

Dear prof. Natascha Töpfer,

Thanks for giving us the opportunity to revise the paper "Options for mitigating global warming potential of a double-rice field in China" (**MS No.: acp-2016-227**) after public discussions. In the following we will consecutively address the points given by two anonymous referees (*in italic*) and, if appropriate, will make suggestions how to modify the manuscript.

**Anonymous Referee #1:**

*This study investigated CH<sub>4</sub> and N<sub>2</sub>O fluxes from a Chinese double-rice field and the responses to drainage and tillage in winter fallow season for 4 years, estimated the mitigation potential of drainage and tillage, and finally suggested the optimal land management strategies for reducing GWPs of CH<sub>4</sub> and N<sub>2</sub>O emissions in the double rice-cropping systems. More importantly, reasons for decreasing CH<sub>4</sub> and N<sub>2</sub>O emissions were well demonstrated by the measurements of total C and N contents and methanogens. The study provided useful agricultural strategies to mitigate global greenhouse gas emissions from Chinese double-rice fields. The experiment is well designed, and the high-quality data are well presented. The main conclusions are supported by the data. In general, this work is timely and very important with respect to our knowledge of options in winter fallow season for mitigating GWPs in the typical Chinese paddy fields. In particular, the paper made a good analysis of available data and discussed in detail. The manuscript is well presented, and the English is generally well written, although it has a potential to be improved. Overall, I do not have any major concerns but recommend it to be accepted by Atmospheric Chemistry and Physics.*

Thanks for your positive comments and useful suggestions. For more clarity, the language has been improved and perfected, and please refer to the revised manuscript for the detailed revisions.

**Anonymous Referee #2:**

*This paper reported 4 years fields experiment results to show how both CH<sub>4</sub> and N<sub>2</sub>O emissions from double-rice paddies affect by drainage and tillage managements in winter*

*fallow season in a typical subtropical climate zone in Southern China. The global warming potentials (GWPs) from CH<sub>4</sub> and N<sub>2</sub>O, greenhouse gas intensity per yield (GHGI) were also estimated in this paper. The data shown in the paper was reliable and calculation and statistical analyses were suitable. Please consider the minor points shown below for improving this manuscript.*

*1. The title should be clear, it can be changed as "Options of drainage and tillage managements in winter fallow season for mitigating global warming potential of a double-rice field in China".*

It is a good idea, thanks. For more clear and concise however, the title is supposed to be changed as "Drainage and tillage in winter fallow season mitigate global warming potential of a double-rice field in China" (Line 1~2 in the revised manuscript).

*2. Line 35, (WMO, 2014) can be renewed to (WMO, 2015). Also (WMO, 2015) should be listed in References.*

Thanks for your suggestions. The data and reference have been updated in the text (Line 34~35 in the revised manuscript). In addition, the Reference has been changed in the list (Line 623~624 in the revised manuscript).

*3. Line 45, (FAOSTAT, 2013), same as above.*

Thanks! The data and reference have been changed (Line 44~45 in the revised manuscript). Also, the Reference is revised in the list (Line 539~540 in the revised manuscript).

*4. Line 47, (Yearbook, 2013), same as above.*

Thanks so much! The Reference is revised both in the text (Line 47 in the revised manuscript) and the list (Line 636~637 in the revised manuscript).

*5. Line 78, before "In addition", one recent paper (Biol. Fertil. Soils (2016) 52:739–748) can be referred here.*

Thanks. It is very useful. The reference is cited in the text (Line 78 in the revised manuscript) and

it is supplemented in the list (Line 614~616 in the revised manuscript).

*6. Line 103, (Soil Survey Staff, 1975) was not found in References.*

Sorry for our carelessness, the reference is supplemented (Line 609~610 in the revised manuscript).

*7. Line 117-125, the management of rice straw from early rice season was not explained here.*

Thanks for your suggestion. A sentence “After early-rice harvest, rice straw and stubble were all moved out of the plots” is supplemented (Line 125~126 in the revised manuscript) to describe the rice straw management in the early-rice season.

*8. Line 146, (Myhre, 2013) as not found in References. It should be (Myhre et al., 2013) or (IPCC, 2013).*

Sorry for our carelessness. It has been revised in the text (Line 148 in the revised manuscript) and supplemented in the list (Line 585~591 in the revised manuscript).

*9. Line 347, "Parashar et al., 1993" should be before "Cai et al., 2003".*

Thanks. It has been revised (Line 350 in the revised manuscript). Additionally, similar problems in the text are all corrected.

*10. Line 467, delete "yr-1" before (Table 3).*

Sorry for our carelessness. It has been deleted (Line 470 in the revised manuscript). Moreover, similar problems in Table 3 and **Abstract** are all revised.

*11. Line 496, please check the subscript of N<sub>2</sub>O and CH<sub>4</sub> in this manuscript.*

*12. Line 497, it should be "Biol. Fertil. Soils," not Biol. Fert. Soils.,". The style of with or without DOI number should be consistent, for example, in line 525.*

Sorry for our carelessness. There (11 and 12) are all changed (Line 499~500 in the revised manuscript). In addition, the DOI numbers in the References are all deleted, please carefully refer to the list, thanks.

*13. Table 3, delete "yr-1" from the unit of GHGI.*

It has been revised (Table 3 in the revised manuscript).

*14. Table 5, (Myhre, 2013) as not found in References. It should be (Myhre et al., 2013) or (IPCC, 2013).*

Thanks. It has been revised (Table 5, Line 815 in the revised manuscript).

*15. Table 6, The C/N ratios of rice stubble increased after winter fallow season was easily understood, but it was confused readers why there were no data for NTD and NTND after late-rice harvest, and for TD and TND before early-rice transplanting.*

Thanks for your valuable suggestion. It should be noted that, firstly, there were two different times of tillage, i.e. tilling the field immediately after late-rice harvest in previous winter fallow season (Treatments TD and TND) and prior to early-rice transplanting during the following rice-growing season (Treatments NTD and NTND). Secondly, the contents of Total C and Total N in rice stubble were sampled and measured before early-rice transplanting. That is to say, rice stubble in Treatments TD and TND were buried under the soil while in Treatments NTD and NTND rice stubble were exposed to the air throughout the whole winter fallow season. Thereby, we can estimate the effect of tillage in winter fallow season on the degradation of rice straw by sampling rice stubble before early-rice transplanting and measuring the Total C and Total N contents. It is thus clear that, the phrases "after late-rice harvest" and "before early-rice transplanting" were just the times of soil tillage, not indicating the times of measurement (or times of data obtained). In addition, the data in Table 6 were from the measurements of rice stubble sampled before early-rice transplanting (Line 113~115 in the original manuscript). **Nevertheless**, Table 6 and its caption are changed for more understandable (Table 6, Line 832~833 in the revised manuscript). Please see below.

*16. Figure 4, putting NTD and NTND, and TD and TND in same bar graph were not suitable, there are independent treatments.*

Thanks! Certainly, it is more reasonable for showing the four of them apart, and in deed it was done before. Nevertheless, the **Figure 4** is presented like this, and it is still supposed to be kept in the

revised manuscript if the figure won't result in any misunderstandings. There are at least two reasons. Firstly, we put the measurements of NTD and NTND, and TD and TND in the same bar graph here mainly for emphasizing the importance of tillage to the abundance of methanogens and methanotrophs populations. Because the effect of drainage on the abundance of methanogens and methanotrophs populations in paddy soil is well known, however, the effect of tillage, particularly the impact of tillage in winter fallow season on the abundance of methanogens and methanotrophs populations in paddy soil during the previous winter fallow and following early- and late-rice seasons are scarcely reported. Secondly, it is more clear and better comparative by putting Treatment tillage (TD and TND) and Treatment non-tillage (NTD and NTND) together.

**Table 6** Measurements of total C ( $\text{g kg}^{-1}$ ) and total N ( $\text{g kg}^{-1}$ ) contents in rice stubble before early-rice transplanting in 2012 and 2013.

Year	Treatment	Total C	Total N	C/N
2012	TD	338	6.9	49
	TND	314	7.8	40
	NTD	356	12.7	28
	NTND	374	10.4	36
2013	TD	368	8.7	42
	TND	364	7.1	51
	NTD	404	12.8	32
	NTND	397	13.4	30

Thanks again! If the current manuscript still need revising, please feel free to let me know.

With best regards,

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1 Drainage and tillage in winter fallow season mitigate Options for mitigating  
2 **global warming potential of a double-rice field in China**

3  
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9  
10 **Abstract.** Traditional land managements (neither drainage nor tillage, NTND) in winter fallow season  
11 result in substantial CH<sub>4</sub> and N<sub>2</sub>O emissions from the double-rice fields in China. For investigating the  
12 effects of drainage and tillage in winter fallow season on global warming potentials (GWPs) of CH<sub>4</sub> and  
13 N<sub>2</sub>O emissions and developing mitigation options, a field experiment with four treatments: NTND,  
14 drainage but non-tillage (NTD), tillage but non-drainage (TND), and both drainage and tillage (TD) were  
15 carried out from 2010 to 2014 in a Chinese double-rice field. In winter fallow season total precipitation  
16 and mean daily temperature had important effects on CH<sub>4</sub> emission, and significant correlations were  
17 observed between them and CH<sub>4</sub> emission. Compared with NTND, drainage and tillage reduced CH<sub>4</sub>  
18 emission in early- and late-rice seasons and decreased annual emission by 54 and 33 kg CH<sub>4</sub> ha<sup>-1</sup> yr<sup>-1</sup>,  
19 respectively. Drainage and tillage increased N<sub>2</sub>O emission in winter fallow season while reduced it in  
20 early- and late-rice seasons, causing annual N<sub>2</sub>O emission unaffected. Accordingly, the GWPs were  
21 decreased by 1.49 and 0.92 t CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>, respectively, and they were far more reduced by  
22 combining drainage with tillage, with a mitigation potential of 1.96 t CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>. Low total C  
23 content and high C/N ratio in rice residues revealed that tillage in winter fallow season reduced CH<sub>4</sub> and  
24 N<sub>2</sub>O emissions in early- and late-rice seasons. Moreover, drainage and tillage significantly decreased the  
25 abundance of methanogens in paddy soil, which was a possible reason for the decrease of CH<sub>4</sub> emission.  
26 Greenhouse gas intensity was significantly decreased by drainage and tillage, and it was much more  
27 reduced by combining drainage with tillage, with a reduction of 0.17 t CO<sub>2</sub>-eq t<sup>-1</sup> yield yr<sup>-1</sup>. The results  
28 indicate that soil drainage combined with tillage in winter fallow season is an effective mitigating  
29 strategy in double-rice fields.

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## 31 **1 Introduction**

32 Methane (CH<sub>4</sub>) and nitrous oxide (N<sub>2</sub>O) are two of the most important greenhouse gases (GHGs) after  
33 carbon dioxide (CO<sub>2</sub>) in the atmosphere. According to the Greenhouse Gas Bulletin of World  
34 Meteorological Organization, the concentrations of atmospheric CH<sub>4</sub> and N<sub>2</sub>O reached at [1824](#) [1833](#) and  
35 [325.97](#) ppb in [2013](#)[2014](#), respectively (WMO, [2014](#)[2015](#)). Paddy fields are considered to be the major  
36 sources of atmospheric CH<sub>4</sub> and N<sub>2</sub>O. Since the 2000s, effective options for mitigating CH<sub>4</sub> and N<sub>2</sub>O  
37 emissions from paddy fields have been continually explored over the world (McCarl and Schneider,  
38 2001; Yan et al., 2005; Hussain et al., 2015), i.e. modifying irrigation and fertilization patterns (Cai et al.,  
39 2003; Hussain et al., 2015; Linquist et al., 2015), setting integrated soil–crop system management  
40 practices ([Zhang et al., 2013](#); Chen et al., 2014; [Zhang et al., 2013b](#)), and selection of suitable rice  
41 cultivar with high production but low GHGs emissions ([Su et al., 2015](#); [Ma et al., 2010](#); Hussain et al.,  
42 2015; [Su et al., 2015](#)[Ma et al., 2010b](#)), etc. Nevertheless, potential mitigating methods might be still  
43 available due to the diversity of rice-based ecosystems and the difference in agronomic management  
44 practices (Weller et al., 2016).

