## 1 In Vitro Exposure to Isoprene-Derived Secondary Organic Aerosol by Direct Deposition

## 2 and its Effects on COX-2 and IL-8 Gene Expression

- 3 Maiko Arashiro<sup>1</sup>, Ying-Hsuan Lin<sup>1,6</sup>, Kenneth G. Sexton<sup>1</sup>, Zhenfa Zhang<sup>1</sup>, Ilona Jaspers<sup>1-5</sup>,
- 4 Rebecca C. Fry<sup>1,3</sup>, William G. Vizuete<sup>1</sup>, Avram Gold<sup>1</sup>, Jason D. Surratt<sup>1,\*</sup>
- <sup>1</sup> Department of Environmental Sciences and Engineering, Gillings School of Global Public
   Health, University of North Carolina at Chapel Hill, Chapel Hill, NC 27599, USA
- <sup>2</sup> Center for Environmental Medicine, Asthma, and Lung Biology, School of Medicine,
   <sup>8</sup> University of North Carolina at Chapel Hill, Chapel Hill, NC 27599, USA
- <sup>3</sup> Curriculum in Toxicology, University of North Carolina at Chapel Hill, Chapel Hill, NC
   27599, USA
- <sup>4</sup> Department of Pediatrics, School of Medicine, University of North Carolina at Chapel Hill,
   Chapel Hill, NC 27599, USA
- <sup>5</sup> Department of Microbiology and Immunology, School of Medicine, University of North
   Carolina at Chapel Hill, Chapel Hill, NC 27599, USA
- <sup>6</sup> Michigan Society of Fellows, Department of Chemistry, University of Michigan, Ann Arbor,
   MI 48109, USA
- 17
- 18 \*To whom correspondence should be addressed:
- Jason D. Surratt, Department of Environmental Sciences and Engineering, Gillings School of
   Global Public Health, University of North Carolina at Chapel Hill, Chapel Hill, North Carolina
- 21 27599, USA. Tel: (919)-966-0470; Fax: (919)-966-7911; Email: <u>surratt@unc.edu</u>
- 22
- 23
- 24
- 25 For Submission To: Atmospheric Chemistry and Physics Discussions
- 26 <u>Manuscript Information</u>: Number of Figures 5.
- 27
- 28
- 29

### 30 Abstract

31 Atmospheric oxidation of isoprene, the most abundant non-methane hydrocarbon emitted into 32 Earth's atmosphere primarily from terrestrial vegetation, is now recognized as a major 33 contributor to the global secondary organic aerosol (SOA) burden. Anthropogenic pollutants 34 significantly enhance isoprene SOA formation through acid-catalyzed heterogeneous chemistry 35 of epoxide products. Since isoprene SOA formation as a source of fine aerosol is a relatively 36 recent discovery, research is lacking on evaluating its potential adverse effects on human health. 37 The objective of this study was to examine the effect of isoprene-derived SOA on inflammation-38 associated gene expression in human lung cells using a direct deposition exposure method. We 39 assessed altered expression of inflammation-related genes in human bronchial epithelial cells 40 (BEAS-2B) exposed to isoprene-derived SOA generated in an outdoor chamber facility. 41 Measurements of gene expression of known inflammatory biomarkers interleukin 8 (IL-8) and 42 cyclooxygenase 2 (COX-2) in exposed cells, together with complementary chemical measurements, showed that a dose of 0.067  $\mu$ g cm<sup>-2</sup> of SOA from isoprene photooxidation leads 43 44 to statistically significant increases in IL-8 and COX-2 mRNA levels. Resuspension exposures 45 using aerosol filter extracts corroborated these findings, supporting the conclusion that isoprene-46 derived SOA constituents induce the observed changes in mRNA levels. The present study is an 47 attempt to examine the early biological responses of isoprene SOA exposure in human lung cells.

- 48
- 49
- 50
- 51

52 **1. Introduction** 

53 Recent work has shown that isoprene (2-methyl-1,3-butadiene) is an important precursor 54 of secondary organic aerosol (SOA), which has potential impacts on climate change and public 55 health (Lin et al., 2013b; Rohr, 2013; Lin et al., 2016). Current understanding of isoprene SOA 56 formation is based on laboratory studies showing that gas-phase photooxidation of isoprene 57 generates key SOA precursors, including isomeric isoprene epoxydiols (IEPOX), methacrylic 58 acid epoxide (MAE), hydroxymethyl-methyl-α-lactone (HMML), and isoprene 59 hydroxyhydroperoxides (ISOPOOH) (Paulot et al., 2009; Surratt et al., 2010; Lin et al., 2012; 60 Lin et al., 2013b; Nguyen et al., 2015; Krechmer et al., 2015). The formation of SOA from these 61 precursors is influenced by controllable anthropogenic emissions such as oxides of nitrogen  $(NO_x)$  and sulfur dioxide  $(SO_2)$ . Atmospheric oxidation of  $SO_2$  contributes to particle acidity, 62 63 which enhances isoprene SOA formation through acid-catalyzed reactive uptake and multiphase 64 chemistry of IEPOX and MAE (Surratt et al., 2007; Surratt et al., 2010; Lin et al., 2012; Gaston 65 et al., 2014; Riedel et al., 2015), while  $NO_x$  determines whether the oxidation pathway leading to 66 IEPOX or MAE/HMML predominates (Lin et al., 2013b; Surratt et al., 2010; Nguyen et al., 2015). Isoprene SOA comprises a large portion of global atmospheric fine particles ( $PM_{2.5}$ , 67 68 aerosol with aerodynamic diameters  $\leq 2.5 \,\mu$ m) (Carlton et al., 2009; Henze et al., 2008) but few 69 studies have focused on its health implications (Lin et al., 2016). Evaluating the health effects of 70 SOA from isoprene oxidation is important from a public health perspective, not only because of 71 its atmospheric abundance, but also because the anthropogenic contribution is the only 72 component amenable to control (Pye et al., 2013; Gaston et al., 2014; Xu et al., 2015; Riedel et 73 al., 2015).

74 Many studies have shown that particulate matter is closely linked to health effects 75 ranging from exacerbation of asthma symptoms to mortality associated with lung cancer and 76 cardiopulmonary disease (Dockery et al., 1993; Schwartz et al., 1993; Samet et al., 2000). PM<sub>2.5</sub>, 77 in particular, has been linked to negative health outcomes with an estimated contribution of 3.2 78 million premature deaths worldwide as reported in the Global Burden of Disease Study 2010 79 (Lim et al., 2012). Despite evidence that particle composition affects toxicity, fewer studies 80 focus on the link between chemical composition and health/biological outcomes (Kelly and 81 Fussell, 2012). Prior work on complex air mixtures has shown that gaseous volatile organic 82 compounds (VOCs) alter the composition and ultimately the toxicity of particles (Ebersviller et 83 al., 2012a, b). SOA resulting from natural and anthropogenic gaseous precursors, such as  $\alpha$ -84 pinene and 1,3,5-trimethylbenzene, have been shown to affect cellular function (Gaschen et al., 85 2010; Jang et al., 2006) and recently isoprene-SOA formed from the reactive uptake of epoxides 86 has been shown to induce the expression of oxidative stress genes (Lin et al., 2016).

87 The objective of this study is to generate atmospherically relevant isoprene-derived SOA 88 and examine its toxicity through *in vitro* exposures using a direct deposition device. Compared to 89 exposure of cells in culture media to resuspended particles, direct particle deposition likely 90 provides a more biologically relevant exposure model and enhances sensitivity of cells to air 91 pollution particle exposures (Volckens et al., 2009; Lichtveld et al., 2012; Hawley et al., 2014a; 92 Hawley et al., 2014b; Zavala et al., 2014; Hawley and Volckens, 2013). The Electrostatic 93 Aerosol in vitro Exposure System (EAVES) used in this study deposits particles generated in our 94 outdoor photochemical chamber directly onto lung cells by electrostatic precipitation (de Bruijne 95 et al., 2009). Similar techniques and devices have been used to expose cells to diesel exhaust particles (Lichtveld et al., 2012; Hawley et al., 2014b), but our study is the first to utilize the 96

97 EAVES to explore the potential adverse effects of isoprene SOA on human lung cells.
98 Additionally, for a more atmospherically relevant exposure, isoprene-SOA was photochemically
99 generated in an outdoor chamber to mimic its formation in the atmosphere.

100 We have recently demonstrated through a chemical assay that isoprene-derived SOA has 101 the potential for inducing reactive oxygen species (ROS) (Kramer et al., 2016), which are linked 102 to oxidative stress and inflammation (Reuter et al., 2010; Li et al., 2003). An in vitro study that 103 followed supported the potential for isoprene-SOA to affect the levels of oxidative stress genes 104 (Lin et al., 2016). In this study we chose to examine the gene expression levels of interleukin-8 105 (IL-8) and cyclooxygenase-2 (COX-2), not only for their links to inflammation and oxidative 106 stress (Kunkel et al., 1991; Uchida, 2008), but because both have been examined in previous 107 studies using the EAVES for fresh and aged diesel exhaust (Lichtveld et al., 2012). Other studies 108 on air pollution mixtures have also examined IL-8 as a biological endpoint due to its involvement 109 with inflammation (Zavala et al., 2014; Ebersviller et al., 2012a, b; Doyle et al., 2004; Doyle et 110 al., 2007). We compared the gene expression levels in cells exposed to SOA generated in an 111 outdoor chamber from photochemical oxidation of isoprene in the presence of NO and acidified 112 sulfate seed aerosol to cells exposed to a dark control mixture of isoprene, NO, and acidified 113 sulfate seed aerosol to isolate the effects of the isoprene-derived SOA on the cells using the 114 EAVES. In addition, we collected SOA onto filters for subsequent resuspension exposure to 115 ensure that effects observed from EAVES exposures were attributable to particle-phase organic 116 products.

