

# Investigation of the $\beta$ -pinene photooxidation by OH in the atmosphere simulation chamber SAPHIR

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## Abstract.

Beside isoprene, monoterpenes are the non-methane volatile organic compounds (VOC) with the highest global emission rates. Due to their high reactivity towards OH, monoterpenes can dominate the radical chemistry of the atmosphere in forested areas. In the present study the photochemical degradation mechanism of  $\beta$ -pinene was investigated in the Jülich atmosphere simulation chamber SAPHIR. One focus of this study is on the OH budget in the degradation process. Therefore the SAPHIR chamber was equipped with instrumentation to measure radicals (OH, HO<sub>2</sub>, RO<sub>2</sub>), the total OH reactivity, important OH precursors (O<sub>3</sub>, HONO, HCHO), the parent VOC  $\beta$ -pinene, its main oxidation products, acetone and nopinone, and photolysis frequencies. All experiments were carried out under low NO conditions ( $\leq 300$  ppt) and at atmospheric  $\beta$ -pinene concentrations ( $\leq 5$  ppb) with and without addition of ozone. For the investigation of the OH budget, the OH production and destruction rates were calculated from measured quantities. Within the limits of accuracy of the instruments, the OH budget was balanced in all  $\beta$ -pinene oxidation experiments. However, even though the OH budget was closed, simulation results from the Master Chemical Mechanism 3.2 showed that the OH production and destruction rates were underestimated by the model. The measured OH and HO<sub>2</sub> concentrations were underestimated by up to a factor of two whereas the total OH reactivity was slightly overestimated because the model predicted a nopinone mixing ratio which was three times higher than measured. A new, theory-derived first-generation product distribution by Vereecken and Peeters (2012) was able to reproduce the measured nopinone time series and the total OH reactivity. Nevertheless the measured OH and HO<sub>2</sub> concentrations remained underestimated by the numerical simulations. These observations together with the fact that the measured OH budget was closed suggest the existence of unaccounted sources of HO<sub>2</sub>. Although the mechanism of additional HO<sub>2</sub> formation could not be resolved, our model studies suggest that an activated alkoxy radical intermediate proposed in the model of Vereecken and Peeters (2012)

generates HO<sub>2</sub> in a new pathway, whose importance has been underestimated so far. The proposed reaction path involves unimolecular rearrangement and decomposition reactions and photolysis of dicarbonyl products, yielding additional HO<sub>2</sub> and CO. Further experiments and quantum chemical calculations have to be made to completely unravel the pathway of HO<sub>2</sub> formation.

## 5 1 Introduction

Thousands of different volatile organic compounds (VOCs) are emitted into the atmosphere (Goldstein and Galbally, 2007). The emissions of biogenic volatile organic compounds BVOCs exceed those of anthropogenic VOCs by a factor of ten (Piccot et al., 1992; Guenther et al., 1995, 2012). On a global scale, isoprene and monoterpenes are the BVOCs with the highest emission rates with the exception of methane. About 44 % of the global BVOC emissions can be attributed to isoprene and  
10 about 11 % to monoterpenes (Guenther et al., 1995). Isoprene and monoterpenes are unsaturated hydrocarbons. Hence, their main atmospheric sink is the addition of hydroxyl radicals (OH), nitrate radicals (NO<sub>3</sub>) or ozone to the double bond (Calogirou et al., 1999; Atkinson and Arey, 2003). During daytime the reaction of isoprene and monoterpenes with the OH radical is the major sink for these VOC species. The subsequent addition of oxygen produces organic peroxy radicals (RO<sub>2</sub>). In the presence of nitrogen oxides (NO<sub>x</sub>), RO<sub>2</sub> is indirectly converted to hydroperoxy radicals (HO<sub>2</sub>) through reaction with NO. HO<sub>2</sub> reacts  
15 further with NO, recycling the OH consumed in the initial reaction step and producing further NO<sub>2</sub>. As a side effect, ozone is produced by NO<sub>2</sub> photolysis. The oxidation of VOCs in the presence of NO is the main photochemical source of ozone in the troposphere (Seinfeld and Pandis, 2006). Moreover, the oxidation processes of isoprene and monoterpenes mainly lead to the production of less reactive polar oxygenated volatile organic compounds (OVOCs) which are significantly involved in the formation of secondary organic aerosols (Kanakidou et al., 2005; Goldstein and Galbally, 2007).

20 During the last decade, the research on the chemical degradation of BVOCs in the atmosphere has made significant progress through laboratory and atmospheric chamber experiments, and theoretical chemistry studies. It was discovered that RO<sub>2</sub> radicals from the reaction of biogenic VOCs with OH can undergo unimolecular reactions which influence the chemistry of HO<sub>x</sub> and OVOCs. In case of the degradation of isoprene and methacrolein, RO<sub>2</sub> was found to regenerate efficiently HO<sub>x</sub> by isomerization and decomposition reactions (Paulot et al., 2009; da Silva et al., 2010; Peeters and Müller, 2010; Crouse et al.,  
25 2011, 2012, 2013; Wolfe et al., 2012; Taraborrelli et al., 2012; Liu et al., 2013; Fuchs et al., 2013, 2014; Peeters et al., 2014). RO<sub>2</sub> radicals from the oxidation of isoprene and some monoterpenes were found to produce low-volatility OVOCs, which contribute substantially to SOA formation in the atmosphere (Paulot et al., 2009; Ehn et al., 2014; Bates et al., 2014). The discovered chemistry is particularly important in forests, which contribute to the global non-methane BVOC emissions with an estimated share of 75 % (Guenther et al., 1995; Wiedinmyer et al., 2004; Guenther et al., 2012). In forests, the unimolecular  
30 RO<sub>2</sub> reactions can effectively compete with the RO<sub>2</sub> + NO reaction, since anthropogenic NO emissions are generally missing.

The above mentioned studies of BVOC oxidation mechanisms were mostly inspired by field observations of unexplained high OH concentrations in isoprene-dominated forests, which have pointed to unknown NO-independent OH recycling processes (Tan et al., 2001; Carslaw et al., 2001; Ren et al., 2008; Lelieveld et al., 2008; Hofzumahaus et al., 2009; Kubistin et al.,

2010; Whalley et al., 2011; Lu et al., 2012). The newly discovered mechanisms for isoprene and methacrolein, however, can explain only part of the observed high OH concentrations. Another possible reason could be OH interferences in the low-pressure laser-induced fluorescence (LIF) instruments that were applied in the above field studies. Artificial OH production was discovered in two similar LIF instruments applying a newly developed chemical modulation technique for OH detection  
5 (Mao et al., 2012; Hens et al., 2014; Novelli et al., 2014; Feiner et al., 2016). The interference seems to be related to organic compounds, but the underlying OH formation mechanism is not known. Experimental tests with other type of LIF instruments have not found such interference (Fuchs et al., 2012, 2016; Griffith et al., 2013; Tan et al., 2017), yet it is difficult to draw firm conclusions for past campaigns as long as the reported artefacts (Mao et al., 2012) are not fully understood.

Due to their abundance and their structural similarity to isoprene, unknown monoterpene chemistry may contribute to the  
10 underestimation of OH concentrations in forests as proposed by da Silva et al. (2010) for open-chain monoterpenes like myrcene and ocimene. During a field campaign in Borneo, Whalley et al. (2011) observed that discrepancies between measured and modeled OH occurred in the morning hours when VOC emissions were dominated by monoterpenes. Moreover, field studies in Greece (Carslaw et al., 2001), in the U.S. (Kim et al., 2013) and in Finland (Hens et al., 2014) indicate that the radical chemistry in forested areas, which are dominated by monoterpene and 2-methyl-3-buten-2-ol (MBO) emissions, is not  
15 well understood.

In this work we investigated the atmospheric degradation of monoterpenes in the atmosphere simulation chamber SAPHIR in Jülich.  $\beta$ -Pinene comprises 17 % of the estimated global monoterpene emission rate (Sindelarova et al., 2014) and was therefore chosen as a representative species for our investigations. To our knowledge it is the first chamber study investigating the  $\beta$ -pinene, or any monoterpene degradation in general, under natural concentration conditions (VOC less than 5 ppb).  
20 In comparison to other chamber studies which focused on the determination of products and SOA yields (Lee et al., 2006; Saathoff et al., 2009; Eddingsaas et al., 2012a, b; Zhao et al., 2015) our main goal was to investigate the radical budget of the monoterpene degradation. For that purpose all critical radical species (OH, HO<sub>2</sub>, RO<sub>2</sub>) were measured. In order to exclude possible measurement artefacts for OH, differential optical absorption spectroscopy (DOAS) was applied for OH measurements in addition to LIF.

## 25 **2 Methods**

### **2.1 SAPHIR atmosphere simulation chamber**

The atmosphere simulation chamber SAPHIR (Simulation of Atmospheric PHotochemistry In a large Reaction Chamber) located in the Forschungszentrum Jülich (Germany) is a tool to investigate complex atmospheric mechanisms under nearly natural conditions. The chamber has a cylindrical shape (18 m length, 5 m diameter, 270 m<sup>3</sup> volume) and consists of a double  
30 walled FEP Teflon foil attached to a steel frame. The Teflon foil guarantees a maximum of inertness of the chamber surface and leads to a minimization of wall effects. In SAPHIR natural sunlight is used as light source for photochemical reactions. About 85 % of the UV-A, UV-B and visible light is transmitted by the FEP foil. A shutter system allows to switch between illuminated and dark chamber conditions within 60s. To investigate photochemical degradation processes in the ppb and sub-

ppb range SAPHIR is operated with ultra pure synthetic air (Linde, N<sub>2</sub> 99.9999 %, O<sub>2</sub> 99.9999 %). A slight overpressure of about 30 Pa in the inner chamber prevents diffusion of outside air into SAPHIR. Due to small leakages and consumption of air by instruments a replenishment flow has to be introduced into the chamber to keep up the pressure difference to the outside. During experimental operation this flow is in a range of 9-12 m<sup>3</sup>h<sup>-1</sup>, leading to a dilution of trace gases at a rate of approximately 3-4 % h<sup>-1</sup>. An installed ventilator guarantees well mixed conditions during the experiments. For more detailed information about the chamber and its properties the reader is referred to previous publications (Poppe et al., 2007; Schlosser et al., 2007, 2009; Wegener et al., 2007; Dorn et al., 2013).

## 2.2 Instrumentation

OH, HO<sub>2</sub> and RO<sub>2</sub> concentrations were measured simultaneously by a laser-induced fluorescence (LIF) system, using three independent low-pressure detection cells. Each cell samples ambient air by gas expansion through an inlet nozzle, producing a fast gas flow through the cell. OH is detected by pulsed laser-excited resonance fluorescence at 308 nm (Holland et al., 1995). RO<sub>2</sub> and HO<sub>2</sub> are detected indirectly by chemical conversion with NO to OH, followed by LIF detection of the formed OH (Fuchs et al., 2008, 2011). The peroxy radicals are distinguished from each other by their different conversion efficiencies, which depend on the amount of added NO and the reaction time between NO addition and OH detection. In the low-pressure HO<sub>x</sub> cell, the addition of NO leads to fast formation of OH in just one reaction step. In contrast, conversion of RO<sub>2</sub> to OH requires at least three reaction steps:



For simple alkyl peroxy radicals, this reaction sequence is relatively slow (especially at reduced O<sub>2</sub> partial pressure) compared to the residence time in the HO<sub>2</sub> detection cell and results in a very low detection efficiency. However,  $\beta$ -hydroxy RO<sub>2</sub> species produced by the reaction of alkenes with OH are converted by NO to highly reactive  $\beta$ -hydroxy alkoxy radicals. Instead of reacting with O<sub>2</sub> directly,  $\beta$ -hydroxy alkoxy radicals nearly exclusively decompose and then react rapidly with O<sub>2</sub> forming thereby HO<sub>2</sub> much faster than other alkoxy radicals. The fact that for  $\beta$ -hydroxy alkyl peroxy radicals the overall conversion to OH is very fast leads to an interference in the HO<sub>2</sub> channel of the LIF instrument (Fuchs et al., 2011). The interference was carefully characterized for RO<sub>2</sub> species formed by the reaction of  $\beta$ -pinene with OH in laboratory experiments following the procedure described by Fuchs et al. (2011). About 25 % of these RO<sub>2</sub> species are detected as an additional signal in the HO<sub>2</sub> channel of the instrument. In the third measurement cell, the sum of atmospheric RO<sub>2</sub> and HO<sub>2</sub> is measured. In this case, RO<sub>2</sub>

radicals are converted by NO in a pre-reactor to HO<sub>2</sub>, which is then further converted together with atmospheric HO<sub>2</sub> to OH in the detection cell (Fuchs et al., 2008). Since the RO<sub>2</sub> concentration is calculated as the difference between the concentration of RO<sub>x</sub> (RO<sub>2</sub> + HO<sub>2</sub>) and measured HO<sub>2</sub>, the interference in the HO<sub>2</sub> measurement also affects indirectly the RO<sub>2</sub> data.

On 27th of August 2012 OH was measured additionally by a differential optical absorption spectrometer (DOAS). In general both instruments showed a good agreement over the past 10 years (Schlosser et al., 2007, 2009; Fuchs et al., 2012). Also for the terpenoid campaign in 2012 on average no significant difference between LIF and DOAS instrument was observed. As the DOAS instrument is the only absolute method for the quantification of OH (Hofzumahaus and Heard, 2016), the DOAS OH data were used for the following evaluation of the OH budget analysis.

The OH reactivity  $k(\text{OH})$  was measured by flash photolysis / laser induced fluorescence (FP/LIF) technique (Lou et al., 2010). The evaluation of the pseudo-first-order decays of OH gives a direct measure of the total rate coefficient of the OH loss.

Besides OH, HO<sub>2</sub>, RO<sub>2</sub> and  $k(\text{OH})$ , HCHO (Hantzsch reaction), HONO (long path absorption photometry, LOPAP), CO (reduction gas analysis, RGA), CO<sub>2</sub>, CH<sub>4</sub>, H<sub>2</sub>O (cavity ring-down spectroscopy, CRDS), as well as NO, NO<sub>2</sub> and O<sub>3</sub> (chemiluminescence, CL) were determined by direct measurements. VOC were measured by a PTR-TOF-MS (proton transfer reaction time of flight mass spectrometer) and two gas chromatographs of the same type coupled with mass spectrometric and flame ionization detectors (GC/MS/FID). Moreover experimental boundary conditions including temperature (ultrasonic anemometer), pressure (capacitive gauge), replenishment flow rate (mass flow controller) and photolysis frequencies (spectroradiometer) were continuously recorded.

