We thank the two anonymous reviewers for their helpful comments. The reviewer comments are in black and our responses are in blue and include page and line numbers where changes were made to the accompanying manuscript.

Response to Reviewer #1

The crux of the puzzle being addressed in this paper is stated on page 2, lines 15 and 16, which states: [There appears to be a mechanism that maintains the neutralization ratio at a low value even as the SO2/NH3 emission ratio decreases.] Ignoring for the moment the ambiguity of the term neutralization ratio (discussed below), this very issue was identified in a recent paper [Weber et al., 2016] where it was shown that a thermodynamic model predicts this precise drop in neutralization ratio (ie, molar ratio) with decreasing SO4= and steady NH3 levels. The cause was identified to be due to the semi-volatile nature of ammonia and the insensitivity of the system to gas phase NH3 concentrations. Equilibrium partitioning of ammonia predicts lower ammonium-sulfate molar ratios as sulfate concentrations drop. The model was verified by comparison to observations [Guo et al., 2016; Guo et al., 2015]; the thermodynamic model predictions of the partitioning (gas- particle concentration ratios) of semi-volatile species, such as ammonia and nitric acid, are consistent with observations. Furthermore, the paper explained why little nitrate aerosol has been observed despite large reductions in sulfate.

We have amended the text (page 2 line 24-page 3 line 1) to address the findings of Weber et al. (2016) and to put our work in context of that paper. This is very useful. Our work goes beyond the Weber paper by pointing out the thermodynamic inconsistency in the simultaneous observations of low ammonium-sulfate ratios (R) and significant gas-phase ammonia. The Weber paper claims consistency but the values of R from their model calculations (Figure 2, "past") are higher than observed. We now point out that low values of nitrate are consistent with low values of R, citing Guo et al. (2015) and Weber et al. (2016) (page 6 lines 2-3).

In this paper, the authors propose a different cause for acidity (actually a NH4+/2SO4= molar ratios less than 1) despite the presence of gas phase ammonia. They suggest that impedance of gas phase ammonia uptake by fine aqueous particles due to an outer organic film is the cause. Simply put, since sulfate is nonvolatile, this increases the time to equilibrium, results in an average (I guess) predicted lower particle ammonium concentration, which lowers the molar ratio. There is no discussion on what happens when equilibrium is reached; presumably this model cannot explain the molar ratios at that point. A question is, why the different conclusions. One explanation may be due to the different analysis approach taken. One group has focused on pH predictions, and then using the predicted pH to assess it's affects on verifiable processes (comparing predicted and measured partitioning of various species), whereas in this work, the question is framed around observed vs predicted ammonium/sulfate molar ratios and using them to interpret ambient data (referred to here as MR). No other verification is provided (ie, comparison of some pH dependent process or species concentration to observations, such as nitrate aerosol).

Our main focus in this paper is to point out the thermodynamic inconsistency between observed low ammonium-sulfate ratios and gas-phase ammonia, clarified now in the text (page 2 line 32-

page 3 line 1). We would be remiss if we didn't come up with a tentative explanation, but we now use the word "tentative" in the abstract and text, because it is certainly not our intention to claim that this crude parameterization of kinetic limitation to uptake is the final word. We point out in the conclusions that this explanation would have a number of other implications but to our knowledge there are no available observations that would rule it out. As requested by the reviewer we have added an evaluation of gas-phase ammonia (page 9 lines 18-22, 29-32 with a new Figure 7, page 24).

First off, do the authors have a possible explanation why simply ammonia volatility is insufficient to explain these observations? It would seem prudent to least include a brief discussion stating that an alternative explanation to what they are proposing has been offered, briefly describe it and state why an alterative explanation is necessary. The logic used here seems odd. The authors use E-AIM or ISORROPIA in the analysis (and seem to accept that they largely agree with each other), but adjust the NH3 uptake to get these model predicted MR to agree with observations. But it has been shown that the same model (ISORROPIA) explains these observations. I conclude the issue is that MRs are not a robust nor reliable way to assess model predicted aerosol pH nor it's effects, and this is the fundamental cause for the different views.

