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This discussion paper is/has been under review for the journal Atmospheric Chemistry and Physics (ACP). Please refer to the corresponding final paper in ACP if available.

Typical synoptic situations and their impacts on the wintertime air pollution in the Guanzhong basin, China

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Received: 12 September 2015 – Accepted: 9 December 2015 – Published: 15 January 2016

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Published by Copernicus Publications on behalf of the European Geosciences Union.

Abstract

Rapid industrialization and urbanization have caused severe air pollution in the Guanzhong basin, northwestern China with heavy haze events occurring frequently in recent winters. Using the NCEP reanalysis data, the large scale synoptic situations influencing the Guanzhong basin during wintertime of 2013 are categorized into six types to evaluate the contribution of synoptic situations to the air pollution, including “north-low”, “southwest-trough”, “southeast-high”, “transition”, “southeast-trough”, and “inland-high”. The FLEXPART model has been utilized to demonstrate the corresponding pollutant transport patterns for the typical synoptic situations in the basin. Except “southwest-trough” and “southeast-high” (defined as favorable synoptic situations), the rest four synoptic conditions (defined as unfavorable synoptic situations) generally facilitate the accumulation of air pollutants, causing heavy air pollution in the basin. In association with the measurement of PM_{2.5} (particulate matter with aerodynamic diameter less than 2.5 μm) in the basin, the unfavorable synoptic situations correspond to high PM_{2.5} mass concentrations or poor air quality and vice versa. The same analysis has also been applied to winters of 2008–2012, which shows that the basin was mainly influenced by the unfavorable synoptic situations during wintertime leading to poor air quality. The WRF-CHEM model has further been applied to simulate the selected six days representing the typical synoptic situations during the wintertime of 2013, and the results generally show a good consistence between the modeled distributions and variations of PM_{2.5} and the corresponding synoptic situations, demonstrating reasonable classification for the synoptic situations in the basin. Detailed meteorological conditions, such as temperature inversion, low-level horizontal wind speed, vertical wind velocity, and convergence all contribute to heavy air pollution events in the basin under unfavorable synoptic conditions. Considering the proportion of occurrence of unfavorable synoptic situations during wintertime, reduction of emissions is the optimum approach to mitigate the air pollution in the Guanzhong basin.

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1 Introduction

Elevated atmospheric pollutants, such as particulate matter (PM) and ozone (O₃), exert deleterious impacts on human health and environment (e.g., Penner et al., 2001; Pope and Dockery, 2006; J. Zhang et al., 2010). Over the past three decades, with tremendous economic growth in China, rapid industrialization and urbanization have caused severe air pollution, as reflected in the heavy haze event that often occurs in the north of China, particularly during wintertime (e.g., Chan and Yao, 2008; Fang et al., 2009; Gao et al., 2011; Liu et al., 2013; Zhao et al., 2013; Huang et al., 2014; Fu et al., 2014; Guo et al., 2014; Han et al., 2014; Zhang et al., 2015; Yang et al., 2015). Guanzhong basin is located in the northwest of China, nestled between the Qinling Mountains in the south and the Loess Plateau in the north. The unique topography facilitates the accumulation of air pollutants, and with the rapid increasing industries and city expansions, heavy air pollution frequently attacks the basin (e.g., Cao et al., 2009; Sheng et al., 2011).

Numerous studies have demonstrated that the meteorological conditions play an important role in the formation, transformation, diffusion, transport, and removal of the atmospheric pollutants (e.g., Seaman, 2000; Solomon et al., 2000; de Foy et al., 2005, 2006; Bei et al., 2008, 2010, 2012, 2013). If the emissions of pollutants remain invariable, transformations in the chemical state of the atmosphere are principally determined by the meteorological conditions. Recent advances in understanding the role of the meteorological conditions in the air pollution formation in China have mainly concentrated on the regions of Beijing-Tianjin-Hebei, the Pearl River Delta, and the Yangtze River Delta (e.g., Wu et al., 2008, 2013; Wang et al., 2009; Q. H. Zhang et al., 2010; Gao et al., 2011; Zhang et al., 2012; L. Wang et al., 2014; H. Wang et al., 2014; Zhang et al., 2015). Wang et al. (2009) have shown that the O₃ decrease at a Beijing rural site during the 2008 Olympics is attributed to the favorable meteorological condition in comparison with the same period in 2006 and 2007. Q. H. Zhang et al. (2010) have proposed that, during the 2008 Olympics, the atmospheric visibility improvements