45 China is one of the largest rice producers in the world, and its harvested area contributes [18.59](#)% of  
46 the world total (FAOSTAT, [2013](#)[2014](#)). In China, total CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy fields were  
47 estimated to be 6.4 Tg yr<sup>-1</sup> and 180 Gg yr<sup>-1</sup>, respectively (Zhang et al., 2014). Double rice is the major  
48 rice-cropping system in China, accounting for over 40% of total rice cultivation area (Yearbook,  
49 [2013](#)[2014](#)) and emitting ca. 50% of the total paddy CH<sub>4</sub> in China (Zhang et al., 2011b; Chen et al., 2013).  
50 Double-rice fields mainly distribute at the south of the Yangtze River where usually has relative large  
51 precipitation and high temperature in winter fallow season. Traditionally, the fields are fallow in winter  
52 season with the soil neither drainage nor tillage after late-rice harvest, and they are usually subjected to  
53 visible floodwater after a heavy or a long-time raining. It is very likely to bring about CH<sub>4</sub> emission from  
54 these fields in winter fallow season and further to promote its emission during the following rice growth  
55 season. Modeling data had shown that CH<sub>4</sub> emission was significantly correlated with simulated soil  
56 moisture and mean precipitation of the preceding non-rice growth season (Kang et al., 2002). Incubation  
57 and pot experiments also affirmed that the higher the soil water contents in the non-rice growth season,  
58 the higher the CH<sub>4</sub> production rates and the more the CH<sub>4</sub> emissions in the subsequent rice season (Xu et  
59 al., 2003). An available mitigating option is hence proposed in this region, that is, the fields are drained  
60 to decrease the accumulation of rainwater in winter fallow season and finally to attenuate the positive

61 effect of winter precipitation on CH<sub>4</sub> emission. However, drainage possibly stimulates N<sub>2</sub>O emission  
62 from paddy field in winter fallow season because soil water content changes more quickly and  
63 intensively. It is well recognized that soil moisture regulates the processes of denitrification and  
64 nitrification and thus N<sub>2</sub>O emission ([Lan et al., 2013](#); Bateman and Baggs, 2005; [Lan et al., 2013](#)). Since  
65 the overall balance between the net exchange of CH<sub>4</sub> and N<sub>2</sub>O emissions constitutes global warming  
66 potentials (GWPs) of rice ecosystem, the effect of soil drainage in winter fallow season on mitigating  
67 GWPs year-round from the double-rice field is not well understood.

68 Soil tillage is a conventional practice in rice cultivation, and considerable reports have shown that  
69 tilling the soil prior to rice transplanting plays a key role in CH<sub>4</sub> and N<sub>2</sub>O emissions (Hussain et al., 2015;  
70 Zhao et al., 2016). Meanwhile, tillage after rice harvest in winter fallow season probably has very  
71 important effects on CH<sub>4</sub> and N<sub>2</sub>O emissions. Firstly, it is beneficial for the rainwater to penetrate into  
72 the subsoil, which won't lead to the accumulation of rainwater in winter fallow season. It is then difficult  
73 to form the strict anaerobic environments in the top soil, which not only reduces CH<sub>4</sub> emission directly  
74 during the non-rice growing season, but also indirectly inhibits CH<sub>4</sub> emission during the following rice  
75 season. On the contrary, tillage makes rice residues fully contact with the soil and microorganism, which  
76 may accelerate the decomposition of organic matters and then in favor of CH<sub>4</sub> production and emission  
77 in the non-rice growth season (Pandey et al., 2012; Hussain et al., 2015). Secondly, it may also play a  
78 key role in CH<sub>4</sub> emission during the following rice season owing to the incompletely decomposed rice  
79 residues ([Tang et al., 2016](#)). In addition, tillage in winter fallow season whether increases N<sub>2</sub>O emission  
80 from the field or not is still not very clear. There are some contradictive lines of evidence asserting the  
81 promotion and reduction in N<sub>2</sub>O emissions from rice fields by soil tillage. For instance, tillage changes  
82 the soil properties (soil porosity and soil moisture, etc.) and then promotes N<sub>2</sub>O emission ([Mutegi et al.,  
83 2010](#); Pandey et al., 2012; [Mutegi et al., 2010](#)) whereas incorporation of rice residues due to tillage may  
84 reduce N<sub>2</sub>O emission as a result of N immobilization (Huang et al., 2004; Ma et al., 2010a). Based on a  
85 3-year field measurement (Shang et al., 2011), the possible agricultural mitigating strategy that is crop  
86 residues incorporated into the soil accompanying with drainage in winter fallow season, has been  
87 proposed in a double-rice field. Nevertheless, the effects of drainage combined with tillage in winter  
88 fallow season on annual CH<sub>4</sub> and N<sub>2</sub>O emissions from double-rice fields, in particular on the  
89 corresponding mitigation potential are scarcely documented.

90 An *in situ* field measurement was conducted year-round for 4 years from 2010 to 2014 to study the

91 CH<sub>4</sub> and N<sub>2</sub>O emissions from a typical double-rice field in China. The objectives of this study are (1) to  
92 investigate the effects of soil drainage and tillage in winter fallow season on CH<sub>4</sub> and N<sub>2</sub>O emissions  
93 from the paddy field, (2) to estimate the mitigation potential of drainage and tillage, and thereby (3) to  
94 suggest the optimal land management strategies in winter fallow season for reducing GWPs of CH<sub>4</sub> and  
95 N<sub>2</sub>O emissions in the double rice-cropping systems in China.

96

## 97 **2 Methods and materials**

### 98 **2.1 Field site and experimental design**

99 The experimental field is located at Yujiang Town, Yingtan City, Jiangxi Province, China (28°15'N,  
100 116°55'E). The region has a typical subtropical monsoon climate with an annual mean temperature of  
101 about 18 °C and an annual precipitation of about 1800 mm. Prior to the experiment, the field was  
102 cultivated with early rice from April to July and late rice from July to November, and then kept in fallow  
103 for the rest of year. The soil type at the experimental field is classified as Typical Haplaquepts (Soil  
104 Survey Staff 1975). The initial properties (0–15 cm) of the soil are pH (H<sub>2</sub>O) 4.74, organic carbon (SOC)  
105 17.0 g kg<sup>-1</sup>, and total N 1.66 g kg<sup>-1</sup>. Daily air temperature (°C) and rainfall (mm) throughout the whole  
106 observational period was provided by Red Soil Ecological Experiment Station, Chinese Academy of  
107 Sciences (Appendix S1).

108 Four treatments, laid out in a randomized block design in triplicate, were conducted in the  
109 experimental field after late-rice harvest from 2010 to 2014: (1) the plots were neither drainage nor  
110 tillage in the whole winter fallow season as Treatment NTND, which is the traditional land management  
111 in the local region; (2) the plots were drainage but non-tillage as Treatment NTD; (3) the plots were  
112 tillage but non-drainage as Treatment TND; (4) and the plots were drainage and tillage simultaneously as  
113 Treatment TD. Rice stubble in all treatments was around 25–35 cm long, about 3.0–4.0 t ha<sup>-1</sup> during the  
114 4 winter fallow seasons, respectively. [Undergone the whole winter fallow season, a A](#) small portion of  
115 rice stubble was collected before early-rice transplanting [in 2012 and 2013](#), and the total C and N  
116 contents were measured by the wet oxidation-redox titration method and the micro-Kjeldahl method,  
117 respectively (Lu, 2000). Soil water content in winter fallow season was determined gravimetrically after  
118 drying at 105 °C for 8 h.

119 Local rice (*Oryza sativa* L.) cultivars, Zhongzao 33 and Nongxiang 98, were planted for the following  
120 early- and late-rice seasons, respectively. The seeds were sown in the seedling nursery and then

121 transplanted into the experimental plots at their 3- to 4-leaf stage. Each season, nitrogen (N) and  
122 potassium (K) fertilizations in form of urea and potassium chloride (KCl) were split into three  
123 applications, namely, basal fertilizers consisting of 90 kg N ha<sup>-1</sup> and 45 kg K ha<sup>-1</sup>, tillering fertilizers  
124 consisting of 54 kg N ha<sup>-1</sup> and 60 kg K ha<sup>-1</sup>, and panicle initiation fertilizers consisting of 36 kg N ha<sup>-1</sup>  
125 and 45 kg K ha<sup>-1</sup>. Phosphorus (P) fertilization in form of phosphorus pentoxide (P<sub>2</sub>O<sub>5</sub>) was applied to all  
126 the treatments as basal fertilizer at a rate of 75 kg P ha<sup>-1</sup>. [After early-rice harvest, rice straw and stubble](#)  
127 [were all moved out of the plots, and more D](#)detailed descriptions about the water management and  
128 fertilization [in early- and late-rice seasons are given are shown](#) in Appendix S2.

129

## 130 **2.2 CH<sub>4</sub> and N<sub>2</sub>O fluxes sampling and measurements**

131 Both CH<sub>4</sub> and N<sub>2</sub>O fluxes were measured once every 2–6 d and 7–10 d during the rice and non-rice  
132 seasons, respectively, using the static chamber technique (Zhang et al., 2011a). The flux chamber was 0.5  
133 × 0.5 × 1 m, and plastic base (0.5 × 0.5 m) for the chamber was installed before the experiment. Four gas  
134 samples from each chamber were collected using 18-mL vacuum vials at 15-min intervals. Soil  
135 temperature and soil redox potential (Eh) at 0.1 m depth were simultaneously measured during gas  
136 collection. Rice grain yields were determined in each plot at early- and late-rice harvests.

137 The concentrations of CH<sub>4</sub> and N<sub>2</sub>O were analyzed with gas chromatographs equipped with a flame  
138 ionization detector (Shimadzu GC-12A, Shimadzu Co., Japan) and with an electron capture detector  
139 (Shimadzu GC-14B, Shimadzu Co., Japan), respectively. Both the emission fluxes were calculated from  
140 the linear increase of gas concentration at each sampling time (0, 15, 30 and 45 min during the time of  
141 chamber closure) and adjusted for area and volume of the chamber. Sample sets were rejected unless  
142 they yielded a linear regression value of  $r^2$  greater than 0.90. The amounts of CH<sub>4</sub> and N<sub>2</sub>O emissions  
143 were calculated by successive linear interpolation of average CH<sub>4</sub> and N<sub>2</sub>O emissions on the sampling  
144 days, assuming that CH<sub>4</sub> and N<sub>2</sub>O emissions followed a linear trend during the periods when no sample  
145 was taken.

146

## 147 **2.3 GWPs and GHGI estimates**

148 The 100-year GWPs (CH<sub>4</sub> and N<sub>2</sub>O) in different treatments were calculated by using IPCC factors  
149 (100-year GWPs (CH<sub>4</sub> + N<sub>2</sub>O) = 28 × CH<sub>4</sub> + 265 × N<sub>2</sub>O) (Myhre [et al.](#), 2013). The greenhouse gas  
150 intensity (GHGI) represented the GWPs per unit rice grain yield (Li et al., 2006): GHGI = GWPs/grain

151 yield.

152

#### 153 **2.4 Soil sampling and DNA extraction**

154 During the 2013–2014 winter fallow and early- and late-rice seasons, soil samples were collected in the  
155 beginning, middle and end of each season from the experimental plots for analyzing the abundances of  
156 methanogens and methanotrophs. Totally, there were 108 soil samples (3 seasons × 3 stages in each  
157 season × 4 treatments × 3 replicates). Each sample was collected at 0–5 cm depth in triplicate and fully  
158 mixed. Subsequently, all samples were stored at 4 °C for analyses of soil characteristics and subsamples  
159 were maintained at –80 °C for DNA extraction.

160 For each soil sample, genomic DNA was extracted from 0.5 g soil using a FastDNA spin kit for soil  
161 (MP Biomedicals LLC, Ohio, USA) according to the manufacturer's instructions. The extracted soil  
162 DNA was dissolved in 50 µl of elution buffer, checked by electrophoresis on 1% agarose, and then  
163 quantified using a spectrophotometer (NanoDrop Technologies, Wilmington, DE, USA) (Fan et al.,  
164 2016).

165

#### 166 **2.5 Real-time PCR quantification of *mcrA* and *pmoA* genes**

167 The abundance of methanogenic *mcrA* gene copies and of methanotrophic *pmoA* genes copies was  
168 determined by quantitative PCR (qPCR) (Fan et al., 2016). Fragments of the *mcrA* and *pmoA* genes,  
169 encoding the methyl coenzyme-M reductase and the  $\alpha$  subunit of the particulate methane  
170 monooxygenase, respectively, were amplified using primers according to Hales et al. (1996) and  
171 Costello and Lidstrom (1999), respectively. Real-time quantitative PCR was performed on a CFX96  
172 Optical Real-Time Detection System (Bio-Rad Laboratories, Inc. Hercules, USA), and for the detailed  
173 descriptions please refer to our previous study (Fan et al., 2016).

174

#### 175 **2.6 Statistical analyses**

176 Statistical analysis was performed using SPSS 18.0 software for Windows (SPSS Inc., USA).  
177 Differences in seasonal CH<sub>4</sub> and N<sub>2</sub>O emissions, 100-year GWPs (CH<sub>4</sub> and N<sub>2</sub>O), and grain yields  
178 among treatments were analyzed with a repeated-measures one-way analysis of variance (ANOVA) and  
179 least significant differences (LSD) test. The significance of the factors (land management and year) was  
180 examined by using a two-way analysis of variance (ANOVA). Statistically significant differences and

181 correlations were set at  $P < 0.05$ .

