

1 **Quantification of PM<sub>2.5</sub> Organic Carbon Sampling Artifacts in U.S. Networks**

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38 **Abstract**

39 Field blanks (bQF) and backup filters (quartz-fiber behind quartz-fiber filter; QBQ) have  
40 been adopted by U.S. long-term air quality monitoring networks to estimate PM<sub>2.5</sub> organic  
41 carbon (OC) sampling artifacts. This study documents bQF and QBQ carbon levels for 1)  
42 Interagency Monitoring of Protected Visual Environments (IMPROVE); 2) Speciation Trends  
43 Network (STN; part of the Chemical Speciation Network [CSN]); and 3) Southeastern Aerosol  
44 Research and Characterization (SEARCH) networks and examines the similarities/differences  
45 associated with network-specific sampling protocols. A higher IMPROVE sample volume and  
46 smaller filter deposit area results in PM<sub>2.5</sub> areal density ( $\mu\text{g}/\text{cm}^2$  on filter) 3–11 times those of  
47 STN/CSN and SEARCH samples for the same ambient PM<sub>2.5</sub> concentrations, thus reducing the  
48 relative contribution of sampling artifacts from passive OC adsorption. A relatively short (1–15  
49 minutes) passive exposure period of STN/CSN and SEARCH bQF OC ( $0.8\text{--}1 \mu\text{g}/\text{cm}^2$ )  
50 underestimates positive and negative OC artifacts resulting from passive adsorption or  
51 evaporation of semi-volatile organic compounds on quartz-fiber filters. This is supported by low  
52 STN/CSN and SEARCH bQF levels and lack of temporal or spatial variability among the sites  
53 within the networks. With a much longer period,  $\sim 7$  days of ambient passive exposure, average  
54 IMPROVE bQF and QBQ OC are comparable ( $2.4 \pm 0.5$  and  $3.1 \pm 0.8 \mu\text{g}/\text{cm}^2$ , respectively) and  
55 more than twice levels found in the STN/CSN and SEARCH networks. Sampling artifacts in  
56 STN/CSN were estimated from collocated IMPROVE samples based on linear regression. At six  
57 of the eight collocated sites in this study, STN/CSN bQFs underestimated OC artifacts by 11–  
58 34%. Using a preceding organic denuder in the SEARCH network minimized passive adsorption  
59 on QBQ, but OC on QBQ may not be attributed entirely to the negative sampling artifact (e.g.,  
60 evaporated or volatilized OC from the front filter deposits after sample collection).

61 **Keywords:** Carbonaceous aerosol, organic sampling artifact, organic carbon

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63 **1. Introduction**

64 PM<sub>2.5</sub> and PM<sub>10</sub> (particulate matter with aerodynamic diameters <2.5 and 10 μm, respectively)  
65 sampling onto quartz-fiber filters is accompanied by positive (e.g., adsorption of organic vapors)  
66 and negative (e.g., volatilization of organic aerosols after sample collection) artifacts. The  
67 positive artifact, as indicated by field blanks and backup filters, is believed to exceed the  
68 negative artifact for most samples (ten Brink et al., 2004; Watson et al., 2009). Without blank or  
69 backup filter subtraction, the artifact inflates organic carbon (OC) concentrations. The artifact  
70 also biases elemental carbon (EC) values by as much as ~50%, especially when measured by  
71 thermal-optical transmittance (TOT), because light attenuation due to charring of the adsorbed  
72 organic gases within the filter has a greater influence than charring of the surface particle deposit  
73 (Chen et al., 2004; Chow et al., 2004a). In addition, OC sampling artifacts could affect PM<sub>2.5</sub> and  
74 PM<sub>10</sub> trends, mass closure, visibility degradation assessment (Chow et al., 2002a; Watson, 2002),  
75 and estimates of radiative forcing (MacCracken, 2008).

76 Composition of the adsorbed/desorbed material, its exchange between gas and particle  
77 phases, the degree to which filters become saturated, and how the sign and amount of artifact  
78 differ among filter media and sampling environments have been studied, but these issues are not  
79 well understood (Arhami et al., 2006; Arp et al., 2007; Cadle et al., 1983; Chow et al., 1994;  
80 1996; 2002b; 2006; 2008a; Eatough et al., 1989; 2003; Fan et al., 2004; Fitz, 1990; Hart and  
81 Pankow, 1994; Kim et al., 2001; 2005; Kirchstetter et al., 2001; Lewtas et al., 2001; Mader and  
82 Pankow, 2000; 2001a; 2001b; Matsumoto et al., 2003; McDow and Huntzicker, 1990; Noll and  
83 Birch, 2008; Olson and Norris, 2005; Salma et al., 2007; Subramanian et al., 2004; ten Brink et  
84 al., 2004; Turpin et al., 1994; Vecchi et al., 2009; Viana et al., 2006; Watson and Chow, 2002;  
85 Watson et al., 2009). Several approaches have been used to estimate the OC sampling  
86 artifact—including passive field blank subtraction, backup filter adjustment, slicing method (e.g.,  
87 examination of artifact distribution homogeneity within quartz-fiber filters), pre-filter organic  
88 denuders, and regression intercepts (Watson et al., 2009).

89 Frank (2006) developed the SANDWICH method to estimate artifact-free OC or OC  
90 mass (OCM). This method assumes that all of the unaccounted PM<sub>2.5</sub> mass measured on a  
91 Teflon®-membrane filter (i.e., when weighted sums of elements and ions are subtracted) can be  
92 associated with the carbonaceous component. This is based on the principle that Teflon®-  
93 membrane filters are inert and their tendency to adsorb organic vapors is low. These filters are

94 expected to have a minimal positive OC artifact, although their negative organic artifact might be  
95 larger than that of quartz-fiber filters.

96 In the U.S., the 1) Interagency Monitoring of Protected Visual Environments  
97 (IMPROVE; Malm et al., 1994) network, 2) Speciation Trends Network (STN; part of the  
98 Chemical Speciation Network [CSN]; Chu, 2004), and 3) Southeastern Aerosol Research and  
99 Characterization study (SEARCH; Hansen et al., 2003) are three long-term PM<sub>2.5</sub> chemical  
100 speciation programs that include OC and EC measurements with different approaches to  
101 sampling, analysis, and OC artifact assessment and correction. Of the 181 IMPROVE sites, more  
102 than 93% (170 sites) are located in National Parks and wilderness areas that represent different  
103 regions of the U.S. These sites are far away from population centers or local pollution sources,  
104 with a 100–1000 km zone of representation (40 CFR part 50; U.S. EPA 2006a). Regional or non-  
105 urban PM<sub>2.5</sub> sites are affected by naturally occurring aerosol from windblown dust, wildfires, and  
106 marine aerosol, as well as by pollution generated in urban and industrial areas that may be more  
107 than 1000 km distant. Urban STN/CSN sites represent a mixture of particles from many sources  
108 within the urban complex, including but not dominated by neighborhood-scale (500 m to 4 km)  
109 sources. Urban-scale (4 to 100 km) sites are usually located on city roof-tops of two- to four-  
110 story buildings—away from highly travelled roads, industries, and residential heating to represent  
111 human exposure—typically in an urban area with population > 200,000 (U.S. EPA, 1997; Chow  
112 et al., 2002c). The SEARCH network was designed to evaluate human exposure at urban versus  
113 rural environments in the southeastern U.S. (Mississippi, Alabama, Georgia, and Florida). Fig 1  
114 shows the sampling site locations, and Table 1 summarizes network characteristics relevant to  
115 the OC artifact.

116 IMPROVE artifact corrections using monthly median OC<sub>QBQ</sub> at six sites (shown in Fig.  
117 1) assume that vapors are adsorbed uniformly throughout the front and backup filters. This  
118 implies that a saturation level is attained. Otherwise, organic vapors would be scavenged  
119 preferentially in the upper layers of QF before the gas is transmitted to QBQ. Since a subset of  
120 filters is used for blank subtraction, it also is assumed that saturated OC artifact values are  
121 invariant with respect to the filter batch, sampled environment, passive/active deposition, and  
122 sampling period. Kirchstetter et al. (2001) suggested that each filter may have a different  
123 capacity for organic vapor adsorption. The slicing method by Watson et al., (2009) further  
124 showed that adsorbed OC is neither uniformly distributed throughout the filter depth, nor does

125 the adsorbed OC on the backup filter always equal that on the front filter. However, the number  
126 of samples examined was too small to draw broad generalizations.

127 Urban environments, where most of the STN/CSN sites are located, contain volatile  
128 organic compounds (VOCs) and semi-volatile organic compounds (SVOCs) adsorbable to  
129 quartz-fiber filters. SVOCs in fresh emission plumes gradually come into equilibrium as the  
130 plumes age. Oxidation of low-volatility hydrocarbons has been suggested to be a main pathway  
131 for secondary organic aerosol (SOA) formation (Robinson et al., 2007). By the time urban  
132 plumes transport to rural and remote atmospheres (e.g., most IMPROVE sites), many SVOCs  
133 could have been scavenged or converted to more stable PM compounds (Yu et al., 2004; Lane et  
134 al., 2008). Average  $OC_{QBQ}$  measurements in the IMPROVE network were ~19% higher than  
135  $OC_{bQF}$  values, but this difference is within the standard deviation of the average (Watson et al.,  
136 2009). The fact that levels of  $OC_{QBQ}$  and  $OC_{bQF}$  are similar reflects relatively low SVOC  
137 concentrations at most of the IMPROVE regional-background environments. In contrast, Watson  
138 et al. (2009) showed that  $OC_{QBQ}$  from an urban site (Fort Meade, MD) contained twice the levels  
139 of  $OC_{bQF}$ .

140 This study examines the methods and results of OC artifact assessment in these networks  
141 by 1) documenting procedures to acquire blank and backup filters; 2) comparing laboratory  
142 blank, field blank (bQF), trip blank (tbQF), and quartz-fiber backup (QBQ) filter OC levels for  
143 the period from 1/1/2005 to 31/12/2006; and 3) assessing blank OC levels from eight collocated  
144 IMPROVE and STN/CSN sites using the SANDWICH method (Frank, 2006). These results  
145 should be of interest to those using data from these and similar networks for various data analysis  
146 purposes.

147 Three hypotheses are tested using data from the three networks:

148 H1: The OC sampling artifact represented by bQF or QBQ depends on sampling  
149 protocol and differs among ambient networks.

150 H2: Sampling artifact and SVOC content are lower at non-urban (rural and remote)  
151 sites than urban sites due to aerosol aging.

152 H3: Artifact-free OC concentrations can be better estimated by the SANDWICH  
153 method (Frank, 2006) than by direct OC measurements.

154 **2. Methods**

155 As shown in Table 1, seven different filter samplers are used among the three networks with  
156 flow rates ranging from 6.7 to 22.8 liters per minute (L/min). The largest variability is in  
157 STN/CSN, which uses five types of samplers, varying from single channel (e.g., URG MASS  
158 [Chapel Hill, NC] and Rupprecht & Patashnick [R&P; now Thermo Scientific] Partisol-Plus  
159 Model 2025 Sequential Federal Reference Method [FRM] sampler [Franklin, MA]) to five  
160 parallel channels [e.g., MetOne Spiral Aerosol Speciation Sampler; SASS; Grants Pass, OR]).  
161 STN sites were originally required to use one of three samplers (i.e., URG MASS, MetOne  
162 SASS, or the Andersen RAAS). In 2005, about 75% of the STN/CSN sites used 6.7 L/min  
163 MetOne SASS samplers. The Texas Commission on Environmental Quality (TCEQ) uses the  
164 R&P 2025 to collect PM<sub>2.5</sub> at non-trends CSN sites.

165 The IMPROVE and SEARCH networks use 25 mm and 37 mm diameter Pallflex®  
166 Tissuquartz (Ann Arbor, MI) quartz-fiber filters, respectively, while STN/CSN used 47 mm  
167 Whatman QMA filters (Clifton, NJ), which contain a 5% borosilicate binder. Deposit areas range  
168 from 3.53 cm<sup>2</sup> (IMPROVE) to 11.78 cm<sup>2</sup> (R&P 2025) and face velocities range from 9.5 cm/sec  
169 (MetOne) to 107.2 cm/sec (IMPROVE). The different filter holder configurations (e.g.,  
170 single/tandem filter packs vs. magazine [R&P 2025, with a stack of 16 filter cassettes]) and  
171 materials (e.g., polycarbonate, aluminum, or Teflon®-coated) also might affect the magnitude of  
172 the OC artifact (Watson and Chow, 2009).

173 Prior to sampling, quartz-fiber filters are treated at 900°C for three to four hours and  
174 acceptance tested. After this treatment, average blank levels are 0.15 ± 0.15 µg OC or total  
175 carbon (TC = OC + EC)/cm<sup>2</sup> and 0 ± 0.02 µg EC/cm<sup>2</sup> for Pallflex® quartz-fiber filters, and 0.10  
176 ± 0.10 µg OC /cm<sup>2</sup> and 0 ± 0.01 µg EC/cm<sup>2</sup> for Whatman QMA quartz-fiber filters.  
177 Approximately 2–3% of laboratory blanks are maintained for each network. Acceptance criteria  
178 are ≤ 2.0, 1.5, and 0.5 µg/cm<sup>2</sup> for TC, OC, and EC, respectively, in the IMPROVE and SEARCH  
179 networks, and <1 µg/cm<sup>2</sup> for TC in STN/CSN.

