

1 ***In Vitro* Exposure to Isoprene-Derived Secondary Organic Aerosol by Direct Deposition**  
2 **and its Effects on *COX-2* and *IL-8* Gene Expression**

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30 **Abstract**

31 Atmospheric oxidation of isoprene, the most abundant non-methane hydrocarbon emitted into  
32 Earth's atmosphere primarily from terrestrial vegetation, is now recognized as a major  
33 contributor to the global secondary organic aerosol (SOA) burden. Anthropogenic pollutants  
34 significantly enhance isoprene SOA formation through acid-catalyzed heterogeneous chemistry  
35 of epoxide products. Since isoprene SOA formation as a source of fine aerosol is a relatively  
36 recent discovery, research is lacking on evaluating its potential adverse effects on human health.  
37 The objective of this study was to examine the effect of isoprene-derived SOA on inflammation-  
38 associated gene expression in human lung cells using a direct deposition exposure method. We  
39 assessed altered expression of inflammation-related genes in human bronchial epithelial cells  
40 (BEAS-2B) exposed to isoprene-derived SOA generated in an outdoor chamber facility.  
41 Measurements of gene expression of known inflammatory biomarkers interleukin 8 (IL-8) and  
42 cyclooxygenase 2 (COX-2) in exposed cells, together with complementary chemical  
43 measurements, showed that a dose of  $0.067 \mu\text{g cm}^{-2}$  of SOA from isoprene photooxidation leads  
44 to statistically significant increases in *IL-8* and *COX-2* mRNA levels. Resuspension exposures  
45 using aerosol filter extracts corroborated these findings, supporting the conclusion that isoprene-  
46 derived SOA constituents induce the observed changes in mRNA levels. The present study is an  
47 attempt to examine the early biological responses of isoprene SOA exposure in human lung cells.

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## 52 **1. Introduction**

53           Recent work has shown that isoprene (2-methyl-1,3-butadiene) is an important precursor  
54 of secondary organic aerosol (SOA), which has potential impacts on climate change and public  
55 health (Lin et al., 2013b; Rohr, 2013; Lin et al., 2016). Current understanding of isoprene SOA  
56 formation is based on laboratory studies showing that gas-phase photooxidation of isoprene  
57 generates key SOA precursors, including isomeric isoprene epoxydiols (IEPOX), methacrylic  
58 acid epoxide (MAE), hydroxymethyl-methyl- $\alpha$ -lactone (HMML), and isoprene  
59 hydroxyhydroperoxides (ISOPOOH) (Paulot et al., 2009; Surratt et al., 2010; Lin et al., 2012;  
60 Lin et al., 2013b; Nguyen et al., 2015; Krechmer et al., 2015). The formation of SOA from these  
61 precursors is influenced by controllable anthropogenic emissions such as oxides of nitrogen  
62 ( $\text{NO}_x$ ) and sulfur dioxide ( $\text{SO}_2$ ). Atmospheric oxidation of  $\text{SO}_2$  contributes to particle acidity,  
63 which enhances isoprene SOA formation through acid-catalyzed reactive uptake and multiphase  
64 chemistry of IEPOX and MAE (Surratt et al., 2007; Surratt et al., 2010; Lin et al., 2012; Gaston  
65 et al., 2014; Riedel et al., 2015), while  $\text{NO}_x$  determines whether the oxidation pathway leading to  
66 IEPOX or MAE/HMML predominates (Lin et al., 2013b; Surratt et al., 2010; Nguyen et al.,  
67 2015). Isoprene SOA comprises a large portion of global atmospheric fine particles ( $\text{PM}_{2.5}$ ,  
68 aerosol with aerodynamic diameters  $\leq 2.5 \mu\text{m}$ ) (Carlton et al., 2009; Henze et al., 2008) but few  
69 studies have focused on its health implications (Lin et al., 2016). Evaluating the health effects of  
70 SOA from isoprene oxidation is important from a public health perspective, not only because of  
71 its atmospheric abundance, but also because the anthropogenic contribution is the only  
72 component amenable to control (Pye et al., 2013; Gaston et al., 2014; Xu et al., 2015; Riedel et  
73 al., 2015).

74 Many studies have shown that particulate matter is closely linked to health effects  
75 ranging from exacerbation of asthma symptoms to mortality associated with lung cancer and  
76 cardiopulmonary disease (Dockery et al., 1993; Schwartz et al., 1993; Samet et al., 2000). PM<sub>2.5</sub>,  
77 in particular, has been linked to negative health outcomes with an estimated contribution of 3.2  
78 million premature deaths worldwide as reported in the Global Burden of Disease Study 2010  
79 (Lim et al., 2012). Despite evidence that particle composition affects toxicity, fewer studies  
80 focus on the link between chemical composition and health/biological outcomes (Kelly and  
81 Fussell, 2012). Prior work on complex air mixtures has shown that gaseous volatile organic  
82 compounds (VOCs) alter the composition and ultimately the toxicity of particles (Ebersviller et  
83 al., 2012a, b). SOA resulting from natural and anthropogenic gaseous precursors, such as  $\alpha$ -  
84 pinene and 1,3,5-trimethylbenzene, have been shown to affect cellular function (Gaschen et al.,  
85 2010; Jang et al., 2006) and recently isoprene-SOA formed from the reactive uptake of epoxides  
86 has been shown to induce the expression of oxidative stress genes (Lin et al., 2016).

87 The objective of this study is to generate atmospherically relevant isoprene-derived SOA  
88 and examine its toxicity through *in vitro* exposures using a direct deposition device. Compared to  
89 exposure of cells in culture media to resuspended particles, direct particle deposition likely  
90 provides a more biologically relevant exposure model and enhances sensitivity of cells to air  
91 pollution particle exposures (Volckens et al., 2009; Lichtveld et al., 2012; Hawley et al., 2014a;  
92 Hawley et al., 2014b; Zavala et al., 2014; Hawley and Volckens, 2013). The Electrostatic  
93 Aerosol *in vitro* Exposure System (EAVES) used in this study deposits particles generated in our  
94 outdoor photochemical chamber directly onto lung cells by electrostatic precipitation (de Bruijne  
95 et al., 2009). Similar techniques and devices have been used to expose cells to diesel exhaust  
96 particles (Lichtveld et al., 2012; Hawley et al., 2014b), but our study is the first to utilize the

97 EAVES to explore the potential adverse effects of isoprene SOA on human lung cells.  
98 Additionally, for a more atmospherically relevant exposure, isoprene-SOA was photochemically  
99 generated in an outdoor chamber to mimic its formation in the atmosphere.