182

### 183 **3 Results**

#### 184 **3.1 CH<sub>4</sub> emission**

185 Obvious CH<sub>4</sub> fluxes were observed over the 4 winter fallow seasons, particularly during the 2011–2012  
186 winter fallow season though a small net sink of CH<sub>4</sub> to the atmosphere was measured occasionally (Fig.  
187 1). Total CH<sub>4</sub> emissions of the 4 treatments were highly lower ( $P < 0.05$ ) in the 2010–2011 winter fallow  
188 season (~0.1–1 kg CH<sub>4</sub> ha<sup>-1</sup>) than the following three winter fallow seasons (~1–11 kg CH<sub>4</sub> ha<sup>-1</sup>), and  
189 they were ranged from 1.73 to 4.91 kg CH<sub>4</sub> ha<sup>-1</sup> on average (Table 1). Seasonal CH<sub>4</sub> emissions varied  
190 significantly with year and field managements (Table 2,  $P < 0.01$ ). Tillage increased CH<sub>4</sub> emissions by  
191 43–69% relative to non-tillage over the 4 winter fallow seasons. In comparison of non-drainage,  
192 drainage reduced CH<sub>4</sub> emissions by 40–50%. Consequently, CH<sub>4</sub> emission was decreased by 14.8%  
193 relative to Treatment NTND with the integrated effects of soil drainage and tillage (Table 1).

194 During the 4 early- and late-rice seasons, the CH<sub>4</sub> fluxes of all treatments dramatically ascended under  
195 continuous flooding, and the highest CH<sub>4</sub> fluxes were observed on about 20–30 days after rice  
196 transplanting in early-rice seasons and about 10–30 days after rice transplanting in late-rice seasons (Fig.  
197 1). Subsequently, they sharply decreased after midseason aeration. An obvious flux peak was observed  
198 again approximately 1–2 weeks after re\_flooding, particularly in the early-rice season. Apparently, the  
199 CH<sub>4</sub> emission always showed a higher flux peak in Treatment NTND than in Treatment TD.

200 Seasonal CH<sub>4</sub> emissions in early-rice season varied significantly with land managements, but it was  
201 not highly impacted by year or their interaction (Table 2). In contrast, total CH<sub>4</sub> emission did  
202 significantly vary with land managements and year in late-rice season (Table 2). In comparison of  
203 Treatment NTND, CH<sub>4</sub> emission was decreased by soil drainage and tillage, and on average, reduced by  
204 22.2% and 17.8% in early- and late-rice seasons, respectively (Table 1). Soil drainage combined with  
205 tillage further reduced CH<sub>4</sub> emission by 35.0% and 29.4% in early- and late-rice seasons, respectively.  
206 Compared with early-rice season (68.3–105.1 kg CH<sub>4</sub> ha<sup>-1</sup>), total CH<sub>4</sub> emission in late-rice season was  
207 8.0–17.9% greater.

208 Annually, total CH<sub>4</sub> emission was ranged from 151 to 222 kg CH<sub>4</sub> ha<sup>-1</sup>, averaged 46.1% and 52.1% of  
209 which came from the early- and late-rice seasons, respectively (Tables 1 and 3). Soil drainage and tillage  
210 played important roles in decreasing CH<sub>4</sub> emission. Relative to Treatment NTND, averaged CH<sub>4</sub>

211 emission was decreased by 24.3% and 14.9% by drainage and tillage, separately, and it was highly  
212 reduced by 32.0% when drainage was combined with tillage simultaneously (Table 3).

213

### 214 **3.2 N<sub>2</sub>O emission**

215 Substantial N<sub>2</sub>O emission was measured in the non-rice growth season though the fields were fallowed  
216 with no N-fertilization (Fig. 2 and Table 1). Total N<sub>2</sub>O emissions over the 4 winter fallow seasons varied  
217 significantly with land management and year while it did not significantly depended on their interaction  
218 (Table 2). Seasonal N<sub>2</sub>O emissions were relatively lower in the 2010–2012 winter fallow seasons than  
219 the following two winter fallow seasons. Compared with Treatment NTND, soil drainage and tillage  
220 generally increased N<sub>2</sub>O emissions, separately, and N<sub>2</sub>O emissions were significantly stimulated when  
221 combined drainage with tillage simultaneously. Over the 4 winter fallow seasons, seasonal N<sub>2</sub>O  
222 emissions averaged 36.4–68.2 g N<sub>2</sub>O–N ha<sup>-1</sup>, being 87.3%, 64.5% and 57.5% higher in Treatment TD  
223 than in Treatments NTND, TND, and NTD, respectively (Table 1).

224 After rice transplanting, pronounced N<sub>2</sub>O fluxes were observed with N-fertilization and midseason  
225 aeration, particularly in the period of dry/wet alternation (Fig. 2). Two-way ANOVA analyses indicated  
226 that seasonal N<sub>2</sub>O emissions during the early- and late-rice seasons were not highly influenced by land  
227 management, and the interactions of land management and year, except that N<sub>2</sub>O emissions depended  
228 significantly on year (Table 2). Compared with Treatments NTND and NTD, tillage increased N<sub>2</sub>O  
229 emission in 2011 early- and late-rice seasons whereas generally reduced N<sub>2</sub>O emission during the  
230 following rice seasons (Table 1).

231 Over the 4 early-rice seasons, drainage increased seasonal N<sub>2</sub>O emissions by 38.9–43.5% while tillage  
232 decreased by 10–12.9%, although no significant difference was observed (Table 1). In contrast, the  
233 effects of drainage and tillage seemed to be more important over the 4 late-rice seasons. For instance,  
234 drainage increased seasonal N<sub>2</sub>O emissions by 41.0–47.8% while tillage decreased by 10.3–14.4%.  
235 Annually, total N<sub>2</sub>O emission was ranged from 113 to 167 g N<sub>2</sub>O–N ha<sup>-1</sup>, averaged 34.4% of which  
236 derived from the winter fallow season (Tables 1 and 3). There was no significant difference in total N<sub>2</sub>O  
237 emission among the 4 treatments (Table 3).

238

### 239 **3.3 Global warming potential (GWP)**

240 Throughout the 4 winter fallow seasons, soil drainage and tillage had important effects on GWPs over

241 the 100-year time, although it was, on average, very small, being from 0.07 to 0.16 t CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>  
242 (Table 1). Compared with Treatment NTND, drainage significantly decreased GWPs while tillage highly  
243 increased it. Consequently, soil drainage combined with tillage played a slightly role in GWPs relative to  
244 Treatment NTND.

245 In contrast, both soil drainage and tillage decreased GWPs in comparison of Treatment NTND over  
246 the 4 early-rice seasons, with 16.0–36.2% and 4.2–36.2% lower in Treatment NTD and Treatment TND,  
247 respectively (Table 1). The GWPs was hence far more decreased by drainage combined with tillage,  
248 being 26.6–42.4% lower in Treatment TD than in Treatment NTND. Totally, drainage significantly  
249 reduced GWPs by 27.4% for Treatment NTD, in particular on Treatment TD by 34.8% with the  
250 integrated effect of drainage and tillage relative to Treatment NTND. Meanwhile, tillage tended to  
251 decrease GWPs relative to Treatment NTND but this effect was not statistically significant.

252 Similar effects of soil drainage and tillage on GWPs were observed over the 4 late-rice seasons (Table  
253 1). Compared with Treatment NTND, GWPs was 7.5–35.4% and 11.7–20.4% lower in Treatments NTD  
254 and TND, respectively. Soil drainage combined with tillage significantly decreased GWPs by  
255 23.7–36.8% for Treatment TD in comparison of Treatment NTND. On average, drainage and tillage  
256 reduced GWPs by 20.6% and 15%, separately, and GWPs was significantly reduced (29.1%) by  
257 combining drainage with tillage simultaneously.

258 Annually, the GWPs averaged 4.29–6.25 t CO<sub>2</sub>-eq ha<sup>-1</sup>, with 46% and 52% of which derived from the  
259 early-rice and late-rice seasons, respectively (Tables 1 and 3). Compared with Treatment NTND, GWPs  
260 was significantly reduced by 0.92–1.49 t CO<sub>2</sub>-eq ha<sup>-1</sup> in Treatments TND and NTD, respectively, and it  
261 was decreased much more by 1.96 t CO<sub>2</sub>-eq ha<sup>-1</sup> in Treatment TD (Table 3).

262

### 263 3.4 Rice grain yields

264 Grain yields of Treatments TND and TD [are were](#) generally higher than those of Treatments NTND and  
265 NTD over the 4 annual cycles (Table 1) though the yields slightly varied with land management and year  
266 as well as their interaction (Table 2). On average, the yields in Treatments TND and TD were over 6.5 t  
267 ha<sup>-1</sup>, 4.8%–7.3% and 3.1%–4.4% higher than those of Treatments NTND and NTD during the early- and  
268 late-rice seasons, respectively. Annually, no significance in the total yields was observed among the  
269 treatments over the 4 years (Table 3). Throughout the 4 late-rice seasons, positive correlation was  
270 observed between grain yields of 4 treatments and the corresponding CH<sub>4</sub> emissions ( $r= 0.733$ ,  $P <$

271 0.01).

272

### 273 **3.5 Greenhouse gas intensity (GHGI)**

274 Annual GHGI ranged from 0.32 to 0.49 t CO<sub>2</sub>-eq t<sup>-1</sup> yield, and it changed significantly among the  
275 treatments owing to GWPs highly controlled while annual rice yields slightly influenced by soil drainage  
276 and tillage (Table 3). Compared with Treatment NTND, drainage and tillage reduced GWPs by 23.8%  
277 and 14.7%, thus causing GHGI significantly decreased by 22.4% and 18.4%, separately. Expectedly, soil  
278 drainage combined with tillage reduced GHGI much more, with a reduction of 34.7% relative to  
279 Treatment NTND.

280

### 281 **3.6 Precipitation, temperature, soil Eh and soil water content in winter fallow season**

282 Over the 4 winter fallow seasons, total precipitation changed remarkably, which was ranged from ~400  
283 mm to ~750 mm during 2010–2012. Subsequently, it was relatively stable around 600 mm in 2012–2014  
284 (Table 4). In contrast, mean daily air temperature varied slightly, with values of ca. 9.0 °C to 10.0 °C.  
285 Soil Eh, on average, fluctuated obviously from the highest (~150 mV) in 2010–2011 to the lowest (~90  
286 mV) in 2013–2014. Soil water content in 2010 winter fallow season was generally higher in Treatment  
287 NTND than in Treatments NTD and TND, and it was lowest in Treatment TD (Fig. 3a), with a mean  
288 value of 55%, 50%, 44% and 38%, respectively. It is easy to see that the higher the precipitation and  
289 temperature, the lower the soil Eh, and thus the more the CH<sub>4</sub> emission in winter fallow season (Table 4).  
290 Statistical analyses show that a significant exponential relationship was observed between mean CH<sub>4</sub>  
291 emission and total precipitation (Fig. 3b,  $P < 0.01$ ), and mean CH<sub>4</sub> emission positively and negatively  
292 correlated with mean temperature (Fig. 3c,  $P < 0.05$ ) and soil Eh (Fig. 3d,  $P < 0.01$ ), respectively.

293

### 294 **3.7 Abundance of methanogens and methanotrophs populations**

295 The abundance of methanogens in paddy soil decreased significantly from winter fallow season to the  
296 following early-rice season, but it increased again during the late-rice season (Fig. 4a). Compared with  
297 non-drainage (Treatments NTND and TND), drainage (Treatments NTD and TD) generally decreased  
298 the abundance of methanogens throughout the winter fallow (Fig. 4a,  $P < 0.001$ ) and following early-  
299 and late-rice seasons (Fig. 4a,  $P < 0.05$ ). Relative to non-tillage (Treatments NTND and NTD), tillage  
300 (Treatments TND and TD) also significantly decreased the abundance of methanogens throughout the

301 winter fallow and following early- and late-rice seasons (Fig. 4a,  $P < 0.001$ ).

302 The abundance of methanotrophs was highest in winter fallow season, and then it decreased gradually  
303 (Fig. 4b). Drainage (Treatments NTD and TD) relative to non-drainage (Treatments NTND and TND)  
304 significantly decreased the abundance of methanotrophs over the winter fallow and early-rice seasons  
305 (Fig. 4b,  $P < 0.05$ ) though no significance during the late-rice season. In addition, tillage (Treatments  
306 TND and TD) significantly decreased the abundance of methanogens during the previous winter (Fig. 4b,  
307  $P < 0.001$ ) and following early-rice seasons (Fig. 4b,  $P < 0.01$ ) in comparison of non-tillage (Treatments  
308 NTND and NTD), except in the late-rice season.

309

#### 310 **4 Discussion**

##### 311 **4.1 CH<sub>4</sub> emission from double-rice fields**

312 It is reported that *in situ* measurement of CH<sub>4</sub> emission in China was firstly carried out from 1987 to  
313 1989 in a double-rice field in Hangzhou City (Shangguan et al., 1993b). Subsequently, more and more  
314 CH<sub>4</sub> emissions from double-rice fields were observed (Cai et al., 2001; Shang et al., 2011). However,  
315 few investigations were referred to related measurements in the non-rice growth season. Fortunately,  
316 Shang et al. (2011) found the double-rice fields in Hunan province China usually acting as a small net  
317 sink of CH<sub>4</sub> emission (as low as  $-6 \text{ kg CH}_4 \text{ ha}^{-1}$ ) in winter fallow season. Although an occasionally  
318 negative CH<sub>4</sub> flux was also observed over the 4 winter fallow seasons (Fig. 1), the double-rice field in  
319 this study was an entire source of CH<sub>4</sub> emission, in particular during the 2011–2012 winter fallow season  
320 (Table 1). On average, around 2% of annual CH<sub>4</sub> emission emitted from the winter fallow season.