180 STN/CSN collects 3% trip blanks (i.e., tbQF), which are loaded into filter holders and  
181 accompany the sampled filters to and from each sampling site. Trip blanks are intended to assess  
182 contamination during shipping and are not installed in the sampler or exposed to ambient air.

183 Field blanks (e.g., dynamic blanks), accompany sample shipments and are placed in the  
184 sampler along with the sampled filters (Chow and Richards, 1990). The only difference between

185 samples and bQF is that air is not drawn through bQF. The bQF fraction of total sample number  
186 varies by tenfold among the networks: ~2% of sample filters for IMPROVE, ~10% for  
187 STN/CSN sites and SEARCH, and ~10–25% for Texas non-trends CSN. The passive period for  
188 bQF has been 1–15 minutes for STN/CSN and SEARCH, and ~7 days for IMPROVE and Texas  
189 non-trends CSN sites.

190 Since the bQF fraction of all samples is only 2–10% of the total number of samples,  
191 average  $OC_{bQF}$  concentrations are used to correct sampled values with the standard deviation of  
192 the average representing the blank precision. Outliers are identified (i.e., values > 3 or 4 times  
193 the standard deviation). The small number of outliers likely results from inadvertent  
194 contamination during filter shipping/receiving or sample loading/unloading and are excluded  
195 from the averages and standard deviations.

196 QBQs are obtained from six IMPROVE and all eight of the SEARCH sites (Watson et  
197 al., 2009). Both networks collect QBQ every third day with the exception of daily sampling at  
198 two SEARCH sites (i.e., Jefferson Street, Atlanta, GA and Birmingham, AL; see Fig. 1). Ten  
199 percent of SEARCH QBQ are randomly selected for analysis. Without preceding organic  
200 denuders, the IMPROVE  $OC_{QBQ}$  represents a combination of positive and negative OC artifacts.  
201 SEARCH corrects the organic sampling artifact by calculating the quarterly mean concentrations  
202 for the QBQ and bQF and attributing them to negative and positive artifacts, respectively.  $OC_{bQF}$   
203 is multiplied by two to account for passive adsorption on both QF and QBQ. Thus,

$$204 \quad OC_{\text{artifact corrected}} = OC_{QF} + OC_{QBQ} - 2OC_{bQF} \quad (1)$$

205 where:

206  $OC_{QF}$  = Quartz-fiber front filter OC

207  $OC_{QBQ}$  = Quartz-fiber behind quartz-fiber filter OC

208  $OC_{bQF}$  = field blank OC from the quartz-fiber front filter

209 To compare carbon measurements between the IMPROVE network and STN/CSN,  
210 collocated  $PM_{2.5}$  data were acquired from three urban vs. non-urban paired sites (see Fig. 1;  
211 Seattle and Mt. Rainier, WA; Phoenix and Tonto National Monument, AZ; and Washington,  
212 D.C. and Dolly Sods Wilderness, WV; Solomon et al., 2004). In addition, collocated  
213 measurements are available from the urban Fresno, CA (Watson et al., 2000) and the non-urban  
214 Big Bend, TX (Chow et al., 2004b) sites. As indicated in Table 2, four types of STN/CSN

215 samplers were collocated with the IMPROVE samplers. The IMPROVE-STN/CSN data pairs  
216 from 2001 to 2006 with complete mass, elements, ions (i.e., a minimum of nitrate [NO<sub>3</sub><sup>-</sup>] and  
217 sulfate [SO<sub>4</sub><sup>=</sup>]) and carbon measurements are included. Prior to May 2007, the STN/CSN used a  
218 customized thermal/optical transmittance (TOT) carbon analysis protocol (Peterson and  
219 Richards, 2002) while the IMPROVE and SEARCH networks followed the IMPROVE  
220 thermal/optical reflectance (TOR) protocol (Chow et al., 1993; 2001; 2004a; 2005; 2007a). Since  
221 blank and backup filter EC levels are expected to be negligible, the analysis protocols should  
222 return equivalent OC and TC results. As noted in the footnote to Table 1, a new STN/CSN  
223 carbon sampling and analysis protocol was fully implemented in October 2009 to be consistent  
224 with the IMPROVE network.

### 225 **3. Results**

#### 226 **3.1 Blank and backup filter levels**

227 Table 3 compares average bQF levels for TC, OC, and EC in terms of areal density  
228 ( $\mu\text{g}/\text{cm}^2$ ) and ambient concentration equivalents ( $\mu\text{g}/\text{m}^3$ ), based on exposed filter areas and 24 hr  
229 sample volumes for each instrument, respectively. EC values are at or near minimum detection  
230 limits (i.e.,  $0.06 \mu\text{g}/\text{cm}^2$ ), accounting for 0 to 5% of TC, indicating that passive PM deposition is  
231 negligible. As a result, TC and OC are not statistically different and will be used  
232 interchangeably. Average bQF levels for individual sampling sites and the number of bQF  
233 acquired for IMPROVE and STN/CSN are available as supplemental information (Tables S1 –  
234 S4) and in more detailed reports (Chow et al., 2008b; Watson et al., 2008).  $\text{OC}_{\text{bQF}}$  at some  
235 sampling locations statistically differ from the network mean, although the small number of bQF  
236 at some sites may not represent the true distribution of  $\text{OC}_{\text{bQF}}$  levels during the two-year  
237 sampling period.

238 IMPROVE bQF TC (i.e.,  $\text{TC}_{\text{bQF}}$ ) areal density levels ( $2.41 \pm 0.48 \mu\text{g}/\text{cm}^2$ ) are 2.5 to 3  
239 times those of the other networks (i.e.,  $0.97 \pm 0.27 \mu\text{g}/\text{cm}^2$  for STN/CSN and  $0.81 \pm 0.61 \mu\text{g}/\text{cm}^2$   
240 for SEARCH). This probably results from the 7-day IMPROVE passive exposure period that  
241 better represents exposure of the sample filter than the 1–15 minute bQF exposure experienced  
242 by STN/CSN and SEARCH. Earlier studies in urban Los Angeles, California, and Pittsburgh,  
243 Pennsylvania, suggested a minimum exposure time for VOC passive adsorption of several hours  
244 (Subramanian et al., 2004; Turpin et al., 1994). Ambient-equivalent  $\text{TC}_{\text{bQF}}$  concentrations,  
245 however, are four times higher for STN/CSN ( $1.03 \pm 0.21 \mu\text{g}/\text{m}^3$ ) than for IMPROVE ( $0.26 \pm$



246 0.05  $\mu\text{g}/\text{m}^3$ ) and SEARCH ( $0.24 \pm 0.18 \mu\text{g}/\text{m}^3$ ) samples. This is attributed to the lower flow rate  
247 (e.g., 6.7 L/min for MetOne SASS, compared to 22.8 L/min for IMPROVE and 16.7 L/min for  
248 SEARCH) and larger exposed area of the filter deposit ( $11.76 \text{ cm}^2$  for SASS, compared to  $3.53$   
249  $\text{cm}^2$  for IMPROVE and  $7.12 \text{ cm}^2$  for SEARCH).

250 Fig 2 shows that most of the site average  $\text{OC}_{\text{bQF}}$  areal densities are 2–2.5  $\mu\text{g}/\text{cm}^2$  for  
251 IMPROVE, 0.5–1  $\mu\text{g}/\text{cm}^2$  for STN/CSN, and  $<0.5 \mu\text{g}/\text{cm}^2$  for SEARCH. For STN/CSN, average  
252  $\text{OC}_{\text{bQF}}$  varies more than twofold among sampler types, from  $0.74 \pm 0.66 \mu\text{g}/\text{cm}^2$  (URG MASS)  
253 to  $1.49 \pm 0.8 \mu\text{g}/\text{cm}^2$  (R&P 2025). Table 3 shows that the two R&P samplers (R&P 2300 and  
254 R&P 2025) reported the highest  $\text{OC}_{\text{bQF}}$  ( $1.3\text{--}1.5 \mu\text{g}/\text{cm}^2$ ). The greased inlet impaction plate and  
255 variable passive exposure periods (e.g., minutes to 7 days) for the R&P 2300 may affect  $\text{OC}_{\text{bQF}}$   
256 levels. Detailed records of bQF exposure periods are not available.

257 There were 3,628 bQF and 2,335 tbQF acquired in STN/CSN during 2005 and 2006.  
258 Average areal densities are the same:  $0.95 \pm 0.25 \mu\text{g}/\text{cm}^2$  for  $\text{OC}_{\text{bQF}}$  and  $0.95 \pm 0.23 \mu\text{g}/\text{cm}^2$  for  
259  $\text{OC}_{\text{tbQF}}$ .  $\text{OC}_{\text{bQF}}$  and  $\text{OC}_{\text{tbQF}}$  areal densities are also similar for a given sampler type, agreeing  
260 within  $\pm 0.05 \mu\text{g}/\text{cm}^2$  (Table 3). Trip blanks (tbQF) are not exposed to ambient air and are  
261 expected to have lower concentrations. The similarity of the STN/CSN  $\text{OC}_{\text{bQF}}$  and  $\text{OC}_{\text{tbQF}}$  and  
262 the SEARCH  $\text{OC}_{\text{bQF}}$  support hypothesis (H1) that the bQF exposure period (1–15 minutes) is  
263 insufficient to represent the passively adsorbed VOCs experienced by the sample filters.

264 Average  $\text{OC}_{\text{QBQ}}$  areal density is  $3.1 \pm 0.8 \mu\text{g}/\text{cm}^2$  for IMPROVE and  $1.2 \pm 0.5 \mu\text{g}/\text{cm}^2$  for  
265 SEARCH, 30 and 60% higher than the corresponding  $\text{OC}_{\text{bQF}}$  reported in Table 3, respectively.  
266 With the denuded SEARCH PCM3 sampler, average  $\text{OC}_{\text{QBQ}}$  is  $0.43 \pm 0.97 \mu\text{g}/\text{cm}^2$  higher than  
267  $\text{OC}_{\text{bQF}}$ . While  $\text{OC}_{\text{QBQ}}$  is intended to quantify negative OC artifacts that should be added to  $\text{OC}_{\text{QF}}$   
268 (see Eq. 1), it also could be interpreted as a better representation of actual bQF levels, since QBQ  
269 spends more passive exposure time in the sampler than bQF.

270 Average ambient-equivalent  $\text{OC}_{\text{QBQ}}$  concentrations are similar:  $0.33 \pm 0.09 \mu\text{g}/\text{m}^3$  for  
271 IMPROVE and  $0.35 \pm 0.15 \mu\text{g}/\text{m}^3$  for SEARCH. These levels are 20–35% higher than  $\text{OC}_{\text{bQF}}$  of  
272  $0.26 \pm 0.05$  and  $0.23 \pm 0.17 \mu\text{g}/\text{m}^3$  for IMPROVE and SEARCH, respectively, but ~65% lower  
273 than  $\text{OC}_{\text{bQF}}$  of  $1.01 \pm 0.21 \mu\text{g}/\text{m}^3$  (for all sampler types) found in the STN/CSN sites.

274 Fig 3 shows that  $\text{OC}_{\text{bQF}}$  seasonal variations are most apparent for the IMPROVE  
275 network, varying by more than 40% from winter ( $1.97 \pm 0.61 \mu\text{g}/\text{cm}^2$ ) to summer ( $2.92 \pm 0.78$   
276  $\mu\text{g}/\text{cm}^2$ ). There are no apparent changes in fractional contributions of the IMPROVE thermal

277 carbon fractions among the four seasons. Seasonal variations of  $OC_{QBQ}$  fractions (Fig. 3b) follow  
278 the same pattern as those of the IMPROVE field blanks with a summer high and winter low.  
279 Short passive exposure times at STN/CSN and SEARCH sites resulted in little to no seasonal  
280 variability:  $OC_{bQF}$  spans 0.8–1.1  $\mu\text{g}/\text{cm}^2$  and 0.52–1.0  $\mu\text{g}/\text{cm}^2$ , respectively.

281 Fig 4 shows little difference between urban and non-urban IMPROVE  $OC_{bQF}$ , but  
282 SEARCH  $OC_{bQF}$  is 17% higher at non-urban compared to urban sites. Average  $OC_{QBQ}$  for the  
283 SEARCH samples was ~25% higher at the urban ( $1.51 \pm 1.50 \mu\text{g}/\text{cm}^2$ ) compared to non-urban  
284 ( $1.18 \pm 0.98 \mu\text{g}/\text{cm}^2$ ) sites (Fig. 5). The urban increment for  $OC_{QBQ}$  is mostly in the OC1  
285 fraction, which is 146% higher at urban than at non-urban sites ( $0.51 \pm 0.84$  vs.  $0.21 \pm 0.35$   
286  $\mu\text{g}/\text{cm}^2$ ). OC2 is ~11% higher ( $0.42 \pm 0.37$  vs.  $0.38 \pm 0.48 \mu\text{g}/\text{cm}^2$ ) at the urban sites, while the  
287 other thermal fraction levels are similar. These results are consistent with hypothesis H2,  
288 indicating more SVOC adsorption at the urban sites. Average  $OC_{QBQ}$  levels from the six non-  
289 urban IMPROVE sites ( $3.1 \pm 0.8 \mu\text{g}/\text{cm}^2$ ) are 2.6 times higher than  $OC_{QBQ}$  from the four non-  
290 urban SEARCH sites ( $1.18 \pm 0.98 \mu\text{g}/\text{cm}^2$ ), consistent with the denuder removing adsorbable  
291 organic vapors.