100 We have recently demonstrated through a chemical assay that isoprene-derived SOA has  
101 the potential for inducing reactive oxygen species (ROS) (Kramer et al., 2016), which are linked  
102 to oxidative stress and inflammation (Reuter et al., 2010; Li et al., 2003). An *in vitro* study that  
103 followed supported the potential for isoprene-SOA to affect the levels of oxidative stress genes  
104 (Lin et al., 2016). In this study we chose to examine the gene expression levels of interleukin-8  
105 (*IL-8*) and cyclooxygenase-2 (*COX-2*), not only for their links to inflammation and oxidative  
106 stress (Kunkel et al., 1991; Uchida, 2008), but because both have been examined in previous  
107 studies using the EAVES for fresh and aged diesel exhaust (Lichtveld et al., 2012). Other studies  
108 on air pollution mixtures have also examined *IL-8* as a biological endpoint due to its involvement  
109 with inflammation (Zavala et al., 2014; Ebersviller et al., 2012a, b; Doyle et al., 2004; Doyle et  
110 al., 2007). We compared the gene expression levels in cells exposed to SOA generated in an  
111 outdoor chamber from photochemical oxidation of isoprene in the presence of NO and acidified  
112 sulfate seed aerosol to cells exposed to a dark control mixture of isoprene, NO, and acidified  
113 sulfate seed aerosol to isolate the effects of the isoprene-derived SOA on the cells using the  
114 EAVES. In addition, we collected SOA onto filters for subsequent resuspension exposure to  
115 ensure that effects observed from EAVES exposures were attributable to particle-phase organic  
116 products.

## 117 **2. Experimental Section**

118 **2.1 Generation of SOA in the Outdoor Chamber Facility.** SOA were generated by  
119 photochemically oxidizing a mixture of acidified sulfate seed aerosol, isoprene, and NO injected

120 into an outdoor smog chamber facility. The outdoor chamber is a 120-m<sup>3</sup> triangular cross-section  
121 Teflon chamber located on the roof of the Gillings School of Global Public Health, University of  
122 North Carolina at Chapel Hill. The chamber facility has been described in detail elsewhere by  
123 Lichtveld et al. (2012). The outdoor chamber facility is equipped with sampling lines that allow  
124 direct deposition exposure of cells, online chemical measurements, and filter collection for  
125 offline chemical analysis. Sampling lines run from the underside of the chamber directly to the  
126 chemistry lab below where online measurement instruments and the direct deposition exposure  
127 device are located. Injection ports are also located on the underside of the chamber.

128 To generate isoprene-derived SOA, the chamber was operated on sunny days, under high  
129 relative humidity, to allow natural sunlight to trigger photochemical reactions. Acidified sulfate  
130 seed aerosols were generated by nebulizing an aqueous solution containing 0.06 M MgSO<sub>4</sub> +  
131 0.06 M H<sub>2</sub>SO<sub>4</sub> into the chamber to a particle concentration of approximately 170 µg m<sup>-3</sup>, which  
132 was allowed to stabilize for 30 min to ensure a well-mixed condition. After stabilization, 3.5  
133 ppmv isoprene (Sigma-Aldrich, 99%) and 200 ppbv NO (AirGas, 1.00%) were injected into the  
134 chamber. Photochemical aging was allowed for approximately one hour to reach the desired  
135 exposure conditions of 30-40 µg m<sup>-3</sup> growth of isoprene-derived SOA on the pre-existing 170 µg  
136 m<sup>-3</sup> of acidified sulfate aerosol. This chamber experiment was replicated on three separate sunny  
137 days with temperatures ranging from 24.9°C to 26.8°C with a relative humidity of approximately  
138 70% in the chamber.

139 **2.2 Control Chamber Experiments.** As a dark chamber control, to isolate the effect of SOA on  
140 exposed cells, mixtures of isoprene, NO, and 170 µg m<sup>-3</sup> of acidified sulfate seed aerosol were  
141 injected into the chamber in the dark (after sunset). Conducting the chamber experiments in the  
142 dark ensured no photochemical oxidation of isoprene. The dark control was replicated on three

143 different nights. Except for the absence of solar radiation (no SOA), all chamber operations and  
144 exposure conditions were similarly maintained.

145 As an added control to ensure that the device itself and the cell handling had no  
146 significant effect on cell cytotoxicity, cells were exposed in the EAVES to a clean chamber and  
147 compared to unexposed cells kept in an incubator for the same duration as the exposure. The  
148 cytotoxicity results ensured that there is no effect of chamber conditions and device operation on  
149 the cells.

150 **2.3 Cell Culture.** Human bronchial epithelial (BEAS-2B) cells were maintained in keratinocyte  
151 growth medium (KGM BulletKit; Lonza), a serum-free keratinocyte basal medium (KBM)  
152 supplemented with 0.004% of bovine pituitary extract and 0.001% of human epidermal growth  
153 factor, insulin, hydrocortisone, and GA-1000 (gentamicin, amphotericin B), and passaged  
154 weekly. Passage number for photochemical exposures and dark control exposures varied  
155 between 52 and 60. Because BEAS-2B are an immortalized line of human bronchial epithelium,  
156 there are limitations with its use such as it being genetically homogeneous, being a single cell  
157 type, and being SV-40 transformed (Reddel et al., 1988). However, BEAS-2B is a stable,  
158 proliferative cell line shown to be useful in airway inflammation studies such as ours (Devlin et  
159 al., 1994).

160 **2.4 Direct Deposition Exposure.** In preparation for air-liquid interface exposures, cells were  
161 seeded onto collagen-coated Millicell cell culture inserts (30 mm diameter, 0.4  $\mu\text{m}$  pore size, 4.2  
162  $\text{cm}^2$  filter area; Millipore, Cambridge, MA) at a density of 200,000 cells/well 24 hours prior to  
163 exposure. At the time of exposure, cells reached  $\sim$ 80% confluence, confirmed through  
164 microscopy. Immediately before exposure, cell medium was removed from the apical and  
165 basolateral sides of 2 seeded Millicell cell culture inserts. One insert was transferred to a titanium

166 dish containing 1.5 mL of keratinocyte basal medium (KBM; Lonza), supplying cells with  
167 nutrients from the basolateral side and constant moisture while allowing exposure to be  
168 performed at an air-liquid interface. The other insert was transferred into a 6 well plate with 2  
169 mL of KBM and placed in the incubator as an unexposed control.

170 Cells were exposed to chamber-generated isoprene SOA using the EAVES located in the  
171 laboratory directly beneath the outdoor chamber (de Bruijne et al., 2009; Lichtveld et al., 2012).  
172 The EAVES, located in an incubator at 37°C, sampled chamber air at 1 L min<sup>-1</sup>. The target  
173 relative humidity (RH) in the chamber during EAVES exposures was approximately 70%.  
174 Exposure time was one hour commencing when target exposure conditions were achieved in the  
175 outdoor chamber for both photochemical and dark control experiments. Detailed description of  
176 the EAVES can be found in de Bruijne et al. (2009).