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2 Data, models, and methodology

The National Centers for Environmental Prediction (NCEP) final operational global gridded analysis (FNL) ($1^\circ \times 1^\circ$) is used to categorize the large-scale synoptic weather systems influencing the Gunazhong basin during the period from 2008 to 2013 through the subjective procedure. The geopotential height and wind fields on 850 hPa are applied to identify the synoptic situations that affect the plume transport patterns in the basin.

Continuous daily $PM_{2.5}$ measurements have been performed at the Institute of Earth Environment, Chinese Academy of Sciences (IEECAS) in Xi'an, China since 2003. Additionally, since January 2013, the China's Ministry of Environmental Protection (China MEP) has commenced to release the real-time hourly concentrations of $PM_{2.5}$. Total 33 monitoring sites are distributed in the Guanzhong basin (Fig. 1b). The daily $PM_{2.5}$ measurement at IEECAS site from 2008 to 2012 and the hourly $PM_{2.5}$ measurement released by China MEP from 2013 to 2014 are used to validate the categorized synoptic situations influencing the basin.

In order to analyze the corresponding pollutant transport patterns under the typical categorized synoptic situations, The FLEXPART model is employed to calculate the forward Lagrangian particle dispersion (Stohl et al., 1998; Fast and Easter, 2006), which is driven by the output from the WRF model (Skamarock et al., 2008). The FLEXPART model is set-up with releases of 6000 computational particles within a grid cell of $10\text{ km} \times 10\text{ km} \times 0.02\text{ km}$ centered at Xi'An urban area in the morning. Tracer particles are released continuously from 04:00 to 10:00 BJT (Beijing Time) of the day, and traced until 04:00 BJT of next day. For the convenience, all the time used hereafter is BJT. The WRF model adopts one grid with horizontal resolution of 3 km and 35 sigma levels in the vertical direction. The grid cells used for the domain are 201×201 (Fig. 1a). The selected six days, representing six categorized typical synoptic situations of the Guanzhong basin during wintertime of 2013, are simulated. They are initialized at 20:00 BJT on each day and integrated for 36 h. The NCEP FNL analysis data ($1^\circ \times 1^\circ$) is used to produce the initial and boundary conditions for the WRF model. The physi-

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cal process parameterization schemes used in simulations included the Grell-Devenyi ensemble scheme for cumulus scheme (Grell and Devenyi, 2002), the WRF Single Moment (WSM) three-class microphysics (Hong et al., 2004), and Mellor–Yamada–Janjic (MYJ) TKE scheme (Janjic, 2002) for the PBL processes.

The WRF-CHEM model has been used to further simulate the selected six days representing the typical categorized synoptic situations and to verify the particular meteorological conditions during the severe air pollution events in the Guanzhong basin. A specific version of the WRF-CHEM model (Grell et al., 2005) is used in the present study, which was developed by Li et al. (2010, 2011a, b, 2012) at the Molina Center for Energy and the Environment, with a new flexible gas phase chemical module and the CMAQ (version 4.6) aerosol module developed by US EPA (Binkowski and Roselle, 2003). The inorganic aerosols are simulated in the WRF-CHEM model using ISORROPIA (“equilibrium” in Greek, here referred to as an improved thermodynamic equilibrium aerosol model) Version 1.7 (<http://nenes.eas.gatech.edu/ISORROPIA/>). The secondary organic aerosols (SOA) formation is simulated using a non-traditional SOA model including the volatility basis-set modeling method in which primary organic components are assumed to be semi-volatile and photochemically reactive and are distributed in logarithmically spaced volatility bins (Li et al., 2011a). Detailed description of the WRF-CHEM model can be found in Li et al. (2010, 2011a, b, 2012). The meteorological setup in the WRF-CHEM model simulations is same as those in the WRF model, except that the spin-up time of the WRF-CHEM model is one day. The chemical initial and boundary conditions for the WRF-CHEM model simulations are interpolated from the 6 h output of a global chemical transport model for O₃ and related chemical tracers (MOZART). The anthropogenic emission inventory (EI) developed by Zhang et al. (2009) is used in the study, including contributions from agriculture, industry, power, residential and transportation sources. The MEGAN model developed by Guenther et al. (2006) is used to calculate on-line biogenic emissions.