117 2. Experimental Section

118 2.1 Generation of SOA in the Outdoor Chamber Facility. SOA were generated by 119 photochemically oxidizing a mixture of acidified sulfate seed aerosol, isoprene, and NO injected

into an outdoor smog chamber facility. The outdoor chamber is a 120-m<sup>3</sup> triangular cross-section 120 121 Teflon chamber located on the roof of the Gillings School of Global Public Health, University of 122 North Carolina at Chapel Hill. The chamber facility has been described in detail elsewhere by 123 Lichtveld et al. (2012). The outdoor chamber facility is equipped with sampling lines that allow 124 direct deposition exposure of cells, online chemical measurements, and filter collection for 125 offline chemical analysis. Sampling lines run from the underside of the chamber directly to the 126 chemistry lab below where online measurement instruments and the direct deposition exposure 127 device are located. Injection ports are also located on the underside of the chamber.

128 To generate isoprene-derived SOA, the chamber was operated on sunny days, under high 129 relative humidity, to allow natural sunlight to trigger photochemical reactions. Acidified sulfate 130 seed aerosols were generated by nebulizing an aqueous solution containing 0.06 M MgSO<sub>4</sub> +  $0.06 \text{ M H}_2\text{SO}_4$  into the chamber to a particle concentration of approximately 170 µg m<sup>-3</sup>, which 131 132 was allowed to stabilize for 30 min to ensure a well-mixed condition. After stabilization, 3.5 133 ppmv isoprene (Sigma-Aldrich, 99%) and 200 ppbv NO (AirGas, 1.00%) were injected into the 134 chamber. Photochemical aging was allowed for approximately one hour to reach the desired exposure conditions of 30-40  $\mu$ g m<sup>-3</sup> growth of isoprene-derived SOA on the pre-existing 170  $\mu$ g 135 m<sup>-3</sup> of acidified sulfate aerosol. This chamber experiment was replicated on three separate sunny 136 137 days with temperatures ranging from 24.9°C to 26.8°C with a relative humidity of approximately 138 70% in the chamber.

139 **2.2 Control Chamber Experiments.** As a dark chamber control, to isolate the effect of SOA on 140 exposed cells, mixtures of isoprene, NO, and 170  $\mu$ g m<sup>-3</sup> of acidified sulfate seed aerosol were 141 injected into the chamber in the dark (after sunset). Conducting the chamber experiments in the 142 dark ensured no photochemical oxidation of isoprene. The dark control was replicated on three different nights. Except for the absence of solar radiation (no SOA), all chamber operations and
exposure conditions were similarly maintained.

As an added control to ensure that the device itself and the cell handling had no significant effect on cell cytotoxicity, cells were exposed in the EAVES to a clean chamber and compared to unexposed cells kept in an incubator for the same duration as the exposure. The cytotoxicity results ensured that there is no effect of chamber conditions and device operation on the cells.

150 **2.3 Cell Culture.** Human bronchial epithelial (BEAS-2B) cells were maintained in keratinocyte 151 growth medium (KGM BulletKit; Lonza), a serum-free keratinocyte basal medium (KBM) 152 supplemented with 0.004% of bovine pituitary extract and 0.001% of human epidermal growth 153 factor, insulin, hydrocortisone, and GA-1000 (gentamicin, amphotericin B), and passaged 154 weekly. Passage number for photochemical exposures and dark control exposures varied 155 between 52 and 60. Because BEAS-2B are an immortalized line of human bronchial epithelium, 156 there are limitations with its use such as it being genetically homogeneous, being a single cell 157 type, and being SV-40 transformed (Reddel et al., 1988). However, BEAS-2B is a stable, 158 proliferative cell line shown to be useful in airway inflammation studies such as ours (Devlin et 159 al., 1994).

**2.4 Direct Deposition Exposure.** In preparation for air-liquid interface exposures, cells were seeded onto collagen-coated Millicell cell culture inserts (30 mm diameter, 0.4  $\mu$ m pore size, 4.2 cm<sup>2</sup> filter area; Millipore, Cambridge, MA) at a density of 200,000 cells/well 24 hours prior to exposure. At the time of exposure, cells reached ~80% confluence, confirmed through microscopy. Immediately before exposure, cell medium was removed from the apical and basolateral sides of 2 seeded Millicell cell culture inserts. One insert was transferred to a titanium

dish containing 1.5 mL of keratinocyte basal medium (KBM; Lonza), supplying cells with nutrients from the basolateral side and constant moisture while allowing exposure to be performed at an air-liquid interface. The other insert was transferred into a 6 well plate with 2 mL of KBM and placed in the incubator as an unexposed control.

170 Cells were exposed to chamber-generated isoprene SOA using the EAVES located in the 171 laboratory directly beneath the outdoor chamber (de Bruijne et al., 2009; Lichtveld et al., 2012). 172 The EAVES, located in an incubator at 37°C, sampled chamber air at 1 L min<sup>-1</sup>. The target 173 relative humidity (RH) in the chamber during EAVES exposures was approximately 70%. 174 Exposure time was one hour commencing when target exposure conditions were achieved in the 175 outdoor chamber for both photochemical and dark control experiments. Detailed description of 176 the EAVES can be found in de Bruijne et al. (2009).

177 Following exposure, the cell culture insert was transferred to a 6-well tissue culture plate 178 containing 2 mL of fresh KBM. The control Millicell was also transferred to 2 mL of fresh 179 KBM. Nine hours post-exposure, extracellular medium was collected and total RNA was isolated 180 using Trizol (Life Technologies), consistent with past studies (de Bruijne et al., 2009). 181 Extracellular medium and the extracted RNA samples were stored at -20°C and -80°C, 182 respectively, until further analysis. For quality assurance purposes, the RNA concentration and 183 integrity were assessed using Nanodrop and Bioanalyzer over the period of storage. No changes 184 were observed under the given storage conditions.

185 2.5 Filter Resuspension Exposure. Chamber particles were collected, concurrently with 186 EAVES sampling, onto Teflon membrane filters (47 mm diameter, 1.0 µm pore size; Pall Life 187 Science) for photochemical (light) and dark chamber experiments to be used for chemical 188 analysis and resuspension exposures. The resuspension experiments served as a control for

189 possible effects of gaseous components such as ozone  $(O_3)$  and  $NO_x$  present in the direct 190 deposition experiments; however, prior studies have shown that gaseous components do not 191 yield cellular responses within the EAVES device (de Bruijne et al., 2009; Ebersviller et al., 192 2012a, b). Mass loadings of SOA collected on the filters were calculated from sampling volumes 193 and average aerosol mass concentrations in the chamber during the sampling period. A density correction of 1.6 g cm<sup>-3</sup> (Riedel et al., 2016) and 1.25 g cm<sup>-3</sup> (Kroll et al., 2006) was applied to 194 195 convert the measured volume concentrations to mass concentrations for the acidified sulfate seed 196 and SOA growth, respectively. The particles collected on Teflon filter membranes for 197 resuspension cell exposure were extracted by sonication in high-purity methanol (LC/MS 198 CHROMASOLV, Sigma-Aldrich). Filter samples from multiple experiments were combined and 199 the combined filter extract was dried under a gentle stream of nitrogen (N<sub>2</sub>). KBM medium was 200 then added into the extraction vials to re-dissolve SOA constituents.

In preparation for filter resuspension exposures, cells were seeded in 24-well plates at a density of  $2.5 \times 10^4$  cells/well in 250 µL of KGM 2 days prior to exposure. At the time of exposure when cells reached ~80% confluence, cells were washed twice with phosphate buffered saline (PBS) buffer, and then exposed to KBM containing 0.01 and 0.1 mg mL<sup>-1</sup> isoprene SOA extract from photochemical experiment and seed particles from dark control experiments.

Following a 9-hour exposure, extracellular medium was collected and total RNA was isolated using Trizol (Life Technologies) and stored alongside samples from direct deposition exposures until further analysis.

209 **2.6 Chemical and Physical Characterization of Exposures.** Online and offline techniques 210 were used to characterize the SOA generated in the chamber. The online techniques measured 211 the gas-phase species NO,  $NO_x$  and  $O_3$  and the physical properties of the aerosol continuously throughout the chamber experiments. Offline techniques measured aerosol-phase species collected onto Teflon membrane filters (47 mm diameter, 1.0 µm pore size; Pall Life Science) from photochemical and dark chamber experiments. Filter samples were stored in 20 mL scintillation vials protected from light at -20°C until analyses.

216 Real-time aerosol size distributions were measured using a Differential Mobility 217 Analyzer (DMA, Brechtel Manufacturing Inc.) coupled to a Mixing Condensation Particle 218 Counter (MCPC, Model 1710, Brechtel Manufacturing Inc.) located in the laboratory directly 219 underneath the chamber. O<sub>3</sub> and NO<sub>x</sub> were measured with a ML 9811 series Ozone Photometer 220 (Teledyne Monitor Labs, Englewood, CO) and ML 9841 series NO<sub>x</sub> Analyzer (American 221 Ecotech, Warren RI), respectively. Data were collected at one-minute intervals using a data 222 acquisition system (ChartScan/1400) interfaced to a computer. The presence of isoprene in the 223 chamber was confirmed and quantified using a Varian 3800 gas chromatograph (GC) equipped 224 with a flame ionization detector (FID).

225 Chemical characterization of SOA constituents was conducted offline from extracts of 226 filters collected from chamber experiments by gas chromatography interfaced with an electron 227 ionization quadrupole mass spectrometer (GC/EI-MS) or by ultra performance liquid 228 chromatography interfaced with a high-resolution quadrupole time-of-flight mass spectrometer 229 equipped with electrospray ionization (UPLC/ESI-HR-QTOFMS). Detailed operating conditions 230 for the GC/EI-MS and UPLC/ESI-HR-QTOFMS analyses as well as detailed filter extraction 231 protocols have been described previously by Lin et al. (2012). For GC/EI-MS analysis, filter 232 extracts were dried under a gentle stream of N<sub>2</sub> and trimethylsilylated by the addition of 100 µL 233 of BSTFA + TMCS (99:1 v/v, Supelco) and 50 µL of pyridine (anhydrous, 99.8%, Sigma-234 Aldrich) and heated at 70 °C for 1 h. For UPLC/ESI-HR-QTOFMS analysis, residues of filter

extracts were reconstituted with 150  $\mu$ L of a 50:50 (v/v) solvent mixture of high-purity water and methanol.

