

Dear Dr. Barbara Ervens,

Thank you very much for your patience and help! Thank you for permission for an extension of the resubmission of the revised manuscript (Ms. Ref. No.:acp-2016-1183). Based on the two reviewers' comments, we have made a substantial revision again to our manuscript, including reanalyzing the model results, enhancing our discussion and changing the figures. Additionally, this resubmitted version has been polished by English editor. We are confident that it is ready for publishing in a high quality journal. Detailed item-by-item responses to the comments are listed below.

Best regards,

Yours sincerely,

Jianhua Qi

Response to Referee report 1

General Comments: First of all, the space was not inserted appropriately in many parts, so it is hard to read and follow. Such a crude revision with low presentation quality should not be sent to reviewers. I have reviewed again this manuscript and found some improvements on the manuscript; however, many replies have not been found in the revised manuscript and/or replied well to my concerns. I feel that the presentation quality is still low as to be published from the high quality journal of Atmospheric Chemistry and Physics. With regret, I have judged to reject this manuscript again.

Response: We are sorry for the space missing problem. We have a double check for our submitted version, but not find this problem. We therefore re-install our software and avoid the problem in the new submission. In the new version, we further improve the quality of the manuscript according to two reviewers' comments. We are confident that it is ready for publishing in a high quality journal.

Q1. I have partly understood my concerning issue regarding the definition of dust event. Further concerning issue is the sampling duration of continuous dust event. Even the dust event continued multiple days, how should we consider the representativeness of the sampling? For instance, sample 20080528 and 20080529 (please note that the sampling time of 20080529 will have typo) had approximately one day interval. Was there large temporal variation of PM10 concentration during continuous dust days? If there was large change on PM10 concentration, why the authors collected on the listed time? The authors should state the reason, and should present the representativeness of 4 hrs sampling. In the revised manuscript, it will be kind for readers to explicitly state that '<http://www-cfors.nies.go.jp/~cfors/>' is for forecast model over Asia, and '<http://www.qepb.gov.cn/m2/>' is for observed concentration at Qingdao.

Response: Due to no dust events lasting over 12 hrs (Lee et al., 2015; Su et al., 2017; Zhang et al., 2007), we collected one dust sample with a 4-hr duration in a day. The sampling for dust particles started only when the measured PM10 mass concentration in Qingdao (<http://www.qepb.gov.cn/m2/>) and the forecasted dust mass over Asia (<http://www-cfors.nies.go.jp/~cfors/>) had greatly increased.

The samples with ID of 20080528 and 20080529 were subject to two different dust events occurring in two days instead of continuous samples for one dust event. On March 20-21, 2010, two dust events subsequently swept Qingdao. The 4 hr dust samples with IDs of 20100320 and 20100321 may not capture the entirety of the two events. However, the on-line data can allow adequate separation of the two dust event samples. The same was true for the dust samples with IDs of 20110501, 20110502.

The link illustration for these two links has been also added properly in the new version.

Lee, Y. G., Ho, C., Kim, J., and Kim, J.: Quiescence of Asian dust events in South Korea and Japan during 2012 spring: Dust outbreaks and transports, *Atmos. Environ.*, 114, 92-101, 2015.

Su X., Wang Q., Li Z., Calvello M., Esposito F., Pavese G., Lin M., Cao J., Zhou C., Li D., Xu H.: Regional transport of anthropogenic pollution and dust aerosols in spring to Tianjin — A coastal megacity in China, *Sci. Total. Environ.*, 584–585, 381–392, 2017.

Zhang, K., and Gao, H. W.: The characteristics of Asian-dust storms during 2000–2002: From the source to the sea, *Atmos. Environ.*, 41, 9136-9145, 2007.

Q2. I have partly agreed, but I have further question on the application of a 3-D air quality model. First, what is the merit of the application of 3-D air quality model? In the revised manuscript, only the spatial distributions of PM₁₀ were shown (Fig. 2 from CFORS model and Figs. S1-S3 with CFORS and WRF-CMA. Can such application reinforce the authors' discussion points? The behavior of IN were discussed in this manuscript, so what is the purpose to show PM₁₀? The authors stated that 'The spatial distribution of PM₁₀ concentrations for each dust event was consistent with the model results of dust by the Chemical Weather Forecast System (CFORS) by Uno et al. (2003)' (L199-201). If the consistency between other models is important, why the author calculated on your own model? I cannot follow this reason from the revised manuscript.

Response: Thank you for the suggestion. We have deleted the results by CFORS. The CMAQ model (v5.0.2) was applied to simulate the concentration of PM₁₀, NO_x, NH₃, NO₃⁻ and NH₄⁺ over the East Asia area for aerosol samples on dust and comparison days. We have revised the discussion on model results. Distribution of PM₁₀ was used to characterize the dust events. Spatial distributions of PM₁₀ during each dust events were consistent with the records in the "Sand-dust Weather Almanac" (CMA, 2009; 2010; 2012; 2013). The model results indicated that CMAQ results reasonably reproduce the mass concentrations of NO₃⁻ (Fig. S6). Simulated NH₄⁺ concentrations in dust samples were severely under-predicted with NMB values at -71%. For reference samples, simulated NH₄⁺ concentrations sometimes can well reproduce the observational values, but sometimes totally off. The external mixing mechanism proposed in this study is urgently needed to be included in the model for accurately predicting the concentrations during dust events.

CMA: Sand-dust weather almanac 2008, China Meteorological Press, Beijing, 10-64, 2009.

CMA: Sand-dust weather almanac 2009, China Meteorological Press, Beijing, 11-59, 2010.

CMA: Sand-dust weather almanac 2010, China Meteorological Press, Beijing, 11-79, 2012.

CMA: Sand-dust weather almanac 2011, China Meteorological Press, Beijing, 10-53, 2013.

Q3. The following specific points also should be revised to clarify the model application.

L189: Centered point is needed because we cannot follow the modeling domain at the current description.

L193: On the INTEX-B emission inventory (Zhang et al., 2009), I suppose that NH₃ emissions have not been provided. If so, this description should be changed.

L195, and Figures 6 and 8: So, all calculations were based on the emission level on 2008? Because the temporal resolution of INTEX-B emission inventory is month, I feel that there are no need to display all emissions on all dust samples. These emissions level should be differed only on month. Therefore, I suppose that the averaged (spring time) emissions of NO_x and NH₃ on each one figure is enough.

Response: We have supplemented the centered point (110 °E, 34 °N) in the new version.

According to the publications of INTEX-B and TRACE-P Asia emission inventories (Zhang Q et al., 2009; Streets et al., 2003.), INTEX-B inventory was developed based on TRACE-P inventory with NH₃ emission considered (the annual emission amount of NH₃ in China was 13.6 Tg). However, due to the low priority and low variability of NH₃ emission during 2000-2006, NH₃ emission was not updated in INTEX-B inventory, and the NH₃ emission in INTEX-B inventory was consistent with TRACE-P.

Agree and revised.

Zhang Q et al., Asian emissions in 2006 for the NASA INTEX-B mission; *Atmos. Chem. Phys.*, 9, 5131-5153, 2009.

Streets D.G. et al., An inventory of gaseous and primary aerosol emissions in Asia in the year 2000, 108 (D21), DOI: 10.1029/2002JD003093, 2003.

Q4. Figure S1: What is the purpose to show the difference between (b) and (c)? In this caption, what is 'WRF-CMA' ?

Response: We have indicated that one exterior dust sample was collected on 1 March when no dust was recorded in Qingdao by MICAPS. However, the MICAPS information over the whole country indeed showed the dust events in China on 1 March. And the modeled spatial distribution of PM₁₀ and TSP mass concentration for this dust event on 1 March implied that the sample should be classified into dust sample. Therefore we listed all the supporting figures in Fig. S1. Fig. S1 (b) was the weather information from the MICAPS at 8:00 on Mar.2, 2008 and (c) was hourly PM₁₀ concentration modeled by the WRF-CMAQ model at 15:00 on Mar.1, 2008. We guessed that the reviewer maybe refer to the difference between (c) and (d), therefore we deleted (d).

We have revised the caption.

Q5. Figure S3: In the main manuscript, it was stated that 'each dust sampling day are shown' in Fig. S3 (L218-219, L895). However, only the hourly concentration of PM₁₀ concentration at 14:00 on 19 Mar 2011 were shown. Please confirm this supplemental figure.

Response: We really modeled the PM₁₀ concentration on each dust sampling day, but

only showed the PM10 concentration at the middle time of the sampling in Fig.S3 (Now Fig.S5 in the new version) due to too many figures. We have revised this section and the sentence has been revised into “The concentrations of PM10 and its major components NO_3^- and NH_4^+ over East Asia on dust days and comparison days were modeled using the WRF-CMAQ model (Fig. S5-6)” in L341-342 in the new version.

Specific comments:

Q6. L35-36: This conclusion does not match to the manuscript contents. The authors stated that input of nitrogen to the ocean depends on the dust events.

Re-comment: cannot find this revision.

Response: The revision in the last round was prepared after the quick response. After a careful consideration, we agreed with the comment and delete the part in the last revised version.

Q7. I have confirmed and understood the meaning. However, is this revision corresponded to L226-230? If so, this revised sentence seems to contain many errors (NOT Table S2 but Table S1?). For example, we can find $410 \mu\text{g}/\text{m}^3$ on dust day sample on 20080315. What is the value of 80-1303%? These increased value were not corresponded to ‘Ratio of DD to CS’ shown in Table S1.

Response: Yes, this revision corresponded to L226-230 in last version. And the times of dust to non-dust day samples were replaced by the ratio according to the former suggestion. To avoid the confusion, we have revised Table S1 to give the increased ratio.

Q8.L171: Again, I cannot follow the calculation of “1.7-21.9 times (mean: 6.9) “.

L175: I cannot follow “10.3 times” for Fe. It can be calculated as 7.90 from the values in Table 2.

L176: “3.6-fold” will not be followed from Fig. 2. It should be listed in Table 2.

Re-comment: So, in this revised manuscript, these statements of the increment ratio on dust-day compared to non-dust day have not been explicitly appeared. In L243, the authors stated ‘Table S1’, but Table S1 contained not only the information of inorganic nitrogen but also TSP, Al, Fe, and nss-Ca. So, it is appropriate to mention on Table S1 in Section 3.1.

Response: Agree and revised.

Q9.L175: So, please state explicitly regarding this point to the readers. In the current form, nss-Ca was suddenly shown in Fig. 3 without any introduction.

Response: Agree and revised.

Q10. First of all, I cannot find the revision of ‘The concentrations of ammonium were increased by 20’ anywhere. Is this corresponded to Table S1? I suppose that the

authors discussed regarding this point in L240-L244. Although ratio was shown in Table S1, percentages are discussed here. So it is hard to follow the manuscript. Why the discussion point have not been arranged on the uniformed unit?

Response: Agree and revised.

Q11.L194-L195: In this sentence, the authors stated “the effect of dust on inorganic nitrogen differed during different types of dust events “. Why the authors suddenly focused on inorganic nitrogen here? In L192-193, it was mentioned “inorganic ion SO_4^{2-} exhibited concentration variations that were similar to those of nitrate” . L197: The figures for inorganic nitrate will be helpful information here, if the authors focused on inorganic nitrogen.

Re-comments: I cannot find this revision.

Response: The revision in the last round was prepared after the quick response. After a careful consideration, we completely rewrote the part to avoid confusion in the last revised version.

Q12. L207: (respectively less than 50 ug/g and 6 ug/g) will be the correct expression for ammonium.

Re-comments: I cannot find this revision.

Response: The revision in the last round was prepared after the quick response. After a careful consideration, we completely rewrote the part into “The ratios of mass concentrations of nitrate and ammonium to the total mass of sand particles were very low, i.e., less than 81 $\mu\text{g/g}$, which are approximately three orders of magnitude less than the corresponding values in our dust samples.” at L230-232.

Q13.I am wondering that the differences of IN concentration between Duolun and Zhurihe. Both are Hunshandake Desert, however, as is shown in Table 5, IN concentration was much higher in Duolun. Are there some emission source?

Response: Sand samples were collected at a remote site in Zhurihe desert. Little anthropogenic influence is expected. Atmospheric aerosol samples were collected at an urban site in Duolun on dust days for comparison. It is not surprised for a strong signal for anthropogenic sources. This has been clarified in the new version.

Q14. Again, only from the dust spatial distribution, it is hard to state the dust intensity.

Response: We had made a substantial revision on the part in the last round revision and didn't consider dust intensity as an important factor for our unique results.

Q15. L214-L216: Without more information of the intensity of dust, the discussion on ‘dilution effect’ seems to be lacked in scientific understanding. This part should be fully revised based on not only dilution effect but also dust intensity.

Re-comments: Again, only from the dust spatial distribution, it is hard to state the dust intensity.

Response: We had made a substantial revision on the part in the last round revision and didn't consider dust intensity as an important factor for our unique results.

Q16. I have confirmed the revision, but if the authors discussed on average (L300-303), the averaged values were also needed.

Response: We really had given the average of TSP in form of average±standard deviations at L300-303 in last revision. Now we had made a substantial revision on this part and didn't discuss TSP average concentration.

Q17. L227-L228: The favorable condition to form ammonium cannot be discussed without the information of NH₃. In addition, Table 3 indicated the aerosol samples in the coastal region of the Yellow Sea. How about the status over air mass path? Is it sufficient to conclude only from the downwind information to the formation of inorganic nitrogen?

Re-comments: Again, I cannot understand the model application results.

Response: We modeled the emission and concentration of NO_x and NH₃ over East Asia on the dust and comparison days. The model results showed that the calculated trajectories of the entire dust air mass passed over those highly polluted regions with strong emissions of NO_x and NH₃ shown in Fig 6 and experienced different residence times therein. The average concentration of NO_x and NH₃ during transport were calculated and discussed according to Categories 1 and 2. The air masses in Category 1 took over 11-39 hrs to cross over the highly polluted area with appreciable concentrations of NO_x (5.7±1.4 ppb) and NH₃ (7.6±3.3 ppb). Except for the exterior samples, air masses in Category 2 took less than 10 hrs to cross over the polluted areas with lower concentrations of NO_x (3.6±3.4 ppb) and NH₃ (4.7±4.7 ppb) and the mixing layer height along the route was 916-1194 m (on average) for each dust event. This further led to the external mixing of anthropogenic particulate matters and dust.

Q18. L230: "strong dust storm" cannot be discussed without any information on dust intensity here.

Re-comments: Again, from the additional information of CFORS, the spatial distribution pattern was found; however, how can we estimate the intensity?

Response: We had made a substantial revision on the part in the last round revision and didn't consider dust intensity as an important factor for our unique results.

Q19. L233-L234: But NO_x concentration was high in Case 3. I cannot follow why the authors concluded on Case 3.

Re-comments: So where did the authors discussed the NO_x concentration in the manuscript?

Response: We had made a substantial revision on the part in the last round revision after the quick response. The NO_x concentration was discussed in Section 4.3 in the new revision.

Q20. L254-L255: The authors simply mentioned “local emissions” here. Because the samples were collected on downwind regions in the coastal region of the Yellow Sea, I guess that the discussion on emission characteristics of each (or, at least, some categorized) air mass should be discussed in detail. The inorganic nitrogen concentrations are highly related to the local conditions both on emissions strength and meteorological parameters, so the discussion only on air mass speed and air mass path over ocean are insufficient.

Re-comments: Again, model is used only for spatial distribution and not inform the chemical production process.

Response: We had made a substantial revision according to the suggestion. The chemical production process was discussed in Section 4.1 “Theoretical analysis of the three categories”. In Category 1, ammonium salt aerosols may externally exist with dust aerosols in these dust day samples and NO_3^- and SO_4^{2-} were almost completely associated with NH_4^+ in these dust day samples; whereas a larger fraction of $\text{NO}_3^- + \text{SO}_4^{2-}$ may exist as metal salts due to reactions of their precursors with dust aerosols in Category 2. The simulated NO_3^- and NH_4^+ concentrations was compared with the observation in Qingdao, and the results indicated that the external mixing mechanism proposed in this study is urgently needed to be included in the model for accurately predicting the concentrations during dust events.

Q21. I have rechecked the discussion of backward trajectories discussed on Section 3.4. There are many points should be clarified.

Figure 5: Please add the explanation of the trajectory of 20110415 was excluded based on the discussion on Fig. 2. Why the authors displayed “non-dust samples” ? What were the differences between non-dust and dust samples trajectories? I feel that these were similar.

Response: Agree and revised.

Q22. L314: What is the ‘remaining one’ ? Please specify the trajectory data. In my opinion, two trajectories of 20110418 and 20110501 originated from northeast China.

Response: Yes, trajectory 20110501 was really from northeast China, however it then passed over the Inner Mongolia, and arrived at Qingdao from north, just like 20110502. Therefore, we grouped the trajectory into the air mass originated from Inner Mongolia, China. However, it was really easy to mislead the readers. Therefore, we accepted the suggestion, and revised the sentences into “The calculated air mass trajectories for 13 out of 14 samples showed that the air mass originated from North and Inner Mongolia, China (Fig. 5), generally consistent with the results of Zhang and Gao (2007). The remaining one, with ID of 20110418 originated from Northeast China.”.

