volatility of the IEPOX-SOA tracer C₅H₆O⁺ in both SE US and Amazon studies was also found
224 (Fig. 2).

225 The volatility distributions of IEPOX-SOA and OA were estimated following the method of
226 Faulhaber et al. (2009), based on calibration of the relationship between TD temperature and
227 organic species saturation concentration at 298 K (C^*). Similar methods have been developed for
228 other thermal desorption instruments (e.g., Chattopadhyay and Ziemann, 2005; Lopez-Hilfiker et
229 al., 2016). The volatility distribution of IEPOX-SOA (Fig. 1b) shows mass peaks at $C^*=10^{-4}$ –
230 10^{-3} $\mu\text{g m}^{-3}$, which are much lower than those of diesel POA ($C^*=10^{-2}$ – 1 $\mu\text{g m}^{-3}$) and biomass-
231 burning POA ($C^*=10^{-2}$ – 100 $\mu\text{g m}^{-3}$, Fig. 1d) at various OA concentrations (1– 100 $\mu\text{g m}^{-3}$). Those
232 types of POA are reported to be semivolatile (Cappa and Jimenez, 2010; Ranjan et al., 2012;
233 May et al., 2013). The estimated distribution implies that very little of the ambient IEPOX-SOA
234 was actively partitioning to the gas phase during SE US study (Fig. 1b). Although we cannot rule
235 out some chemical changes during TD heating, this conclusion is dictated by the data at the
236 lowest TD temperatures, when such chemistry is less likely. Lopez-Hilfiker et al. (2016) have
237 shown that oligomer decomposition for IEPOX-SOA upon heating at ~90°C was important

238 during SE US study, but that process will only make the measured volatility of IEPOX-SOA in
239 TD higher than it should be. This reinforces our conclusion about the low volatility of ambient
240 IEPOX-SOA, consistent with the independent results of Lopez-Hilfiker et al. (2016).

241 Several molecular species (e.g., 2-methyltetrols, C₅-alkene triols, IEPOX organosulfate and
242 its dimer) comprising IEPOX-SOA have been characterized both in field and chamber studies
243 (Surratt et al., 2010; Lin et al., 2012; Budisulistiorini et al., 2013; Liao et al., 2015). At the CTR
244 site during the SE US study, 2-methyltetrols, C₅-alkene triols and IEPOX organosulfate
245 measured by GC/MS and LC/MS in the particle phase accounted for an average of 80%
246 (individually 29%, 28% and 24%, respectively) of total IEPOX-SOA factor mass (Hu et al.,
247 2015). The volatilities of these IEPOX-SOA molecular species was estimated based on SIMPOL
248 group contribution method (Pankow and Asher, 2008). The species reported to comprise most of
249 IEPOX-SOA have relatively high C* (2-methyltetrol=2.7 μg m⁻³; C₅-alkene triols=400 μg m⁻³,
250 and IEPOX organosulfate=0.5 μg m⁻³). The alkene triols in ambient air during SE US study
251 (where average OA mass concentration was 3.6 μg m⁻³) should have been almost completely in
252 the gas phase (>98%), while 43% and 12% of the methyltetrol and organosulfate should have
253 been in the gas-phase, respectively. The C* of those monomer species is much higher than for
254 the bulk IEPOX-SOA (C*= 10⁻⁶ -10⁻² μg m⁻³) that they are thought to comprise. On the other
255 hand, the estimated C* of a hypothetical methyltetrol molecular dimer (~10⁻⁷ μg m⁻³) is
256 significantly lower than that of most of the bulk IEPOX-SOA (Fig. 1d). This suggests that
257 IEPOX-SOA may exist as oligomers in the aerosol phase, but that the oligomers were not
258 evaporating as oligomers, rather decomposing and evaporating as monomer species at
259 temperatures intermediate with those corresponding to the C* of the monomers and the dimers,
260 consistent with results of Lopez-Hilfiker et al. (2016). Although the IEPOX organosulfate may
261 have lower volatility than estimated in Fig. 1D, it only accounts for 24% of total IEPOX-SOA
262 (Hu et al., 2015) and thus it cannot be the only reason for the low volatility of the bulk of
263 IEPOX-SOA. For reference, only 5% of the total sulfate is due to the IEPOX organosulfate, with
264 the rest being inorganic sulfate consistent with other results from the SE US in Summer 2013
265 (Liao et al., 2015). Indeed, the thermogram of total sulfate is very different from that of IEPOX-
266 SOA (Fig. 1a and Fig. 2a).

267 Further evidence supporting low volatility and strong oligomerization of IEPOX-SOA
268 molecular species has also been reported. Lin et al. (2014) showed oligomers as part of IEPOX-

269 SOA in filter-based LC/MS measurement at three sites (including CTR) during SE US study.
270 Some of the oligomers were separated by mass units of 100 (C₅H₈O₂) and 82 (C₅H₆O), which
271 would be consistent with C₅-alkene triol (C₅H₁₀O₃) and methyltetrol (C₅H₁₂O₄) oligomerization
272 though dehydration reactions (-H₂O or 2 H₂O), or with other reactions resulting in similar
273 products. Results from online gas-particle partitioning measurements at the same site during this
274 study have shown that the measured particle-phase fractions (F_p , negatively correlated with C^*)
275 of ambient IEPOX-SOA tracers (e.g., 2-methyltetrols and C₅-alkene triols) are much higher than
276 expected based on the species vapor pressures, consistent with these tracers being formed during
277 GC analysis by decomposition of larger molecules (likely oligomers) (Isaacman-VanWertz et al.,
278 2016). Thus, the low volatility of IEPOX-SOA estimated from our TD data is consistent with
279 multiple other measurements.

280 It is of high interest to estimate the fractional losses for both OA and IEPOX-SOA due to
281 isothermal evaporation upon dilution. These losses can be estimated using the volatility
282 distributions estimated from the TD measurements. This parameter can be quantified as (Cappa
283 and Jimenez, 2010):

$$284 \quad E_{loss} = 100\% \left[1 - \frac{C_{OA}(DF)}{C_{OA}(0)/DF} \right]. \quad (1)$$

285 where E_{loss} is the fractional OA loss due to evaporation; $C_{OA}(0)$ is the initial organic mass
286 concentration before dilution, and DF is the dilution factor applied. $C_{OA}(DF)$ is the OA
287 concentration in equilibrium after dilution. Dilution factors varying from one to thirty were used
288 here. The results are shown in Fig. 1c. After a 30-fold dilution, IEPOX-SOA mass loss due to
289 evaporation is estimated to be ~5%, substantially lower than for total OA (17%). There are two
290 uncertainties affecting this result. One is that the real volatility distribution of IEPOX-SOA is
291 likely even lower, since the TD results are thought to be affected by oligomer decomposition
292 upon heating. The other one is that this calculation neglects the effect of possible decomposition
293 of oligomers into monomers in ambient air. If that process occurs on a timescale of e.g., 1 day, it
294 would lead to higher evaporated fractions than estimated here. The residence time of TD is ~10-
295 15s, which may not be sufficient time for oligomer decomposition, especially at the lower
296 temperatures that determined the upper end of the estimated volatility distribution. E.g. Vaden et
297 al (2011) reported that it took 24 h to evaporate 75% of α -pinene SOA (although it is possible
298 that processes other than oligomer decomposition were important for determining the timescale

299 of those experiments). The kinetics of oligomer decomposition of IEPOX-SOA under ambient
300 conditions should be further investigated to fully constrain its evaporation dynamics.

301 **3.2 Fate of gas-phase IEPOX**

302 IEPOX-SOA loadings exhibited a continuous decrease as OH exposure increases in the
303 OFR. To interpret the observed decay of IEPOX-SOA in the OFR, we first need to understand
304 whether additional IEPOX-SOA was formed in the OFR during SE US study. More details about
305 the IEPOX-SOA decay will be discussed in Sect. 3.3. Here, the box model described above (Fig.
306 3) was used to simulate the fate of gas-phase IEPOX in OFR and ambient conditions, as shown
307 in Fig. 4.