321 Because of the residues (mainly including roots and stubble) of early rice as well as high temperature  
322 resulting in substantial CH<sub>4</sub> production in paddy fields (Shangguan et al., 1993a; Yan et al., 2005), CH<sub>4</sub>  
323 emission of late-rice season was generally higher than that of early-rice season. More importantly, a very  
324 high CH<sub>4</sub> flux peak was usually observed in a couple of days after late-rice transplanting (Cai et al., 2001;  
325 Shang et al., 2011). In the present study, CH<sub>4</sub> emission in late-rice seasons was  $80.1\text{--}113.5 \text{ kg CH}_4 \text{ ha}^{-1}$ ,  
326 being 8.0–17.9% larger than that of early-rice seasons (Table 1) though total CH<sub>4</sub> emission in the last two  
327 early-rice seasons was found to be slight greater than those in late-rice seasons (Fig. 1). Mean annual  
328 CH<sub>4</sub> emission varied between 151 and 222  $\text{kg CH}_4 \text{ ha}^{-1}$  over the 4 years (Table 3), which was much  
329 lower than previous results (Cai et al., 2001; Shang et al., 2011). Great differences in these CH<sub>4</sub>  
330 measurements were probably attributed to different water and rice straw managements.

331 Significant differences in CH<sub>4</sub> emission from the fields in winter fallow and late-rice seasons were  
332 observed (Table 2), indicating large changes in the interannual CH<sub>4</sub> emission. It is believed that the  
333 climatic variation may be the major factor leading to interannual variation of CH<sub>4</sub> emission at the  
334 macroscopic scale (Cai et al., 2009). In this study we found that total winter rainfall had an important  
335 effect on CH<sub>4</sub> emission, and the higher the rainfall, the greater the CH<sub>4</sub> emission throughout the 4 winter  
336 fallow seasons (Table 4). And an exponential relationship was observed between mean CH<sub>4</sub> emission and  
337 total rainfall in winter fallow season (Fig. 3b). The importance of rainfall in controlling CH<sub>4</sub> emission in  
338 winter fallow season, to some extent, also could be demonstrated by the negative relationships between  
339 mean soil Eh and CH<sub>4</sub> emission (Fig. 3d). According to different rice fields from 4 main rice growing  
340 regions in China, similar correlation was found between rainfall in winter fallow season and CH<sub>4</sub>  
341 emission in the rice growth season (Kang et al., 2002).

342 Nevertheless, we did not find any correlations between rainfall in winter fallow season and CH<sub>4</sub> flux  
343 in early-or late-rice season in this study, suggesting that rainfall in winter fallow season just significantly  
344 regulated CH<sub>4</sub> flux on-season, but didn't off-season. In contrast, a significant linear relationship was  
345 found ( $P < 0.01$ ) between CH<sub>4</sub> emissions and corresponding yields over the 4 late-rice seasons,  
346 demonstrating that crop growth benefited rice yield and biomass and thus stimulated CH<sub>4</sub> emission. It is  
347 reported that seasonal CH<sub>4</sub> emission depended greatly on rice biomass based on a long-term fertilizer  
348 experiment (Shang et al., 2011). Furthermore, changes in temperature over the 4 winter fallow seasons  
349 (Table 4) were supposed to play a key role in CH<sub>4</sub> emission, and the positive correlation had  
350 demonstrated this well (Fig. 3c). Many field measurements have shown the importance of temperature to  
351 CH<sub>4</sub> emission (Parashar et al., 1993; Cai et al., 2003; Parashar et al., 1993; Zhang et al., 2011a).

352

#### 353 4.2 Effect of soil drainage in winter fallow season on CH<sub>4</sub> emission

354 Considerable measurements of CH<sub>4</sub> emission as affected by soil drainage in winter fallow season have  
355 been reported from single-rice fields, and most of which were from the permanently flooded fields.  
356 Obviously, drainage significantly decreases CH<sub>4</sub> emission (Table 5). Draining the flooded fields inhibits  
357 CH<sub>4</sub> production and CH<sub>4</sub> emission in winter fallow season directly, and more importantly, it plays an  
358 important role in reducing CH<sub>4</sub> production and its emission in the subsequent rice-growing season  
359 (Zhang et al., 2011a). Compared with non-drainage, drainage in this study significantly decreased CH<sub>4</sub>  
360 emission both in previous winter fallow seasons and following early- and late-rice seasons (Table 1), and

361 over the 4 years, mean annual CH<sub>4</sub> emission was reduced by 38–54 kg CH<sub>4</sub> ha<sup>-1</sup> (Table 3). Such changes  
362 were very likely due to the decrease of methanogens in paddy soils throughout the winter, early- and  
363 late-rice seasons by soil drainage (Fig. 4a) because drainage increases soil aeration and hence effectively  
364 reduces the survival rate and activity of methane-producing bacteria. According to microcosm  
365 experiments, Ma and Lu (2011) found that the total abundance of methanogenic archaeal populations  
366 decreased by 40% after multiple drainages, and quantitative PCR analysis further revealed that both *mcrA*  
367 gene copies and *mcrA* transcripts significantly decreased after dry/wet alternation (Ma et al., 2012).

368

#### 369 4.3 Effect of soil tillage in winter fallow season on CH<sub>4</sub> emission

370 Although CH<sub>4</sub> emission in winter fallow season was increased by soil tillage, it was highly decreased  
371 during the following early- and late-rice seasons (Table 1), and over the 4 years, on average, it was  
372 reduced by 17–33 kg CH<sub>4</sub> ha<sup>-1</sup> yr<sup>-1</sup> (Table 3). Compared to non-tillage, tillage may promote the  
373 decomposition of rice residues, and then stimulates CH<sub>4</sub> production and emission in winter fallow season.  
374 By contrast, as the readily decomposable part of the residues has largely been decomposed after a whole  
375 winter fallow season, the remaining hardly-decomposable part of organic matter doesn't have much  
376 effect on promoting CH<sub>4</sub> emission next year (Watanabe and Kimura, 1998). The content of total C in rice  
377 residues generally lower in Treatments TND and TD than in Treatments NTND and NTD (Table 6) has  
378 well demonstrated that tillage decreased the carbon substrates for methanogenesis. It therefore, relative  
379 to non-tillage, significantly reduced CH<sub>4</sub> emission (Table 3). In a rice-wheat rotation system, our 2-year  
380 field measurements also showed that the carbon content of rice straw incorporated into the soil in winter  
381 fallow season was decreased sharply in comparison of that applied to the field just prior to rice  
382 transplanting (Zhang et al., 2015). In addition, tillage highly reduced the abundance of methanogens  
383 throughout the winter fallow and early- and late-rice seasons (Fig. 4a) should be a probable reason for  
384 the decrease of CH<sub>4</sub> emission.

385

#### 386 4.4 N<sub>2</sub>O emission from double-rice paddy fields

387 Direct N<sub>2</sub>O emission from rice-based ecosystems mainly happens in the periods of midseason aeration  
388 and subsequent dry/wet alternation in rice-growing season, and in winter crop or fallow season (Cai et al.,  
389 1997; Yan et al., 2003; Zheng et al., 2004; Cai et al., 1997; Ma et al., 2013; Yan et al., 2003). It is  
390 estimated that most of croplands N<sub>2</sub>O emission comes from uplands and just 20–25% of which is from

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391 rice fields in China (Zhang *et al.*, 2014). In China, field measurements of N<sub>2</sub>O emission began in 1992  
392 from a single-rice field in Liaoning province (Chen *et al.*, 1995), and considerable observations from  
393 double-rice fields had been performed (Xu *et al.*, 1997; Shang *et al.*, 2011; Zhang *et al.*, 2013a). The total  
394 N<sub>2</sub>O emission of early- and late-rice seasons in this study, on average, varied between 70.6 and 114.7 g  
395 N<sub>2</sub>O-N ha<sup>-1</sup> yr<sup>-1</sup> over the 4 years (Table 1), being significantly lower than those reported by Shang *et al.*  
396 (2011) and Zhang *et al.* (2013a) but similar to our previous measurements Ma *et al.* (2013). Furthermore,  
397 over 1/3 of annual N<sub>2</sub>O emission came from the winter fallow season (Table 1), indicating that N<sub>2</sub>O  
398 emission from paddy fields in winter fallow season was very important. Early field observations even  
399 showed that as high as 60–90% of N<sub>2</sub>O emission occurred in winter fallow season (Shang *et al.*, 2011).  
400 On a national scale, it is found that 41 Gg N<sub>2</sub>O-N yr<sup>-1</sup> emitted in the non-rice growth period, contributing  
401 45% of the total N<sub>2</sub>O emission from rice-based ecosystems (Zheng *et al.*, 2004). Although N<sub>2</sub>O emission  
402 from rice fields significantly affected by year (Table 2), reasons for the interannual variation were still  
403 not well known. In order to specify rules for interannual change in N<sub>2</sub>O emission, it is essential to  
404 maintain all-the-year-round long-term stationary field observations of N<sub>2</sub>O emission from the  
405 double-rice fields.

406

#### 407 4.5 Effect of soil drainage in winter fallow season on N<sub>2</sub>O emission

408 The production of soil N<sub>2</sub>O is mainly by the microbial processes of nitrification and denitrification while  
409 soil water content determines the general direction of the transformation of soil nitrogen. Soil drainage  
410 can cut down the soil water content and accelerate soil dry/wet alternation, thus promoting N<sub>2</sub>O emission  
411 from paddy fields (Davidson, 1992; Cai *et al.*, 1997). It is because that soil dry/wet alternation stimulates  
412 the transformation of C and N in the soil, in particular on the microbial biomass C and N turnover  
413 (Potthoff *et al.*, 2001). Expectedly, drainage usually decreased the soil water content in this study (Fig. 3a)  
414 and then increased N<sub>2</sub>O emission, on average, by 42% relative to non-drainage in winter fallow season  
415 (Table 1). Noted that drainage in previous winter fallow season also had an important effect on N<sub>2</sub>O  
416 emission from paddy fields during the following rice seasons, namely, it increased N<sub>2</sub>O emission both in  
417 early- and late-rice seasons (Table 1). It was possibly attributed to that drainage in winter fallow season  
418 would create soil moisture more beneficial to N<sub>2</sub>O production in the subsequent rice-growing seasons.  
419 Early report had well demonstrated that the production and emission of soil N<sub>2</sub>O was not only related to  
420 the soil moisture regime at the time, but also strongly affected by the previous soil moisture regime

421 (Groffman and Tiedje, 1988). And regardless of how the water conditions were at that time, the previous  
422 soil moisture conditions affected the concentration of reductase or synthetic ability of the enzymes, thus  
423 affecting denitrification (Dendooven and Anderson, 1995; Dendooven et al., 1996). Totally, annual N<sub>2</sub>O  
424 emission was increased by 37–48% compared drainage with non-drainage though there was no  
425 significant difference among the 4 treatments (Table 3).

426

#### 427 **4.6 Effect of soil tillage in winter fallow season on N<sub>2</sub>O emission**

428 Compared to non-tillage, tillage usually increased N<sub>2</sub>O emission in winter fallow season, on average, by  
429 39% over the 4 years (Table 1), which might be ascribed to two reasons. First, tillage increases soil  
430 aeration, which possibly promotes the process of nitrification. A soil column experiment has well  
431 demonstrated that moderate O<sub>2</sub> concentration is conducive to N<sub>2</sub>O production (Khdyer and Cho, 1983).  
432 Second, tillage accelerates rainwater from the plow layer percolating into the subsoil layer, stimulating  
433 the processes of soil dry/wet alternation and then promoting the transformation of N and production of  
434 N<sub>2</sub>O in the soil (Cai et al., 1997; Potthoff et al., 2001). Tillage usually decreased soil water content (Fig.  
435 3a) could validate this to some extent. In contrast, it had negative effects on N<sub>2</sub>O emission during the  
436 following early- and late-rice seasons, and mean N<sub>2</sub>O emission over the 4 years was reduced by 12% and  
437 13%, respectively (Table 1). Compared to non-tillage, tillage decreased the content of total N in rice  
438 residues, which probably reduced the substrates for nitrification and denitrification. More importantly,  
439 the ratio of C/N in rice residues was increased by tillage (Table 6). Because the decomposition of rice  
440 residues with high C/N ratio probably resulted in more N immobilization in the soil and less N available  
441 to nitrification and denitrification for N<sub>2</sub>O production (Huang et al., 2004; Zou et al., 2005). As a whole,  
442 soil tillage played a slight role in annual N<sub>2</sub>O emission over the 4 years (Table 3).