Q23. L317: What is the ‘one exterior sample’ ? Please specify.

Response: Agree and revised.

Q24. L319: I cannot see ‘the air masses crossed over the sea for 94-255km’ from Figure 5, because Fig. 5 showed the whole view of trajectories across China. More detailed figure or explanation will be required.

Response: The distance over sea of the air mass for each sample was measured from the trajectory using TrajStat software (Wang et al., 2009). We have added the explanation in Section 2.3.

Wang, Y. Q., Zhang, X. Y., and Draxler, R. R.: TrajStat: GIS-based software that uses various trajectory statistical analysis methods to identify potential sources from long-term air pollution measurement data, *Environ. Modell. Softw.*, 24, 938-939, 2009.

Q25. L328-329: What is the definition of the ‘average mixing layer’ . I suppose that the altitude of backward trajectories were so high because most of trajectories were originated outside China on 72 hrs. So, where is the averaged region to calculate ‘900m’ in this sentence?

Response: The average mixing layer was calculated as an average of all points on the air mass back trajectory of each sample. This has been clarified in section 2.4 in the new version.

Q26. I cannot still understand the authors’ conclusion here. As was discussed on L355-361, Fig. 8, and Table 7, although the coal combustion have increased on dust days, the contributions of local anthropogenic sources (especially secondary aerosols) have decreased on dust days. According to the discussion on Section 3.3, the authors concluded that ammonium salts were externally co-exist with dust aerosols in Category 1. So, why the contribution of secondary aerosols were decreased from PMF analysis. I feel that these results have contradicted. More careful discussion is required for this conclusion.

Response: The source of coal combustion on dust days became complicated. “mixture of coal combustion and other pollutants” means these compounds present contemporaneously, because that PMF model can’t show the mixing or existing state. We have revised the sentence into “The source profile for coal combustion in dust day samples showed a high percentage of K⁺, Cl⁻, Ca, Mg, Co, Ni, As, Al and Fe, indicating coal combustion presenting contemporaneously with other pollutants emitted along the transport path on dust days.”. Ammonium salts were externally co-exist with dust aerosols in Category 1, but showed lower concentrations in Category 2 likely due to unfavorable conditions for forming ammonium salts. Here the conclusion was a result of source appointment for all dust samples including Category 1 to 3. And we have revised the sentence “In these dust samples, including Categories 1-3, oil combustion, industry, soil dust, secondary aerosols, and coal combustion/other sources were identified as five major sources (Table 6).

Q27.L306: I cannot follow the calculation of “a factor of 1.1-5.8” and “a factor of 1.8-6.3” .

Re-Comments: I cannot find this revision.

Response: We had made a revision on these sentences in the last round revision after the quick response. According to the former suggestion, we changed the factor to ratios. And this sentence was revised to “In Category 1, the dry deposition fluxes of $N_{NH4++NO3-}$ increased by 9-75% with increased TSP flux by 86-252% (Table S3)” at L371-372 in the new revision. And we also revised Table S3 to give increased proportion and the calculation method.

Q28. L309: What is the calculation method of “63%” and “46%” ?

L310: What is the calculation method of “14%” ?

Re-Comments: I cannot find this revision.

Response: We had made a revision on these sentences in the last round revision after the quick response. We have revised Table S3 to give increased proportion and the calculation method.

Q29. L317: For Fe, it seems that the increased ratio were 2.81-11.08 from Table S3.

Response: This sentence has been revised to “However, the dry atmospheric deposition fluxes of Fe increased by a factor of 124-2370% in dust day samples.” at L383-384 in new revision.

Specific comments:

Q30.Table 6: Missing the note of a and b.

Response: Done.

Q31. Table S2: Please align the right-column, it is hard to follow. What is the meaning of *?

Response: Agree and revised.

Response to Referee report 2

This manuscript focuses on an important issue, specifically the relationship between ammonium and nitrate with mineral dust in China. The variability in previously reported relative concentration trends is worth exploring in detail. There are a number of areas where the paper could use improvement prior to publication, particularly with connecting to the literature and placing the work in context.

Response: We thank the reviewer's constructive comments and revise our manuscript accordingly. We are confident that it is ready for publishing in a high quality journal.

Comments:

Q1 - Throughout the manuscript there are odd spacing issues, where two words are together without a space. As an example in the abstract "For these two groups, NH₄⁺ in dust day samples was present in the form of ammonium salts externally co-existing with dust aerosols or the residual of incomplete reactions between ammonium salt and carbonate salts."

Response: We are sorry for the space missing problem. We have a double check for our submitted version, but not find this problem. We therefore re-install our software and avoid the problem in the new submission.

Q2 - Line 51 - 54: The authors state the Asian dust has been transported as far as the north pacific, but this understates what has been observed for Asian Dust. Uno et al 2009 Nat Geosci showed that Asian dust can circumnavigate the globe. showed Asian dust impacting California air quality, while Ault et al. 2011 JGR and Creamean 2013 Science showed impacts of Asian dust on orographic precipitation in the Sierra Nevada (in California). Pratt et al. 2009 Nat Geo showed Asian dust influencing clouds over Wyoming.

Response: Thank you very much for the suggestion. We revised the sentence to "Asian dust has been reported to not only frequently cross over the mainland and the China Seas, but also to occasionally reach the remote northern Pacific Ocean or North America (Creamean 2013; Tan and Wang, 2014; Van Curen and Cahill, 2002; Zhang and Gao, 2007). In an extreme case, Asian dust was found to be transported more than one full circuit around the globe in approximately 13 days (Uno et al 2009)." at Line 45-49.

Creamean, J. M., Suski, K. J., Rosenfeld, D., Cazorla, A., DeMott, P. J., Sullivan, R. C., White, A. B., Ralph, F. M., Minnis, P., Comstock, J. M., Tomlinson, J. M., Prather, K. A.: Dust and Biological Aerosols from the Sahara and Asia Influence Precipitation in the Western U.S., *Science*, 339, 1572-1578, 2013.

Tan, S. C., and Wang, H.: The transport and deposition of dust and its impact on phytoplankton growth in the Yellow Sea, *Atmos. Environ.*, 99, 491-499, 2014.

VanCuren, R., and Cahill, T.: Asian aerosols in North America: Frequency and concentration of fine dust, *J. Geophys. Res.*, 107(D24), 4804, doi:10.1029/2002JD002204, 2002.

Zhang, K., and Gao, H. W.: The characteristics of Asian-dust storms during 2000–2002: From the

source to the sea, *Atmos. Environ.*, 41, 9136-9145, 2007.

Q3 Though the authors note that a native English speaker was utilized for the revision, a considerable improvement in the grammar and proofreading are needed before the writing is at a publishable level.

Response: This resubmitted version has been polished by English editor.

Q4 Line 233 insert comma after “samples”

Response: Done.

Q5 Line 266: Is it really a safe assumption that gas aerosol thermodynamic equilibrium is met for inorganic ions during a dust storm? It would seem that many non - aqueous (i.e. solid) aerosol would be present that would not have normal equilibrium partitioning. It would be nice to see some evidence of this. This would also help support the conclusion that $\text{Ca}(\text{NO}_3)_2$ and CaSO_4 are negligible.

Response: According to this reviewer’s suggestion, the progresses on this issue in literature have been summarized and cited to support our analysis.

Q6 Line 282: The presence of Cu, brings to mind the question of transition metal ions and industrial sources of metal containing particles. How were these accounted for? Particularly since they often have different properties and propensity for generating ROS as Weber and company at Georgia Tech have shown.

Response: Cu was once used as an effective marker of diesel and biodiesel-blend exhaust (Gangwar et al., 2012), while it can also be derived from copper pyrites (CuFeS_2) in Inner Mongolia mines (Huang et al., 2010). The increase of Cu in the mass concentration in dust samples implied dust particles mixed with anthropogenic particles, particularly from industrial emissions, during transport.

Gangwar, J. N., Gupta, T., and Agarwal, A.K.: Composition and comparative toxicity of particulate matter emitted from a diesel and biodiesel fuelled CRDI engine, *Atmos. Environ.*, 46, 472-481, 2012.

Huang, K., Zhuang, G., Li, J., Wang, Q., Sun, Y., Lin Y., and Fu J. S.: Mixing of Asian dust with pollution aerosol and the transformation of aerosol components during the dust storm over China in spring 2007, *J. Geophys. Res-Atmos.*, 115, D00k13, Doi:10.1029/2009jd013145, 2010.

Q7. - It should be noted that there is a great deal of uncertainty regarding aerosol pH, particularly in North China, with estimates ranging from 3 - 7 pH units. This of course will affect nitrate. The authors could comment on this with respect to their data, though keeping in mind Hennigan et al showing the proxy methods such as $\text{NH}_4^+ / (\text{NO}_3^- + \text{SO}_4^{2-})$ are qualitative at best.

Response: We thank the comments. We are not sure whether the estimated aerosols pH from 3 - 7 pH units in north China were valid or not, by considering three types of

aerosols, i.e., ammonium salt aerosol, K_2SO_4 or KNO_3 aerosol, and $CaCO_3$. TSP was collected in this study while $PM_{2.5}$ was used for analysis by Hennigan et al (2015). It is not surprising parts of NO_3^- and SO_4^{2-} to be associated with metals in TSP samples, but we agree that NO_3^- and SO_4^{2-} may overwhelmingly be associated with NH_4^+ in $PM_{2.5}$ as found by Hennigan et al (2015). However, we tried our best to properly interpret the formation of nitrate and sulfate in different Categories in the revised version.

Hennigan, C. J., J. Izumi, A. P. Sullivan, R. J. Weber, and Nenes, A.: A critical evaluation of proxy methods used to estimate the acidity of atmospheric particles, *Atmos. Chem. Phys.*, 15(5), 2775–2790, 2015.

Q8. What is the mineralogy of the Hunshandake Desert? Is it rich in $CaCO_3$? Based on a few assumptions made, documentation of the presence of this mineral from aerosols in the region would be helpful. Perhaps some of Ro and co-workers analysis of transported dust with SEM - EDX?

Response: Thank you for the suggestion. The references results show that mineral dust is relatively enriched with Calcite (Matsuki, et al., 2005; Formenti et al., 2011; Nie et al., 2012) and the carbonate content generally decreases from west to east with exception of the Gurbantunggut desert in China desert (Fig.RS1, Formenti et al., 2011). Except for 20080502, the remaining dust samples in Category 2 were transported from the desert relatively enriched with $CaCO_3$ (1-25% in Wt%) (Formenti et al., 2011). And Huang et al. (2010) found that calcite was one of the main species in the aerosol over Duolun and the high content of $CaCO_3$ (~80% in dust storm days) in the total soluble part of Duolun aerosol. A positive correlation between NO_3^- and SO_4^{2-} in Category 2 against a negative correlation in Category 1 also implied that the dust particles enriched with $CaCO_3$ in Category 2 might play an important role to form SO_4^{2-} and NO_3^- . Ca-rich dust particles coated with highly soluble nitrate were observed at Kanazawa in Japan during Asian dust storm periods using SEM/EDX (scanning electron microscopy equipped with an energy dispersive X-ray spectrometer) (Tobo et al., 2010). The single-particle observation conducted by Hwang and Ro (2006) showed that $CaCO_3$ in dust particles was almost completely consumed to produce mainly $Ca(NO_3)_2$ species.

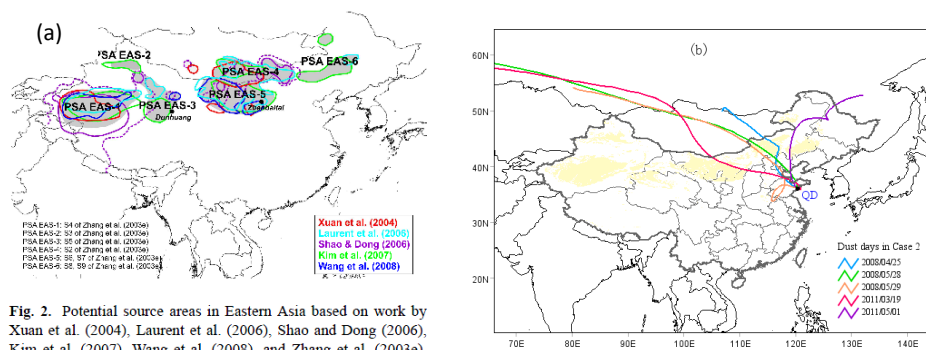


Fig. 2. Potential source areas in Eastern Asia based on work by Xuan et al. (2004), Laurent et al. (2006), Shao and Dong (2006), Kim et al. (2007), Wang et al. (2008), and Zhang et al. (2003e). Outlines of potential source areas (shaded areas) are drawn by hand. PSA EAS-1: Taklamakan; PSA EAS-2: Gurbantunggut; PSA EAS-3: Kumtaq, Qaidam, Hexi corridor; PSA EAS-4: Mongolian (Northern Gobi) deserts; PSA EAS-5: Inner Mongolian (Southern Gobi) deserts: Badain Jaran and Tengger (PSA EAS-5a), Ulan Buh, Hobq, Mu Us (PSA EAS-5b); PSA EAS-6: north-eastern deserts (Otindag Sandy Land, Horquin Sandy Land, Hulun Buir Sandy Land).

Figure RS1. Potential source areas in Eastern Asia adapted from Formenti et al., 2011(a) and The 72-h backward trajectories for samples in Category 2 (b).

Formenti, P., Schütz, L., Balkanski, Y., Desboeufs, K., Ebert, M., Kandler, K., Petzold, A., Scheuven, D., Weinbruch, S., and Zhang, D.: Recent progress in understanding physical and chemical properties of African and Asian mineral dust, *Atmos. Chem. Phys.*, 11, 8231–8256, doi:10.5194/acp-11-8231-2011, 2011.

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Hwang, H. and Ro, C. U.: Direct observation of nitrate and sulfate formations from mineral dust and sea-salts using low-Z particle electron probe X-ray microanalysis, *Atmos. Environ.*, 40, 3869-3880, 2006.

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Nie, W., Wang, T., Xue, L. K., Ding, A. J., Wang, X. F., Gao, X. M., Xu, Z., Yu, Y. C., Yuan, C., Zhou, Z. S., Gao, R., Liu, X. H., Wang, Y., Fan, S. J., Poon, S., Zhang, Q. Z., and Wang, W. X.: Asian dust storm observed at a rural mountain site in southern China: chemical evolution and heterogeneous photochemistry, *Atmos. Chem. Phys.*, 12, 11985-11995.

Tobo, Y., Zhang, D. Z., Matsuki, A., and Iwasaka, Y.: Asian dust particles converted into aqueous droplets under remote marine atmospheric conditions, *PNAS Proceedings of the National Academy of Sciences of the United States of America*, 107, 17905–17910, 2010.

Q8 - Line 321 some evidence for “humid marine conditions might have enhanced particle - particle coagulation” would be helpful. The number concentrations in the marine boundary layer are unlikely to be > 105 #/cm³ where coagulation is prevalent, more likely in the 102–103 #/cm³. Are the authors referring to fog - processing? That would seem to be the primary way this could happen in a marine environment.

Response: We are sorry for the confusion. The sentence has been revised to “On the other hand, the humid marine conditions (the average RH ranged in 50-75% over the Bohai and Yellow Seas in 2006-2012) might have enhanced hetero-coagulation between dust and smaller anthropogenic particles, leading to the release of NH₃ via reactions between preexisting ammonium salts and carbonate salts.”.

Tobo Y., Zhang D., Matsuki A., Iwasaka Y. Asian dust particles converted into aqueous droplets under remote marine atmospheric conditions, PNAS, 2010, 107: 17905–17910.

Q9 - Line 326 The line “ammonium salts mostly co - existed with dust aerosols externally” is confusing as written. Is the population externally mixed with respect to ammonium nitrate and dust? Or are the salts co - existing with dust, but not other particle types? Please rephrase for clarity.

Response: Thank you for the suggestion. We have revised to “ammonium salt aerosols may externally exist with dust aerosols”.

Q10 - Overall many of the conclusions on page 12 appear to mostly be speculation with little data to support it. I would recommend sticking to conclusions with more support from the data in the paper.

Response: We have made major revision on this section in the new revision.

Q11 - Line 357: The source profile for coal, could it have dust mixed in? When the author’ s say that there is a “mixture of coal combustion and other pollutants” are they saying that they are internally mixed or simply present contemporaneously? Clarifying that point would be helpful.

Response: “mixture of coal combustion and other pollutants” means these compounds present contemporaneously, because that PMF model can’t show the mixing or existing state. We have revised the sentence into “The source profile for coal combustion in the dust day samples showed a high percentage of K⁺, Cl⁻, Ca, Mg, Co, Ni, As, Al and Fe, indicating that coal combustion particles may exist contemporaneously with other anthropogenic pollutants emitted along the transport path.”.

Q12 - Overall the Figures could use improvement as portions are hard to read and the take home point of each is not always clear. It seems at times as if the authors are simply showing everything they can, as opposed to targeting their figure to the main points of the paper.

Response: Agree and revised.