Table 1 provides an overview of the key instruments for this study and their specifications. For more detailed information on the analytical instrumentation of SAPHIR the reader is referred to previous publications (Bohn and Zilken, 2005; Bohn et al., 2005; Rohrer et al., 2005; Wegener et al., 2007; Dorn et al., 2013), and references therein.

### 2.3 Experimental procedure

Before every experiment day the chamber was flushed with dry ultra-pure synthetic air over night to purge contaminants of previous experiments under their detection limit. At the beginning of the experiment 20 ppm of CO<sub>2</sub> were injected into SAPHIR as dilution tracer. After that the relative humidity was increased to 75 % by adding water vapour, generated by the vaporisation of ultra-pure water (Milli-Q), to the purge flow. As HONO photolysis is the main source of OH in the SAPHIR chamber it is impossible to conduct experiments in the complete absence of NO. To lower the NO level in the experiment on 27th August 50 ppb of ozone, produced from a silent discharge ozonizer (O3Onia), was injected after humidification. Shortly afterward the shutter system of SAPHIR was opened, exposing the chamber to sunlight.

In the following two hours of the experiments (so-called "zero air phase") no other trace gases were introduced into SAPHIR. During the zero air period HONO was formed from the chamber walls (Rohrer et al., 2005) depending on relative humidity and UV radiation. In addition to the OH production the photolysis of HONO leads to an increase in NO and NO<sub>2</sub> concentration. In addition acetaldehyde, formaldehyde and acetone were formed in the chamber with a rate of 90 – 250 ppt/h. The zero air phase ended with the injection of  $\beta$ -pinene while the SAPHIR chamber was exposed to light. The injection was performed by introducing a high concentration gas mixture of  $\beta$ -pinene (about 50 ppm) from a silcosteel canister (Restek) through a mass

flow controller to the experimental flow. The  $\beta$ -pinene concentration of the mixture was previously determined by oxidizing a part of the  $\beta$ -pinene mixture on a platinum catalyst and quantifying the produced  $\text{CO}_2$ . This absolute method makes it possible to calculate the VOC starting concentration of the experiment very accurately. During the following 6 h of the experiment, the so-called "VOC phase",  $\beta$ -pinene was degraded by OH in the illuminated chamber. In the experiment of 27th August  $\beta$ -pinene was injected for a second and third time into SAPHIR approximately two and four hours after the first VOC injection, respectively. Every experiment ended with closing the louver system of the chamber in the late evening of the experiment day. For all the chamber experiments the fan was running during the whole time ensuring homogeneous mixing of the chamber air.

Table 2 sums up the experimental conditions of the three  $\beta$ -pinene oxidation experiments.

## 2.4 Model calculations

The acquired time series of trace gases and radicals were compared to zero-dimensional box model simulations with the Master Chemical Mechanism (MCM). The MCM is a state-of-the-art chemical mechanism developed by Jenkin et al. (1997) and Saunders et al. (2003). For this publication the MCM version 3.2 was used (available at <http://mcm.leeds.ac.uk/MCMv3.2/>). For the application on modeling chamber experiments the model was extended by some chamber specific processes. As an alternative to the  $\beta$ -pinene chemistry in the MCM, we also applied the reaction mechanism by Vereecken and Peeters (2012), which is based on theoretical-kinetic analyses of the reaction mechanism. The mechanism by Vereecken and Peeters (2012) only describes the first-generation product formation, i.e. the subsequent chemistry of the products formed in the first radical chain is not included in the model. The accumulated yield of primary products in our model runs remains below 20 % compared to the sum of the residual concentration of  $\beta$ -pinene, and the concentrations of reactive primary products whose chemistry is fully described (e.g. nopinone and acetone). As such, it appears that omitting the secondary chemistry of these products does not have an overly large impact on the reaction fluxes, and is therefore unlikely to be the main reason for any discrepancies relative to the measurements.

As mentioned in section 2.1 the required replenishment flow into SAPHIR leads to an additional dilution process for every model species. The applied dilution rate is thereby calculated from the measured  $\text{CO}_2$  loss in the chamber. Previous characterization experiments showed that ozone had a shorter lifetime than  $\text{CO}_2$  in the chamber (dilution corrected ozone lifetime approximately 30 h). This observation was included as additional loss term in the model. The chamber sources of HONO, HCHO and acetone are well known from routine reference experiments in SAPHIR and can be parametrized by empirical equations, depending on temperature, relative humidity and solar radiation in the chamber (Rohrer et al., 2005; Karl et al., 2006; Kaminski, 2014). The source strengths were adjusted to match the time series of  $\text{NO}_x$ , HCHO and acetone during the zero air phases of the experiments. The parametrization of the acetaldehyde source was less satisfactory and so the model was constrained by the measured acetaldehyde concentration.

In all experiments the summed contributions of known chamber sources to the OH reactivity measured in the zero air phase ( $0.1 - 0.7 \text{ s}^{-1}$ ) were not sufficient to explain the measured OH reactivity ( $0.7 - 1.5 \text{ s}^{-1}$ ). Analogous to the procedure applied by Fuchs et al. (2012, 2014) the unexplained part of the measured OH reactivity was modeled as a co-reactant Y, with constant OH reactivity in the model, where the concentration times rate coefficient  $[\text{Y}] \cdot k_{\text{OH}+\text{Y}}$  was set to reproduce the measured OH

reactivity in the chamber after humidification. Analogous to CO the reaction of Y with OH is assumed to form one molecule of HO<sub>2</sub>.

The parameters temperature, pressure, water vapor concentration, the calculated dilution rate and the photolysis frequencies for HONO, HCHO, O<sub>3</sub> and NO<sub>2</sub> were set as fixed boundary conditions in the model. Photolysis frequencies that were not measured were calculated for clear sky conditions by the function included in MCM 3.1 and then corrected for cloud cover and the transmission of the Teflon film by multiplying the clear sky value with the ratio of measured to modeled photolysis frequency of NO<sub>2</sub>. Constrained parameters were re-initialized on a 1 min time grid. The injections of  $\beta$ -pinene and ozone in the chamber were modeled as sources which were only present during the time period of injection. The source strengths were adapted to match the measured ozone concentration and the OH reactivity at the point of injection. The subsequent time series of the concentrations were determined by the kinetic models described above.

Because of described instrumental interferences it is not possible to directly compare the modeled HO<sub>2</sub> concentration [HO<sub>2</sub>] and the sum of the concentrations of the different RO<sub>2</sub> species [RO<sub>2</sub>] against the measured time series of the LIF instrument, [HO<sub>2</sub><sup>\*</sup>] and [RO<sub>2</sub><sup>\*</sup>], for HO<sub>2</sub> and RO<sub>2</sub>, respectively.

$$[\text{HO}_2^*] = [\text{HO}_2] + \sum (\alpha_{\text{RO}_2}^i \cdot [\text{RO}_2]_i) \quad (1)$$

$$[\text{RO}_2^*] = [\text{RO}_2] - \sum (\alpha_{\text{RO}_2}^i \cdot [\text{RO}_2]_i) \quad (2)$$

$\alpha_{\text{RO}_2}^i$ : relative detection sensitivity for RO<sub>2</sub> species i (compared to HO<sub>2</sub> with  $\alpha = 1$ )

$\sum [\text{RO}_2]_i$ : interfering RO<sub>2</sub> radicals of  $\beta$ -pinene

$\sum (\alpha_{\text{RO}_2}^i \cdot [\text{RO}_2]_i)$ : RO<sub>2</sub> interference

For a direct comparison of the measured [HO<sub>2</sub><sup>\*</sup>] against the model, the modeled HO<sub>2</sub> plus an estimated RO<sub>2</sub> interference is combined to yield the model parameter HO<sub>2</sub><sup>\*</sup> (Lu et al., 2012). Depending on the experimental phase, up to 25 % of the modeled HO<sub>2</sub><sup>\*</sup> can be attributed to the interfering RO<sub>2</sub> species [RO<sub>2</sub><sub>i</sub>]. Moreover, note that the MCM and the modifications by Vereecken and Peeters yield different RO<sub>2</sub> species, which results in rather different contributions of RO<sub>2</sub> into the HO<sub>2</sub> signal.

RO<sub>2</sub> radicals are detected in the LIF instrument by a three step conversion of RO<sub>2</sub> to OH. Only species reacting with NO to RO and then decomposing or reacting with O<sub>2</sub> in a second reaction step to HO<sub>2</sub> can be detected with a sufficient sensitivity. Depending on the model used up to 70 % of the modeled RO<sub>2</sub> species of  $\beta$ -pinene are not detectable under these conditions. To account for this, the measured RO<sub>2</sub> signal [RO<sub>2</sub><sup>\*</sup>] is compared to the model parameter RO<sub>2</sub><sup>\*</sup>, which corresponds to the sum of the theoretically detectable RO<sub>2</sub> model species.

The model  $RO_2^*$  must be additionally corrected by the subtraction of the  $RO_2$  species which are already included in the model parameter  $HO_2^*$ . This is again related to the operating conditions of the LIF instrument where in the  $RO_x$  cell the sum of detectable  $RO_2$  plus  $HO_2$  and in the  $HO_x$  cell  $HO_2$  plus interfering  $RO_2$  radicals are measured. As the  $RO_2$  concentration is determined by subtracting the signal of the  $HO_x$  cell from the signal of the  $RO_x$  cell, an  $RO_2$  interference in the  $HO_x$  cell automatically leads to an underestimation of the calculated  $RO_2$  concentration.

### 3 Results and discussion

#### 3.1 Determination of product yields

The formation yields of first-generation degradation products are important information for the understanding of the oxidation mechanism of  $\beta$ -pinene with OH (Fig. 1). By correlating the concentration of the products with the concentration of the degraded  $\beta$ -pinene it is possible to determine the product yield. Because of the lack of suitable reference standards and the low concentration of  $\beta$ -pinene it was only possible to determine the yield of acetone and nopinone in the OH oxidation experiment. The concentrations of  $\beta$ -pinene and nopinone were determined by PTR-TOF-MS whereas interpolated GC/FID data of the acetone concentration were used for the yield determination. This was done to exclude any possible interferences on the quantifier ion of acetone in the PTR-TOF-MS.

As a result of ozone addition in the experiment on 27 Aug 2012 a part of the injected  $\beta$ -pinene was degraded by ozonolysis. The fraction of the ozonolysis in the total conversion of  $\beta$ -pinene was approximately 5 % and can be neglected.

The experiment duration of several hours necessitated the correction of the measured concentration time series to account for reactive losses of acetone and nopinone with OH and chamber effects like dilution (all species) and chamber sources (acetone). This was done using a recursive discrete time equation analogous to Galloway et al. (2011). The correction of the acetone concentration was done by scaling the assumed acetone chamber source to the measured values during the zero air phase of the experiments. The assumed acetone source strength was typically  $70 \text{ ppth}^{-1}$  which was as large as 20 to 30 % of the total amount of acetone produced in the  $\beta$ -pinene experiments. Equations 3 - 7 illustrate all applied corrections on the acetone concentration.

$$[\text{CH}_3\text{COCH}_3]_{\text{corr}(i)} = [\text{CH}_3\text{COCH}_3]_{\text{corr}(i-1)} + \Delta c_{\text{CH}_3\text{COCH}_3} + \Delta c_{\text{RL}} + \Delta c_{\text{DIL}} + \Delta c_{S_{\text{CH}_3\text{COCH}_3}} \quad (3)$$

$$\Delta c_{\text{RL}} = [\text{CH}_3\text{COCH}_3]_{(i-1)} \cdot [\text{OH}]_{(i-1)} \cdot \Delta t \cdot k_{\text{CH}_3\text{COCH}_3+\text{OH}} \quad (4)$$

$$\Delta c_{\text{DIL}} = [\text{CH}_3\text{COCH}_3]_{(i-1)} \cdot \Delta t \cdot k_{\text{DIL}} \quad (5)$$

$$\Delta c_{S_{\text{CH}_3\text{COCH}_3}} = S_{\text{CH}_3\text{COCH}_3} \cdot \Delta t \quad (6)$$

$$S_{\text{CH}_3\text{COCH}_3} = a_{\text{CH}_3\text{COCH}_3} \cdot J_{\text{NO}_2} \cdot (0.21 + 2.6 \cdot 10^{-2} \cdot RH) \cdot e^{(-2876/T)} \quad (7)$$

$[\text{CH}_3\text{COCH}_3]_{\text{corr}}$ : corrected acetone concentration

5  $\Delta c_{\text{RL}}$ : reactive loss

$\Delta c_{\text{DIL}}$ : dilution

$\Delta c_{\text{S}_{\text{CH}_3\text{COCH}_3}}$ : chamber source

10

$\Delta t$ : time interval between between time  $i$  and  $(i-1)$

$S_{\text{CH}_3\text{COCH}_3}$ : source strength

15  $a_{\text{CH}_3\text{COCH}_3}$ : scaling factor

$RH$ : relative humidity

$J_{\text{NO}_2}$ : photolysis frequency  $\text{NO}_2$

20

The results of the yield determination are listed in Table 3. In principle product yields of nonlinear degradation processes depend on the fate of  $\text{RO}_2$  which is governed by multiple physical and chemical boundary conditions such as pressure, temperature,  $\text{H}_2\text{O}$ ,  $\text{O}_3$ , VOC,  $\text{HO}_2$  and  $\text{NO}$  concentration. The discussed  $\beta$ -pinene experiment was conducted at ambient pressure in a temperature range of 298 – 304 K. The relative humidity was about 50 % before the first VOC injection and decreased to 30 % over the course of the experiment, due to the warming of the chamber and the dilution of the chamber air by the replacement flow. It is known for many VOC species that the product yields depend on the VOC to  $\text{NO}$  ratio (Atkinson, 2000). This is why the two  $\beta$ -pinene experiments without and the  $\beta$ -pinene with the addition of 50 ppb ozone are handled separately. During the experiment on 27th August the nopinone yield as well as the acetone yield increased subsequently with the second and third  $\beta$ -pinene addition and are therefore denoted as range. The specified errors consider the errors of measurement of the correlated VOC concentrations as well as the errors originating from the correction of reactive losses, dilution and chamber sources. To reduce the influence of secondary product formation and to facilitate the comparability of the results only the data of the experiment when less than 70 % of  $\beta$ -pinene was used for the yield calculation. To our knowledge, these are the first acetone and nopinone yields measured for reaction mixtures with less than 5 ppb of  $\beta$ -pinene.