443

#### 444 **4.7 Effect of soil drainage and tillage on GWPs and GHGI**

445 Although drainage increased N<sub>2</sub>O emission throughout the winter fallow, and early- and late-rice seasons,  
446 it significantly decreased CH<sub>4</sub> emission from paddy fields (Table 1). As a consequence, it highly reduced  
447 GWPs, with a decrease of 1.49 t CO<sub>2</sub>-eq ha<sup>-1</sup> annually (Table 3). Considerable studies have showed that  
448 drainage results in a trade-off between CH<sub>4</sub> and N<sub>2</sub>O emissions from rice fields (Table 5), and it is widely  
449 considered to be an effective mitigation option. Annually, the mitigation potential of GWPs from paddy  
450 fields by drainage in winter fallow season is over 50%. However, these measurements are mostly related

451 to the single-rice fields with continuous flooding (Table 5), and few information are available about the  
452 effect on GWPs from double rice-cropping systems. In this study, we found that as high as 21–30% of  
453 the GWPs reduced by drainage in winter fallow season throughout the previous winter fallow and  
454 following early- and late-rice seasons, and with 24% of mitigation potential annually (Table 3).

455 In contrast, tillage obviously increased both CH<sub>4</sub> and N<sub>2</sub>O emissions, thus highly increased GWPs in  
456 winter fallow season (Table 1). Indeed, in a single-rice field, Liang et al. (2007) found that it increased  
457 the GWPs of CH<sub>4</sub>, N<sub>2</sub>O and CO<sub>2</sub> emissions in winter fallow season (Table 5). Fortunately, it significantly  
458 decreased CH<sub>4</sub> and N<sub>2</sub>O emissions both in early- and late-rice seasons, and as a result, with a reduction of  
459 GWPs by 17% and 15%, respectively (Table 1). Annually, the GWPs were reduced by 0.92 t CO<sub>2</sub>-eq  
460 ha<sup>-1</sup>, with 15% of mitigation potential (Table 3). As expected, the integrated effects of soil drainage and  
461 tillage decreased GWPs much more, with a further reduction by 1.04 t CO<sub>2</sub>-eq ha<sup>-1</sup> yr<sup>-1</sup>. Moreover, the  
462 annual mitigation potential (as high as 32%) of soil drainage combined with tillage in this study was in  
463 the ranges of previous results reported by Zhang et al. (2012) and Zhang et al. (2015) in single-rice fields  
464 (Table 5). It is obvious that the soil drainage together with tillage simultaneously in winter fallow season  
465 might be an effective option for mitigating the GWPs of CH<sub>4</sub> and N<sub>2</sub>O emissions from the double  
466 rice-cropping systems.

467 More importantly, no significant difference in rice grain yields was observed among the 4 treatments  
468 over the 4 years (Tables 1 and 3). It suggests that we would not risk rice yield loss when we try to  
469 decrease the GWPs of CH<sub>4</sub> and N<sub>2</sub>O emissions by means of soil drainage or tillage in winter fallow  
470 season. So, soil drainage and tillage significantly decreased GHGI by 22.4% and 18.4%, separately, and  
471 the GHGI was decreased much more by combining drainage with tillage, with a reduction of 0.17 t  
472 CO<sub>2</sub>-eq t<sup>-1</sup> yield yr<sup>-1</sup> (Table 3). Based on a long-term fertilizer experiment, balanced fertilizer  
473 management, in particular on P fertilizer supplement, was suggested to be an available strategy in double  
474 rice-cropping systems (Shang et al., 2011). In this study, the effective mitigation option in double-rice  
475 fields we proposed is that soil drainage combined with tillage in winter fallow season.

476 In Conclusion, the study demonstrated that in winter fallow season large differences in CH<sub>4</sub> emissions  
477 were probably due to the changes in total precipitation and temperature. Soil drainage and tillage in  
478 winter fallow season separately, in particular on combining both of them, significantly decreased CH<sub>4</sub>  
479 emission and then GWPs of CH<sub>4</sub> and N<sub>2</sub>O emissions from double-rice field. One possible explanation for  
480 this phenomenon is that drainage and tillage decreased the abundance of methanogens in paddy soil.

481 Moreover, low total C content in rice residues due to tillage was a potential reason for the decrease of  
482 CH<sub>4</sub> emission in the following early- and late-rice seasons. Finally, tillage reduced total N content but  
483 increased C/N ratio in rice residues would be important to the decrease of N<sub>2</sub>O emission. For both  
484 achieving high rice grain yield and low GWPs in double-rice fields, land management strategies in this  
485 study we proposed, including the fields were drained immediately after late-rice harvest, and meanwhile,  
486 the fields were tilled with rice residues incorporated into the soil. The results would benefit the  
487 development of optimal management strategies in the double-rice systems and the interpretation of the  
488 corresponding mechanisms.

489

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#### 500 **Reference**

501 Bateman, E. J., and Baggs, E. M.: Contributions of nitrification and denitrification to N<sub>2</sub>O emissions  
502 from soils at different water-filled pore space, *Biol. Fertil. Soils.*, 41, 379-388,  
503 [10.1007/s00374-005-0858-3](https://doi.org/10.1007/s00374-005-0858-3), 2005.

504 Bayer, C., Zschornack, T., Pedroso, G. M., Rosa, C. M. D., Camargo, E. S., Boeni, M., Elio Marcolin b,  
505 Reis, C. E. S. d., and Santos, D. C. d.: A seven-year study on the effects of fall soil tillage on yield-scaled  
506 greenhouse gas emission from flood irrigated rice in a humid subtropical climate, *Soil Till. Res.*, 145,  
507 118-125, 2015.

508 Cai, Z. C., Xing, G. X., Yan, X. Y., Xu, H., Tsuruta, H., Yagi, K., and Minami, K.: Methane and nitrous  
509 oxide emissions from rice paddy fields as affected by nitrogen fertilisers and water management, *Plant*

带格式的: 下标

510 Soil, 196, 7-14, [10.1023/a:1004263405020](https://doi.org/10.1023/a:1004263405020), 1997.

511 Cai, Z. C., Xu, H., Lu, W. S., Liao, Z. W., Wei, C. F., and Xie, D. T.: Influence of water management in  
512 winter crop season on CH<sub>4</sub> emission during the rice growing seasons, Chinese Journal of Applied Ecology,  
513 9, 171-175 (In Chinese), 1998.

514 Cai, Z. C., Tsuruta, H., Rong, X. M., Xu, H., and Yuan, Z. P.: CH<sub>4</sub> emissions from rice paddies managed  
515 according to farmer's practice in Hunan, China, Biogeochemistry, 56, 75-91, [10.1023/a:1011940730288](https://doi.org/10.1023/a:1011940730288),  
516 2001.

517 Cai, Z. C., Tsuruta, H., Gao, M., Xu, H., and Wei, C. F.: Options for mitigating methane emission from a  
518 permanently flooded rice field, Glob. Change Biol., 9, 37-45, 2003.

519 Cai, Z. C., Xu, H., and Ma, J.: Methane and nitrous oxide emissions from rice-based ecosystems (In  
520 Chinese), Science and Technology University of China Press, Hefei, China, 2009.

521 Chen, G. X., Huang, G. H., Huang, B., Wu, J., Yu, K. W., Xu, H., Xue, X. H., and Wang, Z. P.: CH<sub>4</sub> and  
522 N<sub>2</sub>O emission from a rice field and effect of *Azolla* and fertilization on them, Chinese Journal of Applied  
523 Ecology (In Chinese), 6, 378-382, 1995.

524 Chen, H., Zhu, Q. A., Peng, C. H., Wu, N., Wang, Y. F., Fang, X. Q., Jiang, H., Xiang, W. H., Chang, J.,  
525 Deng, X. W., and Yu, G. R.: Methane emissions from rice paddies natural wetlands, lakes in China:  
526 synthesis new estimate, Glob. Change Biol., 19, 19-32, [Doi 10.1111/Gcb.12034](https://doi.org/10.1111/Gcb.12034), 2013.

527 Chen, X., Cui, Z., Fan, M., Vitousek, P., Zhao, M., Ma, W., Wang, Z., Zhang, W., Yan, X., Yang, J., Deng,  
528 X., Gao, Q., Zhang, Q., Guo, S., Ren, J., Li, S., Ye, Y., Huang, J., Tang, Q., Sun, Y., Peng, X., Zhang, J.,  
529 He, M., Zhu, Y., Xue, J., Wang, G., Wu, L., An, N., Ma, L., and Zhang, F.: Producing more grain with  
530 lower environmental costs, Nature, 514, 486-489, [nature13609 \[pii\]](https://doi.org/10.1038/nature13609)  
531 [10.1038/nature13609](https://doi.org/10.1038/nature13609), 2014.

532 Costello, A. M., and Lidstrom, M. E.: Molecular characterization of functional and phylogenetic genes  
533 from natural populations of methanotrophs in lake sediments, Appl. Environ. Microb., 65, 5066-5074,  
534 1999.

535 Davidson, E. A.: Sources of nitric oxide and nitrous oxide following wetting of dry soil, Soil Sci. Soc.  
536 Am. J., 56, 95-102, 1992.

537 Dendooven, L., and Anderson, J. M.: Maintenance of denitrification potential in pasture soil following  
538 anaerobic events, Soil Biol. Biochem., 27, 1251-1260, [Doi 10.1016/0038-0717\(95\)00067-O](https://doi.org/10.1016/0038-0717(95)00067-O), 1995.

539 Dendooven, L., Duchateau, L., and Anderson, J. M.: Gaseous products of the denitrification process as

540 affected by the antecedent water regime of the soil, *Soil Biol. Biochem.*, 28, 239-245, [Doi](#)  
541 [10.1016/0038-0717\(95\)00132-8](#), 1996.

542 Fan, X., Yu, H., Wu, Q., Ma, J., Xu, H., Yang, J., and Zhuang, Y.: Effects of fertilization on microbial  
543 abundance and emissions of greenhouse gases (CH<sub>4</sub> and N<sub>2</sub>O) in rice paddy fields, *Ecol. Evol.*, 6,  
544 1054-1063, [10.1002/ece3.1879](#), 2016.

545 [Food and Agriculture Organization of the United Nations Statistics Division \(FAOSTAT\):](#)  
546 [http://faostat3.fao.org/download/Q/QC/E](#), 2014. (accessed on 19 July, 2016)

547 Groffman, P. M., and Tiedje, J. M.: Denitrification hysteresis during wetting and drying cycles in soil,  
548 *Soil Sci. Soc. Am. J.*, 52, 1626-1629, 1988.

549 Hales, B. A., Edwards, C., Ritchie, D. A., Hall, G., Pickup, R. W., and Saunders, J. R.: Isolation and  
550 identification of methanogen-specific DNA from blanket bog peat by PCR amplification and sequence  
551 analysis, *Appl. Environ. Microb.*, 62, 668-675, 1996.

552 Huang, Y., Zou, J. W., Zheng, X. H., Wang, Y. S., and Xu, X. K.: Nitrous oxide emissions as influenced  
553 by amendment of plant residues with different C : N ratios, *Soil Biol. Biochem.*, 36, 973-981, [DOI](#)  
554 [10.1016/j.soilbio.2004.02.009](#), 2004.

555 Hussain, S., Peng, S. B., Fahad, S., Khaliq, A., Huang, J. L., Cui, K. H., and Nie, L. X.: Rice  
556 management interventions to mitigate greenhouse gas emissions: a review, *Environ. Sci. Pollut. R.*, 22,  
557 3342-3360, [10.1007/s11356-014-3760-4](#), 2015.

558 Jiang, C. S., Huang, Y. S., Zheng, X. H., Zhu, B., and Huang, Y.: Effects of tillage-cropping systems on  
559 methane and nitrous oxide emissions from permanently flooded rice fields in a central Sichuan hilly  
560 area of south west China, *Environmental Science*, 27, 207-213 (In Chinese), 2006.

561 Kang, G. D., Cai, Z. C., and Feng, Z. H.: Importance of water regime during the non-rice growing period  
562 in winter in regional variation of CH<sub>4</sub> emissions from rice fields during following rice growing period in  
563 China, *Nutr. Cycl. Agroecosys.*, 64, 95-100, [Doi 10.1023/A:1021154932643](#), 2002.

564 Khdyer, I. I., and Cho, C. M.: Nitrification and denitrification of nitrogen fertilizers in a soil column, *Soil*  
565 *Sci. Soc. Am. J.*, 47, 1134-1139, 1983.

566 Lan, T., Han, Y., Roelcke, M., Nieder, R., and Cai, Z. C.: Processes leading to N<sub>2</sub>O and NO emissions  
567 from two different Chinese soils under different soil moisture contents, *Plant Soil*, 371, 611-627,  
568 [10.1007/s11104-013-1721-1](#), 2013.

569 Li, C. S., Salas, W., DeAngelo, B., and Rose, S.: Assessing alternatives for mitigating net greenhouse gas

带格式的：下标

570 emissions and increasing yields from rice production in china over the next twenty years, *J. Environ.*  
571 *Qual.*, 35, 1554-1565, [10.2134/jeq2005.0208](https://doi.org/10.2134/jeq2005.0208), 2006.