The Weber et al. (2016) paper does not show agreement between model and observed R. We now devote a new paragraph of the introduction to point this out – this is indeed important to show how our paper goes beyond Weber et al. so we are very thankful to the reviewer for raising the issue and we believe that we have addressed it in the revision.

Why do the authors use MRs when investigating particle pH? Although commonly used in many different forms (e.g., ion balances, gas ratios, etc) it has been shown that MR do not provide quantitative insight on pH [Guo et al., 2016; Guo et al., 2015; Hennigan et al., 2015; Weber et al., 2016], but instead can lead to confusion and imprecise conclusions (see a list of them in specific comments below). Given this, it is not clear what their utility is. The author's show that the particle MR is not related to pH in any straightforward way (see fig 1) and that pH remains very low despite MR near or approaching 1. Given that pH is what controls aerosol processes related to acidity, such as partitioning of semi-volatile acids, acid catalyzed reactions, etc, and not MR, the use of MR to infer acidity is problematic.

Our focus is not on investigating particle pH. Our focus is on reconciling the joint occurrence of low R and high gas-phase ammonia. We have clarified this in revision, and also completely purged the word "neutralization" from the text at the reviewer's suggestion and to make this clearer.

As an example, what do the authors mean by their commonly used term, sulfate neutralization? Or sometimes, the term is just, aerosol neutralization, such as in the figures and in the text (pg 5 line 32). It is also in the title of the paper. Strictly, is one to interpret this to mean that the MR is 1 or greater? However, it also seems to be used to imply (or at least it gives the impression) that the authors are referring to overall particle acidity (i.e., pH), such that a MR of 1 implies a neutral aerosol. Example, page 8 line 7 and 8 states: quote [The aerosol in the standard model is fully neutralized throughout the eastern US (mean f = 1.00), at odds with observations.], end quote. But than Fig 1 shows that the pH is nowhere near neutral as f approaches 1; a

contradiction. Including with their Fig 1, a plot of specifically f vs pH, would be very insightful, I believe. There are many instances in this paper of similar ambiguous statements (discussed more below).

In order to address the reviewer's concern we have (1) eliminated the "neutralization" terminology throughout the text, and (2) replaced *f* by $R = [NH_4^+]/[S(VI)]$ to facilitate comparison with Weber et al. (1) should help to avoid the impression that we are focusing on aerosol pH, which we are not. See response above regarding distinction between pH and MR. Figure 1 (page 18) shows a plot of *R* and pH.

In any case, molar ratios are used in this work. The author's point is that the molar ratio predicted by the thermodynamic model is higher than what is observed, and so there is either an issue with the thermodynamic model or with the assumptions the model is based on. This paper proposes the latter, that NH3 mass transport limitations due to organic films are the cause and that it takes over a day for equilibrium to be established, thus the strict use of thermodynamic models, which assume equilibrium, give incorrect results. However, maybe other causes for the discrepancy are possible, eg, relating to oversimplification when using the model, such as lack of size dependent composition and the mixing state of the aerosol. What if one simply adds a small amount of other non-volatile cations, such as mineral dust, to the model; a reasonable thing to do since the data are PM2.5. This should lower the predicted MRs. Considering the likely small amount of other cations needed to significantly alter the MR, and measurement uncertainties (discussed more below), it is not clear how the authors can outright concluded that other cations play no role (ie, reference to Kim 2015 and Liao 2015). More details, provided in this paper, such as including sensitivity tests with various amounts of other cations could be very insightful.