35 Within the calculated error the determined nopinone yield in this work agrees well with every literature value except the published yield of Hatakeyama et al. (1991). These authors report nopinone yields a factor of three higher than every other

literature value. Vereecken and Peeters (2012) pointed out that Hatakeyama et al. (1991) measured the nopinone yield by using FTIR absorption at  $1740\text{ cm}^{-1}$ , which includes the absorption of other carbonyl compounds. Taking recent literature and our results into account it seems that the nopinone yield of  $\beta$ -pinene oxidation with OH has no strong dependence on the NO level (see Table 3). The slight increase of the nopinone yield over the three  $\beta$ -pinene injections in the experiment of 27th Aug 2012 can be related to a change of boundary conditions as well as a secondary nopinone source. For example, the MCM 3.2 contains nopinone formation pathways from the degradation of the related hydroperoxides and organic nitrates.

The determined acetone yield is in agreement with the reported literature values of Wisthaler et al. (2001), Librando and Tringali (2005) and Larsen et al. (2001). All reported literature values are smaller than the determined acetone yields in SAPHIR and show a wide range. Similar to nopinone there is no clear evidence of a NO dependence of the acetone yield. Due to the long reaction time the increase of the acetone yield in the experiment of 27 Aug 2012 is most likely related to secondary acetone production. Since the yields in the literature were determined under various boundary conditions (e.g. light source, OH source, relative humidity) it is not possible to determine the reasons for the discrepancy. It could be related to different boundary conditions or measurement errors.

### 3.2 Comparison of trace-gas measurements with MCM 3.2 model calculations

In this section the measured trace gas concentrations of the  $\beta$ -pinene experiment from 27th August are compared to the base model using the unmodified MCM 3.2 (see Fig. 2). From the moment the roof of the SAPHIR chamber was opened, HONO was formed at the chamber walls. Due to the photolysis of HONO, OH and NO were produced in the chamber, leading to a rise in the OH as well as the NO concentration. The parametrized HONO source sufficiently describes the measured nitrogen oxides in the zero air phase. The rise in the NO and  $\text{NO}_2$  concentration is well captured. The modeled OH concentration also agreed well with the measurements.

Beside HONO, also formaldehyde, acetaldehyde and acetone were formed or released from the chamber walls, as can be seen in the case of acetone as a slight concentration rise. These oxygenated VOC species (OVOCs) contributed to the increase of the measured background OH reactivity of  $1.5\text{ s}^{-1}$  during the zero air phase of the experiment. As the sum of the measured OH reactants was not sufficient to explain the measured OH reactivity ( $0.7\text{ s}^{-1}$  unexplained), the modeled OH reactivity was adjusted by a constant source of a species Y, assumed to react like CO, i.e. with similar rate coefficient and  $\text{HO}_2$  formation. Under the assumption of a constant concentration of 120 ppb Y the measured background reactivity is well reproduced by the model. The concentration of OH is well reproduced by MCM in the zero air phase, while  $\text{HO}_2^*$  is slightly overestimated and  $\text{RO}_2^*$  is underestimated by 25 %, each. These deviations are probably caused by the chemistry of the unknown species, which contribute about half of the OH reactivity before  $\beta$ -pinene is injected.

With the beginning of the VOC phase of experiments, the OH reactivity is dominated by known reactants, and good model-to-measurement agreement is expected for the radical concentrations, if the chemistry of the reactants is well understood. The reactants CO and  $\text{CH}_4$ , for example, give agreement better than 15 % for experiments in the SAPHIR chamber (Fuchs et al., 2013).

For the current case, the addition of  $\beta$ -pinene led to a sharp increase in the measured OH reactivity. Directly after the  $\beta$ -pinene injection the increase of the modelled OH reactivity, calculated from the canister injection, corresponded well with the measured  $k(\text{OH})$  increase. The  $\beta$ -pinene concentration measured by PTR-TOF-MS was about 15 % lower than the calculated injection, but still agreed with the canister injection within the instrumental uncertainty. Over the course of the VOC phase, and thereby the consumption of  $\beta$ -pinene, the measured OH reactivity was increasingly overestimated by the model. During this time period nopinone has the highest proportion of modeled OH reactivity beside  $\beta$ -pinene. However the measured nopinone concentration was overestimated by a factor of three by MCM 3.2 whereas the acetone and CO concentration were underestimated by a factor of two. In general the MCM gives a poor description of the first-generation  $\beta$ -pinene degradation products. Simultaneously with the increase of the OH reactivity a sharp decrease of OH radical concentration was observed. At the time of  $\beta$ -pinene injection model and measurement agreed well, but over the course of the experiment OH was increasingly underestimated by the model (30-50 %). The modeled concentration of theoretically measurable  $\text{RO}_2$  radicals  $\text{RO}_2^*$  exceeded the measured concentration by about 40 %. Similar to OH, the modeled  $\text{HO}_2^*$  concentration initially agreed well with the measurements directly after  $\beta$ -pinene injection but was increasingly underestimated by the MCM in the latter part of the experiment. The measured time series of ozone was well captured by the MCM 3.2, whereas from the moment  $\beta$ -pinene was injected the model slightly overestimated the measured concentrations of HCHO, NO and  $\text{NO}_2$ .

### 3.3 Experimental OH budget analysis

In the OH budget analysis, the total OH production rate  $P_{\text{OH}}$  is compared to the OH destruction rate ( $D_{\text{OH}}$ ). Both rates ( $P_{\text{OH}}$  and  $D_{\text{OH}}$ ) were calculated from measurements performed during the experiments.  $P_{\text{OH}}$  is the sum of production rates of all known OH sources in the  $\beta$ -pinene experiments in SAPHIR: the photolysis of ozone and HONO, VOC ozonolysis, plus the OH production by the reaction of  $\text{HO}_2$  with NO and  $\text{O}_3$ . Where  $j_{\text{O}(^1\text{D})}$  and  $j_{\text{HONO}}$  are the measured photolysis frequencies of  $\text{O}_3$  and HONO,  $f_{\text{OH}}$  is the fraction of  $\text{O}(^1\text{D})$  reacting with water to OH and  $\alpha$  defines the OH yield of  $\beta$ -pinene ozonolysis. The OH destruction  $D_{\text{OH}}$  is given by the product of the measured OH reactivity and the measured OH concentration. As the short-lived OH is in steady-state,  $D_{\text{OH}}$  should be balanced by the calculated  $P_{\text{OH}}$ , if all relevant OH source terms are included in  $P_{\text{OH}}$ .

$$P_{\text{OH}} = j_{\text{O}(^1\text{D})}[\text{O}_3] \cdot 2f_{\text{OH}} + j_{\text{HONO}}[\text{HONO}] + \alpha k_1[\text{VOC}][\text{O}_3] + k_2[\text{HO}_2][\text{NO}] + k_3[\text{HO}_2][\text{O}_3] \quad (8)$$

$$D_{\text{OH}} = k(\text{OH}) \cdot [\text{OH}] \quad (9)$$

$$(10)$$

Figure 3 displays the measured OH budget of the  $\beta$ -pinene experiment on 27th August 2012. The lower panel of the plot shows the time series of the calculated OH turnover rates. The OH destruction rate  $D_{\text{OH}}$  is given as black line. The OH production rate  $P_{\text{OH}}$  is shown by the sum of the colored areas. Because of the higher instrumental accuracy the OH concentration measured by the DOAS instrument was used to calculate  $D_{\text{OH}}$ . For  $P_{\text{OH}}$  the OH recycling reaction of  $\text{HO}_2$  with NO is the dominant OH production term followed by the photolysis of HONO. The OH production by the ozonolysis

reaction of  $\beta$ -pinene is of minor importance. As mentioned in the previous section  $\text{HO}_2$  measurements include an interference from specific  $\text{RO}_2$ . For the calculation of the measured OH budget  $\text{HO}_2$  data were not corrected for an  $\text{RO}_2$  interference, as additional sensitivity studies showed that the results of the budget analysis are not affected by an assumed  $\text{RO}_2$  cross sensitivity of 25 %, because the derived  $\text{HO}_2$  concentration would be lowered by less than 10 %. The upper panel of Fig. 3 shows the time series of the ratio of  $D_{\text{OH}}/P_{\text{OH}}$  (red line). The maximum systematic error of  $D_{\text{OH}}/P_{\text{OH}}$  is indicated by the grey area. Over the course of the experiment the measured OH destruction rate is balanced by the sum of the quantifiable OH production terms within the maximum systematic error as calculated from the sum of the uncertainties of the individual measurements. Therefore the existence of a significant unknown OH source can be excluded in the degradation of  $\beta$ -pinene under the experimental conditions. This result is different to previous studies of the photooxidation of isoprene and methacrolein in SAPHIR, where the same experimental setup and similar experimental conditions were applied as in the  $\beta$ -pinene experiments. In case of isoprene and methacrolein, the OH budget analysis revealed significant additional OH sources (Fuchs et al., 2013, 2014), which were linked to OH regeneration by unimolecular reactions of  $\text{RO}_2$  and contributed as much OH as the other OH production mechanisms together. To assure the quality of the measured data used for the evaluation of the OH budget, test experiments were performed in SAPHIR with CO or  $\text{CH}_4$  as main OH reactants. These experiments were performed before and after the  $\beta$ -pinene experiments, and showed a balance between  $P_{\text{OH}}$  (Eq. 8) and  $D_{\text{OH}}$  (Eq. 9) as is expected for the well-known CO and  $\text{CH}_4$  chemistry.

### 3.4 Modifications of the $\beta$ -pinene oxidation mechanism

#### 3.4.1 The $\beta$ -pinene oxidation mechanism by Vereecken and Peeters

As discussed in sections 3.1 (see table 3) and 3.2 the primary product yields of acetone and nopinone, calculated by the MCM 3.2, are not in agreement with the determined product yields under low NO conditions in SAPHIR as well as with yields reported in the literature. For further evaluation of radical chemistry processes a good reproduction of the first-generation  $\beta$ -pinene products is essential. In the MCM 3.2 mechanism the OH radicals initially add onto the double bonds of  $\beta$ -pinene (Reactions a, b and c in Fig. 1). About 85 % of the molecules are transformed into the tertiary radicals BPINAO1. These radicals add oxygen and form peroxy radicals BPINAO2 (MCM specific designation), which react to nopinone. Acetone is a product of a minor pathway in which the four-membered ring of  $\beta$ -pinene is broken and BPINCO2 is formed (Reaction b in Fig. 1). An alternative model was published by Vereecken and Peeters (2012). Still, the addition of OH to the external carbon of the double bond forming BPINO1\* is the main reaction. But in contrast to MCM3.2 Vereecken and Peeters (2012) proposed a fast ring opening of BPINO1\* based on quantum chemical and theoretical kinetic calculations. This adjustment reduces the formation of the stabilized alkyl peroxy radical BPINAO2, the main precursor in the MCM model for nopinone formation, by about 70 %. Instead of BPINAO2 as in the MCM 3.2 mechanism, BPINCO2, is the dominant alkyl peroxy radical. With BPINCO2 as a starting point Vereecken and Peeters developed a new degradation scheme for this branch of the  $\beta$ -pinene oxidation. This leads to an increase of acetone formation at low NO concentrations compared to the MCM 3.2 while the yield of nopinone is predicted to be lower in the model by Vereecken and Peeters (2012). The model of Vereecken and Peeters (2012)

was used without further changes except for the rate constant of  $\beta$ -pinene with OH which was set to the MCM 3.2 value to facilitate model intercomparison. The original rate constant in the Vereecken and Peeters model refers to the published rate constant of Gill and Hites (2002) which is approximately 10 % lower. In the following, the MCM with the revised  $\beta$ -pinene mechanism of Vereecken and Peeters (2012)) is denoted VP2012. The result of the model calculation is shown in Fig. 2 as blue line. In comparison to the MCM 3.2 the alternative  $\beta$ -pinene degradation scheme describes the measured time series of  $k(\text{OH})$  better, assuming  $\beta$ -pinene products with a lower OH reactivity. The time behaviour of the nopinone concentration is reproduced well by Vereecken and Peeters model. The acetone formation which was slightly underestimated by MCM 3.2 is now overestimated by nearly the same amount. It should be noted that the acetone formation in the model by Vereecken and Peeters depends on fate of the radical ROO6R2O. This radical can either release acetone or undergo a hydrogen shift to yield radical ROO6R8. Unfortunately, Vereecken and Peeters could not predict the branching of these reactions accurately and were only estimating that acetone cleavage is the dominant reaction. Still, Vereecken and Peeters explicitly mark acetone formation in the current reaction conditions as a valuable metric to verify this branching ratio. The current implementation assumes 100 % acetone formation; a more balanced value of 65 % would bring the acetone yield in agreement with the experiments.

Table 4 further illustrates the difference of the product yields for acetone and nopinone calculated by the measured and modeled time series. To enable an intercomparison the product yields calculated by modeled time series were also normalized to a  $\beta$ -pinene conversion of 70 %. All the corrections applied to the measured time series were applied in the same way to the modeled data. The measured nopinone yield of the first  $\beta$ -pinene injection is about 20 % lower than the nopinone yield observed for the 2nd and 3rd injection. This feature is well captured by the MCM model even if the total nopinone yield is approximately a factor of 2 too high. The reason for the increase in the nopinone model yield is the secondary nopinone production by the degradation of previously formed hydroperoxides and organic nitrates originating from the same RO<sub>2</sub> radical which is also responsible for nopinone formation. In contrast to the MCM 3.2 the model of Vereecken and Peeters predicts a more stable nopinone yield. However, it does not include all secondary chemistry.

Over the three injections the measured acetone yield increased from 20 to 36 %, showing a clear evidence for secondary acetone production. The MCM 3.2 as well as Vereecken and Peeters model also show an increasing acetone yield over time. In the MCM 3.2 the acetone yield is much too low compared to the measurements, but increases by a factor of three during the course of the experiment due to secondary acetone formations. The acetone yield calculated by Vereecken and Peeters model for the first injection is 70 % higher than the measured value. In contrast to the time behaviour of the measured values the acetone yield is only slightly rising over the three injections, again possibly due to omitted secondary chemistry.

Concerning the agreement between measured and modeled radical concentrations the application of Vereecken and Peeters model does not lead to an improvement (see Fig. 2). The measured OH and HO<sub>2</sub><sup>\*</sup> concentrations are still underestimated in the VOC phase of the experiment. For HO<sub>2</sub><sup>\*</sup> the decrease after the first  $\beta$ -pinene injection is even more pronounced. The reason for that is the RO<sub>2</sub> interference included in the modeled HO<sub>2</sub><sup>\*</sup> data. In Vereecken and Peeters model less first-generation RO<sub>2</sub> radicals, formed by the oxidation of  $\beta$ -pinene by OH, can be theoretically detected by the LIF system. That's why directly after the first  $\beta$ -pinene injection the modeled observable RO<sub>2</sub> concentration by Vereecken and Peeters model is lower than in MCM 3.2. Simultaneously this also means that the modeled RO<sub>2</sub> interference on the HO<sub>2</sub><sup>\*</sup> time series is reduced. Compared to the

measured time series of  $\text{RO}_2^*$  Vereecken and Peeters model still overestimates the measured  $\text{RO}_2^*$  concentration. The behaviour of modeled NO,  $\text{NO}_2$ , CO and  $\text{O}_3$  is similar to the MCM 3.2: NO and  $\text{NO}_2$  concentrations are slightly overestimated by the model, CO is increasingly underestimated over time, and ozone is well captured.