237 The isoprene-derived SOA markers: 2-methyltetrols, isomeric 3-methyltetrahydrofurans-238 3,4-diols (3-MeTHF-3,4-diols), and 2-methylglyceric acid, synthesized according to the 239 published procedures (Lin et al., 2013b; Zhang et al., 2012), were available in-house as authentic 240 standards to quantify the major components of isoprene SOA. 2-Methyltetrol organosulfates, 241 synthesized as a mixture of tetrabutylammonium salts, were also available as a standard. Purity was determined to be >99% by <sup>1</sup>H NMR and UPLC/ESI-QTOFMS analysis (Budisulistiorini et 242 243 al., 2015b). The C<sub>5</sub>-alkene triols and IEPOX dimer were quantified using the response factor 244 obtained for the synthetic 2-methyltetrols.

A representative ambient  $PM_{2.5}$  sample collected from the rural southeastern U.S. (Yorkville, GA) (Lin et al., 2013a) during the summer of 2010 was analyzed in an identical manner to confirm atmospheric relevance of the chamber-generated SOA constituents.

248 2.7 Cytotoxicity Assay. Cytotoxicity was assessed through measurement of lactate 249 dehydrogenase (LDH) released into the extracellular medium from damaged cells using the LDH 250 cytotoxicity detection kit (Takara). To ensure that the EAVES device itself and operation 251 procedure had no effect on cytotoxicity, the LDH release from cells exposed to clean chamber air 252 was measured. LDH release by cells exposed via the EAVES to the photochemically aged (light) 253 and non-photochemically aged (dark) particles was compared to release from unexposed cells 254 maintained in the incubator for the same duration. For the resuspension exposures, LDH release 255 by cells exposed to SOA through resuspended extract of photochemically aged and non-256 photochemically aged particles was compared to release by cells maintained in KBM only. 257 Additionally, LDH release from the light exposures, dark control, and resuspension exposures

was compared to release by positive control cells exposed to 1% Triton X-100 to ensure that celldeath would not affect gene expression results.

260 **2.8 Gene Expression Analysis.** We chose to measure the levels of the inflammation-related 261 mRNA in the BEAS-2B cells exposed to isoprene-derived SOA generated in our outdoor 262 chamber because various particle types are capable of sequestering cytokines (Seagrave, 2008). 263 Other direct deposition studies have also used mRNA transcripts as a proxy for cytokine 264 production (Hawley et al., 2014a; Hawley et al., 2014b; Hawley and Volckens, 2013; Volckens 265 et al., 2009; Lichtveld et al., 2012). Changes in IL-8 and COX-2 mRNA levels were measured using QuantiTect SYBR Green RT-PCR Kit (Qiagen) and QuantiTect Primer Assays for 266 267 Hs\_ACTB\_1\_SG (Catalog #QT00095431), Hs\_PTGS2\_1\_SG (Catalog #QT00040586), and 268 Hs\_CXCL8\_1\_SG (Catalog #QT00000322) for one-step RT-PCR analysis. All mRNA levels 269 were normalized against  $\beta$ -actin mRNA, which was used as a housekeeping gene. The relative 270 expression levels (i.e., fold change) of IL-8 and COX-2 were calculated using the comparative cycle threshold  $(2^{-\Delta\Delta CT})$  method (Livak and Schmittgen, 2001). For EAVES exposures, changes 271 272 in *IL-8* and *COX-2* from isoprene-derived SOA exposed cells were compared to cells exposed to 273 the dark controls. Similarly, for resuspension exposures changes in IL-8 and COX-2 from 274 isoprene-derived SOA exposed cells were compared to cells exposed to particles collected under 275 dark conditions.

276 **2.9 Statistical Analysis.** The software package GraphPad Prism 4 (GraphPad) was used for all 277 statistical analyses. All data were expressed as mean  $\pm$  SEM (standard error of means). 278 Comparisons between data sets for cytotoxicity and gene expression analysis were made using 279 unpaired *t*-test with Welch's correction. Significance was defined as p < 0.05.

#### 280 **3. Results and Discussion**

3.1 Physical and Chemical Characterization of Exposure. Figure 1 shows the change in particle mass concentration and gas (O<sub>3</sub>, NO, NO<sub>x</sub>) concentration over time during typical photochemical and dark control experiments. Under dark control conditions (Fig. 1a) there is no increase in aerosol mass concentration following isoprene injection. Average total aerosol mass concentration was  $155.0\pm2.69 \ \mu g \ m^{-3}$  (1 standard deviation) with no particle mass attributable to organic material.

In contrast, Fig. 1b shows an increase in aerosol mass concentration after 1 h post isoprene injection, which can be attributed to the photochemical oxidation of isoprene and subsequent production and reactive uptake of its oxidation products. The average increase in aerosol mass concentration attributable to SOA formation for three daylight chamber experiments conducted on separate days was  $44.5\pm5.7 \ \mu g \ m^{-3}$ . Average total aerosol mass concentration during particle exposure was  $173.1\pm4.2 \ \mu g \ m^{-3}$ .

O<sub>3</sub> and NO<sub>x</sub> concentrations measured during EAVES exposure were approximately 270 ppb and 120 ppb for photochemical experiments. For dark control experiments (e.g., Fig. 1a), the O<sub>3</sub> and NO<sub>x</sub> concentrations were approximately 15 ppb and 180 ppb. Previous studies characterizing the EAVES device show definitively that gas-phase products do not induce cell response (de Bruijne et al., 2009). However, resuspension exposures were conducted in addition to EAVES exposure to ensure that biological effects were attributable to only particle-phase constituents and not gas-phase products such as O<sub>3</sub> and NO<sub>x</sub>.

The chemical composition of aerosol, collected onto filters concurrently with cell exposure and characterized by GC/EI-MS and UPLC/ESI-HR-QTOFMS, are shown in Fig. 2. No isoprene-SOA tracers were observed in the filters collected from dark control experiments.

303 The dominant particle-phase products of the isoprene-SOA collected from photochemical 304 experiments are derived from the low-NO channel, where IEPOX reactive uptake onto acidic 305 sulfate aerosol dominates, including 2-methyltetrols, C5-alkene triols, isomeric 3-MeTHF-3,4-306 diols, IEPOX-derived dimers, and IEPOX-derived organosulfates. The sum of the IEPOX-307 derived SOA constituents quantified by the available standards accounted for ~80% of the 308 observed SOA mass. The MAE-derived SOA constituents 2-methylglyceric acid and the 309 organosulfate derivative of MAE, derived from the high-NO channel, accounted for 1.4% of the 310 observed SOA mass, confirming that particle-phase products generated were predominantly 311 formed from the reactive uptake of IEPOX onto acidic sulfate aerosols. As demonstrated in 312 Figure 2, all the same particle-phase products are measured in the  $PM_{2.5}$  sample collected in 313 Yorkville, GA (a typical low-NO region), demonstrating that the composition of the chamber-314 generated SOA is atmospherically relevant. Recent SOA tracer measurements from the Southern 315 Oxidant and Aerosol Study (SOAS) campaign at Look Rock, TN, Centerville, AL, and 316 Birmingham, AL, also support the atmospheric relevance of IEPOX-derived SOA constituents 317 that dominate the isoprene SOA mass in summer in the southeastern U.S. (Budisulistiorini et al., 318 2015a; Rattanavaraha et al., 2016).

**3.2 Cytotoxicity.** LDH release for cells exposed using the EAVES device is expressed as a foldchange relative to the unexposed incubator control. For resuspension exposures, LDH release is expressed as fold-change relative to cells exposed to KBM only. Results shown in Fig. 3a confirm that there is no effect of chamber conditions and device operation on the cells when comparing LDH release from cells exposed to a clean air chamber and cells unexposed in an incubator. Additionally, LDH release from all exposure conditions in EAVES exposed cells (Fig. 3b) and resuspension exposed cells (Fig. 3c) is negligible relative to positive controls exposed to 326 1% Triton X-100, confirming that the exposure concentration of isoprene-derived SOA utilized 327 in this study was not cytotoxic. All cytotoxicity results ensured that exposure conditions were not 328 adversely affecting the cells nor their gene expression.

329 **3.3 Pro-inflammatory Gene Expression**. Changes in the mRNA levels of *IL*-8 and *COX-2* 330 from cells exposed to isoprene-derived SOA using the EAVES are shown as fold-changes 331 relative to dark controls in Fig. 4. This comparison, as well as the results of the resuspension 332 experiment discussed below, ensure that all effects seen in the cells are attributable to the 333 isoprene-derived SOA and no other factors. A one-hour exposure to a mass concentration of approximately 45  $\mu$ g m<sup>-3</sup> of organic material was sufficient to significantly alter gene expression 334 335 of the inflammatory biomarkers in bronchial epithelial cells. Based on deposition efficiency characterized by de Bruijne et al. (2009), the estimated dose was 0.29  $\mu$ g cm<sup>-2</sup> of total particle 336 337 mass with 23% attributable to organic material formed from isoprene photooxidation (0.067 µg cm<sup>-2</sup> of SOA). 338

339 Changes in the mRNA levels of *IL-8* and *COX-2* from cells exposed to resuspended 340 isoprene-derived SOA collected from photochemical experiments are shown as fold-changes 341 relative to cells exposed to resuspended particles from dark control experiments in Fig. 5. At a low dose of 0.01 mg mL<sup>-1</sup> of isoprene SOA extract there is no significant increase in *IL*-8 and 342 343 COX-2 mRNA expression. The isoprene SOA extract, however, induces a response at a dose of 0.1 mg mL<sup>-1</sup>. The statistically significant increase in mRNA expression from the resuspension 344 exposure at 0.1 mg mL<sup>-1</sup> confirms that similar fold changes observed for both *IL*-8 and *COX-2* 345 346 from the EAVES exposures are not attributable to gaseous photooxidation products, such as  $O_3$ , 347 and support the characterization of the EAVES as a particle exposure device (de Bruijne et al., 348 2009).