308 In ambient air, gas-phase IEPOX will either react with OH radicals to form more oxidized
309 gas-phase products (e.g. hydroxyacetone) (Bates et al., 2014; Bates et al., 2015), be taken up
310 onto acidic aerosol (Surratt et al., 2010), or be lost from the atmosphere by dry or wet deposition
311 (Nguyen et al., 2015). Photolysis of IEPOX in ambient air should be negligible, since the
312 epoxide and hydroxyl groups in IEPOX are photostable at visible or actinic UV wavelengths
313 (Fleming et al., 1959). A model scenario accounting for organic resistance with slower IEPOX
314 uptake than pure inorganic is applied to simulate the fate of gas-phase IEPOX (Gaston et al.
315 2014; Riva et al. 2016). This scenario is the most realistic assumption, since 67% of ambient
316 aerosol is OA during SE US study (Fig. S15). Results from an alternative model assuming pure
317 inorganic aerosols are shown in Supp. Info. The model predicts that the main pathway of gas-
318 phase IEPOX removal in ambient air is aerosol-phase uptake during SE US study, where about
319 75% of IEPOX was taken up by the aerosol after one day under ambient conditions, because of
320 the efficient uptake of gas-phase IEPOX onto acidic ambient aerosols ($\text{pH}=0.8\pm 0.5$) at the CTR
321 site ($\gamma_{\text{IEPOX}}=0.009$, lifetime ~ 1.8 h). The rest of IEPOX was lost to dry deposition to the surface
322 (16%), according to reported boundary layer of 1200 m and dry deposition rate of 3 cm s^{-1}
323 (Nguyen et al., 2015), or to gas-phase reaction with OH (9%).

324 The fate of IEPOX sampled into the OFR differed from its fate in ambient air. Remaining
325 unreacted and then leaving OFR or destruction in the gas phase completely dominate the fate of
326 IEPOX under OFR conditions (Fig 4b). Negligible amounts of IEPOX (<1%) were taken up into
327 the aerosol phase in the OFR. This is mainly because the lifetime of IEPOX aerosol uptake
328 ($\gamma_{\text{IEPOX}}=0.002$; lifetime=7.0h) was much longer than the OFR residence time (200s). The lower

329 γ_{IEPOX} in OFR (0.002) than in ambient condition (0.008) was because of the higher pH of
330 aerosol leading to a slower IEPOX uptake. Higher pH in OFR (1.35 ± 0.6) than that in ambient
331 (0.8 ± 0.5) was because extra neutralized inorganic aerosol was formed in OFR. Photolysis of
332 IEPOX in OFR is estimated to be very minor (less than 0.2%) (Fig. 4b and Table S3). Loss of
333 IEPOX to the reactor walls is thought to be minor under the conditions of SE US study, given its
334 high vapor pressure (Krechmer et al., 2015; Palm et al., 2016). We can estimate the timescale of
335 IEPOX loss rate to the walls by assuming that the walls are covered by a layer of deposited
336 ambient aerosol. We combine the 1st order rate of collision of gas molecules with the walls (400
337 s; Palm et al., 2016) and the uptake coefficient for IEPOX-SOA in ambient aerosols ($\gamma_{IEPOX} =$
338 0.009) to estimate a timescale of IEPOX loss to the walls of 12.3 h, which is negligible compared
339 to residence time of IEPOX (~ 200 s) in the OFR. Even if the walls were covered by sulfuric acid
340 ($\gamma_{IEPOX} = 0.082$), the timescale of loss would be 1.4 h.

341 IEPOX-SOA mass concentrations formed in both ambient and OFR conditions were
342 calculated as a function of OH exposure. For this estimate the molar mass of IEPOX-SOA and
343 the SOA molar yield (ϕ_{SOA}) of IEPOX, defined as the sum of formed aqueous phase SOA tracer
344 relative to the heterogeneous rate of gas-phase epoxide loss to particles (Riedel et al., 2015), are
345 needed. Using the measured molecular composition of IEPOX-SOA (Hu et al., 2015), and
346 assuming all species were present as dimers as discussed above, yields an average molar mass of
347 bulk IEPOX-SOA of 270 g mol^{-1} . Laboratory uptake experiments showed the SOA molar yield
348 of IEPOX is around 10-12% for acidic $(\text{NH}_4)_2\text{SO}_4$ (Riedel et al., 2015). A molar mass of 270 g
349 mol^{-1} and $\phi_{SOA}=6\%$ (to account for the dimerization) for IEPOX-SOA were applied here. In the
350 OFR, the maximum modeled IEPOX-SOA mass concentrations were less than 12 ng m^{-3} ,
351 peaking at ~ 1 day OH exposure. The model-predicted IEPOX-SOA formation is equivalent to
352 $\sim 1\%$ of the ambient IEPOX-SOA, indicating negligible IEPOX-SOA was formed in the OFR. If
353 the more detailed IEPOX-SOA formation model of Riedel et al. (2016) were used, the modeled
354 IEPOX-SOA formation would be significantly lower, due to the consideration of the kinetics of
355 IEPOX-SOA formation. That reinforces our conclusion that IEPOX-SOA formation in the
356 reactor was negligible. An upper limit of $\sim 6\%$ of the ambient IEPOX-SOA mass being formed in
357 the OFR can be derived assuming that the particles are 100% inorganic, as shown in Fig. S17.

358 In addition to the box model results, we also have experimental evidence demonstrating
359 negligible IEPOX-SOA formation in the OFR. During the Amazon study, standard additions of

360 isoprene (50-200 ppb) were injected into ambient air at the entrance of the OFR, during a period
361 when little SOA was formed from ambient precursors. After isoprene was exposed to varied OH
362 exposures ($\sim 10^9$ - 10^{12} molec. cm^{-3} s) in the OFR in the presence of ambient aerosols, no
363 additional IEPOX-SOA formation was observed in the oxidized air exiting the OFR, as shown in
364 Fig. 5. Even under optimum OH exposures (8 - 11×10^{10} molec. cm^{-3} s), where most of the
365 isoprene and isoprene dicydroxyhydroperoxide (ISOPOOH) are expected to be oxidized and
366 before substantial decay of IEPOX-SOA occurs, no enhancements of IEPOX-SOA tracer
367 $\text{C}_5\text{H}_6\text{O}^+$ ion abundance in OA spectra were observed. Consistent with our results, a laboratory
368 flow tube study (residence time = 1 min) of low-NO isoprene oxidation in the presence of
369 acidified inorganic seeds also reported negligible IEPOX-SOA formation (Wong et al., 2015).
370 Those results highlight a key limitation of this type of OFR: processes that do not scale with OH
371 and thus are not greatly accelerated in the reactor are not captured. This limitation can be
372 removed by seeding the OFR with H_2SO_4 particles, which greatly accelerate IEPOX aerosol
373 uptake. Simulation results (not shown) indicate that adding $\sim 100 \mu\text{g m}^{-3}$ of pure H_2SO_4 to the
374 ambient air allows to efficiently form IEPOX-SOA in the reactor. The increased surface area and
375 acidity from added H_2SO_4 seed both help accelerate IEPOX reactive uptake, although acidity
376 plays a more important role. If we use the ambient surface area and pure H_2SO_4 in the model the
377 lifetime of IEPOX uptake is ~ 10 min, while if we assume the ambient acidity and the same
378 surface area as $100 \mu\text{g m}^{-3}$ of pure H_2SO_4 , the lifetime is 1.1 h.

379 **3.3 Lifetime of IEPOX-SOA against OH oxidation**

380 IEPOX-SOA loadings showed a continuous decrease as OH exposure increases in the OFR
381 (Fig. 6a). Since negligible IEPOX-SOA mass was added in the OFR (Sect. 3.2), this decay
382 should be due to the sum of all IEPOX-SOA loss processes. The loss of IEPOX-SOA is defined
383 empirically here as the loss of the molecular structures that result on AMS spectral features of
384 IEPOX-SOA (e.g., $\text{C}_5\text{H}_6\text{O}^+$ and C_4H_5^+ enhancements, Lin et al., 2012; Hu et al., 2015), such that
385 an IEPOX-SOA component cannot be distinguished in constrained PMF analysis. Evaporation,
386 photolysis and heterogeneous reaction with OH radicals are three possible loss pathways. Note
387 that the rate derived here may be a lower limit for individual molecular components of IEPOX-
388 SOA, if e.g. it takes two or more OH reactions for their AMS spectrum to no longer resemble
389 that of IEPOX-SOA.

390 In principle some IEPOX-SOA could evaporate, if semivolatile molecules in equilibrium
391 with it were oxidized by OH. As discussed above, IEPOX-SOA itself has low volatility and only
392 a small fraction (~5%) may evaporate to the gas phase after dilution of a factor of 30. Oligomer
393 decomposition followed by evaporation is very likely negligible in the flow reactor residence
394 time scale of 3 min. However, this process could be more important in ambient air, and if fast,
395 could influence the IEPOX-SOA lifetime. No results for oligomer decomposition rates or extents
396 for IEPOX-SOA have been reported in the literature, to our knowledge. Thus, further research on
397 this topic is recommended. Thus IEPOX-SOA evaporation is unlikely to contribute to the large
398 observed IEPOX-SOA loss in the OFR (up to 90%).