177 Following exposure, the cell culture insert was transferred to a 6-well tissue culture plate  
178 containing 2 mL of fresh KBM. The control Millicell was also transferred to 2 mL of fresh  
179 KBM. Nine hours post-exposure, extracellular medium was collected and total RNA was isolated  
180 using Trizol (Life Technologies), consistent with past studies (de Bruijne et al., 2009).  
181 Extracellular medium and the extracted RNA samples were stored at -20°C and -80°C,  
182 respectively, until further analysis. For quality assurance purposes, the RNA concentration and  
183 integrity were assessed using Nanodrop and Bioanalyzer over the period of storage. No changes  
184 were observed under the given storage conditions.

185 **2.5 Filter Resuspension Exposure.** Chamber particles were collected, concurrently with  
186 EAVES sampling, onto Teflon membrane filters (47 mm diameter, 1.0 µm pore size; Pall Life  
187 Science) for photochemical (light) and dark chamber experiments to be used for chemical  
188 analysis and resuspension exposures. The resuspension experiments served as a control for



189 possible effects of gaseous components such as ozone (O<sub>3</sub>) and NO<sub>x</sub> present in the direct  
190 deposition experiments; however, prior studies have shown that gaseous components do not  
191 yield cellular responses within the EAVES device (de Bruijne et al., 2009; Ebersviller et al.,  
192 2012a, b). Mass loadings of SOA collected on the filters were calculated from sampling volumes  
193 and average aerosol mass concentrations in the chamber during the sampling period. A density  
194 correction of 1.6 g cm<sup>-3</sup> (Riedel et al., 2016) and 1.25 g cm<sup>-3</sup> (Kroll et al., 2006) was applied to  
195 convert the measured volume concentrations to mass concentrations for the acidified sulfate seed  
196 and SOA growth, respectively. The particles collected on Teflon filter membranes for  
197 resuspension cell exposure were extracted by sonication in high-purity methanol (LC/MS  
198 CHROMASOLV, Sigma-Aldrich). Filter samples from multiple experiments were combined and  
199 the combined filter extract was dried under a gentle stream of nitrogen (N<sub>2</sub>). KBM medium was  
200 then added into the extraction vials to re-dissolve SOA constituents.

201 In preparation for filter resuspension exposures, cells were seeded in 24-well plates at a  
202 density of 2.5×10<sup>4</sup> cells/well in 250 μL of KGM 2 days prior to exposure. At the time of  
203 exposure when cells reached ~80% confluence, cells were washed twice with phosphate buffered  
204 saline (PBS) buffer, and then exposed to KBM containing 0.01 and 0.1 mg mL<sup>-1</sup> isoprene SOA  
205 extract from photochemical experiment and seed particles from dark control experiments.

206 Following a 9-hour exposure, extracellular medium was collected and total RNA was  
207 isolated using Trizol (Life Technologies) and stored alongside samples from direct deposition  
208 exposures until further analysis.

209 **2.6 Chemical and Physical Characterization of Exposures.** Online and offline techniques  
210 were used to characterize the SOA generated in the chamber. The online techniques measured  
211 the gas-phase species NO, NO<sub>x</sub> and O<sub>3</sub> and the physical properties of the aerosol continuously

212 throughout the chamber experiments. Offline techniques measured aerosol-phase species  
213 collected onto Teflon membrane filters (47 mm diameter, 1.0  $\mu\text{m}$  pore size; Pall Life Science)  
214 from photochemical and dark chamber experiments. Filter samples were stored in 20 mL  
215 scintillation vials protected from light at  $-20^{\circ}\text{C}$  until analyses.

216 Real-time aerosol size distributions were measured using a Differential Mobility  
217 Analyzer (DMA, Brechtel Manufacturing Inc.) coupled to a Mixing Condensation Particle  
218 Counter (MCPC, Model 1710, Brechtel Manufacturing Inc.) located in the laboratory directly  
219 underneath the chamber.  $\text{O}_3$  and  $\text{NO}_x$  were measured with a ML 9811 series Ozone Photometer  
220 (Teledyne Monitor Labs, Englewood, CO) and ML 9841 series  $\text{NO}_x$  Analyzer (American  
221 Ecotech, Warren RI), respectively. Data were collected at one-minute intervals using a data  
222 acquisition system (ChartScan/1400) interfaced to a computer. The presence of isoprene in the  
223 chamber was confirmed and quantified using a Varian 3800 gas chromatograph (GC) equipped  
224 with a flame ionization detector (FID).

225 Chemical characterization of SOA constituents was conducted offline from extracts of  
226 filters collected from chamber experiments by gas chromatography interfaced with an electron  
227 ionization quadrupole mass spectrometer (GC/EI-MS) or by ultra performance liquid  
228 chromatography interfaced with a high-resolution quadrupole time-of-flight mass spectrometer  
229 equipped with electrospray ionization (UPLC/ESI-HR-QTOFMS). Detailed operating conditions  
230 for the GC/EI-MS and UPLC/ESI-HR-QTOFMS analyses as well as detailed filter extraction  
231 protocols have been described previously by Lin et al. (2012). For GC/EI-MS analysis, filter  
232 extracts were dried under a gentle stream of  $\text{N}_2$  and trimethylsilylated by the addition of 100  $\mu\text{L}$   
233 of BSTFA + TMCS (99:1 v/v, Supelco) and 50  $\mu\text{L}$  of pyridine (anhydrous, 99.8%, Sigma-  
234 Aldrich) and heated at  $70^{\circ}\text{C}$  for 1 h. For UPLC/ESI-HR-QTOFMS analysis, residues of filter

235 extracts were reconstituted with 150  $\mu$ L of a 50:50 (v/v) solvent mixture of high-purity water and  
236 methanol.

237 The isoprene-derived SOA markers: 2-methyltetrols, isomeric 3-methyltetrahydrofurans-  
238 3,4-diols (3-MeTHF-3,4-diols), and 2-methylglyceric acid, synthesized according to the  
239 published procedures (Lin et al., 2013b; Zhang et al., 2012), were available in-house as authentic  
240 standards to quantify the major components of isoprene SOA. 2-Methyltetrol organosulfates,  
241 synthesized as a mixture of tetrabutylammonium salts, were also available as a standard. Purity  
242 was determined to be >99% by  $^1\text{H}$  NMR and UPLC/ESI-QTOFMS analysis (Budisulistiorini et  
243 al., 2015b). The  $\text{C}_5$ -alkene triols and IEPOX dimer were quantified using the response factor  
244 obtained for the synthetic 2-methyltetrols.

