## **Response to Reviewers' Comments on Manuscript: acp-2016-1038**

## Dear Editor,

We are thankful very much to you and the anonymous reviewers for the profound comments and suggestions. We have revised this manuscript accordingly. Listed below are our point-by-point responses (blue) to each reviewer's comments (black) in the revised manuscript with marks. In addition, we would like to ask you if we can add Yanli Feng as another corresponding author, considering contributions to experiment instruction and manuscript revised?

# Best regards,

Dr. Yingjun Chen

# Referee #1

#### General comments

The manuscript by Cui et al. summarizes emissions measurements from multiple generation diesel excavators and trucks under different operating and driving conditions. These types of measurements are unique in China and much needed. The paper is well organized, but it needs a thorough edit as many words, verbs, etc are not used correctly or are missing. Below I highlight the technical weaknesses, minor clarifications, and instances where sentences are confusing and need to be rephrased. I approve publishing the paper after these concerns are addressed.

**<u>Response:</u>** Thanks for the reviewer's positive approval. Clarifications have been provided and confusing sentences were rephrased in the revised manuscript.

<u>Comments #1</u>: (1) One of the weaknesses of this work is that each truck/excavator was tested only once. Thus it's unknown how representative these results are and how variability in the measurements affect the observed emission factors. I doubt that duplicate runs can now be carried out; however, the authors should at least mention and address this weakness. (2) Another weakness is that driving conditions of the

trucks were not similar (as shown in Figure S2); since driving conditions and engine load can have significant impacts on the emission factors, how can the results be interpreted in a unified manner? This should also be addressed in the discussion and conclusion sections. (3) Related to this is the variety of the engines tested in this work for both excavators and trucks. For example for excavators, engine powers span a range of 35-169 KW and total weights and engine displacements also vary a lot. On one hand, it's good to have a sampling pool of various engine types/sizes. On the other hand, these difference should be kept in mind and referred to when comparisons are made throughout the paper.

**<u>Response:</u>** Thanks for the reviewer's constructive suggestions. This major question was divided into three questions and we would provide a personal response to your comments, separately.

(1)We appreciate the review's comment. Indeed, we are also attached importance to the weaknesses of tested time in this study. However, given the difficulties of field measurements and some important parameters missing in the links of repeat tests, only one relative complete test was chosen for further discussion. In order to evaluate the variability, we had conducted some repeats for individual vehicles, and the results were presented in Tables S3 and S4 in the revised supporting information. As shown in Tables S3 and S4, the variability in test times for the same operational mode was considered acceptable. Moreover, we combined some repeat tests for organic matter analysis for T1 and T2, which could reduce the uncertainty. We confirmed that the weaknesses of repeatability existed in this study, and mentioned this weakness in the revised manuscript (**Page 9 line 16-23**).

	$O_2^{a}(\%)$	$CO_2^{a}(\%)$	CO <sup>a</sup> (ppm)	NOx <sup>a</sup> (ppm)	$PM (mg m^{-3})$	$OC (mg m^{-3})$	EC (mg $m^{-3}$ )
1	16.2	3.4	309	453	11.9	4.3	1.9
2	16.3	3.4	257	457	14.6	6.1	2.9
3	16.3	3.4	262	445	14.4	6.8	2.5
SD	0.08	0.01	28.6	5.68	1.55	1.26	0.53

Table S3 Pollutants mass concentrations emitted from E4 in three idling repeat tests

a: the datum were presented on other unpublished research

Trucks	Roads	1	2	3	SD
	non-highway 1	15.0	16.2	/	0.87
Light duty-China III	highway 1	19.8	30.6	/	7.67
	non-highway 2	21.3	16.1	/	3.68
Heavy duty-China II	non-highway 3	7.87	6.11	6.69	0.89
Madium duty China III	non-highway 4	11.0	10.3	/	0.49
Medium duty-China III	highway 2	8.79	17.1	/	5.85
	non-highway 4	5.29	9.56	6.99	2.15
Heavy duty-China III	highway 2	10.6	7.42	/	2.24

Table S4 PM mass concentrations emitted from trucks in some repeat tests (mg  $m^{-3}$ )

(2) Thanks. As mentioned in the revised manuscript (**Page 9 line 3-6**), different emission standards diesel trucks must run on different roads, which was restricted by traffic rules. For example, "yellow label car" can only run on the particular road and is not allowed running on the highway and arterial road. Therefore, different routes were chosen for different trucks. Although driving conditions of the trucks were not similar shown in Figure S2, the different characteristics of velocity on the highway and non-highway were obviously. Therefore, we just discussed highway and non-highway routes in this study. We have addressed this weakness and interpreted the unified manner in the revised manuscript (**Page 9 line 6-7**).

(3) Thanks for the comment. As we could seen from Figure S5 in the revised supporting information, the average  $EF_{PM}$  was less affected by engine power. It was regretful that the sample size in this study seemed not enough to reflect the impact from engine power. Thus, we just gave  $EF_{PM}$  of different engine power in the revised manuscript, and didn't discuss in-depth (**Page 14 line 19-21**).



Figure S5 PM emission factors for different power excavators

<u>Comments #2</u>:For readers who are not familiar with the standards in China, it will be useful to have a table where major particulate and gaseous emissions of each generation standard for trucks/excavators are listed.

**<u>Response</u>:** Thanks for the suggestion. The major particulate and gaseous emissions of each generation standard for trucks/excavators were listed in Tables S1 and S2 (**Supporting information**).

<u>Comments #3:</u> P7, L23: Although mentioned in Table 2, please indicate in the text the average (or range of) sulfur content of the fuels as well as the limit of GB 252-2015.

**Response:** Thanks. The range of sulfur content and limit of GB 252-2015 have been added in the revised manuscript (**Page 10 line 20**).

<u>Comments #4:</u> P8, L22: what recovery % for each species were achieved? <u>Response:</u> Thanks for the comment. The recoveries of five surrogates have been added in the revised manuscript (Page 11 line 28).

<u>Comments #5:</u>P12, L3: It seems the trucks with China II and China III standards had similar PM emission factors. Why is that so? Do these standards pose similar levels

for PM? or is it that the trucks tested don't necessarily represent the standard? or is this an instance where results from a single measurement from a truck are uncertain? **Response:** We appreciate the review's comment. As we discussed in the manuscript, the most important reason causing this result was different driving conditions for those two trucks. Due to heavy pollutions from China II trucks, traffic laws regulate that China II trucks are forbidden to drive on city center and only allowed to drive on some remote parts of the city, while the roads for China III trucks are always jammed. For evaluating the emission from trucks in the real world, we shouldn't neglect the driving conditions to discuss trucks itself. However, we confirmed that the number of measurement was shortage in this study, and we will lucubrate in the future.

**<u>Comments</u> #6:**P12, L7: unclear what "more volatile" means here

**Response:** Thanks for the comment. "more volatile" refers to highly varied speed (**Page 16 line 15**).

<u>**Comments** #7:</u>P12, L11-13: It doesn't make sense that trucks driven on road with higher grade have lower emissions. Please clarify.

**<u>Response</u>**: Thanks. There was wrong with expression and we have modified in the revised manuscript (**Page 16 line 23**).

<u>**Comments**</u> #8:P. 12, L17: what's the justification for using OM/OC=1.6 for such fresh emissions? How will the result change if a lower factor, more representative of fresh emissions, is used?

**Response:** Thanks for the reviewer's constructive suggestions. Chow et al (2015) showed that a conversion factor used to transform OC to OM was ranged from 1.2 to 2.6, depending on the extend of OM oxidation. Fresh aerosols from different sources had different values, such as 1.4 and 1.6 for diesel engine (Gilardoni et al., 2007, Japar et al., 1984) and 1.7 for biomass burning (Chow et al., 2015). Therefore, we assumed the conversion factor is 1.6 in this study.

**<u>Comments</u> #9:**P13, L9, P14,L2: it is mentioned that diesel sulfur content affected OC/EC. It is unclear to me how fuel sulfur can affect emission of organic compounds and soot. Please explain.

**Response:** We appreciate the review's comment. According to references, we assumed that the formation of organic compounds and soot was obviously affected by diesel sulfur content in two points. On the one hand, organosulfurs constituted up to 62% of the total sulfur content in diesel (Adlakha et al., 2016). Organic compounds existing in diesel were removed simultaneously by process of desulfurization. Therefore, emissions of organic compounds and soot generated by hydrogen abstraction/acetylene addition were reduced (S ánchez et al., 2013). One the other hand, sulfuric acid, the nucleating agent in diesel particle formation, generated by sulfur in diesel (Ruiz et al., 2015). These nucleating agents might provide a place for organic compounds condensation and reaction.

<u>Comments #10:</u>P13, L11-26: It is unclear what the elemental emissions are stemming from: the fuel or bad conditions of the engine or the lubricating oil? Please explain. For example, L22, it is mentioned that diesel quality used in E4 was poor. Was the fuel also tested for elemental content? Were Cu and Zn higher in this fuel as well? <u>Response:</u> Thanks for the comment. Although diesel quality was analyzed in this study, many elemental contents were below the method detection limit. Wang et al. (2003) reported that the concentrations of Fe, Ca and Mg accounted for 50% of the total elements in diesel fuel. Thus, the possible source of elements was diesel, while Cu and Zn were affected by sampling environments for E4. The detail information could be seen in the revised manuscript (**Page 18 line 6-17**).

<u>**Comments#11:**</u>P13, L23, P14, L4-5, P17, L24-26: Authors mention that % of elemental composition in E1 and E6 was higher. How did absolute concentrations or emission factors of the elements compared for these two vs. the others? Since % values depend on concentrations of other components as well, I don't think they're as relevant to be mentioned, especially since the contribute to a very small fraction of the emissions.

**<u>Response</u>:** Thanks for the comment. The average emission factors of elemental were 5.66 mg·kg<sup>-1</sup> for E1+E6 and 4.02 mg·kg<sup>-1</sup> for E2+E3+E4+E5, and were mentioned in the revised manuscript (**Page 18 line 19-20**).

<u>**Comments#12:**</u>P14, L7-10: It is unclear how the authors concluded that alkane/hopane/steranes were influenced by fuel quality and PAHs by combustion. Please explain and clarify.

**<u>Response</u>:** Thanks for the comment. N-alkanes, hopanes and steranes fractions were the highest in excavator E4, while PAHs fraction was the highest in excavator E3. Comparing the fuel quality between E3 and E4, E4 had a poorer diesel quality, which might be the main reason for high n-alkane, hopanes and steranes. Similarly, it was said by Rogge et al. (1993) that n-alkanes, hopanes and steranes were mostly derived from incomplete combustion of fuel and lubricant oil. However, we speculated that PAHs was affected by combustion conditions (e.g. combustion temperature) in this study, due to E3's better performance (stage 2) and relatively superior fuel quality. The distinct explanation was added in the revised manuscript (**Page 19 line 1-13**).

<u>Comments#13:</u>P16, L11: Please explain what reactions in the engine authors refer to. <u>Response:</u> Thanks for the comment. The description of reactions was provided in the revised manuscript (Page 22 line 1).

<u>**Comments#14:**</u>P17, L3: Is it really that presence of metals oxidizes soot?! or do the metals enhance combustion and reduce formation of soot?

**<u>Response</u>**: We appreciate the review's comment. It was said by Kasper et al., (1999) that the action of iron oxide was recognized as a catalyst and burnout rate of soot could promote during combustion process. Therefore, we inferred that metals may enhance combustion of soot. The corresponding expression was added in the revised manuscript (**Page 22 line 27-28**).

Comments#15: Acronyms of PAHs should not be used in the abstract.

**<u>Response</u>**: Thanks, the acronyms of PAHs have been changed to full names (**Page 1 line 24-25; Page 3 line 1-3**).

**<u>Comments#16:</u>**Define BaPeq in the abstract 3.

**<u>Response</u>**: Thanks for the comment. The  $BaP_{eq}$  has been defined in the revised abstract (**Page 3 line 9**).

**<u>Comments#17:</u>** P3, L 7: define PM. Throughout the paper indicate what size PM refers to (PM1, PM2.5, etc).

**<u>Response</u>**: Thanks for the comment. PM referred to total suspended particulate  $(D_p \le 100 \ \mu m)$  in this study. We have remarked in the revised manuscript (**Page 3 line 23**).

Comments#18: P12, L12: consider using "higher road grade".

