



- 1 New insights into the atmospheric mercury cycling in
- 2 Central Antarctica and implications at a continental scale
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19 Abstract

20 Under the framework of the GMOS project (Global Mercury Observation System) 21 atmospheric mercury monitoring has been implemented at Concordia Station on the high-22 altitude Antarctic plateau (75°06'S, 123°20'E, 3220 m above sea level). We report here the 23 first year-round measurements of gaseous elemental mercury (Hg(0)) in the atmosphere and 24 in snowpack interstitial air on the East Antarctic ice sheet. This unique dataset shows 25 evidence of a continuous oxidation of atmospheric Hg(0) in summer (24-hour daylight) due to 26 the high oxidative capacity of the Antarctic plateau atmosphere at this period of the year. 27 Summertime Hg(0) concentrations exhibited a pronounced daily cycle in ambient air with 28 maximal concentrations around midday. Photochemical reactions and chemical exchange at 29 the air/snow interface were prominent, highlighting the role of the snowpack on the





30 atmospheric mercury cycle. Our observations reveal a 20 to 30% decrease of atmospheric 31 Hg(0) concentrations from May to mid-August (winter, 24-h darkness). This phenomenon has 32 never been observed elsewhere and likely results from a gas-phase oxidation and/or 33 heterogeneous reactions. We also reveal the occurrence of multi-day to weeklong atmospheric 34 Hg(0) depletion events in summer, not associated with depletions of ozone, and likely due to a 35 stagnation of air masses above the plateau triggering an accumulation of oxidants within the 36 shallow boundary layer. Our observations suggest that the inland atmospheric reservoir is 37 depleted in Hg(0) in summer. Due to katabatic winds flowing out from the Antarctic plateau 38 down the steep vertical drops along the coast and according to observations at coastal Antarctic stations, the striking reactivity observed on the plateau most likely influences the 39 40 cycle of atmospheric mercury at a continental scale.

41

42 **1** Introduction

43 Mercury biomagnifies in its methylated form in aquatic food webs to elevated levels in 44 freshwater and marine fish, causing adverse health effects to wildlife and humans (Mason et 45 al., 2012). In 2013, the Minamata Convention (UNEP, 2013) was adopted and opened for 46 signature to reduce the exposure of populations to this worldwide contamination. Gaseous 47 elemental mercury (Hg(0)), the most abundant form of mercury in the atmosphere, is 48 efficiently transported around the globe, and even remote areas receive significant inputs of 49 anthropogenic mercury by long-range atmospheric transport, as recently reported in modeling 50 and observational studies (Weiss-Penzias et al., 2007; Lin et al., 2010).

51 Hg(0) can be oxidized into highly reactive and water-soluble gaseous and/or particulate 52 divalent species (Hg(II) and Hg(p), respectively) (Lin and Pehkonen, 1999) leading to the 53 formation and subsequent rapid deposition of reactive mercury onto environmental surfaces 54 (Hedgecock and Pirrone, 2004). Upon deposition mercury can be reemitted back to the 55 atmosphere or may enter the food chain through the conversion of Hg(II) to its methylated 56 form (Driscoll et al., 2013). Effects and toxicity of mercury depends on this complex cycle, 57 which is still not fully understood, and are only indirectly related to regional and global 58 emissions (Driscoll et al., 2013). A better understanding of atmospheric mercury chemistry 59 will lead to improved global transport and deposition models and could help refine pollution-60 control strategies around the world.





61 New oxidation pathways, discovered in 1995 in the Arctic (Schroeder et al., 1998) and 62 highlighting the influence of halogen radicals on Hg(0) oxidation in spring, changed our 63 understanding of the mercury cycle. While the Arctic has been extensively monitored, there is 64 still much to be learned from the Antarctic continent where studies are scarce (Dommergue et 65 al., 2010), especially on the high altitude plateau (see Fig. 1). The Antarctic plateau - ice-66 covered area of ~ 7 million km^2 – was first considered to be chemically-inactive and a giant cold trap (Eisele et al., 2008) for atmospheric species, including mercury. It turned out to be 67 68 highly photochemically active (Davis et al., 2001; Grannas et al., 2007) during the sunlit 69 period with oxidant concentrations approaching those of tropical or urban mid-latitude 70 environments (Eisele et al., 2008; Kukui et al., 2014). Earlier studies (Brooks et al., 2008; 71 Dommergue et al., 2012) – the only two mercury studies ever carried out on the high-altitude 72 Antarctic plateau with modern instruments – also suggested, based on short-term observations 73 (a few weeks) in summer, an intense reactivity of mercury on the plateau at the air/snow 74 interface. In this context, and under the framework of the GMOS project (Global Mercury 75 Observation System, www.gmos.eu), atmospheric mercury was continuously monitored at 76 Concordia Station (see Fig. 1) since 2012 and, for the first time, Hg(0) has been monitored 77 year-round in both the snow interstitial air and the overlying atmosphere in 2013. Given harsh 78 weather conditions (see section 2.1), and technical and logistical limitations, presenting such a 79 record is in itself an important achievement. The main objective of this study is to provide 80 new insights into the year-round cycling of gaseous mercury on the Antarctic plateau.

81

82 2 Experimental Section

83 2.1 Sampling site

84 Year-round measurements of gaseous mercury were conducted in 2012 and 2013 at the 85 French/Italian Concordia Station (75°06'S, 123°20'E, 3220 m above sea level), located on the 86 Antarctic plateau, 1100 km away from the nearest coast of East Antarctica (see Fig. 1). 87 Concordia Station is a regional topographic maximum on the plateau; the surface terrain 88 slopes do not exceed 1% (Genthon et al., 2010). The air temperature ranges between -20 °C in 89 summer and -70 °C in winter, with an annual mean value of -45 °C (Pietroni et al., 2012). 90 There is permanent daylight in summer and permanent darkness in winter. Snow 91 accumulation is ~ 10 cm/year (Genthon et al., 2013).