572 Liang, W., Shi, Y., Zhang, H., Yue, J., and Huang, G. H.: Greenhouse gas emissions from northeast China  
573 rice fields in fallow season, *Pedosphere*, 17, 630-638, [Doi 10.1016/S1002-0160\(07\)60075-7](https://doi.org/10.1016/S1002-0160(07)60075-7), 2007.

574 Linquist, B. A., Anders, M. M., Adviento-Borbe, M. A. A., Chaney, R. L., Nalley, L. L., da Rosa, E. F. F.,  
575 and van Kessel, C.: Reducing greenhouse gas emissions, water use, and grain arsenic levels in rice  
576 systems, *Glob. Change Biol.*, 21, 407-417, [10.1111/gcb.12701](https://doi.org/10.1111/gcb.12701), 2015.

577 Lu, R.: Soil agro-chemical analyses (In Chinese), Agricultural Technical Press, Beijing, 2000.

578 Ma, E., Zhang, G., Ma, J., Xu, H., Cai, Z., and Yagi, K.: Effects of rice straw returning methods on N<sub>2</sub>O  
579 emission during wheat-growing season, *Nutr. Cycl. Agroecosys.*, 88, 463-469,  
580 [10.1007/s10705-010-9369-1](https://doi.org/10.1007/s10705-010-9369-1), 2010a.

581 Ma, J., Ji, Y., Zhang, G. B., Xu, H., and Yagi, K.: Timing of midseason aeration to reduce CH<sub>4</sub> and N<sub>2</sub>O  
582 emissions from double rice cultivation in China, *Soil Sci. Plant Nutr.*, 59, 35-45, [Doi](https://doi.org/10.1080/00380768.2012.730477)  
583 [10.1080/00380768.2012.730477](https://doi.org/10.1080/00380768.2012.730477), 2013.

584 Ma, K., Qiu, Q. F., and Lu, Y. H.: Microbial mechanism for rice variety control on methane emission  
585 from rice field soil, *Glob. Change Biol.*, 16, 3085-3095, [DOI 10.1111/j.1365-2486.2009.02145.x](https://doi.org/10.1111/j.1365-2486.2009.02145.x), 2010b.

586 Ma, K., and Lu, Y. H.: Regulation of microbial methane production and oxidation by intermittent  
587 drainage in rice field soil, *FEMS Microbiol. Ecol.*, 75, 446-456, [DOI 10.1111/j.1574-6941.2010.01018.x](https://doi.org/10.1111/j.1574-6941.2010.01018.x),  
588 2011.

589 Ma, K., Conrad, R., and Lu, Y.: Responses of Methanogen mcrA Genes and Their Transcripts to an  
590 Alternate Dry/Wet Cycle of Paddy Field Soil, *Appl. Environ. Microb.*, 78, 445-454,  
591 [10.1128/aem.06934-11](https://doi.org/10.1128/aem.06934-11), 2012.

592 McCarl, B. A., and Schneider, U. A.: Climate change - Greenhouse gas mitigation in US agriculture and  
593 forestry, *Science*, 294, 2481-2482, [DOI 10.1126/science.1064193](https://doi.org/10.1126/science.1064193), 2001.

594 Mutegi, J. K., Munkholm, L. J., Petersen, B. M., Hansen, E. M., and Petersen, S. O.: Nitrous oxide  
595 emissions and controls as influenced by tillage and crop residue management strategy, *Soil Biol.*  
596 *Biochem.*, 42, 1701-1711, [10.1016/j.soilbio.2010.06.004](https://doi.org/10.1016/j.soilbio.2010.06.004), 2010.

597 [Myhre, G., D. Shindell, F. M. Brón, W. Collins, J. Fuglestedt, J. Huang, D. Koch, J. F. Lamarque, D.](#)  
598 [Lee, B. Mendoza, T. Nakajima, A. Robock, G. Stephens, and Zhang, T. T. a. H.: Anthropogenic and](#)  
599 [Natural Radiative Forcing. In: Climate Change 2013: The Physical Science Basis, Contribution of](#)

600 [Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change.](#)  
601 [Stocker, T.F., D. Qin, G-K. Plattner, M. Tignor, S.K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex and](#)  
602 [P.M. Midgley ed., edited by: Change, I. P. o. C., Cambridge University Press, Cambridge, United](#)  
603 [Kingdom and New York, NY, USA, 659-740 pp., 2013.](#)

604 Pandey, D., Agrawal, M., and Bohra, J. S.: Greenhouse gas emissions from rice crop with different  
605 tillage permutations in rice-wheat system, *Agr. Ecosyst. Environ.*, 159, 133-144,  
606 [10.1016/j.agee.2012.07.008](#), 2012.

607 Parashar, D. C., Gupta, P. K., Rai, J., Sharma, R. C., and Singh, N.: Effect of soil temperature on  
608 methane emission from paddy fields, *Chemosphere*, 26, 247-250,  
609 [http://dx.doi.org/10.1016/0045-6535\(93\)90425-5](http://dx.doi.org/10.1016/0045-6535(93)90425-5), 1993.

610 Potthoff, M., Joergensen, R. G., and Wolters, V.: Short-term effects of earthworm activity and straw  
611 amendment on the microbial C and N turnover in a remoistened arable soil after summer drought, *Soil*  
612 *Biol. Biochem.*, 33, 583-591, [Doi 10.1016/S0038-0717\(00\)00200-5](#), 2001.

613 Shang, Q., Yang, X., Gao, C., Wu, P., Liu, J., Xu, Y., Shen, Q., Zou, J., and Guo, S.: Net annual global  
614 warming potential and greenhouse gas intensity in Chinese double rice --cropping systems: a 3- - year  
615 field measurement in long- - term fertilizer experiments, *Glob. Change Biol.*, 17, 2196-2210, 2011.

616 Shangguan, X. J., Wang, M. X., Chen, D. Z., and Shen, R. X.: Methane production in rice paddy fields,  
617 *Advance in Earth Sciences (In Chinese)*, 8, 1-12, 1993a.

618 Shangguan, X. J., Wang, M. X., and Shen, R. X.: Regularity of methane emission from rice paddy fields,  
619 *Advance in Earth Sciences (In Chinese)*, 8, 23-36, 1993b.

620 Shiratori, Y., Watanabe, H., Furukawa, Y., Tsuruta, H., and Inubushi, K.: Effectiveness of a subsurface  
621 drainage system in poorly drained paddy fields on reduction of methane emissions, *Soil Sci. Plant Nutr.*,  
622 53, 387-400, [10.1111/j.1747-0765.2007.00171.x](#), 2007.

623 [Soil Survey Staff: Soil taxonomy: A Basic System of Soil Classification for Making and Interpreting Soil](#)  
624 [Surveys. in, U.S. Government Printing Office, Handbook No. 436. Washington, DC, 1975.](#)

625 Su, J., Hu, C., Yan, X., Jin, Y., Chen, Z., Guan, Q., Wang, Y., Zhong, D., Jansson, C., Wang, F., Schnurer,  
626 A., and Sun, C.: Expression of barley SUSIBA2 transcription factor yields high-starch low-methane rice,  
627 *Nature*, 523, 602-606, [10.1038/nature14673](#), 2015.

628 [Tang, S., Cheng, W., Hu, R., Guigue, J., Kimani, S. M., Tawaraya, K., and Xu, X.: Simulating the effects](#)  
629 [of soil temperature and moisture in the off-rice season on rice straw decomposition and subsequent CH<sub>4</sub>](#)

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630 [production during the growth season in a paddy soil, Biol. Fertil. Soils, 52, 739-748, 2016.](#)

631 Watanabe, A., and Kimura, M.: Effect of rice straw application on CH<sub>4</sub> emission from paddy fields IV.

632 Influence of rice straw incorporated during the previous cropping period, Soil Sci. Plant Nutr., 44,

633 507-512, 1998.

634 Weller, S., Janz, B., Jürg, L., Kraus, D., Racela, H. S. U., Wassmann, R., Butterbach-Bahl, K., and Kiese,

635 R.: Greenhouse gas emissions and global warming potential of traditional and diversified tropical rice

636 rotation systems, Glob. Change Biol., 22, 432-448, [10.1111/gcb.13099](#), 2016.

637 [World Meteorological Organization \(WMO\): The State of Greenhouse Gases in the Atmosphere Based](#)

638 [on Global Observations through 2014, Greenhouse Gas Bulletin, 1-4, 2015.](#)

639 Xu, H., Guangxi, X., Cai, Z. C., and Tsuruta, H.: Nitrous oxide emissions from three rice paddy fields in

640 China, Nutr. Cycl. Agroecosys., 49, 23-28, [Doi 10.1023/A:1009779514395](#), 1997.

641 Xu, H., Cai, Z. C., and Tsuruta, H.: Soil moisture between rice-growing seasons affects methane

642 emission, production, and oxidation, Soil Sci. Soc. Am. J., 67, 1147-1157, 2003.

643 Yan, X. Y., Akimoto, H., and Ohara, T.: Estimation of nitrous oxide, nitric oxide and ammonia emissions

644 from croplands in East, Southeast and South Asia, Glob. Change Biol., 9, 1080-1096, [DOI](#)

645 [10.1046/j.1365-2486.2003.00649.x](#), 2003.

646 Yan, X. Y., Yagi, K., Akiyama, H., and Akimoto, H.: Statistical analysis of the major variables

647 controlling methane emission from rice fields, Glob. Change Biol., 11, 1131-1141,

648 [10.1111/j.1365-2486.2005.00976.x](#), 2005.

649 Yao, Z. S., Zheng, X. H., Wang, R., Xie, B. H., Butterbach-Bahl, K., and Zhu, J. G.: Nitrous oxide and

650 methane fluxes from a rice-wheat crop rotation under wheat residue incorporation and no-tillage

651 practices, Atmos. Environ., 79, 641-649, 2013.

652 [Yearbook, E. B. o. C. A.: China Agriculture Yearbook 2014 \(In Chinese\), China Agriculture Press,](#)

653 [Beijing, 2014.](#)

654 Zhang, G. B., Zhang, X. Y., Ma, J., Xu, H., and Cai, Z. C.: Effect of drainage in the fallow season on

655 reduction of CH<sub>4</sub> production and emission from permanently flooded rice fields, Nutr. Cycl. Agroecosys.,

656 89, 81-91, [10.1007/s10705-010-9378-0](#), 2011a.

657 Zhang, H. L., Bai, X. L., Xue, J. F., Chen, Z. D., Tang, H. M., and Chen, F.: Emissions of CH<sub>4</sub> and N<sub>2</sub>O

658 under different tillage systems from double-cropped paddy fields in southern China, Plos One, 8,

659 [10.1371/journal.pone.0065277](#), 2013a.

660 Zhang, J. K., Jiang, C. S., Hao, Q. J., Tang, Q. W., Cheng, B. H., Li, H., and Chen, L. H.: Effects of  
661 tillage-cropping systems on methane and nitrous oxide emissions from agro-ecosystems in a purple paddy  
662 soil, *Environmental Science* ([In Chinese](#)), 33, 1979-1986 ([In Chinese](#)), 2012.

663 Zhang, W., Yu, Y. Q., Huang, Y., Li, T. T., and Wang, P.: Modeling methane emissions from irrigated rice  
664 cultivation in China from 1960 to 2050, *Glob. Change Biol.*, 17, 3511-3523, 2011**b**.

665 Zhang, W., Yu, Y., Li, T., Sun, W., and Huang, Y.: Net greenhouse gas balance in China's croplands over  
666 the last three decades and its mitigation potential, *Environ. Sci. Technol.*, 48, 2589-2597, 2014.

667 Zhang, W. F., Dou, Z. X., He, P., Ju, X. T., Powlson, D., Chadwick, D., Norse, D., Lu, Y. L., Zhang, Y.,  
668 and Wu, L.: New technologies reduce greenhouse gas emissions from nitrogenous fertilizer in China,  
669 *Proceedings of the National Academy of Sciences*, 110, 8375-8380, 2013**b**.

670 Zhang, Y. F., Sheng, J., Wang, Z. C., Chen, L. G., and Zheng, J. C.: Nitrous oxide and methane emissions  
671 from a Chinese wheat-rice cropping system under different tillage practices during the wheat-growing  
672 season, *Soil Till. Res.*, 146 261-269, 2015.

673 Zhao, X., Liu, S.-L., Pu, C., Zhang, X.-Q., Xue, J.-F., Zhang, R., Wang, Y.-Q., Lal, R., Zhang, H.-L., and  
674 Chen, F.: Methane and nitrous oxide emissions under no-till farming in China: a meta-analysis, *Glob.*  
675 *Change Biol.*, [22, 1372-1384, n/a-n/a, 10.1111/gcb.13185](#), 2016.

676 Zheng, X. H., Han, S. H., Huang, Y., Wang, Y. S., and Wang, M. X.: Re-quantifying the emission factors  
677 based on field measurements and estimating the direct N<sub>2</sub>O emission from Chinese croplands, *Glob.*  
678 *Biogeochem. Cycle*, 18, [10.1029/2003gb002167](#), 2004.