**Response:** Thanks. We agree with the reviewer's suggestion and changed the word as suggested (**Page 16 line 23**).

**<u>Comments#19:</u>** P20, L3: Do authors mean excavators rather than diesel truck here or should E1, E2,.... be T1, T2, etc?

**Response:** Thanks. The E1,E2... have changed to T1,T2....(**Page 26 line 25**)

**<u>Comments#20</u>**: Figures: Axis labels are all too small and need to be modified for better quality figures.

**Response:** Thanks for the advice. Axis labels in Figure 1, 3, 4, 5 and 7 were modified in the revised manuscript (**Page 40; Page 42; Page 43; Page 44; Page 46**).

**<u>Comments#21:</u>**Fig 7: what do the errors bars represent? Unclear form the caption what the difference between A-B and C-D symbols are.

**Response:** Thanks for the comment. A and B are isomer ratios of PAHs for

excavators and trucks tested in this study, respectively; C and D are average isomer ratios of PAHs for trucks and excavators tested in this study. The vertical and horizontal errors bars represent the standard deviation of values shown in vertical and horizontal axis, respectively (**Page 46**).

**<u>Comments#22:</u>**Fig. S3. What are the crosses and dashed lines in these box and whisker plots?

**<u>Response</u>**: Thanks for the comment. The annotations were shown in the revised Figure S4 (**Supporting information**).

<u>Comments#23:</u>Sentences needing to be rephrased: 1. P3, L 13-15 2. P4, L18-20 3. P7, L12-14 4. P12, L1-3 5. P. 13, L19-20 6. P.18, L3-4.

**<u>Response</u>**: Thanks for the comment. We have made every effort to polish our English and asked a native English speaker to take a proof reading of the revised manuscript.

## Referee #2:

## General comments

Cui et al. present data from measurements of particulate matter emissions and composition from real-world testing of a suite of on- and non-road diesel vehicles. They find that PM emissions, while variable, exhibit trends with fuel quality and emissions standard. Although these data add to the literature and will eventually help build more realistic emissions inventories for China, I do not recommend publication of this version of the manuscript in ACP. I have two major comments and numerous minor comments.

**<u>Response</u>**: Thanks very much for the comments. We have revised this manuscript carefully, and please find our detailed responses below.

# Major comments:

Fit: The manuscript, in my opinion, does not fit the research foci of publications

typically accepted in ACP and I wonder if another journal would offer a better fit for this research.

**Response:** Thanks. But the authors disagree with this comment and consider ACP as the best journal of high quality to publish our precise measurement data. On the one hand, in recent years, PM emission from diesel vehicles drawn more and more attention in China, due to severe air pollution. However, the great uncertainty existing in PM from diesel vehicles exhausts makes those field datum very precious. Although our research is preliminary, as far as we know, this manuscript is the first on-board research in China that focused on PM chemical constituents from on-road and non-road diesel vehicles exhaust. The results of this study could provide basic data for air quality assessment and establishment of emission standard. Therefore, we chose ACP, one of the most influential journals in atmospheric fields, to publish our results for obtaining broader attention. On the other hand, the main subject areas for ACP comprise atmosphere modeling, field measurements, remote sensing, and laboratory studies of gases, aerosols, etc. Nowadays, several researches about emission factors and characteristics of PM from diesel engine have been published in ACP (Dai et al., 2015, Dallmann et al., 2014, Lin et al., 2015, Zhang et al., 2015). Therefore, this manuscript is fit to publish in ACP, because of general implications for source apportionment and health assessment.

## <u>Comments #1</u>:No new methods/instruments were used that make the data novel.

**Response:** Thanks. We have added some descriptions about the progressiveness of methods/instruments used which made the data novel in the present study in the revised manuscript (**Page 9 line 25-26; Page 10 line 5-10**). Briefly, the portable on-board emission measurement and dilute sampling system which was designed and manufactured in our laboratory has good performance (Zhang et al., 2015), and has obvious advancement compared with other on-board instrument for vehicles such as PEMS and FPS4000 (Zheng et al., 2015) by the portability and capability of filter sample collection for further PM chemical analysis in the laboratory. Furthermore, the present result was the first set data of on-board measurement for non-road diesel

#### vehicle exhaust in China.

<u>Comments #2</u>: The measurements were performed on a very small cross-section and are not necessarily representative of the on- and off-road fleet in China. The small sample size, small cross-section, and large variability do not suggest large shifts/trends in emissions (or at least make them hard to observe).

**<u>Response</u>**: Thanks. We admitted that the sample size in this study was small, but wide ranges of vehicle types (including different emission standards and engine powers) were considered in this study. Furthermore, the most important purpose in this manuscript was to analyze the chemical constituents of PM from diesel vehicles exhausts, which needed a heavy workload.

Actually, we had selectively conducted some repeated experiments in this study to evaluate those variability and the results were shown in Tables S3 and S4 (**Supporting information**). As shown in the Tables S3 and S4, the variability was considered acceptable. Because there were some parameters missing in the field measurement, we decided to select an completed test for calculating the emission factors and combine the repeated filters to reduce this uncertain for some diesel vehicles. In the future work, we would increase the sample size to ensure the datum stability after this first attempt (**Page 9 line 16-23**).

<u>**Comments**</u> #3:Comparisons with literature data are not very insightful. While the data add to the literature in terms of quantifying emission factors of PM from a modern set of vehicles under real-world conditions, the scientific contributions in this research effort are lean. The data need to be published but this journal may not be the right target.

**<u>Response</u>:** Thanks. The purpose and the greatest contribution of this study were established characteristics of PM and its constituents emitted from trucks and excavators using on-board measurements. In China, diesel vehicles are facing imperfect emission standards and messy diesel quality, especially for non-road diesel vehicles. The knowledge relative to the characteristics of PM emission from those

diesel vehicles was slim to none. It was extremely difficult to collect literature data and compare with results obtained in this study, due to lacking of researches for characteristics of PM and its constituents by on-board tests. Following the reviewer's suggestion, we have made more interpretation among the comparison in the revised manuscript (**Page 18 line 6-17; Page 19 line 1-13; Page 25 line 1-11**). Finally, we chose ACP to publish our results for obtained broader attention from the perspective of the importance of the datum.

<u>**Comments #4</u>**:Writing: The quality of technical communication is very poor. This suggests one or all of the following: (a) the first author was rushed to write and submit this manuscript, (b) the senior authors have not read through this manuscript, (c) the authors place no emphasis on clear and effective communication. The manuscript needs to be significantly</u>

improved by the senior authors to meet the expectations of an English language publication in a high impact journal. If the manuscript is not heavily edited for English, this would be reason enough for rejecting the manuscript from publication. Here are a few examples from just the first few pages:

a. Page 1, line 24: 'involving wide-range emission standards'

b. Page 2, line 11: 'PM compositions emitted from excavators dominated'

c. Page 2, line 23: 'the complex of operating modes'

d. Page 3, line 7: 'diesel vehicles exhaust is a major source of emissions in ambient PM'

e. Page 3, line 9: '30% of emissions in ambient PM'

f. Page 3, line 18: 'causing severe emission situation'

g. Page 3, line 23: 'almost higher than 90% of PM came from on-road diesel vehicles emission'

h. Page 3, line 27: '349 thousand tons PM emission'

- i. Page 5, line 23: 'organic matters'?
- j. Page 5, line 26: 'impact factors of PM'; what does that mean?

**Response:** Thanks for the comment. We have made every effort to polish our English

and asked a native English speaker to take a proof reading of the final version of the revised manuscript.

a 'Involving wide-range emission standards' was changed to 'involving a range of emission standards.

- b 'PM compositions emitted from excavators dominated' was changed to 'PM composition emitted from excavators was dominated'.
- c 'The complex of operating modes' was changed to 'the complex characteristics of excavator operational modes'.
- d 'Diesel vehicles exhaust is a major source of emissions in ambient PM' was changed to 'Diesel vehicles exhaust is a major source of ambient PM emissions'.
- e '30% of emissions in ambient PM' was changed to '30% of ambient PM emissions'.
- f 'Causing severe emission situation' was changed to 'and have contributed to severe emissions problems'.
- g 'Almost higher than 90% of PM came from on-road diesel vehicles emission' was changed to 'more than 90% of PM resulted from on-road diesel vehicle emissions'.
- h '349 thousand tons PM emission' was changed to '349 Gg of PM emissions'.
- i 'Organic matters' was changed to 'organic compounds'.
- j 'Impact factors of PM' was changed to 'influential factors of PM'.

# **Minor Comments:**

<u>**Comments #1</u>**: Emissions standards: It might be worthwhile to describe the on-road and off-road emissions standards (e.g., Stages and China) and their emissions limits for PM (and other pollutants too) at the beginning of the manuscript through a Table. This would help orient the reader and also allow easy comparison with the EPA and EURO standards.</u>

**<u>Response</u>**: Thanks. We have added the on-road and off-road emission standards in the revised manuscript (**Supporting information**).

<u>**Comments #2:**</u> Page 2, line 9: Did vehicle exhaust contribute to 30% of the PM concentrations or emissions? Unclear; please clarify.

**<u>Response</u>**: Thanks. We have modified the unclear place in the revised manuscript (**Page 3 line 24-26**).

<u>Comments #3</u>: Page 4, line 3: construction equipment might be better word <u>Response:</u> Thanks for the advice. Following the reviewer's suggestion, we have changed the expression in revised manuscript (**Page 4 line 28**).

<u>**Comments**</u> #4: Page 3, line 16 to page 4, line 5: It might be better if the number of vehicles, fuel consumption and PM emissions in China were represented through a table or figure, alongside the relative importance of trucks and excavators to justify the use of those vehicle types in this research.

**<u>Response</u>**: Thanks for the advice. The figure S1 was added in the revised supporting information (**Supporting information**).

<u>**Comments #5:**</u> Page 4, line 18 to page 5, 10: The authors have only cited other people's work but have not paraphrased their findings. Hence, it is unclear what the gaps and motivation for this work is.

**<u>Response:</u>** Thanks. We have rephrased the correspond contents in revised manuscript (**Page 6 line 1-30;Page 7 line 1-6**).

<u>**Comments #6**</u>: Page 6, line 19: I did not understand how the duration of the different modes were determined. Also, what torque-speed ratings do the idling, moving, and working mode correspond to?

**<u>Response</u>**: Thanks. The time of sampling under different modes was not strictly required, as long as assured enough contents of PM to conduct chemical analysis. We have clarified it in the revised manuscript (**Page 8 line 24-28**). Actually, the basis of selecting those modes were not according to torque-speed ratings. The idling mode refers to engine keeps running at low speed (about 600-800 rpm), but not moving or working. The moving mode refers to that excavator moves at low speed (below 3-5

 $km \cdot h^{-1}$ ), but the bucket is not unload. The working mode refers to that bucket scoops the soil, then moves to another location and scoops again.

<u>**Comments #7**</u>: Page 7, line 28: Why did the researchers use quartz-fiber filters? My understanding is that the fibers can tear off during handling and bias the gravimetric measurement. Do the authors mean Teflon-coated quartz fiber filters?

**<u>Response</u>:** Thanks. We used quartz-fiber filters for gravimetric measurements in this study. The quartz-fiber filters losses could be neglected. Because the filters were parceled by aluminum foil after sampling to avoid filters tearing off, and the PM weight of error in quartz-fiber and Teflon filters could acceptance. In adition, quartz-fiber filters were selected to measure PM weight for consistent with those used in the chemical analysis. We have added the reasons in the revised manuscript (**Page 10 line 26-29**).

**<u>Comments</u> #8**: Section 2.4.3: The BaPeq method needs to be discussed in detail for the reader to follow the calculation.

**<u>Response</u>**: Thanks. The detailed BaPeq method was added in the revised manuscript (**Page 13 line 18-30**).

<u>Comments #9</u>: Section 3.1: What fraction of the improvement between pre-stage 1 and stage 2 can be attributed to better quality fuel as opposed to the emission standard?

**Response:** Thanks. We supposed that the fuel quality rather than the emission standards has a more great impact on PM constituents. Although the threshold (total emission) was set in non-road emission standards, constitutes of PM haven't regulated in these standards. Furthermore, it was said that sulfur in fuel translates to sulfuric acid which is the nucleating agent in diesel nanoparticle formation (Ruiz et al., 2015). After sulfuric acid nucleation particles formation, the organic compounds (volatile and low volatile) condense on it. Similarity, the soot was also influenced by this

nucleating agent (Schneider et al., 2005). Considering the limit of sample size of our study, it was difficult to calculate the influence of the fuel quality and the emission standards on PM constituents separately. In our future study, we will continue to focus on this complex issue.