We now address the lack of size dependent composition in the text (page 4 lines 6-7). We cannot conjure a scenario where size-dependent composition would explain the low value of R so it really shouldn't matter to the argument, and indeed Weber et al. didn't bother with it. At the reviewer's suggestion we now specifically quantify the small effect of soil dust cations in the data that we analyze and point out that the effect is small (page 5 lines 25-26).

Another issue that is not clear is the extensive comparisons of parameters, such as aerosol composition and MRs, between measurements of wet deposition versus fine particles. In the abstract it is used to infer that gas phase ammonia is in excess (meaning, presumably, not equal to zero). Fig 3 compares wet deposition neutralization ratios to those of fine particles. Generally the authors find that the wet deposition trends for sulfate and ammonium follow SO2 and NH3 emission trends and that the neutralization ratios (ie, MRs) are close to 1, as expected for excess NH3, whereas neither is true for fine particles. Are the authors inferring this provides evidence for some additional process affecting the fine particles? It is actually thermodynamically consistent and is due to the vastly different concentrations of liquid water (many orders of magnitude) associated with these systems. It is not clear how differences in MRs between fine haze particles and fog/cloud particles support the hypothesis of this manuscript and needs to be clarified? Maybe that was not the intent, but the use of wet deposition is confusing and should be clarified.

We have clarified our use of wet deposition data as it relates to emissions of ammonia and SO_2 on page 5 lines 3-6. We have also now included explicit discussion of gas-phase ammonia (page 4 lines 4-6; page 9 lines 18-22, 29-32; page 18, Figure 1; page 24, Figure 7).

Overall, this paper proposes an explanation for the persistent particle acidity despite reductions in SO2 emissions in the eastern US. Understanding the cause is clearly of great importance. In my view, the proposed explanation is speculative and largely unsupported because the use of MRs in their analysis is highly suspect. Furthermore, a much simpler explanation exists. I do believe the authors could test their hypothesis further; say by predicting concentrations of semivolatile aerosol components that depend on pH, in different settings. I am not against publishing this paper, but believe that significant clarifications are needed throughout prior to publication. More specific detailed questions are provided below

Specific comments (in the following [. . .] indicates a quote from the manuscript)

How was the pH in Fig 1 calculated? Is it the equilibrium predicted pH, (ie, no kinetic limitation due to organic layer), or avg pH in the overall system that includes predicted pH before equilibrium is reached?

We added a clarification to the text that the pH in Figure 1 is for thermodynamic equilibrium (page 3, lines 31-32).

It is often not clear exactly what data were used for gas phase ammonia in the thermodynamic models. Was the wet deposition data used under the assumption that it represents the total sulfur and ammonia, and thus as model input? This is likely reasonable, but leads to much higher uncertainty than using direct measurements of the species involved (sulfate, ammonium, ammonia). Granted the wet deposition data is more widely available then NH3, but still analysis for studies where all the data exists would greatly support the hypothesis of this work.

Yes indeed, and we thank the reviewer for the suggestion. We have added an evaluation of gasphase ammonia measurements compared to model predictions (page 9 lines 18-22, 29-32; Figure 7 page 24) as well as additional discussion of ammonia predicted by thermodynamic models (page 4 lines 4-6; page 18, Figure 1). The wet deposition data shown on page 20, Figure 3 is compared to thermodynamic models under the assumption that it represents total S(VI) and NH_x and is now stated in the text (page 7 lines 5-6).

If this organic film inhibits uptake of NH3, how it is likely to influence other semi-volatile species (nitric acid, HCl) and how will this impact the partitioning of these compounds. Will the model be able to accurately predict partitioning of these other species? Why not use this as a means of testing the hypothesis?

We now stress in the text that the hypothesis is tentative and that our representation of the OA effect is crude. Our focus here is mostly to offer a direction for future work.

One might expect this proposed process to be highly important in biomass burning plumes where the organic aerosol fraction is very high. Thermodynamic models appear to also work well in these cases [Bougiatioti et al., 2016]. Based on the arguments in this work one may expect them

not to. The authors could investigate if there is other published work on thermodynamic predictions in smoke plumes that support their hypothesis.