679 Zou, J. W., Huang, Y., Jiang, J. Y., Zheng, X. H., and Sass, R. L.: A 3-year field measurement of methane  
680 and nitrous oxide emissions from rice paddies in China: Effects of water regime, crop residue, and  
681 fertilizer application, *Glob. Biogeochem. Cycle*, 19, [Doi 10.1029/2004gb002401](#), 2005.

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710 **Figure captions:**

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712 **Figure 1** Seasonal variation of CH<sub>4</sub> emission from 2010 to 2014.

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715 **Figure 2** Seasonal variation of N<sub>2</sub>O emission from 2010 to 2014.

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718 **Figure 3** Soil water content in 2010 winter fallow season (a) and the relationships  
719 between mean CH<sub>4</sub> emission and total winter precipitation (b), and mean daily air

720 temperature (c) and soil Eh (d) over the 4 winter fallow seasons (Data from Table 4).

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723 **Figure 4** The abundance of methanogens and methanotrophs populations in paddy soil  
724 from 2013 to 2014, WS, ES, and LS means winter fallow season, early-rice season, and  
725 late-rice season, respectively.

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737 **Table 1** Seasonal CH<sub>4</sub> and N<sub>2</sub>O emissions, global warming potentials (GWPs), and rice grain yields over the 4 years from 2010 to 2014.

Year	Treatment	Winter fallow season			Early-rice season				Late-rice season			
		CH <sub>4</sub> emission (kg CH <sub>4</sub> ha <sup>-1</sup> )	N <sub>2</sub> O emission (g N <sub>2</sub> O-N ha <sup>-1</sup> )	GWPs (t CO <sub>2</sub> -eq ha <sup>-1</sup> )	CH <sub>4</sub> emission (kg CH <sub>4</sub> ha <sup>-1</sup> )	N <sub>2</sub> O emission (g N <sub>2</sub> O-N ha <sup>-1</sup> )	GWPs (t CO <sub>2</sub> -eq ha <sup>-1</sup> )	Yield (t ha <sup>-1</sup> )	CH <sub>4</sub> emission (kg CH <sub>4</sub> ha <sup>-1</sup> )	N <sub>2</sub> O emission (g N <sub>2</sub> O-N ha <sup>-1</sup> )	GWPs (t CO <sub>2</sub> -eq ha <sup>-1</sup> )	Yield (t ha <sup>-1</sup> )
2010–2011	TD	0.46 ± 0.02	46.4 ± 1.5	0.03 ± 0.01	61.3 ± 12.5	49.0 ± 7.2	1.74 ± 0.39	6.44 ± 0.82	133.9 ± 18.6	98.5 ± 4.3	3.79 ± 0.17	7.13 ± 0.07
	TND	1.05 ± 0.13	30.4 ± 3.1	0.04 ± 0.02	80.6 ± 2.4	46.6 ± 7.1	2.28 ± 0.06	6.29 ± 0.20	158.5 ± 28.3	67.4 ± 2.1	4.46 ± 0.40	7.33 ± 0.09
	NTD	0.11 ± 0.19	42.7 ± 5.3	0.02 ± 0.02	70.6 ± 6.1	45.3 ± 11.1	2.00 ± 0.16	6.08 ± 0.60	147.0 ± 15.6	62.8 ± 5.1	4.14 ± 0.02	6.72 ± 0.22
	NTND	0.38 ± 0.07	32.2 ± 5.1	0.02 ± 0.01	84.9 ± 14.3	38.9 ± 12.3	2.38 ± 0.29	5.82 ± 0.34	179.6 ± 26.2	44.5 ± 11.0	5.05 ± 0.15	6.83 ± 0.84
2011–2012	TD	5.06 ± 1.18	42.0 ± 1.8	0.16 ± 0.04	64.0 ± 12.5	17.7 ± 7.9	1.80 ± 0.35	6.67 ± 0.08	79.6 ± 8.8	45.2 ± 7.8	2.25 ± 0.24	6.63 ± 0.09
	TND	11.1 ± 2.51	35.1 ± 2.7	0.33 ± 0.07	90.6 ± 8.2	16.2 ± 7.2	2.54 ± 0.23	7.03 ± 0.50	103.1 ± 6.0	35.4 ± 8.0	2.90 ± 0.16	6.70 ± 0.21
	NTD	4.54 ± 0.32	27.3 ± 11.3	0.14 ± 0.04	68.1 ± 11.8	28.2 ± 6.1	1.92 ± 0.22	6.36 ± 0.36	81.0 ± 4.3	63.0 ± 9.6	2.30 ± 0.80	6.57 ± 0.35
	NTND	7.09 ± 1.08	14.1 ± 4.4	0.20 ± 0.05	107.1 ± 9.9	23.4 ± 4.8	3.01 ± 0.27	6.67 ± 0.47	126.4 ± 12.2	47.2 ± 11.0	3.56 ± 0.66	6.53 ± 0.14
2012–2013	TD	1.40 ± 0.21	88.2 ± 14.7	0.08 ± 0.02	79.7 ± 15.2	27.5 ± 4.1	2.24 ± 0.49	6.33 ± 0.50	44.3 ± 2.1	32.3 ± 3.7	1.25 ± 0.07	6.46 ± 0.41
	TND	3.75 ± 0.21	59.7 ± 18.0	0.13 ± 0.02	101.1 ± 14.8	17.7 ± 15.0	2.84 ± 0.42	6.48 ± 0.78	52.7 ± 11.1	15.3 ± 3.5	1.48 ± 0.31	6.30 ± 0.23
	NTD	0.73 ± 0.22	52.0 ± 9.1	0.04 ± 0.01	80.6 ± 9.6	36.4 ± 13.1	2.27 ± 0.27	6.05 ± 0.47	60.8 ± 11.8	38.1 ± 2.4	1.72 ± 0.34	6.27 ± 0.50
	NTND	2.11 ± 0.23	56.5 ± 13.0	0.08 ± 0.00	108.7 ± 5.8	24.1 ± 14.9	3.05 ± 0.15	6.38 ± 0.73	65.9 ± 12.9	32.3 ± 6.7	1.86 ± 0.36	6.08 ± 0.24
2013–2014	TD	2.94 ± 0.78	96.1 ± 22.9	0.12 ± 0.04	68.1 ± 7.0	76.0 ± 15.1	1.94 ± 0.29	7.07 ± 0.34	62.6 ± 4.7	49.5 ± 2.8	1.77 ± 0.14	6.64 ± 0.31
	TND	3.73 ± 0.85	44.7 ± 26.0	0.12 ± 0.08	76.2 ± 5.0	42.1 ± 8.0	2.15 ± 0.11	6.43 ± 0.60	72.1 ± 9.2	42.1 ± 12.9	2.04 ± 0.25	6.38 ± 0.47
	NTD	1.52 ± 0.48	52.0 ± 28.4	0.06 ± 0.02	88.4 ± 6.3	85.4 ± 10.9	2.51 ± 0.21	6.19 ± 0.23	70.6 ± 13.6	99.7 ± 7.5	2.02 ± 0.39	6.46 ± 0.61
	NTND	2.01 ± 0.09	42.9 ± 10.6	0.07 ± 0.04	119.7 ± 10.8	49.4 ± 13.6	3.37 ± 0.33	6.16 ± 0.36	82.2 ± 3.1	54.4 ± 9.5	2.32 ± 0.08	6.16 ± 0.12
Mean*	TD	2.47 ± 0.10 bc	68.2 ± 16.4 a	0.10 ± 0.02 b	68.3 ± 11.4 b	42.5 ± 11.2 a	1.93 ± 0.32 b	6.62 ± 0.25 a	80.1 ± 2.7 c	56.4 ± 17.4 ab	2.27 ± 0.08 c	6.71 ± 0.14 a
	TND	4.91 ± 0.43 a	42.5 ± 12.3 ab	0.16 ± 0.02 a	87.2 ± 13 ab	30.6 ± 15.0 a	2.45 ± 0.37 ab	6.56 ± 0.49 a	96.6 ± 8.3 b	40.0 ± 4.3 b	2.72 ± 0.23 b	6.68 ± 0.24 a
	NTD	1.73 ± 0.37 c	43.5 ± 18.4 ab	0.07 ± 0.00 c	76.2 ± 6.9 b	48.8 ± 18.1 a	2.15 ± 0.19 b	6.17 ± 0.27 a	89.9 ± 1.2 bc	65.9 ± 6.6 a	2.54 ± 0.03 bc	6.51 ± 0.39 a
	NTND	2.90 ± 0.21 b	36.4 ± 13.5 b	0.10 ± 0.02 b	105.1 ± 15.5 a	34.0 ± 6.9 a	2.96 ± 0.44 a	6.26 ± 0.33 a	113.5 ± 8.0 a	44.6 ± 8.0 b	3.20 ± 0.22 a	6.40 ± 0.20 a

738 Mean\* ± SD, different letters within the same column indicate statistical differences in variables mean among treatments over the 4 years by LSD's multiple range test ( $P < 0.05$ ).

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**Table 2** A two-way ANOVA for the effects of land management (L) and year (Y) on CH<sub>4</sub> and N<sub>2</sub>O emissions and grain yields in the rice field.

Season	Factors	df	CH <sub>4</sub> (kg CH <sub>4</sub> ha <sup>-1</sup> )			N <sub>2</sub> O (g N <sub>2</sub> O-N ha <sup>-1</sup> )			Yield (t ha <sup>-1</sup> )		
			ss	F	P	ss	F	P	ss	F	P
Early-rice	L	3	3052.7	5.196	0.005	820.1	1.007	0.403	0.603	2.361	0.090
	Y	3	692.3	1.178	0.333	4357.4	5.349	0.004	0.598	3.340	0.092
	L × Y	9	254.2	0.433	0.907	267.0	0.328	0.959	0.161	0.631	0.762
	Model	15	901.5	1.535	0.151	1195.7	1.468	0.176	0.337	1.319	0.248
	Error	32	587.5			814.7			0.256		
Late-rice	L	3	2379.4	4.700	0.008	1635.2	1.528	0.226	0.259	1.522	0.228
	Y	3	22545.7	44.534	0.000	3515.8	3.286	0.033	1.193	7.015	0.001
	L × Y	9	223.0	0.440	0.903	826.9	0.806	0.614	0.057	0.338	0.955
	Model	15	5118.8	10.111	0.000	1547.9	1.447	0.185	0.325	1.910	0.061
	Error	32	506.3			1070.0			0.170		
Winter	L	3	21.582	5.215	0.005	2367.6	4.537	0.009			
	Y	3	86.036	20.788	0.000	3265.9	6.259	0.002			
	L × Y	9	4.020	0.971	0.481	314.4	0.603	0.785			
	Model	15	23.935	5.783	0.000	1315.4	2.521	0.014			
	Error	32	4.139			521.8					

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**Table 3** Mean annual CH<sub>4</sub> and N<sub>2</sub>O emissions, global warming potentials (GWPs) of CH<sub>4</sub> and N<sub>2</sub>O emissions, rice grain yields, and greenhouse gas intensity (GHGI) over the 4 years from 2010 to 2014.

Treatment	CH <sub>4</sub> emission (kg CH <sub>4</sub> ha <sup>-1</sup> yr <sup>-1</sup> )	N <sub>2</sub> O emission (g N <sub>2</sub> O-N ha <sup>-1</sup> yr <sup>-1</sup> )	GWPs (t CO <sub>2</sub> -eq ha <sup>-1</sup> yr <sup>-1</sup> )	Rice yields (t ha <sup>-1</sup> yr <sup>-1</sup> )	GHGI (t CO <sub>2</sub> -eq t <sup>-1</sup> yield yr <sup>-1</sup> )
TD	151 ± 10 d	167 ± 28 a	4.29 ± 0.27 d	13.3 ± 0.3 a	0.32 ± 0.02 c
TND	189 ± 15 b	113 ± 13 a	5.33 ± 0.41 b	13.2 ± 0.6 a	0.40 ± 0.05 b
NTD	168 ± 6 cd	158 ± 27 a	4.76 ± 0.17 cd	12.7 ± 0.6 a	0.38 ± 0.02 b
NTND	222 ± 9 a	115 ± 38 a	6.25 ± 0.26 a	12.7 ± 0.1 a	0.49 ± 0.02 a

Note: different letters within the same column indicate statistical differences among treatments at  $P < 0.05$  level by LSD's test.

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**Table 4** Total precipitation, mean daily temperature, <sup>a</sup> mean soil Eh, CH<sub>4</sub>, and N<sub>2</sub>O fluxes over the 4 winter fallow seasons.