92 2.2 Sampling instrumentation

93 Instrumentation was located in a below-surface shelter at the edge of the "clean area", 800 m 94 away from the main camp and upwind with respect to the dominant wind direction (south 95 west). In 2012, year-round measurements were performed in the atmospheric boundary layer 96 at about 500 cm above the snow surface. In 2013, measurements were performed in both the 97 atmosphere and in snowpack interstitial air for several trace gases including gaseous mercury 98 and ozone (O₃). Sampling instrumentation included one 10 m meteorological tower for above-99 surface gradient sampling and two multi-inlet snow sampling manifolds ("snow towers") for 100 measuring trace gases at various depths in interstitial air (Bocquet et al., 2007; Seok et al., 101 2009). The 10 m meteorological tower was installed ~ 15 m upwind of the underground 102 instrument shelter. It accommodated three gas inlets at 1070 cm, 210 cm, and 25 cm above the 103 snow surface (see Fig. 2a). Trace gas measurements were acquired on each snow tower at six 104 height levels: 50 and 10 cm above the snow surface, and 10, 30, 50, and 70 cm below the 105 snow surface (see Fig. 2b). Sampling lines were purged continuously at 5 L/min on the 106 meteorological tower and intermittently at ~ 2-3 L/min on the snow towers. On each snow 107 tower, inlets were fitted with a small glass fibre filter in PTFE housing (25 mm Acrodisc 108 syringe filters, Pall Life Sciences, Ann Arbor, Michigan, USA) to prevent snow crystals from 109 entering the PFA sampling lines. Sampling lines were inside insulation tubing and the 110 temperature of the sampling lines was maintained at a level 5-10 °C warmer than the 111 snowpack temperature with a heat trace to prevent water vapor from freezing and clogging the 112 lines. An automatic sampling pattern was implemented: trace gases were collected 113 sequentially from the uppermost inlets on the meteorological tower to deepest levels of the 114 snow towers. Measurements were taken for 10 min from each inlet.

115 **2.3 Gaseous mercury measurements**

116 Measurements were performed using a Tekran 2537A analyzer (Tekran Inc., Toronto, 117 Canada) based on the amalgamation of mercury onto a gold cartridge followed by thermal 118 desorption and detection by an integrated cold vapor atomic fluorescence spectrometer 119 (CVAFS) at 253.7 nm (Fitzgerald and Gill, 1979; Bloom and Fitzgerald, 1988). The presence 120 of two gold cartridges allowed alternating sampling and desorption modes and thus a 121 continuous analysis in the pre-filtered (0.45 µm PTFE filter) and unheated sample air stream. 122 The sampling resolution was 5 min with a sampling flow rate of 0.8 L/min. Concentrations 123 are expressed in nanograms per cubic meter at standard temperature and pressure (273.15 K,





124 1013.25 hPa). Using both a 0.45 µm PTFE filter at the entrance inlet of the sample line, and

125 an unheated ¹/₄"PTFE sample line, we assume that only Hg(0) (vs. total gaseous mercury,

126 defined as the sum of gaseous mercury species) was efficiently collected and subsequently

127 analyzed in this study (Steffen et al., 2002; Temme et al., 2003; Steffen et al., 2008).

128 Quality assurance and quality control procedures

129 An automatic calibration step of the Tekran 2537A analyzer was carried out every 25 h with 130 an internal mercury permeation source. External calibrations were performed twice a year by 131 injecting manually saturated mercury vapor taken from a temperature-controlled vessel 132 (Tekran 2505 mercury vapor calibration unit, Hamilton digital syringe). As described by 133 Angot et al. (2014), fortnightly to monthly routine maintenance operations were performed. A 134 software programme was developed at the LGGE (Laboratoire de Glaciologie et Géophysique 135 de l'Environnement) in accordance with quality control practice in well-established North 136 American networks (Steffen et al., 2012). Based on various flagging criteria (Munthe et al., 137 2011; D'Amore et al., 2015), it enabled rapid data processing in order to produce clean time series of Hg(0). The detection limit is estimated at 0.10 ng/m^3 (Tekran, 2011). 138

139 **2.4 Surface snow sampling and analysis**

140 Surface snow samples (first cm) were collected weekly from February 2013 to January 2014 141 using acid cleaned PTFE bottles and clean sampling procedures. Upon collection, samples 142 were stored in the dark at -20 °C. Field blanks, carried out by opening and closing a bottle 143 containing mercury-free water, were regularly conducted. Surface snow samples and field 144 blanks were analyzed for total mercury using a Tekran Model 2600. Quality assurance and 145 quality control included the analysis of analytical blanks, replicates, internal standards, and spiked materials. The limit of quantification - calculated as 10 times the standard deviation of 146 147 a set of 3 analytical blanks – amounted to 0.3 ng/L with a relative accuracy of \pm 8%.

148 2.5 Ancillary parameters

149 Ozone

Measurements were performed using a UV absorption monitor (Thermo Electron
Corporation, Franklin, MA), model 49I in 2012 (Legrand et al., Submitted to ACP) and model
49C in 2013. The instruments were calibrated against the National Oceanic and Atmospheric
Administration Global Monitoring Division, Boulder, Colorado, standard.





154 Air mass back trajectories

- 155 Air mass back trajectories were computed using the Lagrangian model FLEXPART (Stohl et 156 al., 1998; Stohl and Thomson, 1999; Stohl et al., 2005) run in the backward mode and driven 157 by NCEP (National Center for Environmental Predictions) GFS (Global Forecast System) 158 final meteorological fields. Simulations were done every day at 1200 UTC in 2012 and 2013. 159 For each simulation, 20000 pseudo-particles were released by the model around the position of Concordia Station and tracked for 5 days in three layers of altitude (0-0.1, 0.1-4 and 4-10 160 km above ground level). Simulations at an altitude of 4-10 km were computed in order to 161 162 investigate the potential occurrence of upper troposphere/lower stratosphere intrusions. For 163 each 1-h time step, the model produced a normalized particle residence time (in seconds) 164 within an output grid of $0.5 \times 0.5^{\circ}$. The sum of the 5 days outputs provided potential emission sensitivities (PES, in seconds) for the three layers of altitude. PES in a particular grid cell is 165 166 proportional to the particle residence time in that cell. It should be noted that, in Antarctica, 167 the meteorological data driving the FLEXPART transport model rely on sparse 168 measurements. Consequently, the trajectories calculated in this region are often associated 169 with relatively high uncertainties.
- 170 Height of the boundary layer and shortwave radiation

The height of the boundary layer and downwelling shortwave radiation were calculated by the
MAR regional atmospheric model (Modèle Atmosphérique Régional). MAR was developed
at the LGGE for Polar Regions and the simulations have been evaluated against
meteorological observations made at Concordia Station (Gallée and Gorodetskaya, 2010;
Gallée et al., 2015).