Bougiatioti et al. (2016) focused on predicting aerosol pH and on average conditions in that study were more sulfate-rich than is typical of the Southeast US summer (33.8% organic aerosol by mass in Bougiatioti et al., 2016 in contrast to 55% organic aerosol in August-September 2013 in the Southeast US shown in Kim et al., 2015).

In the Abstract it states: [...sulfate aerosol is not fully neutralized even in the presence of excess ammonia, at odds with thermodynamic equilibrium models]. Is it ever shown in this paper that this statement is true? There is only a general reference to Seinfeld and Pandis. A thermodynamic calculation that supports this statement is needed. Also from the Abstract, last line states: [If sulfate aerosol becomes more acidic as OA/sulfate ratios increase, then controlling SO2 emissions to decrease sulfate aerosol will not have the co-benefit of suppressing acid-catalyzed secondary organic aerosol (SOA) formation.] This requires knowing how this proposed process specifically affects pH, which I don't believe is ever discussed. That is, what is the effect of this organic coating on aerosol pH?

Figure 1 (page 18) shows that ISORROPIA predicts *R* approaching 2 in the presence of excess ammonia and is stated in the text (page 4 lines 2-4).

We now state in the Conclusions (page 10 lines 23-25) the implication of the organic coating on the aerosol pH trend and how it could explain the observations of Weber et al. (2016).

Page 3, line 2-6 states: [Nitrate partitions into the aerosol only when ammonia is in excess of sulfate neutralization, limited by the supply of excess ammonia and dependent on temperature.] Can one provide a thermodynamic analysis proving this is true? Fig A above, suggests otherwise.

Figure 1 does not include nitrate, stated in the text (page 3 lines 23-24). Citations of the relevant literature supporting the statement in the text have been added (page 3 lines 21-22; Ansari and Pandis, 1998; Park et al., 2004).

Page 3 last sentence states: [In any case, sufficient ammonia is emitted in the eastern US to fully neutralize the sulfate contributed by SO2 emissions.] Again, can this statement be shown to be true based on a thermodynamic calculation?

This is a statement based on the relative emissions of NH_3 to SO_2 shown in Figure 2 (page 19) and has been clarified in the text (page 4 lines 28-29).

Page 4 it states: [alkaline cations other than ammonium do not contribute significantly to the neutralization (Kim et al., 2015),...] A few lines explicitly stating how this is shown in Kim would be very useful (as noted above).

We have added discussion of the influence of alkaline cations in the text (page 5 lines 22-26).

Page 4 gives f(N) from various studies, including AMS data from SOAS and SEAC4RS. Is the reported +/- the standard deviation of the data or the uncertainty in f(N). If not uncertainty, what is the uncertainty? (As an example, AMS uncertainty of 35% is often reported, which would lead to an uncertainty of about 50% for f(N). Also, in side by side comparisons, such as in SOAS and WINTER studies, systematic differences in AMS and PILS measurements of sulfate, nitrate and ammonium can easily by 20%. I suspect similar levels apply to network data. It seems that measurement uncertainty alone, both random and systematic (the latter more important here), makes asserting that there is a difference between measured and model predicted f(N) tenuous.

The mean values of *R* stated in this manuscript are provided with a propagated uncertainty and we now clarify this in the text (page 5 lines 28-29). SEARCH PM_{2.5} measurements have a 1σ precision of 3% for sulfate and 5% for ammonium (as derived from the precision statistics reported by Edgerton et al., 2005) and CSN measurements have a 1σ precision of 6% for sulfate and 8% for ammonium (Flanagan et al. 2006). We have added to the text the corresponding propagated uncertainties for *R* (page 6 lines 8-9). We also note that while AMS absolute uncertainties of 35% for inorganic species are reported, the uncertainties on ratios of AMS species are lower, since the dominant source of uncertainty (collection efficiency) cancels out in the species ratio.