399 Photolysis of IEPOX-SOA also cannot explain the large decreases of IEPOX-SOA in Fig.
400 6a. Washenfelder et al. (2015) reported that IEPOX-SOA during SOAS contributed negligibly to
401 the aerosol absorption at 365 nm. Lin et al. (2014) reported a wavelength-dependent effective
402 mass absorption coefficient (MAC) value of $\sim 247 \text{ cm}^2 \text{ g}^{-1}$ at 254 nm for laboratory-generated
403 IEPOX-SOA on acidified $(\text{NH}_4)_2\text{SO}_4$ seed. Using the MAC trend vs. wavelength and the
404 measured data down to 200 nm we estimate an MAC of $\sim 5200 \text{ cm}^2 \text{ g}^{-1}$ at 185 nm. Using those
405 absorption efficiencies (and assuming an upper limit quantum yield of 1) we can derive an upper
406 limit photolysis fraction of 1.5% of IEPOX-SOA in the OFR when neglecting other competing
407 effects (e.g. OH oxidation, Table S3 and Fig. S18). In addition, the actual quantum yield may be
408 much less than 1, as IEPOX-SOA molecular species contain mainly hydroxyl and carbonyl
409 groups (Surratt et al., 2010; Lin et al., 2014). Interactions between these groups are thought to
410 result in low quantum yields in the condensed phase (Phillips and Smith, 2014; Sharpless and
411 Blough, 2014; Phillips and Smith, 2015; Peng et al., 2016). Therefore, photolysis of IEPOX-
412 SOA should contribute negligibly to the observed IEPOX-SOA decay.

413 The observed decay of IEPOX-SOA in Fig. 6a must then be the result of heterogeneous
414 reactions with OH radicals. This process can be quantitatively described as:

$$415 \quad [\text{IEPOX} - \text{SOA}]_i / [\text{IEPOX} - \text{SOA}]_0 = e^{-k_{\text{OH}} \times \text{OH}_i \times \Delta t} = e^{-k_{\text{OH}} \times \text{OH}_{\text{exp},i}} \quad (2)$$

416 where $[\text{IEPOX-SOA}]_i$ is the IEPOX-SOA mass concentration after the i^{th} OH exposure step in
417 the OFR. $[\text{IEPOX-SOA}]_0$ is the initial ambient IEPOX-SOA entering the OFR; $[\text{IEPOX-}$
418 $\text{SOA}]_i / [\text{IEPOX-SOA}]_0$ is the mass fraction remaining of IEPOX-SOA in the OFR output, shown
419 on Fig. 6a. OH_i is the average OH concentration of step i in the OFR, Δt is the real exposure

420 time. $OH_{exp,i} = OH_i \times \Delta t$ is the OH exposure of step i . k_{OH} is the heterogeneous reaction rate
421 coefficient between IEPOX-SOA and OH radicals.

422 Fitting the results in Fig. 6a with Eq. (2) results in a k_{OH} of $4.0 \pm 2.0 \times 10^{-13} \text{ cm}^3 \text{ molec.}^{-1} \text{ s}^{-1}$
423 The 1σ uncertainty was obtained by Monte Carlo simulation, from propagation of the errors of
424 $[IEPOX-SOA]_i/[IEPOX-SOA]_0$ (9%) and the uncertainty of OH exposure (35%, Fig. S4). The
425 uncertainty of $[IEPOX-SOA]_i/[IEPOX-SOA]_0$ was estimated as 9% from PMF analysis of OFR
426 data (Hu et al., 2015).

427 A similar k_{OH} value ($4.6 \times 10^{-13} \text{ cm}^3 \text{ molec.}^{-1} \text{ s}^{-1}$) was obtained by fitting the IEPOX-SOA
428 tracer $C_5H_6O^+$ ion decay as a function of OH exposure during a period (June 26th, 14:00-19:00)
429 when 80-90% of ambient OA was composed of IEPOX-SOA (Fig. S19-S20), which confirms the
430 k_{OH} determined above.

431 For comparison, the average mass fraction remaining of IEPOX-SOA vs. OH exposure
432 during the Amazon study is also shown in Fig. 6a. A similar k_{OH} value of $3.9 \pm 1.8 \times 10^{-13} \text{ cm}^3$
433 $\text{molec.}^{-1} \text{ s}^{-1}$ was obtained. Despite differences between the SE and Amazon studies, the similarity
434 of results from both studies increases our confidence in the derived value of the heterogeneous
435 reaction rate coefficient. The higher aerosol concentrations from biomass burning during the
436 Amazon study did not appear to cause any major differences in the observed OH uptake. This
437 may be due to the mostly liquid state of the ambient particles in both studies (Bateman et al.,
438 2015; Pajunoja et al., 2016), which will be discussed in detail below.

439 To investigate k_{OH} of OA, multiple experiments (usually with $RH < 30\%$) with laboratory-
440 generated different types of OA have been conducted. The bulk of those OA in the lab usually
441 had mobility particle sizes ranging from 100-300 nm (Table 1), similar to that of IEPOX-SOA in
442 SE US (wet size=415 nm). The k_{OH} value of IEPOX-SOA determined here is similar to
443 heterogeneous k_{OH} determined in those laboratory studies, including highly-oxidized OA (e.g.
444 citric acid; $3.3 - 7.6 \times 10^{-13} \text{ cm}^3 \text{ molec.}^{-1} \text{ s}^{-1}$) (Kessler et al., 2012), levoglucosan ($1.4 - 4.3 \times 10^{-13}$
445 $\text{cm}^3 \text{ molec.}^{-1} \text{ s}^{-1}$) (Slade and Knopf, 2014), and pure erythritol ($2.5 \times 10^{-13} \text{ cm}^3 \text{ molec.}^{-1} \text{ s}^{-1}$), which
446 has a similar structure to the 2-methyltetrols in IEPOX-SOA (Kessler et al., 2010). A summary
447 of k_{OH} in this study and other laboratory studies with additional experimental information for
448 each study is shown in Table 1.

449 A dependence of k_{OH} on ambient RH was found in both the SE US and Amazon studies,
450 with larger k_{OH} at high RH, especially above 90% RH (Fig. 7). This effect may be due to higher

451 liquid water content, leading to a larger surface area that facilitates faster OH uptake to the
452 aerosol phase and thus resulting in faster k_{OH} values. Accounting for liquid water content, the
453 calculated particle surface areas show similar trends to k_{OH} , in both studies, as shown in Fig. 7.
454 The values of both parameters increase with RH (especially for RH>90%).

455 An alternative explanation for the measured RH dependence would be the influence of
456 diffusion limitations. However, at the RH levels studied here (>40%), diffusion limitations of
457 OH in the aerosol phase are thought to be negligible (calculated lifetime of bulk diffusion of OH
458 radical < 1s). The diffusion coefficient of OH radical in liquid phase ($>10^{-14} \text{ m}^2 \text{ s}^{-1}$) was obtained
459 from other laboratory-generated OA (e.g. isoprene derived SOA, α -pinene derived SOA,
460 levoglucosan particles) (Renbaum-Wolff et al., 2013; Arangio et al., 2015; Li et al., 2015b). Li et
461 al. (2015b) reported that the diffusion of NH_3 on laboratory biogenic SOA is only slowed at
462 much lower transition RH (10-40%) than that for liquid/solid phase transition (50-80%). This
463 supports that under the conditions in SE and Amazon studies diffusion limitations should not
464 play a role. An effect of temperature on k_{OH} was not apparent in our study. Lai et al. (2015)
465 reported a significant effect for a laboratory study with a pure compound. We recommend that
466 this issue is explored further in the laboratory using pure IEPOX-SOA.

467 The ambient lifetime of IEPOX-SOA due to the heterogeneous reaction with OH radicals
468 was estimated to be more than 2 weeks (19 ± 9 days) based on the average k_{OH} ($4.0 \pm 2.0 \times 10^{-13} \text{ cm}^3$
469 $\text{molec.}^{-1} \text{ s}^{-1}$), assuming an average ambient OH concentration of $1.5 \times 10^6 \text{ molec. cm}^{-3}$. A similar
470 lifetime can be estimated for the Amazon study. Longer lifetimes of 48 days in SE US study and
471 99 days in Amazon study were estimated when the observed average 24h OH concentration in
472 both studies ($0.6 \times 10^6 \text{ molecule cm}^{-3}$ in SE US and $0.3 \times 10^6 \text{ molecule cm}^{-3}$ in Amazon) were used
473 (Krechmer et al., 2015). The long lifetime of IEPOX-SOA against heterogeneous oxidation is
474 consistent with the estimated lifetime of total OA in urban and forested areas (Ortega et al.,
475 2016; Palm et al., 2016), and also pure highly-oxidized OA (1-2 weeks) in laboratory studies
476 (Kessler et al., 2010; Kessler et al., 2012).