1 **The concentration, source and deposition flux of**
2 **inorganic nitrogen in atmospheric particles during dust**
3 **events at a coastal site in northern China ~~during dust~~**
4 **events**

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11
12 **Abstract.** Asian dust has been reported to carry anthropogenic reactive nitrogen during ~~the~~
13 transport from source areas to the oceans. In this study, we attempted to characterize ~~the~~-NH₄⁺ and
14 NO₃⁻ in atmospheric particles collected at a coastal site in northern China during spring dust events
15 from 2008 to 2011. Based on the mass concentrations of NH₄⁺ and NO₃⁻ in each total suspended
16 particle (TSP) sample, the samples can be classified into ~~increasing or decreasing two types groups.~~
17 ~~I~~-in Category 1, the concentrations of NH₄⁺ and NO₃⁻ were 20%-440% higher in dust day samples
18 relative to samples collected immediately before or after ~~a~~ dust event, ~~while in Categories 2 and~~
19 ~~3.~~ ~~T~~hese concentrations decreased by ~~10-75%~~ in the dust day samples ~~in Categories 2 and 3.~~
20 ~~Back trajectory analysis suggested showed that multiple factors such as the transport distance~~
21 ~~prior to the reception site, the mixing layer depth on the transport route and the residence time~~
22 ~~across highly polluted regions, might affect affect the concentrations of NH₄⁺ and NO₃⁻ were~~
23 ~~apparently affected by the transport distance over sea prior to the reception site, the mixing layer~~
24 ~~depth and the residence time across highly polluted regions during transport. For these two groups,~~
25 NH₄⁺ in ~~the~~ dust day samples was ~~likely present either~~ in the form of ammonium salts ~~existing~~
26 externally ~~existing co-existing~~ with dust aerosols or ~~as~~ the residual of incomplete reactions
27 between ammonium salts and carbonate salts. ~~The~~-NO₃⁻ in the dust day samples was attributed to
28 ~~various formation processes interactions between anthropogenic air pduring the long-range~~
29 ~~transporttransport~~ pollutants and dust particles during dust transport from the source zone to the
30 ~~reception site. Back trajectory analysis showed that the concentrations of NH₄⁺ and NO₃⁻ were~~

31 ~~apparently affected by the transport distance over sea prior to the reception site, the mixing layer~~
32 ~~depth and the residence time across highly polluted regions during transport.~~ The positive matrix
33 factorization (PMF) receptor model results showed that the contribution of soil dust increased
34 from 23% to 36% (90% of the residuals < 3.0 and $r^2=0.97$) on dust days with decreasing
35 contributions ~~of from~~ local anthropogenic inputs, ~~especially secondary aerosols and~~
36 ~~associated~~ associated secondary aerosols. The dry deposition flux of atmospheric particulates
37 ~~increased from 2,800±700 mg/m²/month on non-dust comparison days to 16,800±15,900~~
38 ~~mg/m²/month on dust days. The dry deposition flux of particulate inorganic nitrogen increased by~~
39 ~~9-285% in Category 1. The average dry deposition flux of nitrate decreased by 46%-73% in~~
40 ~~Category 2, while that of ammonium decreased by 47% in Category 3. The estimated dust~~
41 ~~deposition flux of inorganic nitrogen varied greatly from event to event, e.g., the dry deposition~~
42 ~~flux of particulate inorganic nitrogen increased by 9-285% in Category 1, but decreased by~~
43 ~~46%-73% in Category 2. Overall, a slight increase in dry deposition flux of particulate inorganic~~
44 ~~nitrogen associated with dust events in this study relative to values in the literature may reflect the~~
45 ~~combined effect of anthropogenic nitrogen emissions and the occurrence of natural dust events.~~

46 Keywords: aerosols, nitrogen, dust, source apportionment, dry deposition flux

47 1 Introduction

48 Reactive nitrogen carried in dust particles can be transported over a long distances, and the
49 atmospheric nitrogen deposition in oceans has been recognized as an important external source of the
50 nitrogen supporting phytoplankton growth (Duce et al., 2008; Zhang et al., 2010b). This hypothesis has
51 been evaluated ~~in-through~~ incubation experiments, ~~in-situ- in situ~~ in situ experiments, and ~~through~~ the use of
52 satellite observational data (Banerjee and Kumar 2014; Shi et al., 2012; Guo et al., 2012; Liu et al.,
53 2013; Shi et al., 2012; Banerjee and Kumar 2014; Tan and Wang, 2014). For example, Tan and Wang
54 (2014) found that a phytoplankton bloom with a nearly four fold increase in chlorophyll concentrations
55 occurred 10-13 days after dust deposition. In addition, Banerjee and Kumar (2014) hypothesized that
56 dust induced episodic phytoplankton blooms are important to the interannual variability of chlorophyll
57 ~~in the Arabian Sea. However, the process is dynamic because of due to the worldwide changing~~
58 ~~emission~~ emissions of NO_x and NH₃ ~~worldwidely~~ in the last few decades. For example, China and
59 ~~most of the developing countries in Asia experienced~~ experienced a large ~~Dramatic changes have~~

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60 occurred in the reactive nitrogen in anthropogenic emission in the last three decades, e.g., large
61 increases in emissions of NH₃ and NO_x in emissions in China and other developing countries in Asia
62 and while a substantial decrease in emissions occurred in Europe over the last three decades (Grice et al.,
63 2009; Liu et al., 2017; Ohara et al., 2007; Skjøth and Hertel, 2013). These changes would may greatly
64 affect the nitrogen carried by dust particles to some extent, but and updated few recent studies have are
65 thereby essential examined this issue.

66
67 Asian dust is one of three largest dust sources the main components of dust worldwide on the earth.
68 Asian dust has been reported to not only to to frequently cross over cross over the mainland and the
69 marginal seas of China Seas frequently and reach, but also to occasionally reach —as far as the remote
70 northern Pacific Ocean or the continent of the US North America occasionally— (Creamean 2013; Tan
71 and Wang, 2014; Van Curen and Cahill, 2002; Zhang and Gao, 2007; Tan and Wang, 2014). In an
72 extreme case. And Asian dust is was was found that can to be transported more than one full circuit
73 around the globe in approximately 13 days (Uno et al 2009). During the long-range transport, dust
74 particles may mix with anthropogenic gas and particles air pollutants and, consequently undergo
75 resulting in complicated chemical reactions (Cui et al., 2009; Li et al., 2014; Ma et al., 2012; Wang et
76 al., 2011; Wang et al., 2016, Wang et al., 2017a; Xu et al., 2014; Li et al., 2014; Ma et al., 2012; Yang
77 et al., 2002). However, the extent of these chemical reactions varies widely and depends on the
78 meteorological conditions, such as cloud fraction, wind speed, relative humidity and atmospheric
79 circulation (Yang et al., 2002; Li et al., 2014; Ma et al., 2012; Yang et al., 2002). For example, a few
80 studies have shown that the concentrations of atmospheric particulate NO₃⁻ and NH₄⁺ on dust storm
81 days were 2-5 times larger than those prior to the events in Beijing (Liu et al., 2014; Liu and Bei, 2016).
82 Xu et al. (2014) also reported that concentrations of particulate SO₄²⁻ and NO₃⁻ simultaneously
83 increased during dust storm events along the northern boundary of the Tibetan Plateau. Fitzgerald et al.
84 (2015) found that almost all Asian dust events observed in Korea contained considerable amounts of
85 nitrate and proposed that the dust from the Gobi and Taklamakan Deserts probably mixed and reacted
86 with anthropogenic air pollutants during transport over the Asian continent. However, Although
87 increased concentrations of NO₃⁻ and NH₄⁺ in aerosol particles were observed on dust storm days in
88 northern China relative to those non dust days prior to the dust storm events, Zhang et al. (2010a)
89 reported an interesting result, i.e., also found that the concentrations of the two species were associated

90 with the intensity of the dust storm, i.e., the stronger dust storms corresponded to the smaller increases
91 in these ions. In other words, lower NO_3^- and NH_4^+ concentrations occurred during strong dust storm
92 events than during weak dust events (Zhang et al., 2010a). This raises the complex for carrying reactive
93 nitrogen by dust particles.

94 A few contradictory results were also reported in the literature, which made the scientific issue
95 even more complicated. On the other hand, some studies reported the reverse result. For example, at
96 Yulin, a rural site near the Asian dust source region, the concentration of NO_3^- in atmospheric aerosols
97 on dust days was significantly lower in comparison to the concentration measured immediately before
98 or after the event, as a result of the dilution effect at a rural site at-in Yulin near the Asian dust source
99 region (Wang et al., 2016). The phenomenon was also observed in Shanghai, a mega city at a few
100 thousands of kilometers from dust source zones in China, and more downwind sites. Even in Shanghai,
101 a mega city located at a few thousands kilometers from dust source zones in China, the concentrations
102 of NO_3^- and NH_4^+ were notably lower in the observed dust plumes than in a polluted air parcel observed
103 immediately prior to the dust events (Kang et al., 2013; Li et al., 2014; Wang et al., 2013). Li et al.
104 (2014) also found that the concentrations of nitrate and ammonium decreased on dust storm days with a
105 decreasing ratio of the total soluble inorganic ions to $\text{PM}_{2.5}$ in the Yellow River Delta, China. When
106 dust was rapidly transported from desert regions without passing through major urban areas and
107 lingering over the Yellow Sea, the concentrations and size distributions of nitrate and ammonium had
108 no significant variation in heavy Asian dust (AD) plumes (Kang et al., 2013).

109 To update and improve our knowledge on reactive nitrogen carried by dust particles, the
110 contradictory results highlight the importance of investigating the concentrations of ammonium and
111 nitrate in atmospheric particles during dust events based on a larger database. In this study, we
112 collected atmospheric aerosol particles during and prior to (or post, but after, only when no sample
113 was collected prior to) dust events) at a coastal site adjacent to the Yellow Sea during the in each spring
114 from of 2008 to 2011 when smaller outbreak peaks of dust storms occurred. We measured (the
115 concentrations of inorganic nitrogen in the samples as well as and other components were
116 determined determined for to facilitate our analysis. In this study, we first characterized the
117 concentrations of inorganic nitrogen concentrations in various dust samples by comparing them with
118 events relative to the concentrations the values in atmospheric particles in samples collected measured
119 either prior to or post after the events the event. We then conducted source apportionment to quantify

120 their sources. Finally, we calculated ~~and discussed~~ the deposition flux of atmospheric particulate
121 inorganic nitrogen during dust events ~~and compared the results with the values in the literature in order~~
122 ~~to update the flux values due to dynamic changes in anthropogenic emissions and other factors.~~

123 2 Experimental methods

124 2.1 Sampling

125 Fig. 1 shows the sampling site, which is situated at the top of a coastal hill (Baguanshan) in Qingdao
126 in northern China (36° 6' N, 120° 19' E, 77 m above sea level) and is approximately 1.0 km from the
127 Yellow Sea to the east. A high-volume air sampler (Model KC-1000, Qingdao Laoshan Electronic
128 Instrument Complex Co., Ltd., [China](#)) was set up on the roof of a two-story office building to collect
129 total suspended particle (TSP) samples on quartz microfiber filters (Whatman QM-A) at a flow rate of
130 1 m³/min. Prior to ~~the~~ sampling, the filters were heated at 450 °C for 4.5 hrs to remove organic
131 compounds. Our sample collection strategy involved collecting dust samples representing long-range
132 transported particles. We followed the definition of dust events adopted in the regulations of surface
133 meteorological observations of China (CMA, 2004; Wang et al., 2008) and identified dust events based
134 on the meteorological records (Weather Phenomenon) of Qingdao from [the](#) Meteorological Information
135 Comprehensive Analysis and Process System (MICAPS) of the China Meteorological Administration.
136 ~~Due to no dust events without enhanceive-source lasting over 12 hrs (Lee et al., 2015; Su et al., 2017;~~
137 ~~Zhang et al., 2007), we collected only one dust sample with a 4-hr duration in a day was collected for~~
138 ~~dust events with durations of less than one day. And the sampling for dust particles started only when~~
139 ~~the measured PM₁₀ mass concentration observed in Qingdao on the website~~
140 ~~(<http://www.qepb.gov.cn/m2/>) and the forecasted dust mass over Asia~~
141 ~~(<http://www-cfors.nies.go.jp/~cfors/>) had greatly increased.~~

142 ~~On 20-21 March 20-21, 2010, two dust events subsequently swept Qingdao. The 4- hr dust samples~~
143 ~~with IDs of 20100320 and 20100321 may not capture the entirety of the two whole events. However,~~
144 ~~the on-line data can allow adequate well separating separation of these separating two dust event samples.~~
145 ~~The same was true for the dust samples with IDs of For dust events with durations greater than one~~
146 ~~day due to supplementing from enhanceive source, such as 20100320, 20100321 and 20110501,~~
147 ~~20110502. a 4 hr dust sample was collected once per day according to temporal variation of PM10~~

148 ~~concentration. This can be clearly identified from the on-line data of PM₁₀ concentration. Each dust~~
149 ~~sample was collected over 4 hrs, and the sampling started only when the PM₁₀ and dust mass~~
150 ~~concentration available on the website~~
151 ~~(<http://www.cfors.nies.go.jp/~cfors/>;<http://www.qepb.gov.cn/m2/>) had increased greatly. 4 hrs~~
152 ~~samplings may not full represent these dust event samples since some events lasted over 4 hrs, and in~~
153 ~~fact because off-line samplings are practically hard to exactly cover the whole dust event period.~~
154 ~~However, this approach made the dust sample more representative relative to urban background under~~
155 ~~off-line samplings. This approach made the dust sample more representative relative to urban~~
156 ~~background. For dust events with durations of less than one day, only one sample was collected. For~~
157 ~~dust events with durations greater than one day, a 4 hr dust sample was collected once per day.~~ Table 1
158 lists the sampling information. Based on the forecast, we also collected aerosol particle samples
159 immediately before, ~~which were regarded as the reference samples~~ or after the dust event for
160 ~~comparison.~~ These ~~reference comparison~~ samples were further classified into sunny day samples ~~and~~
161 cloudy day samples. ~~For those events missing sampling prior to dust events, we collected and~~
162 ~~post-dust samples. The post-dust samples were collected under clear and sunny weather conditions~~
163 ~~and low PM₁₀ mass concentrations as early as possible.~~

164 Since Asian dust events ~~were~~ ~~at the sampling site~~ mostly ~~observed~~ ~~occur~~ in the spring at the
165 ~~sampling site.~~ Our intensive samplings were concentrated in the period of ~~we collected samples during~~
166 ~~every spring, i.e., from March to May in from 2008-2011, where when a smaller outbreak peak~~
167 ~~for~~ Asian dust ~~events~~ was observed in ~~northern eastern~~ China in 2008-2011, which followed a larger
168 ~~peak in 2000-2003 during this century~~ (Fig. 6S5). Overall, a total of 14 sets of dust samples and 8 sets
169 of comparison samples were ~~available~~ ~~available for analysis~~ ~~collected for in~~ this study.

170 To facilitate the coastal sampling data analysis, sand samples were collected ~~at the remote site at of~~
171 ~~the~~ Zhurihe ~~site~~ (42°22'N, 112°58'E) in the Hunshandake Desert, one of ~~the~~ main Chinese sand deserts,
172 in April 2012. ~~—~~ Sand samples were packed in clean plastic sample bags and were stored below -20 °C
173 before the transfer. An ice-box was used to store the samples during transport to the lab for chemical
174 analysis.

175 2.2 Analysis

176 The aerosol samples were ~~weighted according to the standard standard protocol~~ ~~protocol~~ ~~allowed to~~

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177 ~~achieve equilibrium in an air conditioned chamber at a constant relative humidity and temperature for~~
178 ~~24hrs before weighing.~~ The sample membranes were then cut into several portions for analysis. One
179 portion of each aerosol sample was ultrasonically extracted with ultra-pure water in an ice water bath
180 ~~and the concentration of for determining~~ inorganic water-soluble ions ~~was determined~~ using ICS-3000
181 ion chromatography (Qi et al., 2011). The sand samples collected at the Zhurihe site were analyzed
182 using the same procedure. We ~~later~~ refer to dissolved inorganic nitrogen (DIN) ~~as~~ the sum of nitrate
183 and ammonium ~~later, in the later discussion, by excluding nitrite because of its due to the~~ very low
184 concentration ~~of nitrite in the samples.~~

185 One portion of ~~the~~ each aerosol filter was cut into 60 cm² pieces and digested with
186 HNO₃+HClO₄+HF (5:2:2 by volume) at 160 °C using an electric heating plate. ~~A blank membrane~~
187 ~~was also analyzed using the same procedure to ensure analytical precision.~~ The concentrations of Cu,
188 Zn, Cr, Sc and Pb were measured using inductively coupled plasma mass spectrometry (Thermo X
189 Series 2), while the concentrations of Al, Ca, Fe, Na and Mg were measured using inductively coupled
190 plasma atomic emission spectroscopy (IRIS Intrepid II XSP). ~~Field bThe membrane blanks have been~~
191 ~~corrected for in the calculation of the metal concentrations. A blank membranes wasere also analyzed~~
192 ~~for correction using the same procedure to ensure analytical precision. T~~

193 One portion of ~~each~~ aerosol sample was digested with an HNO₃ solution (10% HNO₃, 1.6 M) at
194 160 °C for 20 min in a microwave digestion system (CEM, U.S.). The ~~Hg~~ and As in sample extracts
195 were analyzed following the U.S. Environmental Protection Agency method 1631E (U.S. EPA, 2002)
196 using cold vapor atomic fluorescence spectrometry (CVAFS). The detection limits, precisions and
197 recoveries of water-soluble ions and metal elements are listed in Table 2.