**<u>Comments #10</u>**: Section 3.2: Given that there was only one China IV truck, how confident are the authors in their assessment that China IV trucks are better compared to the China III trucks. Similarly, is the China II truck any different than the China III trucks. Can the authors comment on how the small sample size could affect their conclusion?

**<u>Response</u>**: We appreciated this question. Actually, China IV truck is extremely rare, because few trucks could reach this emission standards in China. Therefore, we just found only one truck of China IV to conduct experimental. Furthermore, through comparing our results with references and assessing repeatability in the test results, we considered that our conclusions were credible. The detail explanations were added in the revised manuscript (**Page 9 line 16-23; Page 20 line 14-17**).

<u>**Comments #11</u>**: Section 3.3: Is the lack of a mass closure on the PM filter a result of using a quartz-fiber filter for gravimetric analysis?</u>

**<u>Response</u>**: Thanks. We have replied in the comment 7, using quartz-fiber filter was not the main reason caused poor mass closure. The main reasons might be distribution error from OC and EC, water effect and metal oxidation. As mentioned in the revised manuscript, the distribution error from OC and EC by using IMPROVE could highly affect the results of mass closure (**Page 20 line 13-14**). As shown in table 1, emission factors of OC was lower than those of n-alkanes for T3, which indicated that the OC content was underestimated. For example, emission factors of OC increased to 85.0 mg kg<sup>-1</sup> fuel, the mass closure would almost increase by 10%, correspondingly. For T2, the thick moisture was trapped in the filter, which could increase PM weighing error.

Species	Units	<b>T</b> 1	T2	Т3	T4	T5
OC	mg/kg fuel	22.4	32.7	0.64	153	10.3
EC	mg/kg fuel	337	3.61	200	37.4	186
OM	mg/kg fuel	35.9	52.3	1.02	245	16.5
Water soluble ions	mg/kg fuel	12.0	27.7	14.5	8.80	14.6
Elements	mg/kg fuel	0.77	2.95	2.15	6.34	6.62
N-alkanes	mg/kg fuel	7.19	1.79	4.72	26.2	4.87
PAHs	mg/kg fuel	0.05	0.11	0.17	2.94	0.06
Hopane and sterane	mg/kg fuel	0.01	0.03	0.05	0.12	0.02
РМ	mg/kg fuel	847	200	459	548	436
Mass balance	%	46	43	49	54	53

Table 1 Mass closure on the PM filter for trucks

<u>Comments #12</u>: Pry, Fluo etc.: Repeatedly, the authors have used abbreviated names to refer to various PM species. Using the full name of the species might improve readability.

**<u>Response</u>**: Thanks for the comment. The full name of the individual PAH was displayed in the revised manuscript (**Page 12 line 4-11**). But, considering the concise expression, we also used abbreviated names in the part of discussion.

<u>**Comments #13</u>**: Sections 3.3, 3.4 and 3.5: The authors have compared the PM composition data amongst the excavators and trucks and to literature data. However, it was hard for me to glean anything meaningful from all those comparisons and the ensuing discussion. I recommend that the authors spend some more time trying to make the interpretation more palatable to the reader.</u>

**<u>Response</u>**: We appreciate the review's comment. We also want to do it, but the maneuverability was poor. It is extremely difficult to collect literature data and compare with results obtained in this study, due to lacking of researches for characteristics of PM and its constituents by on-board tests, especially for non-road

diesel engine. Based on our purpose in this manuscript, we presented three parts for further discussion. In section 3.3, we tried to interpret difference in characteristics of PM emission between individual diesel vehicles tested in this study. In section 3.4, we tried to combine our results with those from other references to find some consensus. In section 3.5, through comparing the differences in characteristic of PM emission between excavators and trucks, we emphasized the PM emission difference of two types of vehicles. Following the reviewer's suggestion, we have made more interpretation between the comparison in the revised manuscript (**Page 18 line 6-17; Page 19 line 1-13; Page 25 line 1-11**).

**<u>Comments #14</u>**: Page 18, line 26 to page 19, line 2: The health relevant calculations, comparisons, and following discussion were too hard to follow and seemed like they were added to the manuscript as an afterthought.

**<u>Response</u>:** Thanks. The carcinogenic risks of PAHs emitted from trucks and excavators were the important indicators to evaluate emission situation for those two diesel vehicles. We have enhanced the expression in the revised manuscript (**Page 13 line 19-30; Page 25 line 1-11**).

# **References:**

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# Measurement on of PM and its chemical compositions for in real-world emissions from non-road and on-road diesel vehicles

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**Abstract.** With <u>increasing population the rapid growth in the number</u> of both non-road and on-road diesel vehicles, the adverse effects of <u>particulate matter (PM-)</u> and its <u>compositions constituents</u> (such as element<u>al</u> carbon (EC), <u>and pPolycyclic</u> aromatic hydrocarbons (PAHs))<sub>2</sub> on air quality and human health <u>get have been</u> <u>receiving more and more increasing attention</u>. However, <u>studies on the characteristics</u> of PM and its composition <u>whichs</u> emitted from diesel vehicles <u>are scarce</u>, particularly

those measured under performed in real-world conditions, are scarce. In this study, six excavators and five trucks; involving a wide-range of emissions standards and working in different operating operational modes, were tested to characterize PM constituents, of PM (including organic carbon (OC), EC, water soluble ions (WSIs), elements, and organic species such as PAHs, n-alkanes, hopanes, and steranes). The average emission factors of for PM (EF<sub>PM</sub>) for from excavators and trucks were 829  $\pm$ 806 and  $498_\pm 234$  mg kg<sup>-1</sup> fuel, respectively, which are are similar to values found in comparable with other studies. However, EF<sub>PM</sub> was significantly affected by fuel quality, operating operational mode,s and emission standards. High A significant correlation (R<sup>2</sup>=0.79, p<0.01) existed was found between the EF<sub>PM</sub> for excavators and the sulfur contents in fuel. The highest average EFPM under for working mode for excavators was 904\_±\_979 mg kg<sup>-1</sup> fuel, <u>due tobecause of the</u> high engine load required in this mode. - under this mode. From pre-stage 1 to stage 2 emission standards, the average EF<sub>PM</sub> for excavators with different emission standards decreased by 58%. Similarly, fFor trucks, the average non-highway  $EF_{PM}$  under non-highway condition (548  $\pm$  311 mg kg<sup>-1</sup> fuel) was higher than those under the highway condition- $EF_{PM}$  (497 ± 231 mg kg<sup>-1</sup> fuel). Meanwhile, the reductions when switching from China II and China III to China IV standards were 63.53.5% and 65.65.6%, respectively. Generally, the PM compositions emitted from excavators was dominated by OC (39.2% ± 21.0%) and ,-EC (33.3% ± 25.9%), and while PM for from trucks, PM-was dominated by EC (26.9%  $\pm$  20.8%), OC (9.89%  $\pm$  12%), and WSIs (4.67%  $\pm$  5.74%). Several differences of in compositions were observed among the various operating operational modes, emission standards, and fuel quality qualities. The average OC/EC ratios under idling and working modes for idling and working excavators were 3 and to 4 times higher than those in-for moving modes excavators. Although the EF<sub>PM</sub> for excavators and trucks was reduced by the constraint of with stringent emission standards, the fractions of elemental fractionss for excavators ranged from 0.49% to 3.03% from pre-stage 1 to stage  $2_{37}$  and the fraction of WSIs for the China IV truck was 6--fold higher than those from they were for the other trucks.

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Furthermore, <u>as</u> compared with the results from other diesel vehicles, wide ranges of

in the ratios of benzo[a]anthracene/(benzo[a]anthracene+chrysene) (0.26-0.86), indeno[1,2,3-cd]pyrene/(indeno[1,2,3-cd]pyrene +benzo[ghi]perylene) (0.20-1.0) and fluoranthene/(fluoranthene+pyrene)BaA/(BaA+Chry) (0.26-0.86), IcdP/(IcdP+BghiP) (0.20-1.0) and Flua/(Flua+Pry) (0.24-0.87) for excavators-were found for excavators, which may might be attributed to a result of the complex characteristics \_-of excavator

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operating operational modes for excavators. Although Similar fractions of the total-16 priority –PAHs (as identified by the U.S. Environmental Protection Agency) were found in the exhaust from the for excavators and trucks-were similar, the The equivalent concentrations of total-of benzo[a]pyrene, BaPeq that which was were used to evaluate the carcinogenic risk, was were 31 times higher for excavators than those they were for trucks. Therefore, implying that more attention should be paid to non-road vehicle's emissions.

# Keywords

Diesel vehicles; excavators; trucks; PM; chemical composition; impact-influential factors

# **Copyright statement**

We confirm that the material is original and has not been submitted elsewhere.

# 1. Introduction

Particulate matter (<u>PM</u>) emitted from diesel vehicles <u>have has</u> significantly adverse impacts <u>effects</u> on air <u>pollutionquality</u>, human health, and global climate change, and therefore <u>merit close should be examined examination closely</u> (Aggarwal et al., 2015, 2016). <u>Many Previous</u> studies have reported that diesel vehicles exhaust <u>was is a a</u> major source of <u>ambient PM emissions (D<sub>p</sub> <100 µm) emissions in ambient PM</u> (Oanh et al., 2010, Zhang et al., 2015a). For <u>exampleinstance</u>, it is reported that vehicle exhaust <u>was reported to contributed contribute</u> to almost 30% of <u>ambient PM</u> emissions in <u>ambient PM</u> in 9 cities <u>in of</u> China in 2015 (MEP 2016). The International Agency for Research on Cancer (IARC) found that exposure to diesel exhaust causes lung cancer (IARC 2012). <u>It is Adar et al.</u> (2015) reported that more

than 25 million children breathe polluted air on diesel school\_diesel school\_buses, which then causing causes a disproportionate occurrence of adverse respiratory disease health (Adar et al., 2015). Nearly 34% of element carbon (EC) emission emissions, a major contributor to current global warming and poor\_human health, accounts comes for nearly 34% from off-road diesel vehicles in the USA (USEPA 2015).

The populations numbers of on-road and non-road diesel vehicles have have increased considerably in China, and have contributed to especially for non-road diesel vehicles, causing severe emissions situation problems. On-road diesel vehicles can be classified as light-duty, medium-duty, and heavy-duty trucks. Non-road diesel vehicles mainly include construction machinery and agricultural equipment (MEP 2014). Airplanes, trains, and vessels are not included as non-road diesel vehicles in this study, because the primary fuels used for these vehicles does not include diesel. According to reports, tThe number of on-road diesel vehicles increased from 11.0 million in 2009 to 32.8 million in 2015,, and the number of while the number of non-road diesel vehicles increased from 20.6 million in 2006 to 33.6 million in 2012 (CCCMIY et al., 2013, MEP 2016). According to Based on the China vehicle environmental management annual report for 2015 (MEP 2016), 0.56 million tons of PM were emitted from on-road mobile sources and almost higher more than 90% of PM <u>came\_resulted</u> from on-road diesel vehicles emissions in 2015(Figure S1). However, pollutants emitted from non-road diesel vehicles should not be neglected. In 1991, The U.S. Environmental Protection Agency (USEPA) published a report indicating that PM emitted from non-road diesel vehicles was significantly higher than that emitted from on-road diesel vehicles (USEPA 1991). Wang et al. (2016) estimated the an emission inventory from for non-road equipment (including agricultural equipment, river/ocean-going vessels, locomotives, and commercial airplanes) and found that there are were 349 thousand tons Gg of PM emissions from non-road vehicles in China in-during 2012. Construction equipments was the largest source of PM emissions from non-road diesel vehicles. According to Zhang et al. (2010) reported that Pearl River Delta (PRD) region's PM emissions from

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construction instrumentsequipment —in the Pearl River Delta (PRD) region significantly\_accounted for 26.5% of the total emission from non-road vehicles in 2006.\_As aAn important type of non-road diesel vehicle, the number of construction instruments equipment in China increased from 1.97 million to 5.85 million between during 2006 to-and 2012 in China (CCCMIY 2013). According to Zhang et al. (2010) Pearl\_River\_Delta\_(PRD)\_region's\_PM\_emission\_from\_construction\_instruments significantly accounted for 26.5% of the total non-road vehicles in 2006. Furthermore, As as one of the most abundant types of construction\_instruments\_equipment (Figure S1), excavators contribute diesel consumption and PM emission from excavators were 7450\_and\_34.8\_thousand\_tons\_in\_2007almost\_\_65% of the PM\_emissions from construction equipment (Li et al., 2012).