292 Blank TC areal densities in Fig. 6 show that STN/CSN  $tbQF$  TC (i.e.,  $TC_{tbQF}$ ) areal  
293 densities are similar for urban and non-urban sites, but they differ among samplers, consistent  
294 with two-year average  $tbQF$  levels in Table 3. Using the URG MASS sampler,  $TC_{tbQF}$  areal  
295 densities at the Seattle and Mount Rainier sites are  $0.53 \pm 0.19$  and  $0.67 \pm 0.12 \mu\text{g}/\text{cm}^2$ ,  
296 respectively, lower than the 0.84–1.12  $\mu\text{g}/\text{cm}^2$  found at sites using the Andersen RAAS or  
297 MetOne SASS samplers.  $TC_{bQF}$  and  $TC_{tbQF}$  levels are similar, with a few  $bQF$  levels higher than  
298 those of  $tbQF$ . These blanks were not always acquired together.

299 IMPROVE  $TC_{bQF}$  areal densities are 2–3 times higher than those of STN/CSN  $TC_{bQF}$  or  
300  $TC_{tbQF}$ . In addition to variations in passive exposure time, IMPROVE uses the Pallflex®  
301 Tissuquartz while STN/CSN used QMA quartz-fiber filters prior to 2007, and these filters may  
302 differ in 1) capacity and affinity for VOC and gaseous SVOC adsorption and desorption, and 2)  
303 the rate to reach saturation or equilibrium between gaseous SVOC and particulate OC. The  
304 effects of these differences cannot be determined from available data.

305 For the collocated IMPROVE vs. STN/CSN comparison at the eight sites, IMPROVE  
306  $TC_{bQF}$  is most consistent among the four urban sites (Seattle, Phoenix, Washington, D.C., and  
307 Fresno), ranging from 2.5–2.7  $\mu\text{g}/\text{cm}^2$ , with lower areal densities measured at two non-urban

308 sites: Mount Rainier National Park ( $1.4 \pm 0.4 \mu\text{g}/\text{cm}^2$ ) and Tonto National Monument ( $2.0 \pm 1.1$   
309  $\mu\text{g}/\text{cm}^2$ ). Collocated STN/CSN  $\text{TC}_{\text{bQF}}$  are 40–75% lower than IMPROVE, with larger variability,  
310 ranging from  $0.66 \pm 0.42$  (Mount Rainier using URG MASS) to  $1.44 \pm 0.48 \mu\text{g}/\text{cm}^2$  (Big Bend  
311 using R&P 2025 sequential FRM). This is consistent with hypothesis H1 that longer passive  
312 deposition periods result in higher field blank levels. The number of blanks is insufficient to  
313 evaluate seasonal variability for individual sites.

314 Site-averaged non-blank corrected ambient TC concentrations ( $\mu\text{g}/\text{m}^3$ ) at each collocated  
315 IMPROVE and STN/CSN site are within  $\pm 30$ –50% of each other. STN/CSN site-averaged TC  
316 areal densities ( $\mu\text{g}/\text{cm}^2$ ) are 9–20% of those for collocated IMPROVE samples.  $\text{TC}_{\text{bQF}}$  to  $\text{TC}_{\text{QF}}$   
317 ratios are larger for non-urban than urban sites due to the lower ambient  $\text{TC}_{\text{QF}}$  levels. For  
318 example, the average  $\text{TC}_{\text{bQF}}$  reaches  $\sim 49\%$  of  $\text{TC}_{\text{QF}}$  at the Dolly Sods site for the STN/CSN  
319 sampler, but it is only 12% for the collocated IMPROVE sampler. The actual difference could be  
320 larger if STN/CSN underestimates  $\text{OC}_{\text{bQF}}$  adsorption due to the short passive exposure period.

321

### 322 **3.2 Regression method**

323 A regression method similar to that of White and Macias (1989) is used to evaluate the  
324 relative sampling artifact among collocated samples. If the collocated IMPROVE and STN/CSN  
325 samples measure the same TC, a linear regression of collocated data pairs should yield a slope of  
326 1.0, an intercept of 0, and a correlation of 1.0, within experimental precision. A statistically  
327 significant positive or negative intercept at  $\text{TC}=0$  can be interpreted as the difference in organic  
328 sampling artifacts. A robust perpendicular least squares regression method (Dutter and Huber,  
329 1981) is used to avoid biases caused by a few outliers and to account for the presence of errors in  
330 both variables. Using Phoenix data as an example, Fig. 7 shows a positive STN/CSN TC  
331 sampling artifact of  $1.65 \mu\text{g}/\text{m}^3$  or  $1.34 \mu\text{g}/\text{cm}^2$  (using MetOne SASS sampling volume and  
332 deposit area) relative to the IMPROVE sampler. Reversing the independent and dependent  
333 variables in Fig. 7 does not change the conclusion when using a robust regression.

334 Fig 8 shows that the regression intercepts are positive for each season at the eight sites,  
335 consistent with lower flow rates for the STN/CSN samples. For five of the eight sites, the  
336 intercept is largest during summer, ranging from  $0.22$ – $2.03 \mu\text{g}/\text{m}^3$ . It is highest during spring at  
337 the Mount Rainier and Tonto sites, and highest during fall at the Fresno site. The intercepts in  
338 Table 5 represent the average of four seasons. The largest two intercepts are found at the Phoenix

339 (1.34  $\mu\text{g}/\text{cm}^2$ ) and Big Bend (1.29  $\mu\text{g}/\text{cm}^2$ ) sites using the MetOne SASS and R&P 2025  
 340 samplers, respectively, while the lowest two are found at the Seattle (0.24  $\mu\text{g}/\text{cm}^2$ ) and Mount  
 341 Rainier (0.50  $\mu\text{g}/\text{cm}^2$ ) sites using the URG MASS samplers.

342 Based on the sample volume/deposit area for each sampler type (Table 1), the  
 343 relationship between IMPROVE and STN/CSN sampling artifacts (i.e.,  $TC_{IMP}$  vs.  $TC_{STN}$  in  
 344  $\mu\text{g}/\text{cm}^2$ ) can be expressed as:

$$345 \quad TC_{STN} = TC_{STN_{art}} + b \times TC_{IMP} \quad (2)$$

346 where the intercept,  $TC_{STN_{art}}$  in  $\mu\text{g}/\text{cm}^2$ , represents the additional artifact in  $TC_{STN}$  relative to  
 347  $TC_{IMP}$ . Regression statistics are summarized in Table 5. Table 6 shows that STN/CSN  $TC_{bQF}$  is  
 348 11–34% lower than  $TC_{STN_{art}}$  at all sites except for the non-urban Tonto and Dolly Sod sites.  
 349 Measured STN/CSN  $TC_{bQF}$  is similar to calculated  $TC_{STN_{art}}$  at the Tonto site. The Dolly Sods site  
 350 exhibits low  $TC_{bQF}$  levels (e.g., 0.4 and 0.3  $\mu\text{g}/\text{cm}^2$ ; see Fig. 6), and a lower correlation ( $r = 0.7$ )  
 351 was found between IMPROVE and STN (Andersen RAAS) samples at this site.

### 352 **3.3 Organic carbon Mass (OCM) estimated by the SANDWICH method**

353 The SANDWICH method was applied to 716 collocated filter pairs taken at four urban  
 354 (i.e., Seattle, WA; Phoenix, AZ; Washington, D.C.; and Fresno, CA) sites from 28/4/2001 to  
 355 29/12/2004. The number of sample pairs varied from 27 at the Fresno Supersite to 354 at the  
 356 Seattle site. Total carbonaceous mass (TCM) was calculated by subtracting  $\text{NO}_3^-$ ,  $\text{SO}_4^-$ ,  
 357 ammonium ( $\text{NH}_4^+$ ), an estimate for water ( $\text{H}_2\text{O}$ ), and crustal components from the measured  
 358  $\text{PM}_{2.5}$  mass. The calculated OCM is derived by subtracting measured EC from TCM:

$$359 \quad \text{TCM} = \text{PM}_{2.5} - (\text{SO}_4^- + \text{Retained NO}_3^- + \text{NH}_4^+ + \text{H}_2\text{O} + \text{Crustal Material} + \text{Blank}) \quad (3)$$

$$360 \quad \text{OCM} = \text{TCM} - \text{EC} \quad (4)$$

361 where:

$$362 \quad \text{Crustal Material} = 3.73 \times \text{Si} + 1.63 \times \text{Ca} + 2.42 \times \text{Fe} + 1.94 \times \text{Ti} \quad (5)$$

$$363 \quad \text{Blank} = 0.3 - 1.5 \mu\text{g}/\text{m}^3 \text{ for STN/CSN; } 0 \text{ for IMPROVE}$$

364

365 All IMPROVE data were blank-subtracted (in  $\mu\text{g}/\text{m}^3$ ). For STN/CSN, a nominal  $\text{OC}_{bQF}$   
 366 value of 0.3–1.5  $\mu\text{g}/\text{m}^3$  is used for carbon blank subtraction (Frank, 2006), which varies by  
 367 sampler type. This interval overlaps with the  $\text{OC}_{bQF}$  of  $0.66 \pm 0.94 \mu\text{g}/\text{m}^3$  at the Seattle, Phoenix,  
 368 and Washington, D.C., sites;  $\text{OC}_{bQF}$  for the Fresno site were not available. Retained  $\text{NO}_3^-$  was

369 calculated using the daily average temperature and relative humidity during the sampling period;  
370 and particle-bound water was calculated using the Aerosol Inorganics Model (AIM) as described  
371 by Frank (2006).

372 OCM concentrations from the SANDWICH method are converted to measured OC using  
373 a multiplier that accounts for unmeasured hydrogen, oxygen, and other elements in the organic  
374 compounds (i.e., El Zanan et al., 2005; Turpin and Lim, 2001; White and Roberts, 1977):

$$375 \quad \text{OCM} = X \times \text{OC} \quad (6)$$

376 where:

377 X = unmeasured element multiplier (assumed to be 1.4 for fresh and 1.8 for aged  
378 aerosol)

379 OC = measured particulate organic carbon

380

381 For IMPROVE samples, average OCM concentrations are  $3.99 \pm 2.96 \mu\text{g}/\text{m}^3$ ,  $4.40 \pm 3.45$   
382  $\mu\text{g}/\text{m}^3$ ,  $3.00 \pm 3.16 \mu\text{g}/\text{m}^3$ , and  $6.73 \pm 3.56 \mu\text{g}/\text{m}^3$  at the Seattle, Phoenix, Washington, D.C., and  
383 Fresno sites, respectively (Table 7). Better agreement with measured OC was found for a  
384 multiplier of 1.4 rather than 1.8 for all but the Fresno site. Agreement between  $\text{OC} \times 1.4$  and  
385 OCM for the IMPROVE samples was 95%, 100%, 123%, and 71% at the Seattle, Phoenix,  
386 Washington D.C., and Fresno sites, respectively. For STN/CSN samples, agreement was 90%  
387 (URG MASS) at the Seattle site, 79% (Andersen RAAS) at the Washington, D.C. site, and 123%  
388 and 88% (both using MetOne SASS) at the Phoenix and Fresno sites, respectively.

389 To assess whether low, mid-range, or high concentration samples exhibit differences,  
390 Table 7 compares estimated 10<sup>th</sup>, 50<sup>th</sup>, and 90<sup>th</sup> percentiles, respectively. The percent differences  
391 between the average and median (50% of total) are similar (within  $\pm 25\%$ ) for the sites using  
392 IMPROVE samples for multipliers of 1.4 or 1.8. At low concentrations (the 10<sup>th</sup> percentile),  
393 OCM by the SANDWICH method is 217–279% higher than measured OCM concentrations at  
394 the Washington, D.C. site. Using STN/CSN samples, OCM by the SANDWICH method is also  
395 twofold higher at low concentrations for the Phoenix site; but the agreement is reasonable  
396 (117%) for high concentration samples (90<sup>th</sup> percentile) at this and other STN/CSN sites.

#### 397 **4. Deviations from hypotheses**

398 Findings from this study are used to address the three hypotheses (H1 to H3):

399 H1: The OC sampling artifact represented by bQF or QBQ depends on sampling protocol and  
400 differs among ambient networks.

401 This hypothesis is valid based on observations. The IMPROVE, STN/CSN, and  
402 SEARCH networks use different sampling configurations, flow rates, filter material, and filter  
403 sizes. For bQF, which accompany sample filters to the field and are intended to emulate their  
404 passive deposition and adsorption, only the IMPROVE network provides an adequate (~7 days)  
405 passive exposure period for blank subtraction. The limited exposure times (1–15 minutes) in the  
406 STN/CSN and SEARCH networks are of insufficient duration to represent passive adsorption on  
407 the sampled filter. Based on both the network averages and collocated-site comparisons,  
408 IMPROVE TC<sub>bQF</sub> (or OC<sub>bQF</sub>) areal density ranges from 2.0 to 2.5  $\mu\text{g}/\text{cm}^2$ , while STN/CSN and  
409 SEARCH field blanks are below or close to 1  $\mu\text{g}/\text{cm}^2$ . STN/CSN field and trip blank TC and OC  
410 concentrations are similar ( $\sim 0.95 \pm 0.23 \mu\text{g}/\text{cm}^2$ ), within  $\pm 5\%$  for site averages. Among the five  
411 STN/CSN samplers, URG MASS reports the lowest OC<sub>bQF</sub> levels.