198 2.3 Computational modeling

199 The enrichment factor of metal elements ~~was~~ given by

$$200 \quad EF_i = \frac{(X_i/X_{Re})_{aerosols}}{(X_i/X_{Re})_{crust}} \quad (1)$$

201 where subscripts *i* and *Re* refer to the studied metal and the reference metal, ~~respectively~~; $(X_i/X_{Re})_{aerosols}$
202 is the concentration ratio of metal *i* to metal *Re* in the aerosol samples; and $(X_i/X_{Re})_{crust}$ is the ratio of
203 metal *i* to metal *Re* in the Earth's crust. For the calculation of the enrichment factor of the metal
204 elements, scandium was used as the reference element (Han et al., 2012), and the abundance of

205 elements in the Earth's crust given by Taylor (1964) was adopted.

206 The 72-h air mass back trajectories were calculated for each TSP sample using TrajStat software
207 (Wang et al., 2009) and ~~the~~ NOAA GDAS ([Global Data Assimilation System](http://www.arl.noaa.gov/ready/hysplit4.html)) archive data ([http://](http://www.arl.noaa.gov/ready/hysplit4.html)
208 www.arl.noaa.gov/ready/hysplit4.html). The air mass back trajectories were calculated at an altitude of
209 1500 m to identify the dust origin. In addition, the distance over sea of the air mass for each sample
210 was measured from the trajectory using TrajStat software (Wang et al., 2009).

211 The positive matrix factorization (PMF) is a commonly used receptor modeling method. This model
212 can quantify the contribution of sources to samples based on the composition or fingerprints of the
213 sources (Paatero and Tapper, 1993; Paatero, 1997). The measured composition data can be represented
214 by a matrix X of i by j dimensions, in which i number of samples and j chemical species were
215 measured, with uncertainty u . X can be factorized as a source profile matrix (F) with the number of
216 source factors (p) and a contribution matrix (G) of each source factor to each individual sample, as
217 shown in Equation 2.

$$218 X_{ij} = \sum_{k=1}^p G_{ik} F_{kj} + E_{ij} \quad (2)$$

219 where E_{ij} is the residual for species j of the i -th sample.

220 The aim of the model is to minimize ~~a~~ the objective function Q , which was calculated from the
221 residual and uncertainty of all samples (Equation 3), to obtain the most optimal factor contributions and
222 profiles.

$$223 Q = \sum_{i=1}^n \sum_{j=1}^m (E_{ij}/u_{ij})^2 \quad (3)$$

224 The EPA PMF 3.0 model was used to obtain the source apportionment of atmospheric particulates on
225 dust and comparison days. The correlation coefficient between the predicted and observed
226 concentrations was 0.97.

227 Dry deposition velocities were obtained using Williams' model (Williams, 1982) by accounting for
228 particle growth (Qi et al., 2005). Williams' model is a two-layer model used to calculate the dry
229 velocity of size-segregated particles over the water. In an upper layer below a reference height (10 m),
230 the deposition of aerosols particles is governed by turbulent transfer and gravitational settling. In the
231 deposition layer, the gravitational settling of particles is affected by particle growth due to high relative
232 humidity. To obtain the deposition velocity of different particle sizes, Williams' model needs many
233 input parameters, such as the wind speed at ~~10-10-~~10-m height (U_{10}), air/water temperature, and relative
234 humidity. Relative humidity, air temperature and U_{10} ~~obtained~~ from the National Centers for

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235 Environmental Prediction (NCEP) were used in this study. Surface seawater temperature data was
236 collected from the European Centre for Medium-Range Weather Forecasts (ECMWF). The
237 meteorological and seawater temperature data had a six-hour resolution. According to a previously
238 reported method (Qi et al., 2013), the dry deposition fluxes of the particles and the nitrogen species
239 were calculated for dust and comparison days.

240 The CMAQ model (v5.0.2) was applied over the East Asia area to simulate the concentrations of
241 PM₁₀, NO_x and NH₃ for 14 samples collected during 11 dust events. The simulated domain contains
242 164×97 grid cells with a ~~36-36~~-km spatial resolution, and the centered point was 110°E, 34°N. The
243 vertical resolution includes 14 layers from the surface to the tropopause, with the first model layer at a
244 height of 36_m above the ground level. The meteorological fields were generated by the Weather
245 Research and Forecasting (WRF) Model (v3.7). Considering that the simulated area is connected to the
246 Yellow Sea, the CB05Cl chemical mechanism was chosen to simulate the gas-phase chemistry. The
247 emissions of NO_x and NH₃ over East Asia for each dust event were also modeled using the CMAQ
248 model according to the emission inventory in 2008, which was generated by extrapolating the 2006
249 activity data to the year 2008 using the method described by Zhang et al. (2009). Initial conditions
250 (ICONS) and boundary conditions were generated from a global chemistry model of GEOS-CHEM. All
251 the dust events simulations are performed separately, each with a 1-week spin-up period to minimize
252 the influence of the ICONs. ~~The Validation-validation~~ of the application of the CMAQ model in China
253 has been reported by Liu et al. (2010a, b). ~~The spatial distribution of PM10 concentrations for each~~
254 ~~dust event was consistent with the model results of dust by the Chemical Weather Forecast System~~
255 ~~(CFORS) by Uno et al. (2003).~~

256 2.4 Other data sources and statistical analysis

257 Meteorological data were obtained from the Qingdao Meteorological Administration
258 (<http://qdqx.qingdao.gov.cn/zdz/ystj.aspx>) and the MICAPS of the Meteorological Administration of
259 China. Different weather characteristics, such as sunny days, cloudy days and dust days, were defined
260 according to information from the MICAPS and Qingdao Meteorological Administration. According to
261 the altitude, longitude and latitude of the ~~72-72~~-hr air mass back trajectory of each dust sample, the
262 pressure level, temperature and relative humidity (RH) data along the path of the air mass were derived
263 from the NCEP/NCAR re-analysis system

264 (http://www.esrl.noaa.gov/psd/data/gridded/data.ncep.reanalysis.html) for each sample. The mixed
265 layer depth during the air mass transport of dust samples was obtained from the HYSPLIT Trajectory
266 Model (http://ready.arl.noaa.gov/hypub-bin/trajsrc.pl) using the same method. Then the average
267 mixing layer, transport altitude, air temperature and RH were calculated as an average of all points
268 during the air mass back trajectory of each sample. Spearman correlation analysis was applied to
269 examine the relationships of nitrate and ammonium with transport parameters, and P values of <0.05
270 were considered to be statistically significant.

271 **3 Results and discussion**

272 **3.1 Characterization of aerosol samples collected during dust events**

273 ~~To support our analysis, the spatial distributions of PM₁₀ concentration over East Asia for each dust~~
274 ~~event were modeled by WRF CMAQ model (Fig. S3) are shown in Fig. 2 and Fig. S3. From Fig. S 3,~~
275 ~~we can found that almost all dust events originated in northern or northwestern China and passed over~~
276 ~~the sampling site and then transported eastward to the China Sea and further other ocean region.~~
277 ~~However, Sample 20110415 was judged to be local blowing dust because no corresponding high dust~~
278 ~~concentrations were observed in the dust source areas; this sample was therefore excluded from further~~
279 ~~analysis. All these dust events, affected observed in Qingdao, had a characteristics a characteristic of~~
280 ~~large influence range and relatively strong intensity. Though the dust intensity was judged by visibility~~
281 ~~in China (CMA, 2004), the PM₁₀ concentration during a dust event can also show the dust intensity~~
282 ~~indirectly. Generally, strong dust storm had a high PM₁₀ value, such as 2010320, consistent with the~~
283 ~~dust records (CMA, 2009). The dust intensity and influence range of these dust events varied greatly,~~
284 ~~depending on the dust event. The different dust events were expected to had different impact on the~~
285 ~~composition of aerosols in downwind area intensity and influence range of the dust events modeled by~~
286 ~~Uno et al. (2003) were analyzed, and the spatial distributions of dust concentration over East Asia~~
287 ~~during each dust sampling day are shown in Fig. 2 and Fig. S3. Almost all dust events originated in~~
288 ~~northern or northwestern China and passed over the sampling site. However, Sample 20110415 was~~
289 ~~judged to be local blowing dust because no corresponding high dust concentrations were observed in~~
290 ~~the dust source areas; this sample was therefore excluded from further analysis.~~
291 ~~Before characterizing the inorganic nitrogen in atmospheric particles from the Baguanshan site, We~~

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292 first examined the mass concentrations of TSP samples and the concentrations of crustal and
293 anthropogenic metals ~~therein through a comparison with the reference to compare samples~~
294 ~~reference~~ samples collected on dust days and immediately before or after days, providing the background
295 information for our target species analyzed later. The comparative results are highlighted below. For
296 these ~~reference comparison~~ samples, the TSP mass concentrations ranged from 94 to 275 $\mu\text{g}\cdot\text{m}^{-3}$, with
297 an average of 201 $\mu\text{g}\cdot\text{m}^{-3}$ (Fig. 2, Table S1). The TSP mass concentration increased substantially to
298 ~~501410~~-3857 $\mu\text{g}\cdot\text{m}^{-3}$ in dust day samples, with an average of 1140.~~2-3~~ $\mu\text{g}\cdot\text{m}^{-3}$. In each ~~dust~~
299 ~~day comparison day sample~~ pair of dust day sample ~~as against~~ reference sample, the net increase in
300 the mass concentration of TSPs increased by was 8082-1303%, with a median value of ~~537~~403% (Table
301 S1). A similar increase was present in the crustal elements in each pair of samples. For example, the
302 mean concentrations of Sc, Al, Fe, ~~Mg Ca~~ and ~~nss-Ca~~ (usually used as a typical dust index) ~~Mg~~
303 increased by ~~over more than~~ a factor of ~~four two~~ in dust day samples relative to comparison samples.
304 On the other hand~~In addition~~, the enrichment factors (EF) of Al, Fe, Ca, and Mg were less than three in
305 dust day samples with ~~but were~~ the values ~~less than 14 in the reference comparison samples~~ (Table
306 3). Lower values are indicative of elements from ~~the a~~ primarily crustal origin. ~~Although t~~The average
307 mass concentrations of anthropogenic elements, such as Cu, Pb, Zn, Cr, Hg and As, in dust day samples
308 increased by 107% to 722% ~~relative to against t~~ those in ~~comparison the reference~~ samples; ~~however~~,
309 the EF of the anthropogenic metal elements ~~however~~ decreased in the ~~former~~ dust day samples. ~~This~~
310 ~~pattern indicates a decreasing relative contribution of anthropogenic sources to the total TSP mass in~~
311 ~~dust day samples. This indicates that dust particles likely carried more anthropogenic elements,~~
312 ~~although their relative contribution to the total mass was lower than that in the reference sample. Note d~~
313 ~~that a~~ Sample 20110415 was excluded for further analysis ~~as aforementioned~~. It was judged as a local
314 blowing dust event because of no corresponding dust event existed upwind.

3.2 Concentrations of inorganic nitrogen in dust day samples

316 When the mass concentrations of NH_4^+ and NO_3^- in each pair of TSP samples were compared, the
317 concentrations of NH_4^+ increased by ~~458%~~487473% in some dust day samples (20080301, 20080315,
318 20090316, 20100315, 20100320, ~~20100315~~20100321, ~~20110415~~ 20110418 and 20110502), but
319 decreased by 28-84% in other dust day samples (Fig. 3, Column NH_4^+ and NO_3^- in Table S1 ~~Table S1~~).
320 The same was generally true for the measured concentrations of NO_3^- .

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321 Considering the relative values of NH_4^+ and NO_3^- in dust day samples relative to the reference
322 ~~comparison~~-samples, we classified the dust day samples ~~can be classified~~ into three categories (Table 4).
323 In Category 1, the mass concentrations of NH_4^+ and NO_3^- were larger in dust day samples ~~agaist~~against
324 ~~than in the reference comparison~~-samples. In Category 2, the reverse was true. In Category 3, the mass
325 concentrations of NO_3^- were lower in the dust samples than in the ~~reference comparison~~-samples,
326 whereas the ~~concentrations of reverse was true for~~ NH_4^+ ~~wasere~~ close to the ~~reference comparison ones~~.
327 As reported, ~~Considering that~~ the Yellow Sea ~~encountered~~encountered ~~was mainly affected by~~ dust
328 storms mainly derived from the Hunshandake Desert (Zhang and Gao, 2007). ~~W~~-we thereby compared
329 our observations with the sand particles collected from this desert (Table 5). The ~~ratios relative mass~~
330 ~~concentof mass concentrations ations~~ of nitrate and ammonium to the total mass of sand particles were
331 very low, i.e., less than $81 \mu\text{g/g}$ ~~due to little anthropogenic influence expected~~, which are approximately
332 three orders of magnitude less than the corresponding values in our dust samples. ~~T~~Moreover, ~~the~~
333 values obtained from atmospheric aerosols at ~~an the urban sites of~~ Duolun (Cui, 2009) and Alxa Right
334 Banner (Niu and Zhang, 2000), ~~which are closer to the desert~~, ~~increased on dust days~~increased, but
335 were ~~still also more than over~~ one order of magnitude lower than the corresponding values in this study
336 (Table 5). ~~This suggested that~~ NO_3^- and NH_4^+ ~~observed in the dust day samples were very likely due to~~
337 The mixing and chemical interactions ~~and mixing~~ between anthropogenic air pollutants and dust
338 particles during ~~the~~ transport from the source zone to the reception site likely played an important role
339 ~~to in increaseing~~ the ratios, leading to ~~an extremely larger ratio values at this site relative to those in~~
340 source dust and in upwind atmospheric particles (Cui et al., 2009, Wang et al., 2011; Wu et al., 2016).
341 However, the increase or decrease in the mass concentration absolute of nitrate and ammonium in
342 different dust samples ~~agaist~~against the reference implied the complex ~~for the interactions along the~~
343 ~~different transport paths of Asian dust, air pollutant emissions, meteorological conditions, chemical~~
344 ~~reactions, and other factors can affect the abundance of~~ NH_4^+ and NO_3^- ~~in atmospheric particles. These~~
345 ~~factors can vary greatly among different dust events, hence leading to the three different categories.~~

346 4. Discussion and conclusion

347 34.13 Theoretical analysis of the three categories

348 Ammonium salts are common in atmospheric particles with diameters of less than $2 \mu\text{m}$ (Yao et al.,
349 2003; Yao and Zhang, 2012). Many modeling studies have shown that the Gasgas-aerosol

350 thermodynamic equilibrium is ~~widely~~ assumed to be fully attained for inorganic ions, including
351 ammonium salts in PM_{2.5} (Dentener et al., 1996; Underwood et al., 2001; Wang et al., 2017a; Zhang et
352 al., 1994; Zhang and Carmichael, 1999), ~~in all regional air quality modeling studies~~. Reasonably good
353 agreements between ammonium salt modeling results and observations reported in the literature
354 support the validity of this assumption (Chen et al., 2016; Penrod et al., 2014; Walker et al., 2012).
355 ~~Supposed~~ Assuming that a thermodynamic equilibrium had been attained by the ammonium salts in
356 Category 1, the reactions between carbonate salts and ammonium salts, such as 1) (NH₄)₂SO₄+
357 CaCO₃⇒ CaSO₄ + NH₃ (gas) +CO₂ (gas) +H₂O and 2) 2NH₄NO₃+ CaCO₃⇒Ca(NO₃)₂ + 2NH₃ (gas)
358 +CO₂ (gas) +H₂O, will release NH₃ (gas) until CaCO₃ has been completely used up. During dust events,
359 ~~much-very~~ high concentrations of Ca²⁺ were observed, and high CaCO₃ concentrations were therefore
360 expected. For example, the single-particle characterization showed that Asia dust from the Gobi and
361 the Inner Mongolian Deserts had rich CaCO₃, with a ratio of 4.3-6.7% for reacted CaCO₃ and
362 3.0-4.6% for unreacted CaCO₃ (Hwang et al., 2008). Heterogeneous chemical reactions of mineral
363 dust mostly occurred on CaCO₃ mineral dust (Hwang and Ro, 2006). However, when Category 1
364 was considered alone and one exterior sample was excluded, a good correlation ~~however,~~ was obtained
365 for [NH₄⁺]_{equivalent concentration} = -0.98*[NO₃⁻+SO₄²⁻]_{equivalent concentration} (R²=0.83, P<0.05). The good correlation,
366 together with the slope of 1, strongly indicated that the NO₃⁻ and SO₄²⁻ were almost completely
367 associated with NH₄⁺ in these dust day samples. It was commonly believed ~~generally accepted that~~
368 anthropogenic ammonium nitrate and ammonium sulfate were produced by gas, aqueous phase reaction
369 and thermodynamic equilibrium processes, and these anthropogenic ammonium nitrate and
370 ammonium sulfate were sulfate was externally mixed with dust particles (Wang et al., 2017a). In
371 reverse ~~reverse~~. And the poor correlation of Ca²⁺ to NO₃⁻ and SO₄²⁻ showed that the formation of
372 CaSO₄ and/or Ca(NO₃)₂ was probably negligible. Thus, ammonium salt aerosols may externally
373 ~~co~~ exist with dust aerosols in these dust day samples. Wang et al. (2017a) also found that coarse mode
374 ammonium was quite low and fine mode dust particles were externally mixed with anthropogenic
375 ammonium nitrate and ammonium sulfate. The results on Asia dust samples verified that NO₃⁻ and
376 NH₄⁺ in Asia dust samples were argued to be verified that more physically affected by the dust
377 storm for Asia dust samples, i.e., the dilution effect, rather than the chemical reaction on the dust
378 (Huang et al., 2010). The hypothesis appeared to be valid Note that only samples in Category 1, where
379 showed NH₄⁺ to be negatively correlated with Ca²⁺ (Fig.S4). In the exterior sample collected on

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380 | 21 March 2010, $[\text{NH}_4^+]$ only accounted for ~70% of the observed $[\text{NO}_3^- + \text{SO}_4^{2-}]$ in an equivalent
381 | concentration. This result suggested that ~30% of $(\text{NO}_3^- + \text{SO}_4^{2-})$ may be associated with dust aerosols
382 | via the formation of metal salts of the two species. ~~The~~ This hypothesis was supported by the
383 | correlation result, i.e., NO_3^- was positively correlated with NH_4^+ and Cu, and SO_4^{2-} was correlated with
384 | K^+ , Na^+ and Mg^{2+} (Fig.S4). Cu was once was generally used as an effective markers of vehicular
385 | emission sources from diesel and biodiesel-blend exhaust (Gangwar et al., 2012), while it also can also
386 | be and was also derived from copper pyrites (CuFeS_2) in Inner Mongolia mines (Huang et al., 2010).
387 | The increase of Cu in the mass concentration in dust samples implied ~~indicating~~ dust particles mixed
388 | with the anthropogenic particles, particularly from industrial industrial emissions, during transport. Note
389 | that only samples in Category 1 showed NH_4^+ to be negatively correlated with Ca^{2+} (Fig.S4).