Page 5 Lines 29 to 31 it states: [We see that the observed aerosol neutralization does not follow thermodynamic predictions. . .] A similar statement is made in the last line on page 5. Specifically, what thermodynamic predictions are being referred to? Was a full thermodynamic model run to verify this? Please provide specific evidence to support this statement.

These statements refer to Figure 3 (page 20) as described in the text (page 7 lines 3-8) in which two thermodynamic models (E-AIM and ISORROPIA) are run and compared to observations. We have eliminated the term "neutralization" throughout the paper at the reviewer's suggestion and to clarify our intent.

Page 6 it states: [...the sulfate aerosol is far from neutralized, and one would expect under these conditions that the aerosol neutralization ratio should increase as the supply of sulfate decreases. But this is not observed.] What does, [one would expect], mean? Again, can these statements be proven with a thermodynamic calculation?

The statement refers to the thermodynamic predictions in Figures 1 and 3 that show the ammonium-sulfate ratio should increase as the $NH_x/S(VI)$ ratio increases and the text has been clarified on page 7 lines 17-22. Again, getting rid of "neutralization" should also help.

The proposed mechanism to bring model predictions closer to observed MRs is to lower the NH3 uptake. This results in the equilibrium time scales increasing to roughly a day or more. Presumably this will also affect fine particle nitric acid equilibrium time scales. Even ignoring any effect this film may have on nitric acid uptake. This could have a large impact on predicted fine particle nitrate concentrations, given that in many locations there is a significant mass of cations in the coarse mode (Na+, Ca2+, ...) that can form non-volatile forms of nitrate, if given sufficient time. Thus, it would seem the viability of this proposed mechanism could be tested by predicting nitrate aerosol levels and comparing to observations, for example during cooler

periods when more nitrate partitions to fine particles. It has been noted in a number of publications that ISORROPIA can accurately predict nitrate (without this proposed mechanism). How do the authors explain this if by their mechanism the aerosol is not in equilibrium?

Nitrate concentrations are very low in the Southeast US summer, now stated in the text (page 5 lines 19-22). We have added an evaluation of gas-phase ammonia in GEOS-Chem compared to observations from the SEARCH network (page 9 lines 18-22, 29-32; Figure 7, page 24) as further supporting evidence of a kinetic limitation.

Bougiatioti, A., P. Nikolaou, I. Stavroulas, G. Kouvarakis, R. Weber, A. Nenes, M. Kanakidou, and N. Mihalopoulos (2016), Particle water and pH in the eastern Mediterranean: Sources variability and implications for nutrients availability, Atm. Chem. Phys., 16, 4579-4591.

Guo, H., A. P. Sullivan, P. Campuzano-Jost, J. C. Schroder, F. D. Lopez-Hilfiker, J. E. Dibb, J. L. Jimenez, J. A. Thornton, S. S. Brown, A. Nenes, and R. J. Weber (2016), Fine particle pH and the partitioning of nitric acid during winter in the northeastern United States, J. Geophys. Res., in review.

Guo, H., L. Xu, A. Bougiatioti, K. M. Cerully, S. L. Capps, J. R. Hite, A. G. Carlton, S.-H. Lee, M. H. Bergin, N. L. Ng, A. Nenes, and R. J. Weber (2015), Predicting particle water and pH in the southeastern United States, Atmos. Chem. Phys., 15, 5211–5228.

Hennigan, C. J., J. Izumi, A. P. Sullivan, R. J. Weber, and A. Nenes (2015), A critical evaluation of proxy methods used to estimate the acidity of atmospheric particles, Atm. Chem. Phys., 15, 1-15.

Weber, R. J., H. Guo, A. G. Russell, and A. Nenes (2016), High aerosol acidity despite declining atmospheric sulfate concentrations over the past 15 years, Nature Geoscience, 9(10.1038/ngeo2665), 282-285.