245 A representative ambient  $\text{PM}_{2.5}$  sample collected from the rural southeastern U.S.  
246 (Yorkville, GA) (Lin et al., 2013a) during the summer of 2010 was analyzed in an identical  
247 manner to confirm atmospheric relevance of the chamber-generated SOA constituents.

248 **2.7 Cytotoxicity Assay.** Cytotoxicity was assessed through measurement of lactate  
249 dehydrogenase (LDH) released into the extracellular medium from damaged cells using the LDH  
250 cytotoxicity detection kit (Takara). To ensure that the EAVES device itself and operation  
251 procedure had no effect on cytotoxicity, the LDH release from cells exposed to clean chamber air  
252 was measured. LDH release by cells exposed via the EAVES to the photochemically aged (light)  
253 and non-photochemically aged (dark) particles was compared to release from unexposed cells  
254 maintained in the incubator for the same duration. For the resuspension exposures, LDH release  
255 by cells exposed to SOA through resuspended extract of photochemically aged and non-  
256 photochemically aged particles was compared to release by cells maintained in KBM only.  
257 Additionally, LDH release from the light exposures, dark control, and resuspension exposures

258 was compared to release by positive control cells exposed to 1% Triton X-100 to ensure that cell  
259 death would not affect gene expression results.

260 **2.8 Gene Expression Analysis.** We chose to measure the levels of the inflammation-related  
261 mRNA in the BEAS-2B cells exposed to isoprene-derived SOA generated in our outdoor  
262 chamber because various particle types are capable of sequestering cytokines (Seagrave, 2008).  
263 Other direct deposition studies have also used mRNA transcripts as a proxy for cytokine  
264 production (Hawley et al., 2014a; Hawley et al., 2014b; Hawley and Volckens, 2013; Volckens  
265 et al., 2009; Lichtveld et al., 2012). Changes in *IL-8* and *COX-2* mRNA levels were measured  
266 using QuantiTect SYBR Green RT-PCR Kit (Qiagen) and QuantiTect Primer Assays for  
267 Hs\_ACTB\_1\_SG (Catalog #QT00095431), Hs\_PTGS2\_1\_SG (Catalog #QT00040586), and  
268 Hs\_CXCL8\_1\_SG (Catalog #QT00000322) for one-step RT-PCR analysis. All mRNA levels  
269 were normalized against  $\beta$ -actin mRNA, which was used as a housekeeping gene. The relative  
270 expression levels (i.e., fold change) of *IL-8* and *COX-2* were calculated using the comparative  
271 cycle threshold ( $2^{-\Delta\Delta CT}$ ) method (Livak and Schmittgen, 2001). For EAVES exposures, changes  
272 in *IL-8* and *COX-2* from isoprene-derived SOA exposed cells were compared to cells exposed to  
273 the dark controls. Similarly, for resuspension exposures changes in *IL-8* and *COX-2* from  
274 isoprene-derived SOA exposed cells were compared to cells exposed to particles collected under  
275 dark conditions.

276 **2.9 Statistical Analysis.** The software package GraphPad Prism 4 (GraphPad) was used for all  
277 statistical analyses. All data were expressed as mean  $\pm$  SEM (standard error of means).  
278 Comparisons between data sets for cytotoxicity and gene expression analysis were made using  
279 unpaired *t*-test with Welch's correction. Significance was defined as  $p < 0.05$ .

## 280 3. Results and Discussion

281 **3.1 Physical and Chemical Characterization of Exposure.** Figure 1 shows the change in  
282 particle mass concentration and gas ( $O_3$ ,  $NO$ ,  $NO_x$ ) concentration over time during typical  
283 photochemical and dark control experiments. Under dark control conditions (Fig. 1a) there is no  
284 increase in aerosol mass concentration following isoprene injection. Average total aerosol mass  
285 concentration was  $155.0 \pm 2.69 \mu\text{g m}^{-3}$  (1 standard deviation) with no particle mass attributable to  
286 organic material.

287 In contrast, Fig. 1b shows an increase in aerosol mass concentration after 1 h post  
288 isoprene injection, which can be attributed to the photochemical oxidation of isoprene and  
289 subsequent production and reactive uptake of its oxidation products. The average increase in  
290 aerosol mass concentration attributable to SOA formation for three daylight chamber  
291 experiments conducted on separate days was  $44.5 \pm 5.7 \mu\text{g m}^{-3}$ . Average total aerosol mass  
292 concentration during particle exposure was  $173.1 \pm 4.2 \mu\text{g m}^{-3}$ .

293  $O_3$  and  $NO_x$  concentrations measured during EAVES exposure were approximately 270  
294 ppb and 120 ppb for photochemical experiments. For dark control experiments (e.g., Fig. 1a), the  
295  $O_3$  and  $NO_x$  concentrations were approximately 15 ppb and 180 ppb. Previous studies  
296 characterizing the EAVES device show definitively that gas-phase products do not induce cell  
297 response (de Bruijne et al., 2009). However, resuspension exposures were conducted in addition  
298 to EAVES exposure to ensure that biological effects were attributable to only particle-phase  
299 constituents and not gas-phase products such as  $O_3$  and  $NO_x$ .

300 The chemical composition of aerosol, collected onto filters concurrently with cell  
301 exposure and characterized by GC/EI-MS and UPLC/ESI-HR-QTOFMS, are shown in Fig. 2.  
302 No isoprene-SOA tracers were observed in the filters collected from dark control experiments.

303 The dominant particle-phase products of the isoprene-SOA collected from photochemical  
304 experiments are derived from the low-NO channel, where IEPOX reactive uptake onto acidic  
305 sulfate aerosol dominates, including 2-methyltetrols, C5-alkene triols, isomeric 3-MeTHF-3,4-  
306 diols, IEPOX-derived dimers, and IEPOX-derived organosulfates. The sum of the IEPOX-  
307 derived SOA constituents quantified by the available standards accounted for ~80% of the  
308 observed SOA mass. The MAE-derived SOA constituents 2-methylglyceric acid and the  
309 organosulfate derivative of MAE, derived from the high-NO channel, accounted for 1.4% of the  
310 observed SOA mass, confirming that particle-phase products generated were predominantly  
311 formed from the reactive uptake of IEPOX onto acidic sulfate aerosols. As demonstrated in  
312 Figure 2, all the same particle-phase products are measured in the PM<sub>2.5</sub> sample collected in  
313 Yorkville, GA (a typical low-NO region), demonstrating that the composition of the chamber-  
314 generated SOA is atmospherically relevant. Recent SOA tracer measurements from the Southern  
315 Oxidant and Aerosol Study (SOAS) campaign at Look Rock, TN, Centerville, AL, and  
316 Birmingham, AL, also support the atmospheric relevance of IEPOX-derived SOA constituents  
317 that dominate the isoprene SOA mass in summer in the southeastern U.S. (Budisulistiorini et al.,  
318 2015a; Rattanavaraha et al., 2016).