176 Meteorological data

177 Temperature, wind speed and direction were recorded at six height levels on a 45 m tower.

- 178 The general observation set up is described by Genthon et al. (2010).
- 179 Ice precipitation
- 180 A tropospheric depolarization LIDAR (Light Detection And Ranging) operating at 523 nm
- 181 provided tropospheric profiles of aerosol and clouds every 5 min allowing detection of
- 182 water/ice clouds, snow drift, diamond dust and pollution plumes.
- 183 Tropospheric temperature and integrated water vapor





- A H₂O Antarctica Microwave Stratospheric and Tropospheric Radiometers (HAMSTRAD)
 instrument was used for the detection of the 60-GHz oxygen and the 183-GHz water vapor
 lines allowing measurement of tropospheric temperature and water vapor profiles,
 respectively, together with integrated water vapor (IWV) every 7 min. The instrument is fully
 automated and a liquid nitrogen calibration is performed twice per year (Ricaud et al., 2015).
- 189 Eddy diffusivity and friction velocity
- 190 The Eddy diffusivity was calculated as follows (Xiao et al., 2014):

$$191 \quad K = k \, u_* \, z / \varphi_h \tag{1}$$

where k (set to 0.40) is the von Karman constant, u_* the friction velocity (m/s), z the measurement height (m), and φ_h the Obukhov stability function. According to Frey et al. (2013), the stability function was $\varphi_h = 0.95 + 4.62 \frac{z}{L}$ for stable conditions above snow (King and Anderson, 1994), and $\varphi_h = 0.95 \left(1 - 11.6 \frac{z}{L}\right)^{-0.5}$ for unstable conditions (Hoegstroem, 1988). u_* and L (the Obukhov length (m)) were computed from the three-dimensional wind components (u, v, w) and temperature measured by a Metek sonic anemometer mounted at 8 m above the snow surface.

199

200 3 Results and Discussion

201 3.1 Seasonal variation of Hg(0) in ambient air

The seasonal boundaries were defined according to the transitions in downwelling shortwave
radiation (see Fig. 3b) as follows: winter from May to mid-August, spring from mid-August
to October, summer from November to mid-February, and fall from mid-February to April.

The record of atmospheric Hg(0) over the entire 2012-2013 period is displayed in Fig. 3a. 205 Hg(0) concentrations ranged from below 0.10 to 2.30 ng/m^3 , with average values amounting 206 to 0.76 ± 0.24 ng/m³ in 2012, and to 0.81 ± 0.28 ng/m³, 0.84 ± 0.27 ng/m³, and 0.82 ± 0.26 207 ng/m³ in 2013 at 25, 210, and 1070 cm above the snow surface, respectively. No significant 208 difference was observed between annual averages of Hg(0) concentrations measured at the 209 three inlets of the meteorological tower in 2013 (p value = $3.1.10^{-14}$, Mann-Whitney test). 210 These mean annual Hg(0) concentrations are lower than annual averages reported at coastal 211 Antarctic stations (i.e., $0.93 \pm 0.19 \text{ ng/m}^3$ for Hg(0) at Troll (Pfaffhuber et al., 2012) and 1.06 212 213 ± 0.24 ng/m³ for total gaseous mercury at Neumayer (Ebinghaus et al., 2002)).





Unlike in winter, Hg(0) concentrations were highly variable during the sunlit period with concentrations ranging from below 0.10 ng/m³ to 1.50-2.00 ng/m³, up to twice the average background levels recorded in the Southern Hemisphere mid-latitudes (Slemr et al., 2015). These seasonal features, in good agreement with observations at other Antarctic stations (Ebinghaus et al., 2002; Pfaffhuber et al., 2012), suggest the existence of a photochemicallyinduced reactivity of atmospheric mercury during the sunlit period.

3.2 Summertime continuous oxidation of Hg(0) in ambient air and Hg(II) deposition onto snowpack

In summer, the mean atmospheric Hg(0) concentration was 0.69 ± 0.35 ng/m³. This means 222 223 that Hg(0) concentrations are ~ 25% lower than levels recorded at the same period of the year 224 at coastal Antarctic stations (Pfaffhuber et al., 2012; Ebinghaus et al., 2002; Sprovieri et al., 225 2002). Total mercury concentrations in surface snow samples were highly variable (median 226 value: 4.8 ng/L, range: < detection limit - 73.8 ng/L, see Fig. 4) and were higher in summer 227 (median value: 10.4 ng/L, range: 1.3 - 73.8 ng/L), suggesting that divalent mercury species 228 were preferentially deposited onto the snowpack at this period of the year. The lower Hg(0)229 concentrations in ambient air along with high total mercury concentrations in surface snow 230 samples suggest a continuous oxidation of Hg(0) in ambient air in summer, followed by the 231 deposition of oxidation products on surface snow. This hypothesis is further supported by 232 elevated oxidized mercury concentrations measured on the Antarctic plateau at South Pole in summer $(0.10 - 1.00 \text{ ng/m}^3)$ by Brooks et al. (2008). 233

234 The oxidative capacity of the Antarctic plateau atmosphere is elevated in summer, as 235 evidenced by several studies (Davis et al., 2001; Grannas et al., 2007; Eisele et al., 2008; 236 Kukui et al., 2014), likely explaining this continuous oxidation of Hg(0) in ambient air. 237 Among these oxidants, NO₂, RO₂, and OH are particularly abundant at Concordia Station in 238 summer (Frey et al., 2013; Kukui et al., 2014) and a recent study provided as a first estimate 239 a BrO mixing ratio of 2-3 pptv near the ground during sunlight hours (Frey et al., 2015). 240 Given the current understanding of mercury oxidation and the lack of continuous halogens 241 measurements, we were not able to identify the exact mechanism for the reactivity observed at 242 Concordia Station. A two-step oxidation mechanism, favored at cold temperatures, is worth 243 being considered further. The initial recombination of Hg(0) and Br is followed by the 244 addition of a second radical (e.g., I, Cl, BrO, ClO, OH, NO₂, or HO₂) in competition with 245 thermal dissociation of the HgBr intermediate (Goodsite et al., 2004; Wang et al., 2014).





According to Dibble et al. (2012), HO₂, NO, NO₂, and NO₃ bound Hg(0) too weakly to initiate its oxidation in the gas phase and reactions of the HgBr intermediate with NO₂, HO₂, ClO, and BrO are more important than with Br and OH. Further modeling or laboratory chamber studies investigating the fate of Hg(0) in the presence of various potential oxidants are needed to improve our understanding of the mechanisms.