390 | For Category 2, no correlation between $[\text{NH}_4^+]_{\text{equivalent}}$ concentration and $[\text{NO}_3^- + \text{SO}_4^{2-}]_{\text{equivalent}}$ concentration
391 | existed. When Category 2 was considered alone and one exterior sample was excluded, the equivalent
392 | ratios of NH_4^+ to $\text{NO}_3^- + \text{SO}_4^{2-}$ were generally much smaller than 1, suggesting that a larger fraction of
393 | $\text{NO}_3^- + \text{SO}_4^{2-}$ may exist as metal salts due to reactions of their precursors with dust aerosols. NO_3^- and
394 | SO_4^{2-} showed no correlations with NH_4^+ but did show significant correlations with Pb (Fig.S4). ~~an~~
395 | effective marker of vehicular emission sources, implying that $\text{NO}_3^- + \text{SO}_4^{2-}$ existed as metal salts. The
396 | average concentration of Ca^{2+} in Category 2 ($0.43 \pm 0.40 \mu\text{g}/\text{m}^3$) was ~~evidently~~ evidently ~~clearly~~ higher
397 | than that in Category 1 (Ca^{2+} : $0.17 \pm 0.04 \mu\text{g}/\text{m}^3$), implying the probable formation of CaSO_4 and/or
398 | $\text{Ca}(\text{NO}_3)_2$ and the release of NH_3 (gas), ~~resulting in a decrease in NH_4^+ .~~ Moreover, e However, ~~the~~
399 | concentration of total Ca was $1.11 \pm 0.70 \mu\text{g}/\text{m}^3$ in Category 1 and $0.74 \pm 0.49 \mu\text{g}/\text{m}^3$ in Category 2.
400 | Except for 20080502, the remaining dust rest samples in Category 2 ~~were~~ transported from the
401 | middle desert in China carrying dust relatively enriched with CaCO_3 (1-25% in Wt%) (Formenti et al.,
402 | 2011). A positive correlation between In Category 1, NO_3^- was negatively correlated with SO_4^{2-} (Fig.S4),
403 | suggesting competition for NH_3 under NH_3 poor dust days during long range transport. However, NO_3^-
404 | was positively correlated with and SO_4^{2-} in Category 2 against a negative correlation in Category 1
405 | also implied that the dust particles enriched with CaCO_3 in Category 2 might play an important role
406 | to form. The latter relationship can be explained by the fact that the amount of CaCO_3 was sufficient to
407 | absorb the precursors of both SO_4^{2-} and NO_3^- in these dust days. ~~Tobo et al. (2010) provide field~~
408 | evidence for the existence of Ca-rich dust particles coated with highly soluble nitrate were observed at

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409 Kanazawa in Japan during Asian dust storm periods using SEM/EDX (scanning electron microscopy
 410 equipped with an energy dispersive X-ray spectrometer) ~~Tobo et al.~~ (Tobo et al., 2010). The
 411 single-particle observation conducted by Hwang and Ro (2006) ~~characterization~~ showed that CaCO₃ in
 412 dust species particles was almost completely consumed ~~reacted~~ to produce ~~mostly~~ mainly Ca(NO₃)₂
 413 species (Hwang and Ro, 2006). Due to the absence of TSP concentration data along the transport
 414 pathway, we compared TSP concentrations at the sampling site and found that the average value of
 415 Category 2 (1391±981 µg/m³) was substantially higher than that of Category 1 (591±158 µg/m³). This
 416 implied that dust events in Category 2 were even stronger. ~~Note that the NO₂ concentrations in~~
 417 Category 2 (1.35±2.45 µg/m³) were lower or comparable to those in Category 1 (1.51±2.16 µg/m³).
 418 The potential formation of nitrate metal salts was expected to be similar between the two categories,
 419 while ~~unfavorable~~ formation conditions for ammonium nitrate ~~greatly increased~~ decreased the mass
 420 concentrations of nitrate and the contributions to the TSPs in Category 1. ~~Note that the NO₂~~
 421 concentrations in Category 2 (1.35±2.45 µg/m³) were lower or comparable to those in Category 1
 422 (1.51±2.16 µg/m³).

423 Overall, the higher ammonium concentrations observed in the dust day samples in Category 1 were
 424 likely associated with external co-existence of ammonium salt aerosols and dust particles in some dust
 425 days. However, the lower concentrations in Category 2 were likely due to unfavorable conditions for
 426 forming ammonium salts in some dust days. The observed ammonium was just the residual of
 427 incomplete reactions between preexisting ammonium salt and carbonate salts ~~depending on dust event~~
 428 atmospheric chemical state, etc. More discussion on this issue will be presented in Section 3.4.

429 3.4 Influence of transport pathways on particulate inorganic nitrogen dust samples

430 The calculated air mass trajectories of 13 out of 14 samples showed that the air mass originated from
 431 north and Inner Mongolia, China (Fig. 54), generally consistent with the results by Zhang and Gao
 432 (2007). The remaining one 20110418, originated from Northeast China. Figs. 6 and 75 (taking emission
 433 in 2008 as an example) show a few areas with high emissions of NO_x and NH₃, e.g., Liaoning,
 434 Beijing Tianjin Hebei, Shandong, Henan and Jiangsu in China. The calculated trajectories showed that
 435 all the air mass passed over parts of these highly polluted regions and experienced different residence
 436 time in these regions. In Fig. 54, except for the one exterior sample 20110502, all trajectories in
 437 Category 1 showed that the air masses were transported from either the north or northwest over the

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438 continent. In Category 2, the air masses crossed over the sea for 94–255 km prior to arriving at the
439 reception site. NH_3 poor conditions in the marine atmosphere disfavored the formation and existence of
440 ammonium nitrate. On the other hand, the humid marine conditions (the average RH being ranged in
441 50–75% over the Bohai and Yellow Sea in Qingdao in 2006–2012) might have enhanced
442 hetero-coagulation between dust and smaller anthropogenic particles particle-particle coagulation and
443 might have led to the release of NH_3 via reactions between preexisting ammonium salts and carbonate
444 salts. Tobo et al. (2010) suggested that the conversion of insoluble CaCO_3 to $\text{Ca(NO}_3)_2$ tends to be
445 dominated over urban and industrialized areas of the Asian continent, while relatively moist conditions
446 in the marine boundary layer (usually, $\text{RH} > 60\%$), it is highly likely that the production of CaCl_2
447 exceeds that of $\text{Ca(NO}_3)_2$ by modifying Ca-rich particles in dust storms. Therefore we think the input
448 of marine air during the transport was one reason for the low concentration of $\text{NH}_4^+ + \text{NO}_3^-$ in Category
449 2.

450 Moreover, we also examined the links among the measured concentrations of particulate ammonium
451 and nitrate, the mixing layer along the back trajectories, and the residence time of air masses crossing
452 over the highly polluted zones. The results supported our hypothesis, i.e., ammonium salts mostly
453 co-existed with dust aerosols externally. For example, except for 20080425, most of the time all dust
454 day samples mostly traveled at an altitude above the mixing layer before mixing down to ground level.
455 The transport of dust air masses above the mixing layer reduced the possibility for internal mixing of
456 ammonium salts and reaction with dust aerosols along the long transport path. For most sampling days
457 in Category 1, the average mixing layer was less than 900 m, favoring the trapping of locally emitted
458 anthropogenic air pollutants in the mixing layer. In addition, the air masses in Category 1 at this
459 elevation apparently moved slowly and took over 10–11–39 hr to cross over the highly polluted area.
460 Even lower speeds were expected for air masses at the bottom of the mixing layer, as wind speed
461 decreases with height. Except for exterior samples, the sampling days in Category 2 featured a mixing
462 layer that was higher than 900 m on average (916–1194 m), higher than 900 m. The air masses in this
463 Category at this elevation took less than 10 hr to cross over the highly polluted areas and generally had
464 higher speeds. Theoretically, a lower mixing layer and a lower wind speed favored the accumulation of
465 air pollutants and the formation of ammonium nitrate to some extent. The transport of dust air masses
466 above the mixing layer reduced the possibility for internal mixing of ammonium salts and reaction with
467 dust aerosols along the long transport path. The shorter time for dust air masses mixing down to ground

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468 level before arriving at the reception site and lower wind speed (mean of 2.8m/s at sampling site) also
469 increased the possibility for external co-existence between ammonium salt aerosols and dust aerosols in
470 Category 1. The reverse could be argued to explain the observations for Category 2 (average wind
471 speed being 6.2 m/s at sampling site). The single particle characterization also showed that the Asian
472 dust particles collected in Korea were mixed with sea salts entrained over the Yellow Sea, as well as air
473 pollutants from the eastern China coastal areas for a slow moving, low altitude air mass (Hwang et al.,
474 2008). The correlation analysis results in Table S2 indirectly support these conclusions. In fact,
475 previous studies proposed that nitrate is rarely formed on the surface of dust particles for some dust
476 events (Zhang and Iwasaka, 1999). Therefore, much lower nitrate concentrations were observed in
477 Category 2. Noted that the exterior with ID of 20110415 and 20110502 have not yet been explained.

478 3.5.2 Source apportionment of aerosols during dust and non-dust events

479 The sources of atmospheric aerosols in on-dust and reference samples comparison days were
480 determined by the PMF modeling (Paatero and Tapper, 1993; Paatero, 1997). Fig. 8-6 shows that
481 atmospheric aerosols in the reference samples on comparison days mainly included mainly consisted
482 of six sources, i.e., industry, soil dust, secondary aerosols, sea salt, biomass burning, and coal
483 combustion/other sources, with 90% of the scaled residuals falling between -3 and +3; $r^2=0.97$. In these
484 dust samples, on all dust days including Categories 1-3, the sources of aerosols mainly included
485 oil combustion, industry, soil dust, secondary aerosols, and coal combustion/other sources were
486 identified as five major sources (Table 7). These values are compared in Table 7. The contribution of
487 soil dust evidently increased from 23% to 36% in the dust samples on dust days relative to the
488 reference comparison days, consistent with the high concentrations of TSPs and crustal metals observed
489 on dust days. Liu et al. (2014) also found an even larger increase in the contribution of dust aerosols to
490 PM_{10} , i.e., 31%–40%, on dust days relative to non-dust days. Accordingly, the contributions of local
491 anthropogenic sources decreased on dust days, especially those of secondary aerosols, consistent with
492 EF of anthropogenic metals observed on dust days. The source profile for coal combustion in dust day
493 samples showed a high percentage of K^+ , Cl^- , Ca, Mg, Co, Ni, As, Al and Fe, indicating coal
494 combustion presenting contemporaneously with a mixture of coal combustion and other pollutants
495 emitted along the transport path on dust days. The calculation results also showed that the ed
496 contribution of nitrate plus ammonium from the soil dust source aerosol mass (the sum of nitrate and

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497 ammonium associated with the dust source) to the total aerosol mass (the total of nitrate plus and
498 ammonium) in the dust samples greatly increased on dust days. However, we could not determine the
499 contributions of above sources to aerosols for the different dust Categories because of the limited
500 number of samples. The source profile for coal combustion in the dust day samples showed a high
501 percentage of K⁺, Cl⁻, Ca, Mg, Co, Ni, As, Al and Fe, indicating that coal combustion particles may
502 exist contemporaneously with other anthropogenic pollutants emitted along the
503 transport path. Liu et al. (2014) also found an even larger net increase in the contribution of dust
504 aerosols to the mass of PM₁₀, i.e., 31%-40%, on dust days against relative to non-dust days in Beijing
505 which is approximately 600 km upwind of Qingdao. Accordingly, they reported that the contributions of
506 local anthropogenic sources decreased on dust days, especially those of from secondary aerosols,
507 consistent with the EF of anthropogenic metals observed on dust days.

508 4.3 Influence of transport path ways on particulate inorganic nitrogen in dust samples

509 The calculated air mass trajectories for 13 out of 14 samples showed that the air mass originated from
510 North and Inner Mongolia, China (Fig. 4), generally consistent with the results by Zhang and Gao
511 (2007). The remaining one, with ID of 20110418 originated from Northeast China. The calculated
512 trajectories showed that all the entire dust air mass passed over those highly polluted regions with
513 strong emissions of NO_x and NH₃ shown in Fig 5 and experienced different residence times therein. Fig.
514 4 showed that all air mass trajectories in Category 1 were transported from either the north or
515 northwest over the continent, except for the one exterior sample 20110502. In Category 2, the air
516 masses always took a 94-255 km trip over the sea prior to arriving at the reception site. NH₃-poor
517 conditions in the marine atmosphere disfavored the formation and existence of ammonium nitrate. On
518 the other hand, the humid marine conditions (the average RH ranged in 50-75% over the Bohai and
519 Yellow Seas in 2006-2012) might have enhanced hetero-coagulation between dust and smaller
520 anthropogenic particles, leading to the release of NH₃ via reactions between preexisting ammonium
521 salts and carbonate salts.

522 The average mixing layer was less than 900 m along the air mass transport routes for most sampling
523 days in Category 1 (Table 7), favoring the trapping of locally emitted anthropogenic air pollutants in
524 the mixing layer. The air masses in Category 1 took over 11-39 hrs to cross over the highly polluted
525 area with appreciable concentrations of NO_x (5.7±1.4 ppb) and NH₃ (7.6±3.3 ppb). Except for the

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526 exterior samples, air masses in Category 2 took less than 10 hrs to cross over the polluted areas with
527 lower concentrations of NO_x (3.6 ± 3.4 ppb) and NH_3 (4.7 ± 4.7 ppb) and the mixing layer height along
528 the route was 916-1194 m (on average) for each dust event. Moreover, the averaged wind speed at
529 sampling site was 2.8 m/s in Category 1, but 6.2 m/s in Category 2. The lower wind speed in Category
530 1 was unexpected, implying dust particles very likely traveled at aloft with a high speed and then
531 mixed down to the ground through subsidence. This further led to the external mixing of anthropogenic
532 particulate matters and dust. The correlation analysis results in Table S2 indirectly support these
533 conclusions.

534 The concentrations of PM_{10} and its major components NO_3^- and NH_4^+ over East Asia on dust days and
535 comparison days were modeled using the WRF-CMAQ model (Fig. S3). Spatial distributions of PM_{10}
536 during each dust events were consistent with the records in the "Sand-dust Weather Almanac" (CMA,
537 2009; 2010; 2012; 2013). The dust particles were transported eastward by passing over the sampling
538 site, the China Sea and arriving at the far remote ocean region, except for the local blowing dust sample
539 with ID of 20110415, as mentioned previously. NMB (normalized mean bias) values of NO_3^- were -4%
540 and -12% in dust and non-dust reference samples, respectively, indicating that CMAQ results
541 reasonably reproduce the mass concentrations of NO_3^- (Fig. S4). Simulated NH_4^+ concentrations in dust
542 samples were severely under-predicted with NMB values at -71%. For reference samples, simulated
543 NH_4^+ concentrations sometimes can well reproduce the observational values, but sometimes totally off.
544 The external mixing mechanism proposed in this study is urgently needed to be included in the model
545 for accurately predicting the concentrations during dust events.