Winter fallow season	Precipitation (mm)	Temperature (°C)	Soil Eh (mV)	CH <sub>4</sub> flux (mg CH <sub>4</sub> m <sup>-2</sup> h <sup>-1</sup> )	N <sub>2</sub> O flux (µg N <sub>2</sub> O-N m <sup>-2</sup> h <sup>-1</sup> )
2010 (December 2, 2010 to April 15, 2011)	404	9.1	152 ± 11	0.02 ± 0.01	5.01 ± 0.26
2011 (November 3, 2011 to April 19, 2012)	754	10.0	102 ± 13	0.18 ± 0.08	3.11 ± 0.31
2012 (December 5, 2012 to April 15, 2013)	574	9.7	141 ± 34	0.07 ± 0.04	8.41 ± 0.54
2013 (November 11, 2013 to April 5, 2014)	661	9.4	92 ± 12	0.08 ± 0.03	7.06 ± 0.38

Note: <sup>a</sup> mean soil Eh, CH<sub>4</sub>, and N<sub>2</sub>O fluxes were the average of 4 treatments.

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**Table 5** Relative mitigating GWPs of GHGs emissions from paddy fields with various land management practices as compared to traditional managements in winter crop season.

Type	Traditional management	Suggested practice	GHGs	<sup>a</sup> Mitigation potential (%)				Reference
				WS	ES	LS	Annual	
Double rice	Winter fallow without drainage nor tillage	Drainage	CH <sub>4</sub> and N <sub>2</sub> O	30	27	21	24	This study
		Tillage	CH <sub>4</sub> and N <sub>2</sub> O	-60	17	15	15	
		Drainage combined with tillage	CH <sub>4</sub> and N <sub>2</sub> O	0	35	29	32	
Single rice	Winter wheat with drainage	Tillage	CH <sub>4</sub> and N <sub>2</sub> O	21	14		15	(Zhang et al., 2015)
Single rice	Winter ryegrass with drainage	Tillage	N <sub>2</sub> O	<sup>b</sup> N.m.	22		N.m.	(Bayer et al., 2015)
Single rice	Winter wheat with drainage	Tillage	CH <sub>4</sub> and N <sub>2</sub> O	38	N.m.		N.m.	(Yao et al., 2013)
Single rice	Winter fallow and continuous flooding	Oil-seed rape with drainage and tillage	CH <sub>4</sub> and N <sub>2</sub> O	4	57		43	(Zhang et al., 2012)
Single rice	Winter fallow without drainage nor tillage	Drainage	CH <sub>4</sub>	N.m.	71		>71	(Shiratori et al., 2007)
Single rice	Winter fallow with drainage but non-tillage	tillage	CH <sub>4</sub> , N <sub>2</sub> O, and CO <sub>2</sub>	-21	N.m.		N.m.	(Liang et al., 2007)
Single rice	Winter fallow and continuous flooding	Wheat drainage with	CH <sub>4</sub> and N <sub>2</sub> O	59	55		56	(Jiang et al., 2006)
		Oil-seed rape with drainage	CH <sub>4</sub> and N <sub>2</sub> O	53	57		56	
Single rice	Winter fallow and continuous flooding	Wheat drainage with	CH <sub>4</sub>	100	30		59	(Cai et al., 2003)
Single rice	Winter fallow and continuous flooding	Wheat drainage with	CH <sub>4</sub>	N.m.	68		>68	(Cai et al., 1998)

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Note: WS, ES, and LS means winter fallow season, early-rice season and late-rice season, respectively; annual is the total of winter and rice seasons; <sup>a</sup> Mitigation potential of combined gases was calculated on the basis of CO<sub>2</sub> equivalents by assuming GWPs for CH<sub>4</sub> and N<sub>2</sub>O as 28 and 265 times the equivalent mass of CO<sub>2</sub> over a 100-year period (Myhre et al., 2013): GWPs (CH<sub>4</sub> + N<sub>2</sub>O + CO<sub>2</sub>) = (CH<sub>4</sub> × 28) + (N<sub>2</sub>O × 265) + (CO<sub>2</sub> × 1); <sup>b</sup> N.m. indicates no measurements.

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**Table 6** Contents of total C ( $\text{g kg}^{-1}$ ) and total N ( $\text{g kg}^{-1}$ ) in rice stubble.

<u>Tillage time</u>	<u>Treatment</u>	<u>Total C</u>	<u>Total N</u>	<u>C/N</u>	<u>Tillage time</u>	<u>Treatment</u>	<u>Total C</u>	<u>Total N</u>	<u>C/N</u>
<u>After late-rice harvest in 2011</u>	<u>TD</u>	<u>338</u>	<u>6.9</u>	<u>49</u>	<u>After late-rice harvest in 2012</u>	<u>TD</u>	<u>368</u>	<u>8.7</u>	<u>42</u>
	<u>TND</u>	<u>314</u>	<u>7.8</u>	<u>40</u>		<u>TND</u>	<u>364</u>	<u>7.1</u>	<u>51</u>
<u>Before early-rice transplanting in 2012</u>	<u>NTD</u>	<u>356</u>	<u>12.7</u>	<u>28</u>	<u>Before early-rice transplanting in 2013</u>	<u>NTD</u>	<u>404</u>	<u>12.8</u>	<u>32</u>
	<u>NTND</u>	<u>374</u>	<u>10.4</u>	<u>36</u>		<u>NTND</u>	<u>397</u>	<u>13.4</u>	<u>30</u>

**Table 6** Measurements of total C ( $\text{g kg}^{-1}$ ) and total N ( $\text{g kg}^{-1}$ ) contents in rice stubble before early-rice transplanting in 2012 and 2013.

<u>Year</u>	<u>Treatment</u>	<u>Total C</u>	<u>Total N</u>	<u>C/N</u>
<u>2012</u>	<u>TD</u>	<u>338</u>	<u>6.9</u>	<u>49</u>
	<u>TND</u>	<u>314</u>	<u>7.8</u>	<u>40</u>
	<u>NTD</u>	<u>356</u>	<u>12.7</u>	<u>28</u>
	<u>NTND</u>	<u>374</u>	<u>10.4</u>	<u>36</u>
<u>2013</u>	<u>TD</u>	<u>368</u>	<u>8.7</u>	<u>42</u>
	<u>TND</u>	<u>364</u>	<u>7.1</u>	<u>51</u>
	<u>NTD</u>	<u>404</u>	<u>12.8</u>	<u>32</u>
	<u>NTND</u>	<u>397</u>	<u>13.4</u>	<u>30</u>

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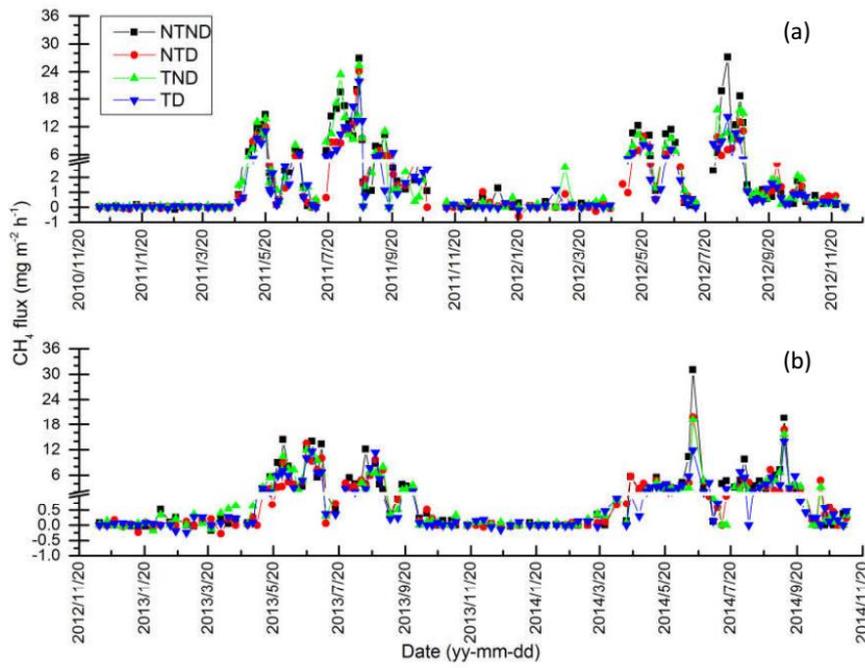


Figure 1 Seasonal variation of CH<sub>4</sub> emission from 2010 to 2014.

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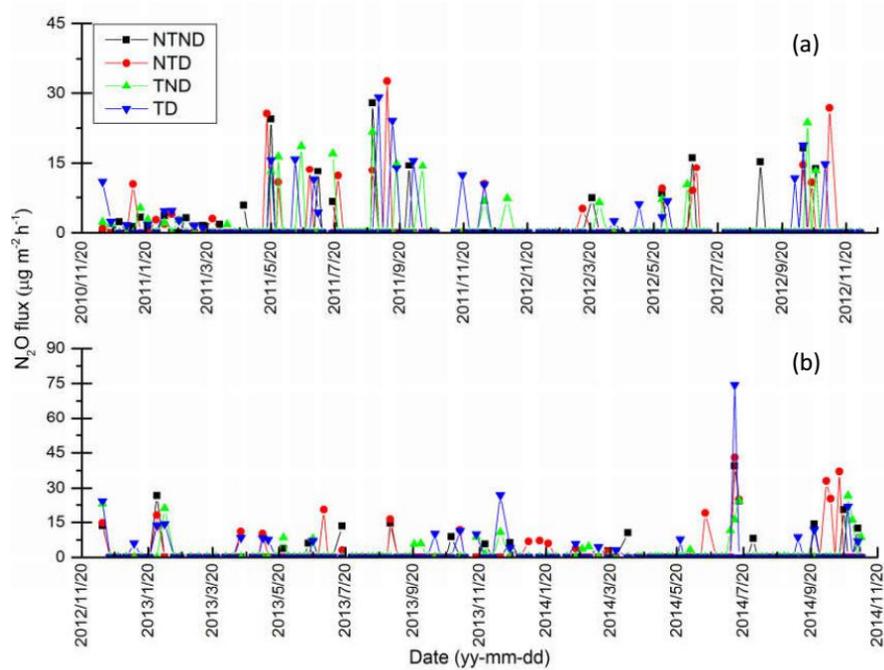
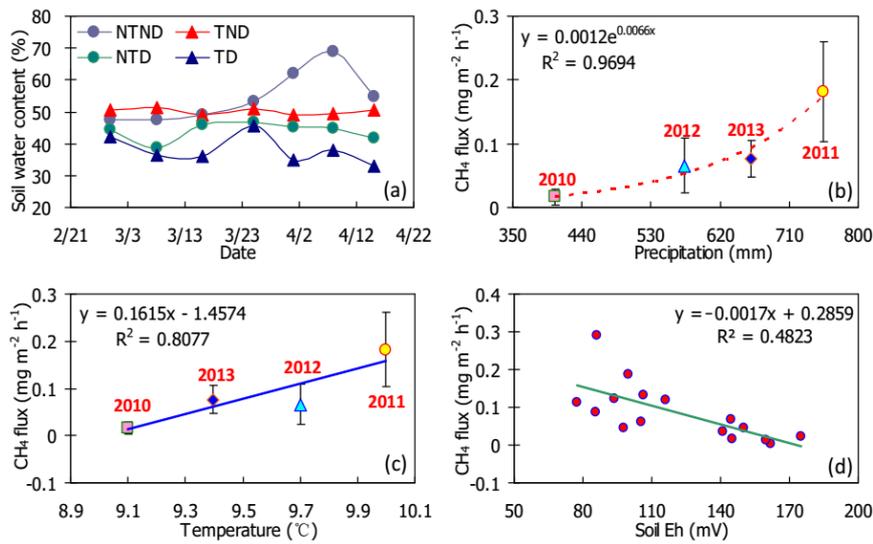


Figure 2 Seasonal variation of N<sub>2</sub>O emission from 2010 to 2014.

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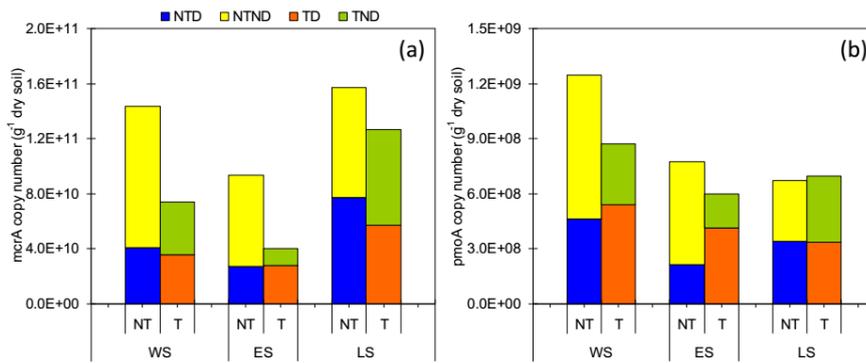
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**Figure 3** Soil water content in 2010 winter fallow season (a) and the relationships between mean CH<sub>4</sub> emission and total winter precipitation (b), and mean daily air temperature (c) and soil Eh (d) over the 4 winter fallow seasons (Data from Table 4).

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**Figure 4** The abundance of methanogens and methanotrophs populations in paddy soil from 2013 to 2014, WS, ES, and LS means winter fallow season, early-rice season, and late-rice season, respectively.

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