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days, indicating the significant contribution from the large-scale meteorological conditions to the poor air quality in the Guanzhong basin. In addition, the percentage of “inland-high” is around 42%, which is the most popular synoptic situation associated with the poor air quality in the basin.

The WRF-CHEM model has been further used to simulate the selected six days representing the typical synoptic situations during the wintertime of 2013, and the results shows that the modeled $PM_{2.5}$ distribution and variations are generally consistent well with the corresponding synoptic conditions, which demonstrates the critical role of the synoptic meteorological conditions in air pollution events in the basin. The WRF-CHEM model simulations also indicate the reasonable classification for the synoptic situations in the basin. In addition, detailed meteorological conditions, including temperature inversion, low-level horizontal wind speed, vertical wind velocity, and convergence are also analyzed for the selected days. Under unfavorable synoptic situations, temperature inversion, weak low-level wind and convergence do not facilitate the dispersion of pollutants in the basin. While in the favorable synoptic situations, low-level divergence, caused by strong horizontal winds or active vertical motions, efficiently evacuate air pollutants in the basin.

During wintertime, 5 year filter $PM_{2.5}$ measurement from 2008 to 2012 and the $PM_{2.5}$ measurement released by China MEP in 2013 and 2014 all show that the Guanzhong basin has experienced heavy air pollution. Even under favorable synoptic situations, the observed $PM_{2.5}$ mass concentrations have barely reached the excellent level due to massive local emissions of air pollutants and the background dust transportation from Loess Plateau. Hence, considering the proportion of occurrence of unfavorable synoptic situations during wintertime, reduction of emissions is a feasible method to reduce the air pollution in the Guanzhong basin.

Given that the synoptic situations categorized are made at 850 hPa that influence the Guanzhong basin, potential uncertainties still exist in the classification results. More quantitative studies are needed in the future to improve the synoptic situation classification. Further, the analysis using local meteorological observations on ground surfaces

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and inside the PBL is also imperative to investigate the role of the local meteorological conditions in the severe pollution events.

Acknowledgements. Naifang Bei is supported by the National Natural Science Foundation of China (no. 41275101) and the Fundamental Research Funds for the Central Universities of China. Guohui Li is supported by “Hundred Talents Program” of the Chinese Academy of Sciences and the National Natural Science Foundation of China (no. 41275153).

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Table 1. Synoptic categories influencing the Guanzhong basin during the wintertime of 2013.

Categories	Date*	Sum	Percentage (%)
North-low	20131217 20131208 20140110 20140111 20140102	16	17.8
	20140106 20140204 20140205 20140206 20140215		
	20140216 20140217 20140224 20140225 20140228		
	20140208		
Southwest-trough	20131201 20131205 20131209 20131211 20131212	13	14.4
	20131215 20131231 20140103 20140104 20140119		
	20140120 20140108 20140112		
Southeast-high	20131210 20131226 20140107 20140227	4	4.4
Transition	20131202 20131207 20131214 20131230 20140101	11	12.2
	20140116 20140127 20140130 20140201 20140202		
	20140207		
Southeast-trough	20131204 20140123 20140129 20140131 20140226	5	5.6
Inland-high	20131203 20131206 20131213 20131219 20131221	41	45.6
	20131222 20131223 20131224 20131216 20131225		
	20131218 20131220 20131227 20131229 20131228		
	20140105 20140113 20140118 20140122 20140125		
	20140115 20140121 20140126 20140128 20140109		
	20140114 20140117 20140124 20140214 20140219		
	20140211 20140210 20140213 20140220 20140221		
	20140222 20140223 20140218 20140212 20140209		
	20140203		

* The format of date is YYYYMMDD, in which YYYY, MM, and DD represent year, month, day, respectively.