251 **3.3** Hg(0)/Hg(II) redox conversions within the snowpack

Hg(0) moves easily between the atmosphere and the cryosphere (Durnford and Dastoor, 253 2011). A fraction of deposited mercury species can be reduced (the reducible pool) and subsequently reemitted to the atmosphere as Hg(0). Reduced mercury can concurrently be reoxidized within the snowpack. In Fig. 5, the redox processes occurring within the snowpack are shown along with several other processes that govern mercury exchange at the air/snow interface. These processes are discussed in details in the following sections.

258 3.3.1 Sunlit period

Fig. 6 depicts the mean Hg(0) concentration at various heights above and below the snow surface (in the interstitial air of the snow) for all seasons. During the sunlit period (summer, spring/fall), Hg(0) concentration peaked in the upper layers of the snowpack and then decreased with depth, with levels in the snow interstitial air (SIA) dropping below atmospheric values.

264 Hg(0) is generally produced in the upper layers of the snowpack – as the result of a photolytically initiated reduction of Hg(II) (Lalonde et al., 2003) - and diffuses upward and 265 266 downward. According to our observations, Hg(0) concentration peaked at ~ 10 cm below the 267 snow surface (see Fig. 6). Similarly, Brooks et al. (2008) reported Hg(0) concentrations peaking at a depth of 3 cm at South Pole. Below the top layer, the actinic flux decreases 268 exponentially with depth (King and Simpson, 2001; Domine et al., 2008). The light 269 penetration depth (e-folding depth) is the depth at which the actinic flux's magnitude is 1/e of 270 271 its incident value (Perovich, 2007). It is estimated that $\sim 85\%$ of the photoreduction occurs in 272 the top two e-folding depths (King and Simpson, 2001). At Concordia Station, the e-folding 273 depth is ~ 10 cm at 400 nm for the windpack layers (France et al., 2011), which supports our 274 observations. As previously mentioned, reduced mercury can concurrently be reoxidized 275 within the snowpack. Below the top layer, Hg(0) concentration in the SIA dropped with depth (see Fig. 6) suggesting that oxidation dominated in the deepest layers - in good agreement 276





with observations within the snowpack at Kuujjuarapik/Whapmagoostui, Québec, Canada
(Dommergue et al., 2003) – leading to the formation of a Hg(II) reservoir.

279 The amount of Hg(0) emitted from the snowpack to the atmosphere depends on the balance of 280 reduction and oxidation processes within the upper layers of the snowpack (Durnford and 281 Dastoor, 2011). Fig. 7 depicts the hourly mean atmospheric and interstitial air Hg(0) 282 concentrations. Our observations indicate that summertime Hg(0) concentration in the upper layers of the snowpack exhibited a diurnal cycle and peaked in the afternoon (see Fig. 7a, 283 lower panel). Conversely, in spring/fall, Hg(0) concentration reached a maximum at night and 284 285 a minimum near midday in the upper layers of the snowpack (see Fig. 7b). The balance of 286 reduction and oxidation processes within the upper layers of the snowpack suddenly shifted in 287 summer. Similarly, Faïn et al. (2008) found that reduction dominated during summer and oxidation in spring in the upper layers of the snowpack at Summit, Greenland. 288

289 It is worth noting that Hg(0) concentration in the SIA was occasionally very high. For 290 instance, on 24 February 2013, Hg(0) concentration reached 3.00 ng/m³ at a depth of 10 cm. 291 During this event, ice precipitation was observed at Concordia Station with depolarization 292 values greater than 30% (see Fig. 8). This suggests that the presence of ice crystals could enhance the dry deposition of Hg(II) species onto the snow surface leading to increased Hg(0) 293 294 formation in the upper layers of the snowpack. Indeed, due to an elevated specific surface 295 area, mercury-capture efficiency of ice crystals is high (Douglas et al., 2008). Unfortunately, due to a low sampling frequency of surface snow samples (weekly), total mercury 296 297 concentrations cannot be used to study further the relationship between the occurrence of ice 298 precipitation events and dry deposition of mercury species.

299 3.3.2 Winter

Contrarily to the sunlit period, Hg(0) concentration increased with depth in the SIA in winter 300 (see Fig. 6). The average (winter) Hg(0) concentration amounted to 3.60 ng/m³ at a depth of 301 302 70 cm and was observed at a temperature of about -60 °C and not related to any change in 303 atmospheric composition. Our observations are in agreement with earlier studies indicating 304 that reduction of Hg(II) species is possible in the dark (Ferrari et al., 2004; Faïn et al., 2007; Ferrari et al., 2008). The production of Hg(0) might be due to the reduction of Hg(II) species 305 accumulated in the deepest layers of the snowpack during the sunlit period (see section 3.3.1). 306 307 This shift from oxidation to reduction in the deepest layers of the snowpack at the beginning 308 of winter remains unexplained.





309 **3.4** Boundary layer dynamics and its influence on Hg(0) in ambient air

Several studies highlighted that the atmospheric turbulence at Concordia Station in summer
influences the vertical flux and concentration profiles of various atmospheric species
(Legrand et al., 2009; Dommergue et al., 2012; Kerbrat et al., 2012; Frey et al., 2013).