477 **3.4 Fate of Oxidized IEPOX-SOA mass**

478 It is of interest to determine whether the mass of IEPOX-SOA continues to be present in
479 the aerosol after OH heterogeneous oxidation, albeit as a different chemical form, or whether it
480 evaporates from the particles. Functionalization reactions would favor the former, while
481 fragmentation reactions would favor the latter (George et al., 2007).

482 At lower OH exposures ($<1 \times 10^{12}$ molec. cm^{-3} s) during daytime, SOA formation (non-
 483 IEPOX-SOA) was observed in the OFR (e.g., from monoterpene and sesquiterpenes oxidation,
 484 Fig. 6b), making it difficult to discern whether functionalization or fragmentation dominated for
 485 IEPOX-SOA losses. However, at OH exposures in the OFR above 1×10^{12} molec. cm^{-3} s, net
 486 SOA formation from ambient air was no longer observed. This is presumably due to organic
 487 vapors undergoing multiple generations of oxidation and fragmenting in the gas phase in the
 488 OFR (Palm et al., 2016). For that OH exposure range, changes of the aerosol phase should be
 489 dominated by heterogeneous reactions. In this regime, OA mass was lost at a rapid rate of $\sim 6\%$
 490 OA mass per 1×10^{12} molec. cm^{-3} s of OH exposure through volatilization. A very similar rate
 491 was observed for the IEPOX-SOA ($\sim 7\%$ per 1×10^{12} molec. cm^{-3} s), which implies that the main
 492 loss mechanism of IEPOX-SOA at higher OH exposures is due to volatilization following
 493 fragmentation. In the period when 80-90% of OA was composed of IEPOX-SOA, the OA also
 494 showed an up to 70% mass loss (Fig. S20), confirming the conclusion that a high fraction of
 495 IEPOX-SOA was volatilized to the gas phase after heterogeneous reaction at higher OH
 496 exposures.

497 The aerosol mass losses of IEPOX-SOA and OA into gas phase are consistent with
 498 laboratory experiments of heterogeneous reaction of pure erythritol particles (a surrogate of the
 499 IEPOX-SOA tracer 2-methyltetrols, see Fig. 6b), which also showed that OH oxidation led to
 500 formation of volatile products escaping to the gas phase (Kessler et al., 2010; Kroll et al., 2015).
 501 We note however that IEPOX-SOA is mostly composed of oligomers, rather than monomers as
 502 with erythritol.

503 3.5 Estimation of reactive uptake coefficient (γ) of OH

504 By quantifying the removal of IEPOX-SOA in the aerosol phase, an effective reactive
 505 uptake coefficient of OH (γ_{OH}) on the aerosol in the OFR can be estimated. To our knowledge,
 506 this is the first time that γ_{OH} has been derived from measurements of ambient SOA aging.

507 The variable γ_{OH} can be calculated from k_{OH} per Smith et al. (2009):

$$508 \gamma_{\text{OH}} = \frac{4 \cdot k_{\text{OH}} \cdot V_{\text{IEPOX-SOA}} \cdot \rho_0 \cdot N_A}{\bar{c} \cdot S_{\text{IEPOX-SOA}} \cdot MW_{\text{IEPOX-SOA}}} = \frac{4 \cdot k_{\text{OH}} \cdot D_{\text{surf}} \cdot \rho_0 \cdot N_A}{\bar{c} \cdot MW_{\text{IEPOX-SOA}}}, \quad (3)$$

509
 510 where k_{OH} is the heterogeneous reaction rate coefficient of IEPOX-SOA discussed above
 511 ($4.2 \pm 2.1 \times 10^{-13}$ cm^3 molec. $^{-1}$ s $^{-1}$); ρ_0 is density of aerosol in OFR, which is estimated as 1.46 ± 0.49
 512

513 g cm^{-3} based on the aerosol composition (Fig. S15). N_A is Avogadro's number; \bar{c} is the mean
514 speed of gas-phase OH radicals, calculated as $(8RT/\pi M)^{0.5}$ (R is the universal gas constant, T is
515 the temperature in K, and M is the molar mass of the OH radical). The calculated \bar{c} for OH (at
516 293 K) is 604 m s^{-1} . $MW_{IEPOX-SOA}$ is the molar mass of IEPOX-SOA. The estimated
517 $MW_{IEPOX-SOA}=270 \text{ g mol}^{-1}$ was used here, which is similar to isoprene-SOA molar mass of 252
518 g mol^{-1} estimated from a separate flow tube study based on CCN measurement (King et al.,
519 2010). An uncertainty of 30% is assigned to $MW_{IEPOX-SOA}$. $V_{IEPOX-SOA}$ and $S_{IEPOX-SOA}$ are the
520 mean volume and surface areas of IEPOX-SOA. We assume IEPOX-SOA is uniformly mixed
521 with the other aerosol species (both in the surface and volume), and independent of particle size.
522 Then for a spherical particle V_{total}/S_{total} is equal to $d_{surf}/6$, where d_{surf} is defined as surface-
523 weighted particle diameter. The dried surface-weighted aerodynamic size distribution of m/z 82
524 (background corrected), tracer of IEPOX-SOA (Hu et al., 2015), peaks around 400 nm (Fig. 8),
525 which is equivalent to mobility size of ~ 274 nm. By applying the average particle size growth
526 factor of 1.5 calculated from average kappa (0.27) and ambient RH (Nguyen et al., 2014b), the
527 average d_{surf} of wet IEPOX-SOA is estimated as 410 nm; Similar method was applied to
528 calculate d_{surf} of wet IEPOX-SOA in Amazon study, which is finally calculated to be 490 nm.

529 The average mass-weighted aerodynamic size distribution of m/z 82 in SOAS and Amazon
530 ($d_{va}=\sim 500$ nm and 600 nm) is consistent with that of sulfate ($d_{va}=\sim 450$ nm and 510 nm), which
531 may indicate sulfate control of the IEPOX uptake formation pathway (Xu et al., 2014; Liao et al.,
532 2015; Marais et al., 2016). Both peaks of m/z 82 and sulfate were systematically larger than of
533 total OA ($d_{va}=\sim 370$ or 400 nm), suggesting the IEPOX-SOA formation in SE US and Amazon
534 studies may be partially contributed by aqueous/cloud processing (Meng and Seinfeld, 1994).
535 The systematically higher oxidation level of IEPOX-SOA in the ambient air than from chamber
536 studies also support this conclusion (Chen et al., 2015; Hu et al., 2015).

537 Finally, γ_{OH} is estimated as 0.59 ± 0.33 under a range of OH concentrations between 5×10^7 -
538 $5\times 10^{10} \text{ molec. cm}^{-3}$, which is consistent with the range of γ_{OH} (0.37-0.77) calculated for highly
539 oxidized OA in laboratory studies (Table 1). The uncertainty of γ_{OH} was estimated by
540 MonteCarlo simulation, propagated from errors of each parameter in equation (2) (50% for k_{OH} ,
541 30% for d_{surf} , 28% for ρ_0 , and 30% for $MW_{IEPOX-SOA}$). When considering the apparent RH effect
542 on k_{OH} , the estimated γ_{OH} varies between 0.34-1.19. The γ_{OH} above 1 at the highest RH range

543 (90-100%) might be due to secondary reactions of IEPOX-SOA in the more dilute liquid phase.
544 The estimated γ_{OH} in Amazon study is around 0.68 ± 0.38 .

545 Ambient particles in both SOAS and GoAmazon were liquid as quantified by particle
546 bounce experiments (Bateman et al., 2015; Pajunoja et al., 2016) and thus kinetic limitations to
547 OH uptake in the OFR should not play a role (Li et al., 2015). In this study, we calculated γ_{OH}
548 based on a wide range of OH concentrations (5×10^7 - 5×10^{10} molec. cm^{-3}). Several laboratory
549 experiments suggest that OH uptake should obey the Langmuir–Hinshelwood (LH) kinetic
550 mechanism, where γ_{OH} tends to lower under higher OH concentrations, because of a saturation of
551 surface reactive sites at higher OH concentrations (George and Abbatt, 2010; Slade and Knopf,
552 2013). We have calculated k_{OH} at different OH exposure ranges (10^{10} to 10^{11} - 10^{13} molec. $\text{cm}^{-3} \text{ s}^{-1}$,
553 Fig. S22). No obvious OH dependence of k_{OH} (γ_{OH}) was found above 3×10^9 molec. cm^{-3}
554 (beyond where k_{OH} calculation is more robust), which suggests the γ_{OH} calculated in this study
555 does not depend on OH concentration. A possible explanation is that in our study OH uptake
556 occurs on liquid particles, resulting on fast OH diffusion into the particle bulk, and causing OH
557 uptake not to be limited by surface adsorption. We note that our experiments do not rule out
558 some dependence of γ_{OH} on OH at lower OH levels in the atmosphere. However, Che et al.
559 (2009) found no effect of OH on γ_{OH} for squalane particles in the range 1 - 7×10^8 molec. cm^{-3} .
560 More consideration of other factors (e.g., surface regeneration due to volatilization; aerosol
561 phase influence) should be explored in future studies of the γ_{OH} for IEPOX-SOA.