349 The similar fold change observed in both the EAVES exposure and resuspension 350 exposure, in addition to confirming that the biological effects can be attributed to the particle-351 phase photochemical products (isoprene-derived SOA), suggests that exposure by resuspension 352 is appropriate for isoprene-derived SOA and may yield results similar to direct deposition 353 exposures. Unlike diesel particulate extracts, which agglomerate during resuspension exposures, 354 isoprene-derived SOA constituents are water-soluble based on reverse-phase LC separations 355 (Surratt et al., 2006; Lin et al., 2012) and remain well mixed in the cell medium used for 356 exposure. Therefore, resuspension exposures do not appear to be a limitation for toxicological 357 assessments of isoprene SOA.

358 **3.4 Biological Implications**. The goal of this study was to initially identify potential biological 359 response associated with exposure to isoprene-derived SOA by using a direct exposure device as 360 a model that has both atmospheric and physiological relevance. With this model, a dose of 0.067  $\mu$ g cm<sup>-2</sup> of isoprene SOA, induced statistically significant increases in *IL*-8 and *COX*-2 mRNA 361 362 levels in exposed BEAS-2B cells. There are many ways to classify in vitro particle dosimetry 363 based on the various properties of particles (Paur et al., 2011). For this direct deposition study, 364 we chose to classify dose as SOA mass deposition per surface area of the exposed cells to mimic 365 lung deposition. Gangwal et al. (2011) used a multiple-path particle dosimetry (MPPD) model to estimate that the lung deposition of ultrafine particles ranges from 0.006 to 0.02  $\mu$ g cm<sup>-2</sup> for a 24-366 hr exposure to a particle concentration of 0.1 mg m<sup>-3</sup>. Based on this estimate, a dose of 0.067  $\mu$ g 367 cm<sup>-2</sup> of isoprene SOA in our study can be considered a prolonged exposure over the course of a 368 369 week. In fact, most other in vitro studies require dosing cells at a high concentration sometimes 370 close to a lifetime exposure to obtain a cellular response. Despite this limitation, in vitro 371 exposures serve as a necessary screening tool for toxicity (Paur et al., 2011).

Our findings are consistent with other studies showing that photochemical oxidation of similar chemical mixtures increases toxicity in cell culture models and elevates expression of inflammatory biomarker genes (Lichtveld et al., 2012; Rager et al., 2011). Previous *in vitro* studies using a gas-phase only exposure system have shown that gas-phase products of isoprene photooxidation significantly enhance cytotoxicity and *IL-8* expression (Doyle et al., 2004; Doyle et al., 2007).

378 By choosing *IL-8* and *COX-2* as our genes of interest, we are able to compare our results 379 to other studies of known harmful particle exposures. In a similar study using the EAVES, normal human bronchial epithelial (NHBE) cells exposed to 1.10 µg cm<sup>-2</sup> diesel particulate 380 381 matter showed less than a 2-fold change over controls in both IL-8 and COX-2 mRNA 382 expression (Hawley et al., 2014b). In another study, A549 human lung epithelial cells were 383 exposed by direct deposition for 1 hour to photochemically-aged diesel exhaust particulates at a dose of 2.65 µg cm<sup>-2</sup> from a 1980 Mercedes or a 2006 Volkswagen (Lichtveld et al., 2012). 384 385 Exposure to aged Mercedes particulates induced a 4-fold change in IL-8 and ~2-fold change in 386 COX-2 mRNA expression, while exposure to aged Volkswagen particulates induced a change of 387 ~1.5-fold in IL-8 and 2-fold in COX-2 mRNA expression (Lichtveld et al., 2012). Although the 388 differences in cell types preclude direct comparisons, the finding of significant increases in COX-389 2 and IL-8 expression at doses much lower than reported for comparable increases in gene 390 expression levels induced by photochemically-aged diesel particulates is notable.

*IL-8* and *COX-2* are both linked to inflammation and oxidative stress (Kunkel et al.,
1991; Uchida, 2008). *IL-8* is a potent neutrophil chemotactic factor in the lung and its expression
by various cells plays a crucial role in neutrophil recruitment leading to lung inflammation
(Kunkel et al., 1991). *COX-2* is the inducible form of the cyclooxygenase enzyme, regulated by

cytokines and mitogens, and is responsible for prostaglandin synthesis associated with inflammation (FitzGerald, 2003). Consistent with the reports that *IL-8* and *COX-2* play important roles in lung inflammation (Nocker et al., 1996; Li et al., 2013), *in vivo* studies have shown that isoprene oxidation products cause airflow limitation and sensory irritation in mice (Rohr et al., 2003). In humans, the role of *IL-8* and *COX-2* in lung inflammation can be associated with diseases such as chronic obstructive pulmonary disease and asthma (Nocker et al., 1996; Peng et al., 2008; Fong et al., 2000).

402 The mechanism by which isoprene-SOA causes elevation of the inflammatory markers 403 IL-8 and COX-2 is not yet fully understood. However, recent work from our laboratory using the 404 acellular dithiothreitol (DTT) assay demonstrated that isoprene-derived SOA has significant 405 ROS generation potential (Kramer et al., 2016). High levels of ROS in cells can overwhelm the 406 antioxidant defense and lead to cellular oxidative stress (Sies, 1991; Bowler and Crapo, 2002; Li 407 et al., 2003). Following the discovery of the potential importance of isoprene-SOA in generating 408 ROS, Lin et al. (2016) showed that isoprene-SOA formed from the reactive uptake of epoxides 409 alters levels of oxidative stress-associated genes, including COX-2 in human lung cells. 410 Oxidative stress caused by ROS plays a major role in lung inflammation and the induction of 411 oxidative stress can lead to IL-8 expression (Tao et al., 2003; Yan et al., 2015). Specifically, 412 oxidants can activate the transcription factor NF-kB, which regulates a wide range of 413 inflammatory genes including IL-8 and COX-2 (Barnes and Adcock, 1997; Schreck et al., 1992). 414 Therefore, isoprene-SOA may cause increases in both IL-8 and COX-2 primarily through an 415 oxidative stress response. Additionally, the relationship between IL-8 and COX-2 can also 416 explain the observed increase in IL-8 gene expression as the production of IL-8 can be stimulated 417 through a COX-2 dependent mechanism in airway epithelial cells (Peng et al., 2008).

418 *In vitro* studies such as this one using a direct deposition model cannot fully elucidate 419 mechanisms of lung inflammation and potential pathogenesis but serve as a necessary part of 420 hazard characterization, particularly for a complex air mixture that has not been fully studied 421 (Hayashi, 2005; Paur et al., 2011). Ozone exposure studies have shown that comparable dose and 422 effect measurements for IL-8 and COX-2 can be found between in vivo and in vitro exposures 423 which add promise to extrapolating effects seen in vitro to effects in vivo (Hatch et al., 2014). In 424 vivo effects associated with isoprene-SOA exposure in vitro cannot be inferred as it is a different 425 system from ozone, so further *in vitro* studies exploring the health implication of the elevation of 426 *IL-8* and *COX-2* due specifically to isoprene-SOA exposure are necessary and may in turn justify 427 further extension to *in vivo* work.

#### 428 **4.** Conclusions

429 This study indicates that an atmospherically relevant composition of isoprene-derived 430 SOA is capable of increasing the expression of *IL-8* and *COX-2* in human bronchial epithelial 431 cells. The present study is an initial step in a long planned analysis of the biological impacts of 432 isoprene SOA exposure on lung cells. The SOA were generated as NO levels approached zero, 433 which represents conditions characteristic of urban locales downwind of rural isoprene sources. 434 As shown in Fig. 2, the aerosol generated for exposures in this study are chemically similar to 435 fine aerosol samples collected from the Southeastern U.S., which indicates that the chamber 436 exposures are representative of exposures that may be encountered by populations in regions 437 where isoprene emissions interact with anthropogenic pollutants. The same particle-phase 438 products found in our photochemical experiments have been measured in significant quantities 439 (accounting on average for 33% of fine organic aerosol mass) in ambient fine organic particles 440 collected in the Southeastern U.S. (Lin et al., 2013b; Budisulistiorini et al., 2013; Rattanavaraha et al., 2016; Budisulistiorini et al., 2016) and in other isoprene-rich environments (Hu et al.,
2015). The results of this study show that, because of its abundance, isoprene SOA may be a
public health concern warranting further toxicological investigation through *in vitro* or *in vivo*work.