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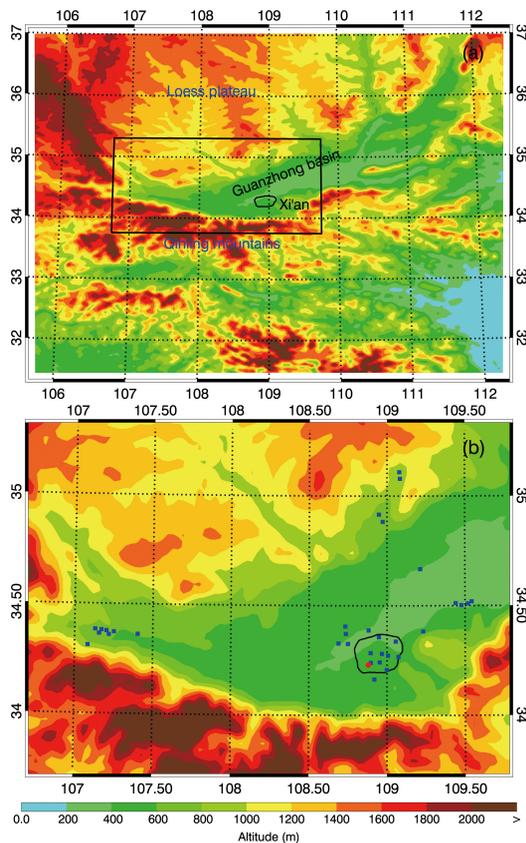


Figure 1. (a) WRF and WRF-CHEM model simulation domain with topography and (b) geographic distributions of ambient monitoring stations. In (b), the blue filled squares are the ambient monitoring sites and the red filled circle is the IEECAS site.

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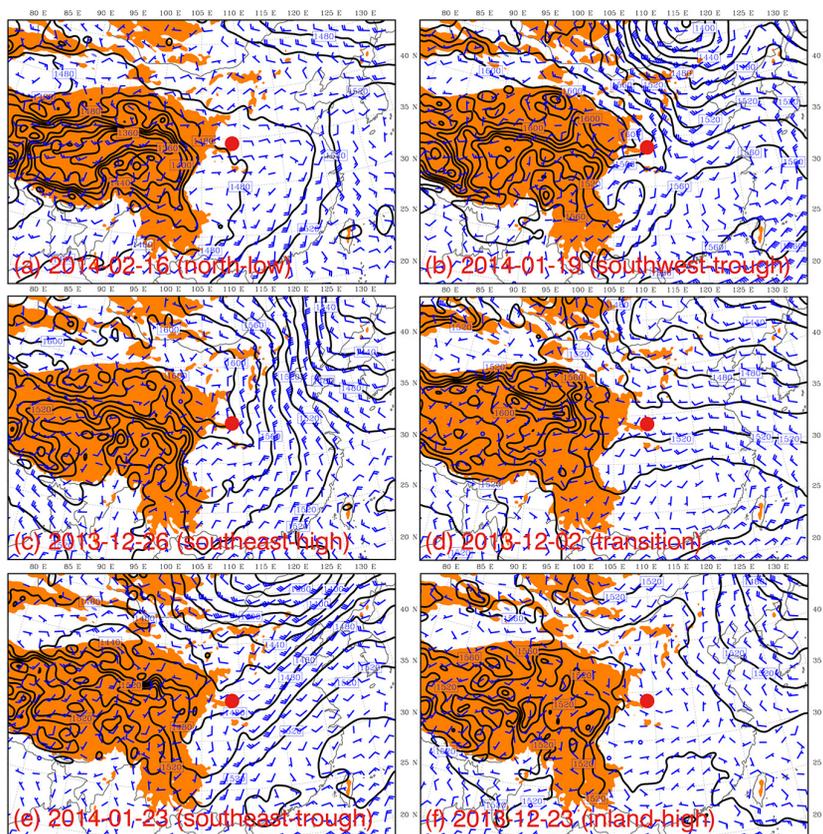


Figure 2. Distributions of winds and geopotential heights on 850 hPa at 08:00 BJT on **(a)** 16 February 2014 (“north-low”), **(b)** 19 January 2014 (“southwest-trough”), **(c)** 26 December 2013 (“southeast-high”), **(d)** 2 December 2013 (“transition”), **(e)** 23 January 2014 (“southeast-trough”), and **(f)** 23 December 2013 (“inland-high”). The red filled circle is Xi’an.