412 Regression analysis using uncorrected TC from collocated IMPROVE-STN/CSN  
413 samples show higher STN/CSN than IMPROVE areal densities ( $\mu\text{g}/\text{cm}^2$ ) at the same site.  
414 Without blank correction, STN/CSN sampling artifacts in  $\mu\text{g}/\text{m}^3$  could be 5–11 times higher than  
415 those in IMPROVE, depending on the sampler type. When corrected with respective field  
416 blanks, STN/CSN TC concentrations are still higher at most sites, indicating that STN/CSN field  
417 blanks underrepresent the organic artifact by ~20–30% (assuming IMPROVE bQF fully  
418 represents the artifact), but the number of bQF available for comparison was limited.

419 QBQ filters stay in the field for more than 24 hours with filtered air drawn through them  
420 for 24 hours. With a similar level of sampling artifact in areal density ( $\mu\text{g}/\text{cm}^2$ ), STN/CSN and  
421 SEARCH TC (or OC) concentration ( $\mu\text{g}/\text{m}^3$ ) would be more influenced than those of IMPROVE  
422 due to smaller sampling volumes and larger filter sizes. The average OC<sub>QBQ</sub> concentration is  $0.35$   
423  $\pm 0.1 \mu\text{g}/\text{m}^3$  for IMPROVE and  $0.38 \pm 0.15 \mu\text{g}/\text{m}^3$  for SEARCH (with preceding denuder).

424 H2: Sampling artifact and SVOC content are lower at non-urban sites than urban sites due to  
425 aerosol aging.

426 Comparisons between urban and non-urban sites in the SEARCH network are consistent  
427 with this hypothesis, but they are not sufficient to prove it. Average OC<sub>QBQ</sub> was ~25% higher at  
428 the urban sites, with  $1.51 \pm 1.50 \mu\text{g}/\text{cm}^2$  at urban sites and  $1.18 \pm 0.98 \mu\text{g}/\text{cm}^2$  at the non-urban  
429 sites in the SEARCH network. The increments between the urban and non-urban sites were  
430 ~146% for OC1 and 11% for OC2. The majority of this low temperature OC is gaseous VOCs.

431 However, during the collocated IMPROVE-STN/CSN comparisons,  $TC_{bQF}$  were not always  
432 lower at non-urban than urban sites, although this depends on the extent of VOC saturation. The  
433 contrast between urban and non-urban sites only can provide indirect indication of aging effect  
434 since the degree of aging is not certain.

435 H3: Artifact-free OC concentrations can be better estimated by the SANDWICH method  
436 (Frank, 2006) than by direct OC measurements.

437 This hypothesis is invalid based on observations. The SANDWICH method (Frank,  
438 2006) is based on  $PM_{2.5}$  mass closure, but many species are not measured on Teflon®-membrane  
439 filters, including carbon,  $NO_3^-$ ,  $SO_4^{2-}$ , and  $NH_4^+$ . Different collection/retention efficiencies of  
440 Teflon®-membrane, quartz-fiber, and nylon-membrane filters with respect to these species have  
441 not been well quantified. In addition, the mass of water and unidentified species may generate  
442 more uncertainties (Tierney and Connor, 1967; Kajino et al., 2006). All of these contribute to  
443 mass closure uncertainties. Even if organic carbon mass (OCM) can be calculated from the  
444 SANDWICH method, this study shows that variation in OCM concentration due to the choice of  
445 OC multiplier (e.g., 1.4 or 1.8) is comparable to the magnitude of the organic sampling artifact  
446 (5–30% of OCM). It is difficult to determine whether the excess OCM mass, if any, is due to  
447 sampling artifact or the correction coefficient used to convert OC to OCM.

448 The SANDWICH method did not work well for samples with low concentrations, for  
449 which the calculated and measured OC ratio exceeded 200% with the STN/CSN sampler (e.g.,  
450 Phoenix, AZ). Even though the SANDWICH method did not provide a better representation of  
451 OC or OC artifact, it is a useful tool to estimate OC when carbon measurements are not  
452 available.

## 453 **5. Conclusions**

454 There is no simple way to correct for sampling artifacts using current measurements. With the  
455 newly implemented STN/CSN carbon measurements (U.S.EPA, 2006b), using the modified  
456 IMPROVE Module C sampler (i.e., URG 3000N sampler), sampling artifacts will be reduced via  
457 a higher flow rate (e.g., 22.8 L/min instead of 6.7 L/min) and a smaller deposit area ( $3.53 \text{ cm}^2$   
458 instead of  $11.76 \text{ cm}^2$ ). In addition, bQF will remain in the sampler for the same period as QF and  
459 QBQ samples at all STN/CSN sites. For each network, blank corrections should be made and  
460 uncertainties propagated, even though the reported OC is under-corrected for adsorbed organic

461 vapors due to inadequate passive deposition period for field blanks. Each network should acquire  
462 bQFs and QBQs at the same frequency and passive deposit duration (e.g., once per month on an  
463 every-sixth-day sampling schedule; expose field blanks for a minimum of three days). More  
464 research, perhaps through controlled experiments, are warranted on 1) sample duration for filter  
465 saturation of adsorbed gases; 2) dependence of adsorbed gas saturation on particle composition,  
466 temperature, relative humidity, and sampling face velocity; 3) evaporation rates of semi-volatile  
467 organic compounds during sampling; and 4) source-specific tests (e.g., diesel, gasoline, and  
468 wood smoke).

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689

**Table 1.** Sampling protocols for carbon in the IMPROVE, STN/CSN, and SEARCH networks from 1/1/2005 to 31/12/2006.

Network <sup>a</sup>	IMPROVE		STN/CSN		CSN		SEARCH
Sampler Type <sup>b</sup>	IMPROVE	Met One SASS	Andersen RAAS	URG MASS400/450	R&P 2300	R&P Partisol Plus 2025 Sequential FRM	PCM3
Number of channels	3 modules	5 channels (3 used)	4 channels (3 used)	2 single channel samplers	12 channels (4 used)	2 single channel modules	3 channels
Carbon denuder	None	None	None	None	None	None	Activated carbon honeycomb denuder
Inlet type <sup>c</sup>	AIHL cyclone	SASSSCC sharp-cut cyclone	AIHL cyclone	Louvered PM <sub>10</sub> inlet/WINS	Harvard Impactor	Louvered PM <sub>10</sub> inlet/WINS or VSSC	WINS impactor
Filter holder and cassette types	Polycarbonate	Aluminum holder and Delrin® cassette	Teflon®-coated in-line holder preceded with a diffuser	Teflon® holder and cassette	Teflon®-coated aluminum holder and Teflon® support screens	Molded plastic cassette in a 16 cassette magazine	Savillex-molded Teflon®
Sampling frequency	3rd day	3rd day	3rd day	3rd day	3rd day/ 6 <sup>th</sup> day	3rd day/6th day	3rd day
Flow rate	22.8 L/min	6.7 L/min	7.3 L/min	16.7 L/min	10.0 L/min	16.7 L/min	16.7 L/min
Filter deposit area	3.53 cm <sup>2</sup>	11.76 cm <sup>2</sup> <sup>d</sup>	11.76 cm <sup>2</sup> <sup>d</sup>	11.76 cm <sup>2</sup> <sup>d</sup>	11.76 cm <sup>2</sup> <sup>d</sup>	11.78 cm <sup>2</sup> <sup>e</sup>	7.12 cm <sup>2</sup>
Filter face velocity	107.2 cm/sec	9.5 cm/sec	10.3 cm/sec	23.7 cm/sec	14.2 cm/sec	23.6 cm/sec	39.1 cm/sec
Sample volume	32.7 m <sup>3</sup>	9.6 m <sup>3</sup>	10.5 m <sup>3</sup>	24 m <sup>3</sup>	14.4 m <sup>3</sup>	24 m <sup>3</sup>	24 m <sup>3</sup>
Quartz-fiber filter pre-fire temperature and duration	900 °C for 4 hr	900 °C for 3 hr	900 °C for 3 hr	900 °C for 3 hr	900 °C for 3 hr	900 °C for 3 hr	900 °C for 3 hr
Quartz filter type	25 mm Pall	47mm Whatman <sup>f</sup>	47mm Whatman <sup>f</sup>	47mm Whatman <sup>f</sup>	47mm Whatman <sup>f</sup>	47mm Whatman <sup>f,g</sup>	37 mm Pall
Quartz filter pack configuration <sup>h</sup>	QF or QBQ	QF	QF	QF	QF	QF	Organic carbon denuder/ QBQ
Sites with backup filters (QBQ)	6	0	0	0	0	0	8%
Passive deposition duration	7 days <sup>i</sup>	variable <sup>j</sup>	1-15 min <sup>j</sup>	1-15 min <sup>j</sup>	variable <sup>j</sup>	5 – 7 days <sup>j</sup>	1-15 min
Laboratory blank <sup>k</sup> frequency	2%	2 - 3%	2 - 3%	2 - 3%	2 - 3%	2 - 3%	2%
Trip blank <sup>l</sup> frequency	0%	3%	3%	3%	3%	0%	0%
Field blank <sup>m</sup> frequency	2%	10%	10%	10%	10%	10-25%	10%
Field blank analysis frequency	100%	100%	100%	100%	100%	100%	10% on QF
Backup filter analysis frequency	100%	0	0	0	NA	NA	10%
Sample shipping method	Second day cardboard box with icepack	Priority overnight cooler with icepack	Priority overnight cooler with icepack	Priority overnight cooler with icepack	Priority overnight cooler with icepack	Priority overnight cooler with icepack	Standard overnight cooler with ice
Temperature for sample storage	4 °C	< -15 °C	< -15 °C	< -15 °C	< -15 °C	< -15 °C	4 °C
Number of sites (2006)	181	179	18	6	14	22	8

<sup>a</sup> Networks  
IMPROVE: Interagency Monitoring of PROtected Visual Environments network (Malm et al., 1994)

## Table 1. Continued.

### <sup>a</sup> Networks, continued

STN/CSN: Speciation Trends Network/Chemical Speciation Network (Flanagan et al., 2006); starting in May 2007, STN/CSN sites collect ~5% bQF and ~30% QBQ. In 2008, all of the STN/CSN sites (except the TCEQ non-trends sites) installed MetOne SASS samplers for mass, elements, and ion analyses. As of October 2009, the modified IMPROVE Module C, URG 3000N sampler (URG Corp; Chapel Hill, NC) is placed at all STN/CSN sites using Pallflex® Tissuquartz at a flow rate of 22.8 L/min on 25 mm filters for organic and elemental carbon (OC and EC) following the IMPROVE\_A thermal/optical protocol (Chow et al., 2007b).

SEARCH: Southeastern Aerosol Research and Characterization study (Hansen et al., 2003); eight sites (Mississippi pair: urban Gulfport [GLF] in Gulfport and non-urban Oak Grove [OAK] near Hattiesburg; Alabama pair: urban Birmingham [BHM] in North Birmingham and non-urban Centreville [CTR] south of Tuscaloosa; Georgia pair: urban Jefferson Street [JST] in Atlanta and non-urban Yorkville [YRK] northwest of Atlanta; and Florida pair: urban Pensacola [PNS] in Pensacola and suburban outlying field [OLF] northwest of Pensacola).

### <sup>b</sup> Sampler Descriptions

IMPROVE (new units manufactured by special order from URG, Inc.[Chapel Hill, NC]): Four parallel filter modules, each with up to four sequential sample sets (Eldred et al., 1990). Module A collects PM<sub>2.5</sub> through an Air and Industrial Hygiene Laboratory (AIHL) cyclone followed by a 25mm Teflon®-membrane filter analyzed for mass by gravimetry and for elements by X-ray fluorescence (XRF). Module B collects PM<sub>2.5</sub> through a sodium carbonate denuder (Ashbaugh et al., 2004) followed by an AIHL cyclone and a 25mm nylon-membrane filter analyzed by ion chromatography (IC) for nitrate (NO<sub>3</sub><sup>-</sup>) and sulfate (SO<sub>4</sub><sup>-</sup>). Module C collects PM<sub>2.5</sub> through an AIHL cyclone followed by a 25mm Pall quartz-fiber filter for OC and EC by the IMPROVE\_A thermal/optical reflectance (TOR) protocol. Module D collects through a louvered PM<sub>10</sub> inlet at 16.7 L/min followed by a 25 mm Pall Teflon®-membrane filter for mass by gravimetry.

SASS (Spiral Aerosol Speciation Sampler, Met One, Grants Pass, OR): Spiral centrifugal impaction inlets were originally used on this sampler (thus the name), but excessive re-entrainment from impaction surfaces caused these to be replaced with sharp-cut cyclones (Watson and Chow, 2009). The Super SASS can contain up to eight parallel channels, but the STN/CSN configuration uses three channels of a five channel version, each channel containing one 47 mm filter with a 6.7 L/min flow rate. For STN/CSN, Channel 1 contains a Whatman Teflon®-membrane filter for mass by gravimetry and elements by XRF; Channel 2 can be used for a field blank; Channel 3 includes a magnesium oxide-coated aluminum (Al) honeycomb after the cyclone followed by a Nylasorb nylon-membrane filter for water-soluble anions (i.e., NO<sub>3</sub><sup>-</sup> and SO<sub>4</sub><sup>-</sup>) and cations (i.e., ammonium [NH<sub>4</sub><sup>+</sup>] and water-soluble sodium [Na<sup>+</sup>] and potassium [K<sup>+</sup>]) by IC; Channel 4 contains a Whatman QMA quartz-fiber filter for OC and EC by the STN transmission (TOT) protocol (Peterson and Richards, 2002); Channel 5 is available for field blanks or special study samples.