445

#### 446 Acknowledgements

447 Research described in this article was conducted under contract to the Health Effects Institute 448 (HEI), an organization jointly funded by the United States Environmental Protection Agency 449 (EPA) (Assistance Award No. R-82811201), and certain motor vehicle and engine 450 manufacturers. The contents of this article do not necessarily reflect the views of HEI, or its 451 sponsors, nor do they necessarily reflect the views and policies of the EPA or motor vehicle and 452 engine manufacturers. M. A. was supported by a graduate fellowship provided by the National 453 Science Foundation (DGE-0646083), from the Center for Faculty Excellence, University of North Carolina at Chapel Hill, and in part by a grant from the National Institute of 454 Environmental Health Sciences (T32-ES007018). 455

456

#### 457 **References**

- Barnes, P. J., and Adcock, I. M.: NF-kB: a pivotal role in asthma and a new target for therapy,
  Am. J. Physiol, 265, C577-506, 1997.
- Bowler, R. P., and Crapo, J. D.: Oxidative stress in allergic respiratory diseases, Journal of
  Allergy and Clinical Immunology, 110, 349-356, http://dx.doi.org/10.1067/mai.2002.126780,
  2002.
- 463 Budisulistiorini, S. H., Canagaratna, M. R., Croteau, P. L., Marth, W. J., Baumann, K., Edgerton,
- 464 E. S., Shaw, S. L., Knipping, E. M., Worsnop, D. R., Jayne, J. T., Gold, A., and Surratt, J. D.:
- 465 Real-Time Continuous Characterization of Secondary Organic Aerosol Derived from Isoprene
- 466 Epoxydiols in Downtown Atlanta, Georgia, Using the Aerodyne Aerosol Chemical Speciation
- 467 Monitor, Environmental science & technology, 47, 5686-5694, 10.1021/es400023n, 2013.

- 468 Budisulistiorini, S. H., Baumann, K., Edgerton, E. S., Bairai, S. T., Mueller, S., Shaw, S. L.,
- 469 Knipping, E. M., Gold, A., and Surratt, J. D.: Seasonal characterization of submicron aerosol
- 470 chemical composition and organic aerosol sources in the southeastern United States: Atlanta,
- 471 Georgia and Look Rock, Tennessee, Atmos. Chem. Phys. Discuss., 2015, 22379-22417,
- 472 10.5194/acpd-15-22379-2015, 2015a.
- 473 Budisulistiorini, S. H., Li, X., Bairai, S. T., Renfro, J., Liu, Y., Liu, Y. J., McKinney, K. A.,
- 474 Martin, S. T., McNeill, V. F., Pye, H. O. T., Nenes, A., Neff, M. E., Stone, E. A., Mueller, S.,
- 475 Knote, C., Shaw, S. L., Zhang, Z., Gold, A., and Surratt, J. D.: Examining the effects of
- 476 anthropogenic emissions on isoprene-derived secondary organic aerosol formation during the
- 477 2013 Southern Oxidant and Aerosol Study (SOAS) at the Look Rock, Tennessee ground site,
- 478 Atmos. Chem. Phys., 15, 8871-8888, 10.5194/acp-15-8871-2015, 2015b.
- 479 Budisulistiorini, S. H., Baumann, K., Edgerton, E. S., Bairai, S. T., Mueller, S., Shaw, S. L.,
- 480 Knipping, E. M., Gold, A., and Surratt, J. D.: Seasonal characterization of submicron aerosol
- 481 chemical composition and organic aerosol sources in the southeastern United States: Atlanta,
- 482 Georgia, and Look Rock, Tennessee, Atmos. Chem. Phys., 16, 5171-5189, 10.5194/acp-16-5171-
- 483 2016, 2016.
- 484 Carlton, A. G., Wiedinmyer, C., and Kroll, J. H.: A review of Secondary Organic Aerosol (SOA)
  485 formation from isoprene, Atmospheric Chemistry and Physics, 9, 4987-5005, 2009.
- 486 de Bruijne, K., Ebersviller, S., Sexton, K. G., Lake, S., Leith, D., Goodman, R., Jetters, J.,
- 487 Walters, G. W., Doyle-Eisele, M., Woodside, R., Jeffries, H. E., and Jaspers, I.: Design and
- 488 Testing of Electrostatic Aerosol In Vitro Exposure System (EAVES): An Alternative Exposure
- 489 System for Particles, Inhalation toxicology, 21, 91-101, 10.1080/08958370802166035, 2009.
- 490 Devlin, R. B., McKinnon, K. P., Noah, T., Becker, S., and Koren, H. S.: Ozone-induced release
- 491 of cytokines and fibronectin by alveolar macrophages and airway epithelial cells, American
- 492 Journal of Physiology Lung Cellular and Molecular Physiology, 266, L612-L619, 1994.
- 493 Dockery, D. W., Pope, C. A., Xu, X. P., Spengler, J. D., Ware, J. H., Fay, M. E., Ferris, B. G.,
- 494 and Speizer, F. E.: An Association between Air-Pollution and Mortality in 6 United-States
- 495 Cities, New England Journal of Medicine, 329, 1753-1759, 10.1056/NEJM199312093292401,
- 496 1993.
- 497 Doyle, M., Sexton, K. G., Jeffries, H., Bridge, K., and Jaspers, I.: Effects of 1,3-butadiene,
- 498 isoprene, and their photochemical degradation products on human lung cells, Environmental
  499 health perspectives, 112, 1488-1495, 10.1289/ehp.7022, 2004.
- 500 Doyle, M., Sexton, K. G., Jeffries, H., and Jaspers, I.: Atmospheric photochemical
- 501 transformations enhance 1,3-butadiene-induced inflammatory responses in human epithelial
- 502 cells: The role of ozone and other photochemical degradation products, Chemico-Biological
- 503 Interactions, 166, 163-169, http://dx.doi.org/10.1016/j.cbi.2006.05.016, 2007.

# Ebersviller, S., Lichtveld, K., Sexton, K. G., Zavala, J., Lin, Y. H., Jaspers, I., and Jeffries, H. E.: Gaseous VOCs rapidly modify particulate matter and its biological effects – Part 1: Simple

- 506 VOCs and model PM, Atmos. Chem. Phys., 12, 12277-12292, 10.5194/acp-12-12277-2012,
  507 2012a.
- 508 Ebersviller, S., Lichtveld, K., Sexton, K. G., Zavala, J., Lin, Y. H., Jaspers, I., and Jeffries, H. E.:
- 509 Gaseous VOCs rapidly modify particulate matter and its biological effects Part 2: Complex
- 510 urban VOCs and model PM, Atmos. Chem. Phys., 12, 12293-12312, 10.5194/acp-12-12293-
- 511 2012, 2012b.
- FitzGerald, G. A.: COX-2 and beyond: approaches to prostaglandin inhibition in human disease,
  Nat Rev Drug Discov, 2, 879-890, 2003.
- 514 Fong, C. Y., Pang, L., Holland, E., and Knox, A. J.: TGF-β1 stimulates IL-8 release, COX-2
- 515 expression, and PGE2release in human airway smooth muscle cells, American Journal of
- 516 Physiology Lung Cellular and Molecular Physiology, 279, L201-L207, 2000.
- 517 Gangwal, S., Brown, J. S., Wang, A., Houck, K. A., Dix, D. J., Kavlock, R. J., and Hubal, E. A.
- 518 C.: Informing Selection of Nanomaterial Concentrations for ToxCast in Vitro Testing Based on
- 519 Occupational Exposure Potential, Environmental Health Perspectives, 119, 1539-1546,
- 520 10.1289/ehp.1103750, 2011.
- 521 Gaschen, A., Lang, D., Kalberer, M., Savi, M., Geiser, T., Gazdhar, A., Lehr, C.-M., Bur, M.,
- 522 Dommen, J., Baltensperger, U., and Geiser, M.: Cellular Responses after Exposure of Lung Cell
- 523 Cultures to Secondary Organic Aerosol Particles, Environmental science & technology, 44,
- 524 1424-1430, 10.1021/es902261m, 2010.
- 525 Gaston, C. J., Riedel, T. P., Zhang, Z. F., Gold, A., Surratt, J. D., and Thornton, J. A.: Reactive
- 526 Uptake of an Isoprene-Derived Epoxydiol to Submicron Aerosol Particles, Environmental
- 527 Science & Technology, 48, 11178-11186, 10.1021/es5034266, 2014.
- 528 Hatch, G. E., Duncan, K. E., Diaz-Sanchez, D., Schmitt, M. T., Ghio, A. J., Carraway, M. S.,
- 529 McKee, J., Dailey, L. A., Berntsen, J., and Devlin, R. B.: Progress in Assessing Air Pollutant
- Risks from In Vitro Exposures: Matching Ozone Dose and Effect in Human Airway Cells,
   Toxicological Sciences, 10.1093/toxsci/kfu115, 2014.
- Hawley, B., and Volckens, J.: Proinflammatory effects of cookstove emissions on human
  bronchial epithelial cells, Indoor air, 23, 4-13, 10.1111/j.1600-0668.2012.00790.x, 2013.
- Hawley, B., L'Orange, C., Olsen, D. B., Marchese, A. J., and Volckens, J.: Oxidative Stress and
- 535 Aromatic Hydrocarbon Response of Human Bronchial Epithelial Cells Exposed to Petro- or
- 536 Biodiesel Exhaust Treated with a Diesel Particulate Filter, Toxicological Sciences, 141, 505-514,
- 537 10.1093/toxsci/kfu147, 2014a.
- 538 Hawley, B., McKenna, D., Marchese, A., and Volckens, J.: Time course of bronchial cell
- 539 inflammation following exposure to diesel particulate matter using a modified EAVES,
- 540 Toxicology in Vitro, 28, 829-837, 10.1016/j.tiv.2014.03.001, 2014b.