313 3.4.1 Summertime Hg(0) diurnal cycle

314 Based on a week of measurements made at Concordia Station in January 2009, Dommergue et al. (2012) reported that atmospheric Hg(0) exhibited a significant and daily cycle with 315 316 maximal concentrations around noon. We show that this daily cycle occurred all along the summer, with low atmospheric Hg(0) concentrations ($\sim 0.50 \text{ ng/m}^3$) when solar radiation was 317 318 minimum and a maximum (~ 0.80 ng/m^3) around noon (see Figs. 7a, upper panel and 9g). 319 Such a pronounced daily cycle has never been observed at other Antarctic stations 320 (Dommergue et al., 2010; Pfaffhuber et al., 2012). Several studies showed that Hg(0) 321 emission from the snowpack maximizes near midday (e.g., Steffen et al., 2002; Ferrari et al., 322 2005; Brooks et al., 2006; Faïn et al., 2007; Ferrari et al., 2008; Johnson et al., 2008). As 323 suggested by Durnford and Dastoor (2011), the noon emission does not necessarily reflect 324 maximum concentrations of cryospheric Hg(0) around midday (Hg(0) concentration peaked 325 in the afternoon at 10 cm below the snow surface, see section 3.3.1) and could be driven by 326 ventilation generated by atmospheric thermal convection. Stable boundary layers are almost 327 ubiquitous in Polar Regions due to radiation cooling (Anderson and Neff, 2008). However, 328 convective boundary layers have been observed in summer at polar domes at Concordia 329 Station (King et al., 2006) and Summit in Greenland (Cohen et al., 2007). Fig. 9 displays the hourly mean variation of several parameters. As illustrated by Figs. 9a and 9c, and in 330 331 agreement with earlier observations (e.g., Argentini et al., 2005; Pietroni et al., 2012; 332 Argentini et al., 2013), there was a strong diurnal cycle in near-surface temperature and wind 333 speed in summer at Concordia Station. These observations are typical for locations where a 334 convective boundary layer develops as a response to daytime heating (King et al., 2006), as 335 can be seen in Fig. 9d. In a convective boundary layer, vertical mixing is enhanced during 336 convective hours (Anderson and Neff, 2008), as shown in Figs. 9e and 9f by increasing values 337 for the Eddy diffusivity (K) and the friction velocity $(u_*, indicative of the strength of the$ 338 mixing processes in the surface layer (Neff et al., 2008)).

In summary, the observed summertime Hg(0) diurnal cycle in ambient air might be due to a combination of factors: i) a continuous oxidation of Hg(0) in ambient air due to the high





341 oxidative capacity on the plateau, ii) important Hg(II) deposition onto snowpack, and iii)
342 important emission of Hg(0) from the snowpack during convective hours following
343 photoreduction of Hg(II) in the upper layers of the snowpack.

344 3.4.2 Elevated Hg(0) concentrations in fall

345 In fall, Hg(0) concentrations in ambient air no longer peaked around midday (see Fig. 9g) and 346 were in average 67% higher than during the summer, exceeding levels recorded at lower 347 latitudes in the Southern Hemisphere (Slemr et al., 2015). At this period of the year, the 348 boundary layer lowered to ~ 50 m in average and no longer exhibited a pronounced diurnal 349 cycle (see Figs. 3c and 9d). We believe that the shallow boundary layer could cause Hg(0)350 concentrations in ambient air to build up to where they exceeded levels recorded at lower 351 latitudes in the Southern Hemisphere because Hg(0) was dispersed into a reduced volume of 352 air, limiting the dilution. Similarly, NOx mixing ratios are enhanced when the boundary layer 353 is shallow (Neff et al., 2008; Frey et al., 2013). Elevated Hg(0) concentrations were also 354 likely favored by the fact that oxidation in ambient air was weaker under lower solar 355 radiation.

356 **3.5 Multi-day depletion events of atmospheric Hg(0)**

357 First discovered in the Arctic (Schroeder et al., 1998), atmospheric Hg(0) depletion events 358 result from an oxidation by reactive bromine species released during springtime explosions in 359 coastal regions (Durnford and Dastoor, 2011 and references therein) and are concurrent with 360 tropospheric O_3 depletion events (Simpson et al., 2007). Despite the distance of Concordia Station from the coast (1100 km), a Hg(0) depletion event was observed on 11 September 361 362 2013 due to a maritime air transport event (see Fig. 10e). During this event, Hg(0)363 concentrations exhibited a strong positive correlation with O_3 mixing ratios (rho = 0.94, p value = 5.10^{-7}). 364

From 19 January to 8 February 2012, and from 5 to 20 February 2013, we, however, observed different depletion events. While atmospheric Hg(0) concentrations dropped and remained low $(0.39 \pm 0.19 \text{ ng/m}^3 \text{ from 19 January to 8 February 2012, } 0.41 \pm 0.21 \text{ ng/m}^3 \text{ from 5 to 20}$ February 2013) for several weeks (see Figs. 3a, 11a, and 11e), O₃ showed no abnormal variability (see Figs. 11d and 11h). These depletion events occurred as air masses stagnated over the Antarctic plateau (see Figs. 10a and 10b) according to our FLEXPART simulations. This stagnation of air masses is confirmed in 2013 (see Figs. 11f and 11g) by a decrease of





temperature at 10 m a.g.l (from -29 ± 3 °C in January to -43 ± 4 °C during the Hg(0) depletion event) and a low integrated water vapor (0.40 ± 0.13 kg/m² during the Hg(0) depletion event vs. 0.77 \pm 0.20 kg/m² in January). In both 2012 and 2013, depletions of Hg(0) ended when air masses started moving out of the plateau (see Figs. 10c and 10d).

376 While previous studies attributed high Hg(II) concentrations in the Antarctic summer to 377 subsiding upper tropospheric air (Holmes et al., 2006; Brooks et al., 2008), potential emission 378 sensitivities suggest that these depletions of Hg(0) were unlikely concomitant with upper 379 troposphere/lower stratosphere intrusions (see Figs. 10a and 10b, PES at 4-10 km). This is 380 also confirmed by stable O_3 mixing ratios. High altitude vertical profiles of Hg(0) should be 381 carried out to rule out this hypothesis of subsiding upper tropospheric air. We suggest that 382 these Hg(0) depletion events observed at Concordia Station result from processes occurring 383 within the shallow boundary layer. Since O_3 was not depleted during these events, Hg(0)384 depletion cannot be accounted for by bromine oxidation alone. FLEXPART simulations along 385 with integrated water vapor and temperature measurements indicate that these Hg(0) depletion 386 events occurred as air masses stagnated over the Antarctic plateau. As highlighted in section 387 3.2, the oxidative capacity is high in summer on the plateau (Davis et al., 2001; Grannas et 388 al., 2007; Eisele et al., 2008; Kukui et al., 2014). This air mass stagnation might favor an 389 accumulation of oxidants within the shallow boundary layer (< 300 m in average), leading to 390 an oxidation of Hg(0) stronger than usual.

391 3.6 Decreasing trend in winter

While stable concentrations were expected in winter given the absence of photochemistry, our observations reveal a 20 to 30% decrease of atmospheric Hg(0) concentrations from May to mid-August (see Fig. 3a). Conversely, Hg(0) concentrations remained stable at Troll and Neumayer from late fall through winter (Ebinghaus et al., 2002; Pfaffhuber et al., 2012). This decreasing trend observed in winter might be due to several mechanisms, including gas-phase oxidation and heterogeneous reactions.