In order to control diesel vehicles PM emissions pollution from diesel vehicles, China has beganstarted to implement emission standards early in early 2001 for light-duty diesel vehicles and heavy-duty diesel vehicles (SEPA et al., 2001). Those These standards have been were tightened in the subsequent 12 years, from the China I to China V standards-in 12 years. Although emission standards for on-road diesel vehicles were were formulated to in China V, insufficient diesel fuel quality slows their retards implementationg of emission standards (Yue et al., 2015). In addition, <u>The the</u> China IV emission standards for on-road diesel vehicles are not fully implemented until now. Compared with on-road diesel vehicles, tThe implementing implementation timeline for-of emission standards for non-road diesel vehicles has lagged behind that of the on-road diesel vehicles. China has implemented two emission standards for new non-road diesel engines, stage 1 and stage 2, in 2007 and 2009, respectively. FurthermoreHowever, this first implemented implementation time in China was 7 years later than implementation in the compared with the USA (USEPA 2003, SEPA et al., 2007). The pollution emissions limits for on-road and non-road diesel vehicles are given in Tables S1 and S2.

The fundamental work of  $EF_{PM}$  that is an important parameter in the compilation of emission inventories for on-road and non-road diesel vehicles in China. However, the foundational work towards quantifying  $EF_{PM}$  is relatively weak and contains large

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uncertainties (Huang et al., 2011). Recently, mMost of the EF<sub>PM</sub> from trucks have been measured using tunnel and dynamometer tests, which do not allow for evaluating influential factors for PM emissions from a single truck in real-world conditions (Alves et al., 2015b, Mancilla et al., 2012, Pio et al., 2013). used in emission inventory research came from developed countries. Several studies have measured PM emissions from trucks using on-board tests in real-world conditions (Wu et al., 2016, Wu et al., 2015, Zhang et al., 2015a). Because the EF<sub>PM</sub> emitted from trucks could change along with improved emission standards, data should be updated frequently (Huo et al., 2012). Wang et al. (2016) estimated emission inventory from non road equipment and suggested that real world measurements of emissions for non-road equipment are desperately needed. Along with increasingly serious environmental problems, PM emission from on road diesel vehicles has been taken seriously in China. There are considerable studies about on-road tests to study PM from on-road diesel vehicles. Wu et al. (2015) tested 17 in-use diesel trucks in Beijing using a portable emission measurement system (PEMS) and calculated the EF<sub>PM</sub> of those vehicles. Moreover, Zhang et al. (2015b) measured the real world PM emission factors from in-use HDDTs using PEMS. In addition, the data for EF<sub>PM</sub> emitted from non-road diesel vehicles on-in real-world conditions was-is scarce in China. To our knowledge, dynamometer test was used by most of studies to research non-road vehicle emission (Liang et al., 2005, Liu et al., 2015, Pietikainen et al., 2015). Liu et al. (2015) measured the PAH and nitro-PAH emission from non-road diesel engine, which was conducted utilizing the dynamometer test cycles required by U.S. EPA Tier 4 Final standards. In 2014, the Ministry of Environmental Protection of the People's Republic of China had issued "Technical guide for the preparation of a single source emission inventory of atmospheric fine particulate matter." However, no measured baseline for emission factors of PM from non-road vehicles, especially construction machinery (6 g km<sup>-1</sup> were predicted for uncontrolled standards) could be found in this technical guide (MEPPRC 2014). Until now, there was only one study in China by Fu et al. (2012), who provided EF<sub>PM</sub> for tested-12 excavators using portable emission measurement system (PEMS) PEMS to determine PM emission factors under for

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different working operational modes. HoweverO, n-board measurements need to be expanded to improve localization of  $EF_{PM}$  for non-road diesel vehicles in China as soon as possible, due tobecause of the complexity of real-world conditions, includingsuch as lagging diesel quality and changing emission standards, the on board measurements need to be expanded to improve localization of  $EF_{PM}$  for on-road and non-road diesel vehicles in China as soon as possible.

Analysis of the Chemical composition of PM constituents are is essentialimportant for studies of source apportionment, human health, and climate change studies. Primary PM emitted from diesel vehicles contains a variety of chemical compositions components, such as including organic carbon (OC), elemental carbon (EC), water soluble ions (WSIs), elements, and organic species (such as n-alkanes, polycyclic aromatic hydrocarbons (PAHs), hopane and sterane). Several previous field studies have have focused on chemical compositions of PM emitted from diesel vehicles. Zhang et al. (2015a) characterized PM compositions (OC, EC, WSIs and elements) emission emitted from heavy-duty diesel trucks (HDDTs). Wu et al. (2016) reported the detailed chemical composition of PM<sub>2.5</sub> emitted from China III and China IV diesel trucks, including the organic carbon (OC), elemental carbon (EC), water soluble ions (WSIs), and element contents, emitted from China III and China IV diesel trucks. In 2012, Fu et al. (2012) tested 12 excavators in the first on-board test for excavators in China, but to determine only optically-based PM emission factorsEF<sub>PM</sub> were given, which was the first on-board test for excavators in China. However, Therefore, the specific characteristics of PM emitted from diesel vehicles and its compositions emitted from diesel vehicles are still largely unknown, especially for organic matterscompounds, are lacking.

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In this study, PM and its composition emitted from on-road and non-road diesel vehicles were was measured in order to (I) test emission factors of PM for excavators and trucks under in real-world conditions; (II) identify influential impact factors of on the emitted PM and its compositions for non- and on road diesel vehicles; and (III) characterize chemical components of present in the emitted PM from excavators and trucks. Our The study results required substantial effort results of this study could

<u>and provide valuable information for use in the development of effective control</u> policies <u>and for reducinge PM emissions</u> from excavators and trucks.

# **Experimental**

# 2.1 Diesel vehicles and operational modes selection

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In this study, six excavators and five trucks were selected to cover a wide range of emission standards, manufacturers and engine loads. The dDetailed information of for the selected excavators and trucks is is shown in Table 1. The tested excavators were divided into two groups based on their emission standards: three pre stage 1 excavators and three stage 2 excavators. As shown in Figure S1S2, the annual productions of excavators have did not changed substantially much in between 2007 and -2009 (an increase from 70,000 to 85,000 excavators), during which stage 1 non-road vehicle emission standards was implemented, varying from 70,000 to 85,000 pieces of excavators. Therefore, excavators conducted produced with pre-stage 1 and stage 2 emission standards were chosen in-for this study. Based on China national standard (SEPA 2007), excavators can arebe divided into five types (0<P<8 kw; 8<P<18 kw; 18<P<37 kw; 37<P<75 kw; 75<P<130 kw; 130<P<560 kw) according to the their rated power rating(P). Thus, each type of eThe excavators were categorized for this study divided by by emission standards and were rated asincluded low (0-75 kw), medium (75-130 kw) and or high (130-560 kw) excavatorspower, which represent the low, medium and high power excavators, respectively. As a way to For reflecting the realactual operation use environments, three operating operational modes were selected for the excavators were selected –idling mode, moving mode and working mode, respectively. Further descriptions of these three modes can be found were listed in Fu et al. (2012). In this study In addition, consistent sampling times for the different modes were not strictly required in this study, as long as sufficient amounts of PM were collected to conduct the subsequent chemical analysis. The average average duration consumption sampling times in-during idling, moving, and working were were 41.7, 24.0, and 28.5 minutes, respectively.

For diesel trucks, there were three <u>Three</u> types of <u>diesel</u> trucks <u>were selected</u> according to emission standards, one China II <u>standards</u> truck, three China III

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standards trucks, and one China IV standards truck. Practicality The, just China III trucks contained included three trucks includingone each of light-duty, medium-duty, and heavy-duty diesel trucks. Based on the traffic control measuresrules and driving conditions of for on-road diesel trucks, different predesigned routes were chosen for different emission standards and size testing the trucks in Yantai, Shandong province in of China (Figure 1). Because different trucks drove on different routes, the selected routes in this study were divided into non-highway and highway categories. The selected routes chosen for China III and China IV light-duty trucks included arterial roadnon-highway 1-(non-highway 1), secondary road (non-highway 2) non-highway 2 and highway 1. The lengths of those three roads were 19, 35 and 17 km, respectively. The route chosen for the China II heavy-duty truck (yellow label) was special used for "yellow label car" (non highway 3) non-highway 3 which was 25 km. The routes chosen for China III medium-duty and heavy-duty trucks included non-highway 4-4 and highway 2. The lengths of those these -two roads were 47 and 23 km, respectively. The detailed velocity and road grade information for all of the tested routes were are shown in Figures S32 and S3S4. Although repeated tests were conducted for some vehicles, it should be noted that only one set of integral data was selected for further discussion, due to the incompleteness of some monitoring data (e.g. CO<sub>2</sub> and CO concentrations). As shown in Tables S3 and S4, the variability in test times for the same operational mode was considered acceptable. Some actions were required to reduce the uncertainty. For example, we combined sampling filters for the repeated experiments for vehicles T1 and T3 to carry out organic compound analysis.

# 2.2 On-board emission measurement system

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The on-board emission measurement system was <u>self-designed</u> and <u>combined</u> <u>constructed</u> in <u>our our-</u>laboratory (Figure 2). <u>The A</u> description of the <u>used-on-board</u> emissions test system was given <u>by in our previous previous study report</u> (Zhang et al., 2015b). Briefly, this system <u>has-consists of</u> two main <u>functional partscomponents,-:</u> <u>including a</u> Photon II <u>analyzer</u>, which was used to analyze the for-flue gas (HC, CO, CO<sub>2</sub>, SO<sub>2</sub>, <u>and NOx</u>) <u>analyzer</u>, and <u>a PM sampler sampling</u> system. The PM

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samplinger system consists consisted of a dilution system followed by, and five exhaust channels behind this dilution system. Two channels were connected with to PM samplers, and the others three were blocked. When Before sample sampling the PM\_-emitted from an excavators, the emission measurement system was put on a truck and connected to the excavators exhaust tube via by a stainless steel pipe. The system showed clear improvements over other on-board instruments, such as PEMSs and FPS4000 (Zheng et al., 2015), with better portability and the ability to collect filter samples for further chemical analysis in the laboratory. The results presented here include the first dataset from on-board measurement of non-road diesel vehicle exhaust in China.

#### 2.3 Chemical analysis

2.3.1 Fuel quality analysis

Fuel quality has a great-large\_effect on PM emissions from vehicles (Cui et al., 2016, Liang et al., 2005, Zhang et al., 2014). Due toSince the poor-fuel quality-used in excavators is often of poor quality, diesels for-was collected from each of the tested excavators were collected and <u>-analyzedtested</u>. The results for of fuel quality analysis were are shown given in Table 2. Comparing the diesel quality used in this study with the diesel quality standards for non-road vehicles (GB 252-2015) (SEPA et al., 2015), it was found that most of the sulfur contents in most in of the diesels used in this study (200-1100 ppm) were higher than those-allowed in-by\_GB 252-2015 (<350 ppm). Additionally, the sulfur content in the diesel used by E4 was 1100 ppm, which was much higher than those-that in diesel-used for-in the other excavators. Furthermore, the ash content of E4's-diesel used by E4 was 4.16%, about-therefore 420 times higher than the limit value given in by\_GB 252-2015.

25 2.3.2 PM and chemical composition analysis

Quartz-fiber filters were used for collecting the PM samples because the weight losses of these filters could be neglected through strict sampling processes, and quartz-fiber filters could be used for both the PM weight measurement and chemical analysis. The The quartz fiber filters were weighed before and after sampling to determine the collected PM mass concentrations of PM. Before each weighing, the

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filters were balanced at 25 \_<sup>o</sup>C and 40% relative humidity for 24 h. The-Each\_filters were-was\_weighed there-three\_times-before and after sampling to insure that the error for each measurement was as low as possible. WSIs were analyzed by-using ion chromatography (Dionex ICS3000, Dionex Ltd., America) following the method of Cui et al. (2016). Elements was-analysis was performed analyzed using inductively coupled plasma-coupled with-mass spectrometerry (ICP-MS,-; ELAN DRC II type, Perkin Elmer Ltd., Hong Kong).