546 ~~All these dust events, observed in Qingdao, had a characteristic of large influence range and relatively~~
547 ~~strong intensity. Though the dust intensity was judged by visibility in China (CMA, 2004), the PM_{10}~~
548 ~~concentration during a dust event can also show the dust intensity indirectly. Generally, strong dust~~
549 ~~storm had a high PM_{10} value, such as 2010320, consistent with the dust records (CMA, 2009). The dust~~
550 ~~intensity and influence range of these dust events varied greatly, depending on the dust event. The~~
551 ~~different dust events were expected to had different impact on the composition of aerosols in~~
552 ~~downwind area.~~

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555 Moreover, we also examined the links among the measured concentrations of particulate ammonium
556 and nitrate, the mixing layer along the back trajectories, concentration of NO_x and NH_3 along the back
557 trajectories, and the residence time of air masses crossing over the highly polluted zones. The results
558 supported our hypothesis, i.e., ammonium salts mostly existed with dust aerosols externally. For
559 example, except for 20080425, most of the time all dust day samples traveled at an altitude above the
560 mixing layer before mixing down to ground level. The transport of dust air masses above the mixing
561 layer reduced the possibility for internal mixing of ammonium salts and reaction with dust aerosols
562 along the long transport path. For most sampling days in Category 1, the average mixing layer was less
563 than 900 m with high concentration of NO_x and NH_3 in range of 13,000-48,000 (Table 5), favoring the
564 trapping of locally emitted anthropogenic air pollutants in the mixing layer. In addition, the air masses
565 in Category 1 took over 11-39 hr to cross over the highly polluted area. Except for exterior samples, the
566 sampling days in Category 2 featured a mixing layer that was higher than 900 m on average
567 (916-1194m) and much low concentration of NO_x and NH_3 in range of 26-13,100 (Table 5). The air
568 masses in this Category took less than 10 hr to cross over the highly polluted areas. Theoretically, a
569 lower mixing layer, high concentration of NO_x and NH_3 and a lower wind speed favored the
570 accumulation of air pollutants and the formation of ammonium nitrate to some extent. The shorter
571 residence time over the highly polluted zones for dust air masses mixing down to ground level before
572 arriving at the reception site and lower wind speed (mean of 2.8m/s at sampling site) also increased the
573 possibility for external existence between ammonium salt aerosols and dust aerosols in Category 1. The
574 reverse could be argued to explain the observations for Category 2 (average wind speed being 6.2 m/s
575 at sampling site). The single particle characterization also showed that the Asian dust particles
576 collected in Korea were mixed with sea salts entrained over the Yellow Sea, as well as air pollutants
577 from the eastern China coastal areas for a slow moving, low altitude air mass (Hwang et al., 2008).
578 The correlation analysis results in Table S2 indirectly support these conclusions. In fact, previous
579 studies proposed that nitrate is rarely formed on the surface of dust particles for some dust events
580 (Zhang and Iwasaka, 1999). Therefore, much lower nitrate concentrations were observed in Category 2.
581 Noted that the exteriors with ID of 20110415 and 20110502 have not yet been explained.

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583 34.6.4 Dry deposition fluxes of TSP, particulate inorganic nitrogen and metals

584 Dust events are known to increase the ~~concentrations and~~ deposition fluxes of aerosol particles
585 ~~during long-range transport~~ along the transport path because of high particle loadings. For example, Fu
586 et al. (2014) found that the long-range transported dust particles increased the dry deposition of PM₁₀ in
587 the Yangtze River Delta region by a factor of approximately 20. In terms of atmospheric deposition in
588 the oceans, ~~a few some~~ studies reported enhancements in oceanic chlorophyll *a* following dust storm
589 events (Banerjee and Kumar, 2014; Tan and Wang, 2014; Banerjee and Kumar, 2014). ~~However, the~~
590 ~~deposition fluxes of dust varied greatly among different dust storms, and only a few dust episodes were~~
591 ~~followed by an increase in oceanic chlorophyll *a* (Banerjee and Kumar, 2014)~~. In addition to those in
592 high-nutrient and low-chlorophyll (HNLC) regions, the input of nitrogen and other nutrients associated
593 with dust deposition is expected to promote the growth of phytoplankton in oceans with varying -
594 ~~However, the extent can vary greatly depending on the~~ nutrient limitation conditions ~~in the oceans~~. A
595 ~~similar principle holds for the occurrence or absence of algal blooms following dust events~~. Thus, we
596 calculated the dry deposition fluxes of aerosols particles, N_{NH₄+NO₃}, and metal elements during dust and
597 ~~comparison~~ reference periods using the measured component concentrations and modeled dry
598 deposition velocities (Table 8). We ~~then also~~ compared the calculated dry deposition flux of TSP and
599 N_{NH₄+NO₃} with ~~the~~ previous observations in the literature.

600 The dry deposition fluxes of atmospheric particulates increased on dust days ~~against~~ against the
601 ~~reference~~ relative to comparison days to some extent. All increases or decreases in this section reflected
602 ~~the value on dust days relative to comparison days, if not specified~~. For example, ~~the dry deposition~~
603 ~~flux of TSP was only 2,800±700 mg/m²/month on comparison days in the coastal region of the Yellow~~
604 ~~Sea~~. The particle deposition fluxes varied over a wide range from 5,200 to 65,000 mg/m²/month ~~under~~
605 in different dust sampling days, with an average of 18,453 mg/m²/month, in comparison with ~~the dry~~
606 ~~deposition flux of TSP to be of 2,800±700 mg/m²/month on~~ the reference periods in the coastal
607 ~~region of the Yellow Sea~~. ~~However,~~ The dry deposition fluxes of N_{NH₄+NO₃} varied, depending on
608 Category 1, 2 or 3 ~~did not follow this pattern~~. In Category 1, the dry deposition fluxes of N_{NH₄+NO₃}
609 increased by 9-~~285.75%~~ with, corresponding to the increased in the TSP flux of by 86-252% (Table S3).
610 In Categories 2 and 3, the dry deposition fluxes of TSP increased by 126% to 2,226% against the
611 ~~reference~~ compared to that on comparison days. ~~Excluded~~ except for ammonium in Category 3, the dry
612 deposition fluxes of particulate N_{NH₄+NO₃} ~~however,~~ decreased by ~~4144%~~ (on average). A larger
613 ~~relative~~ decrease in the concentration of nitrate was present found for in Categories 2 and 3 ~~the~~

614 ~~concentration of nitrate~~, i.e., decreases of ~~-73%~~ and 46% ~~in Category 2 and 3~~, respectively. Note that
615 the average ammonium deposition flux decreased by 47% in Category 2 but increased ~~by 10%~~ in
616 Category 3.

617 ~~The dry atmospheric deposition fluxes of Fe increased by a factor of 124-2370% on dust days.~~
618 ~~Atmospheric inputs of iron to the ocean have been proposed to enhance primary production in HNLC~~
619 ~~areas (Jickells et al., 2005). Moreover, e~~Except for Pb and Zn in Category 2, the dry deposition fluxes
620 of Cu, Pb and Zn increased with those of nitrogen ~~and iron~~ on dust days. Trace metals were found to
621 have a toxic effect on marine phytoplankton and inhibit their growth (Bielmyer et al., 2006; Echeveste
622 et al., 2012). Liu et al. (2013) found that ~~this the~~ inhibition coexisted with the promotion of ~~some~~
623 phytoplankton species in incubation experiments ~~in the southern Yellow Sea in the spring of 2011 by~~
624 ~~adding involving the addition of Asian dust samples to collected seawaters in the southern Yellow Sea~~
625 ~~in the spring of 2014. However, the dry atmospheric deposition fluxes of Fe increased by a factor of~~
626 ~~124-2370% in dust day samples. Wang et al. (2017b) recently reported that Fe can alleviate the toxicity~~
627 ~~of heavy metals. Moreover, Atmospheric inputs of iron to the ocean have been widely proposed to~~
628 ~~enhance primary production in HNLC areas (Jickells et al., 2005).~~

629 ~~3.7 Potential impacts of nitrogen dry deposition flux associated with dust influenced by~~ 630 ~~anthropogenic activity~~

631 Due to anthropogenic activity and economic development, inorganic nitrogen emissions increased in
632 China from 1980 to 2010 (Fig. S5). ~~Accordingly, t~~The $N_{NH_4+NO_3}$ dry deposition flux ~~of $N_{NH_4+NO_3}$~~
633 should have theoretically increased with the increase in ~~inorganic nitrogen the~~ emissions of inorganic
634 ~~nitrogen. However, from the limited data shown in Table 9, we did not find the expected increase in dry~~
635 ~~deposition flux of inorganic nitrogen during the dust days. Considering~~ However, from the limited data
636 ~~shown in Table 9, we did not find the expected increase in dry deposition flux of inorganic nitrogen~~
637 ~~during the dust days. Consider~~ ~~ing the different the~~ uncertainty in dry deposition ~~velocity-velocities to~~
638 ~~be used in various studies, we recalculated~~ ~~normalized~~ the dry deposition flux of $N_{NH_4+NO_3}$ ~~in the~~
639 ~~literature using the concentration of nitrate and ammonium reported in the literature and the~~
640 ~~recommended~~ dry deposition ~~velocity-velocities~~ of 1 cm/s for nitrate and 0.1 m/s for ammonium, ~~as in~~
641 ~~coastal areas~~ reported by Duce et al. (1991). We ~~then thereby~~ found that dry deposition fluxes of
642 $N_{NH_4+NO_3}$ over the Yellow Sea during the dust days increased greatly from 1999 to 2007, ~~but the~~

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643 values. The fluxes of $N_{\text{NH}_4+\text{NO}_3}$ in Qingdao, ~~including during the dust days,~~ varied narrowly within
644 a range of 94.75-99.65 mg N/m²/month during the dust days from 1997 to 2011 (Table 8). The
645 complicated results implied that even more updated works are needed in the future. may reflect the
646 combined effects of NO_x and NH₃ emissions in northern China, the occurrence frequency and intensity
647 of dust events and metrological conditions affecting the transport pathways and moving speeds of dust
648 air masses and chemical reactions occurring therein. For example, dust events commonly exhibited a
649 periodic variation from 2000 to 2011 (Fig.S5). However, from the limited data shown in Table 9, we
650 did not find the expected increase in dry deposition flux of inorganic nitrogen during the dust days.

651 4 Conclusion

652 The concentrations of nitrate and ammonium in TSP samples varied greatly from event to event on dust
653 days. Relative to non dust day samples, the concentrations were both higher in some cases and lower in
654 others. The observed ammonium in dust day samples was explained by ammonium salt aerosols
655 co-existing externally with dust aerosols or the residual of incomplete reactions between ammonium
656 salts and carbonate salts. NO₃⁻ in the dust day samples was partially related to mixing and reactions
657 between anthropogenic air pollutants and dust particles during the transport from the source zone to the
658 reception site. However, this process was generally much less effective and led to a sharp decrease in
659 nitrate in Category 2 TSP samples. The external co-existence of ammonium salt aerosols with dust
660 aerosols and the extent of the reactions between ammonium salts and carbonate salts were apparently
661 associated with the transport pathway, moving speeds and metrological conditions, among other
662 factors.

663 Due to a sharp increase in dust loads on dust days, the contribution of soil dust to the total aerosol
664 mass was higher on dust days than on comparison days, while the contributions from local
665 anthropogenic sources were accordingly lower.

666 Overall, this study strongly suggested that atmospheric deposition of $N_{\text{NH}_4+\text{NO}_3}$ on dust days varied
667 greatly and that no simple linear increase existed with increasing dust load. More observations at
668 various locations are needed to obtain a statistical relationship between dust events and atmospheric
669 deposition of $N_{\text{NH}_4+\text{NO}_3}$. A simple assumption of a linear increase in $N_{\text{NH}_4+\text{NO}_3}$ with increasing dust
670 load, like that in the literature, could lead to considerable overestimation of the dry deposition flux of

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671 ~~nutrients into the oceans and the consequent primary production associated with dust events.~~

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673 *Acknowledgments.* This work was supported by the Department of Science and Technology of the P. R.
674 China through the State Key Basic Research & Development Program under Grant No. 2014CB953701
675 and the National Natural Science Foundation of China (No. 41375143). We thank Prof. Yaqiang Wang
676 and Jinhui Shi for the valuable discussion regarding this research. We also express our appreciation to
677 Tianran Zhang for help with sand sampling, and Qiang Zhang, Yang Yu and Jiuren Lin for data
678 collection.

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Table 1. Sampling information for the aerosol samples collected at the Baguanshan site in the coastal region of the Yellow Sea.

Sampling Year	Sample category	Sampling number	Sampling time	Weather characteristics
2008	Samples on dust days	20080301	From 13:22 a.m. to 17:22 p.m. on Mar. 1st	Floating dust ^a
		20080315	From 13:21 a.m. to 17:21 p.m. on Mar. 15th	Floating dust
		20080425	From 13:14 a.m. to 17:14 p.m. on Apr. 25th	Floating dust
		20080528	From 11:38 a.m. to 15:38 p.m. on May 28th	Floating dust
	Samples on non-dust days	20080529	From 10:15 a.m. to 12:15 p.m. on May 29th ^b	Floating dust
		20080316	From 13:00 a.m. to 17:00 p.m. on Mar. 16th	Sunny day
		20080424	From 13:00 a.m. to 17:00 p.m. on Apr. 24th	Sunny day
		20080522	From 13:00 a.m. to 17:00 p.m. on May 22nd	Cloudy day with mist
2009	Samples on dust days	20090316	From 8:25 a.m. to 12:25 p.m. on Mar. 16th	Floating dust
	Samples on non-dust days	20090306	From 13:00 a.m. to 17:00 p.m. on Mar. 6th	Sunny day
2010	Samples on dust days	20100315	From 11:30 a.m. to 15:30 p.m. on Mar. 16th	Mist after floating dust
		20100320	From 10:30 a.m. to 14:30 p.m. on Mar. 20th	Floating dust
		20100321	From 10:30 a.m. to 14:30 p.m. on Mar. 21st	Floating dust
	Samples on non-dust days	20100324	From 11:30 a.m. to 15:30 p.m. on Mar. 24th	Sunny day
2011	Samples on dust days	20110319	From 12:00 a.m. to 16:00 p.m. on Mar. 19th	Floating dust
		20110415	From 12:00 a.m. to 16:00 p.m. on Apr. 15th	Floating dust
		20110418	From 12:25 a.m. to 16:25 p.m. on Apr. 18th	Floating dust ^c
		20110501	From 12:10 a.m. to 16:10 p.m. on May 1st	Floating dust

		p.m. on May 1st	
	20110502	From 16:00 a.m. to 20:00 p.m. on May 2nd	Floating dust
	20110308	From 12:00 a.m. to 16:00 p.m. on Mar. 8th	Sunny day
Samples on non-dust days	20110416	From 12:00 a.m. to 16:00 p.m. on Apr. 16th	Sunny day
	20110523	From 12:00 a.m. to 16:00 p.m. on May 23rd	Sunny day

935 ^aNote that one exterior dust sample was collected on March 1 when no dust was recorded by the
936 MICAPS. However, the MICAPS information indeed showed dust events in China on March 1. The
937 modeled spatial distribution of the PM₁₀ mass concentration for this dust event on March 1 implies that
938 the sample should be classified as a dust sample. The supporting ~~figure is figures~~ was are shown in Fig.
939 S1.

940 ^bThe sampling duration was reduced to only 2 hrs because of extremely high particle loads. In addition,
941 the samples with IDs of 20080528 and 20080529 were subjected to two different dust events occurring
942 over two days instead of continuous samples for one dust event (CMA, 2009).

943 ^cNote that one exterior dust sample was collected on April 18 when no dust was recorded by the
944 MICAPS. However, blowing dust occurred and was recorded on April 17 by the Sand-dust ~~weather~~
945 Weather almanac Almanac 2011 (CMA, 2013). The modeled spatial distribution of the PM₁₀ mass
946 concentration for this dust event on April 18 implies that the sample should be classified as a dust
947 sample. The supporting figure is Fig. S2.

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978 **Table 2.** Detection limits, precisions and recoveries of water-soluble ions and metal elements.

Component	Measurement method	Detection limit ($\mu\text{g L}^{-1}$)	Precision (RSD%)	Recovery (%)
NO_3^-	IC	2.72	1.54	97
SO_4^{2-}		1.62	1.55	98
NH_4^+		0.4	1.10	97
Ca^{2+}		0.44	0.79	94
Cu	ICP-MS (Xin et al., 2012)	0.006	4.0	106
Zn		0.009	2.5	102
Cr		0.004	3.0	95
Sc		0.002	2.4	97
Pb		0.008	3.9	104
Al	ICP-AES (Lin et al., 1998)	7.9	0.6	103
Ca		5.0	1.2	99
Fe		2.6	0.7	104
Na		3.0	0.6	99
Mg		0.6	0.6	105
Hg	CVAFS	0.0001	6.6	105
As	CVAFS	0.1	5.0	98

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Table 3. The average concentrations and EFs of metal elements on dust and non-dust days.