RAAS (Reference Ambient Air Sampler, Andersen [now Thermo Scientific] Model 25-400; Franklin, MA, no longer manufactured; Watson and Chow, 2002): Contains four parallel channels with two 2.5 µm AIHL cyclones; all filters are 47 mm in diameter. In the STN/CSN configuration, only three channels are used: Channel 1 contains a Whatman QMA quartz-fiber filter at 7.3 L/min for OC and EC by the STN TOT protocol; Channel 2 contains a Whatman Teflon®-membrane filter at 16.7 L/min for mass by gravimetry and elements by XRF; Channel 3 is empty, but can be used for replicates or blanks at a flow of 16.7 L/min; Channel 4 contains a magnesium oxide-coated denuder followed by a Whatman Nylasorb nylon-membrane filter at a flow rate of 7.3 L/min for total nitrate by IC.

URG MASS (URG, Chapel Hill, NC): Uses two parallel modules with 47 mm filters operating at 16.7 L/min. Module 1 includes a louvered PM<sub>10</sub> inlet followed by a PM<sub>2.5</sub> WINS impactor, a magnesium oxide-coated denuder, and a stacked filter pack with a Whatman Teflon®-membrane filter on top for mass by gravimetry and elements by XRF and a Nylasorb nylon-membrane backup filter for anions and cations by IC. Module 2 contains a louvered PM<sub>10</sub> inlet followed by a WINS PM<sub>2.5</sub> impactor, which includes a Whatman QMA quartz-fiber filter for OC and EC by the STN TOT protocol.

## Table 1. Continued.

### <sup>b</sup> Sampler Descriptions, continued

R&P 2300 (Rupprecht & Patashnick [now Thermo Scientific] Model 2300; Franklin, MA): Twelve channels are available that can be programmed to be operated in parallel or sequentially. The non-trends CSN sites in Texas use four parallel channels with 47 mm diameter filters. Channel 1 contains a Whatman Teflon®-membrane filter with 16.7 L/min for mass by gravimetry and elements by XRF; Channel 2 contains an additional Teflon®-membrane filter for anion and cation analyses by IC; Channel 3 contains a quartz-fiber filter, with an optional quartz-fiber backup filter, at 10 L/min for OC and EC by the IMPROVE\_A TOR protocol; Channel 4 contains a sodium carbonate-coated honeycomb denuder followed by a Nylasorb nylon 10 L/min for total NO<sub>3</sub><sup>-</sup> by IC.

R&P 2025 (Rupprecht & Patashnick [now Thermo Scientific] Model 2025; Franklin, MA): Contains two parallel modules operated in a sequential mode using 47 mm diameter filters at 16.7 L/min. Filters are stored in a 16 cassette magazine. Both modules are preceded by a louvered PM<sub>10</sub> inlet followed by a sharp cut cyclone PM<sub>2.5</sub> inlet. Module 1 contains a Pall Teflon®-membrane filter for mass by gravimetry, elements by XRF, and cations and anions by ion chromatography. Module 2 contains a quartz-fiber filter for OC and EC by the IMPROVE\_A TOR protocol.

PCM3 (Particle Composition Monitor, Aerosol Research Associates, Plano, TX; Edgerton et al., 2005): Uses three parallel channels operated at 16.7 L/min with a URG PM<sub>10</sub> cyclone followed by a PM<sub>2.5</sub> WINS impactor. Solenoid valves behind the filter packs allow up to four sample sets to be acquired sequentially. Channel 1 contains sodium carbonate-coated annular denuder followed by a citric acid-coated annular denuder, then followed by a three-stage filter packs in: a 47 mm Teflon®-membrane filter for mass by gravimetry and elements by XRF, followed by a 47 mm Nylasorb Nylon-membrane filter for volatilized NO<sub>3</sub><sup>-</sup> by IC, followed by a 47 mm citric acid-impregnated filter for volatilized NH<sub>4</sub><sup>+</sup> by automated colorimetry (AC). Channel 2 contains a sodium carbonate-coated annular denuder followed by a citric acid-coated annular denuder and a 47 mm Nylasorb nylon-membrane filter for total NH<sub>4</sub><sup>+</sup> and total NO<sub>3</sub><sup>-</sup> by AC and IC, respectively. Channel 3 samples through a URG PM<sub>10</sub> cyclone, followed by an activated carbon honeycomb denuder to remove carbon vapors, then through a WINS PM<sub>2.5</sub> impactor onto a 37 mm Pall quartz-fiber filter followed by a backup quartz fiber filter for OC and EC by the IMPROVE\_A TOR protocol (Chow et al., 2007a).

<sup>c</sup> All inlets are made of anodized aluminum.

<sup>d</sup> RTI uses 11.76 cm<sup>2</sup> for quartz-fiber filters and 11.70 cm<sup>2</sup> exposed area for Teflon®-membrane filters for the STN/CSN sites.

<sup>e</sup> DRI uses 11.78 cm<sup>2</sup> for quartz-fiber and Teflon®-membrane exposed area for Texas non-trends CSN sites.

<sup>f</sup> Whatman QMA filters were switched to Pallflex® Tissuquartz (Ann Arbor, MI) quartz-fiber filters as of May 2007.

<sup>g</sup> Field blank is kept in the inlet and outlet of the 16 filter cassette magazines for as long as 5–7 days depending on the sampling frequency but is in sampling position (without air being drawn through it) for only a few seconds.

<sup>h</sup> QF = quartz-fiber front filter only, QBQ = quartz-fiber behind quartz-fiber filter, with the backup quartz-fiber used to estimate adsorbed organic vapors.

<sup>i</sup> Field blanks usually in samplers for 1–15 minutes, but in some cases for as long as 5–7 days.

<sup>j</sup> Based on the assumption of once per week site visits.

<sup>k</sup> Laboratory blanks are selected from each batch of 100 unexposed filters and submitted for acceptance testing.

<sup>l</sup> Trip blanks accompany batches of shipped filters but are not removed from their storage containers.



<sup>m</sup> Field blanks accompany batches of shipped filters, but are removed from storage containers and left exposed to passive sampling. Only the IMPROVE network exposes field blanks for the same length of times as the sampled filters.

**Table 2.** Collocated IMPROVE and STN/CSN PM<sub>2.5</sub> speciation data from 16/10/2001 to 31/12/2006.

Type	Site Name	Inclusive Period	Number of Samples	IMPROVE <sup>a</sup>		MetOne SASS	STN <sup>a</sup> Sampler Type			# of Field Blanks
				Module C Sampler	# of Field Blanks		Anderson RAAS	URG MASS	R&P 2025	
Special Study	Puget Sound (PUSO), Seattle (Beacon Hill), WA	16/10/2001-29/12/2003	224	X	8			X		25
	Mount Rainier NP (MORA), WA	16/10/2001-1/11/2002	69	X	6			X		12
	Phoenix (PHOE), AZ	16/10/2001-29/12/2003	201	X	6	X				26
	Tonto National Monument (TONT), AZ	16/10/2001-29/12/2003	181	X	8	X				28
	Washington D.C. (WASH)	16/10/2001-29/12/2003	206	X	5			X		25
	Dolly Sods Wilderness (DOSO), WA	16/10/2001-29/12/2003	140	X	5			X		26
<b>Total</b>			<b>1,021</b>		<b>38</b>					<b>142</b>
Long-term Sites	Fresno (FRES), CA	1/1/2005-31/12/2006	227 <sup>b</sup>	X	4			X		23
	Big Bend NP (BIBE), TX	1/1/2005-31/12/2006	81 <sup>b</sup>	X	3				X <sup>c</sup>	26
<b>Total</b>			<b>308</b>							<b>49</b>

<sup>a</sup> See Table 1 for sampler specifications.

<sup>b</sup> Due to missing NO<sub>3</sub><sup>-</sup> data at the Big Bend site and incomplete 2006 data, only 172 data pairs, as well as 7 IMPROVE and 32 STN/CSN field blanks from Fresno and Big Bend National Park (BIBE) were included. For carbon analysis alone, the completed 2005 and 2006 data would provide 308 sample pairs and 49 field blanks.

<sup>c</sup> Big Bend NP (BIBE) is a CSN site.

**Table 3.** Comparison of average field blank (bQF), trip blank (tbQF), and backup (QBQ) filter carbon levels ( $\pm$  standard deviation) among the IMPROVE, STN/CSN, and SEARCH networks for the period from 1/1/2005 to 31/12/2006.

Network	Filter Type	Type of PM <sub>2.5</sub> Speciation Sampler	Site Count	No. of Field Blanks	TC	OC	EC	TC	OC	EC
					$\mu\text{g}/\text{cm}^{2\text{d}}$			$\mu\text{g}/\text{m}^{3\text{e}}$		
IMPROVE <sup>c</sup>	bQF	IMPROVE Module C	181	886	2.41 $\pm$ 0.48	2.37 $\pm$ 0.45	0.04 $\pm$ 0.05	0.26 $\pm$ 0.05	0.26 $\pm$ 0.05	0 $\pm$ 0.01
		QBQ	6	1,401	3.23 $\pm$ 0.96	3.08 $\pm$ 0.83	0.16 $\pm$ 0.13	0.35 $\pm$ 0.1	0.33 $\pm$ 0.09	0.02 $\pm$ 0.01
	bQF	All Samples	239 <sup>a</sup>	3,628	0.97 $\pm$ 0.27	0.95 $\pm$ 0.25	0.02 $\pm$ 0.03	1.03 $\pm$ 0.21	1.01 $\pm$ 0.21	0.01 $\pm$ 0.02
		Andersen RAAS	22	249	0.88 $\pm$ 0.33	0.88 $\pm$ 0.33	0.01 $\pm$ 0.03	0.99 $\pm$ 0.38	0.98 $\pm$ 0.37	0.01 $\pm$ 0.03
		MetOne SASS	185	2,572	0.86 $\pm$ 0.39	0.85 $\pm$ 0.38	0.01 $\pm$ 0.05	1.05 $\pm$ 0.47	1.04 $\pm$ 0.47	0.01 $\pm$ 0.06
		URG MASS	7	150	0.75 $\pm$ 0.66	0.74 $\pm$ 0.66	0 $\pm$ 0.02	0.37 $\pm$ 0.32	0.36 $\pm$ 0.32	0 $\pm$ 0.01
		R&P 2300 Sequential Speciation	15	236	1.33 $\pm$ 0.52	1.3 $\pm$ 0.51	0.03 $\pm$ 0.11	1.09 $\pm$ 0.42	1.06 $\pm$ 0.41	0.02 $\pm$ 0.09
R&P 2025 Sequential FRM	24	421	1.57 $\pm$ 0.77	1.49 $\pm$ 0.76	0.08 $\pm$ 0.12	0.78 $\pm$ 0.38	0.73 $\pm$ 0.37	0.04 $\pm$ 0.06		
STN/CSN	tbQF	All Samples	239 <sup>a</sup>	2,335	0.98 $\pm$ 0.26	0.95 $\pm$ 0.23	0.02 $\pm$ 0.03	0.89 $\pm$ 0.33	0.87 $\pm$ 0.32	0.02 $\pm$ 0.02
		Andersen RAAS	22	241	0.84 $\pm$ 0.38	0.83 $\pm$ 0.34	0.01 $\pm$ 0.05	0.94 $\pm$ 0.42	0.93 $\pm$ 0.38	0.01 $\pm$ 0.05
		MetOne SASS	185	1,832	0.89 $\pm$ 0.45	0.88 $\pm$ 0.45	0.01 $\pm$ 0.03	1.09 $\pm$ 0.56	1.08 $\pm$ 0.55	0.01 $\pm$ 0.04
		URG MASS	7	159	0.81 $\pm$ 0.70	0.80 $\pm$ 0.69	0.01 $\pm$ 0.03	0.4 $\pm$ 0.34	0.39 $\pm$ 0.34	0.00 $\pm$ 0.01
		R&P 2300 Sequential Speciation	15	103	1.36 $\pm$ 0.48	1.30 $\pm$ 0.48	0.06 $\pm$ 0.16	1.11 $\pm$ 0.39	1.06 $\pm$ 0.39	0.05 $\pm$ 0.13
		R&P 2025 Sequential SCC	24	0	N/A <sup>b</sup>	N/A <sup>b</sup>	N/A <sup>b</sup>	N/A <sup>b</sup>	N/A <sup>b</sup>	N/A <sup>b</sup>
SEARCH <sup>c</sup>	bQF	PCM3	8	144	0.81 $\pm$ 0.61	0.76 $\pm$ 0.57	0.04 $\pm$ 0.06	0.24 $\pm$ 0.18	0.23 $\pm$ 0.17	0.01 $\pm$ 0.02
		QBQ	8	257	1.29 $\pm$ 0.52	1.19 $\pm$ 0.52	0.1 $\pm$ 0.06	0.38 $\pm$ 0.15	0.35 $\pm$ 0.15	0.03 $\pm$ 0.02

<sup>a</sup> 253 if counting 14 sites where sampler type changed between 1/1/2005 and 31/12/2006

<sup>b</sup> Data is not available

**Table 3. Continued.**

<sup>c</sup> Carbon analysis follows the IMPROVE thermal/optical reflectance (TOR) protocol (Chow et al., 2007a) for the IMPROVE and SEARCH network and the STN thermal/optical transmittance (TOT) protocol (Chu et al., 2004; Peterson and Richards, 2002) for STN/CSN.