Because the there was not enough organic matters on each filter was insufficient for quantification, we combined merged filters from different operating operational modes or driving routes filters for analysis analyzing for each diesel vehicles according based to on the proportion of samplingtested time during each mode or route. Quartz filter samples were spiked with internal standards (including acenaphthene- $d_{10}$ , benzo[a]anthracene- $d_{12}$ , pyrene- $d_{10}$ , coronene- $d_{12}$ , cholestane- $d_4$ , n-C15- $d_{32}$ , n-C20- $d_{42}$ , n-C24- $d_{50}$ , n-C30- $d_{58}$ , n-C32- $d_{66}$ , n-C36- $d_{74}$ ) were and ultrasonically extracted two times in 30 ml-mL of a 1:1 mixture of hexane and dichloromethane for 10 min. All extracts for from each sample were combined, filtered and concentrated to -approximately 0.5 mlmL.

Organic species including n-alkanes, PAHs, hopane and sterane were analyzed using GC-MS (Agilent 7890A GC-5975C MS)-equipped with a DB-5MS column (length 30 m ×\_i.d. 0.25 mm × thickness 0.25 µm). The GC operating program was as followingfollows: 60  $^{\circ}$ C with static time of for 4 min, increase 5  $^{\circ}$ C min<sup>-1</sup> to 150 °C with 2 min static time to 150 °C with static time of 2 min, then ramped-increase 3 °C min<sup>-1</sup> to 306  $^{\circ}$ C at rate of 3 °C min<sup>-4</sup>-with a 20 min static time of 20 min; \_\_and The GC \_\_conditions: \_had an injector temperature was of 290  $^{\circ}$ C, injector volume of injector was 2 µL, helium carrier gas-was helium, and gas flow rate of gas was-1.2 mlmL min<sup>-1</sup>. The electron impact (EI) mode at 70 eV and selected-ion-monitoring (SIM) mode were selected to determining-determine concentrations of PAHs, hopane, and sterane. For organic matters, the blank samples and recovery rates (66.7%-128% for five surrogates) were measured. The blank concentrations were subtracted from the sample concentrations.

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<u>The PM\_Chemical\_chemical\_</u>constituents of <u>PM</u>\_analyzed in this study are <u>werelisted as follows:</u> OC,-; EC,-; WSIs:  $SO_4^{2-}$ ,  $NO_3^{-}$ ,  $CI^{-}$ ,  $NH_4^{+}$ ; <u>Elementselements</u>: Na, Mg, K, Ca, Ti, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, Pb); n-alkanes: C12<u>to</u>-C40; <u>the</u> sixteen <u>USEPA</u> priority PAHs<u>of</u> - naphthalene (Nap), acenaphthylene (Acy), acenaphthene

5 (Ace), fluorine (Flu), phenanthrene (Phe), anthracene (Ant), fluoranthene (Fluo), pyrene (Pyr), benzo [a]anthracene (BaA), chrysene (Chry), benzo[b]fluoranthene (BbF), benzo[k]fluoranthene (BkF), benzo[a]pyrene (BaP), indeno[1,2,3-cd]pyrene (IcdP), dibenz[a,h]anthracene (DahA) and benzo[ghi]perylene (BghiP); Hopane and sterane: ABB-20R-C27-Cholestane (ABB), AAA-20S-C27-Cholestane (AAA),
10 17A(H)-22,29,30-Trisnorhopane (Tm), 17A(H)-21B(H)-30-Norhopane (30AB), and 17A(H)-21B(H)-Hopane (29AB).

#### 2.4. Data processing

2.4.1 Fueled-based emission factors

Fueled-based emission factors were calculated by <u>using</u> the carbon\_-mass balance formula-:

$$EF_i = \frac{\Delta X_i}{\Delta CO_2} \cdot \frac{M_i}{M_{CO_2}} \cdot EF_{CO_2} \tag{1}$$

Where where  $EF_i$  and  $EF_{co_2}$  (g kg<sup>-1</sup> fuel) are the emission factors for species *i* and CO<sub>2</sub>, respectively,  $\Delta X_i$  and  $\Delta CO_2$  (mol m<sup>-3</sup>) are the background-corrected concentrations of species *i* and CO<sub>2</sub>, respectively, and  $M_i$  and  $M_{co_2}$  (g mol<sup>-1</sup>) represent the molecular weights of species *i* and CO<sub>2</sub>, respectively.

<u>The CO<sub>2</sub> emission factors ( $EF_{CO_2}$ ) were calculated as:</u>

The flue gas emissions were calculated as:

$$EF_{CO_2} = R_{FG} \cdot c(CO_2) \cdot M_{CO_2} \tag{2}$$

Where where  $c(CO_2)$  (mol m<sup>-3</sup>) is the molar concentration of CO<sub>2</sub>, and  $R_{FG}$  (m<sup>3</sup> kg<sup>-1</sup> fuel) represents the flue gas emission rate.

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$$R_{FG} = \frac{c_F}{c(c_{CO}) + c(c_{CO_2}) + c(c_{PM})}$$
(3)

Where where  $C_F$  (g C kg<sup>-1</sup> fuel) represents the mass of carbon in 1 kg of diesel fuel,

and  $c(C_{CO})$ ,  $c(C_{CO_2})$ , and  $c(C_{PM})$  (g C m<sup>-3</sup>) represent the <u>flue gas</u> mass concentrations of carbon as CO, CO<sub>2</sub>, and PM, respectively in the flue gas, respectively.

2.4.2 Average fuel-based emission factors for excavators and trucks

The average fuel-based emission factor for each excavator under-in each relevant different operating operational modes was calculated by followsas:

$$EF_{i,j} = \sum EF_{i,j,g} \times P_{j,g}$$
(4)

Where where  $EF_{i,j}$  (g kg<sup>-1</sup> fuel) is the average emission factor of species *i* for from excavator j,  $EF_{i,j,g}$  (g kg<sup>-1</sup> fuel) represents is the emission factor of species i for from excavator j under in mode g-mode, and  $P_{j,g}$  (%) is the proportion of activity time (Fu et al., 2012) for excavator *j* in mode under g-mode.

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The average fuel-based emission factor for each truck under-in different driving conditions <u>was</u> calculated <u>by followsas</u>:

$$\mathrm{EF}_{i,j} = \sum EF_{i,j,s} \times P_{j,s} \tag{5}$$

Where where  $EF_{i,i}$  (g kg<sup>-1</sup> fuel) is the average emission factor of for species *i* for from excavator truck j,  $EF_{i,j,s}$  (g kg<sup>-1</sup> fuel) represents is the emission factor of species i for excavator truck j under in driving condition s, condition, and  $P_{j,s}$  (%) is the proportion of activity time for truck j under in driving condition s-condition.

2.4.3 Benzo[a]pyrene equivalent concentration (BaPeq)

The various PAHs have a wide range of carcinogenic risks. Therefore, it is not accurate to evaluate the harmful effects of PAHs on human health using the total combined mass concentration. BaPeq is typically used to evaluate the carcinogenic risks associated with individual PAH (Mirante et al., 2013). The BaPeq was calculated <u>as:</u>

$$BaP_{eq} = \sum PAH_i \times PEF$$
(6)

- where  $PAH_i$  is the measured concentration of an individual PAH for excavator *i*, and PEF is Because of different carcinogenic risks for each PAH, the BaPeq for parent PAHs were given. The BaPeq was calculated by multiplication of the measured concentrations by the respective the potency equivalent equivalence factor (PEF)

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(Mirante et al., 2013). The PEF values were for that PAH obtained from Wang et al. (2008).

# 3. Results and discussion

# 3.1 Particulate matter <u>f</u>Fuel-based emission factors <u>of PM</u> for <u>in</u> excavator

5 <u>exhaust</u>s

The EF<sub>PM</sub> values for excavators exhaust are are presented illustrated in Figure 3, with the detailed information shown-given in Table S1S5. The maximum fuel-based PM fuel based emission factor was almost 37 times higher than the minimum-under different operating modes for different vehicles. In general, the average EF<sub>PM</sub> for different excavators ranged from 96.5 to 2323 mg kg<sup>-1</sup> fuel, with an average of 829  $\pm$  806 mg kg<sup>-1</sup> fuel. The EF<sub>PM</sub> values of excavators reported by Fu et al. (2012) were within the range of EF<sub>PM</sub> in this study but in a narrower range. The reason for the more-widely range ind EF<sub>PM</sub> values in this studyhere maycould be that the be due to the difference in the selection of excavators emission standards. The excavators selected by Fu et al. (2012)\_included stage 1 and stage 2 emission standards. Therefore, the range of EF<sub>PM</sub> in this study may reflect the general excavator's PM emission situation in China.

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 $EF_{PM}$  could beis affected by many factors. In this study, the  $EF_{PM}$  range for excavators with different power ratings was 96.5 (35 kw) to 2323 (110 kw) mg kg<sup>-1</sup> fuel, but the correlations between  $EF_{PM}$  and engine power (See Figure S5) were weak. Conversely, fuel quality, emission standard and operational mode significantly affected the  $EF_{PM}$ . Some variation characteristics about the  $EF_{PM}$  values due to the different fuel quality, emission standards and operating modes were summarized as follows. Firstly, fFuel quality hashad a great large impact on  $EF_{PM}$  for the excavators. As shown from in Figure 3, a high-significant correlation ( $R^2$ =\_0.79, P\_<\_0.01) was found between the average emission factors for excavators and the fuel sulfur contents in fuel, which is is consistent with the results studied from reported by Yu et al. (2007). The Secondly,  $EF_{PM}$  also decreased with stricter enhancing of emission standards for the excavators. The  $EF_{PM}$  measured  $EF_{PM}$ -for pre-stage 1 excavators under during

idling, moving and working conditions were  $914_{\pm}393$  mg kg<sup>-1</sup> fuel,  $609_{\pm}38$  mg kg<sup>-1</sup> fuel and  $1258_{\pm}1295$  mg kg<sup>-1</sup> fuel, respectively, whereas. The EF<sub>PM</sub> for stage 2 excavators under idling, moving and working conditions for stage 2, they were 243  $\pm 236$  mg kg<sup>-1</sup> fuel,  $165_{\pm}144$  mg kg<sup>-1</sup> fuel and  $551_{\pm}587$  mg kg<sup>-1</sup> fuel, respectively.

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Compared to pre-stage 1, <u>The EF<sub>PM</sub> of the stage 2 excavators were reduced by 73%</u>, 73% and 56% in from the pre-stage 1 values under idling, moving and working modes, respectively. The average  $EF_{PM}$  for excavators of different emission standards decreased by 58% from pre-stage 1 to stage 2, suggesting the effectiveness of the emissions control policy.

Lastly, <u>The</u> EF<sub>PM</sub> varied sharply between different operating operational modes for the various excavators. Specifically, excavators under working excavators modes have had the highest EF<sub>PM</sub>, which is-was higher-more than double the values for other operating modes idling and moving excavators by more than 1 fold. The average EF<sub>PM</sub> for excavators under different driving conditions were  $578_{\pm}467$ -mg kg<sup>-1</sup> fuel (while idling),  $343_{\pm}264_{-}$ mg kg<sup>-1</sup> fuel while (moving) and  $904_{\pm}979$  mg kg<sup>-1</sup> fuel while (working), respectively. The Working mode produced the highest average EF<sub>PMs</sub> which under working mode might be because the attributed to higher engine load, which causes caused a lower air-fuel ratios and thus then prompted the PM production.