In summary, it can be said that the alternative  $\beta$ -pinene degradation mechanism of Vereecken and Peeters is able to describe the measured time series of nopinone, the measured OH reactivity and with that the OH losses during the experiment much better than the MCM 3.2. However, these improvements do not lead to a satisfying description of the measured radical concentrations by the model, OH and  $\text{HO}_2^*$  are still underestimated.

The good reproduction of the total OH loss together with the underestimation of OH and  $\text{HO}_2^*$  by the model implies the need for an additional radical source to increase the modeled OH and  $\text{HO}_2$  concentration. On the other hand the OH budget analysis clearly showed that the measurable OH sources were able to balance the measured total OH loss in the experiment. With this additional information of the previous OH budget analysis, indicating no significant missing OH source, there is the arising question how the radical production can be increased without overbalancing the OH budget. One option for that is the addition of an  $\text{HO}_2$  source.

### 3.4.2 Oxidation mechanism by Vereecken and Peeters with measured $\text{HO}_2^*$ as model input

To investigate the influence of an additional  $\text{HO}_2$  source, another model run was performed using the VP2012 mechanism and the measured  $\text{HO}_2^*$  taken data as model input. The known  $\text{RO}_2$  interference in the measured  $\text{HO}_2^*$  data was taken into account and corrected in the  $\text{HO}_2$  model input. The result of the model run is displayed by the green curve in Fig. 2. Applying an additional  $\text{HO}_2$  source to the model improves the agreement of the modeled OH concentration with the measured values. In general the modeled OH increases by about 50%. The higher OH level leads to an increase of chemical conversion over time, visible in a stronger decrease of  $\beta$ -pinene, nopinone and  $k(\text{OH})$  as well as in an increase of the modeled  $\text{RO}_2^*$  concentration. Measured  $\beta$ -pinene, nopinone and  $k(\text{OH})$  are now underestimated by the model. A reason for that can be an underestimated  $\text{RO}_2$  interference assumed for the  $\text{HO}_2$  data, leading to a too strong  $\text{HO}_2$  source in the model. In the case of the OH reactivity there is the additional uncertainty of the OH rate constants for the assumed  $\beta$ -pinene oxidation products beside nopinone, causing potentially a disagreement of modeled and measured  $k(\text{OH})$ . For the overestimation of the measured  $\text{RO}_2^*$  concentration one also has to take into account that the displayed time series of modeled  $\text{RO}_2^*$  reflects the maximum  $\text{RO}_2$  concentration which is theoretically detectable by LIF. An overestimation of the measured  $\text{RO}_2^*$  concentration by the model might be related to an overestimation of the theoretically detectable  $\text{RO}_2$  species in model or an incomplete conversion of  $\beta$ -pinene derived  $\text{RO}_2$  radicals in the  $\text{RO}_x$  cell of the LIF system. In addition the increase of the modeled  $\text{HO}_2^*$  concentration leads to an improved description of the measured NO and  $\text{NO}_2$  time series. Especially in the second half of the VOC phase the modeled NO and  $\text{NO}_2$  concentration is reduced. Also the time series of HCHO is improved, whereas CO remains unchanged and is still underpredicted by the model. As in any other model run there is no influence on the modeled ozone time series.

By the application of an  $\text{HO}_2$  source to the model it was shown that the agreement between model and measurement could be improved for important key species like OH, NO and  $\text{NO}_2$ . Discrepancies in the OH lifetime and the  $\text{RO}_2^*$  concentration could

be attributed to uncertainties of the model. Therefore, a missing source of HO<sub>2</sub> in the degradation mechanism of β-pinene seems to be a reasonable hypothesis.

### 3.4.3 Uncertainties in the measured OH concentration

As stated in the previous section the input of the measured HO<sub>2</sub> concentration led to a satisfactory description of the measured OH concentration by the model. On the other hand the elevated OH concentration also resulted in an overestimated decrease of the β-pinene concentration measured by PTR-TOF-MS. From the decay of β-pinene, an OH concentration can be calculated using a reaction rate coefficient of  $7.95 \times 10^{-11} \text{ cm}^3\text{s}^{-1}$  (MCM v3.2) and taking dilution in the chamber into account. The calculated OH concentration is about 31 % lower than measured by the LIF and 24 % lower than measured by the DOAS instrument. Since both direct OH measurements agree well with each other and on the other hand the decay of β-pinene measured by PTR-TOF-MS agrees well with the decay measured by GC/MS/FID there is no clear indication for an instrumental failure or interference which would lead to an exclusion of one or the other dataset. Because this contradiction cannot be solved the implications of a potentially lower OH concentration on the previously discussed results should be elucidated. For the OH budget analysis a 24 % lower OH concentration would lead to a decrease of the calculated OH destruction ( $D_{\text{OH}}$ ) by an equal percentage.  $D_{\text{OH}}$  would be overbalanced by  $P_{\text{OH}}$ , but the mean ratio  $D_{\text{OH}}/P_{\text{OH}}$  would still not be significantly different from unity, as can be seen from its experimental error (see Fig. 3, upper panel). As reported by Nehr et al. (2014) for OH budgets during SAPHIR chamber experiments investigating CO as reference system uncertainties of  $\pm 20\%$  for  $D_{\text{OH}}/P_{\text{OH}}$  are common. For the comparison of the measured OH concentration with the model calculations, a 24 % lower measured OH concentration would result in a reduced underestimation of the measured OH concentration by the models of only 5-25 %, whereas HO<sub>2</sub>\* would still be underestimated by a factor of two. Consequently, taking the corrected HO<sub>2</sub> concentration as model input would result in an overestimation of the OH concentration by the model up to 50 %. The influence of a 24 % lower measured OH concentration on the determined product yields would be negligible because the corrections were small anyways.

## 3.5 Possible reasons for the underestimation of HO<sub>2</sub>\*

### 3.5.1 Field observations

The model simulations in the previous section demonstrated that an unaccounted source of HO<sub>2</sub> is a probable explanation for the disagreement of measured and modeled HO<sub>x</sub> concentrations. A comparison of the acquired results from the SAPHIR experiments with recent field campaigns shows qualitatively the same results as in field studies which were conducted in forested areas dominated by monoterpene emissions. Kim et al. (2013) reported a mismatch of the observed HO<sub>2</sub> concentration and model calculations. As in the SAPHIR experiments the OH budget was nearly balanced. Kim et al. postulated a missing photolytic HO<sub>2</sub> source as the reason for the discrepancy between the measured and modeled HO<sub>2</sub> concentration in a 2-methyl-3-buten-2-ol (MBO) dominated environment. Further investigations of the radical budget by Wolfe et al. (2014) came to the same result. Additionally to the missing HO<sub>2</sub> source previously postulated by Kim et al., Wolfe et al. also suggested a second

peroxy radical source, being a photolytical independent source of RO<sub>2</sub> radicals produced by the ozonolysis of unidentified VOC species. Similar to Wolfe et al. and Kim et al. also Hens et al. (2014) reported that they found an unaccounted primary HO<sub>2</sub> source, when they were comparing the measured time series of OH and HO<sub>2</sub> with model calculations. Under conditions of moderate observed OH reactivity and high actinic flux, an additional RO<sub>2</sub> source was needed to close the radical budget. Again  
5 also in the case of Hens et al. the measured OH budget was nearly balanced. In general it seems that the radical chemistry in a monoterpene dominated biogenic atmosphere in field campaigns or chamber studies recent atmospheric models underpredict the HO<sub>2</sub> production.

### 3.5.2 Model sensitivity studies

From the present study, it is obvious that an unknown HO<sub>2</sub> source is linked to the oxidation of  $\beta$ -pinene. Further model  
10 studies were performed to identify possible mechanisms that could generate additional HO<sub>2</sub>. In atmospheric chemistry, primary sources of HO<sub>2</sub> include the photolysis of aldehydes and ketones, and ozonolysis of VOCs. Furthermore, HO<sub>2</sub> is produced by the reaction of CO, ozone, or formaldehyde with OH. In the chemical degradation of VOCs, HO<sub>2</sub> can be formed by the decomposition of alkoxy radicals, and finally by unimolecular rearrangement reactions of alkyl peroxy radicals (Orlando and Tyndall, 2012). We have investigated two potential sources of HO<sub>2</sub> in separate model runs: firstly, the formation of HO<sub>2</sub> by  
15 photolysis of  $\beta$ -pinene reaction products, in particular aldehydes and ketones, and secondly additional conversion of RO<sub>2</sub> to HO<sub>2</sub> without the involvement of NO. In both cases, generic reactions were added to the chemical mechanism (see details in the Supplement). In case of the photolytical source, it was assumed that every reaction of  $\beta$ -pinene with OH produces one molecule of a carbonyl-type species Z additional to the related RO<sub>2</sub> species. It was further assumed that Z is photolysed with a rate like for formaldehyde and generates six HO<sub>2</sub> and CO molecules per molecule Z, which is in terms of chemical  
20 feasibility a rather unlikely, but not impossible assumption (see Supplement). Based on these assumptions, agreement between measurement and model is found for HO<sub>2</sub>\* and OH in the second half of the VOC phase, but in the first half of the VOC phase a strong underestimation of HO<sub>2</sub>\* remains (Figure 1, Supplement). Compared to all previous model runs, the measured concentration of CO is now well matched by the model. The modeled time series of RO<sub>2</sub>, NO, NO<sub>2</sub>, ozone and the  $\beta$ -pinene products formaldehyde, acetone and nopinone stay nearly unchanged in comparison to the model run using measured HO<sub>2</sub> as  
25 model input. In conclusion, the assumed photolytical HO<sub>2</sub> source gives an improved model description of the observations, but is not capable to regenerate HO<sub>2</sub> sufficiently fast in the first 1-2 hours after the first  $\beta$ -pinene addition.

Next, the possible influence of unimolecular rearrangement of RO<sub>2</sub> yielding HO<sub>2</sub> was studied (see Supplement). For this purpose, the so called X-mechanism published by Hofzumahaus et al. (2009) was used. An NO like species X is thereby reducing RO<sub>2</sub> radicals to RO radicals. The rate constants applied for these reactions are the same as the rate coefficients of  
30 NO with the corresponding RO<sub>2</sub> radical. Contrary to the X-mechanism of Hofzumahaus et al., in case of  $\beta$ -pinene, X is not reacting with HO<sub>2</sub> radicals. With 300 pptv of X, the model gives a significantly improved description of HO<sub>2</sub>\*, but an underprediction of 25 % remains (Figure 1, Supplement). The introduction of X causes a substantial decrease of RO<sub>2</sub>\* and a significant overprediction of NO and NO<sub>2</sub> by the model. Also CO is greatly overestimated. In conclusion, additional RO<sub>2</sub> to HO<sub>2</sub> conversion (without NO) alone is also not capable to describe all observations consistently.

Two additional model sensitivity tests were carried out in order to investigate, if the  $\text{HO}_2^*$  underprediction is caused by too fast  $\text{RO}_2 + \text{HO}_2$  reactions in the Vereecken and Peters model, and how the model measurement comparison is influenced by uncertainties of the  $\text{RO}_2$  interference in the  $\text{HO}_2$  measurements (see details in Supplement).

In accordance with a proposed uncertainty of a factor of two for the rate constants of biogenic  $\text{RO}_2 + \text{HO}_2$  reactions (Orlando and Tyndall, 2012), the rate constants for the formation of ROOH were reduced by 50%. As a result the modeled  $\text{HO}_2$  concentration increases by 30%, but  $\text{HO}_2$  is still underestimated by the model (Figure 1, Supplement). The modeled OH concentration slightly increases and the measured  $\text{RO}_2$  concentration becomes overestimated by a factor of two. The measured concentrations of NO and  $\text{NO}_2$  are well matched by the model, but CO remains underestimated. In conclusion, a reduction of the ROOH production may help to reduce the discrepancy between the modeled and measured  $\text{HO}_2$  concentration, but cannot solely explain the deviations between model and measurements. As the interference of  $\text{RO}_2$  radicals in the measurements of  $\text{HO}_2$  is also a subject of discussions, the maximum influence of the assumed  $\text{RO}_2$  interference on the model results was estimated in a fourth model case (see Supplement). The sensitivity study proved that the interference of the  $\text{RO}_2$  radicals on the measured  $\text{HO}_2$  time series is incapable to explain the observed deviations between modeled and measured  $\text{HO}_2$ . More than 50% of the observed discrepancy cannot be explained by any known interference.

### 3.5.3 Modifications of the $\beta$ -pinene oxidation mechanism by Vereecken and Peeters to explain the missing $\text{HO}_2^*$ source

The major difference of the  $\beta$ -pinene oxidation mechanism by Vereecken and Peeters compared to the MCM 3.2 mechanism is the fast ring opening of the alkoxy radical BPINO1\* which is transformed into the radical BPINCO1 (see Fig. 1). At low NO concentrations the largest fraction of these molecules are expected to react to ROO6R2O. The formation of ROO6R2O is exothermic and the reaction sequence can either proceed via elimination of acetone (path e in Fig. 1 and Fig. 4) or via 1,5-H-migration of the hydrogen at  $\alpha$  position of the aldehyde (path f in Fig. 1 and in Fig. 4).

The branching ratio of path b and path a significantly influences the amount of  $\text{HO}_2$  produced. After acetone is eliminated ROO6R8 radicals add two oxygen molecules. The emerged radical cleaves an OH radical and forms a peracid. No additional  $\text{HO}_2$  radicals are supposed to be produced if degradation of the radical ROO6R2O proceeds via the acetone elimination channel.

If instead the hydrogen atom on  $\alpha$  position of the aldehyde of ROO6R2O migrates, ROO6R8 is formed. This acyl radical is supposed to cleave CO, and after another 1,5-H-migration, also  $\text{HO}_2$ . The resulting molecule is the dicarbonyl compound ROO6R9P whose photolytical cleavage results in the additional production of one molecule CO and one molecule  $\text{HO}_2$  (path f in Fig. 4). Unfortunately, Vereecken and Peeters could not accurately predict the branching ratio of these two reaction channels due to large number of active conformers at higher energies. Instead, the 1,5-H-migration in path f was supposed to be outrun by acetone elimination in path e, and path e was omitted in the model of Vereecken and Peeters.