562 **4. Conclusions**

563 We investigated volatility and aging process of IEPOX-SOA during the late spring and
564 early summer of SE US and the dry season of central Amazonia with field-deployed
565 thermodenuder and oxidation flow reactor. IEPOX-SOA had a volatility distribution much lower
566 than those of the monomer tracers that have been reported as comprising most of its mass. Much
567 of IEPOX-SOA likely exists as oligomers in the aerosol phase. The kinetics of decomposition of
568 oligomers to monomers needs further investigation to fully constrain the lifetime of IEPOX-SOA
569 against evaporation.

570 The formation of IEPOX-SOA in the field and in the OFR flow reactor was investigated. In
571 contrast to the efficient IEPOX uptake in the ambient air, negligible IEPOX-SOA was formed in
572 the OFR under OH oxidation, as the OFR as used here cannot accelerate processes such as
573 aerosol uptake and reactions that do not scale with OH. Simulation results indicate that adding

574 ~100 $\mu\text{g m}^{-3}$ of pure H_2SO_4 to the ambient air would allow to efficiently form IEPOX-SOA in the
575 reactor. Photolysis and evaporation of IEPOX-SOA in the OFR contributed negligibly to
576 IEPOX-SOA loss. From the OFR results, we determined the lifetime of IEPOX-SOA through
577 heterogeneous reaction with OH radicals ($k_{\text{OH}}=4.0\pm 2.0\times 10^{-13}\text{ cm}^3\text{ molecule}^{-1}\text{ s}^{-1}$ in SE US and
578 $3.9\pm 1.8\times 10^{-13}\text{ cm}^3\text{ molecule}^{-1}\text{ s}^{-1}$ in the Amazon) is equivalent to more than a 2-week
579 photochemical aging lifetime (assuming $\text{OH} = 1.5\times 10^6\text{ molec. cm}^{-3}$). The mass lost at high OH
580 exposures is mainly volatilized, rather than transformed into other aerosol species with different
581 composition, which suggests fragmentation plays an important role during ambient aging
582 process.

583 Values of effective γ_{OH} based on the measured IEPOX-SOA k_{OH} and other particle
584 parameters were determined to be 0.59 ± 0.33 in SE US and 0.68 ± 0.38 in Amazon with no
585 dependence on OH concentration over the range 5×10^7 - $5\times 10^{10}\text{ molecule cm}^{-3}$. This is the first
586 time of γ_{OH} was estimated based on ambient SOA. Positive correlation between γ_{OH} and wet
587 particle surface areas (RH dependent) suggest that OH uptake is surface area-limited. The
588 substantially larger size distribution of IEPOX-SOA tracer m/z 82 and sulfate vs. bulk OA
589 suggests that IEPOX-SOA formation in SE US study may be controlled by sulfate and/or
590 influenced by cloud processing. However, the effect of aqueous processing under very dilute
591 conditions relevant to clouds has not been investigated to our knowledge. Our results provide
592 constraints on the sinks of IEPOX-SOA, which are useful to better quantify OA impacts on air
593 quality and climate.

594

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963 **Table 1** Summary of k_{OH} , γ_{OH} and different experiment parameters used in this study and other
 964 lab studies.

Species Name	$k_{OH} \times 10^{12}$ (cm^3 $\text{molec.}^{-1} \text{s}^{-1}$)	γ_{OH}	OH conc.(molec. cm^{-3})	React. time	Particle Size (nm)	RH	Refer ences
IEPOX-SOA in SE US	0.40±0.20	0.59±0.33	10^7-10^{10}	~200s	415	~83%	(1)
IEPOX-SOA	0.32	0.34	10^7-10^{10}	~200s	302	<60%	(1)
In SE US RH dependent	0.33	0.39	10^7-10^{10}	~200s	328	60-80%	(1)
	0.34	0.46	10^7-10^{10}	~200s	380	80-90%	(1)
	0.64	1.19	10^7-10^{10}	~200s	525	90-100%	(1)
IEPOX-SOA in Amazon	0.39±0.19	0.68±0.38	10^7-10^{10}	~200s	490	~86%	(1)
IEPOX-SOA in Amazon RH dependent	0.35	0.45	10^7-10^{10}	~200s	363	<60%	(1)
	0.35	0.46	10^7-10^{10}	~200s	380	60-80%	(1)
	0.37	0.54	10^7-10^{10}	~200s	415	80-90%	(1)
	0.53	1.09	10^7-10^{10}	~200s	576	90%-100%	(1)
Highly oxidized organic species							
<i>BTA^a</i>	0.76	0.51	$\sim 10^9-3 \times 10^{11}$	~37s	~130- 145	30%	(2)
<i>Citric acid</i>	0.43	0.37	$\sim 10^9-3 \times 10^{11}$	~37s	~130- 145	30%	(2)
<i>Tartaric acid</i>	0.33	0.40	$\sim 10^9-3 \times 10^{11}$	~37s	~130- 145	30%	(2)
<i>Erythritol</i>	0.25	0.77	$\sim 1 \times 10^9-2 \times 10^{11}$	~37s	~200	30%	(3)
Motor oil particles							
<i>Diesel particles</i>	0.4-34	0.1-8	$0.6-40 \times 10^6$	4h	~300	10-75%	(4)
<i>Nucleated motor oil particles</i>	N/A	0.72	$0-3 \times 10^{10}$	37s	~170	~30%	(5)
Biomass burning tracers							
<i>Levogluconan</i>	0.31	0.91	$\sim 1 \times 10^9-2 \times 10^{11}$	~37s	~200	30%	(3)
	0.14-0.43	0.21-0.65	10^8 to 10^9	N/A	120-267	0-40%	(6)
	N/A	0.15-0.6	10^7-10^{11}	<1 s	N/A	0%	(7)
<i>Abietic acid</i>	N/A	0.15-0.6	10^7-10^{11}	N/A	N/A	0%	(7)
<i>Nitroguaiacol</i>	N/A	0.2-0.5	10^7-10^{11}	N/A	N/A	0%	(7)
<i>MNC^b</i>	0.04-0.16	0.07-0.22	10^8 to 10^9	N/A	203-307	0-26%	(6)
Other pure organic species							
<i>Squalene</i>	N/A	0.3±0.07	1×10^{10}	~37s	~160	30%	(8)
<i>Squalene</i>	1.8-1.9	0.49-0.54	$1-7 \times 10^8$	1.5-3h	~220	30%	(9)
<i>Palmitic Acid</i>	N/A	0.8-1	$1.4-3 \times 10^{10}$	10-17s	85-220	~16%	(10)