<sup>d</sup> Areal density on filters in  $\mu\text{g}/\text{cm}^2$  is based on sample loading divided by the exposed area (e.g.,  $3.53 \text{ cm}^2$  for IMPROVE samplers,  $11.76 \text{ cm}^2$  for STN/CSN speciation samplers, and  $7.12 \text{ cm}^2$  for SEARCH PCM3 samplers as noted in Table 1)

<sup>e</sup> Equivalent ambient concentration in  $\mu\text{g}/\text{m}^3$  is based on the sample loading divided by the nominal sampler volume (varies from  $9.6 \text{ m}^3$  for MetOne SASS to  $32.7 \text{ m}^3$  for the IMPROVE sampler as noted in Table 1).

**Table 4.** Average blank TC concentrations for the eight collocated IMPROVE–STN/CSN sites.

Site Code	Site Name	Instrument Used	Number of Pairs IMPROVE-STN QF (number)	Field Blank TC Areal Density					
				IMP_bQF <sup>a</sup> ( $\mu\text{g}/\text{cm}^2$ )	IMP_bQF <sup>a</sup> (number)	STN_bQF <sup>b</sup> ( $\mu\text{g}/\text{cm}^2$ )	STN_bQF <sup>b</sup> (number)	STN_tbQF <sup>b</sup> ( $\mu\text{g}/\text{cm}^2$ )	STN_tbQF <sup>b</sup> (number)
PUSO	Seattle, WA	URG MASS	224	2.66 ± 0.54	8	0.68 ± 0.41	25	0.53 ± 0.19	9
MORA	Mount Rainier, WA	URG MASS	69	1.44 ± 0.36	6	0.66 ± 0.42	12	0.67 ± 0.12	4
PHOE	Phoenix, AZ	MetOne SASS	201	2.63 ± 0.58	6	1.40 ± 0.77	26	1.12 ± 0.50	10
TONT	Tonto Monument, AZ	MetOne SASS	181	2.00 ± 1.05	8	0.87 ± 0.31	28	0.86 ± 0.32	9
WASH	Washington, D.C.	Andersen RAAS	206	2.49 ± 0.87	5	0.87 ± 0.40	25	0.84 ± 0.26	10
DOSO	Dolly Sods, WV	Andersen RAAS	140	2.57 ± 0.31	5	1.18 ± 0.68	26	0.97 ± 0.38	8
FRES	Fresno, CA	MetOne SASS	227	2.58 ± 0.50	7	0.74 ± 0.23	18	0.94 ± 0.48	11
BIBE	Big Bend National Park, TX	R&P 2025 Sequential FRM	81	2.40 ± 0.68	7	1.44 ± 0.48	15	N/A ± N/A	N/A

<sup>a</sup> Carbon analysis follows the IMPROVE thermal/optical reflectance (TOR) protocol (Chow et al., 2007a) for the IMPROVE network

<sup>b</sup> Carbon analysis follows the STN thermal/optical transmittance (TOT) protocol (Chu et al., 2004) for STN/CSN.

**Table 5.** Robust regression statistics of uncorrected STN/CSN TC against IMPROVE TC for data from the eight collocated sites.

Site Code	Site Name <sup>a</sup>	Sampling Period	TC Concentration		TC Areal Density		Correlation (r)	N
			Slope	Intercept ( $\mu\text{g}/\text{m}^3$ )	Slope	Intercept ( $\mu\text{g}/\text{cm}^2$ )		
PUSO	Seattle, WA	16/10/2001 - 29/12/2003	0.91	0.12	0.22	0.24	0.98	224
MORA	Mount Rainier, WA	22/10/2001 - 20/10/2002	0.87	0.25	0.22	0.5	0.97	69
PHOE	Phoenix, AZ	16/10/2001 - 29/12/2003	1.08	1.65	0.088	1.34	0.94	201
TONT	Tonto Monument, AZ	16/10/2001 - 29/12/2003	1.06	0.85	0.088	0.69	0.92	181
WASH	Washington, D.C.	16/10/2001 - 26/12/2003	1.08	0.95	0.096	0.85	0.92	206
DOSO	Dolly Sods, WV	16/10/2001 - 29/12/2003	0.87	0.83	0.096	0.74	0.67	140
FRES	Fresno, CA	1/1/2005 - 31/12/2006	1.16	1.1	0.088	0.9	0.95	227
BIBE	Big Bend National Park, TX	1/1/2005 - 31/12/2006	1.22	0.64	0.22	1.29	0.79	81

<sup>a</sup> See Table 4 for STN/CSN sampler specifications.

**Table 6.** Comparison between estimated and measured sampling artifact for the eight collocated IMPROVE/STN sites.

Site Code	Site Name	Sampling Period	IMP $TC_{bQF}$ ( $\mu\text{g}/\text{cm}^2$ ) <sup>a</sup>	STN/CSN $TC_{bQF}$ ( $\mu\text{g}/\text{cm}^2$ ) <sup>c</sup>	Calculated $TC_{STN_{art}}$ ( $\mu\text{g}/\text{cm}^2$ ) <sup>b</sup>	Difference (%) <sup>d</sup>
PUSO	Seattle, WA	16/10/2001 - 29/12/2003	2.66 ± 0.54	0.68	0.83	-0.18
MORA	Mount Rainier, WA	22/10/2001 - 20/10/2002	1.44 ± 0.36	0.66	0.82	-0.19
PHOE	Phoenix, AZ	16/10/2001 - 29/12/2003	2.63 ± 0.58	1.4	1.57	-0.11
TONT	Tonto Monument, AZ	16/10/2001 - 29/12/2003	2 ± 1.05	0.87	0.87	0.01
WASH	Washington, D.C.	16/10/2001 - 26/12/2003	2.49 ± 0.87	0.87	1.09	-0.2
DOSO	Dolly Sods, WV	16/10/2001 - 29/12/2003	2.57 ± 0.31	1.18	0.99	0.19
FRES	Fresno, CA	1/1/2005 - 31/12/2006	2.58 ± 0.5	1.44	1.82	-0.21
BIBE	Big Bend National Park, TX	1/1/2005 - 31/12/2006	2.4 ± 0.68	0.74	1.13	-0.34

<sup>a</sup> IMPROVE field blanks

<sup>b</sup> Estimated STN/CSN artifact

<sup>c</sup> STN/CSN field blanks

<sup>d</sup> 
$$\frac{\text{measured STN/CSN } TC_{bQF} - \text{calculated } TC_{STN_{art}}}{\text{calculated } TC_{STN_{art}}} \times 100$$

**Table 7.** Estimates of organic carbon mass (OCM) based on the SANDWICH method for the four collocated IMPROVE/STN sites.

Site	Seattle, WA	Phoenix AZ	Washington, D.C.	Fresno, CA
Site Type	Urban	Urban	Urban	Urban
Site Code	PUSO	PHOE	WASH	FRES
Number of Collocated Pairs	354	290	45	27
Sampler Type	IMPROVE	IMPROVE	IMPROVE	IMPROVE
Sampling Period	12/7/2001-29/12/2004	28/4/2001-30/9/2004	8/7/2004-29/12/2004	3/9/2004-23/12/2004
<b>SANDWICH OCM <math>\mu\text{g}/\text{m}^3</math></b>				
Average	3.99 ± 2.96	4.40 ± 3.45	3.00 ± 3.16	6.73 ± 3.56
10%tile	1.22	1.48	0.58	2.86
50%tile	3.16	3.27	2.47	6.16
90%tile	8	8.66	6.23	11.29
<b>Measured OC <math>\mu\text{g}/\text{m}^3</math></b>				
Average	2.70 ± 2.06	3.13 ± 2.27	2.63 ± 1.51	3.42 ± 1.66
10%tile	0.87	1.32	0.90	1.47
50%tile	1.91	2.25	2.51	3.18
90%tile	5.28	6.10	4.21	5.52
<b>Measured OC×1.4/OCM</b>				
Average	95%	100%	123%	71%
10%tile	100%	125%	217%	72%
50%tile	85%	96%	142%	72%
90%tile	92%	99%	95%	68%
<b>Measured OC×1.8/OCM</b>				
Average	122%	128%	158%	91%
10%tile	128%	161%	279%	93%
50%tile	109%	124%	183%	93%
90%tile	119%	127%	122%	88%
<b>Sampler Type</b>				
<b>Sampling Period</b>				
<b>SANDWICH OCM <math>\mu\text{g}/\text{m}^3</math></b>				
Average	4.63 ± 3.27	4.48 ± 3.62	4.85 ± 5.14	7.66 ± 4.37
10%tile	1.57	1.03	0.65	3.08
50%tile	3.72	3.64	4.03	6.72
90%tile	9.59	9.12	9.05	14.22
<b>Measured OC <math>\mu\text{g}/\text{m}^3</math></b>				
Average	2.98 ± 2	3.94 ± 2.43	2.75 ± 1.86	4.8 ± 2.71
10%tile	1.16	1.58	0.56	1.81
50%tile	2.38	3.34	2.78	4.12
90%tile	6.06	7.59	5.62	8.91
<b>Measured OC×1.4/OCM</b>				
Average	90%	123%	79%	88%
10%tile	103%	215%	121%	82%
50%tile	90%	128%	97%	86%
90%tile	88%	117%	87%	88%
<b>Measured OC×1.8/OCM</b>				
Average	116%	158%	102%	113%
10%tile	133%	276%	155%	106%
50%tile	115%	165%	124%	110%
90%tile	114%	150%	112%	113%



## 8. Figure Captions

**Fig. 1.** Sampling locations for the 181 Interagency Monitoring of PROtected Visual Environments (IMPROVE) sites (circles, mostly non-urban), 239 Speciation Trends Network (STN)/Chemical Speciation Network (CSN) sites (triangles, mostly urban), and eight paired Southeastern Aerosol Research and Characterization study (SEARCH) sites (squares, urban vs. non-urban/suburban pairs [Mississippi GLF (urban Gulfport) and OAK (non-urban Oak Grove near Hattiesburg); Alabama BHM (urban north Birmingham) and CTR (non-urban Centreville south of Tuscaloosa); Georgia JST (urban Jefferson Street in Atlanta) and YRK (non-urban Yorkville, north of Atlanta); and Florida PNS (urban Pensacola) and OLF (suburban outlying field northwest of Pensacola, classified as non-urban)]). The six IMPROVE locations that include quartz-fiber behind quartz-fiber filters (QBQ) are indicated by purple crosses (Mount Rainier National Park, WA [MORA1]; Yosemite National Park, CA [YOSE1]; Hance Camp, Grand Canyon National Park, AZ [HANC1]; Chiricahua National Monument, AZ [CHIR1]; Okefenokee National Wildlife Reserve, GA [OKEF1]; and Shenandoah National Park, WV [SHEN1]). The eight collocated IMPROVE/CSN sites are indicated by red stars (Puget Sound, WA [PUGO1]; Mount Rainier National Park, WA [MORA1]; Fresno, CA [FRES1]; Phoenix, AZ [PHOE1]; Tonto National Monument, AZ [TONT1]; Big Bend National Park [BIBE1]; Dolly Sods Wilderness [DOSO1]; and Washington, D.C. [WASH1]). MORA1 (Mount Rainier, WA), which is included both in the IMPROVE QBQ and the collocated IMPROVE/CSN sites, is indicated in dark green.

**Fig. 2.** Field blank organic carbon ( $OC_{bQF}$ ) concentration density ( $\mu\text{g}/\text{cm}^2$ ) for a) 181 IMPROVE sites, b) 239 STN/CSN sites, and c) eight SEARCH sites for the period from 1/1/2005 to 31/12/2006 (each bar represents the concentration sector less than or equal to the assigned value).

**Fig. 3.** Seasonal variations of blanks among a) IMPROVE field blanks ( $OC_{bQF}$ ), b) IMPROVE backup filters ( $OC_{QBQ}$ ; six sites), c) STN/CSN field blanks ( $OC_{bQF}$ ), d) STN/CSN trip blanks ( $OC_{tbQF}$ ), e) SEARCH denuded field blanks ( $OC_{dbQF}$ ), and f) SEARCH denuded backup filters ( $OC_{dQBQ}$ ; eight sites). IMPROVE thermal carbon fractions are defined as OC1 (140°C), OC2 (280°C), OC3 (480°C), and OC4 (580°C) in 100% helium (He); and EC1 (580°C), EC2 (740°C), and EC3 (840°C) in 98% He/2% oxygen ( $O_2$ ), and charring/pyrolysis carbon (OP; carbon evolved when reflectance returns to its initial value);  $OC = OC1 + OC2 + OC3 + OC4 + OP$ . EC levels ( $EC1 + EC2 + EC3 - OP$ ) were negligible for blank and backup filters and are not plotted.