The effect of the branching ratio in Fig. 4 on the predicted  $\text{HO}_2^*$  concentration can be evaluated if ROO6R2O is fixed in the model to react exclusively via path f. The respective model run (see the green curve in Fig. 5) predicts a  $\text{HO}_2^*$  concentration

which is 30 % higher than forecasted by the original model of Vereecken and Peeters. Also, the predicted CO, RO<sub>2</sub><sup>\*</sup>, HCHO and nopinone concentration now coincide with the the measured data. Still, the measured HO<sub>2</sub><sup>\*</sup> is 20 % higher than prognosticated.

The gap between measured HO<sub>2</sub><sup>\*</sup> and modeled HO<sub>2</sub><sup>\*</sup> can be closed if the cleavage of a second HO<sub>2</sub> is incorporated into the model (see orange curve in Fig. 5). The time series of HO<sub>2</sub><sup>\*</sup>, RO<sub>2</sub><sup>\*</sup> and OH are now captured by the model. Also the measured  
5 nopinone, CO and HCHO are well described. Only acetone is now underestimated by the model, because acetone is mainly formed via the pathway e in Fig. 4. Although ROO6R9P can potentially cleave acetone, quantum chemical calculations are needed further to pin down the mechanism of acetone cleavage.

#### 4 Summary and Conclusions

A set of three  $\beta$ -pinene oxidation experiments, conducted in the SAPHIR atmosphere simulation chamber, was comprehensively investigated with regard to the involved radical species during the OH oxidation. A special focus was placed on the  
10 identification of possible missing OH production terms in the degradation mechanism (Whalley et al., 2011). The experiments were conducted under nearly ambient  $\beta$ -pinene concentration (4.3-4.7 ppb VOC) and low NO conditions (100-300 ppt NO). The comparatively low VOC concentration allowed for the first time the investigation of the radical budget of  $\beta$ -pinene by parallel measurements of OH, HO<sub>2</sub>, RO<sub>2</sub> and  $k(\text{OH})$ . In a first approach this comprehensive dataset was used for a model  
15 independent analysis of the OH budget. For this purpose the sum of the measurable OH production terms (HONO photolysis, O<sub>3</sub> photolysis, VOC ozonolysis, HO<sub>2</sub>+NO, HO<sub>2</sub>+O<sub>3</sub>) was compared with the measured OH destruction rate ( $k(\text{OH}) \times [\text{OH}]$ ). Contrary to previous studies of isoprene and methacrolein in SAPHIR (Fuchs et al., 2013, 2014) the OH budget was balanced in the  $\beta$ -pinene oxidation experiments, giving no evidence for significant missing OH production terms. In a second approach the measured time series of the atmospheric key species were compared to zero-dimensional box model calculations to investigate  
20 whether the models are able to predict the  $\beta$ -pinene degradation well. The comparison of the measured time series with the MCM 3.2 revealed that the model was not able to reproduce the measured time series of OH, HO<sub>2</sub>,  $k(\text{OH})$  and nopinone. The modeled OH as well as the HO<sub>2</sub> concentration was underestimated by more than 50 %. At the same time the modeled OH reactivity was slightly overestimated. The reason for this disagreement is obviously a biased product distribution of the first-generation degradation products. The measured nopinone concentration was about a factor of three lower than predicted  
25 by the model. A comparison of the experimentally determined nopinone yield with recent literature showed a good agreement but is a factor of two lower than in the MCM model. Hence, for further investigations an updated MCM mechanism published by Vereecken and Peeters (2012) was used. Their model was able to reproduce the measured time series of nopinone and  $k(\text{OH})$  much better than the MCM 3.2, but still significantly underpredicted the measured OH and HO<sub>2</sub> concentration. As the previous analysis of the OH budget showed no evidence of a missing OH source, an additional HO<sub>2</sub> source was introduced into  
30 the model to improve the agreement for OH and HO<sub>2</sub>. A sensitivity study showed that taking the measured HO<sub>2</sub> time series as model input generally improves the overall agreement of the modeled time series with the measurements. OH is now well described by the model. These findings are qualitatively in agreement with recent field studies (Kim et al., 2013; Wolfe et al.,

2014; Hens et al., 2014) reporting that in a monoterpene dominated biogenic atmosphere models were not able to describe OH and HO<sub>2</sub> levels well although the measured OH budget was balanced.

In accordance with the results for  $\beta$ -pinene presented in this paper we propose an additional HO<sub>2</sub> source linked to  $\beta$ -pinene oxidation products as the reason for the underestimation of OH and HO<sub>2</sub> in the model. With additional sensitivity studies  
5 it was possible to rule out photolytical processes or rearrangement reactions of RO<sub>2</sub> as sole HO<sub>2</sub> sources. Also a possible overestimation of the yield of organic hydroperoxides as well an underestimation of the known RO<sub>2</sub> interference on the HO<sub>2</sub> measurements were excluded as explanations for underestimating HO<sub>2</sub> in the model.

The gap between measured and modeled HO<sub>2</sub><sup>\*</sup> concentration can significantly be reduced modifying the mechanism of Vereecken and Peeters such that the radical intermediate ROO6R2O rearranges rather than being cleaved. The resulting acyl  
10 radical produces HO<sub>2</sub>, CO and a dicarbonyl compound which itself is a photolytical source of HO<sub>2</sub> and CO. Still, the exact HO<sub>2</sub> formation mechanism remains uncertain. Additional experiments and quantum chemical calculations have to be made to completely unravel the pathway of HO<sub>2</sub> formation.

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## References

- Arey, J., Atkinson, R., and Aschmann, S. M.: Product study of the gas-phase reactions of monoterpenes with the OH radical in the presence of NO<sub>x</sub>, *J. Geophys. Res. - Atmos.*, 95, 18 539–18 546, doi:10.1029/JD095iD11p18539, 1990.
- Atkinson, R.: Atmospheric chemistry of VOCs and NO<sub>x</sub>, *Atmos. Environ.*, 34, 2063–2101, doi:10.1016/s1352-2310(99)00460-4, 2000.
- 5 Atkinson, R. and Arey, J.: Gas-phase tropospheric chemistry of biogenic volatile organic compounds: a review, *Atmos. Environ.*, 37, Supplement 2, 197–219, doi:10.1016/s1352-2310(03)00391-1, 2003.
- Bates, K. H., Crouse, J. D., St Clair, J. M., Bennett, N. B., Nguyen, T. B., Seinfeld, J. H., Stoltz, B. M., and Wennberg, P. O.: Gas Phase Production and Loss of Isoprene Epoxydiols, *J. Phys. Chem. A*, 118, 1237–1246, doi:10.1021/jp4107958, 2014.
- Bohn, B. and Zilken, H.: Model-aided radiometric determination of photolysis frequencies in a sunlit atmosphere simulation chamber, *Atmos. Chem. Phys.*, 5, 191–206, doi:10.5194/acp-5-191-2005, 2005.
- 10 Bohn, B., Rohrer, F., Brauers, T., and Wahner, A.: Actinometric Measurements of NO<sub>2</sub> Photolysis Frequencies in the Atmosphere Simulation Chamber SAPHIR, *Atmos. Chem. Phys.*, 5, 493503, doi:10.5194/acp-5-493-2005, 2005.
- Brauers, T., Bossmeyer, J., Dorn, H. P., Schlosser, E., Tillmann, R., Wegener, R., and Wahner, A.: Investigation of the formaldehyde differential absorption cross section at high and low spectral resolution in the simulation chamber SAPHIR, *Atmos. Chem. Phys.*, 7, 3579–3586, doi:10.5194/acp-7-3579-2007, 2007.
- 15 Calogirou, A., Larsen, B. R., and Kotzias, D.: Gas-phase terpene oxidation products: a review, *Atmos. Environ.*, 33, 1423–1439, doi:10.1016/s1352-2310(98)00277-5, 1999.
- Carslaw, N., Creasey, D., Harrison, D., Heard, D., Hunter, M., Jacobs, P., Jenkin, M., Lee, J., Lewis, A., Pilling, M., Saunders, S., and Seakins, P.: OH and HO<sub>2</sub> radical chemistry in a forested region of north-western Greece, *Atmos. Environ.*, 35, 4725–4737, doi:10.1016/S1352-2310(01)00089-9, 2001.
- 20 Crouse, J. D., Paulot, F., Kjaergaard, H. G., and Wennberg, P. O.: Peroxy radical isomerization in the oxidation of isoprene, *Phys. Chem. Chem. Phys.*, 13, 13 607–13 613, doi:10.1039/C1CP21330J, 2011.
- Crouse, J. D., Knap, H. C., Ørnsø, K. B., Jørgensen, S., Paulot, F., Kjaergaard, H. G., and Wennberg, P. O.: Atmospheric Fate of Methacrolein. 1. Peroxy Radical Isomerization Following Addition of OH and O<sub>2</sub>, *J. Phys. Chem. A*, 116, 5756–5762, doi:10.1021/jp211560u, 2012.
- 25 Crouse, J. D., Nielsen, L. B., Jørgensen, S., Kjaergaard, H. G., and Wennberg, P. O.: Autoxidation of Organic Compounds in the Atmosphere, *J. Phys. Chem. Lett.*, 4, 3513–3520, doi:10.1021/jz4019207, 2013.
- da Silva, G., Graham, C., and Wang, Z.-F.: Unimolecular  $\beta$ -Hydroxyperoxy Radical Decomposition with OH Recycling in the Photochemical Oxidation of Isoprene, *Environ. Sci. Technol.*, 44, 250–256, doi:10.1021/es900924d, 2010.
- 30 Dorn, H.-P., Brandenburger, U., Brauers, T., and Hausmann, M.: A New In Situ Laser Long-Path Absorption Instrument for the Measurement of Tropospheric OH Radicals, *J. Atmos. Sci.*, 52, 3373–3380, doi:10.1175/1520-0469(1995)052<3373:ANISLL>2.0.CO;2, 1995.
- Dorn, H.-P., Apodaca, R. L., Ball, S., Brauers, T., Brown, S., Crowley, J., Dubé, W., Häsel, R., Heitmann, U., Jones, R., Kiendler-Scharr, A., Labazan, I., Langridge, J., Meinen, J., Mentel, T., Platt, U., Pöhler, D., Rohrer, F., Ruth, A., Schlosser, E., Schuster, G., Schillings, A., Simpson, W., Thieser, J., Tillmann, R., Varma, R., Venebles, D., and Wahner, A.: Intercomparison of NO<sub>3</sub> radical detection instruments in the atmosphere simulation chamber SAPHIR, *Atmos. Meas. Tech.*, 6, 1111–1140, doi:10.5194/amt-6-1111-2013, 2013.
- 35

- Eddingsaas, N. C., Loza, C. L., Yee, L. D., Chan, M., Schilling, K. A., Chhabra, P. S., Seinfeld, J. H., and Wennberg, P. O.:  $\alpha$ -pinene photooxidation under controlled chemical conditions – Part 2: SOA yield and composition in low- and high-NO<sub>x</sub> environments, *Atmos. Chem. Phys.*, 12, 7413–7427, doi:10.5194/acp-12-7413-2012, 2012a.
- Eddingsaas, N. C., Loza, C. L., Yee, L. D., Seinfeld, J. H., and Wennberg, P. O.:  $\alpha$ -pinene photooxidation under controlled chemical conditions – Part 1: Gas-phase composition in low- and high-NO<sub>x</sub> environments, *Atmos. Chem. Phys.*, 12, 6489–6504, doi:10.5194/acp-12-6489-2012, 2012b.
- Ehn, M., Thornton, J. A., Kleist, E., Sipila, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F., Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I.-H., Rissanen, M., Jokinen, T., Schobesberger, S., Kangasluoma, J., Kontkanen, J., Nieminen, T., Kurten, T., Nielsen, L. B., Jorgensen, S., Kjaergaard, H. G., Canagaratna, M., Dal Maso, M., Berndt, T., Petaja, T., Wahner, A., Kerminen, V.-M., Kulmala, M., Worsnop, D. R., Wildt, J., and Mentel, T. F.: A large source of low-volatility secondary organic aerosol, *Nature*, 506, 476+, doi:10.1038/nature13032, 2014.
- Fantechi, G.: Atmospheric oxidation reactions of selected biogenic volatile organic compounds (BIOVOCs): A smog chamber study, Ph.D. thesis, KULeuven, 1999.
- Feiner, P. A., Brune, W. H., Miller, D. O., Zhang, L., Cohen, R. C., Romer, P. S., Goldstein, A. H., Keutsch, F. N., Skog, K. M., Wennberg, P. O., Nguyen, T. B., Teng, A. P., DeGouw, J., Koss, A., Wild, R. J., Brown, S. S., Guenther, A., Edgerton, E., Baumann, K., and Fry, J. L.: Testing Atmospheric Oxidation in an Alabama Forest, *J. Atmos. Sci.*, 73, 4699–4710, doi:10.1175/JAS-D-16-0044.1, 2016.
- Fuchs, H., Holland, F., and Hofzumahaus, A.: Measurement of tropospheric RO<sub>2</sub> and HO<sub>2</sub> radicals by a laser-induced fluorescence instrument, *Rev. Sci. Instrum.*, 79, 084 104, doi:10.1063/1.2968712, 2008.
- Fuchs, H., Bohn, B., Hofzumahaus, A., Holland, F., Lu, K. D., Nehr, S., Rohrer, F., and Wahner, A.: Detection of HO<sub>2</sub> by laser-induced fluorescence: calibration and interferences from RO<sub>2</sub> radicals, *Atmos. Meas. Tech.*, 4, 1209–1225, doi:10.5194/amt-4-1209-2011, 2011.
- Fuchs, H., Dorn, H. P., Bachner, M., Bohn, B., Brauers, T., Gomm, S., Hofzumahaus, A., Holland, F., Nehr, S., Rohrer, F., Tillmann, R., and Wahner, A.: Comparison of OH concentration measurements by DOAS and LIF during SAPHIR chamber experiments at high OH reactivity and low NO concentration, *Atmos. Meas. Tech.*, 5, 1611–1626, doi:10.5194/amt-5-1611-2012, 2012.
- Fuchs, H., Hofzumahaus, A., Rohrer, F., Bohn, B., Brauers, T., Dorn, H. P., Häsel, R., Holland, F., Kaminski, M., Li, X., Lu, K., Nehr, S., Tillmann, R., Wegener, R., and Wahner, A.: Experimental evidence for efficient hydroxyl radical regeneration in isoprene oxidation, *Nature Geosci.*, 6, 1023–1026, doi:10.1038/ngeo1964, 2013.
- Fuchs, H., Acir, I. H., Bohn, B., Brauers, T., Dorn, H. P., Häsel, R., Hofzumahaus, A., Holland, F., Kaminski, M., Li, X., Lu, K., Lutz, A., Nehr, S., Rohrer, F., Tillmann, R., Wegener, R., and Wahner, A.: OH regeneration from methacrolein oxidation investigated in the atmosphere simulation chamber SAPHIR, *Atmos. Chem. Phys.*, 14, 7895–7908, doi:10.5194/acp-14-7895-2014, 2014.
- Fuchs, H., Tan, Z. F., Hofzumahaus, A., Broch, S., Dorn, H. P., Holland, F., Kunstler, C., Gomm, S., Rohrer, F., Schrade, S., Tillmann, R., and Wahner, A.: Investigation of potential interferences in the detection of atmospheric RO<sub>x</sub> radicals by laser-induced fluorescence under dark conditions, *Atmos. Meas. Tech.*, 9, 1431–1447, doi:10.5194/amt-9-1431-2016, 2016.
- Galloway, M. M., Huisman, A. J., Yee, L. D., Chan, A. W. H., Loza, C. L., Seinfeld, J. H., and Keutsch, F. N.: Yields of oxidized volatile organic compounds during the OH radical initiated oxidation of isoprene, methyl vinyl ketone, and methacrolein under high-NO<sub>x</sub> conditions, *Atmos. Chem. Phys.*, 11, 10 779–10 790, doi:10.5194/acp-11-10779-2011, 2011.
- Gill, K. J. and Hites, R. A.: Rate Constants for the Gas-Phase Reactions of the Hydroxyl Radical with Isoprene,  $\alpha$ - and  $\beta$ -Pinene, and Limonene as a Function of Temperature, *J. Phys. Chem. A*, 106, 2538–2544, doi:10.1021/jp013532q, 2002.