# 20 3.2 Particulate matter <u>fF</u>uel<u>ed</u>-based emission factors <u>of PM</u> for trucks

The  $EF_{PM}$  for all measured trucks under different driving patterns-varied from 176 mg kg<sup>-1</sup> fuel-to 951 mg kg<sup>-1</sup> fuel. The maximum  $EF_{PM}$  for trucks was three times more than the minimum. There were just tripled in PM emission factors for trucks from maximum to minimum. The average  $EF_{PM}$  for the tested diesel trucks was  $498_{\pm}234$ 

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Besides, The <u>The</u> average  $EF_{PM}$  of diesel trucks <u>for</u> <u>with</u> different emission standards <u>and</u>, vehicle sizes <u>and while using different</u> driving patterns were provided under real-world conditions (Figure 4). The measured  $EF_{PM}$  for China II, China III,

mg kg<sup>-1</sup> fuel. In comparison, Wu et al. (2016) reported an average  $EF_{PM}$  for diesel trucks of 427 (95.6-1147 mg kg<sup>-1</sup> fuel) mg kg<sup>-1</sup> fuel (95.6-1147 mg kg<sup>-1</sup> fuel) and it is, which was within similar the same to the range foras our results.

and China IV diesel trucks varied from  $200 \text{--mg kg}^{-1}$ -fuel to 548 mg kg<sup>-1</sup> fuel. The EF<sub>PM</sub> for the China II truck measured in this study was lower than the results obtained from<u>reported by</u> Liu et al. (910-2100 mg kg<sup>-1</sup>-fuel) (2009) (910-2100 mg kg<sup>-1</sup> fuel). The average EF<sub>PM</sub> for light-duty, medium-duty and heavy-duty diesel trucks were 524  $\pm$  457 mg kg<sup>-1</sup> fuel, 459, mg kg<sup>-1</sup> fuel and 492 mg kg<sup>-1</sup> fuel, respectively. The average EF<sub>PM</sub> of for trucks under non-highway and highway driving patterns were 548\_±\_311 mg kg<sup>-1</sup>-fuel and 497\_ $\pm$ 231 mg kg<sup>-1</sup> fuel, respectively. As shown in Figure 4\_-shows, reductions of in the measured EF<sub>PM</sub> between the for-China II truck to and China IV trucks and from between the China III truck to and China IV trucks in EF<sub>PM</sub>-were 63.53.5% and 65.65.6%, which indicated indicating that improvements of in the emission standards for diesel trucks significantly decreased could significantly reduce PM emissions. It should be noticed Of particular note was that the EF<sub>PM</sub> for China III and light-duty diesel trucks were higher than the values for the other corresponding trucks. The reason may might be attributed to a result of poor driving conditions, that include i.e., low average speed and more volatilehighly varied speed in speed for those trucks (Figures S2-S3 and Figure S3S4). The Same same tendency is apparent in could be seen from Figure 4, that with diesel trucks emitted emitting more PM while driving on the under-non-highway condition (average speed of: 28.5 km h<sup>-1</sup>) than while driving on the those under highway condition (average speed of :- 60.7 km h<sup>-1</sup>). Furthermore, tThe road grade further was an another aspect effected affected the EF<sub>PM</sub> of the on-road diesel trucks. For example, the EF<sub>PM</sub> for T5 under-driving on the highway road was lower than those for T1 driving on the highway, because of bigger lower road grade for T5 under highway road than those for T1 (Figure S3S4).

# 3.3 Particulate matter composition for individual diesel vehicles

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Four <u>types of constituents</u> were considered for reconstituting PM mass, in this study: (1) organic matter, which was estimated by multiplying the corrected OC <u>byby</u> a factor of 1.6 (Almeida et al., 2006); (2) EC; (3) <u>water soluble ionsWSIs; and</u> (4) elements. The reconstituted mass<u>es</u> for <u>each the</u> excavator sampler<u>s</u> <u>was were</u> 74.7-123% of <u>the</u> measured mass, while <u>the</u> reconstituted mass<u>es</u> for <u>the</u> diesel truck <u>sampler samples was were</u> only 43.2-54.4% of <u>the</u> measured mass (Figure 5). <u>Except</u>

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for<u>In addition to</u> uncalculated components, this discrepancy <u>may might</u> be attributed to uncertainties in the weighing process due to a distribution error from OC and EC, moisture effects, or metal oxidation (Dai et al., 2015).

3.3.1 Particulate matter composition for individual excavator

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The chemical composition of PM for each excavator was is shown in Figure 5 and Table S2S6. For each excavator, the carbonaceous component (OM+EC) was the dominant species, which is consisted consistent with results from a previous study by Liu et al. (2005)from a non-road diesel generator that had found, who reported that the proportions of organic and element carbonOC and EC in PM ranged from 70.2% to 90.61% (Liu et al., 2005). BBecause ecause the OC/EC ratio is also used to identify the source of atmospheric particulate pollution, deeper-further assessment was performed on the discussion about OC/EC ratios under in different operating operational modes for each excavator was conducted (Figure 6). The average OC/EC ratios for <u>during</u> idling, moving, and working modes were 1.57, 0.57, and 2.38, respectively. The OC/EC ratio under-during idling was higher-greater than 1 because soot hardly generated at low temperatures <u>hardly</u> and fuel-rich zone. These results were consistent those in , which is similar to the research done by Liu et al. (2005). Furthermore, Liu et al. (2005) reported that the OC/EC ratios decreased with an increasing increase in the load for non-road engines load. However, this trend was not couldn't be observed in this study. The OC/EC ratio was 2.38 under while working mode, and increasing increased again with load increasing load, which was consistent with the results reported by from Zhang et al. (2014). As shown in Figure 6, the differences between Large-OC/EC ratios differences for excavators under different excavator operating operational modes were profoundseen in Figure 6, which may and could be caused affected by a number of factors, (such asincluding transient working conditions, diesel sulfur content, and extensive OC sources for OC) (Cocker et al., 2004, Liu et al., 2005, Ruiz et al., 2015).

As shown from <u>in</u> Figure 5, <u>WISs-WSIs</u> and elements fractions ranged from 0.335% to 1.21% and from 0.163% to 7.50%, <u>respectively</u>, for all excavators. The total-<u>sum</u> proportion of <u>WISs-WSIs</u> and elements to PM was the highest in excavator E6,

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followed by excavator E1. Generally, the total-sum proportion of WISS-WSIs and elements to PM in exhaust from excavator E1-E6 was 4 to 14 times higher than the corresponding proportions in exhaust from the other excavators. Sulfate and nitrate were the main WISs-WSIs (79.1%-90.0% of WISsWSIs) for almost all of the excavators, except for E1, in which while the proportion of Cl<sup>-</sup> of WISs for excavator E1-(67.2%) was the highest (Table <u>\$2\$6</u>). Fe, Ca, Na, Mg, and K were relatively dominant in elements, but except for E4 excavator, Fe, Zn, and Cu were the most abundant elements. Wang et al. (2003) reported that the concentrations of the crustal elements Fe, Ca, and Mg that account for 50% of the total elements in diesel fuel were significantly higher than anthropogenic elements emitted from diesel vehicle engines, which is consistent with the results from our study. Similarly, diesel was the dominant source for these elements because the sampling tube was placed directly on the tailpipe. It may be attributed to that Zn is known from oil additives and Cu usually emitted from wear debris (Lin et al., 2015, Wu et al., 2015). Table 1 and Table 2 showed that excavator E4 produced in 2004 and the diesel quality used was poor, resulting in high Zn and Cu emission. The abundance of Fe, Zn, and Cu in the exhaust of E4 could have been affected by E4 being used to transport ironstone. BesidesIn addition, the elements fractions for the two excavators produced manufactured in 2013 (E1 (1.42% for E1), \_\_and , E6 (7.50% for E6) and 5.66 mg kg<sup>-1</sup> for E1 + E6) were higher than those for the other excavators (a total of 4.02 mg kg<sup>-1</sup> for E2, E3, E4, and E5), which may. This indicates that elements emissions was were deteriorating and more stringent control technology should be developed to avoid the total elements adverse health effects from the total elements composition of PM in the exhaust.

- 25 the excavators were ranged from 3.6% to 9.6%, from 0.03% to 0.24%, and from 0.001% to 0.09% for excavators, respectively. Liang et al. (2005) characterized diesel particulate matter emitted from non-road engines using a dynamometer test and found that n-alkanes accounted for 0.83% of PM, which was lower than the proportion found in results obtained from this study., The main reasons are the possibly because

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they used low sulfur diesel fuel and different sampling methodsused in Liang's study

In addition, tThe n-alkanes, PAHs, hopane and steranes fractions in exhaust from

and different methods used in obtained the PM. Contrary In contrast to what was observed the fractions for of WISs WSIs and elements, Figure 5 showed that the fractions of n-alkanes, hopane and steranes fractions-were the highest in excavator E4, while the fractions of PAHs fraction was the highest in for the exhaust from excavator E3. In a comparison of the fuel quality between E3 and E4, E4 had poorer diesel quality, which might be the reason for high n-alkane, hopane and steranes concentrations. Similarly, It was said by Rogge et al. (1993) found that n-alkanes, PAHs, hopane and steranes are were mostly derived from the incomplete combustion of fuel and lubricant oil. - By comparing the differences between fuel quality and performance of excavators, it could be deduced that n alkanes, hopane and steranes were influenced by fuel quality and However, we speculated that -PAHs was were affected by combustion conditions (i.e., combustion temperature) in this study, because E3, with the stage standard, had better performance and superior fuel quality. PAH\_s-isomer ratios have have been widely used to distinguish conduct the source apportionment in-for environmental receptors (such as sediments) (Liu et al., 2012). Yunker et al. (2002) found that the ratios of the principal masses of PAH 178, 202, 228 and 276-parent PAHs have had a the best potential to distinguish between natural and anthropogenic sources. For the excavators, the ratios of ratios of for BaA/(BaA+Chry), IcdP/(IcdP+BghiP), and Flua/(Flua+Pry) were 0.26-0.86, 0.20-1.0, and 0.24-0.87, <u>respectively</u>, with averages of  $0.47 \pm 0.27$ ,  $0.44 \pm 0.38$ , and  $0.48 \pm 0.27$ , respectively (Figure 7). The average ratios of PAHs for-in excavator exhausts obtained in this study were similar with to that those from Liu et al. (2015) reported for non-road diesel engines. The E4 excavators had a clear showed an obvious difference in the ratios of BaA/(BaA+Chry), IcdP/(IcdP+BghiP), and Flua/(Flua+Pry) to those from between the other excavators tested in this study. The isomer ratios of BaA/(BaA+Chry), IcdP/(IcdP+BghiP) and Flua/(Flua+Pry) for E4 were 0.86, 1.0 and 0.87, respectively, and it were. These were different with from the ranges for fuel combustion defined by Yunkers et al. (2002).- The ratios of PAHs emitted from diesel vehicles reported by Yunkers et al. (2002) mainly referred to those from on-road diesel vehicles. However, the operating operational modes and fuel quality

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for non-road diesel vehicles <u>are\_are\_</u>more complicated than those <u>from\_for\_</u>on-road diesel vehicles. Therefore, <u>the\_results\_in</u> this study <u>could\_giveprovides</u> references <u>values</u> for <u>the</u> isomer ratios of PAHs <u>for-in\_</u>non-road diesel vehicle<u>exhausts</u>.

3.3.2 Particulate matter composition for individual diesel trucks

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For diesel trucks, the total carbonaceous composition (OM+EC) were accounted for 44.0% (TE1), 27.9% (E2T2), 43.9% (E3T3), 51.6% (E4T4) and 46.3% (E5T5) of PM, which is are all lower than those the values reported in previous studies (Chow et al., 2011, Wu et al., 2015) because of . The reason may be mainly attributed towas -the different OC and EC detection methods used in our studyfor organic carbon and elements carbon. Cheng et al. (2011) collected 333 PM<sub>2.5</sub> samples and analyzed OC and EC by two common thermal-optical methods (NIOSH and IMPROVE). They-and found that NIOSH-defined EC was lower (up to 80%) than that defined by IMPROVE. The <u>IMPROVE</u> thermal-optical method <u>was</u> used in this study was IMPROVE, which would cause under valuation make content of OC-under evaluated.- Except for the T2 and T4 trucks, Almost all of the OC/EC ratios for diesel trucks under different driving conditions calculated in this study were lower than 1, which was is consistent with the conclusions from previous studies (Figure 6), except for the T2 and T4 trucks. The OC/EC ratios for T2 under during highway and non-highway driving conditions were 5.64 and 15.5, respectively,... This result may be attributed to which may be a result of the China IV emission standard for. T2 (China IV). Alves et al. (2015b) reported that modern diesel passenger cars (Euro 4 and Euro 5) exhibit have high OC/EC ratios. As shown in Figure S3, the driving speed for T4 was zero for the first 500 seconds. Cheng et al. (2015) reported that the OC/EC ratios were substantially above 1 while idling or with low load. Furthermore Therefore, the OC/EC ratio for T4 while driving on the under-non-highway condition-was 4.10, which may might be have been caused by the low driving speed. Cheng et al. (2015) reported that the OC/EC ratios were substantially above unity at idling and low load. As shown from Figure S2, the driving speed for T4 was zero in 500 seconds before driving.