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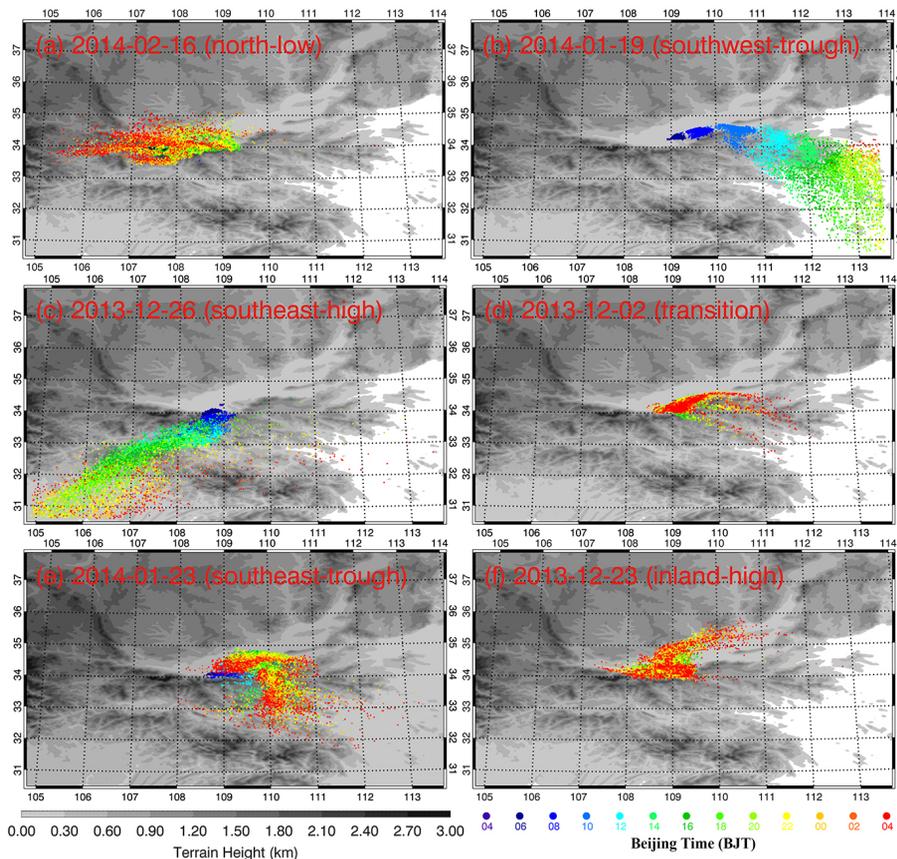


Figure 3. 24 h plume transport patterns initialized from 04:00 BJT on (a) 16 February 2014 (“north-low”), (b) 19 January 2014 (“southwest-trough”), (c) 26 December 2013 (“southeast-high”), (d) 2 December 2013 (“transition”), (e) 23 January 2014 (“southeast-trough”), and (f) 23 December 2013 (“inland-high”).