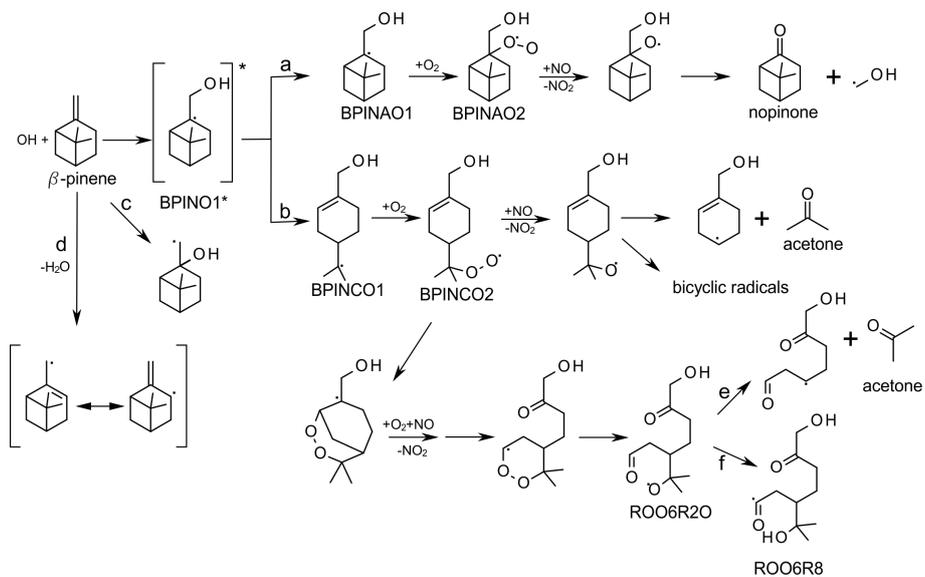
- Goldstein, A. H. and Galbally, I. E.: Known and Unexplored Organic Constituents in the Earth's Atmosphere, *Environ. Sci. Technol.*, 41, 1514–1521, doi:10.1021/es072476p, 2007.
- Griffith, S. M., Hansen, R. F., Dusanter, S., Stevens, P. S., Alaghmand, M., Bertman, S. B., Carroll, M. A., Erickson, M., Galloway, M., Grossberg, N., Hottle, J., Hou, J., Jobson, B. T., Kammrath, A., Keutsch, F. N., Lefer, B. L., Mielke, L. H., O'Brien, A., Shepson, P. B., Thurlow, M., Wallace, W., Zhang, N., and Zhou, X. L.: OH and HO<sub>2</sub> radical chemistry during PROPHET 2008 and CABINEX 2009-Part 1: Measurements and model comparison, *Atmos. Chem. Phys.*, 13, 5403–5423, doi:10.5194/acp-13-5403-2013, 2013.
- Guenther, A., Hewitt, C. N., Erickson, D., Fall, R., Geron, C., Graedel, T., Harley, P., Klinger, L., Lerdau, M., McKay, W. A., Pierce, T., Scholes, B., Steinbrecher, R., Tallamraju, R., Taylor, J., and Zimmerman, P.: A global model of natural volatile organic compound emissions, *J. Geophys. Res. - Atmos.*, 100, 8873–8892, doi:10.1029/94JD02950, 1995.
- 10 Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittay, T., Duhl, T., Emmons, L. K., and Wang, X.: The Model of Emissions of Gases and Aerosols from Nature version 2.1 (MEGAN2.1): An extended and updated framework for modeling biogenic emissions, *Geosci. Model Dev.*, 5, 1471–1492, doi:10.5194/gmd-5-1471-2012, 2012.
- Hakola, H., Arey, J., Aschmann, S., and Atkinson, R.: Product formation from the gas-phase reactions of OH radicals and O<sub>3</sub> with a series of monoterpenes, *J. Atmos. Chem.*, 18, 75–102, doi:10.1007/BF00694375, 1994.
- 15 Hatakeyama, S., Izumi, K., Fukuyama, T., Akimoto, H., and Washida, N.: Reactions of OH with  $\alpha$ -pinene and  $\beta$ -pinene in air: Estimate of global CO production from the atmospheric oxidation of terpenes, *J. Geophys. Res. - Atmos.*, 96, 947–958, doi:10.1029/90JD02341, 1991.
- Hausmann, M., Brandenburger, U., Brauers, T., and Dorn, H.-P.: Detection of tropospheric OH radicals by long-path differential-optical-absorption spectroscopy: Experimental setup, accuracy, and precision, *J. Geophys. Res. - Atmos.*, 102, 16 011–16 022, doi:10.1029/97JD00931, 1997.
- 20 Hens, K., Novelli, A., Martinez, M., Auld, J., Axinte, R., Bohn, B., Fischer, H., Keronen, P., Kubistin, D., Nölscher, A. C., Oswald, R., Paasonen, P., Petäjä, T., Regelin, E., Sander, R., Sinha, V., Sipilä, M., Taraborrelli, D., Tatum Ernest, C., Williams, J., Lelieveld, J., and Harder, H.: Observation and modelling of HO<sub>x</sub> radicals in a boreal forest, *Atmos. Chem. Phys.*, 14, 8723–8747, doi:10.5194/acp-14-8723-2014, 2014.
- 25 Hofzumahaus, A. and Heard, D. H.: Assessment of local HO<sub>x</sub> and RO<sub>x</sub> Measurement Techniques: Achievements, Challenges, and Future Directions. Report of the International HO<sub>x</sub> Workshop 2015, Jülich, Tech. rep., Forschungszentrum Jülich, 2016.
- Hofzumahaus, A., Rohrer, F., Lu, K., Bohn, B., Brauers, T., Chang, C.-C., Fuchs, H., Holland, F., Kita, K., Kondo, Y., Li, X., Lou, S., Shao, M., Zeng, L., Wahner, A., and Zhang, Y.: Amplified Trace Gas Removal in the Troposphere, *Science*, 324, 1702–1704, doi:10.1126/science.1164566, 2009.
- 30 Holland, F., Heßling, M., and Hofzumahaus, A.: *In situ* measurement of tropospheric OH radicals by laser-induced fluorescence – A description of the KFA instrument, *J. Atmos. Sci.*, 52, 3393–3401, doi:10.1175/1520-0469(1995)052<3393:ismoto>2.0.co;2, 1995.
- Häseler, R., Brauers, T., Holland, F., and Wahner, A.: Development and application of a new mobile LOPAP instrument for the measurement of HONO altitude profiles in the planetary boundary layer, *Atmos. Meas. Tech. Discuss.*, 2, 2027–2054, doi:10.5194/amtd-2-2027-2009, 2009.
- 35 Jenkin, M. E., Saunders, S. M., and Pilling, M. J.: The tropospheric degradation of volatile organic compounds: A protocol for mechanism development, *Atmos. Environ.*, 31, 81–104, doi:10.1016/s1352-2310(96)00105-7, 1997.

- Jordan, A., Haidacher, S., Hanel, G., Hartungen, E., Märk, L., Seehauser, H., Schottkowsky, R., Sulzer, P., and Märk, T. D.: A high resolution and high sensitivity proton-transfer-reaction time-of-flight mass spectrometer (PTR-TOF-MS), *Int. J. Mass Spectrom.*, 286, 122–128, doi:10.1016/j.ijms.2009.07.005, 2009.
- Kaminski, M.: Untersuchung des photochemischen Terpenoidabbaus in der Atmosphärensimulationskammer SAPHIR, Forschungszentrum Jülich, 2014.
- Kanakidou, M., Seinfeld, J. H., Pandis, S. N., Barnes, I., Dentener, F. J., Facchini, M. C., Van Dingenen, R., Ervens, B., Nenes, A., Nielsen, C. J., Swietlicki, E., Putaud, J. P., Balkanski, Y., Fuzzi, S., Horth, J., Moortgat, G. K., Winterhalter, R., Myhre, C. E. L., Tsigaridis, K., Vignati, E., Stephanou, E. G., and Wilson, J.: Organic aerosol and global climate modelling: a review, *Atmos. Chem. Phys.*, 5, 1053–1123, 2005.
- 10 Karl, M., Dorn, H. P., Holland, F., Koppmann, R., Poppe, D., Rupp, L., Schaub, A., and Wahner, A.: Product study of the reaction of OH radicals with isoprene in the atmosphere simulation chamber SAPHIR, *J. Atmos. Chem.*, 55, 167–187, doi:10.1007/s10874-006-9034-x, 2006.
- Kim, S., Wolfe, G. M., Mauldin, L., Cantrell, C., Guenther, A., Karl, T., Turnipseed, A., Greenberg, J., Hall, S. R., Ullmann, K., Apel, E., Hornbrook, R., Kajii, Y., Nakashima, Y., Keutsch, F. N., DiGangi, J. P., Henry, S. B., Kaser, L., Schnitzhofer, R., Graus, M., Hansel, A.,  
15 Zheng, W., and Flocke, F. F.: Evaluation of HO<sub>x</sub> sources and cycling using measurement-constrained model calculations in a 2-methyl-3-butene-2-ol (MBO) and monoterpene (MT) dominated ecosystem, *Atmos. Chem. Phys.*, 13, 2031–2044, doi:10.5194/acp-13-2031-2013, 2013.
- Kubistin, D., Harder, H., Martinez, M., Rudolf, M., Sander, R., Bozem, H., Eerdekens, G., Fischer, H., Gurk, C., Klüpfel, T., Königstedt, R., Parchatka, U., Schiller, C. L., Stickler, A., Taraborrelli, D., Williams, J., and Lelieveld, J.: Hydroxyl radicals in the tropical tropo-  
20 sphere over the Suriname rainforest: comparison of measurements with the box model MECCA, *Atmos. Chem. Phys.*, 10, 9705–9728, doi:10.5194/acp-10-9705-2010, 2010.
- Larsen, B. R., Di Bella, D., Glasius, M., Winterhalter, R., Jensen, N. R., and Hjorth, J.: Gas-Phase OH Oxidation of Monoterpenes: Gaseous and Particulate Products, *J. Atmos. Chem.*, 38, 231–276, doi:10.1023/A:1006487530903, 2001.
- Lee, A., Goldstein, A. H., Keywood, M. D., Gao, S., Varutbangkul, V., Bahreini, R., Ng, N. L., Flagan, R. C., and Seinfeld, J. H.:  
25 Gas-phase products and secondary aerosol yields from the ozonolysis of ten different terpenes, *J. Geophys. Res. - Atmos.*, 111, doi:10.1029/2005jd006437, 2006.
- Lelieveld, J., Butler, T. M., Crowley, J. N., Dillon, T. J., Fischer, H., Ganzeveld, L., Harder, H., Lawrence, M. G., Martinez, M., Taraborrelli, D., and Williams, J.: Atmospheric oxidation capacity sustained by a tropical forest, *Nature*, 452, 737–740, doi:10.1038/nature06870, 2008.
- Librando, V. and Tringali, G.: Atmospheric fate of OH initiated oxidation of terpenes. Reaction mechanism of  $\alpha$ -pinene degradation and  
30 secondary organic aerosol formation, *J. Environ. Manage.*, 75, 275–282, doi:10.1016/j.jenvman.2005.01.001, 2005.
- Lindinger, W., Hansel, A., and Jordan, A.: On-line monitoring of volatile organic compounds at pptv levels by means of Proton-Transfer-Reaction Mass Spectrometry (PTR-MS) Medical applications, food control and environmental research, *Int. J. Mass Spectrom. Ion Process.*, 173, 191–241, doi:10.1016/s0168-1176(97)00281-4, 1998.
- Liu, Y. J., Herdlinger-Blatt, I., McKinney, K. A., and Martin, S. T.: Production of methyl vinyl ketone and methacrolein via the hydroperoxyl  
35 pathway of isoprene oxidation, *Atmos. Chem. Phys.*, 13, 5715–5730, doi:10.5194/acp-13-5715-2013, 2013.
- Lou, S., Holland, F., Rohrer, F., Lu, K., Bohn, B., Brauers, T., Chang, C. C., Fuchs, H., Häseler, R., Kita, K., Kondo, Y., Li, X., Shao, M., Zeng, L., Wahner, A., Zhang, Y., Wang, W., and Hofzumahaus, A.: Atmospheric OH reactivities in the Pearl River Delta - China in summer 2006: measurement and model results, *Atmos. Chem. Phys.*, 10, 11 243–11 260, doi:10.5194/acp-10-11243-2010, 2010.