319 **3.2 Cytotoxicity.** LDH release for cells exposed using the EAVES device is expressed as a fold-  
320 change relative to the unexposed incubator control. For resuspension exposures, LDH release is  
321 expressed as fold-change relative to cells exposed to KBM only. Results shown in Fig. 3a  
322 confirm that there is no effect of chamber conditions and device operation on the cells when  
323 comparing LDH release from cells exposed to a clean air chamber and cells unexposed in an  
324 incubator. Additionally, LDH release from all exposure conditions in EAVES exposed cells (Fig.  
325 3b) and resuspension exposed cells (Fig. 3c) is negligible relative to positive controls exposed to

326 1% Triton X-100, confirming that the exposure concentration of isoprene-derived SOA utilized  
327 in this study was not cytotoxic. All cytotoxicity results ensured that exposure conditions were not  
328 adversely affecting the cells nor their gene expression.

329 **3.3 Pro-inflammatory Gene Expression.** Changes in the mRNA levels of *IL-8* and *COX-2*  
330 from cells exposed to isoprene-derived SOA using the EAVES are shown as fold-changes  
331 relative to dark controls in Fig. 4. This comparison, as well as the results of the resuspension  
332 experiment discussed below, ensure that all effects seen in the cells are attributable to the  
333 isoprene-derived SOA and no other factors. A one-hour exposure to a mass concentration of  
334 approximately  $45 \mu\text{g m}^{-3}$  of organic material was sufficient to significantly alter gene expression  
335 of the inflammatory biomarkers in bronchial epithelial cells. Based on deposition efficiency  
336 characterized by de Bruijne et al. (2009), the estimated dose was  $0.29 \mu\text{g cm}^{-2}$  of total particle  
337 mass with 23% attributable to organic material formed from isoprene photooxidation ( $0.067 \mu\text{g}$   
338  $\text{cm}^{-2}$  of SOA).

339 Changes in the mRNA levels of *IL-8* and *COX-2* from cells exposed to resuspended  
340 isoprene-derived SOA collected from photochemical experiments are shown as fold-changes  
341 relative to cells exposed to resuspended particles from dark control experiments in Fig. 5. At a  
342 low dose of  $0.01 \text{ mg mL}^{-1}$  of isoprene SOA extract there is no significant increase in *IL-8* and  
343 *COX-2* mRNA expression. The isoprene SOA extract, however, induces a response at a dose of  
344  $0.1 \text{ mg mL}^{-1}$ . The statistically significant increase in mRNA expression from the resuspension  
345 exposure at  $0.1 \text{ mg mL}^{-1}$  confirms that similar fold changes observed for both *IL-8* and *COX-2*  
346 from the EAVES exposures are not attributable to gaseous photooxidation products, such as  $\text{O}_3$ ,  
347 and support the characterization of the EAVES as a particle exposure device (de Bruijne et al.,  
348 2009).

349           The similar fold change observed in both the EAVES exposure and resuspension  
350 exposure, in addition to confirming that the biological effects can be attributed to the particle-  
351 phase photochemical products (isoprene-derived SOA), suggests that exposure by resuspension  
352 is appropriate for isoprene-derived SOA and may yield results similar to direct deposition  
353 exposures. Unlike diesel particulate extracts, which agglomerate during resuspension exposures,  
354 isoprene-derived SOA constituents are water-soluble based on reverse-phase LC separations  
355 (Surratt et al., 2006; Lin et al., 2012) and remain well mixed in the cell medium used for  
356 exposure. Therefore, resuspension exposures do not appear to be a limitation for toxicological  
357 assessments of isoprene SOA.

358 **3.4 Biological Implications.** The goal of this study was to initially identify potential biological  
359 response associated with exposure to isoprene-derived SOA by using a direct exposure device as  
360 a model that has both atmospheric and physiological relevance. With this model, a dose of 0.067  
361  $\mu\text{g cm}^{-2}$  of isoprene SOA, induced statistically significant increases in *IL-8* and *COX-2* mRNA  
362 levels in exposed BEAS-2B cells. There are many ways to classify in vitro particle dosimetry  
363 based on the various properties of particles (Paur et al., 2011). For this direct deposition study,  
364 we chose to classify dose as SOA mass deposition per surface area of the exposed cells to mimic  
365 lung deposition. Gangwal et al. (2011) used a multiple-path particle dosimetry (MPPD) model to  
366 estimate that the lung deposition of ultrafine particles ranges from 0.006 to 0.02  $\mu\text{g cm}^{-2}$  for a 24-  
367 hr exposure to a particle concentration of 0.1  $\text{mg m}^{-3}$ . Based on this estimate, a dose of 0.067  $\mu\text{g}$   
368  $\text{cm}^{-2}$  of isoprene SOA in our study can be considered a prolonged exposure over the course of a  
369 week. In fact, most other in vitro studies require dosing cells at a high concentration sometimes  
370 close to a lifetime exposure to obtain a cellular response. Despite this limitation, in vitro  
371 exposures serve as a necessary screening tool for toxicity (Paur et al., 2011).



372 Our findings are consistent with other studies showing that photochemical oxidation of  
373 similar chemical mixtures increases toxicity in cell culture models and elevates expression of  
374 inflammatory biomarker genes (Lichtveld et al., 2012; Rager et al., 2011). Previous *in vitro*  
375 studies using a gas-phase only exposure system have shown that gas-phase products of isoprene  
376 photooxidation significantly enhance cytotoxicity and *IL-8* expression (Doyle et al., 2004; Doyle  
377 et al., 2007).

378 By choosing *IL-8* and *COX-2* as our genes of interest, we are able to compare our results  
379 to other studies of known harmful particle exposures. In a similar study using the EAVES,  
380 normal human bronchial epithelial (NHBE) cells exposed to 1.10  $\mu\text{g cm}^{-2}$  diesel particulate  
381 matter showed less than a 2-fold change over controls in both *IL-8* and *COX-2* mRNA  
382 expression (Hawley et al., 2014b). In another study, A549 human lung epithelial cells were  
383 exposed by direct deposition for 1 hour to photochemically-aged diesel exhaust particulates at a  
384 dose of 2.65  $\mu\text{g cm}^{-2}$  from a 1980 Mercedes or a 2006 Volkswagen (Lichtveld et al., 2012).  
385 Exposure to aged Mercedes particulates induced a 4-fold change in *IL-8* and  $\sim$ 2-fold change in  
386 *COX-2* mRNA expression, while exposure to aged Volkswagen particulates induced a change of  
387  $\sim$ 1.5-fold in *IL-8* and 2-fold in *COX-2* mRNA expression (Lichtveld et al., 2012). Although the  
388 differences in cell types preclude direct comparisons, the finding of significant increases in *COX-*  
389 *2* and *IL-8* expression at doses much lower than reported for comparable increases in gene  
390 expression levels induced by photochemically-aged diesel particulates is notable.