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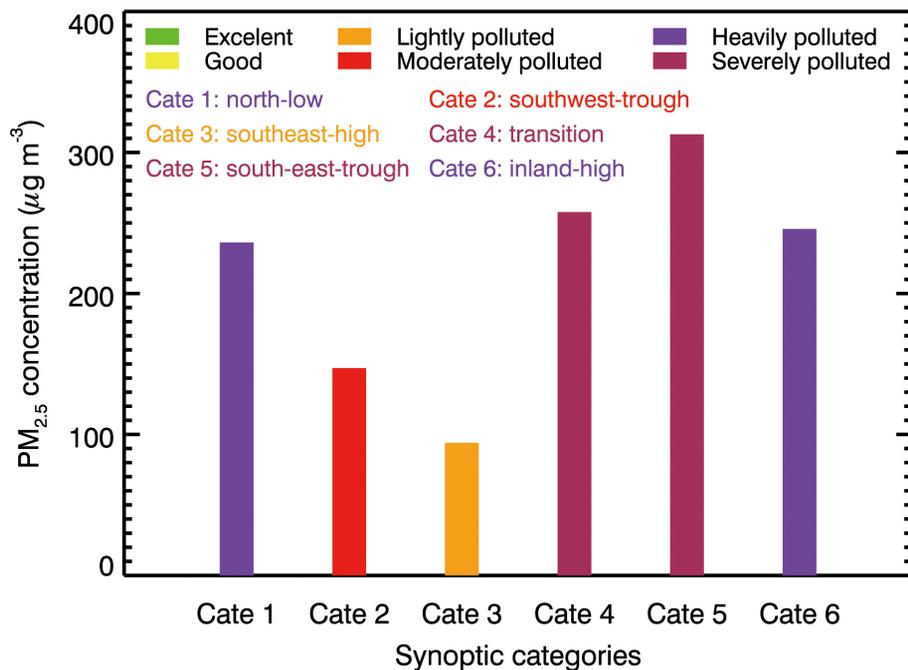


Figure 5. Daily mean PM_{2.5} mass concentration averaged during the six typical synoptic situations from 2008 to 2012 at the IEECAS site.

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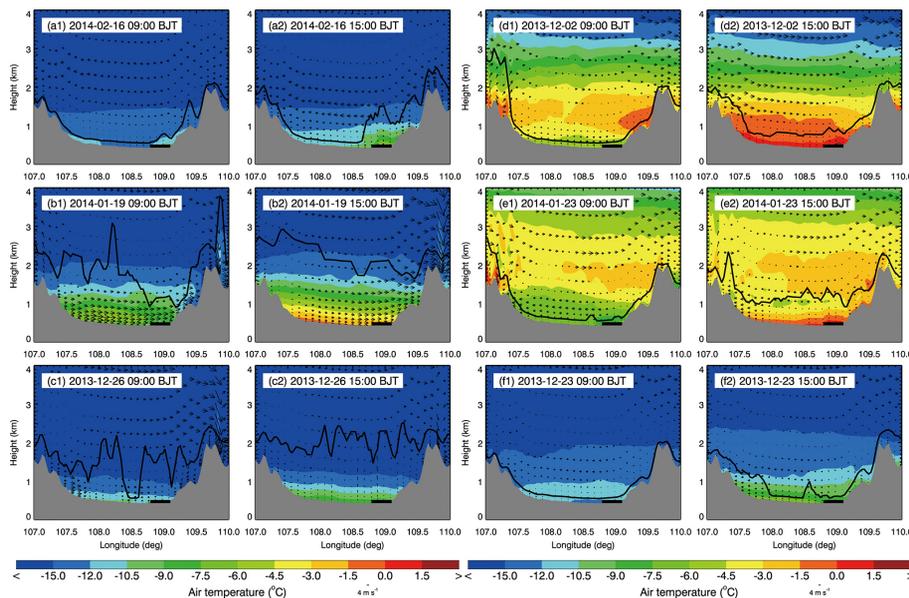


Figure 6. Vertical distributions of temperature, wind vectors, and PBL height through Xi'an along the east–west direction at 09:00 and 15:00 BJT on **(a)** 16 February 2014 (“north-low”), **(b)** 19 January 2014 (“southwest-trough”), **(c)** 26 December 2013 (“southeast-high”), **(d)** 2 December 2013 (“transition”), **(e)** 23 January 2014 (“southeast-trough”), and **(f)** 23 December 2013 (“inland-high”). The black filled rectangle represents the urban area of Xi'an, China.