- Lu, K. D., Rohrer, F., Holland, F., Fuchs, H., Bohn, B., Brauers, T., Chang, C. C., Häsel, R., Hu, M., Kita, K., Kondo, Y., Li, X., Lou, S. R., Nehr, S., Shao, M., Zeng, L. M., Wahner, A., Zhang, Y. H., and Hofzumahaus, A.: Observation and modelling of OH and HO<sub>2</sub> concentrations in the Pearl River Delta 2006: a missing OH source in a VOC rich atmosphere, *Atmos. Chem. Phys.*, 12, 1541–1569, doi:10.5194/acp-12-1541-2012, 2012.
- 5 Mao, J., Ren, X., Zhang, L., Van Duin, D. M., Cohen, R. C., Park, J.-H., Goldstein, A. H., Paulot, F., Beaver, M. R., Crounse, J. D., Wennberg, P. O., DiGangi, J. P., Henry, S. B., Keutsch, F. N., Park, C., Schade, G. W., Wolfe, G. M., Thornton, J. A., and Brune, W. H.: Insights into hydroxyl measurements and atmospheric oxidation in a California forest, *Atmospheric Chemistry and Physics*, 12, 8009–8020, doi:10.5194/acp-12-8009-2012, 2012.
- Nehr, S., Bohn, B., Dorn, H. P., Fuchs, H., Häsel, R., Hofzumahaus, A., Li, X., Rohrer, F., Tillmann, R., and Wahner, A.: Atmospheric photochemistry of aromatic hydrocarbons: OH budgets during SAPHIR chamber experiments, *Atmos. Chem. Phys.*, 14, 6941–6952, doi:10.5194/acp-14-6941-2014, 2014.
- 10 Novelli, A., Vereecken, L., Lelieveld, J., and Harder, H.: Direct observation of OH formation from stabilised Criegee intermediates, *Physical Chemistry Chemical Physics*, 16, 19941–19951, doi:10.1039/c4cp02719a, 2014.
- Orlando, J. J. and Tyndall, G. S.: Laboratory studies of organic peroxy radical chemistry: an overview with emphasis on recent issues of atmospheric significance, *Chem. Soc. Rev.*, 41, 6294–6317, doi:10.1039/C2CS35166H, 2012.
- 15 Orlando, J. J., Nozière, B., Tyndall, G. S., Orzechowska, G. E., Paulson, S. E., and Rudich, Y.: Product studies of the OH- and ozone-initiated oxidation of some monoterpenes, *J. Geophys. Res. - Atmos.*, 105, 11 561–11 572, doi:10.1029/2000JD900005, 2000.
- Paulot, F., Crounse, J. D., Kjaergaard, H. G., Kürten, A., St. Clair, J. M., Seinfeld, J. H., and Wennberg, P. O.: Unexpected Epoxide Formation in the Gas-Phase Photooxidation of Isoprene, *Science*, 325, 730–733, doi:10.1126/science.1172910, 2009.
- 20 Peeters, J. and Müller, J.-F.: HO<sub>x</sub> radical regeneration in isoprene oxidation via peroxy radical isomerisations. II: experimental evidence and global impact, *Phys. Chem. Chem. Phys.*, 12, 14 227–14 235, doi:10.1039/C0CP00811G, 2010.
- Peeters, J., Müller, J.-F., Stavrou, T., and Nguyen, V. S.: Hydroxyl Radical Recycling in Isoprene Oxidation Driven by Hydrogen Bonding and Hydrogen Tunneling: The Upgraded LIM1 Mechanism, *J. Phys. Chem. A*, 118, 8625–8643, doi:10.1021/jp5033146, 2014.
- Piccot, S. D., Watson, J. J., and Jones, J. W.: A global inventory of volatile organic compound emissions from anthropogenic sources, *J. Geophys. Res. - Atmos.*, 97, 9897–9912, doi:10.1029/92JD00682, 1992.
- 25 Poppe, D., Brauers, T., Dorn, H.-P., Karl, M., Mentel, T., Schlosser, E., Tillmann, R., Wegener, R., and Wahner, A.: OH-initiated degradation of several hydrocarbons in the atmosphere simulation chamber SAPHIR, *J. Atmos. Chem.*, 57, 203–214, doi:10.1007/s10874-007-9065-y, 2007.
- Reissell, A., Harry, C., Aschmann, S. M., Atkinson, R., and Arey, J.: Formation of acetone from the OH radical- and O<sub>3</sub>-initiated reactions of a series of monoterpenes, *J. Geophys. Res. - Atmos.*, 104, 13 869–13 879, doi:10.1029/1999JD900198, 1999.
- 30 Ren, X., Olson, J. R., Crawford, J. H., Brune, W. H., Mao, J., Long, R. B., Chen, G., Avery, M. A., Sachse, G. W., Barrick, J. D., Diskin, G. S., Huey, L. G., Fried, A., Cohen, R. C., Heikes, B., Wennberg, P., Singh, H. B., Richard, D. R. B., and Shetter, E.: HO<sub>x</sub> Chemistry during INTEX-A 2004: Observation, Model Calculations and comparison with previous studies, *J. Geophys. Res.*, 113, D05 310, doi:10.1029/2007JD009166, 2008.
- 35 Ridley, B. A., Grahek, F. E., and Walega, J. G.: A Small High-Sensitivity, Medium-Response Ozone Detector Suitable for Measurements from Light Aircraft, *J. Atmos. Oceanic Technol.*, 9, 142–148, doi:10.1175/1520-0426(1992)009<0142:ASHSMR>2.0.CO;2, 1992.
- Rohrer, F. and Brüning, D.: Surface NO and NO<sub>2</sub> mixing ratios measured between 30° N and 30° S in the Atlantic region, *J. Atmos. Chem.*, 15, 253–267, doi:10.1007/BF00115397, 1992.

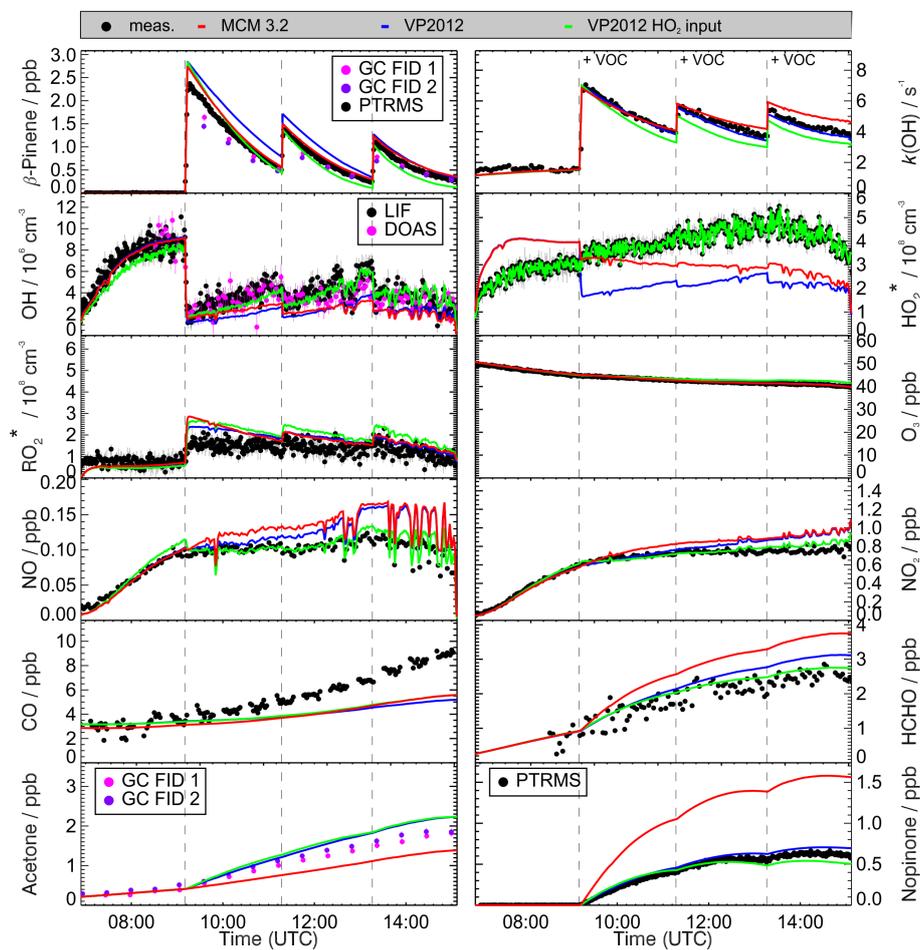
- Rohrer, F., Bohn, B., Brauers, T., Brüning, D., Johnen, F.-J., Wahner, A., and Kleffmann, J.: Characterisation of the photolytic HONO-source in the atmosphere simulation chamber SAPHIR, *Atmos. Chem. Phys.*, 5, 2189–2201, 2005.
- Saathoff, H., Naumann, K.-H., Möhler, O., Jonsson, M., Hallquist, M., Kiendler-Scharr, A., Mentel, T. F., Tillmann, R., and Schurath, U.: Temperature dependence of yields of secondary organic aerosols from the ozonolysis of  $\alpha$ -pinene and limonene, *Atmos. Chem. Phys.*, 9, 1551–1577, doi:10.5194/acp-9-1551-2009, 2009.
- 5 Saunders, S. M., Jenkin, M. E., Derwent, R. G., and Pilling, M. J.: Protocol for the development of the Master Chemical Mechanism, MCM v3 (Part A): tropospheric degradation of non-aromatic volatile organic compounds, *Atmos. Chem. Phys.*, 3, 161–180, doi:10.5194/acp-3-161-2003, 2003.
- Schlosser, E., Bohn, B., Brauers, T., Dorn, H.-P., Fuchs, H., Häsel, R., Hofzumahaus, A., Holland, F., Rohrer, F., Rupp, L. O., Siese, M., Tillmann, R., and Wahner, A.: Intercomparison of Two Hydroxyl Radical Measurement Techniques at the Atmosphere Simulation Chamber SAPHIR, *J. Atmos. Chem.*, 56, 187–205, doi:10.1007/s10874-006-9049-3, 2007.
- 10 Schlosser, E., Brauers, T., Dorn, H.-P., Fuchs, H., Häsel, R., Hofzumahaus, A., Holland, F., Wahner, A., Kanaya, Y., Kajii, Y., Miyamoto, K., Nishida, S., Watanabe, K., Yoshino, A., Kubistin, D., Martinez, M., Rudolf, M., Harder, H., Berresheim, H., Elste, T., Plass-Dülmer, C., Stange, G., and Schurath, U.: Technical Note: Formal blind intercomparison of OH measurements: results from the international campaign HOxComp, *Atmos. Chem. Phys.*, 9, 7923–7948, doi:10.5194/acp-9-7923-2009, 2009.
- 15 Seinfeld, J. H. and Pandis, S. N.: *Atmospheric Chemistry and Physics From Air Pollution to Climate Change*, John Wiley & Sons, Inc., 2 edn., 2006.
- Sindelarova, K., Granier, C., Bouarar, I., Guenther, A., Tilmes, S., Stavrou, T., Müller, J.-F., Kuhn, U., Stefani, P., and Knorr, W.: Global data set of biogenic VOC emissions calculated by the MEGAN model over the last 30 years, *Atmos. Chem. Phys.*, 14, 9317–9341, doi:10.5194/acp-14-9317-2014, 2014.
- 20 Tan, D., Faloon, I., Simpas, J. B., Brune, W., Shepson, P. B., Couch, T. L., Sumner, A. L., Carroll, M. A., Thornberry, T., Apel, E., Riemer, D., and Stockwell, W.: HO<sub>x</sub> budgets in a deciduous forest: Results from the PROPHET summer 1998 campaign, *J. Geophys. Res. - Atmos.*, 106, 24407–24427, doi:10.1029/2001jd900016, 2001.
- Tan, Z., Fuchs, H., Lu, K., Hofzumahaus, A., Bohn, B., Broch, S., Dong, H., Gomm, S., Hler, R., He, L., Holland, F., Li, X., Liu, Y., Lu, S., Rohrer, F., Shao, M., Wang, B., Wang, M., Wu, Y., Zeng, L., Zhang, Y., and Wahner, A.: Radical chemistry at a rural site (Wangdu) in the North China Plain: observation and model calculations of OH, HO<sub>2</sub> and RO<sub>2</sub> radicals, *Atmos. Chem. Phys.*, 17, 663–690, doi:10.5194/acp-17-663-2017, 2017.
- 25 Taraborrelli, D., Lawrence, M. G., Crowley, J. N., Dillon, T. J., Gromov, S., Gross, C. B. M., Vereecken, L., and Lelieveld, J.: Hydroxyl radical buffered by isoprene oxidation over tropical forests, *Nature Geosci.*, 5, 190–193, doi:10.1038/NGEO1405, 2012.
- 30 Vereecken, L. and Peeters, J.: A theoretical study of the OH-initiated gas-phase oxidation mechanism of  $\beta$ -pinene (C<sub>10</sub>H<sub>16</sub>) first generation products, *Phys. Chem. Chem. Phys.*, 14, 3802–3815, doi:10.1039/C2CP23711C, 2012.
- Wegener, R., Brauers, T., Koppmann, R., Rodríguez Bares, S., Rohrer, F., Tillmann, R., Wahner, A., Hansel, A., and Wisthaler, A.: Simulation chamber investigation of the reactions of ozone with short-chained alkenes, *J. Geophys. Res. - Atmos.*, 112, doi:10.1029/2006jd007531, 2007.
- 35 Whalley, L. K., Edwards, P. M., Furneaux, K. L., Goddard, A., Ingham, T., Evans, M. J., Stone, D., Hopkins, J. R., Jones, C. E., Karunaharan, A., Lee, J. D., Lewis, A. C., Monks, P. S., Moller, S. J., and Heard, D. E.: Quantifying the magnitude of a missing hydroxyl radical source in a tropical rainforest, *Atmos. Chem. Phys.*, 11, 7223–7233, doi:10.5194/acp-11-7223-2011, 2011.