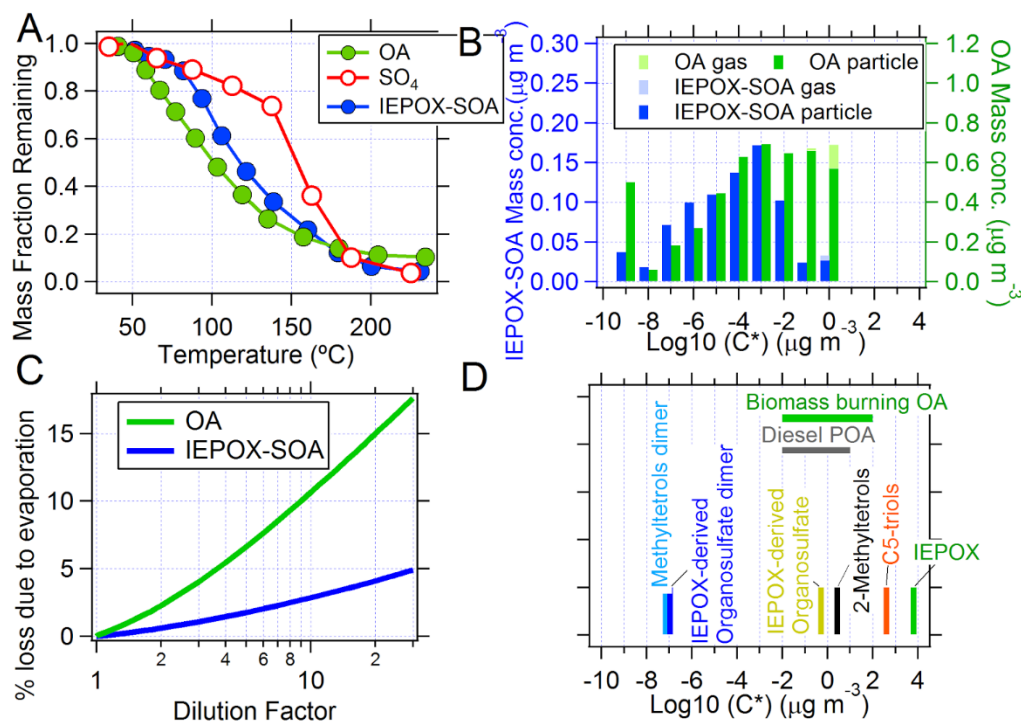
965 ^a 1, 2, 3, 4-Butanetetracarboxylic acid; ^b 4-methyl-5-Nitrocatechol

966 (1) This study; (2) (Kessler et al., 2012); (3) (Kessler et al., 2010); (4) (Weitkamp et al., 2008); (5) (Isaacman et al., 2012); (6)

967 (Slade and Knopf, 2014); (7) (Slade and Knopf, 2013); (8) (Smith et al., 2009); (9) (Che et al., 2009); (10) (McNeill et al., 2008).

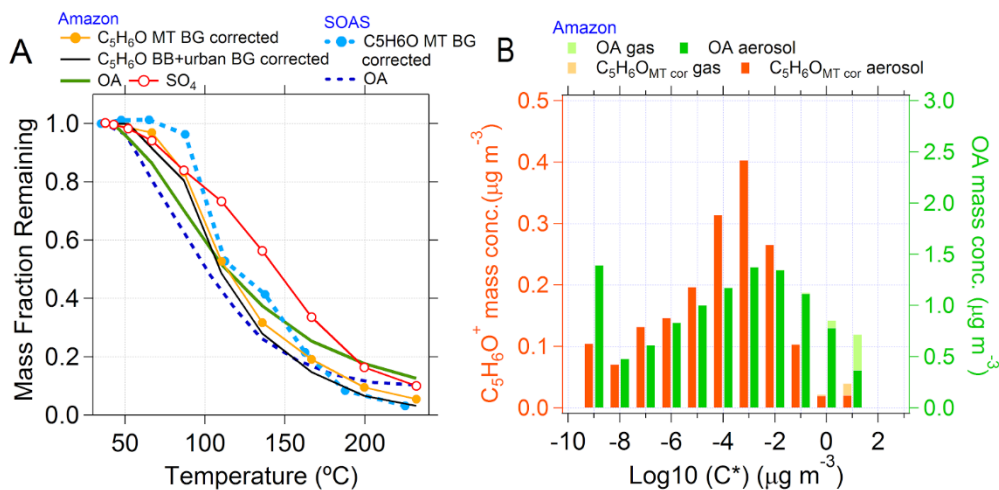
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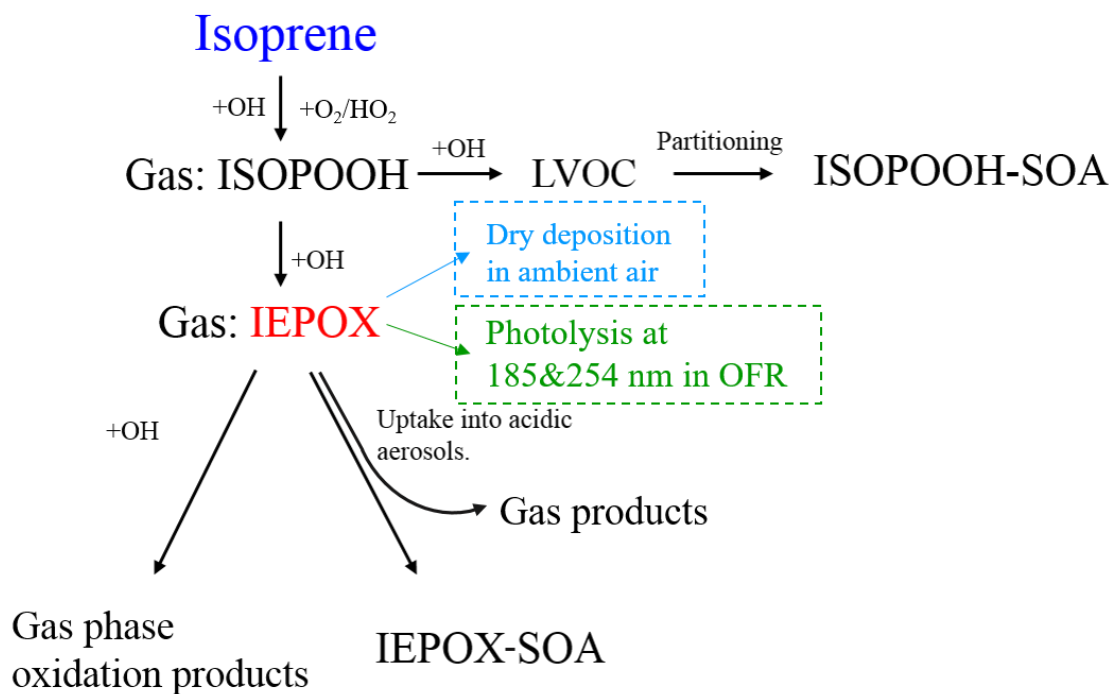
971 **Figure 1.** (a) Mean mass fraction remaining of IEPOX-SOA, OA and SO₄ versus temperature in
 972 TD (“thermograms”) during SE US study. (b) Volatility distributions of IEPOX-SOA and OA
 973 estimated from TD thermograms (see text). Bars are offset for clarity and were both calculated
 974 for integer log(C*) values. (c) Evaporation losses of IEPOX-SOA and OA as a function of
 975 dilution factors. (d) Volatility of typical IEPOX-SOA molecular species in the aerosol phase
 976 based on the on SIMPOL group contribution method (Pankow and Asher, 2008). The reduction in
 977 vapor pressure upon addition of a nitrate group was used to estimate the effect of the sulfate
 978 group, due to lack of SIMPOL parameters for the latter, and the derived C* may be
 979 overestimated for this reason.



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981 **Figure 2.** (a) Thermogram of OA, SO₄ and background-corrected C₅H₆O⁺ ion in the SE US and
 982 Amazon studies. (b) Volatility distributions of C₅H₆O⁺ and OA estimated based on TD
 983 thermograms from the Amazon study.

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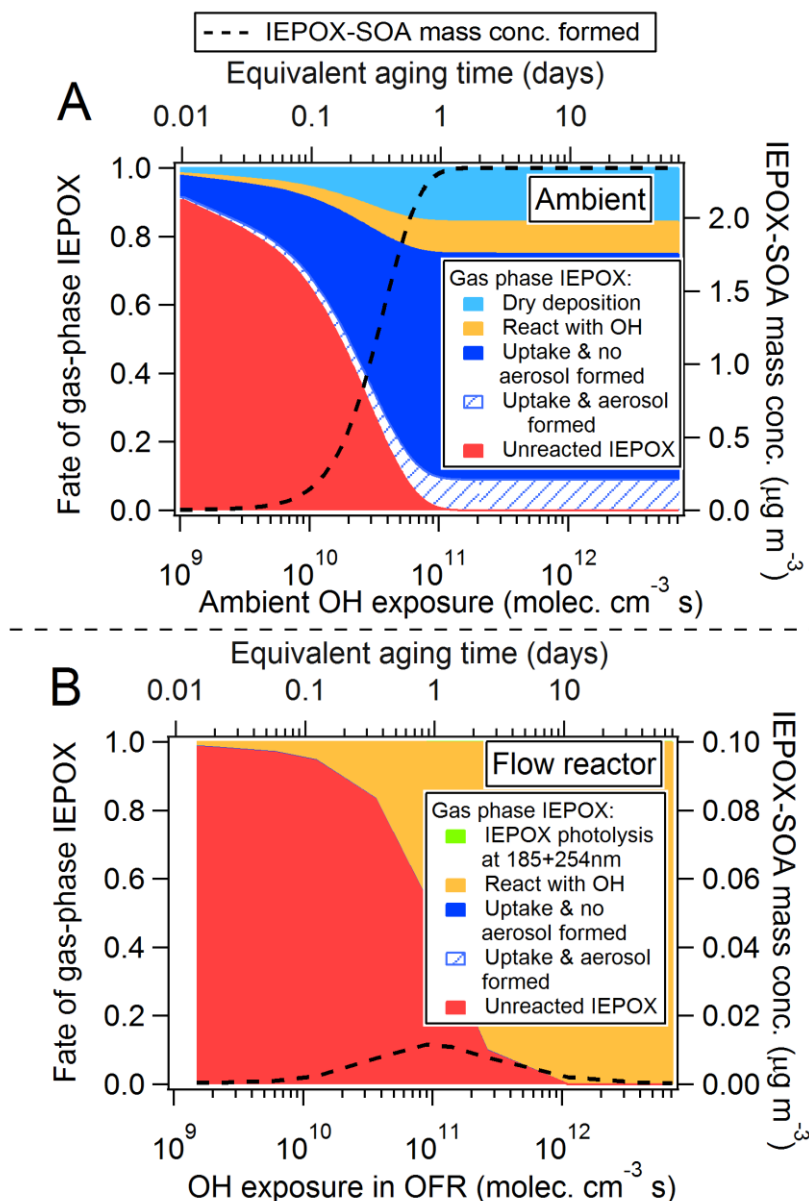