Element	Concentration (ng/m ³)		EF*	
	Non-dust days	Dust days	Non-dust days	Dust days
Sc	1.11	13.90	-	-
Al	8.53×10 ³	6.86×10 ⁴	3.8	1.4
Fe	4.91×10 ³	3.88×10 ⁴	3.	1.2
Ca	1.05×10 ⁴	4.29×10 ⁴	14.0	2.1
Mg	1.62×10 ³	1.58×10 ⁴	3.5	1.1
Cu	50.2	124.5	36.3	6.1
Pb	127.9	221.0	389.4	56.1
Zn	340.0	457.7	248.9	20.6
Cr	33.8	244.0	44.0	11.1
Hg	0.26	0.36	176.0	13.8
As	25.5	27.4	707.2	43.9

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*EF values less than 10 indicate that the studied element is mainly derived ~~mainly~~ from crustal sources, whereas EF values much higher than 10 indicate an anthropogenic source.

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Table 4. Average concentrations of inorganic nitrogen (DIN), TSP, NO_x, relative humidity (RH) and air temperature for each aerosol sample_category_in Qingdao.

	Sample number	TSP ($\mu\text{g}\cdot\text{m}^{-3}$)	NO ₃ ⁻ ($\mu\text{g}\cdot\text{m}^{-3}$)	NH ₄ ⁺ ($\mu\text{g}\cdot\text{m}^{-3}$)	RH (%)	T (°C)	NO _x ($\mu\text{g}\cdot\text{m}^{-3}$)	Summary
Category 1	20080301	527	20.5	12.7	57	7.0	36	DIN concentration on dust days higher than that on non-dust days
	20080315	410	19.5	29.9	62	11.0	59	
	20090316	688	15.9	17.2	27	16.0	75	
	20100321	519	16.5	9.4	51	8.8	76	
	20110502	810	21.0	11.0	49	17.7	62	
Category 2	20080425	622	6.8	2.0	30	18.0	40	DIN concentration on dust days lower than that on non-dust days
	20080528	2579	9.2	2.7	17	27.0	34	
	20080529	2314	17.5	4.8	60	20.0	29	
	20110319	939	12.3	9.4	16	12.6	93	
	20110501	502	4.5	5.3	23	21.6	66	
Category 3	20100315	501	5.4	4.3	30	7.2	73	NO ₃ ⁻ concentration on dust days lower than that on non-dust days; NH ₄ ⁺ close to that on non-dust days
	20100320	3857	5.5	3.4	35	10.6	92	
	20110418	558	3.8	6.6	33	12.6	47	
Non-dust a	20080316	225	12.6	8.4	28	11.0	60	
	20080424	137	21.7	7.2	49	18.0	53	
	20080522	206	27.4	16.6	78	20.0	60	
	20090306	94	2.9	3.0	29	7.00	51	
	20100324	275	7.2	2.4	23	9.0	82	

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20110308	194	13.0	13.1	20	11.5	111
20110416	252	5.6	5.4	26	14.1	55
20110523	224	15.2	10.2	42	20.6	49

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^aFor the corresponding non-dust day for each dust event, see Table 1.

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Table 5. Comparison of the inorganic nitrogen (DIN) content in sand and aerosol particles on dust days or close to the dust source region (unit: $\mu\text{g/g}$).

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Study region and data source	Sands sampled in dust source regions		Aerosols in or close to dust source region on dust days			Aerosols in the coastal region of the Yellow Sea	
	Relative concentration ^a		Study region and data source	Relative concentration ^a		NO_3^-	NH_4^+
	NO_3^-	NH_4^+		NO_3^-	NH_4^+		
Zhurihe (This study)	25.46±22.87	4.21±1.03	Duolun (Cui, 2009)	1200	900	Non-dust: 28,200±24,819	Non-dust: 24,063±21,515
Alxa Left Banner, Inner Mongolia (Niu and Zhang, 2000)	62.1±7.4	79.1±1.1	Alxa Right Banner, Inner Mongolia (Niu and Zhang, 2000)	1975 ^b	4091 ^b	Category 1: 34,892±9570	Category 1: 22,571±7,016
Yanchi, Ningxia (Niu and Zhang, 2000)	46.4±2.2	80.9±1.3	Hinterland of the Taklimakan Desert, Xinjiang (Dai et al., 2016)	142-233	2-15	Category 2: 5,542±5,117	Category 2: 4,758±5,698
			Average of Sonid Youqi, Huade (Inner Mongolia), Zhangbei (Hebei) (Mori et al., 2003)	253	710	Category 3: 6,359±4,697	Category 3: 7,059±5,591
			Yulin, the north edge of Loess Plateau (Wang et al., 2011)	216.4	80.6		

Golmud,
 Qinghai(Sheng et al., 892.9 -^c
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 Mongolia (Yang et al., 588.1 No data
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1045 ^aRelative concentration of DIN per aerosol particle mass
 1046 ^b-Samples collected on a floating dust day (~~Horizontal~~horizontal visibility less than 10000 m and very
 1047 low wind speed)
 1048 ^c-The ammonium concentration was lower than the detection limit of the analytical instrument.

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1053 **Table 6. Sources and source contributions (expressed in%) calculated for aerosol samples collected**
 1054 **during dust and non-dust events**

<u>Dust event</u>		<u>Comparison days</u>	
<u>Source</u>	<u>% of TSP</u>	<u>Source</u>	<u>% of TSP</u>
<u>Soil dust</u>	<u>36</u>	<u>Soil dust</u>	<u>23</u>
<u>Industrial</u>	<u>21</u>	<u>Industrial</u>	<u>24</u>
<u>Secondary aerosol</u>	<u>6</u>	<u>Secondary aerosol</u>	<u>23</u>
<u>Oil combustion</u>	<u>6</u>	<u>Biomass burning</u>	<u>16</u>
<u>Coal combustion and other uncertain sources</u>	<u>31</u>	<u>Coal combustion</u>	<u>5</u>
		<u>Sea salt</u>	<u>9</u>

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1072 | **Table 67.** Concentrations of TSP, NO₃⁻, and NH₄⁺; transport speed; transport distance over the sea;
 1073 | transport distance; air temperature; RH; average mixed layer during transport and transport time in
 1074 | polluted region for atmospheric aerosol samples on dust days.

Group	Sample number	TSP (μg/m ³)	NO ₃ ⁻ (μg/g)	NH ₄ ⁺ (μg/g)	Speed (km/h)	Distance over the sea (km)	Transport altitude (m)	Mixed layer depth (m)	Residence time ^a (h)	T ^b (°C)	RH (%)
Category 1 IN>ND	080301	527	38,984	24,107	40.1	0	1,160±702	864±745	39	-2.9±11.7	29±10
	080315	410	47,611	34,130	79.1	0	4,921±1,870	950±525	13	-32.5±16.4	34±16
	090316	688	23,050	25,012	86.2	0	3,739±1,083	702±665	11	-19.1±11.7	42±17
	100321	519	31,741	18,155	87.2	0	3,407±1,249	1,113±760	19	-23.0±13.6	42±22
	110502	810	25,995	13,632	30.2	177	3,666±1,371	747±957	26	-13.2±15.8	31±13
Category 2 IN<ND	080425	256	4,089	372	29.6	0	887±656	1,161±1,040	10	-2.7±6.1	66±13
	080528	2579	232	72	88.2	244	4,336±1,461	1,064±830	8	-15.5±13.6	31±16
	080529	2314	26	166	63.7	94	2,148±1,725	1,194±816	43	3.6±18.4	25±17
	110319	939	13,088	10,067	70.6	132	4,271±1,867	790±719	27	-26.3±20.0	48±32
	110501	502	8,924	10,631	35.1	252	3,212±810	916±1,114	5	-13.4±8.5	39±13
Category 3 NO ₃ ⁻ <ND NH ₄ ⁺ ≧ND	100315	501	10,767	8,515	57.3	0	5,009±1,410	1,110±365	7	-40.4±13.3	45±29
	100320	3857	1,418	884	76.9	0	1,284±401	525±371	10	-12.2±6.3	61±16
	110418	558	6,891	11,778	35.6	931	1,344±780	695±672	2	-0.1±8.2	52±28

1075 | ^aResidence time of the air mass passing over parts of highly polluted regions according to the
 1076 | trajectories of samples.

1077 | ^bAverage air temperature with the definition in Section 2.4.

1078 | ^cAverage relative humidity with the definition in Section 2.4.

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Table 7. Sources and source contributions (expressed in%) calculated for aerosol samples collected during dust and non dust events

Dust event		Comparison days	
Source	% of TSP	Source	% of TSP
Soil dust	36	Soil dust	23
Industrial	21	Industrial	24
Secondary aerosol	6	Secondary aerosol	23
Oil combustion	6	Biomass burning	16
Coal combustion and other uncertain sources	31	Coal combustion	5
		Sea salt	9

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Table 8. Dry deposition of TSP (mg/m²/month), particulate inorganic nitrogen (mg N/m²/month) and some toxic trace metals (mg/m²/month) on dust and non-dust days.

	Dry deposition flux							
	TSP	NO ₃ ⁻ -N	NH ₄ ⁺ -N	N _{NH4++NO3-}	Fe	Cu	Pb	Zn
Category 1 ^a	8,000± 1800	65±9	24±14	90±17	533±179	2±0.3	0.3±0.3	6±2
Category 2 ^a	18000± 11,000	13±18	8±4	21±22	1300±100 0	3±2	0.08±0.04	4±1
Category 3 ^a	29,000± 31,000	26±6	17±8	42±12	2100±220 0	6±1	0.20±0.02	5±3
Non-dust	2,800± 700	48±33	15±8	63±39	190±110	1±1	0.09±0.1	5±4

^aFor the characterization of $N_{NH4++NO3-}$ concentration and sample information of the category, see Table 3.

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Table 9. Comparison of dry deposition flux and normalized flux of TSP (mg/m²/month) and N_{NH4++NO3-} (mg N/m²/month) with observations from other studies (mg N/m²/month)

Source	Year	Area		TSP	N _{NH4++NO3-}	Normalized average flux of N _{NH4++NO3-} ^a
This work	2008-2011	Qingdao, coastal region of the Yellow Sea	Non-dust day	2,800±700	63±39	93.90
			Dust day	10,138±15,940	58±36	101.39
			Average of dust and non-dust			97.64
Qi et al., 2013	2005-2006	Qingdao, coastal region of the Yellow Sea	Average of nine months samples	159.2 - 3,172.9	1.8-24.5	94.75
Zhang et al., 2011	1997-2005	Qingdao	Average of annual samples		132	99.65
Zhang et al., 2007	1999-2003	The Yellow Sea			11.43	9.91
Shi et al., 2013	2007	The Yellow Sea	Non-dust day		19.2	132.17
			Dust day		104.4	227.07
			Average of dust and non-dust			179.62

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^a-The calculation method of the normalized flux of N_{NH4++NO3-} was discussed in section-Section 3.7.

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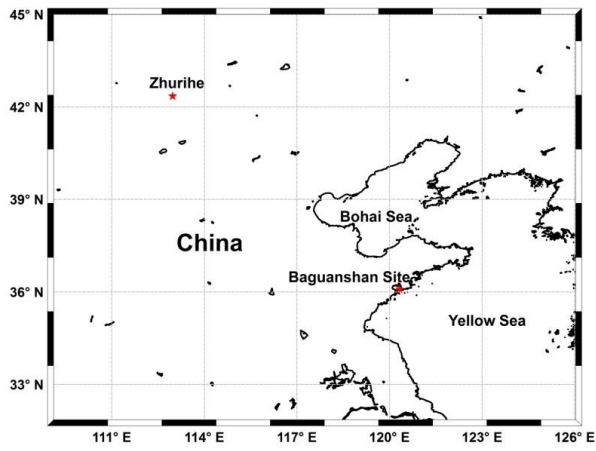
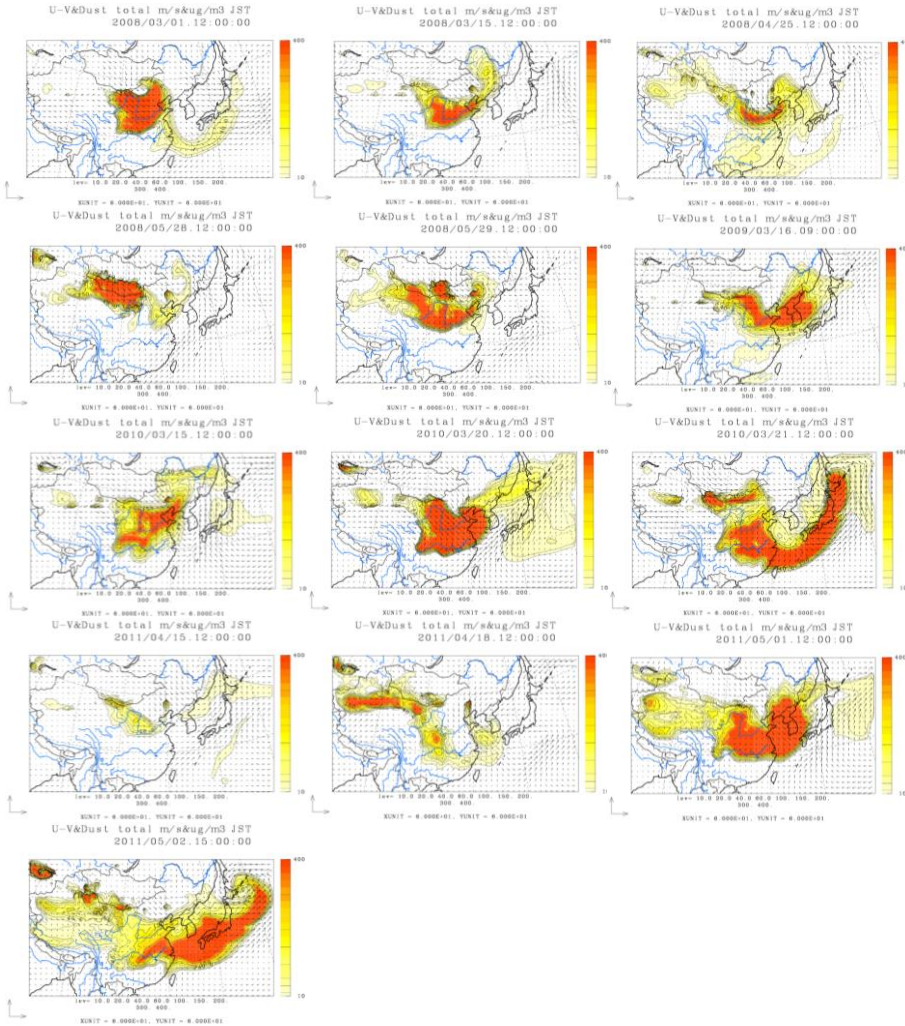


Figure 1. Location of the aerosol and dust sampling sites.

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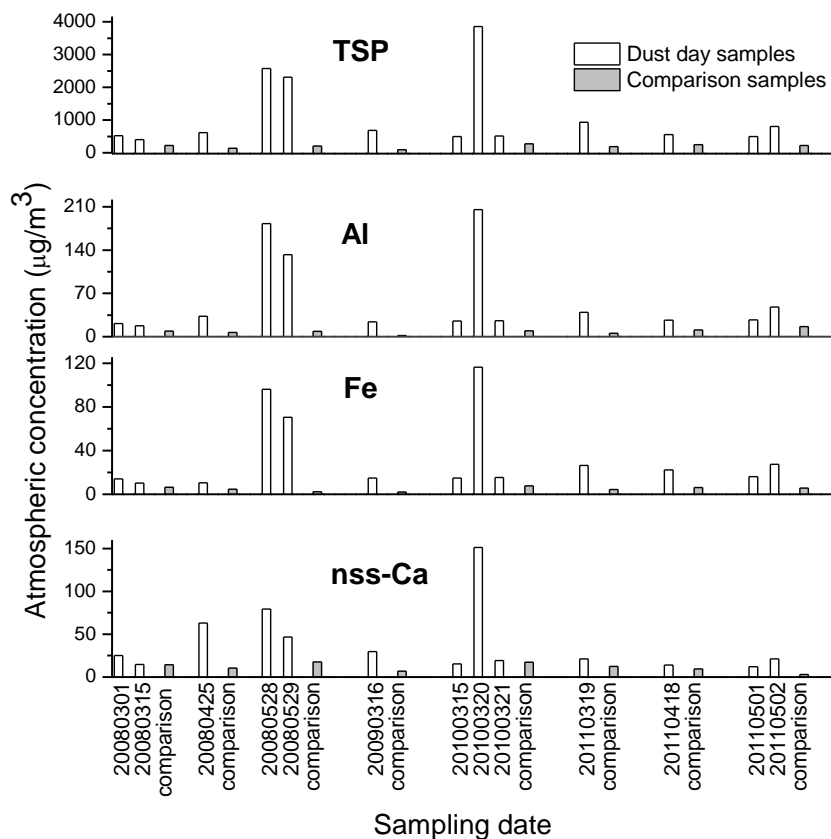
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1204 **Figure 2.** Modeled dust concentrations over East Asia by CFORS model during each dust sampling day
1205 from 2008 to 2011 (<http://www.cfors.nies.go.jp/>). (The figures show the modeled dust concentration
1206 in the middle of each sampling duration). No data are available for Mar. 19, 2011, because of the
1207 earthquake in Japan. Hourly PM10 concentrations were modeled by the WRF CMAQ model for
1208 each sampling day, and the results are shown in Fig. S3.