- 541 Hayashi, Y.: Designing in vitro assay systems for hazard characterization. Basic strategies and
- related technical issues, Experimental and Toxicologic Pathology, 57, Supplement 1, 227-232,
- 543 http://dx.doi.org/10.1016/j.etp.2005.05.012, 2005.
- 544 Henze, D. K., Seinfeld, J. H., Ng, N. L., Kroll, J. H., Fu, T. M., Jacob, D. J., and Heald, C. L.:
- 545 Global modeling of secondary organic aerosol formation from aromatic hydrocarbons: high- vs. 546 low-yield pathways, Atmospheric Chemistry and Physics, 8, 2405-2420, 2008.
- 547 Hu, W. W., Campuzano-Jost, P., Palm, B. B., Day, D. A., Ortega, A. M., Hayes, P. L.,
- 548 Krechmer, J. E., Chen, Q., Kuwata, M., Liu, Y. J., de Sá, S. S., McKinney, K., Martin, S. T., Hu,
- 549 M., Budisulistiorini, S. H., Riva, M., Surratt, J. D., St. Clair, J. M., Isaacman-Van Wertz, G.,
- 550 Yee, L. D., Goldstein, A. H., Carbone, S., Brito, J., Artaxo, P., de Gouw, J. A., Koss, A.,
- 551 Wisthaler, A., Mikoviny, T., Karl, T., Kaser, L., Jud, W., Hansel, A., Docherty, K. S., Alexander,
- 552 M. L., Robinson, N. H., Coe, H., Allan, J. D., Canagaratna, M. R., Paulot, F., and Jimenez, J. L.:
- 553 Characterization of a real-time tracer for isoprene epoxydiols-derived secondary organic aerosol
- 554 (IEPOX-SOA) from aerosol mass spectrometer measurements, Atmos. Chem. Phys., 15, 11807-
- 555 11833, 10.5194/acp-15-11807-2015, 2015.
- Jang, M., Ghio, A. J., and Cao, G.: Exposure of BEAS-2B Cells to Secondary Organic Aerosol
- 557 Coated on Magnetic Nanoparticles, Chemical Research in Toxicology, 19, 1044-1050,
- 558 10.1021/tx0503597, 2006.
- 559 Kelly, F. J., and Fussell, J. C.: Size, source and chemical composition as determinants of toxicity
- attributable to ambient particulate matter, Atmospheric Environment, 60, 504-526,
- 561 10.1016/j.atmosenv.2012.06.039, 2012.
- 562 Kramer, A. J., Rattanavaraha, W., Zhang, Z., Gold, A., Surratt, J. D., and Lin, Y.-H.: Assessing
- the oxidative potential of isoprene-derived epoxides and secondary organic aerosol, Atmospheric
- 564 Environment, 130, 211-218, http://dx.doi.org/10.1016/j.atmosenv.2015.10.018, 2016.
- 565 Krechmer, J. E., Coggon, M. M., Massoli, P., Nguyen, T. B., Crounse, J. D., Hu, W., Day, D. A.,
- 566 Tyndall, G. S., Henze, D. K., Rivera-Rios, J. C., Nowak, J. B., Kimmel, J. R., Mauldin, R. L.,
- 567 Stark, H., Jayne, J. T., Sipilä, M., Junninen, H., Clair, J. M. S., Zhang, X., Feiner, P. A., Zhang,
- 568 L., Miller, D. O., Brune, W. H., Keutsch, F. N., Wennberg, P. O., Seinfeld, J. H., Worsnop, D.
- 569 R., Jimenez, J. L., and Canagaratna, M. R.: Formation of Low Volatility Organic Compounds
- and Secondary Organic Aerosol from Isoprene Hydroxyhydroperoxide Low-NO Oxidation,
- 571 Environmental Science & Technology, 49, 10330-10339, 10.1021/acs.est.5b02031, 2015.
- 572 Kroll, J. H., Ng, N. L., Murphy, S. M., Flagan, R. C., and Seinfeld, J. H.: Secondary organic
- aerosol formation from isoprene photooxidation, Environmental Science & Technology, 40,
- 574 1869-1877, 10.1021/es0524301, 2006.
- 575 Kunkel, S. L., Standiford, T., Kasahara, K., and Strieter, R. M.: Interleukin-8 (II-8) the Major
- 576 Neutrophil Chemotactic Factor in the Lung, Experimental lung research, 17, 17-23,
- 577 10.3109/01902149109063278, 1991.
- 578 Li, H., Edin, M. L., Bradbury, J. A., Graves, J. P., DeGraff, L. M., Gruzdev, A., Cheng, J.,
- 579 Dackor, R. T., Wang, P. M., Bortner, C. D., Garantziotis, S., Jetten, A. M., and Zeldin, D. C.:

- 580 Cyclooxygenase-2 Inhibits T Helper Cell Type 9 Differentiation during Allergic Lung
- 581 Inflammation via Down-regulation of IL-17RB, American Journal of Respiratory and Critical
- 582 Care Medicine, 187, 812-822, 10.1164/rccm.201211-2073OC, 2013.
- 583 Li, N., Hao, M., Phalen, R. F., Hinds, W. C., and Nel, A. E.: Particulate air pollutants and
- asthma: A paradigm for the role of oxidative stress in PM-induced adverse health effects,
- 585 Clinical Immunology, 109, 250-265, http://dx.doi.org/10.1016/j.clim.2003.08.006, 2003.
- 586 Lichtveld, K. M., Ebersviller, S. M., Sexton, K. G., Vizuete, W., Jaspers, I., and Jeffries, H. E.:
- 587 In Vitro Exposures in Diesel Exhaust Atmospheres: Resuspension of PM from Filters versus
- 588 Direct Deposition of PM from Air, Environmental science & technology, 46, 9062-9070,
  589 10.1021/es301431s, 2012.
- 590 Lim, S. S., Vos, T., Flaxman, A. D., Danaei, G., Shibuya, K., Adair-Rohani, H., AlMazroa, M. 591 A., Amann, M., Anderson, H. R., Andrews, K. G., Arvee, M., Atkinson, C., Bacchus, L. J., 592 Bahalim, A. N., Balakrishnan, K., Balmes, J., Barker-Collo, S., Baxter, A., Bell, M. L., Blore, J. 593 D., Blyth, F., Bonner, C., Borges, G., Bourne, R., Boussinesq, M., Brauer, M., Brooks, P., Bruce, 594 N. G., Brunekreef, B., Bryan-Hancock, C., Bucello, C., Buchbinder, R., Bull, F., Burnett, R. T., Byers, T. E., Calabria, B., Carapetis, J., Carnahan, E., Chafe, Z., Charlson, F., Chen, H., Chen, J. 595 596 S., Cheng, A. T.-A., Child, J. C., Cohen, A., Colson, K. E., Cowie, B. C., Darby, S., Darling, S., 597 Davis, A., Degenhardt, L., Dentener, F., Des Jarlais, D. C., Devries, K., Dherani, M., Ding, E. L., 598 Dorsey, E. R., Driscoll, T., Edmond, K., Ali, S. E., Engell, R. E., Erwin, P. J., Fahimi, S., Falder, 599 G., Farzadfar, F., Ferrari, A., Finucane, M. M., Flaxman, S., Fowkes, F. G. R., Freedman, G., 600 Freeman, M. K., Gakidou, E., Ghosh, S., Giovannucci, E., Gmel, G., Graham, K., Grainger, R., 601 Grant, B., Gunnell, D., Gutierrez, H. R., Hall, W., Hoek, H. W., Hogan, A., Hosgood Iii, H. D., 602 Hoy, D., Hu, H., Hubbell, B. J., Hutchings, S. J., Ibeanusi, S. E., Jacklyn, G. L., Jasrasaria, R., 603 Jonas, J. B., Kan, H., Kanis, J. A., Kassebaum, N., Kawakami, N., Khang, Y.-H., Khatibzadeh, 604 S., Khoo, J.-P., Kok, C., Laden, F., Lalloo, R., Lan, Q., Lathlean, T., Leasher, J. L., Leigh, J., Li, 605 Y., Lin, J. K., Lipshultz, S. E., London, S., Lozano, R., Lu, Y., Mak, J., Malekzadeh, R., 606 Mallinger, L., Marcenes, W., March, L., Marks, R., Martin, R., McGale, P., McGrath, J., Mehta, 607 S., Memish, Z. A., Mensah, G. A., Merriman, T. R., Micha, R., Michaud, C., Mishra, V., Hanafiah, K. M., Mokdad, A. A., Morawska, L., Mozaffarian, D., Murphy, T., Naghavi, M., 608 609 Neal, B., Nelson, P. K., Nolla, J. M., Norman, R., Olives, C., Omer, S. B., Orchard, J., Osborne, 610 R., Ostro, B., Page, A., Pandey, K. D., Parry, C. D. H., Passmore, E., Patra, J., Pearce, N., 611 Pelizzari, P. M., Petzold, M., Phillips, M. R., Pope, D., Pope Iii, C. A., Powles, J., Rao, M., 612 Razavi, H., Rehfuess, E. A., Rehm, J. T., Ritz, B., Rivara, F. P., Roberts, T., Robinson, C., 613 Rodriguez-Portales, J. A., Romieu, I., Room, R., Rosenfeld, L. C., Roy, A., Rushton, L., 614 Salomon, J. A., Sampson, U., Sanchez-Riera, L., Sanman, E., Sapkota, A., Seedat, S., Shi, P., Shield, K., Shivakoti, R., Singh, G. M., Sleet, D. A., Smith, E., Smith, K. R., Stapelberg, N. J. 615 616 C., Steenland, K., Stöckl, H., Stovner, L. J., Straif, K., Straney, L., Thurston, G. D., Tran, J. H., Van Dingenen, R., van Donkelaar, A., Veerman, J. L., Vijayakumar, L., Weintraub, R., 617 618 Weissman, M. M., White, R. A., Whiteford, H., Wiersma, S. T., Wilkinson, J. D., Williams, H. 619 C., Williams, W., Wilson, N., Woolf, A. D., Yip, P., Zielinski, J. M., Lopez, A. D., Murray, C. J. 620 L., and Ezzati, M.: A comparative risk assessment of burden of disease and injury attributable to 67 risk factors and risk factor clusters in 21 regions, 1990-2010: a systematic analysis for the 621
- 622 Global Burden of Disease Study 2010, The Lancet, 380, 2224-2260,
- 623 http://dx.doi.org/10.1016/S0140-6736(12)61766-8, 2012.