398 **3.6.1** Hypothesis on a gas-phase reaction

399 Several studies suggested the involvement of nitrate radicals in the night-time oxidation of 400 Hg(0) (Mao and Talbot, 2012; Peleg et al., 2015). However, as previously mentioned, Dibble 401 et al. (2012) indicated that NO₃ bound Hg(0) too weakly to initiate its oxidation in the gas 402 phase. Another potential oxidant is O_3 , with this reactant reaching a maximum in the winter





403 (Legrand et al., 2009). However, according to some theoretical studies (e.g., Hynes et al., 404 2009), reaction (R1) is unlikely to proceed as a homogeneous reaction. Several experimental 405 studies confirmed the major product of reaction (R1) to be solid mercuric oxide, HgO (s) and 406 not HgO (g) (e.g., Pal and Ariya, 2004; Ariya et al., 2009), suggesting that pure gas phase 407 oxidation of elemental mercury by O₃ may not occur in the atmosphere. However, Calvert and 408 Lindberg (2005) proposed an alternative mechanism that would make this reaction potentially 409 viable in the atmosphere (Subir et al., 2011). The reaction may start with the formation of a 410 metastable HgO_3 (g) molecule which then decomposes to OHgOO (g) and thereafter 411 transforms to HgO (s) and O_2 (g).

412 $Hg(0)(g) + O_3(g) \rightarrow HgO(g) + O_2(g)(R1)$

413 **3.6.2** Influence of heterogeneous surfaces

414 As suggested by Subir et al. (2011), the influence of heterogeneous surfaces of water droplets, 415 snow, ice and aerosols should be taken into account when attempting to describe mercury 416 chemistry in the atmosphere. O'Concubhair et al. (2012) showed that freezing an acidic 417 solution containing nitrite or hydrogen peroxide can oxidize dissolved gaseous mercury in the 418 dark. Nitrous acid and hydrogen peroxide are present on the Antarctic plateau (Huey et al., 419 2004; Hutterli et al., 2004). As suggested by Dommergue et al. (2012), similar processes 420 could occur in the snow or on surface hoar at Concordia Station in winter. In 2013, the height of measurement had a significant influence on the decline over time of Hg(0) concentrations 421 (ANCOVA test, p value $< 2.10^{-16}$), with a steeper decrease at 25 cm than at 1070 cm. This 422 result suggests that snowpack may act as a sink for mercury, enhancing the deposition rate 423 424 due to heterogeneous reactions, through absorption of oxidation products, and/or physical 425 sorption/condensation of Hg(0) on surface snow.

In spite of the observed decreasing trend of Hg(0) concentrations in ambient air, total mercury concentrations in surface snow samples did not significantly increase over time in winter (see Fig. 4). Using a snow density of 300 kg/m³ a loss of 0.30 ng/m³ over a period of three months in a mixing layer of 50 m high would lead to a 5.0 ng/L increase in the first cm of the snowpack. Given the variability of chemical species deposition onto the snow surface, and the occurrence of either fresh snowfall or blowing snow, this 5.0 ng/L increase over a period of three months could not be detected in our weekly surface snow samples.

433 Despite the overall decreasing trend in winter, Hg(0) concentration exhibited abrupt increases434 when moist and warm air masses from lower latitudes occasionally reached Concordia





435 Station. This is, for example, evidenced on 13 June 2012 by an increase of 0.25 ng/m^3 of the 436 Hg(0) concentration, an increase of temperature at 10 m a.g.l. from -63 to -26 °C, and a high

437 integrated water vapor column (see Fig. 12).

438 **3.7** Implications at a continental scale

439 Depletion events of atmospheric Hg(0) that have been observed in the Artic and at various 440 coastal Antarctic stations have been associated with O_3 depletions, where Hg(0) and O_3 concentrations are positively correlated (Simpson et al., 2007). Increases in both Hg(II) and 441 442 Hg(p) have been reported in conjunction with decreases of Hg(0) (Lu et al., 2001; Lindberg et 443 al., 2002; Aspmo et al., 2005). Conversely, low Hg(0) concentrations that were not correlated 444 or anti-correlated with O₃ were observed at Neumayer and Troll (Temme et al., 2003; Pfaffhuber et al., 2012), while elevated Hg(II) concentrations (up to 0.33 ng/m³) were 445 446 recorded at Terra Nova Bay in the absence of $Hg(0)/O_3$ depletion (Sprovieri et al., 2002). The 447 continuous oxidation of Hg(0) in summer (see section 3.2) and multi-day Hg(0) depletion 448 events observed at Concordia Station in January/February (see section 3.5) are expected to 449 result in the build-up of an inland atmospheric reservoir enriched in Hg(II) and depleted in 450 Hg(0) in the summer. Due to strong katabatic winds flowing out from the Antarctic Plateau – 451 generated by the negative buoyant force that develops in the stable cooled layer along the ice 452 sheet slopes (Gallée and Pettré, 1998) -, a fraction of this inland atmospheric reservoir can be 453 transported toward the coastal margin. The influence of the flows from the Antarctic plateau 454 on coastal locations varies depending on the location. As demonstrated by Parish and 455 Bromwich (1987) and Parish and Bromwich (2007), the volume of air moving off inland 456 Antarctica toward the coastal margin displays significant spatial variability due to the topographic slope and orientation of the underlying ice sheets. Northward transport of air 457 from the plateau is enhanced in a few locations called confluence zones - e.g., the broad 458 459 region upslope from the Ross Ice Shelf at 175°E and the area near Adélie Land at 142°E (Parish and Bromwich, 1987, 2007) - but can be sporadically observed elsewhere explaining 460 461 observations at Neumayer, Troll, or Terra Nova Bay (Temme et al., 2003; Sprovieri et al., 462 2002; Pfaffhuber et al., 2012). Monitoring atmospheric mercury at a coastal station situated 463 close to a confluence zone could provide new insights regarding the extent of the transport of 464 reactive air masses from the Antarctic plateau. This topic will be addressed in a companion papier (Angot et al., this issue). The Antarctic continent shelters unconventional atmospheric 465 pathways of mercury reactivity both in winter and in summer. Its role should be taken into 466 467 account in the modeling of the global geochemical cycle of mercury.