References

Ansari, A. S., and Pandis, S. N.: Response of inorganic PM to precursor concentrations, Environmental Science & Technology, 32, 2706-2714, 10.1021/es971130j, 1998.

Edgerton, E. S., Hartsell, B. E., Saylor, R. D., Jansen, J. J., Hansen, D. A., and Hidy, G. M.: The southeastern aerosol research and characterization study: Part II. Filter-based measurements of fine and coarse particulate matter mass and composition, Journal of the Air & Waste Management Association, 55, 1527-1542, 2005.

Flanagan, J. B., Jayanty, R. K. M., Rickman, E. E., and Peterson, M. R.: PM2.5 speciation trends network: Evaluation of whole-system uncertainties using data from sites with collocated samplers, Journal of the Air & Waste Management Association, 56, 492-499, 2006.

Kim, P. S., Jacob, D. J., Fisher, J. A., Travis, K., Yu, K., Zhu, L., Yantosca, R. M., Sulprizio, M. P., Jimenez, J. L., Campuzano-Jost, P., Froyd, K. D., Liao, J., Hair, J. W., Fenn, M. A., Butler, C. F., Wagner, N. L., Gordon, T. D., Welti, A., Wennberg, P. O., Crounse, J. D., St Clair, J. M., Teng, A. P., Millet, D. B., Schwarz, J. P., Markovic, M. Z., and Perring, A. E.: Sources, seasonality, and trends of southeast US aerosol: an integrated analysis of surface, aircraft, and satellite observations with the GEOS-Chem chemical transport model, Atmospheric Chemistry and Physics, 15, 10411-10433, 10.5194/acp-15-10411-2015, 2015.

Park, R. J., Jacob, D. J., Field, B. D., Yantosca, R. M., and Chin, M.: Natural and transboundary pollution influences on sulfate-nitrate-ammonium aerosols in the United States: Implications for policy, Journal of Geophysical Research-Atmospheres, 109, 10.1029/2003jd004473, 2004.

Response to Reviewer #2

The authors present data suggesting that the sulfate in the Eastern US is not fully neutralized by ammonia despite the availability of the latter in the gas phase and argue that this could be due to mass transfer delays between the gas and particulate phase caused by organics. This is clearly an important issue that has received a lot of attention in the literature during the last few years. The paper is well written, however, the evidence provided to support its major conclusion is rather weak and it neglects other simpler explanations. It also neglects important work that has been published recently explaining similar observations with existing aerosol thermodynamics models. These problems are discussed below.

(1) Evidence that the aerosol is in equilibrium in the Southeast US The authors conclude from their data analysis that the aerosol in the eastern US is not in equilibrium probably due to limitations of the mass transfer of ammonia due to the organics. However, they neglect at least two studies that have used high quality gas-phase ammonia measurements in the same area and season to investigate this issue and reached the opposite conclusion. Weber et al. (2016) showed that the measured gas-phase ammonia was in equilibrium with the particulate phase during SOAS (see for example Figure 1 in that paper). They used 1-hr as averaging time therefore the equilibration timescale should be much less than this. If there was a significant delay in the mass transfer of ammonia between the two phases there should be significant observed discrepancies. Similar conclusions were reached by Nowak et al. (2006) for the 2002 ANARChE study in Atlanta. They concluded that the agreement between the measured and predicted ammonia within the uncertainties suggests that the assumption of thermodynamic equilibrium on the 7.5 min timescale is appropriate for most of the ANARChE data examined here.

We have amended the text to discuss the findings of Weber et al. (2016) (page 2 line 24-page 3 line 1) and identify the remaining need to reconcile simultaneous observations of low ammonium-sulfate ratios (*R*) and significant gas-phase ammonia. We have added a discussion of gas-phase ammonia predicted by thermodynamic models (page 4 lines 4-6; Figure 1, page 18) as well as model comparisons to ammonia observations (Figure 7, page 24). We address the findings of Nowak et al. (2006) on page 4 lines 9-11.