391 *IL-8* and *COX-2* are both linked to inflammation and oxidative stress (Kunkel et al.,  
392 1991; Uchida, 2008). *IL-8* is a potent neutrophil chemotactic factor in the lung and its expression  
393 by various cells plays a crucial role in neutrophil recruitment leading to lung inflammation  
394 (Kunkel et al., 1991). *COX-2* is the inducible form of the cyclooxygenase enzyme, regulated by

395 cytokines and mitogens, and is responsible for prostaglandin synthesis associated with  
396 inflammation (FitzGerald, 2003). Consistent with the reports that *IL-8* and *COX-2* play important  
397 roles in lung inflammation (Nocker et al., 1996; Li et al., 2013), *in vivo* studies have shown that  
398 isoprene oxidation products cause airflow limitation and sensory irritation in mice (Rohr et al.,  
399 2003). In humans, the role of *IL-8* and *COX-2* in lung inflammation can be associated with  
400 diseases such as chronic obstructive pulmonary disease and asthma (Nocker et al., 1996; Peng et  
401 al., 2008; Fong et al., 2000).

402         The mechanism by which isoprene-SOA causes elevation of the inflammatory markers  
403 *IL-8* and *COX-2* is not yet fully understood. However, recent work from our laboratory using the  
404 acellular dithiothreitol (DTT) assay demonstrated that isoprene-derived SOA **has significant**  
405 **ROS generation potential** (Kramer et al., 2016). High levels of ROS in cells can overwhelm the  
406 antioxidant defense and lead to cellular oxidative stress (Sies, 1991; Bowler and Crapo, 2002; Li  
407 et al., 2003). Following the discovery of the potential importance of isoprene-SOA in generating  
408 ROS, Lin et al. (2016) showed that isoprene-SOA formed from the reactive uptake of epoxides  
409 alters levels of oxidative stress-associated genes, including *COX-2* in human lung cells.  
410 Oxidative stress caused by ROS plays a major role in lung inflammation and the induction of  
411 oxidative stress can lead to *IL-8* expression (Tao et al., 2003; Yan et al., 2015). **Specifically,**  
412 **oxidants can activate the transcription factor NF- $\kappa$ B, which regulates a wide range of**  
413 **inflammatory genes including *IL-8* and *COX-2* (Barnes and Adcock, 1997; Schreck et al., 1992).**  
414 Therefore, isoprene-SOA may cause increases in both *IL-8* and *COX-2* primarily through an  
415 oxidative stress response. Additionally, the relationship between *IL-8* and *COX-2* can also  
416 explain the observed increase in *IL-8* gene expression as the production of *IL-8* can be stimulated  
417 through a *COX-2* dependent mechanism in airway epithelial cells (Peng et al., 2008).

418 *In vitro* studies such as this one using a direct deposition model cannot fully elucidate  
419 mechanisms of lung inflammation and potential pathogenesis but serve as a necessary part of  
420 hazard characterization, particularly for a complex air mixture that has not been fully studied  
421 (Hayashi, 2005; Paur et al., 2011). **Ozone exposure studies have shown that comparable dose and**  
422 **effect measurements for *IL-8* and *COX-2* can be found between *in vivo* and *in vitro* exposures**  
423 **which add promise to extrapolating effects seen *in vitro* to effects *in vivo* (Hatch et al., 2014). *In***  
424 ***vivo* effects associated with isoprene-SOA exposure *in vitro* cannot be inferred as it is a different**  
425 **system from ozone, so further *in vitro* studies exploring the health implication of the elevation of**  
426 ***IL-8* and *COX-2* due specifically to isoprene-SOA exposure are necessary and may in turn justify**  
427 **further extension to *in vivo* work.**

#### 428 **4. Conclusions**

429 This study indicates that an atmospherically relevant composition of isoprene-derived  
430 SOA is capable of increasing the expression of *IL-8* and *COX-2* in human bronchial epithelial  
431 cells. The present study is an initial step in a long planned analysis of the biological impacts of  
432 isoprene SOA exposure on lung cells. The SOA were generated as NO levels approached zero,  
433 which represents conditions characteristic of urban locales downwind of rural isoprene sources.  
434 As shown in Fig. 2, the aerosol generated for exposures in this study are chemically similar to  
435 fine aerosol samples collected from the Southeastern U.S., which indicates that the chamber  
436 exposures are representative of exposures that may be encountered by populations in regions  
437 where isoprene emissions interact with anthropogenic pollutants. The same particle-phase  
438 products found in our photochemical experiments have been measured in significant quantities  
439 (accounting on average for 33% of fine organic aerosol mass) in ambient fine organic particles  
440 collected in the Southeastern U.S. (Lin et al., 2013b; Budisulistiorini et al., 2013; Rattanavaraha

441 et al., 2016; Budisulistiorini et al., 2016) and in other isoprene-rich environments (Hu et al.,  
442 2015). The results of this study show that, because of its abundance, isoprene SOA may be a  
443 public health concern warranting further toxicological investigation through *in vitro* or *in vivo*  
444 work.