- 624 Lin, Y.-H., Zhang, Z., Docherty, K. S., Zhang, H., Budisulistiorini, S. H., Rubitschun, C. L.,
- 625 Shaw, S. L., Knipping, E. M., Edgerton, E. S., Kleindienst, T. E., Gold, A., and Surratt, J. D.:
- 626 Isoprene Epoxydiols as Precursors to Secondary Organic Aerosol Formation: Acid-Catalyzed
- 627 Reactive Uptake Studies with Authentic Compounds, Environmental science & technology, 46,
- 628 250-258, 10.1021/es202554c, 2012.
- 629 Lin, Y.-H., Knipping, E. M., Edgerton, E. S., Shaw, S. L., and Surratt, J. D.: Investigating the
- 630 influences of SO2 and NH3 levels on isoprene-derived secondary organic aerosol formation
- 631 using conditional sampling approaches, Atmos. Chem. Phys., 13, 8457-8470, 10.5194/acp-13-
- 632 8457-2013, 2013a.
- 633 Lin, Y.-H., Zhang, H., Pye, H. O. T., Zhang, Z., Marth, W. J., Park, S., Arashiro, M., Cui, T.,
- 634 Budisulistiorini, S. H., Sexton, K. G., Vizuete, W., Xie, Y., Luecken, D. J., Piletic, I. R., Edney,
- 635 E. O., Bartolotti, L. J., Gold, A., and Surratt, J. D.: Epoxide as a precursor to secondary organic
- 636 aerosol formation from isoprene photooxidation in the presence of nitrogen oxides, Proceedings
- 637 of the National Academy of Sciences of the United States of America, 110, 6718-6723,
- 638 10.1073/pnas.1221150110, 2013b.
- 639 Lin, Y.-H., Arashiro, M., Martin, E., Chen, Y., Zhang, Z., Sexton, K. G., Gold, A., Jaspers, I.,
- 640 Fry, R. C., and Surratt, J. D.: Isoprene-Derived Secondary Organic Aerosol Induces the
- 641 Expression of Oxidative Stress Response Genes in Human Lung Cells, Environmental Science &
- 642 Technology Letters, 3, 250-254, 10.1021/acs.estlett.6b00151, 2016.
- 643 Livak, K. J., and Schmittgen, T. D.: Analysis of Relative Gene Expression Data Using Real-
- 644 Time Quantitative PCR and the  $2-\Delta\Delta CT$  Method, Methods, 25, 402-408,
- 645 http://dx.doi.org/10.1006/meth.2001.1262, 2001.
- 646 Nguyen, T. B., Bates, K. H., Crounse, J. D., Schwantes, R. H., Zhang, X., Kjaergaard, H. G.,
- 647 Surratt, J. D., Lin, P., Laskin, A., and Seinfeld, J. H.: Mechanism of the hydroxyl radical
- oxidation of methacryloyl peroxynitrate (MPAN) and its pathway toward secondary organic 648
- 649 aerosol formation in the atmosphere, Physical Chemistry Chemical Physics, 17, 17914-17926, 650
- 2015.
- 651 Nocker, R. E. T., Schoonbrood, D. F. M., vandeGraaf, E. A., Hack, C. E., Lutter, R., Jansen, H.
- M., and Out, T. A.: Interleukin-8 in airway inflammation in patients with asthma and chronic 652
- 653 obstructive pulmonary disease, Int. Arch. Allergy Immunol., 109, 183-191, 1996.
- 654 Paulot, F., Crounse, J. D., Kjaergaard, H. G., Kuerten, A., St Clair, J. M., Seinfeld, J. H., and
- 655 Wennberg, P. O.: Unexpected Epoxide Formation in the Gas-Phase Photooxidation of Isoprene,
- 656 Science, 325, 730-733, 10.1126/science.1172910, 2009.
- 657 Paur, H.-R., Cassee, F. R., Teeguarden, J., Fissan, H., Diabate, S., Aufderheide, M., Kreyling, W.
- 658 G., Hänninen, O., Kasper, G., Riediker, M., Rothen-Rutishauser, B., and Schmid, O.: In-vitro
- cell exposure studies for the assessment of nanoparticle toxicity in the lung-A dialog between 659
- 660 aerosol science and biology, Journal of Aerosol Science, 42, 668-692,
- 661 http://dx.doi.org/10.1016/j.jaerosci.2011.06.005, 2011.

- Peng, H., Chen, P., Cai, Y., Chen, Y., Wu, Q.-h., Li, Y., Zhou, R., and Fang, X.: Endothelin-1
- 663 increases expression of cyclooxygenase-2 and production of interlukin-8 in hunan pulmonary
- 664 epithelial cells, Peptides, 29, 419-424, http://dx.doi.org/10.1016/j.peptides.2007.11.015, 2008.

665 Pye, H. O. T., Pinder, R. W., Piletic, I. R., Xie, Y., Capps, S. L., Lin, Y.-H., Surratt, J. D., Zhang,

- Context States Contex
- 667 Lewandowski, M., and Edney, E. O.: Epoxide Pathways Improve Model Predictions of Isoprene
- 668 Markers and Reveal Key Role of Acidity in Aerosol Formation, Environmental science &
- 669 technology, 47, 11056-11064, 10.1021/es402106h, 2013.
- 670 Rager, J. E., Lichtveld, K., Ebersviller, S., Smeester, L., Jaspers, I., Sexton, K. G., and Fry, R.
- 671 C.: A Toxicogenomic Comparison of Primary and Photochemically Altered Air Pollutant
- 672 Mixtures, Environmental health perspectives, 119, 1583-1589, 10.1289/ehp.1003323, 2011.
- 673 Rattanavaraha, W., Chu, K., Budisulistiorini, S. H., Riva, M., Lin, Y. H., Edgerton, E. S.,
- Baumann, K., Shaw, S. L., Guo, H., King, L., Weber, R. J., Neff, M. E., Stone, E. A., Offenberg,
- J. H., Zhang, Z., Gold, A., and Surratt, J. D.: Assessing the impact of anthropogenic pollution on
- 676 isoprene-derived secondary organic aerosol formation in PM2.5 collected from the Birmingham,
- Alabama, ground site during the 2013 Southern Oxidant and Aerosol Study, Atmos. Chem.
- 678 Phys., 16, 4897-4914, 10.5194/acp-16-4897-2016, 2016.
- 679 Reddel, R. R., Ke, Y., Gerwin, B. I., McMenamin, M. G., Lechner, J. F., Su, R. T., Brash, D. E.,
- 680 Park, J.-B., Rhim, J. S., and Harris, C. C.: Transformation of Human Bronchial Epithelial Cells
- by Infection with SV40 or Adenovirus-12 SV40 Hybrid Virus, or Transfection via Strontium
- 682 Phosphate Coprecipitation with a Plasmid Containing SV40 Early Region Genes, Cancer
- 683 Research, 48, 1904-1909, 1988.
- 684 Reuter, S., Gupta, S. C., Chaturvedi, M. M., and Aggarwal, B. B.: Oxidative stress,
- inflammation, and cancer: How are they linked?, Free Radical Biology and Medicine, 49, 1603 1616, http://dx.doi.org/10.1016/j.freeradbiomed.2010.09.006, 2010.
- 687 Riedel, T. P., Lin, Y.-H., Budisulistiorini, S. H., Gaston, C. J., Thornton, J. A., Zhang, Z.,
- 688 Vizuete, W., Gold, A., and Surratt, J. D.: Heterogeneous Reactions of Isoprene-Derived
- 689 Epoxides: Reaction Probabilities and Molar Secondary Organic Aerosol Yield Estimates,
- 690 Environmental Science & Technology Letters, 10.1021/ez500406f, 2015.
- 691 Riedel, T. P., Lin, Y. H., Zhang, Z., Chu, K., Thornton, J. A., Vizuete, W., Gold, A., and Surratt,
- 692 J. D.: Constraining condensed-phase formation kinetics of secondary organic aerosol
- 693 components from isoprene epoxydiols, Atmos. Chem. Phys., 16, 1245-1254, 10.5194/acp-16-
- 6941245-2016, 2016.
- Rohr, A. C., Shore, S. A., and Spengler, J. D.: Repeated exposure to isoprene oxidation products
- causes enhanced respiratory tract effects in multiple murine strains, Inhalation toxicology, 15,
  1191-1207, 10.1080/08958370390229870, 2003.
- Rohr, A. C.: The health significance of gas- and particle-phase terpene oxidation products: A
   review, Environment International, 60, 145-162, 10.1016/j.envint.2013.08.002, 2013.