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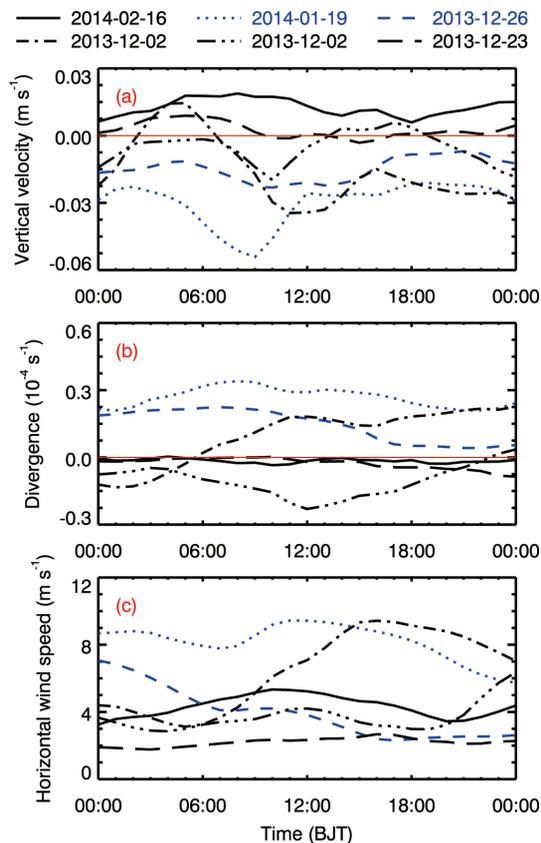


Figure 7. Temporal variations of the area averaged low-level **(a)** vertical velocity, **(b)** divergence, and **(c)** horizontal wind speed over the basin.

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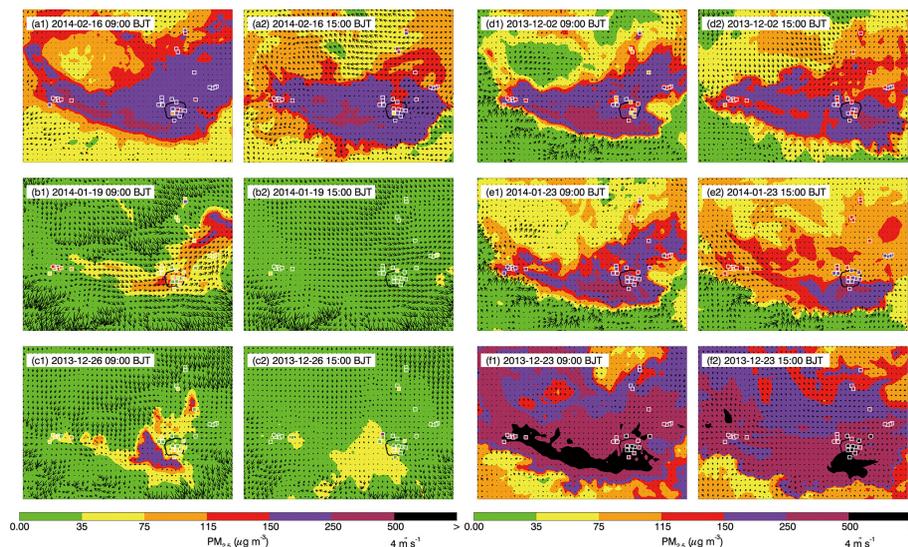


Figure 8. Pattern comparison of simulated vs. observed near-surface $PM_{2.5}$ mass concentrations at 09:00 and 15:00 BJT on **(a)** 16 February 2014 (“north-low”), **(b)** 19 January 2014 (“southwest-trough”), **(c)** 26 December 2013 (“southeast-high”), **(d)** 2 December 2013 (“transition”), **(e)** 23 January 2014 (“southeast-trough”), and **(f)** 23 December 2013 (“inland-high”). Colored squares: $PM_{2.5}$ observations; color contour: $PM_{2.5}$ simulations; black arrows: simulated surface winds.