468

469 **4 Conclusion**

Mean summertime atmospheric Hg(0) concentration was ~ 25% lower compared to values 470 471 recorded at other Antarctic stations at the same period of the year, suggesting a continuous 472 oxidation of atmospheric Hg(0) within the shallow boundary layer as a result of the high 473 oxidative capacity of the Antarctic plateau atmosphere at this period of the year. This 474 hypothesis is further supported by high total mercury concentrations in surface snow samples 475 measured at the station (up to 74 ng/L). Our results confirm short-term observations by 476 Brooks et al. (2008) and Dommergue et al. (2012) of chemical exchange at the air/snow 477 interface. During the sunlit period, Hg(0) concentration peaked in the upper layers of the 478 snowpack. Summertime Hg(0) concentration in ambient air exhibited a pronounced diurnal 479 cycle likely due to large emissions from the snowpack as a response to daytime heating. Our 480 observations also reveal a decrease of atmospheric mercury concentrations in winter (24-h 481 darkness) likely due to a gas-phase oxidation and/or heterogeneous reactions. Interestingly, 482 this decreasing trend has never been observed elsewhere. Finally, we reveal the occurrence of 483 multi-day to weeklong depletion events of Hg(0) in ambient air in summer, that are not 484 associated with depletion of O_3 , and likely result from a stagnation of air masses on the 485 plateau triggering an accumulation of oxidants in the shallow boundary layer. This behaviour 486 is radically different from what is usually observed in the Arctic where only mercury depletion events that were associated with O_3 depletion (and with a Hg(0)/O₃ correlation) 487 488 have been highlighted so far. According to observations at coastal Antarctic stations, the 489 reactivity observed at Concordia Station can be transported at a continental scale by strong 490 katabatic winds. Our understanding of the atmospheric mercury chemistry on the Antarctic 491 plateau is currently limited by the lack of continuous halogens measurements. Our findings 492 point out new directions for future kinetic, observational, and modeling studies.

493

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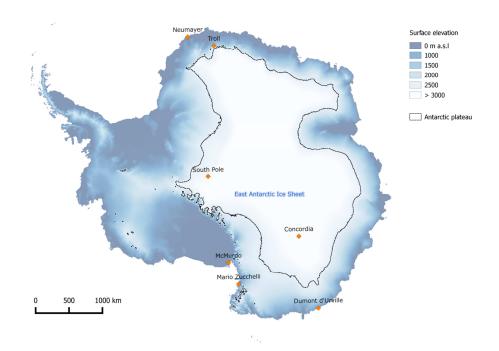


Figure 1: Map of Antarctica showing surface elevation (meters above sea level, m a.s.l) and the position of stations where atmospheric mercury measurements have been performed with modern online instruments. The black line shows the periphery of the high altitude plateau (> 2500 m a.s.l).







b)



Figure 2: Photographs showing **a**) the meteorological tower with the three gas inlets (red arrows) at 1070 cm, 210 cm and 25 cm above the snow surface (photo credit: B. Jourdain), and **b**) one of the snow towers with the two sampling inlets above the snowpack at 50 and 10 cm (photo credit: D. Helmig).





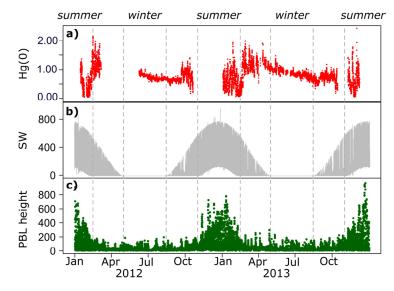


Figure 3: Annual variation in 2012 and 2013 of **a**) hourly-averaged Hg(0) concentrations (in ng/m³) at 500 cm and 25 cm above the snow surface in 2012 and 2013, respectively, **b**) downwelling shortwave (SW) radiation (in W/m²), and **c**) planetary boundary layer (PBL) height (in m). The vertical dashed lines represent seasonal boundaries.





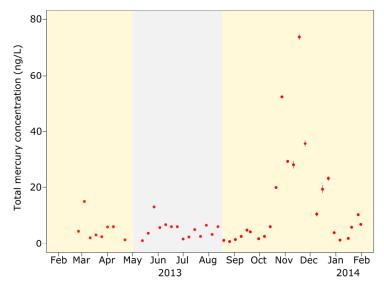


Figure 4: Total mercury concentration (ng/L), along with standard errors, in surface snow samples collected weekly at Concordia Station from February 2013 to January 2014. Dark period (winter) highlighted in grey, sunlit period highlighted in yellow. Total mercury concentrations were elevated (up to 74 ng/L) in November-December 2013 (summer).





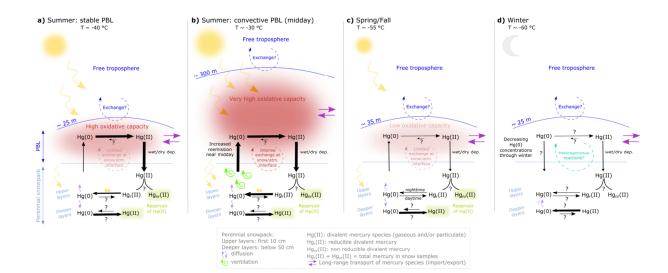


Figure 5: Schematic diagram illustrating the processes that govern the Hg(0) budget at Concordia Station **a**) in summer under stable Planetary Boundary Layer (PBL) conditions, **b**) in summer under convective PBL conditions, **c**) in spring/fall, and **d**) in winter. In summer, Hg(0) is continuously oxidized due to the high oxidative capacity of the boundary layer and a large amount of divalent mercury species deposit onto the snowpack. A fraction of deposited mercury can be reduced (the reducible pool, Hg_r(II)) in the upper layers of the snowpack and subsequently reemitted to the atmosphere as Hg(0). Hg(0) emission from the snowpack maximizes near midday likely due to increased ventilation as a response to daytime heating. Oxidation of Hg(0) dominates in the deeper layers of the snowpack leading to the formation of a Hg(II) reservoir. In spring/fall, the balance of reduction and oxidation processes within the upper layers of the snowpack differs from summertime: oxidation dominates during the day, reduction at night. In winter, Hg(0) is produced in the deeper layers of the snowpack likely as a result of the reduction of Hg(II) species accumulated during the sunlit period. Ambient Hg(0) concentrations exhibit a 20 to 30% decrease through winter.





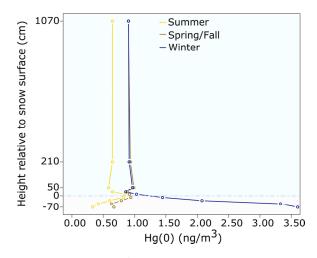


Figure 6: Mean Hg(0) concentration (ng/m³) measured at various heights above and below the snow surface (cm) at Concordia Station in summer (yellow), spring/fall (brown), and winter (dark blue) The horizontal light blue dashed line represents the snow surface.