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The <u>Sum sum of WISs WSIs</u> and elements fractions were lower than 5% for <u>the</u>

exhaust from almost all of the diesel trucks, except for that from T2-truck, which is consistent with the results gained from *in* Zhang et al. (2015a).  $SO_4^{2-}$  was the most abundant ions for trucks T2 and T5, while  $NO_{\overline{3}}$  was the most abundant ions for trucks T1, T3 and T4. For T2-diesel truck, WISs-WSIs (13.8%) was were the most significant component of PM, after followed by OC, and which it was higher by a factor of 4 to 10 times higher than it was for the those in other trucks (Table S2S6). T2 truck is was a China IV diesel vehicle and with well-controlled combustion conditions caused leading to more water emissions, which accelerates the translation transformation from the gas phase to WISs WSIs (such ase.g., SO2 translate the transformation of SO<sub>2</sub> to SO<sub>4</sub><sup>2-</sup>). As we cancan be seen from in Table S2S6, Fe was the most abundant element for trucks T1, T3 and T5, while Ca was the most abundant the most abundant element for trucks T1T2, T3, and T4. The total element fraction of T2 (China IV) was 16 times higher than that of T1 exhaust (China III) Compared with elements fractions in T2 (China IV) and T1 (China III) trucks, fractions changed from 0.09% (T2) to 1.5% (T1). Although the PM emission factors <u>EF<sub>PM</sub></u> for diesel trucks decreased with stricter emission standards, the WISs-WSIs and elements contents increasing increased along with promoting the emission standards for diesel trucks. In consideration of Because acid rain is causing caused by sulfate and nitrate and adverse health effects are caused by elements, great attention should needs to be pay paid to this phenomenon.

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The n-alkanes, PAHs, hopane and steranes fractions ranged fromwere 0.85%-to 4.78%, from 0.01% to 0.54% and from 0.002% to 0.024%, for the trucks, respectively. As shown in Table S2S6, C20 was the most abundant species in n-alkanes for truckin exhaust from T1, T2 and T4, while C19 was the most abundant n-alkane in exhaust from species for truck T3 and T5. For Of the PAHs, the most notable-abundant species was Pyrenepyrene, which was substantially higher than all other PAHs for all trucks. The proportions of nN-alkanes, PAHs, hopane and steranes accounted for the highest proportions to of PM were highest for the exhaust from truck T3, which and may might be affected by many factors, such asincluding differences in the engine rate power rating, complex reactions in the engine

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(combustion process and pyrolysis reactions related to temperature, humidity, etc.), and driving conditions. As shown from in Figure 7, scatters of the isomer ratios for diesel trucks were covered from were 0.28 to -0.35 for BaA/(BaA+Chry), from 0.08 to -0.22 for IcdP/(IcdP+BghiP) and from 0.08 to -0.39 for Fluas/(Flua+Pry), with averages of  $0.31 \pm 0.03$ ,  $0.15 \pm 0.06$  and  $0.23 \pm 0.12$ , respectively. There These were are similar to the results from reported by Schauer et al. (1999).

# 3.4 Average chemical constituent composition of PM emitted from diesel vehicles

3.4.1 Average chemical composition constituent of PM for in excavator exhausts

The average PM chemical component compositions of PM for excavator exhausts was are listed in Table 3. It appeared that cCarbonaceous matter was was the 10 dominant component and accounted for 72.5% of the PM for excavators, whereas OC was the most abundant species (39.2%) for PM. Total The total element fraction was the second largest group and contributed 1.76% of PM. For Of the elements, the emissions was were obviously dominated by Fe at which accounted for 46.3% of the elements. In addition, Table 3 showed that the proportion of n-alkanes in PM for from excavator exhausts (5.14%) was higher than the proportions of the those for other organic matter typess (PAHs were :- 0.098% while ;- hopane and sterane were :-0.026%), and C20 and / C19 were was the most abundant maximum carbon in n-alkanes. For-the parent PAHs, the emissions were dominated by Pry and Fluo,

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20 followed by Nap and Chry.

> To compare our results with other studies, Table 3 summarizess the average source profiles of PM for-in excavator exhaust ass derived in this study, as well as those-ones previously reported by others for comparison. As shown in Table 3, the average fraction of total carbonaceous components for the excavators tested in this study are was consistent with thatose for a measured from marine engine, while the element fraction of elements was lower than that for a marine engine (Sippula et al., 2014). Iron oxide is recognized as a catalyst and can promote soot burnout during combustion processes (Kasper et al., 1999). It is said that oxidation of soot was enhanced during increasing of transition metals for diesel engines (Kasper et al., 1999). The EC fraction of <u>PM in the PM \_\_for excavator exhausts</u> was higher than that

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those from-reported by Sippula et al. (2014), which may might be attributed be the result of a to lower metal fraction in the excavators used for their study. Comparing results from this study with other references showed that tThe proportions of n-alkanes measured in this study is were significantly higher than those emitted from a marine engine (4-fold) and non-road generators (6-fold) in another study (Liang et al., 2005). (Liang et al., 2005), which could be the result of \_\_\_\_\_\_The reason may be attributed to different contents of aliphatic compounds existing in the diesel fuels used for those non road vehicles (Sippula et al., 2014). For the marine engine and non-road generators, C22 and C17 were the most abundant n-alkane\_species. \_\_\_\_\_\_in n-alkanes, respectively. PAHs emission was were dominanted by Phe for a marine engine and Fluo for non-generators, which was different with from the result obtained from for the excavators. This could indicate that the PM emitted from different types of non-road diesel vehicles has various varying source profiles because of based on the diverse-operational conditions.

# 3.4.2 Average source profile of PM for trucks

As shown in Table 3, average emission of PM from trucks was dominated by carbonaceous <u>matter</u> (36.8%)<sub>27</sub> and followed by <u>WISs-WSIs</u> (4.67%) and elements (0.941%). For individual species, sulfate and nitrate were the most abundant in water soluble ionsWSIs, and Fe was the most abundant dominated in elements. Moreover, for organic matters, the average proportions of n-alkanes, PAHs, hopanes and steranes was-were 1.73%, 0.130%, and 0.011%, respectively. C20 was the maximum-most abundant earbon in n-alkanes, and the emission of PAHs was were dominated by Pry. In comparison, emission of total carbon emissions from in this study was-were lower than those in previous studies, whereas, the WISs-WSIs and elements fractions were relatively higher than results obtained from other research groups (Alves et al., 2015a, Cui et al., 2016, Schauer et al., 1999, Wu et al., 2016). There are sSeveral reasons factors could have influenced these differing be used to explain the results, including fuel quality, driving condition, parameters of engine parameters (fuel injector design) and experimental

As shown in Table 3, Fe was <u>the\_</u>dominant <u>in\_</u>elements <u>from\_in\_</u>results\_measured <u>bystudies using</u> on-road tests and tunnels, which <u>was\_is\_</u>similar <u>with\_to\_</u>our results, while Zn and Na were dominant in elements from results obtained by <u>a</u>\_dynamometer. Therefore, <u>the results obtained from real world (on-road tests and tunnels) were</u> different from those obtained in a laboratory. on road test and tunnels measured in real world would reflect real PM emission better. For organic matters, the proportion of PAHs, hopane and sterane to PM were consistent with the results from Schauer et al. (1999) and Cui et al. (2016). <u>Similar\_withAs in</u> this study, the <u>maximum</u> earbonmost abundant in n-alkanes was C20 <u>as</u> measured by Schauer et al. (2016). <u>Thus, the</u> average profile of PM for on-road diesel trucks <u>was-is</u> relatively stable <u>and consistent</u> <u>across studies</u>.

# 3.5 <u>Comparing average s</u>ource profiles <u>comparison from for</u> excavators <del>with</del> those from and trucks

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Compared with the average  $EF_{PM}$  of excavators and diesel trucks obtained in this study, a<u>A</u>verage  $EF_{PM}$  for excavators (836\_±\_801 mg kg<sup>-1</sup> fuel) was higher than those\_that\_for diesel trucks (498\_±\_234 mg kg<sup>-1</sup> fuel). The\_This\_result was\_is understandable because state\_the\_operations for excavators were\_are\_more transient than those for trucks. Sarvi et al. (2010) reported that particulate matter emission emitted\_from diesel engines was typically low during steady state operation. Although the\_average  $EF_{PM}$  of excavators was\_higher than that emitted\_fromof\_trucks, the average  $EF_{PM}$  of the\_stage 2 excavators was 477 mg kg<sup>-1</sup> fuel, which was lower than those for the\_that\_emitted\_by\_China\_II and China\_III trucks. Thus, appropriate regulations formulated for non-road diesel vehicles could improve the<u>ir</u> PM emissions situation.

When we compared the average percentages of chemical components in PM for excavators with those for trucks, we found that there were someseveral differences were foundbetween excavators and trucks. In general, the carbonaceous composition (95.9%) and elements (1.76%) fractions for excavators were higher than those for diesel trucks (42.8% for carbonaceous composition and 0.94%, for

elements respectively). As shown in Figure 8, the the structures of different ring PAHs in the exhaust from excavators and trucks varied sharply, especially for 5 and 6-ring PAHsBaPed levels for excavators and trucks were absolutely difference, although the average percentage of total PAHs average percentages of in the PM were consistent between the excavators and trucks. Almost all of the parent PAHs's BaPeer calculated in this study for trucks and excavators were higher than the datum from WHO that concentration caused 1/10000 risk of carcinogenic. Due to their lipophicity, high molecular weight (5+6 ring) PAHs are considered to be more harmful to human health than the other PAHs. For further distinction, BaPeq was used in this study. The range of total BaPeq for trucks was 5.32 (T5) to 155 (T3) ng m<sup>-3</sup>, while for excavators, the range of total BaPeq was 38.3 (E1) to 3637 (E4) ng m<sup>-3</sup>. Moreover, the total- average of BaPeq for the excavators was 31 fold times larger than that of those for the diesel trucks. Almost all of the parent PAHs's BaPeq values calculated in this study for trucks and excavators were higher than the datumconcentrations from WHO that concentration caused 1/10000 of the risk of carcinogenic risk, according to the World Health Organization (WHO). Due to the some adverse environmental effects and health hazards caused by for carbonaceous composition, elements, and PAHs, the PM emissions from excavators require urgent should be controlled urgently.