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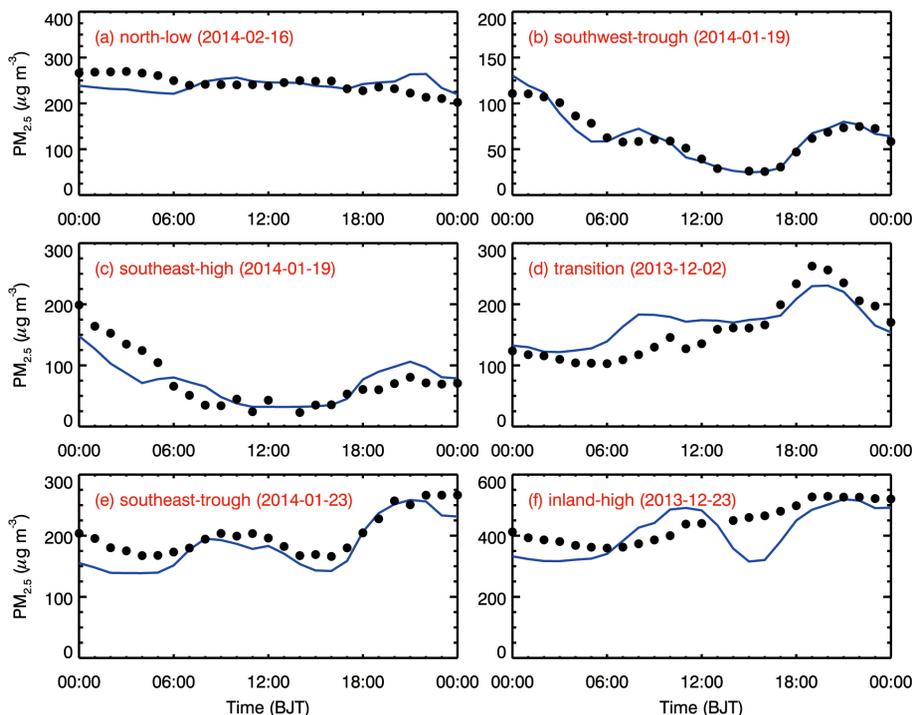


Figure 9. Comparison of observed and predicted diurnal profiles of the $PM_{2.5}$ mass concentrations averaged over the monitoring sites in the Guanzhong basin on (a) 16 February 2014 (“north-low”), (b) 19 January 2014 (“southwest-trough”), (c) 26 December 2013 (“southeast-high”), (d) 2 December 2013 (“transition”), (e) 23 January 2014 (“southeast-trough”), and (f) 23 December 2013 (“inland-high”).

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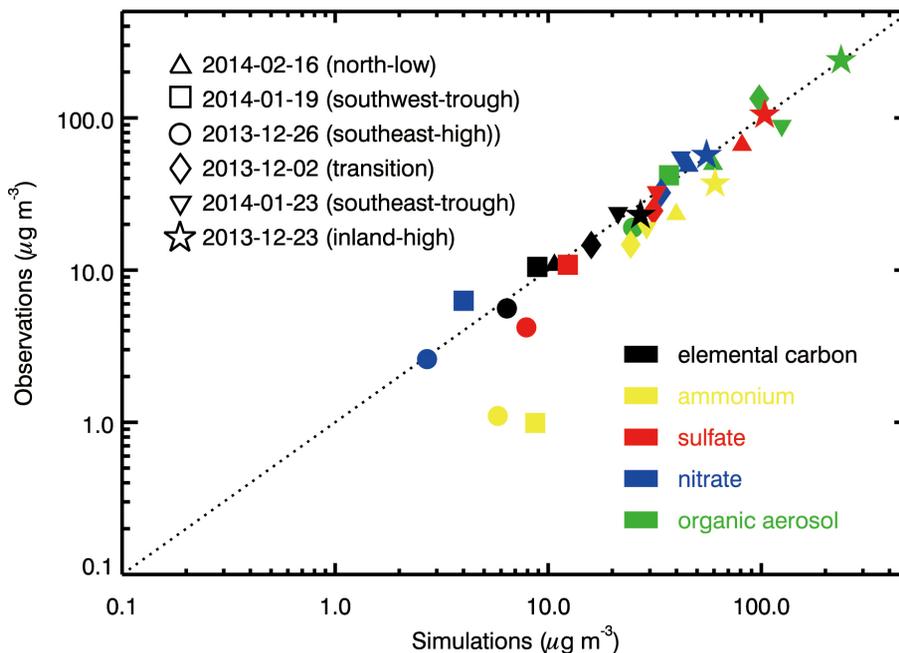


Figure 10. Scatter plot of the measured vs. modeled daily mean mass concentration of aerosol constituents at IEECAS site on the selected six days.

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