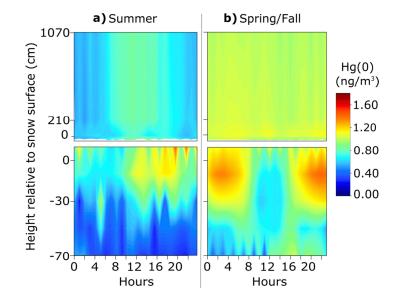


Figure 7: Hourly (local time) mean atmospheric and interstitial air Hg(0) concentrations in **a**) summer, and **b**) spring/fall. The vertical axis is the height of measurement relative to the snow surface (in cm). Color contours show Hg(0) concentrations (in ng/m³).





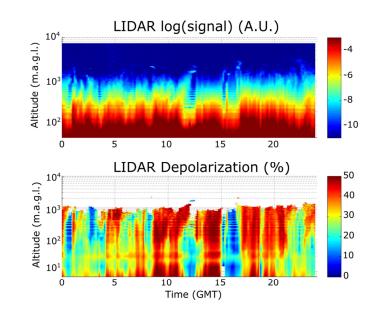


Figure 8: Lidar backscatter (upper panel) and depolarization ratio (lower panel) measured at Concordia Station on 24 February 2013. Ice precipitation was observed in the first 1000 m of the atmosphere, with only two small liquid-water clouds around 8 and 12 UTC.





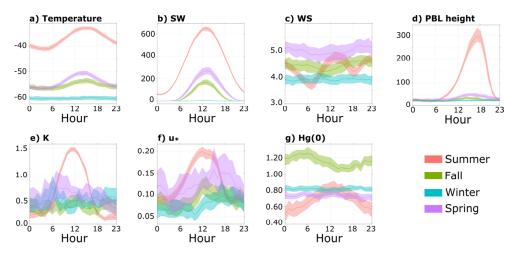


Figure 9: Hourly (local time) mean variation, along with the 95% confidence interval for the mean, of: **a**) temperature (in °C) at 3 m above the snow surface, **b**) downwelling shortwave (SW) radiation (in W/m²) according to the MAR model simulations, **c**) wind speed at 3 m above the snow surface (in m/s), **d**) planetary boundary layer (PBL) height (in m) according to the MAR model simulations, **e**) Eddy diffusivity (*K*, in m²/s), **f**) friction velocity (u_* , in m/s), and **g**) Hg(0) concentration (in ng/m³), in summer (red), fall (green), winter (blue), and spring (purple).





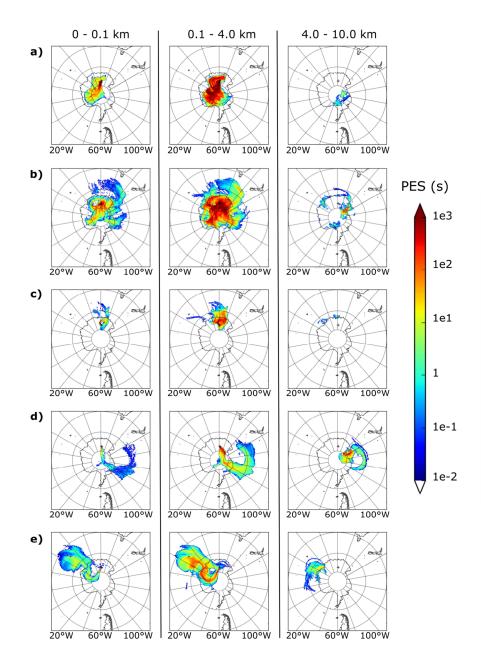


Figure 10: Back trajectories for the 3 layers of altitude colored according to the potential emission sensitivity (PES, in seconds) **a**) from 19 January to 8 February 2012, **b**) from 5 to 20 February 2013, **c**) on 10 February 2012, **d**) on 22 February 2013, and **e**) on 11 September 2013. Note that PES in a particular grid cell is proportional to the particle residence time in that cell.





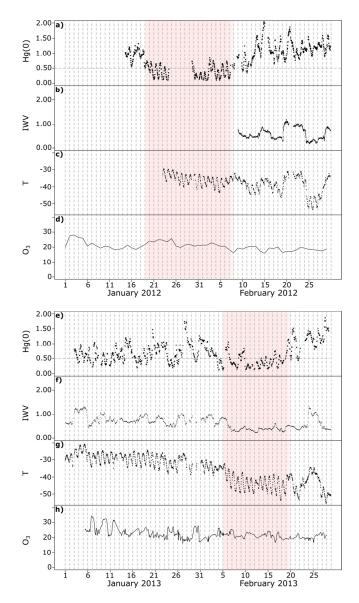


Figure 11: Top: January and February 2012 cycle of: **a**) hourly-averaged Hg(0) concentrations (in ng/m^3), **b**) Integrated Water Vapor (IWV, kg/m²), **c**) Temperature (in °C) at 10 m above ground level, and **d**) ozone (O₃, daily mean) mixing ratios (ppbv). Hg(0) was low from 19 January to 8 February (period highlighted in red) while O₃ showed no abnormal variability. Bottom: January and February 2013 cycle of: **e**) hourly-averaged Hg(0) concentrations (in ng/m^3), **f**) Integrated Water Vapor (IWV, kg/m²), **g**) Temperature (in °C) at 10 m above ground level, and **h**) ozone (O₃) mixing ratio (ppbv). Hg(0), IWV, and temperature were low from 5 to 20 February (period highlighted in red) while O₃ showed no abnormal variability.





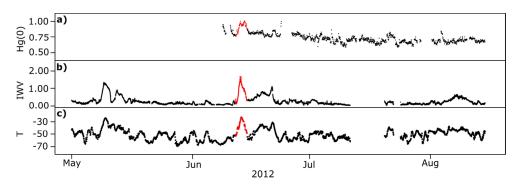


Figure 12: Year 2012 wintertime record of: **a**) hourly-averaged Hg(0) concentrations (in ng/m³), **b**) Integrated Water Vapor (IWV, kg/m²), and **c**) Temperature (T, $^{\circ}$ C) at 10 m above ground level. Hg(0), temperature, and IWV increased from June 12 to 15 (in red) suggesting transport of moist and warm air masses originating from lower latitudes