# Conclusions

This study report<u>eds</u> the characteristics of PM source profiles for excavators and the EF<sub>PM</sub>\_trucks. Above all, PM emission factors values for exhaust emitted from excavators and trucks with different emission standards and used <u>under in</u> different operating operational modes, emission standards andor road conditions were obtained. <u>The The EF<sub>PM</sub> for different excavators ranged from 96.5 to 2323 mg kg<sup>-1</sup> fuel, with an</u> average of 810 mg kg<sup>-1</sup> fuel and <u>showed a high correlation</u> (R<sup>2</sup>=0.79, P<0.01) with the fuel sulfur contents in the fuel. The highest average EF<sub>PM</sub> for excavators that are in working mode-(904\_±\_979 mg kg<sup>-1</sup> fuel) might be attributed to the result of higher engine load, which caused causing lower air-fuel ratios. The average EF<sub>PM</sub> for the tested diesel trucks of with different emission standards and vehicle sizes under different driving conditions was  $498_\pm 234$  mg kg<sup>-1</sup> fuel. The average EF<sub>PM</sub> for

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excavators with different emission standards excavators decreased by 58% from pre-stage 1 to stage 2. Moreover, the reductions in EF<sub>PM</sub> from the China II truck-to the China IV truck and from the China III truck to the China IV truck in EF<sub>PM</sub>-were 63.5% and 65.6%, respectively., Those lindicateing that improvements of to the emission standards for diesel trucks and excavators have significantly decreased PM emissions significantly. It should be noticed that the  $EF_{PM}$  for China III and light-duty diesel trucks were higher than those for the other corresponding trucks., which The could be a result of reasons may be attributed to poor driving conditions that included a low average and highly variable speed and more volatile in speed for those trucks. For each excavator, the carbon component (OM+EC) was the dominant species fraction and accounted for approximately 74.1-123% of the PM. The average ranges of WISSWSIS, elements, n-alkanes, PAHs, hopane and sterane fractions for each excavator were 0.335%-1.21%, 0.163%-7.50%, 3.6%-9.6%, 0.03%-0.24% and 0.001%-0.09%, respectively. In contrast to the other excavators, Zn and Cu were the second and third most abundant elements in excavators exhaust from E4, which may might be attributed toto the result of poor fuel quality and the old vehicles vehicle age. BesidesAdditionally, the elements fractions for the two excavators produced in 2013 (E1 (1.42%) and E6 (7.50%)) were higher than other excavators, which may might indicate that elements emissions control was deteriorating deteriorated and more stringent control technology should be developed to avoid the total elements adverse health effects. For excavators, the ranges of the ratios of BaA/(BaA+Chry), IcdP/(IcdP+BghiP) and Flua/(Flua+Pry) were 0.26-0.86, 0.20-1.0 and 0.24-0.87, respectively, with average of  $0.47 \pm 0.27$ ,  $0.44 \pm 0.38$  and  $0.48 \pm 0.27$ , respectively. For diesel trucks, the total carbonaceous composition (OM+EC) were accounted for 44.0% (E1T1), 27.9% (E2T2), 43.9% (E3T3), 51.6% (E4T4) and 46.3% (E5T5) of PM. For T2-diesel truck, wSIs-WSIs (13.8%) was were the most significant component-fraction of PM after OC, and it was higher than those in-for the other trucks by, within a factor of 4 to 10. The n-alkanes, PAHs, hopane and steranes fractions ranged from 0.85% to 4.78%, from 0.01% to 0.54% and from 0.002% to 0.024% for trucks, respectively. In comparison with the results from other

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**literatures**<u>studies</u>, the characteristics of <u>the</u> average source profiles for different types of non-road diesel vehicles varied sharply, while <u>those</u> for on-road diesel vehicles, those characteristics showed more stability. Although <u>the PAHs</u> fractions of PAHs for <u>the</u> excavators and trucks were <u>identicalsimilar</u>, the total of  $BaP_{eq}$  that was used to evaluate the carcinogenic risk was 31 <u>times greater for excavators than</u> fold of those for trucks.

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# **Table captions**

Table 1 Specifications of for the tested excavators and trucks

Table 2 Diesel contents from excavators

Table 3 Comparison of average chemical constituents of PM for different diesel

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ID		Model	Emission	Powers	Total weights	Displacements	Working hours	Mileages
ID	manufacturers <u>Manufacturers</u>	years	standards	(kw)	(kg)	(L)	(h)	(km)
E1	Volvo	2013	stage 2	169	30,500	7.1	2,751	/
E2	Hitachi	2007	pre-stage 1	162	30,200	9.8	16,166	/
E3	Sany	2012	stage 2	128	22,900	/	5,598	/
E4	Doosan	2004	pre-stage 1	110	22,000	8.1	12,000	/
E5	Doosan	2007	pre-stage 1	40	5,250	2.8	/	/
E6	Komatsu	2013	stage 2	35	5,300	2.4	780	/
T1	Futian	2010	China III	68	4,495	2.6	/	100,238
T2	JAC	2014	China IV	88	4,495	2.8	/	/
T3	Futian	2011	China III	70	11,190	3.9	/	99,000
T4	Chunlan	2002	China II	125	15,480	/	/	/
T5	JAC	2011	China III	105	15,590	4.3	/	130,000

Table 1 Specifications of tested excavators and trucks

ID	E1	E2	E3	E4	E5	E6	GB 252-2015
Gross thermal value	45 1	45 1	15.2	45.2	45.2	45.2	,
(MJ/kg)	45.1	45.1	45.5	45.5	45.5	45.5	/
Net thermal value	42.4	40.4	40.7	42.9	10 6	40 5	,
(MJ/kg)	42.4	42.4	42.7	42.0	42.0	42.5	/
Kinematic viscosity	4.02	4.02	2.00	4.16	1.00	4 20	2 00 9 00
(20 <u>°</u> C)(mm <sup>2</sup> /s)	4.23	4.23	5.09	4.10	4.60	4.39	3.00-8.00
Moisture (%)	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	/
Ash content (%)	0.04	0.04	0.05	4.16	0.03	0.05	0-0.01
C (%)	86.3	86.3	86.4	86.8	85.9	85.9	/
H (%)	11.6	11.6	11.5	11.2	12.0	12.1	/
O (%)	1.99	1.99	2.01	1.85	2.07	1.86	/
N (%)	0.05	0.05	0.05	0.04	0.06	0.05	/
S (ppm)	400	400	700	1100	200	200	<350

Table 2 Diesel contents from excavators

 $\underline{n.d.} = not detected$ 

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Vehicle	Excavators	Trucks	Trucks	Medium-duty	Diesel	Light-duty	Marine	Non-road
types				trucks	vehicles	engines	engine	generator
Methods	On-road		On-road	Dynamometer	Tunnel	Dynamometer	Dynamometer	Dynamometer
Reference	This stu	ıdy	(Wu et al., 2016)	(Schauer et al., 1999)	(Cui et al., 2016)	(Alves et al., 2015b)	(Sippula et al., 2014)	(Liang et al., 2005)
EC	33.3	26.9	55.3	30.8	39.5	69.9	14.1	
OC	39.2	9.89	31.8	19.7	27.2	12.7	60.0	
Ions	0.614	4.67	1.49	1.96	11.7	0.638		
$\mathbf{NH}_4^+$	0.044	0.215	0.188	0.730	2.06	0.005		
Cl	0.098	0.110	0.247		1.06	0.115		
$NO_3^-$	0.278	1.08	0.529	0.230	3.81	0.459		
$\mathrm{SO}_4^{2\text{-}}$	0.193	3.27	0.529	1.00	4.80	0.059		
Elements	1.76	0.941	0.493	0.200	12.8	0.069	3.17	
Na	0.245	0.047			0.287	0.041	0.564	
Mg	0.106	0.079			1.71	0.008	0.422	
Κ	0.197	0.028			0.872	0.002	0.671	
Ca	0.241	0.211		0.030	5.69	0.017	1.01	
Ti	0.008	0.011	0.145		0.206	0.0001	0.005	
V	0.001	0.000	0.001		0.008		0.044	
Cr	0.035	0.039	0.011	0.010	0.013		0.010	
Mn	0.013	0.009	0.002	0.010	0.064		0.006	

Ta	ble	e 3	6 C	Comparison of	average cl	nemical	consti	tuents	of PM	for	different	diesel	vehicles (	(%)	)

Continued It								
Fe	0.815	0.276	0.247	0.050	3.71	0.0003	0.138	
Co	0.001	0.005	0.0002	0.010	0.002		0.006	
Ni	0.015	0.006	0.002	nd			0.016	
Cu	0.042	0.107	0.004	0.010	0.013		0.130	
Zn	0.027	0.111	0.076	0.070	0.213	0.0001	0.130	
Pb	0.011	0.010	0.005	0.010	0.008		0.013	
Alkanes	5.14	1.73		0.222			1.37	0.831
C12	0.003	0.020						0.003
C13	0.003	nd						0.006
C14	0.019	0.0003						0.020
C15	0.057	0.013		0.001				0.056
C16	0.201	0.062		0.005				0.116
C17	0.107	0.144		0.003				0.265
C18	0.587	0.215		0.002			0.049	0.148
C19	0.777	0.308		0.002			0.120	0.126
C20	0.977	0.311		0.052			0.260	0.074
C21	0.516	0.290		0.022				0.014
C22	0.769	0.143		0.028			0.264	0.001
C23	0.349	0.099		0.025			0.177	0.001
C24	0.245	0.061		0.022			0.128	0.001
C25	0.197	0.032		0.014			0.083	0.0004

Continued Table 3											
C26	0.119	0.016	0.019		0.075						
C27	0.031	0.009	0.014		0.056						
C28	0.023	0.004	0.011		0.058						
C29	0.013	0.002	0.003		0.046						
C30	0.007	0.001			0.025						
C31	0.010	0.002			0.017						
C32	0.010	0.001			0.007						
C33	0.010	0.00001			0.002						
C34	0.010	0.0004									
C35	0.013	0.00004									
C36	0.016	nd									
C37	0.018	nd									
C38	0.025	nd									
C39	0.031	nd									
C40	0.003	nd									
PAHs	0.098	0.130		0.251	0.021	0.021					
Nap	0.008	0.001		0.014		0.0004					
Acy	0.005	0.0003		0.006		0.0002					
Ace	0.001	0.00004		0.001		0.0003					
Flu	0.002	0.0001				0.001					
Phe	0.005	0.021		0.007		0.008					
Ant	0.001	0.001		0.002		0.0004					

Fluo	0.026	0.010		0.027	0.009	0.002
Pyr	0.028	0.088		0.052	0.008	0.007
BaA	0.007	0.001		0.014	0.001	0.0005
Chry	0.008	0.002		0.025	0.003	0.0005
BbF	0.002	0.001		0.016		0.0003
BkF	0.001	0.0001		0.003		0.0002
BaP	0.0004	0.00001		0.009		0.0004
IcdP	0.001	0.00002		0.013	0.0004	0.001
DahA	0.000	0.001		0.001		0.0002
BghiP	0.003	0.004		0.062	0.0003	0.0003
Hopane,						
sterane	0.026	0.011	0.014	0.167	0.143	
ABB	0.001	0.0005	0.0004	0.007		
AAA	0.002	0.001	0.001	0.006		
Tm	0.001	0.001	0.001	0.014	0.012	
30AB	0.011	0.005	0.006	0.065	0.069	
29AB	0.011	0.004	0.006	0.075	0.061	

5 <u>n.d. = not detected</u>

# **Figure captions**

Figure 1The routes for diesel trucks

Figure 2 Particulate matter sampling system

Figure 3 EF<sub>PM</sub> for excavators with different operating operational modes and emission

standards and the correlation with sulfur contents

Figure 4 Diesel trucks  $EF_{PM}$  for different emission standards, vehicle sizes and driving conditions

Figure 5 PM compositional constituents for individual vehicles

Figure 6 OC/EC ratios under in different operating operational modes and driving

10 conditions for excavators and trucks

Figure 7 Cross plots for the ratios of BaA/(BaA+Chry) vs IcdP/(IcdP+BghiP) and BaA/(BaA+Chry) vs Flua/(Flua+Pry) and comparison with those from other diesel vehicle sources.

Figure 8 Percentages of each ring PAHs to total PAHs; BaPeq for parent PAHs in each

15 tested trucks and excavators

20

5

25

30



Figure 1The routes for diesel trucks; a was the site of Yantai; b was the route for China III and China IV light-duty diesel trucks; c was the rout for China II heavy-duty diesel truck ; d was route for China III medium-duty and heavy-duty trucks



Figure 2 Particulate matter sampling system; 1 <u>was\_is the</u> flowmeter; 2 <u>was\_is the</u> dilute tunnel; 3 <u>was\_is the</u> filtrator; 4 <u>was\_is the</u> activated carbon; 5 <u>was\_is the</u> fan; 6 <u>was\_is the</u> valve; 7 <u>was\_is the</u> flow divider; 8 <u>was\_is the</u> filter membrane sampler; <u>and</u> 9 <u>was\_is the</u> exhaust analyzer



Figure 3 EF<sub>PM</sub> for excavators with different <u>operating operational</u> modes and emission standards (A) and the correlation with sulfur contents (B)



Figure 4 Diesel trucks  $EF_{PM}$  for different emission standards (a), vehicle sizes (b) and driving conditions (c)



Figure 5 Compositional constituents of PM for individual vehicles (%)



Figure 6 OC/EC ratios <u>under in different operating operational</u> modes and driving conditions for excavators and trucks



Figure 7 Cross plots for the ratios of BaA/(BaA+Chry) vs IcdP/(IcdP+BghiP) and BaA/(BaA+Chry) vs Flua/(Flua+Pry) and comparison with those from other diesel vehicle sources. A and B were are the isomer ratios of the PAHs for from the excavators and trucks-, respectively, tested in this study, respectively; C and D were are the average isomer ratios of PAHs for trucks and excavators tested in this study; E, F, G, H, I were are results obtained from Liu et al. (2015), Wang et al. (2015), Shah et al. (2005), Schauer et al. (1999), Chen et al. (2013)



each tested trucks (B) and excavators (C)