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987 **Figure 3.** Mechanism diagram of gas-phase IEPOX model in ambient and OFR conditions.
 988 ISOPOOH-SOA refers to SOA formed through gas-particle partitioning of low-volatile VOCs
 989 from oxidation of isoprene 4-hydroxy-3-hydroperoxide (4,3-ISOPOOH) under low-NO
 990 conditions (Krechmer et al., 2015).

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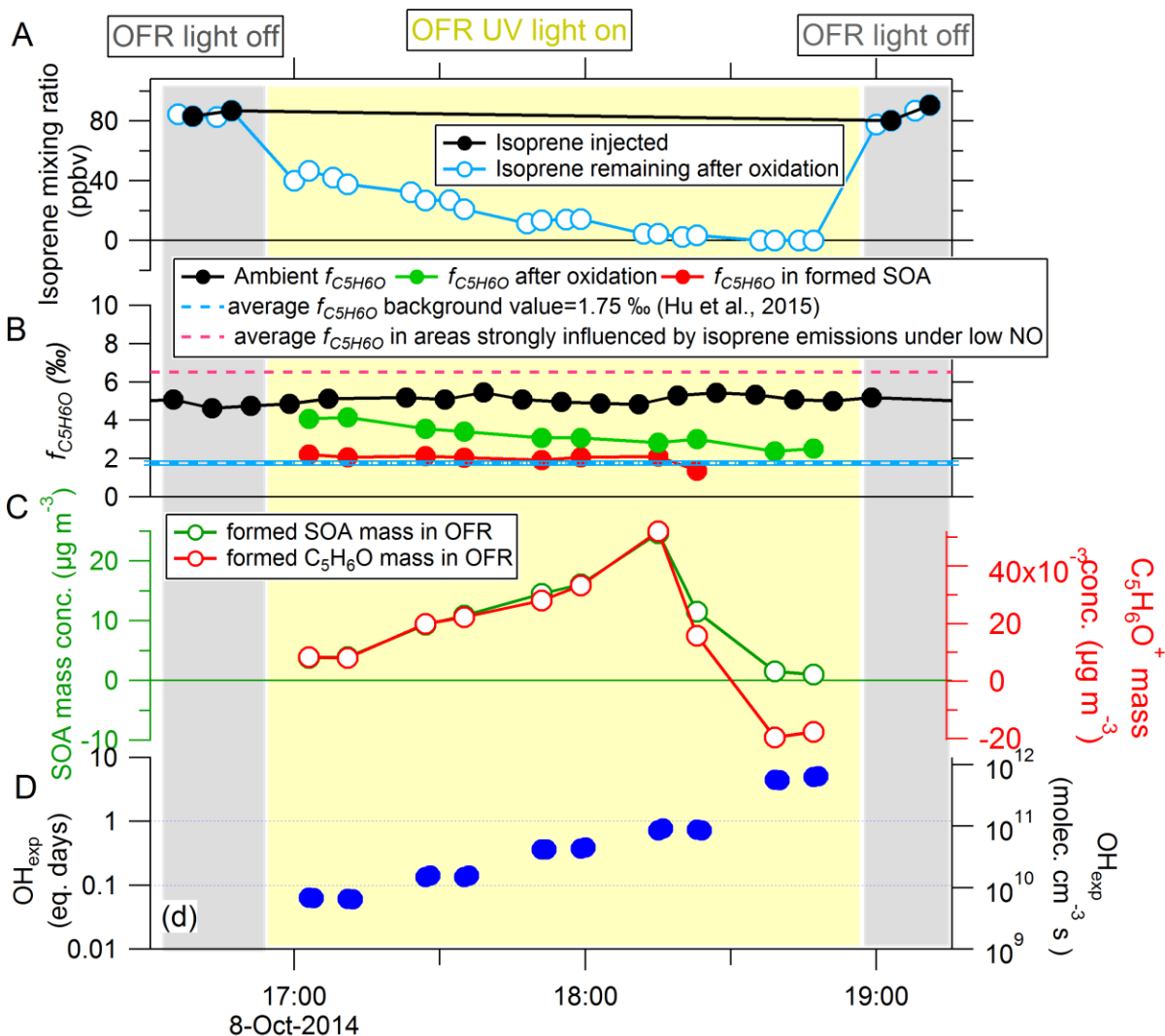
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995 **Figure 4.** Modeled IEPOX fate (a) in ambient air and (b) oxidation flow reactor (OFR)
 996 conditions in SE US study. The uptake rate of gas-phase IEPOX onto aerosol is calculated by
 997 using the model of Gaston et al. (2014), and is mainly influenced by aerosol pH (estimated as 0.8
 998 and 1.35 for ambient and OFR aerosol, respectively) and aerosol surface areas (300 and 350
 999 $\mu\text{m}^2/\text{cm}^3$ for ambient and OFR aerosol, respectively). The calculated IEPOX-SOA mass
 1000 concentrations are shown in Fig. 3. The OH exposures for both panels range 15 min-2 months of
 1001 atmospheric equivalent age (at OH concentration= 1.5×10^6 molec. cm^{-3}).