445

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456

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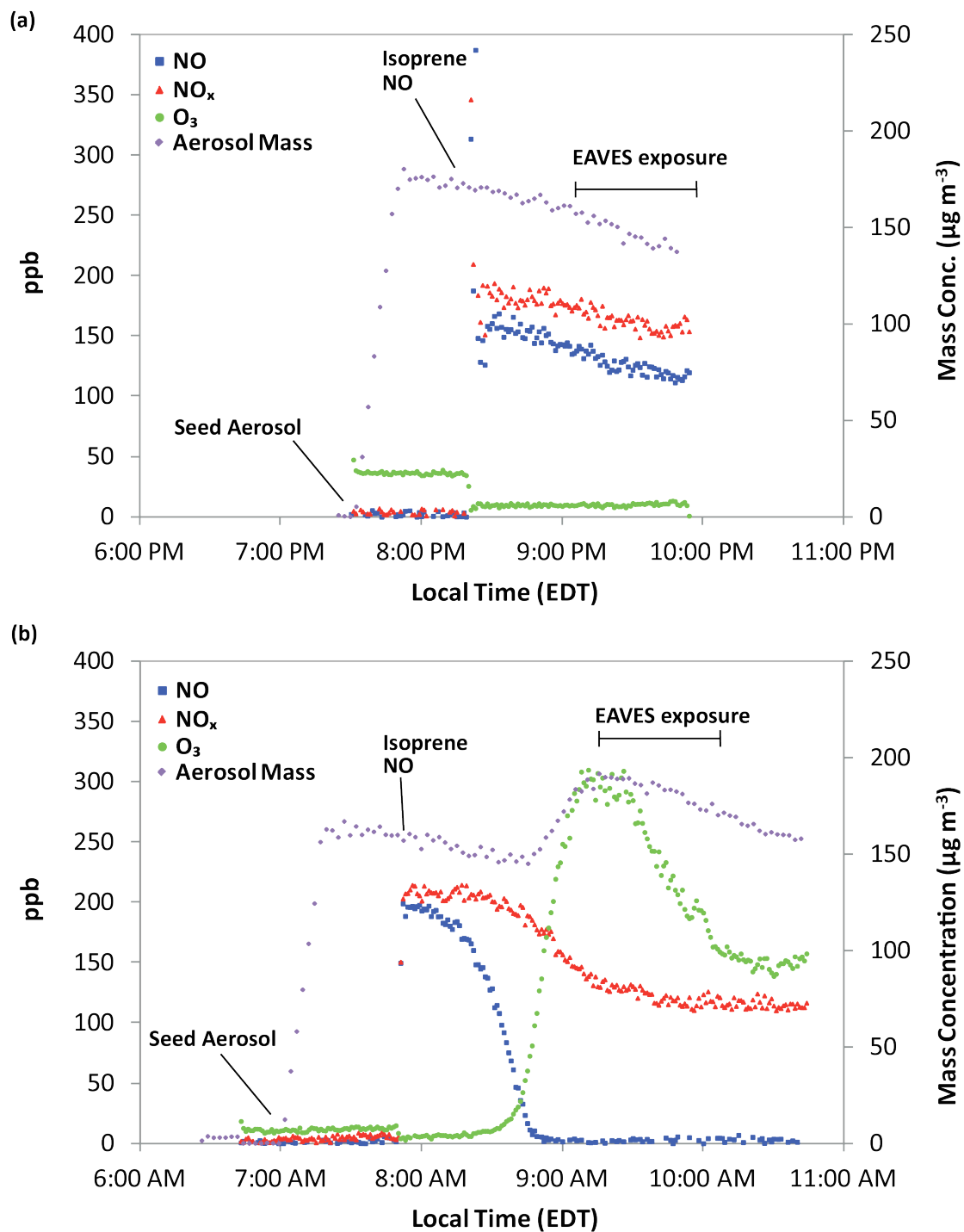
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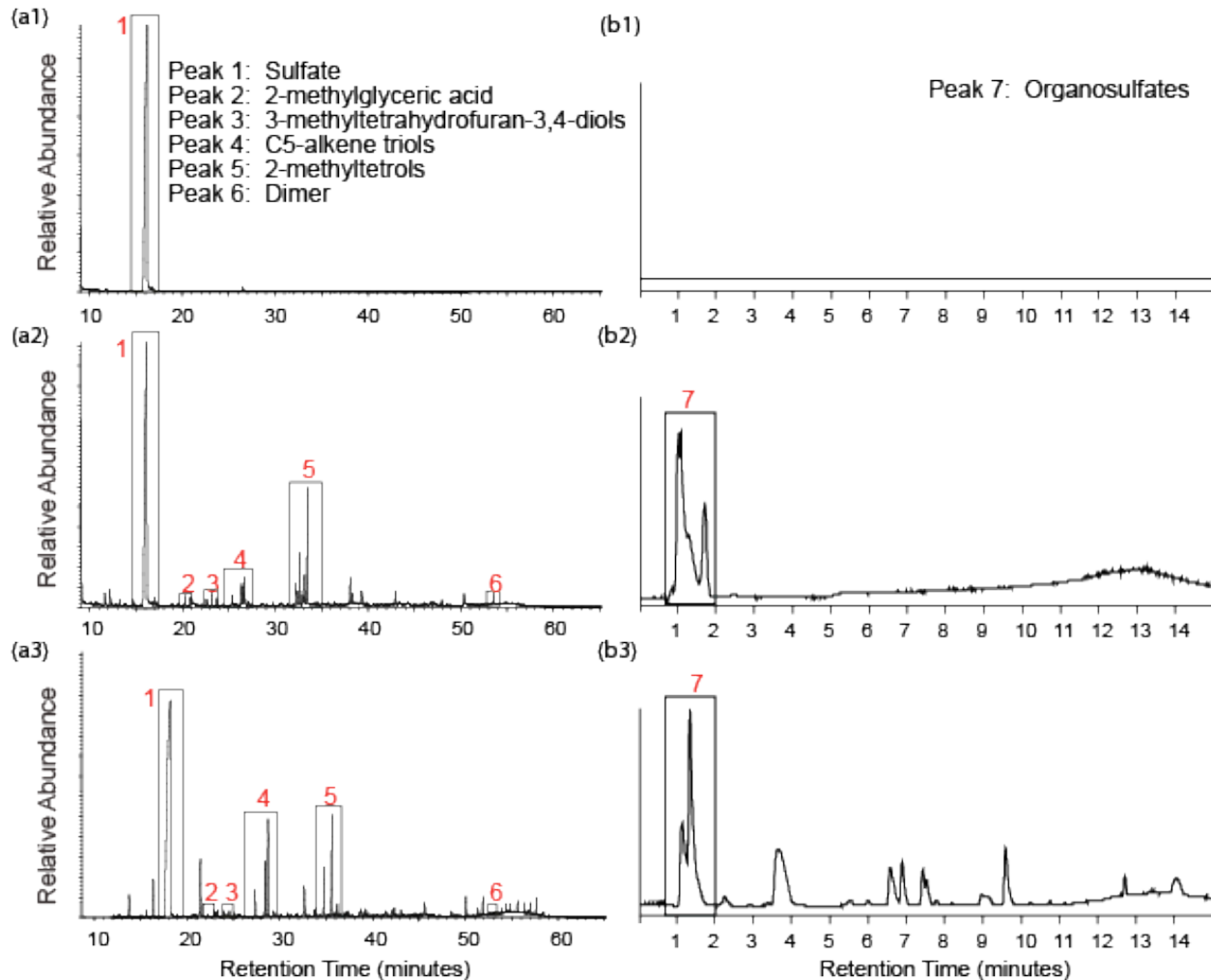


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754 **Figure 1.** Aerosol mass concentration and gas-phase product concentrations over time for (a)