(2) Alternative explanations Weber et al. (2016) have argued that the incomplete neutralization of sulfate in the southeast US and the corresponding changes during the last 15 years are in general consistent with our current understanding of inorganic aerosol thermodynamics. They did not need to invoke delays in mass transfer and lack of equilibrium between the two phases. The authors appear to try something similar in Figure 3 but in a rather convoluted way using wet deposition data (please see comment 6 below).

See response to comment 1 above. Weber et al. (2016) cannot explain the low observed values of R and we have now clarified this in the text. See our responses to reviewer 1.

(3) Model evaluation The use of GEOS-Chem to test the hypothesis of delays in the mass transfer of ammonia is the most original aspect of the current work. However, the evaluation of the corresponding GEOS-Chem predictions is rather superficial given the availability of the SOAS measurements during the same summer. The ability of GEOS-Chem to reproduce observed gas-phase ammonia concentrations is critical for the authors' argument. For example, the problems of the base case simulation could be due to the overprediction of ammonia availability (e.g., due to errors in ammonia emissions) and the assumed delay in mass transfer could be correcting one error by introducing another.

In response to the reviewer, we have added an evaluation of gas-phase ammonia in GEOS-Chem in the standard model and model with limited NH_3 uptake compared to observations from the SEARCH network (page 9 lines 18-22, 29-32; Figure 7, page 24). We point out that different ammonia emission inventories for the US agree to within 20% in summer (page 4 lines 27-28; Paulot et al., 2014). We now emphasize the thermodynamic inconsistency between low *R* and high $NH_3(g)$ as a key original component of this work.

(4) Terminology and details about calculated quantities I found the terminology used in the paper quite confusing. For example, the terms "sulfate neutralization" and "aerosol neutralization" are used throughout the paper instead of the ammonium to sulfate molar or equivalent ratio. Sometimes the nitrate is also used sometimes it is not. In all cases the other cations that could, in principle at least, be neutralizing sulfate are not included in this neutralization ratio.

At the reviewer's suggestion we have eliminated the "neutralization" terminology throughout the text and have replaced it with "ammonium-sulfate ratios" defined as $R = [NH_4^+]/[S(VI)]$. When nitrate is included in this ratio, it is defined as $R_N = ([NH_4^+]-[NO_3^-])/[S(VI)]$ and stated explicitly. Discussion of the influence of other cations has been amended in the text (page 5 lines 22-26).

It is not clear if the different ratios shown throughout the paper correspond to PM1, PM2.5, PM10 or something else. This detail is critical given the importance of calcium, magnesium, etc., for the coarse particles.

The particle size of the measurements shown is now given on page 19 line 8 and the focus throughout the paper on fine particulate matter is now stated on page 2 lines 5-6.

It is also not clear how the authors estimate the averages of the different ratios. Do they average the concentrations and then estimate the ratios or do they estimate the ratios with some averaging time (daily?) and then average them?

Mean ratios are presented as the ratio of mean quantities, now stated in the text (page 4 line 25) and in the caption of Figure 2 (page 19 lines 12-13).

(5) Originality of hypothesis The same hypothesis regarding the organic aerosol role in mass transfer of ammonia in the southeast US has been presented Kim et al. (ACP, 2015) with a number of common authors in the two studies. This should be discussed in the introduction of the paper.

The text has been amended on page 3 lines 8-9 to address this point.

(6) Use of wet deposition data The authors use wet deposition measurements as practically a surrogate of total (gas and particulate) ammonia. This is rather tricky given that rainfall takes place during specific meteolorogical conditions, clouds can produce sulfate, and there are of course different wet removal efficiencies for ammonia and sulfate. Despite these problems, the paper does not address the potential biases that could be introduced in the analysis because of the use of these data. This could be one of the reasons for the differences between the conclusions here and those of other studies. I realize that gas-phase ammonia measurements exist only during specific field campaigns, but there are enough of them available both in the US and Europe. Use of these measurements is clearly preferable to the wet deposition data.