- Samet, J. M., Dominici, F., Curriero, F. C., Coursac, I., and Zeger, S. L.: Fine particulate air
- pollution and mortality in 20 US Cities, 1987-1994, New England Journal of Medicine, 343,
   1742-1749, 10.1056/NEJM200012143432401, 2000.
- 703 Schreck, R., Albermann, K., and Baeuerle, P. A.: Nuclear Factor Kb: An Oxidative Stress-
- 704 Responsive Transcription Factor of Eukaryotic Cells (A Review), Free Radical Research
- 705 Communications, 17, 221-237, 10.3109/10715769209079515, 1992.
- 706 Schwartz, J., Slater, D., Larson, T. V., Pierson, W. E., and Koenig, J. Q.: Particulate Air-
- 707 Pollution and Hospital Emergency Room Visits for Asthma in Seattle, American Review of
- 708 Respiratory Disease, 147, 826-831, 1993.
- 709 Seagrave, J.: Mechanisms and implications of air pollution particle associations with
- chemokines, Toxicology and Applied Pharmacology, 232, 469-477, 10.1016/j.taap.2008.08.001,
  2008.
- 712 Sies, H.: Oxidants And Antioxidants: Pathophysiologic Determinants and Therapeutic
- 713 AgentsOxidative stress: From basic research to clinical application, The American Journal of
- 714 Medicine, 91, S31-S38, http://dx.doi.org/10.1016/0002-9343(91)90281-2, 1991.
- 715 Surratt, J. D., Murphy, S. M., Kroll, J. H., Ng, N. L., Hildebrandt, L., Sorooshian, A.,
- 716 Szmigielski, R., Vermeylen, R., Maenhaut, W., and Claeys, M.: Chemical composition of
- secondary organic aerosol formed from the photooxidation of isoprene, The Journal of Physical
- 718 Chemistry A, 110, 9665-9690, 2006.
- Surratt, J. D., Kroll, J. H., Kleindienst, T. E., Edney, E. O., Claeys, M., Sorooshian, A., Ng, N.
- L., Offenberg, J. H., Lewandowski, M., Jaoui, M., Flagan, R. C., and Seinfeld, J. H.: Evidence
- for organosulfates in secondary organic aerosol, Environmental science & technology, 41, 517-
- 722 527, 10.1021/es062081q, 2007.
- 723 Surratt, J. D., Chan, A. W. H., Eddingsaas, N. C., Chan, M., Loza, C. L., Kwan, A. J., Hersey, S.
- P., Flagan, R. C., Wennberg, P. O., and Seinfeld, J. H.: Reactive intermediates revealed in
- secondary organic aerosol formation from isoprene, Proceedings of the National Academy of
- 726 Sciences of the United States of America, 107, 6640-6645, 10.1073/pnas.0911114107, 2010.
- Tao, F., Gonzalez-Flecha, B., and Kobzik, L.: Reactive oxygen species in pulmonary
- inflammation by ambient particulates, Free Radical Biology and Medicine, 35, 327-340,
- 729 http://dx.doi.org/10.1016/S0891-5849(03)00280-6, 2003.
- Uchida, K.: A Lipid-derived Endogenous Inducer of COX-2: a Bridge Between Inflammationand Oxidative Stress, Mol. Cells, 25, 347-351, 2008.
- Volckens, J., Dailey, L., Walters, G., and Devlin, R. B.: Direct Particle-to-Cell Deposition of
- 733 Coarse Ambient Particulate Matter Increases the Production of Inflammatory Mediators from
- Cultured Human Airway Epithelial Cells, Environmental science & technology, 43, 4595-4599,
- 735 10.1021/es900698a, 2009.

- 736 Xu, L., Guo, H. Y., Boyd, C. M., Klein, M., Bougiatioti, A., Cerully, K. M., Hite, J. R.,
- 737 Isaacman-VanWertz, G., Kreisberg, N. M., Knote, C., Olson, K., Koss, A., Goldstein, A. H.,
- Hering, S. V., de Gouw, J., Baumann, K., Lee, S. H., Nenes, A., Weber, R. J., and Ng, N. L.:
- 739 Effects of anthropogenic emissions on aerosol formation from isoprene and monoterpenes in the
- 740southeastern United States, Proceedings of the National Academy of Sciences of the United
- 741 States of America, 112, 37-42, 10.1073/pnas.1417609112, 2015.
- Yan, Z., Wang, J., Li, J., Jiang, N., Zhang, R., Yang, W., Yao, W., and Wu, W.: Oxidative stress
- and endocytosis are involved in upregulation of interleukin-8 expression in airway cells exposed
- 744 to PM2.5, Environmental Toxicology, n/a-n/a, 10.1002/tox.22188, 2015.
- 745 Zavala, J., Lichtveld, K., Ebersviller, S., Carson, J. L., Walters, G. W., Jaspers, I., Jeffries, H. E.,
- 746 Sexton, K. G., and Vizuete, W.: The Gillings Sampler An electrostatic air sampler as an
- alternative method for aerosol in vitro exposure studies, Chemico-Biological Interactions, 220,
- 748 158-168, http://dx.doi.org/10.1016/j.cbi.2014.06.026, 2014.
- 749 Zhang, Z., Lin, Y.-H., Zhang, H., Surratt, J. D., Ball, L. M., and Gold, A.: Technical Note:
- 750 Synthesis of isoprene atmospheric oxidation products: isomeric epoxydiols and the
- rearrangement products cis- and trans-3-methyl-3,4-dihydroxytetrahydrofuran, Atmospheric
- 752 Chemistry and Physics, 12, 8529-8535, 10.5194/acp-12-8529-2012, 2012.

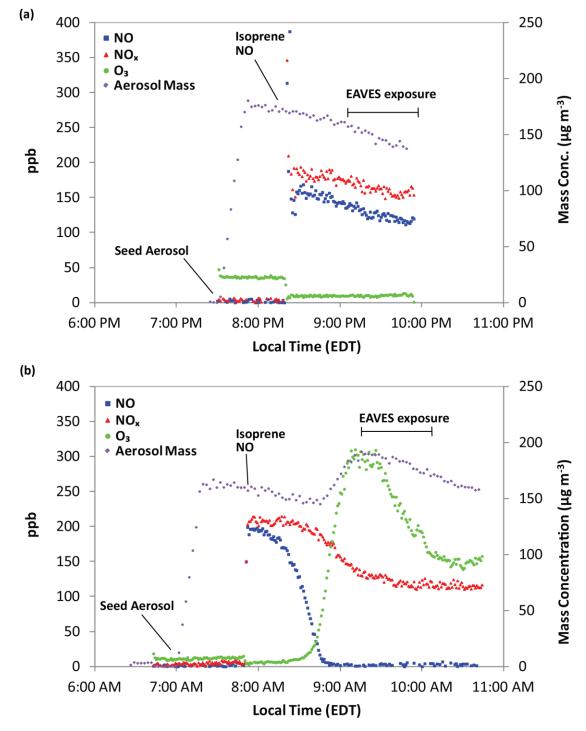


Figure 1. Aerosol mass concentration and gas-phase product concentrations over time for (a)
dark control chamber experiment and (b) photochemically produced isoprene-derived SOA
exposure chamber experiment.

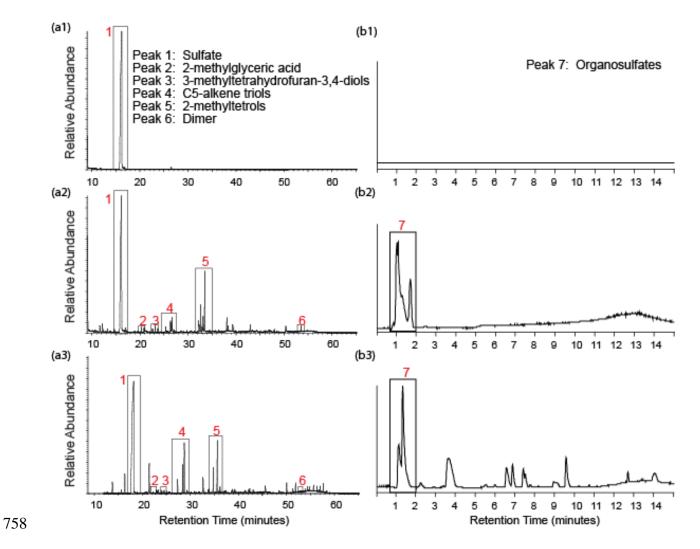


Figure 2. (a) GC/EI-MS total ion chromatograms (TICs) and (b) UPLC/ESI-HR-QTOFMS base
peak chromatograms (BPCs) from a (1) dark control chamber experiment, (2) isoprene-derived
SOA exposure chamber experiment, and (3) PM<sub>2.5</sub> sample collected from Yorkville, GA during
summer 2010.

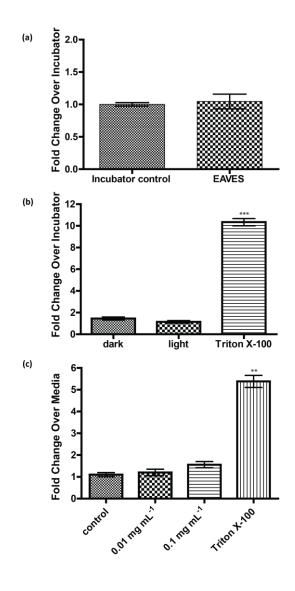


Figure 3. LDH release for (a) clean air controls, (b) EAVES exposures, normalized to incubator
control, and (c) resuspension exposures, normalized to KBM only control. \*\*p<0.005 and</li>
\*\*\*p<0.0005.</li>

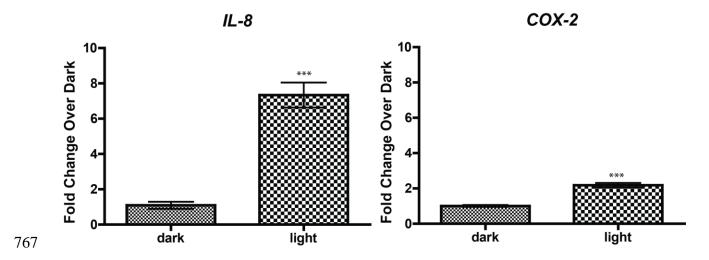


Figure 4. *IL-8* and *COX-2* mRNA expression induced by exposure to isoprene-derived SOA

vising EAVES device all normalized to dark control experiments and against housekeeping gene,

- 770 β-actin. All experiments conducted in triplicate. \*\*\*p<0.0005.
- 771

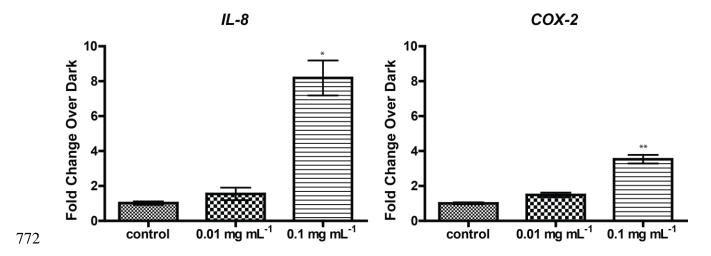


Figure 5. *IL-8* and *COX-2* expression induced by exposure to isoprene-derived SOA using

resuspension method all normalized to dark control experiments and against housekeeping gene,