**Fig. 4.** Comparison of field blank (bQF) carbon concentrations between the urban and non-urban sites among the IMPROVE, STN/CSN, and SEARCH networks for the period from 1/1/2005 to 31/12/2006. There are 13 urban and 169 non-urban IMPROVE sites, 239 STN/CSN urban sites, and four urban and four non-urban SEARCH sites. The urban IMPROVE sites are Atlanta, GA (ATLA1); Baltimore, MD (BALT1); Birmingham, AL (BIRM1); Chicago, IL (CHIC1); Detroit, MI (DETR1); Fresno, CA (FRES1); Houston, TX (HOUS1); New York, NY (NEYO1); Old Town, ME (OLTO1); Phoenix, AZ (PHOE1); Pittsburgh (PITT1); Rubidoux, CA (RUBI1); Washington, D.C. (WASH1) [<http://vista.cira.colostate.edu/improve/>].

**Fig. 5.** Comparison of quartz-fiber backup filter (QBQ) carbon fractions between urban and non-urban sites in the IMPROVE and SEARCH networks for the period from 1/1/2005 to

31/12/2006. Carbon fractions follow the IMPROVE\_A thermal/optical reflectance (TOR) protocol (Chow et al., 2007a).

**Fig. 6.** Time series of IMPROVE and STN/CSN blank TC concentrations at eight collocated sites from 1/1/2005 to 31/12/2006 (IMP\_bQF: IMPROVE field blanks; STN\_bQF: STN/CSN field blanks; STN\_tbQF: STN/CSN trip blanks).

**Fig. 7.** Linear regression of uncorrected STN/CSN TC vs. IMPROVE TC acquired from the Phoenix, AZ site (PHOE1). The non-zero intercept indicates sampling artifacts between STN/CSN and the IMPROVE network.

**Fig. 8.** The uncorrected STN–IMPROVE TC regression intercept for the entire data set and seasonally-segregated data from the eight collocated sites. Site names are PUSO1 (Seattle, WA); MORA1 (Mount Rainier, WA); PHOE1 ([Phoenix, AZ); TONT1 (Tonto National Monument, AZ); WASH1 (Washington, D.C.); DOSO1 (Dolly Sods Wilderness, WV); FRES1 (Fresno, CA); and BIBE1 (Big Bend National Park, TX).

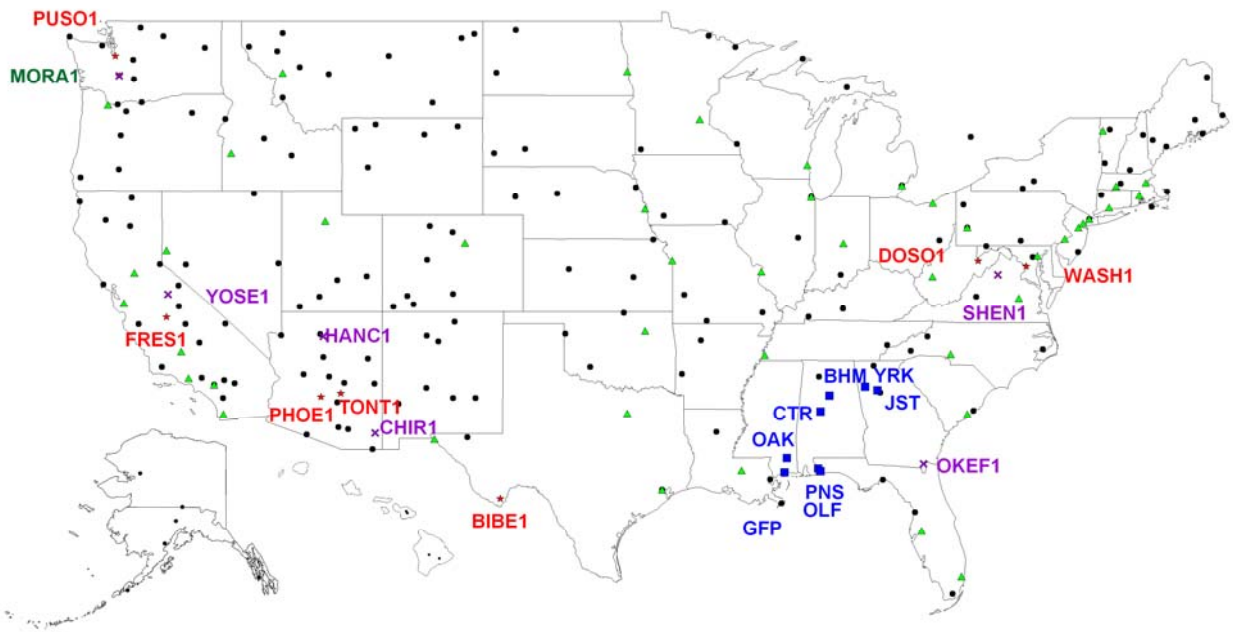
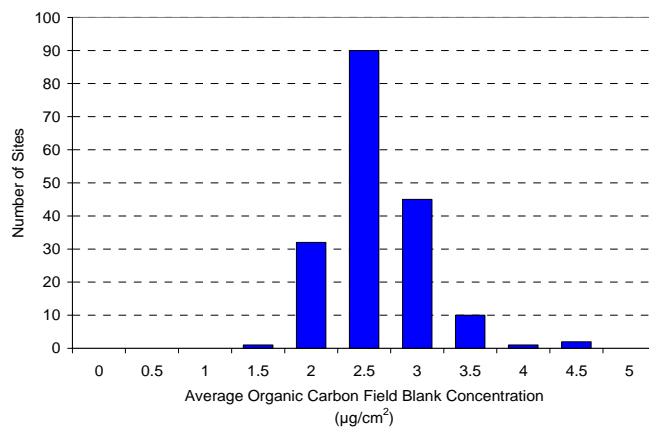
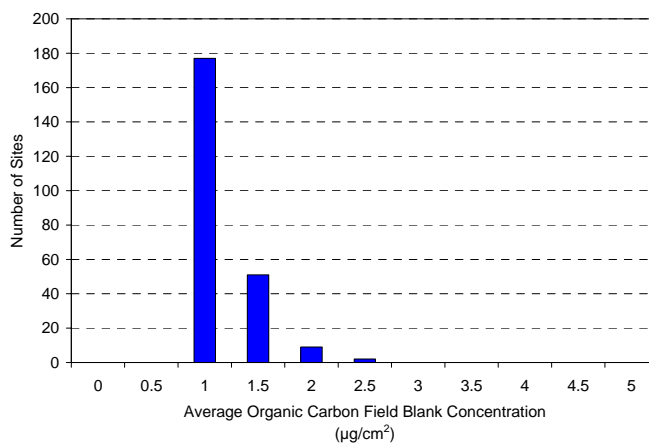


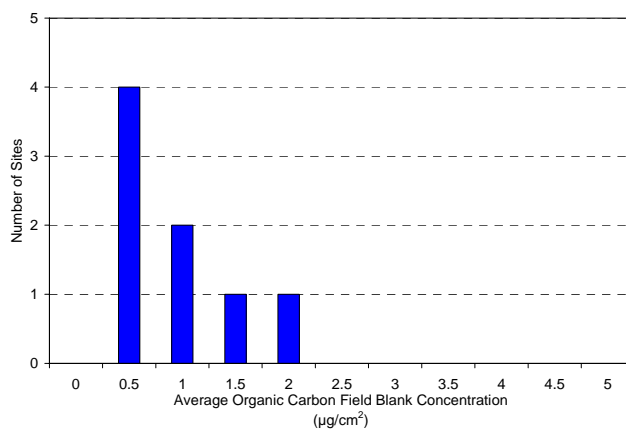
Figure 1.



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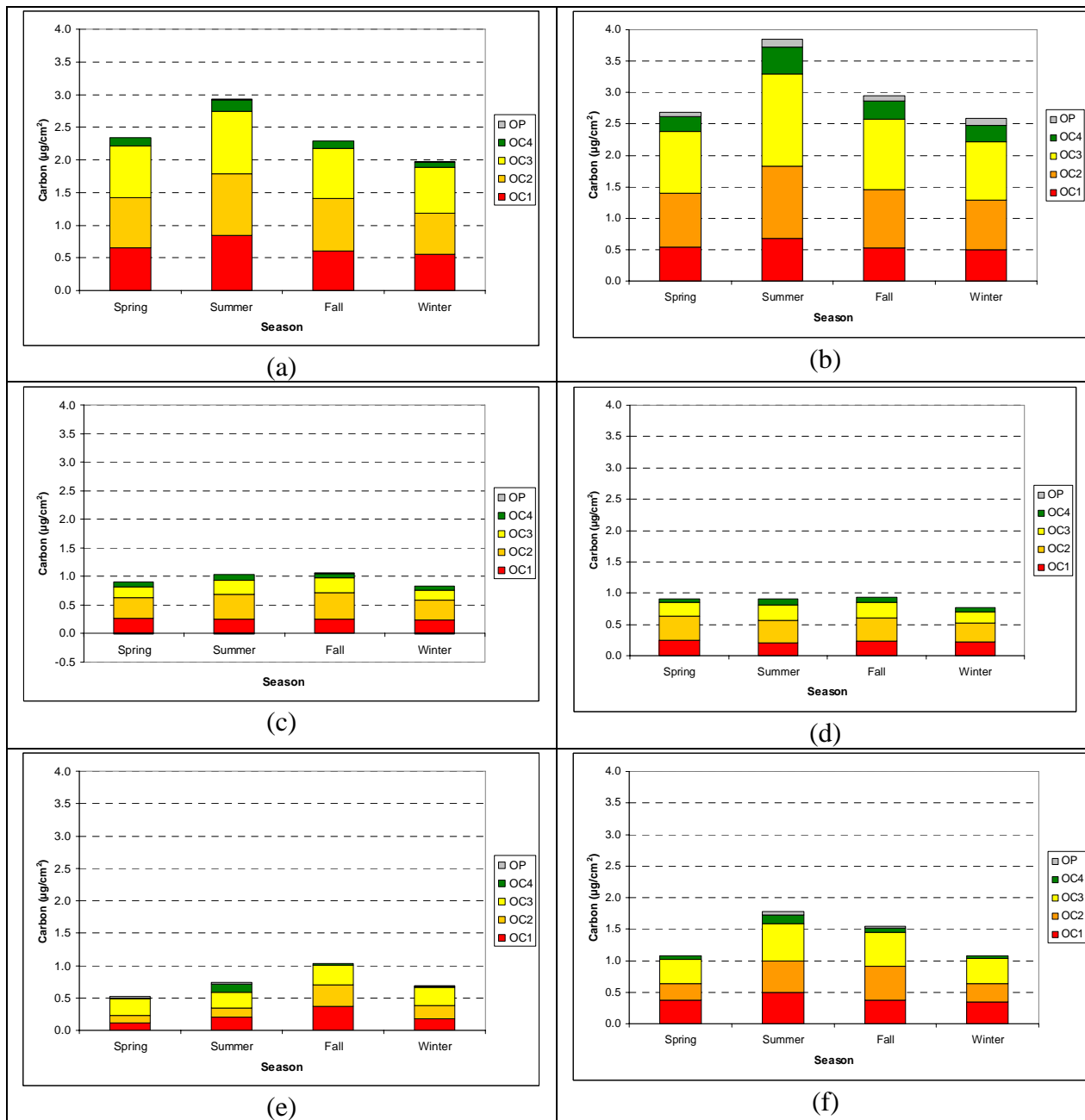


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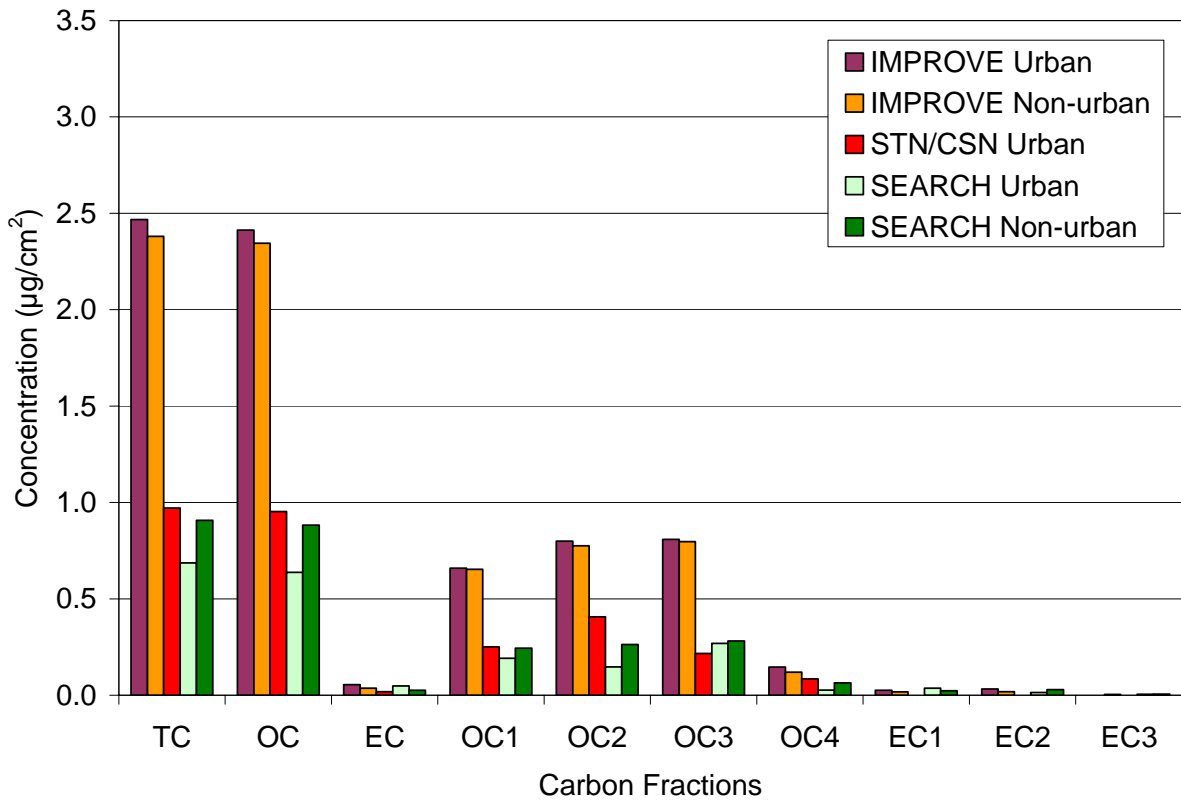