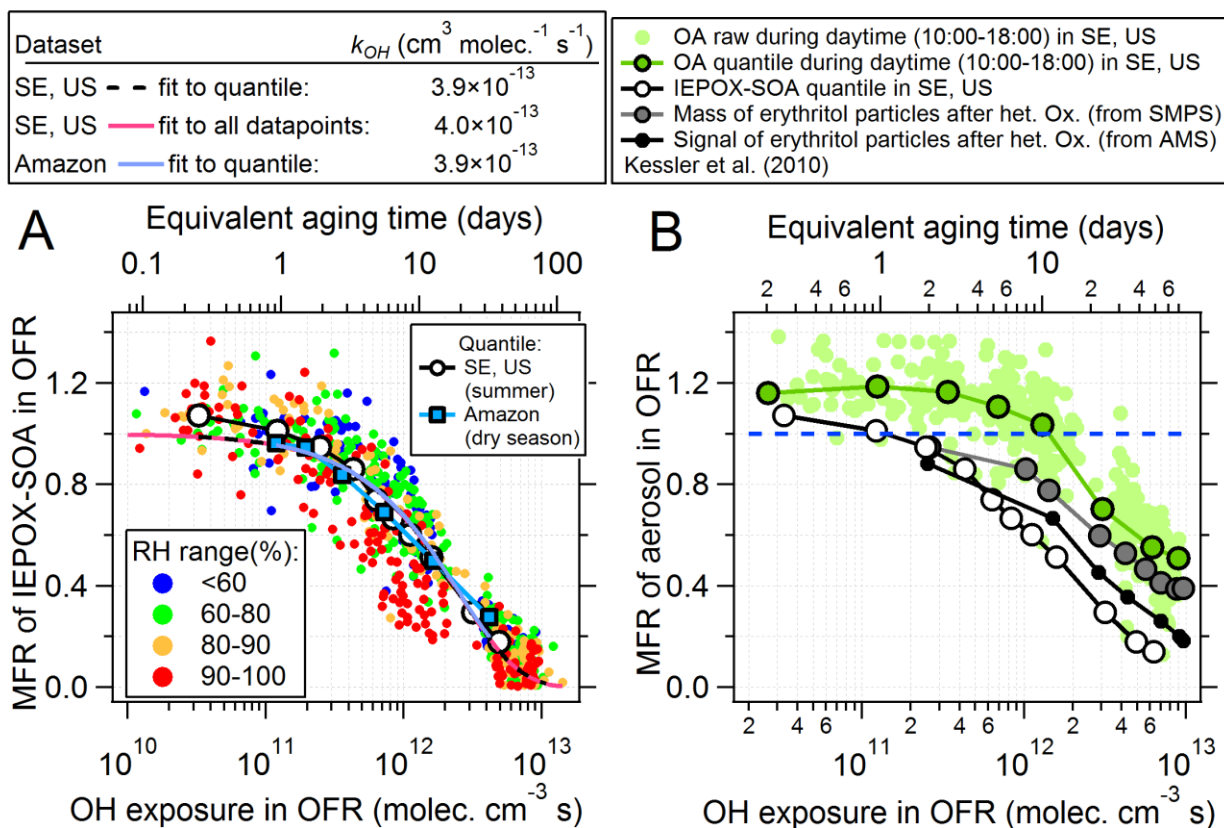
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1004 **Figure 5.** Isoprene standard addition experiment in ambient air during the GoAmazon2014/5
 1005 study. (a) Isoprene concentration injected and remaining after OFR. (b) Time series of ambient
 1006 $f_{C_5H_6O}$, $f_{C_5H_6O}$ in OA after oxidation and $f_{C_5H_6O}$ in newly formed SOA from OFR oxidation. The
 1007 average background value $f_{C_5H_6O}=1.75\%$ from urban and biomass burning emissions and
 1008 $f_{C_5H_6O}=6.5\%$ from aerosol strongly influenced by isoprene emissions are also shown (Hu et al.,
 1009 2015). (c) Time series of mass concentration of newly formed SOA (left axis) and $C_5H_6O^+$ (right
 1010 axis). (d) Time series of equivalent aging time (left axis) and OH exposure in OFR (right axis).
 1011 OH concentration= 1.5×10^6 molec. cm^{-3} was assumed here to calculate equivalent OH aging
 1012 times. The grey background indicates OFR light off period and light yellow is OFR light on
 1013 period. Different OH exposures were achieved by varying the UV light intensity. Residence time
 1014 in the OFR was about 200 s.

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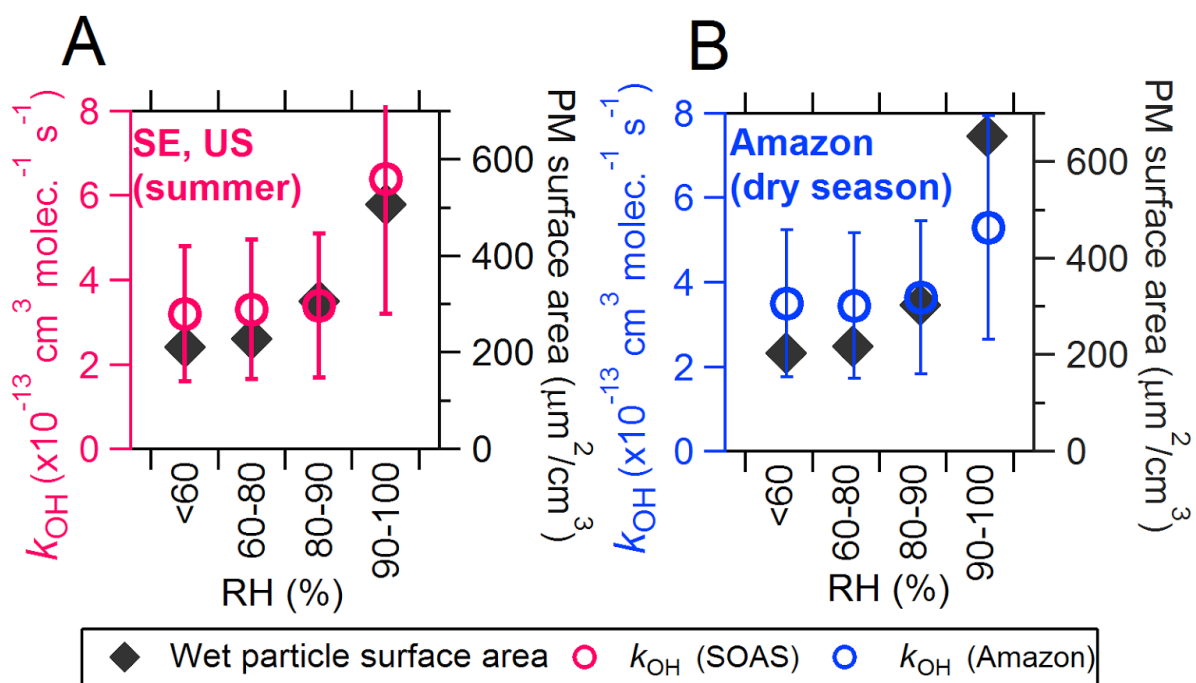
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1017 **Figure 6.** (a) Mass fraction remaining (MFR) of IEPOX-SOA in OFR output as a function of
 1018 OH exposure during the entire SOAS and GoAmazon 2014/5 (dry season) studies. Individual
 1019 datapoints from SOAS are color-coded by ambient RH. Similar data for GoAmazon 2014/5 are
 1020 shown in Fig. S24. (b) Mass fraction of OA remaining in OFR output as a function of OH
 1021 exposure in daytime (12:00-18:00) during SOAS. Also shown is the MFR of pure erythritol
 1022 particles after heterogeneous oxidation as detected by SMPS and by AMS for reference (Kessler
 1023 et al., 2010). Erythritol has a similar structure to the IEPOX-SOA tracers 2-methyltetrols.

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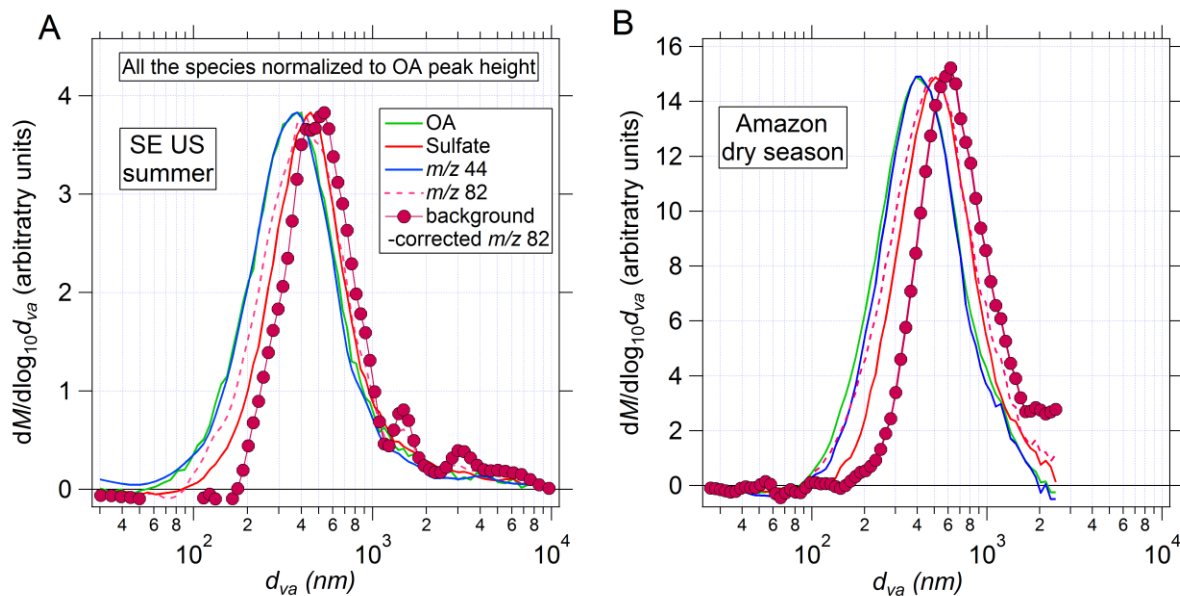
1028 **Figure 7.** Estimated k_{OH} of IEPOX-SOA vs. ambient RH during the SOAS and Amazon studies.

1029 The ambient wet particle surface areas in both studies are shown on the right axis.

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1034 **Figure 8.** Average mass-weighted aerodynamic size distribution of OA, sulfate, m/z 44 and m/z
1035 82 in (a) SE US and (b) Amazon. The mass size distribution of m/z 82 with background
1036 correction is also shown. The background correction method was introduced in Hu et al.(2015).
1037 Heights of all the size distributions are set to the same value for ease of visual comparison.

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