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1220 **Figure 32.** Mass concentrations of TSP, Al, Fe and nss-Ca in aerosol samples collected at the
 1221 Baguanshan site on dust and comparison days from 2008-2011.

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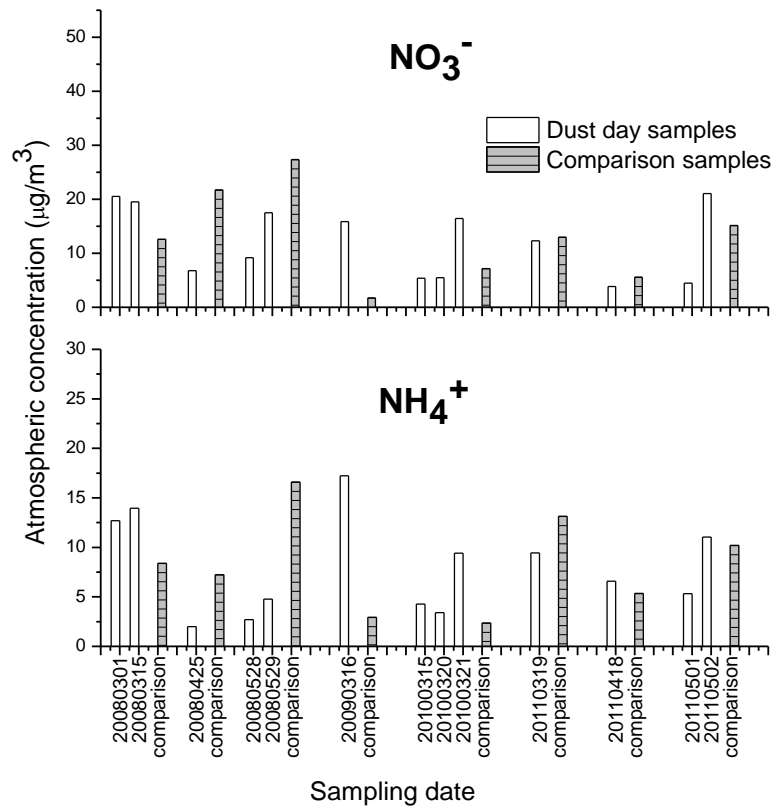
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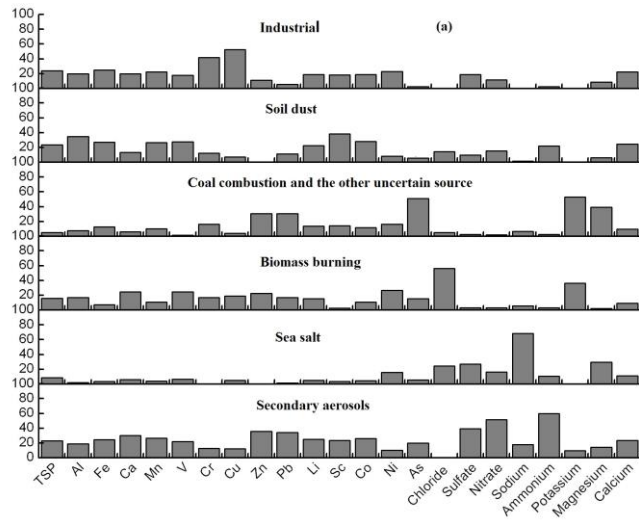
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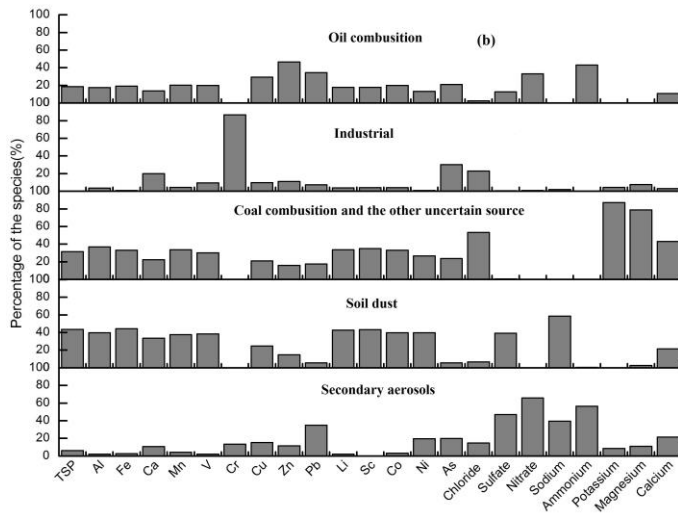


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Figure 43. Mass concentrations of NH₄⁺ and NO₃⁻ in aerosol samples collected at the Baguanshan site on dust and comparison days during March-May in 2008 to 2011.



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Figure 4. Source profiles of atmospheric aerosol samples collected on non-dust (a) and dust (b) days using the PMF model.

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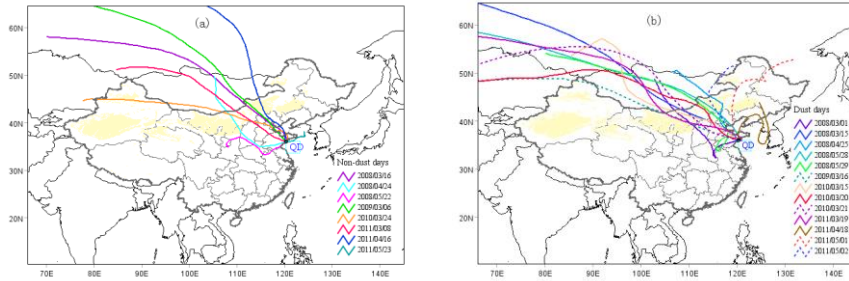
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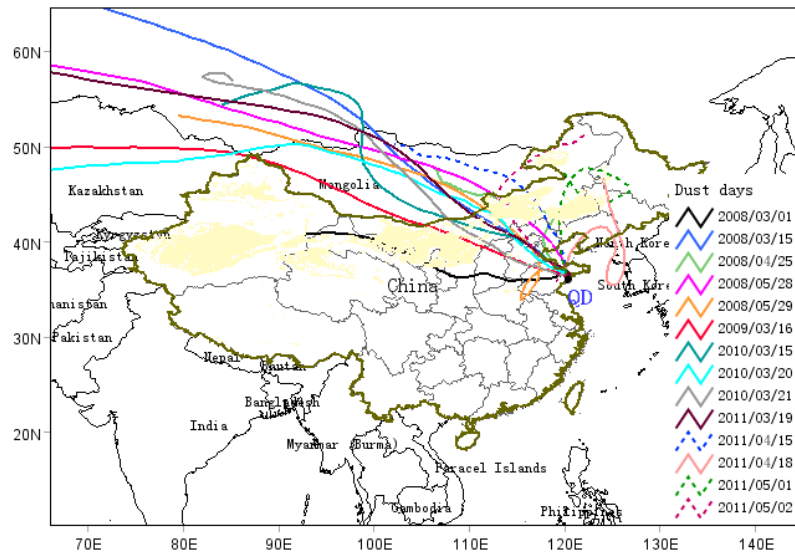
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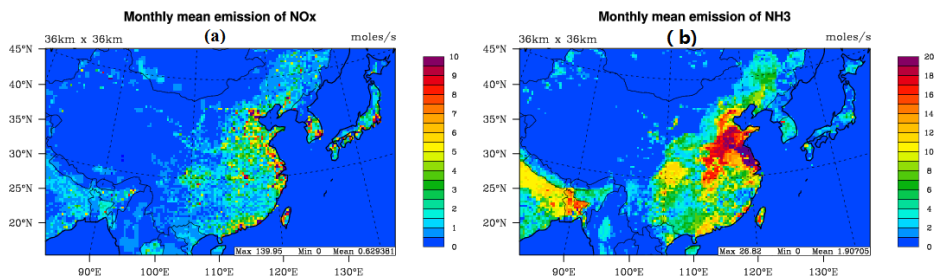
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Figure 55. The 72-h backward trajectories for non-dust (a) and dust (b) samples from 2008 to 2011 (the yellow domains in the maps represent the dust source regions in China).

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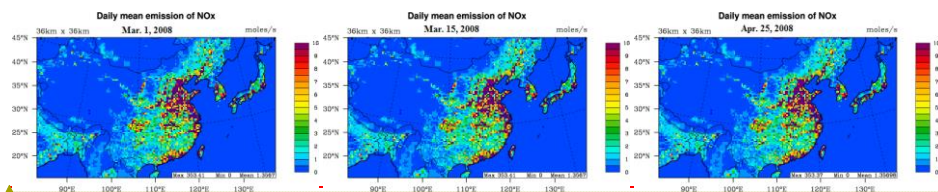
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Figure 6. Seasonal mean emissions of NO_x (a) and NH₃ (b) over East Asia from March-May 2008.

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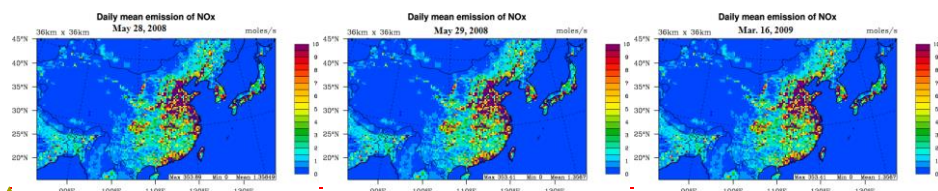
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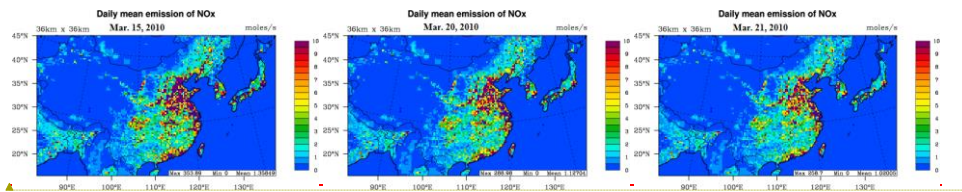
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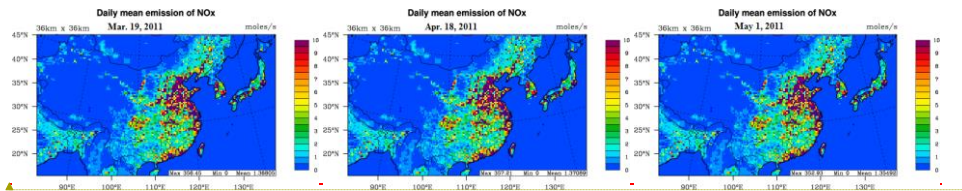
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Figure 6. Daily mean emission of NOx over East Asia on the dust days from 2008 to 2011.

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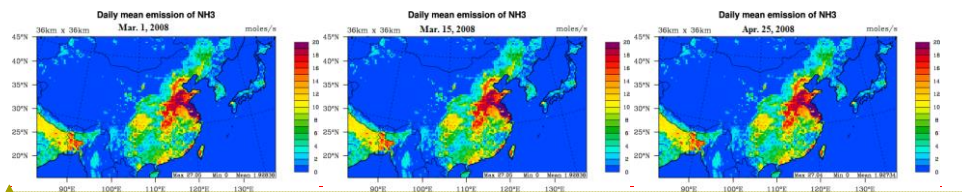
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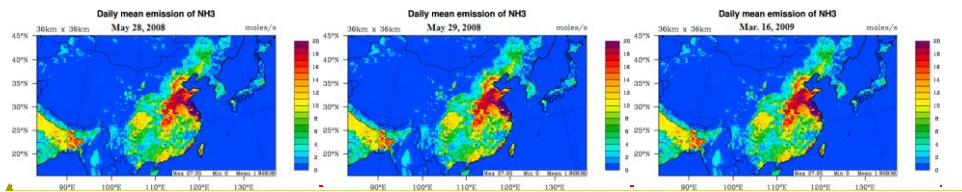
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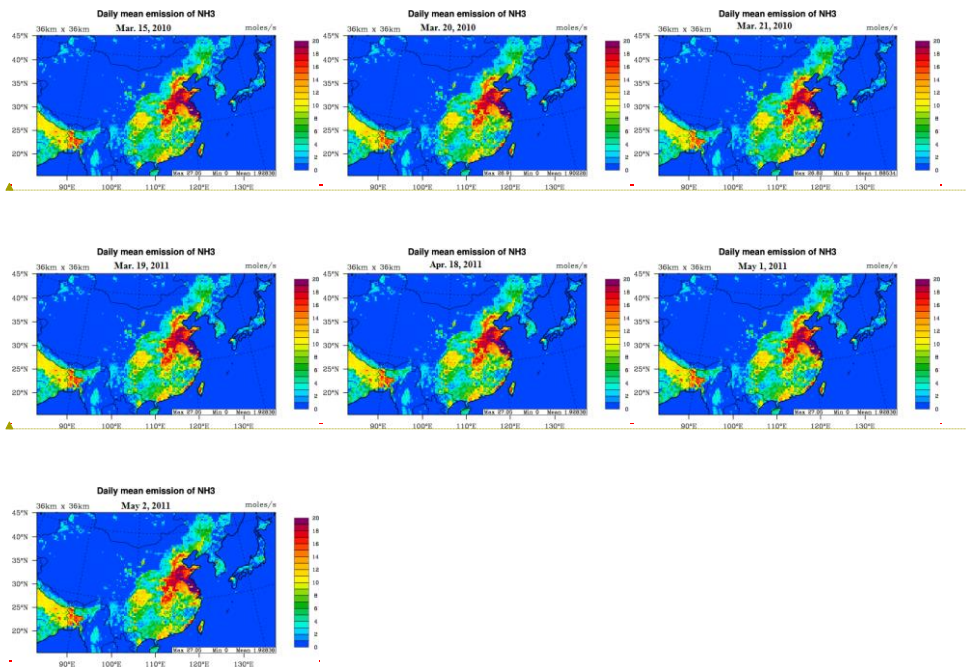


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Figure 7. Daily mean emission of NH_3 over East Asia on the dust days from 2008 to 2011.

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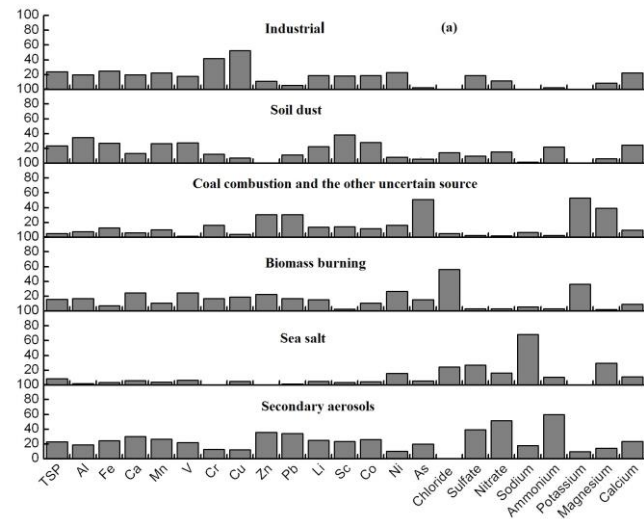
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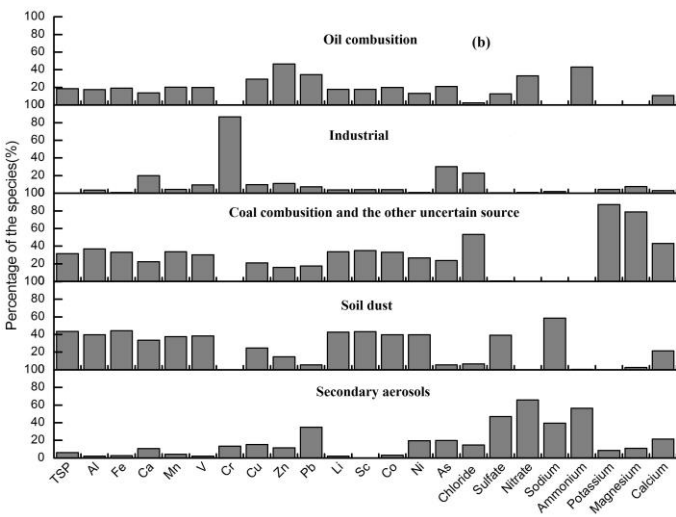


Figure 8. Source profiles of atmospheric aerosol samples collected on non-dust (a) and dust (b) days using the PMF model