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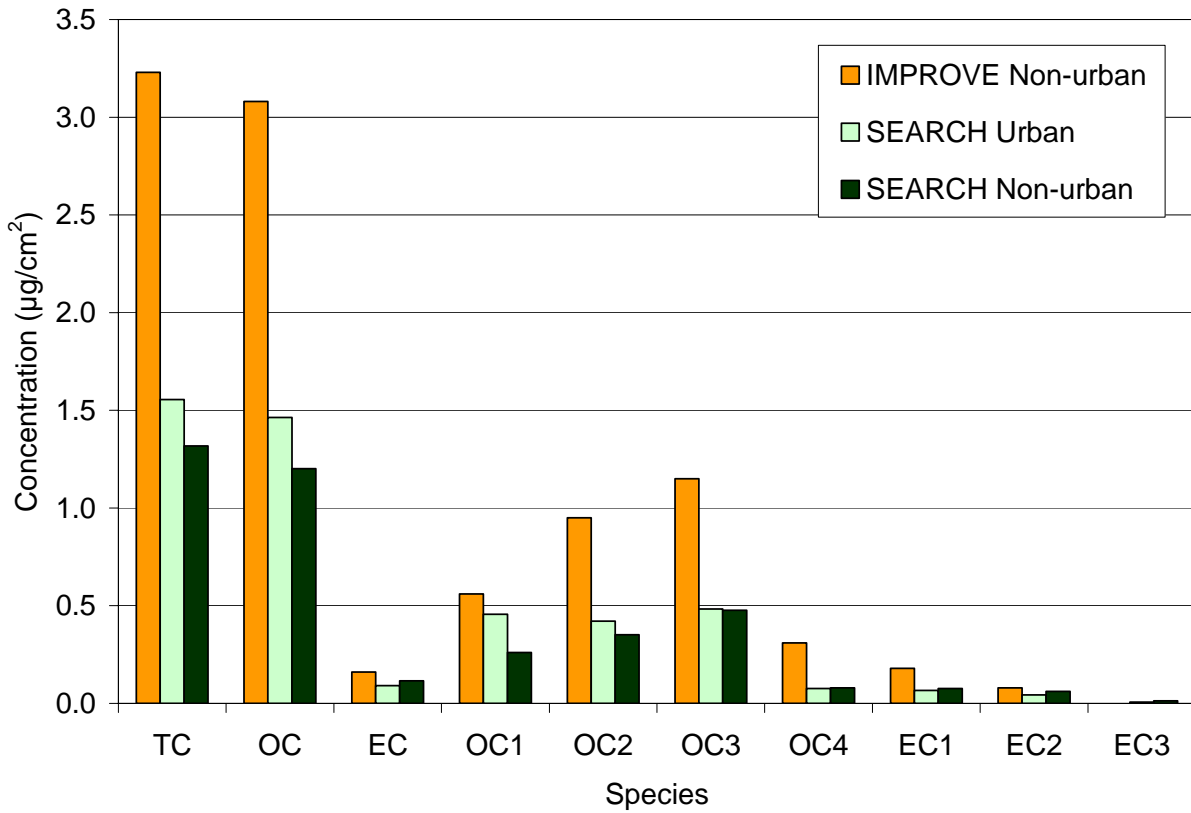
Figure 2.



**Figure 3.**



**Figure 4.**



**Figure 5.**

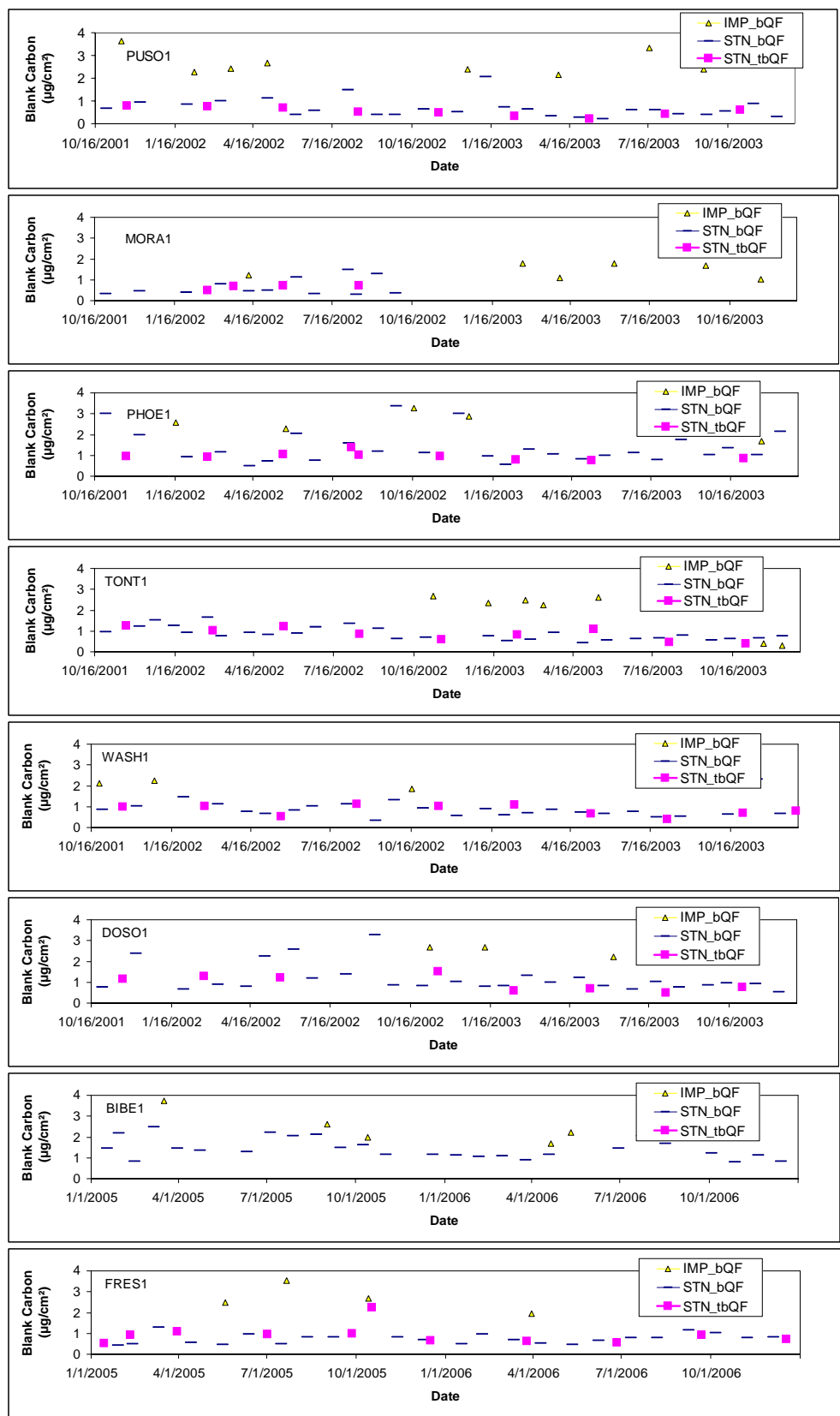
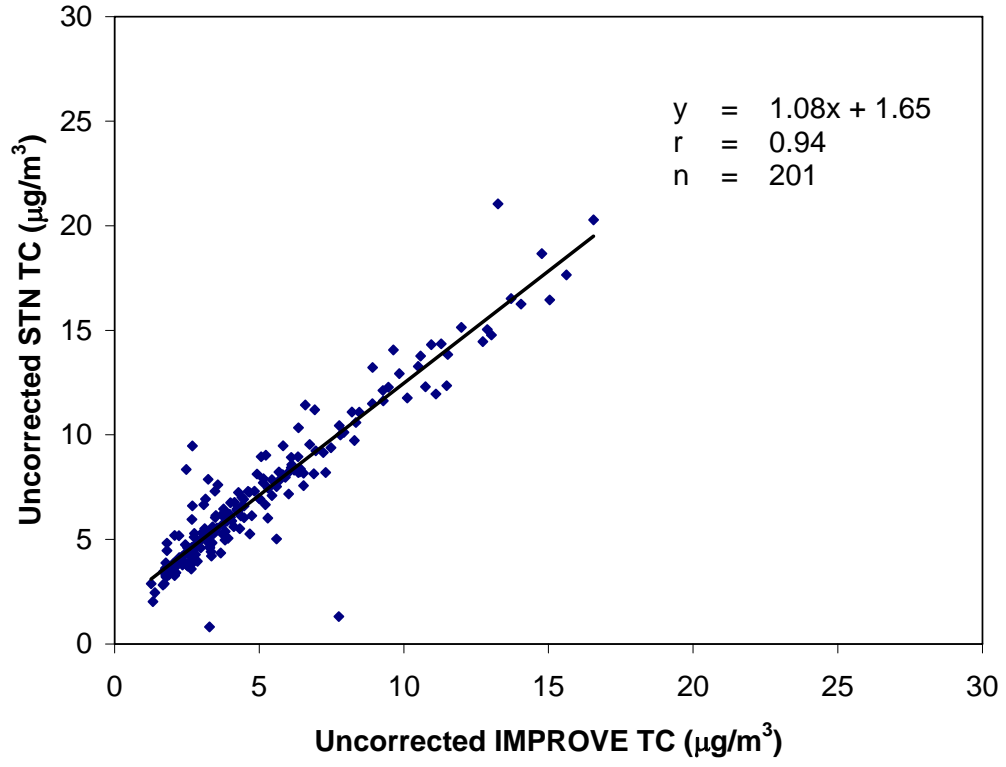


Figure 6.





**Figure 7.**

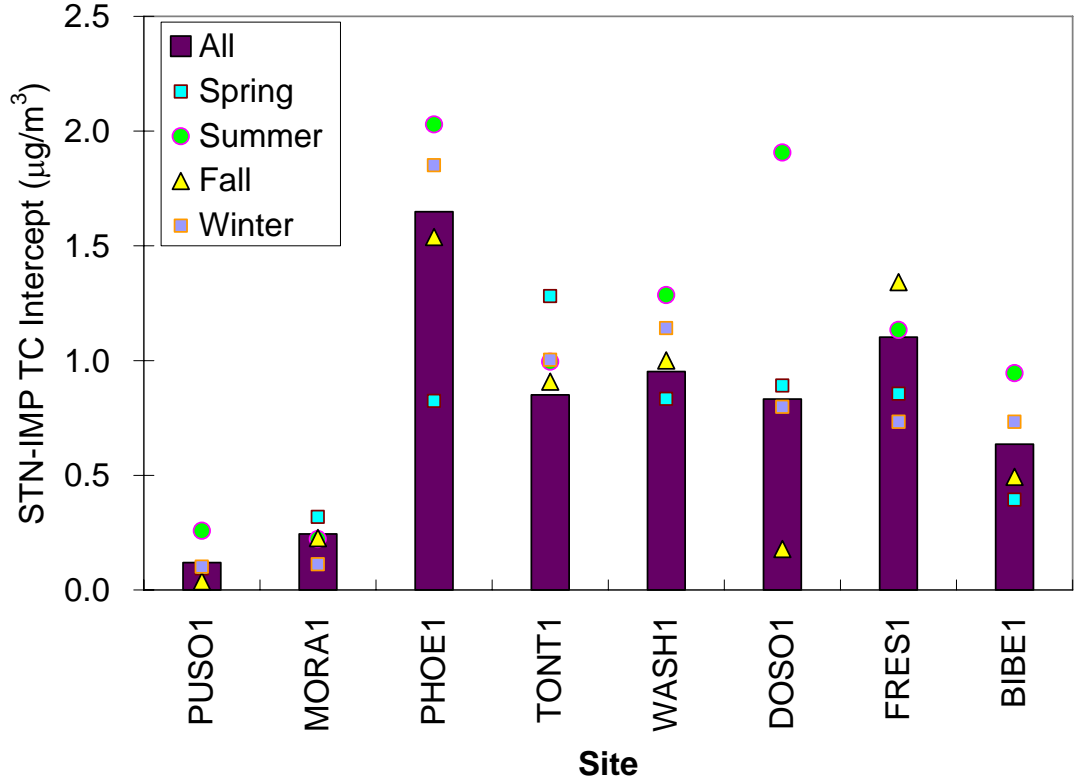


Figure 8.