- Wiedinmyer, C., Guenther, A., Harley, P., Hewitt, N., Geron, C., Artaxo, P., Steinbrecher, R., and Rasmussen, R.: Global Organic Emissions from Vegetation, vol. 18, pp. 115–170, Springer Netherlands, doi:10.1007/978-1-4020-2167-1\_4, 2004.
- Wisthaler, A., Jensen, N. R., Winterhalter, R., Lindinger, W., and Hjorth, J.: Measurements of acetone and other gas phase product yields from the OH-initiated oxidation of terpenes by proton-transfer-reaction mass spectrometry (PTR-MS), *Atmos. Environ.*, 35, 6181–6191, doi:10.1016/s1352-2310(01)00385-5, 2001.
- 5 Wolfe, G. M., Crouse, J. D., Parrish, J. D., St. Clair, J. M., Beaver, M. R., Paulot, F., Yoon, T. P., Wennberg, P. O., and Keutsch, F. N.: Photolysis, OH reactivity and ozone reactivity of a proxy for isoprene-derived hydroperoxyenals (HPALDs), *Phys. Chem. Chem. Phys.*, 14, 7276–7286, doi:10.1039/C2CP40388A, 2012.
- Wolfe, G. M., Cantrell, C., Kim, S., Mauldin III, R. L., Karl, T., Harley, P., Turnipseed, A., Zheng, W., Flocke, F., Apel, E. C., Hornbrook, R. S., Hall, S. R., Ullmann, K., Henry, S. B., DiGangi, J. P., Boyle, E. S., Kaser, L., Schnitzhofer, R., Hansel, A., Graus, M., Nakashima, Y., Kajii, Y., Guenther, A., and Keutsch, F. N.: Missing peroxy radical sources within a summertime ponderosa pine forest, *Atmos. Chem. Phys.*, 14, 4715–4732, doi:10.5194/acp-14-4715-2014, 2014.
- 10 Zhao, D. F., Kaminski, M., Schlag, P., Fuchs, H., Acir, I.-H., Bohn, B., Häseler, R., Kiendler-Scharr, A., Rohrer, F., Tillmann, R., Wang, M. J., Wegener, R., Wildt, J., Wahner, A., and Mentel, T. F.: Secondary organic aerosol formation from hydroxyl radical oxidation and ozonolysis of monoterpenes, *Atmos. Chem. Phys.*, 15, 991–1012, doi:10.5194/acp-15-991-2015, 2015.
- 15

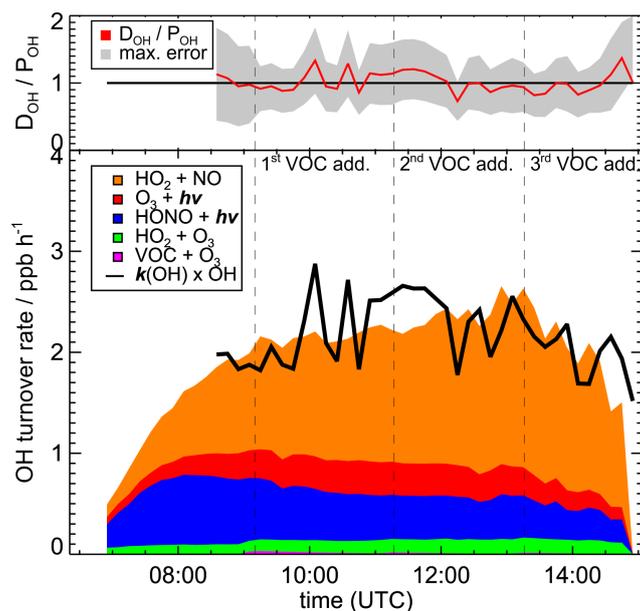


**Figure 1.** Acetone and nopinone formation from OH initiated  $\beta$ -pinene oxidation after Vereecken and Peeters (2012). For simplification only the major reactions are shown

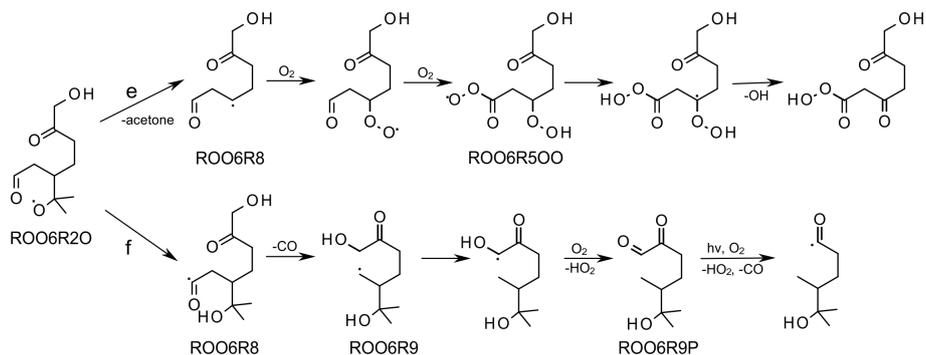
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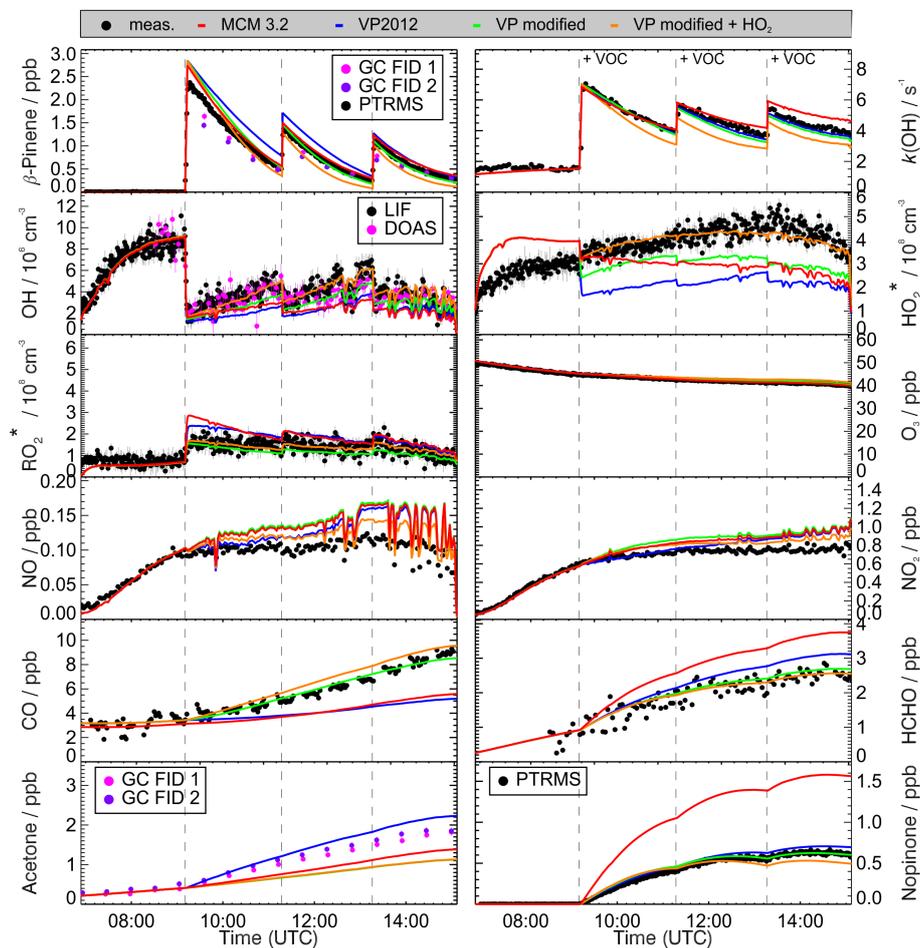
**Figure 2.** Comparison of the measured and modeled time series of  $\beta$ -pinene,  $k(\text{OH})$ ,  $\text{OH}$ ,  $\text{HO}_2^*$ ,  $\text{RO}_2^*$ ,  $\text{NO}$ ,  $\text{NO}_2$ ,  $\text{CO}$ ,  $\text{HCHO}$ , acetone and nopinone in the  $\beta$ -pinene oxidation experiment from 27<sup>th</sup> August. Red: MCM 3.2 Blue: modified MCM model by Vereecken and Peeters (2012) with changed product yields Green: modified MCM model by Vereecken and Peeters (2012) constrained by the measured  $\text{HO}_2$  concentration



**Figure 3.** OH budget for the experiment on 27th August 2012. The OH destruction rate  $D_{\text{OH}}$  calculated from the measured OH reactivity  $k(\text{OH})$  and the measured OH concentration (DOAS) is given as black line. The coloured areas display the OH production rate  $P_{\text{OH}}$  calculated from measurements. The upper panel of the diagram shows the ratio of  $D_{\text{OH}}/P_{\text{OH}}$  as red line. The maximum systematic error of the ratio is indicated by the grey area. For reasons of clarity all data in the upper as well as the lower panel of the diagram are shown as 5 min average values. During the course of the experiment the OH destruction rate is balanced by the sum of the measurable OH production terms. The reaction of  $\text{HO}_2$  with  $\text{NO}$  and the photolysis of  $\text{HONO}$  are the dominant OH production terms.  $\text{HO}_2$  measurements were not corrected for the interference from specific  $\text{RO}_2$  species.



**Figure 4.** Possible  $\text{HO}_2$  formation pathway in the oxidation of  $\beta$ -pinene modified after Vereecken and Peeters (2012)



**Figure 5.** Comparison of the measured and modeled time series of  $\beta$ -pinene, OH, OH,  $\text{HO}_2^*$ ,  $\text{RO}_2^*$ , NO,  $\text{NO}_2$ , CO, HCHO, acetone and nopinone in the  $\beta$ -pinene oxidation experiment from 27<sup>th</sup> August. Red: MCM 3.2, Blue: model by Vereecken and Peeters (2012), Green: model by Vereecken and Peeters (2012) with 1,5-H-migration of ROO6R2O, Orange: Modified model by Vereecken and Peeters (2012) with 1,5-H-migration of ROO6R2O and an additional production term for one molecule of  $\text{HO}_2$  and CO.

**Table 1.** Instrumentation for radical and trace gas detection during the  $\beta$ -pinene oxidation experiments.

	Technique	Time Resolution	$1\sigma$ Precision	$1\sigma$ Accuracy
OH	DOAS <sup>a</sup> (Dorn et al., 1995; Hausmann et al., 1997; Schlosser et al., 2007)	205 s	$0.8 \times 10^6 \text{ cm}^{-3}$	6.5 %
OH	LIF <sup>b</sup> (Lu et al., 2012)	47 s	$0.3 \times 10^6 \text{ cm}^{-3}$	13 %
HO <sub>2</sub> , RO <sub>2</sub>	LIF <sup>b</sup> (Fuchs et al., 2011, 2008)	47 s	$1.5 \times 10^7 \text{ cm}^{-3}$	16 %
$k(\text{OH})$	Laser-photolysis + LIF <sup>b</sup> (Lou et al., 2010)	180 s	$0.3 \text{ s}^{-1}$	$0.5 \text{ s}^{-1}$
NO	Chemiluminescence (Rohrer and Brüning, 1992)	180 s	4 pptv	5 %
NO <sub>2</sub>	Chemiluminescence (Rohrer and Brüning, 1992)	180 s	2 pptv	5 %
O <sub>3</sub>	Chemiluminescence (Ridley et al., 1992)	180 s	60 pptv	5 %
VOCs	PTR-TOF-MS <sup>c</sup> (Lindinger et al., 1998; Jordan et al., 2009)	30 s	15 pptv	14 %
	GC <sup>d</sup> (Kaminski, 2014)	30 min	4-8 %	5 %
CO	RGA <sup>e</sup> (Wegener et al., 2007)	3 min	4 %	10 %
HONO	LOPAP <sup>f</sup> (Häseler et al., 2009)	300 s	1.3 pptv	10 %
HCHO	BB-DOAS <sup>g</sup> (Brauers et al., 2007)	100 s	20 %	6 %
Photolysis frequencies	Spectroradiometer (Bohn and Zilken, 2005)	60 s	10 %	10 %

<sup>a</sup>Differential Optical Absorption Spectroscopy.<sup>b</sup>Laser Induced Fluorescence.<sup>c</sup>Proton-Transfer-Reaction Time-Of-Flight Mass-Spectrometry.<sup>d</sup>Gas Chromatography.<sup>e</sup>Reactive Gas Analyzer.<sup>f</sup>Long Path Absorption Photometer.<sup>g</sup>Broadband Differential Optical Absorption Spectroscopy.

**Table 2.** Experimental conditions of the  $\beta$ -pinene oxidation experiments. Maximum values are given for  $\beta$ -pinene and averaged values for the part of the experiment, when  $\beta$ -pinene was present, for the other parameters.

$\beta$ -pinene ppbv	OH $10^6 \text{ cm}^{-3}$	NO <sub>x</sub> ppbv	NO pptv	O <sub>3</sub> ppbv	RH %	$j(\text{NO}_2)$ $10^{-3} \text{ s}^{-1}$	$T$ K	date
4.3	6.0	1.0	300	10	45	5	295	12 Aug 2012
4.3	4.5	0.9	200	10	45	4	299	15 Aug 2012
4.7	3.5	0.9	100	40	40	4.5	293	27 Aug 2012

**Table 3.** Product yields from the reaction of  $\beta$ -pinene with OH radicals under various NO and VOC concentrations

Product	Yield OH reaction	Reference	consumed VOC ppbv	NO ppbv
Nopinone	0.35 $\pm$ 0.13	This work	3	0.4
	0.28-0.37 $\pm$ 0.13		3	0.1
	0.79 <sup>a</sup> $\pm$ 0.08	Hatakeyama et al. (1991)	700	1800
	0.30 $\pm$ 0.045	Arey et al. (1990)	960	960
	0.27 $\pm$ 0.04	Hakola et al. (1994)	1000	9600
	0.25 $\pm$ 0.05	Larsen et al. (2001)	1300-1600	0
	0.25 $\pm$ 0.03	Wisthaler et al. (2001)	1000-3000	1000-2000
	0.24	Librando and Tringali (2005)	4100-13200	0
Acetone	0.19 $\pm$ 0.06	This work	3	0.4
	0.20-0.36 $\pm$ 0.07		3	0.1
	0.13 $\pm$ 0.02	Wisthaler et al. (2001)	1000-3000	1000-2000
	0.11 $\pm$ 0.03	Larsen et al. (2001)	1300-1600	0
	0.03-0.06	Fantechi (1999)		
	0.02 $\pm$ 0.002	Orlando et al. (2000)	1800-12000	800-8000
	0.085 $\pm$ 0.018	Reissell et al. (1999)	880-920	9600
	0.14	Librando and Tringali (2005)	4100-13200	0

<sup>a</sup>Yield measured by FTIR absorption at  $1740 \text{ cm}^{-1}$

**Table 4.** Comparison of measured and modeled product yields from the reaction of  $\beta$ -pinene with OH radicals for the three  $\beta$ -pinene injections during the experiment on 27 th August 2012

Product	Injection	Yield measured	Yield MCM 3.2	Yield Vereecken and Peeters
Nopinone	1 st	0.28	0.53	0.27
	2 nd	0.37	0.61	0.28
	3 rd	0.35	0.65	0.30
Acetone	1 st	0.20	0.07	0.37
	2 nd	0.24	0.16	0.47
	3 rd	0.36	0.21	0.49