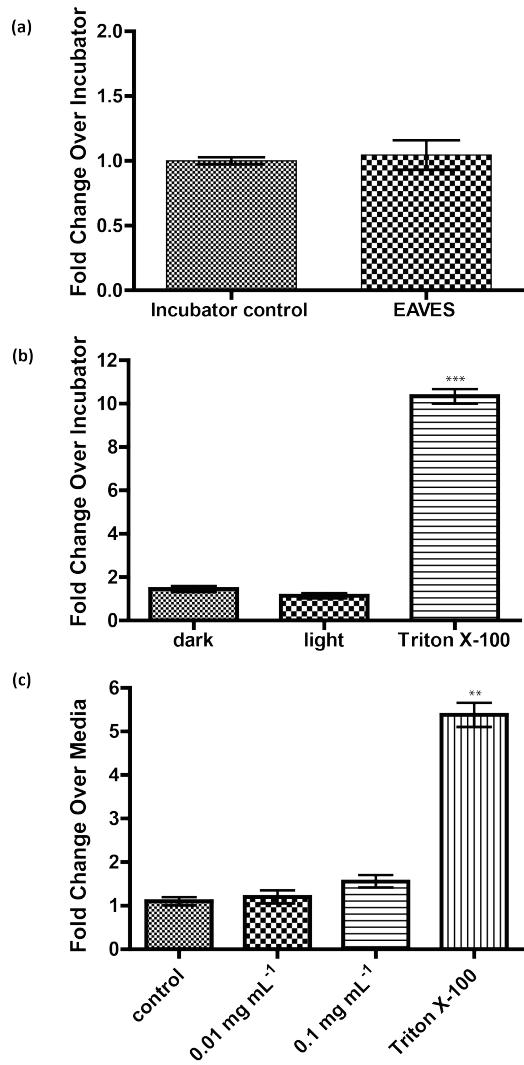
755 dark control chamber experiment and (b) photochemically produced isoprene-derived SOA

756 exposure chamber experiment.



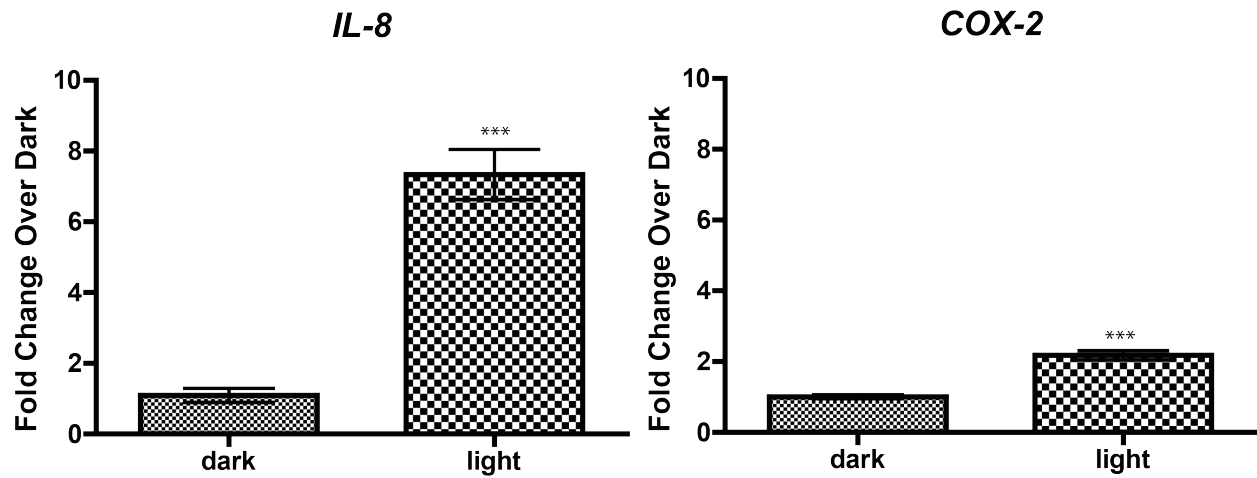
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759 **Figure 2.** (a) GC/EI-MS total ion chromatograms (TICs) and (b) UPLC/ESI-HR-QTOFMS base  
 760 peak chromatograms (BPCs) from a (1) dark control chamber experiment, (2) isoprene-derived  
 761 SOA exposure chamber experiment, and (3) PM<sub>2.5</sub> sample collected from Yorkville, GA during  
 762 summer 2010.



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764 **Figure 3.** LDH release for (a) clean air controls, (b) EAVES exposures, normalized to incubator  
 765 control, and (c) resuspension exposures, normalized to KBM only control. \*\*p<0.005 and  
 766 \*\*\*p<0.0005.

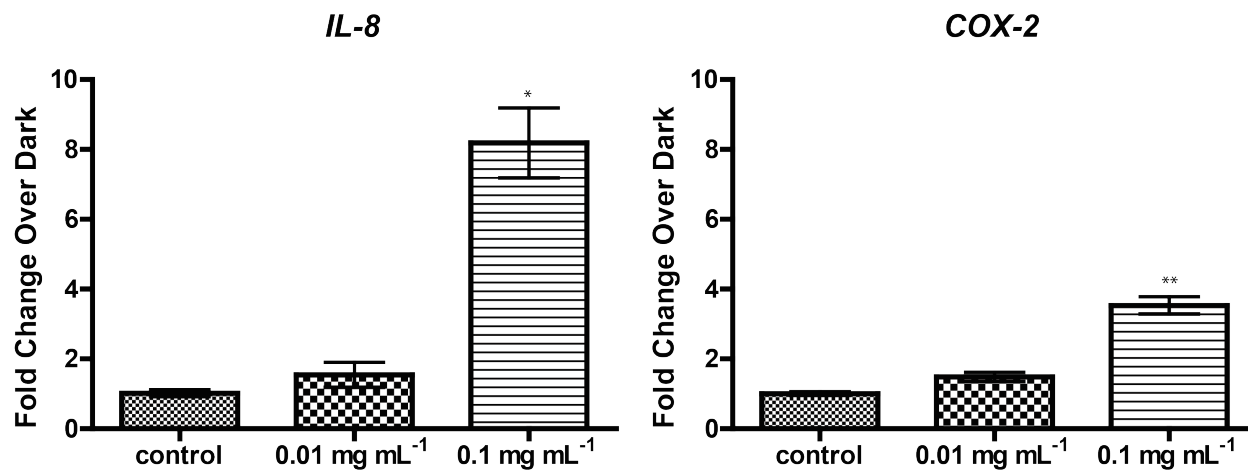


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768 **Figure 4.** *IL-8* and *COX-2* mRNA expression induced by exposure to isoprene-derived SOA  
 769 using EAVES device all normalized to dark control experiments and against housekeeping gene,  
 770  $\beta$ -actin. All experiments conducted in triplicate. \*\*\* $p < 0.0005$ .

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771  
 772 **Figure 5.** *IL-8* and *COX-2* expression induced by exposure to isoprene-derived SOA using  
 773 resuspension method all normalized to dark control experiments and against housekeeping gene,  
 774  $\beta$ -actin. All experiments conducted in triplicate. \* $p < 0.05$  and \*\* $p < 0.005$ .