We have clarified our use of wet deposition data as it relates to emissions of ammonia and SO₂ on page 5 lines 3-6. We have also added a model comparison to gas-phase ammonia measurements (Figure 7, page 24).

(7) Role of organosulfates and crustal elements Some additional information is needed regarding the potential role of organosulfates and crustal elements for the present analysis. The authors appear to assume that they are negligible for the purposes of this work. However, some quantitative arguments are needed taking for example advantage of the SOAS measurements.

As stated in the text (page 3 lines 1-5, page 5 lines 22-26) organosulfates and crustal elements are present in low concentrations in the Southeast US summer based on observations. Organosulfates were also present in low concentrations during SOAS (Budisulistiorini et al., 2015; Hettiyadura et al., 2015; Rattanavaraha et al., 2016) and these references have been added in the text (page 3 lines 4-5). We have added a calculation of the effect of crustal components specifically for the data set that we analyze (page 5, lines 25-26).

(8) Some additional minor points:

Page 2, lines 2-4. This is clearly not true when there are other cations present.

Other cations are not present in high concentrations under the conditions examined in this paper as discussed in the response to the comment above.

Figure 2a is rather misleading given that most of the sulfur dioxide is emitted by point sources. It should probably be replaced with a table with the emissions in different regions of the US (e.g., southeast, northeast, etc.).

Emissions in the eastern US are stated in the text (page 4 line 30) as well as ratios for the Southeast and Northeast (page 6 line 32).

References

Budisulistiorini, S. H., Li, X., Bairai, S. T., Renfro, J., Liu, Y., Liu, Y. J., McKinney, K. A., Martin, S. T., McNeill, V. F., Pye, H. O. T., Nenes, A., Neff, M. E., Stone, E. A., Mueller, S., Knote, C., Shaw, S. L., Zhang, Z., Gold, A., and Surratt, J. D.: Examining the effects of anthropogenic emissions on isoprene-derived secondary organic aerosol formation during the 2013 Southern Oxidant and Aerosol Study (SOAS) at the Look Rock, Tennessee ground site, Atmospheric Chemistry and Physics, 15, 8871-8888, 10.5194/acp-15-8871-2015, 2015.

Hettiyadura, A. P. S., Stone, E. A., Kundu, S., Baker, Z., Geddes, E., Richards, K., and Humphry, T.: Determination of atmospheric organosulfates using HILIC chromatography with MS detection, Atmospheric Measurement Techniques, 8, 2347-2358, 10.5194/amt-8-2347-2015, 2015.

Paulot, F., Jacob, D. J., Pinder, R. W., Bash, J. O., Travis, K., and Henze, D. K.: Ammonia emissions in the United States, European Union, and China derived by high-resolution inversion of ammonium wet deposition data: Interpretation with a new agricultural emissions inventory (MASAGE_NH3), Journal of Geophysical Research-Atmospheres, 119, 4343-4364, 10.1002/2013jd021130, 2014.

Rattanavaraha, W., Chu, K., Budisulistiorini, H., Riva, M., Lin, Y. H., Edgerton, E. S., Baumann, K., Shaw, S. L., Guo, H. Y., King, L., Weber, R. J., Neff, M. E., Stone, E. A., Offenberg, J. H., Zhang, Z. F., Gold, A., and Surratt, J. D.: Assessing the impact of anthropogenic pollution on isoprene-derived secondary organic aerosol formation in PM2.5 collected from the Birmingham, Alabama, ground site during the 2013 Southern Oxidant and Aerosol Study, Atmospheric Chemistry and Physics, 16, 4897-4914, 10.5194/acp-16-4897-2016, 2016.